



# Climate presses and pulses mediate the decline of a migratory predator

T. J. Clark-Wolf<sup>a,1</sup> , P. Dee Boersma<sup>a</sup>, Ginger A. Rebstock<sup>a</sup>, and Briana Abrahms<sup>a</sup>

Edited by Eric Post, University of California, Davis, CA; received June 7, 2022; accepted November 15, 2022 by Editorial Board Member Pablo A. Marquet

Long-term climate changes and extreme climate events differentially impact animal populations, yet whether and why these processes may act synergistically or antagonistically remains unknown. Disentangling these potentially interactive effects is critical for predicting population outcomes as the climate changes. Here, we leverage the “press–pulse” framework, which is used to describe ecological disturbances, to disentangle population responses in migratory Magellanic penguins to long-term changes in climate means and variability (presses) and extreme events (pulses) across multiple climate variables and life history stages. Using an unprecedented 38-y dataset monitoring 53,959 penguins, we show for the first time that the presses and pulses of climate change mediate the rate of population decline by differentially impacting different life stages. Moreover, we find that climate presses and pulses can work both synergistically and antagonistically to affect animal population persistence, necessitating the need to examine both processes in concert. Negative effects of terrestrial heat waves (pulses) on adult survival, for example, were countered by positive effects of long-term changes in oceanographic conditions in migratory grounds (presses) on juvenile and adult survival. Taken together, these effects led to predicted population extirpation under all future climate scenarios. This work underscores the importance of a holistic approach integrating multiple climate variables, life stages, and presses and pulses for predicting the persistence of animals under accelerating climate change.

animal conservation | climate change | extreme events | Magellanic penguins

Global climate change is causing changes in climate means, variability, and extremes at an alarming pace (1). Such changes drive a broad array of ecological responses, with often dire consequences for animal populations. Research in the past few decades has highlighted the impact of long-term changes in climate means and variability on animals (2–5). Recently, the increasing frequency of extreme events has been found to be as or more important than changes in climate means and variability in driving population-level responses (6–10). Moreover, changes in climate means, variability, and extremes likely do not act in isolation from one another, but rather may interact synergistically or antagonistically to influence population outcomes (8, 11). Despite the importance of understanding the influence of climate change on Earth's ecosystems, however, it is poorly understood how interactive changes in long-term climate means, variability, and extreme events coalesce to impact animal populations.

Because the frequency of extreme events is inextricably linked to the mean and variability of climate conditions, a unifying framework is needed to disentangle the effects of extreme climate events from long-term changes in climate means and variability. The “press–pulse” framework, which has been used to describe ecological disturbances (12), can help to explain ecosystem responses to long-term changes in climate means and variability (presses) and extreme events (pulses) (11) (Fig. 1). Under this framework, a long-term change in the mean or variability of a climate variable like precipitation is considered a long-term perturbation, or press, on an animal population. An extreme event, which is typically defined as the outer 5% in climate conditions (6), is considered a one-time discrete perturbation, or pulse, which can be imposed in concert with long-term presses. Presses and pulses, exerted simultaneously, are more likely to push systems past their tipping point and cause regime shifts (13). As climate change accelerates, both the magnitude and frequency of pulse events are expected to increase, which when superimposed on the increasing press of long-term climate change, can have complex and unexpected consequences for animal populations, persistence, and conservation outcomes.

A similarly concerning problem is that animals may be differentially impacted by multiple, co-occurring presses and pulses associated with different climate factors. Thus, an observed effect of a climate variable on a vital rate may be outweighed or complicated by the effects of potentially unconsidered, opposing climatic variables on other vital rates.

## Significance

Anthropogenic climate change has caused gradual, long-term changes in climate conditions, along with more frequent extreme climate events. Yet it is unclear how the ecological effects of long-term changes and extreme events interact, and what that means for our ability to mitigate climate threats to animal populations. Using a multidecadal dataset on Magellanic penguins, we show that both long-term changes and extreme events impact population persistence in interacting and opposing ways by differentially affecting different life stages, ultimately leading to predicted population extirpation. Such findings highlight the complexity of predicting population responses to climate change and demonstrate that accounting for the interactive effects of long-term climate changes and extreme events is essential to accurately predict outcomes for animal populations under increasing climate change.

Author contributions: T.J.C.-W., P.D.B., G.A.R., and B.A. designed research; T.J.C.-W., P.D.B., G.A.R., and B.A. performed research; T.J.C.-W. analyzed data; and T.J.C.-W., P.D.B., G.A.R., and B.A. wrote the paper.

The authors declare no competing interest.

This article is a PNAS Direct Submission. E.P. is a guest editor invited by the Editorial Board.

Copyright © 2023 the Author(s). Published by PNAS. This article is distributed under [Creative Commons Attribution-NonCommercial-NoDerivatives License 4.0 \(CC BY-NC-ND\)](https://creativecommons.org/licenses/by-nc-nd/4.0/).

<sup>1</sup>To whom correspondence may be addressed. Email: [tc130053@uw.edu](mailto:tc130053@uw.edu).

This article contains supporting information online at <https://www.pnas.org/lookup/suppl/doi:10.1073/pnas.2209821120/-/DCSupplemental>.

Published January 9, 2023.

For example, song sparrows (*Melospiza melodia*) in coastal California have higher adult survival after warm, dry winters but also higher juvenile survival after rainy winters the year prior (14). Moreover, climate presses and pulses can affect demography in opposing ways, highlighting the need to consider how climate varies among seasons and life cycles (15). In Eurasian oystercatchers (*Haematopus ostralegus*), for instance, rising winter temperatures have negative effects on its fecundity which are offset by positive effects on adult survival, ultimately improving its population persistence (16).

Migratory animals pose particular concern for such impacts because they are susceptible to climate changes across disparate geographies and life history stages (17). For example, black-throated blue warblers (*Setophaga caerulescens*) are subject to intensifying El Niño conditions that limit food availability. Such climate changes both lower adult survival in warblers' wintering range in the Greater Antilles as well as reduce fecundity in their breeding range in eastern North America (18). Given that migratory species are largely experiencing population declines worldwide (17, 19), there is an urgent need to understand how the presses and pulses of climate change impact migratory populations. However, the differential impacts of climate presses and pulses, operating across different climate variables and on multiple life stages of organisms, obscure a holistic understanding of how climate change affects the persistence of animal populations. We address this using a 38-y demographic dataset on a migratory predator, the Magellanic penguin (*Spheniscus magellanicus*).

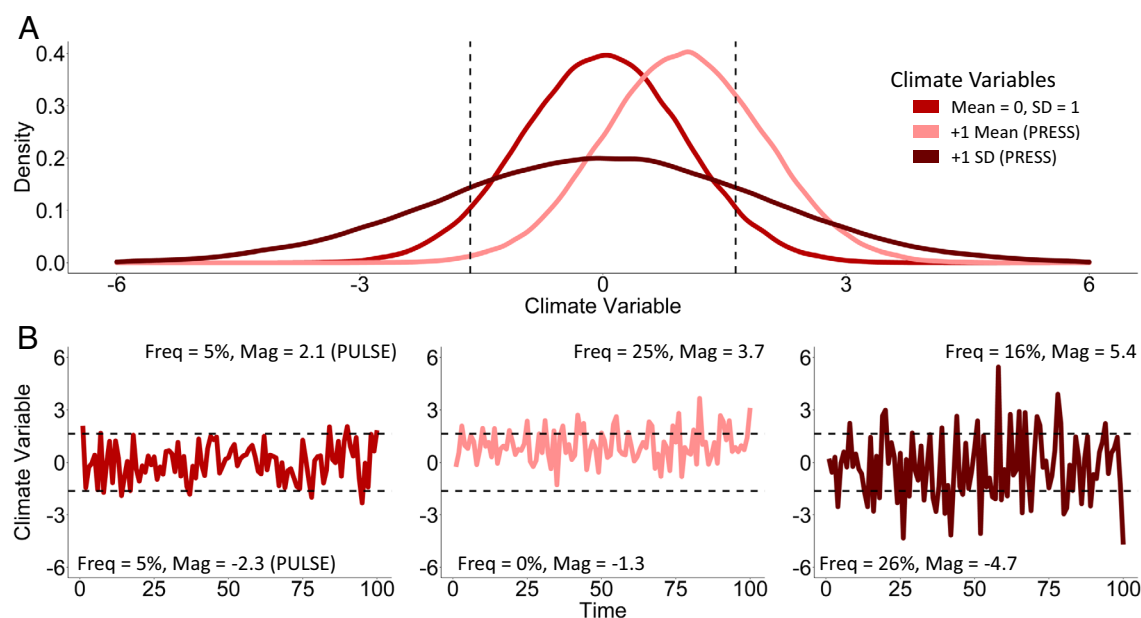
Magellanic penguins are a migratory species that breed in South America (20, 21). Historically, the largest population of Magellanic penguins bred at Punta Tombo, southern Argentina (Fig. 2*A*) (22). Penguins at Punta Tombo arrive in the colony in September–October, breed from October–February, and then migrate beginning February–April. Individuals migrate as far as 3,000 km to reach their wintering grounds off the coast of northern Argentina, Uruguay, and southern Brazil (Fig. 2*A*) (21). Because penguins

