

BEECH LEAVES BRIQUETTES' AND STANDARD BRIQUETTES' COMBUSTION: COMPARISON OF FLUE GAS COMPOSITION

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ABSTRACT

Biomass stoves are not only popular, widespread and important sources of heat but are also not negligible sources of pollutants. The present study had two objectives in this field of research. The first one was to determine the difference between standard wooden and beech leaves briquettes flue gas composition during similar, standard home combustion conditions. The second objective was to determine the possibility of decreasing the mass concentration of pollutants contained in the flue gas produced by standard and alternative fuel combustion, i.e. wooden briquettes and beech leaves briquettes, by an oxidation catalyst. Significantly higher mass concentration of nitrogen oxides (NO_x), almost 2.5 times higher, in the flue gas was observed during the beech leaves combustion. Both fuels reached the edge of actual legislation limit (European Standard Commission regulation [EU] 2015/1185) in case of mass concentration of carbon monoxide (CO). This issue was solved by a palladium-based catalyst with average degree of conversion around 82%. The catalyst also influences flue gas composition from mass concentration of propane point of view with average degree of conversion around 15%. The mass fraction of sulphur, occurring in the beech leaves briquettes, did not cause any issue to the catalyst in terms of its degree of CO conversion. Due to the test results from the beech leaves briquettes, i.e. high mass fraction of ash and high mass concentration of NO_x in the flue gas, it is appropriate to use this kind of fuel as secondary fuel during the co-combustion process.

Keywords: beech leaves, biomass combustion, catalyst, flue gas purification.

1 INTRODUCTION

Despite the increasing search for modern and nearly unattended ways (e.g. using heat pumps) of heating in whole European Union (EU), the popularity of biomass combustion heating is still considerable [1,2]. Biomass combustion units for household heating are popular, widespread and important for reaching European renewable energy share targets [3]. Nowadays there are more than 70 million of solid fuel appliances for household heating in Europe [4]; in Czech republic alone there are approximately 500 thousand houses that are heated by solid fuels combustion [5]. These numbers include all kinds of biomass and coal appliances. Coal combustion for the household heating is decreasing because of many well-known reasons, namely fossil fuel with relatively high sulphur content and mercury content, but on the other hand, biomass consumption for the household heating in EU is still increasing. For example, in year 2018, the consumption of wood pellets was approximately 26.1 million tonnes, which was higher by about 2.1 million tonnes in comparison to that of 2017 [6]. The significantly increasing trend of pressed biomass combustion is obvious in many European markets, e.g.

in year 2017, the consumption of wooden briquettes and pellets reached 210 thousand tonnes in Czech Republic and there was year-on-year increase by 20 tonnes (10%) [7].

The reasons for the rising of popularity of pressed biomass (pellets and briquettes) combustion could be the high energy density, lesser storage place demand and no additional drying need after transportation to the customer. Nowadays, production of pressed biomass is limited by sawmills' capacity and consequently by the amount of produced sawdust [8].

The facts mentioned above encourage the search for alternative sources of suitable (waste) biomass for briquettes and pellets production. From the energy point of view, collected fallen leaves from the town's greenery are considered a relatively homogeneous biomass, suitable for further use, e.g. briquetting and pelletizing.

The potential of usage of different kinds of leaves briquettes was observed in recent research works of Deepak *et al.* [9] (areca leaves), Khorasgani *et al.* [10] (oak leaves), Angono *et al.* [11] (amboyna wood leaves), Biswas [12] (sugar cane leaves), Hardianto *et al.* [13] (amboyna pine leaves), Shiferaw *et al.* [14] (eucalyptus globulus leaves residues), Ushakov *et al.* [15] (black poplar leaves), Shuma *et al.* [16] (mopani leaves and sugar cane leaves), Malak *et al.* [17] (willow leaves) and Jittabut [18] (sugarcane leaves). In the mentioned research papers, ultimate and proximate analyses of different types of leaves were carried out, as well as investigations on the dependence of calorific value on carbonization temperature, physical properties and ignition temperature.

