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Maximum rooting depth of vegetation types at the global scale

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Abstract The depth at which plants are able to grow roots has important implications for the whole ecosystem hydrological balance, as well as for carbon and nutrient cycling. Here we summarize what we know about the maximum rooting depth of species belonging to the major terrestrial biomes. We found 290 observations of maximum rooting depth in the literature which covered 253 woody and herbaceous species. Maximum rooting depth ranged from 0.3 m for some tundra species to 68 m for Boscia albitrunca in the central Kalahari; 194 species had roots at least 2 m deep, 50 species had roots at a depth of 5 m or more, and 22 species had roots as deep as 10 m or more. The average for the globe was 4.6±0.5 m. Maximum rooting depth by biome was 2.0±0.3 m for boreal forest, 2.1±0.2 m for cropland, 9.5±2.4 m for desert, 5.2±0.8 m for sclerophyllous shrubland and forest, 3.9±0.4 m for temperate coniferous forest, 2.9±0.2 m for temperate deciduous forest, 2.6±0.2 m for temperate grassland, 3.7±0.5 m for tropical deciduous forest, 7.3±2.8 m for tropical evergreen forest, 15.0±5.4 m for tropical grassland/savanna, and 0.5±0.1 m for tundra. Grouping all the species across biomes (except croplands) by three basic functional groups: trees, shrubs, and herbaceous plants, the maximum rooting depth was 7.0±1.2 m

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E.-D. Schulze Lehrstuhl Pflanzenökologie, Universität Bayreuth, Postfach 101251, D-954440 Bayreuth, Germany for trees, 5.1±0.8 m for shrubs, and 2.6±0.1 m for herbaceous plants. These data show that deep root habits are quite common in woody and herbaceous species across most of the terrestrial biomes, far deeper than the traditional view has held up to now. This finding has important implications for a better understanding of ecosystem function and its application in developing ecosystem models.

Key words Deep roots function \cdot Terrestrial vegetation \cdot Biomes \cdot Plant forms \cdot Root depth

Introduction

There is good evidence that some plant species are able to send roots very deep in the soil. This pattern is indicated by plants that grow well into the summer drought and by desert plants that grow for years with minimal or no rainfall (Batanouny and Abdel Wahab 1973; Poole and Miller 1975). In fact, survivorship of some species in arid systems has been shown to depend completely on a plant's ability to tap water from permanent water tables, which are sometimes located at depths of 18 m or more (Rawitscher 1948; Lewis and Burghy 1964). In addition, there have been direct observations of roots at depths below 2-3 m in caves, road cuts, mine shafts and trenches, and in some instances, roots of woody species have been seen exceptionally deep in the soil. This is the case of Boscia albitrunca and Acacia erioloba whose roots have been found at a depth of 68 m and 60 m, respectively, in the central Kalahari, Botswana (Jennings 1974), and the case of mesquite roots (Prosopis juliflora) found at 53 m deep in the Sonoran Desert, United States (Phillips 1963). Similarly, Stone and Kalisz (1991) reported 11 tree species rooted below 20 m depth. Hence, we know of the potential of some species to have very deep roots at few sites, yet very little is known about how common the habit of deep rooting is across species and environments.

There are two main reasons why this below-ground aspect of ecosystem structure, with its important functional implications, has been under-emphasized. First of all, there are a number of studies on root biomass distri-

bution that show that most of the root biomass occurs within the first 50 cm of the soil, and that only a minimal fraction reaches depths below that depth (for a recent review see Jackson et al. 1996). Therefore, it has been assumed that a good understanding of the role of the root system regarding structure and function at the ecosystem level can be achieved by studying only the first 0.5 m of soil. Secondly, after a whole century of research on root systems, the means of obtaining data on root distribution and structure has not changed substantially: methods include the manual digging of trenches, the use of various mechanical excavation devices, dynamite, or high pressure water. When it comes to looking at patterns of maximum rooting depth, some of that technology is not even sufficient to provide access to deeper soil layers.

The functional significance of deep roots and their contribution to whole-ecosystem processes is still poorly understood. However, there is an increasing body of research in this field that shows the major role of deep roots, particularly for ecosystem water fluxes, as well as for carbon and nutrient cycling (Nepstad et al. 1994; Fisher et al. 1994; Richter and Markewitz 1985; Trumbore et al. 1995; Dawson 1996; Schulze et al. 1996).

The main objective of this review is to summarize what we know about the maximum rooting depth of the major terrestrial biomes ranging from tundra to tropical forest. The data set presented here provides information on plant structure which is relevant for a better mechanistic understanding of ecosystem function.

Methods

We selected references which had species- or community-level information on root depth below 1.0 m, except for the tundra biome for which we considered all depths because permafrost usually limits root growth beyond 30-50 cm. Here we included references from journal papers, books, reports, and unpublished data when relevant, which cover all continents except Antarctica. The major biomes we considered were: boreal forest, croplands, desert, sclerophyllous shrubland and forest, temperate coniferous forest, temperate deciduous forest, temperate grassland, tropical deciduous forest, tropical evergreen forest, tropical grassland/savanna, and tundra. The species were grouped by biome which means that in some instances two different functional groups, such as grasses and shrubs, may be in the same biome category. This was the case of the tropical grasslands and savannas where both herbaceous and woody species occur together. Similarly, in the temperate grassland we also found a few common shrub species along with the bulk of herbaceous plants. Finally, root data for the commonest agricultural crops were collected, including wheat, soybean, alfalfa, barley, and a few other species.

For each rooting depth observation, we recorded the species from which the observation was made, and the community's dominant species when roots were not identified at the species level. For most of the references, the maximum root depth observed corresponded with the depth of the trench, road cut, mine pit, or other excavation, and it is safe to say that roots probably reached much deeper layers than those recorded. Almost all the data presented here came from direct observations of roots in road cuts, mine shafts, open-cut mines and trenches, and only a few values were inferred from the results of isotopic trace studies or plant and soil water potential measurements. Finally, we also recorded the soil type or any soil textural attribute available to characterize the soil environment in which roots were growing.

Results and discussion

Maximum rooting depth across biomes

We compiled a total of 290 observations of rooting depth which covered 253 different plant species from 11 biomes around the world. From this data set, 194 species had roots at least 2 m deep, 50 species had roots at a depth of 5 m or more, and 22 species had roots as deep as 10 m or more (Appendix 1). The average maximum rooting depth for the globe was 4.6±0.5 m, and the individual maximum rooting depth was 68 m for Boscia albitrunca, the roots of which were found during well drilling in deep sandy soils in the central Kalahari, Botswana (Jennings 1974). The ten deepest rooting species were in decreasing order: Boscia albitrunca (68 m), Acacia erioloba (60 m), Prosopis juliflora (53 m), Eucalyptus marginata (40 m), Retama raetam (20 m), Tamarix aphylla (20 m), Andira humilis (18 m), Alhagi maurorum (15 m), Prosopis farcta (15 m), and Prosopis glandulosa (15 m).

