MixFishSim: highly resolved spatiotemporal simulations for exploring mixed fishery dynamics

Paul J. Dolder^{a,b,*}, Cóilín Minto^a, Jean-Marc Guarini^c, Jan Jaap Poos^d

^a Galway-Mayo Institute of Technology (GMIT), Dublin Road, Galway, Ireland
^b Centre for Environment, Fisheries and Aquaculture Science (Cefas), Pakefield Road, Lowestoft, UK

^cSorbonne Université, Faculty of Sciences Université Pierre et Marie Curie JM, 4 Place Jussieu, 75005 Paris, France

Abstract

Most fisheries Fishing JJ exploits JJ spatially and temporally hetergenous fish populations, using species-unselective gear that can result in unintended, unwanted catch of low quota or protected species. Reducing these unwanted catches is crucial for biological and economic sustainability of 'mixed fisheries' and implementation of an ecosystem approach to fishing.

To implement effective spatial measures to reduce discards a good understanding of spatiotemporal fishery dynamics is required. However, traditional scientific advice is limited by a lack of highly resolved knowledge of population distribution, movement and how fishers interact with different fish populations. This reflects that data on fish location at high temporal and spatial resolutions is expensive and difficult to collect and therefore proxies inferred from either scientific surveys or commercial catch data are often used to model distributions, often with limited spatial and temporal resolution.

To understand how resolution impacts mixed fisheries inference, we develop a highly resolved spatiotemporal simulation model incorporating: i) delay-

Email address: paul.dolder@gmit.ie (Paul J. Dolder)

^d Wageningen Marine Research, Haringkade 1 1976 CP IJmuiden, Netherlands

^{*}Corresponding author

difference population dynamics, ii) population movement using Gaussian Ran-

dom Fields to simulate patchy, hetergenously distributed populations, and iii)

fishery dynamics for multiple fleet characteristics based on targetting via corre-

lated random walk movement and learned behaviour.

We simulate 20 years of exploitation of the fish populations and use the

results from the fishing model to draw inference on the underlying population

structures. We compare this inference to i) a simulated fixed-site sampling

design commonly used for fisheries monitoring purposes, and ii) the true un-

derlying population structures input to the simulation, to establish the poten-

tial and limitations of fishery-dependent data - an inherently biased sampling method due to fisher's targeting- to provide a robust picture of spatiotemporal

distributions. Finally, we simulate an area closure based on areas defined from

commercial the known ("real-population") distribution, commercial catch data

and survey data at different temporal and spatial resolutions and assess their

effectiveness on reducing catches of a fish population.

We conclude from our simulations that commercial data, while not unbiased,

provides a useful tool for managing catches in mixed fisheries if applied at the

correct spatiotemporal scale.

[333 words]

Keywords: Some, keywords, here. Max 6

2010 MSC: 00-01, 99-00

1. Introduction

Fishers exploit fish populations that are heterogenously distributed in space

and time with verying knowledge of species distributions using species-unselective

fishing gear. Fisheries that catch an assemblage of species, known as mixed fish-

eries may, when managed by single-species quotas can end up^{JJ} discarding^{JJ}

2

overquota catch when managed by single species quotas, ^{JJ} leading to overexploitation of fish populations (Ulrich et al., 2011; Batsleer et al., 2015) ^{JJ}. This
discarding of fish in excess of quota hampers the ability to limit fishing mortality to within sustainable limits (Alverson et al., 1994; Crowder and Murawski,
1998; Rijnsdorp et al., 2007). ^{JJ} Reducing discarding is crucial to ensure biological and economic sustainability of fisheries—and implementation of an ecosystem
approach to fisheries ^{JJ} and. As such PD there is increasing interest in technical
solutions such as gear and spatial closures as ways of avoiding discarding of
fish (Kennelly and Broadhurst, 2002; Catchpole and Revill, 2008; Bellido et al.,
2011) s^{JJ}.

16

Use of spatial management as a tool has been proposed as a method to reduce 17 discards (Holmes et al., 2011; Little et al., 2014; Dunn et al., 2014) PD. However, its implementation is hampered by lack of knowledge of fish and fishery 19 spatiotemporal dynamics and understanding of the scale at which processes are 20 important for management. Understanding the correct scale for spatial man-21 agement is crucial in order to implement measures at a resolution that ensures effective management (Dunn et al., 2016) while minimising economic impact. For example, a scale that promotes species avoidance for vulnerable or low 24 quota species while allowing continuance of sustainable fisheries for available quota species. 26

27

Ensuring measures are implemented at an appropriate scale has been a challenge in the past that has led to ineffectual measures with unitended consequences such as limited impact towards the management objective or increased benthic impact on previously unexploited areas (e.g. the cod closure in the North Sea (Rijnsdorp et al., 2001; Dinmore et al., 2003)). Since then more refined spatial information has become available through the combination of logbook and Vessel Monitoring System (VMS) data (Lee et al., 2010; Bastardie et al., 2010; Gerritsen et al., 2012; Mateo et al., 2016) and more real-time spatial management has been possible (e.g. Holmes et al., 2011). Such information is,

however, patchy and derived from an inherently biased sampling programme (i.e. targeted fishing). Further, fishers generally only recorded landings (not catch) on a daily basis. This leads to questions about the validity of inference that can be drawn from landings data assigned to VMS activity pings.

In order to understand challenges that face VMS-linked landings to draw inference on the underlying population structure we develop a simulation model where population dynamics are highly-resolved in space and time and are known rather than inferred from sampling or commercial catches. Population movement is driven by a random (diffusive) and directed (advective) process and we incorporate characterisation of a number of different fisheries exploiting four fish populations with different spatial and population demographics.

what the problem is: landings collected on daily basis or landings recorded rather than catch ^{J,J}

Its not clear

This comes as a surprise: I thought this was going to be about discards^{JJ}

Using our model we simulate 20 years of exploitation of the fish populations and use the results from the fishing model to draw inference on the underlying population structures. We compare this inference to: i) a stratified fixed-site sampling survey design commonly used for fisheries monitoring purposes, otherwise know as a fisheries-independent survey, and ii) the underlying population structures input to the simulation.

56 57

55

62 63

41

42

47

50

51

52

We simulate a fishery closure to protect one species based on the fishery-dependent inferred distributions at a spatial and temporal scale typical in fisheries management, and assess a theoretical "benefit" to the population, and effect on the other three populations. Further, we extend our analysis to a range of spatial and temporal scales to assess the impact of these processes on the success of the management measure.

