MixFishSim: highly resolved spatiotemporal simulations for exploring mixed fishery dynamics

Paul J. Dolder^{a,b,*}, Cóilín Minto^a, Jean-Marc Guarini^c, Jan Jaap Poos^d

 a Galway-Mayo Institute of Technology (GMIT), Dublin Road, Galway, Ireland b Centre for Environment, Fisheries and Aquaculture Science (Cefas), Pakefield Road, Lowestoft, UK

^cSorbonne Université, Faculty of Sciences Université Pierre et Marie Curie JM, 4 Place Jussieu, 75005 Paris, France

Abstract

Most fisheriesFishing^{JJ} exploits^{JJ} spatially and temporally heterogeneous fish populations, using species-unselective gear that can result in unintended, unwanted catch of low quota or protected species. Reducing these unwanted catches is crucial for biological and economic sustainability of 'mixed fisheries' and implementation of an ecosystem approach to fishing.

If fisheries are to avoid unwanted catch To implement effective spatial measures to reduce discards PD a good understanding of spatiotemporal fishery dynamics is required. However, traditional scientific advice is limited by a lack of highly resolved knowledge of population distribution, movement and how fishers interact with different fish populations. This reflects that data on fish location at high temporal and spatial resolutions is expensive and difficult to collect and therefore proxies inferred from either scientific surveys or commercial catch data are often used to model distributions, often with limited spatial and temporal resolution.

To understand how resolution impacts mixed fisheries inference, we develop

Email address: paul.dolder@gmit.ie (Paul J. Dolder)

^d Wageningen Marine Research, Haringkade 1 1976 CP IJmuiden, Netherlands

^{*}Corresponding author

a highly resolved spatiotemporal simulation model incorporating: i) delay-

difference population dynamics, ii) population movement using Gaussian Ran-

dom Fields to simulate patchy, heterogeneously distributed populations, and iii)

fishery dynamics for multiple fleet characteristics based on targeting via corre-

lated random walk movement and learned behaviour.

We simulate 20 years of exploitation of the fish populations and use the

results from the fishing model to draw inference on the underlying population

structures. We compare this inference to i) a simulated fixed-site sampling

design commonly used for fisheries monitoring purposes, and ii) the true un-

derlying population structures input to the simulation, to establish the poten-

tial and limitations of fishery-dependent data - an inherently biased sampling

method due to fisher's targeting- to provide a robust picture of spatiotemporal

distributions. Finally, we simulate an area closure based on areas defined from

commercial the known ("real-population") distribution, commercial catch data

and survey data at different temporal and spatial resolutions and assess their

effectiveness on reducing catches of a fish population.

We conclude from our simulations that commercial data, while not unbiased,

provides a useful tool for managing catches in mixed fisheries if applied at the

correct spatiotemporal scale.

[333 words]

Keywords: Some, keywords, here. Max 6

2010 MSC: 00-01, 99-00

1. Introduction

Fishers exploit fish populations that are heterogeneously distributed in space

and time with varying knowledge of species distributions using species-unselective

fishing gear. In doing soPD fisheries thatPD catch an assemblage of species

and and and and and and an anixed fisheries may, when managed by single species quotas can end up JJ discarding JJ overquota catch when managed by single species quotas, JJ leading to overexploitation of fish populations (Ulrich et al., 2011; Batsleer et al., 2015) JJ. This discarding of fish in excess of quota hampers the ability to limit fishing mortality to within sustainable limits (Alverson et al., 1994; Crowder and Murawski, 1998; Rijnsdorp et al., 2007) JJ; reducing discarding is crucial to ensure biological and economic sustainability of fisheries and implementation of an ecosystem approach to fisheries JJ and. As such pd there is increasing interest in technical solutions such as gear and spatial closures as ways of reducing unwanted catchavoiding discarding of fish JJPD (Kennelly and Broadhurst, 2002; Catchpole and Revill, 2008; Bellido et al., 2011).

Changes to spatial fishing patterns have Use of spatial management as a tools PD been proposed as a method to reduce discards (Holmes et al., 2011; Little et al., 2014; Dunn et al., 2014) PD. However, its PD implementation is hampered by lack of knowledge of fish and fishery spatiotemporal dynamics and understanding of the scale at which processes are important for management. Understanding the correct scale for spatial management is crucial in order to implement measures at a resolution that ensures effective management (Dunn et al., 2016) while minimising economic impact. For example, a scale that promotes species avoidance for vulnerable or low quota species while allowing continuance of sustainable fisheries for available quota species.