Figure 1 shows the maximum rooting depth for all species across biomes in which only the deepest rooting depth is plotted when a given species has more than one observation. Maximum rooting depth by biome was 2.0 ± 0.3 m (n=6; highest value = 3.3 m) for boreal forest, 2.1 ± 0.2 m (n=17; highest value = 3.7 m) for cropland, 9.5 ± 2.4 m (n=22; highest value = 53 m) for desert, 5.2 ± 0.8 m (n = 57; highest value = 40) for sclerophyllous shrubland and forest, 3.9 ± 0.4 m (n = 17; highest value = 7.5 m) for temperate coniferous forest, 2.9 ± 0.2 m (n = 19; highest value = 4.4 m) for temperate deciduous forest, 2.6 ± 0.2 m (n = 82; highest value = 6.3 m) for temperate grassland, 3.7 ± 0.5 m (n = 5; highest value = 4.7 m) for tropical deciduous forest, 7.3 ± 2.8 m (n = 5; highest value = 18 m) for tropical evergreen forest, 15.0±5.4 m (n = 15; highest value = 68 m) for tropical grassland/savanna, and 0.5 ± 0.1 m (n = 8; highest value = 0.9 m) for tundra.

Grouping all the species across biomes (except croplands) by three basic functional groups: trees, shrubs, and herbaceous plants, the maximum rooting depth was 7.0 ± 1.2 m (n=82) for trees, 5.1 ± 0.8 m (n=69) for shrubs, and 2.6 ± 0.1 m (n=85) for herbaceous plants (Fig. 2).

Although differences are large among biomes, there are also important departures from the mean rooting depth pattern within a biome. In the boreal forest, for instance, the water table usually limits the downward growth of roots of *Larix laricina* and *Picea mariana*, whose roots are commonly found no deeper than 0.3 m. Other species, however, do have the capacity to grow below the water table down to a depth of 2 m (Strong and La Roi 1983).

Plants from arid environments or from environments with a long dry season showed the deepest rooting habits of all. The presence of water at deep layers makes it possible for some plants to survive in the rainshadow environments by tapping water from layers as deep as 53 m in the desert of the southwestern United States (Phillips

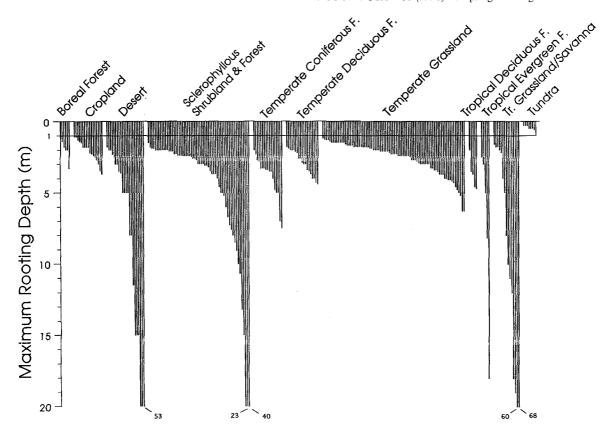


Fig. 1 Reported species maximum rooting depth (m) grouped by terrestrial biome. When there are more than one observations for a given species, only the maximum value is plotted

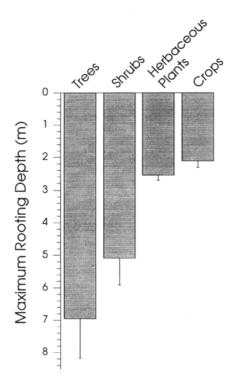


Fig. 2 Mean and SE of reported maximum rooting depth (m) by three major functional groups (trees, shrubs, and herbaceous plants) and crops

1963), and from 68 m deep, possibly even from 140 m deep where the water table was located, in the dry savanna of the central Kalahari (Jennings 1974). Likewise a group of species which also has a consistent pattern of deep rooting is that of the sclerophyllous trees, mostly made up of *Eucalyptus* spp. and *Quercus* spp. from the various Mediterranean regions of the world. The mean maximum rooting depth for sclerophyllous trees is 12.6 ± 3.4 m (n=11), with *E. marginata* in Australia the deepest of all at about 40 m (Dell et al. 1983). Sclerophyllous shrubs, although with deep rooting habits, has a shallower rooting pattern with a maximum rooting depth of 3.5 ± 0.3 m (n=48).

It is generally thought that roots in the evergreen tropical forest tend to be very shallow, but in this review the mean maximum rooting depth of 6 observations is 6.5±2.5 m. The only study that presented data from a depth beyond a few meters in the tropical forests of Brazil found roots all the way down to 18 m deep (Nepstad et al. 1994).

Another surprising result is the depth to which roots of herbaceous plants can descend, that was 2.4 ± 0.1 m as average in this review. Weaver (1919) has published the most complete study to date, on rooting depth habits of herbaceous plants in a prairie in Nebraska, United StatesA. Of 33 species he studied, 18 species have roots that extend beyond depths of 1.5 m, most of them between 2.1 m and 2.7 m, and few to a maximum depth of from 4 m to 6.1 m.

These results offer plenty of evidence that many plant species have the capacity for deep rooting in the soil, and

they provide enough data to challenge the dogma that plants are shallow rooted. Here we have presented, however, data on maximum rooting depth for individuals with the greatest depth. This value represents the observed maximum capacity of a given species to send roots deep into the soil, depths which may be reached by a small number of species and/or individuals within a community. In addition, an average or community weighted maximum rooting depth would also be functionally significant, yet data regarding this are hardly available for any biome. To illustrate the differences between absolute and average maximum rooting depth we shall present data from the root atlas published by Kutschera (1960). A random selection of 69 dicotyledonous species from grasslands in Mid Europe have an average maximum rooting depth of 1.1 m, the average of the 10 deepest species is 4.2 m and the absolute maximum rooting depth is 6.3 m. The average maximum rooting depth, which is the measurement most relevant to ecosystem functioning, will depend on species composition and density, and soil characteristics, all of which are fairly variable in space.