A recently published research of Yunfei *et al.* [19] describes the possibility of tobacco straw usage as the fuel for co-combustion with coal and finds this combination of fuels to be advantageous. A study of Hatmojo *et al.* [20] describes the possibility of leaf litter briquettes as an alternative energy source for cooking by combustion. The mentioned research showed that leaf litter in the form of briquettes has the potential to be considered as a reliable energy source for cooking in household at rural areas.

Possibilities of beech leaves briquettes usage as the stove fuel for household heating was observed by previous research of Ryšavý *et al.* [21]. In this work, beech leaves were collected, spontaneously dried, crushed and pressed into briquettes. These leaves/briquettes were subjected to proximate and ultimate analysis, thermogravimetric analysis and combustion tests in few years old, regularly used stove (Storch ESTELLI KERAMIK). The combustion tests demonstrated the possibility of meeting actual legislative limits, but the Commission regulation (EU) 2015/1185 [22] is much stricter; therefore, the described combustion equipment was not able to meet those limits with beech leaves briquettes, used as a fuel. This article expands on the earlier research by Ryšavý *et al.* [21].

All mentioned investigations confirm the relevance and importance of the topic of thermic utilization of alternative biomass fuels, especially the waste ones from the forestry and agriculture, which was not fully utilized before.

The thermic utilization of the solid fuels also includes gasification and pyrolysis processes. There is a potential to get higher overall efficiency of energy transformation from all kinds of biomass-based solid fuels. Also, there are some other advantages connected with the possibilities of syngas utilization, but there are also some issues with tar formation during the gasification process [23,24].

The mentioned (unconventional) methods of solid fuels thermic utilization are especially suitable for middle and big scale energy systems (>100 kW) and require sophisticated devices and trained operators. The kind of connection of combustion and gasification processes is represented by wood gasification boilers and stoves, which partially transfer the issue of solid fuel combustion to gas fuel combustion. These kinds of devices can ensure higher efficiency of energy transformation under standard home conditions; they also ensure lower mass concentration of pollutants formed during the incomplete combustion process in the flue gas [25].

The main cause of high mass concentration of pollutants in the flue gas from the beech leaves briquettes' combustion was primarily a disintegration of briquettes on the grate. Disintegration of briquettes leads to non-sufficient approach of combustion air to the fuel and to creation of local high temperature anomaly.

The only method for decreasing the mass concentration of pollutants (from an incomplete combustion process) in the flue gas, without any construction intervention in the combustion equipment itself, is by using a catalyst as a secondary method for a flue gas purification. The flue gas temperature, measured during the beech leaves combustion tests in a previous study, ranged between 300 and 500°C, which is a suitable operating temperature for oxidation catalysts based on precious metal active elements.

The positive influence of catalysts' usage on flue gas composition from wooden pellets combustion was observed by Ryšavý *et al.* in two studies [26,27]. The degrees of conversion of mentioned catalysts were up to 78% of carbon monoxide (CO) and 64% of organic gaseous carbon (OGC) in concrete conditions. Catalysts based on precious metals and metal oxides were tested. Both types proved to have significant influence on the mass concentration of pollutants, produced during an incomplete combustion process (CO and OGC). Precious metal-based catalysts were also observed by Reichert *et al.* [28]. These tested catalysts proved with the potential of 95% reduction of CO and 60% reduction of OGC in the flue gas, produced by wood log combustion in real-life operated stoves.

The novelty of the present study lies in a straight comparison of certified, widely used fuel with new (alternative) one from the mass concentrations of pollutants point of view. A novelty in this study lies also in the oxidation, palladium-based catalyst usage for the flue gas purification formed during an alternative fuel combustion which contains sulphur.

This study was focused on a combination of unconventional renewable fuel with unconventional solution for the flue gas purification to provide sustainable, cheap and environmentally friendly option for household heating.

2 MATERIALS AND METHODS

2.1 Fuel

For the combustion tests, two fuels were chosen. The first one was the beech leaves briquettes, which have been described in detail by Ryšavý *et al.* [21]. The second used fuel was standard wooden briquettes (with ÖNORM certification) made only from pure dry spruce shavings (bark and additional binders free). The results of proximate and ultimate analyses are shown in Table 1. The biggest difference between these two tested types of briquettes is in mass fraction of ash. The high amount of ash in the beech leaves briquettes (comparable to brown coal) and mass fraction of sulphur, occurring in their composition, are the biggest disadvantages. The mass fraction of nitrogen is also higher (approximately 3.6 times) than in the case of standard wooden briquettes. Lower mass fraction of oxygen in the beech leaves briquettes worsens the ignition process.