Identifying Ensuring measures are implemented at PD an appropriate scale has been a challenge in the past that has led to ineffectual measures with unintended consequences such as limited impact towards the management objective or increased benthic impact on previously unexploited areas (e.g. the cod closure in the North Sea (Rijnsdorp et al., 2001; Dinmore et al., 2003)). MSince
then mPD ore refined spatial information has since PD become available through
the combination of logbook and Vessel Monitoring System (VMS) data (Lee et al., 2010; Bastardie et al., 2010; Gerritsen et al., 2012; Mateo et al., 2016)

and more real-time spatial management has been possible (e.g. Holmes et al., 2011). Such information is, however, patchy and derived from an inherently biased sampling programme (i.e. targeted fishing). Further, fishers generally only recorded landings (not eatch) on a daily basis. This leads to questions about the validity of inference that can be drawn from landings data assigned to VMS activity pings. PD

a surprise: I
thought this
was going
to be about
discards JA Agree
have removed
this to avoid
confusion PD

In order to understand the consequences of usingehallenges that face^{PD}
VMS-linked landings to draw inference on the underlying population structure
we develop a simulation model where population dynamics are highly-resolved
in space and time. Being-and-are^{PD} known directly^{PD} rather than inferred from
sampling or commercial catches, we can use the population model to evaluate
how inference from fisheries-dependent and fisheries independent sampling relates to the real population structure^{PD}. In our model system pP^{PD}opulation
movement is driven by random (diffusive) and directed (advective) processes and
we incorporate characterisation of a number of different fisheries dynamics^{PD}
exploiting four fish populations with different spatial and population demographics.

Using our model we simulate 2040^{PD} years of exploitation of the fish populations. We and PD use the results from the fishing model: PD

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- to understand how sampling-derived data reflects the underlying population structures. We compare at different spatial and temporal aggregations of data the real population to:
 - (a) the inferred population from a stratified fixed-site sampling survey design commonly used for fisheries monitoring purposes, otherwise know as a fisheries-independent survey,
 - (b) the inferred population from our fishery-dependent model which includes fishery-induced sampling dynamics.

- 2. to understand the impact of data aggregation and source on spatial fisheries management measures we simulate a fishery closure to protect a species based on different spatial and temporal data aggregations:
 - (a) as if the real spatial population structure were known,
 - (b) the fishery-independent inferred population structure
 - (c) the fishery-dependent inferred population structure

We evaluate the theoretical "benefit" to the population of the closure(s), and effect on the other three populations. Further, we extend our analysis

to a range of spatial and temporal scales to assess the impact of these
processes on the success of the management measure. PD

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2. Materials and Methods

AWe developed and implemented a simulation model with a^{PD} modular event-based simulation model was developed with approach, where sub-^{PD} modules are ^{PD} implemented on independent time-scales appropriate to capture the characteristic of the different processes process modelled ^{PD} (Figure 1). The following sub-modules were included to capture the full system: 1) Population dynamics, 2) Recruitment dynamics, 3) Population movement, 4) fishery dynamics. ^{PD}

dynamics (fishing and natural mortality, growth) operate on a daily time-step,
while p.—P^{PD}opulation movement occurs on a weekly time-step. R, while

The fishing model operated on a tow-by-tow basis, while PD PpPD opulation

^{PD} ecruitment takes placeoceurs periodcally each year for a set time durationperiod (e.g. 3 weeks) at at specified point individual to a species. while the fish-

90 ing module operates on a tow-by-tow basis (i.e. multiple events a day)^{PD}. The

 $_{\rm 91}$ $\,$ simulation framework is implemented in the statistical software package R (R

Core Team, 2017) and; PD available as an R package from the authors github

93 (www.github.com/pdolder/MixFishSim).

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If the paper has two goals would like to keep both parts but have made clearer in how its set out. closure scenarios form validation of the data aggregation. rather than of the closures themselves its a continuation of the same question in my eyesPD

Here we describe each of the model components; 1) Population dynamics, 2)
Recruitment dynamics, 3) Population movement dynamics, 4) fishery dynamics. PD

97 2.1. Population dynamics

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The basic population level processes are simulated using a modified twostage Deriso-Schnute delay difference model (Deriso, 1980; Schnute, 1985; Dichmont et al., 2003) occurring at a daily time-step. A daily time-step was chosen as to discretise continuous population processes on a biologically relevent and computationally tractable timescale. PD Under the population dynamics module Here, PD population biomass growth and depletion for pre-recruits and fish PD recruited fish PD to the fishery PD are modelled separately as a function of previous recruited biomass, intrinsic population growth and recruitment:

$$\begin{split} B_{c,d+1} = & (1+\rho)B_{c,d} \cdot e^{-Z_{c,d}} - \rho \cdot e^{-Z_{c,d}} & \times \\ & (B_{c,d-1} \cdot e^{-Z_{c,d-1}} + Wt_{R-1} \cdot \alpha_{d-1} \cdot R_{\tilde{y}(c,y,d-1)}) & + \\ & Wt_R \cdot \alpha_d \cdot R_{\tilde{y}(c,y,d)} \end{split}$$

where ρ is Brody's coefficient, shown to be approximately equal to exp(-K),
where K is the growth rate from a von bertalanffy logistic growth model (Schnute,
1985). Wt_{R-1} is the weight of fish prior to recruitment, while Wt_R is the recruited weight. α_d represents the proportion of fish recruited during that day
for the year, while $R_{c,y}$ is the annual recruits in cell c for year y.

