



Population-level Habitat Selection Analyses Produce Consistent Effect Direction, but Strength is Sensitive to Researcher Choice

Benjamin Michael Marshall^{*1} and Alexander Bradley Duthie^{**1}

¹ Biological and Environmental Sciences, Faculty of Natural Sciences, University of Stirling, Stirling, FK9 4LA, Scotland, UK

^{*}benjaminmichaelmarshall@gmail.com

^{**}alexander.duthie@stir.ac.uk

Abstract

Science strives to be repeatable and reproducible, but we know that different researchers will reach different conclusions even when presented with the same data and questions. This variation due to different choices researchers make during study and analysis may play a role in increasing variation in results. Variation in results may be particularly pronounced in disciplines with high task uncertainty and difficult-to-control experimental environments, such as ecology. Ecological research is often concerned with conservation, and a key component of animal conservation is determining animals' habitat requirements that can steer effective habitat protection. Insights into habitat requirements can be gained via the examination of animal movement, but the translation of movement data to usable estimates of habitat selection can be approached in many ways. We use a multiverse approach to explore a suite of possible habitat selection approaches and determine the chances of incorrectly identifying habitat selection in two simulated scenarios. We find that the majority of different analysis pathways correctly identify selection, but the strength of that selection is susceptible to analysis choice. We identify several choices that have disproportionate impacts on selection estimates, with some choices having contrasting impacts depending on the analysis approach. Overall our results suggest that while we can be broadly confident in the conclusions drawn from habitat selection studies, we cannot ignore analysis choices, especially when the strength of selection is of particular concern.

Keywords

Movement ecology, simulation, compositional, step selection function, two-step, poisson, habitat preference, habitat selection, animal movement, multiverse, research choice, researcher degrees for freedom



1 Introduction

The successful application of science to policy relies on results being repeatable, reliable, and generalisable. Efforts to achieve this often consist of removing idiosyncrasies of studies; we remove confounding variables, control for non-target variables; and conduct replicants. Despite our best efforts, we can never create a fully generalisable study.

The vast majority of the time our efforts to account for study-specific variation appear adequate, resulting in answers that can form the foundations for further studies. However, the re-examination of foundational work reveals the need to repeatedly question and retest the assumptions subsequent work relies upon. For example, this need for reassessment can be seen in studies of bird plumage and its role in mate selection, where false positives rates in published studies may have elevated the importance of plumage colour and pattern size in mate selection (Seguin & Forstmeier, 2012; Parker, 2013; Sánchez-Tójar et al., 2018). Subsequent meta-analyses have highlighted the need to be more cautious with interpreting initial findings, and that future hypotheses require refinement.

While other disciplines' replication rates have been more closely examined (Freedman, Cockburn & Simcoe, 2015; Open Science Collaboration, 2015; see Kelly, 2019 for ecological replication rates), there is evidence that ecology and evolution studies require similar scrutiny with regards to reliability as many of the same incentive structures seen as responsible exist, such as a prioritisation of novel findings and clear significant research narratives (Fanelli, 2010; Ware & Munafò, 2015; Smaldino & McElreath, 2016; Brembs, 2018). Replication studies and meta-analyses have shown that results in ecology and evolution may need further scrutiny (Seguin & Forstmeier, 2012; Parker, 2013; Sánchez-Tójar et al., 2018). In general, it appears that initial findings are more extreme than subsequently findings, and that their initial momentum in regards to research direction is slow to reorient (Jennions & Møller, 2002; Barto & Rillig, 2012).

For disciplines where direct replications are more common and feasible, replication can play a larger feedback role (Freedman, Cockburn & Simcoe, 2015; Open Science Collaboration, 2015; Peterson & Panofsky, 2021). Ecological systems are complex and impossible to fully control leading to ecologists' frequent reliance on "natural experiments".

Natural experiments by their nature are impossible to replicate perfectly, forcing ecology to rely on conceptual or quasi replications (Palmer, 2000; Fraser et al., 2020). These replications may have subjects and methods that differ from the original study, or where the research questions are conceptually similar (Nakagawa & Parker, 2015; Kelly, 2019). These replication efforts are still valuable, and valued by the scientific community (Fraser et al., 2020), as they aid gauging the reliability of proposed patterns, theories, or laws.

Ecology and evolution, due to the high task uncertainty and limited options of direct replication in the wild, tends towards a system of organic self-correction, where the complexity of study systems can explain away the differences between studies (i.e., inconsistencies replication), leaving the integration of past and subsequent findings more open to interpretation (Peterson & Panofsky, 2021) . .

Animal movement, like many aspects of ecology, is a field of study in which replications are difficult to conduct. Tracking animals is costly and comes with serious ethical considerations, especially for animals where the tracking equipment can impact their health, mobility, and behaviour (Sperry et al., 2009; Homberger et al., 2021; Portugal & White, 2022). Therefore, there is a strong incentive to extract the maximum value for every tracked animal and to ensure that all results are as reliable as possible.

The findings from animal movement studies can be linked with landscape management and integrated into conservation plans (Fraser et al., 2018; Doherty & Driscoll, 2018). A key component for protecting species is identifying that species' requirements; the examination of movement can reveal those requirements (Mueller et al., 2011; Doherty, Fist & Driscoll, 2019). Suitable habitat (as defined in various ways) is a fundamental requirement, and the protection/restoration of suitable habitat is an actionable conservation response. Reliable determinations of suitable habitat are therefore required the correct conservation decisions.

As demonstrated by many analyst projects, different answers can originate from the same dataset, and even the same question (e.g., Silberzahn et al., 2018; Huntington-Klein et al., 2021). Such "researcher degrees of freedom" [or flexibility in analysis; Forstmeier, Wagenmakers & Parker (2017)] can be the result of differences in question interpretation (Auspurg & Brüderl, 2021) and the analysis approach taken (Gelman & Loken, 2013; Bastiaansen et al., 2020). When drawing information to be used in a conservation plan, it would be advisable to know whether decisions would be different had the data been examined in a different way. Previous explorations have shown that researcher degrees of freedom are sufficiently large to alter effect strength and the final conclusions (Salis, Lena & Lengagne, 2021; Desbureaux, 2021).

While many analysts projects represent an ideal, actually capable of capturing the variation in results stemming



from the researchers themselves, the logistics of conducting one make them less feasible on broader scales. An alternative, albeit with its own set of limitations, would be to conduct a multiverse analysis (Steegen et al., 2016). Multiverse analyses consist of a suite of analysis pathways, where diverging analysis decisions create a compounding set of possible answers to a given question. Multiverses cannot provide evidence for the correct answer, but can allow researchers to explore the sensitivity of their conclusions to any number of analytical variables (Rijnhart et al., 2021). Part of this exploration can disentangle the variation stemming from different sampling or analysis choices, compared to variation stemming directly from the effect of interest (Simonsohn, 2015; Steegen et al., 2016; Forstmeier, Wagenmakers & Parker, 2017). This can be particularly true when we use the multiverse in a synthetic or simulated environment where we are able to hold the effect of interest constant.

Previously we explored how individual habitat selection estimates were impacted by decisions regarding sampling and analysis (Marshall & Duthie, 2024). Broadly we found that sampling (data quantity) was more important than analysis decisions when estimating habitat selection, and that more modern analysis methods tend to produce less variable results. Here we build upon those findings to target population level habitat selection analyses. We use simulated animal data with a known habitat preference to explore the impacts of sampling and analysis choice in recovering the direction of that predefined preference.

2 Methods

2.1 Simulating the Scenarios

We used the abmAnimalMovement v.0.1.3.0 (Marshall & Duthie, 2022) R package to simulate the movement data of an animal with a predefined (i.e., known) habitat preference. The abmAnimalMovement package uses an agent-based approach to simulate terrestrial animal movement using raster environmental data to guide the animals decisions. We used the NLMR v.1.1.1 package (Sciaini et al., 2018) to generate the three required resource/environmental rasters: movement resistance, foraging quality, and shelter site quality. The abmAnimalMovement package has systems for simulating activity cycles, three separate behavioural states (resting, foraging, exploring; differing in movement characteristics and resource prioritisation), and site fidelity. For the purposes of this study, we used the pre-created pseudo-species presented in Marshall & Duthie (2022). 1. Badger. The badger is a terrestrial species subject to differing movement resistance across the landscape occupying several shelter sites, with an 8-12 hour activity cycle and minor seasonal variation. 2. Vulture. The vulture is treated as a terrestrial species but does not experience any movement resistance across the landscape. It uses three shelter sites, with an 12 hour activity cycle and minor seasonal variation. Unlike the randomly uniform resource landscape of the badger, the vulture's landscape is divided into two distinct areas of higher and lower foraging quality. 3. King Cobra. The king cobra is a terrestrial species subject to differing movement resistance across the landscape similar to the badger. It uses 12 shelter sites, and has a more complex activity cycle with an 8-12 hour diel cycle, a approximately weekly cycle of resting to mimic long digestion periods, and a weak seasonal cycle all working additively to define resting behaviour. The landscape is universally random similar to the badger, but contains two lines of heightened movement resistance serving as movement barriers.

For the purposes of the analysis we simplified the landscape information into categories –akin to the sort of land-use information more frequently available to researchers of animal movement. We focused on foraging quality because it influences the greatest amounts of movement (in terms of time) compared to sheltering or exploratory movements. We converted the continuous foraging quality raster into a binary, where higher quality areas (greater than 0.5) are classed as 2, and lower quality areas as 0. This classified landscape raster served as the basis for assessing habitat selection, in a scenario with a known preference for habitat 2. We also ran all described multiverse analyses (see below) with a second landscape classification. In this second scenario we scrambled the values of the cells, thereby breaking the connection between the underlying raster layers used to simulated the movement and the classification used in the analysis. This scrambled habitat classification allowed us to examining the variation in estimates in a zero-selection scenario, while serving as a direct comparison for the selection scenario.

We used that simulated landscape and abmAnimalMovement to simulate a population of 25, which was later sampled from for collecting movement data. All individuals in this population had the same simulation settings apart from starting location. Therefore, any variation between individuals was due to stochasticity rather than variation in the predefined habitat preference.

2.2 Sampling and Analysis Options

To manage the sampling of the population and the compounding growth of subsequent analysis decisions, we used the targets v.1.6.0 and tarchetypes v.0.9.0 R packages (Landau, 2021a,b). These packages allowed a branching



workflow pipeline, while keeping track of object creation thereby optimising the compute time required to explore the multiverse of analysis choice (Fig. 1).

2.2.1 Sampling

The first decision in most animal movement studies will concern tracking regime, consisting of decisions regarding tracking duration (the overall time the animal's movements are tracked) and tracking frequency (how frequently the animal is relocated). This decision is frequently dictated by more practical considerations such as anatomy and behaviour of the animal, cost of the tracking devices, and environmental factors. Here we covered a range of tracking regimes that vary in the frequency of location fixes (1 to 0.08 points per hour), and the total duration of tracking (7 to 60 days). We created sub-sampled datasets based on every combination of tracking frequencies and durations, provided they would result in greater than 30 data points per individual.

An important component of assessing population level habitat selection is the number of individuals included in analysis. Therefore, we randomly generated a number of samples from our population of 25 simulated individuals. We varied these sample sizes from 3 to 20 individuals, and ran 2 repeats for each size. A sample never mixed tracking regimes.

2.2.2 Analysis

Building on the decisions concerning tracking regime and population sampling, our multiverse expanded dramatically by exploring four primary analysis routes. These routes included an area-based approach, using Compana analysis (Compositional Analysis of Habitat Use), and three step-based approaches including averaged individual step-selection models, two-step conditional regression models, and a Poisson model.

We ran the Compana analysis using the *adehabitatHS* v.0.3.17 (Johnson, 1980; Aebischer, Robertson & Kenward, 1993; Calenge & Mathieu Basille, 2023) R package. Compana allows for the assessment of habitat selection of multiple animals in defined habitat types; it also allows habitat selection to be estimated at different scales. We explored two scales in the multiverse: selection within the home range (Type III), and selection within an overall population available area (Type II). The former design (Type III) requires availability habitats to be defined on an individual-by-individual basis; therefore, we are comparing individual use to individual's defined availability. While the latter (Type II) design uses a summarised population level availability; therefore, we are comparing individual use to the population defined availability.

