

30 Years of Post-disturbance Recruitment in a Neotropical Forest

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Abstract

Questions

Tree biodiversity is central for tropical forests functioning and services. In the global change context it is urgent to clarify the response of tree community diversity and composition to disturbance. The long-term community response to disturbance rely on tree recruitment, long seen as following deterministic successional pathways. These pathways however might be altered due to the hyper-diversity tropical forests and might not apply in cases of slight but recurrent disturbances induced by global changes. Examine post-disturbance recruitment trajectories would (i) elucidate the recruitment determinism and disentangle stochastic from deterministic, selection processes, and (ii) elucidate of tropical forests taxonomic and functional resilience.

Methods

We examined the trajectories over 30 years of recruited trees taxonomic and functional diversity in 75 ha of a neotropical forest following a disturbance gradient. Specifically, we analysed taxonomic richness, evenness, and turnover, and functional diversity and composition (regarding 7 leaf, stem and life-history functional traits) of recruited trees.

Results

We highlighted a three-phased successional pathway defined by the interplay of stochastic and deterministic recruitment processes. The successional pathway translated into (i) the growth of saplings mirroring the pre-disturbance community, (ii) the selective recruitment of light-demanding species entailing a high dominance of pioneers above a disturbance intensity threshold, and (iii) a return towards pre-disturbance taxonomic and functional characteristics with the recovery of stochastic recruitment processes. Both recruited trees functional and taxonomic characteristics seemed resilient, but the recovery time proved decades-long.

Conclusions

Community recruitment response to disturbance was driven by the emergence of deterministic competition processes for light, balancing the stochastic processes that ruled undisturbed communities. Recruitment taxonomic and functional characteristics seemed resilient but remained altered in the long-term which called cautions in terms of forest management.

Keywords

Taxonomic Diversity, Functional Diversity, Recruitment, Resilience, Neotropical Forests, Disturbance Dynamics, Tree Community

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1. Introduction

Determining the response of tropical forests to disturbance is key to predict their fate in the global changing context. In the last decades, tropical forests experienced a wide range of disturbance, from radical land-use changes for agriculture or mining (Dezécache *et al.*, 2017a,b) to more insidious changes of communities structure, diversity and functioning following climatic changes (Aubry-Kientz *et al.*, 2015) or anthropogenic activities like selective logging (Baraloto *et al.*, 2012). In that respect a vast literature successfully modeled community response to disturbance in terms of tree growth, tree height and fluxes of carbon, water and nutrients (Gourlet-Fleury & Houllier, 2000; Putz *et al.*, 2012; Pioniot *et al.*, 2016; Rutishauser *et al.*, 2016). Similar mod-

eling approaches regarding forest diversity and composition remain hindered by the scarcity of long-term monitoring and by studies' restriction to common or commercial species imposed by forest huge biological diversity (Sebbenn *et al.*, 2008; Vinson *et al.*, 2015). The template of community response to disturbance is set by recruitment processes that determine the new species joining the community. Focusing on long-term recruitment trajectories therefore give valuable insight into post-disturbance recovery and hence into the adjustment of exploitation and conservation guidelines (Díaz *et al.*, 2005; Schwartz *et al.*, 2017).

The traditional view of community response to disturbance relies on successional vegetation models (Clements, 1916) based on changes in resources availability and interactions among species. Adapted to forest ecosystems the successional framework translates into (Denslow & Guzman, 2000) (i) the recruitment of pre-disturbance surviving saplings benefiting from the high resources availability and low competition, (ii) the progressive exclusion of species with low competitive ability because of increased competition for resources following stand maturation and (iii) the recovery of pre-disturbance composition and diversity due to the senescence of early-successional pioneers and the emergence of late-successional species. This highly-deterministic successional pathway proved relevant in temperate forests but remain questioned in tropical rainforests (Norden *et al.*, 2015). Indeed, the classical successional pathway may be altered by the huge biological diversity of tropical rainforests and their high functional redundancy that lead up to more stochastic processes. Moreover, the successional pathway proved well-adapted to system trajectories following clear cutting or very intense disturbance, but might be less robust following more insidious global changes. In those cases, community trajectories would depend on the interplay between the stochastic processes, driven by recruitment and dispersal limitations (Hubbell, 2001), and deterministic processes, driven by niche-based competition and biotic interactions (Adler *et al.*, 2007). Stochastic processes, in the neutral theory spirit, build recruited communities as random samples of the surrounding communities (Hubbell, 2001; Chave, 2004). In contrast under deterministic processes, species are selected with respect to their ecological strategies and competitive ability. The relative importance of stochastic and deterministic processes in shaping the post-disturbance trajectories would also change with time, along with the recovery of pre-disturbance environmental conditions.

