MixFishSim: highly resolved spatiotemporal simulations for exploring mixed fishery dynamics

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Abstract

Most fisheries Fishing JJ exploits JJ spatially and temporally hetergenous fish populations, using species-unselective gear that can result in unintended, unwanted catch of low quota or protected species. Reducing these unwanted catches is crucial for biological and economic sustainability of 'mixed fisheries' and implementation of an ecosystem approach to fishing.

If fisheries are to avoid unwanted catch To implement effective spatial measures to reduce discards PD a good understanding of spatiotemporal fishery dynamics is required. However, traditional scientific advice is limited by a lack of highly resolved knowledge of population distribution, movement and how fishers interact with different fish populations. This reflects that data on fish location at high temporal and spatial resolutions is expensive and difficult to collect and therefore proxies inferred from either scientific surveys or commercial catch data are often used to model distributions, often with limited spatial and temporal resolution.

To understand how resolution impacts mixed fisheries inference, we develop

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a highly resolved spatiotemporal simulation model incorporating: i) delay-

difference population dynamics, ii) population movement using Gaussian Ran-

dom Fields to simulate patchy, hetergenously distributed populations, and iii)

fishery dynamics for multiple fleet characteristics based on targetting via corre-

lated random walk movement and learned behaviour.

We simulate 20 years of exploitation of the fish populations and use the

results from the fishing model to draw inference on the underlying population

structures. We compare this inference to i) a simulated fixed-site sampling

design commonly used for fisheries monitoring purposes, and ii) the true un-

derlying population structures input to the simulation, to establish the poten-

tial and limitations of fishery-dependent data - an inherently biased sampling

method due to fisher's targeting- to provide a robust picture of spatiotemporal

distributions. Finally, we simulate an area closure based on areas defined from

commercial the known ("real-population") distribution, commercial catch data

and survey data at different temporal and spatial resolutions and assess their

effectiveness on reducing catches of a fish population.

We conclude from our simulations that commercial data, while not unbiased,

provides a useful tool for managing catches in mixed fisheries if applied at the

correct spatiotemporal scale.

[333 words]

Keywords: Some, keywords, here. Max 6

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1. Introduction

Fishers exploit fish populations that are heterogeneously distributed in space

and time with varying knowledge of species distributions using species-unselective

fishing gear. In doing so^{PD} fisheries that^{PD} catch an assemblage of species

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and and and and and and an anixed fisheries may, when managed by single species quotas can end up JJ discarding JJ overquota catch when managed by single species quotas, JJ leading to overexploitation of fish populations (Ulrich et al., 2011; Batsleer et al., 2015) JJ. This discarding of fish in excess of quota hampers the ability to limit fishing mortality to within sustainable limits (Alverson et al., 1994; Crowder and Murawski, 1998; Rijnsdorp et al., 2007) JJ; reducing discarding is crucial to ensure biological and economic sustainability of fisheries and implementation of an ecosystem approach to fisheries JJ and. As such pd there is increasing interest in technical solutions such as gear and spatial closures as ways of reducing unwanted catchavoiding discarding of fish JJPD (Kennelly and Broadhurst, 2002; Catchpole and Revill, 2008; Bellido et al., 2011).

Changes to spatial fishing patterns have Use of spatial management as a tools PD been proposed as a method to reduce discards (Holmes et al., 2011; Little et al., 2014; Dunn et al., 2014) PD. However, its PD implementation is hampered by lack of knowledge of fish and fishery spatiotemporal dynamics and understanding of the scale at which processes are important for management. Understanding the correct scale for spatial management is crucial in order to implement measures at a resolution that ensures effective management (Dunn et al., 2016) while minimising economic impact. For example, a scale that promotes species avoidance for vulnerable or low quota species while allowing continuance of sustainable fisheries for available quota species.

Identifying Ensuring measures are implemented at PD an appropriate scale has been a challenge in the past that has led to ineffectual measures with unintended consequences such as limited impact towards the management objective or increased benthic impact on previously unexploited areas (e.g. the cod closure in the North Sea (Rijnsdorp et al., 2001; Dinmore et al., 2003)). MSince
then mPD ore refined spatial information has since PD become available through
the combination of logbook and Vessel Monitoring System (VMS) data (Lee et al., 2010; Bastardie et al., 2010; Gerritsen et al., 2012; Mateo et al., 2016)

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and more real-time spatial management has been possible (e.g. Holmes et al., 2011). Such information is, however, patchy and derived from an inherently biased sampling programme (i.e. targeted fishing). Further, fishers generally only recorded landings (not eatch) on a daily basis. This leads to questions about the validity of inference that can be drawn from landings data assigned to VMS activity pings. PD

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discards JA Agree
have removed
this to avoid
confusion PD

In order to understand the consequences of usingehallenges that face^{PD}
VMS-linked landings to draw inference on the underlying population structure
we develop a simulation model where population dynamics are highly-resolved
in space and time. Being-and-are^{PD} known directly^{PD} rather than inferred from
sampling or commercial catches, we can use the population model to evaluate
how inference from fisheries-dependent and fisheries independent sampling relates to the real population structure^{PD}. In our model system pP^{PD}opulation
movement is driven by random (diffusive) and directed (advective) processes and
we incorporate characterisation of a number of different fisheries dynamics^{PD}
exploiting four fish populations with different spatial and population demographics.

Using our model we simulate 2040^{PD} years of exploitation of the fish populations. We and PD use the results from the fishing model: PD

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- to understand how sampling-derived data reflects the underlying population structures. We compare at different spatial and temporal aggregations of data the real population to:
 - (a) the inferred population from a stratified fixed-site sampling survey design commonly used for fisheries monitoring purposes, otherwise know as a fisheries-independent survey,
 - (b) the inferred population from our fishery-dependent model which includes fishery-induced sampling dynamics.