If the paper has two goals this should be clear from the start, but may be better over MSs^{JJ}

2. Materials and Methods

AWe developed and implemented a simulation model with a^{PD} modular

```
event-based simulation model was developed with approach, where sub-PD modules
   are PD implemented on independent time-scales appropriate to capture the char-
   acteristic of the different processes process modelled PD (Figure 1). The following
   sub-modules were included to capture the full system: 1) Population dynamics,
   2) Recruitment dynamics, 3) Population movement, 4) fishery dynamics. PD
71
      The fishing model operated on a tow-by-tow basis, while PD PpPD opulation
   dynamics (fishing and natural mortality, growth) operate on a daily time-step,
   while p. PPD opulation movement occurs on a weekly time-step. R, while
74
   r<sup>PD</sup> ecruitment takes place<del>occurs</del><sup>PD</sup> periodcally each year for a set time duration<del>period</del>
   (e.g. 3 weeks)<sup>PD</sup> at at specified point individual to a species.<sup>PD</sup>, while the fish-
   ing module operates on a tow-by-tow basis (multiple events a day)<sup>PD</sup>. The
   simulation framework is implemented in the statistical software package R (R
   Core Team, 2017) and available as an R package from the authors github
   (www.github.com/pdolder/MixFishSim).
81
      Here we describe each of the model components; 1) Population dynamics, 2)
   Recruitment dynamics, 3) Population movement dynamics, 4) fishery dynamics. PD
   2.1. Population dynamics
      The basic population level processes are simulated using a modified two-
   stage Deriso-Schnute delay difference model (Deriso, 1980; Schnute, 1985; Dich-
   mont et al., 2003) occurring at a daily time-step. A daily time-step was cho-
87
   sen as to discretise continuous population processes on a biologically relevent
   and computationally tractable timescale. PD Under the population dynamics
89
   moduleHere, PD population biomass growth and depletion for pre-recruits and
   fish<sup>PD</sup> recruited fish<sup>PD</sup> to the fishery<sup>PD</sup> are modelled separately as a function
   of previous recruited biomass, intrinsic population growth and recruitment:
```

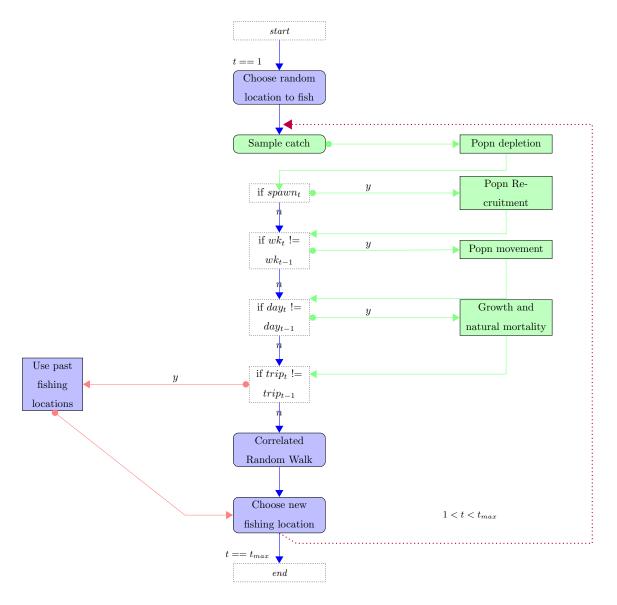


Figure 1: Overview Schematic of simulation model. The blue boxes indicate fleet dynamics processes, the green boxes population dynamics processes while the white boxes are the timesteps at which processes occur; $t=\mathrm{tow},\,tmax$ is the total number of tows, wk is a weekly timestep, day is a day timestep, trip is a trip time step. I NEED TO REDO THIS TO MAKE NOTATION MORE CONCISE AND CONSISTENT

$$\begin{split} B_{c,d+1} &= \\ & (1+\rho)B_{c,d} \cdot e^{-Z_{c,d}} - \rho \cdot e^{-Z_{c,d}} & \times \\ & (B_{c,d-1} \cdot e^{-Z_{c,d-1}} + Wt_{R-1} \cdot \alpha_{d-1} \cdot R_{\tilde{y}(c,y,d-1)}) & + \\ & Wt_R \cdot \alpha_d \cdot R_{\tilde{y}(c,y,d)} \end{split}$$

where ρ is Brody's coefficient, shown to be approximately equal to exp(-K),
where K is the growth rate from a von bertalanffy logistic growth model (Schnute,
1985). Wt_{R-1} is the weight of fish prior to recruitment, while Wt_R is the recruited weight. α_d represents the proportion of fish recruited during that day
for the year, while $R_{c,y}$ is the annual recruits in cell c for year y.

Mortality $Z_{c,d}$ can be decomposed to natural mortality, $M_{c,d}$, and fishing mortality, $F_{c,d}$, where both $M_{c,d}$ and $F_{c,d}$ are instantaneous rates with $M_{c,d}$ fixed and $F_{c,d}$ calculated by solving the Baranov catch equation (Hilborn and Walters, 1992) for $F_{c,d}$:

$$C_{c,d} = \frac{F_{c,d}}{F_{c,d} + M_{c,d}} * (1 - e^{-(F_{c,d} + M_{c,d})}) * B_{c,d}$$

where $C_{c,d}$ is the summed catch from the fishing model across all fleets and vessels in cell c for the population during the day d, and $B_{c,d}$ the daily biomass for the population in the cell.

2.2. Recruitment dynamics

106

108

109

Recruitment is modelled through a function relating the mature biomass to recruits at time of recruitment. In *mixfishsim*, it can be modelled either either as a stochastic Beverton-Holt stock-recruit form (Beverton and Holt, 1957):

$$\bar{R}_{c,d} = \frac{(\alpha * B_{c,d})}{(\beta + B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

and catchability - as I think we have F as an emergent property of the fleets rather than something we solve for (I could be wrong though!) - catch for a vessel is a product of catchability and biomass, i.e. C = qB, but this catch is summed to solve for F. So its both

Where α is the maximum recruitment rate, β the spawning stock biomass (SSB) required to produce half the maximum, B current SSB and σ^2 the variability in the recruitment due to stochastic processes.

or a stochastic Ricker form (Ricker, 1954):

114

118

127

134

135

$$\bar{R}_{c,d} = B_{c,d} * e^{(\alpha - \beta * B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

where α is the maximum productivity per spawner and β the density dependent reduction in productivity as the SSB increases.

