

Model structure conditions ecological inference in food web reconstruction

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Abstract: Aim

Ecological networks are widely used to compare community structure, stability, and responses to disturbance across environmental gradients. However, many networks (particularly those assembled from incomplete interaction data) require model-based reconstruction. We test how alternative reconstruction frameworks condition ecological inference by quantifying their effects on network structure and disturbance dynamics.

Location

Cleveland Basin, United Kingdom.

Time period

Toarcian extinction event (Early Jurassic, late Pliensbachian–late Toarcian, ~183 Ma).

Major taxa studied

Marine animal communities.

Methods

We reconstructed four successive communities from an identical species pool using six contrasting food-web models spanning feasible (trait-based), realised (allometric and energetic), and structural (topological) network representations. For each community and model, 100 replicate networks were generated. We quantified macro-, meso-, and micro-scale network properties and assessed differences among models using multivariate analyses. Pairwise interaction turnover was measured using link-based beta diversity. We then simulated species loss under multiple disturbance scenarios, allowing cascading extinctions, and compared predicted community states using mean absolute differences and rank concordance metrics between models.

Results

Reconstruction framework strongly influenced inferred network topology (MANOVA, $p < 0.001$), generating distinct structural signatures independent of species composition. Models that were similar in global metrics often diverged in species-level interactions, with high -turnover among inferred link sets. During disturbance simulations, species-level vulnerability rankings were broadly consistent across models, but interaction-level outcomes and cascade dynamics varied substantially. Concordance in extinction-scenario rankings was scale dependent, with higher agreement at the species level than at the interaction level.

Main conclusions

Network reconstruction functions as a structural prior that conditions ecological inference. While some aggregate patterns are robust across modelling frameworks, detailed interaction-level dynamics are highly model contingent. Comparative network studies across spatial or environmental gradients should therefore align reconstruction framework with inferential goals and explicitly evaluate sensitivity to modelling assumptions.

Keywords: Ecological networks, Biotic interactions, Community assembly, Environmental gradients, Interaction turnover, Trophic organisation, Ecosystem resilience, Macroecology

¹ 1 Introduction

² Understanding how biodiversity is organised across space and time is a central goal of macroecology and
³ biogeography. While early efforts focused primarily on species richness and composition, there is growing
⁴ recognition that ecological communities are structured not only by which species occur, but by how they
⁵ interact (Thuiller et al., 2024). Interaction networks are increasingly treated as macroecological state variables
⁶ where they are used to compare community organisation across environmental gradients, to quantify -diversity
⁷ in interaction structure, to evaluate stability-complexity relationships, and to infer vulnerability under global
⁸ change (Carstensen et al., 2014; Gravel et al., 2019; Poisot et al., 2015; Trøjelsgaard & Olesen, 2016; Tylianakis
⁹ & Morris, 2017). As a result, ecological networks now play a central role in comparative analyses spanning
¹⁰ latitudinal gradients, disturbance regimes, and deep-time environmental transitions (Dunhill et al., 2024;
¹¹ Hao et al., 2025; Michalska-Smith & Allesina, 2019; Poisot & Gravel, 2014; Roopnarine, 2006). Implicit in
¹² this expansion is the critical assumption that network properties estimated across systems are structurally
¹³ comparable, and that differences among them reflect ecological signal rather than methodological artefact
¹⁴ (Fründ et al., 2016; Jordano, 2016).

¹⁵ Yet most ecological networks are not fully observed as interaction data are incomplete and sampling is
¹⁶ uneven across historical and biogeographic contexts, across both contemporary and deep-time (Catchen et
¹⁷ al., 2023; Poisot et al., 2021; Sandra et al., 2025). Interactions must often be inferred indirectly from traits,
¹⁸ phylogeny, body size, co-occurrence, or theoretical constraints (Morales-Castilla et al., 2015; Strydom et
¹⁹ al., 2021). Network construction therefore constitutes a model-based inference step rather than a purely
²⁰ descriptive exercise. Different reconstruction frameworks encode distinct ecological assumptions about how
²¹ interactions arise - whether as biologically feasible combinations of traits, energetically optimised realised diets,
²² or topological structures constrained by macroecological regularities. These assumptions act as structural
²³ priors over network architecture (Gauzens et al., 2025; Guimarães, 2020; Petchey et al., 2011; Strydom et
²⁴ al., 2026). If alternative reconstruction models systematically generate different trophic configurations, then
²⁵ comparative analyses risk conflating ecological differences among communities with artefacts introduced by
²⁶ modelling choice. The reliability of macroecological inference therefore depends not only on ecological data,
²⁷ but on the structural assumptions embedded in network reconstruction.

²⁸ Despite rapid methodological development in interaction inference, few studies have directly evaluated how
²⁹ alternative reconstruction frameworks condition macroecological conclusions when applied to the same species
³⁰ pool. This gap is particularly consequential for comparative research, where network metrics are routinely
³¹ interpreted as indicators of environmental filtering, disturbance intensity, evolutionary history, or community
³² stability (Allesina & Tang, 2012; Delmas et al., 2018; Poisot et al., 2015). If reconstruction models encode

33 distinct structural priors over interaction topology, then differences among communities may reflect modelling
34 assumptions rather than ecological processes. We therefore test whether macroecological inference derived from
35 ecological networks is robust to variation in reconstruction framework, asking which aspects of network-based
36 inference are stable across plausible representations of interaction structure and which are intrinsically model
37 dependent.

38 Deep-time ecosystems provide an especially stringent test of this issue because interactions are not observed
39 directly and must be reconstructed explicitly (Dunhill et al., 2024; Dunne et al., 2008; Dunne et al., 2014;
40 Roopnarine, 2006), rendering modelling assumptions transparent. We re-evaluate inferences made by Dunhill
41 et al. (2024) on community structure and extinction dynamics during the early Toarcian extinction event
42 (~183 Ma), a volcanic-driven hyperthermal and marine crisis in the Early Jurassic (Kemp et al., 2024).
43 Using four successive communities drawn from an identical taxon pool, we reconstruct ecological networks
44 under six contrasting model classes spanning feasible, realised, and structural representations. For each
45 reconstruction framework, we quantify emergent topology across scales, measure interaction turnover, and
46 simulate disturbance-driven collapse. By holding species composition constant while varying the food web
47 model used, this design isolates the influence of model constrained structure on inferred food web organisation
48 and extinction dynamics, allowing us to distinguish ecological signals that are robust from those that are
49 reconstruction contingent.

