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## **Landscape Ecology: A Helicopter View**

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### **Authors' contributions**

*This work was carried out in collaboration among all authors. Author KUE is the main author, structured the write-up. Author CBE is the main internal editor. Author TNO managed the literature searches. All authors read and approved the final manuscript.*

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### **ABSTRACT**

This article reviews some of the main areas of landscape ecology. It includes the ideas and views of authors to the knowledge and understanding of the interplay between spatial heterogeneity and ecological processes, both in terrestrial and aquatic habitats at their varying scales. Moreover, the ecological consequences of landscape disturbance and landscape fragmentation are reviewed. Man and fire are discussed as the main agents of alternation of the natural landscape both at local and global scales. Studies on landscape restoration and management that emphasize the integration of various stakeholders for effective landscape restoration and management for sustainable biodiversity and socioeconomic benefits to man are as well reviewed. This paper further highlights the integration of the innovative remote sensing and Geographic Information System (GIS) technologies in the generation and analyses of spatial data. The current integration of population genetics in landscape ecology is also visited in this review.

**Keywords:** *Landscape ecology; heterogeneity; fragmentation; remote sensing; GIS; landscape genetics.*

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## 1. INTRODUCTION

Heterogeneity in ecological systems has been observed and described by scientists for quite a long time. However, an explicit focus on understanding spatial heterogeneity, revealing its myriads abiotic and biotic causes and its ecological consequences emerged in the 1980s as landscape ecology developed and spatial data analysis methods became more widely available [1]. Landscape ecology, a term coined by German biogeographer, Carl Troll and elaborated in 1950 [2], arose from the European traditions of regional geography and vegetation science and was motivated by the new perspective offered by aerial photography. Landscape ecology has since been defined in various ways [3,4,5], but common to all definition is focus on understanding the reciprocal interplay between spatial heterogeneity and ecological processes [6]. In other words, the fundamental difference between landscape ecology and broader ecological theory is that the later does not explicate the importance of space. The goal of a landscape ecologist is to understand and describe landscape structure; how this structure influences the movement of organisms, material, or energy across the landscape, and how and why landscape structure changes over time. Landscape ecological approaches are not limited to land, but are also applied in aquatic and marine ecosystems [7].

A landscape structure can be quantified by describing characteristics of patches, such as their number, size, shape, position, and composition [4]. Patch a term fundamental to landscape ecology, is defined as the relatively homogeneous area that differs from its surroundings [4]. Patches are the basic unit of the landscape that change and fluctuate, a process called patch dynamics. Landscape patches have a boundary between them which can be defined or fuzzy [8]. The quantification of spatial heterogeneity is necessary to elucidate relationships between ecological processes and spatial patterns, thus the measurement, analysis, and interpretation of spatial patterns receive much attention in landscape ecology. A number of studies have related landscape patterns to variable sets that include both biophysical and socioeconomic factors and their surrogates. Interactions between landownership and landscape position have emerged as strong determinants of land cover patterns and changes [9,10].

Landscape spatial structure is important in understanding the influence of habitat fragmentation on population survival. [11] referred to fragmentation as simply the disruption of continuity. Habitat fragmentation implies a loss of habitat, reduced patch size and an increasing distance between patches, although this can create new habitat. Local extinctions of fragmented populations are common. It is apparently clear that researches in landscape ecology have broadened perceptions of the causes and consequences of spatial heterogeneity and their varying scales, and this has influenced restoration and management of both the natural and human-inhabited landscapes.

Landscape ecology studies often employ remote sensing data together with field measurements, and undertakes geospatial analyses (using Geographic Information Systems, GIS) or simulation modeling [1]. These tools play a critical role in the data acquisition and processing components of spatial analysis [12]. Also, the development of landscape ecology as a discipline has been particularly simulated by technological developments in remote sensing and GIS.

## 2. LANDSCAPE HETEROGENEITY AND BIODIVERSITY

Landscape heterogeneity has long been considered a key determinant of biodiversity [13,14] Landscape ecologists emphasize how organisms use resources that are spatially heterogeneous and how they live, reproduce, disperse, and interact in landscape mosaic. Understanding the influence of large and small heterogeneity on species distribution and abundance is one of the major foci of landscape ecology research [15]. Studies have been carried on to understand the effects of landscape composition and landscape configuration (i.e., how is it spatially arranged) on biodiversity. Previous studies have reported contrasting associations between landscape heterogeneity and species richness, from positive to negative, and the negative effect is often reported to be the result of landscape fragmentation [16,17,14]. Simulation studies also suggest that changes in landscape composition are likely to have a greater effect on population persistence.

Analyses conducted at multiple scales have demonstrated the importance of landscape context for a wide range of taxa [18,19,20],

although the influence may be less if the focal habitat is abundant and well connected [21]. Much has been learned from studies that have evaluated factors that explain variation in the landscape. Patch size has a strong effect on edge and interior species but is negligible for generalist species [22]. However, local habitat conditions may be inadequate to explain species presence or abundance. Many studies have demonstrated that habitat connectivity is scale dependent; that is, whether a given pattern of habitat is connected depends on the mobility of the species and the pattern of the habitat [23]. Organisms may respond to multivariate habitat heterogeneity at multiple scales, and identification of the factors and scales that best explain variation in the presence or abundance of organisms remains a key goal in landscape ecology.

Two mechanisms which operate at different time scales can initiate varied response of organisms to landscapes. The first mechanism is evolutionary adaptation of species to landscape heterogeneity or homogeneity over longtime scales [24,25], and the second is selective extinction of species inhabiting homogenous landscapes, such as forests or grasslands, by anthropogenic land-cover change over shorter time scale [26,27]. The latter mechanism is called an 'extinction filter' [28]. The magnitude of human-induced environmental change at the global scale is considered to be enormous. Furthermore, the main force driving the global transformation of the biosphere is human population growth, together with increasing resource consumption and sociocultural change. Extinction is a natural event and from geological perspective routine, that is, most species that have ever lived have gone extinct during different geological times.

### 3. HETEROGENEITY AND HIERARCHY OF SCALE

Scale is an important concept in both natural and social sciences, and has been defined in several different ways [29,30]. From landscape perspective, scale refers primarily to grain (or resolution) and extent in space or/and time. Scale may be absolute (measured in spatial or times units) or relative (denoted as ratio). Scale may be the observer's measuring stick or viewing window size, a spatial or temporal characteristic of an ecological pattern or process, or a fundamental framework in which diverse ecological phenomena can be more effectively

studied and understood individually and collectively.

Scaling, on the other hand usually defined as the process of extrapolating or translating information from one scale to another, including scaling up and scaling down [31,32]. Scale and scaling have become buzzwords in ecology in recent years as the research emphasis of the field has shifted from local to increasingly broader scales [33]. As spatial heterogeneity becomes a major theme in wide range of ecological studies, the concepts of scale, scaling and hierarchy become increasingly crucial in ecology in general. A literature survey by [33] reveals that the number of papers that contain words, "scaling", "hierarchy", "hierarchies", or "hierarchy theory", has increased exponentially in four of the major ecology journals since 1930s. This confirms, and is indicative of, the rapidly rising awareness of the importance of scale and hierarchy of ecologists.

In dealing with scale in ecological studies and applications, three related but distinctive tasks stand out. First, we need to appreciate and understand how changing the scale of observation affects research results and their interpretation. Second, if ecological systems are multiple-scaled or hierarchically structured, identifying characteristic scales and hierarchical levels become extremely important for understanding and predicting ecological phenomena. Third, theories, models, procedures for extrapolating information across scales need to be developed for understanding and managing heterogeneous landscapes. Although simple "scaling laws" do exist in ecology, extrapolating information over a wide range of scales may often require a hierarchical approach.

### 4. AQUATIC ECOSYSTEMS IN LANDSCAPE ECOLOGY

Despite the traditional focus on land, landscape ecologists have not entirely ignored aquatic systems. Generally, they have considered rivers as element of a landscape mosaic; rivers linked with their surroundings by boundary dynamics; and rivers as internally heterogeneous landscapes (see Fig. 1). This is the viewpoint that is fostered by remote sensing geographical information systems (GIS), or landscape mapping [34]. Water is an increasing valuable resource to humans in most parts of the world, and rivers and streams have been the focus of human culture and activities since the dawn of civilization.