2.3. Population movement dynamics

To simulate how^{JJ} fish populations might be^{JJ} distributioned^{JJ} in space and time, we employed^{JJ} a Gaussian spatial process was employed^{JJ} to model habitat suitability for each of the populations. An, with an JJ advection-diffusion process to JJ controlled JJ how the JJ populations movementd JJ over time with a moving temperature covariate to capture temporal dependencies. This was intended to balance realism in population movement, capturing the main directed and random processes, and practicality of modelling the population rather than individual fish. JJ

For the PD habitat we defined PD a Gaussian random field process, $\{S(c): c \in \mathbb{R}^2\}$, that is a stochastic process PD where for PD any set of cells c_1, \ldots, c_n where for each $c_i \in \mathbb{R}^{2\text{PD}}$, the joint distribution of $S = \{S(c1), \ldots S(c_n)\}$ is multivariate Gaussian. The distribution is specified by its mean function, $\mu(c) = E[S(c)]$ and its covariance function, $\gamma(c,c') = Cov\{S(c),S(c')\}$ (Diggle and Ribeiro, 2007).

The covariance structure affects the smoothness of the surfaces which the process generates; and PD we used the $Mat\acute{e}rn$ family of PD covariance structure.

What have a temperature covariate? Could just use time JJ Was intended as some biological meaning - species thermal tolerances load onto the temperature effect PD

Not clear how habitat/GRF affect local abundances, only have $B_{y,a}^{\text{JJ}}$ Have included cell reference, c to make spatial link explicit PD

ntroduce the
amma funcion, and why
his covariance
tructure? Why
orrelate vales in the ran-

tures^{PD}, asone where^{PD} the correlation strength weakens the further the distance apart (i.e. the correlation between S(x) and S(x') decreases as the distance u = ||x - x'|| increases)^{PD}. The Matérn covariance structure models the spatial autocorrelation observed with animal distributions (Tobler, 1970; F. Dormann et al., 2007)^{PD} and The Matérn correlation^{PD} is a two-parameter family where:

$$\rho(u) = \{2^{\kappa - 1} \Gamma \kappa\}^{-1} (u/\phi)^{\kappa} K_{\kappa}(u/\phi)$$

 $K_{\kappa}(.)$ is a modified Bessel function of order κ , $\phi > 0$ is a scale parameter with the dimensions of distance, and $\kappa > 0$, called the order, is a shape parameter which determines the smoothness of the underlying process.

TIn the simulation model, t^{PD} he habitat for each of the populations wasis^{PD} generated withthrough^{PD} the *RFSimulate* function of the *RandomFields* R package (Schlater et al., 2015), implementing different parameter settings to affect the patchiness of the populations. Each population wasis^{PD} initialised at a single location, and subsequently moveeds^{PD} according to a probabilistic distribu^{PD} tion based on habitat suitability, temperature and distance from current cell: PD

$$Pr(J|I) = \frac{e^{-\lambda * d_{IJ}} \cdot (Hab_{J,p}^2 \cdot Tol_{J,p,wk})}{\sum\limits_{c=1}^{C} e^{-\lambda * d} \cdot (Hab_{c,p}^2 \cdot Tol_{c,p,wk})}$$
(1)

Where d_{IJ} is the euclidean distance between cell I and cell J, λ is a given rate of decay, $Hab_{J,p}^2$ is the squared index of habitat suitability for cell J and population p, with $Tol_{J,p,wk}$ the temperature tolerance for cell J by population p in week wk.

During specified weeks of the year, the habitat quality wasis^{PD} modified for user-defined^{PD} spawning habitats^{PD}, resulting inmeaning^{PD} each population hadhas^{PD} a concentrated area where spawning takes place and the population

moveds^{PD} towards these cellsthis^{PD} in the weeks prior to spawning.

164

167

168

The temperature field wasis^{PD} simulated to be on a gradient from a South-Westerly to North-Easterly direction, with temperature in each cell changing gradually on a week-by-week basis so that initially high temperature areas cycled^{PD} to lower temperatures and low temperature areas vice versa. Each population p wai^{PD}s assigned a thermal tolerance with mean, μ_p^{PD} and variance, σ_p^{PD} so that each cell and population temperature suitability is defined that:

What does it mean concisely? Areas are assigned? JN Yes, the areas are predefined - I have amended to reflect and tried to clarify PD

$$Tol_{c,p,wk}PD = \frac{1}{\sqrt{(2\pi \cdot \sigma_p^2)}} \cdot \exp(-\frac{(T_{c,wk}PD - \mu_p)^2}{2 \cdot \sigma_p^2})$$
 (2)

Where $Tol_{c,p,wk}^{PD}$ is the tolerance of population p in cell c in week wk^{PD} , $T_{c,wk}^{PD}$ is the temperature in the cell given the week pD and pD and pD and pD the mean and standard deviation of the population temperature tolerance.

173 174

177

178

179

180

181

182

183

184

185

186

187

188

171

The final process resulted in independent populations structure and movement patterns, with population movement occurring on a weekly basis. This process approximated the demographic shifts in fish populations throughout a year while maintaining seasonal patterns for spawning. PD

2.4. Fleet dynamics

The fleet dynamics can be broadly categorised into three components; fleet targeting - which determineds^{PD} the fleet catch efficiency and preference towards a particular species; trip-level decisions, which determined^{PD} the initial location to be fished at the beginning of a trip; and within-trip decisions, determining movement from one fishing spot to another within a trip.

2.4.1. Fleet targeting

Each fleet of n vessels wasis^{PD} characterised by both a general efficiency, Q, and a population specific efficiency, Q_p . Thus, the product of these parameters affectsed^{PD} the overall catch rates for the fleet and the preferential targeting of one population over another. This, in combination with the parameter choice

for the step-function defined below^{PD} (as well as some randomness from the exploratory fishing process) determineds^{PD} the preference of fishing locations for the fleet. All species prices wereare^{PD} kept the same, across fleetsand seasons, though can be made to vary seasonally^{PD}.

2.4.2. Trip-level decisions

Several studies (e.g. Hutton et al., 2004; Tidd et al., 2012; Girardin et al., 195 2015) have confirmed past activity and past catch rates are strong predictors 196 of fishing location choice. For this reason, the fleet dynamics sub-model includeds^{PD} a learning component, where a vessel's initial fishing location in a 198 trip wai^{PD}s based on selecting from previously successful fishing locations. This 199 wai^{PD}s achieved by sorting all previous fishing events in the previous trip as well 200 as the previous time periods in past years, and choosing randomly from the top 75 % of fishing events as defined by the revenue gained in value PD. Simulation 202 testing indicated that this learning increased the mean value of catches for the 203 vessels, over just relying on the correlated random walk function as described 204 for the 'within trip' decisions below^{PD}.