50 2 Methods

51 2.1 Study system and fossil data

52 We used fossil occurrence data from the Cleveland Basin spanning the upper Pliensbachian to the upper
53 Toarcian. This interval encompasses a major volcanic-driven hyperthermal and marine extinction event. To
54 capture network dynamics across this transition, we defined four successive paleo-communities: Pre-extinction
55 (Pliensbachian), Post-extinction (Lower Toarcian), Early recovery, and Late recovery (Middle/Upper Toarcian).
56 Each taxon was characterized using their size and Bambach's ecospace framework (Bambach et al., 2007),
57 coding for tiering, motility, and feeding mode as per Dunhill et al. (2024). Each assemblage was treated
58 as a community of potentially interacting taxa. The dataset includes 57 taxa across diverse groups (*e.g.*,
59 cephalopods, bivalves, and gastropods). By restricting our analysis to a single basin with consistent lithofacies,
60 we controlled for biogeographic noise, ensuring that structural shifts across the four time-bins reflect localised
61 ecological responses to environmental stress rather than sampling or facies bias.

⁶² **2.2 Network reconstruction approaches**

⁶³ **2.2.1 Conceptual classification of network types**

⁶⁴ Most paleo-specific approaches (e.g., Fricke et al. (2022); Roopnarine (2006); Shaw et al. (2024)) currently
⁶⁵ operate within the feasibility space. That is, these approaches identify and encode the entire diet of a species
⁶⁶ and thus represent all feasible interactions within the network. Although well suited for reconstructing
⁶⁷ feasible interactions, these methods represent only a subset of the broader spectrum of network construction
⁶⁸ approaches. Here, we present a suite of models (Table 1) that enable the construction of a wider range of
⁶⁹ ecological networks and the exploration of a broader set of ecological questions, provided that their underlying
⁷⁰ assumptions are compatible with the constraints of fossil data. For example, some tools require quantitative
⁷¹ estimates of body size, which must often be inferred from size classes or functional morphology in the fossil
⁷² record. Structural models, such as the niche model, rely only on species richness and estimates/specification of
⁷³ connectance, but their species-agnostic nature limits their applicability to trait-based or diet-specific questions,
⁷⁴ although they do still accurately recover network structure (Stouffer et al., 2005) . Mechanistic approaches,
⁷⁵ by contrast, depend on accurate assignment of feeding traits or robust phylogenetic support. Recognizing
⁷⁶ how these methodological requirements intersect with the limits of fossil evidence is essential for selecting an
⁷⁷ appropriate modelling framework.

Table 1: Six different models that can be used to construct food webs for both this specific community but are also broadly suited to paleo network prediction. These models span all facets of the network representation space (feasibility, realised, and structural network) and are suitable for an array of different paleo communities as the data requirements fall within the limitations set by the fossil record.

| Model | Assumptions | Data needs | Limitation | Network type | Key reference | Usage examples |
|--------------------------------------|-------------------------------------------------------------------------|------------------------------------|----------------------------------------------------------------|--------------|----------------------------|----------------------------------------------------------------------------------------------------------------------------|
| Random | Links assigned randomly | Species richness, number of links | Parameter assumptions, species agnostic | Structural | Erdős & Rényi (1959) | Null-model comparisons; testing whether observed network structure (connectance, motifs) deviates from random expectations |
| Niche | Species ordered along a ‘niche axis’; interactions interval-constrained | Species richness, connectance | Parameter assumptions, species agnostic | Structural | Williams & Martinez (2008) | Evaluating trophic hierarchy and motif structure; baseline structural predictions |
| Allometric diet breadth model (ADBM) | Energy-maximizing predator diets | Body mass, abundance/dance/biomass | Assumes optimal foraging; does not account for forbidden links | Realised | Petchey et al. (2008) | Predicting realized predator diets; exploring secondary extinctions |

| Model | Assumptions | Data needs | Limitation | Network type | Key reference | Usage examples |
|---------------------------------------|----------------------------------------------------------------------|------------------------------------|-------------------------------------------------------|--------------|--------------------------------------------|------------------------------------------------------------------------|
| Allometric trophic network (ATN) | Links constrained by body-size ratios and functional response | Body mass, number of basal species | Assumes only mechanico/energetic constraints | Realised | Brose et al. (2006); Gauzens et al. (2023) | Simulating species loss; evaluating network collapse dynamics |
| Paleo food web inference model (PFIM) | Interactions inferred using trait-based mechanistic rules | Feeding traits | Assumes feeding mechanisms; trait resolution required | Feasibility | Shaw et al. (2024) | Mapping feasible trophic interactions; assessing secondary extinctions |
| Body-size ratio model | Probabilistic assignment of links based on predator-prey size ratios | Body mass | Does not account for forbidden links | Realised | Rohr et al. (2010) | Estimating likely interactions; simulating cascading effects. |

78 The three body mass-based models (ADBM, ATN, Body-size ratio) differ primarily in their underlying
 79 ecological assumptions. Although all three models use body mass to infer food web structure, they differ in
 80 their ecological assumptions. The ADBM is based on energy maximization under optimal foraging theory,
 81 the ATN constrains interactions via mechanically optimal consumer-resource size ratios, and the Body-size
 82 ratio model defines links probabilistically within a fixed allometric niche. Together, these approaches span
 83 bioenergetic, mechanical, and statistical interpretations of size-structured interactions.

