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Human facilitated metapopulation dynamics in an emerging pest species, *Cimex lectularius*

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1 HUMAN FACILITATED METAPOPOPULATION DYNAMICS IN AN EMERGING PEST SPECIES, *CIMEX*
2 *LECTULARIUS*

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19 Running title: Metapopulation dynamics of *Cimex lectularius*

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23

24 **ABSTRACT**

25

26 The number and demographic history of colonists can have dramatic consequences
27 for the way genetic diversity is distributed and maintained in a metapopulation. Here, using
28 21 newly characterised microsatellite markers and Approximate Bayesian Computation
29 (ABC) in a metapopulation framework we estimate the number and genetic composition of
30 founders in a re-emerging pest species, the common bed bug (*Cimex lectularius*). The bed
31 bug's close association with humans has led to metapopulation dynamics of frequent local
32 extinction and colonisation. Pest control limits the lifespan of sub-populations, causing
33 frequent local extinctions, and human-facilitated dispersal allows the colonisation of empty
34 patches. Founder events often result in drastic reductions in diversity and an increased
35 influence of genetic drift. Coupled with restricted migration this can lead to rapid population
36 differentiation. We therefore predicted strong population structuring. We found very limited
37 diversity within-infestations but high degrees of structuring across the city of London, with
38 extreme levels of genetic differentiation between infestations ($F_{ST} = 0.59$). ABC results
39 suggest a common origin of founders and that the numbers of colonists were low; implying
40 that even a single mated female is enough to found a new infestation successfully. These
41 patterns of colonisation are close to the predictions of the propagule pool model, where all
42 founders originate from the same parental infestation. These results not only formalise the
43 applicability of a metapopulation approach with this system, but also provide a valuable
44 resource for the future targeted control of bed bug infestations.

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46

47 **INTRODUCTION**

48

49 With increased travel and connectivity there are now many opportunities for the
50 human-facilitated dispersal of organisms, with disease vectors and pests of economic
51 importance presenting a particular concern (Estoup *et al.* 2004; Grapputo *et al.* 2005; Tatem
52 *et al.* 2006; Niggemann *et al.* 2009; Lawson Handley *et al.* 2011). Passive dispersal, coupled
53 with high population turnover, can lead to organisms existing as highly structured
54 metapopulations (De Meester *et al.* 2002; Haag *et al.* 2005; Walser & Haag 2012) with
55 discrete breeding groups, frequent and independent local extinctions, and potential for
56 patches to be recolonised (see Hanski 1999).

57

58 The population dynamics of metapopulations greatly affect how genetic diversity is
59 distributed, both within and between local populations (Slatkin 1977; Wade & McCauley
60 1988; Hastings and Harrison 1994). Founder events can shift populations out of equilibrium
61 by reducing genetic diversity and increasing genetic drift. Coupled with restricted migration
62 this can lead to rapid genetic differentiation of sub-populations (Wade & McCauley 1988;
63 Whitlock & McCauley 1990). Frequent local extinction events can make differentiation even
64 more extreme by limiting the ability of gene flow to equalise allele frequencies (Whitlock &
65 McCauley 1990; Hastings & Harrison 1994; Pannell & Charlesworth 1999; Ray 2001).

66

67 Theoretical studies predict that the number and origin of colonists of each sub-population
68 has a strong influence on the degree of sub-population differentiation. When colonists of
69 any given sub-population have originated from a unique parental population (propagule

70 pool model), differentiation between sub-populations is always predicted to increase
71 relative to the equilibrium with no extinction. This is in tandem with a reduction in neutral
72 genetic variation both within sub-populations and throughout the entire metapopulation
73 (Wade & McCauley 1988; Pannell & Charlesworth 1999). In contrast, a mixed origin of
74 colonists (migrant pool model) may lead to an increase or decrease in differentiation (Wade
75 & McCauley 1988). The numbers of founders and their relatedness to each other therefore
76 directly influence the genetic variance among sub-populations, with kin-structured
77 colonisation and subsequent inbreeding leading to substantial differentiation (Whitlock
78 1992; Ingvarsson & Olsson 1997; Ingvarsson 1998; Johannesen & Lubin 1999; Torimaru *et al.*
79 2007).

80

81 Metapopulation dynamics are not only important for understanding many evolutionary
82 processes but also have practical implications for conservation (Frankham *et al.* 2002; Ray
83 2001), control of invasive species (Lawson Handley *et al.* 2011) and integrated pest
84 management (Collins *et al.* 2000; Rinkevich *et al.* 2007; Yakob *et al.* 2008). For example,
85 selection for resistance alleles could reduce the chance of a sub-population going extinct
86 after a control treatment, thus increasing the period during which it acts as a source of
87 migrants. This would facilitate the spread of resistance alleles into new sub-populations and
88 severely hamper control. Knowledge of the nature of population structure would help
89 predict the spread of resistance alleles (Churcher *et al.* 2008).

90

91 The common bed bug (*Cimex lectularius*) is re-emerging as a significant economic and public
92 health pest, precipitated by a sudden global resurgence in its populations (Boase 2001;

93 Doggett *et al.* 2004; Kilpinen *et al.* 2008; Potter *et al.* 2008; Richards *et al.* 2009). The causes
94 of this sudden population expansion have remained a mystery whose resolution has been
95 hampered by a lack of research on the bed bug's basic population and dispersal biology
96 (Reinhardt & Siva-Jothy 2007). Bed bugs are flightless, obligate blood-sucking insects that
97 can form infestations comprising thousands of individuals (Wang *et al.* 2010; Reinhardt *et al.*
98 2010). Infestations typically consist of aggregations of individuals located in discrete refugia.
99 Bed bugs walk to the host and return to a refuge when feeding is complete (Reinhardt &
100 Siva-Jothy 2007). Being flightless, individuals can only move actively over limited distances
101 and much of their recent spread is attributed to long-distance passive dispersal facilitated by
102 human movement (Doggett *et al.* 2004; Reinhardt & Siva-Jothy 2007; Potter *et al.* 2008;
103 Szalanski *et al.* 2008).

