

Strong self-regulation and widespread facilitative interactions in phytoplankton communities

Coralie Picoche^{1,2}, Frédéric Barraquand^{1,2*}

1 – Integrative and Theoretical Ecology, LabEx COTE, University of Bordeaux, Pessac, France

2 – Institute of Mathematics of Bordeaux, CNRS, University of Bordeaux, Talence, France

Correspondence:

Frédéric Barraquand, CNRS, Institute of Mathematics of Bordeaux, 351 Cours de la Libération,
33405 Talence, France. * frederic.barraquand@u-bordeaux.fr

Abstract

1. The persistence of phytoplanktonic diversity in spite of competition for basic resources has long been a source of wonder and inspiration to ecologists. To sort out, among the many coexistence mechanisms suggested by theory and experiments, which ones actually maintain diversity in natural ecosystems, long-term field studies are paramount.
2. We analysed a large dataset of phytoplankton abundance time series using dynamic, multivariate autoregressive models. Phytoplankton was counted and identified down to the genus level, every two weeks over twenty years, at ten sites along the French coastline. Multivariate autoregressive models allowed to estimate biotic interaction networks, while also accounting for abiotic variables that may drive part of the phytoplankton fluctuations. We then analysed the ratio of intra- to inter-taxa interactions (a measure of niche differentiation), the frequency of negative vs positive interactions, and how stability metrics (both at the network and genus level) relate to network complexity and genus self-regulation or abundance.
3. We showed that a strong self-regulation, with competition strength within a taxon (genus) an order of magnitude higher than between taxa, was present in all phytoplanktonic interaction networks. This much stronger intragenus competition suggests that niche differentiation - rather than neutrality - is commonplace in phytoplankton. Furthermore, interaction networks were dominated by positive net effects between phytoplanktonic taxa (on average, more than 50% of interactions were positive). While network stability (*sensu* resilience) was unrelated to complexity measures, we unveiled links between self-regulation, intergenera interaction strengths and abundance. The less common taxa tend to be more strongly self-regulated and can therefore maintain in spite of competition with more abundant ones.
4. *Synthesis:* We demonstrate that strong niche differentiation, widespread facilitation between phytoplanktonic taxa and stabilizing covariances between interaction strengths should be common features of coexisting phytoplankton communities in the field. These are structural properties that we can expect to emerge from plausible mechanistic models of phytoplankton

communities. We discuss mechanisms, such as predation or restricted microscale movement, that are consistent with these findings, which paves the way for further research.

Keywords: phytoplankton; coexistence; facilitation; mutualism; niche theory; time series; networks

Second language abstract

1. La persistance de la biodiversité du phytoplancton malgré la compétition pour les ressources est depuis longtemps une source de questionnement pour les écologues. Afin d'identifier, parmi les nombreux mécanismes de coexistence suggérés par la théorie et les expérimentations, ceux qui maintiennent la diversité dans les écosystèmes naturels, des études à long-terme sont essentielles.
2. Nous avons analysé des séries temporelles longues de comptages de phytoplancton. Le phytoplancton a été identifié au niveau du genre, toutes les deux semaines pendant vingt ans sur dix sites, le long du littoral français. Des modèles autorégressifs multivariés ont permis d'inférer des réseaux d'interactions biotiques, tout en prenant en compte les variables abiotiques susceptibles d'affecter les fluctuations du phytoplancton. Nous avons ensuite calculé le ratio entre les interactions intra et inter-taxa (une mesure de la différenciation des niches), la fréquence des interactions positives et négatives ainsi que la façon dont les métriques de stabilité (au niveau du réseau et du genre) sont liées à la complexité du réseau et à l'autorégulation ou l'abondance de chaque genre.
3. Nous avons montré qu'une autorégulation forte, c'est-à-dire une force de compétition au sein d'un même taxon (ici, le genre) un ordre de grandeur supérieure aux forces d'interaction entre taxa, était omniprésente dans les réseaux d'interactions phytoplanctoniques. Cette compétition intragénre beaucoup plus forte suggère que la différenciation des niches, plutôt que la neutralité, est fréquente pour le phytoplancton. En outre, les réseaux d'interactions étaient dominés par des effets nets positifs d'un taxon sur un autre (en moyenne, plus de la moitié des interactions étaient positives). Bien que la stabilité (au sens de la résilience des réseaux) ne soit pas liée aux mesures de complexité, nous avons révélé des relations entre l'autorégulation, la force des interactions entre genres et l'abondance. Les genres les moins communs tendent à être plus fortement autorégulés et peuvent ainsi se maintenir malgré la compétition avec les genres les plus abondants.

4. *Synthèse* : Nous démontrons qu'une différenciation des niches, une facilitation fréquente parmi les taxa de phytoplancton, et des covariances stabilisant les forces d'interaction, devraient être des facteurs relativement communs dans les communautés phytoplanctoniques qui coexistent in situ. Ces propriétés structurelles peuvent donc être attendues de modèles mécanistes multi-taxa pour le phytoplancton. Nous discutons des mécanismes, tels que la prédation ou des mouvements limités à l'échelle microscopique, qui pourraient expliquer ces résultats, et ainsi ouvrir de nouvelles pistes de recherche.

1 Introduction

2 How species or close genera can coexist together in spite of competition is one of the main puzzles of
3 community ecology, especially for primary producers that seemingly share the same basic resources
4 ([Hutchinson, 1961](#)). Many theoretical studies of competition models have shown that competitive
5 exclusion is likely in those circumstances, unless mechanisms involving spatial or temporal variation
6 are at play ([Armstrong & McGehee, 1976, 1980](#); [Chesson & Huntly, 1997](#); [Huisman & Weissing, 2001](#);
7 [Li & Chesson, 2016](#); [Chesson, 2018](#)). Neutral theory models, assuming that all individuals have equal
8 birth and death rates and exert equal competitive pressure on conspecifics and heterospecifics alike,
9 produce instead a non-equilibrium coexistence maintained by dispersal from a regional pool. They
10 have been proposed as a solution to the puzzle presented by highly diverse communities ([Hubbell,](#)
11 [2001](#); [Rosindell *et al.*, 2011](#)).

12 However, the evidence gathered from terrestrial plant communities starts to suggest that, in fact,
13 niche rather than neutral processes may be paramount to explain coexistence, with intraspecific
14 competition dwarfing interspecific competition in most cases ([Adler *et al.*, 2010, 2018b](#)); see also
15 [Volkov *et al.* \(2009\)](#). Whether these conclusions drawn mostly from studies of terrestrial plants
16 apply to other ecosystems and taxa is currently little known (but see [Mutshinda *et al.* 2009](#)).

17 Moreover, competition may not be the rule: the meta-analysis by [Adler *et al.* \(2018b\)](#) reported a
18 large number of facilitative interactions (30%) and several reviews ([Brooker *et al.*, 2008](#); [McIntire](#)
19 [& Fajardo, 2014](#); [Kinlock, 2019](#)) have highlighted that facilitation may be much more widespread
20 than ecologists usually tend to think. Although some theoretical studies suggest that facilitative
21 interactions can be destabilizing (*sensu* resilience) and therefore undermine coexistence in Lotka-
22 Volterra models ([Coyte *et al.*, 2015](#)), multiple other modelling ([Gross, 2008](#); [Qian & Akçay, 2020](#))
23 and empirical ([Brooker *et al.*, 2008](#); [Cavieres & Badano, 2009](#)) studies have suggested that facilitative
24 interactions can to a large degree benefit coexistence, especially when multiple interaction types
25 are considered simultaneously ([Mougi & Kondoh, 2012](#); [García-Callejas *et al.*, 2018](#); [Qian & Akçay,](#)
26 [2020](#)).

27 Here, we analyse a spatially replicated, long-term community-level dataset, consisting of ten

multivariate time series of phytoplankton abundance along the French coastline. We do so using multivariate autoregressive (MAR) models, that allow to estimate interactions between genera. Although many ecological studies focus on interactions between species, competition has been shown experimentally to occur between different genera of phytoplankton (Titman, 1976; Descamps-Julien & Gonzalez, 2005). The genus level is also a rather fine taxonomic scale for phytoplankton interaction studies, as most studies are restricted to interactions between different classes or even phyla (Ives *et al.*, 2003; Hampton *et al.*, 2008; Griffiths *et al.*, 2015). Studying interactions between different genera of phytoplankton therefore both makes empirical sense in light of competition experiments and allows to estimate better-resolved networks. We focus here on genera that belong mostly to diatoms and dinoflagellates. To put our results into a more general context, we then compare our interaction strength estimates to previously published interaction networks produced under the same statistical framework, both in plankton and other empirical systems.