Getting very deep

Plants show a variety of root types through which they have access to deep soil layers. The most common are tap roots, sinker roots and obliquely descending lateral roots, all of them important adaptations for reaching deep soils. The phenotypic expression of these root types is species dependent, but environmental conditions may completely change root structure, architecture, and depth to which roots are able to descend (Feldman 1984). Tap roots are probably the most specialized root type to access and transport water from deep soil horizons. Tap roots are very common across species and they were found in up to 75% of tropical trees (Klinge 1973), in 73 of 100 Mediterranean woody species (Canadell and Zedler 1995), and in 19 of 30 herbaceous species in the Rocky Mountains foothills, United States (Holch et al. 1941).

The downward growth of roots can be limited by a variety of factors, such as soil bulk density or shallow bedrock, but probably the most efficient barriers are horizontally stratified layers of shale or clay, permafrost, and water table (Dennis et al. 1978; Bennie 1991). There is a common notion that deep roots are mainly limited to sandy loose soils where mechanical impedance to root penetration is least. On the contrary, we have reported in this review a number of examples in which plants have found their way down to very deep layers, even in compact clay and rocky soils, and through hard pans (Appendix 1).

Bedrock and heavy clay soils allow varying degrees of deep root penetration through highly weathered material or through a network of cracks, fissures and channels. Channels, or low resistance pathways, are permanent features of the soil profile, and it has been suggested that they result from dissolution of laterite by humic acid produced by the root itself (Plum and Gosting 1973). Gaiser (1952) found more than 10,000 cavities and root channels per hectare in a hardwood forest in Ohio, United States, pathways that can be reused and expanded by each new generation of trees. Hence, the soil volume should be viewed as a complex network of fissures, cracks and channels on which new root growth largely depends. It has even been suggested that soil compaction in forests may not affect the overall forest productivity, provided that sufficient low resistance pathways allow adequate root development (Nambiar and Sands 1992).

Roots have also been observed penetrating through hard pans and caliche layers in a variety of systems (Silva et al. 1989; Dawson 1993; Day 1994), and into rocks through fissures and cracks (Hellmers et al. 1955, Davis and Pase 1977). Pre-existing old tree channels and earthworm tunnels have also been shown to be important in the downward root development in crop systems (Nambiar and Sands 1992; Nicoullaud et al. 1994).

Finally, some plants find their way deep into the soil by penetrating directly through the bedrock. This phenomenon has been reported for several Mediterranean woody species growing on porous calcareous soils in Israel (Oppenheimer 1958; Orshansky 1951).

Ecological significance of deep roots

Although a small fraction of root biomass might be found at depths below 1 m, the functional significance of those roots may nevertheless be most important for ecosystem water and carbon fluxes, and nutrient cycling.

The water extracted by plants during the wet season comes from shallow layers where the root density is highest. However, as those layers dry there is a progressive shift towards using deeper water, which allows plants to keep stomata open and extend growth far into the dry season (for review see Gardner 1983). Although we know of the differential water sources in the soil profile, there are very few studies which have quantified the contribution of deep water to the whole ecosystem fluxes. Gregory et al. (1978) showed for winter wheat that few roots below 1 m (about 3% of the total root weight) were responsible for supplying 20% of the transpired water during dry periods. In an Amazonian tropical forest, Nepstad et al. (1994) found that had not considered roots deeper than 2 meters they would have underestimated evapotranspiration by >60\% during the dry season. The water available to plants stored below 2 m in the soil provided >75% of the water extracted from the entire soil profile.

There is also plenty of evidence that plants with different rooting habits show different seasonal courses of water potential, and that the duration of water stress and the distribution of soil moisture with depth will determine whether a species can succeed in a particular environment (Davis and Mooney 1986; Crombie et al. 1988; Sala et al. 1989; Hodgkinson 1992).

For some species (e.g., phreatophytes), survival in arid systems depends exclusively on the capacity to send roots to permanent water tables, as in the case of Prosopis tamarugo in the virtually rainless Atacama Desert in Chile (Mooney et al. 1980). Stone and Kalisz (1991) gathered thirty references of plants having contact with water tables at depths from 1.5 to 35 m. In these cases, even if a very small fraction of the roots are tapping water from the water table, the amount of water transferred into the plant may be large. Reicosky et al. (1964) showed that roots tapping water from the water table are hundreds of times more efficient in absorbing it than roots in drier soil. Furthermore, tap roots often show cross sections with a high number of vessels per unit area, indicating a major water transport function (Higgins et al. 1987; see also Pate et al. 1995).

The functional significance of deep roots for water flux in ecosystems under high evaporative conditions has been shown regarding the "hydraulic lift" mechanism which has been reported for several species (Richards and Caldwell 1987; Caldwell and Richards 1989; Dawson 1993). During the night roots take up water from deep soil layers which is released from shallow roots back to the soil in the upper layers. The water is reabsorbed during the next day by the same plants and by shallow-rooted neighbors with no access to deep water. This mechanism has important ecological significance, allowing plants to maintain high transpiration rates during dry periods. Caldwell and Richards (1989) showed that hydraulic lift was responsible for a 30–50% increase of the daytime canopy water flux in artificial mixtures of Artemisia tridentata and Agropyron desertorum (see also Dawson 1996).

Unlike water relations, much less is known about the contribution of deep soil nutrients to the overall plant nutritional demands. Richter and Markewitz (1995) showed the importance of biological processes in weathering materials in a 8 m-soil profile of a Pinus taeda forest in South Carolina, United States; the biological processes were tightly associated with soil influenced by root activity (rhizosphere) all along the soil profile. The importance of deep roots for ecosystem nutrient cycling has also been shown for tropical soils with seasonal drought, Cerrado (Schachtschabel et al. 1992). Nitrate salts from mineralization of organic matter cannot be fully utilized by the vegetation early in the rainy season, and so, are washed out of the top soil down to deep soil horizons. There, nitrate is immobilized by the positive charge balance of Fe³⁺ and Al³⁺ found at depths of 1.6 m or more; deep roots will then have access to this nitrate store later in the growing season.

In the deep rhizosphere of *Prosopis glandulosa* in the Chihuahuan desert, United States, a variety of microarthropod taxa has been found down to a depth of 13 m (Silva et al. 1989). The abundance of microarthropods

was positively correlated with root biomass, which suggests that deep rhizosphere processes such as decomposition and mineralization operate in a similar way to those processes in shallow layers. It is also known that plant-feeding nematodes, which are found deep in the rhizosphere, increase nodulation and nitrogen fixation (Huang 1987), and provide infection sites for vesicular-arbuscular mycorrhizal fungi (Freckman and Virginia 1989). In fact, Jenkins et al. (1988) found N₂-fixing root nodules at a depth of 7 m in the Chihuahuan desert.

