# Estimating Population Trends with Stratified Random

# <sup>2</sup> Sampling Under the Pressures of Climate Change

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- 9 [[Chris comments in double square brackets, search for them to see comments]]

- 10 Abstract
- 11 An Abstract
- 12 Keywords
- 13 keyword 1, keyword 2

## 14 Introduction

18

27

much of below is from https://apps-nefsc.fisheries.noaa.gov/nefsc/ecosystem-ecology/
or https://www.fisheries.noaa.gov/data-tools/fisheries-economics-united-states-data-andvisualizations

• The eastern continental shelf is ecologically diverse and economically important

The Northeast United States continental shelf spans from the Outer Banks of North Carolina to the Gulf of Maine. The region covers over 250,000 km<sup>2</sup> of ocean, extending over 200 km from shore in the largest areas in New England to just 30 km off shore in the southern regions. This ecologically diverse region contains approximately 18,000 vertebrate marine species. Commercial fisheries have been an important part of local economies for centuries. In 2019, New England fisheries produced \$22 billion in sales, which sustained over 200,000 jobs. Maintaining a healthy ecosystem is therefore vital to sustained ecological health and economic prosperity of the region.

• Bottom trawl survey is important for monitoring population trends

Fish stocks in this highly productive and economically important region are managed by the
National Oceanic and Atmospheric Administration's (NOAA) Northeast Fisheries Science
Center (NEFSC) in Woods Hole, Massachusetts. Federal scientists assess the health and
abundance of each commercial fish stock using fishery-independent bottom trawl survey
data that has been collected by NOAA throughout the region since 1963 (Azarovitz 1981;
Politis et al. 2014). The survey uses a stratified random design where bottom trawl sampling
takes place in predefined strata along the eastern continental shelf. The survey has created
a rich time series data set with many uses including species-specific habitat identification,
analysis of how environmental conditions influence species abundance, and estimating yearly
species abundance trends to help inform stock assessments and ultimately quota limits

The survey takes place twice each year- once in the spring and again in the fall. Since most spatial analyses and projections of future distributions typically assume a constant survey catchability and/or availability over time, NOAA's survey design includes sampling during approximately the same 2-3 week time period in each season. [[the survey for the entire region takes ~8 weeks to complete each season, the GB strata get covered in a similar time period each season because the survey starts in the south and moves north. But there are times when logistics, e.g., storms, Covid, cause a change in the timing of any particular stratum being sampled. This timing issue has come up in the past and we've looked at it using the dates of sampled stations within a particular stock definition. See for example GBYT\_survey\_timing\_cdf.png taken from the 2018 assessment of Georges Bank yellowtail flounder that I copied into a newly created "miscellaneous" folder on the Google Drive. Given all this, I suggest modifying the wording to something like "NOAA's survey design includes sampling Georges Bank during approximately the same 3-4 week time period in each season"]

#### • Climate change is happening

Due to a combination of climate change and shifts in circulation, the Northeast United States continental shelf has experienced rapid warming in recent decades, resulting in a shift in spatial distributions of many species. Since stock assessment models rely on accurate descriptions of population dynamics and contemporary patterns of spatial abundance, there is concern that rapid undocumented changes in spatial distributions of species will bias future stock assessments. More specifically, as fish populations shift their distributions over time, catchability and/or availability in the survey will change, altering the relationship between the index and the true population. We are therefore interested in analyzing the impact of climate change on the accuracy of future stock assessment models as measured by NOAA's ongoing bottom-trawl survey along the East coast.

#### use more info from initial proposal

• Briefly describe our study to test this

To test the ability of the bottom trawl survey to track population trends under shifting environmental conditions, we construct spatial models for fish where movement depend on temperature preferences. We can then consider the impact of climate change by simulating scenarios with repeating temperature patterns and those where temperature increases on average over time. In both cases we analyze the ability of stratified random sampling to track population trends.

## $_{\scriptscriptstyle 1}$ Methods

• Describe simulation study

We construct spatial models for Yellowtail Flounder, Atlantic Cod, and Haddock on Georges
Bank, where movement of each species combine static species-specific habitat preferences
with temperature preferences. Model dynamics are driven by a time series of temperature
gradients that were estimated from data to create simulated data sets for each population
where the true biomass is known. Using temperature gradients that repeat each year creates
data sets with predictable, repeating spatial patterns, whereas using a temperature gradient
that increases on average throughout the simulation leads to spatial distributions that shift
over time, approximating the future distribution in the region under climate change. We
conducting stratified random sampling on our simulation output to mimic the bottom trawl
survey and compare the ability of contemporary indexing methods to track population trends.

# 83 Population Model Formulation

84 – Used MixFishSim. Describe edits made to package

We use the R package MixFishSim (MFS) to model our populations (Dolder et al. 2020).

MFS is a discrete spatiotemporal simulation tool where users can model multiple species

under varying environmental conditions. The package uses a delay-difference population

model with discrete processes for growth, death, and recruitment of the population. We for-

mulate the following inputs for the MFS package to address our research question. [[perhaps

of for discussion, but we should mention somewhere that MFS only tracks biomass not age

structure, the latter is important for stock assessments but not considered in this paper.