2.2 Combustion equipment

A small-scale stove Storch ESTELLI KERAMIK was chosen as a combustion equipment. The construction of chosen stove (simple design without any flue gas/water heat exchanger and without flue gas or air fan) ensures relatively high flue gas temperature at the flue gas outlet, which is crucial for proper catalyst operation. Additional heat exchangers (flue gas/water or flue gas/air) may be placed in a flue gas duct behind the catalyst for additional

Table 1: The results of proximate and ultimate analyses of beech leaves briquettes and standard wooden briquettes used for the combustion test.

Element	Chemical symbol	Unit	Beech leaves briquettes	Standard wooden briquettes with ÖNORM certification
Carbon	C ^r	%	42.56	46.17
Hydrogen	H ^r	%	4.62	5.37
Nitrogen	N ^r	%	0.73	0.20
Oxygen	O ^r	%	30.25	38.87
Sulphur	S ^r	%	0.09	0
Water	W ^r	%	8.87	8.58
Ash	A ^r	%	12.88	0.79
Net calorific value	Q _i	MJ kg ⁻¹	15.64	17.00

increase of the thermal efficiency. For this test, no additional heat exchanger was used. The nominal heat output of the chosen stove is 11 kW. A sketch of the used stove with the location of installed catalysts is shown in Fig. 1.

The chosen combustion equipment was connected to the 7.5 m high chimney placed inside the laboratory. The chimney was constructed from polished brick chimney fittings and isostatic liners inside the fittings. The active part of the chimney was approximately 5.7 m high. The chimney was described in detail by Ryšavý *et al.* [29]. During the combustion test, only natural chimney draught was causing the flue gas suction from the stove, which was completely different from standard stove certification process. In the course of the certification process of the stove, an artificial chimney draught is caused by a flue gas fan. This test should have shown a real possibility of catalyst usage during the standard home operation.

2.3 Measuring system for a flue gas analysis

Flue gas was sucked from two sampling places of the flue gas duct (at the catalyst inlet and at the catalyst outlet). The same places served for flue gas relative pressure measurements. At first, flue gas samples were purified from solid particles on a glass wool filler. Then the flue gas went through a heated tube into two separated identical analysers MRU VARIO luxx. The mass concentrations of carbon monoxide (CO), carbon dioxide, oxygen (O₂), nitrogen oxides (NO_x) and propane (C₃H₈) in the flue gas were determined by these devices. Prior to the combustion test, the analysers were justified by calibration gases. During the combustion test, OGC was not measured, except for C₃H₈ mentioned above. Propane is not a negligible part of OGC, but its absolute values could not be compared with the legislative OGC limits.

2.4 Data evaluation

The mass concentrations of each flue gas component, as well as flue gas temperature and its relative pressure, were recorded as average per each minute. All final values were counted as average values of mass concentration of the components (reported as mass concentrations in

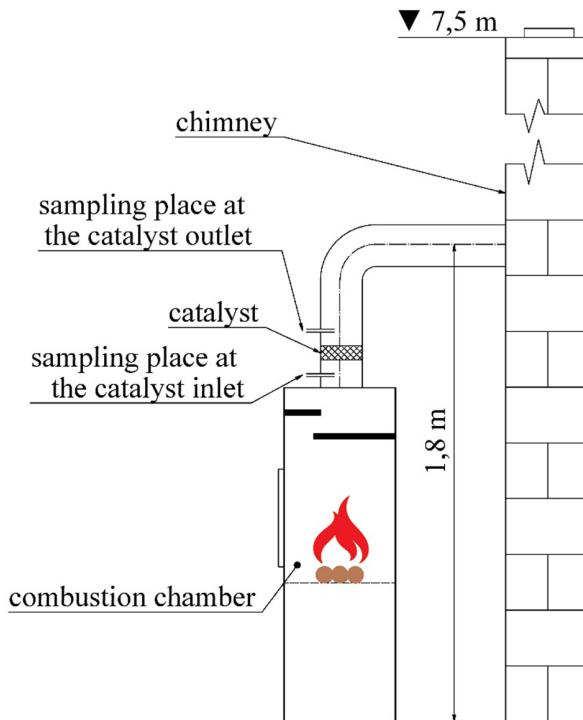


Figure 1: Sketch of used stove.