Deep roots, in addition to extract water and contribute to the cycling of nutrients, also provide carbon to the soil. In an Amazonian tropical forest Nepstad et al. (1994) found that deep soil layers below 1 m contain large active carbon stocks, 15% of which turns over on annual to decadal timescale. The possession of an active carbon cycle at depth seems to be fairly common in the highly wheathered soils in terra firme tropical forest of Amazonia (Trumbore et al. 1995), but almost nothing is known about how common it might be in other biome types.

Ecosystem models which predict carbon sequestration have conventionally used root functional depths between 0.3 m and 2.0 m, which are usually used as fixed factors that do not change or only change for different ecosystem types. The depth at which roots will decay and decompose is essential for determining the ultimate fate of that carbon, and therefore, the capacity of carbon sequestration by different ecosystems. Fisher et al. (1994) showed that increased abundance of introduced deeprooted grasses in the tropical South American savannas account for an increased sequestration of 100–507 Mt carbon per year, which could explain a substantial part of the missing carbon-sink (Siegenthaler and Sarmiento 1993).

In this review we have shown that deep root habits are quite common in woody and herbaceous species across most of the terrestrial biomes. Roots commonly reach far deeper into the soil than the traditional view has held up to now. This structural trait has important implications for ecosystem water fluxes, as well as for carbon and nutrient cycling, and hence should be appropriately taken into account in the development of ecosystem models.

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Appendix 1 Reported maximum rooting depth (m) by species with soil type, country and reference grouped by biome

| Species/ dominant species ^a | Maximum rooting depth (m) | Soil type | Country | Reference |
|--|---------------------------|---|-------------------------|---|
| BOREAL FOREST | | | | |
| Latrix laricina | 1.2 | medium-coarse sand/podzol | S-Canada | Bannan 1940 |
| Latrix sibirica | 1.8 | medium-loamy | Russia | Verzunov 1980 |
| Picea glauca | 1.8 | medium-loamy | Russia | Verzunov 1980 |
| Pinus banksia | 1.2 | medium-coarse sand/podzol | S-Canada | Bannan 1940 |
| Pinus banksiana | 2.0 | aeolian sands/Eutric brunisol | S-Canada | Strong and La Roi 1983 |
| Pinus contorta | 3.3 | _ | S-Canada | Horton 1958 |
| Populus tremuloides | 2.0 | sandy substrate | S-Canada | Strong and La Roi 1983 |
| CROPS | | | | *** |
| Andropogon sorghum | 1.1 | lowland silt loam | Nebraska, USA | Weaver 1926 |
| Avena sativa | 1.8 | | Kansas, USA | Weaver 1926 |
| Beta vulgaris | 1.8 | sandy loam | Nebraska, USA | Weaver 1926 |
| Bromus inermis | 1.1 3.5 | silty-clay to clay-loam alluvial | Canada S-Canada | Leyshon 1991 Lawrence 1975 |
| Elymus angustus | 1.8 | _ | S-Canada S-Canada | Lawrence 1975 |
| Elymus junceus Glycine max | 1.8 | – Muir silt loam | Kansas, USA | Mayaki et al. 1976 |
| Helianthus annuus | 2.7 | lowland silt loam | Nebraska, USA | Weaver 1926 |
| Helianthus annuus | 2.7 | Muir silt loam from alluvium | Kansas, USA | Jaafar et al. 1993 |
| Hordeum vulgare | 1.3 | Lowland silt loam | Nebraska, USA | Weaver 1926 |
| Hordeum sp. | 2.2 | loamy sand/Xeric Psamment | W-Australia | Hamblin and Tennant 1987 |
| Lupinus angustifolius | 2.5 | loamy sand/Xeric Psamment | W-Australia | Hamblin and Tennant 1987 |
| Medicago sativa | 3.7 | _ | Nebraska, USA | Weaver 1926 |
| Secale cereale | 1.5 | silt loam/hard clayey subsoil | Nebraska, USA | Weaver 1926 |
| Solanum tuberosum | 1.4 | mellow loess soil | Nebraska, USA | Weaver 1926 |
| Triticum aestivum | 1.0 | _ | England | Welbank et al. 1974 |
| Triticum aestivum | 1.4 | Muir silt loam | Kansas, USA | Chaudhuri et al. 1990 |
| Triticum aestivum | 1.5 | lowland silt loam | Nebraska, USA | Weaver 1926 |
| Triticum aestivum | 1.8 | _ | Mid Europe | Kutschera 1960 |
| Triticum aestivum | 3.0 | loamy sand/Xeric Psamment | W-Australia | Hamblin and Tennant 1987 |
| Triticum durum | 2.3 | loess soil | Nebraska, USA | Weaver 1926 |
| Zea mays | 1.3 | deep clay loam | E-France | Pages and Pellerin 1994 |
| Zea mays | 2.4 | _ | Nebraska, USA | Weaver 1926 |
| DESERT | | | · 1 | Cl 1: 1049 |
| Alhagi maurorum | 15.0 | river banks | Israel | Shmueli 1948 |
| Artemisia monosperma | 5.0 | sand dunes | Israel | Zohary and Fahn 1952 Branson et al. 1976 |
| Artemisia tridentata | 1.8 | shale/sandstone bedrock | Colorado, USA | Richards and Caldwell 1987 |
| Artemisia tridentata | 2.2 2.3 | loamy-skeletal/Haploxerolls | Utah, USA Idaho, USA | Reynolds and Fraley 1989 |
| Artemisia tridentata | 8.0 | aeolian sandy loam alluvia soils/run-on habitats | Israel | Zohary 1961 |
| Atriplex halimus Chrysothamnus vicidiflorus | 2.0 | aeolian sandy loam | Idaho, USA | Reynolds and Fraley 1989 |
| Franseria deltoidea | | wash with hardpan (caliche) | Arizona, USA | Cannon 1911 |
| Hammada salicornica | 2.5 | sand dunes | Israel | Zohary and Orshan 1949 |
| Leptadenia pyrotechnica | 11.5 | sandy and silty/clay at depth | Egytp | Batanouny and Wahab 1973 |
| Leymus cinereus | 2.0 | aeolian sandy loam | Idaho, USA | Reynolds and Fraley 1989 |
| Mulinum spinosuma | 3.0 | fine sand/caliche layer at 0.6 m | S-Argentina | Schulze et al. 1996 |
| Nassauvia glomerulosaª | 3.0 | sandy loam/caliche at 0.7 m | S-Argentina | Schulze et al. 1996 |
| Nitraria retusa | 5.0 | sandy | Israel | Ginzburg 1966 |
| Ochradenus baccatus | 5.0 | sandy | Israel | Ginzburg 1966 |
| Prosopis farcta | 15.0 | river banks | Israel | Zohary and Orshan 1949 |
| Prosopis glandulosa | 2.0 | Nuvalde clay loam | Texas, USA | Heitschmidt et al. 1988 |
| Prosopis glandulosa | 6.0 | clay loam/sand, clay at depth | California, USA | Nilsen et al. 1983 |
| Prosopis glandulosa | 12.0 | sandy/Torrifluvent | New Mexico, USA | Freckman and Virginia 1989 |
| Prosopis glandulosa | 15.0 | clay loam | New Mexico, USA | Silva et al. 1989 |
| Prosopis juliflora | 53.0 | _ | Arizona, USA | Phillips 1963 Mooney et al. 1980 |
| Prosopis tamarugo Prosopis velutine | 3.5 8.0 | _ | N-Chile Arizona, USA | Mooney et al. 1980 Cannon 1911 |
| Prosopis vetutine Retama raetam | 20.0 | - sand dunes | Israel | Zohary and Fahn 1952 |
| Tamarix aphylla | 20.0 | alluvial soils/run-on habitats | Israel | Zohary 1961 |
| Tamarix apnyna Tamarix pentantra | 3.6 | alluvial banks | Arizona, USA | Gary 1963 |
| Zilla spinosa | 5.0 | - | Israel | Ginzburg 1966 |