Rivers are used as transport corridors, as water sources for settlements or farmlands, as fisheries, as waste disposal conduits, and so on, and rivers in a landscape can be differentiated according to these differing uses [34]. But again, the view of a river is a channel separated from the other elements of the landscape by its edges. The river has neither dynamics nor internal structure of its own. Another view of rivers considers them as functional parts of landscapes that are connected by boundary flows, by exchanges of materials, organisms, energy, or information across boundaries between adjacent landscape elements [35]. Although terrestrial ecologists may think of such flows in rivers simplistically, in terms of downstream hydrology alone, any riverine ecologist knows that a wide array of exchanges occurs across river boundaries. All of the structural and functional features that can be used to characterize a river as a part of a broader terrestrial landscape also apply to the landscape within a river. In many cases, this within-river landscape is also quite dynamic, varying in patch composition and configuration in response to changes in hydrologic flow regimes [36].

Recognizing that patches differ in quality is the first step in transforming a descriptive map of a mosaic into something that can represent the spatial component of ecological processes. Several studies in streams and rivers illustrate the effects of variation in patch quality. For

example, [37] documented that larval chironomids and adult copepods were more abundant in patches of leaves than in sand patches in a fine scale streambed mosaic. The recognition that patches in a stream differ in quality and that organisms respond to these spatial variations is not new, of course in 1920s, [38] noted the selection of patches of high velocity, where respiration is facilitated by lotic invertebrates. Patch quality changes over time, especially in such dynamic systems as streams and rivers. [39] documented the shifting nature of patch quality by experimentally determining how the distribution of stream invertebrates among patches varied under different flow regimes.

## 5. WETLANDS IN LANDSCAPE ECOLOGY

A wetland is a distinct ecosystem that is flooded by water, either permanently or seasonally, where oxygen-free processes prevail [40]. The primary factor that distinguishes wetlands from other land forms or water bodies is the characteristic vegetation of aquatic plants, adapted to the unique hydric soil. Wetlands play a number of functions, including water purification, water storage, processing of carbon and other nutrients, stabilization of shorelines, and support of plants and animals. Wetlands are also considered the most biologically diverse of all ecosystems, serving as home to a wide range of plant and animal life. Wetlands occur naturally on every continent [41].



**Fig. 1. A river surrounded by the other elements of the landscape by its edges**

Physico-chemical constituents of the water column (e.g. pH, nutrients, conductivity) are regarded as potentially important determinants of biotic assemblage composition in wetlands and other freshwater ecosystems e.g. [42,43]. More specifically, in the south-western Cape, South Africa, [44] established that physico-chemical factors exert a significant structuring effect on invertebrate assemblage composition in temporary depression wetlands. Alternation of these factors through anthropogenic disturbance has potential to mediate ecosystem changes in these wetlands, through bottom-up effects on biota such as aquatic invertebrates and amphibians. Studies have focused on permanent wetland types, from which various authors have reported significant effects of habitat fragmentation on an array of individual physico-chemical variables including turbidity, pH, nutrients, conductivity and dissolved oxygen [45,46,47], while very few studies have specifically addressed relationships between terrestrial habitat transformation and physico-chemical conditions within temporary wetlands [48].

## 6. DISTURBANCE AND LANDSCAPE HETEROGENEITY

Studies of disturbance and succession continue to generate new understanding about the interactions between ecological processes and landscape pattern [1]. A disturbance is any relatively discrete event in time that disrupts ecosystem, community, or population structure and changes resources, substrate availability, or the physical environment [49]. The type of disturbance is broad, encompassing such diverse events as wildfire, insect outbreaks, hurricanes, coral bleaching, and floods [50]. Disturbance may either increase or decrease heterogeneity, may enhance or inhibit the spread of disturbance. Disturbance can play an important role in these dynamics, by initiating cycles of secondary succession and generating opportunities for communities of long-lived organisms to reorganize in alternative configuration.

The relationship between disturbance and heterogeneity in a landscape is complex and depends on the scale of disturbance and important underlying environmental gradient [51]. The occurrence or effects of disturbance may depend on the system's state before disturbance occurred. Thus, disturbances are particularly interesting in landscape ecology because they both respond to and create spatial heterogeneity.

In all cases, disturbances result in functional, structural, biological, and biogeochemical legacies [50,52,53], a template left on the ecosystem with which subsequent disturbances may interact in some fashions. The legacies take the form of altered functioning (e.g., rapid growth of surviving individuals, altered nutrient cycling), dead material, residual (e.g., survivors) and regenerating individuals and communities, and unique spatial patterns in community and unique spatial patterns in cover types, ranging from undisturbed to highly disrupted.

A number of studies have documented significant influences of landscape heterogeneity on the spread or effects of disturbances. Effects of hurricane, wind events, and fires can vary with spatial location on the landscape. Researches have frequently found a strong influence of landforms on these effects. For example, the severity of hurricanes on vegetation varies with the exposure of the sites [54]. In general, landscape position influences disturbance when the disturbance has a distinct directionality or locational specificity such that some locations are exposed more than others.

## 7. LANDSCAPE MODELS

Quantification of ecological processes and formulation of the mathematical expressions that describe those processes in computer models has been a cornerstone of landscape ecology research and its application. Consequently, the body of publications on simulation models in landscape ecology has grown rapidly in recent decade. This trend is also evident in the subfield of forest landscape ecology, particularly in relation to the topic of disturbance [55].

Spatially, when a study is expanded to the order of  $10^3$ - $10^6$  ha, experimental studies become limited and additional complexities such as environmental heterogeneity and natural disturbances may further complicate the study. Thus, computer models become useful tools for landscape scale experiments [56,57]. With modeling techniques, knowledge of physiological factors and their effects on the modeled processes and interactions within a population system can be explicitly represented using mathematical equations and logical sequences. Those data can be used in models to deduce results, especially at broad spatial and temporal scales, that cannot otherwise be investigated [58,59].

Forest landscape models share common features, including: simulation (1) forest vegetation response at large spatial and temporal scales (e.g. in excess of 100,000 ha and 100 years) and (2) the outcomes of repeated, stochastic spatial processes (e.g. seed dispersal, fire, wind, insects, diseases, harvests, and fuel treatments). Depending on the model's purpose and design limitations, they may differ in the key ecological processes incorporated, the extent to which mechanistic details are simulated for each process, and the type and scope of applications [60].

Over the last 15 years, we have seen rapid development in the field of forest landscape modeling, fueled by both technological and theoretical advances. Forest landscape models have benefited greatly from technological advances, including increased computing capacity, the development of GIS, remote sensing, and software engineering. Ecological processes and their interactions in forest landscape models can be represented by well-designed computer software [61]. The core of landscape ecology provides a conceptual basis for forest landscape modeling from a theoretical perspective: the interaction of spatial patterns and ecological processes under various spatiotemporal scales, theories of disturbance, and equilibrium and non-equilibrium approaches to vegetation and ecosystems [60].

## 8. LANDSCAPE FRAGMENTATION

Landscape fragmentation is increasingly considered an important environmental indicator in the fields of sustainable land use and biodiversity. The habitat loss and fragmentation associated with human use in many regions is well described in landscape ecology and conservation biology [62,63,64]. A dominant effect of increasing landscape fragmentation is increase in patch areas, with resulting declines in population density and species richness, and significant alternations to community composition, species interactions and ecosystem functioning. For species restricted to the original type of habitat, fragmentation means disintegration into small, spatially disjunct patches, separated by land which is unsuitable to reproduce or find food or shelter [65].

Habitat fragmentation is one of the most cited threats to species extinction and an ensuing loss of biological diversity, making it perhaps the most important contemporary conservation issue [66].

[11] referred to fragmentation as simply the disruption of continuity. Increases in the number of endangered species and the patterns of global extinction are in part the result of habitat loss and landscape fragmentation. These threats isolate populations by disrupting the migration and dispersal of species vital life processes for sustaining population health and species persistence, and alter ecological elements such as fire, stream flow, drought, and species interactions.

## 9. LANDSCAPE MOSAIC AND CONNECTIVITY

Landscape connectivity is a product of boundaries defined by the distinct communities, called patches, which form its element. Landscape connectivity can be defined as the degree to which the landscape facilitates or impedes movement between resources patches [67]. Landscape connectivity is considered a crucial element of landscape structure because of its importance to population survival. Movement of individuals, materials, nutrients, energy, or disturbances through a landscape involve more than boundary configuration, permeability, and context. If a landscape is indeed a mosaic of patches of different types, then these movements are affected by how the patches are in the mosaic. Connectivity includes both structural connectivity (the physical arrangements of disturbance and/or patches) and functional connectivity (the movement of individuals across contours of disturbance and/or among patches) [68,69].

