Supplementary Material

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3 Predator phylogeny

- 4 We used information from timetree.org to add node ages to our tree. This web service provides age estimates
- 5 from the literature for all available pairs of taxa from the same groups as the searched taxa. Thus deeper
- 6 nodes are estimated from more taxa and studies than shallower nodes.
- ⁷ Table S1 Studies used to date nodes on our predator phylogeny. When more than one study was available,
- 8 we used the median value. Papers reporting phylogenetic time estimates include (Foley et al. 1998; Aris-
- 9 Brosou and Yang 2002; Gaunt and Miles 2002; Otsuka and Sugaya 2003; Wiegmann et al. 2003; Peterson
- 10 et al. 2004; Blair and Hedges 2005; Blair et al. 2005; Regier et al. 2005; Peterson et al. 2008).

| Study taxon 1 | Study taxon 2 | Taxon.A | Taxon.B | Time | Reference |
|-----------------|---------------|-------------------|-------------------------|-------|-------------------------|
| Ceratopogonidae | Chironomidae | Ceratopogonidae | Chironomidae | 213 | Bertone et al. 2008 |
| Culicidae | Chironomidae | Chaoborinae | Chironomoidea | 220 | Bertone et al. 2008 |
| Dolichopodidae | Tabanidae | Nemestrinoidea | Pelecorhynchidae | 216 | Wiegmann et al. 2003 |
| Dolichopodidae | Empididae | Gloma | Liancalus | 86 | Wiegmann et al. 2003 |
| Insecta | Hirudinea | Anophelinae | Capitellidae | 610 | Peterson et al. 2008 |
| Insecta | Hirudinea | Arthropoda | Chaetopteridae | 561 | Foto et al. 2005 |
| Insecta | Hirudinea | Arthropoda | Annelida | 560 | Peterson et al. 2004 |
| Insecta | Hirudinea | Crustacea | Annelida | 700 | Otsuka et al. 2003 |
| Insecta | Hirudinea | Culicidae | Lumbricidae | 543 | Aris-Brosou et al. 2002 |
| Odonata | Tabanidae | Coenagrionidae | Anophelinae | 385 | Peterson et al. 2008 |
| Odonata | Tabanidae | Hexagenia | Forficulidae | 151 | Regier et al. 2005 |
| Odonata | Tabanidae | Enallagma | gambiae species complex | 543 | Blair et al. 2005 |
| Odonata | Tabanidae | Hexagenia | Forficulidae | 376.5 | Regier et al. 2004 |
| Tabanidae | Culidicdae | Trichoceroidea | Oestroidea | 196 | Bertone et al. 2008 |
| Tabanidae | Culidicdae | Anopheles | Drosophila | 280 | Peterson et al. 2008 |
| Tabanidae | Culidicdae | Anopheles gambiae | Drosophila melanogaster | 419 | Blair et al. 2005 |
| Tabanidae | Culidicdae | Anopheles | Drosophila | 474 | Blair et al. 2005 |
| Tabanidae | Culidicdae | Aedes | Drosophila | 184 | Aris-Brosou et al. 2002 |
| Tabanidae | Culidicdae | Anopheles | Ceratitidini | 265.2 | Gaunt et al. 2002 |
| Tabanidae | Culidicdae | Culicini | Drosophila | 106.2 | Foley et al. 1998 |

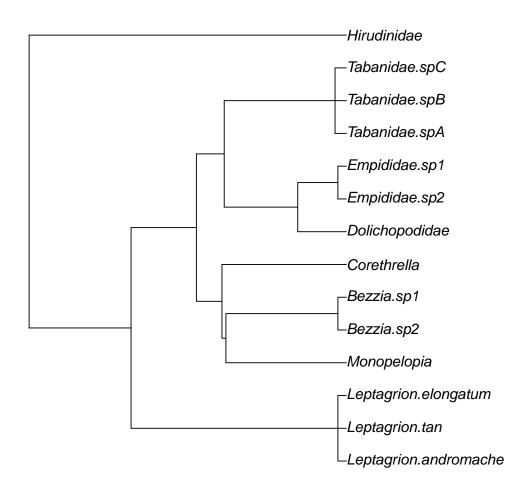


Figure S1 Predator phylogeny, with dated nodes derived from data in Table 1. When multiple time estimates were available for the same internal node, we used the median estimate.