forage and migrate at sea but breed on land, they are subject to diverse climate impacts across marine and terrestrial environments in disparate geographies. Moreover, these climate conditions impact penguins throughout multiple life stages.

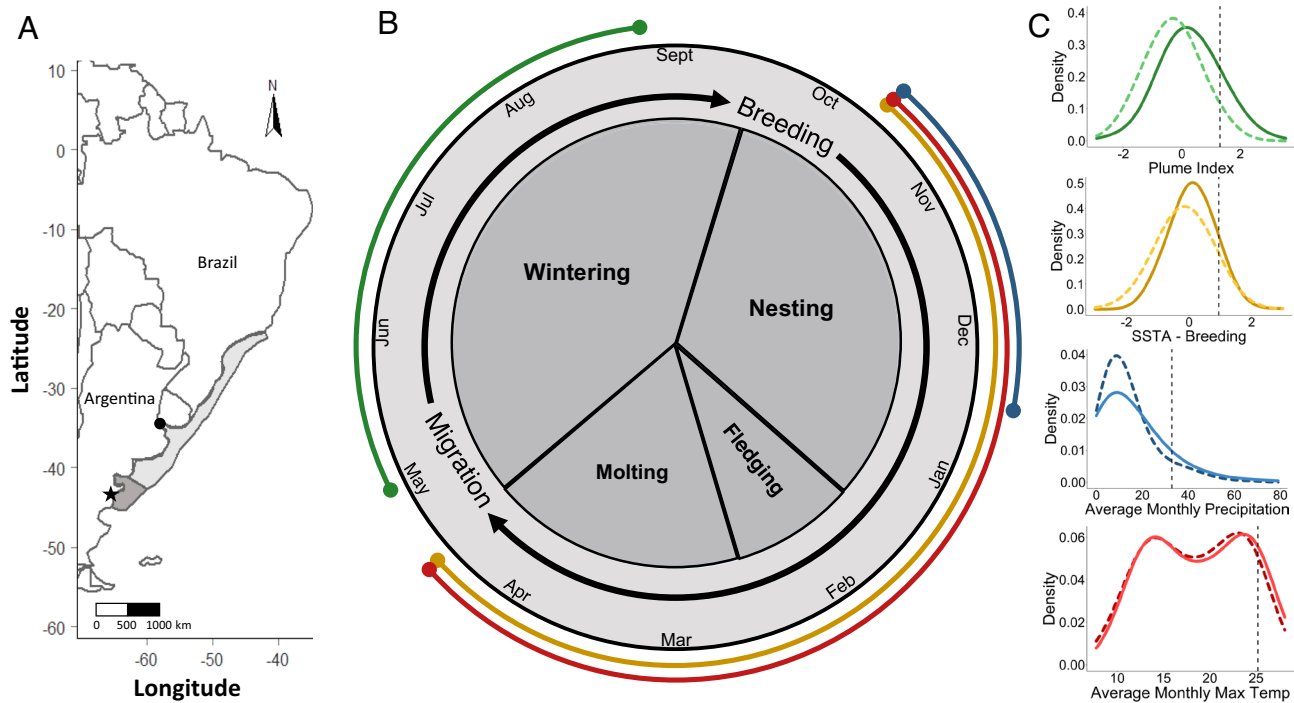
Magellanic penguins exhibit a wide range of behavioral and demographic responses to several climate presses and pulses. Warm, unproductive oceans in the colony's foraging area during the breeding season cause penguins to forage further out from the colony in search of food, leading to lower reproductive success in warmer ocean conditions (23, 24). During migration, the Rio de la Plata in northern Argentina, which drains the second largest basin in South America, carries water from the river to the penguin's wintering area (Fig. 2*A*) (25). In years when southwesterly winds drive stronger Rio de la Plata plumes, prey are more dispersed and harder to encounter, causing declines in adult body condition and fledging success (26, 27). Climate change has driven warmer waters in the colony's foraging area (affecting reproductive success) and weaker plumes in the colony's wintering area (affecting survival), which can manifest as both oceanic presses and pulses on the penguin population (26, 28, 29).

In addition to these oceanic variables, two terrestrial climatic variables have been shown to significantly affect penguins at Punta Tombo (30). Increased average rainfall (presses) and intense storms (pulses) in the breeding colony kill chicks due to exposure, increasing reproductive failure (30, 31). In addition, rising temperatures (presses) can negatively affect the physiology of these penguins, which is particularly a concern with recent heat waves (pulses) that lead to large mortality events of both chicks and adults (32). For example, the largest single-day penguin mortality event occurred in 2019 when air temperatures exceeding 44 °C caused adults and chicks to die due to heat exposure (32).

Thus, these four climate variables—ocean temperature, plume strength, rainfall, and air temperature—manifest as both climate presses and pulses which differentially affect penguin demographic rates like fecundity, recruitment, and adult survival. Moreover,



**Fig. 1.** Presses and pulses of climate change are linked. (A) A normally distributed climate variable with mean = 0 and SD = 1, undergoing a press of climate change that increases its mean, or increases its variability. (B) Time series randomly generated from their respective frequency distributions in A, showing that increasing the long-term mean or variability (press) of climate variables also increases the number of extreme events, or pulses. Dotted black lines in both panels represent 5% and 95% CIs for the normally distributed climate variables, and climate values above the 95% CI or below the 5% CI are considered extreme events (pulses). Text in B shows the frequency and magnitude of pulses, where the frequency represents the proportion of pulses in that time series above the 95% CI (Top) and below the 5% CI (Bottom), and the magnitude represents the maximum positive and negative pulse values of that time series.



**Fig. 2.** The annual range, life cycle, and climate experienced by Magellanic penguins at Punta Tombo, Argentina. (A) The average breeding and wintering ranges. The star symbol shows the location of the breeding colony at Punta Tombo, and the circle symbol shows the Rio de la Plata plume source, which drains the Rio de la Plata and carries cool and productive water to the penguin's wintering area. The dark shaded polygon near the star represents the maximum foraging extent of penguins foraging to feed chicks during the breeding season, and the lighter shaded polygon near the plume represents the migratory range of penguins. (B) The annual life cycle of penguins, beginning with arrival to the colony in September–October. (C) Frequency distributions of the major climate variables impacting penguins in their breeding and wintering range, where the colors of the climate variables correspond to the period of life cycle shown in B. SSTA (sea surface temperature anomaly) and plume index (strength of the Rio de la Plata plume) in C show climate from 1982 to 2000 (dotted lines) and 2001 to 2020 (solid lines), whereas the precipitation and maximum air temperature show climate from 1958 to 1988 (dotted lines) and 1989 to 2019 (solid lines). Dashed vertical lines show the 95% CI of the earlier periods for the climate variables.

the combination and interaction of these presses and pulses are likely contributing to a decline in population size at Punta Tombo since the 1980s (33). Disentangling which presses and pulses are the biggest drivers of population decline in the colony may help to illustrate population responses to climate changes and target conservation interventions.