| Species/ dominant species ^a | Maximum rooting depth (m) | Soil type | Country | Reference |
|--|---------------------------|---|------------------------------------|--|
| SCLEROPHYLLOUS SHRU | BLAND AND | FOREST | | |
| Shrubs | | | | |
| Adenostoma fasciculatum | 2.4 | silt sandy | California, USA | Hanes 1965 |
| Adenostoma fasciculatum | 7.6 | sandy loam on anorthosiste | California, USA | Hellmers et al. 1955 |
| Adenostoma sparsifolium | 2.4 | silt sandy | California, USA | Hanes 1965 |
| Arbutus unedô | 3.5 | sandy loam | NE Spain | J Canadell, unpublished work |
| Arctostaphylos glandulosa | 5.2 | sandy loam on granodiorite | California, USA | Hellmers et al. 1955 |
| Arctostaphylos glauca | 2.6 | sandy loam on granodiorite | California, USA | Hellmers et al. 1955 |
| Arctostaphylos glutinosa | 2.5 | shallow on fractured shales | California, USA | Davis 1972 |
| Arctostaphylos pallida Banksia marginata | 4.0 2.4 | shallow on fractured shales sandy | California, USA SE-Australia | Davis 1972 Specht and Rayson 1957 |
| Banksia marginala Banksia ornata | 2.4 | sandy | SE-Australia | Specht and Rayson 1957 Specht and Rayson 1957 |
| Baccharis pilularis | 3.2 | packed sand like a rock | California, USA | Wright 1928 |
| Banksia spp. | 5.0 | podsolized sand | SW-Australia | Low and Lamont 1990 |
| Calytrix flavescens | 2.0 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 |
| Casuarina muelleriana | 2.0 | sandy | SE-Australia | Specht and Rayson 1957 |
| Casuarina pusilla | 2.4 | sandy | SE-Australia | Specht and Rayson 1957 |
| Ceanothus leucodermis | 3.7 | sandy loam on granodiorite | California, USA | Hellmers et al. 1955 |
| Ceanothus megacarpus | 2.4 | sandstone with fissures | California, USA | Thomas and Davis 1989 |
| Ceanothus oliganthus | 1.8 | clay loam on diorite | California, USA | Hellmers et al. 1955 |
| Ceanothus spinosus Daviesia brevifolia | 3.1 2.0 | sandstone with fissures sandy | California, USA | Thomas and Davis 1989 |
| Eremaea beaufortioides | 6.0 | alluvial sand with colluvium | SE-Australia SW-Australia | Specht and Rayson 1957 |
| Eremaea pauciflora | 2.4 | grey sands with hardpan | SE-Australia | Hnatiuk and Hopkins 1980 Dodd et al. 1984 |
| Erica arborea | 2.0 | sandy loam | NE-Spain | J. Canadell, unpublished wor |
| Hibbertia hypericoides | 2.1 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 |
| Jacksonia floribunda | 3.1 | grey sands with hardpan | SW Australia | Dodd et al. 1984 |
| Jacksonia furcellata | 2.0 | grey sands with hardpan | SW Australia | Dodd et al. 1984 |
| Laudonia behrii | 2.0 | sandy | SE-Australia | Specht and Rayson 1957 |
| Leptospermum myrsinoides | 2.3 | sandy | SE-Australia | Specht and Rayson 1957 |
| Leucadendron salignum Lithraea caustica | 3.0 5.0 | loamy medium sand | South Africa | Higgins et al. 1987 |
| Melaleuca scabra | 2.0 | grey sands with hardpan | Central Chile SW-Australia | Giliberto and Estay 1978 |
| Melaleuca seriata | 2.1 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 Dodd et al. 1984 |
| Petrophile linearis | 2.0 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 |
| Photinia arbutifolia | 2.1 | clay loam on diorite | California, USA | Hellmers et al. 1955 |
| Phyllota pleurandroides | 2.3 | sandy | SE-Australia | Specht and Rayson 1957 |
| Phyllota remota | 2.4 | sandy | SE-Australia | Specht and Rayson 1957 |
| Protea neriifolia | 3.0 | loamy medium sand | South Africa | Higgins et al. 1987 |
| Protea repens | 3.0 | loamy medium sand | South Africa | Higgins et al. 1987 |
| Quercus calliprinosa | 4.6 | terra-rossa on limestone | Israel | Shachori et al. 1967 |
| Quercus dumosa Quercus turbinella | 8.5 6.4 | clay loam on diorite fracturated granite | California, USA | Hellmers et al. 1955 |
| Quercus turbinella | 9.1 | alluvial and redish brown | Arizona, USA Arizona, USA | Davis and Pase 1977 |
| Quillaja saponaria | 8.0 | - | Central Chile | Saunier and Wagle 1967 Giliberto and Estay 1978 |
| R̃hus glabra | 6.7 | loess hills | Nebraska, USA | Weaver 1919 |
| Rhus laurina | 5.4 | sandstone with fissures | California, USA | Thomas and Davis 1989 |
| Rhus laurina | 13.2 | _ | California, USA | DeSouza et al. 1986 |
| Salvia apiana | 1.5 | coarse, loose gravel | California, USA | Hellmers et al. 1955 |
| Scholtzia involucrata | 1.9 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 |
| Spyridium subochreatum Stirlingia latifolia | 1.9 | sandy | SE-Australia | Specht and Rayson 1957 |
| Xanthorrhoea australis | 2.6 2.4 | grey sands with hardpan | SW-Australia | Dodd et al. 1984 |
| Trees | 2.4 | sandy | SE-Australia | Specht and Rayson 1957 |
| Eucalyptus marginata | 15.0 | lateritic, sandy-clay at depth | SW-Australia | Kimber 1974 |
| Eucalyptus marginata | 20.0 | | SW-Australia | Carbon et al. 1980 |
| Eucalyptus marginata | 40.0 | fissured granite, clay subsoil | SW-Australia | Dell et al. 1983 |
| Eucalyptus regnans | 2.7 | - | SW-Australia | Incoll 1969 |
| Eucalyptus signata | 3.0 | sandy | NE-Australia | Westman and Rogers 1977 |
| Eucalyptus sp. | 10.0 | sand dunes | NE-Australia | Westman and Rogers 1977 |
| Quercus agrifolia Quercus chrysolepis | 10.7 7.3 | eandy loam on aroundiquite | California, USA | Cannon 1914 |
| Quercus chrysotepts Quercus douglasii | 7.3 3.7 | sandy loam on granodiorite alluvial loam | California, USA California, USA | Hellmers et al. 1955 |
| Quercus douglasti Quercus ilex | 3.7 | sandstone | NE-Spain | Cannon 1914 J. Canadell, unpublished worl |
| Quercus wislizenii ^a | 22.9 | fractured rock | California, USA | Lewis and Burgy 1964 |