Mortality $Z_{c,d}$ can be decomposed to natural mortality, $M_{c,d}$, and fishing mortality, $F_{c,d}$, where both $M_{c,d}$ and $F_{c,d}$ are instantaneous rates with $M_{c,d}$ fixed and $F_{c,d}$ calculated by solving the Baranov catch equation (Hilborn and Walters, 1992) for $F_{c,d}$:

$$C_{c,d} = \frac{F_{c,d}}{F_{c,d} + M_{c,d}} * (1 - e^{-(F_{c,d} + M_{c,d})}) * B_{c,d}$$

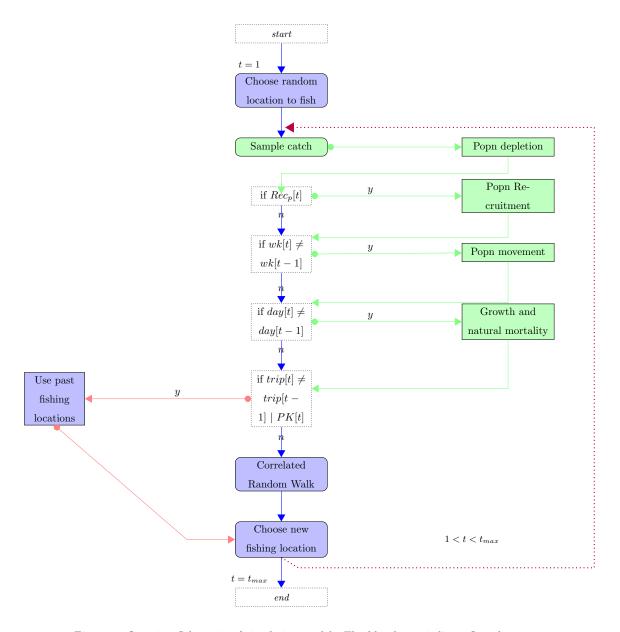


Figure 1: Overview Schematic of simulation model. The blue boxes indicate fleet dynamics processes, the green boxes population dynamics processes while the white boxes are the timesteps at which processes occur; Rec is a recruitment period for population p, t = tow, tmax is the total number of tows, wk is a weekly timestep, day is a day timestep, trip is a trip time step.

where $C_{c,d}$ is the summed catch from the fishing model across all fleets and 116 vessels in cell c for the population during the day d, and $B_{c,d}$ the daily biomass for the population in the cell.

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2.2. Recruitment dynamics

Recruitment is modelled through a function relating the mature biomass to recruits at time of recruitment. In mixfishsim, it can be modelled either either as a stochastic Beverton-Holt stock-recruit form (Beverton and Holt, 1957):

$$\bar{R}_{c,d} = \frac{(\alpha * B_{c,d})}{(\beta + B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

Where α is the maximum recruitment rate, β the spawning stock biomass (SSB) 124 required to produce half the maximum, B current SSB and σ^2 the variability 125 in the recruitment due to stochastic processes. 126

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or a stochastic Ricker form (Ricker, 1954):

$$\bar{R}_{c,d} = B_{c,d} * e^{(\alpha - \beta * B_{c,d})}$$

$$R_{c,d} \sim \log N[(\log(\bar{R}_{c,d}), \log(\sigma^2))]$$

where α is the maximum productivity per spawner and β the density depen-129 dent reduction in productivity as the SSB increases. 130

2.3. Population movement dynamics

time, we employed JJ a Gaussian spatial process was employed JJ to model habi-133 tat suitability for each of the populations. An, with an JJ advection-diffusion 134 process to JJ controlled JJ how the JJ populations JJ movement d JJ over time with a moving temperature covariate to capture temporal dependencies. This was 136

To simulate how^{JJ} fish populations might be^{JJ} distributioned^{JJ} in space and

 $[eally]^{CM}$

time^{JJ}Was intended as some biological mean ing - species thermal tolerances load onto the temperature intended to balance realism in population movement, capturing the main directed and random processes, and practicality of modelling the population rather than individual fish. ^{JJ}

For the PD habitat we defined PD a Gaussian random field process, $\{S(c): c \in \mathbb{R}^2\}$, that is a stochastic process PD where for PD any set of cells c_1, \ldots, c_n where for each $c_i \in \mathbb{R}^{2\text{PD}}$, the joint distribution of $S = \{S(c1), \ldots S(c_n)\}$ is multivariate Gaussian. The distribution is specified by its mean function, $\mu(c) = E[S(c)]$ and its covariance function, $\gamma(c, c') = Cov\{S(c), S(c')\}$ (Diggle and Ribeiro, 2007).

The covariance structure affects the smoothness of the surfaces which the process generates;, and PD we used the $Mat\acute{e}rn$ family of PD covariance structures PD, as one where PD the correlation strength weakens the further the distance apart (i.e. the correlation between S(x) and S(x') decreases as the distance u = ||x - x'|| increases) PD. The Mat\acute{e}rn covariance structure models the spatial autocorrelation observed with animal distributions (Tobler, 1970; F. Dormann et al., 2007) PD and The $Mat\acute{e}rn$ correlation PD is a two-parameter family where:

$$\rho(u) = \{2^{\kappa - 1} \Gamma \kappa\}^{-1} (u/\phi)^{\kappa} K_{\kappa}(u/\phi)$$

 $K_{\kappa}(.)$ is a modified Bessel function of order κ , $\phi > 0$ is a scale parameter with the dimensions of distance, and $\kappa > 0$, called the order, is a shape parameter which determines the smoothness of the underlying process.