92 Study Area

We obtained a shapefile for the 15 strata that comprise Georges Bank, where strata were partitioned based primarily on depth and secondarily by latitude (Politis et al. 2014). The region was discritized into a raster with 88 rows and 144 columns to use as our modeling environment. A fish stock is considered to be a subpopulation of a species that has similar intrinsic parameters. Each of the species being modeled has multiple distinct stocks along the Atlantic coast resulting from local environmental conditions. Biological differences be-

tween species results in each stock inhabiting a different number of strata on Georges Bank.

Haddock inhabit all 15 strata in the domain, Cod inhabit 13 strata, and Yellowtail exist in

9 strata. Figure 1 shows the regions used in our models.

102 Population Dynamics and Recruitment

The time step for our models is one week. MFS uses a modified two-stage Deriso-Schnute 103 delay difference equation that models the biomass in each cell in our study area (Dolder et 104 al. 2020). Individual terms in the formulation account for growth of mature adults, natural 105 and fishing mortality, and the addition of new recruits. [[not sure we need to highlight the 106 recruitment model, might be able to just use the table because Bev-Holt formulation is well 107 known in fisheries]BL: MFS has 2 recruitment options and I edited the package 108 to use previous years biomass We chose to represent recruitment in the model using a 109 Beverton-Holt formulation cite. Recruitment is a function of the adult biomass that existed 110

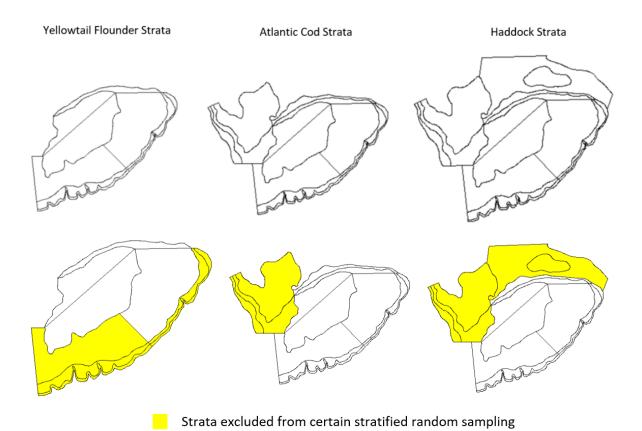


Figure 1: Strata inhabited by each species in our population models.

in the previous year and is added to the population incrementally throughout each species'
predefined spawning period. Parameter inputs were either obtained from the literature or
chosen to produce desired model dynamics. A full list of parameters used in our model can
be seen below in Tables ?? and 4.

115 Movement

The package was designed to generate theoretical habitat preferences using Gaussian Random Fields that combine with hypothetical temperature gradients to drive the probability
of movement from cell I to cell J using the formulation

$$Pr(C_{wk+1} = J | C_{wk} = I) = \frac{e^{-\lambda \cdot d_{I,J}} \cdot (Hab_{J,s}^2 \cdot Tol_{J,s,wk})}{\sum_{c=1}^{C} e^{-\lambda \cdot d} \cdot (Hab_{c,s}^2 \cdot Tol_{c,s,wk})}, \tag{1}$$

119 where

 $e^{-\lambda \cdot d_{I,J}}$  accounts for distance between cells I and J,

 $_{121}$   $Hab_{J,s}^2$  is the static habitat value for species s in cell J, and

 $Tol_{c,s,wk}$  is the value from normally distributed temperature tolerance for species s in cell c in week wk.

The following sections describe how we formulated the habitat and temperature components to model real species on the northeast continental shelf.

126 Habitat Input

Species-specific habitat preferences were derived using the *lrren* tool from the R package *envi*(Buller 2022) to create a niche model for each species. The *lrren* tool estimates an ecological
niche using the relative risk function by relating presence/absence data to two covariate
predictors. We used bottom trawl point data in from 2009-2021 as our presence/absence
input by using a value of 0 for any tow that failed to catch the given species and weighting

a successful catch by the biomass of the given tow [I don't think we need to cite the trawl data because we said we were using trawl data above]] BL: Does trawl data ever get 133 cited? It was not cited above. cite trawl data?. We combined data from both the 134 fall and spring surveys to obscure the influence of temperature so that the niche model 135 would instead infer habitat preferences. Depth and mean sediment size were used as our 136 covariate predictors. Estimated depth for the region was obtained from FVCOM (Chen et 137 al. 2006). The mean sediment size raster was interpolated in ArcMap using the natural 138 neighbor interpolation method **cite arcmap** using point data collected by the United States 139 Geologic Survey (USGS) (McMullen et al. n.d.). Since the values in  $Hab_{J,s}^2$  are required 140 to be between 0 and 1, we transform the spatial estimates from *lrren* to fall between these 141 bounds. See Figure 2 for a visual representation of this process being applied to Cod. Figure 142 3 depicts habitat preferences  $Hab_{J,s}^2$  for each species. 143

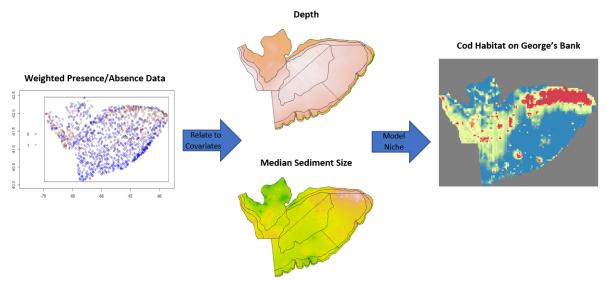


Figure 2: Visual representation of niche model for Cod.