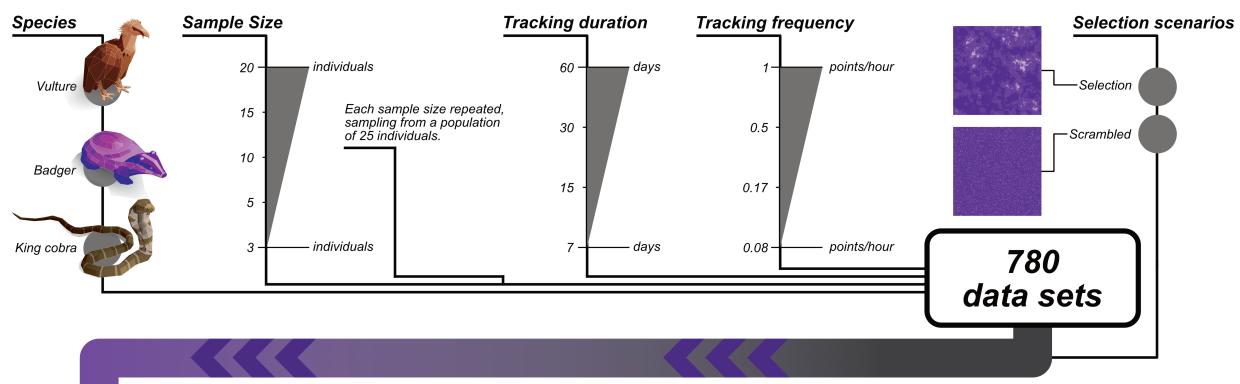
To define availability, we created home range polygons, then sampled points within, recovering the corresponding habitat at those points. The home range (AKA availability) polygons can be generated via many different processes. We explored MCP and AKDE (Minimum Convex Polygons, Autocorrelated Kernel Density Estimators). We used the *ctmm* v.1.2.0 (Fleming & Calabrese, 2023) package for the creation of AKDEs, and the *adehabitatHR* v.0.4.21 (Calenge & Scott Fortmann-Roe, 2023) package for the MCPs. The AKDEs required the fitting of a movement model; we fitted multiple models and selected the top scoring model by AIC. For the best model, we followed recommendations of Silva et al. (2022), opting to run AKDEs as weighted and using perturbative hybrid residual maximum likelihood method (pHREML), as both these options tend to be the most robust in scenarios of lower tracking frequencies. There are many options involved in the creation of AKDEs that could shape habitat availability and therefore estimations of selection. We avoided a deeper exploration of the impact of the internal AKDE choices, as they are likely smaller than those between area methods (e.g., compared to MCPs), and the relatively greater computational cost to generate many AKDEs. By comparison the MCPs have very few options impacting their creation. Mainly MCP size varies based on the percentage of outliers excluded from the polygon area. For all area methods, we vary this percentage: 95%, 99%, that correspond to increasingly large areas included in the availability polygon (i.e., covering areas that are less used by the individual). We selected these contour values as they are the most frequently used in spatial studies (Crane et al., 2021).

Once we had created all area polygons for each individual, for every combination of tracking regime (duration x frequency, total regimes: 13), we created the population available areas for the type II design analysis by combining all polygons into a single polygon. We used this combined polygon for the type II design, where all individuals have the same availability.

We used each availability polygon to define the extent within which we sampled habitat characteristics. Within those defined polygons we generated a number of random points per animal location to determine the available habitats. We varied both the number of points generated relative to the number of animal locations (1 to 10 times the number of known animal locations), as well as the pattern of how they were distributed within the polygon



Sampling



Analysis

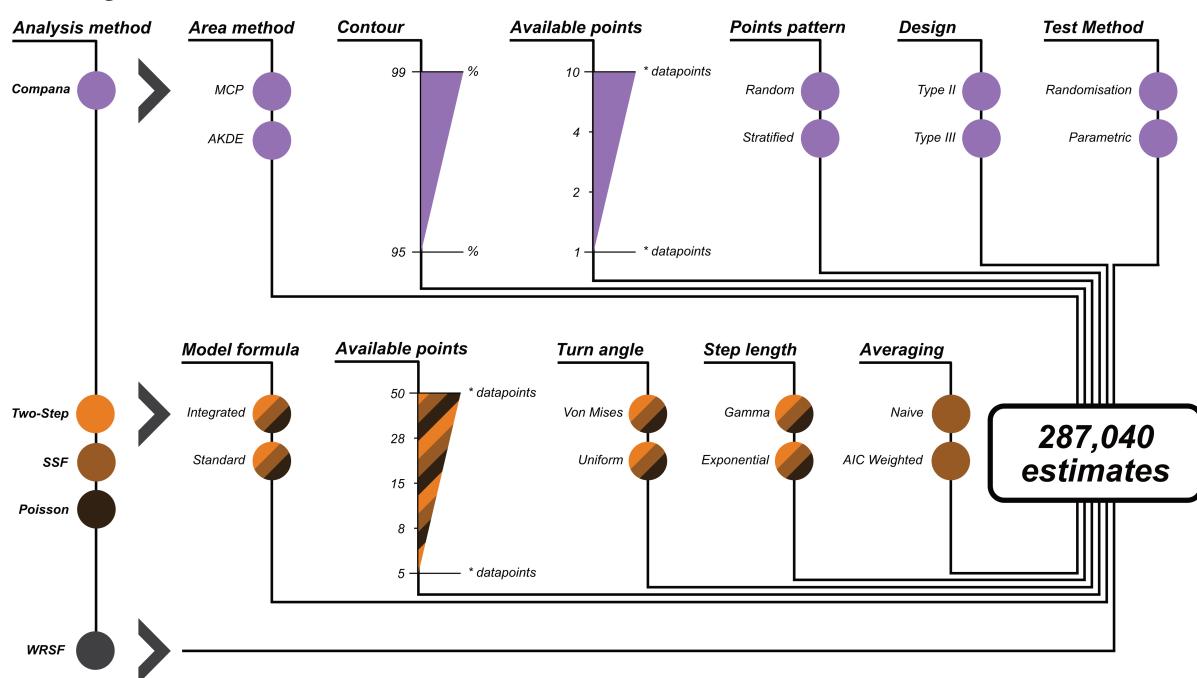


Figure 1. An illustration of the different decisions creating a multiverse examining animal movement. Species: three different movement archetypes simulations were loosely based upon. Sample Size: the number of simulated individuals per each species, with each random sample being repeated twice. Tracking duration: the length in days of the tracking data sampled from the overall simulated year. Tracking frequency: the frequency of the sampling of data points from the 1-per-minute simulated data. Selection scenarios: the two classified habitat landscapes used during analysis. Purple choices concerned with analysis pathways using area-based estimations of availability (Compana: Compositional Analysis). Area method: the method used to generate the available area polygon for the individual (MCP: Minimum Convex Polygon, and AKDE: Autocorrelated Kernel Density Estimator). Contour: the threshold used to convert the utilisation/occurrence distribution into a polygon. Available points: the number of generated points per data point to extract available habitat within the polygon. Points pattern: whether the random points were generated in a stratified or random pattern. Light orange choices pertain to the Two-Step analysis pathways. Medium orange-brown choices pertain to the SSF (Step Selection Function) analysis pathways. Dark brown choices pertain to the Poisson analysis pathways. Model formula: whether the step and turn angles were added as interactions to the habitat variable (integrated if added). Available points: the number of randomly generated available points per step/used point. Turn angle: the distribution used to generate the turn angles of the available points. Step distribution: the distribution used to generate the step distances of the available points. Dark orange is the WRSF (Weighted Resource Selection Function) analysis. Averaging: the method of model averaging used for the SSF models to gain an overall estimate. Decisions that required discrete choices are illustrated with circles, whereas those that continuous are depicted as triangles.



(random or stratified). Once the points were generated, we used them to extract the underlying habitat from classified landscape raster.

The final decision in the area-based method approach was whether the Compana analysis was tested using randomisation or in a parametric fashion. The former uses randomisation tests to estimate the habitat selection, while the latter uses a chi-squared distribution.

The other analysis pathways do not rely on an available area; instead they focus on randomly generated steps that mirror the observed movement of the animal. As such, the decisions involved in running the step-selection, two-step conditional regression, and Poisson models are largely the same.

The first decision is the number of random steps generated per observed step; we ranged this from 5 to 50. The generation of random steps is guided by two distributions, one describing the step lengths from the last observed location, and another describing the turn angle from the last observed direction of travel. We chose to explore the impacts of different distributions. For the step length, we used Gamma and exponential distributions; for the turn angle we used Von Mises and uniform distributions.

All the step-based methods require the definition of a formula, where use/availability is predicted by the habitat characteristics at those locations. Model formulation opens up many possible alternative routes of analysis, but for this exploration we focused on the impacts of integrating step and turn angle interactions with habitat characteristics. We ran two versions of all models, one with no interactions and one where both step lengths and turn angles interact with the habitat classification. Previous work has suggested that the integrated formulation (i.e., model with interactions) performs better, producing less biased estimates of habitat selection (Forester, Im & Rathouz, 2009).

A key component of all the population approaches is how they translate a highly structured dataset into an overall summary of habitat selection. We ran individual step selection models using the *amt* v.0.2.2.0 (Signer, Fieberg & Avgar, 2019) package, to explore summary methods (i.e., averaging regression coefficients Thurfjell, Ciuti & Boyce (2014)). The individual step selection models need to be averaged in some manner to extract a population mean selection. We explored two simple options for achieving this: a naive mean of the final estimates (i.e., a mean of the estimated betas), and a model average using the *MuMin* v.1.47.5 package (Bartoń, 2023). The *MuMin* model average approach is weighted based on AICc. However, AICc is not directly comparable between models fitted with different datasets so this does not necessarily present a solution for real-world scenarios as it is unlikely multiple animals would have identical amounts of locations recorded. In the simulated scenario here, as the model formulas are identically complex and the datasets equal in data quantity, the differences in AIC are purely describing the goodness of fit to the different individuals. Therefore, the model average produced by *MuMin* is weighted by model goodness of fit.

The difficulties in model averaging based on different data have led to the creation of models that account for individual variability in the model formula. A Two-Step Conditional Regression is one such solution, where strata (time steps) and clusters (individual animals) are accounted for (Craiu et al., 2011). We implemented these models using the *TwoStepCLogit* v.1.2.5 (Craiu et al., 2016) package, again using the same decisions as applied in the SSF approach (except model averaging). The Two-Step implementation in *TwoStepCLogit* required no NAs in landscape covariates (i.e., all locations had associated classified habitat information). In the rare instances where a simulated animal reached the edge of the landscape and available points were generated beyond the boundary (i.e., return a habitat class of NA), we repeated the random generation of points until fewer than 10 NAs existed in the data that were then excluded.

More recently Muff, Signer & Fieberg (2020) have suggested a reformulation of the step selection models to provide faster, reliable estimations of population habitat selection. The suggestion is to reformulate the model as a Poisson model, but critically with stratum-specific fixed effects. As this would result in a large number of fixed effects, the computational cost would be large; so instead they are treated as a random effect, but to avoid excessive shrinkage to the mean, a large variance prevents this. To avoid the slow convergence of MCMC estimation, Muff, Signer & Fieberg (2020) made use of integrated nested Laplace approximation (INLA) to approximate the Bayesian inference required for such a reformulation. We followed the code provided by Muff, Signer & Fieberg (2020) and made use of *INLA* v.23.4.24 (Rue, Martino & Chopin, 2009; Lindgren, Rue & Lindström, 2011; Martins et al., 2013; Rue et al., 2017; Kourounis, Fuchs & Schenk, 2018) to run Poisson models to estimate habitat selection. We applied the same suite of choices as applied to the SSF and Two-Step approaches.



2.3 Assessing the multiverse

Specification curves provide an overview of the estimates of a given range of analyses. Tighter more steep curves suggest greater agreement between all the analysis end points. Here we use specification curves and plotted the estimates against the different decisions that results in those estimates, allowing direct comparison on how the sampling and analysis decision impacts the variation in the estimates.

To better detect the impact of decisions, while accounting for the random variation stemming from the differences in individuals/samples, we ran a number of Bayesian Regression Models. The Bayesian Regression Models aimed to describe how much of the deviation from a median answer could be explained by the various sampling and analysis decisions. For each analysis route, we ran a model that included tracking frequency, tracking duration, sample size, and all the corresponding analysis choices. All continuous variables (x) were scaled to help determine their relative importance to each other: $(x - \text{mean}(x))/\text{sd}(x)$.

For the area-based Compana approach, population effects included: the continuous variable contour (continuous); and the categoric predictors sampling pattern (random, stratified), test (randomisation, parametric), area method (AKDE, MCP), and design type (II, III). For the step-based approaches they all included: model formula (integrated, non-integrated), step distribution (gamma, exponential), turn distribution (Von Mises, uniform). The step-selection model approach also included: averaging method (naive, MuMIn average).

We ran the Bayesian Regression Models using the *brms* v.2.21.0 (Bürkner, 2021) package. Once complete, we checked model convergence using R-hat values, ACF, and trace plots. After assessment of the model convergence measures, we modified the running parameters to 4000 iterations, a warmup of 750, and a thinning rate of 4 for all models bar the Poisson models that instead used 20000 iterations, a warmup of 8000, and a thinning rate of 20. We also examined the R^2 values, to determine how much of the variation in the results could be explained by the decisions.

3 Results

3.1 Specification Curves

The specification curves provide an overview of all the analysis end points ($n = 287,040$; Fig. 1). The area-based method – Compositional Analysis (Compana) – returned a median estimate that agreed with the underlying known simulation preference (i.e., preference for habitat 2; Fig. 2), and a near zero effect for the scrambled selection scenario. In the positive selection scenario very few estimates suggested negative selection, and those that did appeared to be extreme outliers. When broken down by decision, all decisions except higher sample sizes and the test method, appear to have a large range in final estimates. The sample size decisions had a clear reduction in possible answers when the sample is 15 or 20 individuals, but also a potential interaction with the test method when including 20 individuals. Unexpectedly, the lowest sample size of 3 also led to visibly reduced estimate variation, and the higher sample sizes had a lower rate of significant positive estimates (e.g., 6902/9984 significant results with 3 individuals, compared to only 254/9984 with 20 individuals). The test method presented two different outcomes, with the randomisation method presenting a far tighter grouping of estimates in both selection scenarios, and with higher rates of significance in the positive selection scenario (11663/24960 compared to 5432/24960). In the scrambled scenario, the grouping was reduced, but rates of significance were very similar (13985/24960 compared to 13175/24960; Table. 1).