Connectivity is a fundamental feature of most natural landscapes [70]. The theoretical works of [71,72,73,74] coupled with considerable empirical evidence [75,76] suggest that physically connected patches of habitat on the landscape are one means to support and maintain a great richness of indigenous species and system integrity [77]. Baudry and Merriam [78] noted how connectivity is a parameter receiving value from processes moving across landscape elements while connectedness is demonstrated through structural landscape features. Thus, landscape connectivity can be defined as the degree to which the landscape facilitates or impedes movement between resources patches [67].

The role of environmental corridors in facilitating seed dispersal and the movement of animals and other phenomena such as disease has been

reviewed in depth [79]. The degree to which a landscape is connected determines the amount of dispersal there is among patches, which influences gene flow, local adaptation, extinction risk, colonization probability, and the potential for organisms to move as they cope with climate change [80,81].

## 10. HUMAN-INDUCED LANDSCAPES

Landscapes are continuously altered by humans to make them better adapt to their social, economic or ecological needs. With such developing landscapes, the ecosystem network is bound to change in harmony, thus the capacity of landscapes to sustain biodiversity is questioned. Agricultural dynamics and high rate of urbanization are the chief human-induced factors that adversely affect the natural landscapes.

## 11. AGRICULTURAL LAND USE

The world's natural landscapes have greatly been interfered with by man, and the concomitant effects are progressively increasing in tempo, as a result of man's inherent dominion over the natural ecosystems. The consequences of world ecosystem structure of man's cultural

progression from a predator to a domesticator and farmer, and the associated diffusion of agricultural concepts throughout the world, have really been profound [82]. Exploitative and seriously destructive agricultural systems have been recently extended in scales in the recent times by the widespread adoption of evermore intensive methods of food crop production, especially monoculture. As a result of forest clearance, the extent to which forest ecosystem interrelationships are further altered depends mainly on the type of agricultural system adopted and on the human population density and these have altered the landscape, with serious ecological consequences for flora and fauna world-wide, although at different degrees and rates among regions.

Agricultural dynamics and associated alternations in the structure of habitat patches affect species composition and distribution in the landscape. There is empirical evidence that the proportion of land uses and their spatial arrangement can affect the long-term dynamics of bird species in agro-landscape [17]. Piha et al. [83] observed from their research that landscape structure and agricultural land use were the principal determinants of the bird assemblage.



**Fig. 2. o-MARCELLUS-SHALE-DRILL-PAD-facebook (1)**

*Man's interference in the natural forests through urbanization has led to alternation in the structure of the natural landscape, with serious consequences for species composition and distribution*



## 12. URBANIZATION EXPANSION

The problems of landscape pattern change and region ecological risk under urbanization expansion have drawn the increasing attention of the ecologists. From the view point of landscape ecology, urbanization is the process in which landscape/cover changes from natural landscape which is mainly made of water, soil, and vegetation to man-made landscape which is mainly composed of cement, asphalt, chemical materials and metals [84,85]. Urbanization is mainly viewed from social and economic viewpoints while paying little or no attention to influence on ecological impacts. Urbanization causes profound changes in the ecological functioning of the landscape and gradually results in a changing spatial structure, that is, forms new landscape patterns (see Fig. 2) Studies have shown relationship between urbanization and the influence on the landscape. [86] showed that significant positive relation existed between urbanization intensity and regional ecological risk in Chinese coastal city. Meanwhile, the change of land-use/cover landscape pattern could be interpreted as the changes of patch shape, area quality, and spatial combination [87,88].

## 13. THE INFLUENCE OF FIRE ON LANDSCAPE PATTERNS

The influence of fire in controlling landscape composition and structure both at local and global scales is apparent. Fire contributes to the

structure of landscapes ranging from tropical savannas to boreal forests. However, fire plays a particularly prominent role in regions with a Mediterranean climate. Mediterranean landscapes are dynamic systems that undergo temporal changes in composition and structure in response to disturbances such as fire [89] (see Fig. 3). Fire regimes are determined by complex interactions between climate, land use, vegetation attribute and the pattern of ignition [90,91]. However, at larger temporal spatial scales, fire regimes appear to be more determined by climatic variability with periods of high fire risk linked to particular weather conditions accounting for more fire events [92]. Fire frequency and severity vary with the amount, spatial distribution and conditions of available fuels, as well as moisture, temperature and ignition sources [93].

For thousands of years, natural and man-caused fires have shaped landscapes across the globe, affecting structure and connectivity of biomass, liberating and recycling stored nutrients and converting natural landscapes for human uses [94]. Conversion of forests and shrublands to agricultural or pastoral uses through burning, suppression of fire in fire-adapted landscapes, and continual expansion of the interface between urban population and flammable forests are directly influencing fire in boreal [95,96,97], temperate [98,99] and grassland systems, and introducing fire to desert [100] and tropical [101,102,103,104] biomes. Research conducted by [89] on the influence of topography and fire in



**Fig. 3.**The deleterious influence of fire on the forest landscape



controlling landscapes composition and structure in Sierra de Gredos (Central Spain) revealed that the impact of fire on landscape patterns was high variable among regions due to the different regeneration abilities to main landscapes, the topographic constraints and the fire histories of each region. Although climate and ecosystem properties continue to be important component of global fire regimes, in recent decades a shift from climate controlled fire regimes towards human-controlled fire regimes has taken place over much of the globe [105,106].

#### 14. LANDSCAPE LEGACIES

The persistent influence of land-use history on explaining the vegetation and biogeochemical characteristics of contemporary ecosystems has become increasingly apparent [107]. Landscape legacies reveal the evolutionary history of contemporary landscapes. A landscape history exposes the evolutionary patterns of a specific landscape by revealing its ecological stages, cultural periods, and keystone processes. Such history can be a valuable tool as it has the potential to improve description, prediction, and prescription in landscape planning [108]. Ecologists, conservationists, and natural resource policy makers now recognize that the legacies of land-use activities continue to influence ecosystem structure, or even longer after those activities have ceased. As a result, environmental history emerges as an integral part of ecological science and conservation planning.

Recognition of the importance of land-use history and its legacies in most ecological systems has been a major factor driving the recent focus on human activity as a legitimate and essential subject of environmental science. At stand to landscape scales, differences in land-use history strongly controls modern vegetation patterns [109]. Studies from Harvard forest show that another enduring legacy of land use in New England is the homogenization of tree species composition at a regional scale [110]. Because of the broadly similar history of agriculture, logging, and reforestation, a subset of the regional tree flora with disturbance-adapted life history traits has been favoured. The result is a shift from pre European patterns of forest variation that correspond to subtle gradients in regional climate to a more homogenous condition [111].

Land-use legacies have different relevance in context of varying landscape and management objectives. Recognition that the history of

disturbance shapes the long-term structure, composition, and function of most ecosystems and landscapes can increase the effectiveness of management [112]. In contrast, ignoring historical legacies may lead to the development of ill-conceived conservation and management schemes [113]. Historical perspectives aid the interpretation of landscapes that we wish to manage and contribute to the identification of realistic goals and appropriate tools and approaches to achieve those ends [114].

#### 15. FOREST LANDSCAPE RESTORATION

Understanding how to restore and maintain spatial ecological processes can be considered the most important element of landscape ecology [115,116]. Landscape restoration is defined as a planned process to regain ecological integrity and enhance human well-being in deforested or degraded landscape. Forest landscape restoration recognizes that forest restoration has social and economic functions [117]. It aims to achieve the best possible compromise between meeting with conservation goals and the needs of rural communities [118]. Forest landscape restoration (FLR) includes both the planning and implementation of measures to restore degraded forests within the perspective of the wider landscapes.

Forest can be restored in wide range of circumstances, but degraded sites within protected areas are a high priority, especially where some climax forest remains as a seed source within the landscape. Landscape restoration is a specialized form of reforestation, but it differs from conventional tree plantations in that its primary goals are biodiversity recovery and environmental protection [119,120]. As human pressure on landscapes increases, forest restoration will most commonly be practiced within a mosaic of other forms of forest management to meet the economic needs of local people.

It is becoming increasingly recognized that landscape restoration requires the involvement of multiple stakeholders operating in multiple sectors, and at multiple scales. Forest restoration is any inclusive process, which depends on collaboration among a wide range of stakeholders including local community, government officials, non-governmental organizations, scientists and funding agencies. This type of stakeholders' involvement in design, planning and decision-making of forest

landscape restoration programme is increasingly referred to by the term 'landscape governance' [121,122]. During the last decade the concept of landscape governance has become generally accepted as referring to the multi-stakeholder process of negotiation and decision making about policies and programmes for effective conservation and sustainable use of forests, and for implementing the planned measures within spatial landscape units [123,124]. Despite the general acceptance there is still divergence in the way landscape governance is perceived and implemented in different restoration programmes [125]. [126] identify different modes of governance with respect to three different dimensions of politics, polity and policy.

