*Biological Sub-model*

*Larkin model*

The Larkin model (Larkin REF) is a modified version of the Ricker that accounts for delayed density dependence between cycle lines. As a result, it includes multiple parameters and lagged spawner abundances.

Equation A1

where *i* represents a CU, *y* is a given year, *R* the number of recruits (number of offspring that return to spawn or are captured in the fishery), and *S* the number of spawners. The parameter represents the number of recruits produced per spawner at low abundance and the parametersrepresent density dependent interactions at different time lags. Like the Ricker model, the Larkin is generally linearized to account for normally distributed process error with mean 0 and standard deviation . Unlike the Ricker model used in our analysis, however, we did not generate autocorrelated process variance in Larkin stocks because appropriate parameter values for the autocorrelation coefficient are unavailable in the literature and validating a Larkin-model equivalent was beyond the scope of this study.

*Variation in age-at-maturity*

Although the majority of Fraser River sockeye salmon mature at age 4 (i.e. one year in the gravel, one year of lake residence, and two years of ocean residence), smaller proportions mature at ages 2, 3, and 5, with age structure varying among CUs. We modeled this process by calculating the number of recruits *R* spawning in year *t* in CU *i* as a function of the total number of adult recruits *R’* generated in previous years, multiplied by the mean proportion *p* of fish that return at a given age *g*:

Equation A2

We incorporated multivariate logistic variation into the proportion of mature fish returning at each age as:

Equation A4

where *y* is the brood year (equal to *t*-2, *t*-3, *t*-4 or *t*-5 in Eqn. A3), the summation in the denominator is over ages 2 to 5, is the CU-specific mean proportion of adult fish that return at a given age, is a parameter controlling interannual variability in the proportion returning at each age, and are standard normal deviates (Holt and Bradford 2011 REF). We identified CU-specific parameters using time series of age-specific returns and a grid search.

*Management Sub-model*

*Harvest control rule*

Fraser River sockeye salmon are managed using a harvest control rule that adjusts total allowable catch (TAC) based on two fishery reference points (FRP). Both TACs and FRPs are defined at the management unit (MU) level (i.e. aggregates of conservation units) because MUs exhibit relatively consistent differences in migration timing that moderate their exposure to commercial marine fisheries (Grant REF). The overarching framework for this harvest control rule is referred to as a Total Allowable Mortality (TAM) rule because TACs are regularly reduced based on environmental conditions. Specifically, in-season estimates of recruit abundance are adjusted downwards to account for anticipated mortality, which regularly occurs during in-river migrations to spawning grounds. This management adjustment is set as a proportion of the escapement target (referred to as a pMA) and attempts to ensure that a sufficiently large number of spawners “escape” the fishery to spawn, even if considerable en route mortality occurs. Given uncertainty in how pMAs are generated in any given year, we assumed they were stable in our forward simulation and used median values since 2000 (Table A1).

The Fraser River Panel of the Pacific Salmon Commission meets weekly to assess each MU’s abundance relative to its FRPs, resulting in one of three harvest strategies:

1. If a MU is below its lower FRP the TAC is calculated using a minimum exploitation rate (0.10 for all MUs except for the Late Run MU), which is intended to account for mortality due to test fishing and bycatch in mixed stock fisheries (even though MUs differ in run timing, substantial overlap persists).
2. If a MU is between its lower and upper FRP, a constant escapement harvest strategy is used to calculate TAC. The escapement target is the lower FRP, adjusted upwards based on estimates of en route mortality (i.e. the pMA). For example, if the FRP is 100,000 individuals and the pMA is 0.5 that year, reflecting relatively high levels of loss en route, the TAC will be calculated assuming an escapement target of 150,000 spawners. The exception to this rule is the target exploitation rate must be at least the minimum noted above and cannot exceed 0.6.
3. If a MU is above its upper FRP (after incorporating the pMA), the TAC is calculated using a maximum target exploitation rate of 0.6.

The in-season abundance estimates necessary to generate TACs are provided by test fisheries conducted at regular intervals as adult salmon migrate into nearshore areas (i.e. Johnston and Juan de Fuca straits). MU-specific abundance is estimated using genetic stock identification techniques conducted on a subsample of test fishery catches (Beacham et al. 2005).

We simulated the in-season estimation process as

Equation A1

where the estimated abundance of recruits *Ȓ* in MU *m* and *y* is assumed to be a function of true recruit abundance *R* plus normally distributed observation error with mean and standard deviation 0.15.

Observation error was parameterized using deviations between in-season and post-season estimates of salmon abundance from 2005-2011 (Fraser River Panel reports). Given that estimates of in-season abundance are updated throughout the migration period, multiple in-season TACs are produced for each MU and each year. Therefore when parameterizing forecast uncertainty, we compared the final in-season run size estimate generated after the estimate of migration timing was fixed (i.e. 50% migration date had been finalized) to post-season estimates of abundance, which incorporate data collected in freshwater migration corridors and on the spawning grounds. MU-specific FRPs, which may vary by cycle line, are listed in Table A2 and an example TAM rule calculation is shown in Figure A1. Most MUs exhibited similar deviations and was set to 1.2 for all MUs except Early Summers, which were frequently underestimated ( = 0.85).