- 2. to understand the impact of data aggregation and source on spatial fisheries management measures we simulate a fishery closure to protect a species based on different spatial and temporal data aggregations:
 - (a) as if the real spatial population structure were known,
 - (b) the fishery-independent inferred population structure
 - (c) the fishery-dependent inferred population structure

We evaluate the theoretical "benefit" to the population of the closure(s), 72 and effect on the other three populations. Further, we extend our analysis 73

to a range of spatial and temporal scales to assess the impact of these 74 processes on the success of the management measure. PD 75

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2. Materials and Methods

AWe developed and implemented a simulation model with a^{PD} modular event-based simulation model was developed with approach, where sub-PD modules 79 are PD implemented on independent time-scales appropriate to capture the char-80 acteristic of the different processesprocess modelled PD (Figure 1). The following 81 sub-modules were included to capture the full system: 1) Population dynamics, 82 2) Recruitment dynamics, 3) Population movement, 4) fishery dynamics. PD

The fishing model operated on a tow-by-tow basis, while PD PpPD opulation dynamics (fishing and natural mortality, growth) operate on a daily time-step, 86 while p. PPD opulation movement occurs on a weekly time-step. R, while 87

 r^{PD} ecruitment takes placeoccurs PD periodcally each year for a set time duration r^{PD} periodcally each year for a set time duration r^{PD} (e.g. 3 weeks)^{PD} at at specified point individual to a species-PD, while the fish-

ing module operates on a tow-by-tow basis (multiple events a day)^{PD}. The 90

simulation framework is implemented in the statistical software package R (R 91

Core Team, 2017) and PD available as an R package from the authors github

(www.github.com/pdolder/MixFishSim).

would like to keep both parts but have made clearer in how its set out. The closure scenarios form validation of the data aggregation. rather than of the closures themselves - so its a continua-

tion of the same

question in my

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If the paper

has two goals

- Here we describe each of the model components; 1) Population dynamics, 2)
- 96 Recruitment dynamics, 3) Population movement dynamics, 4) fishery dynamics. PD

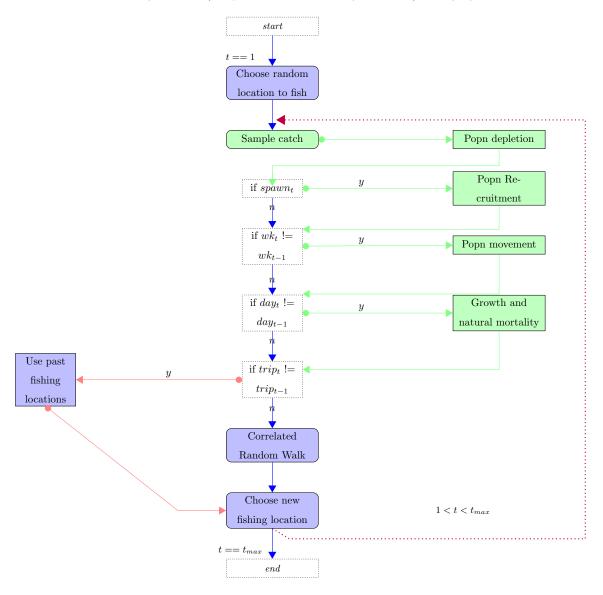


Figure 1: Overview Schematic of simulation model. The blue boxes indicate fleet dynamics processes, the green boxes population dynamics processes while the white boxes are the timesteps at which processes occur; $t=\mathrm{tow},\,tmax$ is the total number of tows, wk is a weekly timestep, day is a day timestep, trip is a trip time step. I NEED TO REDO THIS TO MAKE NOTATION MORE CONCISE AND CONSISTENT

2.1. Population dynamics

The basic population level processes are simulated using a modified twostage Deriso-Schnute delay difference model (Deriso, 1980; Schnute, 1985; Dichmont et al., 2003) occurring at a daily time-step. A daily time-step was chosen as to discretise continuous population processes on a biologically relevent and computationally tractable timescale. PD Under the population dynamics module Here, PD population biomass growth and depletion for pre-recruits and fish PD recruited fish PD to the fishery PD are modelled separately as a function of previous recruited biomass, intrinsic population growth and recruitment:

$$\begin{split} B_{c,d+1} &= \\ & (1+\rho)B_{c,d} \cdot e^{-Z_{c,d}} - \rho \cdot e^{-Z_{c,d}} & \times \\ & (B_{c,d-1} \cdot e^{-Z_{c,d-1}} + Wt_{R-1} \cdot \alpha_{d-1} \cdot R_{\tilde{y}(c,y,d-1)}) &+ \\ & Wt_R \cdot \alpha_d \cdot R_{\tilde{y}(c,y,d)} \end{split}$$

where ρ is Brody's coefficient, shown to be approximately equal to exp(-K), where K is the growth rate from a von bertalanffy logistic growth model (Schnute, 1985). Wt_{R-1} is the weight of fish prior to recruitment, while Wt_R is the recruited weight. α_d represents the proportion of fish recruited during that day for the year, while $R_{c,y}$ is the annual recruits in cell c for year y.

Mortality $Z_{c,d}$ can be decomposed to natural mortality, $M_{c,d}$, and fishing mortality, $F_{c,d}$, where both $M_{c,d}$ and $F_{c,d}$ are instantaneous rates with $M_{c,d}$ fixed and $F_{c,d}$ calculated by solving the Baranov catch equation (Hilborn and

Walters, 1992) for $F_{c,d}$:

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$$C_{c,d} = \frac{F_{c,d}}{F_{c,d} + M_{c,d}} * (1 - e^{-(F_{c,d} + M_{c,d})}) * B_{c,d}$$

where $C_{c,d}$ is the summed catch from the fishing model across all fleets and vessels in cell c for the population during the day d, and $B_{c,d}$ the daily biomass for the population in the cell.

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catch for a ves

2.2. Recruitment dynamics

Recruitment is modelled through a function relating the mature biomass to recruits at time of recruitment. In *mixfishsim*, it can be modelled either either as a stochastic Beverton-Holt stock-recruit form (Beverton and Holt, 1957):

$$\bar{R}_{c,d} = \frac{(\alpha * B_{c,d})}{(\beta + B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

Where α is the maximum recruitment rate, β the spawning stock biomass (SSB) required to produce half the maximum, B current SSB and σ^2 the variability in the recruitment due to stochastic processes.

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or a stochastic Ricker form (Ricker, 1954):

$$\bar{R}_{c,d} = B_{c,d} * e^{(\alpha - \beta * B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

where α is the maximum productivity per spawner and β the density dependent reduction in productivity as the SSB increases.