## **16. RIVERINE LANDSCAPE RESTORATION**

The intensive use and alternation of riverine landscapes by humans have led to severe degradation of river-flood plain, especially in highly industrialized countries. Recent water-related regulations and legislation focusing on high standards of ecological integrity back efforts to restore or rehabilitate these systems [126]. Restoration tests the feasibility of recreating complex ecosystems from more simple and degraded states, thereby presenting a major challenge to ecological science. Therefore, close cooperation between practitioners and scientists would be beneficial, but most river restoration projects are currently performed with little or no scientific involvement. Ecological restoration is a recent discipline that should be conducted scientifically and rigorously to move from trial-and-error process to a predictive science to increase its success and the self-sustainability of restored ecosystems [127].

Restoration of important ecological processes often implies improving connectivity of the stream. For example, longitudinal and lateral connectivity can be enhanced by restoring fluvial dynamics on flood-suppressed rivers and by increasing water availability in rivers subject to water diversion or withdrawal, thereby increasing habitat and species diversity. Restoring links between surface and ground water flow enhances vertical connectivity and communities associated with the hyporheic zone.

## **17. FOREST LANDSCAPE MANAGEMENT STRATEGIES**

To maintain biodiversity over long period of time, natural landscape patterns of heterogeneity and

other emergent ecological processes must be recognized and addressed at the level of regional ecological planning [127,128,77]. To effectively manage heterogeneous landscape to maintain biodiversity, we need to understand the mechanisms of such variability in the associations between landscape heterogeneity and species richness at multiple spatial scales [129,130,131]. One reason for such context-dependent patterns may be the difference among biomes in the sensitivity of species pools to fragmented landscapes, with lower sensitivity in the temperate zones of the Northern Hemisphere than in Oceania and tropical regions [26,132].

Acting on the spatial arrangements of land uses to increase heterogeneity of landscapes without altering the proportion of land uses, could help to reconcile production and biodiversity in agro-landscapes. Modifying the proportion of land uses, through the conversion of some intensive land uses into extensive ones often involves a trade-off for production [133].

Recent studies emphasize the need for integration of various stakeholders for effective and result-oriented forest management. This method is termed integrated landscape management. Integrated landscape management is a way of managing a landscape that brings together multiple stakeholders, who collaborate to integrate policy and practice for their different land-use objectives, with the purpose of achieving sustainable landscape [134,135]. Reed et al. [135] elaborated systematic ways to implement integrated landscape management. According to them five elements of landscape management describe the implementation cycle of a landscape approach, namely (1) interested stakeholders in the landscape come together for cooperative dialogue and action in a multi-stakeholder platform (2) they undertake a systematic process to exchange information and discuss perspective to achieve a shared understanding of the landscape conditions, challenges and opportunities (3) this enables collaborative planning to develop an agreed action plan (4) stakeholders then implement the plan, with attention to maintaining collaborative commitments (5) and finally, monitoring for adaptive management and accountability, feeds into subsequent rounds of dialogue and the design of new collaborative action. These five elements can be presented graphically in a circle, similar to the project management cycle.

## 18. FIRE MANAGEMENT STRATEGIES

Management strategies to restore forest landscapes are often designed to concurrently reduce fire risk. Both fire risk reduction forest restoration objectives will benefit from spatially coordinated landscape level planning among landowners [136]. Understanding the role and the relative weight of different factors leading to changes in fire regime is thus of critical importance to anticipate the fate of biodiversity, or to implement management strategies aiming at mitigating or modulating the impact of fires arising from such changes [137].

Fire suppression has long been used to mitigate fire, especially in the Mediterranean basin. Fire suppressing is a direct anthropogenic activity altering fire regime, even though its significance has been a point of debate [138]. Fire suppression efforts are aimed at limiting fire impact by decreasing fire severity or decreasing fire effective fire size. In fact, because a fire suppression policy is conducive to fuel build up and more intense fires, it can ultimately contribute to larger and more severe fires [139,140]. The ultimate goal of fire management is to modify the fire regime, which results from the interactions between ignitions and the fire environment, that is, topography, weather and vegetation (fuels).

Because of the debates on the usefulness of fire suppression as a fire management strategy, especially in the Mediterranean basin, researches show that fire suppression gives room to more fire occurrence. Fire management policies in the Mediterranean basin rely heavily on fire suppression and do not sufficiently address the socioeconomic and land management issues behind the inception and spread of fires [139]. As the Mediterranean becomes a more fire-prone environment due to continuing rural abandonment and climate change, it will be advisable to consider the management fires, both as a fuel treatment and an ecological process [141].

Innovational strategies aimed at fire management have emphasized the efficiency of integrated approach. In line with this, [142] proved that both ecological benefits and socioeconomic gains can be achieved using fire-smart forest management approach. They defined fire-smart forest management as an integrated approach primarily based on fuel treatments through which the socioeconomic

impacts of fire are minimized while its ecological benefits are maintained and maximized, by lowering ignition likelihood and fire behaviour potential, fire suppression capacity is increased and forests and landscapes become more resistant to fire spread and more resilient to its occurrence.

## 19. REMOTE SENSING DATA AND TECHNIQUES IN LANDSCAPE STUDIES

As with many areas in physical geography and interrelated fields, remote sensing is a key technology for quantifying landscape patterns and processes in the twenty-first century. The interpretation and classification of data generated from remote sensors has matured as a discipline with growing specialist literature. Advanced tools and algorithms can be used to pre-process (radiometric corrections, atmospheric corrections, image registration), analyze (pixel classification), post-process (generalize, error report) and visualize imagery. These tools play a critical role in the data acquisition and processing components of spatial analysis [143]. The remote sensing community has developed rich and varied traditional, cartographic products [144].

Satellite remote sensing imagery revolutionized land cover mapping with the introduction of Landsat Thematic Mapper in the 1970s, which provided global coverage at high temporal and spatial resolution. Over the years, sensor and data processing technologies have progressed significantly and the methods for mapping the Earth's surface have evolved. The availability of continuous spatial information provided by remote sensing and advancement in spatial analysis techniques supported by powerful personal computing began to address the need for synthesized natural resources information at an operational level. Methods and techniques have been gradually developed into standards for terrestrial needs and have begun to adapt to coastal marine needs [145,146]. This ability to collect information remotely is advantageous to coastal marine research as access to the sea floor by boat is usually complicated and costly [147].

Satellite imagery in digital format allows for the acquisition of environmental data and land occupation patterns and features over large areas [148]. Sensors in satellites record multispectral data from different wave bands in digital format. Different features of the terrain

reflect differently in each waveband, allowing for their recognition in the images. The main limitations of satellite images are cloud cover and resolution. Even with the best resolution available (pixels < 30 m), it is not possible to see houses, to adequately classify some types of agricultural practice, or to locate some breeding sites [148]. Some of the problems may be circumvented by using satellite navigation receivers, which enables a user to obtain an instantaneous three-dimensional position, anywhere on the earth, at any time, under any weather condition.

## 20. GEOGRAPHIC INFORMATION SYSTEM (GIS) IN LANDSCAPE ECOLOGY

One of the most exciting and rapidly growing technologies for the 1990s is that of Geographic Information System (GIS). The computerized retrieval, manipulation, analysis and display of geographic information allow experts in a variety of geographic disciplines to improve their effectiveness and efficiency when addressing location-based problems and issues [149]. A Geographic Information System (GIS) is a computer-based tool for mapping and analyzing things that exist and events that happen on the earth. GIS integrates common database operations, such as query and statistical analysis, with the unique visualization and geographic analysis offered by maps (<http://www.hgac.cog.tx.us/geography/cep/whatis.html>). It consists of a powerful set of automated tools for collecting, retrieving, analyzing and communicating spatial data.

Using GIS to explore the spatial relationships of animal populations is a relatively new field for ecologists. GIS is well-established in habitat-based studies of animal populations to analyze remotely-sensed databases [150] and as a predictive tool for animal or plant species distributions [151,152]. In addition, GIS is now used to create databases, manipulate spatially explicit surfaces to represent specific parameters, and to displace spatial relationships through simulation modeling, hydrologic constructs, and species relationships [153,154].

Geographic Information System (GIS) technology can help establish cross-sectoral communication by providing not only powerful tools for storage and analysis of multisectoral spatial and statistical data, but also by integrating databases of different sectors in the same format, structure and map projection in the GIS system [155].