The specification curve generated from summarised Step Selection Functions (SSFs) similarly revealed a median estimate that agrees well with the simulated parametrisation for both selection and scrambled scenarios (Fig. 3). Also like the Compana curve, the SSF curve had far fewer results that indicated very low habitat preference in the positive selection scenario. The majority of these lower estimates of habitat selection appeared connected to the decision to use an non-integrated (standard) model formulation, and the decision to model average weighted by AIC. The naively model averaged estimates were also much more tightly grouped with fewer estimates failing to correctly detect positive preference; the tighter grouping was also reflected in the scrambled selection scenario.

Outside the lower estimates, the clearest structure in the estimates results from different decisions in sample size and tracking regime. The tracking regime decisions, frequency and duration, appeared to shift the overall range of estimates but with limited impact on the variation. By comparison, sample size increases appeared to reduce the spread of estimates produced.

Compared to the Compana analysis there was a cleaner pattern of significance, with the selection scenario having much higher rates of significant selection compared to the scrambled scenario (Table. 2). We also saw a clear increase in significant answers when we increased the sample size in the selection scenario (e.g., sample size of 3

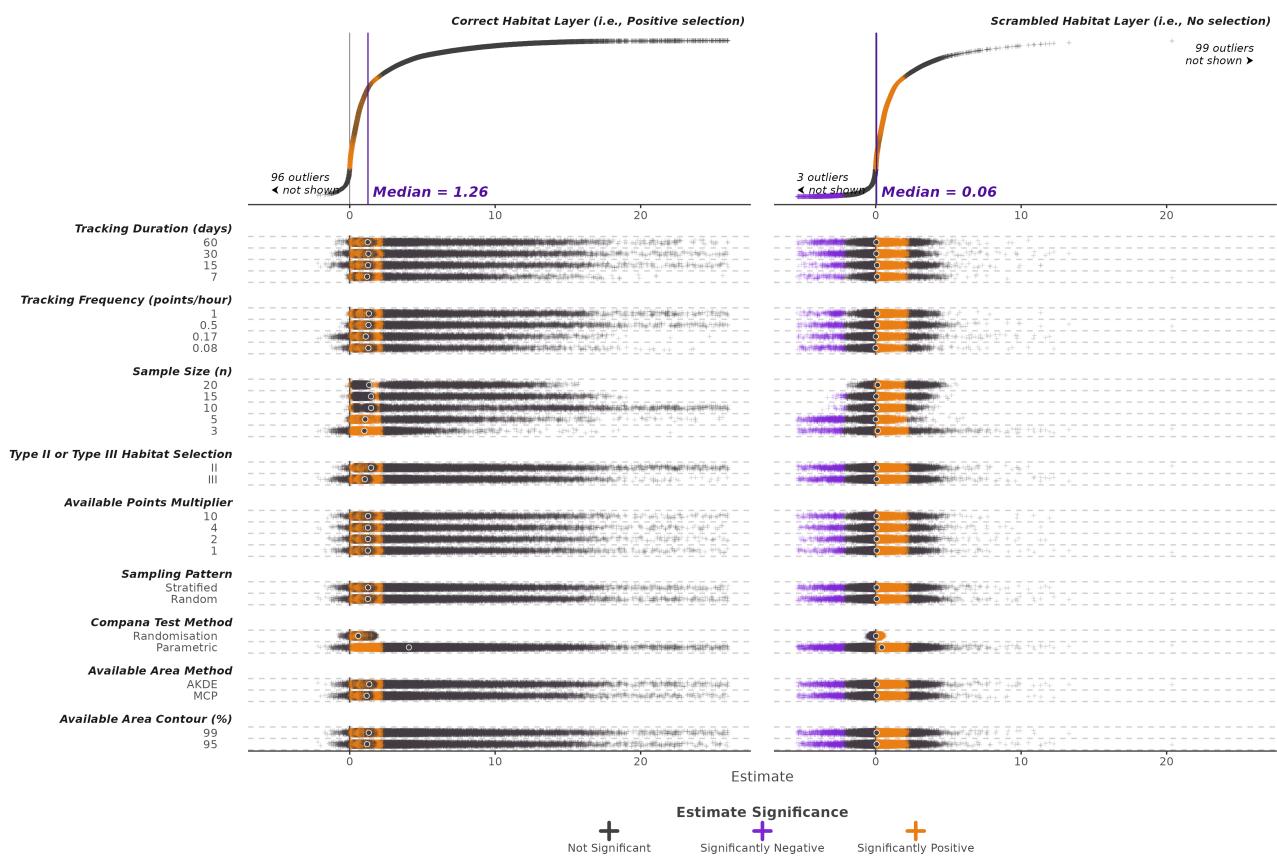


Figure 2. A specification curve showing habitat selection estimates resulting from the Compana (area-based) analysis pathways when applied to all simulated animal movement datasets. Every point represents a separate estimate of habitat selection. The top plot shows all estimates organised by estimate strength, with purple vertical line depicting the median estimate. Instances of outlying high or low estimates are labelled instead of plotted. The lower plot shows the same estimates, but split by the analysis or sampling decisions. Solid circles are the medians for each choice.

1687/6240 compared to 20 4322/6240).

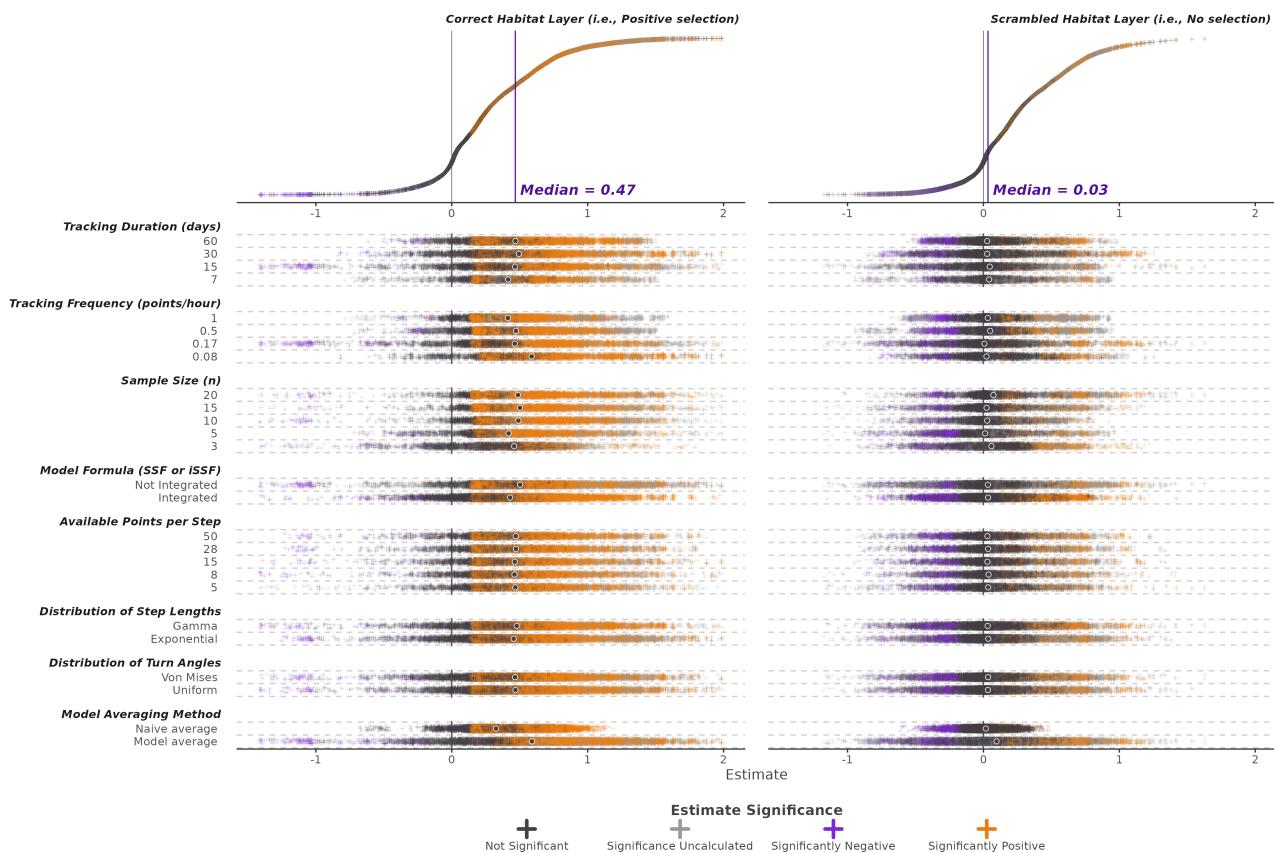


Figure 3. A specification curve showing habitat selection estimates resulting from the Step-selection Function (SSF) analysis pathways when applied to all simulated animal movement datasets. Every point represents a separate estimate of habitat selection. The top plot shows all estimates organised by estimate strength, with purple vertical line depicting the median estimate. Instances of outlying high or low estimates are labelled instead of plotted. The lower plot shows the same estimates, but split by the analysis or sampling decisions. Solid circles are the medians for each choice.

The two-step model approach appeared to have had the greatest difficulty consistently detecting the positive selection, but the median selection in the positive selection scenario was still greater than the median from the scrambled scenario (Fig. 4). Tracking frequency appeared to be a critical decision in the two-step model, with the 1 point per hour choice resulting in more detections of positive selection, and a tighter grouping of estimates. The decision to integrate step and turn angle into the model formula appeared similarly critical. The choice not to include the step and turn angle resulted in estimates more consistently positive in the positive selection scenario, and close to zero in the scrambled selection scenario. The integrated model formulation appeared to be largely responsible for the majority of the variation in estimates. Unlike the SSF results, the Two-Step approach appeared to provide significant answers (both positively and negatively) very readily (Table. 4). The balance of positively significant did appear more positive skewed in the selection scenario (128429/327600: 39%) compared to the scrambled scenario (41069/327600: 13%), but all choices resulted in numbers of false positives and negatives bar the non-integrated model formulation choice.

The Poisson based approach, like those before, showed that the vast majority of analysis pathways will result in the correct identification of positive preference for habitat 2 (Fig. 5), and centre on zero for the scrambled scenario. Unlike the previous analyses, the tracking duration did not have as clear an impact on the spread of estimates. Sample size and tracking frequency did; larger sample sizes and higher frequency led to overall decreases in estimate spread. Of particular note is that the majority of failures to detect positive preference occurred when the sample size is three and at lower tracking frequencies. The analysis decisions appeared largely inconsequential, particularly the number of points per step. Model formulation did have an impact, with the integrated formulation being more consistent at retrieving positive preference in the positive selection scenario, but the spread of estimates was much larger. This spread increase associated with the integrated model formulation was similarly seen in the scrambled selection scenario, but resulted in near identical median estimates across model formulations. Unexpectedly, despite