Here, we build an integrated population model (IPM) to disentangle the impacts of presses and pulses across multiple climate variables and life history stages on the population trajectory of Magellanic penguins, following evidence showing the impacts of climate presses and pulses for climate and oceanographic variables on penguin demography. Using an unprecedented long-term dataset that follows the fates of 53,959 individuals, we aim to address the following: 1) What are the demographic responses of penguins to climatic and oceanographic presses and pulses; 2) What is the sensitivity of the population growth rate to these presses and pulses; and 3) What are the expected trends and outcomes for penguins given future climate change scenarios? We show how the presses and pulses of climate change mediate the population decline of Magellanic penguins by differentially impacting different life stages, and ultimately leading the population to potentially approach extirpation in the next 50 y.

## Results

**Climate Is Changing in Both Penguin Breeding and Wintering Grounds.** The climate in southern Argentina surrounding Punta Tombo is becoming warmer and wetter (34, 35). Regional average monthly precipitation extremes, defined as months with total precipitation greater than the 95th percentile, have increased by 41.6% in the last 30 y compared to the preceding 30 y, and

average monthly maximum air temperatures have increased by 2.4% (Fig. 2C). In addition, mean sea surface temperature anomalies (SSTA) in the ocean surrounding the penguin's foraging area during breeding warmed over the past 38 y ( $\bar{x} = -0.15$  to 0.11) (Fig. 2C), which correlates with reduced chlorophyll-a and declining ocean productivity (36). Lastly, average plume strength has weakened over the past 38 y ( $\bar{x} = -0.31$  to 0.29; positive values correspond to weaker plumes) (Fig. 2C), consistent with climate-driven changes in regional and local wind conditions (26, 28, 29). The increasing presses of climate and ocean variables drive more frequent and higher magnitude pulses of these variables (Figs. 1 and 2). Higher means and SDs in climate presses resulted in greater frequency and severity of climate pulses (SI Appendix, Fig. S1).

**Climate Presses and Pulses Differentially Affect Penguin Demographic Rates.** We collected survival data from 53,959 Magellanic penguins at Punta Tombo, as well as reproductive and population count data, between 1982 and 2019. To estimate the fecundity, juvenile and adult survival, and population size (number of breeding pairs) of the colony, we developed an IPM to fit the data. Adult survival has become more variable in the past 10 y than previously (ranging from 0.65 to 0.96 and SD from 2008 to 2019 = 0.13; SD from 1982 to 2007 = 0.051), whereas juvenile survival has become less variable during that time (ranging from 0.01 to 0.11 and SD from 2008 to 2019 = 0.037; SD from 1982 to 2007 = 0.13) (SI Appendix, Fig. S2), consistent with previous findings on survival trends in the colony (37). The penguin colony at Punta Tombo has fallen by approximately 50% in total size since the 1980s, with a population growth rate of 0.98 (95% CI: 0.977, 0.985) (SI Appendix, Fig. S2). Population growth was most sensitive to adult survival ( $r = 0.66$ , 95% CI: 0.434,



0.813;  $P < 0.001$ ), followed by juvenile survival ( $r = 0.39$ , 95% CI: 0.079, 0.636;  $P = 0.01$ ) and immigration ( $r = 0.33$ , 95% CI: 0.021, 0.686;  $P = 0.01$ ), but not significantly sensitive to fecundity ( $r = -0.07$ , 95% CI:  $-0.383$ ,  $0.262$ ;  $P = 0.69$ ) (SI Appendix, Fig. S3). This indicates that strong fluctuations in adult survival and low juvenile survival may contribute to the population decline of penguins at the colony.

In order to estimate the effects of presses and pulses on penguin population persistence, we need to first calculate the relationship between climate variables and demographic rates. To estimate the general effect of climate variables on demographic rates, we used regressions in the IPM to relate climatic (rain and air temperature in the breeding grounds) and oceanographic (SSTA in the breeding grounds and plume strength in the wintering grounds) variables to fecundity, juvenile, and adult survival. Climate variables differentially impacted different life stages and demographic rates (Fig. 3). Rain negatively impacted fecundity ( $\bar{x} = -0.34$ , 95% credible intervals (CRI):  $-0.391$ ,  $-0.286$ ). Air temperature had a small negative effect on adult survival ( $\bar{x} = -0.09$ , 95% CRI:  $-0.129$ ,  $-0.058$ ), but had no influence on fecundity. SSTA in the breeding grounds negatively affected juvenile survival ( $\bar{x} = -0.40$ , 95% CRI:  $-0.440$ ,  $-0.360$ ) but had only very weak effects on fecundity and adult survival. The plume strength in the wintering grounds strongly influenced adult and juvenile survival, with higher juvenile and adult survival in years with weak plumes ( $\bar{x} = 0.61$ , 95% CRI:  $0.555$ ,  $0.674$ ; and  $\bar{x} = 0.29$ , 95% CRI:  $0.222$ ,  $0.345$ ; respectively). However, the prior winter's plume did not impact penguin fecundity.

To determine how presses and pulses differentially impacted penguin population growth rates, we performed a sensitivity analysis of the IPM by varying the means and SDs (long-term presses) and frequency and magnitude of extreme events (pulses). We first varied means and SDs to examine effects if only presses, but not pulses, were considered. Under this scenario, mean plume strength had the greatest impact on population growth ( $\bar{x} = 0.96$ , 95% CI:  $0.954$ ,  $0.963$ ), followed by the shape and rate of rain ( $\bar{x} = -0.14$ , 95% CI:  $-0.202$ ,  $-0.065$ ; and  $\bar{x} = 0.38$ , 95% CI:  $0.333$ ,  $0.443$ ; respectively) and mean SSTA ( $\bar{x} = 0.20$ , 95% CI:  $0.130$ ,  $0.269$ ) (Fig. 4A). Importantly, however, incorporating pulses changed the results of our sensitivity analysis. When both presses and pulses were considered, the means of SSTA and air temperature (presses) no longer affected population growth (Fig. 4A and B), whereas

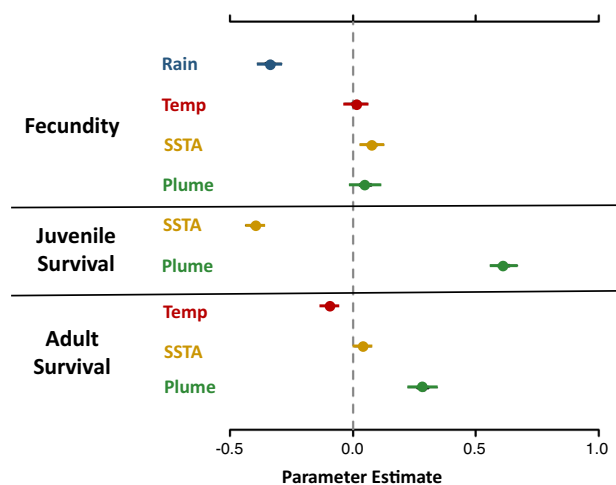
the frequency of air temperature extremes (pulse) had a strong negative effect ( $\bar{x} = -0.15$ , 95% CI:  $-0.225$ ,  $-0.090$ ) (Fig. 4B). In addition, years with higher frequency and magnitudes of weak plumes significantly improved population growth ( $\bar{x} = 0.24$ , 95% CI:  $0.182$ ,  $0.299$ ; and  $\bar{x} = 0.12$ , 95% CI:  $0.057$ ,  $0.177$ ; respectively). Therefore, presses and pulses, operating across multiple climate variables and life stages, differentially impacted penguin population growth.