#### 144 Temperature Input

Each species is assumed to have normally distributed temperature preferences  $(N(\mu, \sigma))$ .

We assume Yellowtail Flounder's preferences are N(8.75, 4.25), while Haddock and Cod

have preferences N(9,4). We chose these values by combining information in the literature

with temperatures recorded in the bottom trawl survey. Weekly estimated temperature data

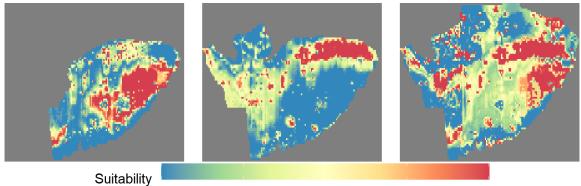


Figure 3: Static habitat preferences for each species in our population models (Yellotwtail, Cod, Haddock).

for the region for 2012 was obtained from FVCOM (Chen et al. 2006). The 2012 data was repeated yearly in the constant temperature scenario because displayed an average tem-150 perature pattern that consistently oscillated between maximum and minimum temperature 151 values. We chose to repeat estimates for a single year rather than use data for consecutive 152 years to reduce the number of factors impacting model dynamics while still incorporating 153 real data. The 2012 temperature data was also transform to create an oscillating pattern 154 that increases 5 degrees Celsius on average over the duration of the simulation. We chose a 5 155 degree increase over a 20 year simulation to allow temperature change to have a meaningful 156 impact on dynamics while remaining within reasonable computational limits in terms of the 157 length of the simulation. Figure 4 depicts mean trends for the temperature scenarios used 158 in our models. dont forget to include gif in final submission 159

60 —Describe difference between increasing and constant temperature scenarios (images?)

In equation (1),  $Hab_{J,s}^2$  is constant for the duration of the simulation, while  $Tol_{c,s,wk}$  changes each week. Using a temperature gradient that repeats every 52 weeks produces the same spatial preferences in a given week each year, resulting in consistent spatial biomass patterns. Scenarios where the temperature increases over time creates spatial preferences that evolve as the water warms, producing spatial biomass patterns that shift in a given week over the duration of the simulation. Thus, stratified random samples in scenarios with a repeating temperature pattern will have constant survey catchability and availability over time, which

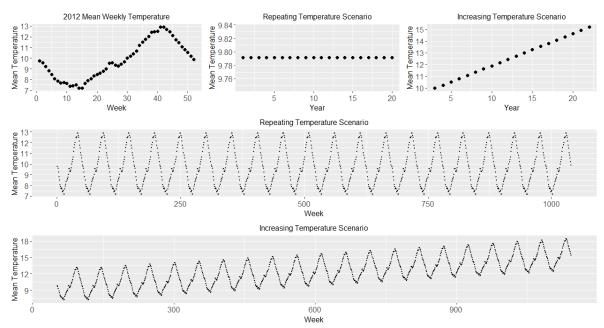


Figure 4: Mean trends of temperature data used in our model.

may not be true for increasing temperature scenarios due to evolving spatial preferences.

[[BL: included in above paragraph may want to mention here or in the discussion that
we used repeating 2012 to simplify the simulations and reduce the number of factors explored
that would have occurred with different annual cycles]

- Describe each scenario that is considered

We consider 20 year simulations under three population parameter scenarios for each of our 173 three species- a scenario where parameters result in each population increasing over time, one 174 where the populations are relatively constant over time, and a scenario where the parameter 175 combination results in each population decreasing over time. Each of these three scenarios 176 is paired with a temperature gradient that repeats as well as one that increasing roughly 5 177 degrees Celsius [BL: included in above paragraph need some justification for this large 178 change, wanted to be extreme because if the extreme doesn't make a difference, then more 179 reasonable increases wouldn't make a difference] over the duration of the 20 year simulation. 180 We therefore simulate 6 scenarios for each population. 181

Simulating Bottom Trawl Survey and Population Indexing

-Describe post hoc sampling process and how data is used

After each simulation is complete, we mimic the bottom trawl survey by conducting stratified 184 random sampling in each inhabited strata twice each year. We sample each strata in the same 185 weeks in which the Spring and Fall surveys take place and the number of the samples taken 186 reflect true target values for each strata. Most strata contain enough cells that we were able 187 to generate a list of sampling locations without replacement, allowing us to sample a unique 188 random cell in each survey over the duration of the simulation. For strata that contain 189 small numbers of cells, we generate a random sampling locations with replacement of cells 190 previously sampled. We then use the biomass collected from our samples in contemporary 191 population indexing methods to estimate population trends. Knowing the true population 192 values in our simulations allows us to compare the error calculated from each estimation 193 method. BL: added both to above paragraph [should we mention here or later that we did both with and without measurement error analyses?]] [[need to make sure that we don't 195 lose sight of random sampling with the following sentences. 196