$\text{mg}\cdot\text{m}^{-3}$) in a dry flue gas, during the one combustion period. The final values were recalculated for standard conditions for temperature T and pressure p (STP conditions: $T = 273.15 \text{ K}$; $p = 101,325 \text{ Pa}$) and a reference volume fraction of oxygen in the flue gas $\varphi_{\text{ref},[\text{O}_2]} = 13\%$.

Degrees of conversion of the catalysts were calculated from a comparison of the mass concentrations of pollutants in the flue gas at the catalyst inlet and at the catalyst outlet present according to eqn (1):

$$X = \frac{\rho_{BO} - \rho_B}{\rho_{BO}} \quad (1)$$

where X is the degree of conversion of a pollutant as flowing through the catalyst (%), ρ_{BO} the mass concentration of a pollutant at the catalyst inlet ($\text{mg}\cdot\text{m}^{-3}$) and ρ_B the mass concentration of a pollutant at the catalyst outlet ($\text{mg}\cdot\text{m}^{-3}$).

2.5 Combustion test

For ignition and heating the stove, 1.5 kg of standard wooden briquettes was used. The first test starts when the basic firebed was made. The combustion tests were divided into six consecutive modes, three modes with standard wood briquettes (mode 1, mode 3 and mode 5) and three modes with beech leaves briquettes (mode 2, mode 4 and mode 6). This fuel changing was crucial for determination of the flue gas composition difference between these two chosen fuels during comparable combustion conditions. Before each addition of the fuel, the basic firebed was raked through, causing part of the ash to fall into the ashtray. This act was

necessary due to high mass fraction of ash in the beech leaves briquettes. According to the shape of beech leaves briquettes, which is given by the briquetting machine, and lower tendency to its internal strength (which caused the individual briquettes to be relatively small), it was necessary to add at least 12 pieces to reach overall weight of one batch of approximately 1,500 g (weights of all batches ranged around 1,500 g ± 10 g). In case of standard wooden briquettes, two, seven and eight pieces of briquettes were added to one batch. These pieces were cut off from standard briquettes shape (~0.1 m diameter; ~0.3 m length). The main reasons for cutting the standard briquettes were the small area of the grate (two pieces) and induction of similar conditions of the combustion process, as during the beech leaves briquettes' combustion (seven and eight pieces).

Always about 2 min after batch adding, a combustion air inlet flap was set to the operating position which was chosen according to operator experiences with this type of stove, to get the best flue gas composition from mass concentration of pollutants point of view. This position was not changed until the end of the mode. Each mode was terminated when only a hot basic layer of carbon without visible flames was left on the grate (approximately of the same layer thickness as before the previous fuel adding). Subsequently, the process was repeated in the next mode.

The test time of each mode was affected by the number of pieces of the fuel batch since it is closely related to the reacting surface of the fuel batch.

2.6 Tested catalyst

Detailed description of tested catalyst (referred to only as CAT A in what follows) is given in Table 2. A picture of tested catalyst CAT A is shown in Fig. 2. The producer of the catalyst

Table 2: Detailed description of used catalyst.

Marking	–	CAT A
Producer	–	Whitebeam
Catalyst's type	–	Monolithic – honeycomb
Body (substrate) material	–	cordierite
Active compound	–	Pd
Precious metal loading	g dm ⁻³	0.619
Shape	–	cylindrical
Main dimension	m	ø 0.144
Height	m	0.051
Volume	m ³	0.00083
Active surface	m ²	0.56373
Number of cells	cells cm ⁻²	4.14
Pressure loss ^a	pa	<1
Cell inlet surface	m ²	0.0000168
Number of cells	–	674

^a The average catalyst's pressure loss during the combustion tests.