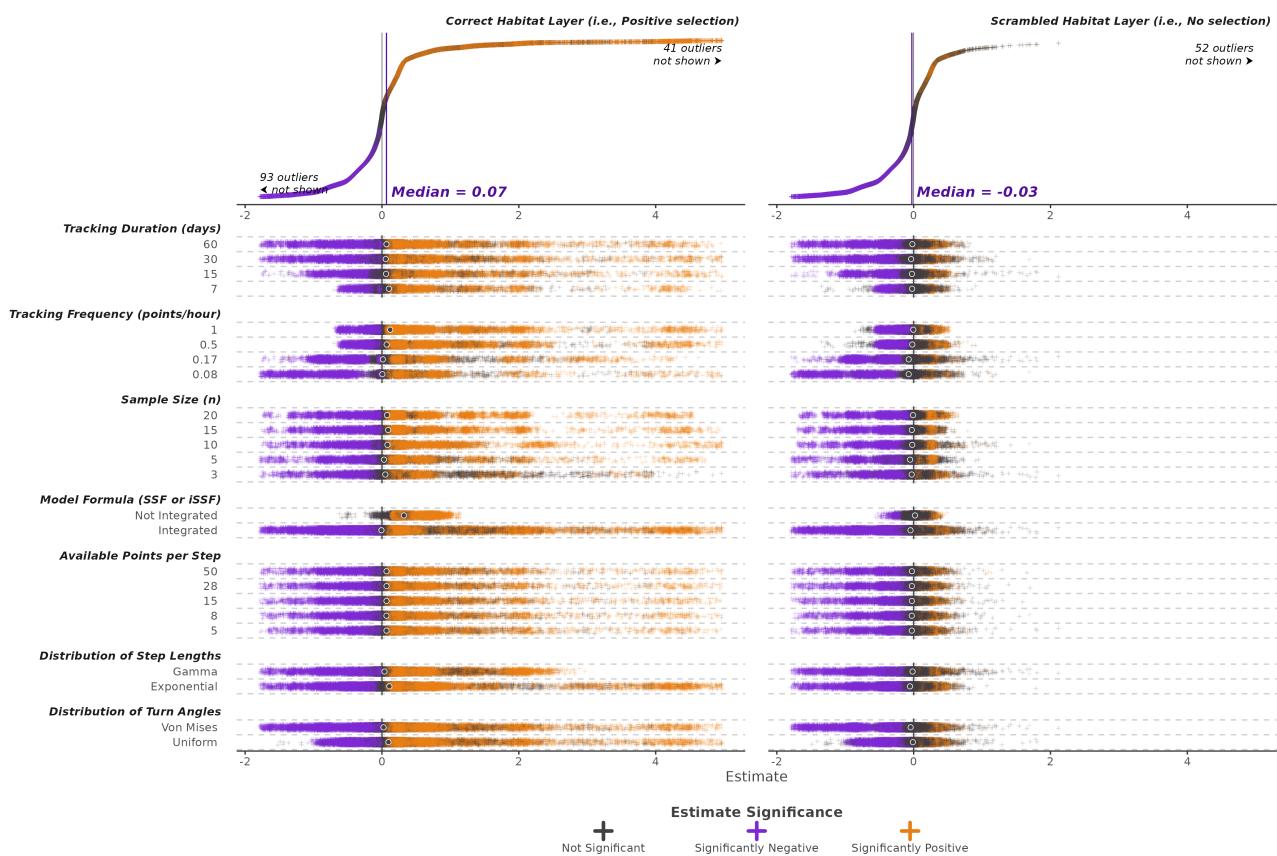


Figure 4. A specification curve showing habitat selection estimates resulting from the Two-step analysis pathways when applied to all simulated animal movement datasets. Every point represents a separate estimate of habitat selection. The top plot shows all estimates organised by estimate strength, with purple vertical line depicting the median estimate. Instances of outlying high or low estimates are labelled instead of plotted. The lower plot shows the same estimates, but split by the analysis or sampling decisions. Solid circles are the medians for each choice.

quite clear positive selection in many estimates in the selection scenario, all estimates (regardless of scenario) were deemed not to be significant, with 95% confidence intervals overlapping zero (Table. 3).

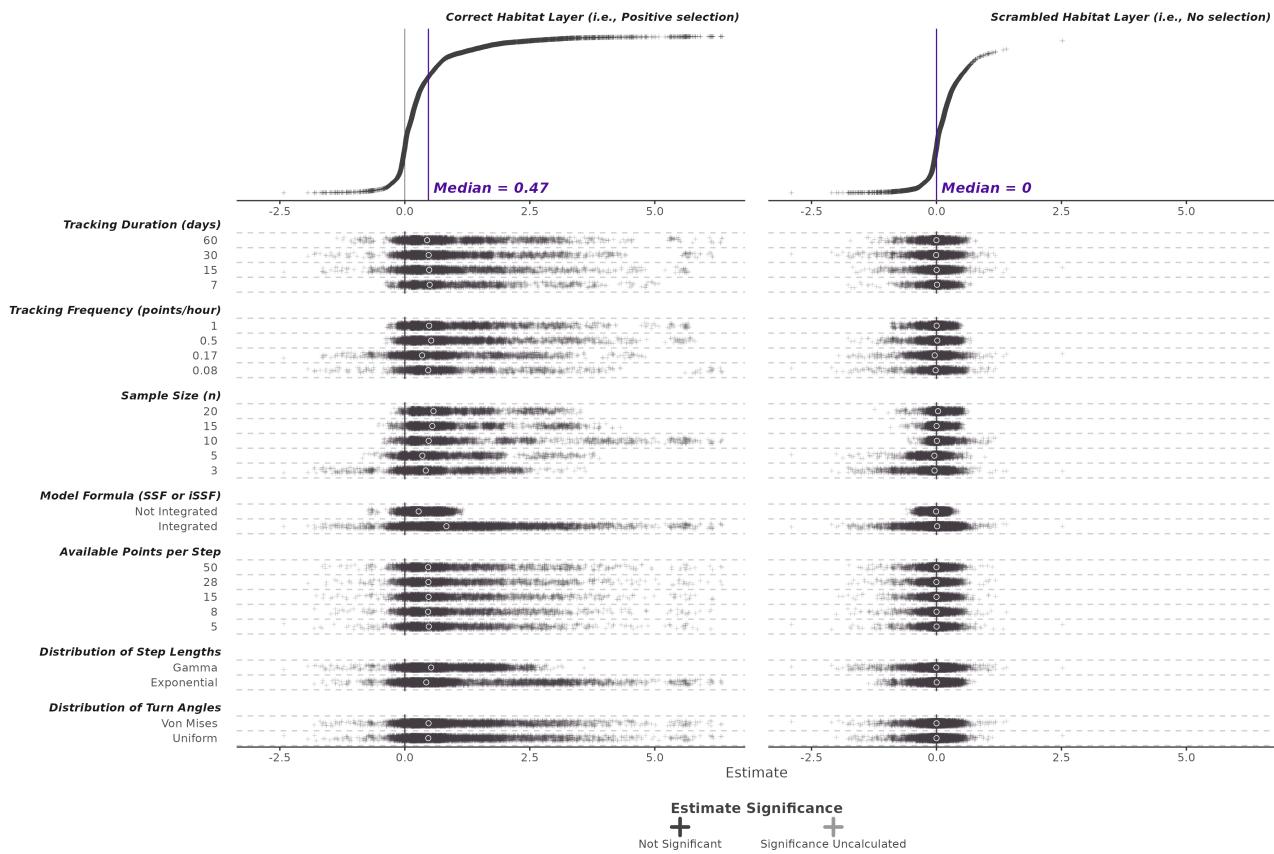


Figure 5. A specification curve showing habitat selection estimates resulting from the Poisson analysis pathways when applied to all simulated animal movement datasets. Every point represents a separate estimate of habitat selection. The top plot shows all estimates organised by estimate strength, with purple vertical line depicting the median estimate. Instances of outlying high or low estimates are labelled instead of plotted. The lower plot shows the same estimates, but split by the analysis or sampling decisions. Solid circles are the medians for each choice.

3.2 Model Results

We also used four Bayesian Regression Models to explore the impact of different decisions (Fig. 6), and whether those decisions could predict deviation from the median estimate (i.e., as a proxy for seeing which decisions lead to outlying extreme results). Greater variation would indicate greater scope for researcher choice to have a significant impact on the final results. The models provided a means to account for the variation that originates from stochasticity in the simulation and sampling captured by the group effects and reflected in the difference between marginal and conditional R^2 values, as the sheer number of final estimates can obscure patterns in specification curves.

The conditional R^2 values differed for the four models, ranging from 0.14 to 0.32, indicating a sizeable amount of variation in estimates being uncounted for by either the analysis discretions of random/group effects of sample ID, selection scenario, and species. The Compana results model had a conditional R^2 of 0.32; whereas the SSF model returned 0.29; the Two-Step model returned 0.14; and the Poisson model returned 0.3.

The marginal R^2 represents the bulk of the conditional R^2 for Compana and SSF models, suggesting an important role for the fixed/population effects. The Compana results model had a conditional R^2 of 0.18; whereas the SSF model returned 0.18. The Two-Step model returned marginal R^2 of 0.05; and the Poisson model returned 0.11. The low marginal R^2 in the Two-Step model highlights the greater sensitivity to variation between species, individuals and samples, compared to the variation resulting from analysis decisions. The same is true for the Poisson model, but to a lesser degree.

The Poisson model approach showed the most consistent benefits from more data (Fig. 6), where all increases in



sampling intensity led to marked reductions in deviation from the median estimate: sample size ($\beta = -0.02$; 95% HDCI -0.06 - 0.02), tracking duration ($\beta = -0.02$; 95% HDCI -0.02 - -0.01), and tracking frequency ($\beta = -0.02$; 95% HDCI -0.03 - -0.02). Similarly SSF methods benefited from increasing sampling duration ($\beta = -0.04$; 95% HDCI -0.04 - -0.04) and frequency, but the impact on increasing sample size was not as apparent ($\beta = -0.04$; 95% HDCI -0.04 - -0.04). Two-Step methods appeared to benefit from increases in tracking frequency ($\beta = -0.08$; 95% HDCI -0.09 - -0.08), but not duration ($\beta = 0.00$; 95% HDCI 0.00 - 0.01) or sample size ($\beta = 0.00$; 95% HDCI -0.02 - 0.01). Potentially the benefits from some of these decisions are masked by the extreme variation in the integrated model formulation decision. The area-based Compana approach appeared to vary more in its estimates when provided more data. Sample size ($\beta = 0.28$; 95% HDCI 0.10 - 0.47), tracking duration ($\beta = 0.06$; 95% HDCI 0.04 - 0.07) and tracking frequency ($\beta = 0.12$; 95% HDCI 0.10 - 0.13) all led to increased variation from the median estimate.

The points per step decision made little impact on the variation of estimates in any of the step-based approaches (Fig. 6), potentially due to the simplified nature of the landscape used (i.e., no rare habitat that would require high numbers of random points to detect). Use of the gamma distribution, in contrast to the exponential distribution, in all the step-based approaches led to reduced deviation from the median (Poisson: ($\beta = -0.07$; 95% HDCI -0.08 - -0.06); SSF: ($\beta = -0.01$; 95% HDCI -0.01 - -0.01); Two-Step: ($\beta = -0.09$; 95% HDCI -0.10 - -0.09)). Whereas the use of Von Mises, in contrast to a uniform distribution, failed to have a sizeable impact in Poisson ($\beta = 0.00$; 95% HDCI -0.01 - 0.01) and SSF ($\beta = 0.00$; 95% HDCI 0.00 - 0.00) methods and led to greater variation in Two-Step estimates ($\beta = 0.09$; 95% HDCI 0.08 - 0.10). The underlying simulation used a gamma distribution to generate step lengths and Von Mises for turn angles, which may partially explain this result.

The most mixed decision was the use of a model formulation that includes the step and turn angles (i.e., integrated versus not integrated; Fig. 6). Here we saw including the step and turn angles in the model formula for the summarised SSF approach led to greater variation from the median estimate ($\beta = 0.04$; 95% HDCI 0.04 - 0.05), but a contrasting effect for both the Two-step ($\beta = -0.17$; 95% HDCI -0.17 - -0.16) and Poisson approaches ($\beta = -0.37$; 95% HDCI -0.38 - -0.36) where the non-integrated formula tended to reduce the spread of estimates.

The greater deviation in averaged SSFs when using the integrated formulation may be a result of its interaction with the model averaging approach. The naive model averaging approach led to a dramatically lower spread of estimates ($\beta = -0.18$; 95% HDCI -0.18 - -0.17), and avoided underestimating selection in the positive selection scenario (unlike the AIC model average approach; (Fig. 3).

Finally, the area-based approach had a number of unique decisions. Largely the decisions associated with defining the available area had a larger impact than those linked to generating the random points (Fig. 6). In brief, the use of MCPs ($\beta = -0.16$; 95% HDCI -0.19 - -0.13) and type III designs ($\beta = -0.29$; 95% HDCI -0.32 - -0.26) tended to lead to less variable results. Larger contours (areas) ($\beta = 0.05$; 95% HDCI 0.03 - 0.06) and more points ($\beta = 0.02$; 95% HDCI 0.01 - 0.04) unexpectedly increased variation of estimates from the median. The effect of larger contours and type III designs appeared to contradict each other, as both would tend towards a larger area that would presumably result in similar results. By far, the clearest result in the area-based approach was the Compana test method, where randomisation dramatically reduced the spread of estimates ($\beta = -2.32$; 95% HDCI -2.35 - -2.29).

4 Discussion

The multiverse approach highlights the diverse array of answers one can obtain even when analysing the same data, simulated using the same parameters. Fortunately the vast majority of the answers from the 287,040 analysis end points agree, and correctly recover the positive habitat selection programmed into the simulated animals in the positive selection and correctly recover the absence of habitat selection in the no selection scenarios. This broad agreement is cause for optimism, suggesting that the potentially deviating analysis choices made by researchers largely converge to an agreement –albeit within this given simulated scenario and given habitat preference strength. Looking back at past studies reporting habitat preference, we have considerable confidence that different (or reported) choices during analysis likely would not have changed the overall conclusions.