The processes shaping recruitment trajectories may differently affect communities taxonomic characteristics, that refer to neutral species assemblages, and functional characteristics, that account for species ecology and ecosystem functioning (Violle *et al.*, 2007; Kunstler *et al.*, 2016). The correlations, or not, between community taxonomic and functional trajectories are therefore insightful of the main ecological processes underlying species recruitment (Fukami *et al.*, 2005). Among these processes, competition depends on the species competitive ability for a given niche, defined by the functional differences regarding the use of limited shared resources (Perronne *et al.*, 2017). In tropi-

cal forests where light is the limiting resource, community response to disturbance would translate in a shift from slow-growing, long-lived species with “conservative” resource use, to fast-growing species with “acquisitive” resource use (Denslow, 1980; Molino & Sabatier, 2001; Bongers *et al.*, 2009).

The competition processes at stake would be grasped by shifts in key leaf, wood and life-history functional traits assessing species resources acquisition strategy and ecology (Wright *et al.*, 2004; Chave *et al.*, 2009; Hérault *et al.*, 2011).

Balancing between determinism and stochasticity, the post-disturbance recruitment trajectories might thus show either a random divergence or a deterministic convergence towards stable taxonomic and functional characteristics, likely defined by the environment (Clements, 1916; Diamond, 1975). Both views were reconciled under the hypothesis that communities might diverge in the taxonomic space, having different taxonomic composition and diversity, while they converge in the functional space, but this remains to be tested in tropical forests (Fukami *et al.*, 2005; Li & Shipley, 2018).

In this paper we followed recruitment trajectories over 30 years of 75 ha of neotropical forest plots set up on a gradient of disturbance intensity, from 10 to 60% of forest biomass removed. We examined the recruited trees (i) taxonomic composition, richness and evenness, (ii) taxonomic turnover compared to pre-disturbance community, and (iii) functional composition and diversity based on seven major leaf, stem and life-history traits. We compared the recruitment trajectories to neutral models corresponding to a stochastic recruitment and a randomization of species functional traits. Specifically, we (i) elucidated the successional pathway shaping community response to disturbance and the underlying ecological processes and (ii) clarified the extent of community taxonomic and functional resilience and its consequences for tropical forest management.

2. Material and Methods

2.1 Study Site

The Paracou station is located in a lowland tropical rainforest in French Guiana (518°N and 5253°W). Climate is tropical wet with mean annual precipitation averaging 2980 mm.yr⁻¹ (30-yr period) and a 3-months dry season (< 100 mm.mo⁻¹) from mid-August to mid-November, and a one-month dry season in March (Wagner *et al.*, 2011). Elevation ranges from 5 to 50 m and mean annual temperature is 26°C. Soils are thin acrisols over a layer of transformed saprolite with low permeability generating lateral drainage during heavy rains. The experiment is a network of twelve 6.25 ha plots (Table 1) that underwent three disturbance treatments in 1987 according to a randomized plot design with three replicate blocks of four plots (Hérault & Piponiot, 2018).

2.2 Inventories Protocol and Dataset Collection

Dominant families in the study site are Fabaceae, Chrysobalanaceae, Lecythidaceae and Sapotaceae. All trees above 10 cm DBH were mapped and measured annually since 1984. Trees are first identified with a vernacular name assigned by

Table 1. Intervention table, summary of the disturbance intensity for the 4 plot treatments in Paracou.

Treatment	Timber	Thinning	Fuelwood	%AGB lost
Control				0
T1	DBH \geq 50 cm, commercial species, \approx 10 trees/ha			[12% – 33%]
T2	DBH \geq 50 cm, commercial species, \approx 10 trees/ha	DBH \geq 40 cm, non-valuable species, \approx 30 trees/ha		[33% – 56%]
T3	DBH \geq 50 cm, commercial species, \approx 10 trees/ha	DBH \geq 50 cm, non-valuable species, \approx 15 trees/ha	40 cm \leq DBH \leq 50 cm, non-valuable species, \approx 15 trees/ha	[35% – 56%]

the forest worker team, and afterward with a scientific name assigned by botanists during regular botanical campaigns. Botanical campaigns have been carried out every five to six years from 2003 onwards but identification levels varied between campaigns.

These variability of protocols in time raised methodological issues as vernacular names usually correspond to different botanical species. It resulted in significant taxonomic uncertainties that had to be propagated to composition and diversity metrics. The uncertainty propagation was done through a Bayesian framework reconstituting complete inventories at genus level from real incomplete ones on the basis of vernacular/botanical names association. Vernacular names were replaced through multinomial trials based on the association probability $[\alpha_1, \alpha_2, \dots, \alpha_N]$ observed across all inventories between each vernacular name v and the species $[s_1, s_2, \dots, s_N]$:

$$M_v([s_1, s_2, \dots, s_N], [\alpha_1, \alpha_2, \dots, \alpha_N])$$

See appendix 1 and Aubry-Kientz *et al.* (2013) for the detailed methodology.