2.3. Population movement dynamics

To simulate how^{JJ} fish populations might be^{JJ} distributioned^{JJ} in space and time, we employed^{JJ} a Gaussian spatial process was employed^{JJ} to model habitat suitability for each of the populations. An, with an JJ advection-diffusion process to JJ controlled JJ how the JJ populations movementd JJ over time with a moving temperature covariate to capture temporal dependencies. This was intended to balance realism in population movement, capturing the main directed and random processes, and practicality of modelling the population rather than individual fish. JJ

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What have a temperature covariate? Could just use time J Was intended as some biological meaning - species thermal tolerances load onto the temperature effect PD

For the PD habitat we defined PD a Gaussian random field process, $\{S(c): c \in \mathbb{R}^2\}$, that is a stochastic process PD where for PD any set of cells c_1, \ldots, c_n where for each $c_i \in \mathbb{R}^{2}$ PD, the joint distribution of $S = \{S(c1), \ldots S(c_n)\}$ is multivariate Gaussian. The distribution is specified by its mean function, $\mu(c) = E[S(c)]$ and its covariance function, $\gamma(c,c') = Cov\{S(c),S(c')\}$ (Diggle and Ribeiro, 2007).

The covariance structure affects the smoothness of the surfaces which the process generates; and PD we used the Matérn family of PD covariance structures tures PD, as one where PD the correlation strength weakens the further the distance apart (i.e. the correlation between S(x) and S(x') decreases as the distance u = ||x - x'|| increases) PD. The Matérn covariance structure models the spatial autocorrelation observed with animal distributions (Tobler, 1970; F. Dormann et al., 2007) PD and The Matérn correlation PD is a two-parameter family where:

Not clear how Introduce the gamma function, and why this covariance structure? Why correlate values in the random field? JM to allow populations to have different aggregation densities: have tried to clarify PD

$$\rho(u) = \{2^{\kappa - 1} \Gamma \kappa\}^{-1} (u/\phi)^{\kappa} K_{\kappa}(u/\phi)$$

 $K_{\kappa}(.)$ is a modified Bessel function of order κ , $\phi > 0$ is a scale parameter with the dimensions of distance, and $\kappa > 0$, called the order, is a shape parameter which determines the smoothness of the underlying process.

TIn the simulation model, t^{PD}he habitat for each of the populations wasis^{PD} generated withthrough^{PD} the *RFSimulate* function of the *RandomFields* R package (Schlater et al., 2015), implementing different parameter settings to affect the patchiness of the populations. Each population wasis^{PD} initialised at a single location, and subsequently moveeds^{PD} according to a probabilistic distribu^{PD} tion based on habitat suitability, temperature and distance from current cell: ^{PD}

$$Pr(J|I) = \frac{e^{-\lambda * d_{IJ}} \cdot (Hab_{J,p}^2 \cdot Tol_{J,p,wk})}{\sum\limits_{c=1}^{C} e^{-\lambda * d} \cdot (Hab_{c,p}^2 \cdot Tol_{c,p,wk})}$$
(1)

Where d_{IJ} is the euclidean distance between cell I and cell J, λ is a given rate of decay, $Hab_{J,p}^2$ is the squared index of habitat suitability for cell J and population p, with $Tol_{J,p,wk}$ the temperature tolerance for cell J by population p in week wk.

During specified weeks of the year, the habitat quality wasis^{PD} modified for user-defined^{PD} spawning habitats^{PD}, resulting inmeaning^{PD} each population hadhas^{PD} a concentrated area where spawning takes place and the population moveds^{PD} towards these cellsthis^{PD} in the weeks prior to spawning.

The temperature field wasis^{PD} simulated to be on a gradient from a South-Westerly to North-Easterly direction, with temperature in each cell changing gradually on a week-by-week basis so that initially high temperature areas cycled^{PD} to lower temperatures and low temperature areas vice versa. Each population p wai^{PD}s assigned a thermal tolerance with mean, μ_p^{PD} and variance, σ_p^{PD} so that each cell and population temperature suitability is defined that:

What does it mean concisely? Areas are assigned? JM Yes, the areas are predefined - I have amended to reflect and tried to clarify PD

$$Tol_{c,p,wk}^{\text{PD}} = \frac{1}{\sqrt{(2\pi \cdot \sigma_p^2)}} \cdot \exp\left(-\frac{(T_{c,wk}^{\text{PD}} - \mu_p)^2}{2 \cdot \sigma_p^2}\right)$$
(2)

Where $Tol_{c,p,wk\text{PD}}$ is the tolerance of population p in cell c in week wk^{PD} , $T_{c,wk\text{PD}}$ is the temperature in the cell given the week $^{\text{PD}}$ and μ_p^{PD} and σ_p^{PD} the mean and standard deviation of the population temperature tolerance.

The final process resulted in independent populations structure and movement patterns, with population movement occuring on a weekly basis. This process approximated the demographic shifts in fish populations throughout a year while maintaining seasonal patterns for spawning. PD

2.4. Fleet dynamics

The fleet dynamics can be broadly categorised into three components; fleet targeting - which determineds^{PD} the fleet catch efficiency and preference towards

a particular species; trip-level decisions, which determined PD the initial location to be fished at the beginning of a trip; and within-trip decisions, determining 196 movement from one fishing spot to another within a trip.

2.4.1. Fleet targeting

Each fleet of n vessels wasis^{PD} characterised by both a general efficiency, Q, 199 and a population specific efficiency, Q_{ν} . Thus, the product of these parameters 200 affectsed^{PD} the overall catch rates for the fleet and the preferential targeting of 201 one population over another. This, in combination with the parameter choice 202 for the step-function defined below^{PD} (as well as some randomness from the 203 exploratory fishing process) determineds the preference of fishing locations for the fleet. All species prices wereare PD kept the same, across fleetsand seasons, though can be made to vary seasonally PD. 206

2.4.2. Trip-level decisions 207

Several studies (e.g. Hutton et al., 2004; Tidd et al., 2012; Girardin et al., 208 2015) have confirmed past activity and past catch rates are strong predictors 209 of fishing location choice. For this reason, the fleet dynamics sub-model in-210 cludeds^{PD} a learning component, where a vessel's initial fishing location in a trip wai^{PD}s based on selecting from previously successful fishing locations. This 212 wai^{PD}s achieved by sorting all previous fishing events in the previous trip as well 213 as the previous time periods in past years, and choosing randomly from the top 214 75 % of fishing events as defined by the revenue gainedin valuePD. Simulation 215 testing indicated that this learning increased the mean value of catches for the 216 vessels, over just relying on the correlated random walk function as described 217 for the 'within trip' decisions below^{PD}. 218

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2.4.3. Within-trip decisions

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Fishing locations within a trip are determined by a modified random walk 220 process. A random walk type was chosen as it is the simplest assumption commonly used in ecology to describe optimal^{PD} animal movement which^{PD} search strategying^{PD} for exploiting^{PD} homogeneously distributed prey about which there is uncertain knowledge (Viswanathan et al., 1999). In a random walk, movement is a stochastic process through a series of steps. These steps have a length, and a direction JJ that can either be equal in length or take some other functional form. The direction of the random walk can be correlated, a characteristic known as 'persistence', providing some overall location of PD directional movement (Codling et al., 2008) or uncorrelated PD.