TIn the simulation model, t^{PD}he habitat for each of the populations wasis^{PD} generated withthrough^{PD} the *RFSimulate* function of the *RandomFields* R package (Schlater et al., 2015), implementing different parameter settings to affect the patchiness of the populations. Each population wasis^{PD} initialised at a single location, and subsequently moveeds^{PD} according to a probabilistic distribu^{PD}tion based on habitat suitability, temperature and distance from current

Not clear how habitat/GRF affect local abundances, only have $B_{y,d}^{\text{JJ}}$ Have included cell reference, c to make spatial link explicit PD

Introduce the gamma function, and why this covariance structure? Why correlate values in the random field? JM to allow populations to have different aggregation densities have tried to clarify PD

67 cell:-PD

$$Pr(J|I) = \frac{e^{-\lambda * d_{IJ}} \cdot (Hab_{J,p}^2 \cdot Tol_{J,p,wk})}{\sum\limits_{c=1}^{C} e^{-\lambda * d} \cdot (Hab_{c,p}^2 \cdot Tol_{c,p,wk})}$$
(1)

Where d_{IJ} is the euclidean distance between cell I and cell J, λ is a given rate of decay, $Hab_{J,p}^2$ is the squared index of habitat suitability for cell J and population p, with $Tol_{J,p,wk}$ the temperature tolerance for cell J by population p in week wk.

During specified weeks of the year, the habitat quality wasis^{PD} modified for user-defined^{PD} spawning habitats^{PD}, resulting inmeaning^{PD} each population hadhas^{PD} a concentrated area where spawning takes place and the population moveds^{PD} towards these cellsthis^{PD} in the weeks prior to spawning.

The temperature field was is PD simulated to be on a gradient from a South-Westerly to North-Easterly direction, with temperature in each cell changing gradually on a week-by-week basis so that initially high temperature areas cycled PD to lower temperatures and low temperature areas vice versa. Each population p wai PDs assigned a thermal tolerance with mean, μ_p and variance, σ_p^2 so that each cell and population temperature suitability is defined that: it mean concisely? Areas are assigned? In Yes, the areas are predefined - I have amended to reflect and tried to clarify PD

$$Tol_{c,p,wkPD} = \frac{1}{\sqrt{(2\pi \cdot \sigma_p^2)}} \cdot \exp\left(-\frac{(T_{c,wkPD} - \mu_p)^2}{2 \cdot \sigma_p^2}\right)$$
(2)

Where $Tol_{c,p,wk}^{\text{PD}}$ is the tolerance of population p in cell c in week wk^{PD} , $T_{c,wk}^{\text{PD}}$ is the temperature in the cell given the week $^{\text{PD}}$ and μ_p^{PD} and σ_p^{PD} the mean and standard deviation of the population temperature tolerance.

The final process resulted in independent populations structure and movement patterns, with population movement occuring on a weekly basis. This process approximated the demographic shifts in fish populations throughout a year while maintaining seasonal patterns for spawning.^{PD}

2.4. Fleet dynamics

The fleet dynamics can be broadly categorised into three components; fleet targeting - which determineds^{PD} the fleet catch efficiency and preference towards a particular species; trip-level decisions, which determined^{PD} the initial location to be fished at the beginning of a trip; and within-trip decisions, determining movement from one fishing spot to another within a trip.

2.4.1. Fleet targeting

Each fleet of n vessels wasis^{PD} characterised by both a general efficiency, Q, and a population specific efficiency, Q_p . Thus, the product of these parameters affectsed^{PD} the overall catch rates for the fleet and the preferential targeting of one population over another. This, in combination with the parameter choice for the step-function defined below^{PD} (as well as some randomness from the exploratory fishing process) determineds^{PD} the preference of fishing locations for the fleet. All species prices wereare^{PD} kept the same, across fleetsand seasons, though can be made to vary seasonally^{PD}.

2.4.2. Trip-level decisions

Several studies (e.g. Hutton et al., 2004; Tidd et al., 2012; Girardin et al., 208 2015) have confirmed past activity and past catch rates are strong predictors 200 of fishing location choice. For this reason, the fleet dynamics sub-model in-210 cludeds^{PD} a learning component, where a vessel's initial fishing location in a 211 trip wai^{PD}s based on selecting from previously successful fishing locations. This 212 wai^{PD}s achieved by calculating expected profit from locations fished during^{PD} previous fishing events in the previous trip as well as the previous time periods 214 in past years, and choosing randomly from the top 75 % of fishing events as de-215 fined by the expected profitin value PD. Expected profit was estimated from the 216 revenue from previous times fished at a location minus the fuel cost of travelling 217 to the location. PD Simulation testing indicated that this learning increased the mean value of catches for the vessels, over just relying on the correlated random 219 walk function as described for the 'within trip' decisions below^{PD}. 220

dom walk of what JJ

2.4.3. Within-trip decisions

Fishing locations within a trip are initially PD determined by a modified random walk process. As the simulation progresses, the within-trip decision become gradually more influenced by past locations fished, based on the same process as the initial location, influenced by expected profit at a fishing location. PD A random walk was chosen for the exploratory fishing process as it is the simplest assumption commonly used in ecology to describe optimal PD animal movement which PD search strategying PD for exploiting PD homogeneously distributed prey about which there is uncertain knowledge (Viswanathan et al., 1999). In a random walk, movement is a stochastic process through a series of steps. These steps have a length, and a direction JJ that can either be equal in length or take some other functional form. The direction of the random walk can be correlated, (known as 'persistence'), providing some overall location of PD directional movement (Codling et al., 2008) or uncorrelated PD.