Correlated ran dom walk of what^{JJ}

2.4.3. Within-trip decisions

206

217

218

Fishing locations within a trip are determined by a modified random walk 207 process. A random walk type was chosen as it is the simplest assumption com-208 monly used in ecology to describe optimal^{PD} animal movement which^{PD} search 209 strategying^{PD} for exploiting^{PD} homogeneously distributed prey about which 210 there is uncertain knowledge (Viswanathan et al., 1999). In a random walk, 21 movement is a stochastic process through a series of steps. These steps have a 212 length, and a direction JJ that can either be equal in length or take some other 213 functional form. The direction of the random walk can be correlated, a charac-214 teristic known as 'persistence', providing some overall location of PD directional movement (Codling et al., 2008) or uncorrelated PD. 216

A Lévy flightlévy walk J is a particular form of random walk characterised by

a heavy-tailed distribution of step-length . The Lévy flight and JJ has received a lot of attention in ecological theory in recent years as having shown to have very 220 similar characteristics as those observed by animals in nature, and being a near optimum searching strategy for predators pursuing patchily distributed prey 222 (Viswanathan et al., 1999; Bartumeus et al., 2005; Sims et al., 2008). Bertrand 223 et al. (2007) showed that Peruvian anchovy fishermen have a stochastic search 224 pattern similar to that observed with a lévy walk. However, it remains a subject 225 of debate (e.g. see Edwards, 2011; Reynolds, 2015)^{PD}, with the contention that search patterns may be more simply characteristed as random walks (Sakiyama 227 and Gunji, 2013) with specific patterns related to the characteristics of the prey 228 field (Sims et al., 2012). 229

230

We use a modified random walk where directional change is based on a correlated circular distribution where a favourable fishing ground is likely to be "fished back over" by the vessel returning in the direction it came from and step length (i.e. the distance travelled from the current to the next fishing location) is determined by relating $^{\rm JJ}$ recent fishing success, measured as the summed value of fish caught (revenue, Rev),

$$Rev = \sum_{p=1}^{P} \underline{LC}^{PD}_{p} \cdot Pr_{p}$$

where $L\underline{C}^{\mathrm{PD}}_{p}$ is landingseatch of a population p, and Pr_{p} price of a population, to step distance. Here, when fishing is successful vessels remain in a similar location and continue to exploit the local fishing grounds. When unsuccessful, they move some distance away from the current fishing location. The movement distance retains some degree of stochasticity, which can be controlled separately.

The step function takes the form:

$$StepL = e^{log(\beta_1) + log(\beta_2) - (log(\frac{\beta_1}{\beta_3}))} * Rev$$

237

236

So that, a step from (x1,y1) to (x2, y2) is defined by:

So step length ncreases with increasngly gross revenue? JJ No, the opposite PD

$$(x2, y2) = x1 + StepL \cdot \cos(\frac{\pi \cdot Br}{180}),$$

$$y1 + StepL \cdot \sin(\frac{\pi \cdot Br}{180})$$

$$with \quad Br_{t-1} < 180, Br_t = 180 + \sim vm[(0, 360), k]$$

$$Br_{t-1} > 180, Br_t = 180 - \sim vm[(0, 360), k]$$

with k the concentration parameter from the von Mm^{JJ} is distribution which we correlate with the revenue so that $k = (Rev + 1/RefRev) * max_k$, where max_k is the maximum concentration value, k, and RefRev is parameterised as for β_3 in the step length function.

2.4.4. Local population depletion

Where several fishing vessels are exploiting the same fish population competition is known to play an important role in local distribution of fishing effort 245 (Gillis and Peterman, 1998). If several vessels are fishing on the same patch of 246 fish, local depletion and interference competition JJ will affect fishing location 247 choice of the fleet as a whole (Rijnsdorp, 2000; Poos and Rijnsdorp, 2007). In 248 order to account for this behaviour, the fishing sub-model operates spatially on a daily time-step so that for future days the biomass available to the fishery 250 is reduced in the areas fished. The cumulative effect is to make heavily fished 251 areas less attractive as future fishing opportunities. 252

2.5. Fisheries independent survey

A fisheries-independent survey is simulated where fishing on a regular grid
begins each year at the same time for a given number of stations (a fixed station
survey design). Catches of the populations at each station present JJ are recorded
but not removed from the population. This provides a fishery independent
snapshot of the populations at a regular spatial intervals distribution JJ each
year, similar to scientific surveys undertaken by fisheries research agencies.

3. Calculation

261

270

271

274

275

28

283

284

285

287

288

3.1. Population parameterisation

We parameterised the simulation model for four populations with differing 262 habitat preference, and temperature tolerances (Figures S1, S3, S4, S5, S6, S7), 263 population demographic, and recruitment functions. In addition, each of the 264 populations has two defined spawning areas which result in the populations 265 moving towards these areas in given weeks (Figure S2) and population-specific movement rates (Table 4). The realised movement of the populations for a num-267 ber of weeks is shown in Figure S9 while the realised daily fishing mortality are 268 shown in Figure S10. 269

3.2. Fleet parameterisation

The fleets were parameterised to reflect five different characteristics based 272 on targeting preference and exploitation dynamics (Table 5). Setting a tar-273 geting parameter (Q) that differed across fleets ensured different spatial dynamics, due to preferential targeting of populations that differ in their spatial distributions This ensures that different fleets have different spatial dynamics, 276 preferentially targeted different fish populations^{PD}. The stochasticity in the 277 random walk process ensures that different vessels within a fleet have slightly 278 different spatial distributions based on individual experience, while the step 279 function was parameterised dynamically so that vessels take smaller steps where the fishing location yields in a top quartile of the value available in that year (as defined per fleet in Table 5). 282

Each fleet was set so that, after the first year, fishing locations were chosen based on experience built up in the same month from previous years and from past trip fishing success. 'Success' in this context was defined as the locations where the top 75 % of revenue from was found in previous trips.

An example of the realised fleet movements for a single vessel during a single trip are given in Figure S11, while Figure S12 shows multiple trips for a single vessel, Figure^{PD} S13 the vessel movements for some trips overlaid on the value field, Figure^{PD} S14 shows fishing locations for an entire fleet of 20 vessels for a single trip, and Figurewhile^{PD} S15 shows an example of the step function realisation and turning angles from the correlated random walk.