Figure 2: Tested catalyst CAT A.

did not provide information about suitable operating temperature of the catalyst. A thickness of the catalyst's active element layer was not observed and no information was provided by the producer. The producer also did not provide a conversion efficiency diagram of the observed catalyst for oxidation reaction of CO and OGC.

3 RESULTS AND DISCUSSION

The most important overall measured values are presented in Table 3. The final values of the mass concentrations of pollutants were recalculated for STP conditions and a reference volume fraction of oxygen in the flue gas ($\varphi_{\text{ref};[\text{O}_2]} = 13\%$). The final value of mass concentration of NO_x is recalculated to NO_2 .

3.1 Fuels comparison

From the mass concentration of CO point of view, the flue gas produced during the standard wooden briquettes' combustion reached (on average) lower values (by approximately 28%) than during the beech leaves briquettes' combustion. An average value of mass concentration of CO during the modes 1, 3 and 5 was $1,256 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $1,052 \text{ mg}\cdot\text{m}^{-3}$ and $1,618 \text{ mg}\cdot\text{m}^{-3}$), and an average value of mass concentration of CO during the modes 2, 4 and 6 was $1,751 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $1,089 \text{ mg}\cdot\text{m}^{-3}$ and $2,390 \text{ mg}\cdot\text{m}^{-3}$). The most significant difference was made between mode 5 and mode 6. Modes 1 and 2, and 3 and 4 were comparable.

Values in accordance to Commission regulation (EU) 2015/1185 are marked green and those not in accordance with this regulation are marked red. BLB: beech leaves briquettes; nm: not measured; WB: wooden briquettes.

As can be seen in measured values of modes 3, 4 and 5, the three lowest concentrations were reached during the combustion of 2 and 8 pieces of wooden briquettes and 12 pieces of

Table 3: Overall results.

Mode number			1	2	3	4	5	6
Test time		h	0.7	0.8	1.0	0.8	0.8	0.8
Number of briquettes pieces		pcs	7 pcs	12 pcs	2 pcs	12 pcs	8 pcs	12 pcs
Type of fuel		-	WB	BLB	WB	BLB	WB	BLB
Average heat output during the mode		kW	7.5	6.2	4.8	5.9	6.8	6.1
Fuel consumption		kg/h	2.22	2.01	1.48	1.88	2.01	2.00
Power input		kW	10.5	8.7	7.0	8.2	9.5	8.7
Seasonal space heating energy efficiency		%	61.5	61.1	59.3	61.5	61.2	59.9
Flue gas temerature		°C	347	351	323	334	319	309
Volume fraction of oxygen in the flue gas		% _{vol.}	11.6	11.9	13.4	12.3	12.9	13.5
Air excess ratio		-	2.22	2.30	2.77	2.41	2.60	2.79
Flue gas volume flow		m ³ /h	51.65	45.51	40.69	43.39	51.78	50.74
Space velocity		1/h	62 192	54 799	48 987	52 246	62 345	61 092
Reference volume fraction of O ₂ in flue gas		% _{vol.}	13	13	13	13	13	13
Mass concentration of pollutants in the flue gas at the catalyst inlet	CO	mg/m ³	1 618	1 774	1 099	1 089	1 052	2 390
	NO _X	mg/m ³	nm	nm	nm	nm	nm	nm
	SO ₂	mg/m ³	nm	nm	nm	nm	nm	nm
	C ₃ H ₈	mg/m ³	27	12	19	7	18	64
	CO ₂	g/m ³	157	161	161	160	161	161
Reference volume fraction of O ₂ in flue gas		% _{vol.}	13	13	13	13	13	13
Mass concentration of pollutants in the flue gas at the catalyst outlet	CO	mg/m ³	172	293	232	244	188	419
	NO _X	mg/m ³	106	299	167	344	142	376
	SO ₂	mg/m ³	0	19	0	34	0	26
	C ₃ H ₈	mg/m ³	21	10	17	5	18	54
	CO ₂	g/m ³	159	164	163	161	163	164
Degree of conv. of CO		%	89.4	83.5	78.9	77.6	82.1	82.4
Degree of conv. of C ₃ H ₈		%	22.0	12.4	6.3	32.4	2.0	15.8

beech leaves briquettes in one batch. This result suggests that the shape of beech leaves briquettes (small cylinders – 0.055 m diameter; 0.04 ÷ 0.07 m length) is not the main reason of higher average mass concentrations of CO in the flue gas.