Material and methods

Sampling methods

All phytoplankton samples were collected by Ifremer coastal laboratories as part of the National Phytoplankton and Phycotoxin Monitoring Network (REPHY, 2017). Since 1987, this monitoring program has required the sampling of 26 sites along the French coastline every 2 weeks within 2 hours of high tide to document both biotic (phytoplankton counts) and abiotic (water temperature, salinity) variables. We focused on sites which had the longest time series. We also excluded time series which had missing data for over 6 months or an average delay between sampling dates above 20 days. This reduced the number of study sites to 10 sites nested within 4 regions (Brittany, Oléron, Arcachon and the Mediterranean Sea; Fig. S1 and Table S1 in the Supporting Information).

Abiotic variables (temperature, salinity) were measured directly from the boat during the sampling process while water samples for biotic analyses were fixed with a Lugol's solution and examined later. Phytoplankton cells above 20 μm were identified at the lowest possible taxonomic level and counted with the Utermöhl method using an optical microscope (Utermöhl, 1958). Throughout the years

and sites, more than 600 taxa were identified at different taxonomic levels. We aggregated them at the genus (or group of genera when not possible) level based on previous work (Table S2; [Hernández Fariñas *et al.* 2015](#); [Barraquand *et al.* 2018](#)), except for cryptophytes and euglenophytes in Arcachon, which could not be identified below the family level. Although the taxonomic resolution used here may seem coarse in comparison to land plants, it is in fact more refined than 86% of the MAR(1) studies of phytoplankton listed in Table S4.

For each region, the MAR(1) analysis focused on the most abundant and most frequently observed genera to avoid most of the gaps in the time series. When gaps did not exceed a month, missing values were linearly interpolated; remaining missing values were replaced by a random number between 0 and half of the lowest observed abundance ([Hampton *et al.*, 2006](#)). Time series are plotted in Fig. S2. We tested extensively this and other methods to deal with missing data in a previous publication on a subset of this dataset ([Barraquand *et al.*, 2018](#)). All time series were scaled and centered before MAR analyses.

MAR(1) models

Multivariate autoregressive (MAR) models are used to determine the interspecific interactions and abiotic effects shaping a community's dynamics ([Ives *et al.*, 2003](#)). MAR(1) models are based on a stochastic, discrete-time Gompertz equation which relates the log-abundance of each of the S taxa at time $t + 1$ to log-abundances of the whole community at time t , with possible interactions between taxa, and effects of V abiotic variables at time $t + 1$. These assumptions are encapsulated in eq. 1:

$$\mathbf{n}_{t+1} = \mathbf{B}\mathbf{n}_t + \mathbf{C}\mathbf{u}_{t+1} + \mathbf{e}_t, \mathbf{e}_t \sim \mathcal{N}_S(0, \mathbf{Q}) \quad (1)$$

where \mathbf{n}_t is the $1 \times S$ vector of log-abundance of phytoplankton taxa, \mathbf{B} is the $S \times S$ community (interaction) matrix, \mathbf{C} is the $S \times V$ environment matrix describing the effects of V variables (stacked in vector \mathbf{u}_{t+1}) on growth rates, with $V = 2$ in our case (temperature and salinity). The noise \mathbf{e}_t is a $1 \times S$ noise vector, following a multivariate normal distribution with a variance-covariance matrix \mathbf{Q} . \mathbf{Q} is diagonal and we have previously showed that this parsimonious choice did not affect

78 qualitatively the results (Barraquand *et al.*, 2018). We used the MARSS package v3.9 (Holmes *et al.*,
79 2012, 2014), in R v3.3.2 (R Core Team, 2016), to estimate parameters with a maximum likelihood
80 procedure.

81 Our previous analysis of the Arcachon region, for which more covariables were available (Bar-
82 raquand *et al.*, 2018), revealed that hydrodynamics and hydrology had more influence on phyto-
83 plankton dynamics than nutrients on the two-week timescale. Because temperature and salinity, in
84 addition to their direct effects, sum up seasonal changes in light and hydrology (salinity is inversely
85 related to freshwater inflow), they represent the two key drivers needed to account for abiotic
86 influences (Scheef *et al.*, 2013). They are therefore used to summarize the abiotic environment in
87 the remainder of the article.

88 The analysis of real data in Barraquand *et al.* (2018) was complemented by that of simulated
89 data mimicking the study design, which confirmed the ability of MAR(1) models to infer biotic
90 interactions and abiotic forcings. Fitting a more sophisticated model (threshold autoregressive model)
91 did not reveal extra non-linearities or a storage effect in the Arcachon subset of the data (Barraquand
92 *et al.*, 2018). Other aspects of the MAR(1) modelling are likewise quite robust: using two abiotic
93 variables (temperature and salinity) in this study rather than the full set used in Barraquand *et al.*
94 (2018) led to almost identical covariate effects and interaction estimates for the Arcachon study
95 sites. Even if some departures from the true data-generating model may not always be detectable
96 through MAR(1) diagnostics (e.g., residuals), the analysis of nonlinear simulations has showed that
97 MAR(1) models are in general robust to nonlinearities if the inference focuses on interaction sign
98 and order of magnitude of model coefficients (Certain *et al.*, 2018), which is how these models are
99 used here. For ease of interpretation of MAR(1) interaction coefficients, we also highlight how
100 intra- and inter-taxa interaction strengths in a MAR(1) model map to their counterparts in a
101 multispecies Beverton-Holt model, i.e., a discrete-time Lotka-Volterra model (Cushing *et al.*, 2004),
102 in the Supporting Information.

103 In this study, the number of phytoplankton taxa (S) and the community composition vary
104 slightly between regions but sites share on average 67% of their taxa. In order to have comparable
105 models, we also keep the same 2 covariates, i.e., water temperature and salinity, that were measured

at all study sites. Therefore, the dimension of the dynamical system depends on the (square of the) number of phytoplankton taxa we study, which ranges between 7 (Mediterranean Sea) and 14 (Brittany). The smallest system still requires 63 parameters to be estimated (49 for the 7×7 interaction matrices and 14 for the 7×2 environment matrices) if we consider all possible interactions between taxa. To reduce this dimensionality and remove unnecessary parameters, we built different ‘interaction scenarios’ based on known phylogenetic information (as suggested in [Violle *et al.*, 2011](#); [Narwani *et al.*, 2017](#)). The null interaction scenario assumed no interaction between genera (diagonal interaction matrix) and was compared to four other interaction scenarios. The first interaction scenario assumed that interactions could only occur between phylogenetically close organisms, i.e., within a class (groups were then diatoms, dinoflagellates, and other phytoplanktonic organisms) while the second interaction scenario further differentiated pennate and centric diatoms. The third interaction scenario considered the reverse hypothesis, that only unrelated organisms could interact (i.e., a diatom could only interact with a dinoflagellate or a cryptophyte, but not with another diatom), and the last interaction scenario did not constrain the interactions at all (full interaction matrix). We selected the best scenario by comparing BIC (Fig. S3), which proved to be satisfactory in our previous analyses of both real data and similar simulated datasets ([Barraquand *et al.*, 2018](#), Appendix 2). The second interaction scenario, hereafter called the pennate-centric scenario, had the lowest BIC for all sites (Fig. S3). This parsimonious scenario was therefore chosen as the basis for further investigations of network structure.

Analysis of interaction strengths

The interaction matrix obtained from MAR(1) analyses can be used to determine the stability of a discrete-time dynamical system ([Ives *et al.*, 1999, 2003](#)). To investigate stability-complexity relationships, we compared the maximum modulus of the eigenvalues of the pennate/centric matrices for each site to network descriptors. The maximum modulus is analogous to the real part of the leading eigenvalue for continuous time models, and measures resilience while still accounting for some variability properties ([Ives *et al.*, 1999](#)). However, because most theory on stability-complexity has been developed in continuous time (e.g., [Allesina & Tang, 2015](#)), we numerically checked that the

133 maximum modulus of the eigenvalues in a discrete-time interaction matrix and its continuous-time
 134 model counterpart yield similar information in the Supporting Information. We then compared
 135 this resilience measure to complexity metrics, such as the interaction strength distribution (sign,
 136 mean and variance) and weighted connectance (Bersier *et al.*, 2002). Weighted connectance is a
 137 measure of the proportion of realized links compared to all possible links, taking into account the
 138 shape of the flux distribution. This metric is adapted to weighted interaction matrices but cannot
 139 accommodate for both positive and negative coefficients: we therefore chose to focus on interaction
 140 strength only (absolute values of the coefficients), irrespective of interaction sign. In contrast, mean
 141 and variance of the off-diagonal coefficients, which can affect the stability of a community (Allesina
 142 & Tang, 2015), are computed on raw values of the coefficients. Interaction coefficient variance is
 143 multiplied by the number of taxa, according to theory (Allesina & Tang, 2015).