-Stratified mean vs VAST with and without covariates [[could use an intro sentence here to distinguish design-based (stratified mean) and model-based (VAST) estimators]] [[may want to highlight the potential of VAST with covariates to address the climate change issue here as the reason we are doing these simulations]]

Stock assessment scientists use samples from the bottom trawl survey to estimate the abundance dance for each stock stock. There are a number of approaches to obtaining abundance estimates ranging from traditional design-based estimates such as the stratified mean to model-based estimates that range in complexity. The stratified mean calculates the geometric mean catch per tow and does not account for spatial variation in samples or allow the inclusion of environmental covariates. Model-based abundance estimates are often derived from model framewroks [[frameworks]] such as General Linear Models (GLM), General Additive Model (GAM), and General Linear Mixed Models (GLMM) that can incorporate both

random and mixed effects. The Vector Autoregressive Spatio-Temporal (VAST) model is a spatial delta-generalized linear mixed model (Thorson 2019). VAST is a spatio-temporal 210 statistical framework that models both abundance (biomass) and probability of occurrence 211 (presence/absence). If desired, VAST also allows users to include covariate data to better 212 inform the model. Covariates can be static (eg. habitat preferences) or dynamics (eg. tem-213 perature). We explore whether including environmental predictors can help inform models 214 and provide better abundance estimates, which is particularly relevant as climate change pro-215 gresses. The stratified mean calculations are straightforward and quick, while VAST models 216 require numerous user inputs and [[can?]] take on the order of hours to complete. We 217 compare the yearly estimated of abundance obtained from the stratified mean to estimates 218 obtained from the Vector-Autoregressive Spatio-temporal (VAST) model. 219

We follow the advice given in (Thorson 2019) to build VAST models to estimate biomass 220 on Georges Bank using stratified mean samples from our model output. In addition to exploring different link functions and assumed distributions, our VAST model-building pro-222 cess included testing the impact of including spatial and/or spatio-temporal variation in our models, considering varying number of knots in our mesh, and testing different forms of temporal correlation. We also carried out the same process both including covariates in our 225 model as well as running models without covariate information. We considered covariates in 226 the form of dynamic temperature values and/or static habitat values from our population 227 model. When using covariates we ultimately decided to provide the most information to 228 the model by including both covariates for both linear predictors. Since we know the true 229 population values in our models we calculate the absolute error of each VAST estimate to 230 compare between potential settings. Through this process, and in consultation with the 231 VAST package creator, we determined setting that allowed VAST models to converge for all 232 of our scenarios while also providing the lowest absolute error values. Settings for our VAST 233 models can be seen in Table???. 234

Our goal is to determine indexing approaches and settings that are robust to future environ-

mental conditions and resulting spatial biomass patterns. An underlying assumption in all indexing methods is that individual random samples combine to accurately represent true 237 abundance by a) containing a low enough noise level in the samples to allow for a discernible 238 pattern and b) sampling all strata in which the population exists. These assumptions can be 239 questioned given enough noise in the sampling process cite? and/or climate change causing 240 a population to move into previously uninhabited strata. To simulate the impact of noise, 241 we compare indexing estimates after adding noise to our samples versus those using the 242 true sampling values. BL: Help with correct notation for adding noise. We simulate 243 the effect of populations moving into new habitat by comparing indexing estimates using 244 samples from all strata inhabited by each species on Georges bank [Bank] to those that 245 only include a subset of the full spatial domain for each species. We chose strata to exclude 246 for each species by reviewing how spatial preferences evolved in our increasing temperature 247 scenarios and removing strata that each species either shifted into, or away from. Tabl XXX 248 [[needs to be added]] lists all strata inhabited by each species, those that are removed from 249 certain calculations, and the explanation of why these strata were removed. 250

BL: I tried to provid emore detail above and also added another table [[may need to spell the aspect of reduced spatial domain a bit more because it might be confusing to readers who are expected us to just add areas around the current ones instead of reducing the strata]

When combining population trends for each species, differing temperature scenarios, altering seasons, and sampling possibilities (noise, strata, covariates) there are a large number of scenario combinations to consider. The columns in Table 1 show the choices that define each scenario.

Table 1: Each index estimate chooses one condition from each of the following 7 columns. There are 3\*3\*2\*2\*2\*2\*2 = 288 VAST model combinations and 3\*3\*2\*2\*2\*2 = 144 stratified mean estimates.