104

105 In a bed bug metapopulation, human dwellings form habitat patches and it is expected that
106 limited numbers of individuals found new infestations. Bed bugs have been observed
107 actively dispersing throughout buildings (Doggett & Russell 2008; Wang *et al.* 2010) and it is
108 likely that this not only accounts for propagation of infestations within buildings but is also
109 the mechanism by which individuals move into portable items, leading to passive dispersal.
110 Research on the composition of founders has been limited and recent studies have given
111 contrasting results. For example, Szalanski *et al.* (2008), found up to six mitochondrial
112 haplotypes within a single infestation and Booth *et al.* (2012) showed evidence of multiple
113 introductions within an apartment complex. These results are consistent with a migrant pool
114 model of colonisation, with a more genetically diverse group of founders. In contrast, other
115 infestations have been shown to have very limited within-infestation diversity (Booth *et al.*

2012, Saenz *et al.* 2012), which more closely follows the predictions of a propagule pool model. Local extinctions are especially frequent as infested properties are often treated with insecticides, giving the majority of occupied patches a relatively short lifespan. Reports of widespread insecticide resistance (Doggett *et al.* 2004; Romero *et al.* 2007; Potter *et al.* 2008) suggest that selection is further shaping the genetic diversity of bed bug metapopulations. In combination, these factors predict low diversity within and very high levels of structuring between sub-populations of bed bugs.

Despite its importance in the maintenance of genetic diversity, there has been limited research on the number and genetic composition of colonists in natural metapopulations (Gaggiotti *et al.* 2004 but see Whitlock 1992; Austin *et al.* 2010). Here we use a metapopulation framework to provide detailed insight into bed bug population dynamics across two hierarchical levels of structure. We first characterized within-infestation and city-wide genetic diversity before estimating the number and demographic history of bed bug colonists using Approximate Bayesian Computation (ABC). Therefore this study, for the first time, uses a model-based approach to test, in a robust statistical framework, two alternative hypotheses on the genetic composition of founders in a natural bed bug metapopulation. We discuss the implications of these results for the future integrated control of bed bugs.

139 **MATERIALS AND METHODS**

140

141 *Sample collection*

142 To investigate within-infestation diversity, individuals were sampled from multiple
 143 refugia in five properties. These properties are subsequently referred to as AUS (sourced
 144 from New South Wales, Australia), BIR1 and BIR2 (separate properties located within
 145 Birmingham, UK) and LON1 & LON2 (separate properties from London, UK) (Table S1).
 146 Detailed spatial surveys of properties LON1 and LON2 are described in Naylor (2012), case
 147 studies 4 and 3 respectively). These infestations were selected from available samples on the
 148 basis of having individuals sampled separately from multiple refugia, rather than based on
 149 their geographic location.

150

151 To assess diversity at the city scale, 13 infestations from across London, UK, were sampled
 152 (see Figure S1, Table 3 for names and spatial locations). Pest control operatives, who
 153 obtained individuals prior to the treatment of affected properties, provided the majority of
 154 samples. Once received, all samples were stored in screw-topped rubber sealed microfuge
 155 tubes containing 1.5 ml of absolute ethanol (analytical reagent grade) at room temperature.

156

157 *DNA extraction and genotyping*

158 The ammonium acetate precipitation method described by Nicholls *et al.* (2000) was
 159 used to extract DNA. Individuals were then genotyped using 21 newly isolated microsatellite
 160 markers. The microsatellite loci were isolated from either a microsatellite-enriched genomic
 161 library or a recently available transcriptome assembly (O. Otti and K. Reinhardt, in prep; see

Supporting Information for a description of microsatellite isolation and characterisation). All 21 loci used were confirmed as autosomal by the observed presence of heterozygotes in male and female individuals. The same PCR conditions were used as described for primer testing (see Supporting Information), with individuals genotyped at 19 autosomal loci for the within-infestation study and at 21 autosomal loci for the city-wide study. Amplified products were analysed using an ABI3730 48-well capillary DNA analyser and allele sizes were assigned using GENEMAPPER v.3.7 (Applied Biosystems, California). Sequences were searched against the NCBI nr nucleotide database using BLASTn, which confirmed that these markers did not overlap with previously published bed bug microsatellites (Booth et al. 2012).

Genetic diversity

For each dataset descriptive summary statistics including number of alleles and expected (H_E) and observed heterozygosities (H_O) were obtained using Microsatellite Analyser version 4.05 (Dieringer & Schlötterer 2003). Allelic richness was calculated using F_{STAT} version 2.9.3.2 (Goudet 1995). We tested for deviations from Hardy-Weinberg equilibrium (HWE) and estimated the frequency of null alleles with Cervus version 3.0.3 (Kalinowski *et al.*, 2007). Evidence of linkage disequilibrium was assessed using GENEPOP version 4.1.0 (Raymond & Rousset 1995; Rousset 2008). For analyses of deviation from HWE and evidence of linkage disequilibrium, a Bonferroni correction was applied to allow for multiple tests (Rice 1989).

185 *Population structure*

186 We used F_{STAT} to calculate global F_{IS} and F_{ST} values (Weir & Cockerham 1984), both
187 within and among infestations. Values were jack-knifed over loci to give means and standard
188 errors and were bootstrapped over loci to give 95% confidence intervals. 10,000
189 permutations were used to generate significance values.

190

191 The within-city dataset was tested for isolation by distance amongst individuals in SPAGeDi
192 ver 1.3 (Hardy & Vekemans 2002) using the kinship coefficient (Loiselle *et al.* 1995). Distance
193 was partitioned into 10 intervals, with a uniform number of pairwise comparisons per
194 interval. The mean distance value of each interval was log transformed (Rousset 1997). We
195 used 10,000 permutations to test if the slope of the relationship between geographic and
196 genetic distance was significantly negative.

197

198 For both datasets we performed a Discriminant Analysis of Principal Components (DAPC)
199 (Jombart *et al.* 2010) using the ADEGENET 1.3-4 (Jombart 2008) package in R (R Core Team
200 2012) to examine evidence for genetic clusters, using infestation as a grouping prior. DAPC is
201 an ideal clustering method for this dataset as it does not make some commonly required
202 assumptions (e.g. Hardy-Weinberg equilibrium) (Jombart *et al.* 2010), which are unlikely to
203 hold for bed bug infestations. The first step in DAPC is to transform the raw data into
204 principal components (PCs). There is a subsequent trade-off in the number of retained PCs,
205 with a higher number of PCs increasing the ability to discriminate between groups at the
206 cost of the reduced stability of membership probabilities (Jombart *et al.* 2010). We used α -
207 score as a measure for judging the optimal number of retained principal components. The α -

10

score is the difference between the proportions of successful observed discriminations and values obtained from random discrimination. This was calculated with 100 permutations for each increasing number of retained principal components using the *optim.a.score* function in ADEGENET. Due to the low number of sampled individuals in each group, we were conservative with the number of retained principal components (PCs) but in both cases the number of retained PCs still incorporated $\geq 75\%$ of the variance in the data. The *dapc* function was then used to perform the clustering analysis and results are presented as ordination plots.