A Lévy flightlévy walk^{JJ} is a particular form of random walk characterised by a heavy-tailed distribution of step-length. The Lévy flightand^{JJ} has received a lot of attention in ecological theory in recent years as having shown to have very similar characteristics as those observed by animals in nature, and being a near optimum searching strategy for predators pursuing patchily distributed prey (Viswanathan et al., 1999; Bartumeus et al., 2005; Sims et al., 2008). Bertrand et al. (2007) showed that Peruvian anchovy fishermen have a stochastic search pattern similar to that observed with a lévy walk. However, it remains a subject of debate (e.g. see Edwards, 2011; Reynolds, 2015)^{PD}, with the contention that search patterns may be more simply characteristed as random walks (Sakiyama and Gunji, 2013) with specific patterns related to the characteristics of the prey field (Sims et al., 2012).

We use a modified random walk where directional change is based on a correlated circular distribution where a favourable fishing ground is likely to be "fished back over" by the vessel returning in the direction it came from and step length (i.e. the distance travelled from the current to the next fishing location) is determined by relating $^{\rm JJ}$ recent fishing success, measured as the summed value of fish caught (revenue, Rev),

$$Rev = \sum_{p=1}^{P} \underline{L}\underline{C}_{p}^{PD} \cdot Pr_{p}$$

where $L\underline{C}^{\mathrm{PD}}_{p}$ is landingseatch of a population p, and Pr_{p} price of a population, to step distance. Here, when fishing is successful vessels remain in a similar location and continue to exploit the local fishing grounds. When unsuccessful, they move some distance away from the current fishing location. The movement distance retains some degree of stochasticity, which can be controlled separately.

The step function takes the form:

$$StepL = e^{log(\beta_1) + log(\beta_2) - (log(\frac{\beta_1}{\beta_3}))} * Rev$$

So that, a step from (x1,y1) to (x2, y2) is defined by:

 $(x2, y2) = x1 + StepL \cdot \cos(\frac{\pi \cdot Br}{180}),$ $y1 + StepL \cdot \sin(\frac{\pi \cdot Br}{180})$ $with \quad Br_{t-1} < 180, Br_t = 180 + \sim vm[(0, 360), k]$

with k the concentration parameter from the von Mm^{JJ}ises distribution which we correlate with the revenue so that $k = (Rev + 1/RefRev) * max_k$, where max_k is the maximum concentration value, k, and RefRev is parameterised as for β_3 in the step length function.

 $Br_{t-1} > 180, Br_t = 180 - \sim vm[(0, 360), k]$

2.4.4. Local population depletion

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Where several fishing vessels are exploiting the same fish population competition is known to play an important role in local distribution of fishing effort (Gillis and Peterman, 1998). If several vessels are fishing on the same patch of fish, local depletion and interference competition JJ will affect fishing location choice of the fleet as a whole (Rijnsdorp, 2000; Poos and Rijnsdorp, 2007). In order to account for this behaviour, the fishing sub-model operates spatially on a daily time-step so that for future days the biomass available to the fishery is reduced in the areas fished. The cumulative effect is to make heavily fished areas less attractive as future fishing opportunities.

2.5. Fisheries independent survey

A fisheries-independent survey is simulated where fishing on a regular grid
begins each year at the same time for a given number of stations (a fixed station
survey design). Catches of the populations at each station present are recorded
but not removed from the population. This provides a fishery independent
snapshot of the populations at a regular spatial intervals distribution year, similar to scientific surveys undertaken by fisheries research agencies.

278 3. Calculation

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3.1. Population parameterisation

We parameterised the simulation model for four populations with differing habitat preference, and temperature tolerances (Figures S1, S3, S4, S5, S6, S7), population demographic, and recruitment functions. In addition, each of the populations has two defined spawning areas which result in the populations moving towards these areas in given weeks (Figure S2) and population-specific movement rates (Table 4). The realised movement of the populations for a number of weeks is shown in Figure S9 while the realised daily fishing mortality are shown in Figure S10.

3.2. Fleet parameterisation

The fleets were parameterised to reflect five different characteristics based on targeting preference and exploitation dynamics (Table 5). Setting a tar-

geting parameter (Q) that differed across fleets ensured different spatial dynamics, due to preferential targeting of populations that differ in their spatial
distributions This ensures that different fleets have different spatial dynamics,
preferentially targeted different fish populations PD. The stochasticity in the
random walk process ensures that different vessels within a fleet have slightly
different spatial distributions based on individual experience, while the step
function was parameterised dynamically so that vessels take smaller steps where
the fishing location yields in a top quartile of the value available in that year
(as defined per fleet in Table 5).

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Each fleet was set so that, after the first year, fishing locations were chosen based on experience built up in the same month from previous years and from past trip fishing success. 'Success' in this context was defined as the locations where the top 75 % of revenue from was found in previous trips.