84 **2.2.2 Network generation and replication**

85 We evaluated six models spanning this space Table 1: random and niche models (structural networks);
86 allometric diet breadth (ADBM), allometric trophic network (ATN), and Body-size ratio models (realised
87 networks); and a paleo food web inference model (PFIM; feasibility network). Expanded descriptions of
88 model assumptions, parameterisation, and link-generation rules are provided in Supplementary Material
89 S1. For each community, 100 networks were generated per model per successive community ($n = 2400$) to
90 capture stochastic variation in link assignment. Where models required species body mass or trait values,
91 these were sampled within biologically reasonable ranges to preserve relative differences among species. We
92 adopted uniform sampling by default, as alternative distributions (lognormal, truncated lognormal) have
93 negligible impact on topology (Supplementary Material S2; Figure S1). Structural models were parameterized
94 using connectance values drawn from an empirically realistic range (0.07 – 0.34), with species richness held
95 constant. Identical parameter draws were applied across comparable models within each replicate to ensure
96 comparability. For the Body-size ratio model, we followed the approach of Yeakel et al. (2014) and excluded
97 latent trait terms as opposed fitting the full model, which introduces additional inference and assumptions.

98 **2.3 Network metrics and structural analyses**

99 We quantified network structure using a suite of macro-, meso-, and micro-scale metrics Table 2, capturing
100 global properties, motif structure, and species-level variability. Differences among reconstruction approaches
101 were assessed using a multivariate analysis of variance (MANOVA), with model identity as a fixed factor
102 and the full set of network metrics as response variables. Pairwise interaction turnover was quantified using
103 link-based beta diversity, which measures dissimilarity in the identity of trophic links between networks,
104 capturing differences due to species turnover or changes in interactions among shared species (Poisot et al.,
105 2012).

Table 2: Network properties used for analysis.

| Metric | Definition | Scale | Reference (for maths), can make footnotes probs |
|----------|--------------------------------|-------|-------------------------------------------------------|
| Richness | Number of nodes in the network | Macro | |

| Metric | Definition | Scale | Reference (for maths), can make footnotes probs |
|-------------------|---------------------------------------------------------------------------------------|-------|-------------------------------------------------------|
| Links | Normalized standard deviation of links (number of consumers plus resources per taxon) | Micro | |
| Connectance | L/S^2 , where S is the number of species and L the number of links | Macro | |
| Max trophic level | Prey-weighted trophic level averaged across taxa | Macro | Williams & Martinez (2004) |
| S1 | Number of linear chains, normalised | Meso | Milo et al. (2002); Stouffer et al. (2007) |
| S2 | Number of omnivory motifs, normalised | Meso | Milo et al. (2002); Stouffer et al. (2007) |
| S4 | Number of apparent competition motifs, normalised | Meso | Milo et al. (2002); Stouffer et al. (2007) |
| S5 | Number of direct competition motifs, normalised | Meso | Milo et al. (2002); Stouffer et al. (2007) |
| Generality | Normalized standard deviation of generality of a species, standardised by L/S | Micro | Williams & Martinez (2000) |
| Vulnerability | Normalized standard deviation of vulnerability of a species, standardised by L/S | Micro | Williams & Martinez (2000) |

106 **2.4 Extinction simulations and model evaluation**

107 Following Dunhill et al. (2024) and using the pre-extinction and post-extinction networks, we simulated
108 species loss under multiple extinction scenarios. We simulated species loss under multiple scenarios, including
109 trait-based, network-position-based, and random removals, allowing cascading extinctions to propagate.
110 Simulated outcomes were compared to observed or expected community states using mean absolute differences
111 (MAD) and modified true skill statistics (TSS) at node and link levels. Kendall's rank correlation coefficient
112 was used to evaluate concordance in scenario rankings across reconstruction models, providing a measure of
113 robustness in inferred community responses.

114 **3 Results**

115 Across six network reconstruction approaches, inferred food web structure, species interactions, and extinction
116 dynamics differed consistently. Multivariate analyses showed pronounced separation among models in network
117 metric space. Reconstruction approach explained most of the variance in structural properties, leaving a distinct
118 signature independent of community composition. Notably, agreement among models depended on scale -
119 approaches that were statistically similar in multivariate structural space often diverged in inferred interactions
120 or extinction dynamics. This demonstrates that structural similarity does not guarantee concordance in species-
121 level diets or trophic roles. Model choice also substantially influenced inferred extinction dynamics. Temporal
122 trajectories of network collapse, interaction loss, and motif reorganization differed among approaches. Although
123 species-level extinction rankings were often broadly consistent, link-level outcomes and extinction inferences
124 were highly sensitive to reconstruction assumptions. Together, these results show that ecological inferences
125 drawn from networks depend critically on the reconstruction framework employed. Importantly, agreement
126 among models was not consistent across analytical scales - models that were statistically indistinguishable in
127 multivariate structural space often diverged in inferred interactions or extinction dynamics. Together these
128 results show that reconstruction approaches that appear similar when evaluated using global network metrics
129 can yield fundamentally different ecological narratives when interrogated at the level of interactions and
130 extinction dynamics.

131 **3.1 Network structure differs among reconstruction approaches**

132 Across six reconstruction approaches, network structure (network properties listed in Table 2) differed
133 significantly (MANOVA, Pillai's trace = 3.84, approximate $F_{40,11955} = 987.35$, $p < 0.001$), indicating that
134 model choice systematically alters inferred food web topology. Canonical discriminant analysis identified two

135 dominant axes of variation, explaining 86% of between-model variance. LD1 correlated with vulnerability,
136 direct competition motifs, and connectance. LD2 correlated with maximum trophic level and apparent
137 competition motifs, reflecting vertical trophic structure (Figure 1; Table S1, Figure S1). All higher-order
138 canonical variates each explained less than 9% of the remaining variance.

139 [Figure 1 about here.]

140 **3.1.1 Variance partitioning of network structure**

141 Permutational multivariate analysis of variance revealed that reconstruction framework accounted for the
142 majority of variation in multivariate network structure ($R^2 = 0.795$, $p < 0.001$), whereas temporal turnover
143 across extinction phases explained a comparatively small proportion of variance ($R^2 = 0.064$, $p < 0.001$).
144 The model \times time interaction contributed a further 7.1% of variance ($R^2 = 0.071$, $p < 0.001$), indicating
145 limited but significant time-dependent divergence among reconstruction frameworks. Thus, differences among
146 modelling approaches were more than an order of magnitude greater than structural differences associated
147 with ecological turnover through the extinction sequence, even if the Toarcian dataset was characterized with
148 a significant community turnover.