144 In addition to these network-level metrics, we also computed the average vulnerability (average
 145 effect of other taxa on a focal taxon, eq. S5) and impact (average effect of a focal taxon on other
 146 taxa, eq. S6) on both raw and absolute values of the interaction coefficients. Vulnerability and
 147 impact can be related to in-strength and out-strength in the meta-analysis of Kinlock (2019). We
 148 then compared these to the regulation a focal species exerted on itself. Vulnerability computed
 149 on raw coefficient values indicate the average effect that can be expected on the growth rate of
 150 a taxon from the rest of the community (i.e., is the effect of others mostly positive or negative?),
 151 while vulnerability computed on absolute coefficient values characterise the strength of all types of
 152 interactions on a taxon (i.e., is a taxon strongly affected by the others?).

153 Finally, we compared the observed ratio between mean self-regulation (intrataxon interaction
 154 strength) and mean intertaxa interaction strength to other published studies based on a MAR(1)
 155 model. A list of references is given in Table S4. Authors usually reported only coefficients that
 156 were significant with a 5% significance level, thus ignoring potentially many weak effects, which we
 157 had to set to 0. There are therefore two ways of computing the mean intertaxa interactions, i.e.,
 158 taking the mean value of all coefficients outside of the matrix diagonal, including zeroes (which
 159 decreases the estimated mean intertaxa interaction strength, Fig. 4), or taking the mean value of
 160 statistically significant intertaxa coefficients only (which increases the estimated mean intertaxa

161 interaction strength, Fig. S9). We considered both; a detailed description of these different ways to
162 compare intra- and inter-taxa interactions can be found in the Supporting Information.

163 Results

164 Interaction estimates

165 Using MAR(1) autoregressive models, we produced interaction matrices (Ives *et al.*, 2003; Hampton
166 *et al.*, 2013) – i.e., Jacobian community matrices on the logarithmic abundance scale (Ives *et al.*,
167 2003). Best-fitting models corresponded to a phylogenetically-structured interaction scenario, where
168 interactions only occurred between closely related genera (Fig. S3). This led to sparse, modular
169 matrices that have two main features. First, we observed a strong self-regulation for all sites (Fig. 1,
170 diagonal elements of all matrices), a feature that we had previously highlighted in a more detailed
171 analysis on one of the considered study regions (Barraquand *et al.*, 2018). The ratio of mean
172 intragenus to intergenera interaction coefficients varied between 6 and 10, not counting coefficients
173 set to 0 before the estimation process. When we included the zeroes in the interaction matrix in
174 the computation of the intra/inter mean interaction strength (see the Supporting Information for
175 details of that computation), the ratio rose to 21-43. Therefore, intragenus interactions were on
176 average one order of magnitude stronger than intergenera interactions.

177 Second, although the percentage of facilitative interactions varied among sites (between 40%
178 and 71% of interactions in the selected models), facilitation remained predominant in 9 sites out
179 of 10 (only Lazaret, in the Mediterranean Sea, has 60% negative interactions). Our observational
180 setup being nested, with sites within regions, we could examine whether locally positive interactions
181 remain positive in a regional context: the percentage of consistently positive interactions at the
182 regional level varied between 30% and 53%, higher than the percentage of similarly defined negative
183 interactions (between 15% and 40%), except for sites in the Mediterranean Sea.

184 We found that the percentage of true mutualism (+/+) was substantial: averaged over all sites,
185 32% of all interactions were (+/+) while only 12% of them were (-/-), see also Fig. S5. The sign
186 correspondence was not always maintained between regions: the only interaction that was non-zero

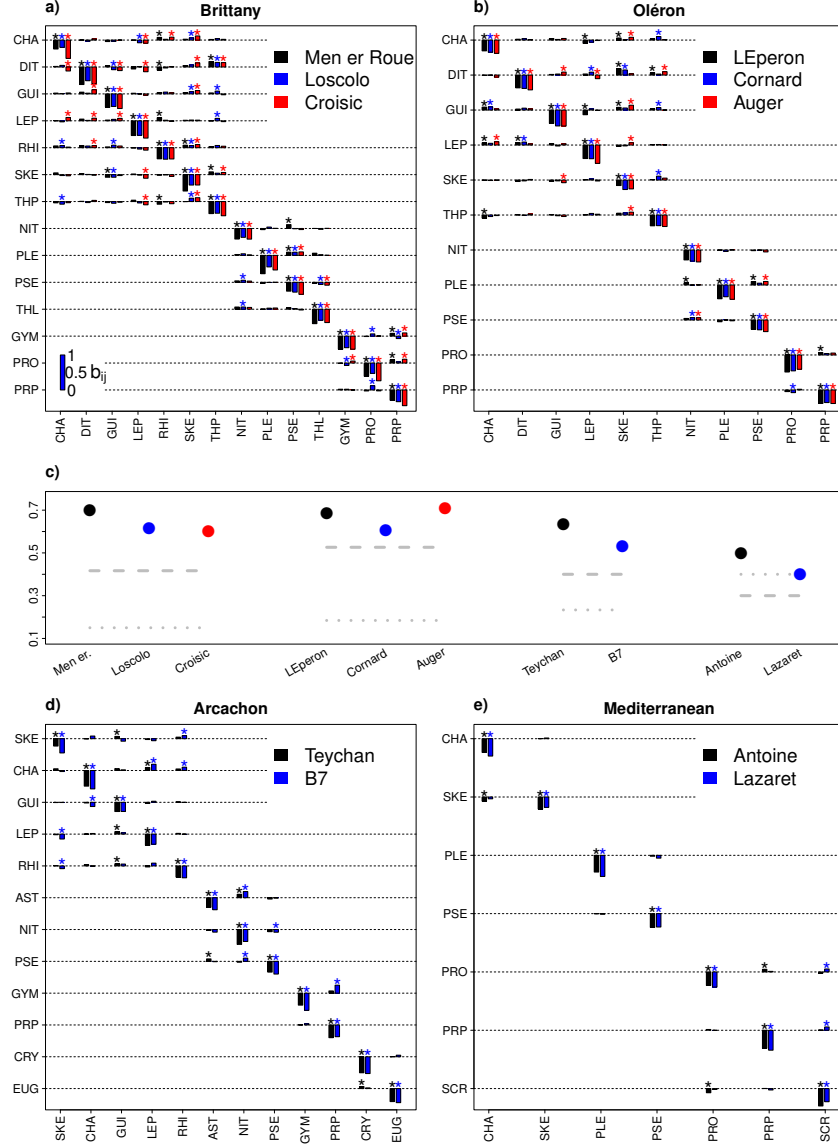


Figure 1: **Interaction matrices estimated at 10 sites along the French coastline.** Taxon j (in columns) has an effect on taxon i 's growth rate (in rows) proportional to the bar height, which corresponds to the $\mathbf{B} - \mathbf{I}$ matrix (community composition in Table S2, most parsimonious interaction scenario presented). The scale for the coefficient values is given at the bottom left of panel a). Coefficients significantly different from 0 ($\alpha = 5\%$) are marked by asterisks (*). The fraction of positive interactions in each matrix is given by points in c) while the dashed (resp., dotted) line represents the ratio of interactions remaining positive (resp., negative) for all sites of a given region.

187 in the 10 sites (CHA/SKE) was mutualistic in Men er Roue only (Brittany) and mixed (+/-) in
 188 all other sites. Within the same region, however, interactions measured in different sites tended to
 189 keep the same sign. In the 3 sites of Oléron, for instance, there were 4 interactions which remained
 190 positive for both taxa involved (CHA/GUI, DIT/GUI, LEP/THP, SKE/THP), 3 of them being also
 191 mutualistic in some of the Brittany sites. This contradicts previous observations that mutualistic
 192 interactions tend to be more context-dependent than competitive interactions (Chamberlain *et al.*,
 193 2014).

