

# Strong self-regulation and widespread facilitative interactions in phytoplankton communities

Coralie Picoche<sup>1,2</sup>, Frédéric Barraquand<sup>1,2\*</sup>

February 1, 2020

**1** University of Bordeaux, Integrative and Theoretical Ecology, LabEx COTE, Bât. B2 - Allée  
Geoffroy St-Hilaire, 33615 Pessac, France; coralie.picoche@u-bordeaux.fr

**2** CNRS, Institute of Mathematics of Bordeaux, 351 Cours de la Libération, 33405 Talence, France;  
frederic.barraquand@u-bordeaux.fr

\* corresponding author: Email: frederic.barraquand@u-bordeaux.fr

## Abstract

1. The persistence of phytoplanktonic diversity in spite of competition for basic resources has long been a source of wonder and inspiration to ecologists. To sort out, among the many coexistence mechanisms suggested by theory and experiments, which ones actually maintain diversity in natural ecosystems, long-term field studies are paramount.
2. We analysed a large dataset of phytoplankton abundance time series using dynamic, multivariate autoregressive models. Phytoplankton was counted and identified down to the genus level, every two weeks over twenty years, at ten sites along the French coastline. Multivariate autoregressive models allowed to estimate biotic interaction networks, while also accounting for abiotic variables that may drive part of the phytoplankton fluctuations. We then analysed the ratio of intra- to inter-taxa interactions (measuring self-regulation, itself a measure of niche differentiation), the frequency of negative vs positive interactions, and how stability metrics (both at the network and genus level) relate to network complexity and genus self-regulation or abundance.
3. We showed that a strong self-regulation, with competition strength within a taxon (genus) an order of magnitude higher than between taxa, was present in all phytoplanktonic interaction networks. This much stronger intragenus competition suggests that niche differentiation - rather than neutrality - is commonplace in phytoplankton. Furthermore, interaction networks were dominated by positive net effects between phytoplanktonic taxa (on average, more than 50% of interactions were positive). While network stability (*sensu* resilience) was unrelated to complexity measures, we unveiled links between self-regulation, intergenus interaction strengths and abundance. The less common taxa tend to be more strongly self-regulated and can therefore maintain in spite of competition with more abundant ones.
4. *Synthesis:* We demonstrate that strong niche differentiation, widespread facilitation between phytoplanktonic taxa and stabilizing covariances between interaction strengths should be common features of coexisting phytoplankton communities in the field. These are structural

properties that we can expect to emerge from plausible mechanistic models of phytoplankton communities. We discuss mechanisms, such as predation or restricted microscale movement, that are consistent with these findings, which paves the way for further research.

**Keywords:** phytoplankton; coexistence; facilitation; mutualism; niche theory; time series; networks

# 1 Introduction

2 How species or close genera can coexist together in spite of competition is one of the main puzzles of  
3 community ecology, especially for primary producers that seemingly share the same basic resources  
4 ([Hutchinson, 1961](#)). Many theoretical studies of competition models have shown that competitive  
5 exclusion is likely in those circumstances, unless mechanisms involving spatial or temporal variation  
6 are at play ([Armstrong & McGehee, 1976, 1980](#); [Chesson & Huntly, 1997](#); [Huisman & Weissing, 2001](#);  
7 [Li & Chesson, 2016](#); [Chesson, 2018](#)). Neutral theory, that assumes a non-equilibrium coexistence  
8 maintained by dispersal and equal competitive abilities for all species ([Hubbell 2001](#), though there  
9 are exceptions, see [Volkov \*et al.\* 2003, 2007](#)), has been proposed as a solution to explain highly  
10 diverse communities ([Hubbell, 2001](#); [Rosindell \*et al.\*, 2011](#)).

11 However, the evidence gathered from terrestrial plant communities starts to suggest that, in fact,  
12 niche rather than neutral processes may be paramount to explain coexistence, with intraspecific  
13 competition dwarfing interspecific competition in most cases ([Adler \*et al.\*, 2010, 2018b](#)). Whether  
14 these conclusions drawn mostly from studies of terrestrial plants apply to other ecosystems and taxa  
15 is currently little known (but see [Mutshinda \*et al.\* 2009](#)).