## 21. LANDSCAPE GENETICS

Landscape genetics is an interdisciplinary field combining tools and concepts from both landscape ecology and population genetics to relate landscape structure to patterns and genetic variations [156,157]. It broadly encompasses any study that analyses plant or animal population genetic data in conjunction with data on the landscape features and matrix quality where the sampled population lives. This allows for the analyses of microevolutionary processes affecting the species in light of landscape spatial patterns, providing a more realistic view of how populations interact with their environments. [157]. The field has evolved tremendously since Manel's landmark paper in 2003 [156], moving from descriptive assignment tests used to define population boundaries to a more explicit analytical framework including landscape variables as predictors in genetic models [158]. A recent review categorized the types of questions and methods used in landscape genetic studies compared to papers published in its predecessor fields and found that most of self-identifying landscape genetic studies fall more into the realm of population genetics (e.g. using terms such as "genetic", "gene", and "barrier" [161]. Landscape genetics attempts to determine which landscape features are barriers to dispersal and gene flow, how human-induced landscape changes after the evolution of populations, the source-sink dynamics of a given population, and how diseases or invasive species spread across landscapes [159].

## 22. SUMMARY AND CONCLUSIONS

Landscape genetics differs from the fields of biogeography and phylogeography by providing information on finer temporal and spatial scales (i.e. at the level of individual genetic variation within a population). Because it focuses on sampling individuals, landscape genetics has the advantage of not having to subjectively define discrete populations prior to analysis. Genetic tools are used to detect abrupt genetic differences between individuals within a population and statistical tools are used to correlate these genetic discontinuities with landscape and environmental features. As a tool, molecular genetics can make hard-to-observe processes visible and thus should be useful for landscape ecologists working on species whose movements are hard to track. For example, genetic markers have been used to estimate both contemporary and historical effective

population size [160], assess sex-biased dispersal [161,162,163], identify population bottlenecks [164], and characterize meta-population dynamics [165]. Genetics can be used to quantify actual functional connectivity either directly or indirectly, and thus provides the means to test hypotheses about how aspects of the intervening landscape matrix support or inhibit dispersal and gene flow [166,167,168]. The results of landscape genetics studies have potentially important applications to conservation biology and land management practices [156].

As a new and fast growing interdisciplinary field with explicitly identified practices, landscape genetics has been subject to a number of flaws in both study design and interpretation [169]. [159] identified four common pitfalls of landscape genetics research that should be targeted for correction. These include assuming gene flow is always advantageous, over-generalizing results, failing to consider the practices that affect the genetic structure of populations, and mistaking quantitative methods for robust study design [169].

## COMPETING INTERESTS

Authors have declared that no competing interests exist.

## REFERENCES

1. Turner MG. Landscape ecology in North America: Past, present and future. *Ecology*. 2005;86:1967-74165.
2. Tucker JM, Schwartz MK, Truex RL, Allendorf FW. Historical and contemporary DNA indicate fisher decline and isolation occurred prior to the European settlement of California. *Plos One*. 2012;7(12): e52803.
3. Troll C. Die geographische Landschaft und ihre Erforschung, *Studium Generale*. 1950;3:163-81.
4. Urban DL, O'Neill RV, Shugart HH. Landscape ecology. *Bioscience*. 1987;37: 119-127.
5. Forman TT. Some general principles of landscape and regional ecology. *Landscape Ecology*. 1995;10(3):133-142.
6. Bell SS, Robbins BO, Jensen SL. Gap dynamics in a sea grass landscape. *Ecosystems*. 1999;2:493-504.
7. Urban DL. Landscape ecology. *Encyclopedia of Environmetrics*. 2006;3.
8. Ward JV, Malard F, Tockner K. Landscape ecology: A framework for integrating pattern and process in river corridors. *Landscape Ecology*. 2002;7(Suppl.):35-45.
9. Sanders J, Harris LO. Landscape ecology: a top-down approach. Lewis publishers, Boca Raton, Florida, USA; 2000.
10. Spies TA, Ripple WJ, Bradshaw GA. Dynamics and pattern of a managed coniferous forest landscape in Oregon. *Ecol. Appl*. 1994;4:555-568.
11. Wear DN, Bolstad P. Land-use changes in southern Appalachian landscapes: spatial analysis and forecast evaluation. *Ecosystems*. 1998;1:575-94.
12. Lord JM, Norton DA. Scale and the spatial concept of fragmentation. *Conservation Biology*. 1990;4(2):197-202.
13. Jensen JR. Remote sensing of the environment: An earth resource perspective (Upper Saddle River, N.J.: Prenmtice-Hall); 2000.
14. Tews JU, Grimm BV, Tielborger K, Wichmann MC, Schwager M et al. Animal species: diversity driven by habitat heterogeneity/diversity: the importance of keystone structures. *J. Biogeogr*. 2004;31: 79-192.
15. Fahrig L, Baudry J, Brotons L, Burel FG, Crist TV, et al. Functional landscape heterogeneity and animal biodiversity in agricultural landscapes. *Ecol. Lett*. 2011; 14:101-112.
16. Thoronton DH, Branch LC, Sunquist ME. The influence of landscape, patch and within-patch factors on species presence and abundance: A review of focal patch studies. *Landscape Ecology*. 2011;26(1): 7-18.
17. Andren H. Effects of habitat fragmentation on birds and mammals in landscapes with different proportions of suitable habitats. *Oikos*. 1994;7:355-366.
18. Benton TG, Vickery JA, Wilson JD. Farmland biodiversity: is habitat heterogeneity the key? *Trends Ecology and Evolution*. 2003;18:182-188.
19. Lindenmayer DB, Cunningham RB, Pope ML, Donnelly CF. The response of arboreal marsupials in landscape context: A large scale fragmentation study. *Ecol. Appl*. 1999;9:594-611.
20. Pearson SM, Turner MG, Wallace LL, Romme WH. Winter habitat use by large ungulates following fires in northern Yellowstone National Park. *Ecol. Appl*. 1995;5:744-55.
21. Stoner KJL, Joern A. Landscape versus local habitat influences to insect

- communities from tallgrass prairie remnants. *Ecol. Appl.* 2004;14:1300-20.
22. Miller JR, Dixon MD, Turner MG. Response of avian communities in large-river flood plains to environmental variation at multiple scales. *Ecol. Appl.* 2004a;14: 1394-10.
23. Bender DJ, Contreras TA, Fahrig L. Habitat loss and population decline: A meta-analysis of the patch size effect. *Ecology.* 1998;79:517-33.
24. Vos CC, Verboom J, Opdam PFM, TerBraak CJF. Toward ecological scaled landscape indices. *Am., Nat.* 2001;157:24-4124. Futuyma DJ, Moreno G. The evolution of ecological specialization. *Ann. Rev. Ecol. Syst.* 1988;19:207-233.
25. Futuyma DJ, Moreno G. The evolution of ecological specialization. *Ann. Rev. Ecol. Syst.* 1988;19:207-233.
26. Sandel B, Arge L, Dalsgaard B, Davies R, Gaston K, et al. The influence of late quarterly climate change velocity on species endemism. *Science.* 2011;334: 660-664.
27. Henle K, Davies KF, Kleyer M, Margules C, Settle J. Predictions of species sensitivity to fragmentation. *Biodiverse Conserv.* 2004;13:207-251.
28. Clavel J, Julliard R, Devictor V. Worldwide decline of specialist species: Towards a global functional homogenization? *Front. Ecol. Environ.* 2011;9:222-228.
29. Balmford A. Extinction filters and current resilience: The significance of past selection pressures for conservation biology. *Tends Ecol. Evol.* 1996;11:193-196.
30. Marceau DJ. The scale issue in social and natural science. *Canadian Journal of Remote Sensing.* 1994;25:247-366. Marcucci DJ. Landscape history as a planning tool. *Landscape and Urban Planning.* 2000;49(1):67-81.
31. Van Gardngen PRG, Foody GM, Curran, PJ. (eds). *Scaling-up from cell in landscape.* Cambridge University Press. Cambridge.1997;166.
32. Wu J. Hierarchy and scaling: extrapolating information along a scaling ladder. *Canadian Journal of Remote Sensing.* 1999;25:367-380.
33. King AW, Johnson AR, O'Neill RV. Transmutation and functional representation of heterogeneous landscapes. *Landscape Ecology.* 1991; 5(4);239-253
34. Wu J, Qi Y. Dealing with scale in landscape analysis. An overview. *Geographic Information Services.* 2000; 6(1):1-5.
35. Wiens JA. Riverine landscape: taking landscape ecology into the water. *Forest Water Biology.* 2002;47(4):501-515.
36. Hansen AJ, di Castri F. (eds.) *Landscape boundaries: Consequences for biotic diversity and ecological flows.* Springer-Verlage, Berlin; 1992.
37. Malard F, Tockner K, Dole-Oliver MJ, Ward JV. A landscape perspective of surface-subsurface hydrological exchanges in river corridors. *Freshw. Biol.* 2002;47(4):621-640.
38. Palmer MA, Lettenmaier DP, Poff NL, Postel SL, Richter B, Warner R. Climate change and river ecosystems: Protection and adaptation options. *Environmental Management.* 2009;44(6):1053-1068.
39. Rutter F. Bermer kungen uber den saver stoffghall der Gewasser und dessen respiratoris chen Wert. *Naturwissenschaften.* 1926;14:237-1239.
40. Lancaster J. Geometric scaling of microhabitat patches and their efficacy as refugia during disturbance. *Journal of Animal Ecology.* 2000;69:442-457.
41. Keddy PA. *Wetland ecology: Principles and conservation* (2<sup>nd</sup> ed). New York. Cambridge University Press; 2010.
42. Hollis T, Bedding J. Can we stop the wetlands from drying up? *New Scientist;* 1994.
43. Blinn DW, Haise SA, Pinder AM, Shiel RJ, Mc Rae JW. Diatom and microinvertebrate communities and experimental determinants in the Western Australian wheat belt: A response to salinization. *Hydrobiologia.* 2004;528:229-248.
44. Batzer DP, Cooper R, Wissinger SA. *Wetland animal ecology.* In: Batzer, D.P., Sharitz RR. (eds). *Ecology of freshwater and estuarine wetlands.* Berkeley, USA: University of California Press. 2006;242-284.
45. De Rock ER. Status and ecology of temporary wetlands in the Western Cape, South Africa: PhD Thesis, Laboratory of Aquatic Ecology and Evolutionary Biology, Catholic University of Leuven, Belgium. 2008;143.
46. Declerck S, De Bie T, Ercksen D, Hampel H, Schrijvers S et al. Ecological characteristics of small farmland ponds. Associations with landuse practices at