Unlike previous multiverse investigations into individual selection (Marshall & Duthie, 2024), the largest effect on whether an estimate was close to the median estimate did not come as a result of sampling (either tracking regime or number of individuals). This is possibly due to the overall increase in data quantity that came from running a multi-individual sample, perhaps past the point of benefiting as dramatically from more data. There was no programmed individual variation (which is a key consideration in real-world scenarios Stuber, Carlson & Jesmer (2022)), therefore any variation between individuals of the same pseudo-species was the result of stochasticity. This lack of individuality may dampen the variation, reducing effects of increased sample sizes. However, a failure to reduce the deviation from the median estimate does not mean that increased sample sizes were not beneficial. Increased sample sizes may have actually led to estimates that were closer to the “true” selection of the animal, and

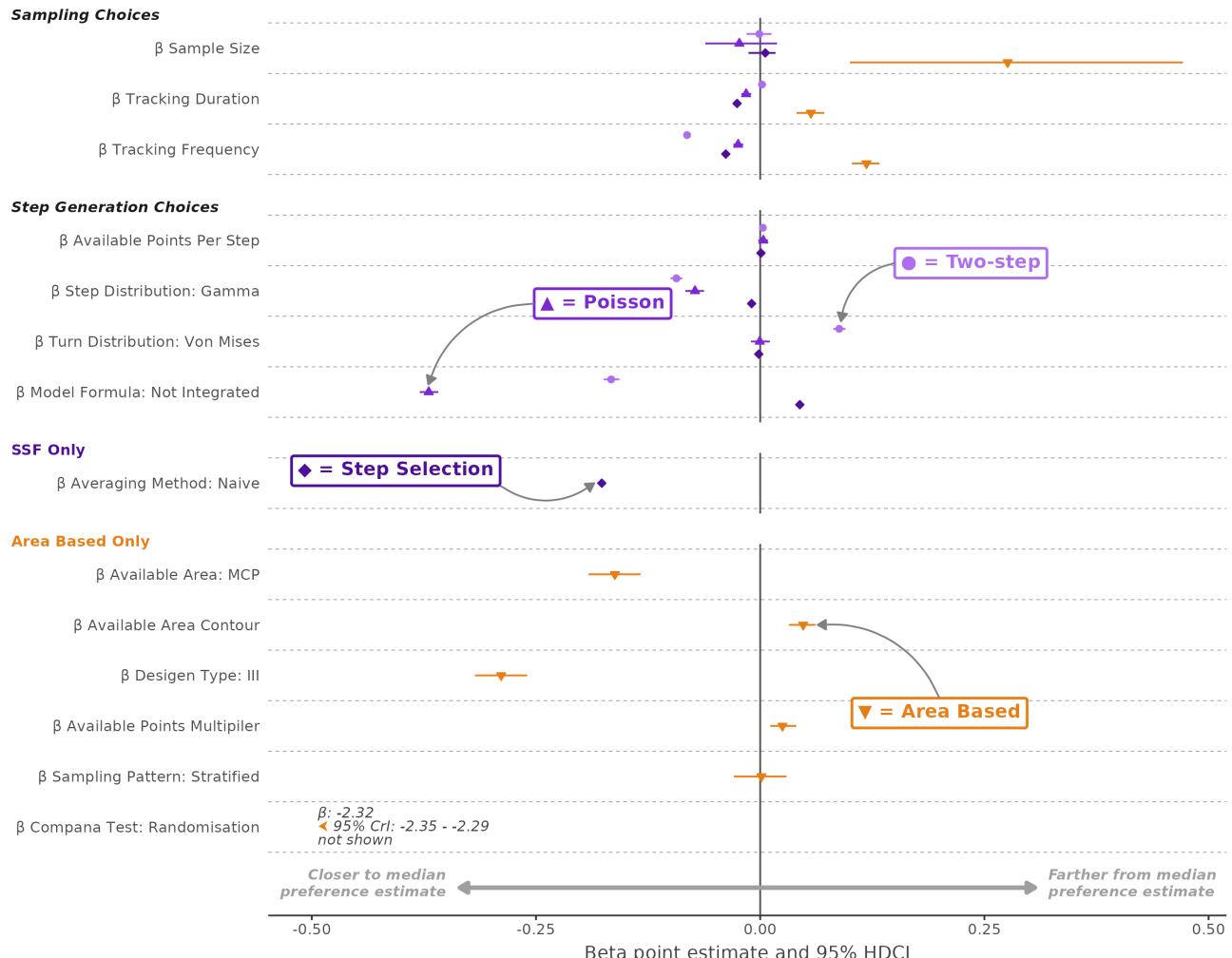


Figure 6. Point estimates and 95% credible intervals associated with all population effects from the Bayesian Regression models aiming to predict deviation from the median estimate using the information on analysis choices. Negative effects indicate that the decisions tends to decrease the deviation from the median estimate, whereas positive effects suggest those decisions are increasingly connected to estimates far from the median estimate. The effects are grouped into those pertaining to sampling choices and tracking, those associated with the step-based analysis approaches (Two-Step, SSF, and Poisson), those connected just to the SSF approach, and those connected to the area-based Compana approach.



Craiu et al. (2011) suggests greater stability in estimates with more data.

When examining the impact of individual analysis choices, we see a few that appear to be key. In the area-based (Compana) analysis, the decision to test selection using randomisation led to dramatically more consistent estimates. We suspect this is due to the parametric method making more stringent assumptions regarding the data's distribution, resulting in a greater frequency of unstable outlying estimates. For the SSF and Poisson models, the decision on integrating step and turn angle into the model formulation appears critical. For SSF, is it suggested that integrating step and turn is ideal as it limits bias by accounting for the movement tendencies of the animal, particularly for high selection scenarios (Forester, Im & Rathouz, 2009). We saw reduced extreme high and low estimates when using the integrated formulation that supports this.

Poisson models are similarly affected, but in the opposite direction. Muff, Signer & Fieberg (2020) previously highlighted the potential contrasting impact of integration in the Poisson models, and our multiverse mirrors this with the integrated formulations leading to estimates of greater strengths and that varied more. In the context of Muff, Signer & Fieberg (2020), we suspect the more outlying variable estimations are more bias results and ultimately further from the true selections strength. The most deviated estimations appear at low or intermediate sample sizes, and low tracking frequencies for the outlying low estimates. This could indicate that any bias resulting from the integrated formula could be mitigated with more data, but further investigation would be required to confirm this as we do not see greater tracking durations eliminating outlier estimates.

Outside of approach-specific decisions we can compare the different approaches. While we should be cautious comparing the different approaches as target and interpret selection in different ways, we can see that SSF, Poisson, and Compana approaches do better than the Two-Step approach when correctly identifying selection and no-selection. The Two-Step approach appears to be overly keen to consider selection significant (in either direction), and could frequently encounter convergence issues. The latter is likely caused by data inadequacies in either the strata or clusters (Craiu et al., 2011). Compana did appear to be more consistent than Two-Step, the SSF and Poisson approaches appear to react more predictability to improvements in data quantity and sample size. The SSF and Poisson approaches appear to be the stronger choices and offer increased flexibility with regards to habitat classification (i.e., allow for continuous habitat variables).

4.1 Multiverse in context

While multiverses can explore the possible answers that can come from a given dataset, they cannot provide guidance on which of those answers is correct. The creation of the multiverse is subject to the same decision making procedure that any single given pathway is: what decisions to include, and how to vary the choices (Simonsohn, Simmons & Nelson, 2020; Del Giudice & Gangestad, 2021). Therefore, the overall median or mean answer is heavily dependent on the construction of the multiverse. Additionally, the choices and resulting analysis pathways are not equally as valid, nor equally likely to be undertaken by a researcher. This means the multiverse of answers may provide some insight on spread of estimates (Steegen et al., 2016), but should not be directly interpreted as providing more or less support for a given estimate based on the distribution of estimates. We can take this issue further, reminding ourselves that agreement is not necessarily a reliable proxy for accuracy (Devezzer et al., 2021). A better model reflecting the mechanistic reality of the study system is preferable to a suite of models that agree but converge on a less accurate answer. An example in the conducted multiverse: we used AKDEs and MCPs as two alternative definitions of availability for the area-based method. Arguably we should have excluded the MCP choice in favour of the AKDE choice, as the latter does a better job at capturing the movement processes/patterns underlying animal space-use (Fleming et al., 2015). Despite this, we felt, the frequency of MCP use in spatial ecology (Crane et al., 2021), warranted its inclusion as a conceivable choice. Judgement calls such as this are unavoidable, and determining whether a model better reflects the underlying system can be exceptionally difficult outside of simulated “known-truth” scenarios. It very well may be ideal to conduct multiple analysis as standard procedure to explore selection at different scales (Aebischer, Robertson & Kenward, 1993).

When we are uncertain of the mechanism, or unable to truly model the vast complexity of the system (as is very frequent in ecology), then informed agreement may be our best option. Recent efforts looking at how many researchers answer the same question from the same dataset reveals that these judgement calls result in different answers (Gould et al., 2023). In many ways the more organic real-world results from Gould et al. (2023) reflect the multiverse results presented here –a general agreement on direction, but variation in uncertainty and magnitude. What differs, is that here we know the simulated correct answer is a positive effect, while in Gould et al. (2023), the truth is unknown, so we rely on the good agreement of many independent informed analyses. The logistics of organising 246 researchers to assess uncertainty in analyses is infeasible on a broader scale. The multiverse approach presents a complementary method for assessing the scope of researcher choice on final results –albeit more vulnerable to the individual researcher's biases and potentially limited to analyses with lower computational



costs.

We have avoided suggesting rules or guides based on these results, instead reporting broad trends. The diverse array of species tracked and the habitats in which they reside makes specific guidance nigh impossible without informed assumptions tailored to a given study system. We simulated a three species to try and broaden the generalisability of our findings. However, they all shared core assumptions stemming from the agent-based simulation, namely: site fidelity, consistent cycadian cycles, and distinct behavioural states (Marshall & Duthie, 2022). For example, the routine use of shelter sites in a the preferred habitat would mean that missing those sites/locations is unlikely with a reasonably frequent tracking regime. This may have dampened the benefits of tracking frequency in our scenarios, and comparatively in scenarios where the use of preferred habitat is more fleeting or temporally inconsistent tracking frequency may be more crucial. Other studies have highlighted that certain decisions may have increased importance in different scenarios. The number of available points in SSF analysis becomes more important when the analysis is required to identify the use of smaller/rarer habitat types (Thurfjell, Ciuti & Boyce, 2014). Some methods, such as Resource Selection Functions, have the means of estimating the required data quantity to detect a given strength of selection in a given landscape, and provide a guide to how consistent results would be from that dataset (Street et al., 2021). These methods still require informed decisions to be made concerning the strength of selection that is of interest as well as how to define the landscape, but complement the goal of the multiverse assessment with a more rigorously mathematical approach.

Habitat selection analyses continue to expand and develop further, for example, to better tackle the non-independence of observed animal locations while building agreement between movement (i.e., local selection) and overall habitat selection (Michelot, Blackwell & Matthiopoulos, 2019; Michelot et al., 2020), integrate non-linear relationships (Klappstein et al., 2024), and develop methods less influenced by data missingness (Michelot et al., 2019; Hofmann, Cozzi & Fieberg, 2024). Repeating multiverse-like assessments may present a valuable context for these analyses, which is independent of the specific modelling approach.

4.2 Conclusions

Overall our multiverse exploration of habitat selection analyses shows that the majority of analysis pathways correctly identify the pre-programmed selection present in the simulated scenarios. While there was broad agreement, which was enhanced in most cases with increased data, several analysis choices highlight the continued need for attention when analysing animal movement data as the impact of those choices will not be the same under different analysis approaches.

5 Acknowledgements

This work was supported by the Natural Environment Research Council (NERC) via the IAPETUS2 Doctoral Training Partnership held by Benjamin Michael Marshall (grant reference NE/S007431/1).

6 Software availability

In addition to packages already mentioned in the methods we also used the following.

We used *R* v.4.2.2 (R Core Team, 2023) via *RStudio* v.2024.09.0 (RStudio Team, 2022). We used *here* v.1.0.1 (Müller, 2020) and *qs* v.0.26.3 (Ching, 2023) to manage directory addresses and saved objects.

We used *raster* v.3.6.26 (Hijmans, 2023) and *RandomFields* v.3.3.14 (Schlather et al., 2015) to aid landscape raster creation alongside *NLMR* v.1.1.1 (Sciaini et al., 2018).

We used *ggplot2* v.3.5.1 for creating figures (Wickham, 2016), with the expansions: *patchwork* v.1.2.0 (Pedersen, 2022), *ggridges* v.0.5.6 (Wilke, 2022), *ggdist* v.3.3.2 (Kay, 2023a), and *ggtext* v.0.1.2 (Wilke & Wiernik, 2022).

We used *brms* v.2.21.0 (Bürkner, 2021) to run Bayesian models, with diagnostics generated used *bayesplot* v.1.11.1 (Gabry et al., 2019), *tidybayes* v.3.0.6 (Kay, 2023b), and *performance* v.0.11.0 (Lüdecke et al., 2021).

We used the *dplyr* v.1.1.4 (Wickham et al., 2023), *tibble* v.3.2.1 (Müller & Wickham, 2023), and *stringr* v.1.5.1 (Wickham, 2022) packages for data manipulation.

We used *sp* v.2.1.4 (Bivand, Pebesma & Gomez-Rubio, 2013), *move* v.4.2.4 (Kranstauber, Smolla & Scharf, 2023) for manipulation of spatial data and estimation of space use not otherwise mentioned in the methods.



