

Methods for Reconstructing Paleo Food Webs

Tanya Strydom ¹; Baran Karapunar ²; Andrew P. Beckerman ¹; Alexander Dunhill ²

Abstract: Food webs represent the feeding relationships between species and can help infer ecosystem-level processes. Alongside the development of food web theory, methods for constructing food webs have been developed to infer species interactions when empirical data is lacking. Food web construction methods are diverse, each utilising different approaches to infer species interactions —such as the use of traits to infer mechanistic relationships vs using gut content as a proxy for species diets. These methods have distinct theories, mechanisms, and data requirements. In paleoecology, where direct evidence of feeding interactions are rare, food web construction methods are especially valuable and affords us the opportunity to make inferences about paleo communities beyond simply a record of species composition. However, the limitations of paleontological data (e.g., information of species traits is limited to that which can be preserved) restrict which methods can reliably be used. By considering both ecological theory and the constraints of what can be derived from the fossil record, we identify the methods best suited for the construction of paleo food webs. Specifically, we focus on how these methods differ in the networks they produce and what these networks can reveal about species interactions. In doing so we hope to clarify the ecological nuances of network prediction and help prevent the accidental misuse or misinterpretation of paleo food webs.

Keywords: food web, network construction

1 There has been a growing interest in looking at community responses to environmental changes across deep
2 time events as a means to help understand current and future biodiversity changes (Dillon et al., 2022; Kiessling
3 et al., 2019). Species interactions and the resulting networks have gained an interest in contemporary
4 settings as a means to help us to understand aspects of community composition and biodiversity (eg Thuiller
5 et al. (2024) and ??) and so it is perhaps unsurprising that there has been a growing interest in using paleo
6 food webs in a similar manner (*e.g.*, Dunhill et al., 2024 looked at...; Hao et al., 2025 looked at...; Yeakel et al.,
7 2014 looked at...). However, one of the core challenges and limitations of being able to effectively *use* food webs
8 is the challenge of *creating* them (Jordano, 2016), although this is a challenge within contemporary settings
9 it is compounded in paleo contexts as we are dependent on the fossil record (and the inherent limitation it
10 imposes) to infer interactions from. As a way to address the challenges with recording species interactions
11 there has been the development of a large number of models and tools that can be used to infer either species
12 interactions (see *e.g.*, Morales-Castilla et al., 2015; Pichler & Hartig, 2023; Strydom et al., 2021 for broader
13 reviews) or networks (see *e.g.*, Allesina et al., 2008). Although there has been the development of models and
14 tools that are specific for inferring paleo food webs (Fricke et al., 2022; Roopnarine, 2006; *e.g.*, Shaw et al.,
15 2024), it should be noted that these models only occupy a subset of the broader family of approaches that are
16 used to predict networks as they typically use only one mechanism for determining interactions (the feasibility
17 of the interaction being able to occur). Being able to only construct one ‘type’ of network means that we
18 are limited in the scope of questions that we can appropriately answer with those networks [see Strydom
19 in prep; Gauzens et al. (2025)]. However there is scope that models and tools that have been developed in
20 contemporary settings have the potential to be used for paleo settings (*e.g.*, Yeakel et al., 2014), which opens
21 the door for researchers to ask a broader and more complete range of questions about community responses
22 to environmental change.

23 Here we want to identify the differences between models that predict interactions, and models that predict
24 network structure. Specifically we want to look at 1) the structural difference between all models (*i.e.*, do we
25 see a difference in the distribution of links between networks that have the same number of nodes?) and 2)
26 the identity of pairwise links between species pairs (*i.e.*, do different models differ in which links are present
27 (or absent) between species pairs?) Additionally we want to establish if using networks that are constructed
28 using different models will change the downstream inferences that are made for this we use the work from
29 Dunhill et al. (2024) as a case study.

30 **1 Contextualising paleo web prediction within the contemporary**

31 **toolbox**

32 Although there is an evolving body of work focused on the development of food web prediction tools it is
33 important that we understand how the underlying philosophy on which a model was built will result in
34 different assumptions being embedded within the network [Strydom in prep]. Broadly we can think about
35 models that are nested within two different schools of thought (and thus methodological approaches, Figure 1),
36 models that focus on assessing the *mechanistic* feasibility of an interaction being able to occur between two
37 species or models that are more closely married to specific bodies of ecological *theory* - such as niche theory
38 or foraging ecology.

39 [Figure 1 about here.]

40 Models that have specifically been developed in the paleo space tend to be mechanistic models that focus on
41 using a trait-based approach to formalise feeding interactions (*e.g.*, Shaw et al. (2024); Roopnarine (2006)),
42 are assembled by expert opinion (*e.g.* Dunne et al. (2014)) or make assumptions based on the evolutionary
43 signals of interactions (*e.g.*, Fricke et al. (2022)). However, there is an argument that the theoretical models
44 that have been developed in contemporary settings should hold even for paleo communities as we expect
45 the fundamental ‘currencies of life’ to have remained constant - *e.g.*, the energetic constraints of foraging or
46 foraging niches. Along with constructing different Additionally there are models that allow one to construct
47 structurally sound networks that require very little input data. These are methods that are amenable to the
48 data constraints that are prevalent in paleo communities in terms of both the completeness of fossil records as
49 well as how the deeper in time we move the further away we might be moving from contemporary analogs.

50 Not all contemporary models may actually be suitable for paleo contexts as the assumptions that they make
51 (or the data that they require) may actually introduce uncertainty/errors into the resulting network rendering
52 them of little use. Similarly not all paleo methods will be suitable for all paleo communities. As a simple
53 example the framework developed by Fricke et al. (2022) uses phylogenetic relatedness as a way to infer
54 interactions of Pleistocene mammals by looking at how their extant relatives interact. Although this approach
55 is ecologically sound (phylogenetic relatedness is also used in other approaches *e.g.*, Strydom et al. (2022))
56 there is also an argument that the further back in evolutionary time we go (and the greater the phylogenetic
57 distance between extant and extinct communities become) there is more uncertainty introduced by the
58 phylogenetic tree than what is introduced by assuming that interactions will be phylogenetically conserved.