149 To determine whether the dominance of reconstruction framework reflected absolute mean shifts among time
150 bins, we repeated the analysis after centring network metrics within each extinction phase. This procedure
151 removes between-phase differences while retaining within-phase structural variation. Even after temporal
152 bin-standardised centring, reconstruction framework explained 84.8% of multivariate variance ($R^2 = 0.848$,
153 $p < 0.001$). These results demonstrate that the influence of model choice is not driven by temporal mean
154 differences, but reflects intrinsic divergence among reconstruction frameworks in how ecological interactions
155 are organised.

156 **3.1.2 Inferred pairwise interactions vary widely among models**

157 Despite some models showing similar global metrics, specific pairwise interactions often differed. Pairwise
158 –turnover revealed that certain model pairs shared very few links Figure 2. Size-based models (ADBM,
159 ATN) were broadly similar due to shared sole reliance on body-size constraints, whereas the Body-size ratio
160 model exhibited consistently higher differences to other models. PFIM showed intermediate overlap with
161 theoretical models. These results demonstrate that agreement in global network structure does not guarantee
162 concordance in species-level interactions.

163 [Figure 2 about here.]

164 **3.2 Model choice influences inferred extinction dynamics**

165 To quantify how network structure changed over time during extinction simulations and whether these
166 dynamics differed among reconstruction models, we fit generalized additive models (GAMs) to time series
167 of network-level metrics. GAMs capture non-linear temporal trajectories, allowing formal tests of whether
168 the shape of these trajectories differs among models. These model-specific temporal trajectories are shown
169 in Figure 3. For all metrics examined, the inclusion of model-specific smooth terms significantly improved
170 model fit (ANOVA model comparison: $p < 0.001$ for all metrics). Model-specific smooths differed not only
171 in magnitude but also in the timing and abruptness of change, indicating distinct modes of collapse across
172 reconstruction approaches (Tables S3–S4). Deterministic, data-driven approaches (PFIM) and allometric
173 models (ADBM, ATN) exhibited highly non-linear trajectories, showing structural shifts in connectivity
174 and motif frequency. In contrast, the Niche model produced the most consistent and gradual trajectories,
175 effectively smoothing the perceived magnitude of structural change during community collapse. These results
176 demonstrate that inferred pathways of collapse, trophic bottlenecks, and secondary extinctions are highly
177 sensitive to model choice. Corresponding raw temporal trajectories are shown in Figure S2.

178 [Figure 3 about here.]

179 To evaluate how model choice affects inferred extinction dynamics, we compared simulated post-extinction
180 networks to observed networks using mean absolute differences (MAD) for network-level metrics and true
181 skills statistics (TSS) for node- and link-level outcomes Figure 4. Across models, MAD-based rankings were
182 generally positively correlated (Kendall's $\tau = 0.13$ across structural metrics), indicating broad agreement on
183 the relative importance of extinction drivers despite substantial differences in reconstructed network structure.
184 However, agreement within the allometric models differed from patterns observed for reconstructed network
185 structure.

186 Node-level TSS rankings were similarly consistent across models ($\tau = 0.26\text{--}0.90$), reflecting broadly comparable
187 species removal sequences. In contrast, link-level outcomes were far more variable ($\tau = -0.48\text{--}0.29$), highlighting
188 that inferences about which interactions are lost, retained, or re-established during collapse and recovery are
189 highly model contingent. Together, these results suggest that while alternative models converge on similar
190 species-level extinction patterns, the inferred pathways of interaction loss and cascading dynamics depend
191 strongly on both reconstruction approaches.

192 [Figure 4 about here.]

193 **4 Discussion**

194 **4.1 Network reconstruction is not neutral: structural priors shape ecological
195 theory**

196 Food web ecology has long treated network reconstruction as a technical step preceding ecological analysis.
197 Once a network is assembled (whether from observation, inference, or simulation) its properties are typically
198 analysed as reflections of underlying ecological organisation. Implicit in this workflow is a powerful assumption
199 - that reconstructed networks provide structurally comparable representations of ecological communities, such
200 that differences in connectance, trophic structure, motif composition, or robustness primarily reflect biological
201 variation. Our results challenge this assumption. Across an identical species pool, reconstruction frameworks
202 explained the majority of variance among reconstructed network topologies, generating distinct structural
203 signatures independent of community composition. Although some model classes converged on similar global
204 metrics, pairwise interaction turnover revealed substantial divergence in inferred trophic links. Structural
205 similarity at the level of connectance or trophic height therefore does not guarantee concordance in species
206 roles or interaction architecture.

207 Reconstruction framework explained the majority of variance in inferred food web topology, far outweighing
208 the influence of temporal turnover across extinction phases. These patterns demonstrate that the divergence
209 among reconstruction approaches reflects intrinsic differences in how models organise interactions, not temporal
210 shifts in network structure. In other words, different reconstruction approaches impose distinct structural
211 priors on inference of ecological organisation. These priors are not subtle - they propagate into emergent
212 topology, species roles, and dynamical predictions. Network structure is therefore not solely a property
213 of ecological communities; it is jointly determined by ecological data, modelling assumptions, and level of
214 organisation (Gauzens et al., 2025; Guimarães, 2020; Pichler & Hartig, 2023; Strydom et al., 2021; Strydom
215 et al., 2026).

216 This has direct implications for the interpretation of comparative network studies. Feasible, realised, and
217 structural models encode different assumptions about constraint, optimisation, and topology, and these
218 assumptions propagate into emergent metrics and dynamical predictions (Allesina & Tang, 2012; Dunne et
219 al., 2002; Michalska-Smith & Allesina, 2019; Poisot & Gravel, 2014; Solé & Montoya, 2001). When networks
220 reconstructed under different model classes are compared across spatial gradients, disturbance regimes, or
221 evolutionary transitions, part of the observed variation may derive from structural priors rather than ecological
222 process. Without explicit standardisation or sensitivity analysis, methodological heterogeneity can be mistaken
223 for biological signal. Food web ecology has devoted substantial effort to understanding how topology shapes

224 dynamics; comparatively less attention has been paid to how reconstruction method shapes topology. Our
225 findings indicate that these two questions cannot be separated.

