Are we there yet? The impact of reduced data on the ability to monitor rebuilding for overfished fish stocks

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ABSTRACT

The impacted ability to collect fishery data can be especially problematic when there are limited to no fishery independent data available. Select rockfish stocks off the U.S. west coast are below target biomass and are managed under rebuilding plans that limit the allowable harvest, reducing the opportunity to collect fishery data, where the fishery is the primary source of data for assessment. A simulation was performed with two alternative operating models that applied time-invariant or time-varying parameters to evaluate the impact of reduced data to estimate spawning biomass and biological parameters during rebuilding. Decreased data during rebuilding resulted in increased variation in estimates of spawning biomass and relative spawning biomass, reduced average catch, and increased variation in catch during rebuilding compared to the simulations where data collection was maintained during rebuilding. The inclusion of time-varying parameters that were not accounted for within the estimation method resulted in an increased variation spawning biomass and relative spawning biomass estimates, with the highest increase in variance among estimates occurring during rebuilding when data were reduced or eliminated. Time-varying annual deviations in fishery selectivity resulted in consistently positively biased estimates of size at maximum selectivity by the estimation method.

*Keywords*:

Rebuilding, overfished, simulation, limited data

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# Introduction

Rebuilding overfished stocks requires a reduction in fishing mortality to a level that allows stock biomass to increase. In the United States, federally-managed stocks that fall below a minimum stock size threshold (MSST) are declared overfished and are mandated to be rebuilt to target biomass levels in the shortest amount of time, accounting for biological and environmental conditions (SFA, 1996). This can lead to substantial reductions in fishing effort relative to historical levels. The severity of restrictions during rebuilding can, for some stocks, lead to a situation where the ability to collect data becomes limited over the period when managers are likely most concerned about stock size (i.e. when the stock is under the rebuilding plan).

Data are necessary to determine the extent to which a stock is on track to rebuild. A rebuilding stock should have harvests limited to levels below the replacement yield. However, being able to measure the rate of recovery is still crucial to management, and increased uncertainty due to limited data can impede monitoring. Additionally, biological data are critical to improve estimates of key parameters within stock assessments (e.g. natural mortality, growth, recruitment compensation termed steepness) and can indicate incoming poor or strong year-classes (recruitment), which can impact estimates of relative stock biomass (the ratio of current biomass to unfished biomass) and rebuilding rates. Potential improvements in parameter estimates, and the ability to detect incoming fluctuations in recruitment during rebuilding could be restricted when collection of new biological data is limited due to harvest restrictions.

Overfished rockfish species off the U.S. west coast have experienced large reductions in harvest during rebuilding. One example is yelloweye rockfish (*Sebastes ruberrimus*), which was declared overfished in 2002 (Methot and Piner, 2002). Similar to other rockfish species off the U.S. west coast, catches of yelloweye rockfish were large and unsustainable during the 1980s and early 1990s. Catches of yelloweye rockfish were cut dramatically following the overfished declaration relative to the historical catches, where the allowable catch in the first year of rebuilding was reduced to approximately 10% of the catch observed four years earlier (Stewart *et al*., 2009).

Many species of rockfish are not sampled well by the main fishery-independent survey off the U.S. west coast, either due to the survey’s inability to sample rocky habitat using trawl gear or other restrictions on sampling locations (e.g. rockfish conservation areas or near-shore habitat). Yelloweye rockfish fall into the category of poorly sampled rockfish by the survey, resulting in the majority of historical information (e.g. index of abundance, length, and age data) available for assessment, albeit somewhat limited, coming primarily from recreational and commercial fishery samples. During rebuilding, retention has been prohibited in the recreational fishery, and the limited allowed catch has led to an avoidance behavior by the commercial fishery (Stewart *et al.*, 2009). The most recent assessment cited limited fishery data during rebuilding as a challenge to ‘produce conclusive information about the stock for the foreseeable future’ (Stewart *et al.*, 2009). The limited survey data, along with the reduction of fishery data during rebuilding, presents a challenge for assessment and management.

There have been numerous simulation studies evaluating the impact of data quality and quantity on the performance of stock assessment methods (e.g. Hilborn, 1979; Chen *et al.*, 2003; Yin and Sampson, 2004; Magnusson and Hilborn, 2007; Wetzel and Punt, 2011; Lee *et al.*, 2012). Studies often focus on the ability to estimate either management quantities or biological parameters. However, to date, these studies have not addressed the long-term impact of reduced data on the ability to monitor a stock during rebuilding. This paper simulates an overfished long-lived rockfish stock, such as those that are common to the U.S. west coast, where harvest and the collection of fishery data are restricted during rebuilding. The simulation study addresses three main questions; 1) does limited data result in increased uncertainty impacting the ability to detect when an overfished stock has recovered to management target stock size (i.e. it is rebuilt), 2) are the limited data from the fishery able to detect a shift in fishery selectivity resulting from changing fishing behavior during rebuilding, and 3) how are model estimates of stock size and biological parameters affected during periods of limited data?

# Material and Methods

## General approach

A rockfish life history type common to the U.S. west coast was simulated (Table 1), based on the life history for yelloweye rockfish. Yelloweye rockfish are assumed to have slow population dynamics based on very low natural mortality and recruitment compensation (steepness). However, the operating model was parameterized using higher natural mortality and steepness values to allow for shorter recovery periods (< 100 years) for computation efficiency while still maintaining the characteristics of a rockfish life history.

Two alternative cases were simulated using the operating model to account for the potential impacts of time-invariant versus time-variation in natural mortality and fishery selectivity. The first case, “time-invariant”, involved a fixed natural mortality rate over time and a fixed fishery selectivity curve during the historical, the overfished period, and after the stock rebuilt to target biomass levels. The historical period involved asymptotic fishery selectivity (Fig. 1a). The simulated stocks were reduced to an overfished state (below MSST) at the time of the first assessment in year 50. Subsequent to the stock being estimated overfished, the fishery selectivity shifted to a dome-shaped selectivity curve while the stock rebuilt to the target biomass to reflect a change in fisher behavior due to harvest restrictions induced by an overfished declaration (e.g. avoidance behavior, closed-areas) (Fig. 1b). The fishery selectivity returned to asymptotic after the stock was estimated to be rebuilt (Fig. 1a).

The second case, “time-varying”, involved autocorrelated annual deviations in natural mortality and uncorrelated normally distributed annual deviations in the parameters on which the fishery selection pattern was based during the historical, overfished, and rebuilt periods (Fig. 1c and 1d). Annual deviations in selectivity were applied to two fishery selectivity parameters: 1) the length (in cm) at which the fishery selectivity ascending limb reached maximum selectivity (termed size at maximum selectivity, Fig. 1c), and 2) the slope of the descending limb of the fishery selectivity curve resulting in dome-shaped selectivity (termed width at maximum selectivity, Fig. 1d) during the years the stock was overfished. A standard error of 0.50 was applied annually about the size at maximum selectivity for all years and a standard error of 0.20 was applied for the width at maximum selectivity during the years the stock was estimated to be overfished.

The operating model was based on a single-sex age-structured model. An annual fishery catch-per-unit effort (CPUE) index was observed with error, length- and age-composition data were collected for selected years, and used by the estimation method to estimate population size and a corresponding catch level. The catches were then removed without error from the simulated stock. The data generation, catch estimation and stock updating was conducted in an iterative fashion for 100 years (termed the management period), a length of time that would allow for recovery (stock growth to at or greater of the target biomass) to occur.

## The operating model

The numbers-at-age at the start of the year are computed using the equation:

 (1)

where *Nt,a*  is the number of fish of age *a* at the start of the year *t*, *Rt* is the number of age-0 fish at the start of year *t*, *St,a* is the selectivity during year *t* for fish of age *a*, *A* is the plus group, *F*t is the instantaneous fishing mortality rate during year *t*, and *Mt* is the instantaneous rate of natural mortality during year *t*.

Autocorrelated annual deviations in natural mortality are defined as:

 (4)

where *M* is the mean value of natural mortality,  is the standard error of the annual variations in natural mortality, and  is the autocorrelated lognormal deviation in natural mortality for year *t*:

 (5)

where  is the level of autocorrelation associated with natural mortality and  is the deviation in natural mortality for year *t*. The time-invariant natural mortality case assumed  and hence .

The number of age-0 fish is related to spawning biomass according to the Beverton-Holt stock recruitment relationship:

 (2)

where *SB0* is the unfished spawning biomass, *SBt* is the spawning biomass at the start of the spawning season in year *t*,  is the standard deviation of recruitment in log space, and *h* is steepness.

A non-equilibrium starting condition was created by applying equations (1) and (2) for the number of years equal to the maximum age prior to the start of fishing with variation in recruitment. Historical catches for years 1-50 were generated so that the populations were at 0.15*SB*0 in year 50. This ratio of spawning biomass to unfished spawning biomass (relative spawning biomass) was selected to allow for correct detection by the estimation method that the stocks were in an overfished state, and that it would require an extended number of years for the stock to rebuild to the target biomass (where the loss of data could impact the long-term performance of the estimation method). The catch of fish of age *a* during year *t* in numbers is given by:

 . (3)

The observation model was used to generate a fishery CPUE index for each year *t*:



where *Q*is the catchability coefficient,  is the standard deviation of catchability in log space, and the  is the expected biomass index given by:

 (4)

where *wa* is the weight of a fish at age *a*. The length- and age-composition data for the fishery were assumed to be multinomially distributed (see *Data Scenarios* for details). Ageing error was assumed to be normally distributed with ages subject to a 5% standard deviation by age.

The fishery selectivity during the historical period (years 1-50) were assumed asymptotic (Fig. 1a and 1c). The fishery selectivity shifted to a dome-shaped (compared to the historical asymptotic) within the operating model during the period that the stock was estimated to be below the target biomass (0.40*SB*0) (Fig. 1b and 1d). Once the population was estimated to have recovered to above the target biomass, fishery selectivity reverted to the asymptotic form. The shift in selectivity was designed as a way to mimic a change in fisher behavior resulting from an overfished declaration. The change in selectivity depended on the estimated stock status rather than the true operating model status, i.e. changes in fisher behavior modeled by a change in selectivity were assumed to be driven by management restrictions based on the estimation method’s perception of the stock rather than the true unobservable state of the simulated stock.

## The estimation method

Stock synthesis (SS), an integrated statistical catch-at-age model (Methot and Wetzel, 2013), was the estimation method used to assess the simulated stocks. SS was applied for the first time in year 50 and then every 6th year thereafter. Assessment frequency of U.S. west coast groundfish varies as a consequence of commercial importance, which is used as an indicator of exploitation, the time since last assessment, and dynamics of the stock (Methot, 2015). Long-lived rockfish species generally have slow dynamics, resulting in minimal fluctuations in biomass from year to year (assuming non-extreme harvesting). Assessing the stock every 6th year was selected as an interval of time that would balance the need for re-evaluation based upon the slow dynamics of the stock and to minimize computational time per simulation.