16 Moreover, competition may not be the rule: the meta-analysis by [Adler \*et al.\* \(2018b\)](#) reported a  
17 large number of facilitative interactions (30%) and several reviews ([Brooker \*et al.\*, 2008](#); [McIntire  
18 & Fajardo, 2014](#); [Kinlock, 2019](#)) have highlighted that facilitation may be much more widespread  
19 than ecologists usually tend to think. Although some theoretical studies suggest that facilitative  
20 interactions can be destabilizing (*sensu* resilience) and therefore undermine coexistence in Lotka-  
21 Volterra models ([Coyte \*et al.\*, 2015](#)), multiple other modelling ([Gross, 2008](#)) and empirical ([Brooker  
22 \*et al.\*, 2008](#); [Cavieres & Badano, 2009](#)) studies have suggested that facilitative interactions can  
23 to a large degree benefit coexistence, especially when multiple interaction types are considered  
24 simultaneously ([Mougi & Kondoh, 2012](#); [García-Callejas \*et al.\*, 2018](#)).

25 Here, we analyse a spatially replicated, long-term community-level dataset, consisting of ten  
26 multivariate time series of phytoplankton abundance along the French coastline. We do so using  
27 multivariate autoregressive (MAR) models, that allow to estimate interactions between genera.

28 Although many ecological studies focus on interactions between species, competition has been shown  
29 experimentally to occur between different genera of phytoplankton ([Titman, 1976](#); [Descamps-Julien  
30 & Gonzalez, 2005](#)). The genus level is also a rather fine taxonomic scale for phytoplankton interaction  
31 studies, as most studies are restricted to interactions between different classes or even phyla ([Ives  
32 et al., 2003](#); [Hampton et al., 2008](#); [Griffiths et al., 2015](#)). Studying interactions between different  
33 genera of phytoplankton therefore both makes empirical sense in light of competition experiments  
34 and allows to estimate better-resolved networks. We focus here on genera that belong mostly to  
35 diatoms and dinoflagellates. To put our results into a more general context, we then compare our  
36 interaction strength estimates to previously published interaction networks produced under the same  
37 statistical framework, both in plankton and other empirical systems.

## 38 Material and methods

### 39 Sampling methods

40 All phytoplankton samples were collected by Ifremer coastal laboratories as part of the National  
41 Phytoplankton and Phycotoxin Monitoring Network ([REPHY, 2017](#)). Since 1987, this monitoring  
42 program has required the sampling of 26 sites along the French coastline every 2 weeks within 2  
43 hours of high tide to document both biotic (phytoplankton counts) and abiotic (water temperature,  
44 salinity) variables. We focused on sites which had the longest time series. We also excluded time  
45 series which had missing data for over 6 months or an average delay between sampling dates above  
46 20 days. This reduced the number of study sites to 10 sites nested within 4 regions (Brittany, Oléron,  
47 Arcachon and the Mediterranean Sea; Fig. S1 and Table S1 in the Supporting Information).

48 Abiotic variables (temperature, salinity) were measured directly from the boat during the sampling  
49 process while water samples for biotic analyses were fixed with a Lugol's solution and examined later.  
50 Phytoplankton cells above 20  $\mu\text{m}$  were identified at the lowest possible taxonomic level and counted  
51 with the Utermöhl method using an optical microscope ([Utermöhl, 1958](#)). Throughout the years  
52 and sites, more than 600 taxa were identified at different taxonomic levels. We aggregated them at  
53 the genus (or group of genera when not possible) level based on previous work (Table S2; [Hernández](#)

54 [Fariñas \*et al.\* 2015](#); [Barraquand \*et al.\* 2018](#)), except for cryptophytes and euglenophytes in Arcachon,  
 55 which could not be identified below the family level. Although the taxonomic resolution used here  
 56 may seem coarse in comparison to land plants, it is in fact more refined than 86% of the MAR(1)  
 57 studies of phytoplankton listed in Table S4.

58 For each region, the MAR(1) analysis focused on the most abundant and most frequently observed  
 59 genera to avoid most of the gaps in the time series. When gaps did not exceed a month, missing  
 60 values were linearly interpolated; remaining missing values were replaced by a random number  
 61 between 0 and half of the lowest observed abundance ([Hampton \*et al.\*, 2006](#)). Time series are plotted  
 62 in Fig. S2. We tested extensively this and other methods to deal with missing data in a previous  
 63 publication on a subset of this dataset ([Barraquand \*et al.\*, 2018](#)). All time series were scaled and  
 64 centered before MAR analyses.