226 4.2 Scale-dependent robustness in network-based inference

227 Importantly, reconstruction sensitivity was not uniform across network scales (macro-, meso-, micro- level
228 properties). Species-level extinction rankings were broadly consistent among model classes, whereas interaction-
229 level outcomes and cascade trajectories were highly contingent on structural prior. The predominance of
230 reconstruction framework over temporal turnover (~80% vs. 6% variance explained) illustrates why coarse-
231 grained patterns like species-level extinction rankings are more robust. Model-imposed structure dominates
232 the overall topology, leaving finer-scale interaction dynamics highly contingent on framework choice. This
233 asymmetry reveals a context-dependent pattern of robustness. Coarse-grained macroecological patterns (such
234 as the vulnerability of a community to collapse) can emerge from multiple plausible interaction architectures.
235 By contrast, fine-grained inferences about which links are lost, retained, or reorganised depend strongly on
236 how interactions are inferred.

237 This distinction challenges a central ambition of food web ecology - the use of detailed interaction structure
238 to diagnose mechanisms of stability and collapse (*e.g.*, Dunhill et al., 2024). If interaction-level cascade
239 pathways vary substantially across equally plausible reconstructions, then mechanistic narratives derived
240 from a single inferred topology may overstate their precision (Allesina & Tang, 2012; Curtsdotter et al., 2011;
241 Dunne et al., 2002). The apparent determinism of extinction cascades or trophic bottlenecks may partly
242 reflect model-imposed structure rather than ecological inevitability.

243 For macroecology, this metric dependence clarifies where network-based inference is accurate. Aggregate
244 properties may be comparatively robust to reconstruction assumptions, whereas conclusions about interaction
245 turnover, motif reorganisation, or fine-scale trophic dynamics are intrinsically uncertain. Recognising this
246 asymmetry is essential if network analyses are to inform comparative synthesis across space and time.

247 Taken together, these results underscore that network reconstruction is not a neutral preprocessing step but
248 part of a hypothesis-generating process in which each model encodes a distinct set of ecological assumptions.
249 The inferred topology and dynamics of a food web therefore reflect not only ecological data, but the theoretical
250 assumptions embedded in the reconstruction framework. Disagreement among models does not imply that
251 any single approach is ‘wrong’, but rather that different models capture different facets of ecological reality
252 (Petchey et al., 2011; Stouffer, 2019). Disagreement among models does not imply that any single approach
253 is ‘incorrect’. Rather, different models capture different facets of ecological constraint—trait compatibility,
254 energetic optimisation, or topological regularity. The critical point is that these facets are not interchangeable.

255 This perspective reframes reconstruction choice as part of hypothesis specification. Researchers must align
256 reconstruction approaches with the ecological signals of interest (whether potential interactions, realised
257 diets, or macro-scale structural properties) rather than treating model selection as a technical convenience.
258 Viewed through the lens of accuracy and precision, our results indicate that some network-based inferences
259 are relatively robust across reconstruction approaches, whereas others remain intrinsically uncertain. High-
260 level extinction rankings were broadly convergent, suggesting relative accuracy at coarse resolution, but
261 interaction-level details and temporal cascade dynamics diverged substantially, indicating limited precision
262 in reconstructing the fine structure of collapse. Recognising and explicitly accounting for this distinction
263 is essential if food web ecology is to move beyond descriptive reconstruction toward rigorous comparative
264 inference.

265 **4.3 Implications for comparative biogeography and global change research**

266 Network approaches are increasingly applied to examine how ecological organisation varies across latitudinal
267 gradients, environmental filters, disturbance regimes, and climate-driven transitions. These studies frequently
268 interpret variation in connectance, trophic height, interaction -diversity, or robustness as indicators of
269 ecological differentiation among regions or time intervals. Yet reconstruction frameworks often differ among
270 systems due to variation in sampling intensity, trait resolution, or modelling preference. Our results show
271 that such differences can systematically alter inferred topology and disturbance dynamics even when species
272 composition is held constant.

273 This finding has consequences for interaction biogeography. Apparent differences in network turnover across
274 spatial gradients may reflect divergence in structural priors rather than genuine ecological reorganisation.
275 Similarly, projected differences in vulnerability under climate change may depend on how realised interactions
276 are modelled, not solely on community composition. Without explicit treatment of reconstruction assumptions,
277 comparative network analyses risk conflating modelling artefact with macroecological process.

278 If ecological networks are to function as reliable macroecological state variables, reconstruction framework must
279 be treated as a formal component of hypothesis specification. Cross-system comparisons should standardise
280 modelling approaches wherever possible, or explicitly evaluate sensitivity to alternative structural priors.
281 Ensemble reconstruction—treating networks as distributions rather than single topologies—offers a principled
282 pathway toward incorporating structural uncertainty into comparative synthesis.

283 **4.4 Toward a more explicit modelling paradigm in food web ecology**

284 The broader implication is not that any single reconstruction framework is ‘correct’ or ‘incorrect’. Rather, each
285 model class represents a distinct hypothesis about how interactions are constrained—by trait compatibility,
286 energetic optimisation, or topological regularity (Petchey et al., 2011). Food web reconstruction is therefore
287 theory-laden. Making this explicit shifts reconstruction from a preparatory step to a central component of
288 ecological modelling.

289 A mature modelling paradigm in food web ecology would treat structural priors as testable assumptions,
290 incorporate probabilistic link inference where possible, and quantify the sensitivity of macroecological
291 conclusions to alternative representations of interaction structure. Such an approach aligns with recent
292 advances in probabilistic and ensemble network modelling and would strengthen the interpretability of
293 network-based inference under global change (Banville et al., 2025; Baskerville et al., 2011; Elmasri et al.,
294 2020; Perez-Lamarque et al., 2026; Poisot et al., 2016).