*R0*, steepness, growth, annual recruitment deviations, initial age-structure deviations, the size and width at maximum selectivity for the fishery were estimated. Natural mortality, the variation of growth at age, weight-at-length, the fecundity relationship, and the variation of recruitment () were assumed known. The relative spawning biomass in the assessment year was estimated and the forecasted catches were determined using the Pacific Fishery Management Council rockfish harvest control rule. The catches were removed from the operating population without error, fishery CPUE index, length- and age-composition data were generated for the subsequent six years.

The Pacific Fishery Management Council harvest control rule applies a linear reduction in catch when a stock falls below 0.40*SB*0, with no fishing when the stock falls below 0.10*SB*0. The overfishing level was set equal to the target spawner-per-recruit harvest rate (*F*0.50) multiplied by *SBt*. The overfished level was reduced by a management buffer (0.956) that accounts for the uncertainty about current biomass for well-assessed stocks to determine the acceptable biological catch level (i.e. acceptable biological catch = 0.956 overfishing level, Ralston *et al.*, 2011). The annual catch limit was set equal to the acceptable biological catch when the stock was above the target biomass, 0.40*SB*0,or reduced from the acceptable biological catch according to the harvest control rule when the stock fell below 0.40*SB*0.

One major difference in this simulation design and actual management practice of U.S. west coast groundfish was the omission of rebuilding plans that are implemented when a stock is assessed to have fallen below the MSST (defined as 0.25*SB*­0 for U.S. west coast rockfish). Harvest for stocks below MSST would no longer be determined based on the standard harvest control rule, but rather a rebuilding plan would be used to determine catches until the stock rebuilt to the target biomass. To reduce simulation complexity, a rebuilding plan was not imposed if a simulated stock was assessed to have fallen below the threshold and the catches were determined based on the harvest control rule for all estimated stock sizes.

## Data scenarios

Three data scenarios were created to explore the impact of data availability on the ability to monitor rebuilding of an overfished stock (Fig. 2). The data scenarios were designed to emulate a stock that is infrequently encountered by a fishery-independent survey (e.g. due to depth or habitat) and only fishery data were available. The historical sample sizes were generally based on the effective sample sizes observed for yelloweye rockfish. Following the first assessment in year 50, the three scenarios have different data availability based on estimated stock status (e.g. overfished vs. rebuilt) in the assessment year.

The “full data” scenario maintained the fishery CPUE index and length- and age-composition data at the historical levels (prior to the stock being declared overfished in year 50) during rebuilding (Fig. 2). The “reduced data” scenario decreased the amount of data available from the fishery during rebuilding (Fig. 2). The length and age-composition data were reduced to 20% of the historical sample sizes during rebuilding and the fishery CPUE index was eliminated during the rebuilding period. The CPUE index resumes and composition sample sizes reverted to historical levels when the stock was estimated to have rebuilt to the target biomass. The “eliminated data” scenario had no fishery data during rebuilding (Fig. 2). The fishery CPUE index and composition data resumes at historical samples sizes when the stock was estimated to be rebuilt.

The operating model applied a shift from asymptotic to dome-shaped fishery selectivity when the stock was under rebuilding. The estimation method in the full and reduced data scenarios were allowed to estimate a potential change in selectivity during the rebuilding period through the application of a time block on selectivity. However, the eliminated data scenario was forced to assume constant asymptotic selectivity over all years because no fishery composition data were available to detect a potential shift in selectivity.

## Performance measures

The outcomes of the simulations for each case and data scenario were summarized using five metrics that were selected to evaluate the impact of data on estimation of indicators of stock status (e.g. relative spawning biomass) and management quantities (e.g. rebuilding catch):

1. The relative errors (REs) for estimated parameters, calculated as:

 (5)

where *E* is the estimated quantity of interest and *T* is the true value from the operating model.

1. The percent root mean square error (RMSE), a measure of precision and bias, was calculated to assess the overall level of error given the amount of data available:

 (6)

1. The average (over simulations) of the total catch while the stock was recovering to the target biomass.
2. The annual average variability of the catches (abbreviation AAV), defined as:

(7)

where C*y* is the catch during year *y*.

# Results

## Assessment performance with time-invariant parameters

The trends of the relative error about spawning biomass and relative spawning biomass were generally consistent among the full and reduced data scenarios (Figs. 3a-b and 4a-b). The median spawning biomass and relative spawning biomass was negatively biased while was rebuilding for both scenarios (Figs. 3a-b and 4a-b). As expected, the estimates of spawning biomass and relative spawning biomass for the full data scenario were less variable during the rebuilding period compared to the reduced and eliminated data scenarios (Figs. 3a-c and 4a-c). However, by the end of the management period, the variance of estimated spawning biomass and relative spawning biomass were similar among the full and reduced data scenarios. The eliminate data scenario resulted in median estimates of spawning biomass and relative spawning biomass with limited bias but were highly imprecise at the start of the management period (years 50-74) (Figs. 3c and 4c). As stocks begin to be estimated rebuilt and data collection resumed for the eliminate data scenario, the median estimates become negatively biased and remained imprecise. In contrast to the full and reduced data scenarios, the estimates of spawning biomass and the relative spawning biomass for the eliminated data scenario showed little improvement in precision by the end of the management period (Figs. 3c and 4c).

The RMSE for the estimated relative spawning biomass for each assessment year shows the increased precision of the full data scenario during the rebuilding period compared to the reduced and eliminated data scenarios (Fig. 5a). However, the RMSE for the reduced data scenario improved over the management period as stock began to be assessed rebuilt, and data sample sizes returned to historical levels. The eliminate data scenario resulted in the highest RMSE over the entire management period (Fig. 5a). The lack of improvement in the RMSE for the eliminated data scenario was driven by the number of simulations that were never estimated to have rebuilt to the target biomass (35 out of 100 simulations).

Examining the eliminated data scenario closer revealed a pattern in the performance of the estimation method. The eliminated data scenario simulations were divided and plotted based on whether the estimation method estimated the simulation rebuilt (65 simulations) or failed to rebuild (35 simulations). To allow comparison, the estimates from the full data scenario were divided according to the eliminated data scenario status (rebuilt vs. failed to rebuild) and plotted. The estimated spawning biomass were strongly negatively biased in the first assessment year (Figs. 6b [white]) for the 35 simulations that were estimated not to be rebuilt by the end of the management period. The biased estimates about spawning biomass (Fig. 6b [white]) were driven by negatively biased estimates of steepness in the first assessment (Fig. 6d [white]). Data collection for the eliminated data scenario did not resume until the stock was estimated rebuilt, and in the absence of new data, the negatively biased estimates of steepness resulted in the estimation method perceived a less productive stock requiring an extended period. However, with full data present, estimated quantities (spawning biomass and steepness) improved for this subset of simulations and were median unbiased by the end of the management period (Fig. 6a and 6c [white]). In contrast, the simulations that were estimated to have successfully rebuilt for the eliminate data scenario estimated steepness values that were positively biased relative to the operating model (Figs. 6d [grey]).

The estimates of steepness varied across data scenarios. The full data scenario resulted in median unbiased estimates by the end of the management period (Fig. 7a). In contrast, the reduced data scenario estimated positively biased median value of steepness across simulations (Fig. 7b). The eliminated data scenario had the highest variance among estimates of steepness during the management period (Fig. 7c) due to the mixture of rebuilt and not rebuilt stocks.

The median number of years estimated for the stocks to recover to the target biomass for the full data scenario was longer the median recovery year within the operating model (Table 2). In contrast, both the reduced and eliminated data scenarios estimated a shorter median recovery time compared to the operating model (Table 2). The contrast in estimated recovery times across the data scenarios was related to the average catch obtained during rebuilding along with the bias and variability of estimates. The full data scenario estimated a negatively biased median relative spawning biomass with low inter-annual variability across assessment years, resulting in estimates that predicted constant rebuilding but at a slower rate than the true stock (Fig. 4a). In contrast, the reduced data scenario had increased variability across the estimates of the relative spawning biomass (Fig. 4b). The variability of estimates between assessments resulted in stocks being estimated recovered earlier than the operating model due to estimation error driven by the limited composition samples during rebuilding.

The reduced data scenario had the lowest median average catch during rebuilding (Table 3) with median rebuilding time estimated shorter than the true time to recovery with the operation model (Table 2). The eliminated data scenario which was entirely dependent upon historical data until the stocks were estimated rebuilt, essentially projected the population forward with each assessment based on the initial parameter estimates from the historical data, resulting in high median average catches during rebuilding, and the lowest median AAV during rebuilding and across the entire management period (Table 3).

Reduction or elimination of data during rebuilding increased the variance about estimates of the size at maximum fishery selectivity which defined the ascending limb of the selectivity curve (see Fig. 1a), with the estimates generally equal to the true value in median terms for all data scenarios (Fig. 8a-c). The variability of the estimates for the reduced and eliminated scenarios improved when the majority of the stocks had rebuilt, and fishery composition sample sizes returned to historical levels. The full and reduced data scenarios were allowed to estimate dome-shaped selectivity during the rebuilding period (the eliminated data scenario did not allow for estimation of dome-shaped selectivity due to the absence of fishery composition data), and resulted in positively biased median estimates that were highly variable at start of the management period (Fig. 9a-b). The positively biased estimates for this parameter indicated that the data available were not sufficient to inform the estimation method about the severity of the dome-shape in selectivity (a higher estimated value implies a less severe dome in selectivity, e.g. decreased slope for the selectivity descending limb) during rebuilding. The full data scenario resulted in markedly improved estimates in the shape of the dome over the management period, compared to the reduced data scenario (Fig. 9a-b).

## The impact of time-varying parameters

The case that assumed time-varying annual deviations in natural mortality and fishery selectivity generally resulted in increased variation in estimates compared to the time-invariant case. The median of spawning biomass estimates at the time of the first assessment were positively biased and highly variable (Fig. 3d-f). The variance about spawning biomass decreased markedly for the full data scenario after the first assessment in year 50 (Fig. 3d). However, the variance of estimated spawning biomass remained high for approximately the first twenty-five years of the management period for both the reduced and eliminated data scenarios, until approximately 50% of the simulated stocks were estimated recovered and the fishery sample sizes increased to historical levels (Fig. 3e-f). The full and reduced data scenarios resulted in median spawning biomass estimates that were generally negatively biased (Fig. 3d-e). However, the median relative spawning biomasses were variable over the management period (Fig. 4d-e). The estimates of the median relative spawning biomass for eliminated data scenario were positively biased at the start of the management period, but became negatively biased as stocks rebuilt to target biomass and data collection resumed (Fig. 4f).

Compared to the case with time-invariant parameters the RMSE was higher for all data scenarios when time-varying parameters were present within the operating model (Fig. 5). The RMSE about the estimated relative spawning biomass for the full data scenario was lower relative to the other scenarios for the entire management period (Fig. 5b). Similar to the time-invariant results, the RMSE of relative spawning biomass for the eliminate data scenario was the highest among the scenarios across the entire management period peaking in assessment year 92 at 1783%.

The time-varying results for the eliminated data scenario were qualitatively similar to time-invariant case where a high number of simulations failed to be estimated rebuilt (32 simulations). As was observed in the time-invariant case (Fig. 6), the simulations that failed to be estimated rebuilt had negatively biased median estimates of spawning biomass and relative spawning biomass at the time of the time of the first assessment which were driven by negatively biased estimates of steepness (not shown).