Diet similarity

- Table S2 Proportion of predation in feeding trials. Each cell of the graph represents a predator-prey pair;
- note that some pairs are represented more than others (not shown). Composition of these trials was limited
- by the abundance of organisms in the field.

| | L.elong | L.tan | L.andro | leech | Stibasoma sp. | Monopelopia | Tab.B | Tab.C |
|------------------------|---------|-------|---------|-------|---------------|-------------|-------|-------|
| Tipulid | 0.60 | 1.00 | 1.00 | 1.00 | 1.00 | 0.40 | 0.50 | 1.00 |
| Culex | 0.80 | 0.57 | 1.00 | 0.17 | | 0.00 | | |
| Polypedilum.1 | 1.00 | 1.00 | 1.00 | | 0.50 | 0.60 | | |
| Scirtes.A | 0.82 | 0.75 | 0.80 | 0.67 | 0.20 | | | |
| Chironomus.detriticula | 1.00 | 1.00 | 1.00 | | 0.80 | | | |
| Psychodid | 1.00 | | 0.80 | 0.50 | 0.25 | | | |
| Hirudinidae | 0.17 | 0.00 | | 0.00 | | | | |
| Ostracoda | 0.27 | 0.62 | | | | 0.00 | | |
| Phylloicus.bromeliarum | 0.00 | 0.12 | | | | 0.00 | | |
| Polypedilum.2 | 1.00 | | 1.00 | | 0.80 | | | |
| Scirtes.B | 0.57 | 0.57 | | 0.67 | | | | |
| Monopelopia | 1.00 | 0.75 | | | | | | |

18 Predator diversity experiment

19 Prey community composition

- We estimated the densities of these common prey species based on their abundances in our 2008 observational
- 21 dataset. We estimated how many of each taxa would be found in a bromeliad of the same size as those we used
- 22 in our experiment, and adjusted our estimates slightly to account for interannual variation in invertebrate
- 23 abundance.
- Table S4 Densities of prey species used in the 2010 predator diversity experiment.

| Species | density |
|------------------------|---------|
| Chironomus detriticula | 10 |
| Polypedium marcondesi | 4 |
| Polypedium kaingang | 2 |
| Psychodid sp. 1 | 1 |
| Scyrtes sp. A | 5 |
| Culex spp. | 4 |
| Trentepholia sp. | 1 |

25 Experimental responses

- ²⁶ We calculated means and standard error for every experimental treatment and response variable.
- Table S3: Means and standard errors for all response variables for each predator treatment in our manipu-
- lative experiment. n=5 for every pair of numbers, with only two exceptions due to missing values: elong
- + leech x growth, and leech x decomp.

| treatment | decomp | fine | growth | N | total.surv |
|---------------|-------------------|------------------|----------------|--------------|----------------|
| andro | 0.37 ± 0.028 | 0.26 ± 0.041 | 4 ± 0.45 | 45 ± 6.2 | 5.8 ± 1.9 |
| control | 0.33 ± 0.023 | 0.38 ± 0.081 | 4.7 ± 1.2 | 53 ± 2.9 | 14 ± 2.7 |
| elong | 0.36 ± 0.0077 | 0.19 ± 0.054 | 2.9 ± 1.3 | 45 ± 6 | 7.8 ± 1.7 |
| elong + andro | 0.33 ± 0.024 | 0.58 ± 0.14 | 3.3 ± 1.2 | 52 ± 4.8 | 5.2 ± 1.4 |
| elong + leech | 0.32 ± 0.0083 | 0.32 ± 0.044 | 5.4 ± 0.62 | 43 ± 7.5 | 9 ± 1.1 |
| elong + tab | 0.37 ± 0.024 | 0.36 ± 0.25 | 4 ± 0.98 | 54 ± 7.8 | 10 ± 1.5 |
| leech | 0.35 ± 0.019 | 0.29 ± 0.092 | 2.6 ± 1.6 | 36 ± 5.2 | 5.2 ± 0.58 |
| tabanid | 0.31 ± 0.027 | 0.24 ± 0.12 | 5.1 ± 0.95 | 56 ± 7 | 5.8 ± 1.4 |