⁵⁹ 2 Challenges specific to building paleo networks

⁶⁰ Although there has been a push for the development of tools and methods that allow us to predict species
⁶¹ interactions and networks they will not all be suitable for the prediction of paleo communities. This is
⁶² primarily due to limitations that we are faced with in terms of the information that can be inferred from
⁶³ the fossil record (such as species traits, abundances, and assemblages), which is needed as input data for the
⁶⁴ different models. The limited information available from the fossil record is compounded by the incomplete
⁶⁵ and biased preservation of species [REF], which part of a species is preserved (part vs whole), the ambiguity of
⁶⁶ the ‘true’ community composition [were communities conserved *in situ* or were they there owing to geological
⁶⁷ processes?; REF], as well as the availability/accessibility of different rock layers (and thus the completeness of
⁶⁸ data we might have for a specific era in time). Additionally there is an increasing degree of ‘fuzziness’ around
⁶⁹ the ecology and life histories of species the further one moves back in geological time [REF]. This is not to say
⁷⁰ that because we have imperfect data we should not be attempting to construct paleo food webs but rather
⁷¹ that we need to be aware of what the uncertainties are and how these might impact the assumptions that we
⁷² need to make when constructing a network (as well as how this will intersect with the intended end use of
⁷³ the network). This will allow us to best identify an approach that minimises the assumption and potential
⁷⁴ uncertainties within the data while still constructing a suitable network. This includes thinking about both
⁷⁵ assumptions you are making about the actual data *e.g.*, trying to extrapolate body size from fossil data but
⁷⁶ also assumptions across time *e.g.*, assuming modern trait-feeding modes are the same or that assumptions
⁷⁷ about network structure will hold across deep time.

⁷⁸ 2.1 Approaches to food web prediction

⁷⁹ Here we present six different models (Table 1) that can be used to construct food webs for both this specific
⁸⁰ community but are also broadly suited to paleo network prediction. These models span all facets of the
⁸¹ network representation space (metaweb, realised, and structural network) and are suitable for an array of
⁸² different paleo communities as the data requirements are ‘paleo friendly’.

Table 1: A summary of the different families of tools that can be used to generate paleo food webs.

Model family	Assumptions	Data needs	'Limitation'	Network type	Key reference
random	Links are randomly distributed within a network	richness, number of links	parameter assumptions, species agnostic	structural network	Erdős & Rényi (1959)
niche	Networks are interval, species can be ordered on a 'niche axis'	richness, connectance	parameter assumptions, species agnostic	structural network	Williams & Martinez (2008)
allometric diet	Interactions are determined by energetic costs (foraging ecology)	body mass, biomass (abundance)	does not account for forbidden links in terms of trait compatibility, assumptions on body size and biomass (abundance) from fossil data	theoretical	Petchey et al. (2008)
breadth model (ADBM)					

Model family	Assumptions	Data needs	'Limitation'	Network type	Key reference
l-matrix	Interactions inferred using allometric rules (ratio of body sizes between predator and prey), with links being constrained by a Ricker function	body mass, number of producer species	does not account for forbidden links in terms of trait compatibility, assumptions on body size from fossil data, assumptions as to the number of producer species	theoretical network	Schneider et al. (2016)
paleo food web inference model (PFIM)	Interactions can be inferred by a mechanistic framework/relationships	feeding traits for taxa, mechanistic feeding rules	Assumption made as to the feeding mechanisms, need to elucidate traits (although this is a way smaller issue)	mechanistic web	Shaw et al. (2024)

Model family	Assumptions	Data needs	'Limitation'	Network type	Key reference
body size ratio model	Interactions inferred using allometric rules (ratio of body sizes between predator and prey). Logit of the linking probability used to further constrain links to an 'optimal size range' for prey.	body mass	does not account for forbidden links in terms of evolutionary compatibility, assumptions on body size from fossil data	theoretical network	Rohr et al. (2010)

3 Case study: Toarcian mass extinction event

3.1 Dataset overview

3.1.1 Species occurrence

Here we use the fossil occurrence data over an interval extends from the upper Pliensbachian (~185 Ma) to the upper Toarcian (~175 Ma) of the Cleveland Basin (see Dunhill et al., 2024 for a more comprehensive overview). The data set consists of a subset of four broad time periods (pre-extinction, post-extinction, early recovery, and late recovery). The assemblages are treated as communities of interacting organisms. Something about the total number of taxa as well as numbers per a time period? Probably also make a comment that this is a 'deep time' community we are looking at.

3.1.2 Defining modes of life (traits)

We used the modes of life (traits) as identified in Dunhill et al. (2024), who defined four traits: motility (fast, slow, facultative, non-motile), tiering (pelagic, erect, surficial, semi-infaunal, shallow infaunal, deep infaunal), feeding (predator, suspension feeder, deposit feeder, mining, grazer), and size: gigantic (>500 mm), very large

96 ($>300\text{--}500$ mm), large ($>100\text{--}300$ mm), medium ($>50\text{--}100$ mm), small ($>10\text{--}50$ mm), tiny (10 mm), for each
97 fossil species based on the ecological traits defined in the Bambach ecospace model (Bambach et al., 2007).

98 3.1.3 Constructing networks

99 For each paleo community (time bin) we constructed **100** networks for each model (so $6 * 100$) networks.
100 These networks were ‘simplified’ to removed any disconnected species. In total 2400 networks were constructed.
101 When a quantitative measure of body size is needed (ADBM, bodymassratio, lmatrix) we drew a body mass
102 for each species from a uniform distribution. The ranges were defined by the different size classes as discussed
103 in insert cross ref to correct subsection here *e.g.*, a species classed as ‘very large’ would have a body mass
104 drawn from $U(300, 500)$. This was repeated for each run in order to add variation to the networks constructed,
105 however the same body sizes were kept consistent for the relevant models (adbm, bodymassratio, l-matrix)
106 *i.e.*, an ADBM and bodymassratio network from the same rep number would have used the same bodysizes.
107 The PFIM networks were downsampled (see relevant section is S1). For both the random and niche model
108 the desired connectance was randomly selected between the range 0.07 - 0.15 for each repetition but kept
109 consistent for both models. For each network we calculated the properties listed in Table 2

110 3.2 Models capture different network structure but in unexpected ways

111 Why is structure important and what can it tell us? Broadly when we talk about quantifying the structure
112 of a network we are interesting in capturing some aspect of how the links are distributed between nodes, or
113 alternatively about properties of the nodes (specifically in terms of the number of links coming in to (prey)
114 or out of (predators) the node). What are some things we can learn/infer from network structure: energy
115 flows and fluxes [REF], propagation of stress [REF], roles of species in the community [REF, think trophic
116 levels]. Some closing statement about how thus there are different facets of network structure and the value
117 of understanding generally how different models differ in terms of the structure that they recover - link to
118 Table 2 maybe.

Table 2: Network properties used fo analysis.