**Climate Presses and Pulses Mediate Penguin Population Persistence.** Climatic and oceanographic presses and pulses differentially impacted penguin demographic rates—but what are the expected outcomes for the population given the trajectory of future climate change? Using the empirically parameterized IPM, we simulated penguin population growth over the next 50 y under future climate scenarios, including examining the sole effects of the strongest presses and pulses (Fig. 4). We explored four scenarios: 1) no change in climate variables from present; 2) projected climate change according to the current rate of change in climate variables ( $\pm$ SD); 3) change in the mean strength of the plume (press;  $\pm$ SD); and 4) change in the frequency of air temperature extremes (pulse;  $\pm$ SD) (Fig. 5). Under the current projected rate of climate change for all variables, the population projected rate of climate change for all variables, the population growth rate will be higher (but still negative) than if the climate did not change from present ( $\bar{x} = 0.95$ , 95% CI:  $0.951$ ,  $0.961$ ; and  $\bar{x} = 0.93$ , 95% CI:  $0.929$ ,  $0.937$ ; respectively), due to the positive effects of the weakening plume on adult and juvenile survival. Confirming this finding, changes in solely the mean strength of the plume (press) had an even greater positive effect on population growth rate ( $\bar{x} = 0.97$ , 95% CI:  $0.969$ ,  $0.975$ ). In contrast, changes in solely the most impactful pulse, the frequency of air temperature extremes, drove growth rates even lower than other scenarios ( $\bar{x} = 0.92$ , 95% CI:  $0.911$ ,  $0.928$ ). In all scenarios, the population trajectory eventually approached extirpation (defined as population size near 0), given the population growth rate  $< 1$ . In sum, opposing effects of presses and pulses of different climate variables lead to the mediation of penguin population decline.

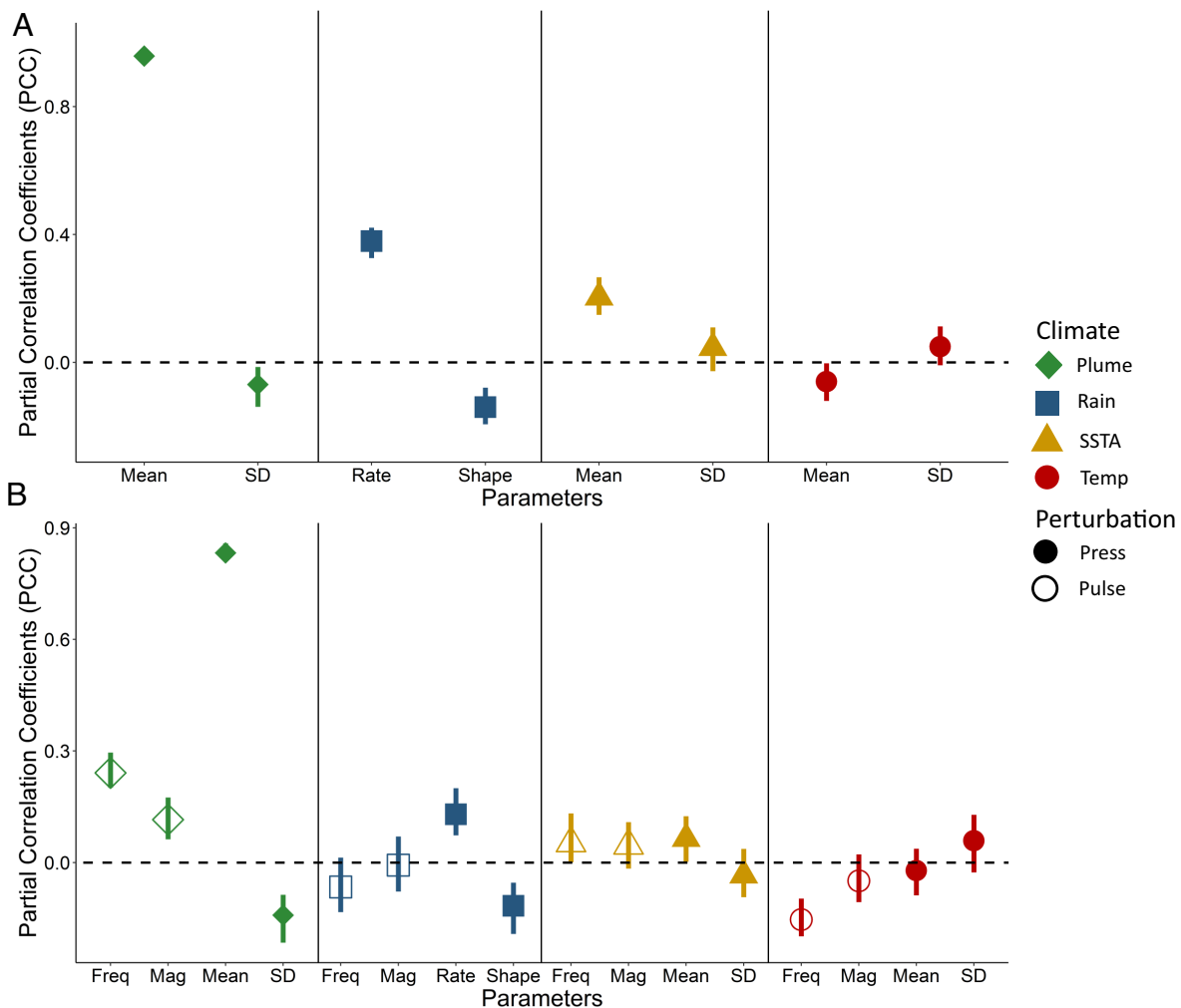
## Discussion

**Understanding and Conserving Migratory Species under the Presses and Pulses of Climate Change.** Our work highlights the importance of disentangling how animals are differentially impacted by multiple, co-occurring climate factors across their life cycle. Yet, many studies focus on the effect of single climate factors on one component of demography across an animal's life cycle (in contrast, see ref. 8, 38, and 39). Further, our work highlights the challenges faced by migratory species that are exposed to multiple presses and pulses throughout their migration cycle and across broad geographies. We showed that migratory Magellanic penguins experienced contrasting effects of climatic and oceanic presses and pulses on demography and population persistence that mediated the rate of population decline. For example, we found that both the presses of the Rio de la Plata plume and the pulses of air temperature extremes affected penguin population persistence in opposing ways. Such results underscore the complexities animal populations face under climate change, as well as both the challenges and opportunities for researchers in accurately predicting population responses to future climate scenarios.

Understanding which climate variables most affect population growth, and where these impacts occur geographically, not only aids prediction of population responses to climate change but can also inform conservation planning. For example, our results revealed that adult survival, followed by juvenile survival, contributes most



**Fig. 3.** Relationship between climate variables and demography. Coefficient means and 95% CRIs from our IPM showing the effects of the respective climate variables on different components of penguin demography. A higher plume index represents a weaker plume. Means and 95% CRIs also shown in SI Appendix, Table S1.