Approximate Bayesian Computation (ABC) analysis

ABC allows likelihood-free inference in complex scenarios with many parameters by comparing summary statistics from observed data to summary statistics from data simulated using various prior distributions (Beaumont 2010). In population genetics, where it is becoming a widely used tool, ABC analysis usually involves the construction of historical population models, simulation of many data sets using a coalescent approach and comparison of simulated to observed data using summary statistics describing genetic diversity. This provides a rigorous framework to compare different demographic models and then infer demographic parameters of interest (for a review of the global ABC procedure see Csilléry *et al.* (2010)).

In order to understand the dynamics of a bed bug metapopulation, we simulated the sampling of 20 microsatellites from 13 infestations under two demographic scenarios, mimicking the two theoretical models of metapopulation colonisation (Figure 1). Note that

from our 21 original microsatellite, the locus Cle001 was omitted, due to a large range of allele sizes, which reduced our ability to fit the models. Firstly, a migrant pool model was constructed where random individuals from the metapopulation found new infestations (scenario one) and, secondly, a propagule pool model was designed such that colonists all come from the same source infestation (scenario two) (Figure 1). In both cases, we would expect a severe bottleneck at the point of founding, followed by a period of rapid growth. We used ABC to test the posterior probabilities of these two hypotheses and to estimate parameters for the preferred scenario. All ABC analysis was conducted with the software package DIYABC (ver. 1.0.4.46) (Cornuet *et al.* 2008).

240

241 *Prior distributions and summary statistics used in ABC*

In order to specify informative priors we used information from case studies of infestations. Across 83 sampled infestations there was a range from eight to approximately 100,000 individuals (How & Lee 2010; Wang *et al.* 2010; Naylor 2012), with a geometric mean of 93 individuals. We therefore initially selected a prior range for the average effective size of an infestation at the time of sampling (N_e) bounded between 10 and 3,000 with a loguniform distribution. However, initial model runs suggested this was an overestimate so we reduced the prior range to a loguniform distribution bounded between 10 and 100 (Table 1). Based on observed female fecundity, infestations can reach over 3,000 individuals in as little as between two and six generations. Due to this rapid growth we fixed the bottleneck at one generation and selected a loguniform prior for time of founding (t_2) bounded between two and 10 generations. Previous studies and our own data have suggested that infestations are started with small numbers of founders (Doggett *et al.* 2004; Saenz *et al.*

12

2012); we therefore set a uniform prior for this parameter (N_b) bounded between two and 14 individuals. This captures the range of possibilities from a single, once-mated founding female through a single, multiply-mated female to a larger group of individuals. Population growth was simulated as a stepwise increase from N_b to N_e at time t_2-1 . For parameters where data were not available, we used a two-step approach starting with wide, uninformative priors and then using narrower ranges that still captured whole posterior distributions (Table 1). As there is no prior knowledge on the mutation rate of short tandem repeats in bed bugs, we made the assumption that microsatellite loci followed a generalised stepwise mutation model (Estoup *et al.* 2002). This model is defined by two parameters: μ , the mean mutation rate across loci and P , the parameter of the geometric distribution describing the number of repeat changes per mutation event. As implemented in DIYABC, each locus has its own μ_i and P_i drawn from a gamma distribution of mean μ and P , respectively, with shape parameter set to $k=2$. All loci had a possible range of 40 contiguous allelic states and we used DIYABC's default values for mean μ and P prior ranges (Table 1). We expected global diversity to depend on the product of the source population size, N_s , and the mutation rate μ , because the vast majority of mutations must have occurred before the first modelled colonisation events. We did not expect to be able to estimate N_s and μ independently but this does not impact our ability to estimate the parameters describing recent colonisation and expansion events.

273

The type and number of summary statistics used in ABC is important to the analysis (Beaumont 2010) but there is still no clear consensus on determining the optimal set of statistics (Csilléry *et al.* 2010). We selected summary statistics that had been used

276

277 successfully in previous population genetic ABC studies, providing evidence that they contain
278 useful information, and that were available in DIYABC, a practical constraint (Cornuet *et al.*
279 2008; Lombaert *et al.* 2011; Dutech *et al.* 2012). These were mean number of alleles, mean
280 genic diversity (Nei 1978) and mean allele-size variance for each population and pairwise
281 population comparisons of F_{ST} (Weir & Cockerham 1984), giving a total of 39 single sample
282 statistics and 78 pairwise comparisons. Goodness-of-fit analyses were later used to check
283 the robustness of the results obtained using these summary statistics (see parameter
284 estimation and model checking).

285

286 *Simulation and model posterior probabilities.*

287 We simulated 1×10^6 genetic datasets for each of the two scenarios. As an initial
288 check that these scenarios could simulate datasets close to our observed data, we
289 performed a principal components analysis (PCA) on the summary statistics of the first
290 100,000 simulated datasets, and evaluated the position of our observed data. To estimate
291 the posterior probabilities of both scenarios a polychotomous weighted regression was
292 performed on the closest 1% of simulated data to the observed data (Cornuet *et al.* 2008).
293 Confidence in scenario choice was calculated by estimating false discovery rate as in Cornuet
294 *et al.* (2010) using 500 pseudo-observed datasets (POD), simulated under the scenario with
295 the lower posterior probability.