294 3.3. Survey settings

The survey simulation was set up with follow a fixed gridded station design with 100 stations fished each year, starting on day 92 and ending on day 112 (5 stations per day)^{PD} with same catchability parameters for all populations $(Q_p = 1)$.

299 3.4. Simulation settings

300

301

302

303

305

307

309

310

311

312

313

314

315

316

317

To illustrate the capabilities on *MixFishSim*, we investigate the influence of the temporal and spatial resolution of different data sources on the reduction in catches of a population given spatial closures. To do so, we first set up with simulation to run for 10 years based on a 100 X 100 square grid, with five fleets of 20 vessels each and four fish populations. Fishing takes place four times a day per vessel and five days a week, while population movement is every week.

We allow the simulation to run unrestricted for 5 years, and subsequently close areas for the last 5 years of the simulation based on data (either derived from the commercial catches, fisheries-independent survey or the 'real population' - the underlying populations assumed to be known perfectly) used at different spatial and temporal scales.

The following steps are undertaken to determine closures:

- 1. Extract data source
- 2. Aggregate according to resolution
- 3. Interpolate across entire area at desired resoltion
 - 4. Close top 5 % of areas

Move some of the supplementary figures to the manuscript JJ

move to start
of methods
section JI think
ecological modelling wants
the 'calculations' section here...will
check PD

Is there equilibrium after
5 years or still some trend in
B^{JJ}Not at equilibrium yet...I
need to rerun
until steady
state, looks
20 years. Will
update^{PD}

Procedure unclear. Refer to symbols in methods section or switch order starting with description of data type etc.. J^JYes, will redo PD

In total 56 closure scenarios were run which represent combinations of

- data types: commercial logbook data, survey data and 'real population',
- temporal resolutions: weekly, monthly and yearly closures,
- spatial resolutions: 1 x 1 grid, 5 x 5 grid, 10 x 10 grid and 20 x 20 grid.

Survey closures were on an annual basis only, as this was the most temporally resolved survey data available.

324 4. Results

The species distribution themselves

The consequences of different spatial aggregations of the data are shown in Figure 2, which represents the aggregation of catch from each of the data sources over a year at different spatial resolutions.

329

325

319

The finer spatial grid for the treal population (top left) and commercial 330 data (top middle) show similar patterns, though there are unsampled gaps in 331 the commercial data from a lack of fishing activity (particularly in the lower left 332 part of the sampling domain). The survey data at this spatial resolution shows very sparse and uninformative information about the spatial distributions of the 334 populations. The slightly aggregated data on a 5 x 5 grid shows similar patterns, 335 and while losing some of the spatial detail there remains good consistency be-336 tween the 'real population' and the commercial data. Survey data starts to pick 337 out some of the similar patterns as the other data sources, but lacks coverage. The spatial catch information on a 10 x 10 and 20 x 20 grid loses a signifi-339 cant amount of information about the spatial resolutions for all data sources, 340 and some differences between the commercial and 'real population' data emerge. 341

342

Figure 3 shows the consequences of different temporal aggregations of the data, with 156 weekly (top), 36 monthly (middle) and 3 yearly (bottom) catch

5 compositions across a 20 x 20 area.

As can be seen from the 'real population', the monthly aggregation captures the major patterns seen in the weekly data, albeit missing more subtle differences. The yearly data results in a constant catch pattern due to the aggregation process (sometimes known as an aggregation bias). The commercial data on a weekly basis shows some of the same patterns as the 'real population', though the first species (in red) is less well represented and some weeks are missing catches from the area. The monthly data. The monthly data shows some consistency between the 'real population' and commercial data for species 2 - 4, though species 1 remains underrepresented. On an annual basis, interestingly the commercial data underrepresents the first species (in red) while the survey overrepresents species 1. This is likely due to the biases in commercial sampling, with the fisheries not targeting the areas where species 1 are present, and the biases in the survey sampling from overrepresentation of the spatial distribution.

We implemented a spatial closure using the different data sources and spatial and temporal aggregations as outlined in the protocol in Section 3.4. We used this to assess the efficacy of a closure in reducing fishing mortality on species 1, given availability of data and its use at different resolutions in order to evaluate the trade-offs in data sources. Figure 4 shows the trend in fishing mortality for each species simulated (columns) given the data sources (rows), temporal aggregations (colour lines) and spatial aggregations (linestyles), while Figure 5 shows the change in fishing mortality from before the closure (average F years 2 - 4) to after the closure (average F years 8 - 10).

For the closures based on 'real population' (bottom row), the most disaggregated data (a weekly timescale and 1 x 1 resolution) was most effective, reducing fishing mortality on species 1 (left) by ~ 60 %. Next was the monthly closures ($< \sim 30$ %). The least effective were the yearly closures (blue lines) at all spatial resolutions, which resulted in increased fishing mortalities (> 30

% - N.B. Note though, this is consistent with the increasing trends in F, which is probably more related to the fact that Fs hadn't stabilised in the simulation from the fishing vessels "learning" the best locations - I will rerun the sims for a longer time (20 - 30 years). 379

For the survey data, which can only be implemented on a yearly timescale, the closures had no effect at any data resolution. The results are identical for the different data resolutions except 20 x 20, which is why you can't see more than 2 points. This is because of the sparsity of the sampling locations.

381

382

384 385

396

399

401

402

For the commercial data, the most effective closure scenario was based on 1 386 x 1 data at a monthly temporal resolution. This results in ~ 10 % reduction in F for species 1. This was the only closure scenario to have positive effect according to Figure 5, though looking at the trend in Figure 4 this looks more 389 related to the continued increased in F trend, as other scenarios had an initial 390 effect. Interestingly the monthly data scenario was more effective than weekly 391 data, which I'd posit is due to the increase amount of data available from the 392 commercial sampling across a month compared to a week.i Commercial data 393 used at an annual timestep was ineffective in bringing fishing mortality down 394 for species 1. 395

Given the scenarios above, it seems clear that spatial disaggregation is more important than the temporal disaggregation of the commercial data, except when its used at an annual timeframe, which is the scenario that gave the worst 400

For the other species in the simulation (population 2 - 4) there was little difference in fishing mortalities across scenarios.

Note: The monthly commercial data scenario is the most effective of the 403 realistic scenarios, as the 'real population' can only be seen as a baseline com-404 parison. 405

- 5. Discussion
- 407 6. Conclusions
- 408 Appendices
- 409 Abbreviations
- Detail any unusual ones used.
- 411 Acknowledgements
- those providing help during the research..
- $_{413}$ Funding
- This work was supported by the MARES doctoral training program; and the
- Centre for Environment, Fisheries and Aquaculture Science seedcorn program.