We used rmarkdown v.2.27 (Xie, Allaire & Grolemund, 2018; Xie, Dervieux & Riederer, 2020; Allaire et al., 2023), bookdown v.0.39 (Xie, 2016, 2022), tinytex v.0.51 (Xie, 2019, 2023a), and knitr v.1.47 (Xie, 2014, 2015, 2023b) packages to generate type-set outputs.

We generated R package citations with the aid of *grateful* v.0.2.4 (Francisco Rodríguez-Sánchez, Connor P. Jackson & Shaurita D. Hutchins, 2023).

7 Data availability

All code used to simulate the movement data, implement the multiverse of analysis, and model the variation in resulting estimates is available at <https://github.com/BenMMarshall/multiverseHabitatPop>. We have also archived this code at [ZENDO REPO HERE WHEN PUBLIC]. We have included further model diagnostic plots not included here in the repository.



8 Supplementary Material

Table 1. An overall summary of all Compana analyses and associated decisions in relation to the frequency of significant results.

Analysis	Selection Scenario	Decision	Value	Significantly Negative	Not Significant	Significantly Positive	Total Estimates	Significance Uncalculated
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	7	0	4852	2828	7680	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	15	0	7654	3866	11520	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	30	0	10090	5270	15360	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	60	0	10229	5131	15360	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.08	0	4871	2809	7680	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.17	0	7198	4322	11520	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.5	0	10288	5072	15360	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	1	0	10468	4892	15360	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	3	0	3082	6902	9984	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	5	0	3716	6268	9984	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	10	0	7039	2945	9984	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	15	0	9258	726	9984	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	20	0	9730	254	9984	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Type II or Type III Habitat Selection	III	0	15353	9607	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Type II or Type III Habitat Selection	II	0	17472	7488	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Points Multiplier	1	0	8093	4387	12480	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Points Multiplier	2	0	8231	4249	12480	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Points Multiplier	4	0	8213	4267	12480	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Points Multiplier	10	0	8288	4192	12480	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sampling Pattern	Random	0	16376	8584	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Sampling Pattern	Stratified	0	16449	8511	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Compana Test Method	Parametric	0	19528	5432	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Compana Test Method	Randomisation	0	13297	11663	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Area Method	MCP	0	15950	9010	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Area Method	AKDE	0	16875	8085	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Area Contour (%)	95	0	16144	8816	24960	0
Compana	Correct Habitat Layer (i.e., Positive selection)	Available Area Contour (%)	99	0	16681	8279	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	7	147	2868	4665	7680	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	15	206	4885	6429	11520	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	30	406	7020	7934	15360	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	60	338	6890	8132	15360	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.08	283	4394	3003	7680	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.17	250	6353	4917	11520	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.5	294	5730	9336	15360	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	1	270	5186	9904	15360	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	3	310	3978	5696	9984	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	5	677	4359	4948	9984	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	10	46	4036	5902	9984	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	15	64	4589	5331	9984	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	20	0	4701	5283	9984	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Type II or Type III Habitat Selection	III	528	10757	13675	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Type II or Type III Habitat Selection	II	569	10906	13485	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Points Multiplier	1	259	5383	6838	12480	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Points Multiplier	2	269	5388	6823	12480	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Points Multiplier	4	295	5435	6750	12480	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Points Multiplier	10	274	5457	6749	12480	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sampling Pattern	Random	557	10791	13612	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Sampling Pattern	Stratified	540	10872	13548	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Compana Test Method	Parametric	1027	10758	13175	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Compana Test Method	Randomisation	70	10905	13985	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Area Method	MCP	543	10859	13558	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Area Method	AKDE	554	10804	13602	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Area Contour (%)	95	571	10704	13685	24960	0
Compana	Scrambled Habitat Layer (i.e., No selection)	Available Area Contour (%)	99	526	10959	13475	24960	0

**Table 2.** An overall summary of all SSF analyses and associated decisions in relation to the frequency of significant results.

Analysis	Selection Scenario	Decision	Value	Significantly Negative	Not Significant	Significantly Positive	Total Estimates	Significance Uncalculated
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	7	5	1663	1966	4800	1166
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	15	84	2104	3494	7200	1518
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	30	22	2600	5178	9600	1800
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	60	44	2014	5946	9600	1596
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.08	4	1397	2782	4800	617
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.17	102	2576	3481	7200	1041
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.5	34	2342	5141	9600	2083
SSF	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	1	15	2066	5180	9600	2339
SSF	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	3	62	3176	1687	6240	1315
SSF	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	5	28	2066	2845	6240	1301
SSF	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	10	28	1504	3509	6240	1199
SSF	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	15	12	852	4221	6240	1155
SSF	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	20	25	783	4322	6240	1110
SSF	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Integrated	95	5480	10025	15600	0
SSF	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Not Integrated	60	2901	6559	15600	6080
SSF	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	5	17	1740	3160	6240	1323
SSF	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	8	28	1689	3246	6240	1277
SSF	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	15	30	1667	3322	6240	1221
SSF	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	28	41	1640	3407	6240	1152
SSF	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	50	39	1645	3449	6240	1107
SSF	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Exponential	87	4437	8036	15600	3040
SSF	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Gamma	68	3944	8548	15600	3040
SSF	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Uniform	64	4146	8329	15600	3061
SSF	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Von Mises	91	4235	8255	15600	3019
SSF	Correct Habitat Layer (i.e., Positive selection)	Model Averaging Method	Model average	155	3440	5925	15600	6080
SSF	Correct Habitat Layer (i.e., Positive selection)	Model Averaging Method	Naive average	0	4941	10659	15600	0
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	7	81	3178	375	4800	1166
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	15	191	4777	714	7200	1518
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	30	415	6440	945	9600	1800
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	60	410	6250	1344	9600	1596
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.08	174	3508	501	4800	617
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.17	211	5283	665	7200	1041
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.5	340	6116	1061	9600	2083
SSF	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	1	372	5738	1151	9600	2339
SSF	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	3	253	4050	622	6240	1315
SSF	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	5	529	4140	270	6240	1301
SSF	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	10	91	4561	389	6240	1199
SSF	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	15	129	4237	719	6240	1155
SSF	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	20	95	3657	1378	6240	1110
SSF	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Integrated	837	12437	2326	15600	0
SSF	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Not Integrated	260	8208	1052	15600	6080
SSF	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	5	210	4120	587	6240	1323
SSF	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	8	211	4111	641	6240	1277
SSF	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	15	218	4121	680	6240	1221
SSF	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	28	229	4125	734	6240	1152
SSF	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	50	229	4168	736	6240	1107
SSF	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Exponential	495	10407	1658	15600	3040
SSF	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Gamma	602	10238	1720	15600	3040
SSF	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Uniform	485	10405	1649	15600	3061
SSF	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Von Mises	612	10240	1729	15600	3019
SSF	Scrambled Habitat Layer (i.e., No selection)	Model Averaging Method	Model average	680	7019	1821	15600	6080
SSF	Scrambled Habitat Layer (i.e., No selection)	Model Averaging Method	Naive average	417	13626	1557	15600	0

Table 3. An overall summary of all Poisson analyses and associated decisions in relation to the frequency of significant results.

Analysis	Selection Scenario	Decision	Value	Significantly Negative	Not Significant	Significantly Positive	Total Estimates	Significance Uncalculated
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	7	0	2400	0	2400	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	15	0	3600	0	3600	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	30	0	4800	0	4800	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	60	0	4799	0	4800	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.08	0	2400	0	2400	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.17	0	3599	0	3600	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.5	0	4800	0	4800	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	1	0	4800	0	4800	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	3	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	5	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	10	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	15	0	3119	0	3120	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	20	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Integrated	0	7799	0	7800	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Not Integrated	0	7800	0	7800	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	5	0	3119	0	3120	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	8	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	15	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	28	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	50	0	3120	0	3120	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Exponential	0	7800	0	7800	0
Poisson	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Gamma	0	7799	0	7800	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Uniform	0	7799	0	7800	1
Poisson	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Von Mises	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	7	0	2400	0	2400	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	15	0	3600	0	3600	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	30	0	4800	0	4800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	60	0	4800	0	4800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.08	0	2400	0	2400	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.17	0	3600	0	3600	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.5	0	4800	0	4800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	1	0	4800	0	4800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	3	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	5	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	10	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	15	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	20	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Integrated	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Not Integrated	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	5	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	8	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	15	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	28	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	50	0	3120	0	3120	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Exponential	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Gamma	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Uniform	0	7800	0	7800	0
Poisson	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Von Mises	0	7800	0	7800	0

**Table 4.** An overall summary of all Two-Step analyses and associated decisions in relation to the frequency of significant results.

Analysis	Selection Scenario	Decision	Value	Significantly Negative	Not Significant	Significantly Positive	Total Estimates	Significance Uncalculated
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	7	1699	2716	2785	7200	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	15	3115	3636	4049	10800	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	30	4687	4252	5461	14400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Duration (days)	60	4859	3489	6052	14400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.08	2692	2421	2087	7200	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.17	3853	3748	3199	10800	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	0.5	4156	4301	5943	14400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Tracking Frequency (points/hour)	1	3659	3623	7118	14400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	3	2274	4567	2519	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	5	2869	3603	2888	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	10	2953	2362	4045	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	15	3069	1883	4408	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Sample Size (n)	20	3195	1678	4487	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Integrated	14360	12421	12219	39000	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Model Formula (SSF or iSSF)	Not Integrated	0	1672	6128	7800	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	5	2802	2927	3631	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	8	2851	2866	3643	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	15	2876	2809	3675	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	28	2910	2753	3697	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Available Points per Step	50	2921	2738	3701	9360	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Exponential	7409	6017	9974	23400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Distribution of Step Lengths	Gamma	6951	8076	8373	23400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Uniform	6080	7618	9702	23400	0
TwoStep	Correct Habitat Layer (i.e., Positive selection)	Distribution of Turn Angles	Von Mises	8280	6475	8645	23400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	7	1928	4663	609	7200	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	15	3565	6016	1219	10800	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	30	5230	7397	1773	14400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Duration (days)	60	5297	6837	2266	14400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.08	2921	3560	719	7200	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.17	4221	5235	1344	10800	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	0.5	4679	7997	1724	14400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Tracking Frequency (points/hour)	1	4199	8121	2080	14400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	3	2535	5642	1183	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	5	3392	5188	780	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	10	3263	5068	1029	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	15	3310	4903	1147	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Sample Size (n)	20	3520	4112	1728	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Integrated	15737	18348	4915	39000	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Model Formula (SSF or iSSF)	Not Integrated	283	6565	952	7800	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	5	3118	5083	1159	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	8	3162	5047	1151	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	15	3211	4958	1191	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	28	3257	4915	1188	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Available Points per Step	50	3272	4910	1178	9360	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Exponential	9593	9498	4309	23400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Distribution of Step Lengths	Gamma	6427	15415	1558	23400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Uniform	6825	13267	3308	23400	0
TwoStep	Scrambled Habitat Layer (i.e., No selection)	Distribution of Turn Angles	Von Mises	9195	11646	2559	23400	0