Metric	Definition	Scale	Reference (for maths), can make footnotes probs
Richness	Number of nodes in the network	Macro	

Metric	Definition	Scale	Reference (for maths), can make footnotes probs
Links	Normalized standard deviation of links (number of consumers plus resources per taxon)	Micro	
Connectance	L/S^2 , where S is the number of species and L the number of links	Macro	
Max trophic level	Prey-weighted trophic level averaged across taxa	Macro	Williams & Martinez (2004)
Diameter	Diameter can also be measured as the average of the distances between each pair of nodes in the network	Macro	Delmas et al. (2018)
Complexity	SVD complexity of a network, defined as the Pielou entropy of its singular values	Macro	Strydom et al. (2021)
Redundancy	$(L - (S - 1))/S$, where S is the number of species and L the number of links. Indicates the number of edges beyond what is needed for a minimum-connected tree	Macro	
S1	Number of linear chains, normalised	Meso	Milo et al. (2002); Stouffer et al. (2007)
S2	Number of omnivory motifs, normalised	Meso	Milo et al. (2002); Stouffer et al. (2007)

Metric	Definition	Scale	Reference (for maths), can make footnotes probs
S4	Number of apparent competition motifs, normalised	Meso	Milo et al. (2002); Stouffer et al. (2007)
S5	Number of direct competition motifs, normalised	Meso	Milo et al. (2002); Stouffer et al. (2007)
Generality	Normalized standard deviation of generality of a species standardized by L/S	Micro	Williams & Martinez (2000)
Vulnerability	Normalized standard deviation of vulnerability of a species standardized by L/S	Micro	Williams & Martinez (2000)

119 In terms of wanting to asses and compare across the different models it is beneficial to approach this task by
 120 thinking about the different aspects of the network as well as interactions that are being predicted by the
 121 different models. It is perhaps beneficial to think of these across different ‘scales’ of organisation within the
 122 network, namely macro (the entire network), meso (smaller interacting units within the network), and micro
 123 (species-level attributes). Although there are a myriad of possible ways to ‘measure’ and analyse ecological
 124 networks (Delmas et al., 2018) we do still lack a clear set of guidelines for assessing how well models recover
 125 network structure (Allesina et al., 2008) and it is beneficial to use a small subset of metrics that can clearly
 126 be tied to broader aspects of network function or capturing a ecological process.

127 Here we used a Multivariate Analysis Of Variance (MANOVA) to assess the differences between networks
 128 generated by different models based on the combined information of the multiple structural network measures.
 129 Model defined as `network structure values ~ model + time period` additionally we used a Linear Dis-
 130 criminant Analysis (LDA) to determine if different models produced networks with differing structure. In
 131 order to do the MANOVA and LDA we had to create within model variation for the different networks, with
 132 the exception of the PFIM model all models have some inherent variation. In order to generate variation

₁₃₃ within the PFIM metawebs we applied a *minimal* degree of downsampling following the protocol described in
₁₃₄ [ROOPNARINE]. This downsampling approach uses a power law distribution to essentially ‘prune’ links from
₁₃₅ the most generalist species (See SUPP MATT for a more detailed overview).

₁₃₆ **Need to report the relevant effect of time in driving observed differences???**

₁₃₇ [Figure 2 about here.]

₁₃₈ should I report some more ‘quantitative results’ i.e. a summary table...

₁₃₉ What is perhaps the most striking result in Figure 2 is that although there are clear structural differences
₁₄₀ between the different models the differences are not distinct between the broader model families but rather
₁₄₁ that there is a degree of overlap between them (specifically the log ratio, PFIM, and niche models). Although
₁₄₂ the log ratio and niche models are classified as different families they are built on similar ecological background
₁₄₃ and theory and so it is perhaps not surprising that these networks capture a similar structure (the same holds
₁₄₄ for the ADBM and l-matrix models). The fact that the random model occupies a completely different space is
₁₄₅ unsurprising as it has clearly been shown that networks are non-random in nature [REF] and so we expect
₁₄₆ random models to be constructing ecologically illogical networks. What is perhaps the most interesting result
₁₄₇ is that the PFIM model constructs networks that are very similar to those that are rooted in niche-based
₁₄₈ processes despite the model being more mechanistic in nature. (Not sure how to articulate but this is cool
₁₄₉ because there is *something* in network structure constraints that is straddling the trait-niche space of ecology -
₁₅₀ but also see my next point about it being ‘correct’ is still up for debate)

₁₅₁ Although it is not possible to confidently identify the models that are predicting the ‘*correct*’ network structure
₁₅₂ the fact that a models from different families are able to recover similar structures is reassuring as it suggests
₁₅₃ that it might be possible to substitute one model for another if the required data are not available. Although
₁₅₄ it is still important that an effort is made to match the ‘best’ model/network to the question being asked.

₁₅₅ It would be remiss to not think about the fact that niche models are supposed to work with trophic
₁₅₆ species not taxonomic species - I wonder if we should run the niche (and possibly random) models
₁₅₇ with trophic species as the starting richness and not the taxo richness...

₁₅₈ **TODO** Is it sound to try and unpack the ‘pairwise differences’ between the different structural
₁₅₉ metrics as well (or some) as this will allow us to say e.g. Niche and PFIM might recover the same
₁₆₀ connectance but differ in vulnerability.

161 **3.3 Some networks don't share any interactions and some share a lot**

162 In addition to wanting to measure network structure researchers may also be interested in understanding
163 aspects about the diets and predators of *specific* species in a community. In this instance the interest should
164 be in understanding how the pairwise links predicted between species pairs differ between models. Here we
165 look at the interaction turnover (Poisot et al., 2012) both within and between the different models. This can
166 be thought of as the equivalent of species β turnover and tells us which interactions are ‘conserved’ (shared)
167 across the networks but only between species pairs that are shared - *i.e.*, this turnover is only driven by
168 interaction and not species turnover. Here we only compared networks that we constructed for the same
169 period (as our interest is only in between model differences) and excluded the random and niche networks
170 from consideration as these two models are essentially species agnostic.

171 [Figure 3 about here.]

172 In Figure 3 it is clear that some models share a large percentage of interactions *e.g.*, between ADBM and
173 l-matrix networks and others share nothing *e.g.*, ADBM and PFIM networks. This result is unsurprising as the
174 mechanisms that determine interactions in ADBM and l-matrix (a single trait (bodysize) + parameterisation
175 of links by ecological theory) is very different from the PFIM model that makes assumptions on a trait-based,
176 mechanistic hierarchy.

177 The key takeaway that this needs to lead into is thinking about diet related questions and the model that
178 is best suited to get you there. It makes sense to contextualise this in the feasible vs realised interaction
179 spectrum - specifically that from a ‘philosophical’ basis if you are asking questions about possible diets of
180 species then it makes sense to use models that fall firmly in the ‘feasible’ space *e.g.*, PFIM model or even
181 something like the Fricke et al. (2022) model. How these results support that is that we can see the ADBM
182 and PFIM are recovering (almost) totally different pairwise links and on a conceptual basis it means that
183 we cannot use something like the ADBM to ask questions about diet breadth (note here this makes the
184 assumption that something like the PFIM is truly aligning with diet breadth, which I am inclined to say it is).

