

Preregistration

The effects of wolves on moose presence and abundance in south-central Sweden.

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Study Information

Title	The effects of wolves on moose presence and abundance in south-central Sweden.
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Description	<p>Predator-prey interactions often have strong effects on lower trophic levels in a given environment. For example, upon the re-establishment of wolves in a national park in North America, aspen trees saw markedly increased recruitment (Ripple et al., 2014). In general, apex predator reintroduction is often considered to benefit ecosystems by limiting grazer populations and their consumption of lower trophic levels, and by providing meals for scavengers (Ripple & Beschta, 2004).</p> <p>To date, most research on the effects of predator reintroduction has been conducted in natural areas minimally affected by human activities (Kuijper et al., 2016). However,</p>
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due to climate-driven relocation and controlled reintroduction programs, modern and future predator populations may become established in areas of strong anthropogenic influence (Kuijper et al., 2016). Consequently, when hoping to predict the effects of future predator reintroduction events, it is important to study areas with nearby human activity.

Recent studies in Sweden have found that wolf reintroduction does not impart biodiversity benefits to lower trophic levels, contrasting observations made in American National Parks. For example, Gicquel, Sand, Månsson, Wallgren, & Wikenros (2020) reported that wolf reintroduction slightly increased moose browsing damage to Scots Pine trees in Sweden, as did Ausilio, Sand, Månsson, Mathisen, & Wikenros (2021). By surveying moose pellets, Ausilio, Sand, Månsson, Mathisen, & Wikenros (2021) also found that wolf presence and wolf territory establishment had slightly positive effects on moose presence and abundance. Other factors, including distance from roads (an indirect measure of human activities), had much stronger effects on moose populations than wolves (Ausilio, Sand, Månsson, Mathisen, & Wikenros, 2021).

Future populations of predators will continue to establish themselves in areas of anthropogenic influence. It is important to critically investigate the effects that they may, or may not, have on prey populations and lower trophic levels in these areas. To increase certainty in their paradigm-questioning results, we will replicate Ausilio, Sand, Månsson, Mathisen, & Wikenros (2021)'s research, conducting parallel analyses on the original study area over subsequent years, and asking whether or not wolves affect moose presence and abundance.

Hypotheses

We hypothesize that:

- 1) Wolf presence will moderately increase probability of moose presence, and overall moose abundance.
- 2) Time since wolf territory establishment will positively effect probability of moose presence, and overall moose abundance.
- 3) 'Distance from roads' will effect moose presence and abundance more strongly than wolf presence or time since wolf territory establishment.

Design Plan

Study type and design This study is a replication of original research by Ausilio et al., described in *Ecological Effects of Wolves on Anthropogenic Landscapes: The Potential for Trophic Cascades is Context-Dependent* (Ausilio, Sand, Månsson, Mathisen, & Wikenros, 2021).

Our hypotheses will be tested using a form of regression analysis. Data on 6 predictor variables and one response variable will be collected. Through modelling, we will predict the influence that the predictor variables have on both a) moose presence, and b) moose abundance. Both outcomes will be informed by response variable values.

Data collection The study will be conducted using data retrieved from various government and private sector institutions beginning in 2021 and ending in 2034, to replicate the 13-year original study by Ausilio et al. The study area will be the approximately 100,000 km² range of south-central Sweden's breeding range of wolves, replicating the region assessed in the original study. The predominant trees in this landscape are Scots pine, Norway spruce, and birch. The average moose winter density in the study range is approximately 1.3 per km² (Zimmermann, Nelson, Wabakken, Sand, & Liberg, 2014). Variables to be collected include: moose pellet counts, tree cover, roads, and wolf presence and time since wolf territory establishment. The sample size will be determined by the data available within the study region. To replicate the original study as closely as possible, approximately 10,350 circular sample plots (area = 38 m²) will be surveyed, with the number sampled per year dependent on data availability.