An example of the realised fleet movements for a single vessel during a single trip are given in Figure S11, while Figure S12 shows multiple trips for a single vessel, Figure^{PD} S13 the vessel movements for some trips overlaid on the value field, Figure^{PD} S14 shows fishing locations for an entire fleet of 20 vessels for a single trip, and Figurewhile^{PD} S15 shows an example of the step function realisation and turning angles from the correlated random walk.

3.3. Survey settings

The survey simulation was set up with follow a fixed gridded station design with 100 stations fished each year, starting on day 92 and ending on day 112 (5 stations per day)^{PD} with same catchability parameters for all populations $(Q_p = 1)$.

3.4. Simulation settings

To illustrate the capabilities on *MixFishSim*, we investigate the influence of the temporal and spatial resolution of different data sources on the reduction in catches of a population given spatial closures. To do so, we first set up

Move some of the supplementary figures to the manuscript JJ with simulation to run for 10 years based on a 100 X 100 square grid, with five fleets of 20 vessels each and four fish populations. Fishing takes place four times a day per vessel and five days a week, while population movement is every week.

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We allow the simulation to run unrestricted for 5 years, and subsequently close areas for the last 5 years of the simulation based on data (either derived from the commercial catches, fisheries-independent survey or the 'real population' - the underlying populations assumed to be known perfectly) used at different spatial and temporal scales.

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The following steps are undertaken to determine closures:

- 1. Extract data source
- 2. Aggregate according to resolution
- 3. Interpolate across entire area at desired resoltion
- 4. Close top 5 % of areas

In total 56 closure scenarios were run which represent combinations of

- data types: commercial logbook data, survey data and 'real population',
- temporal resolutions: weekly, monthly and yearly closures,
- spatial resolutions: 1 x 1 grid, 5 x 5 grid, 10 x 10 grid and 20 x 20 grid.

Survey closures were on an annual basis only, as this was the most temporally resolved survey data available.

4. Results

The species distribution themselves

The consequences of different spatial aggregations of the data are shown in Figure 2, which represents the aggregation of catch from each of the data sources over a year at different spatial resolutions.

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move to start
of methods
section JJ I think
ecological modelling wants
the 'calculations' section here..will
check PD

Is there equilibrium after
5 years or still some trend in
B^{JJ}Not at equilibrium yet...I
need to rerun
until steady
state, looks
20 years. Will
update^{PD}

Procedure unclear. Refer to symbols in methods section or switch order starting with description of data type etc...^{JJ}Yes, will redo^{PD}

The finer spatial grid for the the real population (top left) and commercial data (top middle) show similar patterns, though there are unsampled gaps in the commercial data from a lack of fishing activity (particularly in the lower left part of the sampling domain). The survey data at this spatial resolution shows very sparse and uninformative information about the spatial distributions of the populations. The slightly aggregated data on a 5 x 5 grid shows similar patterns, and while losing some of the spatial detail there remains good consistency between the 'real population' and the commercial data. Survey data starts to pick out some of the similar patterns as the other data sources, but lacks coverage. The spatial catch information on a 10 x 10 and 20 x 20 grid loses a significant amount of information about the spatial resolutions for all data sources, and some differences between the commercial and 'real population' data emerge.

Figure 3 shows the consequences of different temporal aggregations of the data, with 156 weekly (top), 36 monthly (middle) and 3 yearly (bottom) catch compositions across a 20×20 area.

As can be seen from the 'real population', the monthly aggregation captures the major patterns seen in the weekly data, albeit missing more subtle differences. The yearly data results in a constant catch pattern due to the aggregation process (sometimes known as an aggregation bias). The commercial data on a weekly basis shows some of the same patterns as the 'real population', though the first species (in red) is less well represented and some weeks are missing catches from the area. The monthly data. The monthly data shows some consistency between the 'real population' and commercial data for species 2 - 4, though species 1 remains underrepresented. On an annual basis, interestingly the commercial data underrepresents the first species (in red) while the survey overrepresents species 1. This is likely due to the biases in commercial sampling, with the fisheries not targeting the areas where species 1 are present, and the biases in the survey sampling from overrepresentation of the spatial distribution.

We implemented a spatial closure using the different data sources and spatial 379 and temporal aggregations as outlined in the protocol in Section 3.4. We used 380 this to assess the efficacy of a closure in reducing fishing mortality on species 1, given availability of data and its use at different resolutions in order to evaluate 382 the trade-offs in data sources. Figure 4 shows the trend in fishing mortality 383 for each species simulated (columns) given the data sources (rows), temporal 384 aggregations (colour lines) and spatial aggregations (linestyles), while Figure 5 385 shows the change in fishing mortality from before the closure (average F years 2 - 4) to after the closure (average F years 8 - 10). 387

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For the closures based on 'real population' (bottom row), the most disag-389 gregated data (a weekly timescale and 1 x 1 resolution) was most effective, reducing fishing mortality on species 1 (left) by ~ 60 %. Next was the monthly closures ($< \sim 30$ %). The least effective were the yearly closures (blue lines) at all spatial resolutions, which resulted in increased fishing mortalities (> 30 393 % - N.B. Note though, this is consistent with the increasing trends in F, which 394 is probably more related to the fact that Fs hadn't stabilised in the simulation 395 from the fishing vessels "learning" the best locations - I will rerun the sims for a longer time (20 - 30 years).