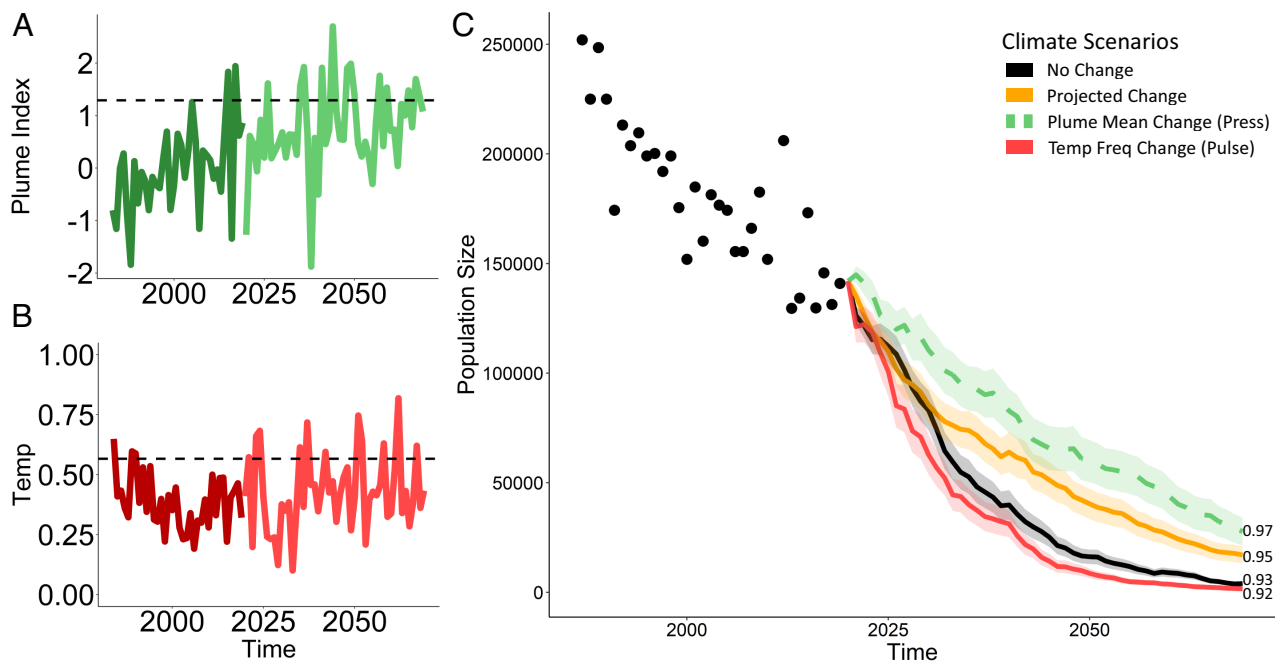


**Fig. 4.** Relationship between climate presses and pulses and penguin population growth. (A) Means and 95% CIs for the sensitivity analysis of solely climate presses on penguin population growth. (B) Means and 95% CIs for the sensitivity analysis integrating both climate presses and pulses on penguin geometric lambda. Symbols on or near the horizontal dashed line mean that the population growth rate did not correlate with that climate press or pulse. Means and 95% CRIs of the full results are also shown in *SI Appendix, Tables S2 and S3*.

to population growth (*SI Appendix, Fig. S3*) and that each of these processes was heavily affected by Rio de la Plata plume conditions which affect prey availability during energy-intensive foraging migrations (Fig. 3). These findings suggest the need to protect migratory foraging areas near the plume, where currently no habitat protection exists, to help protect the penguins' forage base during migration and buffer against strong plume years (26). Such understanding may also inform prioritization of limited conservation resources. For example, although penguin fecundity is negatively impacted by rainstorms (Fig. 3), these processes do not strongly impact population persistence (Fig. 4) because the population growth rate is relatively insensitive to fecundity (*SI Appendix, Fig. S3*). In contrast, heat waves had a significant impact on adult survival, which strongly affected population growth, suggesting that conservation resources that mitigate the effects of heat waves would be relatively more cost-effective. Therefore, following our framework, conservation strategies can identify which environmental processes are demographically more important (e.g., heat waves and plume strength) and which may be less important (e.g., rainstorms) to better concentrate conservation efforts.

This research also highlights the relevance of disentangling the importance of climate presses versus pulses on population persistence. Understanding the relative impacts of presses versus pulses may help identify the timescales of conservation strategies (e.g.,

short term vs. long term) needed to promote persistence. For example, it may be necessary to mitigate short-term exposure of penguins to heat stress during heat waves, but also necessary for long-term management of ocean habitat or fisheries along migratory corridors (40). Furthermore, recognizing the impacts of increasing climate variability, manifested as both presses and pulses, on population persistence indicates the potential value of implementing dynamic management strategies. Dynamic management, whereby management is adjusted dynamically based on changing environmental, biological, or socioeconomic conditions, is increasingly called upon as a means of better coping with environmental variability (41). In the context of conserving migratory predators, proactive seasonal risk assessments based on projected climate conditions could be merged with seasonal management zoning to improve population persistence (42). For example, in seasons or years with predicted poorer ocean productivity and stronger plumes, regulations of fisheries and/or petroleum extraction or transport along penguin migratory corridors could be spatially or temporally adjusted to offset poor foraging conditions and enhance protection of penguins. Thus, identifying how presses and pulses across multiple climate variables differentially impact population growth may help target the types, timescales, and locations of conservation actions that will maximally improve population persistence in a changing climate.



**Fig. 5.** Projected population trajectory of Magellanic penguins under climate change scenarios. (A) The projected plume index based on the current rate of change in mean plume presses. (B) The projected air temperature based on the current rate of change in the frequency of air temperature pulses. The dark parts of the time series represent measured values, whereas light parts of the time series represented projected values. Dashed lines show 95% CIs of the measured climate variables. (C) The population trajectory of simulated penguin populations under climate scenarios. Black dots show the observed penguin data, where lines show the mean and 95% CI (shading) of averaged simulations in the future. Numbers to the right of lines show the average geometric lambda for simulations.

## Conclusion

Animal populations are undoubtedly impacted by multiple climate variables, operating as presses and pulses across multiple life stages and geographic locations. Few studies account for multiple climate variables in this way (15), but we suggest that incorporating a press–pulse framework (11) will help to disentangle and ultimately help mitigate the effects of climate change on animal populations. This framework is necessary because animal populations are likely to be sensitive to both presses and pulses, as demonstrated here. Moreover, our results highlight the importance of considering the effects of both presses and pulses in tandem, as including pulses uncovered additional influences of climate variables on population growth that ultimately influenced projections of population persistence. Thus, we found that important demographic responses to climate change were masked if only climate presses or pulses were examined in isolation, which may obscure predictions of population trends in the future. We suggest that future studies could advance our understanding by exploring population responses to fluctuating environments; for example by decomposing the population growth rate into components driven by nonlinear relationships between the environment and demography (43). We recommend that a holistic approach considering multiple climate variables, life stages, and presses and pulses will help to predict and manage the persistence of animals under accelerating climate change.

## Materials and Methods

**Study System.** The Punta Tombo colony of Magellanic penguins in Argentina (44°03' S, 65°13' W) was established in the 1920s and was once the largest colony of its species, reaching peak size of ~400,000 breeding pairs in the 1960s–1980s (20, 30, 33). Since then, the population has declined to ~150,000 breeding pairs as of 2019. Magellanic penguins in the south Atlantic are migratory from April to August, following the seasonal migration of Argentine anchovy (*Engraulis anchoita*)

to northern Argentina up to southern Brazil (21). Breeding adults arrive at Punta Tombo in September or early October and lay two eggs in mid-October to early November. Chicks hatch in mid-November to early December, and fledge in mid-January to mid-February (22, 27).

To measure annual survival, we banded a total of 45,359 fledglings and 8,600 adults at the Punta Tombo colony from 1982 to 2019. We used stainless-steel bands marked with a unique identification number, which we fit to the left flipper of penguins, or we used small metal tags attached to the foot webbing (44). We note that there is little to no evidence that stainless-steel bands significantly affected the foraging, reproduction, or survival of penguins in our study (44, 45), so bands are unlikely to affect the relationship between climate variables and demographic parameters. We conducted daily searches for banded individuals throughout the breeding season (37). To measure reproduction, we followed individual nests at the same colony from 1983 to 2019 in an area of ~7,200 km<sup>2</sup> from mid-September to late February, and marked and monitored chicks. We counted chicks as fledged if they were last seen after January 10, weighed ≥ 1,800 g, and had not lost weight during the last few measurements (46). We calculated reproductive success in each year as the number of fledged chicks divided by the number of nests with eggs (maximum eggs per nest = 2).

To measure breeding population density, we established a permanent grid with lines running 1.5 km north to south and 2 km east to west, and counted nests, penguins, eggs, and chicks within a 100-m<sup>2</sup> circle every 100 m along the grid from 1987 to 2019 (33). We multiplied the estimated breeding pair density in each year by 3.524 km<sup>2</sup>, which is the estimated occupied penguin habitat at Punta Tombo (33). Prior to 2016, many surveys were conducted in early October, when unpaired, reproductively unsuccessful males could have been incorrectly counted as part of a breeding pair. In 1993 (33), 2017, and 2019, we conducted replicate surveys in both early and late October to determine the difference in density estimates during early October when extraneous males could be counted, and late October, when breeding pairs are established in the colony and most pairs have eggs. The average difference between replicate surveys in early vs. late October was 1.36, which we used to adjust breeding pair density estimates prior to 2016. Afterward, population surveys were conducted in late October.