Appendix 1 (continued)

| Species/ dominant species ^a | Maximum rooting depth (m) | Soil type | Country | Reference |
|---|---------------------------|--|--------------------------------|--|
| TEMPERATE CONIFERO | US FOREST | | | |
| Abies firma | 3.3 | sandy soil | Japan | Karizumi 1979 |
| Picea excelsa | 2.3 | silt loam | Japan | Karizumi 1979 |
| Pinus densiflora | 3.4 | silt loam | Japan | Karizumi 1979 |
| Pinus echinata | 3.3 | sandy soil | New Jersey, USA | Lull and Axley 1958 |
| Pinus elliottii | 3.3 | _ | Florida, USA | van Rees and Comerford 1986 |
| Pinus halepensis ^a | 7.3 | terra-rossa on limestone | Israel | Shachori et al. 1967 |
| Pinus halepensis | 7.5 | weathered granite | NE-Spain | J. Canadell, unpublished work |
| Pinus luchuensis | 3.5 | sandy loam | Japan | Karizumi 1979 |
| Pinus palustris | 4.8 | Norfold sand deep phase | Florida, USA | Heyward 1933 |
| Pinus pinaster | 7.0 5.0 | - waatharad arenita | Australia | Butcher and Havel 1976 |
| Pinus pinea | 3.5 | weathered granite clay loam soil | NE-Spain | J. Canadell, unpublished work Zwieniecki and Newton 1994 |
| Pinus ponderosa Pinus radiata | 2.0 | sandy soil | Oregon, USA S-Australia | Nambiar and Sands 1992 |
| Pinus resinosa | 2.7 | Hinckley coarse sand | New York, USA | White and Wood 1958 |
| Pinus resinosa | 5.0 | sandy outwash | New York, USA | Leaf et al. 1955 |
| Pinus rigida | 2.7 | sandy soil | New Jersey, USA | McQuilkin 1935 |
| Pinus rigida | 3.4 | sandy soil | Japan | Karizumi 1979 |
| Pinus strobus | 2.8 | sandy soil | Japan | Karizumi 1979 |
| Pinus sylvestris | 2.7 | sand overlying chalky drift | United Kingdom | Roberts 1976 |
| Pinus taeda | 2.0 | fullerton and bodine | Tennessee, USA | Harris et al. 1977 |
| Pinus taeda | 4.0 | granite wheathered/Ultisol | S-Carolina, USA | Richter and Markewitz 1995 |
| TEMPERATE DECIDUOU | JS FOREST | | | |
| Acer negundo | 4.0 | upland clay | Missouri, USA | Biswell 1935 |
| Acer saccharum | 3.7 | silty loams with hardpan | New York, USA | Dawson 1993 |
| Carya spp. | 1.8 | sandstone | Ohio, USA | Gaiser 1952 |
| Corylus americana | 3.5 | loess hills | Nebraska, USA | Weaver 1919 |
| Fraxinus japonica | 2.0 | fine texture clay | Japan | Karizumi 1979 Karizumi 1979 |
| Juglans nigra | 3.0 | silt loam | Japan New York, USA | White and Wood 1958 |
| Latrix decidua | 3.4 2.0 | fine silty sand at depth | S-Argentina | Schulze et al. 1996 |
| Nothofagus pumila | 2.6 | orange loam/rocks at depth medium texture | Japan Japan | Karizumi 1979 |
| Platanus orientalis | 2.0 1.9 | silt loam | Japan | Karizumi 1979 |
| Populus nigra Populus sargentii | 2.6 | loam underlain with clay | Missouri, USA | Biswell 1935 |
| Populus tremula | 2.0 | clay subsoil | Sweden | Persson 1975 |
| Populus tremuloides | 2.3 | grey clay | Michigan, USA | Day 1944 |
| Populus tremuloides | 2.9 | sandy loam | Utah, USA | Gifford 1966 |
| Prunus yedoensis | 2.1 | fine texture clay | Japan | Karizumi 1979 |
| Quercus dentata | 4.3 | silt loam | Japan | Karizumi 1979 |
| Quercus macrocarpa | 4.3 | fine-textured loams | Nebraska, USA | Weaver and Kramer 1932 |
| Quercus macrocarpa | 4.4 | upland clay | Missouri, USA | Biswell 1935 |
| Quercus sp-Carya spa | 4.0 | silt loam on sandstone/shale | Virginia, USA | Kochenderfer 1973 Karizumi 1979 |
| Quercus velutina Salix babylonica | 3.0 2.2 | medium texture silt loam | Japan Japan | Karizumi 1979 Karizumi 1979 |
| • | | 3320 204311 | 1 | |
| TEMPERATE GRASSLAN. Agropyron repens | 2.4 | loose sandy | Nebraska, USA | Weaver 1919 |
| Agropyron repens Agropyron smithii | 2.7 | silt loam | Colorado, USA | Weaver 1958 |
| Agropyron spicatum | 1.4 | med. textur. Benge series | Washington, USA | Harris 1967 |
| Agropyron spicatum | 1.5 | silt loam | Washington, USA | Weaver 1919 |
| Amorpha canescens | 5.0 | loose sandy | Nebraska, USA | Weaver 1919 |
| Andropogon furcatus | 1.5 | Judson silt loam | Nebraska, USA | Weaver and Darland 1949 |
| Andropogon furcatus | 2.8 | clay loam | Nebraska, USA | Weaver 1919 |
| Andropogon gerardi | 2.1 | lilt loam | Iowa, USA | Weaver 1958 |
| Andropogon hallii | 1.8 | sandy | Nebraska, USA | Tolstead 1942 |
| Andropogon hallii | 3.0 | sandy | Colorado, USA | Weaver 1958 Weaver 1958 |
| Andropogon scoparius | 1.5 | silt loam | Iowa, USA Colorado, USA | Weaver 1958 Weaver 1919 |
| Andropogon scoparius | 1.8 1.4 | loam sandy loam sandy | Colorado, USA Colorado, USA | Weaver 1919 Weaver 1919 |
| Aragallus lambertii | 3.7 | loam sandy | Colorado, USA Colorado, USA | Weaver 1919 Weaver 1919 |
| Argemone platyceras | 3.7 1.7 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Artemisia frigida Artemisia cana | 2.4 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Atriplex nuttallii | 1.8 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Astragalus crassicarpus | 2.0 | loam soil on hard joint clay | Nebraska, USA | Weaver 1919 |
| Berberis repens | 3.0 | silt loam | Washington, USA | Weaver 1919 |
| Biscutella laevigata | 2.1 | - | Mid Europe | Kutschera 1960 |