Table A2. MU-specific fishery reference points (in millions of fish) across cycle lines.

|  |  |  |  |
| --- | --- | --- | --- |
| **Management Unit** | **Cycle Line** | **Lower FRP** | **Upper FRP** |
| Early Stuart | All cycle lines | 0.108 | 0.1512 |
| Early Summer | 1 | 0.11 | 0.154 |
| 2 | 0.18 | 0.252 |
| 3 and 4 | 0.1 | 0.14 |
| Summer | 1 | 0.885 | 1.239 |
| 2 | 1.02 | 1.428 |
| 3 | 0.76 | 1.064 |
| 4 | 0.64 | 0.896 |
| Late | 1 | 0.35 | 0.49 |
| 2 | 1.1 | 1.54 |
| 3 and 4 | 0.3 | 0.42 |

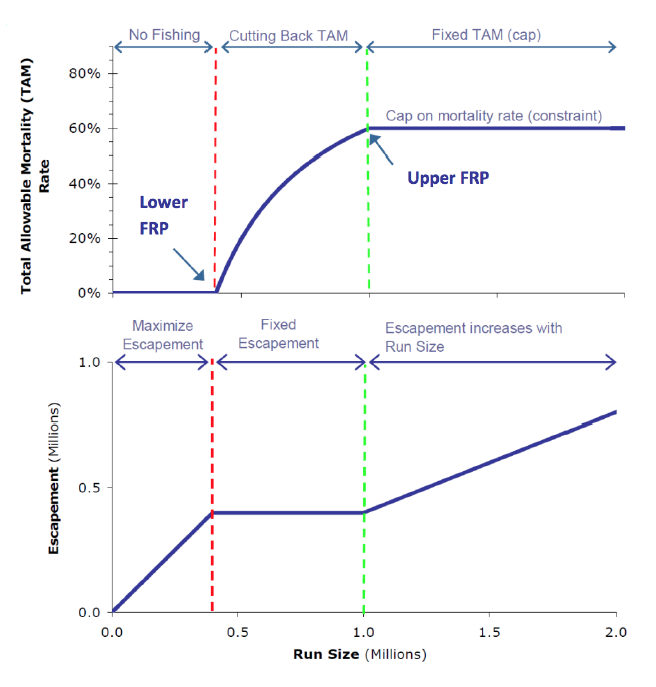


Figure A1. Changes in total allowable mortality (upper panel) and escapement target (lower panel) as a function of run size when using TAM rule harvest strategy. Here run size has been adjusted using pMA.

*Fisheries and en route mortality*

Since the majority of fishing mortality occurs at sea, while en route mortality occurs in-river, we modeled each process sequentially. We note that Fraser River sockeye salmon are also harvested in a variety of in-river fisheries, preliminary analyses indicated our results were not strongly influenced by the inclusion of CU-specific in-river fisheries.

Realized exploitation rates can deviate from targets substantially due to variation in catchability, enforcement, or unreported catch. These processes collectively result in outcome uncertainty and can strongly influence the efficacy of management strategies. We incorporated outcome uncertainty in our model by generating realized harvest rates *H* for each CU *i* within MU *m* as

Equation A3

where is the target TAC, *R* the true abundance of recruits, and an error term representing CU-specific outcome uncertainty. Thus each CU within an MU had the same target harvest rate, but realized harvest rates would differ. Since there is no evidence to suggest a persistent bias in sockeye salmon catches, the error distribution had mean zero and standard deviation .

Similarly, en route mortality *D* was modeled as a stochastic, CU-specific process.

Equation A5

where *E* represents the median and the standard deviation of observed en route mortality since 2000 for each MU.

*Biological benchmarks*

Biological benchmarks are commonly used to assess population status relative to a desired state (REF). In this study, we calculated benchmarks derived from stock-recruit relationships and referenced in Canada’s Wild Salmon Policy (REF). The upper benchmark is the estimated spawner abundance necessary to achieve maximum sustainable yield (*S*MSY), estimated using the Lambert W function following Scheuerell (2016)

Equation 8

The lower benchmark is the estimated spawner abundance necessary to recover to *S*MSY in one generation in the absence of fishing mortality (*S*gen), which was solved numerically according to the following equation (Holt et al. 2009)

Equation 9

SMSY is intended to represent an abundance at which a CU can sustain harvest and provide its full suite of ecosystem services indefinitely. Sgen is intended to be precautionary lower benchmark. Therefore, it represents an abundance at which a CU is at increased risk of extirpation, particularly if additional mortality is introduced by harvest; however Sgen is greater than the abundance that would trigger protection under at-risk species legislation. Within Canada’s Wild Salmon Policy framework, CUs with an abundance greater than SMSY are considered green status, those with abundances below SMSY and above Sgen are amber, and those below Sgen are red (WSP REF).