To minimize the remaining identification uncertainties, the simulated botanical inventories were reported at genus level.

Six functional traits representing the leaf economics (leaves thickness, toughness, total chlorophyll content and specific leaf area) and stem economics spectra (wood specific gravity and bark thickness), and life-history traits (maximum specific height and seed mass) were considered. Traits were extracted from the BRIDGE project (<http://www.ecofog.gf/Bridge/>) where trait values were assessed from a selection of individuals located in nine permanent plots in French Guiana, including two in Paracou, and comprised 294 species pertaining to 157 genera. Missing trait values (10%) were filled by multivariate imputation by chained equation (van Buuren & Groothuis-Oudshoorn, 2011). Imputations were restricted within genus or family when samples were too scarce, in order to account for the phylogenetic signal.

As seed mass information was classified into classes, no data filling process was applied and analyses were restricted to the 414 botanical species recorded.

All composition and diversity metrics were obtained after 50 iterations of the uncertainty propagation framework.

2.3 Recruitment trajectories

Communities were split into surviving trees of pre-disturbance communities and trees recruited afterward in 2-years intervals.

Taxonomic diversity trajectories were assessed through species richness and evenness (the Hill number translation of the Simpson index) (Chao & Jost, 2015; Marcon, 2015). The two diversities belong to the set of HCDT or generalized entropy, respectively corresponding to the zero and two order of diversity (q), which grasps the balance between richness and evenness in the community through the value of q that emphasizes common species.

Functional diversity trajectories were assessed through the Rao index of quadratic entropy, which combines species abundance distribution and average pairwise dissimilarity based on all functional traits. Functional composition trajectories were assessed through the functional traits community weighted means (CWM), representing the average trait value in a community weighted by species relative abundance (Díaz *et al.*, 2007). Seed mass trajectories were reported by the proportion of each class recorded in the inventories (Annexe I).

The taxonomic similarity between recruited trees and pre-disturbance forest was measured with the turnover metrics detailed in Podani *et al.* (2013).

The taxonomic and functional recruitment trajectories were compared to null trajectories obtained after 50 iterations of the null models. The taxonomic null model was a random sampling of recruited trees within the living communities, with the maintenance of species abundance and tree density. The functional null model was a reassignment of species trait values that randomized traits abundances but maintained communities abundance distribution (Mason *et al.*, 2013). The null trajectories were similarly obtained after 50 iterations of the random sampling.

3. Results

3.1 Taxonomic richness and evenness and functional diversity

In undisturbed communities the recruitment taxonomic richness and evenness remained stable over the 30 years and with values equivalent to those of the taxonomic null model (Figure 1).

In disturbed communities the taxonomic richness followed hump-shaped trajectories first increasing until a maximum reached after around 15 years and positively correlated to the disturbance intensity ($\rho_{\text{spearman}}^{\text{Richness}} = 0.93$). Afterward the taxonomic richness decreased and recovered the pre-disturbance values after 30 years. The observed taxonomic richness was increasingly lower than this of null model for 15 years, then the difference started to shrink but the observed richness remained negative until after 30 years. The taxonomic evenness decreased independently of the disturbance intensity over the 30 years ($\rho_{\text{spearman}}^{\text{simpson}} = -0.35$). The observed taxonomic evenness was increasingly lower than this of the null model until 15 years after disturbance, when the difference stabilized.

The functional diversity in the undisturbed plots remained stable and equivalent to this of the functional null model over the 30 years. In the lowest disturbance plots the functional diversity remained stable or slightly increasing, and was higher than this of the null model for two of the T1 plots. In the disturbed plots of higher disturbance intensity (T2 and T3) the functional diversity decreased until 15 years after disturbance, when it started to recover towards initial values. The observed functional diversity remained lower than this of the null model over the 30 years.

3.2 Functional composition

In undisturbed plots functional traits values remained stable over the 30 years while it followed hump-shaped trajectories in all disturbed plots, to the exception of the leaf chlorophyll content. Trajectories of SLA and bark thickness first increased before decreasing towards initial values. Conversely, trajectories of leaf thickness, leaf toughness, wood specific gravity, and maximum height first decreased and then started returning towards initial values but their recovery remained unachieved after 30 years (Figure 2).

3.3 Recruitment Turnover

Over the 30 years in control plots the turnover of recruited species compared to initial community remained low (Figure 3). In disturbed plots the recruited species turnover followed a marked hump-shaped trajectory, with a maximum reached around 15 years after disturbance. The maximum turnover was positively correlated to the disturbance intensity ($\rho_{\text{spearman}} = 0.93$). Thirty years after disturbance the turnover had returned to low values.