Appendix 1 (continued)

| Species/ dominant species ^a | Maximum rooting depth (m) | Soil type | Country | Reference |
|---|---------------------------|------------------------------------|--------------------------------|---|
| Bouteloua curtipendula | 1.7 | silt loam | Colorado, USA | Weaver 1958 |
| Bouteloua gracilis | 1.7 | Colby silt loam | Nebraska, USA | Weaver and Darland 1949 |
| Bouteloua gracilis | 1.8 | silt loam | Colorado, USA | Weaver 1958 |
| Bouteloua gracilisa | 2.1 | _ | Kansas, USA | Albertson et al. 1953 |
| Brauneria pallida | 2.4 | clay loam | Nebraska, USA | Weaver 1919 |
| Buchloe dactyloides | 1.8 | silt loam | Iowa, USA | Weaver 1958 |
| Buchloe dactyloides | 2.0 | Wabash silt loam | Nebraska, USA | Weaver and Darland 1949 |
| Bulbilis dactyloides | 1.9 | alluvial | Nebraska, USA | Weaver 1919 |
| Calamovilfa longifolia | 1.8 | dark borwn soil on shales | S-Canada | Coupland and Johnson 1965 |
| Calamovilfa longifolia | 3.0 | sandy | Colorado, USA | Weaver 1958 |
| Carex arenaria | 1.8 | - ailt laam | Mid Europe | Kutschera 1960 Weaver 1958 |
| Carex filifolia | 1.5 | silt loam | Colorado, USA Germany | Kutschera 1960 |
| Carlina acaulis | 4.1 3.3 | ~ | Germany | Kutschera 1960 Kutschera 1960 |
| Centaurea scabiosa | 3.3 2.4 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Chrysopis villosa | 3.0 | sandy | Canada | Coupland and Johnson 1965 |
| Equisetum arvense Equisetum palustre | 2.5 | sandy | Mid Europe | Kutschera 1960 |
| Eriogonum heracleoides | 2.4 | silt loam | Washington, USA | Weaver 1919 |
| Eriogonum neracteotaes Eriogonum jamesii | 2.3 | loam with some sand | Colorado, USA | Weaver 1919 |
| Eriogonum junesti Eriogonum microthecum | 3.0 | sandy | Colorado, USA | Weaver 1919 |
| Eriogonum microinecum Erodium botrys | 1.3 | gravelly clay loam | California, USA | McKell et al. 1962 |
| Eryngium campestre | 4.2 | - | Germany | Kutschera 1960 |
| Eurotia lanata | 1.8 | dark brown soil on shales | S-Canada | Coupland and Johanson 1965 |
| Festuca arizonica | 1.2 | sandy loam-sandy clay | Colorado, USA | Currie and Hammer 1979 |
| Festuca arizonica | 1.3 | | Colorado, USA | Schuster 1964 |
| Festuca arundinacea | 2.7 | | Germany | Kutschera 1960 |
| Festuca pallescensa | 2.0 | alluvial sandy loam & gravel | S-Argentina | Schulze et al. 1996 |
| Gaillardia aristata | 1.7 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Geranium viscosissimum | 2.9 | silt loam | Washington, USA | Weaver 1919 |
| Grindelia squarrosa | 1.9 | loose sand | Nebraska, USA | Weaver 1919 |
| Heracleum sphondyleum | 2.0 | _ | Mid Europe | Kutschera 1960 |
| Hieracium scouleri | 2.2 | silt loam | Washington, USA | Weaver 1919 |
| Hoorebekia racemosa | 3.4 | silt loam | Washington, USA | Weaver 1919 |
| Kochia prostrata | 6.3 | _ | Germany | Kutschera 1960 |
| Kuhnia glutinosa | 5.2 | - | Nebraska, USA | Weaver 1919 |
| Lepachys pinnata | 1.5 | brown silt loam | Illinois, USA | Sperry 1935 |
| Lespedeza capitata | 2.4 | lower slopes of loess hills | Nebraska, USA | Weaver 1919 |
| Liatris punctata | 2.1 | gravelly | S-Canada | Coupland and Johnson 1965 |
| Liatris punctata | 4.8 | clay | Nebraska, USA | Weaver 1919 |
| Lithospermum gmelini Lupinus ornatus | 2.1 4.0 | sandy silt loam | Nebraska, USA | Tolstead 1942 Weaver 1919 |
| Lupinus ornaius Lygodesmia juncea | 3.0 | | Washington, USA | |
| | 3.0 | sandy dark brown soil on shales | Nebraska, USA S-Canada | Tolstead 1942 Coupland and Johnson 1965 |
| Lygodesmia juncea Lygodesmia juncea | 6.3 | loess | Nebraska, USA | Weaver 1919 |
| Medicago falcata | 4.3 | 10035 | Germany | Kutschera 1960 |
| Muhlenbergia montana | 1.3 | sandy clay loam subsoil | Colorado, USA | Schuster 1964 |
| Onobrychis natrix | 2.3 | - | Mid Europe | Kutschera 1960 |
| Ononis natrix | 2.3 | _ | Germany | Kutschera 1960 |
| Panicum virgatum | 2.7 | loose sand | Nebraska, USA | Weaver 1919 |
| Parthenium integrifolium | 1.8 | brown silt loam | Illinois, USA | Sperry 1935 |
| Petalostemum purpureum | 1.8 | brown silt loam | Illinois, USA | Sperry 1935 |
| Peucedanum cervaria | 4.1 | _ | Germany | Kutschera 1960 |
| Phalaris aquatica | 1.2 | granite | Spain | Joffre et al. 1987 |
| Pimpinella saxifraga | 3.7 | _ | Germany | Kutschera 1960 |
| Potentilla blaschkeana | 2.3 | silt loam | Washington, USA | Weaver 1919 |
| Potentilla fruticosa | 3.0 | gravelly | S-Canada | Coupland and Johnson 1965 |
| Potentilla concinna | 1.8 | dark brown soil on shales | S-Canada | Coupland and Johnson 1965 |
| Psoralea tenuiflora | 1.8 | loose sand | Nebraska, USA | Weaver 1919 |
| Psoralea tenuiflora | 3.7 | loam sandy, silt loam | Colorado, USA | Weaver 1919 |
| Redfieldia flexuosa | 1.5 | sandy | Colorado, USA | Weaver 1958 |
| Ruellia ciliosa | 1.5 | brown silt loam | Illinois, USA | Sperry 1935 |
| Rumex crispus | 3.3 | - , ., | Germany | Kutschera 1960 |
| Senecio riddellii | 1.5 | sandy soil | Nebraska, USA | Tolstead 1942 |
| Silphium integrifolium | 1.7 1.8 | brown silt loam | Illinois, USA | Sperry 1935 |
| | ı x | brown silt loam | Illinois, USA | Sperry 1935 |
| Silphium laciniatum | | | | |
| Silphium laciniatum Solidago canadensis Solidago rigida | 3.4 1.4 | loose sandy brown silt loam | Nebraska, USA Illinois, USA | Weaver 1919 Sperry 1935 |