A Lévy flightlévy walk^{JJ} is a particular form of random walk characterised by a heavy-tailed distribution of step-length. The Lévy flightand^{JJ} has received a lot of attention in ecological theory in recent years as having shown to have very similar characteristics as those observed by animals in nature, and being a near optimum searching strategy for predators pursuing patchily distributed prey (Viswanathan et al., 1999; Bartumeus et al., 2005; Sims et al., 2008). Bertrand et al. (2007) showed that Peruvian anchovy fishermen have a stochastic search pattern similar to that observed with a lévy walk. However, it remains a subject of debate (e.g. see Edwards, 2011; Reynolds, 2015)^{PD}, with the contention that search patterns may be more simply characteristed as random walks (Sakiyama and Gunji, 2013) with specific patterns related to the characteristics of the prey field (Sims et al., 2012).

We use a modified random walk where directional change is based on a correlated circular distribution where a favourable fishing ground is likely to be "fished back over" by the vessel returning in the direction it came from and step length (i.e. the distance travelled from the current to the next fishing location) is determined by relating JJ recent fishing success, measured as the summed value of fish caught (revenue, Rev),

$$Rev = \sum_{p=1}^{P} \underline{L}\underline{\underline{C}}^{\text{PD}}{}_{p} \cdot Pr_{p}$$

where $L\underline{C}^{PD}_{p}$ is landingseatch^{PD} of a population p, and Pr_{p} price of a population, to step distance. Here, when fishing is successful vessels remain in a similar location and continue to exploit the local fishing grounds. When unsuccessful, they move some distance away from the current fishing location. The movement distance retains some degree of stochasticity, which can be controlled separately. The step function takes the form:

$$StepL = e^{log(\beta_1) + log(\beta_2) - (log(\frac{\beta_1}{\beta_3}))} * Rev$$

So that, a step from (x1,y1) to (x2, y2) is defined by:

So step length increases with increasingly gross revenue? JJ No, the opposite PD

$$(x2, y2) = x1 + StepL \cdot \cos(\frac{\pi \cdot Br}{180}),$$

$$y1 + StepL \cdot \sin(\frac{\pi \cdot Br}{180})$$

$$with \quad Br_{t-1} < 180, Br_t = 180 + \sim vm[(0, 360), k]$$

$$Br_{t-1} > 180, Br_t = 180 - \sim vm[(0, 360), k]$$

with k the concentration parameter from the von Mm^{JJ}ises distribution which we correlate with the revenue so that $k = (Rev + 1/RefRev) * max_k$, where max_k is the maximum concentration value, k, and RefRev is parameterised as for β_3 in the step length function.

2.4.4. Local population depletion

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Where several fishing vessels are exploiting the same fish population compe-257 tition is known to play an important role in local distribution of fishing effort 258 (Gillis and Peterman, 1998). If several vessels are fishing on the same patch of 259 fish, local depletion and interference competition JJ will affect fishing location choice of the fleet as a whole (Rijnsdorp, 2000; Poos and Rijnsdorp, 2007). In 261 order to account for this behaviour, the fishing sub-model operates spatially on 262 a daily time-step so that for future days the biomass available to the fishery 263 is reduced in the areas fished. The cumulative effect is to make heavily fished 264 areas less attractive as future fishing opportunities. 265

2.5. Fisheries independent survey

A fisheries-independent survey is simulated where fishing on a regular grid
begins each year at the same time for a given number of stations (a fixed station
survey design). Catches of the populations at each station present JJ are recorded
but not removed from the population. This provides a fishery independent
snapshot of the populations at a regular spatial intervals distribution JJ each
year, similar to scientific surveys undertaken by fisheries research agencies.

273 3. Calculation

274 3.1. Population parameterisation

We parameterised the simulation model for four populations with differing habitat preference,—and temperature tolerances (Figures S1, S3, S4, S5, S6, S7), population demographic,—and recruitment functions. In addition, each of the populations has two defined spawning areas which result in the populations moving towards these areas in given weeks (Figure S2) and population-specific movement rates (Table 4). The realised movement of the populations for a number of weeks is shown in Figure S9 while the realised daily fishing mortality are shown in Figure S10.

3.2. Fleet parameterisation

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The fleets were parameterised to reflect five different characteristics based on targeting preference and exploitation dynamics (Table 5). Setting a targeting parameter (Q) that differed across fleets ensured different spatial dynamics, due to preferential targeting of populations that differ in their spatial distributions This ensures that different fleets have different spatial dynamics, preferentially targeted different fish populations PD. The stochasticity in the random walk process ensures that different vessels within a fleet have slightly different spatial distributions based on individual experience, while the step function was parameterised dynamically so that vessels take smaller steps where

the fishing location yields in a top quartile of the value available in that year (as defined per fleet in Table 5). 295

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Each fleet was set so that, after the first year, fishing locations were chosen based on experience built up in the same month from previous years and from past trip fishing success. 'Success' in this context was defined as the locations where the top 75 % of revenue from was found in previous trips.

An example of the realised fleet movements for a single vessel during a single trip are given in Figure S11, while Figure S12 shows multiple trips for a single vessel, Figure^{PD} S13 the vessel movements for some trips overlaid on the value field, Figure^{PD} S14 shows fishing locations for an entire fleet of 20 vessels for a single trip, and Figurewhile^{PD} S15 shows an example of the step function realisation and turning angles from the correlated random walk.