Variables Variable data will be collected within the study area of south-central Sweden. Data sources will be consistent with those used in the original study in order to maintain as close a replication study as possible. Seven variables will be use in statistical analysis of the effect of wolves on the presence and abundance of moose.

- 1) *Moose pellet counts*: Moose pellet counts will be used to gather presence and absence data on moose. Yearly surveys are conducted between the months of May and September throughout south-central Sweden (Fridman et al., 2014). The data source for this variable is the Swedish National Forest Inventory.

- 2) *Forest age stages*: There are four forest age stages considered when moose pellet count surveys are conducted. These stages classified as clear-cut, young, thinned, and mature. The data source for this variable is the Swedish National Forest Inventory.
- 3) *Pine proportion*: Sample plots, the same as those used for moose pellet count surveys, will be surveyed to assess tree cover. The cover (m^2) of lodgepole pine and Scots pine, as a proportion of all tree cover, will be generated as an index of food availability. The data source for this variable is the Swedish National Forest Inventory.
- 4) *Distance to the nearest forest road (km)*: Distance (km) between each plot and the nearest forest road, as classified by the database, will be calculated as a proxy representing human influence on the landscape. The data source for this variable is the Swedish national road database from the Swedish Transport Administration.
- 5) *Distance to the nearest main road (km)*: Distance (km) between each plot and the nearest main road, as classified by the database, will be calculated as a proxy representing human influence on the landscape. The data source for this variable is the Swedish national road database from the Swedish Transport Administration.
- 6) *Wolf presence*: Presence or absence of a wolf territory, as determined by monitoring obtained by snow tracking, DNA-samples, and GPS locations of collared wolves. A buffered radius will be used based on the average wolf territory size in Scandinavia (Liberg et al., 2011). The data source for this variable is the National wolf-monitoring system from County Administrative Boards.
- 7) *Time since wolf territory establishment*: A continuous range from 1 year to 27 years since establishment, as noted by annually conducted surveys. The data source for this variable is the National wolf-monitoring system from County Administrative Boards.

Analysis Plan

**Statistical models -
should add criteria
for inferences**

In the original data (Ausilio, Sand, Månsson, Mathisen, & Wikenros, 2021), the response variable was over-dispersed and had a large number of zeros (90.68%). We anticipate that we will collect similar data, and thus we plan to use a zero inflated negative binomial and hurdle model to investigate the effect of wolves and roads on moose. Hurdle models are composed of two components - zero inflated and conditional. The zero-inflated component runs a generalized linear mixed model (GLMM) with a binomial distribution and logit link using moose pellet presence/absence as the response variable. The conditional component of hurdle modelling runs a GLMM with truncated negative binomial distribution and log link using moose pellets ($n \geq 1$) as the response variable. For both the zero-inflated and conditional components of the hurdle model, TAXAR (year) is set as a random effect. To validate our data analysis procedures, we ran our scripts using simulated data (Figure 1).