4. Discussion

4.1 A three-phased deterministic successional pathway

Post-disturbance recruitment trajectories relied on a three-phased successional pathway defined by the emergence

of deterministic competition processes for light gradually balancing the stochastic recruitment specific to undisturbed communities.

A first phase (0-8 years), corresponded to the recruitment of pre-disturbance surviving saplings (DBH < 10 cm) that immediately benefited from the increased enlightenment and alleviated competition induced by disturbance (Denslow & Guzman, 2000; Herault *et al.*, 2010). The taxonomic and functional characteristics of recruited trees mirrored the pre-disturbance communities and recruitment processes matched the null stochastic recruitment model.

A second phase (8-15 years) was marked by a shift in community functional composition towards more “acquisitive” functional strategies and the dominance of a restricted set of species. The recruitment then involved true recruits, *i.e.* trees germinating from the seeds bank, representing the main part of the whole post-disturbance recruitment (Lawton & Putz, 1988). The recruitment was dominated by short-lived, fast growing hard pioneers displaying efficient light acquisition (Wright *et al.*, 2004; Chave *et al.*, 2009; Hérault *et al.*, 2011). As already demonstrated in temperate forests, the pool of recruited species was restricted by trait-based selection processes favoring species with efficient light acquisition (high SLA and leaf chlorophyll content) and inexpensive, short-lived tissues (low leaf thickness and toughness, small Hmax and low wood specific gravity and bark thickness) (Chave, 2004; Kunstler *et al.*, 2016). This emergence of trait-based deterministic processes balanced the stochastic recruitment observed in the first place, and the relative importance of both processes was determined by the disturbance intensity. After low intensity disturbance (T1 plots) recruited species still mirrored pre-disturbance taxonomic composition, but included more long-lived pioneers and light-demanding species (Bongers *et al.*, 2009). For intense disturbance in contrast (T2 and T3 plots), the composition of recruited trees rapidly differed from pre-disturbance community and with the high dominance of hard pioneers, such as *Cecropia spp.* or *Vismia spp.*, likely entailing significant changes in communities functioning (Díaz *et al.*, 2005).

A third recruitment phase (15-30 years) corresponded to the recovery of pre-disturbance taxonomic and functional characteristics. Although the recruits remained mainly light-demanding species their functional diversity increased and they increasingly resembled the pre-disturbance taxonomic composition. The deterministic recruitment processes then gradually left room to stochastic recruitment processes specific to undisturbed forest (Lawton & Putz, 1988; Chave, 2004).

4.2 The achievement of communities recovery

After disturbance the stochastic recruitment specific to undisturbed communities was progressively restored and drove community taxonomic and functional recovery. This confirmed previous results from the Paracou experiment, conducted 10 years (Molino & Sabatier, 2001) and 20 years (Baraloto *et al.*, 2012) after disturbance, where the early signs of the resilience of taxonomic and functional composition had been detected.