**IPM.** We built a pre-breeding, three-stage, female-only IPM with immigration for the penguin population following Kery & Schaub (47). Males were excluded

from this model as the sex ratio at the colony at Punta Tombo is increasingly male-biased and males are not a limiting factor for reproduction or population growth (37, 48). Although Magellanic penguins do not breed until age 4+, and sometimes skip breeding due to poor body condition (33), we assumed a simple stage structure due to evidence that survival among adults of all ages is very similar (37). Additionally, models with additional stage structure (e.g., pre-breeding and non-breeding stages) exhibited similar dynamics to simpler models. Stages included juveniles ( $N_{juv}$ ), adults ( $N_{ad}$ ), and immigrants ( $N_{imm}$ ). We modeled the number of individuals in these stages with binomial and Poisson distributions to account for demographic stochasticity:

$$N_{juv,t+1} \sim \text{Poisson}(0.5 * f_t * \phi_{juv,t} * N_{ad,t}),$$

$$N_{ad,t+1} \sim \text{Binomial}(\phi_{ad,t}, N_{juv,t} + N_{ad,t} + N_{imm,t}),$$

$$N_{imm,t+1} \sim \text{Poisson}((N_{juv,t} + N_{ad,t} + N_{imm,t}) * \omega_t),$$

where  $f_t$  is the fecundity (number of fledged chicks per nest with eggs),  $\phi_{juv,t}$  is the apparent juvenile (from fledging to first year) survival probability,  $\phi_{ad,t}$  is the apparent adult survival probability of individuals over one year of age, and  $\omega_t$  is the immigration rate, which is a latent (unobserved) variable expressed as a per capita rate relative to the total population size the year before.  $f_t$  is multiplied by 0.5 to reflect an even sex ratio at hatching (22). We assumed immigration and emigration (incorporated within the apparent survival rates) due to evidence of metapopulation dynamics among Magellanic penguin colonies (49, 50). However, we believe that the emigration rate is very small relative to survival. Only 226 banded penguins at our study site have been found alive at other surveyed colonies, equating to a rough estimate of 0.04% emigration.

The observation equation links the observed population count ( $y$ ) with the true population size ( $N$ ), assuming Normal observation error:

$$y_t \sim \text{Truncated Normal}((N_{juv,t} + N_{ad,t} + N_{imm,t}), \tau_y),$$

where  $\tau_y$  is the observation error, and the distribution is truncated to prevent population values falling below zero. The likelihood for the population count data is denoted as  $L_{CO}(y | \phi_{juv}, \phi_{ad}, \omega, N, \tau_y, f)$ .

For capture-recapture data, we used the standard Cormack-Jolly-Seber model with a multinomial likelihood that allows for estimating adult and juvenile survival and recapture ( $p$ ) probabilities. We did not conduct surveys or recapture during 2011, so we manually set the recapture probability,  $p$ , for this year to be approximately zero. The likelihood for the capture-recapture data is denoted as  $L_{CR}(m | \phi_{juv}, \phi_{ad}, p)$ , where  $m$  is the capture-recapture data summarized in the  $m$ -array format. We modeled fecundity data (i.e., the number of fledglings produced) using a Poisson distribution. The likelihood for the fecundity data is denoted as  $L_{FP}(J, R | f)$ , where  $R$  is the total number of surveyed nests and  $J$  is the total number of fledglings. We assumed independence among our population, capture-recapture, and fecundity datasets. Therefore, we constructed the joint likelihood as a product of the likelihoods for the datasets ( $y, m, J, R$ ):

$$L_{IPM}(y, m, J, R | N, \phi_{juv}, \phi_{ad}, p, \omega, \tau_y, f).$$

**Climatic and Oceanographic Variables.** At Punta Tombo, we recorded daily precipitation and air temperature data. We recorded daily precipitation ( $\pm 0.1$  mm) using a manually emptied plastic rain gauge and daily maximum air temperature using a Sper Scientific min/max thermometer. We calculated the RAIN covariate as the total precipitation between October 15 and December 15, as this was found to have a strong impact on reproductive success (30, 31). We calculated the TEMP covariate as the % of days per breeding season when maximum air temperature was greater than 25 °C, which drove heat deaths (32). Because the locally collected covariates were not measured throughout the year, we also estimated the monthly rate of precipitation and air temperature change at Punta Tombo using TerraClimate, a global ~4 km resolution dataset of climate and climatic water balance (51).

We extracted SSTA for the foraging area during breeding season using the Coral Reef Temperature Anomaly Database (52), which contains global ~4 km resolution SSTA data on a weekly time scale. We extracted these data for an area

bounded from 65.6 to 61° W, 45 to 40° S, which takes up >95% of maximum foraging distances for breeding penguins and is important habitat for adults feeding chicks (23, 24, 40). We calculated the SSTA covariate by spatially and temporally averaging the weekly SSTA data from October–February to align with the breeding season.

We also extracted SSTAs during migration for an area from 40 to 23° S and from the coast to the 1,000 m isobath. We excluded pixels representing water deeper than 1,000 m because Magellanic penguins generally migrate over the continental shelf (21, 26). We then conducted a principal component analysis of the SSTA data to quantify the spatial and temporal patterns in the sea surface temperature and extend the index developed by Rebstock et al. (26). The principal pattern of the PCA shows the extent and strength of the Rio de la Plata plume, with positive values indicating a weak, contracted river plume, and negative values indicating a strong, extended river plume. The first three principal components represented 30.4, 14.0, and 6.3% of the variance in the SSTAs. Following Rebstock et al. (26), we only used the first principal component, and temporally averaged the monthly plume index from May–August to align with the migration season.

**Assessing the Effect of Climatic and Oceanographic Variables.** We fitted our IPM using log-linear and logit-linear regressions to estimate the effect of climatic (rain, air temperature) and oceanographic (SSTA and plume index) variables on demographic parameters (fecundity, juvenile and adult survival). We modeled fecundity as a function of rain, air temperature, SSTA, and plume index with a log-link:

$$\log(f_t) = \mu_f + \alpha_1 * \text{RAIN}_t + \alpha_2 * \text{TEMP}_t + \alpha_3 * \text{SSTA}_t + \alpha_4 * \text{PLUME}_t,$$

where  $\mu_f$  is the intercept,  $\alpha_1$  is the slope for the climatic covariate RAIN,  $\alpha_2$  is the slope for the climatic covariate TEMP,  $\alpha_3$  is the slope for the oceanographic covariate SSTA, and  $\alpha_4$  is the slope for the oceanographic covariate PLUME, prior to the breeding season.

We modeled juvenile and adult survival as a function of climatic and oceanographic variables with a logit-link:

$$\text{logit}(\phi_{juv,t}) = \mu_{juv} + \alpha_5 * \text{SSTA}_t + \alpha_6 * \text{PLUME}_{-1,t},$$

$$\text{logit}(\phi_{ad,t}) = \mu_{ad} + \alpha_7 * \text{TEMP}_t + \alpha_8 * \text{SSTA}_t + \alpha_9 * \text{PLUME}_{-1,t},$$

where  $\mu_{juv}$  and  $\mu_{ad}$  are the intercepts, and  $\text{PLUME}_{-1,t}$  is the oceanographic covariate PLUME, but lagged forward to account for the fact that the Rio de la Plata Plume will affect juvenile and adult survival during the migratory stage of their life cycle. SSTA<sub>t</sub> was included to account for indirect effects on fledgling body condition and effects during the beginning stages of migration, where fledglings are near the breeding–foraging grounds of colony. Each covariate was standardized to have zero mean and unit variance to compare the effect sizes (slopes) of  $\alpha$  to each other. We computed the mean and 95% CRI for the regression coefficients  $\alpha$ .

**Model Implementation.** We modeled the intercepts of the demographic parameters ( $\mu_f$ ,  $\mu_{juv}$ , and  $\mu_{ad}$ ) and slopes of covariates ( $\alpha$ ) using vaguely informative prior distributions of  $N(0, 0.34)$ . Observation error ( $\tau_y$ ) was given a vague prior of  $U(0, 4 * 10^{-10})$ . Bayesian posterior distributions were approximated via Markov chain Monte Carlo (MCMC) algorithms. Following standard practice (47), three independent MCMC chains of 30,000 iterations were used with a burn-in period of 10,000. Final inferences were derived from a sample of  $3 * 20,000$  iterations that resulted from merging the three chains. Gelman–Rubin convergence diagnostics were below 1.1 for each parameter and chains were well mixed (53). We assessed goodness-of-fit graphically using posterior predictive checks, and calculated Bayesian p values, all of which were near 0.5, indicating that the model fit the data well (47). We performed the analyses using Nimble (54, 55) and program R (56).