30 Differences among treatments

Number of predator species

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- We divided our experimental design into three separate hypothesis tests.
 - 1. **Predator species number:** We compared the five control replicates with the means of the four single species treatments and the means of the three two-species treatments. This tests whether species number *per se* has an effect.
 - 2. **Predator species identity:** We compared the four different predator taxa to see if any of these differed in their effect on the community.
 - 3. **Predator phylogenetic diversity:** We contrast our two predator species treatments to see how increasing PD of the predator assemblage effects our response variables.
- Because the first and third hypotheses suggest a ranking of the independent variables (increasing species number and phylogenetic diversity) we used ordinal contrasts in the corresponding linear models.
- Table S4: Number of predator species in each experimental treatment.

| Nspp | treatment |
|------|---------------|
| none | control |
| one | andro |
| one | tabanid |
| one | leech |
| one | elong |
| two | elong + andro |
| two | elong + leech |
| two | elong + tab |

Table S5: The effect of predator species number on all 5 response variables. We used ordinal contrasts to investigate how an increasing number of predator species influenced the response. Treatment order was none < one < two species. (See Table S6). For each non-control treatment, replicates were the mean of all bromeliads with the same predator treatment. Control (no predator) bromeliads were used as independent replicates. Thus there are 5 replicates for the "none" factor level, four replicates of "one predator" and three of "two predators". Note that we model prey survival using OLS, rather than poisson regression, because we are modeling the distribution of means.

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| | | $Dependent\ variable:$ | | | | | |
|--------------------------------|---------------------|------------------------|---------------------|----------------------|---------------------|--|--|
| | Decomposition (g) | FPOM (g) | Bromeliad growth | Nitrogen cycling | Total prey survival | | |
| | (1) | (2) | (3) | (4) | (5) | | |
| Linear | 0.007 | 0.029 | -0.317 | -2.418 | -4.384^{*} | | |
| | (0.020) | (0.072) | (1.015) | (3.634) | (2.228) | | |
| Quadratic | -0.011 | 0.127 | 0.664 | 4.638 | 4.205^{*} | | |
| · | (0.020) | (0.070) | (0.993) | (3.557) | (2.181) | | |
| Constant | 0.339*** (0.011) | 0.349*** (0.041) | 4.175*** (0.580) | 49.277*** (2.076) | 9.583*** (1.273) | | |
| Observations | 12 | 12 | 12 | 12 | 12 | | |
| \mathbb{R}^2 | 0.055 | 0.267 | 0.064 | 0.214 | 0.497 | | |
| Adjusted R^2 | -0.156 | 0.104 | -0.143 | 0.039 | 0.385 | | |
| Residual Std. Error $(df = 9)$ | 0.039 | 0.139 | 1.965 | 7.037 | 4.315 | | |
| F Statistic (df = 2 ; 9) | 0.260 | 1.636 | 0.310 | 1.226 | 4.450^{**} | | |

Note: *p<0.1; **p<0.05; ***p<0.01

 $_{53}$ Table S6: Tukey post-hoc tests for the effects of predator species number on each response variable.

| response | comparison | diff | lwr | upr | p.adj |
|---------------------|------------|-------|--------|-------|-------|
| Decomposition (g) | one-none | 0.02 | -0.05 | 0.09 | 0.76 |
| Decomposition (g) | two-none | 0.01 | -0.07 | 0.09 | 0.94 |
| Decomposition (g) | two-one | -0.01 | -0.09 | 0.07 | 0.95 |
| FPOM (g) | one-none | -0.14 | -0.40 | 0.13 | 0.36 |
| FPOM (g) | two-none | 0.04 | -0.24 | 0.32 | 0.92 |
| FPOM (g) | two-one | 0.18 | -0.12 | 0.47 | 0.27 |
| Bromeliad growth | one-none | -1.04 | -4.72 | 2.64 | 0.72 |
| Bromeliad growth | two-none | -0.45 | -4.46 | 3.56 | 0.95 |
| Bromeliad growth | two-one | 0.59 | -3.60 | 4.78 | 0.92 |
| Nitrogen cycling | one-none | -7.39 | -20.57 | 5.79 | 0.31 |
| Nitrogen cycling | two-none | -3.42 | -17.77 | 10.93 | 0.79 |
| Nitrogen cycling | two-one | 3.97 | -11.04 | 18.98 | 0.75 |
| Total prey survival | one-none | -8.25 | -16.33 | -0.17 | 0.05 |
| Total prey survival | two-none | -6.20 | -15.00 | 2.60 | 0.18 |
| Total prey survival | two-one | 2.05 | -7.15 | 11.25 | 0.81 |