295 **5 Conclusions**

296 Ecological network reconstruction is not a neutral technical procedure but a theoretical act that shapes
297 ecological inference. By applying six contrasting reconstruction frameworks to an identical species pool, we
298 show that structural priors systematically influence inferred food-web topology, interaction identity, and
299 disturbance dynamics. Some coarse-grained patterns, such as relative species vulnerability, are comparatively
300 robust across representations. In contrast, fine-scale interaction structure and cascade pathways are highly
301 contingent on modelling assumptions. The reliability of network-based inference is therefore scale dependent.

302 These results challenge the implicit assumption that reconstructed networks are structurally commensurable
303 across systems. When reconstruction frameworks differ, variation in connectance, trophic organisation,
304 robustness, or interaction turnover may reflect embedded modelling assumptions as much as ecological
305 processes. Network reconstruction should thus be treated as an explicit component of hypothesis specification
306 in comparative macroecology and biogeography.

307 No single model captures the full complexity of ecological organisation, but neither are alternative models
308 interchangeable. Aligning reconstruction framework with inferential goals, standardising approaches across
309 comparative studies, and incorporating ensemble or probabilistic representations will be essential for strength-
310 ening the interpretability of network analyses across spatial and temporal gradients. As ecological networks
311 increasingly inform global change research, recognising structural priors as fundamental determinants of
312 inference is critical for advancing food web ecology from descriptive reconstruction toward rigorous comparative

³¹³ synthesis.

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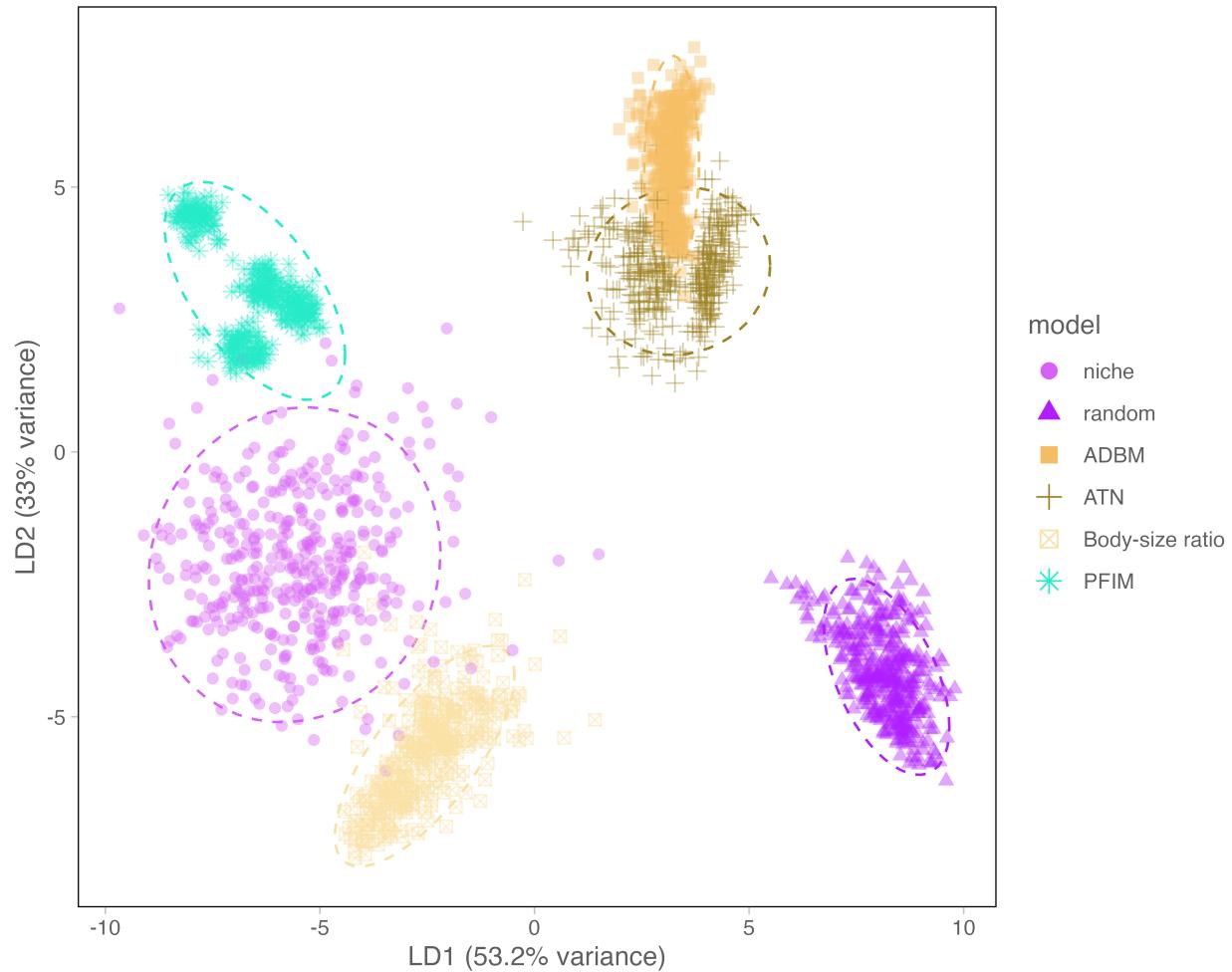


Figure 1: Linear discriminant analysis (LDA) of ecological network metrics for six model types. Each point represents a replicate, and ellipses indicate 95% confidence regions for each model. The second column represents the correlation of the various network metrics with the respective LDA axes.