## 65 **MAR(1) model**

66 Multivariate autoregressive (MAR) models are used to determine the interspecific interactions and  
 67 abiotic effects shaping a community’s dynamics ([Ives \*et al.\*, 2003](#)). MAR(1) models are based on a  
 68 stochastic, discrete-time Gompertz equation which relates the log-abundance of each of the  $S$  taxa  
 69 at time  $t + 1$  to log-abundances of the whole community at time  $t$ , with possible interactions between  
 70 taxa, and effects of  $V$  abiotic variables at time  $t + 1$ . These assumptions are encapsulated in eq. 1:

$$\mathbf{n}_{t+1} = \mathbf{B}\mathbf{n}_t + \mathbf{C}\mathbf{u}_{t+1} + \mathbf{e}_t, \mathbf{e}_t \sim \mathcal{N}_S(0, \mathbf{Q}) \quad (1)$$

71 where  $\mathbf{n}_t$  is the  $1 \times S$  vector of log-abundance of phytoplankton taxa,  $\mathbf{B}$  is the  $S \times S$  community  
 72 (interaction) matrix,  $\mathbf{C}$  is the  $S \times V$  environment matrix describing the effects of  $V$  variables (stacked  
 73 in vector  $\mathbf{u}_{t+1}$ ) on growth rates, with  $V = 2$  in our case (temperature and salinity). The noise  
 74  $\mathbf{e}_t$  is a  $1 \times S$  noise vector, following a multivariate normal distribution with a variance-covariance  
 75 matrix  $\mathbf{Q}$ .  $\mathbf{Q}$  is diagonal and we have previously showed that this parsimonious choice did not  
 76 affect qualitatively the results ([Barraquand \*et al.\*, 2018](#)). We used the MARSS package ([Holmes  
 77 \*et al.\*, 2014](#)) v3.9, in R v3.3.2 ([Venables & Smith, 2013](#)), to estimate parameters with a maximum

likelihood procedure.

Our previous analysis of the Arcachon region, for which more covariables were available (Barraquand *et al.*, 2018), revealed that hydrodynamics and hydrology had more influence on phytoplankton dynamics than nutrients on the two-week timescale. Because temperature and salinity sum up seasonal changes in light as well as hydrology (salinity is inversely related to freshwater inflow), these represent the two key drivers needed to account for abiotic influences (Scheef *et al.*, 2013). They are therefore used to summarize the abiotic environment in the remainder of the article.

The analysis of real data in Barraquand *et al.* (2018) was complemented by that of simulated data mimicking the study design, which confirmed the ability of MAR(1) models to infer biotic interactions and abiotic forcings. Fitting a more sophisticated model (threshold autoregressive model) did not reveal extra non-linearities or a storage effect in the Arcachon subset of the data (Barraquand *et al.*, 2018). Other aspects of the MAR(1) modelling are likewise quite robust: using two abiotic variables (temperature and salinity) in this study rather than the full set used in Barraquand *et al.* (2018) led to almost identical covariate effects and interaction estimates for the Arcachon study sites. Even if some departures from the true data-generating model may not always be detectable through MAR(1) diagnostics (e.g., residuals), the analysis of nonlinear simulations has showed that MAR(1) models are in general robust to nonlinearities if the inference focuses on interaction sign and order of magnitude of model coefficients (Certain *et al.*, 2018), which is how these models are used here. For ease of interpretation of MAR(1) interaction coefficients, we also highlight how intra- and inter-taxa interaction strengths in a MAR(1) model map to their counterparts in a multispecies Beverton-Holt model, i.e., a discrete-time Lotka-Volterra model (Cushing *et al.*, 2004), in the Supporting Information.

In this study, the number of phytoplankton taxa ( $S$ ) and the community composition vary slightly between regions but sites share on average 67% of their taxa. In order to have comparable models, we also keep the same 2 covariates, i.e., water temperature and salinity, that were measured at all study sites. Therefore, the dimension of the dynamical system depends on the (square of the) number of phytoplankton taxa we study, which ranges between 7 (Mediterranean Sea) and 14 (Brittany). The smallest system still requires 63 parameters to be estimated (49 for the  $7 \times 7$