The inclusion of time-varying parameters in the operating model resulted in shorter median estimated recovery time relative to the time-invariant case for all data scenarios (except the eliminated scenario which were equal) (Table 2). However, the median number of rebuilding years for the operating model stocks were similar between the time-varying and invariant cases. The estimation method estimated earlier recovery times for the time-varying case due to the increased variability in the estimates of relative spawning biomass resulting in the estimation method having an increased frequency of erroneously estimating the biomass above the target stock size (Fig. 4).

The median average catch obtained during the recovery period was the highest for the eliminated data scenario due to the positively biased estimates of the relative spawning biomass at the start of the management period (Table 3 and Fig. 4f). Additionally, the eliminated data scenario had the lowest median AAV during the overfished period (Table 3). In contrast, the eliminated data scenario, resulted in the highest number of stocks that were never estimated to exceed the target biomass (Table 3).

Inclusion of time-varying selectivity resulted in a persistent positive bias for the median estimated size at maximum selectivity across all data scenarios (Fig. 8d-f), although the full data scenario resulted in the lowest variation. The full and reduced data scenarios which were allowed to estimate dome-shaped selectivity (width at maximum selectivity) during the recovery period, resulted in highly variable estimates at the start of the management, period with the variability for the estimates decreasing earlier for the full data scenario (Fig. 9c-d).

# Discussion

Maintaining fishery data at historical levels during rebuilding reduced the variation in estimates between assessments. While the full data scenario had a reduction in variation, the median estimates of spawning biomass and relative spawning biomass were consistently negatively biased for much of the management period. This result is contrary to what might be expected when additional data are available. Explorations with simulations where there was a fishery independent survey that provided an index of abundance and composition data (length and age), determined that this negative bias was eliminated if survey composition data were available along with fishery composition data (see *Appendix A*). The observed bias was driven by two key factors; the fishery selectivity curve and data quantity. The fishery selectivity curve was specified to be greater than the maturity curve, resulting in only mature larger fish being selected. The fishery data were informative about recruitment, but with a lag between recruitment to population and being observed in the fishery. However, a fishery independent survey that selected fish at smaller sizes yields information about recruitment earlier. Additionally, an increase in the length- and age-composition samples from multiple data sources can improve estimates of recruitment, spawning biomass, and relative spawning biomass (Yin and Sampson, 2004; Wetzel and Punt, 2011).

The negative bias in median estimated relative spawning biomass by the full data scenario resulted in the estimation method failing to determine that the operating model population was at or above the target biomass (median number of rebuilding years greater than the operating model, Table 2). an outcome that would produce extended harvest restrictions that were not warranted given the true state of the population, a situation fishery management would like to avoid. However, the reduced estimation variability offered by the full data scenario resulted in an improvement in the consistency of estimates by subsequent assessments which offers a level of stability for fisheries managers and stakeholders. In contrast, the higher between-assessment variation in estimates of spawning biomass for the reduced data scenario resulted in stocks being estimated rebuilt when the true population was still below the target biomass which could have undesirable outcomes for fisheries management. Overly optimistic estimates of relative spawning biomass can result in overfishing when catch limits are set too high, leading to further reductions in biomass, potentially requiring a subsequent overfished declaration by a future assessment.

The high number of simulations that failed to be estimated rebuilt by the eliminated data scenario highlighted the importance of continued data collection during rebuilding. In the absence of new data, the first and subsequent assessments are entirely dependent on the quality of the historical data to inform parameter estimates. The simulations that failed to be estimated rebuilt were driven by negatively biased estimates of steepness at the time of the first assessment. Estimating a stock less productive than the true population resulted in reduced estimates of spawning biomass and relative spawning biomass, with the assessment setting harvest at levels well below the true acceptable biological catch. The reduced harvest allowed the population in the operating model to rebuild to or above the target biomass. However, in the absence of new (and informative data) the estimation method was unable to detect the correct stock size. The operating model population represented a two-way trend of abundance (decline and increase in biomass) with fishery data available during the fishing down and recovery period, which previous studies have found informative in estimating steepness (Magnusson and Hilborn, 2007; Conn *et al.*, 2010). This work showed that a one-way trip with limited data may not be adequate to correctly estimate steepness, but the inclusion of even limited data can, with contrast in stock size, improve the estimation of steepness even if the initial assessment produced a poor estimate (Figs. 6c and 7).

The general trend in results when the operating model included time-varying natural mortality and fishery selectivity were similar to the time-invariant case, although the estimates were more variable across all data scenarios. The estimation method assumed a single natural mortality across all years equal to the mean value which was used to generate the autocorrelated annual deviations in the operating model. This setup was a strategic choice that allowed variation in the composition data that the estimation method would not be able to account for, but would not be anticipated to result in strongly biased estimates due to model misspecification. Time-varying natural mortality is likely more complex with extended periods of high or low mortality based upon external factors (e.g. predator abundance, climate conditions), which could result in large biases in estimated quantities if not accounted for in an assessment (Johnson *et al.* 2014).

Shifts in the form of selectivity over time and the impact of annual deviations in selectivity led to mixed results. The estimation method consistently estimated a positively biased median size at maximum selectivity for all data scenarios with time-varying selectivity. The operating model selectivity involved normally distributed deviations to generate the annual shifts in selectivity and one would not *a priori* predict the estimation method to have a consistent bias in estimates. However, the estimation method was able to identify the change in the selectivity form (asymptotic to dome-shaped through a reduction in the width at peak selectivity) during the rebuilding years with similar bias to the time-invariant case. Each case led to estimates that under estimated the decline in the descending limb, defining the dome in selectivity. This evaluation applied time blocks defined by the status of the stock to allow for shifts in selectivity, ignoring the annual deviations in the selectivity curve. Studies have evaluated other ways of estimating time-varying selectivity using state-space models (Nielsen and Berg, 2014), or the implications of applying time blocks vs. allowing a random walk component in selectivity parameters or catchability (Wilber and Bence, 2006; Martell and Stewart, 2014). Further exploration should be conducted to evaluate if allowing a random walk or applying an alternative estimation method eliminates the bias observed in the estimated selectivity here and how data quantity and quality impacts these estimates.

This work highlights the benefits of continued data collection during rebuilding on the precision of estimates, but there are many additional reasons why retaining data streams is important. The first benefit of continued data collection is the potential ability to identify misspecification in the model assumptions. The estimation method and operating models applied here, generally made similar structural assumptions. However, the true state of nature is never known with confidence in real-world assessments. Continued data collection may allow for the identification of model misspecification in the structural assumptions (e.g. growth, recruitment) that will allow for better approximations to reality through models. Specifically, when stocks are driven to depleted stock sizes there could be long-term impacts to the population that could negatively impact the ability of the stock to rebuild (e.g. Hixon *et al*., 2014) that additional data would be required to detect. The second benefit of continued data collections is that ability to identify potential long-term changes in stock dynamics (e.g. changes in climate-driven forces). There has been much work identifying potential links between climate and recruitment (Hollowed *et al.*, 2011; Ianelli *et al.*, 2011; Mueter *et al.*, 2011; Strachura *et al.*, 2014). Additionally, long-term varying climate conditions may result in changes in biological parameters (Swain and Benoit, 2015) which will impact the productivity of stocks (Legault and Palmer, 2015) that needs to be accounted for when setting harvest limits.

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# Tables

Table 1. Life history and observation parameters used in the operating model and the treatment within the estimation model.

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| Parameter | Time-invariant | Time-varying | | Treatment in estimation model |
| Natural mortality(*M*) yr-1 | 0.08 | | 0.08 | Fixed |
| Natural mortality standard error (σm) | 0 | | 0.10 |  |
| Natural mortality autocorrelation (ρ) | 0 | | 0.707 |  |
| Steepness(*h*) | 0.65 | |  | Estimated |
| Maximum length (*L∞*)(cm) | 64 | |  | Estimated |
| Growth coefficient (*kt*) (yr-1) | 0.05 | |  | Estimated |
| Weight at length (kg) | α =1.50x10-5β = 3 | |  | Fixed |
| Length at 50% maturity (cm) | 37 | |  | Fixed |
| Recruitment variation (σR) | 0.50 | |  | Fixed |
| Fishery CPUE standard error (σf) | 0.30 | |  | Fixed |
| Fishery CPUE catchability coefficient (*Qf)* | 0.01 | |  | Analytically Solved |

Table 2. The median and 90% simulation interval of the estimated number of years needed to rebuild to target biomass, the operating model number of years needed to rebuild to target biomass, and the number of stocks that failed to rebuild to the target biomass determined by the estimation method (EM) and the operating model (OM) for each case and data scenario.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
| Selectivity / data scenario | Estimated num. of rebuilding years | | Operating model num. of rebuilding years | | Num. of stocks that failed to rebuild | |
|  | Median | 90% SI | Median | 90% SI | EM | OM |
| Time-invariant |  |  |  |  |  |  |
| full data | 43 | (13 - 87) | 34 | (16 - 73) | 7 | 4 |
| reduced data | 31 | (19 - 61) | 34 | (14 - 83) | 1 | 5 |
| eliminated data | 25 | (14 - 72) | 37 | (14 - 87) | 35 | 4 |
| Time-varying |  |  |  |  |  |  |
| full data | 31 | (13 - 91) | 35 | (13 - 85) | 13 | 4 |
| reduced data | 25 | (13 - 79) | 32 | (12 - 74) | 8 | 2 |
| eliminated data | 25 | (13 - 77) | 36 | (12 - 79) | 32 | 5 |

Table 3. The median and 90% simulation intervals of the average catch during rebuilding, the AAV during rebuilding, and the AAV over all years for each case and data scenario.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
| Selectivity / data scenario | Average catch during rebuilding | | AAV during rebuilding | | AAV all years | |
|  | Median | 90% SI | Median | 90% SI | Median | 90% SI |
| Time-invariant |  |  |  |  |  |  |
| full data | 44.0 | (15.3 - 78.9) | 6.0 | (3.7 - 11.5) | 3.2 | (2.1 - 4.7) |
| reduced data | 28.1 | (14.6 - 57.9) | 7.7 | (4.0 - 14.5) | 3.5 | (2.3 - 5.3) |
| eliminated data | 41.3 | (19.9 - 83.8) | 2.6 | (1.3 - 4.4) | 2.2 | (1.3 - 3.9) |
| Time-varying |  |  |  |  |  |  |
| full data | 31.7 | (11.0 - 75.4) | 7.3 | (4.4 - 17.5) | 4.2 | (2.7 - 5.9) |
| reduced data | 25.1 | (15.6 - 68.0) | 8.9 | (4.5 - 20.7) | 4.5 | (2.6 - 9.8) |
| eliminated data | 36.3 | (15.7 - 79.4) | 2.3 | (1.2 - 4.8) | 2.8 | (1.3 - 5.3) |