Recruitment taxonomic richness and evenness recov-

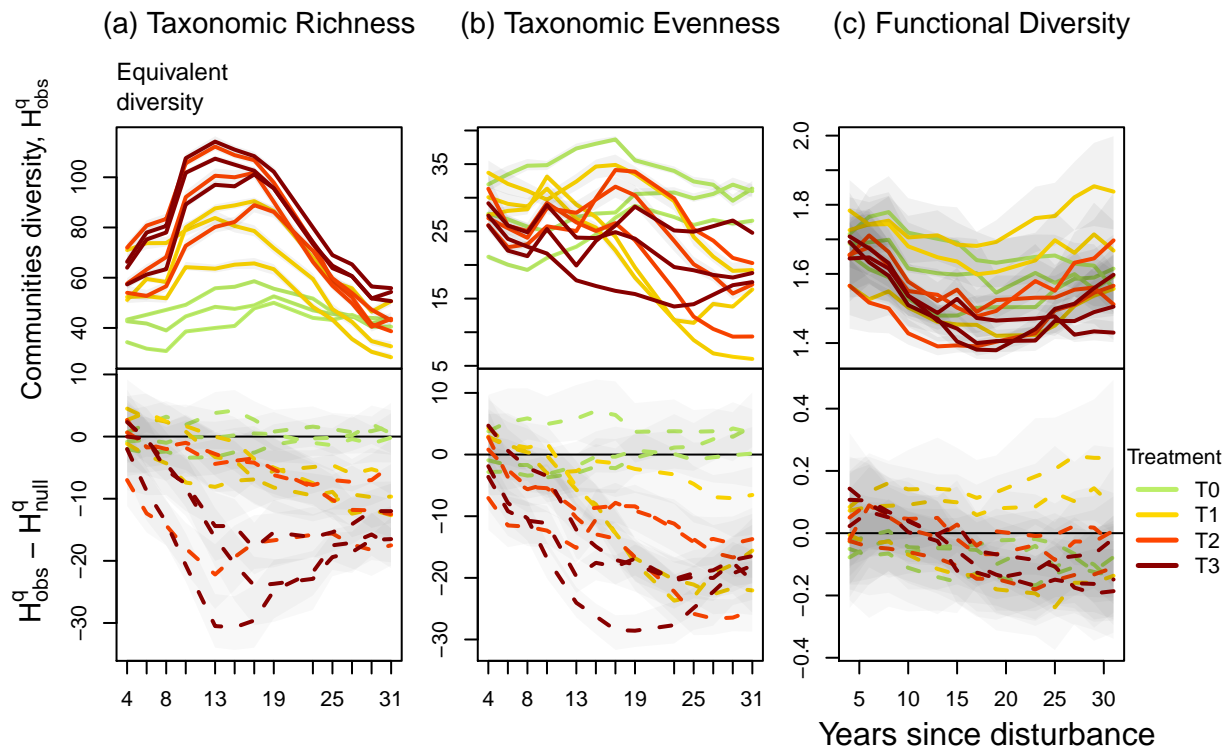


Figure 1. Upper panels, trajectories over 30 years of taxonomic richness (a), taxonomic evenness (b) and functional diversity (c) of 2-years laps recruitment. Lower panels, diversity differences to null models. Colors are treatments: green (control), yellow (T1), orange (T2), red (T3) with shaded areas the credibility intervals.

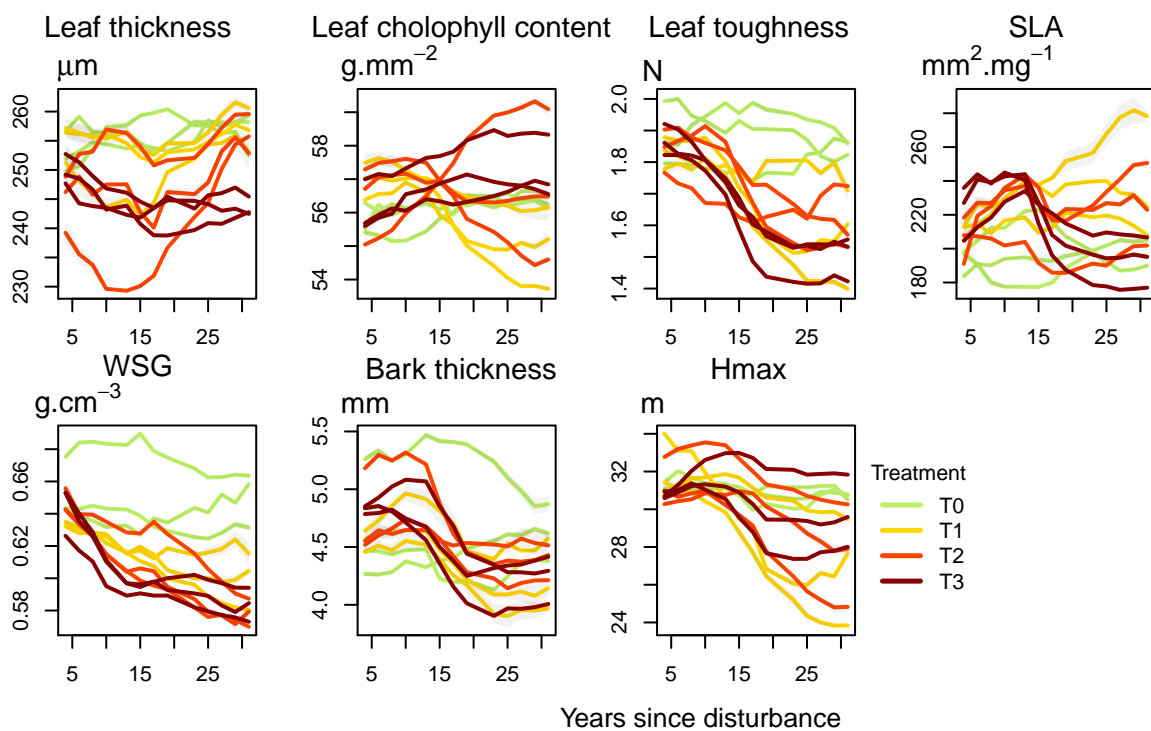


Figure 2. Community weighted means (CWM) of the leaf, the two stem and specific maximum height. Colors are treatments: green (control), yellow (T1), orange (T2), red (T3) with shaded areas the credibility intervals.