3.3. Survey settings

The survey simulation was set up with follow a fixed gridded station design with 100 stations fished each year, starting on day 92 and ending on day 112 (5 stations per day)^{PD} with same catchability parameters for all populations 310 $(Q_p = 1).$ 311

3.4. Simulation settings

To illustrate the capabilities on MixFishSim, we investigate the influence of the temporal and spatial resolution of different data sources on the reduction in catches of a population given spatial closures. To do so, we first set up with simulation to run for 10 years based on a 100 X 100 square grid, with five fleets of 20 vessels each and four fish populations. Fishing takes place four times a day per vessel and five days a week, while population movement is every week.

We allow the simulation to run unrestricted for 5 years, and subsequently close areas for the last 5 years of the simulation based on data (either derived

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section JJI think ecological modelling wants the 'calculations' section here..will $\mathrm{check}^{\mathrm{PD}}$

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from the commercial catches, fisheries-independent survey or the 'real population' - the underlying populations assumed to be known perfectly) used at different spatial and temporal scales.

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The following steps are undertaken to determine closures:

- 1. Extract data source
- 2. Aggregate according to resolution
- 3. Interpolate across entire area at desired resoltion
- 4. Close top 5 % of areas

In total 56 closure scenarios were run which represent combinations of

- data types: commercial logbook data, survey data and 'real population',
- temporal resolutions: weekly, monthly and yearly closures,
- spatial resolutions: 1 x 1 grid, 5 x 5 grid, 10 x 10 grid and 20 x 20 grid.

Survey closures were on an annual basis only, as this was the most temporally resolved survey data available.

337 4. Results

The species distribution themselves

The consequences of different spatial aggregations of the data are shown in Figure 2, which represents the aggregation of catch from each of the data sources over a year at different spatial resolutions.

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The finer spatial grid for the the real population (top left) and commercial
data (top middle) show similar patterns, though there are unsampled gaps in
the commercial data from a lack of fishing activity (particularly in the lower left
part of the sampling domain). The survey data at this spatial resolution shows
very sparse and uninformative information about the spatial distributions of the

Procedure unclear. Refer to symbols in methods section or switch order starting with description of data type etc.. JJ Yes, will redo PD

populations. The slightly aggregated data on a 5 x 5 grid shows similar patterns, and while losing some of the spatial detail there remains good consistency between the 'real population' and the commercial data. Survey data starts to pick out some of the similar patterns as the other data sources, but lacks coverage. The spatial catch information on a 10 x 10 and 20 x 20 grid loses a significant amount of information about the spatial resolutions for all data sources, and some differences between the commercial and 'real population' data emerge.

Figure 3 shows the consequences of different temporal aggregations of the data, with 156 weekly (top), 36 monthly (middle) and 3 yearly (bottom) catch compositions across a 20×20 area.

As can be seen from the 'real population', the monthly aggregation captures the major patterns seen in the weekly data, albeit missing more subtle differences. The yearly data results in a constant catch pattern due to the aggregation process (sometimes known as an aggregation bias). The commercial data on a weekly basis shows some of the same patterns as the 'real population', though the first species (in red) is less well represented and some weeks are missing catches from the area. The monthly data. The monthly data shows some consistency between the 'real population' and commercial data for species 2 - 4, though species 1 remains underrepresented. On an annual basis, interestingly the commercial data underrepresents the first species (in red) while the survey overrepresents species 1. This is likely due to the biases in commercial sampling, with the fisheries not targeting the areas where species 1 are present, and the biases in the survey sampling from overrepresentation of the spatial distribution.

We implemented a spatial closure using the different data sources and spatial and temporal aggregations as outlined in the protocol in Section 3.4. We used this to assess the efficacy of a closure in reducing fishing mortality on species 1, given availability of data and its use at different resolutions in order to evaluate the trade-offs in data sources. Figure 4 shows the trend in fishing mortality

for each species simulated (columns) given the data sources (rows), temporal aggregations (colour lines) and spatial aggregations (linestyles), while Figure 5 shows the change in fishing mortality from before the closure (average F years 2 - 4) to after the closure (average F years 8 - 10).

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For the closures based on 'real population' (bottom row), the most disaggregated data (a weekly timescale and 1 x 1 resolution) was most effective, reducing fishing mortality on species 1 (left) by ~ 60 %. Next was the monthly closures ($< \sim 30$ %). The least effective were the yearly closures (blue lines) at all spatial resolutions, which resulted in increased fishing mortalities (> 30% - N.B. Note though, this is consistent with the increasing trends in F, which is probably more related to the fact that Fs hadn't stabilised in the simulation from the fishing vessels "learning" the best locations - I will rerun the sims for a longer time (20 - 30 years).

For the survey data, which can only be implemented on a yearly timescale, the closures had no effect at any data resolution. The results are identical for the different data resolutions except 20×20 , which is why you can't see more than 2 points. This is because of the sparsity of the sampling locations.

For the commercial data, the most effective closure scenario was based on 1 399 x 1 data at a monthly temporal resolution. This results in \sim 10 % reduction in F for species 1. This was the only closure scenario to have positive effect 401 according to Figure 5, though looking at the trend in Figure 4 this looks more 402 related to the continued increased in F trend, as other scenarios had an initial 403 effect. Interestingly the monthly data scenario was more effective than weekly 404 data, which I'd posit is due to the increase amount of data available from the 405 commercial sampling across a month compared to a week.i Commercial data used at an annual timestep was ineffective in bringing fishing mortality down 407 for species 1. 408

- Given the scenarios above, it seems clear that spatial disaggregation is more important than the temporal disaggregation of the commercial data, except when its used at an annual timeframe, which is the scenario that gave the worst results.
- For the other species in the simulation (population 2 4) there was little difference in fishing mortalities across scenarios.
- Note: The monthly commercial data scenario is the most effective of the realistic scenarios, as the 'real population' can only be seen as a baseline comparison.

419 5. Discussion

420 6. Conclusions

421 Appendices

422 Abbreviations

Detail any unusual ones used.