296

297 *Parameter estimation and model checking*

298 Assessing the goodness-of-fit (GoF) of the chosen model is a critical step in ABC
299 analyses (Gelman *et al.* 1995). We therefore ran posterior predictive simulations by creating

14

300 500 PODs from the posteriors of the chosen model. A PCA was then performed on summary
301 statistics from (i) these 500 PODs, (ii) 500 datasets of each model randomly obtained from
302 their priors and (iii) our observed data (Cornuet *et al.* 2010). As in Cornuet *et al.* (2010), we
303 used a different set of summary statistics to perform the model checking compared to those
304 used to calculate the posterior distributions of parameters to avoid overestimation of fit. We
305 used Garza-Williamson's *M* (Garza and Williamson 2001) as a single sample statistic, and
306 mean genic diversity and classification index (Rannala and Mountain 1997) between
307 populations. The probability that the observed and the simulated data were significantly
308 different was calculated by ranking the observed value of each test statistic against those
309 obtained from the simulated data. *P*-values were corrected for multiple comparisons
310 (Benjamini & Hochberg 1995).

311

312 To evaluate the posterior distribution of parameters, we performed a local linear regression
313 on the closest 1% of logit-transformed simulated data. To assess confidence in our
314 parameter estimates, 500 PODs were created using values drawn from our prior distribution.
315 Using these datasets we calculated the relative median of the absolute error (RMAE) to
316 estimate differences in the point estimates from the true values (Cornuet *et al.* 2010).

317

318 **RESULTS**

319

320 *Characterising C. lectularius infestations*

321 In total, 21 microsatellite markers were found to be polymorphic and were
322 assembled into five ABI four-dye multiplex sets using the program Multiplex Manager

(Holleley & Geerts 2009) (Supporting Information, Table S2). To assess within-infestation genetic structure, 154 samples were genotyped from within five infestations at 19 of these loci. At the level of the whole infestation only LON2 showed a significant departure from HWE after Bonferroni correction. Overall, low genetic diversity was detected within infestations with a high number of monomorphic loci, low allelic richness, and low observed heterozygosity (Table 2). An F_{IS} value of -0.216 suggested an excess of heterozygotes within the AUS infestation (Table 2), which may be a signature of a recent bottleneck (Cornuet & Luikart 1996). Overall, estimates of F_{IS} were variable between infestations. This variability probably also reflects small numbers of founders. Note that the estimates are based on variable numbers of loci because of the occurrence of monomorphic loci within each infestation, consistent with low overall diversity, especially within BIR1, BIR2, LON1 and LON2. Within infestations, significant differentiation between refugia was only observed within LON2 ($F_{ST} = 0.144$, $P=0.008$; Table 2). Five out of 55 pairwise F_{ST} comparisons between LON2 refugia were significant, and no refuge was significantly differentiated from more than 27% of the other refugia (Table S3). In contrast, pairwise F_{ST} comparisons between infestations ranged from 0.492 to 0.834 (Table S4) and DAPC (Figure S2) showed very high levels of differentiation between infestations.

340

341 *City-wide genetic diversity*

342 In total, 63 individuals from 13 infestations across London were genotyped at 21 loci.
343 Due to the low sample sizes per infestation (three to seven individuals) HWE could not be
344 rejected. Within infestations, effective allele numbers ranged from 1.28 to 2.10 across all loci
345 (Table 3) and were therefore comparable to more thoroughly sampled infestations (see

above and Booth *et al.* 2012; Saenz *et al.* 2012). Within infestations, mean kinship was high (0.566 \pm 0.030) and across infestations F_{IS} was 0.084 (95% CI = -0.052 – 0.226). The global F_{ST} value across all 13 infestations showed significant population differentiation (F_{ST} = 0.592, SE = 0.026, $p < 0.001$). However, we found no significant pattern of isolation by distance (Figure 2, slope = 0.013 \pm 0.016). DAPC analysis further supported high differentiation between infestations with a defined genetic cluster for each infestation (Figure S3).

ABC analysis

A principal components analysis showed that our propagule pool model produced datasets that closely matched the observed data (Figure S4). The model comparison gave very strong support to scenario two. In fact, the closest 1% of datasets were all generated by scenario two. We therefore had to use 2% of simulated datasets to perform the logistic regression, which again gave almost complete support to scenario two (propagule pool model: posterior probability 0.9998, 95% Highest Posterior Density (HPD95): 0.9987, 1.0000) over scenario one (migrant pool model: 0.0002, HPD95: 0.0000, 0.0013). False discovery rate was low (5.2%) indicating the robustness of our model selection. Posterior predictive simulations showed that scenario two provides simulated datasets reasonably close to observed data (Figure S5). The deviation on the 2nd axis can be explained by the deviation of Garza-Williamson's M test statistics (the significant deviations are shown in Table S5).

With a posterior modal estimation of 3, the number of founders (N_b) starting each new infestation seems to be very limited (Median = 5.27, HPD90 = 2.09, 13.1 ; see Table 1). There was limited information regarding effective size of either current infestations or the source

369 population (N_e and N_s in Figure 3 and Table 1). The time since founding (t_2) was low,
370 suggesting that the sampled infestations were detected early (Median = 3.13, HPD90= 1.75,
371 9.36). Whilst the mutation rate parameter was not well estimated, additional simulations
372 showed that adjusting the prior on this parameter did not have a strong effect on the other
373 parameters of interest and resulted in an estimate of mutation rate within the range of the
374 original prior (Table S6, Figure S6).

375

376

377 **DISCUSSION**

378

379 Using a combination of descriptive and ABC genetic analyses within a
380 metapopulation framework, we have, for the first time, explicitly tested two competing
381 models of bed bug colonisation dynamics. First, we found very low within-infestation
382 diversity but despite this there is often rapid population growth. Second, ABC estimation
383 favours a propagule pool model that, in combination with low observed diversity within
384 infestations and very high differentiation between them, suggests that founders are related
385 (Whitlock & McCauley 1990). This is also consistent with strong genetic homogeneity
386 between refugia of the same dwelling, which is indicative of a single founding event per
387 infestation. Low F_{ST} within infestations suggests that the infestation is the lowest level of
388 population structure and not the refuge. Very high differentiation between infestations
389 suggests limited connectivity and the lack of isolation by distance at the city scale fits our
390 prediction that passive dispersal has weakened the relationship between genetic and

geographic distance (Johannesen & Lubin 1999; De Meester *et al.* 2002; Colson & Hughes 2004).