185 **3.4 Model choice changes the narrative**

186 **3.4.0.1 Simulating Extinctions**

187 Extinctions were simulated using different plausible mechanisms based on species traits (size, motility), their
188 position within the network (generality, vulnerability), as well as randomly. Each network was subjected
189 to 50 extinction runs for each extinction mechanism. The extinctions themselves were cascading in nature
190 meaning that after the target species was removed all species that no longer had any prey were also deemed

as extinct (secondary extinction), checking for secondary extinctions was then repeated until there were no longer any species without prey. This represents one extinction event and only then would the proceeding target species be removed from the network and cascading extinctions assessed again. Note that for extinction simulations which use the network position of a species to determine extinction order we follow the protocol from Curtsdotter et al. (2011) and reassess the vulnerability/generality of each species after each extinction event to ‘redetermine’ the extinction order based on their position in the newly generated network.

As we are using Dunhill et al. (2024) as a case study we followed their approach when simulating extinctions as well as assessing which extinction mechanism results in a simulated network that most closely matches the real post extinction network. Extinction simulations were only run on the pre extinction networks whereby species were removed until they reached the ‘target richness’, which is the richness of the post extinction community. In order to determine which extinction mechanism creates a network most similar to the post extinction network we used the (get full name of score) TSS (Gupta et al., 2022) to assess how different the pairwise interactions are between ‘simulated’ and ‘real’ post extinction communities as well as looking at the absolute differences in network structure metrics.

TODO not sure if we also want to unpack/showcase robustness R_{50} (Jonsson et al., 2015)

[Figure 4 about here.]

3.4.1 Trends over time

TODO Not sure statistically speaking what the best way to unpack this is... 2-way ANOVA/AN-COVA explanation is valuable? There are intercept differences (*e.g.*, baseline average values are different; are the rankings among all three response variables the same?) and there are shape differences/similarities (*e.g.*, motifs are all the same shape but Co and Gen show some among model differences in pattern.)

Visual take-away seems to suggest that we see that the values (intercepts) of the different summary statistics are different but (broadly) they are capturing the same trends. This might suggest that although we observe differences in structure (Figure 2) the general patterns still remain the same. This is good news because it means that at least the models that we have used here tend to tell us the same general story - which is worth contextualising in the space of ‘right’ vs ‘wrong’ and as long as we are not fixated on the point value but rather on understanding the trends.

219 **3.4.2 Inferred extinction drivers**

220 Points of discussion one will be to point to the mean absolute distance and how generally the ADBM/l-matrix
221 do really badly - high mean absolute value. And this maybe makes sense though because of how we specify
222 extinction mechanisms (trait-based) and so it sets the body-size models are not ‘talking’ the same language.
223 In terms of the TSS scores - not sure how we should unpack it. Individually by model family to see which
224 model agrees with which approach and see if different mechanisms come out stronger?

225 **4 Discussion (need a catchier heading)**

226 I want this section to be more about contextualising model choice within the bigger research question discussion
227 - i.e. mapping question and model choice more tightly...

228 Points to discuss:

- 229 • Guidelines - as a box? Can we give something concrete?? Something like Figure 5
230 • How to we synthesise these results? As in should we give clear directives or is it enough to do a bit
231 more handwaving and have the bigger message be that model choice matters?

232 [Figure 5 about here.]

233 **References**

- 234 Allesina, S., Alonso, D., & Pascual, M. (2008). A general model for food web structure. *Science*, 320(5876),
235 658–661. <https://doi.org/10.1126/science.1156269>
- 236 Bambach, R. K., Bush, A. M., & Erwin, D. H. (2007). Autecology and the Filling of Ecospace: Key Metazoan
237 Radiations. *Palaeontology*, 50(1), 1–22. <https://doi.org/10.1111/j.1475-4983.2006.00611.x>
- 238 Curtsdotter, A., Binzer, A., Brose, U., De Castro, F., Ebenman, B., Eklöf, A., Riede, J. O., Thierry, A., & Rall,
239 B. C. (2011). Robustness to secondary extinctions: Comparing trait-based sequential deletions in static and
240 dynamic food webs. *Basic and Applied Ecology*, 12(7), 571–580. <https://doi.org/10.1016/j.baae.2011.09.008>
- 241 Delmas, E., Besson, M., Brice, M.-H., Burkle, L. A., Dalla Riva, G. V., Fortin, M.-J., Gravel, D., Guimarães, P.
242 R., Hembry, D. H., Newman, E. A., Olesen, J. M., Pires, M. M., Yeakel, J. D., & Poisot, T. (2018). Analysing
243 ecological networks of species interactions. *Biological Reviews*, 112540. <https://doi.org/10.1111/brv.12433>
- 244 Dillon, E. M., Pier, J. Q., Smith, J. A., Raja, N. B., Dimitrijević, D., Austin, E. L., Cybulski, J. D., De
245 Entrambasaguas, J., Durham, S. R., Grether, C. M., Haldar, H. S., Kocáková, K., Lin, C.-H., Mazzini, I.,
246 Mychajliw, A. M., Ollendorf, A. L., Pimienta, C., Regalado Fernández, O. R., Smith, I. E., & Dietl, G. P.