- multiple spatial scales. *Biological Conservation*. 2006;131:523-532.
47. Skagen SK, Melcher CP, Haukos DA. Reducing sedimentation of depressional wetlands in agricultural landscapes. *Wetlands*. 2008;28:594-604.
48. Akasaka M, Takamura N, Mitsuhashi H, Kadono Y. Effects of land use on aquatic macrophyte diversity and water quality of ponds. *Fresh Water Biology*. 2010;155: 909-922.
49. Bird M, Day J, Rebeto A. Physico-chemical impacts of terrestrial alien vegetation on temporary wetlands in a sclerophyllous sand fynbos ecosystem. *Hydrobiologia*. 2013;711:115-128.
50. Pickett STA, White PS. (eds.). *The ecology of natural disturbance and patch dynamics*. Academia. 1985;472. New York.
51. Peters DPC, Lugo AE, Chapin FS, Pickett, STA. Cross-system comparisons elucidate disturbance complexities and generalities. *Online library.wiley.com*; 2011.
52. Risser PG, Karr JR, Forman RTT. *Landscape ecology: Directions and approaches*. Spec. Publ. No. 2, 111. Nat. Hist. Surv., Champaign. 1984;111.
53. McLauchlan KK, Higuera PE, Gavin DG, Perakis SS, Mack MC, Alexander H, Enders SK. Reconstructing disturbances and their biogeochemical consequences over multiple timescales. *BioScience*. 2014;64(2):105-116.
54. Seidl R, Rammer W, Spies TA. Disturbance legacies increase the resilience of forest ecosystem structure, composition, and functioning. *Ecological Applications*. 2014;24(8):2063-2077.
55. Boose ER, Foster DE, Fluet M. Hurricane impacts to tropical and Temperate forest landscapes. *Ecol. Monogr*. 1994;64:369-400.
56. Perera AH, Sturtevant BR, Buse LJ. Simulation modeling of forest landscape disturbances: An overview. In *Simulation Modeling of Forest Landscape Disturbances*. 2015;1-15. Springer, Cham.
57. Mladenoff DJ. LANDIS and forest landscape models. *Ecol. Model*. 2004;180: 7-19.
58. Shifley SR, Thompson ER, Dijak WD, Larson MA, Millsaugh JJ. Simulated effects of forest management alternatives on landscape structure and habitat suitability in the mid western United States. *For. Ecol. Manage*. 2006;36:1740-1748.
59. Mladenoff DJ, Baker WL. *Advances in spatial modeling of forest landscape change: Approaches and applications*. Cambridge University Press, Cambridge, UK; 1999.
60. Urban DL. Modeling ecological processes across scales. *Ecology*. 2005;86:1996-2006.
61. He HS. *Forest Landscape models: definitions, characterization, and classification*. *Forest Ecology and Management*. 2008;254(2008):484-498.
62. Heilman GE, Strittholt JR, Slosser NC, Dellasala DA. Forest fragmentation of the conterminous United States; assessing forest intactness through road density and spatial characteristics. *Bioscience*. 2002; 52:411-422.
63. Riitters K, Wickham J, O'Neil R, Jones B, Smith E. Global-scale patterns of forest fragmentation. *Conserv. Ecol*. 2000;4(2). Available:<http://www.cnsecol.org/vol14/iss2/art3/>
64. Saunders D, Hobbs RJ, Margules CR. Biological consequences of ecosystem fragmentation: A review. *Conserv. Biol*. 1991;5:18-32.
65. Opdam P, Van Aeldoom R, Schotman AGM, Kalkhoven J. Population responses to landscape fragmentation. In: *Landscape Ecology of a Stressed Environment*. 1993;147-171.
66. Taylor PD, Fahrig L, Henein K. and Merriam, G. Connectivity is a vital element of landscape structure. *Oikos*. 1993;68:571-572.
67. Brooks CP. A scalar analysis of landscape connectivity. *Oikos*. 2003;102:433-439.
68. Baguette M, Blanchet S, Legrad D, Stevens VM, Turlure C. Individual dispersal landscape connectivity and ecological networks. *Biological Reviews of the Cambridge philosophical Society*. 2013;88:310-326.
69. Merriam G. Connectivity: A fundamental ecological characteristic of landscape pattern. In: Brandt J, Agger P (eds) *Proceedings of first international seminar on methodology in landscape ecology research and planning*, vol I. Roskilde Universitessforlag GeoRue, Roskilde, Denmark. 1984;5-15.
70. Preston FW. The canonical distribution of commonness and rarity. *Ecology*. 1962; 43:185-215,410-432.