# Figures

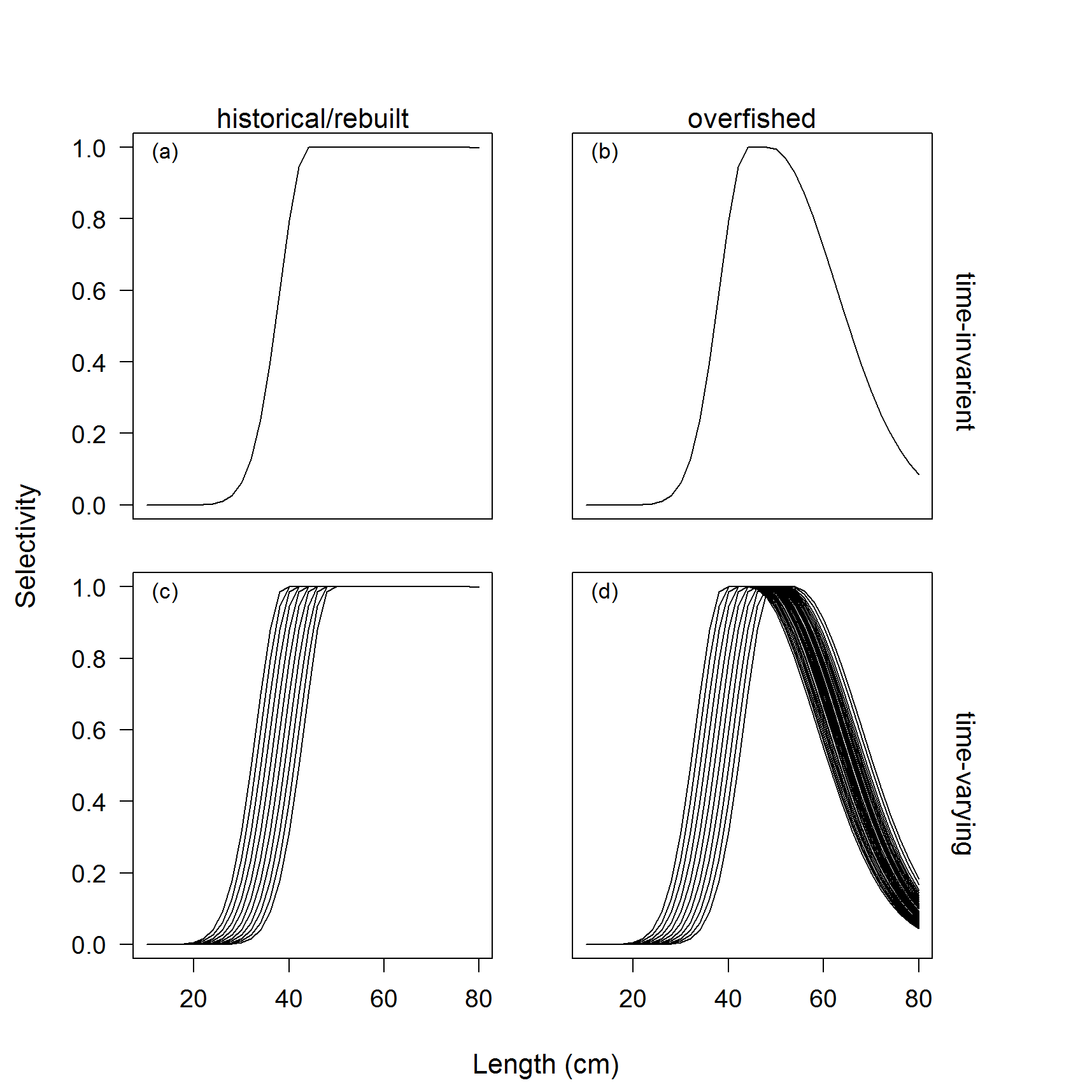


Figure 1. The fishery selectivity for either time-invariant or time-varying selectivity for the historical/rebuilt (a and c) and overfished periods (b and d). A standard error of 0.50 was applied annually about the size at maximum selectivity which defined the variability among the ascending limb of the selectivity curve (c and d), and a standard error of 0.20 was applied for the width at maximum selectivity that defined the slope of the descending limb creating dome-shaped selectivity while the stock was estimated overfished (d) (see Methot and Wetzel, 2013 for additional details on double normal selectivity).

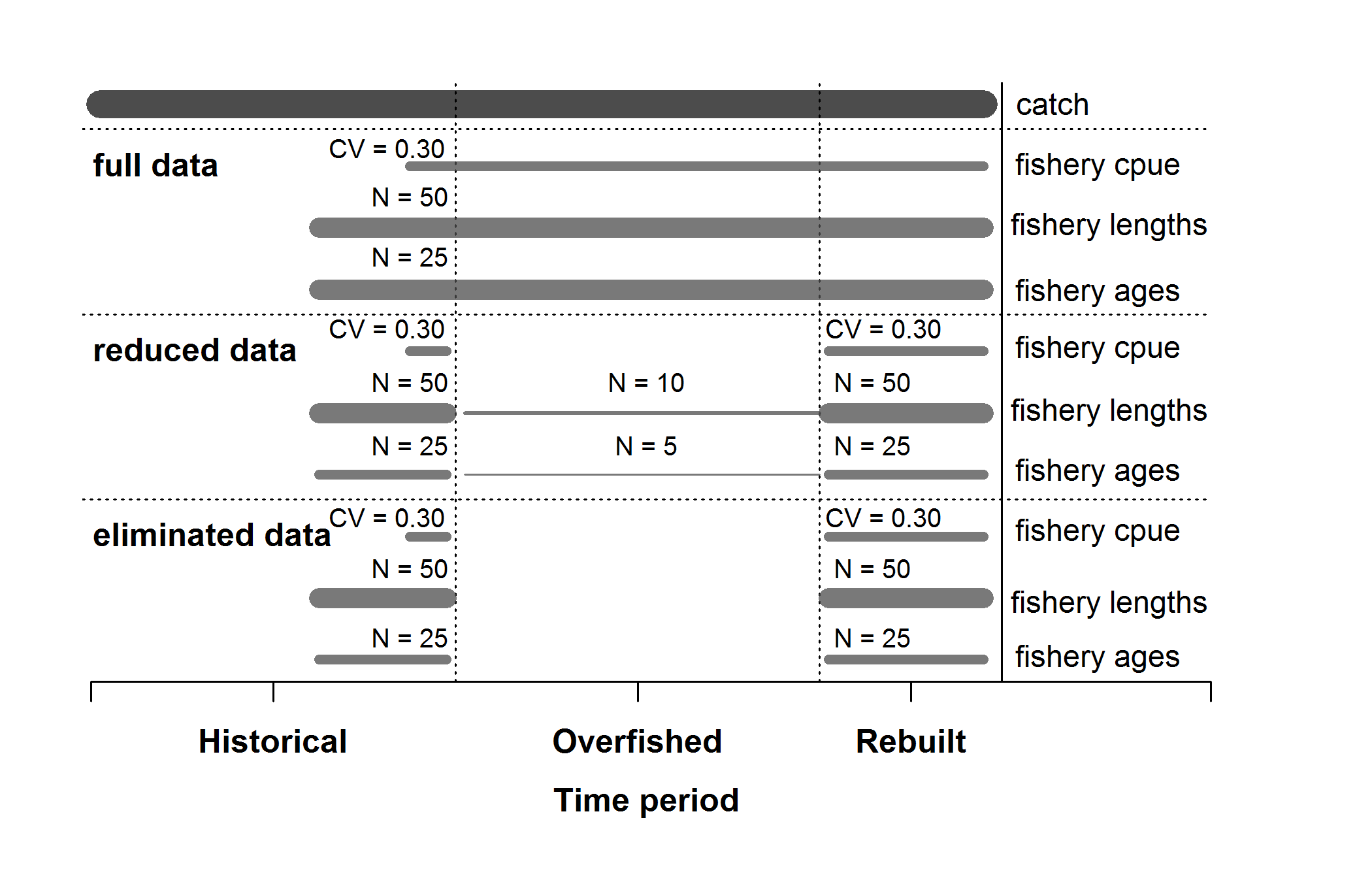


Figure 2. Summary of the data available for each of the data scenarios. Catches were known without error and were available across all data scenarios.

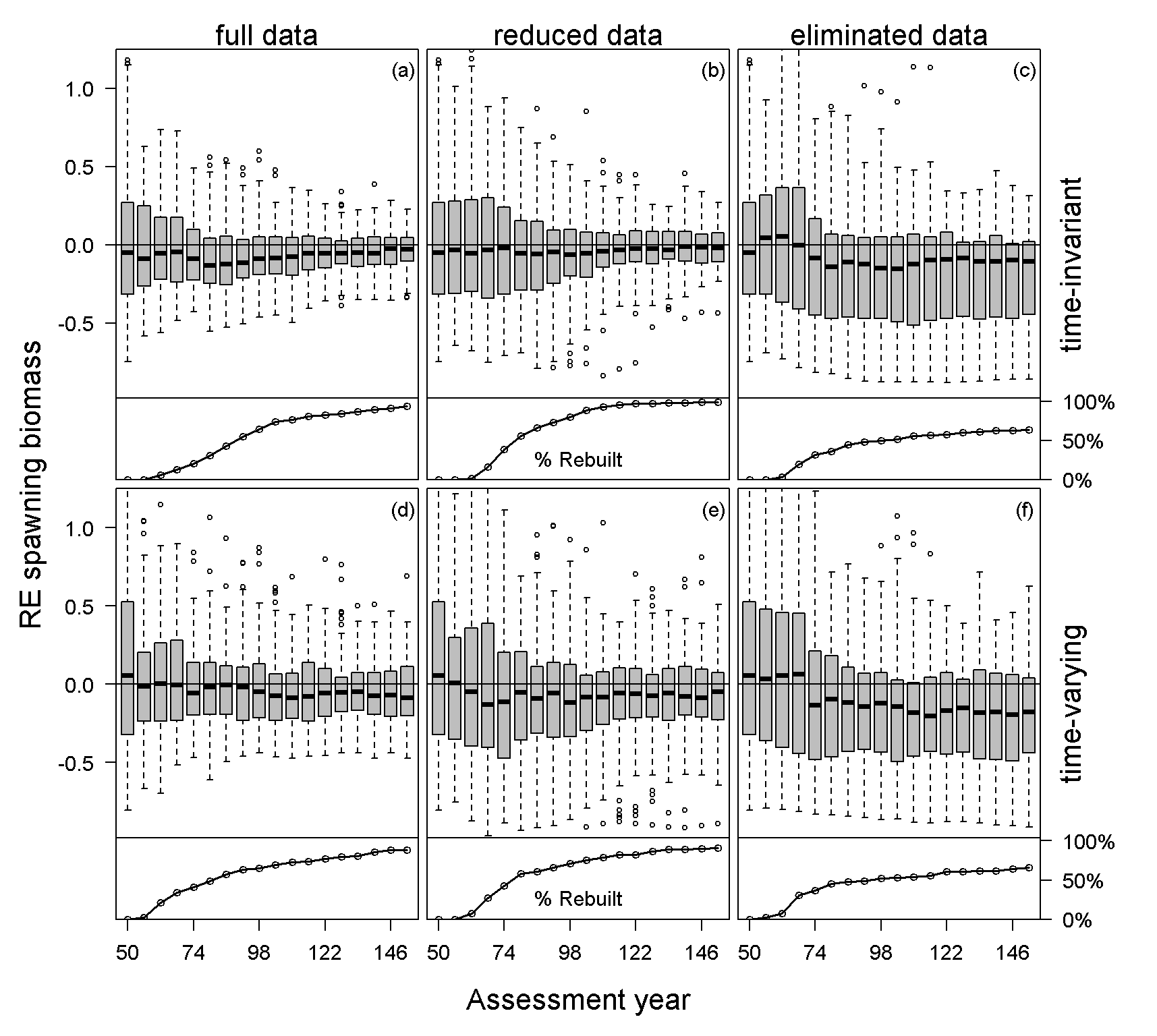


Figure 3. The relative error of estimated spawning biomass in each assessment year for each case and data scenario summarized in each assessment year for all simulations (top panel) and the percentage of stocks that had rebuilt to the target biomass during the management period (bottom panel), with data collection consequently returning to historical levels. The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