References

- Aebischer NJ, Robertson PA, Kenward RE. 1993. Compositional Analysis of Habitat Use From Animal Radio-Tracking Data. *Ecology* 74:1313–1325. DOI: [10.2307/1940062](https://doi.org/10.2307/1940062).
- Allaire J, Xie Y, Dervieux C, McPherson J, Luraschi J, Ushey K, Atkins A, Wickham H, Cheng J, Chang W, Iannone R. 2023. *rmarkdown: Dynamic documents for r*.
- Auspurg K, Brüderl J. 2021. Has the Credibility of the Social Sciences Been Credibly Destroyed? Reanalyzing the “Many Analysts, One Data Set” Project. *Socius: Sociological Research for a Dynamic World* 7:237802312110244. DOI: [10.1177/23780231211024421](https://doi.org/10.1177/23780231211024421).
- Bartoń K. 2023. *MuMIn: Multi-model inference*.
- Barto EK, Rillig MC. 2012. Dissemination biases in ecology: Effect sizes matter more than quality. *Oikos* 121:228–235. DOI: [10.1111/j.1600-0706.2011.19401.x](https://doi.org/10.1111/j.1600-0706.2011.19401.x).
- Bastiaansen JA, Kunkels YK, Blaauw FJ, Boker SM, Ceulemans E, Chen M, Chow S-M, Jonge P de, Emerencia AC, Epskamp S, Fisher AJ, Hamaker EL, Kuppens P, Lutz W, Meyer MJ, Moulder R, Oravec Z, Riese H, Rubel J, Ryan O, Servaas MN, Sjobeck G, Snippe E, Trull TJ, Tschacher W, Veen DC van der, Wicher M, Wood PK, Woods WC, Wright AGC, Albers CJ, Bringmann LF. 2020. Time to get personal? The impact of researchers choices on the selection of treatment targets using the experience sampling methodology. *Journal of Psychosomatic Research* 137:110211. DOI: [10.1016/j.jpsychores.2020.110211](https://doi.org/10.1016/j.jpsychores.2020.110211).
- Bivand RS, Pebesma E, Gomez-Rubio V. 2013. *Applied spatial data analysis with R, second edition*. Springer, NY.
- Brembs B. 2018. Prestigious Science Journals Struggle to Reach Even Average Reliability. *Frontiers in Human Neuroscience* 12:1–7. DOI: [10.3389/fnhum.2018.00037](https://doi.org/10.3389/fnhum.2018.00037).
- Bürkner P-C. 2021. Bayesian item response modeling in R with brms and Stan. *Journal of Statistical Software* 100:1–54. DOI: [10.18637/jss.v100.i05](https://doi.org/10.18637/jss.v100.i05).
- Calenge C, Mathieu Basille contributions from. 2023. *adehabitatHS: Analysis of habitat selection by animals*.
- Calenge C, Scott Fortmann-Roe contributions from. 2023. *adehabitatHR: Home range estimation*.
- Ching T. 2023. *qs: Quick serialization of r objects*.
- Craiu RV, Duchesne T, Fortin D, Baillargeon S. 2011. Conditional Logistic Regression With Longitudinal Follow-up and Individual-Level Random Coefficients: A Stable and Efficient Two-Step Estimation Method. *Journal of Computational and Graphical Statistics* 20:767–784. DOI: [10.1198/jcgs.2011.09189](https://doi.org/10.1198/jcgs.2011.09189).
- Craiu RV, Duchesne T, Fortin D, Baillargeon S. 2016. *TwoStepCLogit: Conditional logistic regression: A two-step estimation method*.
- Crane M, Silva I, Marshall BM, Strine CT. 2021. Lots of movement, little progress: A review of reptile home range literature. *PeerJ* 9:e11742. DOI: [10.7717/peerj.11742](https://doi.org/10.7717/peerj.11742).
- Del Giudice M, Gangestad SW. 2021. A Traveler’s Guide to the Multiverse: Promises, Pitfalls, and a Framework for the Evaluation of Analytic Decisions. *Advances in Methods and Practices in Psychological Science* 4:251524592095492. DOI: [10.1177/2515245920954925](https://doi.org/10.1177/2515245920954925).
- Desbureaux S. 2021. Subjective modeling choices and the robustness of impact evaluations in conservation science. *Conservation Biology* 35:1615–1626. DOI: [10.1111/cobi.13728](https://doi.org/10.1111/cobi.13728).
- Devezer B, Navarro DJ, Vandekerckhove J, Ozge Buzbas E. 2021. The case for formal methodology in scientific reform. *Royal Society Open Science* 8:rsos.200805, 200805. DOI: [10.1098/rsos.200805](https://doi.org/10.1098/rsos.200805).
- Doherty TS, Driscoll DA. 2018. Coupling movement and landscape ecology for animal conservation in production landscapes. *Proceedings of the Royal Society B: Biological Sciences* 285:20172272. DOI: [10.1098/rspb.2017.2272](https://doi.org/10.1098/rspb.2017.2272).
- Doherty TS, Fist CN, Driscoll DA. 2019. Animal movement varies with resource availability, landscape configuration and body size: A conceptual model and empirical example. *Landscape Ecology* 34:603–614. DOI: [10.1007/s10980-019-00795-x](https://doi.org/10.1007/s10980-019-00795-x).
- Fanelli D. 2010. Do Pressures to Publish Increase Scientists’ Bias? An Empirical Support from US States Data. *PLoS ONE* 5:e10271. DOI: [10.1371/journal.pone.0010271](https://doi.org/10.1371/journal.pone.0010271).
- Fleming CH, Calabrese JM. 2023. *ctmm: Continuous-time movement modeling*.
- Fleming CH, Fagan WF, Mueller T, Olson KA, Leimgruber P, Calabrese JM. 2015. Rigorous home range estimation with movement data: A new autocorrelated kernel density estimator. *Ecology* 96:1182–1188.
- Forester JD, Im HK, Rathouz PJ. 2009. Accounting for animal movement in estimation of resource selection functions: Sampling and data analysis. *Ecology* 90:3554–3565. DOI: [10.1890/08-0874.1](https://doi.org/10.1890/08-0874.1).
- Forstmeier W, Wagenmakers E-J, Parker TH. 2017. Detecting and avoiding likely false-positive findings – a practical guide: Avoiding false-positive findings. *Biological Reviews* 92:1941–1968. DOI: [10.1111/brv.12315](https://doi.org/10.1111/brv.12315).
- Francisco Rodríguez-Sánchez, Connor P. Jackson, Shaurita D. Hutchins. 2023. *grateful: Facilitate citation of r packages*.
- Fraser H, Barnett A, Parker TH, Fidler F. 2020. The role of replication studies in ecology. *Ecology and Evolution* 10:5197–5207. DOI: [10.1002/ece3.6330](https://doi.org/10.1002/ece3.6330).
- Fraser KC, Davies KT, Davy CM, Ford AT, Flockhart DTT, Martins EG. 2018. Tracking the conservation promise of movement ecology. *Frontiers in Ecology and Evolution* 6:150. DOI: [10.3389/FEVO.2018.00150](https://doi.org/10.3389/FEVO.2018.00150).



- Freedman LP, Cockburn IM, Simcoe TS. 2015. The Economics of Reproducibility in Preclinical Research. *PLOS Biology* 13:e1002165. DOI: [10.1371/journal.pbio.1002165](https://doi.org/10.1371/journal.pbio.1002165).
- Gabry J, Simpson D, Vehtari A, Betancourt M, Gelman A. 2019. Visualization in bayesian workflow. *J. R. Stat. Soc. A* 182:389–402. DOI: [10.1111/rssa.12378](https://doi.org/10.1111/rssa.12378).
- Gelman A, Loken E. 2013. The garden of forking paths: Why multiple comparisons can be a problem, even when there is no "fishing expedition" or "p-hacking" and the research hypothesis was posited ahead of time. :17.
- Gould E, Fraser H, Parker T, Nakagawa S, Griffith S, Vesk P, Fidler F, Abbey-Lee R, Abbott J, Aguirre L, Alcaraz C, Altschul D, Arekar K, Atkins J, Atkinson J, Barrett M, Bell K, Bello S, Berauer B, Bertram M, Billman P, Blake C, Blake S, Bliard L, Bonisoli-Alquati A, Bonnet T, Bordes C, Bose A, Botterill-James T, Boyd M, Boyle S, Bradfer-Lawrence T, Brand J, Brengdahl M, Bulla M, Bussière L, Camerlenghi E, Campbell S, Campos L, Caravaggi A, Cardoso P, Carroll C, Catanach T, Chen X, Chik HYJ, Choy E, Christie A, Chuang A, Chunco A, Clark B, Cox M, Cressman K, Crouch C, D'Amelio P, De Sousa A, Döbert T, Dobler R, Dobson A, Doherty T, Drobniak S, Duffy A, Dunn R, Dunning J, Eberhart-Hertel L, Elmore J, Elsherif M, English H, Ensminger D, Ernst U, Ferguson S, Ferreira-Arruda T, Fieberg J, Finch E, Fiorenza E, Fisher D, Forstmeier W, Fourcade Y, Francesca Santostefano F, Frank G, Freund C, Gandy S, Gannon D, García-Cervigón A, Géron C, Gilles M, Girndt A, Glikzman D, Goldspiel H, Gomes D, Goslee S, Gosnell J, Gratton P, Grebe N, Greenler S, Griffith D, Griffith F, Grossman J, Güncan A, Haesen S, Hagan J, Harrison N, Hasnain S, Havird J, Heaton A, Hsu B-Y, Iranzo E, Iverson E, Jimoh S, Johnson D, Johnsson M, Jorna J, Jucker T, Jung M, Kačergyté I, Ke A, Kelly C, Keegan K, Keppeler F, Killion A, Kim D, Kochan D, Korsten P, Kothari S, Kuppler J, Kusch J, Lagisz M, Larkin D, Larson C, Lauck K, Lauterbur M, Law A, Léandri-Breton D-J, Lievens E, Lima D, Lindsay S, Macphie K, Mair M, Malm L, Mammola S, Manhart M, Mäntylä E, Marchand P, Marshall B, Martin D, Martin J, Martin C, Martinig A, McCallum E, McNew S, Meiners S, Michelangeli M, Moiron M, Moreira B, Mortensen J, Mos B, Muraina T, Nelli L, Nilsonne G, Nolazco S, Nooten S, Novotny J, Olin A, Organ C, Ostevik K, Palacio F, Paquet M, Pascall D, Pasquarella V, Payo-Payo A, Pedersen K, Perez G, Perry K, Pottier P, Proulx M, Proulx R, Pruitt J, Ramananjato V, Randimbiarison F, Razafindratsima O, Rennison D, Riva F, Riyahi S, Roast M, Rocha F, Roche D, Román-Palacios C, Rosenberg M, Ross J, Rowland F, Rugemalila D, Russell A, Ruuskanen S, Saccone P, Sadeh A, Salazar S, Sales K, Salmón P, Sanchez-Tojar A, Santos L, Schilling H, Schmidt M, Schmoll T, Schneider A, Schrock A, Schroeder J, Schtickzelle N, Schultz N, Scott D, Shapiro J, Sharma N, Shearer C, Sitvarin M, Skupien F, Slinn H, Smith J, Smith G, Sollmann R, Stack Whitney K, Still S, Stuber E, Sutton G, Swallow B, Taff C, Takola E, Tanentzap A, Thawley C, Tortorelli C, Trlica A, Turnell B, Urban L, Van De Vondel S, Van Oordt F, Vanderwel M, Vanderwel K, Vanderwolf K, Verrelli B, Vieira M, Vollering J, Walker X, Walter J, Waryszak P, Weaver R, Weller D, Whelan S, White R, Wolfson D, Wood A, Yanco S, Yen J, Youngflesh C, Zilio G, Zimmer C, Zitomer R, Villamil N, Tompkins E. 2023. Same data, different analysts: Variation in effect sizes due to analytical decisions in ecology and evolutionary biology. *EcoEvoRxiv*. DOI: [10.32942/X2GG62](https://doi.org/10.32942/X2GG62).
- Hijmans RJ. 2023. *raster: Geographic data analysis and modeling*.
- Hofmann DD, Cozzi G, Fieberg J. 2024. Methods for implementing integrated step-selection functions with incomplete data. *Movement Ecology* 12:37. DOI: [10.1186/s40462-024-00476-8](https://doi.org/10.1186/s40462-024-00476-8).
- Homberger B, Jenni L, Duplain J, Lanz M, Schaub M. 2021. Strong effects of radio-tags, social group and release date on survival of reintroduced grey partridges. *Animal Conservation* 24:677–688. DOI: [10.1111/acv.12673](https://doi.org/10.1111/acv.12673).
- Huntington-Klein N, Arenas A, Beam E, Bertoni M, Bloem JR, Burli P, Chen N, Grieco P, Ekpe G, Pugatch T, Saavedra M, Stopnitzky Y. 2021. The influence of hidden researcher decisions in applied microeconomics. *Economic Inquiry* 59:944–960. DOI: [10.1111/ecin.12992](https://doi.org/10.1111/ecin.12992).
- Jennions MD, Møller AP. 2002. Relationships fade with time: A meta-analysis of temporal trends in publication in ecology and evolution. *Proceedings of the Royal Society of London. Series B: Biological Sciences* 269:43–48. DOI: [10.1098/rspb.2001.1832](https://doi.org/10.1098/rspb.2001.1832).
- Johnson DH. 1980. The Comparison of Usage and Availability Measurements for Evaluating Resource Preference. *Ecology* 61:65–71. DOI: [10.2307/1937156](https://doi.org/10.2307/1937156).
- Kay M. 2023a. *ggdist: Visualizations of distributions and uncertainty*. DOI: [10.5281/zenodo.3879620](https://doi.org/10.5281/zenodo.3879620).
- Kay M. 2023b. *tidybayes: Tidy data and geoms for Bayesian models*. DOI: [10.5281/zenodo.1308151](https://doi.org/10.5281/zenodo.1308151).
- Kelly CD. 2019. Rate and success of study replication in ecology and evolution. *PeerJ* 7:e7654. DOI: [10.7717/peerj.7654](https://doi.org/10.7717/peerj.7654).
- Klappstein NJ, Michelot T, Fieberg J, Pedersen EJ, Mills Flemming J. 2024. Step selection functions with non-linear and random effects. *Methods in Ecology and Evolution* 15:1332–1346. DOI: [10.1111/2041-210X.14367](https://doi.org/10.1111/2041-210X.14367).
- Kourounis D, Fuchs A, Schenk O. 2018. Towards the next generation of multiperiod optimal power flow solvers. *IEEE Transactions on Power Systems* PP:1–10.
- Kranstauber B, Smolla M, Scharf AK. 2023. *move: Visualizing and analyzing animal track data*.
- Landau WM. 2021b. *Tarchetypes: Archetypes for targets*.
- Landau WM. 2021a. *The targets r package: A dynamic make-like function-oriented pipeline toolkit for reproducibility and high-performance computing*. *Journal of Open Source Software* 6:2959.
- Lindgren F, Rue H, Lindström J. 2011. An explicit link between Gaussian fields and Gaussian Markov random fields: The stochastic partial differential equation approach (with discussion). *Journal of the Royal Statistical Society B*