71. McArthur RH, Wilson EO. An equilibrium theory of insular zoogeography. *Evolution*. 1963;17:373-387.
72. McArthur RH, Wilson EO. *The theory of Island Biogeography*. Princeton University Press, Princeton; 1967.
73. Diamond JM. The island dilemma: lessons of modern biogeographical studies for the design of nature reserves. *Biological Conservation*. 1975;7:129-146.
74. Beier P. Determining minimum habitat area and habitat corridors for cougars. *Conservation Biology*. 1993;12:1241-1252.
75. Beier P. Dispersal of juvenile cougars in fragmented habitat. *Journal of Wildlife Management*. 1995;59:228-23.
76. Harris LD, Scheck J. From implications to applications: the dispersa corridor principle applied to wildlife corridors. In: Saunders D, Hobbs RJ. (eds.). *Nature conservation 2: The Role of Corridors*, Surely; 1991.
77. Baudry J, Merriam HG. Connectivity and connectedness: Functional versus structural patterns in landscapes. In: proceedings of the 2<sup>nd</sup> international seminar of the international association for landscape ecology (ed). *Connectivity in landscap: Eecology munstersche geographische arbeeiten*, Germany, Münster. 1988;43-51.
78. Forman RTT, Godron M. *Landscape Ecology*. John Wiley and Sons. New York; 1986.
79. Hodgson JA, Thomas CD, Wintle BA, Moilanen A. Climate change, connectivity and conservation decision-making: Back to basics. *Journal of Applied Ecology*. 2009; 46:964-969.
80. McRae BH, Hall SA, Beier P, Theobald DM. Where to restore ecological connectivity? Detecting barriers and quantifying restoratiuon benefits. *PLoS ONE*. 2012;7:e52604.
81. Echereme CB, Mbaekwe EI. Floristic diversity in the tropics: Ensuring sustainability of this natural heritage is a matter of life and death: A review. *International Journal of Advanced Research*. 2015;3(11):1466-1479.
82. Piha M, Tiainen J, Vespsalainem V et al. Effects of land use and landscape characteristics on avian diversity and abundance in a boreal agricultural landscape with organic and conventional farms. *Biological Conservation*. 2007; 140(1):50-61.
83. Al-Manni AA, Abdu ASA, Mohammed NA, Al-Sheeb AE. Urban growth and land-use change detection using remote sensing and geographic information system techniques in Doha city state of Qatar. *Arab Gulf Journal of Scientific Research*. 2007;25(4):190-198.
84. Merlotto A, Piccolo MC, Bertola GR. Land use/cover change at Necochea and Quequen cities, Buenos Aires Province, Argentina *Revista de Geografia Norte Grande*. 2012;53:159-176.
85. Zhou D, Shi P, Wu X, Ma J, Yu J. Effects of urbanization expansion on landscape pattern and region ecological risk in Chinese coastal city: A case study of Yantai city. *The Scientific World Journal*. 2014;9.  
[Article ID 821781]
86. Li Y, Zhu X, Sun X, Nang F. Landscape effects on environmental impact of bay-area wetlands under rapid urban expansion and development policy: A case study of Lianyungang China. *Landscape and Urban Planning*. 2010;94(3-4):218-227.
87. Zhou N, Zhao S. Urbanization process and induced environmental geological hazards in China. *Natural Hazards*. 2013;64(2):797-810.
88. Viedma O Influence of topography and fire in controlling landscape composition and structure in Sierra de Gredos (central Spain). *Landscape Ecology*. 2008;23:657. DOI: 10.1007/s10980-008-9228-5
89. Hessel AE. Pathways for climate change effects on fire: Models, data, and uncertainties. *Progress in Physical Geography*. 2011;35(3):393-407.
90. Moreira F, Viedma O, Arianoutsou M, Curt T, Koutsias N, Rigolot E, Mouillot F. Landscape-wildfire interactions in southern Europe: implications for landscape management. *Journal of Environmental Management*. 2011;92(10): 2389-2402.
91. Westerling AL, Bryant BP. Climate change and wildfire in California. *Climatic Change*. 2008;87(1):231-249.
92. O'Connor CD, Garfin GM, Falk AD, Swetnam TW. Human pyrogeography: A new synergy of fire, climate and people is reshaping ecosystems across the globe. *Geopgraphy Compass*. 2011;516(2011): 329-350.
93. Bowman DMJS et al. Fire in the earth system. *Science*. 2009;324(5926):481-484.

94. Schurr EAG et al. Vulnerability of permafrost global carbon cycle. *Bioscience*. 2008;58(8):701-714.
95. Schurr FM, Midgley GF, Rebelo AG, Reeves G, Poschod P, Higgins SI. Colonization and persistence ability explain the extent to which plant species fill their potential range. *Global Ecology and Biogeography*. 2007;16(4):449-459.
96. Wickland KP, Stiegl RG, Neff JC, Sachs T. Effects of permafrost melting on CO<sub>2</sub> and CH<sub>4</sub> exchange of a poorly drained black spruce lowland. *Journal of Geographical Research-Biogeosciences*. 2006;111(G2): P.G 02011.
97. Alcammo J, Floerke M, Maerker M. Europe. In: Parry ML, Canziani OF, Palutikof JP, van der Linden PJ, Hanson CE. (eds.). *Climate change: Impacts, adaptation and vulnerability. Contribution of working group II to the fourth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge, UK, Cambridge University Press. 2007;541-580.
98. Fields CB, Mortsch LD, Brklacich M, Forbes DL, Kovacs P, Patz JA, Demuth M; 2007.
99. North America. In: Parry ML, Canziani OF, Palutikof JP, van der Linden PJ, Hanson CE (eds.). *Climate change: Impacts, adaptation and vulnerability. Contribution of working group II to the fourth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge, UK, Cambridge University Press; 617-652.
100. Fischlin A, Midgley GF, Price JT, Leemans R, Gopal B, Turley C, Velichko AA. Ecosystems, their properties, goods and services. In: Parry ML, Canziani OF, Palutikof JP, van der Linden PJ, Hanson CE. (eds.). *Climate change: Impacts, adaptation and vulnerability. Contribution of working group II to the fourth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge, UK, Cambridge University Press. 2007;211-272.
101. Cochrane MA (ed). *Fire, land use, land cover dynamics, and climate change in the Brazilian Amazon*. In: *Tropical tree ecology: Climate change, land use and ecosystem dynamics*. Heidelberg, Germany, Springer-praxis. 2009;389-426.
102. Cochrane MA, Laurance WF. Synergisms among fire, land use, and climate change in the Amazonian. *Am Bio*. 2008;37:522-527.
103. Cox PM, Betts RA, Collins M, Harris PP, Huntingford C, Jones CD. Amazonian forest dieback under climate-carbon cycle projections for the 21st century. *Theoretical and Applied Climatology*. 2004; 78(1-3):137-15.
104. Page SE, et al. Tropical peatland fires in southeast Asia. In: Cochrane, M. (ed). *Tropical fire ecology*. Chichester. Springer. 2009;263-287.
105. Marlon JR, et al. Climate and human influences on global biomass burning over the past two millennia. *Nature Gloscience*. 2008;1(10):697-702.
106. Pechony O, Schindell DT. Driving forces of global wildfires over the past millennium and the forth coming century. *Proceedings of the National Academy of Science of the United States of America*. 2010;107(45): 19167.
107. Compton JE, Boone RD. Long-term impacts of agriculture on soil carbon and nitrogen in New England forests. *Ecology*. 2000;81:2314-233.
108. Marcucci DJ. Landscape history as a planning tool. *Landscape and Urban Planning*. 2000;49(1):67-81.
109. Zimmerman JKM, Aide M, Herrera L, Mayde R, Serrano M. Forest restoration recovery in abandoned tropical pastures on Puerto Rico. *For. Eco. Manage*. 1995; 77:77-86.
110. Foster DR, Knight DH, Franklin JF. *Landscape patterns and legacies resulting from large infrequent disturbances*. *Ecosystems*. 1998;1:497-510.
111. Fuller JL, Foster DR, McLachlan TS, Drake N. Impact of human activity on regional forest composition and dynamics in central New England. *Ecosystems*, 1998;1:76-95.
112. Swetnam TW, Allen CD, Betancourt JL. *Applied historical ecology: using the past to manage for the future*. *Ecological Application*. 1999;9:1189-1206.
113. Foster DR. Insights from historical geography to ecology and conservation: Census from the New England landscape. *Journal of Biogeography*. 2000;29:1269-1275.
114. Landres PB, Morgan P, Swanson FJ. Overview of the use of natural variability concepts in managing ecological systems. *Ecol. Appl*. 1999;9:1179-88.
115. Turner MG. *Landscape ecology: The effect of pattern on process*. *Annual Review of*