393

394 *Limitations and promise of our ABC approach*

395 Although theoretical studies have shown that the propagule and the migrant pool
396 models may produce somewhat similar differentiation patterns (Wade & McCauley 1988;
397 Pannell & Charlesworth 1999), our results clearly show that an ABC approach based on
398 multiple summary statistics can easily distinguish them. Using a high number of populations
399 with ABC means there may be a high dimensionality to the analysis, as it can inflate the
400 number of summary statistics used. This can be problematic as it often limits the ability to
401 simulate data close enough to the observed data, thus leading to inaccurate and potentially
402 biased scenario choices and parameter estimates (Beaumont *et al.* 2002; Blum *et al.* 2012).
403 In our case, the large number of summary statistics was primarily due to the use of pairwise
404 F_{ST} statistics. However, our PCA clearly shows that scenario two had the potential to simulate
405 data that were representative of our observed data (Figure S4: 90% of the variance
406 explained by both axes). The GoF analysis (Figure S5) further suggests that our choice of
407 summary statistics, whilst perhaps not optimal, was still good enough to make valid
408 inferences. It should also be noted that whilst our models are relatively simple they are
409 biologically relevant to this system in that they capture the key phases of colonisation,
410 growth and extinction. A more “realistic” model may have come at the cost of information
411 loss for our parameters of interest. In fact, we simulated a 3rd scenario in addition to the two
412 presented here. This scenario was similar to model two but with an additional bottleneck
413 after the initial divergence from the source population (i.e. a bottleneck between times t_3

414 and t_2). Despite this being more realistic, we lost information about our demographic
415 parameters and thus we chose to disregard this model.

416

417 Scenario two, arguably the more biologically realistic of the two scenarios, was favoured and
418 posterior predictive simulations gave good predictions of the observed data. The deviations
419 observed for some test statistics between their predicted distributions and their observed
420 values are likely a result of this being a simplified model. As noted above, we did not expect
421 to be able to obtain robust estimates of N_s and μ , which together determine global
422 population diversity. Adjusting the mutation parameter priors had little effect on the key
423 parameters of interest (see Supplementary Information). Within-infestation diversity was
424 expected to depend on N_e and N_b as well as the time parameters. The lack of information
425 for N_e is likely due to the information loss caused by the strong bottlenecks and very fast
426 subsequent population expansions within infestations. As there was a maximum of 10
427 generations between bottlenecks and sampling, there was not enough time for new
428 mutations to arise, and so there was little constraint on the estimate of infestation size.
429 Whilst care must be taken when using the ABC estimates for interpretation, it is important
430 to note that when taken in combination with the genetic analysis our models produce
431 biologically realistic estimates. This confirms the ability of ABC to make estimates of
432 demographic parameters when using multiple populations and whilst working within a very
433 short evolutionary time scale.

434

435

436

437 *Bed bug infestation population dynamics*

438 The very low levels of genetic diversity observed within infestations in this study,
439 suggest low numbers of colonists. These patterns are reflected in other descriptive surveys
440 of the genetic structure of bed bug populations (Booth *et al.* 2012; Saenz *et al.* 2012). High
441 null allele frequency estimates and departures from Hardy-Weinberg equilibrium have been
442 found in other genetic surveys of highly structured metapopulations (Johannesen & Lubin
443 1999; Massonnet *et al.* 2002; Kankare *et al.* 2004; Orsini *et al.* 2008). Despite several
444 markers having high null allele frequency estimates, from our knowledge of bed bug life
445 history in combination with previously published studies it is likely that these estimates are
446 an artefact of the lower than expected heterozygosity resulting from founder events and
447 inbreeding.

448
449 Due to living in aggregations and a high male mating rate, it is likely that most adult female
450 bed bugs have been multiply mated (Stutt & Siva-Jothy 2001; Reinhardt *et al.* 2011). With
451 low levels of genetic diversity a female that has been mated over 10 times may have only
452 “effectively” mated once or twice, as she will likely only encounter related males. The ABC
453 estimates are consistent with this. The asymmetry of the posterior distribution of Nb results
454 in a wide HPD but there is a strong mode at Nb = 3. This modal value potentially represents a
455 single female mated several times by genetically similar males; a version of the propagule
456 pool model. This can explain the genetic data where high relatedness and kinship are
457 detected within infestations. F_{IS} values have varied considerably across studies, but with few
458 significant departures from zero (Booth *et al.* 2012; Saenz *et al.* 2012). This is likely a result
459 of a low number of colonists with chance allele frequency differences between male and

female founders resulting in variation in the proportion of heterozygotes relative to Hardy-Weinberg equilibrium. Despite low diversity, infestations can rapidly expand, suggesting there are limited costs to inbreeding, and this is supported by preliminary experiments performed by Johnson (1941). This may be due to the continuous purging of deleterious alleles through repeated founder events (Hedrick 1994; Facon *et al.* 2011). Further work is needed to examine the true cost of inbreeding in bed bugs.

Whilst generally consistent patterns of low diversity have been observed within sub-populations, there is some variation in the genetic composition of infestations. For example, some studies have reported higher genetic diversity within infestations, which would be more consistent with the migrant pool model. Szalanski *et al.* (2008) reported one to six different mitochondrial haplotypes within infestations in a survey of 11 different properties. Booth *et al.* (2012) surveyed multiple apartments from residential blocks and found evidence of sub-structuring within two of the three buildings. In contrast, the survey by Saenz *et al.* (2012) showed that only one out of 21 properties had evidence of a multiple introduction, therefore more closely following the predictions of the propagule pool model. This variation suggests that the chance of multiple introductions is largely dependent on the type of property (i.e. single residence or multi-dwelling), with a greater turnover of humans increasing the likelihood of multiple founders. Due to the stochastic distribution of genetic diversity generated by high levels of population turnover, large sample sizes are often required to detect patterns of isolation by distance (Giles & Goudet 1997; Massonnet *et al.* 2002; Haag *et al.* 2005). As this study was not particularly designed to test for patterns of isolation by distance we may have lacked the resolution to detect it and in fact some weak

483 patterns have been found in other studies (Saenz *et al.* 2012). Further work is still needed to
484 determine the overall level of connectivity between sub-populations, and whether several
485 large infestations are acting as sources for multiple patches or whether all infestations can
486 be traced back to a larger mixed source population in an area where bed bug numbers have
487 been consistently high.