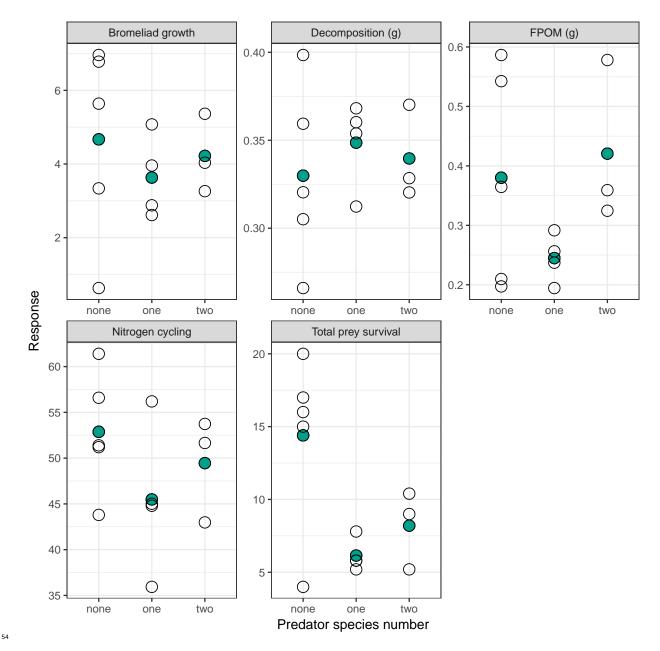


Figure S2: The effect of predator species number on each of our response variables. Each dot represents the mean (n = 5) for different predator treatments. Green dots represent group (i.e. species number) means.

Table S7: Within the single-predator treatments, we tested for difference in our response variables among different predator species. These linear models use treatment contrasts, which contrast all treatments with L.

andromache. Coefficents (standard error) are given for L. andromache, and differences between coefficients (standard error of difference) for all other predators.

| | | | Dependent variable: | | |
|---------------------|---------------------|--------------------|---------------------|--------------------|--------------------|
| | Total prey survival | FPOM (g) | Decomposition (g) | Bromeliad growth | Nitrogen cycling |
| | (1) | (2) | (3) | (4) | (5) |
| L. elongatum | 2.000 | -0.062 | -0.008 | -1.080 | -0.220 |
| | (2.069) | (0.109) | (0.031) | (1.622) | (8.656) |
| Hirudinidae | -0.600 | 0.035 | -0.014 | -1.344 | -9.080 |
| | (2.069) | (0.109) | (0.033) | (1.622) | (8.656) |
| Tabanidae | -0.000 | -0.019 | -0.056^{*} | 1.116 | 11.180 |
| 61 | (2.069) | (0.115) | (0.031) | (1.622) | (8.656) |
| $L.\ and romache$ | 5.800*** | 0.256*** | 0.368*** | 3.960*** | 45.020*** |
| | (1.463) | (0.077) | (0.022) | (1.147) | (6.121) |
| Observations | 20 | 19 | 19 | 20 | 20 |
| R^2 | 0.102 | 0.052 | 0.205 | 0.153 | 0.256 |
| Adjusted R^2 | -0.067 | -0.137 | 0.046 | -0.006 | 0.117 |
| Residual Std. Error | 3.271 (df = 16) | 0.172 (df = 15) | 0.049 (df = 15) | 2.564 (df = 16) | 13.686 (df = 16) |
| F Statistic | 0.603 (df = 3; 16) | 0.277 (df = 3; 15) | 1.289 (df = 3; 15) | 0.961 (df = 3; 16) | 1.838 (df = 3; 16) |