194 Interaction network analysis

195 The stability (*sensu* resilience, Ives & Carpenter 2007) of all interaction matrices was not strongly
 196 affected by the percentage of positive interactions or the mean and variance of the interactions
 197 between taxa (Fig. 2). There was a slight increase in stability with weighted connectance, with a
 198 drop in eigenvalue modulus for weighted connectances between 0.09 and 0.1. The maximum modulus
 199 of the interaction matrix eigenvalues remained between 0.65 and 0.80.

200 Given that a direct complexity-stability (*sensu* resilience) link was not obvious, we investigated
 201 whether the matrix coefficients had some particular structure that could help theoretical ecology to
 202 make better null models of joint community dynamics and interactions (James *et al.*, 2015). We
 203 defined two scores, vulnerability (summed effect of others on the focal taxon growth rate, eq. S5)
 204 and impact (summed effect of the focal taxon onto other taxa's growth rates, eq. S6). Relations
 205 between intra- and inter-taxa interactions emerged (Fig. 3): genera that were more self-regulating
 206 also had also a higher vulnerability score and a lower impact score. Those two influences are likely
 207 to trade-off: a high degree of self-regulation somehow buffers the effect of outside influences on
 208 population dynamics. Taxa that were less self-regulating were also more likely to have a stronger
 209 effect onto other taxa. As these genera tended to be more abundant (Fig. S7), this could be mediated
 210 by the average density of a genus. It is important to note, however, that these trends are weak and
 211 there is therefore a considerable amount of randomness dominating the interaction matrix: many
 212 scenarios of self-regulation vs limitation by others are therefore possible.

213 Aside from the trade-offs of Fig. 3, we found no remarkable patterns of covariation between

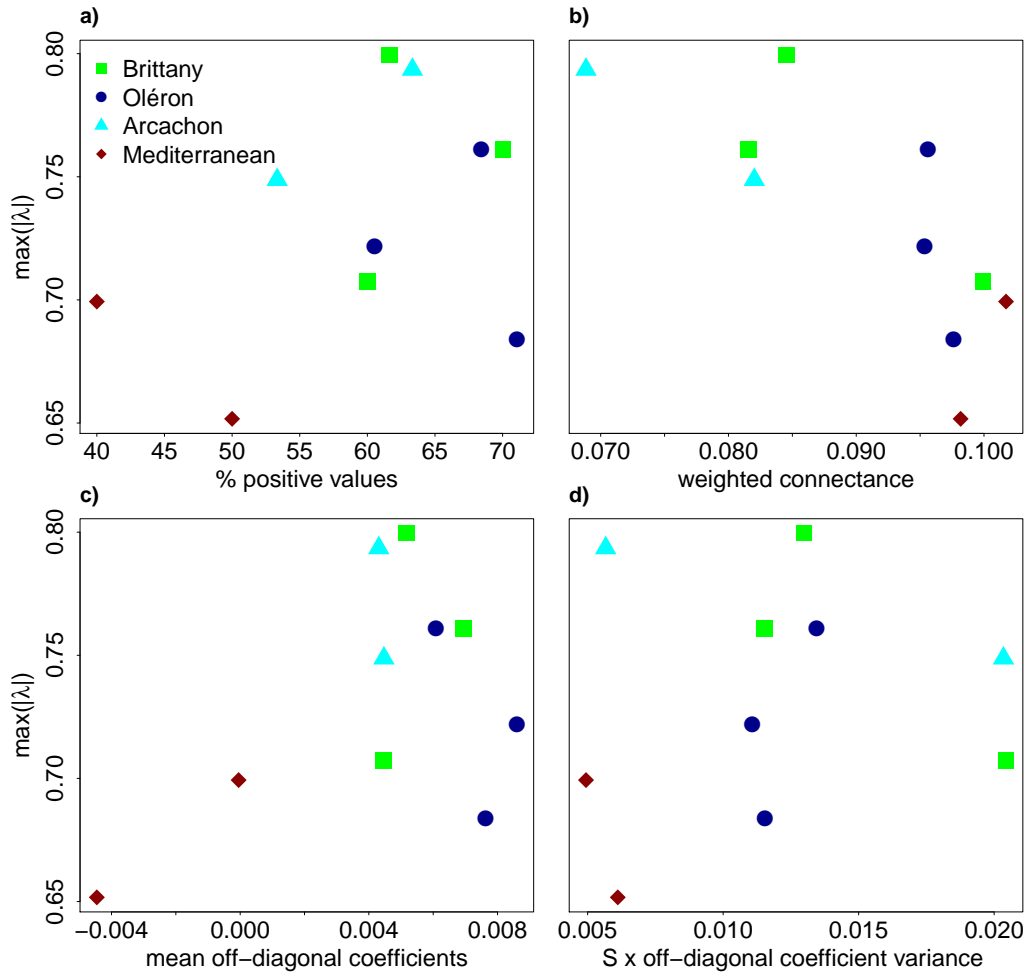


Figure 2: **Relation between stability and complexity of the interaction networks.** The maximum modulus of the eigenvalues of the interaction matrix \mathbf{B} indicates stability *sensu* resilience. Off-diagonal coefficient variance is multiplied by the dimension of the network, that is the number of species in the region. Each color or shape corresponds to a given region. The formula for weighted connectance is given in the Supporting Information.

214 matrix elements (Fig. S5) other than a mean-variance scaling of interaction coefficients (Fig. S6).

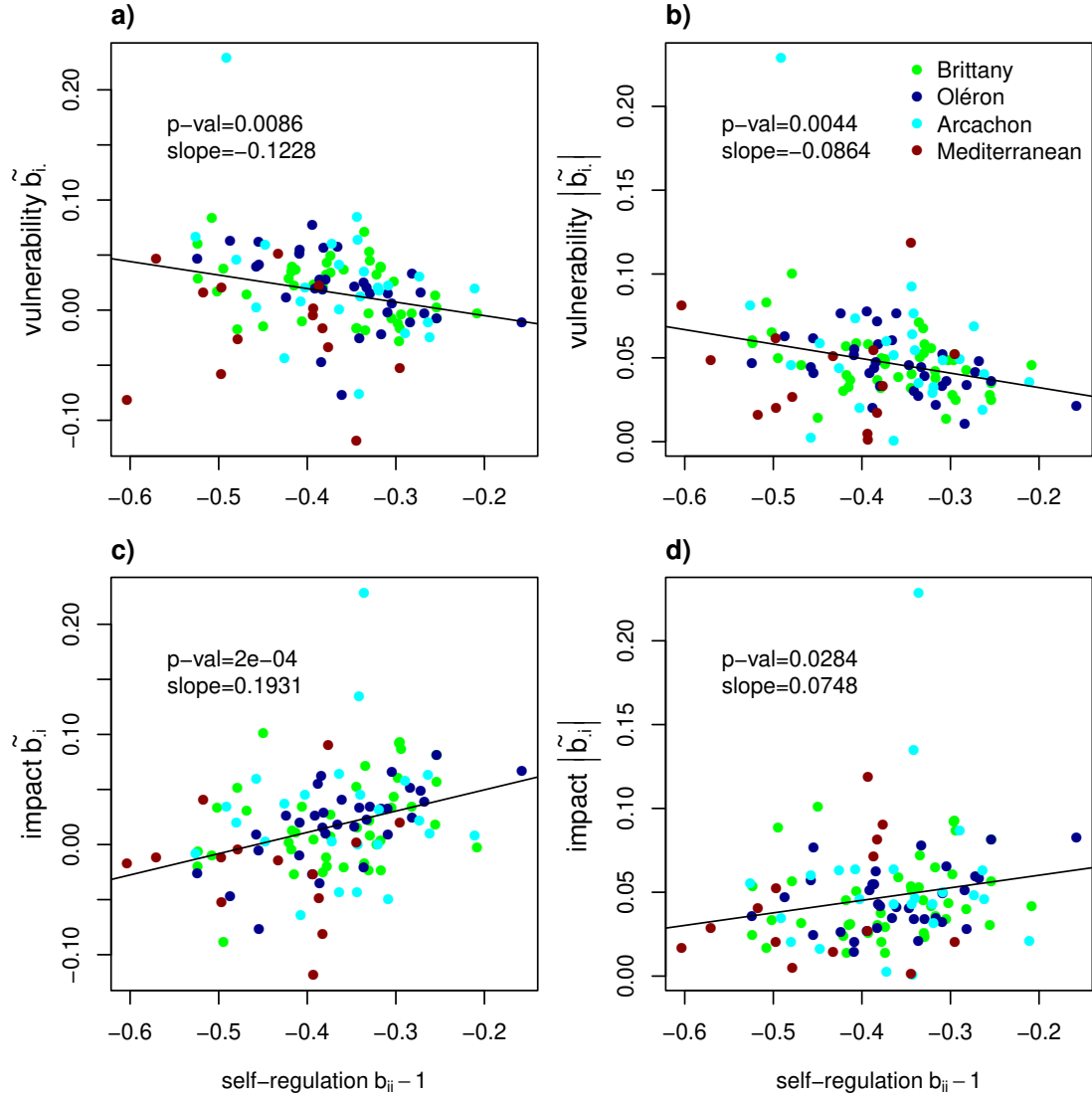


Figure 3: **Relation between vulnerability/impact and self-regulation.** Average vulnerability (effects of others on the focal taxon growth rate, a-b) and impact (effects of the focal taxon on others' growth rates, c-d), as well as self-regulation, are computed for untransformed (a-c) or absolute (b-d) values of the coefficients of the interaction matrix ($\mathbf{B} - \mathbf{I}$) for the 10 study sites. Each color corresponds to a given region (Fig S1). Linear regressions are shown as black lines.