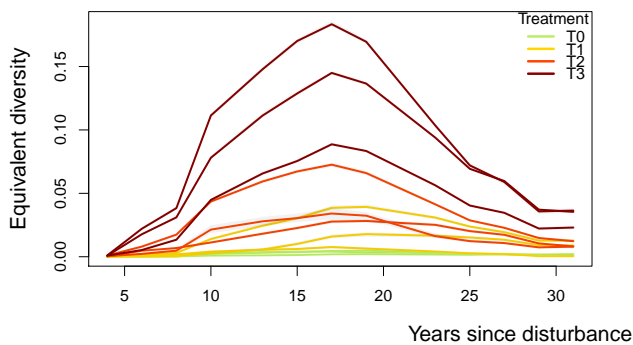


Figure 3. Trajectories over 30 years of the abundance-based turnover between 2-years laps recruited trees and initial communities before disturbance. Colors are treatments: green (control), yellow (T1), orange (T2), red (T3) with shaded areas the credibility intervals.

ered pre-disturbance values and the taxonomic composition converged towards the pre-disturbance community, thus maintaining the initial differences among communities for all disturbance intensity. Community taxonomic convergence to the local pre-disturbance recruitment composition revealed the scarce recruitment of species that did not belong to pre-disturbance community, due to the commonness of dispersal limitation among tropical tree species (Svenning & Wright, 2005).

Functional composition and diversity trajectories converged similarly in the functional space towards the recovery of pre-disturbance values, suggesting a common and resilient functioning despite communities' taxonomic divergence (Fukami *et al.*, 2005).

Trait-based selection processes made deterministic the community functional response to disturbance but dispersal limitation and steady-state stochastic recruitment made community taxonomic response historically contingent. Although resilient, the functional and taxonomic composition of recruited trees remained altered 30 years after disturbance by the dominance of light-demanding species. This long-term impact specifically raises questions for the management of exploited forests, as most valuable species are late-successional and would thus require cutting cycles of more than 30 years (Putz *et al.*, 2012).

5. Conclusion

The post-disturbance recruitment trajectories highlighted a three-phased deterministic successional pathway shaped by the emergence of niche selection for light balancing the stochastic recruitment of undisturbed communities. The successional pathway first corresponded to the enhanced growth of pre-disturbance surviving saplings mirroring the taxonomic and functional characteristics of pre-disturbance communities. Second, recruitment trajectories were shaped by true recruits from the seeds bank selected through the emergence of competitive exclusion for light fostering pioneer species. Above a disturbance intensity threshold the

second recruitment phase was dominated by short-lived hard pioneers that drastically changed community composition, diversity and likely functioning. A third phase eventually corresponded to the return towards pre-disturbance recruitment composition and taxonomic and functional diversity, through the recovery of stochastic recruitment processes specific to undisturbed communities. Besides, repeated disturbance might have increasingly strong impacts, as community recovery involved the seeds bank and probably altered the composition and diversity of the seeds stock (Norden *et al.*, 2009).

References

- Adler, P.B., HilleRisIambers, J. & Levine, J.M. (2007) A niche for neutrality. *Ecology Letters*, **10**, 95–104.
- Aubry-Kientz, M., Rossi, V., Wagner, F. & Hérault, B. (2015) Identifying climatic drivers of tropical forest dynamics. *Biogeosciences*, **12**, 5583–5596.
- Aubry-Kientz, M., Hérault, B., Ayotte-Trépanier, C., Baraloto, C. & Rossi, V. (2013) Toward Trait-Based Mortality Models for Tropical Forests. *PLoS ONE*, **8**.
- Baraloto, C., Hérault, B., Paine, C.E.T., Massot, H., Blanc, L., Bonal, D., Molino, J.F.F., Nicolini, E.a. & Sabatier, D. (2012) Contrasting taxonomic and functional responses of a tropical tree community to selective logging. *Journal of Applied Ecology*, **49**, 861–870.
- Bongers, F., Poorter, L., Hawthorne, W.D. & Sheil, D. (2009) The intermediate disturbance hypothesis applies to tropical forests, but disturbance contributes little to tree diversity. *Ecology Letters*, **12**, 798–805.
- Chao, A. & Jost, L. (2015) Estimating diversity and entropy profiles via discovery rates of new species. *Methods in Ecology and Evolution*, **6**, 873–882.
- Chave, J. (2004) Neutral theory and community ecology. *Ecology Letters*, **7**, 241–253.
- Chave, J., Coomes, D., Jansen, S., Lewis, S.L., Swenson, N.G. & Zanne, A.E. (2009) Towards a worldwide wood economics spectrum. *Ecology Letters*, **12**, 351–366.
- Clements, F.E. (1916) *Plant succession: an analysis of the development of vegetation*. 242. Carnegie Institution of Washington.
- Denslow, J.S. (1980) Gap Partitioning among Tropical Rainforest Trees. *Biotropica*, **12**, 47–55.
- Denslow, J.S. & Guzman, G.S. (2000) Variation in stand structure, light and seedling abundance across a tropical moist forest chronosequence, Panama. *Journal of Vegetation Science*, **11**, 201–212.
- Dezécache, C., Faure, E., Gond, V., Salles, J.M., Vieilledent, G. & Hérault, B. (2017a) Gold-rush in a forested El Dorado: Deforestation leakages and the need for regional cooperation. *Environmental Research Letters*, **12**.