- 247 (2022). What is conservation paleobiology? Tracking 20 years of research and development. *Frontiers in*
248 *Ecology and Evolution*, 10. <https://doi.org/10.3389/fevo.2022.1031483>
- 249 Dunhill, A. M., Zarzyczny, K., Shaw, J. O., Atkinson, J. W., Little, C. T. S., & Beckerman, A. P. (2024).
250 Extinction cascades, community collapse, and recovery across a Mesozoic hyperthermal event. *Nature*
251 *Communications*, 15(1), 8599. <https://doi.org/10.1038/s41467-024-53000-2>
- 252 Dunne, J. A., Labandeira, C. C., & Williams, R. J. (2014). Highly resolved early eocene food webs show
253 development of modern trophic structure after the end-cretaceous extinction. *Proceedings of the Royal*
254 *Society B: Biological Sciences*, 281(1782), 20133280. <https://doi.org/10.1098/rspb.2013.3280>
- 255 Erdős, P., & Rényi, A. (1959). On random graphs. i. *Publicationes Mathematicae Debrecen*, 6(3-4), 290–297.
256 <https://doi.org/10.5486/pmd.1959.6.3-4.12>
- 257 Fricke, E. C., Hsieh, C., Middleton, O., Gorczynski, D., Cappello, C. D., Sanisidro, O., Rowan, J., Svenning,
258 J.-C., & Beaudrot, L. (2022). Collapse of terrestrial mammal food webs since the Late Pleistocene. *Science*,
259 377(6609), 1008–1011. <https://doi.org/10.1126/science.abn4012>
- 260 Gauzens, B., Thouvenot, L., Srivastava, D. S., Kratina, P., Romero, G. Q., Berti, E., O’Gorman, E. J.,
261 González, A. L., Dézerald, O., Eisenhauer, N., Pires, M., Ryser, R., Farjalla, V. F., Rogy, P., Brose, U.,
262 Petermann, J. S., Geslin, B., & Hines, J. (2025). Tailoring interaction network types to answer different
263 ecological questions. *Nature Reviews Biodiversity*, 1–10. <https://doi.org/10.1038/s44358-025-00056-7>
- 264 Gupta, A., Furrer, R., & Petchey, O. L. (2022). Simultaneously estimating food web connectance and structure
265 with uncertainty. *Ecology and Evolution*, 12(3), e8643. <https://doi.org/10.1002/ece3.8643>
- 266 Hao, X., Holyoak, M., Zhang, Z., & Yan, C. (2025). Global Projection of Terrestrial Vertebrate Food
267 Webs Under Future Climate and Land-Use Changes. *Global Change Biology*, 31(2), e70061. <https://doi.org/10.1111/gcb.70061>
- 268
- 269 Jonsson, T., Berg, S., Pimenov, A., Palmer, C., & Emmerson, M. (2015). The reliability of R50 as a
270 measure of vulnerability of food webs to sequential species deletions. *Oikos*, 124(4), 446–457. <https://doi.org/10.1111/oik.01588>
- 271
- 272 Jordano, P. (2016). Chasing Ecological Interactions. *PLOS Biology*, 14(9), e1002559. <https://doi.org/10.1371/journal.pbio.1002559>
- 273
- 274 Kiessling, W., Raja, N. B., Roden, V. J., Turvey, S. T., & Saupe, E. E. (2019). Addressing priority questions
275 of conservation science with palaeontological data. *Philosophical Transactions of the Royal Society B:*
276 *Biological Sciences*, 374(1788), 20190222. <https://doi.org/10.1098/rstb.2019.0222>
- 277 Milo, R., Shen-Orr, S., Itzkovitz, S., Kashtan, N., Chklovskii, D., & Alon, U. (2002). Network motifs: Simple
278 building blocks of complex networks. *Science*, 298(5594), 824–827. <https://doi.org/10.1126/science.298.5594.824>
- 279

- 280 Morales-Castilla, I., Matias, M. G., Gravel, D., & Araújo, M. B. (2015). Inferring biotic interactions from
281 proxies. *Trends in Ecology & Evolution*, 30(6), 347–356. <https://doi.org/10.1016/j.tree.2015.03.014>
- 282 Petchey, O. L., Beckerman, A. P., Riede, J. O., & Warren, P. H. (2008). Size, foraging, and food web structure.
283 *Proceedings of the National Academy of Sciences*, 105(11), 4191–4196. <https://doi.org/10.1073/pnas.0710672105>
- 285 Pichler, M., & Hartig, F. (2023). Machine learning and deep learning—A review for ecologists. *Methods in
286 Ecology and Evolution*, 14(4), 994–1016. <https://doi.org/10.1111/2041-210X.14061>
- 287 Poisot, T., Canard, E., Mouquet, N., & Hochberg, M. E. (2012). A comparative study of ecological
288 specialization estimators. *Methods in Ecology and Evolution*, 3(3), 537–544. [https://doi.org/10.1111/j.2041-210x.2011.00174.x](https://doi.org/10.1111/j.2041-
289 210x.2011.00174.x)
- 290 Rohr, R., Scherer, H., Kehrli, P., Mazza, C., & Bersier, L.-F. (2010). Modeling food webs: Exploring
291 unexplained structure using latent traits. *The American Naturalist*, 176(2), 170–177. <https://doi.org/10.1086/653667>
- 293 Roopnarine, P. D. (2006). Extinction cascades and catastrophe in ancient food webs. *Paleobiology*, 32(1),
294 1–19. <https://www.jstor.org/stable/4096814>
- 295 Schneider, F. D., Brose, U., Rall, B. C., & Guill, C. (2016). Animal diversity and ecosystem functioning in
296 dynamic food webs. *Nature Communications*, 7(1), 12718. <https://doi.org/10.1038/ncomms12718>
- 297 Shaw, J. O., Dunhill, A. M., Beckerman, A. P., Dunne, J. A., & Hull, P. M. (2024). *A framework for
298 reconstructing ancient food webs using functional trait data* (p. 2024.01.30.578036). bioRxiv. <https://doi.org/10.1101/2024.01.30.578036>
- 300 Stouffer, D. B., Camacho, J., Jiang, W., & Nunes Amaral, L. A. (2007). Evidence for the existence of a robust
301 pattern of prey selection in food webs. *Proceedings of the Royal Society B: Biological Sciences*, 274(1621),
302 1931–1940. <https://doi.org/10.1098/rspb.2007.0571>
- 303 Strydom, T., Bouskila, S., Banville, F., Barros, C., Caron, D., Farrell, M. J., Fortin, M.-J., Hemming, V.,
304 Mercier, B., Pollock, L. J., Runghen, R., Dalla Riva, G. V., & Poisot, T. (2022). Food web reconstruction
305 through phylogenetic transfer of low-rank network representation. *Methods in Ecology and Evolution*,
306 13(12). <https://doi.org/10.1111/2041-210X.13835>
- 307 Strydom, T., Catchen, M. D., Banville, F., Caron, D., Dansereau, G., Desjardins-Proulx, P., Forero-Muñoz,
308 N. R., Higino, G., Mercier, B., Gonzalez, A., Gravel, D., Pollock, L., & Poisot, T. (2021). A roadmap
309 towards predicting species interaction networks (across space and time). *Philosophical Transactions of the
310 Royal Society B: Biological Sciences*, 376(1837), 20210063. <https://doi.org/10.1098/rstb.2021.0063>
- 311 Thuiller, W., Calderón-Sanou, I., Chalmandrier, L., Gaüzère, P., O'Connor, L. M. J., Ohlmann, M., Poggiato,
312 G., & Münkemüller, T. (2024). Navigating the integration of biotic interactions in biogeography. *Journal*