Literature comparison

Finally, we sought to put these results in a broader context by compiling the intra vs inter taxa estimates of previous MAR(1) studies of long-term observational count data (listed in Table S4).

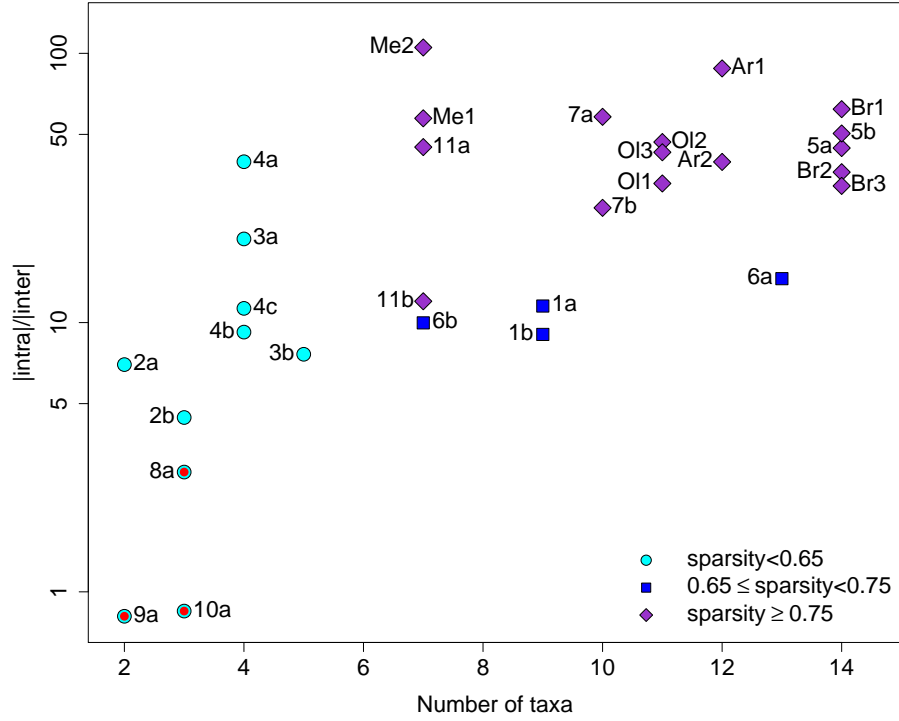


Figure 4: **Ratio of intra- to inter-taxa interaction strengths in Multivariate AutoRegressive (MAR) models.** The reference for each study is given in Table S4. Codes beginning with letters correspond to the present study (Ar: Arcachon; Ol: Oléron; Br: Brittany; Me: Mediterranean Sea). The symbol color and shape correspond to the sparsity of the interaction matrix (e.g., the proportion of null interactions in the matrix). Red dots correspond to terrestrial and/or low dimension predator-prey systems, giving a lower bound for the intra/inter ratio. Intertaxa interactions were set to 0 when they were not specified in the articles (in most cases, authors removed non-significant interactions at the 5% level; Fig. S9 is the same figure taking into account only significant interactions).

218 We found that the order of magnitude of intra/inter interaction strengths considered here is not
 219 particularly above those found for most planktonic systems to which MAR(1) models have been
 220 fitted, considering that our systems are relatively high-dimensional and that the higher the number
 221 of taxa, the larger the intraspecific regulation (Barabás *et al.*, 2017). We included in Fig. 4 not only
 222 plankton studies but also a couple of vertebrate or insect studies on less diverse communities, where
 223 interactions are stronger, in order to provide lower bounds for the intra/inter ratio. The conclusion
 224 from this comparison seems to be that, unlike small communities that can be tight-knit, any diverse

field system of competitors and facilitators has evolved large niche differences making on average intrataxon competition much larger in magnitude than intertaxa interactions.

Discussion

Strong self-regulation and facilitation

We found very large niche differences between genera, translating into much higher intragenus than intergenera effects on growth rates (i.e., strong self-regulation), together with a high degree of facilitative net interactions.

The intra/intertaxa interaction strength ratio (Levine & HilleRisLambers, 2009) that we found, from 6–10 to above 20, depending on whether one includes interactions set to zero before the estimation process, could appear very high in light of previous intra/interspecific competition strength estimates of 4 to 5 by Adler *et al.* (2018b). Additional estimates using the unconstrained interaction matrix yielded ratios between 8 and 11 depending on the site (Table S3 and Fig. S8 in the Supporting Information), but weak intertaxa effects are likely to be inflated in the full model. Therefore, a intra/inter ratio of 10 seems like a conservative estimate. It is twice that of Adler *et al.* (2018b) who use a different model, i.e., a Lotka-Volterra competition model. We outline how to relate a MAR(1) model to a discrete-time Lotka-Volterra equivalent in the Supporting Information; even though there is a relationship between intra/inter ratios in both models, the relationship is not trivial when abundances vary greatly between species. Hence, to some degree, intra/inter ratios can differ between model frameworks or ways of measuring density-dependencies (e.g., a high measurement error due to using proxies of densities for plants can result in bias in interaction coefficient estimates, Detto *et al.*, 2019). However, a ratio intra/inter at least twice larger than the ones previously found may call for other explanations. One could also argue that our high intra/inter ratio arises because we consider the genus as our baseline taxonomic unit, rather than the species. It is logical that niche differentiation increases as one gets up the phylogenetic tree, and that getting down to the species level could slightly decrease that ratio (but see Narwani *et al.*, 2017, in which phylogenetic closeness decreases competition strength). However, taxonomic resolution is unlikely to be the sole

251 explanation for the high intra/inter ratio of interaction strength found here, for two reasons. First,
 252 phytoplankton species belonging to different genera are often found to compete in experiments
 253 (Titman, 1976; Tilman *et al.*, 1982; Descamps-Julien & Gonzalez, 2005). In the field-based dataset
 254 studied here, the same genera that are considered in experiments are found not to compete (or only
 255 weakly), hence there must be some niche differentiation occurring in the field but not in the lab.
 256 Second, the only other study that managed to provide MAR(1) estimates down to the species level
 257 for phytoplankton, that of Huber & Gaedke (2006), provides an intra/interspecific strength ratio
 258 similar to ours (point 7a in Fig. 4). Strong self-regulation seems therefore a genuine feature of field
 259 phytoplanktonic communities. We discuss below possible mechanistic interpretations.

260 Another main finding of our study is the large frequency of positive interactions, with 30% truly
 261 mutualistic (+/+) interactions and between 40 and 70% facilitative effects. Although a seasonal
 262 environment can generate some positive covariation between taxa, those effects have already been
 263 filtered out by the inclusion of our 2 abiotic covariates (Fig. S4). The facilitative effects shown
 264 here are therefore residual effects, once abiotic trends are accounted for. Between 40 and 70%
 265 facilitation can be compared to the meta-analysis by Adler *et al.* (2018b) who also found facilitative
 266 interactions, but less than here ($\approx 30\%$). However, Adler *et al.* (2018b)'s review contains many
 267 experiments while the plant literature is replete with field examples of facilitation (Brooker *et al.*,
 268 2008; McIntire & Fajardo, 2014), so that plant facilitation could be higher in the field. At the
 269 moment, it is therefore unknown how the predominance of facilitative interactions that we found
 270 in phytoplankton compares to facilitation in terrestrial plants. We note that several authors using
 271 MAR(1) models previously forbade positive interactions within the same trophic level, so that the
 272 fraction of facilitative interactions in plankton cannot be computed from literature-derived MAR(1)
 273 estimates.