**Sensitivity Analyses and Population Projection Scenarios.** We conducted sensitivity analyses to quantify the effects of the presses and pulses of climatic and oceanographic variables on the population persistence of Magellanic penguins, as measured by the geometric mean lambda, which is the product of population growth rates, and more useful for characterizing fluctuating populations than using the arithmetic mean lambda. We ran 1,000 simulations for 100 y each of penguin population trajectories using Nimble and the parameter values we estimated previously. In these simulations, we generated climatic and oceanographic



data using normal and gamma distributions which were fit to the prior observed climate data. Normal distributions best fit the air temperature, SSTA, and plume index distributions, whereas gamma distributions best fit the precipitation data as it was positively skewed (Fig. 2C). We did not find any evidence of autocorrelation in the climatic and oceanographic variables based on calculating correlation coefficients at multiple time lags, so we did not add autocorrelation to the randomly drawn values. In each simulation, we randomly drew 100 values from these distributions to represent the yearly climate variables that penguins experienced. We ran sensitivity analyses by varying the mean and standard deviation (for normal distributions) and shape and rate (for gamma distributions) according to  $\pm$  the past rate of change in the past climate variables. We did this to generate climate variables within the realistic rate of climate change in the sensitivity analysis. For each simulation in the sensitivity analysis, we also recorded the frequency and magnitude of extreme events in the 100 randomly drawn climate values by calculating the frequency of values greater than the 95% quantile and the maximum value, respectively (8). We did not calculate the frequency of values less than the 5% quantile and the minimum value because all climatic and oceanographic variables were changing towards positive extreme events.

We then calculated partial correlation coefficients (PCC) via linear regression to determine the relationship between the climate parameters in our sensitivity analysis (mean, SD, frequency of extreme events, magnitude of extreme events) and geometric lambda. The sign of the PCC determines the qualitative relationship between the parameters and population growth, and the magnitude indicates the importance of climate variables in contributing to population growth (57). To explore the effects of including extreme events, we calculated the PCC with and without the frequency and magnitude of extreme events variables.

We then projected the population forward under four climate scenarios by running 100 simulations for 50 y. Scenario 1 (no change in climate variables from present) was run using generated climatic and oceanographic data from normal and gamma distributions based on the prior observed climate data. Scenario 2 (projected climate change based on current rate of change) was run using generated data from distributions based on the observed rate of increase of the climate data ( $\pm$ SD) between 1958 and 1988 and 1989 and 2019 for precipitation

and air temperature data, and between 1982 and 2000 and 2001 and 2020 for oceanographic data. Scenario 3 (change in the mean strength, or press, of the plume) was run by using generated data for the mean plume from Scenario 2, and then generated data for all other variables from Scenario 1. Scenario 4 (change in the frequency, or pulse, of air temperature extremes) was run by using generated data for frequency of air temperature extremes based on the rate of increase in air temperature extremes, and then generated data for all other variables from Scenario 1. For all scenarios, we calculated the mean and 95% CI across the 100 simulations and calculated geometric lambda.

**Data, Materials, and Software Availability.** The code and data needed to reproduce this analysis can be found on Github (<https://doi.org/10.5281/zenodo.7349059>) (58).

**ACKNOWLEDGMENTS.** We thank the following funders for financial support for field work at Punta Tombo: Wildlife Conservation Society, ExxonMobil Foundation, the Pew Fellows Program in Marine Conservation, the Disney Worldwide Conservation Fund, the Chase, Cunningham, CGMK, Offield, Peach, Thorne, Tortuga, and Kellogg Foundations, the Wadsworth Endowed Chair in Conservation Science, Friends of Penguins, and many private donors. We thank the Wildlife Conservation Society and the Global Penguin Society for logistical and permit support, particularly Pablo García-Borboroglu, William Conway, Graham Harris, and Pat Harris. We are grateful to the La Regina family for allowing us to work on their land. All research was conducted under approval of the University of Washington IACUC (Protocol #2213-02) and with permission of the Province of Chubut, Offices of Turismo and Flora and Fauna. We also thank the many field volunteers and students who collected the data used in this study. T.J.C.-W. was supported by the UW Data Science Postdoctoral Fellowship at the eScience Institute. P. Wellington provided database help, and K. Rafiq, A. Nisi, E. Johansson, L. West, and K. Holt provided valuable feedback on this research. The editors, N. Yoccoz, and one anonymous reviewer greatly improved this article.

Author affiliations: <sup>a</sup>Center for Ecosystem Sentinels, Department of Biology, University of Washington, Seattle, WA 98195