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Table 1.	Description	of wariables	for population	dynamics of	h modulo
rable i:	Describtion	or variables	тог роршалоп	ovnamics si	то-тоате

Variable	Meaning	Units		
Population dynamics				
Delay-difference model				
$B_{c,d}$	Biomass in cell c and day d	kg		
$Z_{c,d}$	Total mortality in cell c for day d	-		
$R_{c,\tilde{y}}$	Annualy recruited fish in cell	$\mathrm{yr}^{\text{-}1}$		
ho	Brody's growth coefficient	$\mathrm{yr}^{\text{-}1}$		
Wt_R	Weight of a fully recruited fish	kg		
Wt_{R-1}	Weight of a pre-recruit fish	kg		
$lpha_d$	Proportion of annually recruited fish recruited during	-		
	$\mathrm{day}\ d$			
Baranov co	atch equation			
$C_{c,d}$	Catch from cell c for day d	kg		
$F_{c,d}$	Instantaneous rate of fishing mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$M_{c,d}$	Instantaneous rate of natural mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$B_{c,d}$	Biomass in cell c on day d	kg		
Recruitment dynamics				
$\tilde{R}_{c,d}$	is the recruitment in cell c for day d	d^{-1}		
$B_{c,d}$	is the Biomass in cell c for day d	d^{-1}		
α	the maximum recruitment rate	kg		
β	the biomass required to produce half the maximum	kg		
	rate of recruitment			

Table 2: Description of variables for population movement sub-module						
Variable	Meaning	Units				
	Population movement dynamics					
Habitat me	odel					
a	b	С				
Thermal to	lerance					
$T_{c,wk}$	Temperature for cell in week	$^{\circ}\mathrm{C}$				
μ_p	Mean of the thermal tolerance for population	$^{\circ}\mathrm{C}$				
σ_p^2	Standard deviation of thermal tolerance for the pop-	$^{\circ}\mathrm{C}$				
	ulation					
Population	Population movement model					
λ	decay rate for population movement	-				
$Hab_{c,p}^2$	Square of habitat suitability for cell \boldsymbol{c} and population	-				
	p					
$Tol_{c,p,wk}$	Thermal tolerance for population p in cell c at week	-				
	wk					
d_{IJ}	euclidean distance between cell ${\cal I}$ and cell ${\cal J}$	-				

Table 3: Description of variables for fleet dynamics sub-module

Variable	Meaning	Units			
Short-term fleet dynamics					
Rev	Revenue from fishing tow	€			
L_p	Landings of population p	kg			
Pr_p	Average price of population p	$\in \ \mathrm{kg}^{-1}$			
StepL	Step length for vessel	euclidean			
		distance			
Br	Bearing	degrees			
k	Concentration parameter for Von mises distribution	-			
eta_1	shape parameter for step function	-			
eta_2	shape parameter for step function	-			
β_3	shape parameter for step function	-			

Table 4: Population dynamics and movement parameter setting

Parameter	Pop 1	Pop 2	Pop 3	Pop 4	
Habitat quality					
Matérn ν	1/0.15	1/0.05	1/0.55	1/0.05	
Matérn κ	1	2	1	1	
Anisotropy	1.5,3,-3,4	1,2,-1,2	2.5,1,-1,2	0.1,2,-1,0.2	
Spawning areas (bound	40,50,40,50;	50,60,30,40;	30,34,10,20;	50,55,80,85;	
box)	80,90,60,70	80,90,90,90	60,70,20,30	30,40,30,40	
Spawning multiplier	10	10	10	10	
Movement λ	0.3	0.3	0.3	0.3	
Population dynamics					
Starting Biomass	1e5	2e5	1e5	1e4	
Beverton-Holt Recruit 'a'	60	100	80	2	
Beverton-Holt Recruit 'b'	250	250	200	50	
Beverton-Holt Recruit σ^2	0.4	0.3	0.4	0.3	
Recruit week	13-16	12-16	14-16	16-20	
Spawn week	16-18	16-19	16-18	18-20	
K	0.3	0.3	0.3	0.3	
wt	1	1	1	1	
wt_{d-1}	0.1	0.1	0.1	0.1	
M (annual)	0.2	0.2	0.2	0.1	

Table 5: Fleet dynamics parameter setting					
Parameter	Fleet	Fleet	Fleet	Fleet	Fleet
	1	2	3	4	5
Targeting preferences					
Price Pop1	100	100	100	100	100
Price Pop2	200	200	200	200	200
Price Pop3	600	600	600	600	600
Price Pop4	1600	1600	1600	1600	1600
Q Pop1	0.01	0.02	0.02	0.01	0.01
Q Pop2	0.02	0.01	0.02	0.01	0.03
Q Pop3	0.01	0.02	0.02	0.01	0.02
Q Pop4	0.02	0.01	0.02	0.05	0.01
Exploitation dynamics					
step function β_1	1	2	1	2	3
step function β_2	10	10	8	12	7
step function β_3	Q90	Q90	Q85	Q90	Q80
step function $rate$	10	20	15	25	10
Past Knowledge	Τ	${f T}$	${f T}$	Τ	${ m T}$
Past Year & Month	T	T	T	T	${f T}$
Past Trip	T	T	T	Τ	${f T}$
Threshold	0.75	0.75	0.75	0.75	0.75