274 The large niche differences and facilitative interactions that arise when considering a single
 275 trophic level are an emergent property, resulting from hidden effects of resource or predator
 276 partitioning/sharing (Chesson, 2018). In our previous publication investigating in detail the Arcachon
 277 study sites (Barraquand *et al.*, 2018), we have argued that for phytoplankton, the strong intrataxon
 278 density-dependence could arise from effects of natural enemies (Haydon, 1994). Natural enemies

could also very well create apparent mutualism between prey species (Abrams *et al.*, 1998; de Ruiter & Gaedke, 2017). We believe this to be likely for the present study, given that the study regions (Arcachon, Oléron, Brittany, Mediterranean) have similar predators (zooplankton, e.g., Jamet *et al.*, 2001; Modéran *et al.*, 2010; Tortajada *et al.*, 2012) and parasites (viruses, e.g., Ory *et al.*, 2010; fungi). Though natural enemies are good candidates to explain the observed niche differences and emerging facilitation, one must bear in mind that other known drivers of phytoplankton dynamics such as allelopathy (Felpeto *et al.*, 2018), auxotrophy (Tang *et al.*, 2010) or hydrodynamics (Lévy *et al.*, 2018) can all, in theory, help create different niches and an emerging facilitation (see last subsection of the Discussion). Finally, resources that are usually considered limiting for all species might in fact not always be: Burson *et al.* (2018) show that phytoplanktonic taxa specialize on different components of the light spectrum. This constitutes an example of fine-scale resource partitioning of one resource, light, that all species and genera are usually thought to compete for.

No complexity-stability relationship but connections between self-regulation and interactions between taxa

There was no relation between the complexity of the communities (measured as either the weighted connectance or the interaction coefficient variance) and their stability (measured by the largest modulus of the eigenvalues, which quantifies the return time to a point equilibrium, i.e., resilience). This result is conditional upon our model being a good approximate description of the system (i.e., no multiyear limit cycles or chaotic attractors as the mapping between eigenvalues and actual stability is distorted in that case, Certain *et al.*, 2018). However, we already showed on a subset of this data that a fixed point in a MAR(1) model, perturbed by seasonality and abiotic variables, is an accurate description of the system (Barraquand *et al.*, 2018). Therefore, we are confident that the absence of complexity-resilience relationship found here is not a mere artefact of an inadequate model. This absence of direct link between complexity and stability could be an actual feature of empirical systems, as shown previously by Jacquet *et al.* (2016) using a different technique. This result seems to contradict theory based on random matrices, especially for competitive and/or mutualistic networks (Allesina & Tang, 2012). However, one must bear in mind that such result could also be generated

by the limited size of our networks, as random matrix theory relies on asymptotics (Allesina & Tang, 2015). We should also mention that our interaction matrices (based on a discrete-time model) are not strictly analogous to the ones used most frequently in theoretical ecology (continuous-time model), though the spectral radius (largest modulus) is here tightly related to the real part of the lead eigenvalue in equivalent continuous-time models (see Supporting Information). Thus while the jury is still out regarding the absence of complexity-resilience relation found here, it may well be a genuine absence. In addition to complexity metrics, we also found that the percentage of mutualistic interactions, that is thought to affect the stability of a network, either positively or negatively (Mougi & Kondoh, 2012; Coyte *et al.*, 2015; García-Callejas *et al.*, 2018), does not in fact have a major impact on our networks' resilience.

In addition to weighted connectance and interaction variance, indices at the genus level (vulnerability and impact) approximate the average effects exerted and sustained by any given taxa in the different study sites. While, at the network level, network structure (either complexity measures or the percentage of mutualistic interactions) did not affect resilience, a relation emerged between self-regulation, necessary for coexistence, and genus-level indices. We found that the more a genus is self-regulated, the more it tends to be vulnerable to other genera's impacts and the less it impacts other genera. We examined whether vulnerability and impact could be affected by phylogenetic correlations; they were not, as on Fig. 3, points were not clustered according to genus, family or phylum. High self-regulation usually indicates large niche differences with the rest of the community, and it makes therefore sense that a species/genus whose needs strongly differ from the others only marginally impacts the resources of the other coexisting species. This is what we expect under strong niche partitioning. A low self-regulation was also correlated with high average abundance, which echoes findings by Yenni *et al.* (2017) who demonstrated that rare species usually show stronger self-regulation. This correlation between relative rarity and self-regulation could explain the lesser impact of highly self-regulated species/genus: a taxon which dominates the community composition can have a major effect on the others, especially as they usually cover more space, while it is harder for the less common taxa to have large impacts. In contrast, it was more difficult to explain the relationship between self-regulation and vulnerability: a genus that is more self-regulated and less

334 common was found here to be on average more vulnerable to other genera’s increases in densities.
 335 Such relation implies greater stability (*sensu* resilience, Ives *et al.* 2003, and also invariability,
 336 Arnoldi *et al.* 2019) for the network as a whole, because the taxa that are the more vulnerable
 337 to other taxa’s impacts are also those whose dynamics are intrinsically more buffered. By which
 338 mechanisms this could happen is so far unclear and open to speculation. It could just be a “mass
 339 effect”: common taxa are in high enough numbers to deplete resources or change the environment in
 340 ways that affect the less common ones, but the reverse is not true. As a final note on relationships
 341 between interaction matrix coefficients, we caution that the trends evidenced are all relatively weak:
 342 considerable stochasticity still dominates the distribution of interaction matrix coefficients.

343 Ghosts of competition past and present

344 Overall, the dominance of niche differentiation in observational plankton studies – based on our
 345 analysis of the REPHY dataset and re-analysis of the MAR(1) literature – is similar to what has
 346 been recently found in plant community studies (Adler *et al.*, 2018b) or empirically parameterized
 347 food webs including horizontal diversity (Barabás *et al.*, 2017). Large niche differences might be
 348 due to the ghost of competition past, i.e., competition has occurred in the past, leading to strong
 349 selection and subsequent evolution, and then to progressive niche separation. In this scenario,
 350 species have evolved niches that allow them not to compete or to interact only weakly (very strong
 351 facilitative effects might be likewise destabilizing, Coyte *et al.*, 2015). The likely predator effects that
 352 we highlighted above could be comprised within such niche differentiation *sensu largo*: specialized
 353 predators can make strong conspecific density-dependence emerge (Bagchi *et al.*, 2014; Comita *et al.*,
 354 2014), while switching generalists can also promote diversity (Vallina *et al.*, 2014). Both predators
 355 and resources have often symmetrical effects and can therefore contribute almost equally to such
 356 past niche differentiation (Chesson, 2018).

357 An intriguing new possibility, dubbed the “ghost of competition present” (Tuck *et al.*, 2018),
 358 suggests by contrast that spatial distributions in relation to abiotic factors might have a large
 359 impact on the interaction strengths inferred from temporal interaction models such as ours. Recent
 360 combinations of model fitting and removal experiments have shown that model fitting usually

underestimates the effect of competitors that are uncovered by removal experiments (Tuck *et al.*, 2018; Adler *et al.*, 2018a). This could occur for instance if species are spatially segregated (at a small scale) because each species only exists within a domain where it is relatively competitive (Pacala’s spatial segregation hypothesis, chapter 15 in Pacala & Levin 1997), while a focal species could spread out if competitors were removed. This means that a species can be limited by competitors, but act so as to minimize competition (a little like avoidance behaviour in animals) and maximize opportunities for positive interactions, which implies that competition is in effect hard to detect when all species are present. This mechanism would require some spatial segregation between phytoplankton species at the scale of interactions, i.e., at the microscale. At the moment, while it is known that fine-scale hydrodynamics generate inhomogeneities at the microscale (Barton *et al.*, 2014; Breier *et al.*, 2018) it is yet quite unclear how they might affect multivariate spatial patterns of species distributions (*sensu* Bolker & Pacala 1999 or Murrell & Law 2003). Moreover, even with some microscale spatial segregation between species, a “ghost of competition present” mechanism might not work in phytoplankton as in terrestrial plants, because the turbulent, ever-changing aquatic environment imposes additional constraints on the spatial distribution of organisms.