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For the survey data, which can only be implemented on a yearly timescale, the closures had no effect at any data resolution. The results are identical for the different data resolutions except 20 x 20, which is why you can't see more than 2 points. This is because of the sparsity of the sampling locations.

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For the commercial data, the most effective closure scenario was based on 1 404 x 1 data at a monthly temporal resolution. This results in ~ 10 % reduction 405 in F for species 1. This was the only closure scenario to have positive effect according to Figure 5, though looking at the trend in Figure 4 this looks more 407 related to the continued increased in F trend, as other scenarios had an initial 408 effect. Interestingly the monthly data scenario was more effective than weekly data, which I'd posit is due to the increase amount of data available from the commercial sampling across a month compared to a week.i Commercial data used at an annual timestep was ineffective in bringing fishing mortality down for species 1.

414

Given the scenarios above, it seems clear that spatial disaggregation is more important than the temporal disaggregation of the commercial data, except when its used at an annual timeframe, which is the scenario that gave the worst results.

For the other species in the simulation (population 2 - 4) there was little difference in fishing mortalities across scenarios.

Note: The monthly commercial data scenario is the most effective of the realistic scenarios, as the 'real population' can only be seen as a baseline comparison.

5. Discussion

425 6. Conclusions

426 Appendices

427 Abbreviations

Detail any unusual ones used.

429 Acknowledgements

those providing help during the research..

431 Funding

This work was supported by the MARES doctoral training program; and the
Centre for Environment, Fisheries and Aquaculture Science seedcorn program.

Table 1.	Description	of wariables	for population	dynamics of	h modulo
rable i:	Describtion	or variables	тог роршалоп	ovnamics si	то-тоате

Variable	Meaning	Units		
Population dynamics				
Delay-difference model				
$B_{c,d}$	Biomass in cell c and day d	kg		
$Z_{c,d}$	Total mortality in cell c for day d	-		
$R_{c,\tilde{y}}$	Annualy recruited fish in cell	$\mathrm{yr}^{\text{-}1}$		
ho	Brody's growth coefficient	$\mathrm{yr}^{\text{-}1}$		
Wt_R	Weight of a fully recruited fish	kg		
Wt_{R-1}	Weight of a pre-recruit fish	kg		
$lpha_d$	Proportion of annually recruited fish recruited during	-		
	$\mathrm{day}\ d$			
Baranov co	atch equation			
$C_{c,d}$	Catch from cell c for day d	kg		
$F_{c,d}$	Instantaneous rate of fishing mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$M_{c,d}$	Instantaneous rate of natural mortality in cell \boldsymbol{c} on	-		
	$\mathrm{day}\ d$			
$B_{c,d}$	Biomass in cell c on day d	kg		
Recruitment dynamics				
$\tilde{R}_{c,d}$	is the recruitment in cell c for day d	d^{-1}		
$B_{c,d}$	is the Biomass in cell c for day d	d^{-1}		
α	the maximum recruitment rate	kg		
β	the biomass required to produce half the maximum	kg		
	rate of recruitment			

Table 2: Description of variables for population movement sub-module						
Variable	Meaning	Units				
	Population movement dynamics					
Habitat me	odel					
a	b	С				
Thermal to	lerance					
$T_{c,wk}$	Temperature for cell in week	$^{\circ}\mathrm{C}$				
μ_p	Mean of the thermal tolerance for population	$^{\circ}\mathrm{C}$				
σ_p^2	Standard deviation of thermal tolerance for the pop-	$^{\circ}\mathrm{C}$				
	ulation					
Population	Population movement model					
λ	decay rate for population movement	-				
$Hab_{c,p}^2$	Square of habitat suitability for cell \boldsymbol{c} and population	-				
	p					
$Tol_{c,p,wk}$	Thermal tolerance for population p in cell c at week	-				
	wk					
d_{IJ}	euclidean distance between cell ${\cal I}$ and cell ${\cal J}$	-				

Table 3: Description of variables for fleet dynamics sub-module

Variable	Meaning	Units			
Short-term fleet dynamics					
Rev	Revenue from fishing tow	€			
L_p	Landings of population p	kg			
Pr_p	Average price of population p	$\in \ \mathrm{kg}^{-1}$			
StepL	Step length for vessel	euclidean			
		distance			
Br	Bearing	degrees			
k	Concentration parameter for Von mises distribution	-			
eta_1	shape parameter for step function	-			
eta_2	shape parameter for step function	-			
β_3	shape parameter for step function	-			

Table 4: Population dynamics and movement parameter setting

Parameter	Pop 1	Pop 2	Pop 3	Pop 4	
Habitat quality					
Matérn ν	1/0.15	1/0.05	1/0.55	1/0.05	
Matérn κ	1	2	1	1	
Anisotropy	1.5,3,-3,4	1,2,-1,2	2.5,1,-1,2	0.1,2,-1,0.2	
Spawning areas (bound	40,50,40,50;	50,60,30,40;	30,34,10,20;	50,55,80,85;	
box)	80,90,60,70	80,90,90,90	60,70,20,30	30,40,30,40	
Spawning multiplier	10	10	10	10	
Movement λ	0.3	0.3	0.3	0.3	
Population dynamics					
Starting Biomass	1e5	2e5	1e5	1e4	
Beverton-Holt Recruit 'a'	60	100	80	2	
Beverton-Holt Recruit 'b'	250	250	200	50	
Beverton-Holt Recruit σ^2	0.4	0.3	0.4	0.3	
Recruit week	13-16	12-16	14-16	16-20	
Spawn week	16-18	16-19	16-18	18-20	
K	0.3	0.3	0.3	0.3	
wt	1	1	1	1	
wt_{d-1}	0.1	0.1	0.1	0.1	
M (annual)	0.2	0.2	0.2	0.1	