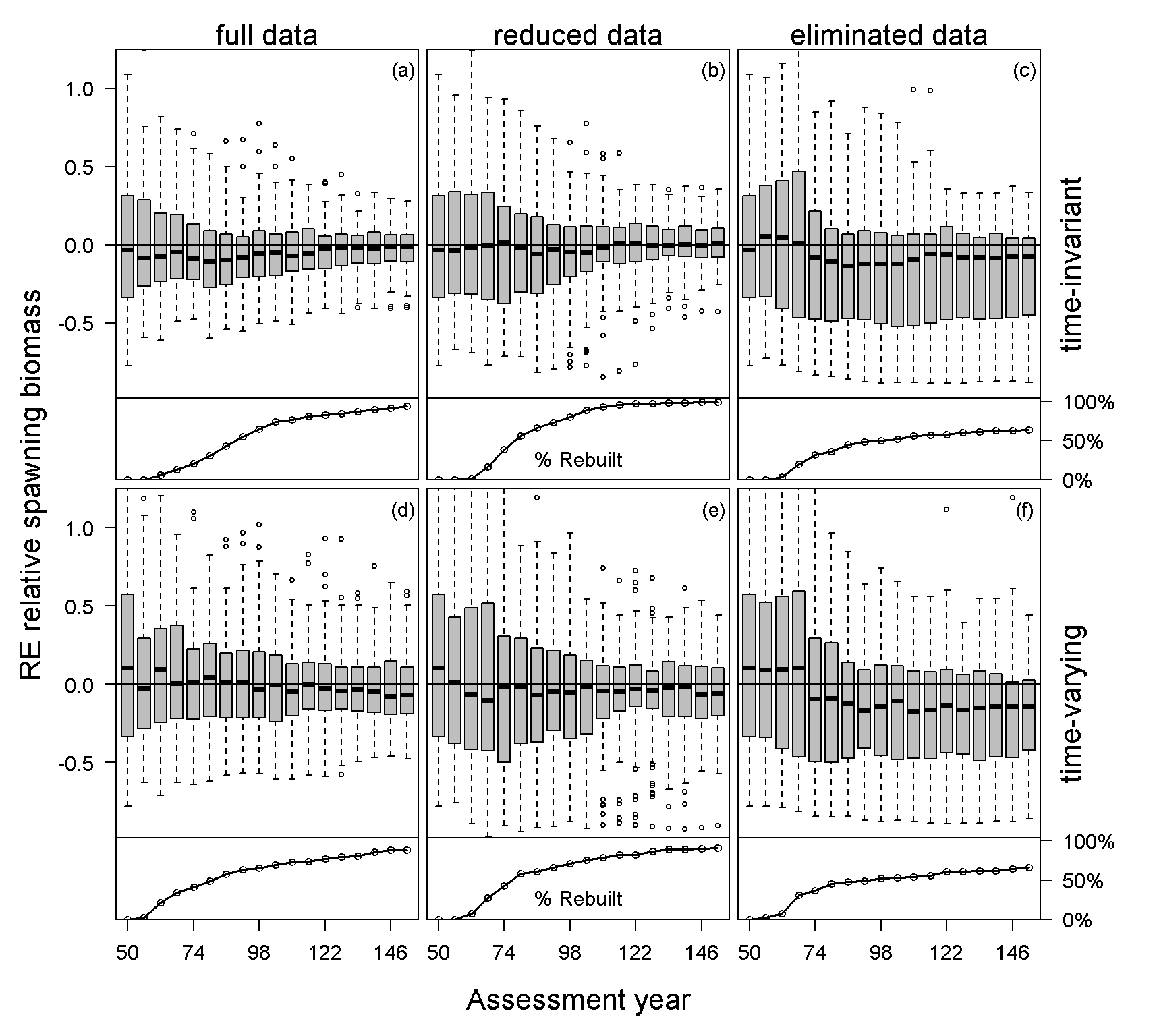


Figure 4. The relative error of estimated relative spawning biomass in each assessment year for each case and data scenario summarized in each assessment year for all simulations (top panel) and the percentage of stocks that had rebuilt to the target biomass during the management period, with data collection consequently returning to historical levels (bottom panel). The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

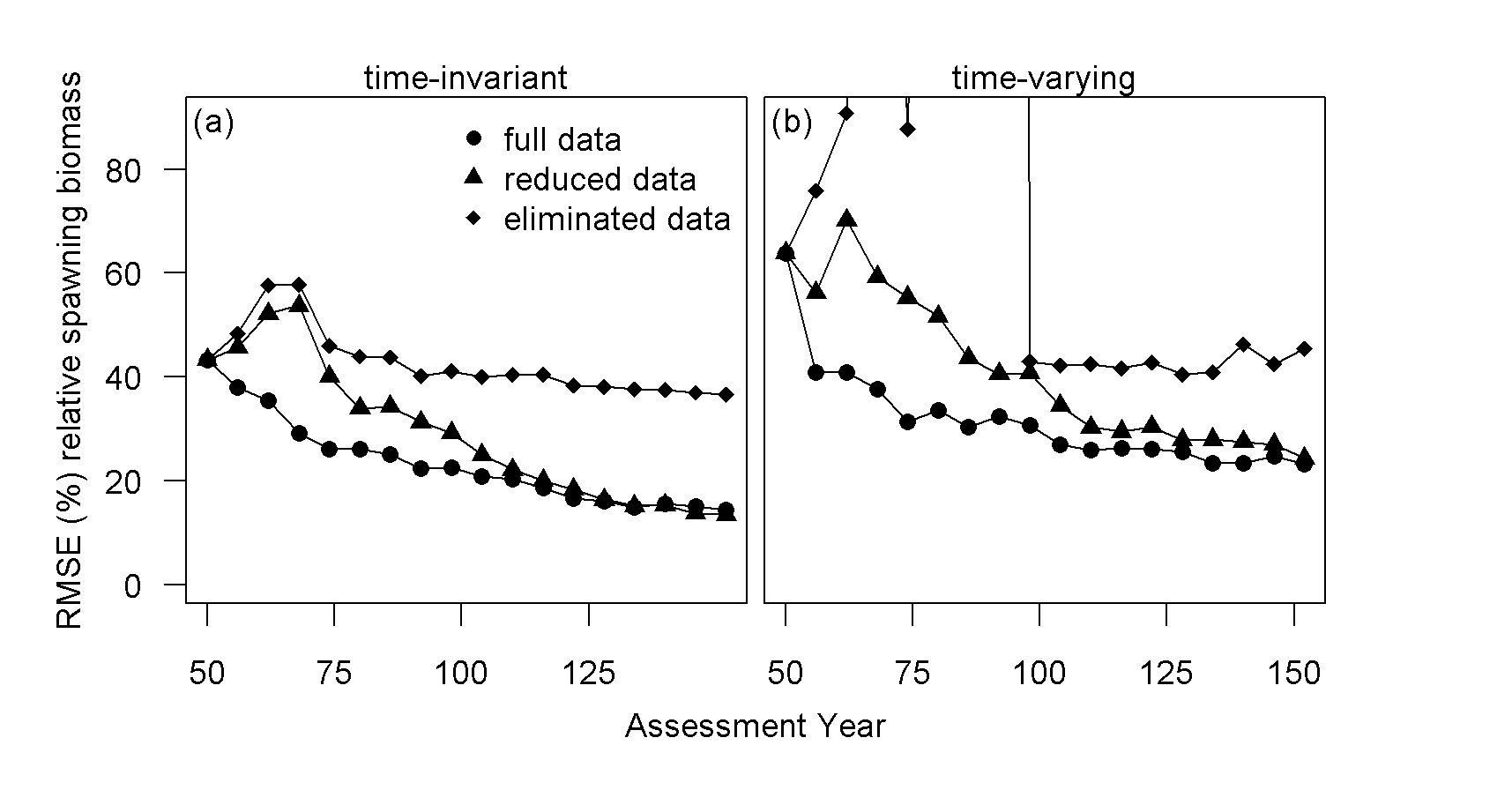


Figure 5. The root mean square error about relative spawning biomass in the assessment year for each cases and data scenario.