488

489

490 *Metapopulation dynamics*

491 Due to discontinuous habitat, and small local population sizes, metapopulation
492 dynamics should be fairly common in insects. However, classic metapopulations appear to
493 be comparatively rare (Driscoll 2008; Driscoll *et al.* 2010), potentially because of low
494 population turnover (Fronhofer *et al.* 2012). The close association between humans and bed
495 bugs has led them to fulfil at least three of the four conditions required to be a classical
496 metapopulation. Each separate infestation can be considered a discrete breeding patch, with
497 bed bugs aggregating around a host food source, and infestations are likely to have similar
498 effective population sizes. Pest control causes the independent local extinction of sub-
499 populations and human facilitated dispersal gives the opportunity for recolonisation.
500 However, reports of bed bug population expansion suggest that colonisation and extinction
501 are not at equilibrium (Reinhardt & Siva-Jothy 2007). At present there is no precise
502 information on the rate of population expansion, so the degree to which colonisation is
503 outweighing extinction is not known.

504

We have demonstrated that bed bugs experience repeated founder events, with a severe genetic bottleneck due to very low estimated numbers of founders. As predicted, the likely common origin of colonists causes particularly strong differentiation (Whitlock & McCauley 1990). The relatively short life span of each infestation makes it unlikely that there will be an introduction of novel alleles via gene flow, thus maintaining these levels of differentiation. The low diversity within sub-populations may be further shaped by selective sweeps from a strong insecticide selection pressure. Examples of species that can be considered metapopulations have been shown to exist on a continuum, ranging from migrant pool colonisers (Giles & Goudet 1997; Colson & Hughes 2004; Yang *et al.* 2008), to intermediates (Whitlock 1992; Austin *et al.* 2010) and examples of high likelihood of a common origin of founders (Ingvarsson *et al.* 1997). From our ABC estimates, alongside knowledge of the bed bug life history, it is highly likely that in the majority of cases bed bug founders within an infestation have a common origin. In combination these dynamics make bed bugs an excellent model system to further investigate human impact on population structure and its implication for diversity, as well as pest control.

520

521 *Implications for control*

Metapopulation frameworks have long been used to formulate efficient pest management strategies (e.g. Levins, 1969; Cloarec 1999; Booth *et al.* 2011). Here, the general low observed diversity within infestations suggests a single introduction to each infestation. Type of dwelling, however, is likely to strongly influence the chances of having multiple introductions. Hotels, apartment blocks and hospitals, which have the highest turnover of visitors, are likely to be the most at risk from multiple introductions (Doggett &

24

Russell 2008). However, from the reported data so far it seems these cases may, for the moment at least, be in the minority and reported repeat infestations may be the result of failure to fully eradicate existing infestations (Boase 2008). The novel microsatellite markers described here, in conjunction with those described in Booth et al. (2012), could be used to determine whether a repeat infestation is a pest management failure or a recolonisation event by surveying properties before or immediately after treatment, and keeping specimens in case of a repeat infestation. This study also shows that due to the low diversity within and high differentiation between infestations, only a relatively small number of individuals is required to test for kinship between samples. Another factor to consider is that bed bugs are becoming rapidly resistant to insecticides. Infestations containing resistant individuals are more likely to avoid extinction, prolonging their time as a source. Resistant alleles would be quickly selected for and become fixed in populations. These markers, in combination with candidate resistance loci, can now be used to track resistance alleles moving through metapopulations.

542

543

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781 **Figure 1** Demographic scenarios for ABC. *Scenario one: Migrant pool model of colonisation:*
782 All populations originate from a single hypothetical source population (N_s), which represents
783 the metapopulation as a whole. At time t_2 these populations diverge signifying the founding
784 of new infestations, which includes a severe bottleneck (N_b). The populations subsequently
785 grow and after a bottleneck of one generation reach effective population size (N_e). It is at
786 this size at which the infestations are sampled. *Scenario two: Propagule pool model of*
787 *colonisation:* In this scenario founders diverge from N_s at t_3 and maintain a population size
788 of N_e until t_2 where there is a founding event and a severe bottleneck (N_b). The sub-
789 population then grows to a size of N_e after a bottleneck of 1 generation before being
790 sampled. In both scenarios only infestations a and b are shown, but models incorporate 13
791 sampled infestations, represented by the dotted line.

792

793 **Figure 2** Kinship plotted against distance with standard error bars. The first point represents
794 within-infestation kinship, the following 10 points represent geographic distance, which was
795 broken down into 10 distance intervals by SPAGeDi.

Figure 3 Prior (Grey) and Posterior (Black) distributions of parameters obtained under the better-supported model (scenario two). The x-axis shows the range of parameter values, and the y-axis the probability density.

Data accessibility

The 329 microsatellite sequences isolated during the development of the genomic library including those of the 21 loci characterised in this study have been submitted to the EMBL database (HF969864–HF970194) and microsatellite genotyping data are available in DRYAD, doi:XXXXXXXXX.

Author contributions

TF, KR and RKB designed the study. TF and GH performed the experiments. TF and LD performed the analysis. TF, LD, KR and RKB wrote the paper.

32

Table 1 Details of prior and posterior distributions of model parameters. Parameters constrained such that $N_s > N_e > N_b$.

Parameters	Prior range	Mean	Median	Mode	HPD90 low	HPD90 high	RMAE
N_s	Loguniform [100 - 50000]	6320	4950	2460	1520	15200	0.416
N_e	Loguniform [10 - 100]	33.5	32.6	35.6	12.7	57.1	0.258
N_b	Uniform [2 - 14]	6.21	5.27	3.00	2.09	13.1	0.345
t_2	Loguniform [2 - 10]	3.97	3.13	2.00	1.75	9.36	0.456
t_3	Uniform [11 - 100]	52.2	49.6	26.8	18.3	94.1	0.266
Mean μ	Loguniform [10^{-4} - 10^{-3}]	3.02×10^{-4}	2.22×10^{-4}	1.00×10^{-4}	1.07×10^{-4}	7.86×10^{-4}	0.382
Mean P	Uniform [0.1 - 0.3]	0.115	0.103	0.100	0.100	0.167	0.243

RMAE = Relative Median of the Absolute Error.

RMAE computed using 500 pseudo-observed datasets taking the median of posterior distribution as point estimates

Table 2 Genetic diversity and structure within five *C. lectularius* infestations for which multiple refugia were sampled. Total is the value obtained when individual for all localities were pooled together. Heterozygosity and F-statistics were calculated within and among *C. lectularius* infestations at 19 loci. Significance of F_{ST} values was calculated after 10,000 permutations.