*p<0.1; **p<0.05; ***p<0.01

 62 Table S8: Post-hoc Tukey tests for differences between predator species in all 5 responses.

| | • | 1.0 | 1 | | 1. |
|-------------------|---------------|-------|--------|-------|-------|
| response | comparison | diff | lwr | upr | p.adj |
| FPOM (g) | elong-andro | -0.06 | -0.37 | 0.25 | 0.94 |
| FPOM (g) | leech-andro | 0.04 | -0.28 | 0.35 | 0.99 |
| FPOM (g) | tabanid-andro | -0.02 | -0.35 | 0.31 | 1.00 |
| FPOM (g) | leech-elong | 0.10 | -0.22 | 0.41 | 0.81 |
| FPOM (g) | tabanid-elong | 0.04 | -0.29 | 0.37 | 0.98 |
| FPOM (g) | tabanid-leech | -0.05 | -0.39 | 0.28 | 0.96 |
| Decomposition (g) | elong-andro | -0.01 | -0.10 | 0.08 | 0.99 |
| Decomposition (g) | leech-andro | -0.01 | -0.11 | 0.08 | 0.97 |
| Decomposition (g) | tabanid-andro | -0.06 | -0.15 | 0.03 | 0.31 |
| Decomposition (g) | leech-elong | -0.01 | -0.10 | 0.09 | 1.00 |
| Decomposition (g) | tabanid-elong | -0.05 | -0.14 | 0.04 | 0.44 |
| Decomposition (g) | tabanid-leech | -0.04 | -0.14 | 0.05 | 0.60 |
| Bromeliad growth | elong-andro | -1.08 | -5.72 | 3.56 | 0.91 |
| Bromeliad growth | leech-andro | -1.34 | -5.98 | 3.30 | 0.84 |
| Bromeliad growth | tabanid-andro | 1.12 | -3.52 | 5.76 | 0.90 |
| Bromeliad growth | leech-elong | -0.26 | -4.90 | 4.38 | 1.00 |
| Bromeliad growth | tabanid-elong | 2.20 | -2.44 | 6.84 | 0.54 |
| Bromeliad growth | tabanid-leech | 2.46 | -2.18 | 7.10 | 0.45 |
| Nitrogen cycling | elong-andro | -0.22 | -24.98 | 24.54 | 1.00 |
| Nitrogen cycling | leech-andro | -9.08 | -33.84 | 15.68 | 0.72 |
| Nitrogen cycling | tabanid-andro | 11.18 | -13.58 | 35.94 | 0.58 |
| Nitrogen cycling | leech-elong | -8.86 | -33.62 | 15.90 | 0.74 |
| Nitrogen cycling | tabanid-elong | 11.40 | -13.36 | 36.16 | 0.57 |
| Nitrogen cycling | tabanid-leech | 20.26 | -4.50 | 45.02 | 0.13 |
| | | | | | |

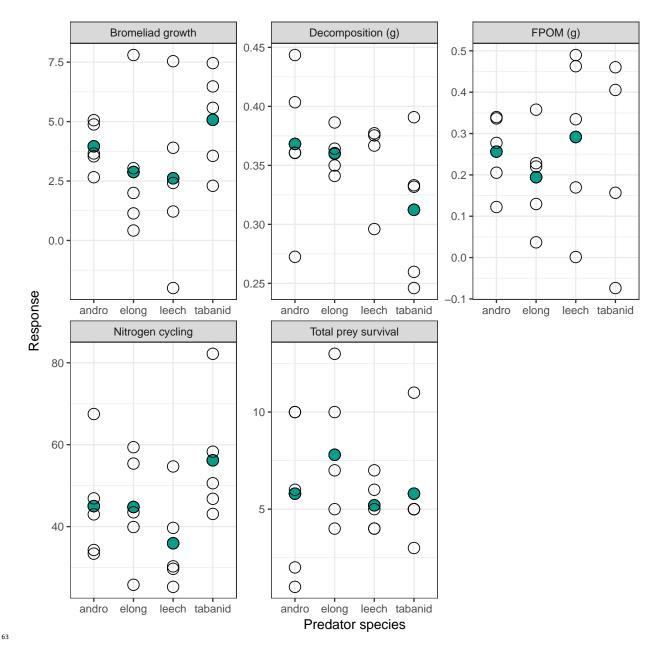


Figure S3: The effect of predator species identity on each of our response variables. Each dot represents the observed value for a different predator treatment. Green dots represent group means.