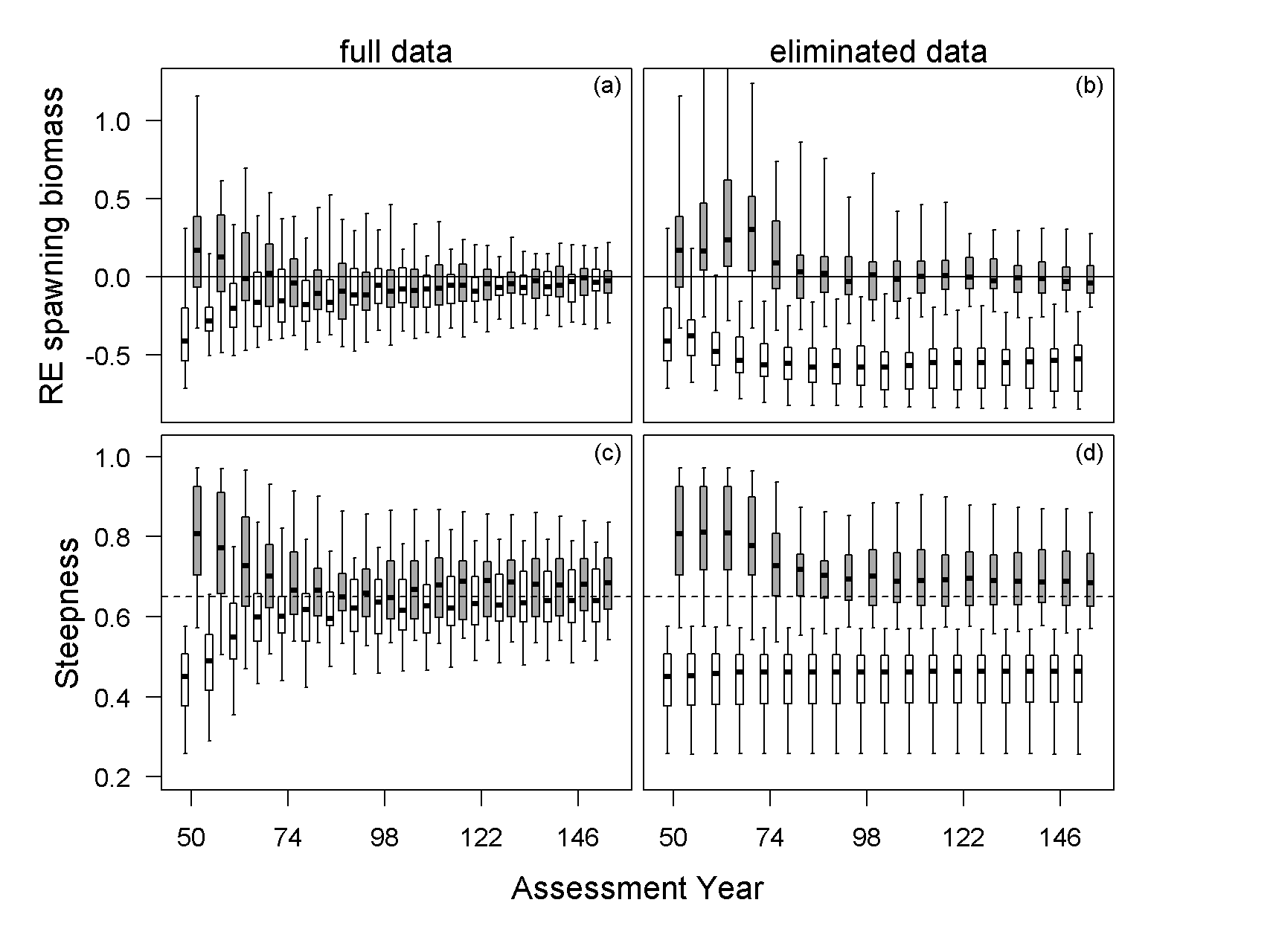


Figure 6. Comparison of the relative error about spawning biomass and the estimates of steepness between the full and eliminated data scenarios for the time-invariant case with the results divided by the simulated stocks that failed to be estimated rebuilt (35 simulations [white]) vs. the simulated stocks that were estimated rebuilt (65 simulations [grey]) by the eliminated data scenario. The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

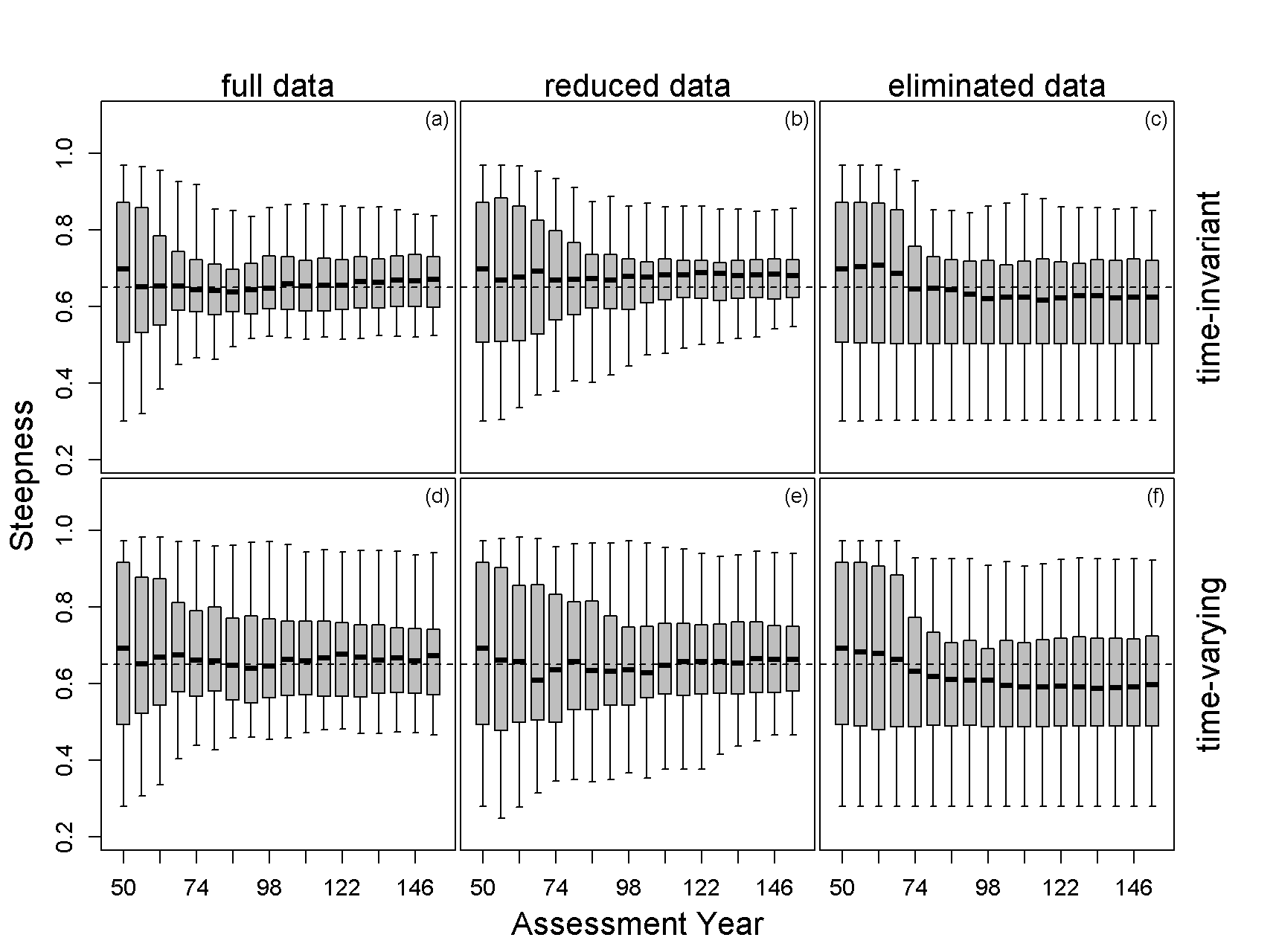


Figure 7. The estimates of steepness in each assessment year for each case and data scenario. The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

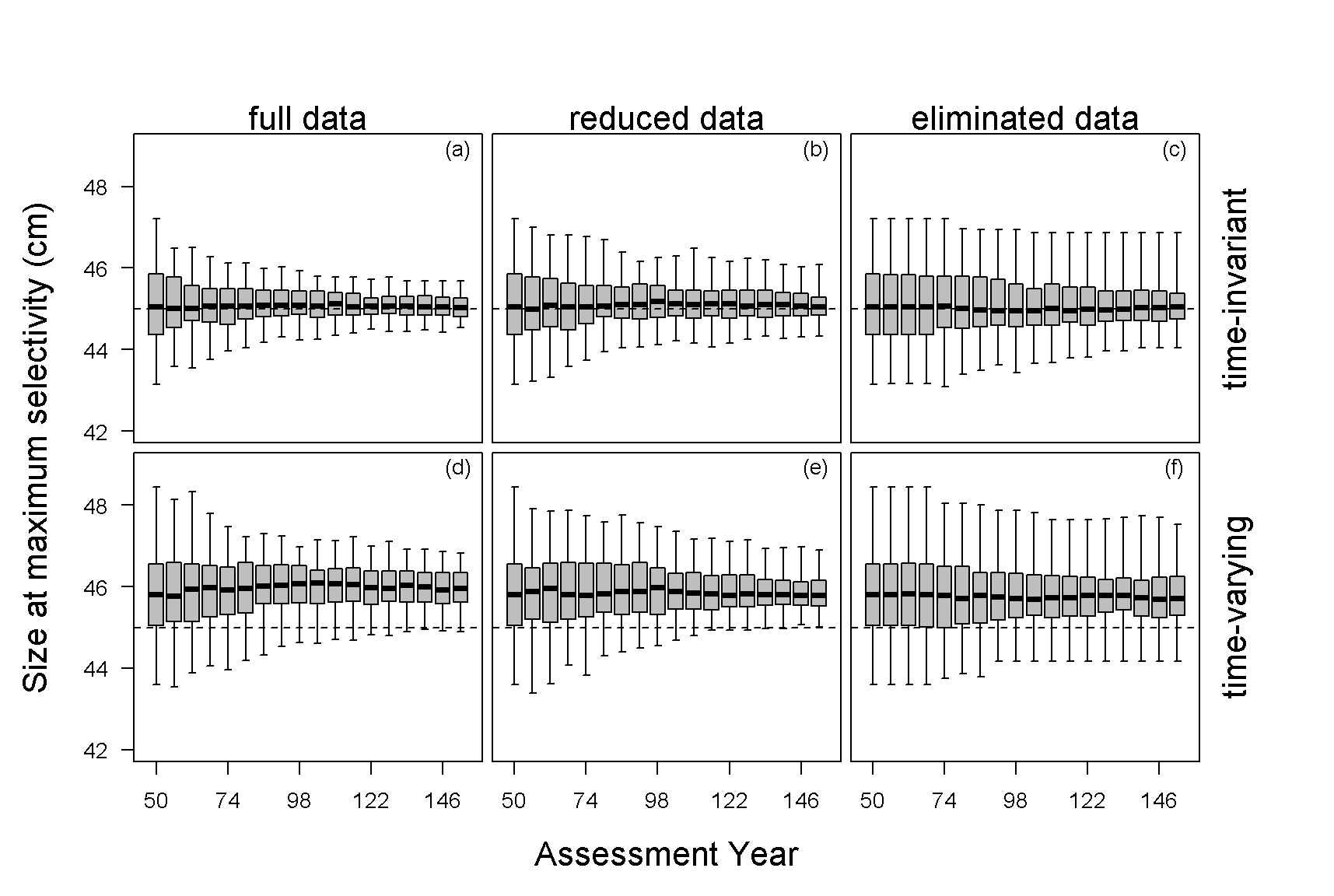


Figure 8. The estimated size at maximum fishery selectivity for each data scenario and case. The black dashed line indicates the operating model parameter value. The estimates from the data scenarios with time-varying selectivity were compared against the mean of the distribution from the operating model. The solid black line indicates the median value and the boxplot whiskers cover the 95% simulation interval for each assessment year.