Table 1.	Description	of wariables	for population	dynamics of	h modulo
rable i:	Describtion	or variables	тог роршалоп	ovnamics si	то-тоате

Variable	Meaning	Units		
Population dynamics				
Delay-difference model				
$B_{c,d}$	Biomass in cell c and day d	kg		
$Z_{c,d}$	Total mortality in cell c for day d	-		
$R_{c,\tilde{y}}$	Annualy recruited fish in cell	$\mathrm{yr}^{\text{-}1}$		
ho	Brody's growth coefficient	$\mathrm{yr}^{\text{-}1}$		
Wt_R	Weight of a fully recruited fish	kg		
Wt_{R-1}	Weight of a pre-recruit fish	kg		
$lpha_d$	Proportion of annually recruited fish recruited during	-		
	$\mathrm{day}\ d$			
Baranov co	atch equation			
$C_{c,d}$	Catch from cell c for day d	kg		
$F_{c,d}$	Instantaneous rate of fishing mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$M_{c,d}$	Instantaneous rate of natural mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$B_{c,d}$	Biomass in cell c on day d	kg		
Recruitment dynamics				
$\tilde{R}_{c,d}$	is the recruitment in cell c for day d	d^{-1}		
$B_{c,d}$	is the Biomass in cell c for day d	d^{-1}		
α	the maximum recruitment rate	kg		
β	the biomass required to produce half the maximum	kg		
	rate of recruitment			

Table 2: Description of variables for population movement sub-module						
Variable	Meaning	Units				
	Population movement dynamics					
Habitat me	odel					
a	b	С				
Thermal to	lerance					
$T_{c,wk}$	Temperature for cell in week	$^{\circ}\mathrm{C}$				
μ_p	Mean of the thermal tolerance for population	$^{\circ}\mathrm{C}$				
σ_p^2	Standard deviation of thermal tolerance for the pop-	$^{\circ}\mathrm{C}$				
	ulation					
Population	Population movement model					
λ	decay rate for population movement	-				
$Hab_{c,p}^2$	Square of habitat suitability for cell \boldsymbol{c} and population	-				
	p					
$Tol_{c,p,wk}$	Thermal tolerance for population p in cell c at week	-				
	wk					
d_{IJ}	euclidean distance between cell ${\cal I}$ and cell ${\cal J}$	-				

Table 3: Description of variables for fleet dynamics sub-module

Variable	Meaning	Units			
Short-term fleet dynamics					
Rev	Revenue from fishing tow	€			
L_p	Landings of population p	kg			
Pr_p	Average price of population p	$\in \ \mathrm{kg}^{-1}$			
StepL	Step length for vessel	euclidean			
		distance			
Br	Bearing	degrees			
k	Concentration parameter for Von mises distribution	-			
eta_1	shape parameter for step function	-			
eta_2	shape parameter for step function	-			
β_3	shape parameter for step function	-			

Table 4: Population dynamics and movement parameter setting

Parameter	Pop 1	Pop 2	Pop 3	Pop 4	
Habitat quality					
Matérn ν	1/0.15	1/0.05	1/0.55	1/0.05	
Matérn κ	1	2	1	1	
Anisotropy	1.5,3,-3,4	1,2,-1,2	2.5,1,-1,2	0.1,2,-1,0.2	
Spawning areas (bound	40,50,40,50;	50,60,30,40;	30,34,10,20;	50,55,80,85;	
box)	80,90,60,70	80,90,90,90	60,70,20,30	30,40,30,40	
Spawning multiplier	10	10	10	10	
Movement λ	0.3	0.3	0.3	0.3	
Population dynamics					
Starting Biomass	1e5	2e5	1e5	1e4	
Beverton-Holt Recruit 'a'	60	100	80	2	
Beverton-Holt Recruit 'b'	250	250	200	50	
Beverton-Holt Recruit σ^2	0.4	0.3	0.4	0.3	
Recruit week	13-16	12-16	14-16	16-20	
Spawn week	16-18	16-19	16-18	18-20	
K	0.3	0.3	0.3	0.3	
wt	1	1	1	1	
wt_{d-1}	0.1	0.1	0.1	0.1	
M (annual)	0.2	0.2	0.2	0.1	

Table 5: Fleet dynamics parameter setting					
Parameter	Fleet	Fleet	Fleet	Fleet	Fleet
	1	2	3	4	5
Targeting preferences					
Price Pop1	100	100	100	100	100
Price Pop2	200	200	200	200	200
Price Pop3	600	600	600	600	600
Price Pop4	1600	1600	1600	1600	1600
Q Pop1	0.01	0.02	0.02	0.01	0.01
Q Pop2	0.02	0.01	0.02	0.01	0.03
Q Pop3	0.01	0.02	0.02	0.01	0.02
Q Pop4	0.02	0.01	0.02	0.05	0.01
Exploitation dynamics					
step function β_1	1	2	1	2	3
step function β_2	10	10	8	12	7
step function β_3	Q90	Q90	Q85	Q90	Q80
step function $rate$	10	20	15	25	10
Past Knowledge	Τ	${f T}$	${f T}$	Τ	${ m T}$
Past Year & Month	T	T	T	T	${f T}$
Past Trip	T	T	T	Τ	${f T}$
Threshold	0.75	0.75	0.75	0.75	0.75