Species	Population Trend	Temperature Scenario	Strata Included	Noise Added	Season	Covariates (VAST)
Yellowtail Cod Haddock	Increasing Constant Decreasing	Repeating Increasing 5°C	All strata Subset	No Yes	Spring Fall	No Yes

## Results

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The goal of our project was to analyze how well contemporary population indexing methods 260 can track population trends under a variety of conditions, as depicted in Table 1 Histori-261 cally, Atlantic Cod has seen significant decline over the last XXX years while Haddock has 262 increased in abundance in recent year [can cite the 2022 management track assessments, see 263 https://apps-nefsc.fisheries.noaa.gov/saw/sari.php for when the document becomes avail-264 able cite. For this reason we compare indexing estimates using stratified random samples 265 from decreasing population scenarios for Cod (see Table???) and increasing population 266 scenarios for Haddock (see Table???). To provide a comprehensive analysis of population 267 indexing methods we consider all possible scenario combinations for Yellowtail Flounder (see 268 Table ???). The specific population trends used in our analyses can be see in Figure 5 269 General themes that exist in Tables X, Y and Z are that VAST estimates provide lower errors relative to those derived from the stratified mean, with VAST models that include covariate 271 information providing the lowest overall errors. We also see individual cases where the 272 stratified mean produced the lowest absolute error and instances where including covarites 273 in VAST models actually increase the absolute error. When we reduce the number of strata 274 that are included in indexing calculations to simulation species shifting into new territory, 275 we typically see an increase in absolute error (as expected), though there are some scenarios 276 where the impact is minimal. 277 In the Cod results, when using reduced strata adding covariates produces worst VAST results

(though still better than stratified mean). VAST without covariates much worse than strat-

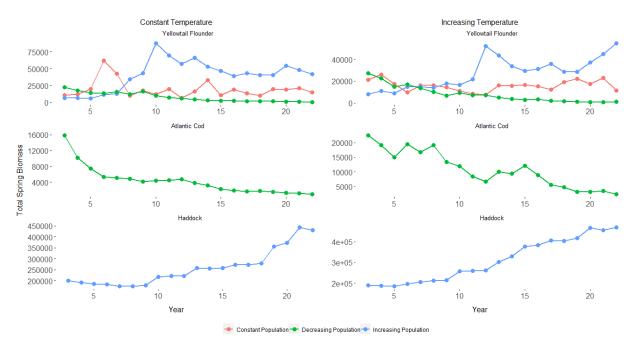


Figure 5: True population trends used in indexing analyses. Spring biomass plots are shown with fall values being very similar.

280 ified mean in fall with increasing temperature and all strata, but adding covariates corrects
281 this.

For Haddock, VAST has a particularly hard time in spring regularly producing larger errors
than the stratified mean with added covariates only improving to the level of the stratified
mean. VAST shows improved results in fall relative to the stratified mean with added
covariates producing extremely low errors in some cases.

In considering the Yellowtail Flounder results in Tables 6 and 7, we can see VAST estimates
generally provide lower errors relative to those derived from the stratified mean, with models
that include covariate information typically providing the lowest errors. However, there are
several instances in which VAST failed to provide improved abundance estimates during
the fall season without covariate information, producing the largest errors seen in the Table.
These errors are corrected by including covariate information allowing for an improved VAST
estimate that are significantly lower than their stratified mean counterparts.

Notes on differences between VAST settings: For YTF,  $29/96 \sim 30\%$  of VAST runs with

new settings were better (70% were worse).  $80/96 \sim 83\%$  of scenarios had a VAST run with a better error than the stratified mean. 15/16 of the times the stratified mean was better 295 VAST was not using covariates. The 1 time that VAST was using covariates and was still 296 worse than the stratified mean was IncPop IncTemp Allstrata WCov WNoise in the Fall 297 (VAST had strong overestimate in the fall during IncPOP\_IncTemP\_allstrata for Had and 298 YT). The overestimate in the fall only seems to be related to the way the population shifts 290 between season. Looking at the percent shift plots, the population is shifting out of strata 300 16 in both seasons (large east strata). In the spring the population is shifting mostly into 301 strata 17 (thin strata adjacent to 16), but in the fall they are shifting into both 17 and 18 302 (18 thin one adjacent to 17). Thus it seems like with much of the population concentrated in 303 the really small outer most strata, VAST produces an underestimate, even with covariates 304 For Haddock (IncPop),  $18/32 \sim 57\%$  of VAST runs better with new settings. 24/32 = 75%305 of scenarios had a VAST run with a better error than the stratified mean. 14/16 VAST 306 were better with covariates (the two that were worse were essentially the same) while only 307 10/16 were better without covariates. Stratified mean tended to perform better when all strata were included in the calculation, while VAST tended to perform better when strata 309 are removed. 310

For Cod (DecPop), only  $11/32 \sim 34\%$  of VAST runs were better with new settings.  $28/32 \sim$ 311 88% of scenarios had a VAST run with a better estimate than the stratified mean. All 16/16 312 = 100% of VAST runs with a covariate were better than the stratified mean. 12/16 VAST 313 were better without covariates. The 4 runs without covariates that were worse were all 314 extremely large errors and all from increasing temp in the fall (with and without all strata 315 and noise). In the fall the population is moving out of strata 16 (large eastern strata) and 21 316 (north large) and into a number of strata (22 (tiny north), 24 (large north), 18 (tiny north)). 317 Might be having trouble tracking in to smaller strata again. 318

[We'll want to expand the results section to focus on the impact of each factor one at at

time within each species. This will be a bit dull, but it is important to walk through the results in words to ensure that readers get the message. They can of course examine the tables in detail and draw their own conclusions, but we should put our interpretation down on paper.]