Appendix 1 (continued)

| Species/ dominant species ^a | Maximum rooting depth (m) | Soil type | Country | Reference |
|---|---------------------------|----------------------------------|----------------------------|---|
| Sporobolus cryptandrus Sporobolus heterolepsis | 1.5 1.5 | deeply eroded loess silt loam | Nebraska, USA Iowa, USA | Weaver 1919 Weaver 1958 |
| Ŝtipa spartea | 1.8 | silt loam | Iowa, USA | Weaver 1958 |
| Taraxacum serotinum | 4.6 | | Germany | Kutschera 1960 |
| Thermopsis rhombifolia | 2.1 | gravelly | S-Canada | Coupland and Johnson 1965 |
| Tradescantia reflexa | 1.6 | brown | Illinois, USA | Sperry 1935 |
| Vernonia baldwinii | 3.5 | loose sand | Nebraska, USA | Weaver 1919 |
| TROPICAL DECIDUOUS F | OREST | | | |
| Antiaris toxicaria | 3.5 | red soil | China | Bang-Xing 1991 |
| Baccaurea ramiflora | 3.7 | red soil | China | Bang-Xing 1991 |
| Gironniera subaequalis | 4.7 | red soil | China | Bang-Xing 1991 |
| Symplocos cochinchinensis | 2.0 | red soil | China | Bang-Xing 1991 |
| Xauthophyllum siamense | 4.6 | red soil | China | Bang-Xing 1991 |
| TROPICAL EVERGREEN F | OREST | | | |
| Apodytes dimidiata | 8.2 | sandy loam on schists | Kenya | Kerfoot 1963 |
| Chlorophora excelsa | 2.0 | ferralitic | Ghana | Mensah and Jenik 1968 |
| Chlorophora excelsa | 3.0 | ferralitic | Ghana | Jenik 1971 |
| Community | 18.0 | clay | Brazil | Nepstad et al. 1994 |
| Community | 5.0 | _ ` | Brazil | Poels 1987 |
| Community | 2.5 | Turraeantho on sandy soil | Ivory Coast | Huttel 1975 |
| TROPICAL GRASSLAND A | ND SAVANNA | | | |
| Acacia erioloba | 60.0 | Kalahari sands | Botswana | Jennings 1974 |
| Anacardium pumilum | 10.0 | _ | Brazil | Ferri 1961 |
| Andira humilis | 18.0 | reddish loamy earth | Brazil | Rawitscher 1948 |
| Andira spp. | 19.0 | - | Brazil | Rawitscher et al. 1943 |
| Aristolachia giberti | 1.8 | redish loamy earth | Brazil | Rawitscher 1948 |
| Boscia albitrunca | 68.0 | Kalahari sands | Botswana | Jennings 1974 |
| Brachiaria brizanthaª | 8.0 | clay | Brazil | Nepstad et al. 1994 |
| Brachystegia sp. | 1.8 | | Zimbabwe | Strang 1969 |
| Capparis sp. | 1.6 | sandy clay loam | Ghana | Okali et al. 1973 |
| Curatella americana | 4.0 | _ | Venezuela | Foldats and Rutkis 1975 |
| Jacaranda decurrensa | 11.0 | | Brazil | Rawitscher et al. 1943 Rutherford 1983 |
| Ochna pulchra | 2.2 12.0 | structureless sand | South Africa Brazil | Nepstad et al. 1994 |
| Panicum maximum ^a | 5.0 | clay Kalahari sands | South Africa | J. Canadell, unpublished work |
| Stipagrostis amabilis ^a Stryphnodendron sp. | 2.0 | reddish loamy earth | Brazil | Rawitscher 1948 |
| • | 2.0 | readish loanly cardi | Diuzn | rawitseller 15 to |
| TUNDRA | 0.3 | | Alaska, USA | Dennis et al. 1978 |
| Cares aquatilisa | 0.5 | permafrost at 40–55 cm | N-Russia | Khodachek 1971 |
| Dryas punctata ^a Dupontia fischeri ^a | 0.3 | organic matter on sediments | Alaska, USA | Dennis 1977 |
| Eriophorum vaginatum ^a | 0.5 | silty soil on permafrost | Alaska, USA Alaska, USA | Wein and Bliss 1974 |
| Eriopnorum vaginaium" Betula nana | 0.5 | permafrost at 50 cm | Alaska, USA | S. Hobbie, unpublished work |
| ьении папа Luzula confusa | 0.3 | loams | N-Canada | Bliss and Svoboda 1984 |
| Luzuta conjusa Salix glauca | 0.5 | permafrost at 45–60 cm | W-Russia | Ignatenko and Khakimzy 1971 |
| Salix planifolia | 0.9 | coarse textured/bottom pit | Colorado, USA | Webber and May 1977 |

^a Maximum rooting depth is not linked to the species name but to the dominant species in the community

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