- Dezécache, C., Salles, J.M., Vieilledent, G. & Hérault, B. (2017b) Moving forward socio-economically focused models of deforestation. *Global Change Biology*, **23**, 3484–3500.
- Diamond, J.M. (1975) Assembly of species communities. *Ecology and evolution of communities*, pp. 342–444.
- Díaz, S., Lavorel, S., de Bello, F., Quétier, F., Grigulis, K. & Robson, T.M. (2007) Incorporating plant functional diversity effects in ecosystem service assessments. *Proceedings of the National Academy of Sciences*, **104**, 20684–20689.
- Díaz, S., Tilman, D., Fargione, J., Chapin III, F.S., Dirzo, R., Kitzberger, T., Gemmill, B., Zobel, M., Vilà, M., Mitchell, C., Wilby, A., Daily, G.C., Galetti, M., Laurance, W.F., Pretty, J., Naylor, R., Power, A., Harvell, A., Potts, S., Kremen, C., Griswold, T. & Eardley, C. (2005) Biodiversity Regulation of Ecosystem Services. *Trends and conditions*, pp. 297–329.
- Fukami, T., Bezemer, T.M., Mortimer, S.R. & Van Der Putten, W.H. (2005) Species divergence and trait convergence in experimental plant community assembly. *Ecology Letters*, **8**, 1283–1290.
- Gourlet-Fleury, S. & Houllier, F. (2000) Modelling diameter increment in a lowland evergreen rain forest in French Guiana. *Forest Ecology and Management*, **131**, 269–289.
- Hérault, B., Bachelot, B., Poorter, L., Rossi, V., Bongers, F., Chave, J., Paine, C.E.T., Wagner, F. & Baraloto, C. (2011) Functional traits shape ontogenetic growth trajectories of rain forest tree species. *Journal of Ecology*, **99**, 1431–1440.
- Hérault, B., Ouallet, J., Blanc, L., Wagner, F. & Baraloto, C. (2010) Growth Responses of Neotropical Trees to Logging Gaps. *Journal of Applied Ecology*, **47**, 821–831.
- Hérault, B. & Piponiot, C. (2018) Key drivers of ecosystem recovery after disturbance in a neotropical forest. *Forest Ecosystems*, **5**, 2.
- Hubbell, S.P. (2001) *The Unified Neutral Theory of Biodiversity and Biogeography*. Princeton University Press.
- Kunstler, G., Falster, D., Coomes, D.A., Hui, F., Kooyman, R.M., Laughlin, D.C., Poorter, L., Vanderwel, M., Vieilledent, G., Wright, S.J., Aiba, M., Baraloto, C., Caspersen, J., Cornelissen, J.H.C., Gourlet-Fleury, S., Hanewinkel, M., Hérault, B., Kattge, J., Kurokawa, H., Onoda, Y., Peñuelas, J., Poorter, H., Uriarte, M., Richardson, S., Ruiz-Benito, P., Sun, I.F., Ståhl, G., Swenson, N.G., Thompson, J., Westerlund, B., Wirth, C., Zavala, M.A., Zeng, H., Zimmerman, J.K., Zimmermann, N.E. & Westoby, M. (2016) Plant functional traits have globally consistent effects on competition. *Nature*, **529**, 204–207.
- Lawton, R.O. & Putz, F.E. (1988) Natural disturbance and gap-phase regeneration in a wind-exposed tropical cloud forest. *Ecology*, **69**, 764–777.
- Li, Y. & Shipley, B. (2018) Community divergence and convergence along experimental gradients of stress and disturbance. *Ecology*, **99**, 775–781.
- Marcon, E. (2015) Practical Estimation of Diversity from Abundance Data. *HAL archives-ouvertes*, p. 9.
- Mason, N.W.H., De Bello, F., Mouillot, D., Pavoine, S. & Dray, S. (2013) A guide for using functional diversity indices to reveal changes in assembly processes along ecological gradients. *Journal of Vegetation Science*, **24**, 794–806.
- Molino, J.F. & Sabatier, D. (2001) Tree diversity in tropical rain forests: a validation of the intermediate disturbance hypothesis. *Science*, **294**, 1702–1704.
- Norden, N., Angarita, H.A., Bongers, F., Martínez-Ramos, M., Granzow-de la Cerda, I., van Breugel, M., Lebrija-Trejos, E., Meave, J.A., Vandermeer, J., Williamson, G.B., Finegan, B., Mesquita, R. & Chazdon, R.L. (2015) Successional dynamics in Neotropical forests are as uncertain as they are predictable. *Proceedings of the National Academy of Sciences*, **112**, 8013–8018.
- Norden, N., Chazdon, R.L., Chao, A., Jiang, Y.H. & Vélchez-Alvarado, B. (2009) Resilience of tropical rain forests: Tree community reassembly in secondary forests. *Ecology Letters*, **12**, 385–394.
- Perronne, R., Munoz, F., Borgy, B., Reboud, X. & Gaba, S. (2017) How to design trait-based analyses of community assembly mechanisms: Insights and guidelines from a literature review. *Perspectives in Plant Ecology, Evolution and Systematics*, **25**, 29–44.
- Piponiot, C., Sist, P., Mazzei, L., Peña-Claros, M., Putz, F.E., Rutishauser, E., Shenkin, A., Ascarrunz, N., de Azevedo, C.P., Baraloto, C., França, M., Guedes, M.C., Coronado, E.N., D'Oliveira, M.V., Ruschel, A.R., da Silva, K.E., Sotta, E.D., de Souza, C.R., Vidal, E., West, T.A. & Hérault, B. (2016) Carbon recovery dynamics following disturbance by selective logging in amazonian forests. *ELife*, **5**, e21394.
- Podani, J., Ricotta, C. & Schmera, D. (2013) A general framework for analyzing beta diversity, nestedness and related community-level phenomena based on abundance data. *Ecological Complexity*, **15**, 52–61.
- Putz, F.E., Zuidema, P.A., Synnott, T., Peña-Claros, M., Pinard, M.A., Sheil, D., Vanclay, J.K., Sist, P., Gourlet-Fleury, S., Griscom, B., Palmer, J. & Zagt, R. (2012) Sustaining conservation values in selectively logged tropical forests: the attained and the attainable. *Conservation Letters*, **5**, 296–303.
- Rutishauser, E., Hérault, B., Petronelli, P. & Sist, P. (2016) Tree Height Reduction After Selective Logging in a Tropical Forest. *Biotropica*, **48**, 285–289.
- Schwartz, G., Falkowski, V. & Peña-Claros, M. (2017) Natural regeneration of tree species in the Eastern Amazon: Short-term responses after reduced-impact logging. *Forest Ecology and Management*, **385**, 97–103.