Acknowledgements

This study was only made possible by the dedication of the members of the REPHY program by Ifremer (REPHY, 2017), providing invaluable data through years of fieldwork. We are grateful to David Murrell for his careful reading and suggestions, and to Peter Adler for helpful exchanges. We also want to thank Bérangère Péquin for constructive comments, as well as Gyuri Barabás whose thoughtful suggestions improved the connection made to continuous-time stability theory. This study was supported by the French ANR through LabEx COTE (ANR-10-LABX-45).

Supporting Information

This article contains supporting information.

385 **Authors' contributions**

386 CP and FB contributed equally to the project design and the methodology. The computer code was
387 written by CP. The authors jointly interpreted the results and co-wrote the manuscript after an
388 early draft by FB.

389 **Data accessibility**

390 The REPHY dataset has already been published (REPHY, 2017). All scripts for MAR models and
391 subsequent network analyses are available online in a GitHub repository ([https://github.com/](https://github.com/CoraliePicoche/REPHY-littoral)
392 [CoraliePicoche/REPHY-littoral](https://github.com/CoraliePicoche/REPHY-littoral)) and Zenodo [address here XXXX].

References

- Abrams, P., Holt, R. & Roth, J. (1998). Apparent competition or apparent mutualism? Shared predation when populations cycle. *Ecology*, **79**, 201–212.
- Adler, P., Ellner, S. & Levine, J. (2010). Coexistence of perennial plants: an embarrassment of niches. *Ecology Letters*, **13**, 1019–1029.
- Adler, P., Kleinhesselink, A., Hooker, G., Taylor, J., Teller, B. & Ellner, S. (2018a). Weak interspecific interactions in a sagebrush steppe? Conflicting evidence from observations and experiments. *Ecology*, **99**, 1621–1632.
- Adler, P., Smull, D., Beard, K., Choi, R., Furniss, T., Kulmatiski, A., Meiners, J., Tredennick, A. & Veblen, K. (2018b). Competition and coexistence in plant communities: intraspecific competition is stronger than interspecific competition. *Ecology Letters*, **21**, 1319–1329.
- Allesina, S. & Tang, S. (2012). Stability criteria for complex ecosystems. *Nature*, **483**, 205–208.
- Allesina, S. & Tang, S. (2015). The stability-complexity relationship at age 40: a random matrix perspective. *Population Ecology*, **57**, 63–75.

- Armstrong, R. & McGehee, R. (1976). Coexistence of species competing for shared resources. *Theoretical Population Biology*, **9**, 317–328.
- Armstrong, R. & McGehee, R. (1980). Competitive exclusion. *The American Naturalist*, **115**, 151–170.
- Arnoldi, J.F., Loreau, M. & Haegeman, B. (2019). The inherent multidimensionality of temporal variability: how common and rare species shape stability patterns. *Ecology Letters*, **22**, 1557–1567.
- Bagchi, R., Gallery, R., Gripenberg, S., Gurr, S., Narayan, L., Addis, C., Freckleton, R. & Lewis, O. (2014). Pathogens and insect herbivores drive rainforest plant diversity and composition. *Nature*, **506**, 85–88.
- Barabás, G., Michalska-Smith, M. & Allesina, S. (2017). Self-regulation and the stability of large ecological networks. *Nature Ecology & Evolution*, **1**, 1870–1875.
- Barraquand, F., Picoche, C., Maurer, D., Carassou, L. & Auby, I. (2018). Coastal phytoplankton community dynamics and coexistence driven by intragroup density-dependence, light and hydrodynamics. *Oikos*, **127**, 1834–1852.
- Barton, A., Ward, B., Williams, R. & Follows, M. (2014). The impact of fine-scale turbulence on phytoplankton community structure: Phytoplankton and turbulence. *Limnology and Oceanography: Fluids and Environments*, **4**, 34–49.
- Bersier, L.F., Banasek-Richter, C. & Cattin, M.F. (2002). Quantitative descriptors of food-web matrices. *Ecology*, **83**, 2394.
- Bolker, B. & Pacala, S. (1999). Spatial moment equations for plant competition: understanding spatial strategies and the advantages of short dispersal. *The American Naturalist*, **153**, 575–602.
- Breier, R., Lalescu, C., Waas, D., Wilczek, M. & Mazza, M. (2018). Emergence of phytoplankton patchiness at small scales in mild turbulence. *Proceedings of the National Academy of Sciences*, **115**, 12112–12117.

- Brooker, R., Maestre, F., Callaway, R., Lortie, C., Cavieres, L., Kunstler, G., Liancourt, P., Tielbörger, K., Travis, J., Anthelme, F., Armas, C., Coll, L., Corcket, E., Delzon, S., Forey, E., Kikvidze, Z., Olofsson, J., Pugnaire, F., Quiroz, C., Saccone, P., Schiffrers, K., Seifan, M., Touzard, B. & Michalet, R. (2008). Facilitation in plant communities: the past, the present, and the future. *Journal of Ecology*, **96**, 18–34.
- Burson, A., Stomp, M., Greenwell, E., Grosse, J. & Huisman, J. (2018). Competition for nutrients and light: testing advances in resource competition with a natural phytoplankton community. *Ecology*, **99**, 1108–1118.
- Cavieres, L. & Badano, E. (2009). Do facilitative interactions increase species richness at the entire community level? *Journal of Ecology*, **97**, 1181–1191.
- Certain, G., Barraquand, F. & Gårdmark, A. (2018). How do MAR(1) models cope with hidden nonlinearities in ecological dynamics? *Methods in Ecology and Evolution*, **9**, 1975–1995.
- Chamberlain, S., Bronstein, J. & Rudgers, J. (2014). How context dependent are species interactions? *Ecology Letters*, **17**, 881–890.
- Chesson, P. (2018). Updates on mechanisms of maintenance of species diversity. *Journal of Ecology*, **106**, 1773–1794.
- Chesson, P. & Huntly, N. (1997). The roles of harsh and fluctuating conditions in the dynamics of ecological communities. *The American Naturalist*, **150**, 519–553.
- Comita, L., Queenborough, S., Murphy, S., Eck, J., Xu, K., Krishnadas, M., Beckman, N. & Zhu, Y. (2014). Testing predictions of the Janzen-Connell hypothesis: a meta-analysis of experimental evidence for distance- and density-dependent seed and seedling survival. *Journal of Ecology*, **102**, 845–856.
- Coyte, K., Schluter, J. & Foster, K. (2015). The ecology of the microbiome: Networks, competition, and stability. *Science*, **350**, 663–666.

- Cushing, J., Levarge, S., Chitnis, N. & Henson, S.M. (2004). Some discrete competition models and the competitive exclusion principle. *Journal of Difference Equations and Applications*, **10**, 1139–1151.
- Descamps-Julien, B. & Gonzalez, A. (2005). Stable coexistence in a fluctuating environment: an experimental demonstration. *Ecology*, **86**, 2815–2824.
- Detto, M., Visser, M., Wright, S. & Pacala, S. (2019). Bias in the detection of negative density dependence in plant communities. *Ecology Letters*, **22**, 1923–1939.
- Felpeto, A., Roy, S. & Vasconcelos, V. (2018). Allelopathy prevents competitive exclusion and promotes phytoplankton biodiversity. *Oikos*, **127**, 85–98.
- García-Callejas, D., Molowny-Horas, R. & Araújo, M. (2018). The effect of multiple biotic interaction types on species persistence. *Ecology*, **99**, 2327–2337.
- Griffiths, J., Hajdu, S., Downing, A., Hjerne, O., Larsson, U. & Winder, M. (2015). Phytoplankton community interactions and environmental sensitivity in coastal and offshore habitats. *Oikos*, **125**, 1134–1143.
- Gross, K. (2008). Positive interactions among competitors can produce species-rich communities. *Ecology Letters*, **11**, 929–936.
- Hampton, S., Holmes, E., Scheef, L., Scheuerell, M., Katz, S., Pendleton, D. & Ward, E. (2013). Quantifying effects of abiotic and biotic drivers on community dynamics with multivariate autoregressive (MAR) models. *Ecology*, **94**, 2663–2669.
- Hampton, S., Izmet’Eva, L., Moore, M., Katz, S., Dennis, B. & Silow, E. (2008). Sixty years of environmental change in the world’s largest freshwater lake - Lake Baikal, Siberia. *Global Change Biology*, **14**, 1947–1958.
- Hampton, S., Scheuerell, M. & Schindler, D. (2006). Coalescence in the lake washington story: Interaction strengths in a planktonic food web. *Limnology and Oceanography*, **51**, 2042–2051.