Infestation	n	Allele Range	Allelic Richness ^a (\pm SE, n=17)	L_M	H_e	H_o	F_{IS} (95% CI)	F_{ST}
AUS	41	1-4	1.71 (0.15)	3	0.280	0.334	-0.216 (-0.349, -0.035)	0.017 ^{NS}
BIR1	8	1-3	1.61 (0.12)	7	0.173	0.054	0.727 (0.476, 0.926)	-0.184 ^{NS}
BIR2	9	1-5	2.10 (0.24)	6	0.380	0.337	0.141 (-0.104, 0.394)	-0.044 ^{NS}
LON1	52	1-3	1.19 (0.09)	13	0.075	0.084	-0.135 (-0.232, -0.013)	0.010 ^{NS}
LON2	46	1-3	1.62 (0.16)	8	0.250	0.163	0.219 (0.079, 0.475)	0.144 ^{**}
Total	156	1-5	2.98 (0.15)	7.4	0.566	0.194	0.052 (-0.072, 0.198)	0.709 ^{***}

n = Number of individuals, ^aSamples standardised to size of the smallest number of individuals for a locus

Two loci were omitted because they failed to amplify for any individual in one infestation

L_M = Number of monomorphic loci, H_e = Expected heterozygosity, H_o = Observed heterozygosity

^{NS} non-significant ($p > 0.05$), ^{**} significant ($p < 0.01$), ^{***} highly significant ($p < 0.001$)

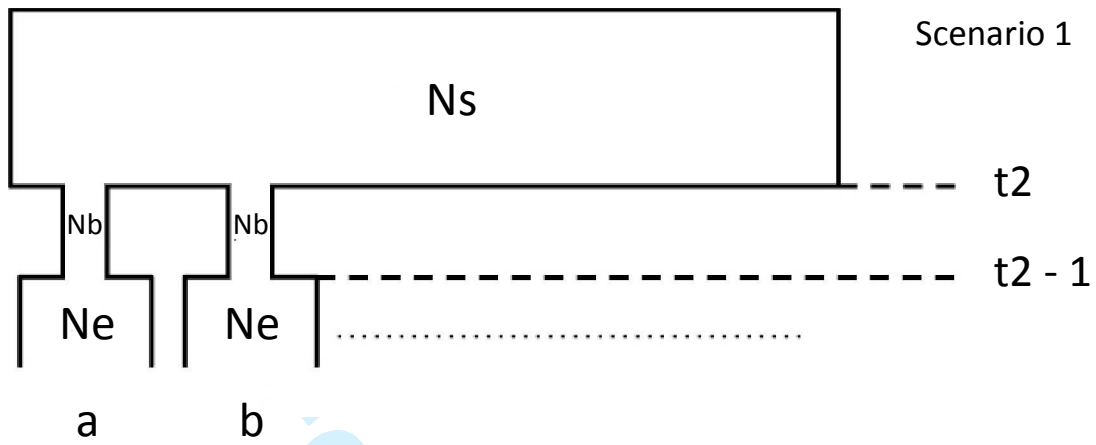
838 **Table 3** Indices of genetic diversity of 13 infestations from across London, UK. Total is the
839 value obtained when individual for all localities were pooled together

Infestation	Co-ordinates		n	E _A	Allelic Richness ^a (± SE, n=20)	H _E	H _O
	Lat	Long					
a	51.4924	-0.2294	3	1.35	1.41 (0.12)	0.209	0.270
b	51.4693	-0.1138	3	1.62	1.65 (0.11)	0.384	0.333
c	51.5293	-0.0218	3	1.28	1.37 (0.10)	0.197	0.270
d	51.4924	-0.1674	7	1.37	1.41 (0.10)	0.223	0.211
e	51.5851	-0.2602	4	1.28	1.37 (0.10)	0.189	0.163
f	51.5333	-0.1681	7	1.54	1.59 (0.11)	0.303	0.213
g	51.4491	-0.1215	5	2.10	2.07 (0.13)	0.510	0.279
h	51.6905	-0.0338	4	1.51	1.58 (0.12)	0.287	0.298
i	51.5840	-0.1171	5	1.62	1.67 (0.09)	0.370	0.412
j	51.4799	-0.0296	4	1.49	1.55 (0.10)	0.308	0.345
k	51.3843	-0.4207	6	1.41	1.41 (0.10)	0.242	0.216
l	51.5113	-0.2679	6	1.42	1.43 (0.11)	0.228	0.213
m	51.5561	-0.1739	6	1.46	1.56 (0.11)	0.275	0.317
Total			63	3.26	2.53 (0.06)	0.680	0.266

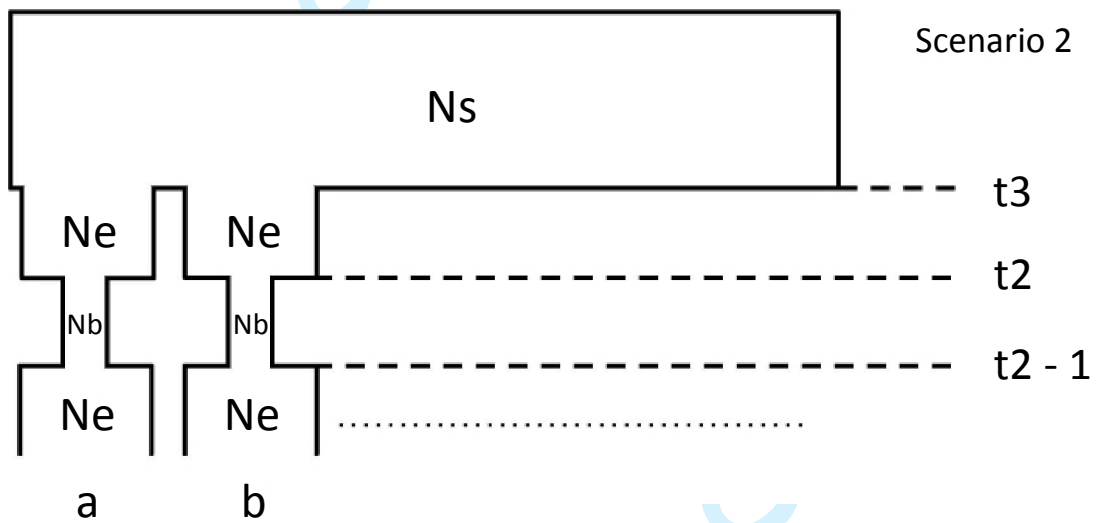
^aSamples standardised to size of the smallest number of individuals for a locus
One locus was omitted as for one infestation no individuals amplified for that locus
E_A = Effective number of alleles n = Number of individuals
H_E = Expected heterozygosity H_O = Observed heterozygosity