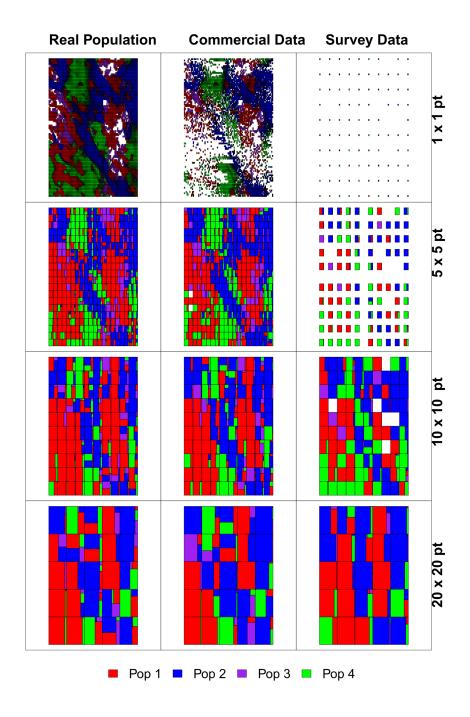


Figure 2: Data aggregation at different spatial resolutions

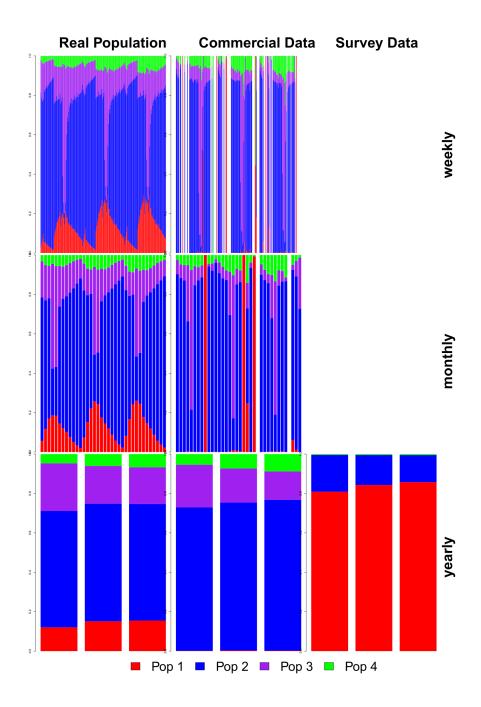


Figure 3: Data aggregation at different temporal resolutions

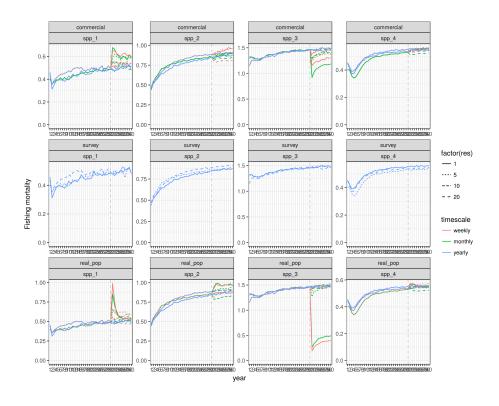


Figure 4: Comparison of closure scenarios - F trends

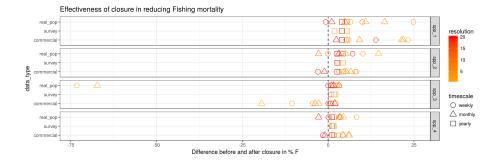


Figure 5: Comparison of closure scenarios

416 References

- Alverson, D.L., Freeberg, M.H., Murawski, S.A., Pope, J., 1994. A global assessment of fisheries bycatch and discards.
- Bartumeus, F., Da Luz, M.G.E., Viswanatham, G.M., Catalan, J., 2005. Animal
- 420 Search Strategies: A Quantitative Random Walk Analysis. Ecological Society
- of America 86, 3078–3087.
- Bastardie, F., Nielsen, J.R., Ulrich, C., Egekvist, J., Degel, H., 2010. De-
- tailed mapping of fishing effort and landings by coupling fishing logbooks
- with satellite-recorded vessel geo-location. Fisheries Research 106, 41–53.
- Batsleer, J., Hamon, K.G., Overzee, H.M.J., Rijnsdorp, A.D., Poos, J.J., 2015.
- High-grading and over-quota discarding in mixed fisheries. Reviews in Fish
- Biology and Fisheries 25, 715–736.
- Bellido, J.M., Santos, M.B., Pennino, M.G., Valeiras, X., Pierce, G.J., 2011.
- Fishery discards and by catch: Solutions for an ecosystem approach to fisheries
- 430 management?
- 431 Bertrand, S., Bertrand, A., Guevara-Carrasco, R., Gerlotto, F., 2007. Scale-
- invariant movements of fishermen: The same foraging strategy as natural
- predators. Ecological Applications 17, 331–337.
- Beverton, R.J., Holt, S.J., 1957. On the Dynamics of Exploited Fish Populations
- 435 , 533.
- 436 Catchpole, T.L., Revill, A.S., 2008. Gear technology in Nephrops trawl fisheries.
- Reviews in Fish Biology and Fisheries 18, 17–31.
- ⁴³⁸ Codling, E.A., Plank, M.J., Benhamou, S., Interface, J.R.S., 2008. Random
- walk models in biology. Journal of the Royal Society, Interface / the Royal
- society 5, 813–34.
- ⁴⁴¹ Crowder, L.B., Murawski, S.A., 1998. Fisheries Bycatch: Implications for Man-
- agement. Fisheries 23, 8–17.

- Deriso, R.B., 1980. Harvesting Strategies and Parameter Estimation for an Age-
- Structured Model. Canadian Journal of Fisheries and Aquatic Sciences 37,
- 268-282. arXiv:1410.7455v3.
- Dichmont, C.M., Punt, A.E., Deng, A., Dell, Q., Venables, W., 2003. Applica-
- tion of a weekly delay-difference model to commercial catch and effort data
- for tiger prawns in Australia 's Northern Prawn Fishery. Fisheries Research
- 449 65, 335–350.
- Diggle, P.J., Ribeiro, P.J., 2007. Model-based Geostatistics (Springer Series in
- Statistics). volume 1.
- Dinmore, T.A., Duplisea, D.E., Rackham, B.D., Maxwell, D.L., Jennings, S.,
- ⁴⁵³ 2003. Impact of a large-scale area closure on patterns of fishing disturbance
- and the consequences for benthic communities. ICES Journal of Marine Sci-
- ence 60, 371–380.
- 456 Dunn, D.C., Boustany, A.M., Roberts, J.J., Brazer, E., Sanderson, M., Gardner,
- B., Halpin, P.N., 2014. Empirical move-on rules to inform fishing strategies:
- A New England case study. Fish and Fisheries 15, 359–375.
- 459 Dunn, D.C., Maxwell, S.M., Boustany, A.M., Halpin, P.N., 2016. Dynamic
- ocean management increases the efficiency and efficacy of fisheries manage-
- ment. Proceedings of the National Academy of Sciences, 201513626.
- 462 Edwards, A.M., 2011. Overturning conclusions of Lévy flight movement patterns
- by fishing boats and foraging animals. Ecology 92, 1247–1257.
- F. Dormann, C., M. McPherson, J., B. Araújo, M., Bivand, R., Bolliger, J.,
- 465 Carl, G., G. Davies, R., Hirzel, A., Jetz, W., Daniel Kissling, W., Kühn, I.,
- Ohlemüller, R., R. Peres-Neto, P., Reineking, B., Schröder, B., M. Schurr,
- 467 F., Wilson, R., 2007. Methods to account for spatial autocorrelation in the
- analysis of species distributional data: A review. Ecography 30, 609–628.
- Gerritsen, H.D., Lordan, C., Minto, C., Kraak, S.B.M., 2012. Spatial patterns
- in the retained catch composition of Irish demersal otter trawlers: High-