From the mass concentration of C_3H_8 point of view, flue gas produced during the standard wooden briquettes' combustion reached (on average) lower values (by approximately 23%) than during the beech leaves briquettes' combustion. An average value of mass concentration of CO during the modes 1, 3 and 5 was $21 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $18 \text{ mg}\cdot\text{m}^{-3}$ and $27 \text{ mg}\cdot\text{m}^{-3}$), and an average value of mass concentration of CO during the modes 2, 4 and 6 was $28 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $7 \text{ mg}\cdot\text{m}^{-3}$ and $64 \text{ mg}\cdot\text{m}^{-3}$). It can be seen that the average value of beech leaves briquettes was increased only by the highest value ($64 \text{ mg}\cdot\text{m}^{-3}$) measured during the mode 6. The rest of measured values during the beech leaves briquettes' combustion (modes 2 and 4) were the lowest of all tests.

From the mass concentration of NO_x point of view (measured only at the catalyst outlet – an oxidation catalyst should not have any influence at this value), flue gas produced during the standard wooden briquettes' combustion reached (on average) lower values (by approximately 60%) than during the beech leaves briquettes' combustion.

The average value of mass concentration of NO_x during the modes 1, 3 and 5 was $138 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $106 \text{ mg}\cdot\text{m}^{-3}$ and $167 \text{ mg}\cdot\text{m}^{-3}$), and an average value of mass concentration of NO_x during the modes 2, 4 and 6 was $340 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $299 \text{ mg}\cdot\text{m}^{-3}$ and $376 \text{ mg}\cdot\text{m}^{-3}$). The first reason for this difference was the higher mass fraction of nitrogen in the beech leaves briquettes than in standard wooden briquettes. The second reason was the disintegration of the beech leaves briquettes on the grate during the combustion process (this caused a partial clogging of the air inlet holes on the grate), which probably caused local heat maxima, responsible for excessive origin of thermic NO_x .

The average of mass concentration of SO_2 during modes 1, 3 and 5 was under the limit of detection and the average of mass concentration of SO_2 during modes 2, 4 and 6 was $26 \text{ mg}\cdot\text{m}^{-3}$ (values ranged between $19 \text{ mg}\cdot\text{m}^{-3}$ and $34 \text{ mg}\cdot\text{m}^{-3}$). The reason for non-detection under modes 1, 3 and 5 was the small mass fraction of sulphur in the beech leaves.

Regarding the seasonal energy efficiency, its average values were the same for both tested fuels (61%). The major heat lost, which mainly affects the efficiency, was the heat loss through sensible heat of the products of combustion, caused by high flue gas outlet temperature and high air excess ratio (high amount of flue gas).

The average heat output ranged between 4.8 and 7.5 kW, according to the time of the mode before adding the next batch and according to the fuel net calorific value. All heat outputs achieved were in the regulatory range according to manufacturer information.

Chimney draught measured at the catalyst inlet and outlet were approximately 12 Pa on average.

3.2 Degree of pollutants' conversion by tested catalysts

As can be seen, from the degree of CO conversion point of view, CAT A worked with approximately 82.3 percentage of conversion on average of the whole combustion test (83.5% in case of wooden briquettes' combustion and 81.2% in case of beech leaves briquettes' combustion). According to the results, the type of combusted fuel has no significant effect on degree of conversion of CO. In comparison to other catalysts, similar results were obtained during the Pd-based catalyst testing by Ryšavý *et al.* [26,27]. The small difference has been

caused by different space velocity and different temperature at the catalyst inlet, which affects the catalyst's degree of conversion.

As can be seen in Table 3, from the degree of C_3H_8 conversion point of view, CAT A worked with approximately 15 percentage of conversion in average of the whole combustion test (10.1% in case of wooden briquettes' combustion and 20.2% in case of beech leaves briquettes' combustion). According to the results, the type of combusted fuel has no significant effect on degree of conversion of C_3H_8 . The result values cannot be compared with some previous results, because previous catalyst's tests were aimed at a whole OGC degree of conversion. The share of C_3H_8 in the OGC was not known and also may vary during the combustion process.