1. C. B. Field, V. Barros, T. F. Stocker, Q. Dahe, Eds., *Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation: Special Report of the Intergovernmental Panel on Climate Change* (Cambridge University Press, 2012), 10.1017/CBO9781139177245, November 12, 2021.
2. G.-R. Walther *et al.*, Ecological responses to recent climate change. *Nature* **416**, 289–307 (2002).
3. C. Parmesan, G. Yohe, A globally coherent fingerprint of climate change impacts across natural systems. *Nature* **421**, 37–42 (2003).
4. M. C. Urban, Accelerating extinction risk from climate change. *Science* **348**, 571–573 (2015).
5. C. Le Coeur, J. Storkey, S. Ramula, Population responses to observed climate variability across multiple organismal groups. *Oikos* **130**, 476–487 (2021).
6. L. D. Bailey, M. van de Pol, Tackling extremes: Challenges for ecological and evolutionary research on extreme climatic events. *J. Anim. Ecol.* **85**, 85–96 (2016).
7. G. Palmer *et al.*, Climate change, climatic variation and extreme biological responses. *Philos. Trans. R. Soc. B Biol. Sci.* **372**, 20160144 (2017).
8. D. Pardo, S. Jenouvrier, H. Weimerskirch, C. Barbraud, Effect of extreme sea surface temperature events on the demography of an age-structured albatross population. *Philos. Trans. R. Soc. B Biol. Sci.* **372**, 20160143 (2017).
9. T. Reierson *et al.*, Inter-population synchrony in adult survival and effects of climate and extreme weather in non-breeding areas of Atlantic puffins. *Mar. Ecol. Prog. Ser.* **676**, 219–231 (2021).
10. B. B. Hansen *et al.*, More frequent extreme climate events stabilize reindeer population dynamics. *Nat. Commun.* **10**, 1616 (2019).
11. R. M. B. Harris *et al.*, Biological responses to the press and pulse of climate trends and extreme events. *Nat. Clim. Change* **8**, 579–587 (2018).
12. E. A. Bender, T. J. Case, M. E. Gilpin, Perturbation experiments in community ecology: Theory and practice. *Ecology* **65**, 1–13 (1984).
13. Z. Ratajczak *et al.*, The interactive effects of press/pulse intensity and duration on regime shifts at multiple scales. *Ecol. Monogr.* **87**, 198–218 (2017).
14. K. E. Dybala, J. M. Eadie, T. Gardali, N. E. Seavy, M. P. Herzog, Projecting demographic responses to climate change: Adult and juvenile survival respond differently to direct and indirect effects of weather in a passerine population. *Glob. Change Biol.* **19**, 2688–2697 (2013).
15. S. Jenouvrier, Impacts of climate change on avian populations. *Glob. Change Biol.* **19**, 2036–2057 (2013).
16. M. van de Pol *et al.*, Effects of climate change and variability on population dynamics in a long-lived shorebird. *Ecology* **91**, 1192–1204 (2010).
17. D. S. Wilcove, M. Wikelski, Going, going, gone: Is animal migration disappearing. *PLoS Biol.* **6**, e188 (2008).
18. T. S. Sillett, R. T. Holmes, T. W. Sherry, Impacts of a global climate cycle on population dynamics of a migratory songbird. *Science* **288**, 2040–2042 (2000).
19. M. A. Tucker *et al.*, Moving in the Anthropocene: Global reductions in terrestrial mammalian movements. *Science* **359**, 466–469 (2018).
20. P. D. Boersma, D. L. Stokes, P. M. Yorio, "Reproductive variability and historical change of Magellanic penguins (*Spheniscus magellanicus*) at Punta Tombo, Argentina" in *Penguin Biology*, (Academic Press, 1990), pp. 13–41.
21. D. L. Stokes, P. D. Boersma, J. Lopez de Casenave, P. García-Borboroglu, Conservation of migratory Magellanic penguins requires marine zoning. *Biol. Conserv.* **170**, 151–161 (2014).
22. P. D. Boersma *et al.*, "Magellanic Penguin (*Spheniscus magellanicus*)" in *Penguins: Natural History and Conservation*, P. García-Borboroglu, P. D. Boersma, Eds. (University of Washington Press, 2013), pp. 233–264.
23. P. D. Boersma, G. A. Rebstock, Foraging distance affects reproductive success in Magellanic penguins. *Mar. Ecol. Prog. Ser.* **375**, 263–275 (2009).
24. P. D. Boersma, G. A. Rebstock, E. Frere, S. E. Moore, Following the fish: Penguins and productivity in the South Atlantic. *Ecol. Monogr.* **79**, 59–76 (2009).
25. E. Acha *et al.*, An overview of physical and ecological processes in the Rio de la Plata Estuary. *Cont. Shelf Res.* **28**, 1579–1588 (2008).
26. G. A. Rebstock, P. D. Boersma, Oceanographic conditions in wintering grounds affect arrival date and body condition in breeding female Magellanic penguins. *Mar. Ecol. Prog. Ser.* **601**, 253–267 (2018).
27. C. D. Cappelto, P. D. Boersma, Consequences of phenological shifts and a compressed breeding period in Magellanic penguins. *Ecology* **102**, e03443 (2021).
28. R. L. Fogt, G. J. Marshall, The Southern annular mode: Variability, trends, and climate impacts across the Southern Hemisphere. *WIREs Clim. Change* **11**, e652 (2020).
29. B. C. Franco *et al.*, Climate change impacts on the atmospheric circulation, ocean, and fisheries in the southwest South Atlantic Ocean: A review. *Clim. Change* **162**, 2359–2377 (2020).
30. D. Boersma, Penguins as marine sentinels. *BioScience* **58**, 597–607 (2008).
31. P. D. Boersma, G. A. Rebstock, Climate change increases reproductive failure in magellanic penguins. *PLoS One* **9**, e85602 (2014).
32. K. A. Holt, P. D. Boersma, Unprecedented heat mortality of Magellanic Penguins. *Ornithol. Appl.* **125**, duab052 (2022).
33. G. A. Rebstock, P. D. Boersma, P. García-Borboroglu, Changes in habitat use and nesting density in a declining seabird colony. *Popul. Ecol.* **58**, 105–119 (2016).
34. L. A. Vincent *et al.*, Observed trends in indices of daily temperature extremes in South America 1960–2000. *J. Clim.* **18**, 5011–5023 (2005).
35. M. R. Haylock *et al.*, Trends in total and extreme South American rainfall in 1960–2000 and links with sea surface temperature. *J. Clim.* **19**, 1490–1512 (2006).
36. M. J. Behrenfeld *et al.*, Climate-driven trends in contemporary ocean productivity. *Nature* **444**, 725–755 (2006).
37. N. Gownaris, P. Boersma, Sex-biased survival contributes to population decline in a long-lived seabird, the Magellanic Penguin. *Ecol. Appl.* **29**, 1–17 (2019).
38. L. E. Loe *et al.*, The neglected season: Warmer autumns counteract harsher winters and promote population growth in Arctic reindeer. *Glob. Change Biol.* **27**, 993–1002 (2021).



39. J.-M. Gaillard *et al.*, How does climate change influence demographic processes of widespread species? Lessons from the comparative analysis of contrasted populations of roe deer. *Ecol. Lett.* **16**, 48–57 (2013).
40. P. D. Boersma, G. A. Rebstock, P. García-Borboroglu, Marine protection is needed for Magellanic penguins in Argentina based on long-term data. *Biol. Conserv.* **182**, 197–204 (2015).
41. W. K. Oestreich, M. S. Chapman, L. B. Crowder, A comparative analysis of dynamic management in marine and terrestrial systems. *Front. Ecol. Environ.* **18**, 496–504 (2020).
42. J. A. Santora *et al.*, Habitat compression and ecosystem shifts as potential links between marine heatwave and record whale entanglements. *Nat. Commun.* **11**, 536 (2020).
43. C. Le Coeur, N. G. Yoccoz, R. Salguero-Gómez, Y. Vindenes, Life history adaptations to fluctuating environments: Combined effects of demographic buffering and lability. *Ecol. Lett.* **25**, 2107–2119 (2022).
44. P. D. Boersma, G. A. Rebstock, Effects of double bands on Magellanic Penguins. *J. Field Ornithol.* **81**, 195–205 (2010).
45. P. D. Boersma, G. A. Rebstock, Flipper bands do not affect foraging-trip duration of Magellanic Penguins. *J. Field Ornithol.* **80**, 408–418 (2009).
46. G. A. Rebstock, P. Boersma, Comparing reproductive success of a colonial seabird, the Magellanic Penguin, estimated by coarse-and fine-scale temporal sampling. *Condor Ornithol. Appl.* **119**, 225–238 (2017).
47. M. Kéry, M. Schaub, *Bayesian Population Analysis Using WinBUGS* (Elsevier Inc., 2012).
48. N. J. Gownaris, P. García Borboroglu, P. D. Boersma, Sex ratio is variable and increasingly male biased at two colonies of Magellanic Penguins. *Ecology* **101**, 1–11 (2020).
49. J. L. Bouzat, B. G. Walker, P. D. Boersma, Regional genetic structure in the Magellanic Penguin (*Spheniscus magellanicus*) suggests metapopulation dynamics. *The Auk* **126**, 326–334 (2009).
50. L. M. Pozzi, P. G. Borboroglu, P. D. Boersma, M. A. Pascual, Population regulation in Magellanic Penguins: What determines changes in colony size? *PLoS One* **10**, e0119002 (2015).
51. J. T. Abatzoglou, S. Z. Dobrowski, S. A. Parks, K. C. Hegewisch, TerraClimate, a high-resolution global dataset of monthly climate and climatic water balance from 1958–2015. *Sci. Data* **5**, 1–12 (2018).
52. S. Korak, *et al.*, The Coral Reef Temperature Anomaly Database (CoRTAD) Version 6 - Global, 4 km Sea Surface Temperature and Related Thermal Stress Metrics for 1982 to 2019 NOAA National Centers for Environmental Information. Dataset. <https://doi.org/10.25921/ffw7-cs39>. Accessed 1/4/22.
53. A. Gelman, J. Hill, *Data Analysis Using Regression and Multilevel/Hierarchical Models* (Cambridge University Press, 2012).
54. P. de Valpine *et al.*, Programming with models: Writing statistical algorithms for general model structures with NIMBLE. *J. Comp Graph Stat* **26**, 403–413 (2017).
55. P. de Valpine *et al.*, NIMBLE: MCMC, Particle Filtering, and Programmable Hierarchical Modeling (Version 0.12.1, R package, <https://doi.org/10.5281/zenodo.1211190>), <https://cran.r-project.org/package=nimble>.
56. R Core Team, R: A language and environment for statistical computing (R version 4.1.1, R Foundation for Statistical Computing, Vienna, Austria, 2021), <https://www.R-project.org/>.
57. S. M. Blower, H. Dowlatabadi, Sensitivity and uncertainty analysis of complex models of disease transmission: An HIV model, as an example. *Int. Stat. Rev.* **62**, 229 (1994).
58. T. J. Clark-Wolf, teejclark/Press\_Pulse: teejclark/Press\_Pulse: archive of data and code (v1.0). Zenodo. <https://doi.org/10.5281/zenodo.7349059> (2022) Deposited on 22 November 2022.