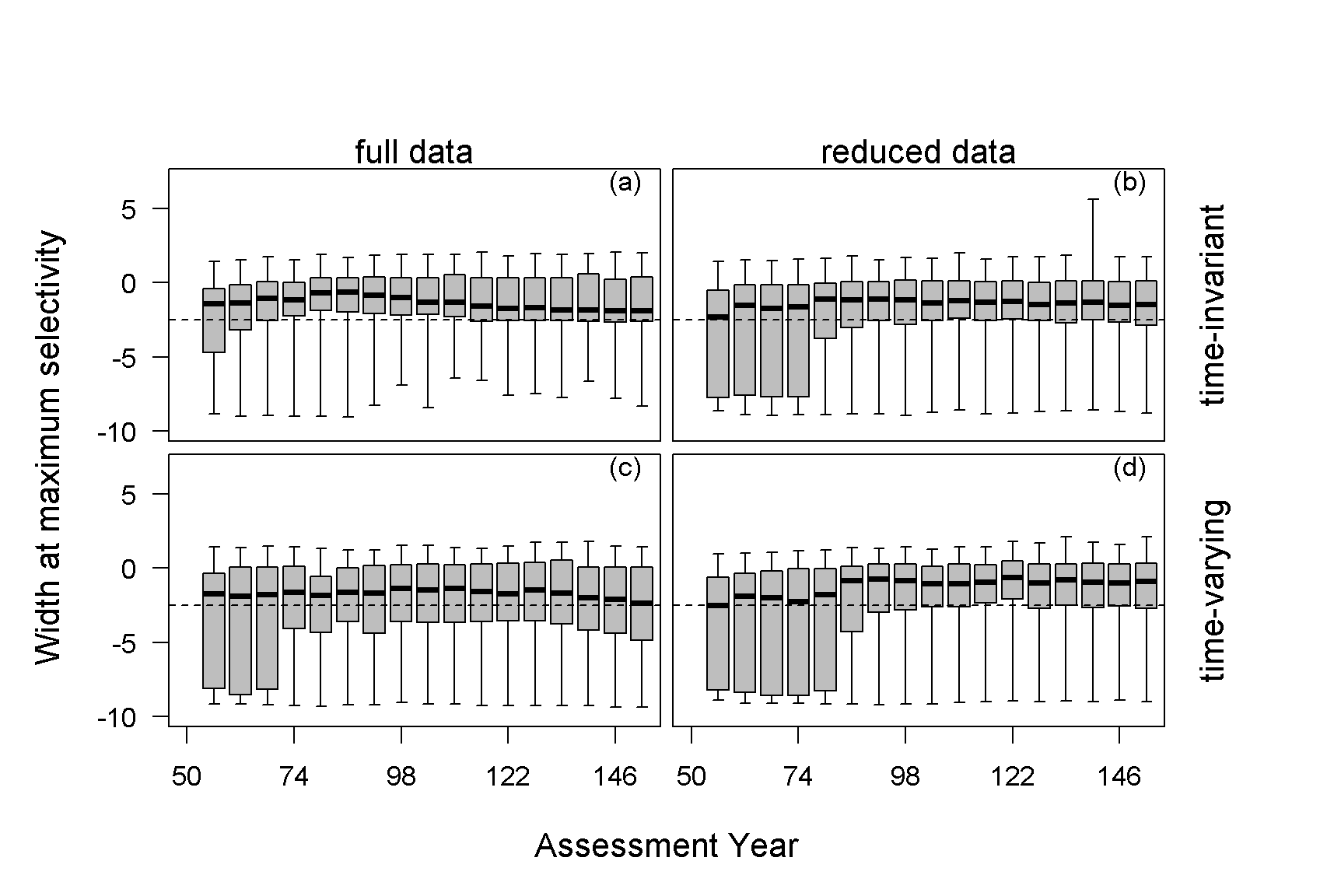


Figure 9. The estimated width at maximum selectivity (i.e. defines the extent of dome-shaped selectivity) for the fishery that occurred while the stock was overfished for each data scenario and case. The black dashed line indicates the operating model parameter value. The estimates from the data scenarios with time-varying were compared against the mean of the distribution from the operating model. The solid black line indicates the median value and the boxplot whiskers cover the 95% simulation interval for each assessment year.

# Appendix A

Additional simulations were conducted to evaluate the impact of only having fishery information versus indices of abundance and length- and age-composition data available from both a fishery-independent survey and a fishery. The operating model generated a highly uncertain survey (CV = 0.40) that was conducted on a biennial basis with low length- and age-composition sample sizes, representative of a survey that poorly sampled the simulated stock (e.g. due to habitat and gear restrictions or restricted sampling areas) (Fig. A.1). The survey selectivity was assumed to be fixed at an asymptotic shape, selecting fish at smaller sizes relative to the fishery selectivity. All other specifications for the fishery within the operating model and the assumptions applied by the estimation method were retained according to those detailed in the Material and Methods section.

The estimates of spawning biomass (Fig. A.2a-c) and relative spawning biomass (Fig. A.3a-c) for the time-invariant case were median unbiased at the time of the first assessment in year 50. The addition of a survey index and composition data for all data scenarios led to less variable and reduced median bias over the management period relative to the simulations without survey data (Figs. 4a-c and 5a-c). The presence of survey data when fishery data were eliminated (eliminated data scenario) allowed the majority of the simulated stocks being estimated rebuilt by the end of the management period (Fig. A.2c) compared to the large fraction of simulations that failed to be estimated rebuilt when only historical data were available from the fishery (Fig. 4c).

The inclusion of survey data, in addition to fishery data when time-varying parameters were present led to reduced variability in the estimates of spawning biomass and relative spawning biomass (Figs. A.2d-f and A.3d-f). The full data scenario, for both cases (time-invariant and time-varying), had the lowest RMSE for relative spawning biomass during the early portion of the management period, when the majority of simulations were rebuilding for both cases (Fig. A.4). However, midway through the management period, after a majority of the simulated stocks had rebuilt and data restrictions were removed, the data scenarios resulted in similar RMSE (Fig. A.4). The inclusion of survey data for all data scenarios resulted in similar estimates of the median number of years to recover to the target biomass, which were similar to the median rebuilding time with the operating model (Table A.1).

Table A.1 The median and 90% simulation interval of the estimated number of years needed to rebuild to target biomass, the operating model number of years needed to rebuild to target biomass, and the number of stocks that failed to rebuild to the target biomass determined by the estimation method (EM) and the operating model (OM) for each case and data scenario.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
| Selectivity / data scenario | Estimated num. of rebuilding years | | Operating model num. of rebuilding years | | Num. of stocks that failed to rebuild | |
|  | Median | 90% SI | Median | 90% SI | EM | OM |
| Time-invariant |  |  |  |  |  |  |
| full data | 37 | (18 - 79) | 39 | (18 - 79) | 2 | 2 |
| reduced data | 37 | (19 - 67) | 38 | (16 - 78) | 0 | 4 |
| eliminated data | 37 | (18 - 74) | 38.5 | (16 - 84) | 2 | 4 |
| Time-varying |  |  |  |  |  |  |
| full data | 31 | (13 - 85) | 41 | (14 - 85) | 15 | 3 |
| reduced data | 31 | (13 - 85) | 39 | (13 - 80) | 9 | 3 |
| eliminated data | 31 | (13 - 88) | 39 | (13 - 80) | 8 | 3 |

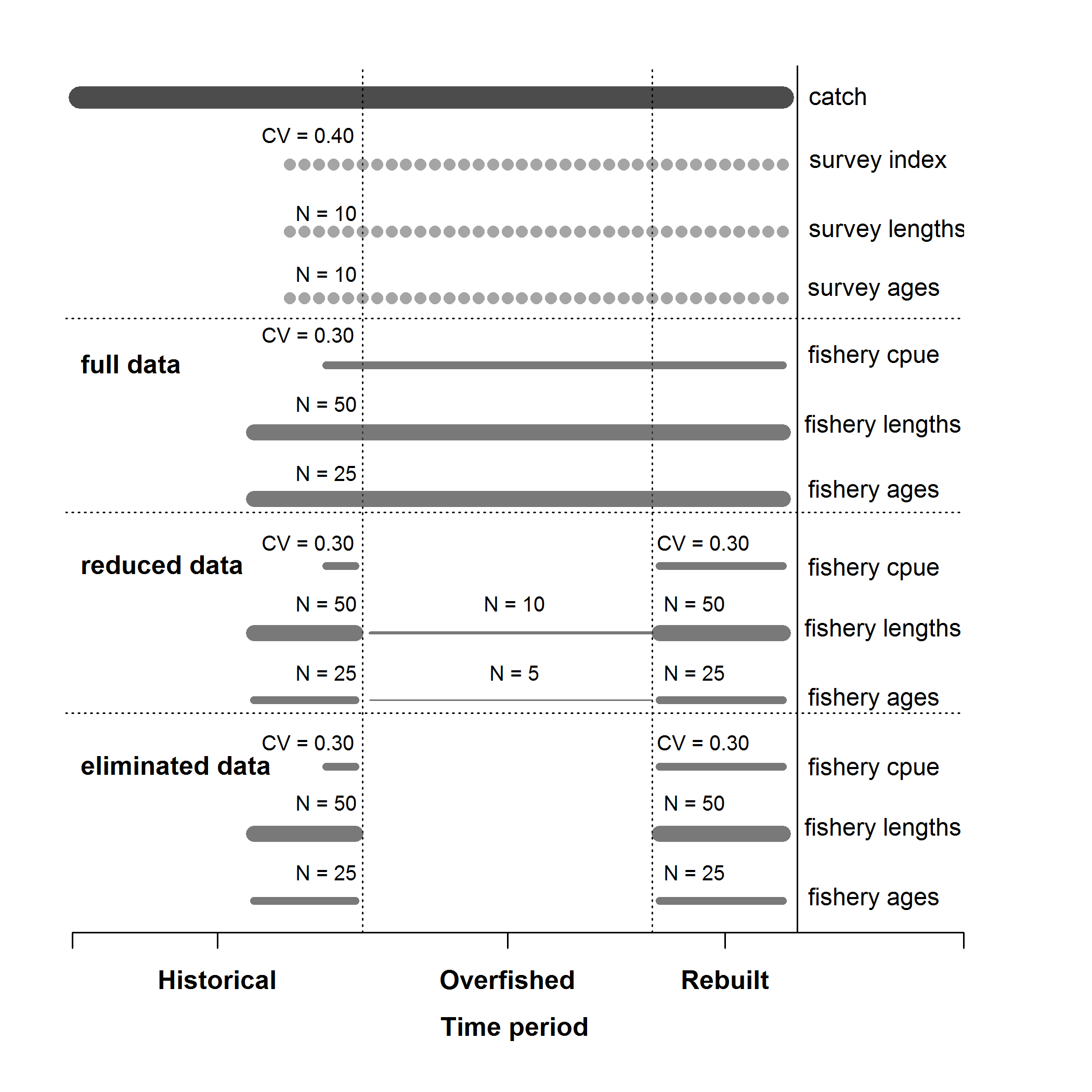


Figure A.1. Summary of the data available for each of the data scenarios. Catches, a fishery independent survey with length- and age-composition data were available across all data scenarios.