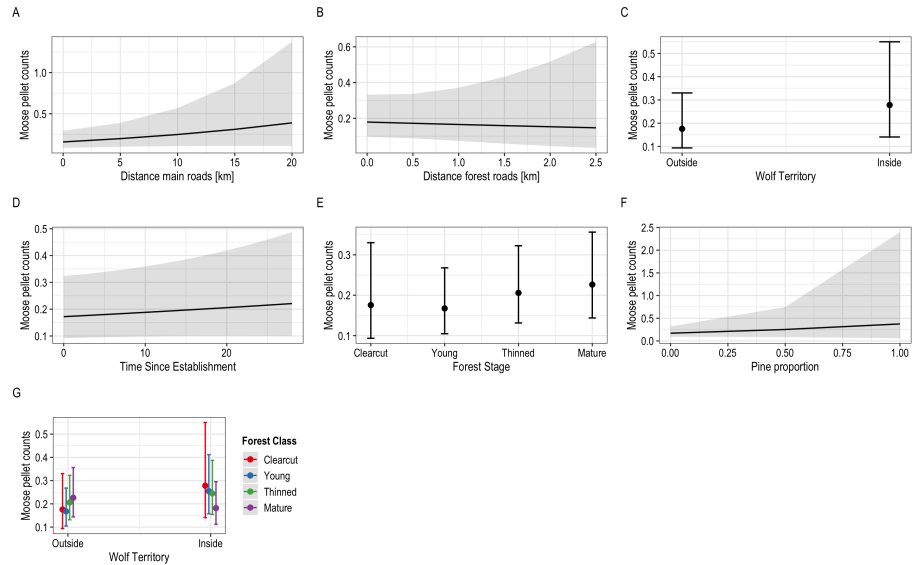


Figure 1: Marginal effects plot for predicted moose pellet counts in relation to (A) distance to main roads (km); (B) distance to forest roads (km); (C) wolf presence; (D) years since wolf territory establishment; (E) forest stage; (F) pine proportion; and (G) the interaction between wolf presence and forest stage. Shaded regions indicate 95% confidence intervals.

Exploratory analyses Enter your response here.

References - We have different citations for the original paper currently

- Ausilio, G., Sand, H., Månsson, J., Mathisen, K. M., & Wikenros, C. (2021). Ecological effects of wolves in anthropogenic landscapes: The potential for trophic cascades is context-dependent. *Frontiers in Ecology and Evolution*, 8, 481. doi:[10.3389/fevo.2020.577963](https://doi.org/10.3389/fevo.2020.577963)
- Fridman, J., Holm, S., Nilsson, M., Nilsson, P., Ringvall, A., & Ståhl, G. (2014). Adapting National Forest Inventories to changing requirements the case of the Swedish National Forest Inventory at the turn of the 20th century. *Silva Fennica*, 48(3). doi:[10.14214/sf.1095](https://doi.org/10.14214/sf.1095)
- Gicquel, M., Sand, H., Månsson, J., Wallgren, M., & Wikenros, C. (2020). Does recolonization of wolves affect moose browsing damage on young Scots pine? *Forest Ecology and Management*, 473, 118298. doi:[10.1016/j.foreco.2020.118298](https://doi.org/10.1016/j.foreco.2020.118298)
- Kuijper, D. P. J., Sahlén, E., Elmhagen, B., Chamaillé-Jammes, S., Sand, H., Lone, K., & Cromsigt, J. P. G. M. (2016). Paws without claws? Ecological effects of large carnivores in anthropogenic landscapes. *Proceedings of the Royal Society B: Biological Sciences*, 283(1841), 20161625. doi:[10.1098/rspb.2016.1625](https://doi.org/10.1098/rspb.2016.1625)
- Liberg, O., Aronson, Å., Sand, H., Wabakken, P., Maartmann, E., Svensson, L., & Åkesson, M. (2011). Monitoring of wolves in Scandinavia. *Hystrix, the Italian Journal of Mammalogy*, 23(1). doi:[10.4404/hystrix-23.1-4670](https://doi.org/10.4404/hystrix-23.1-4670)
- Ripple, W. J., & Beschta, R. L. (2004). Wolves and the ecology of fear: Can predation risk structure ecosystems? *BioScience*, 54(8), 755–766. doi:[10.1641/0006-3568\(2004\)054\[0755:WATEOF\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2004)054[0755:WATEOF]2.0.CO;2)
- Ripple, W. J., Estes, J. A., Beschta, R. L., Wilmers, C. C., Ritchie, E. G., Hebblewhite, M., ... Wirsing, A. J. (2014). Status and ecological effects of the world's largest carnivores. *Science*, 343(6167), 1241484. doi:[10.1126/science.1241484](https://doi.org/10.1126/science.1241484)
- Zimmermann, B., Nelson, L., Wabakken, P., Sand, H., & Liberg, O. (2014). Behavioral responses of wolves to roads: Scale-dependent ambivalence. *Behavioral Ecology*, 25(6), 1353–1364.

doi:[10.1093/beheco/aru134](https://doi.org/10.1093/beheco/aru134)