- resolution fisheries data as a management tool. Fisheries Research 129-130,
- 472 127-136.
- Gillis, D.M., Peterman, R.M., 1998. Implications of interference among fishing
- vessels and the ideal free distribution to the interpretation of CPUE. Canadian
- Journal of Fisheries and Aquatic Sciences 55, 37–46.
- Girardin, R., Vermard, Y., Thébaud, O., Tidd, A., Marchal, P., 2015. Predicting
- fisher response to competition for space and resources in a mixed demersal
- fishery. Ocean & Coastal Management 106, 124–135.
- Hilborn, R., Walters, C., 1992. Quantitative fisheries stock assessment: Choice,
- dynamics and uncertainty. volume 2. arXiv:1011.1669v3.
- 481 Holmes, S.J., Bailey, N., Campbell, N., Catarino, R., Barratt, K., Gibb, A., Fer-
- nandes, P.G., 2011. Using fishery-dependent data to inform the development
- and operation of a co-management initiative to reduce cod mortality and cut
- discards. ICES Journal of Marine Science 68, 1679–1688.
- Hutton, T., Mardle, S., Pascoe, S., Clark, R.a., 2004. Modelling fishing location
- choice within mixed fisheries: English North Sea beam trawlers in 2000 and
- ⁴⁸⁷ 2001. ICES Journal of Marine Science 61, 1443–1452.
- ⁴⁸⁸ Kennelly, S.J., Broadhurst, M.K., 2002. By-catch begone: Changes in the phi-
- $_{\mbox{\tiny 489}}$ $\,$ losophy of fishing technology. Fish and Fisheries 3, 340–355.
- Lee, J., South, A.B., Jennings, S., 2010. Developing reliable, repeatable, and
- accessible methods to provide high-resolution estimates of fishing-effort distri-
- butions from vessel monitoring system (VMS) data. ICES Journal of Marine
- science 67, 1260–1271.
- Little, A.S., Needle, C.L., Hilborn, R., Holland, D.S., Marshall, C.T., 2014.
- Real-time spatial management approaches to reduce by catch and discards:
- experiences from Europe and the United States. Fish and Fisheries , n/a-
- 497 n/a.

- Mateo, M., Pawlowski, L., Robert, M., 2016. Highly mixed fisheries: fine-scale
- spatial patterns in retained catches of French fisheries in the Celtic Sea. ICES
- Journal of Marine Science: Journal du Conseil, fsw129.
- Poos, J.J., Rijnsdorp, A.D., 2007. An "experiment" on effort allocation of fishing
- vessels: the role of interference competition and area specialization. Canadian
- Journal of Fisheries and Aquatic Sciences 64, 304–313.
- R Core Team, 2017. R Core Team (2017). R: A language and environment for
- statistical computing. R Foundation for Statistical Computing, Vienna, Aus-
- tria. URL http://www.R-project.org/., R Foundation for Statistical Com-
- 507 puting.
- $_{508}$ Reynolds, A., 2015. Liberating Lévy walk research from the shackles of optimal
- foraging.
- Ricker, W.E., 1954. Stock and recruitment. Journal of the Fisheries Research
- ₅₁₁ Board of Canada 11, 559 623.
- Rijnsdorp, A., 2000. Competitive interactions among beam trawlers exploiting
- local patches of flatfish in the North Sea. ICES Journal of Marine Science 57,
- 514 894-902.
- Rijnsdorp, a.D., Daan, N., Dekker, W., Poos, J.J., Van Densen, W.L.T., 2007.
- Sustainable use of flatfish resources: Addressing the credibility crisis in mixed
- fisheries management. Journal of Sea Research 57, 114–125.
- Rijnsdorp, A.D., Piet, G.J., Poos, J.J., 2001. Effort allocation of the Dutch
- beam trawl fleet in response to a temporarily closed area in the North Sea.
- Ices Cm 2001/N: 01, 1–17.
- 521 Sakiyama, T., Gunji, Y.P., 2013. Emergence of an optimal search strategy from
- a simple random walk. Journal of the Royal Society, Interface 10, 20130486.
- 523 Schlater, M., Malinowski, A., Menck, P.J., 2015. Analysis, Simulation and Pre-
- diction of Multivariate Random Fields with Package RandomFields. Journal
- of Statistical Software 63, 1–25. arXiv:1501.0228.

- Schnute, J., 1985. A genera theory for analysis of catch and effort data. Canadian Journal of Fisheries and Aquatic Sciences 42, 414–429.
- Sims, D.W., Humphries, N.E., Bradford, R.W., Bruce, B.D., 2012. Lévy flight
- and Brownian search patterns of a free-ranging predator reflect different prey
- field characteristics. Journal of Animal Ecology 81, 432–442.
- Sims, D.W., Southall, E.J., Humphries, N.E., Hays, G.C., Bradshaw, C.J.A.,
- Pitchford, J.W., James, A., Ahmed, M.Z., Brierley, A.S., Hindell, M.A., Mor-
- ritt, D., Musyl, M.K., Righton, D., Shepard, E.L.C., Wearmouth, V.J., Wil-
- son, R.P., Witt, M.J., Metcalfe, J.D., 2008. Scaling laws of marine predator
- search behaviour. Nature 451, 1098–U5.
- Tidd, A.N., Hutton, T., Kell, L.T., Blanchard, J.L., 2012. Dynamic prediction
- of effort reallocation in mixed fisheries. Fisheries Research 125-126, 243–253.
- Tobler, W.R., 1970. A Computer Movie Simulating Urban Growth in the Detroit
- Region. Economic Geography 46, 234. arXiv:1011.1669v3.
- Ulrich, C., Reeves, S.a., Vermard, Y., Holmes, S.J., Vanhee, W., 2011. Rec-
- onciling single-species TACs in the North Sea demersal fisheries using the
- 542 Fcube mixed-fisheries advice framework. ICES Journal of Marine Science 68,
- 1535-1547.
- Viswanathan, G.M., Buldyrev, S.V., Havlin, S., Da Luz, M.G.E., Raposo, E.P.,
- 545 Stanley, H.E., 1999. Optimizing the success of random searches. Nature 401,
- ₅₄₆ 911–914.