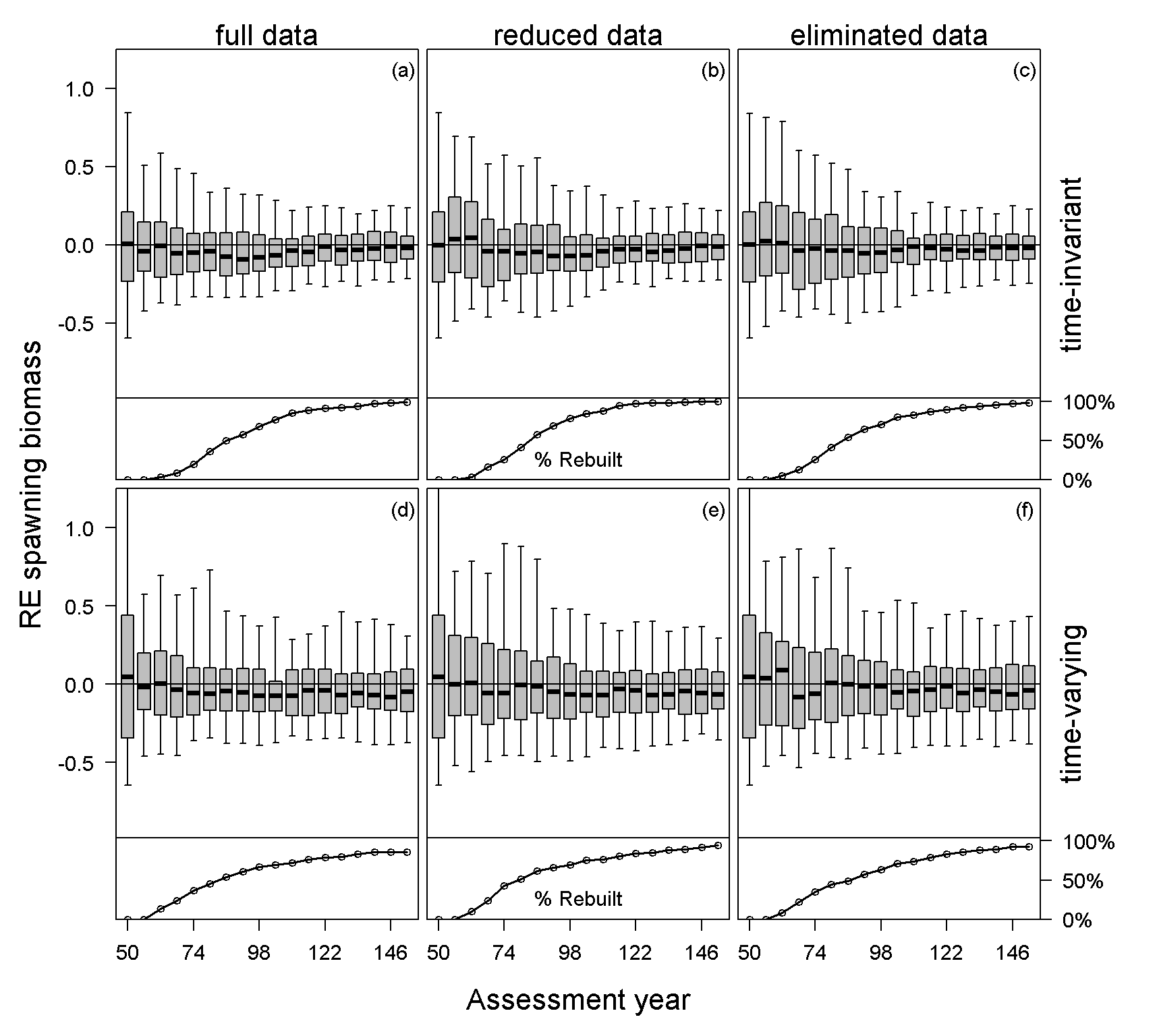


Figure A.2. The relative error of estimated spawning biomass in each assessment year for each data scenario and case summarized in each assessment year for all simulations (top panel) and the percentage of stocks that had rebuilt over time (bottom panel), with data collection consequently returning to historical levels, shown along the bottom of each panel. The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

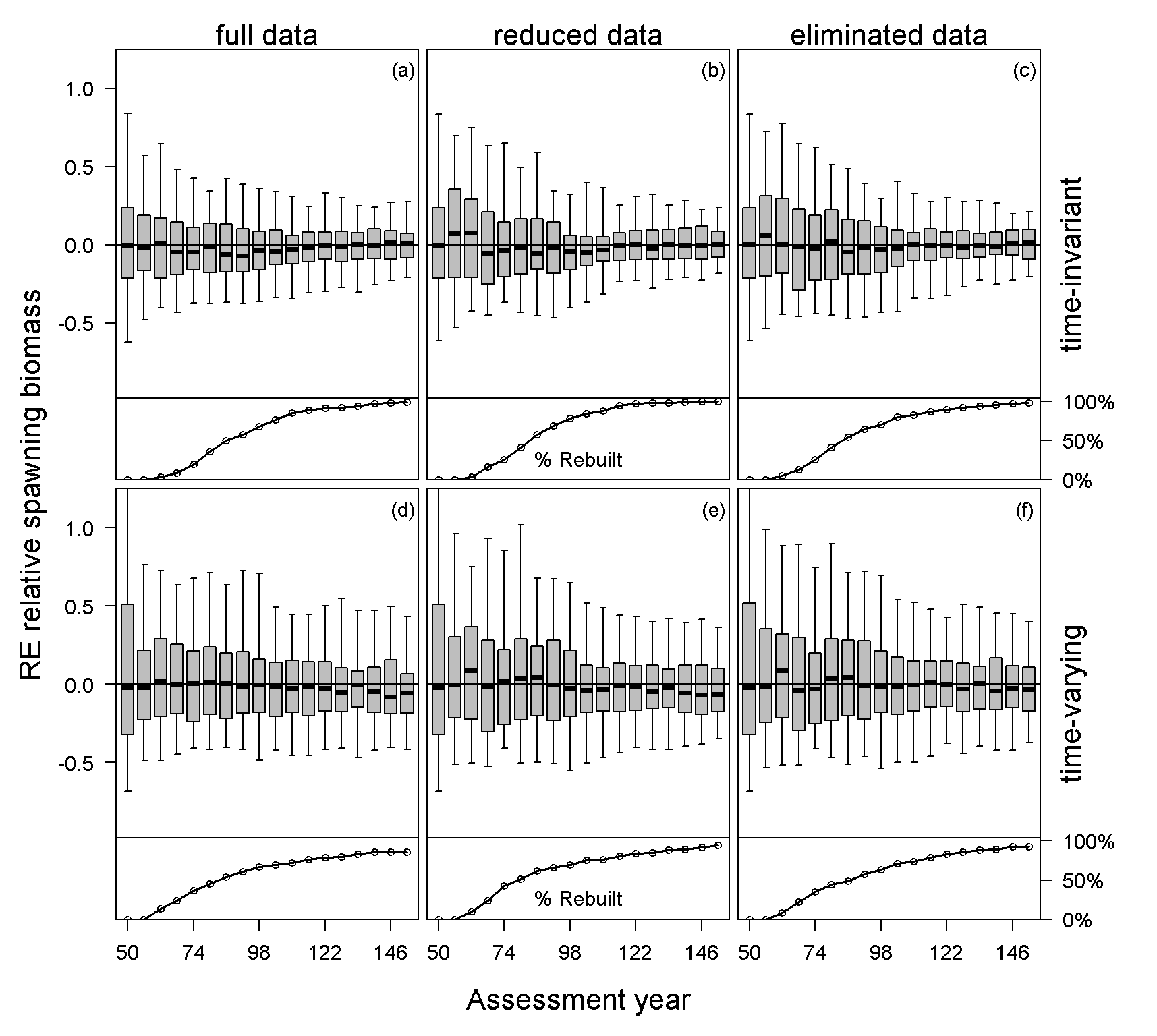


Figure A.3. The relative error of estimated relative spawning biomass in each assessment year for each data scenario and case summarized in each assessment year for all simulations (top panel) and the percentage of stocks that had rebuilt over time, with data collection consequently returning to historical levels (bottom panel). The median is denoted by the black line and the boxplot whiskers cover the 95% simulation interval for each assessment year.

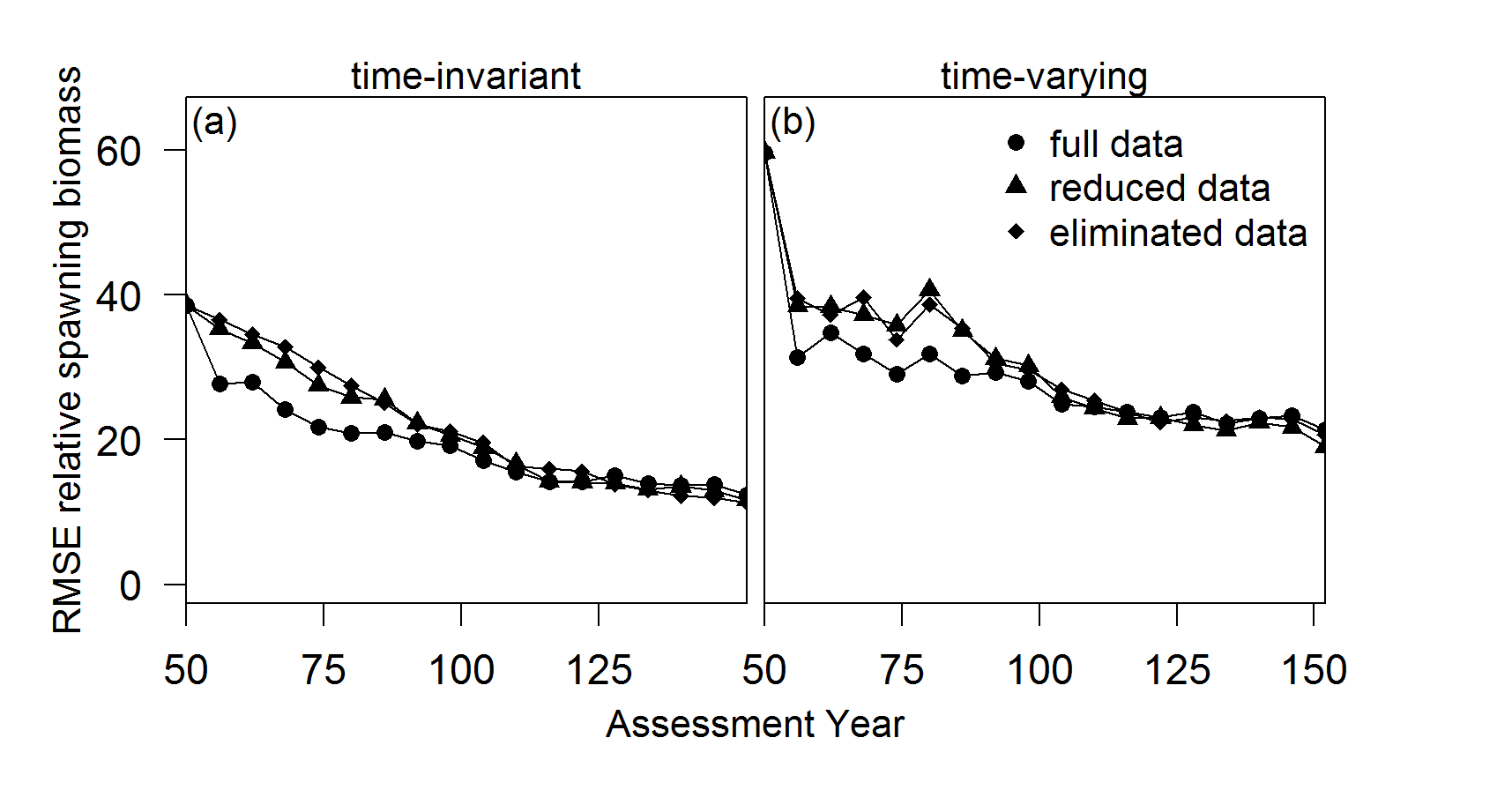


Figure A.4 The root mean square error about relative spawning biomass in the assessment year for each cases and data scenario.