Table 5: Fleet dynamics parameter setting					
Parameter	Fleet	Fleet	Fleet	Fleet	Fleet
	1	2	3	4	5
Targeting preferences					
Price Pop1	100	100	100	100	100
Price Pop2	200	200	200	200	200
Price Pop3	600	600	600	600	600
Price Pop4	1600	1600	1600	1600	1600
Q Pop1	0.01	0.02	0.02	0.01	0.01
Q Pop2	0.02	0.01	0.02	0.01	0.03
Q Pop3	0.01	0.02	0.02	0.01	0.02
Q Pop4	0.02	0.01	0.02	0.05	0.01
Exploitation dynamics					
step function β_1	1	2	1	2	3
step function β_2	10	10	8	12	7
step function β_3	Q90	Q90	Q85	Q90	Q80
step function $rate$	10	20	15	25	10
Past Knowledge	Τ	${f T}$	${f T}$	Τ	${ m T}$
Past Year & Month	T	T	T	T	${f T}$
Past Trip	T	T	T	Τ	${f T}$
Threshold	0.75	0.75	0.75	0.75	0.75

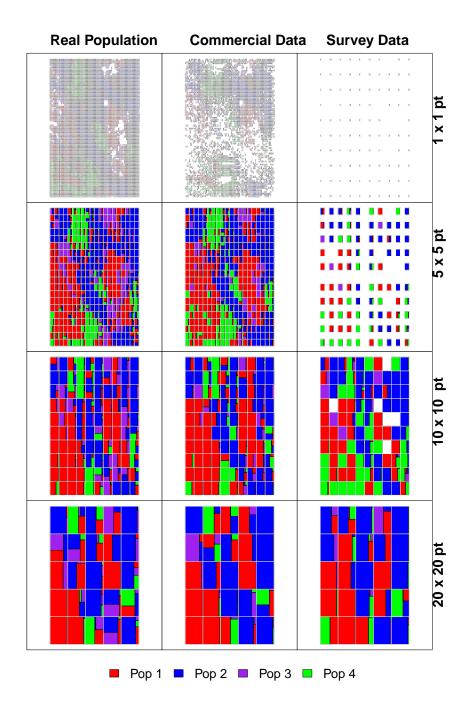


Figure 2: Data aggregation at different spatial resolutions

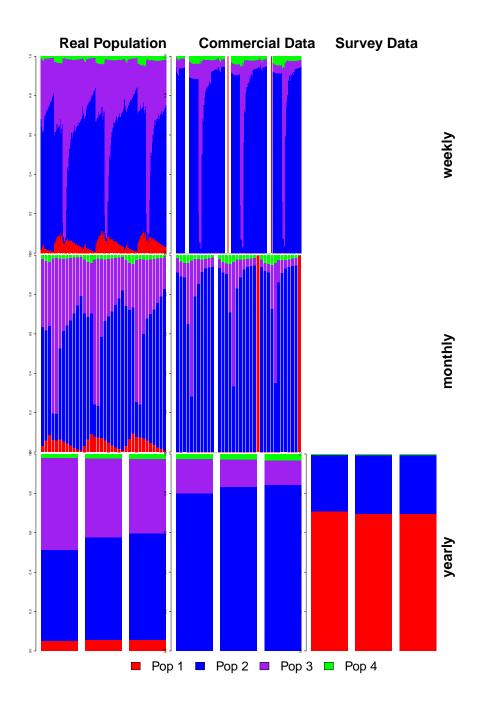


Figure 3: Data aggregation at different temporal resolutions

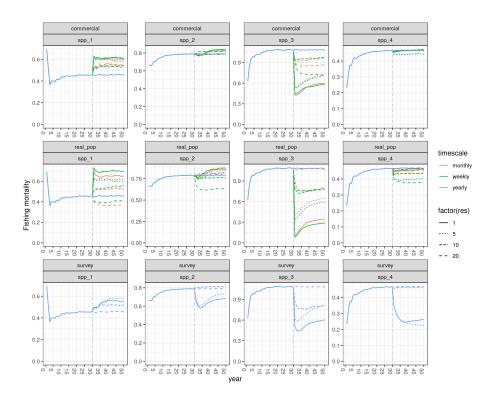


Figure 4: Comparison of closure scenarios - F trends

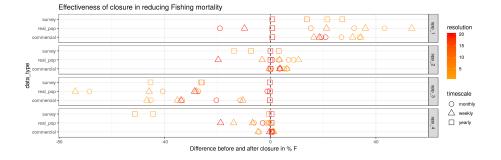


Figure 5: Comparison of closure scenarios

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