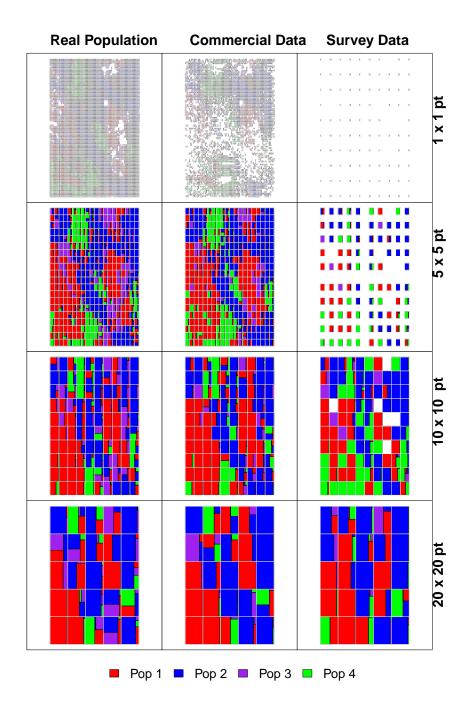


Figure 2: Data aggregation at different spatial resolutions

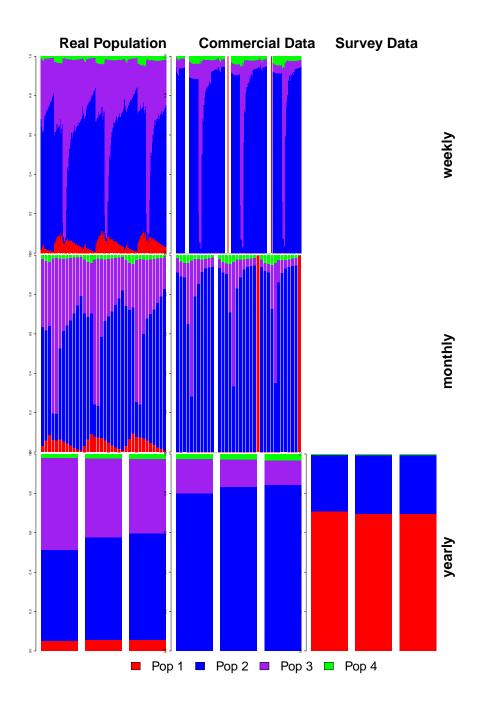


Figure 3: Data aggregation at different temporal resolutions

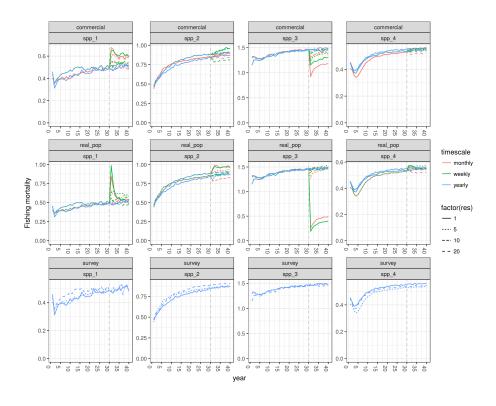


Figure 4: Comparison of closure scenarios - F trends

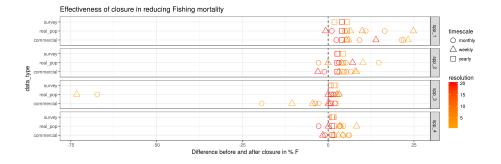


Figure 5: Comparison of closure scenarios

References

- 430 Alverson, D.L., Freeberg, M.H., Murawski, S.A., Pope, J., 1994. A global as-
- sessment of fisheries bycatch and discards.
- Bartumeus, F., Da Luz, M.G.E., Viswanatham, G.M., Catalan, J., 2005. Animal
- Search Strategies: A Quantitative Random Walk Analysis. Ecological Society
- of America 86, 3078–3087.
- Bastardie, F., Nielsen, J.R., Ulrich, C., Egekvist, J., Degel, H., 2010. De-
- tailed mapping of fishing effort and landings by coupling fishing logbooks
- with satellite-recorded vessel geo-location. Fisheries Research 106, 41–53.
- Batsleer, J., Hamon, K.G., Overzee, H.M.J., Rijnsdorp, A.D., Poos, J.J., 2015.
- High-grading and over-quota discarding in mixed fisheries. Reviews in Fish
- Biology and Fisheries 25, 715–736.
- Bellido, J.M., Santos, M.B., Pennino, M.G., Valeiras, X., Pierce, G.J., 2011.
- Fishery discards and by catch: Solutions for an ecosystem approach to fisheries
- 443 management?
- Bertrand, S., Bertrand, A., Guevara-Carrasco, R., Gerlotto, F., 2007. Scale-
- invariant movements of fishermen: The same foraging strategy as natural
- predators. Ecological Applications 17, 331–337.
- Beverton, R.J., Holt, S.J., 1957. On the Dynamics of Exploited Fish Populations
- 448 , 533.
- 449 Catchpole, T.L., Revill, A.S., 2008. Gear technology in Nephrops trawl fisheries.
- Reviews in Fish Biology and Fisheries 18, 17–31.
- ⁴⁵¹ Codling, E.A., Plank, M.J., Benhamou, S., Interface, J.R.S., 2008. Random
- walk models in biology. Journal of the Royal Society, Interface / the Royal
- society 5, 813–34.
- ⁴⁵⁴ Crowder, L.B., Murawski, S.A., 1998. Fisheries Bycatch: Implications for Man-
- agement. Fisheries 23, 8–17.

- Deriso, R.B., 1980. Harvesting Strategies and Parameter Estimation for an Age-
- Structured Model. Canadian Journal of Fisheries and Aquatic Sciences 37,
- 268-282. arXiv:1410.7455v3.
- 459 Dichmont, C.M., Punt, A.E., Deng, A., Dell, Q., Venables, W., 2003. Applica-
- tion of a weekly delay-difference model to commercial catch and effort data
- for tiger prawns in Australia 's Northern Prawn Fishery. Fisheries Research
- 462 65, 335–350.
- Diggle, P.J., Ribeiro, P.J., 2007. Model-based Geostatistics (Springer Series in
- Statistics). volume 1.
- Dinmore, T.A., Duplisea, D.E., Rackham, B.D., Maxwell, D.L., Jennings, S.,
- 2003. Impact of a large-scale area closure on patterns of fishing disturbance
- and the consequences for benthic communities. ICES Journal of Marine Sci-
- ence 60, 371–380.
- Dunn, D.C., Boustany, A.M., Roberts, J.J., Brazer, E., Sanderson, M., Gardner,
- B., Halpin, P.N., 2014. Empirical move-on rules to inform fishing strategies:
- A New England case study. Fish and Fisheries 15, 359–375.
- Dunn, D.C., Maxwell, S.M., Boustany, A.M., Halpin, P.N., 2016. Dynamic
- ocean management increases the efficiency and efficacy of fisheries manage-
- ment. Proceedings of the National Academy of Sciences, 201513626.
- 475 Edwards, A.M., 2011. Overturning conclusions of Lévy flight movement patterns
- by fishing boats and foraging animals. Ecology 92, 1247–1257.
- F. Dormann, C., M. McPherson, J., B. Araújo, M., Bivand, R., Bolliger, J.,
- Carl, G., G. Davies, R., Hirzel, A., Jetz, W., Daniel Kissling, W., Kühn, I.,
- Ohlemüller, R., R. Peres-Neto, P., Reineking, B., Schröder, B., M. Schurr,
- 480 F., Wilson, R., 2007. Methods to account for spatial autocorrelation in the
- analysis of species distributional data: A review. Ecography 30, 609–628.
- Gerritsen, H.D., Lordan, C., Minto, C., Kraak, S.B.M., 2012. Spatial patterns
- in the retained catch composition of Irish demersal otter trawlers: High-