## Discussion

Estimation methods that have large, inaccurate swings (stratified mean) can lead to changes 325 in quotas that do not correspond to the true population trend, which could have a com-326 pounding effect (can lead to quotas that are too high/low given an incorrect assumption of 327 increase/decrease in biomass). Our model has a constant assumed mortality that accounts for fishing and natural death and will not account for impacts of these decisions. VAST may provide more consistent biomass predictions(?) 330 [ some things that I think we'll need to address, in addition to those mentioned above are 331 (in no particular order): number of simulations done for each scenario - time required to do 332 them and analyses pros and cons of using VAST with covariates in a changing environment 333 implications for current practice of using stratified mean (not horrible, but might be able 334 to do better with model-based using covariates if enviro changing, note we don't want to 335 say stratified mean should not be used because that will draw a lot of tomato throwing) 336 future work - some of the ideas we've discussed already could be mentioned limitations to 337 our study perhaps for discussion, but we should mention somewhere that MFS only tracks 338 biomass not age structure, the latter is important for stock assessments but not considered 339 in this paper]]

# ${f Acknowledgements}$

- [we should thank Jim (obviously), but also those who helped you with the habitat data
- 343 (robyn, david Chev.) and others? also should note that funding provided by Climate-
- Groundfish source (I'll dig up the official name of the funding source)

# Data and Code Availability

- All data and code used in this work are available at https://github.com/Blevy2/READ-
- PDB-blevy2-MFS2.

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367 Tables

Table 3: Parameters used in all population models.

Parameter	Description	Unit	Yellowtail	$\operatorname{Cod}$	Haddock	Source
<u+03c1> M W1 W2 sigma</u+03c1>	Ford's growth coefficient Natural mortality Weight of fully recruited fish Weight of pre-recruit fish Variance in recruited fish	1/wk 1/wk kg kg kg*kg	4.48 0.2064 0.39 0.13 0.55	4.43 0.2728 2.95 0.39 0.55	4.49 0.334 1.12 0.19 0.55	
lambda spwn rec	Decay rate for movement Spawning weeks for species s Recruitment weeks for species s	- wks wks	0.7 9-12 9-12	0.7 8-13 8-13	0.7 11-14 11-14	

Table 1. Parameters used in all population models.

Parameter	Description	Unit	Yellowtail	Cod	Haddock	Source
ρ	Ford's growth coefficient	${\rm wk}^{-1}$	4.48	4.43	4.49	
M	Natural Mortality	$\rm wk^{-1}$	0.2064	0.2728	0.3340	
$W_R$	Weight of fully recruited fish	kg	0.39	2.95	1.12	
$W_{R-1}$	Weight of pre-recruit fish	kg	0.13	0.39	0.19	
$\sigma^2$	Variance in recruited fish	$kg^2$	0.55	0.55	0.55	
$\lambda$	Decay rate for movement	-	0.7	0.7	0.7	
$Spwn_s$	Spawning weeks for species $s$	wk	9-12	8-13	11-14	
$Rec_s$	Recruitment weeks for species $s$	wk	9-12	8-13	11-14	

Table XX. Parameters used in all VAST models.

Parameter	Description	Input/Value
ObsModel	Link function and assumed distribution	c(10,2)
FieldCOnfig	Specified spatial and/or spatio-temporal	c(Omega1=0, Epsilon1=0,
	variation in predictors	Omega2=1, Epsilon2=1)
RhoConfig	Specifying whether intercepts or	c(Beta1=3, Beta2=3,
	spatio-temporal variation is structured among	Epsilon1=0, Epsilon2=4)
	time intervals	

Parameter	Description	Input/Value
X1_formula	Right-sided formula affecting the 1st linear	$X1_{formula} = \sim$
	predictor	poly(Temp, degree=2)
X2_formula	Right-sided formula affecting the 2nd linear	$X2$ _formula = $\sim$
	predictor	poly(Temp, degree= $2$ ) +
		$\operatorname{poly}(\operatorname{Habitat},\operatorname{degree}{=}2\ )$

Table 4: Parameters used in population models for each scenario.

Parameter	Description	Unit	Yellowtail	Cod	Haddock
Constant Population					
M+F	Adjusted Mortality (Natural + Fishing)	1/wk	0.764	0.83	0.309
P0	Initial Biomass	kg	3190	21500	180000
a	Max recruitment rate	kg	30400	27900	73600
В	Recruitment half saturation value	kg	4300	10500	40500
Decreasing Population					
M+F	Adjusted Mortality (Natural + Fishing)	1/wk	0.764	0.623	0.334
P0	Initial Biomass	kg	50000	21500	180000
a	Max recruitment rate	kg	1.07e + 12	3.89e + 08	4.97e + 08
В	Recruitment half saturation value	kg	$2.3e{+}12$	9.8e + 08	2.08e + 09
Increasing Population					
M+F	Adjusted Mortality (Natural + Fishing)	1/wk	0.564	0.372	0.134
P0	Initial Biomass	kg	3190	21500	180000
a	Max recruitment rate	kg	40000	45000	1e + 05
В	Recruitment half saturation value	kg	43000	62800	405000

Table 6: Yellowtail flounder error results with all strata included in calculations. Row colors correspond to the same settings applied in different seasons.