Increasing predator phylogenetic diversity

- Our two-species combinations represent a range of increasing phylogenetic distance. By examining differences
- between these three treatments, we can test how increasing PD per se influences our response variables.
- Table S9: Effects of increasing phylogenetic diversity of the predator assemblage on 5 response variables.
- We used linear contrasts, ranking the independent variable by increasing predator PD (see Figure S4). Linear
- model coefficients (standard error) are shown for Linear, Quadratic and Constant contrasts.

| | | | $Dependent\ variable:$ | | |
|---------------------|---------------------|---------------------|------------------------|----------------------|---------------------------|
| | Decomposition (g) | FPOM (g) | Bromeliad growth | Nitrogen cycling | Total prey survival |
| | (1) | (2) | (3) | (4) | (5) |
| treatment.L | -0.006 | -0.179 | 1.486 | -6.138 | 2.687* |
| | (0.020) | (0.167) | (1.027) | (6.841) | (1.354) |
| treatment.Q | -0.037^{*} | 0.075 | 0.227 | -5.242 | -2.694^{*} |
| v | (0.020) | (0.167) | (0.988) | (6.841) | (1.354) |
| Constant | 0.340*** (0.012) | 0.421*** (0.096) | 4.222*** (0.582) | 49.460*** (3.950) | 8.200*** (0.782) |
| Observations | 15 | 15 | 14 | 15 | 15 |
| \mathbb{R}^2 | 0.229 | 0.102 | 0.161 | 0.104 | 0.397 |
| Adjusted R^2 | 0.101 | -0.048 | 0.009 | -0.045 | 0.296 |
| Residual Std. Error | 0.045 (df = 12) | 0.372 (df = 12) | 2.164 (df = 11) | 15.297 (df = 12) | 3.028 (df = 12) |
| F Statistic | 1.787 (df = 2; 12) | 0.681 (df = 2; 12) | 1.056 (df = 2; 11) | 0.696 (df = 2; 12) | $3.949^{**} (df = 2; 12)$ |

Note:

*p<0.1; **p<0.05; ***p<0.01

- Table S10: Tukey post-hoc tests for differences between predator phylogenetic diversity treatments on all
- response variables.

| rognongo | comparison | diff | lwr | unr | p.adj |
|---------------------|-----------------------------|--------|--------|-------|-------|
| response | comparison | | | upr | |
| Decomposition (g) | elong + tab-elong + andro | 0.04 | -0.03 | 0.12 | 0.34 |
| Decomposition (g) | elong + leech-elong + andro | -0.01 | -0.08 | 0.07 | 0.96 |
| Decomposition (g) | elong + leech-elong + tab | -0.05 | -0.13 | 0.03 | 0.22 |
| FPOM (g) | elong + tab-elong + andro | -0.22 | -0.85 | 0.41 | 0.63 |
| FPOM (g) | elong + leech-elong + andro | -0.25 | -0.88 | 0.37 | 0.55 |
| FPOM (g) | elong + leech-elong + tab | -0.03 | -0.66 | 0.59 | 0.99 |
| Bromeliad growth | elong + tab-elong + andro | 0.77 | -2.92 | 4.47 | 0.84 |
| Bromeliad growth | elong + leech-elong + andro | 2.10 | -1.82 | 6.02 | 0.35 |
| Bromeliad growth | elong + leech-elong + tab | 1.33 | -2.59 | 5.25 | 0.64 |
| Nitrogen cycling | elong + tab-elong + andro | 2.08 | -23.73 | 27.89 | 0.97 |
| Nitrogen cycling | elong + leech-elong + andro | -8.68 | -34.49 | 17.13 | 0.65 |
| Nitrogen cycling | elong + leech-elong + tab | -10.76 | -36.57 | 15.05 | 0.53 |
| Total prey survival | elong + tab-elong + andro | 5.20 | 0.09 | 10.31 | 0.05 |
| Total prey survival | elong + leech-elong + andro | 3.80 | -1.31 | 8.91 | 0.16 |
| Total prey survival | elong + leech-elong + tab | -1.40 | -6.51 | 3.71 | 0.75 |