- Haydon, D. (1994). Pivotal assumptions determining the relationship between stability and complexity: an analytical synthesis of the stability-complexity debate. *The American Naturalist*, **144**, 14–29.
- Hernández Fariñas, T., Bacher, C., Soudant, D., Belin, C. & Barillé, L. (2015). Assessing phytoplankton realized niches using a French national phytoplankton monitoring network. *Estuarine, Coastal and Shelf Science*, **159**, 15–27.
- Holmes, E., Ward, E. & Scheuerell, M. (2014). *MARSS: Multivariate Autoregressive State-Space Modeling*. R package version 3.9 <https://CRAN.R-project.org/package=MARSS>.
- Holmes, E.E., Ward, E.J. & Wills, K. (2012). Marss: Multivariate autoregressive state-space models for analyzing time-series data. *The R Journal*, **4**, 30.
- Hubbell, S. (2001). *The unified neutral theory of biodiversity and biogeography*. Princeton University Press.
- Huber, V. & Gaedke, U. (2006). The role of predation for seasonal variability patterns among phytoplankton and ciliates. *Oikos*, **114**, 265–276.
- Huisman, J. & Weissing, F. (2001). Biological conditions for oscillations and chaos generated by multispecies competition. *Ecology*, **82**, 2682–2695.
- Hutchinson, G. (1961). The paradox of the plankton. *The American Naturalist*, **95**, 137–145.
- Ives, A., Gross, K. & Klug, J. (1999). Stability and variability in competitive communities. *Science*, **286**, 542–544.
- Ives, A.R. & Carpenter, S.R. (2007). Stability and diversity of ecosystems. *Science*, **317**, 58–62.
- Ives, A.R., Dennis, B., Cottingham, K.L. & Carpenter, S.R. (2003). Estimating community stability and ecological interactions from time-series data. *Ecological Monographs*, **73**, 301–330.
- Jacquet, C., Moritz, C., Morissette, L., Legagneux, P., Massol, F., Archambault, P. & Gravel, D. (2016). No complexity-stability relationship in empirical ecosystems. *Nature Communications*, **7**.

- James, A., Plank, M., Rossberg, A., Beecham, J., Emmerson, M. & Pitchford, J. (2015). Constructing random matrices to represent real ecosystems. *The American Naturalist*, **185**, 680–692.
- Jamet, J.L., Boge, G., Richard, S., Geneys, C. & Jamet, D. (2001). The zooplankton community in bays of Toulon area (northwest Mediterranean Sea, France). *Hydrobiologia*, **557**, 155–165.
- Kinlock, N.L. (2019). A meta-analysis of plant interaction networks reveals competitive hierarchies as well as facilitation and intransitivity. *The American Naturalist*, **194**, 640–653.
- Levine, J. & HilleRisLambers, J. (2009). The importance of niches for the maintenance of species diversity. *Nature*, **461**, 254–257.
- Lévy, M., Franks, P. & Smith, K. (2018). The role of submesoscale currents in structuring marine ecosystems. *Nature Communications*, **9**, 4758.
- Li, L. & Chesson, P. (2016). The effects of dynamical rates on species coexistence in a variable environment: the paradox of the plankton revisited. *The American Naturalist*, **188**, E46–E58.
- McIntire, E. & Fajardo, A. (2014). Facilitation as a ubiquitous driver of biodiversity. *New Phytologist*, **201**, 403–416.
- Modéran, J., Bouvais, P., David, V., Le Noc, S., Simon-Bouhet, B., Niquil, N., Miramand, P. & Fichet, D. (2010). Zooplankton community structure in a highly turbid environment (Charente estuary, France): Spatio-temporal patterns and environmental control. *Estuarine, Coastal and Shelf Science*, **88**, 219–232.
- Mougi, A. & Kondoh, M. (2012). Diversity of interaction types and ecological community stability. *Science*, **337**, 349–351.
- Murrell, D. & Law, R. (2003). Heteromyopia and the spatial coexistence of similar competitors. *Ecology Letters*, **6**, 48–59.
- Mutshinda, C.M., O’Hara, R.B. & Woivod, I.P. (2009). What drives community dynamics? *Proceedings of the Royal Society B: Biological Sciences*, **276**, 2923–2929.

- Narwani, A., Bentlage, B., Alexandrou, M., Fritschie, K., Delwiche, C., Oakley, T. & Cardinale, B. (2017). Ecological interactions and coexistence are predicted by gene expression similarity in freshwater green algae. *Journal of Ecology*, **105**, 580–591.
- Ory, P., Hartmann, H., Jude, F., Dupuy, C., Del Amo, Y., Catala, P., Mornet, F., Huet, V., Jan, B., Vincent, D., Sautour, B. & Montanié, H. (2010). Pelagic food web patterns: do they modulate virus and nanoflagellate effects on picoplankton during the phytoplankton spring bloom? Effects of viruses and nanoflagellates on picoplankton. *Environmental Microbiology*, pp. 2755–2772.
- Pacala, S. & Levin, S. (1997). *Spatial ecology: the role of space in population dynamics and interspecific interactions*. Princeton University Press, Princeton, NJ.
- Qian, J.J. & Akçay, E. (2020). The balance of interaction types determines the assembly and stability of ecological communities. *Nature Ecology & Evolution*, **4**, 356–365.
- R Core Team (2016). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna, Austria.
- REPHY (2017). *REPHY dataset - French Observation and Monitoring program for Phytoplankton and Hydrology in coastal waters. 1987-2016 Metropolitan data*. SEANOE, <https://www.seanoe.org/data/00361/47248/>. Doi:10.17882/47248 (2017 version).
- Rosindell, J., Hubbell, S. & Etienne, R. (2011). The unified neutral theory of biodiversity and biogeography at age ten. *Trends in Ecology & Evolution*, **26**, 340.
- de Ruiter, P. & Gaedke, U. (2017). Emergent facilitation promotes biological diversity in pelagic food webs. *Food Webs*, **10**, 15–21.
- Scheef, L., Hampton, S. & Izmet'Eva, L. (2013). Inferring plankton community structure from marine and freshwater long-term data using multivariate autoregressive models. *Limnology and Oceanography: Methods*, **11**, 475–484.
- Tang, Y.Z., Koch, F. & Gobler, C.J. (2010). Most harmful algal bloom species are vitamin B1 and B12 auxotrophs. *Proceedings of the National Academy of Sciences*, **107**, 20756–20761.

- Tilman, D., Kilham, S. & Kilham, P. (1982). Phytoplankton community ecology: the role of limiting nutrients. *Annual Review of Ecology and Systematics*, **13**, 349–372.
- Titman, D. (1976). Ecological competition between algae: experimental confirmation of resource-based competition theory. *Science*, **192**, 463–465.
- Tortajada, S., Niquil, N., Blanchet, H., Grami, B., Montanié, H., David, V., Glé, C., Saint-Béat, B., Johnson, G., Marquis, E., Del Amo, Y., Dubois, S., Vincent, D., Dupuy, C., Jude, F., Hartmann, H. & Sautour, B. (2012). Network analysis of the planktonic food web during the spring bloom in a semi enclosed lagoon (Arcachon, SW France). *Acta Oecologica*, **40**, 40–50.
- Tuck, S., Porter, J., Rees, M. & Turnbull, L. (2018). Strong responses from weakly interacting species. *Ecology Letters*, **21**, 1845–1852.
- Utermöhl, H. (1958). Zur Vervollkommnung der quantitativen Phytoplankton-Methodik. *Internationale Vereinigung für theoretische und angewandte Limnologie: Mitteilungen*, **9**, 1–38.
- Vallina, S., Ward, B., Dutkiewicz, S. & Follows, M. (2014). Maximal feeding with active prey-switching: A kill-the-winner functional response and its effect on global diversity and biogeography. *Progress in Oceanography*, **120**, 93–109.
- Violle, C., Nemergut, D., Pu, Z. & Jiang, L. (2011). Phylogenetic limiting similarity and competitive exclusion. *Ecology Letters*, **14**, 782–787.
- Volkov, I., Banavar, J., Hubbell, S. & Maritan, A. (2009). Inferring species interactions in tropical forests. *Proceedings of the National Academy of Sciences*, **106**, 13854–13859.
- Yenni, G., Adler, P. & Ernest, S.K.M. (2017). Do persistent rare species experience stronger negative frequency dependence than common species? Asymmetric NFD and persistent rare species. *Global Ecology and Biogeography*, **26**, 513–523.