The temperature of the flue gas inside the catalyst ranged between 309 and 351°C during all modes. The water vapor content in the flue gas ranged between 7.1 and 8.6% according to the type of fuel (due to different mass concentration of hydrogen), according to actual relative humidity of the combustion air and according to air excess ratio.

Due to the relatively high volume of catalyst itself and due to low flue gas flow, the flue gas flow ranged between approximately 40.7 and 51.8 $m^3 h^{-1}$. A low value of flue gas flow, causing low space velocity, was the reason for relatively high degree of conversion of CO. The flue gas flow through the catalysts was always laminar during all modes. Space velocity is an equivalent value to a hydraulic retention time of gas molecule inside of catalyst. It is obtained by flue gas volumetric flow rate divided by catalyst volume. Space velocity was calculated according to eqn (2):

$$\text{space velocity} = \frac{\text{flue gas volumetric flow rate}}{\text{catalyst volume}} \left[\frac{1}{h} = \frac{\frac{m^3}{h}}{m^3} \right] \quad (2)$$

There was a thin soot layer observed at the catalyst surface after the combustion tests were performed. The mass fraction of sulphur occurring in the beech leaves briquettes' composition did not cause any decrease of pollutants degree of conversion to catalyst.

3.3 Comparison of results with current legislation

According to the standard EN 13240:2001 [30] (currently in effect), from the mass concentration of pollutants point of view, there is a limit only aimed at CO ($12,500 \text{ mg}\cdot\text{m}^{-3}$). The minimum efficiency limit (50%) is also in the mentioned standard. All modes of both fuels were in accordance with the standard.

According to stricter Commission regulation (EU) 2015/1185, from the mass concentration of pollutants point of view, there are limits for CO ($1,500 \text{ mg}\cdot\text{m}^{-3}$), OGC ($120 \text{ mg}\cdot\text{m}^{-3}$), NO_x ($200 \text{ mg}\cdot\text{m}^{-3}$) and PM ($40 \text{ mg}\cdot\text{m}^{-3}$). This regulation also includes limit value for seasonal heat efficiency (65%). Values that are in accordance to this regulation are marked in green in Table 3, and values that are not in accordance with this regulation are marked in red. According to this standard only three of six modes were in accordance with CO limit (two modes with standard wooden briquettes and one mode with beech leaves briquettes), in the case when the catalyst was not used. In case of catalyst usage, the final values of all modes were well under the mentioned CO limit. The OGC limits could not be compared because

only C₃H₈ values were measured. According to this standard, only three out of six modes were in accordance with the NO_x limit (the modes with standard wooden briquettes as the fuel). Mass concentration of PM in the flue gas was not observed during the tests.

4 CONCLUSIONS

The results of this research proved the possibility of beech leaves briquettes usage as an alternative fuel for standard wood log stove. In the comparison of flue gas from standard wooden briquettes' combustion and beech leaves briquettes' combustion, under the same conditions, it is obvious that the biggest notable difference is in the higher mass concentration of NO_x in case of the beech leaves briquettes' combustion. This disadvantage causes the mass concentration of NO_x to exceed its limit value according to the current, valid legislature. A possible solution of this issue could be a co-combustion of both fuels, or briquettes production from mixed materials. The second notable deference is in the mass concentration of sulphur dioxide occurring in the flue gas produced during the beech leaves briquettes' combustion.

Both fuels struggled with the problem to meet the CO mass concentration limit. Oxidation palladium-based catalyst installed in the flue gas duct showed itself as a suitable solution for this problem. The low seasonal space heating efficiency reached is related mainly to stove construction (high flue gas temperature at the stove outlet). This issue could be solved by an additional heat exchanger.

In connection with catalysts, it is necessary to mention that every object inside the flue gas duct could mean a danger in a form of clogging, and the catalyst has to be regularly maintained.

Further research could be aimed at 3D CFD analysis in order to obtain additional information about the connection between the air fuel ratio and the local pollutant formation.

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