- ³¹³ of *Biogeography*, 51(4), 550–559. <https://doi.org/10.1111/jbi.14734>
- ³¹⁴ Williams, R. J., & Martinez, N. D. (2000). Simple rules yield complex food webs. *Nature*, 404(6774), 180–183.
- ³¹⁵ <https://doi.org/10.1038/35004572>
- ³¹⁶ Williams, R. J., & Martinez, N. D. (2004). Stabilization of chaotic and non-permanent food-web dynamics.
- ³¹⁷ *The European Physical Journal B - Condensed Matter*, 38(2), 297–303. <https://doi.org/10.1140/epjb/e2004-00122-1>
- ³¹⁸ Williams, R. J., & Martinez, N. D. (2008). Success and its limits among structural models of complex food webs. *The Journal of Animal Ecology*, 77(3), 512–519. <https://doi.org/10.1111/j.1365-2656.2008.01362.x>
- ³¹⁹ Yeakel, J. D., Pires, M. M., Rudolf, L., Dominy, N. J., Koch, P. L., Guimarães, P. R., & Gross, T.
- ³²⁰ (2014). Collapse of an ecological network in ancient egypt. *PNAS*, 111(40), 14472–14477. <https://doi.org/10.1073/pnas.1408471111>

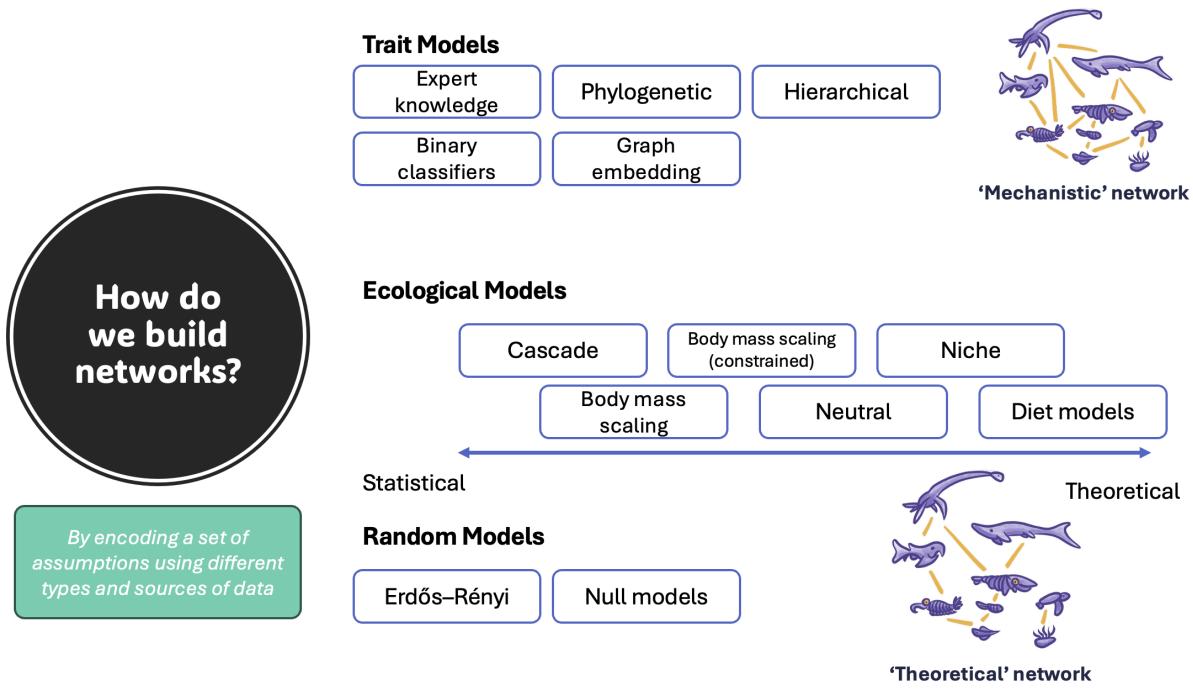


Figure 1: This obviously needs work but a variation on this to try and articulate the different approaches and broadly how they may differ.