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846 **Figure 1**

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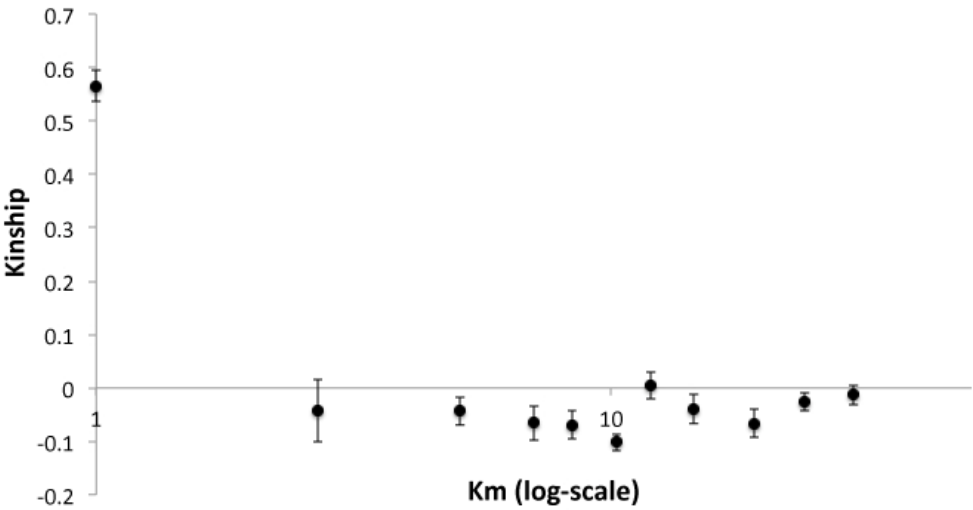
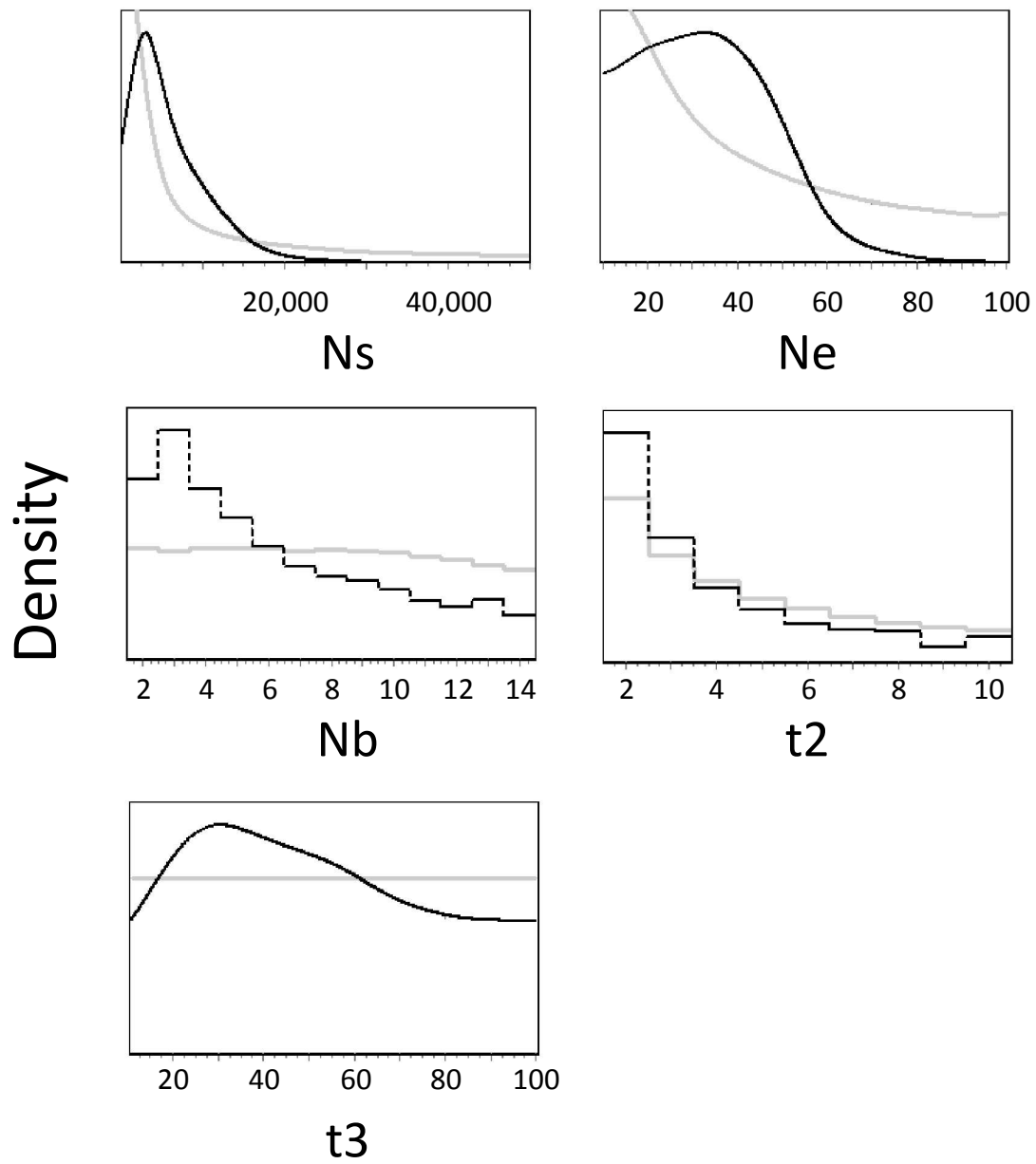


Figure 2

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859 **Figure 3**