- resolution fisheries data as a management tool. Fisheries Research 129-130,
- 485 127-136.
- Gillis, D.M., Peterman, R.M., 1998. Implications of interference among fishing
- vessels and the ideal free distribution to the interpretation of CPUE. Canadian
- Journal of Fisheries and Aquatic Sciences 55, 37–46.
- Girardin, R., Vermard, Y., Thébaud, O., Tidd, A., Marchal, P., 2015. Predicting
- fisher response to competition for space and resources in a mixed demersal
- fishery. Ocean & Coastal Management 106, 124–135.
- ⁴⁹² Hilborn, R., Walters, C., 1992. Quantitative fisheries stock assessment: Choice,
- dynamics and uncertainty. volume 2. arXiv:1011.1669v3.
- 494 Holmes, S.J., Bailey, N., Campbell, N., Catarino, R., Barratt, K., Gibb, A., Fer-
- nandes, P.G., 2011. Using fishery-dependent data to inform the development
- and operation of a co-management initiative to reduce cod mortality and cut
- discards. ICES Journal of Marine Science 68, 1679–1688.
- Hutton, T., Mardle, S., Pascoe, S., Clark, R.a., 2004. Modelling fishing location
- choice within mixed fisheries: English North Sea beam trawlers in 2000 and
- ⁵⁰⁰ 2001. ICES Journal of Marine Science 61, 1443–1452.
- Kennelly, S.J., Broadhurst, M.K., 2002. By-catch begone: Changes in the phi-
- losophy of fishing technology. Fish and Fisheries 3, 340–355.
- Lee, J., South, A.B., Jennings, S., 2010. Developing reliable, repeatable, and
- accessible methods to provide high-resolution estimates of fishing-effort distri-
- butions from vessel monitoring system (VMS) data. ICES Journal of Marine
- Science 67, 1260–1271.
- Little, A.S., Needle, C.L., Hilborn, R., Holland, D.S., Marshall, C.T., 2014.
- $_{508}$ Real-time spatial management approaches to reduce by catch and discards:
- experiences from Europe and the United States. Fish and Fisheries , n/a-
- n/a.

- Mateo, M., Pawlowski, L., Robert, M., 2016. Highly mixed fisheries: fine-scale
- spatial patterns in retained catches of French fisheries in the Celtic Sea. ICES
- Journal of Marine Science: Journal du Conseil, fsw129.
- Poos, J.J., Rijnsdorp, A.D., 2007. An "experiment" on effort allocation of fishing
- vessels: the role of interference competition and area specialization. Canadian
- Journal of Fisheries and Aquatic Sciences 64, 304–313.
- R Core Team, 2017. R Core Team (2017). R: A language and environment for
- statistical computing. R Foundation for Statistical Computing, Vienna, Aus-
- tria. URL http://www.R-project.org/., R Foundation for Statistical Com-
- 520 puting.
- Reynolds, A., 2015. Liberating Lévy walk research from the shackles of optimal
- 522 foraging.
- Ricker, W.E., 1954. Stock and recruitment. Journal of the Fisheries Research
- ₅₂₄ Board of Canada 11, 559 623.
- Rijnsdorp, A., 2000. Competitive interactions among beam trawlers exploiting
- local patches of flatfish in the North Sea. ICES Journal of Marine Science 57,
- 527 894-902.
- Rijnsdorp, a.D., Daan, N., Dekker, W., Poos, J.J., Van Densen, W.L.T., 2007.
- Sustainable use of flatfish resources: Addressing the credibility crisis in mixed
- fisheries management. Journal of Sea Research 57, 114–125.
- Rijnsdorp, A.D., Piet, G.J., Poos, J.J., 2001. Effort allocation of the Dutch
- beam trawl fleet in response to a temporarily closed area in the North Sea.
- $_{533}$ Ices Cm 2001/N: 01 , 1-17.
- 534 Sakiyama, T., Gunji, Y.P., 2013. Emergence of an optimal search strategy from
- a simple random walk. Journal of the Royal Society, Interface 10, 20130486.
- 536 Schlater, M., Malinowski, A., Menck, P.J., 2015. Analysis, Simulation and Pre-
- diction of Multivariate Random Fields with Package RandomFields. Journal
- of Statistical Software 63, 1–25. arXiv:1501.0228.

- Schnute, J., 1985. A genera theory for analysis of catch and effort data. Canadian Journal of Fisheries and Aquatic Sciences 42, 414–429.
- Sims, D.W., Humphries, N.E., Bradford, R.W., Bruce, B.D., 2012. Lévy flight
- and Brownian search patterns of a free-ranging predator reflect different prey
- field characteristics. Journal of Animal Ecology 81, 432–442.
- 544 Sims, D.W., Southall, E.J., Humphries, N.E., Hays, G.C., Bradshaw, C.J.A.,
- Pitchford, J.W., James, A., Ahmed, M.Z., Brierley, A.S., Hindell, M.A., Mor-
- ritt, D., Musyl, M.K., Righton, D., Shepard, E.L.C., Wearmouth, V.J., Wil-
- son, R.P., Witt, M.J., Metcalfe, J.D., 2008. Scaling laws of marine predator
- search behaviour. Nature 451, 1098–U5.
- Tidd, A.N., Hutton, T., Kell, L.T., Blanchard, J.L., 2012. Dynamic prediction
- of effort reallocation in mixed fisheries. Fisheries Research 125-126, 243–253.
- Tobler, W.R., 1970. A Computer Movie Simulating Urban Growth in the Detroit
- Region. Economic Geography 46, 234. arXiv:1011.1669v3.
- Ulrich, C., Reeves, S.a., Vermard, Y., Holmes, S.J., Vanhee, W., 2011. Rec-
- onciling single-species TACs in the North Sea demersal fisheries using the
- 555 Fcube mixed-fisheries advice framework. ICES Journal of Marine Science 68,
- 1535-1547.
- Viswanathan, G.M., Buldyrev, S.V., Havlin, S., Da Luz, M.G.E., Raposo, E.P.,
- Stanley, H.E., 1999. Optimizing the success of random searches. Nature 401,
- 911-914.