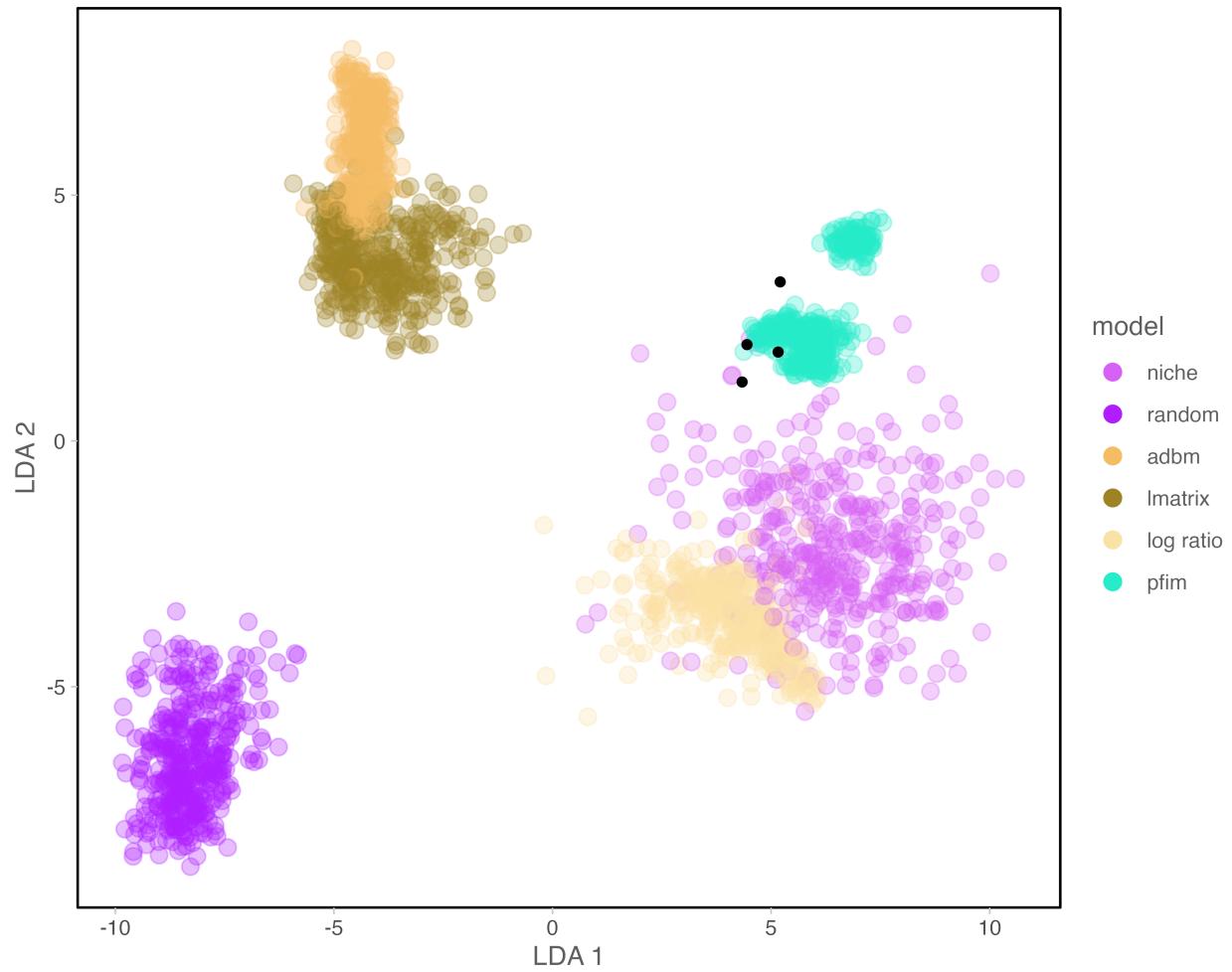


Figure 2: Linear Discriminant Analysis using thirteen different metrics of network structure. The colours represent the different models. The black points represent the predicted metaweb co-ordinates for the four different time periods.

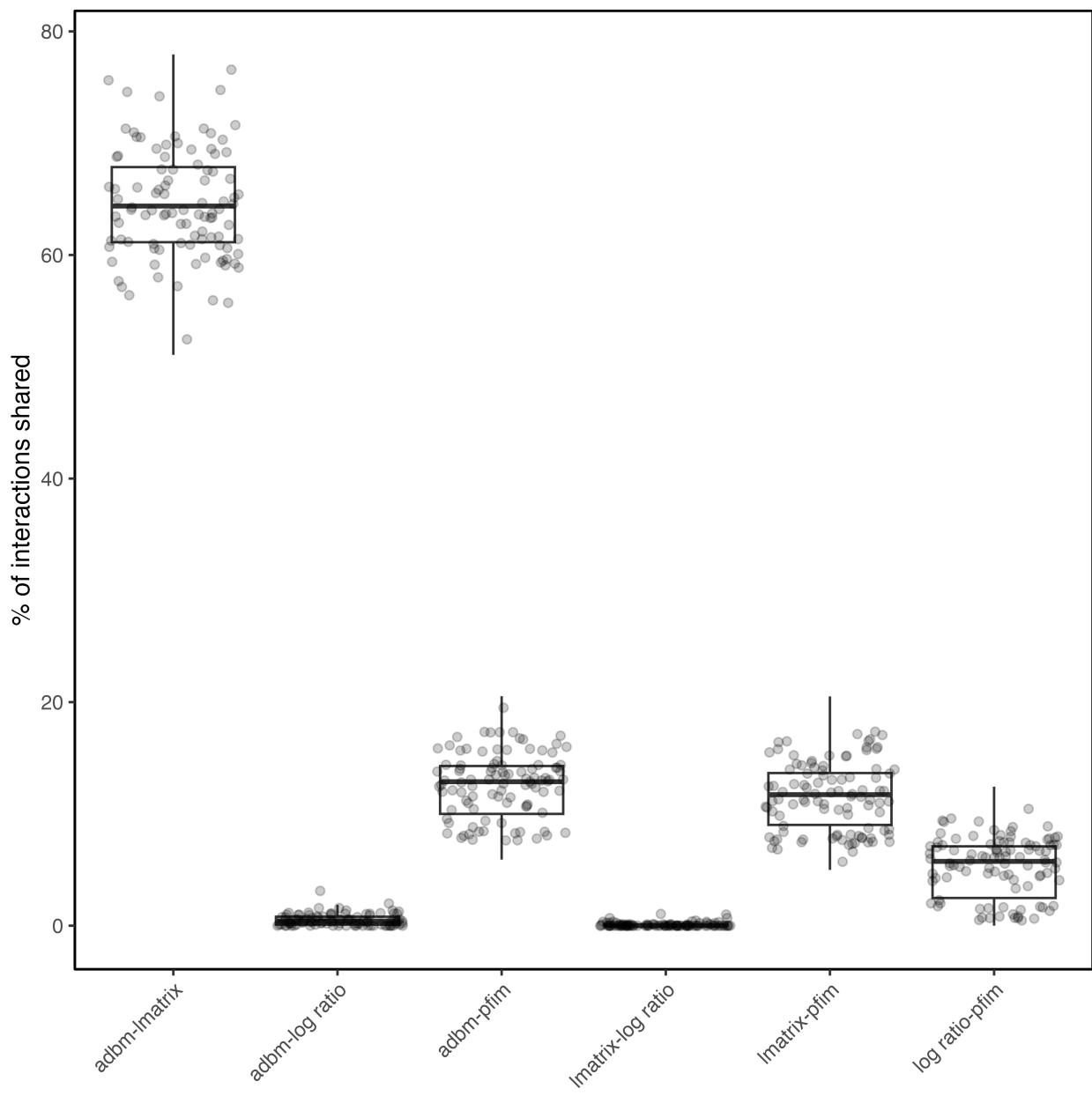


Figure 3: stuff... % interaction shared is calculated as number shared interactions / ((number interactions left - shared interactions) + (number interactions right - shared interactions) + shared interactions). Additionally niche and random models are excluded as it is illogical since both of these models are fundamentally species agnostic