Temperature Scenario Constant Population	Season	Covariate	Noise	VAST A	VAST B	Stratified Mean
Constant	spring	no cov	no	0.13	0.11	0.21
Constant	spring	no cov	yes	0.14	0.16	0.25
Constant	spring	w/ cov	no	0.07	0.07	0.21
Constant	spring	w/ cov	yes	0.08	0.08	0.25
Constant	fall	no cov	no	0.63	0.68	0.32
Constant	fall	no cov	yes	0.80	0.77	0.31
Constant	fall	w/ cov	no	0.14	0.08	0.32
Constant	fall	w/ cov	yes	0.17	0.11	0.31
Increasing	spring	no cov	no	0.14	0.11	0.28
Increasing	spring	no cov	yes	0.18	0.15	0.28
Increasing	spring	w/ cov	no	0.05	0.06	0.28
Increasing	spring	w/ cov	yes	0.10	0.12	0.28
Increasing	fall	no cov	no	1.46	1.26	0.51
Increasing	fall	no cov	yes	1.40	1.38	0.50
Increasing	fall	w/ cov	no	0.21	0.23	0.51
Increasing	fall	w/ cov	yes	0.30	0.28	0.50
Decreasing Population		,		3733		
Constant	spring	no cov	no	0.11	0.08	0.23
Constant	spring	no cov	yes	0.12	0.11	0.27
Constant	spring	w/ cov	no	0.07	0.06	0.23
Constant	spring	w/ cov	yes	0.11	0.07	0.27
Constant	fall	no cov	no	0.97	0.81	0.41
Constant	fall	no cov	yes	0.99	1.09	0.37
Constant	fall	w/ cov	no	0.16	0.08	0.41
Constant	fall	w/ cov	yes	0.29	0.18	0.37
Increasing	spring	no cov	no	0.17	0.15	0.22
Increasing	spring	no cov	yes	0.15	0.17	0.26
Increasing	spring	w/ cov	no	0.08	0.07	0.22
Increasing	spring	w/ cov	yes	0.16	0.10	0.26
Increasing	fall	no cov	no	1.17	1.06	0.28
Increasing	fall	no cov	yes	1.14	1.10	0.25
Increasing	fall	w/ cov	no	0.40	$\begin{array}{c} 1.10 \\ \hline 0.15 \end{array}$	0.28
Increasing	fall	w/ cov	yes	0.40	$0.15 \\ \hline 0.20$	0.25
Increasing Population	lan	W/ COV	yes	0.40	0.20	0.20
Constant	spring	no cov	no	0.46	0.13	0.16
Constant	spring	no cov	yes	0.43	0.15 $0.21$	0.22
Constant	spring	w/ cov	no	0.45	0.21	0.16
Constant	spring	w/ cov	yes	0.08	0.07	0.22
Constant	fall	no cov	no	0.40	0.36	0.34
Constant	fall			0.40 $0.38$	0.30	0.46
Constant	fall	no cov	yes	0.38	0.44	0.34
Constant	fall	w/ cov	no			0.46
		no cov	yes	$\begin{array}{c} 0.24 \\ \hline 0.16 \end{array}$	$\begin{array}{c} 0.17 \\ \hline 0.13 \end{array}$	0.32
Increasing	spring		no	0.10	0.16	0.32 $0.32$
Increasing	spring	no cov	yes			
Increasing	spring	w/ cov	no	0.06	0.07	0.32
Increasing	spring	w/ cov	yes	0.12	0.10	0.32
Increasing	fall	no cov	no	0.71	0.66	0.30
Increasing	fall	no cov	yes	1.03	0.71	0.39
Increasing	fall	w/ cov	no	0.43	0.21	0.30
	fall	w/ cov	yes			0.30

Table 7: Yellowtail flounder error results with certain strata excluded from calculations. Row colors correspond to the same settings applied in different seasons.