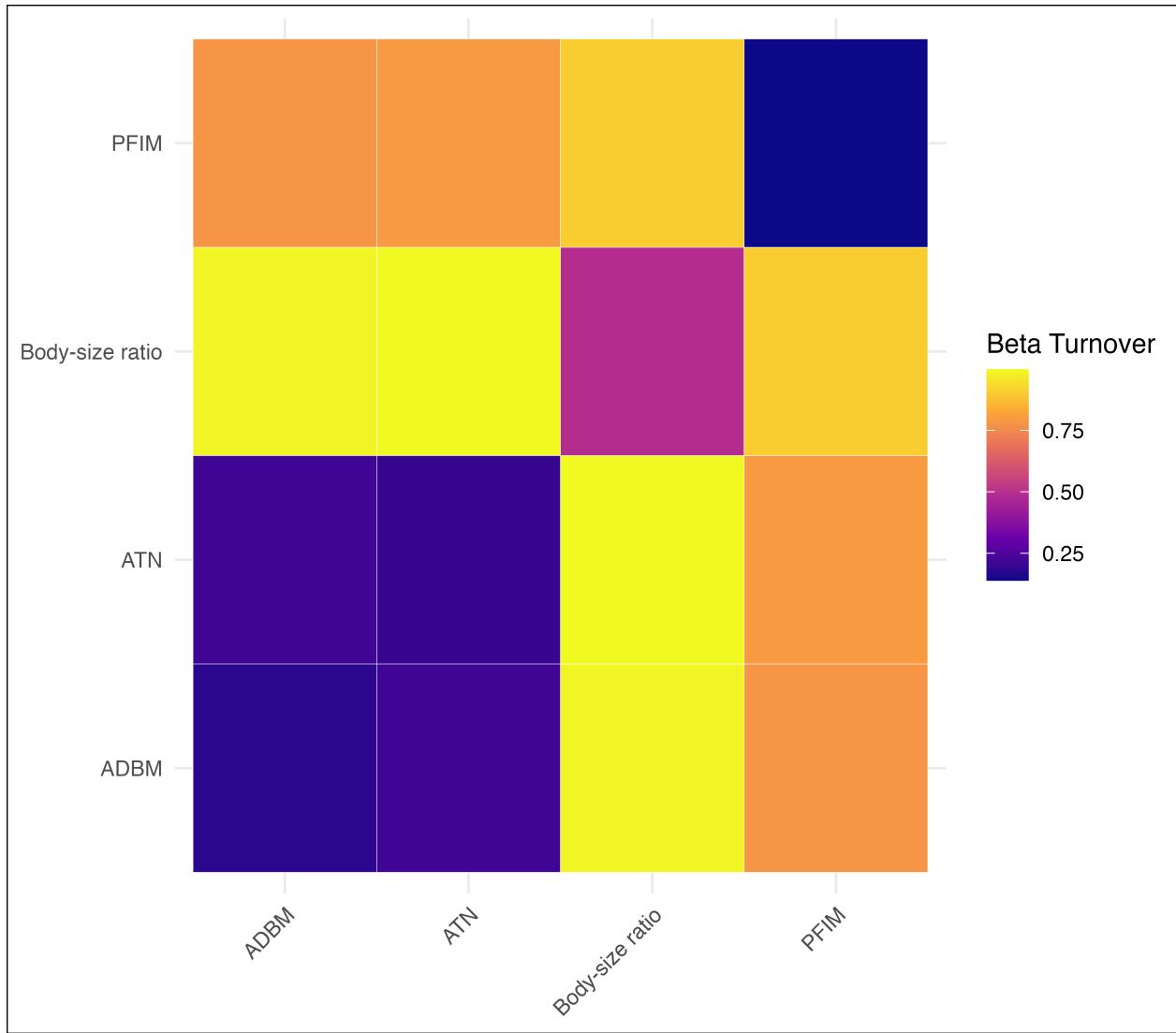


Figure 2: Pairwise β -turnover in species interactions among four ecological network models (ADBM, ATN, Body-size ratio, and PFIM). Each cell represents the mean turnover value between a pair of models, with warmer colours indicating greater dissimilarity in inferred interactions. The diagonal is omitted. High turnover values (yellow) indicate strong disagreement in network structure between models, whereas lower values (blue–purple) indicate greater similarity.