- 73:423–498.
- Lüdecke D, Ben-Shachar MS, Patil I, Waggoner P, Makowski D. 2021. *performance: An R package for assessment, comparison and testing of statistical models*. *Journal of Open Source Software* 6:3139. DOI: [10.21105/joss.03139](https://doi.org/10.21105/joss.03139).
- Marshall BM, Duthie AB. 2022. *abmAnimalMovement: An r package for simulating animal movement using an agent-based model*. *F1000* 0:0.
- Marshall BM, Duthie AB. 2024. A Habitat Selection Multiverse Reveals Largely Consistent Results Despite a Multitude of Analysis Options. *bioRxiv*. DOI: [10.1101/2024.06.19.599733](https://doi.org/10.1101/2024.06.19.599733).
- Martins TG, Simpson D, Lindgren F, Rue H. 2013. Bayesian computing with INLA: New features. *Computational Statistics and Data Analysis* 67:68–83.
- Michelot T, Blackwell PG, Chamaillé-Jammes S, Matthiopoulos J. 2020. Inference in MCMC step selection models. *Biometrics* 76:438–447. DOI: [10.1111/biom.13170](https://doi.org/10.1111/biom.13170).
- Michelot T, Blackwell PG, Matthiopoulos J. 2019. Linking resource selection and step selection models for habitat preferences in animals. *Ecology* 100:e02452. DOI: [10.1002/ecy.2452](https://doi.org/10.1002/ecy.2452).
- Michelot T, Gloaguen P, Blackwell PG, Étienne M. 2019. The Langevin diffusion as a continuous-time model of animal movement and habitat selection. *Methods in Ecology and Evolution* 10:1894–1907. DOI: [10.1111/2041-210X.13275](https://doi.org/10.1111/2041-210X.13275).
- Mueller T, Olson KA, Dressler G, Leimgruber P, Fuller TK, Nicolson C, Novaro AJ, Bolgeri MJ, Wattles D, DeStefano S, Calabrese JM, Fagan WF. 2011. How landscape dynamics link individual- to population-level movement patterns: A multispecies comparison of ungulate relocation data: Population-level movement patterns. *Global Ecology and Biogeography* 20:683–694. DOI: [10.1111/j.1466-8238.2010.00638.x](https://doi.org/10.1111/j.1466-8238.2010.00638.x).
- Muff S, Signer J, Fieberg J. 2020. Accounting for individual-specific variation in habitat-selection studies: Efficient estimation of mixed-effects models using bayesian or frequentist computation. *Journal of Animal Ecology* 89:80–92. DOI: [10.1111/1365-2656.13087](https://doi.org/10.1111/1365-2656.13087).
- Müller K. 2020. *here: A simpler way to find your files*.
- Müller K, Wickham H. 2023. *tibble: Simple data frames*.
- Nakagawa S, Parker TH. 2015. Replicating research in ecology and evolution: Feasibility, incentives, and the cost-benefit conundrum. *BMC Biology* 13:88. DOI: [10.1186/s12915-015-0196-3](https://doi.org/10.1186/s12915-015-0196-3).
- Open Science Collaboration. 2015. Estimating the reproducibility of psychological science. *Science* 349:aac4716–aac4716. DOI: [10.1126/science.aac4716](https://doi.org/10.1126/science.aac4716).
- Palmer AR. 2000. Quasi-Replication and the Contract of Error: Lessons from Sex Ratios, Heritabilities and Fluctuating Asymmetry. *Annual Review of Ecology and Systematics* 31:441–480. DOI: [10.1146/annurev.ecolsys.31.1.441](https://doi.org/10.1146/annurev.ecolsys.31.1.441).
- Parker TH. 2013. What do we really know about the signalling role of plumage colour in blue tits? A case study of impediments to progress in evolutionary biology: Case study of impediments to progress. *Biological Reviews* 88:511–536. DOI: [10.1111/brv.12013](https://doi.org/10.1111/brv.12013).
- Pedersen TL. 2022. *Patchwork: The composer of plots*.
- Peterson D, Panofsky A. 2021. Self-correction in science: The diagnostic and integrative motives for replication. *Social Studies of Science*.
- Portugal SJ, White CR. 2022. Externally attached biologists cause compensatory body mass loss in birds. *Methods in Ecology and Evolution* 13:294–302. DOI: [10.1111/2041-210X.13754](https://doi.org/10.1111/2041-210X.13754).
- R Core Team. 2023. *R: A language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing.
- Rijnhart JJM, Twisk JWR, Deeg DJH, Heymans MW. 2021. Assessing the Robustness of Mediation Analysis Results Using Multiverse Analysis. *Prevention Science*. DOI: [10.1007/s11121-021-01280-1](https://doi.org/10.1007/s11121-021-01280-1).
- RStudio Team. 2022. *RStudio: Integrated development environment for r*. Boston, MA: RStudio, PBC.
- Rue H, Martino S, Chopin N. 2009. Approximate Bayesian inference for latent Gaussian models using integrated nested Laplace approximations (with discussion). *Journal of the Royal Statistical Society B* 71:319–392.
- Rue H, Riebler AI, Sørbye SH, Illian JB, Simpson DP, Lindgren FK. 2017. Bayesian computing with INLA: A review. *Annual Reviews of Statistics and Its Applications* 4:395–421.
- Salis A, Lena J-P, Lengagne T. 2021. How Subtle Protocol Choices Can Affect Biological Conclusions: Great Tits' Response to Allopatric Mobbing Calls. *Animal Behavior and Cognition* 8:152–165. DOI: [10.26451/abc.08.02.05.2021](https://doi.org/10.26451/abc.08.02.05.2021).
- Sánchez-Tójar A, Nakagawa S, Sánchez-Fortún M, Martin DA, Ramani S, Girndt A, Bókony V, Kempenaers B, Liker A, Westneat DF, Burke T, Schroeder J. 2018. Meta-analysis challenges a textbook example of status signalling and demonstrates publication bias. *eLife* 7:e37385. DOI: [10.7554/eLife.37385](https://doi.org/10.7554/eLife.37385).
- Schlather M, Malinowski A, Menck PJ, Oesting M, Strokorb K. 2015. *Analysis, simulation and prediction of multivariate random fields with package RandomFields*. *Journal of Statistical Software* 63:1–25.
- Sciaini M, Fritsch M, Scherer C, Simpkins CE. 2018. *NLMR and landscapetools: An integrated environment for simulating and modifying neutral landscape models in r*. *Methods in Ecology and Evolution* 00:1–9.
- Seguin A, Forstmeier W. 2012. No Band Color Effects on Male Courtship Rate or Body Mass in the Zebra Finch: Four Experiments and a Meta-Analysis. *PLoS ONE* 7:e37785. DOI: [10.1371/journal.pone.0037785](https://doi.org/10.1371/journal.pone.0037785).
- Signer J, Fieberg J, Avgar T. 2019. *Animal movement tools (amt): R package for managing tracking data and*



- conducting habitat selection analyses. *Ecology and Evolution* 9:880–890.
- Silberzahn R, Uhlmann EL, Martin DP, Anselmi P, Aust F, Awtrey E, Bahník Š, Bai F, Bannard C, Bonnier E, Carlsson R, Cheung F, Christensen G, Clay R, Craig MA, Dalla Rosa A, Dam L, Evans MH, Flores Cervantes I, Fong N, Gamez-Djokic M, Glenz A, Gordon-McKeon S, Heaton TJ, Hederos K, Heene M, Hofelich Mohr AJ, Högden F, Hui K, Johannesson M, Kalodimos J, Kaszubowski E, Kennedy DM, Lei R, Lindsay TA, Liverani S, Madan CR, Molden D, Molleman E, Morey RD, Mulder LB, Nijstad BR, Pope NG, Pope B, Prenoveau JM, Rink F, Robusto E, Roderique H, Sandberg A, Schlüter E, Schönbrodt FD, Sherman MF, Sommer SA, Sotak K, Spain S, Spörlein C, Stafford T, Stefanutti L, Tauber S, Ullrich J, Vianello M, Wagenmakers E-J, Witkowiak M, Yoon S, Nosek BA. 2018. Many Analysts, One Data Set: Making Transparent How Variations in Analytic Choices Affect Results. *Advances in Methods and Practices in Psychological Science* 1:337–356. DOI: [10.1177/2515245917747646](https://doi.org/10.1177/2515245917747646).
- Silva I, Fleming CH, Noonan MJ, Alston J, Folta C, Fagan WF, Calabrese JM. 2022. Autocorrelation-informed home range estimation: A review and practical guide. *Methods in Ecology and Evolution* 13:534–544. DOI: [10.1111/2041-210X.13786](https://doi.org/10.1111/2041-210X.13786).
- Simonsohn U. 2015. Small Telescopes: Detectability and the Evaluation of Replication Results. *Psychological Science* 26:559–569. DOI: [10.1177/0956797614567341](https://doi.org/10.1177/0956797614567341).
- Simonsohn U, Simmons JP, Nelson LD. 2020. Specification curve analysis. *Nature Human Behaviour* 4:1208–1214. DOI: [10.1038/s41562-020-0912-z](https://doi.org/10.1038/s41562-020-0912-z).
- Smaldino PE, McElreath R. 2016. The natural selection of bad science. *Royal Society Open Science* 3:160384. DOI: [10.1098/rsos.160384](https://doi.org/10.1098/rsos.160384).
- Sperry JH, Butler LK, Romero LM, Weatherhead PJ. 2009. Effects of parasitic infection and radio-transmitters on condition, hematological characteristics and corticosterone concentrations in Texas ratsnakes. *Journal of Zoology* 278:100–107. DOI: [10.1111/j.1469-7998.2009.00549.x](https://doi.org/10.1111/j.1469-7998.2009.00549.x).
- Steegen S, Tuerlinckx F, Gelman A, Vanpaemel W. 2016. Increasing Transparency Through a Multiverse Analysis. *Perspectives on Psychological Science* 11:702–712. DOI: [10.1177/1745691616658637](https://doi.org/10.1177/1745691616658637).
- Street GM, Potts JR, Börger L, Beasley JC, Demarais S, Fryxell JM, McLoughlin PD, Monteith KL, Prokopenko CM, Ribeiro MC, Rodgers AR, Strickland BK, Beest FM, Bernasconi DA, Beumer LT, Dharmarajan G, Dwinnell SP, Keiter DA, Keuroghlian A, Newediuk LJ, Oshima JEF, Rhodes O, Schlichting PE, Schmidt NM, Vander Wal E. 2021. Solving the sample size problem for resource selection functions. *Methods in Ecology and Evolution* 12:2421–2431. DOI: [10.1111/2041-210X.13701](https://doi.org/10.1111/2041-210X.13701).
- Stuber EF, Carlson BS, Jesmer BR. 2022. Spatial personalities: A meta-analysis of consistent individual differences in spatial behavior. *Behavioral Ecology* 33:477–486. DOI: [10.1093/beheco/arab147](https://doi.org/10.1093/beheco/arab147).
- Thurfjell H, Ciuti S, Boyce MS. 2014. Applications of step-selection functions in ecology and conservation. *Movement Ecology* 2. DOI: [10.1186/2051-3933-2-4](https://doi.org/10.1186/2051-3933-2-4).
- Ware JJ, Munafò MR. 2015. Significance chasing in research practice: Causes, consequences and possible solutions: Significance chasing. *Addiction* 110:4–8. DOI: [10.1111/add.12673](https://doi.org/10.1111/add.12673).
- Wickham H. 2016. *ggplot2: Elegant graphics for data analysis*. Springer-Verlag New York.
- Wickham H. 2022. *stringr: Simple, consistent wrappers for common string operations*.
- Wickham H, François R, Henry L, Müller K, Vaughan D. 2023. *dplyr: A grammar of data manipulation*.
- Wilke CO. 2022. *Ggridges: Ridgeline plots in 'ggplot2'*.
- Wilke CO, Wiernik BM. 2022. *Ggtext: Improved text rendering support for 'ggplot2'*.
- Xie Y. 2014. knitr: A comprehensive tool for reproducible research in R. In: Stodden V, Leisch F, Peng RD eds. *Implementing reproducible computational research*. Chapman; Hall/CRC,.
- Xie Y. 2015. *Dynamic documents with R and knitr*. Boca Raton, Florida: Chapman; Hall/CRC.
- Xie Y. 2016. *bookdown: Authoring books and technical documents with R markdown*. Boca Raton, Florida: Chapman; Hall/CRC.
- Xie Y. 2019. *TinyTeX: A lightweight, cross-platform, and easy-to-maintain LaTeX distribution based on TeX live*. *TUGboat* 40:30–32.
- Xie Y. 2022. *Bookdown: Authoring books and technical documents with r markdown*.
- Xie Y. 2023b. *knitr: A general-purpose package for dynamic report generation in r*.
- Xie Y. 2023a. *tinytex: Helper functions to install and maintain TeX live, and compile LaTeX documents*.
- Xie Y, Allaire JJ, Grolemund G. 2018. *R markdown: The definitive guide*. Boca Raton, Florida: Chapman; Hall/CRC.
- Xie Y, Dervieux C, Riederer E. 2020. *R markdown cookbook*. Boca Raton, Florida: Chapman; Hall/CRC.