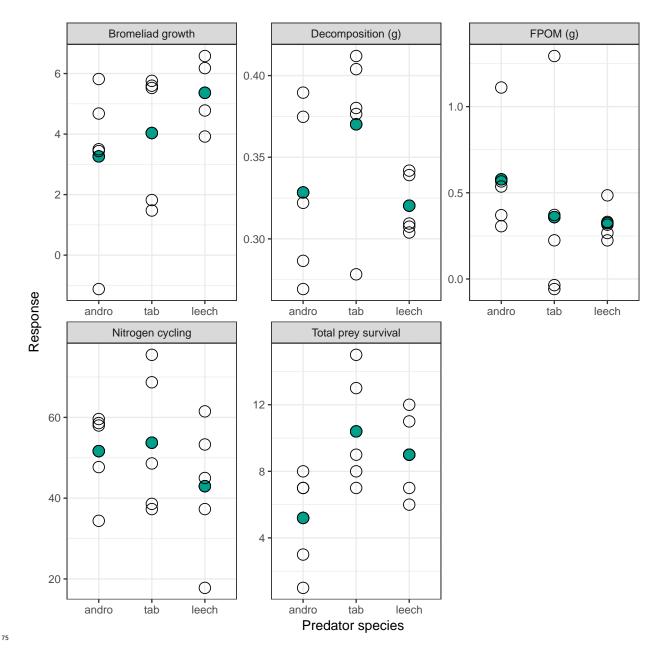


Figure S4: The effect of increasing phylogenetic signal on each of our response variables in our community experiment. In each treatment category, *Leptagrion elongatum* is paired with another predator; these treatments are ordered by increasing phylogenetic diversity. For clarity the axes are labelled with abbreviations for species names: andro = *Leptagrion andromache*, tab = Tabanidae, leech = Hirudinidae.

80 Correlation matrix

Table S11: Correlation matrix of all variables in our experiment. n = 30 for most of these values. Bold type indicates values of Pearson's product moment correlation with p < 0.05.

| | decomp | fine | growth | N | total.surv |
|-------------------------|--------|-------|--------|-------|------------|
| decomp | | 0.08 | -0.33 | 0.13 | 0.15 |
| fine | 0.08 | | 0.02 | -0.12 | -0.09 |
| growth | -0.33 | 0.02 | | 0.1 | 0.11 |
| N | 0.13 | -0.12 | 0.1 | | 0.2 |
| total.surv | 0.15 | -0.09 | 0.11 | 0.2 | |

33 References

- Aris-Brosou, S., and Z. Yang. 2002. Effects of models of rate evolution on estimation of divergence dates with special reference to the metazoan 18S ribosomal RNA phylogeny. Systematic Biology 51:703-714.
- Blair, J. E., and S. B. Hedges. 2005. Molecular clocks do not support the Cambrian explosion. Molecular Biology and Evolution 22:387–390.
- Blair, J. E., P. Shah, and S. B. Hedges. 2005. Evolutionary sequence analysis of complete eukaryote genomes.

 BMC bioinformatics 6:53.
- Foley, D. H., J. H. Bryan, D. Yeates, and A. Saul. 1998. Evolution and systematics of Anopheles: Insights from a molecular phylogeny of Australasian mosquitoes. Molecular Phylogenetics and Evolution 9:262–275.
- Gaunt, M. W., and M. A. Miles. 2002. An Insect Molecular Clock Dates the Origin of the Insects and Accords with Palaeontological and Biogeographic Landmarks. Molecular Biology and Evolution 19:748–761.
- Otsuka, J., and N. Sugaya. 2003. Advanced formulation of base pair changes in the stem regions of ribosomal RNAs; its application to mitochondrial rRNAs for resolving the phylogeny of animals. Journal of Theoretical Biology 222:447–460.
- Peterson, K. J., J. A. Cotton, J. G. Gehling, and D. Pisani. 2008. The Ediacaran emergence of bilaterians:
 Congruence between the genetic and the geological fossil records. Philosophical Transactions of the Royal
 Society of London. Series B, Biological Sciences 363:1435–1443.
- Peterson, K. J., J. B. Lyons, K. S. Nowak, C. M. Takacs, M. J. Wargo, and M. A. McPeek. 2004. Estimating metazoan divergence times with a molecular clock. Proceedings of the National Academy of Sciences of the United States of America 101:6536–6541.
- Regier, J. C., J. W. Shultz, and R. E. Kambic. 2005. Pancrustacean phylogeny: Hexapods are terrestrial crustaceans and maxillopods are not monophyletic. Proceedings. Biological Sciences / The Royal Society 272:395–401.
- Wiegmann, B. M., D. K. Yeates, J. L. Thorne, and H. Kishino. 2003. Time flies, a new molecular time-scale
 for brachyceran fly evolution without a clock. Systematic Biology 52:745–756.