- Ecology and Systematics. 1989;20(1):171-197.
116. Harris LD, Hootor TS, Gergel SE. Landscape processes and their significance to biodiversity conservation. Population Dynamics in Ecological Space and Time. 1996;1:319-47.
117. Mansourian S, Vallauri D, Dudley N. (eds) In: Cooperation in landscapes: Beyond planting trees. Springer, New York; 2005.
118. Reitbergen-McCracken J, Maginnis S, Sarre A. The forest landscape restoration handbook. London. Earthscan; 2007.
119. Lamb David. Regreening the bare hills. Springer. 2011;547.
120. Stanturf JA. What is forest restoration? Restoration of boreal and temperate forests. Boca Raton. CRC Press. 2005;3-11.
121. van Oosten CJ. Restoring landscapes-governing place: A learning approach to forest landscape restoration. J. Sustain. For. 2013;2:659-676
122. Colfer GJP. Collaborative governance of tropical landscapes. Earthscan, London, UK; 2011.
123. Gorg C. Landscape governance: The politics of scale and natural conditions of places. Geoforum. 2007;38:954-966.
124. Van Oosten C, Gunarsop P, Koesoetjahjo I, Wiersum F. Governing forest landscape restoration: Case from Indonesia. Forest. 2014;5:1143-1162.
125. Treib O, Bähr H, Falkner G. Modes of governance: Towards a conceptual clarification. Journal of European Public Policy. 2007;14:1-20.
126. Jungwirth et al. Re-establishing and assessing ecological integrity in riverine landscapes. Freshwater Biology. 2002; 45(4):867-887.
127. Henry CP, Amoros C. Restoration ecology of riverine wetlands. I.A. Scientific base. Environmental Management. 1995;19:891-902.
128. Harris LD, Atkins K. Faunal movement corridors in Florida. Landscape Linkages and Biodiversity. 1991;117-38.
129. Fahrig L, Baudry J, Brotons L, Burel FG, Crist TO, Fuller RJ, Martin JL. Functional landscape heterogeneity and animal biodiversity in agricultural landscapes. Ecology Letters. 2011;14(2): 101-112.
130. Schindler S, Von Wehrden H, Poirazidis K, Wrba T, Kati V. Multi scale performance of landscape metrics as indication of species richness of plants, insect and vertebrates. Ecol. Indic. 2013; 31:41-48.
131. Morelli F, Pruscini F, Santolini R, Perna P, Benedetti Y, Sisti D. Landscape heterogeneity metrics as indicators of bird diversity: Determining the optimal spatial scales in different landscapes. Ecological Indicators. 2013;34:372-379.
132. Baldi A. Edge effects in tropical versus temperate forest bird communities: three alternative hypotheses for the explanation of differences. Acta Zool Hung. 1996;42: 163-72.
133. Sabatier R, Doyen L, Tichit M. Reconciling production and conservation in agrolandscapes: Does landscape heterogeneity help. Innovation and Sustainable Development in Agriculture and Food (ISDA). Montpellier, France; 2010.
134. Denier L, Scherr S, Shames S, Chatterton, P, Hovani L, Stam N. The little sustainable landscapes book. Oxford. Global Company Programme; 2015.
135. Reed J, Deakin E, Sunderland T. What are integrated landscape approaches and how effectively have they been implemented in the tropics: A systematic map protocol. Environmental Evidence. 2015;4:2.
136. Shinneman DJ, Palik B, Cornett MW. Can landscape-level explicit assessment of a northern-southern boreal forest landscape. Forest Ecology and Management. 2012; 274:126-135.
137. Brotons L, Aquilue N, de Caceres M, Fortin MJ, Fall A. How fire history, fire suppression practices and climate change affect wildlife regimes in Mediterranean landscapes, PLoS OME. 2013;8(5): e62392. DOI:10.1371/Journal.pone.0062392
138. Ward PC, Tithecott AG, Wotton BM. Reply: A re-examination of the effects of fire at fire suppression in the boreal forest. Can. T. For. Res. 2011;31:1467-1480.
139. Fernandez PM. Forest fires in Galicia (Spain): The outcome of unbalanced fire management. Journal of Forest Economics. 2008;14:155-157.
140. Pinol J, Castellnou M, Beven KJ. Conditioning uncertainty in ecological models: assessing the impact of fire management strategies. Ecological Modeling. 2007;207:34-44.
141. Fernandez PM. Forest fires in Galicia (Spain): The outcome of unbalanced fire

- management. *Journal of Forest Economics*. 2008;14:155-157
142. Hirsch K, Kafka V, Tymstra C, McAlpine R, Hawkes B, Stegenhuis H, Quintilio S, Gauthier S, Peck K. Fire smart forest management: A pragmatic approach to sustainable forest management in fire-dominated ecosystems. *The Forestry Chronicle*. 2001;77:1-7.
143. Jensen JR. Remote sensing of the environment: An earth resource perspective (Upper Saddle River, N.J.: Prenmtice-Hall); 2000.
144. Newton AC, Cayuela L, Echeverria C, Armesto JJ, Dio Cstillo RF et al. Towards integrated analysis of human impacts on forest biodiversity: Lessons from Latin America. *Ecology and Society*. 2009;14(2): 2.
145. Robbins BD, Bell SS. Seagrass landscapes: a terrestrial approach to the marine subtidal environment. *Trends in Ecology & Evolution*. 1994;9(8):301-304.
146. Hinsley SA, Hill RA, Ballamy PE, Harrison NM, Speakman JR, Wilson AK, Forns PN. Effects of structural and functional habitat gaps on wetland birds: Working harder for less. *Landscape Ecology*. 2008;23:615-626.
147. Baumstark RD. Remote sensing and spatial metrix for quantifying seagrass landscape changes: A study of the 2011 Indian River Lagoon Florida Seagrass Die-off events. Graduate Theses and Dissertation; 2008.  
Avaiable:<http://scholarcommons.usf.edu/etd/7124>
148. Chekuimo GH. Integrating Ecological Tools With Geographic Information Systems (GIS). (ISPRS 2008). 2008;3404-3409.
149. Huxhold WE. Long range planning for a multidisciplinary gis environment in an academic setting. In: URISA (1994). Urban and Regionary Information Association. 1994;732-744.
150. Johnson CA, Naiman RJ. The use of a geographic information system to analyze long-term landscape alteration by beaver. *Landscape Ecology*. 1990;4:5-9.
151. Scott JM, Davis F, Cauti B, Noss R, Butterfield B, Groves C, Anderson H, Caicco S et al. GAP analysis: A geographic approach to protection of biological diversity. *Wildlife Monographs* 1993;123.
152. Jensen JR, Narumalani S, Weatherbee O, Morris KS. Predictive Aquatic Macrophyte Modeling of a New Freshwater Lake Using GIS Techniques. In *acsm asprs annual convention american soc photogrammetry & remote sensing+ amer cong on*. 1992; 1:177-177.
153. Keller JK. Using aerial photography to model species-habitat relationships: The importance of habitat size and shape. . In: Mitchell RSCJ, Sheviak DJ. Leopold (eds). *Ecosystem Management: Rare species and significant habitats*. vol bull 471. New York, New York State Museum. 1990;34-46.
154. Aspinall R, Veitch N. Habitat mapping from satellite imagery and wildlife survey data using a bayesian modeling procedure in a GIS. *Photogrammetric Engineering and Remote Sensing*. 1993;59(4):537-543.
155. SDRN (Geographic Information Systems Group Environment and Natural Resources Service) *Geographic Information Systems in Sustainable Development* .FAO Research; 1999.
156. Manel S, Schwartz MK, Luikart G, Taberlet P. Landscape genetics: combining landscape ecology and population genetics. *Trends in Ecology and Evolution*. 2003;18(4):189-197.
157. Holderegger R, Wagner HH. A brief guide to Landscape Genetics. *Landscape Ecology*. 2006;2196:793-796.
158. Storfer A, Murphy MA, Evans JS, Goldgerg CS, Robinson S, Spear SF, Dezzani R, Delmelle E, Vierling L. Putting the landscape in landscape genetics. *Heredity*. 2007;98(3):128-142.
159. Storfer A, Murphy MA, Spear SF, Holderegger R, Waits LP. Landscape genetics: Where are we now? *Molecular Ecology*. 2010;10(17):3496-3514.
160. Chuicchi JE, Gibbs HI. Similarity of contemporary and historical gene flow among highly fragmented populations of an endangered rattlesnake. *Mol. Ecol*. 2010;19(24):5345-5358.
161. Goudet J, Perrin N, Waser P. Tests for sex-biased dispersal using bi-parentally inherited genetic markers. *Mol. Ecol*. 2002;11(6):1103-1114.
162. Wang Y, Lane A, Ding P. Sex-biased dispersal of a frog (*odorrana schmackeri*) is affected by patch isolation and resource limitation in a fragmented landscape. *Plos One*. 2012;7(10):e47683.
163. Vangestel C, Callens T, Vandomme V, Lens L. Sex-biased dispersal at different geographical scales in a cooperative

- breeder from fragmented rainforest. Plos One. 2013;8(8):e71624.
164. Tucker JM, Schwartz MK, Truex RL, Allendorf FW. Historical and contemporary DNA indicate fisher decline and isolation occurred prior to the European settlement of California. Plos One. 2012;7(12): e52803.
  165. Andreasen AM, Stewart KM, Longland WS, Beckmann JP, Forister ML. Identification of source-sink dynamics in mountain lions of the Great Basin. Mol. Ecol. 2012;21(23):5689-701.
  166. Jaquiere J, Broquet T, Hirzel AH, Yearsley J, Perrin N. Inferring landscape effects on dispersal from genetic distances: How far can we go? Mol. Ecol. 2011;20(4):692-705.
  167. Wang IJ, Savage WK, Shaffer HB. Landscape genetics and least-cost path analysis reveal unexpected dispersal routes in the California tiger salamander (*Ambystoma californiense*). Mol. Ecol. 2009;18(7):1365-1374.
  168. Lowe WH, Allendorf FW. What can genetics tell us about population connectivity? Mol. Ecol. 2010;19(23):5320.
  169. Richardson JL, Brady SP, Wang IJ, Spear SF. Navigating the pitfalls and promise of landscape genetics. Mol. Ecol. 2016;25: 849– 863.

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