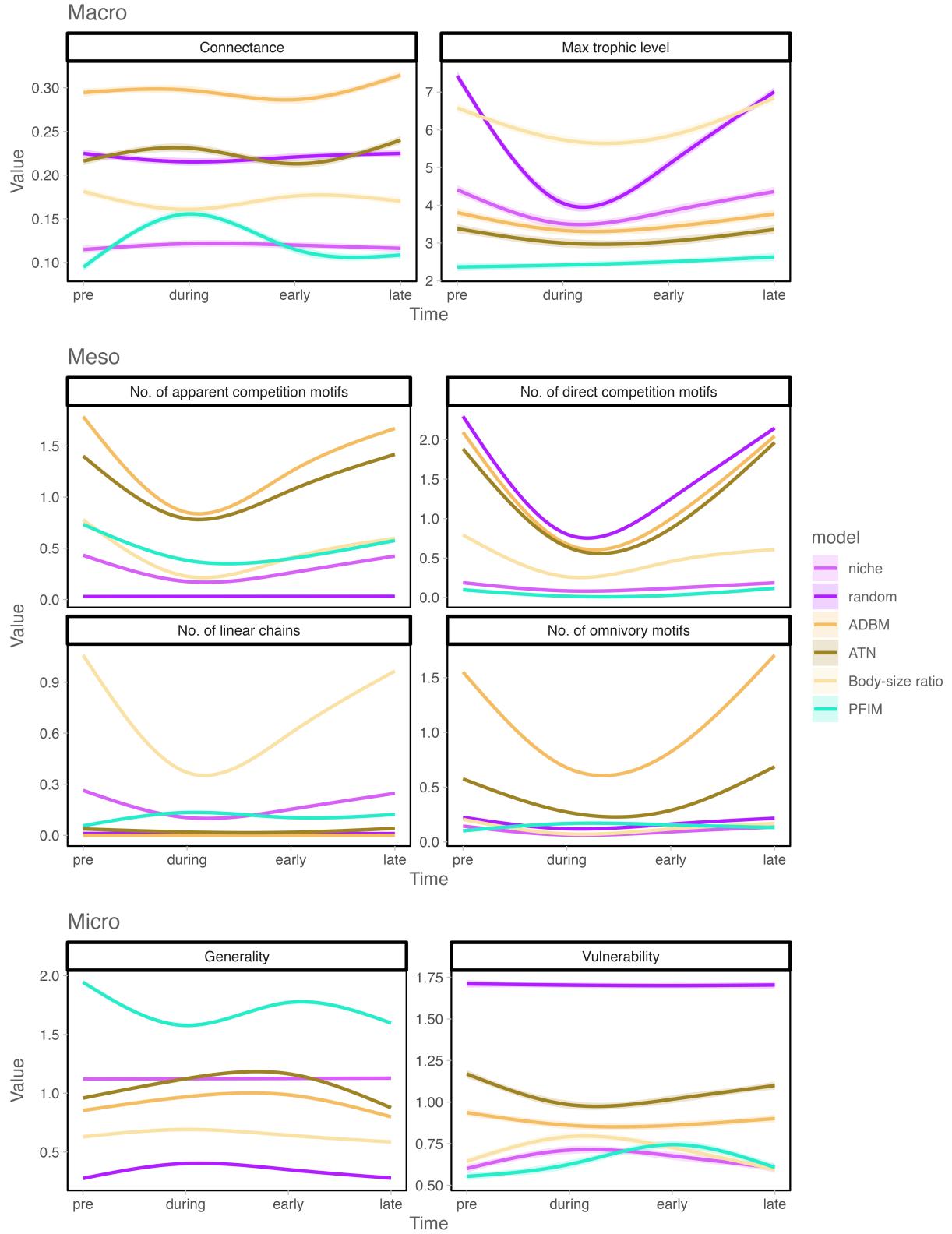


Figure 3: GAM-predicted trajectories of network structure during extinction simulations reveal pronounced differences in the timing and magnitude of change across reconstruction models. Lines show model-specific smooths and shaded areas indicate 95% confidence intervals. Deterministic approaches produce smoother, more consistent dynamics, whereas stochastic models exhibit greater variability, underscoring the sensitivity of inferred collapse pathways to reconstruction assumptions.

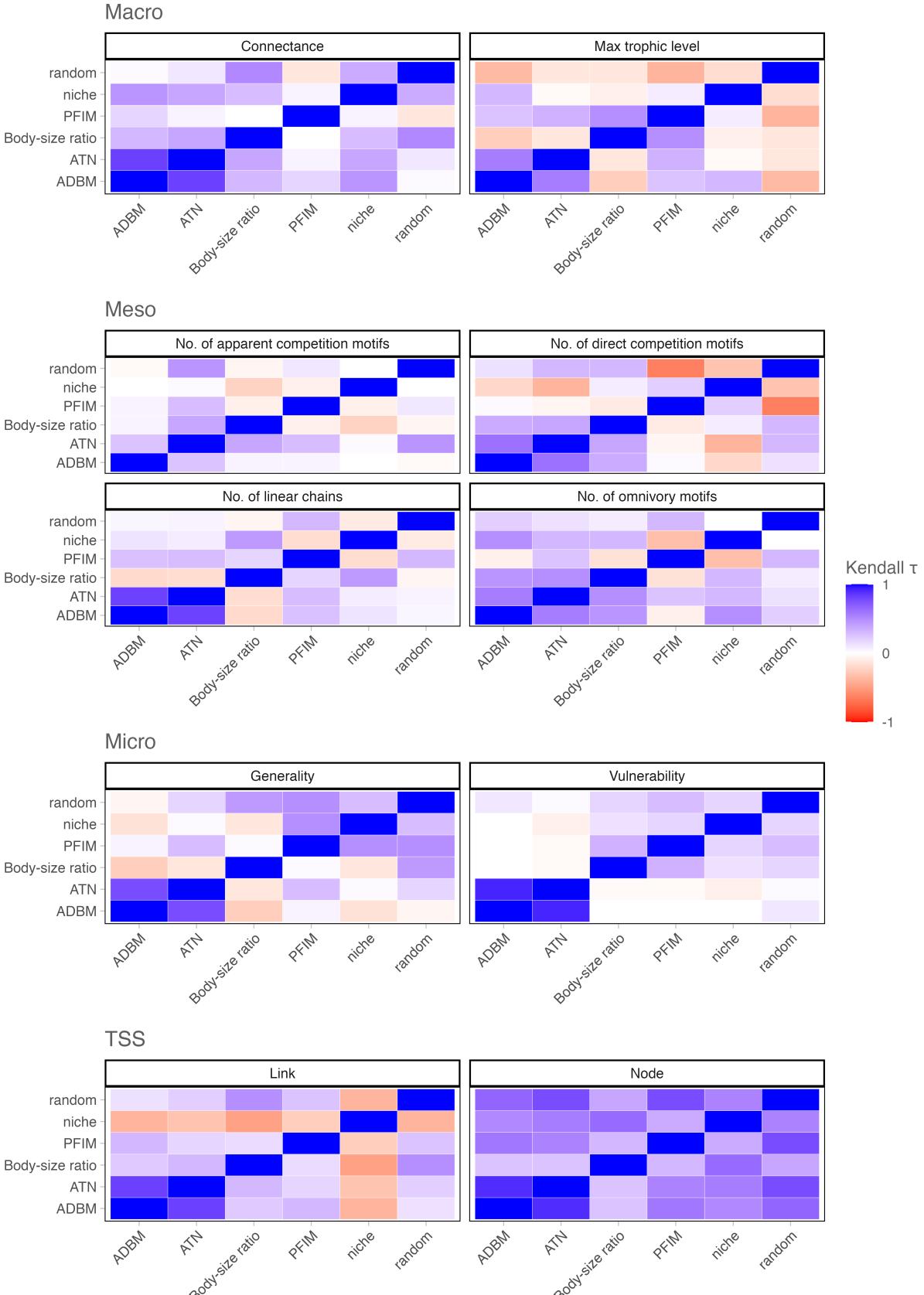


Figure 4: Heatmaps showing pairwise Kendall rank correlation coefficients (τ) between models for each network metric. Each panel corresponds to a different metric and displays the degree of agreement in extinction-scenario rankings across models based on mean absolute differences (MAD) between observed and predicted network values. Positive τ values (blue) indicate concordant rankings between models, whereas negative τ values (red) indicate opposing rankings. Warmer colours approaching zero represent little or no agreement. Panels