- Sebbenn, A.M., Degen, B., Azevedo, V.C., Silva, M.B., de Lacerda, A.E., Ciampi, A.Y., Kanashiro, M., Carneiro, F.d.S., Thompson, I. & Loveless, M.D. (2008) Modelling the long-term impacts of selective logging on genetic diversity and demographic structure of four tropical tree species in the Amazon forest. *Forest Ecology and Management*, **254**, 335–349.
- Svenning, J.C. & Wright, S.J. (2005) Seed limitation in a panamanian forest. *Journal of ecology*, **93**, 853–862.
- van Buuren, S. & Groothuis-Oudshoorn, K. (2011) mice: Multivariate imputation by chained equations in r. *Journal of Statistical Software*, **45**, 1–67.
- Vinson, C.C., Kanashiro, M., Harris, S.A. & Boshier, D.H. (2015) Impacts of selective logging on inbreeding and gene flow in two Amazonian timber species with contrasting ecological and reproductive characteristics. *Molecular Ecology*, **24**, 38–53.
- Violle, C., Navas, M.L.L., Vile, D., Kazakou, E., Fortunel, C., Hummel, I. & Garnier, É. (2007) Let the concept of trait be functional! *Oikos*, **116**, 882–892.
- Wagner, F., Hérault, B., Stahl, C., Bonal, D. & Rossi, V. (2011) Modeling water availability for trees in tropical forests. *Agricultural and Forest Meteorology*, **151**, 1202–1213.
- Wright, I.J., Reich, P.B., Westoby, M., Ackerly, D.D., Baruch, Z., Bongers, F., Cavender-Bares, J., Chapin, T., Cornelissen, J.H.C., Diemer, M., Flexas, J., Garnier, E., Groom, P.K., Gulias, J., Hikosaka, K., Lamont, B.B., Lee, T., Lee, W., Lusk, C., Midgley, J.J., Navas, M.L., Niinemets, &, Oleksyn, J., Osada, N., Poorter, H., Poot, P., Prior, L., Pyankov, V.I., Roumet, C., Thomas, S.C., Tjoelker, M.G., Veneklaas, E.J. & Villar, R. (2004) The worldwide leaf economics spectrum. *Nature*, **428**, 821–827.