Temperature Scenario	Season	Covariate	Noise	VAST A	VAST B	Stratified Mea
Constant Population	•	1	l.		II.	I.
Constant	spring	no cov	no	0.24	0.19	0.27
Constant	spring	no cov	yes	0.16	0.15	0.22
Constant	spring	w/ cov	no	0.19	0.19	0.27
Constant	spring	w/ cov	yes	0.12	0.14	0.22
Constant	fall	no cov	no	0.30	0.25	0.47
Constant	fall	no cov	yes	0.78	0.36	0.44
Constant	fall	w/ cov	no	0.22	0.19	0.47
Constant	fall	w/ cov	yes	0.24	0.17	0.44
Increasing	spring	no cov	no	0.23	0.17	0.31
Increasing	spring	no cov	yes	0.25	0.19	0.29
Increasing	spring	w/ cov	no	0.17	0.17	0.31
Increasing	spring	w/ cov	yes	0.62	0.15	0.29
Increasing	fall	no cov	no	2.22	0.75	0.64
Increasing	fall	no cov	yes	1.75	0.89	0.59
Increasing	fall	w/ cov	no	0.24	0.20	0.64
Increasing	fall	w/ cov	yes	0.59	0.13	0.59
Decreasing Population						
Constant	$\mathbf{spring}$	no cov	no	0.31	0.19	0.25
Constant	spring	no cov	yes	0.27	0.16	0.27
Constant	spring	w/ cov	no	0.19	0.19	0.25
Constant	spring	w/ cov	yes	0.15	0.16	0.27
Constant	fall	no cov	no	0.53	0.24	0.55
Constant	fall	no cov	yes	0.53	0.36	0.53
Constant	fall	w/ cov	no	0.18	0.23	0.55
Constant	fall	w/ cov	yes	0.16	0.24	0.53
Increasing	spring	no cov	no	0.18	0.14	0.32
Increasing	spring	no cov	yes	0.37	0.15	0.29
Increasing	spring	w/ cov	no	0.21	0.22	0.32
Increasing	spring	w/ cov	yes	0.16	0.21	0.29
Increasing	fall	no cov	no	0.90	0.60	0.54
Increasing	fall	no cov	yes	0.84	0.62	0.48
Increasing	fall	w/ cov	no	0.32	0.31	0.54
Increasing	fall	w/ cov	yes	0.36	0.32	0.48
ncreasing Population						
Constant	spring	no cov	no	0.22	0.15	0.20
Constant	spring	no cov	yes	0.19	0.11	0.22
Constant	spring	w/ cov	no	0.17	0.17	0.20
Constant	spring	w/ cov	yes	0.11	0.13	0.22
Constant	fall	no cov	no	0.19	0.11	0.41
Constant	fall	no cov	yes	0.26	0.21	0.46
Constant	fall	w/ cov	no	0.21	0.22	0.41
Constant	fall	w/ cov	yes	0.17	0.19	0.46
Increasing	spring	no cov	no	0.31	0.33	0.40
Increasing	spring	no cov	yes	0.30	0.26	0.38
Increasing	spring	w/ cov	no	0.30	0.30	0.40
Increasing	spring	w/ cov	yes	0.31	0.25	0.38
Increasing	fall	no cov	no	0.56	0.49	0.70
Increasing	fall	no cov	yes	0.58	0.48	0.69
Increasing	fall	w/ cov	no	0.48	0.53	0.70
Increasing	fall	w/ cov	yes	0.47	0.50	0.69

Table 8: Cod error results.

Strata	Noise	Season	VAST NC A	VAST NC B	VAST WC A	VAST WC B	Stratified Mean
Constant Temp.							
all	no	spring	0.11	0.11	0.13	0.12	0.36
all	yes	spring	0.14	0.12	0.09	0.15	0.35
all	no	fall	0.23	0.19	0.09	0.05	0.49
all	yes	fall	0.34	0.30	0.20	0.23	0.41
reduced	no	spring	0.25	0.17	0.22	0.24	0.41
reduced	yes	spring	0.25	0.20	0.14	0.23	0.46
reduced	no	fall	0.16	0.21	0.26	0.33	0.60
reduced	yes	fall	0.16	0.18	0.26	0.31	0.58
Increasing Temp.							
all	no	spring	0.12	0.12	0.16	0.15	0.25
all	yes	spring	0.16	0.19	0.23	0.19	0.27
all	no	fall	0.86	0.76	0.47	0.13	0.45
all	yes	fall	1.13	0.89	0.55	0.33	0.44
reduced	no	spring	0.29	0.26	0.22	0.21	0.34
reduced	yes	spring	0.32	0.19	0.19	0.11	0.33
reduced	no	fall	1.41	0.79	0.37	0.26	0.62
reduced	yes	fall	2.09	1.37	0.40	0.26	0.57

Table 9: Haddock error results.

Strata	Noise	Season	VAST NC A	VAST NC B	$VAST\ WC\ A$	VAST WC B	Stratified Mean
Constant Temp.							
all	no	spring	0.45	0.49	0.13	0.18	0.18
all	yes	spring	0.55	0.73	0.18	0.43	0.14
all	no	fall	0.31	0.28	0.05	0.05	0.26
all	yes	fall	0.45	0.41	0.15	0.06	0.27
reduced	no	spring	0.34	0.34	0.30	0.35	0.45
reduced	yes	spring	0.31	0.30	0.45	0.33	0.44
reduced	no	fall	0.34	0.36	0.46	0.48	0.54
reduced	yes	fall	0.29	0.33	0.41	0.46	0.50
Increasing Temp.							
all	no	spring	0.28	0.25	0.11	0.05	0.26
all	yes	spring	0.35	0.30	0.11	0.06	0.31
all	no	fall	0.82	0.89	0.23	0.23	0.40
all	yes	fall	1.01	1.04	0.29	0.35	0.39
reduced	no	spring	0.35	0.32	0.41	0.40	0.44
reduced	yes	spring	0.33	0.38	0.39	0.37	0.36
reduced	no	fall	0.48	0.44	0.61	0.64	0.72
reduced	yes	fall	0.49	0.42	0.60	0.62	0.70