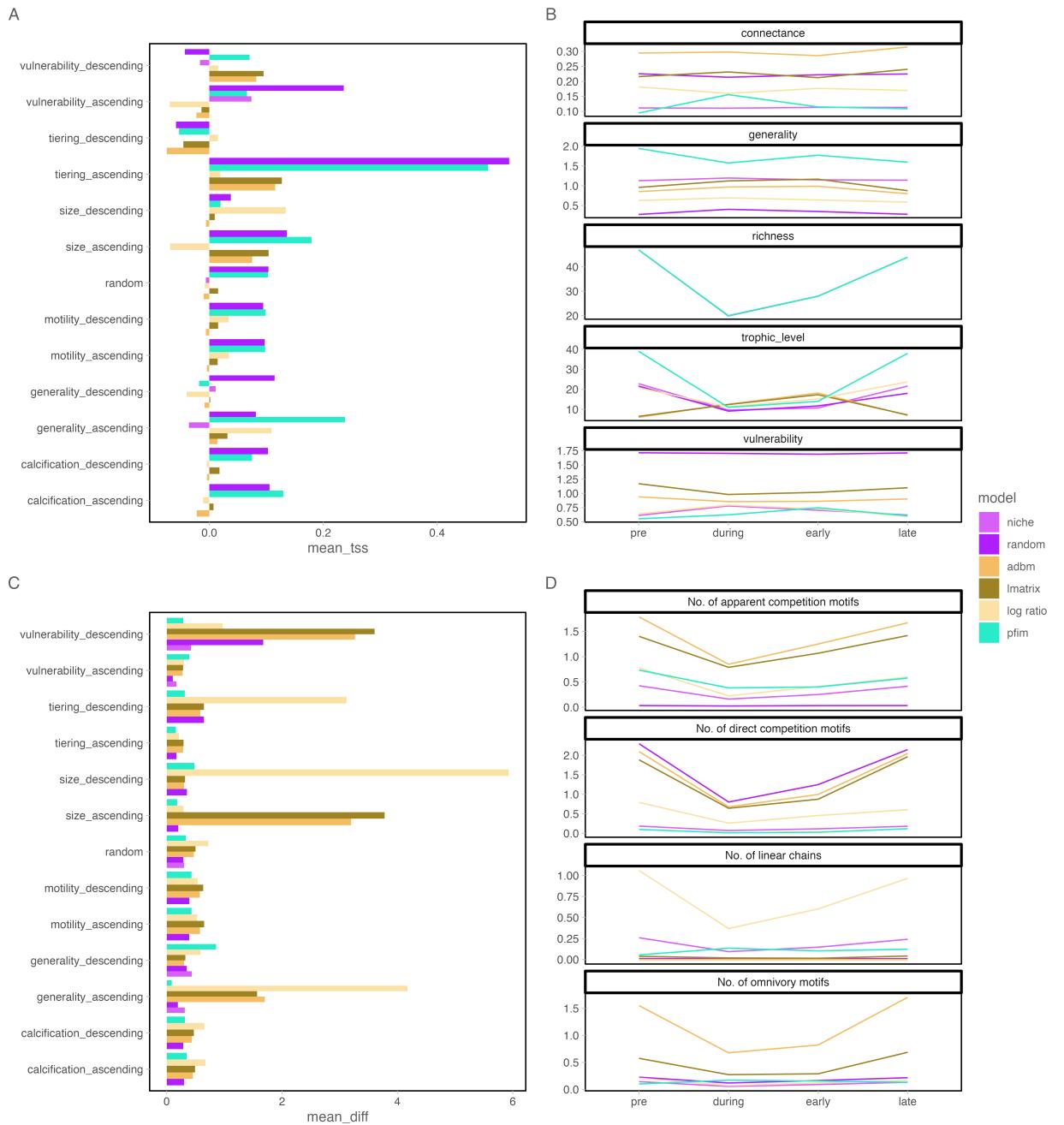


Figure 4: stuff... Recreation of the figure from Dunhill 2024. I think it makes sense to split and divide into the below sections (depending on potential figure limitations of target journal)

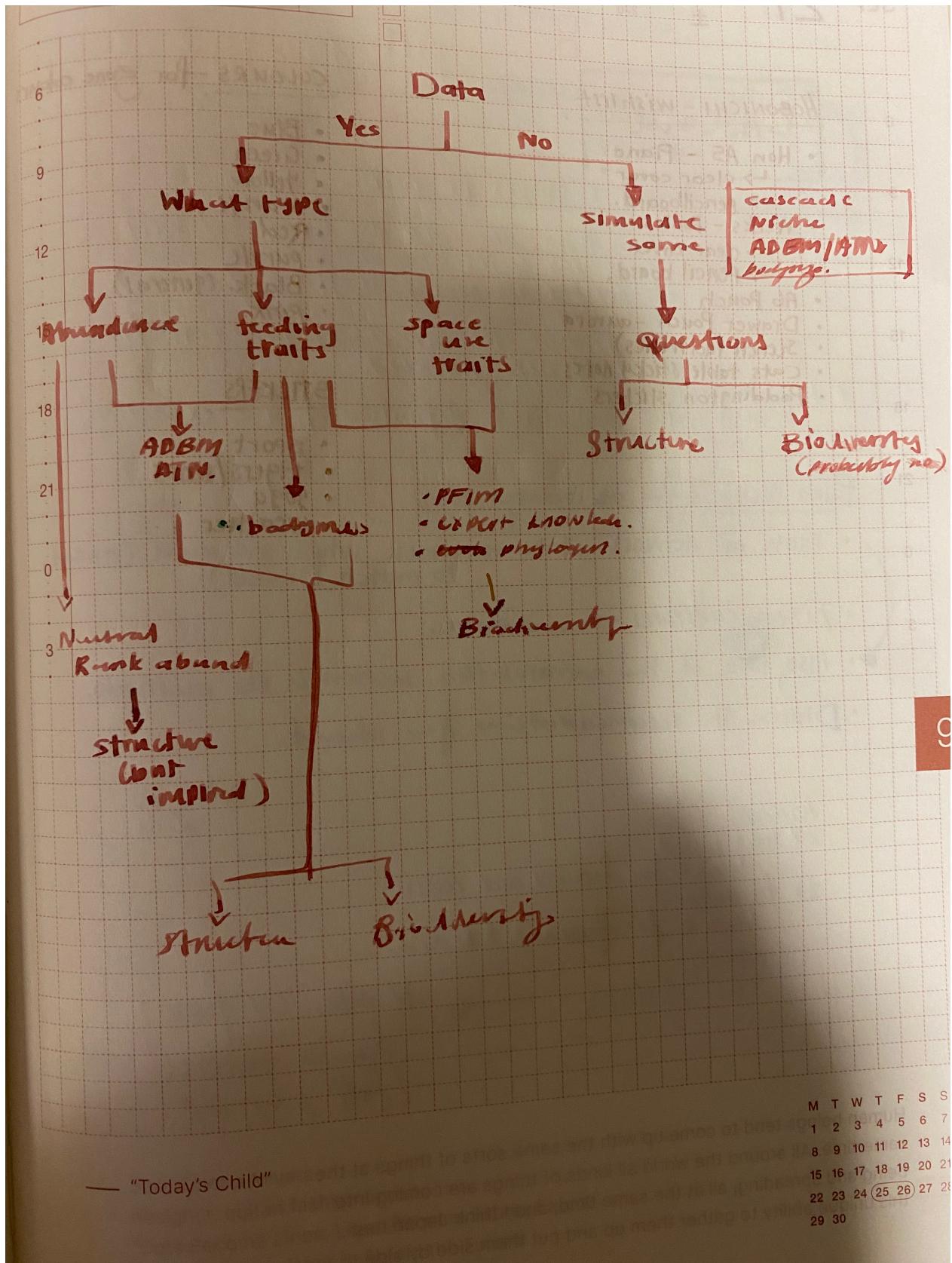


Figure 5: TODO.