106 interaction matrices and 14 for the  $7 \times 2$  environment matrices) if we consider all possible interactions  
 107 between taxa. To reduce this dimensionality and remove unnecessary parameters, we built different  
 108 ‘interaction scenarios’ based on known phylogenetic information (as suggested in [Violle \*et al.\*, 2011](#);  
 109 [Narwani \*et al.\*, 2017](#)). The null interaction scenario assumed no interaction between genera (diagonal  
 110 interaction matrix) and was compared to four other interaction scenarios. The first interaction  
 111 scenario assumed that interactions could only occur between phylogenetically close organisms, i.e.,  
 112 within a class (groups were then diatoms, dinoflagellates, and other phytoplanktonic organisms)  
 113 while the second interaction scenario further differentiated pennate and centric diatoms. The third  
 114 interaction scenario considered the reverse hypothesis, that only unrelated organisms could interact  
 115 (i.e., a diatom could only interact with a dinoflagellate or a cryptophyte, but not with another  
 116 diatom), and the last interaction scenario did not constrain the interactions at all (full interaction  
 117 matrix). We selected the best scenario by comparing BIC (Fig. S3), which proved to be satisfactory  
 118 in our previous analyses of both real data and similar simulated datasets ([Barraquand \*et al.\*, 2018](#),  
 119 Appendix 2). The second interaction scenario, hereafter called the pennate-centric scenario, had the  
 120 lowest BIC for all sites (Fig. S3). This parsimonious scenario was therefore chosen as the basis for  
 121 further investigations of network structure.

## 122 **Analysis of interaction strengths**

123 The interaction matrix obtained from MAR(1) analyses can be used to determine the stability  
 124 of a discrete-time dynamical system ([Ives \*et al.\*, 1999, 2003](#)). To investigate stability-complexity  
 125 relationships, we compared the maximum modulus of the eigenvalues of the pennate/centric matrices  
 126 for each site to network descriptors. The maximum modulus is analogous to the real part of the  
 127 leading eigenvalue for continuous time models, and measures resilience while still accounting for  
 128 some variability properties ([Ives \*et al.\*, 1999](#)). However, because most theory on stability-complexity  
 129 has been developed in continuous time (e.g., [Allesina & Tang, 2015](#)), we numerically checked that the  
 130 maximum modulus of the eigenvalues in a discrete-time interaction matrix and its continuous-time  
 131 model counterpart yield similar information in the Supporting Information. We then compared  
 132 this resilience measure to complexity metrics, such as the interaction strength distribution (sign,



mean and variance) and weighted connectance (Bersier *et al.*, 2002). Weighted connectance is a measure of the proportion of realized links compared to all possible links, taking into account the shape of the flux distribution. This metric is adapted to weighted interaction matrices but cannot accommodate for both positive and negative coefficients: we therefore chose to focus on interaction strength only (absolute values of the coefficients), irrespective of interaction sign. In contrast, mean and variance of the off-diagonal coefficients, which can affect the stability of a community (Allesina & Tang, 2015), are computed on raw values of the coefficients. Interaction coefficient variance is multiplied by the number of taxa, according to theory (Allesina & Tang, 2015).

In addition to these network-level metrics, we also computed the average vulnerability (average effect of other taxa on a focal taxon, eq. S5) and impact (average effect of a focal taxon on other taxa, eq. S6) on both raw and absolute values of the coefficients. Vulnerability and impact can be related to in-strength and out-strength in the meta-analysis of Kinlock (2019). We then compared these to the regulation a focal species exerted on itself. Raw values indicate the average effect that can be expected on the growth rate of a taxon from the rest of the community (i.e., is the effect of others mostly positive or negative?), while absolute effects characterise the strength of all types of interactions on a taxon (i.e., is a taxon strongly affected by the others?).

Finally, we compared the observed ratio between mean self-regulation (intrataxon interaction strength) and mean intertaxa interaction strength to other published studies based on a MAR(1) model. A list of references is given in Table S4. Authors usually reported only coefficients that were significant with a 5% significance level, thus ignoring potentially many weak effects, which we had to set to 0. There are therefore two ways of computing the mean intertaxa interactions, i.e., taking the mean value of all coefficients outside of the matrix diagonal, including zeroes (which decreases the estimated mean intertaxa interaction strength, Fig. 4), or taking the mean value of statistically significant intertaxa coefficients only (which increases the estimated mean intertaxa interaction strength, Fig. S9). We considered both; a detailed description of these different ways to compare intra- and inter-taxa interactions can be found in the Supporting Information.

## Results

### Interaction estimates

Using MAR(1) autoregressive models, we produced interaction matrices (Ives *et al.*, 2003; Hampton *et al.*, 2013) – i.e., Jacobian community matrices on the logarithmic abundance scale (Ives *et al.*, 2003). Best-fitting models corresponded to a phylogenetically-structured interaction scenario, where interactions only occurred between closely related genera (Fig. S3). This led to sparse, modular matrices that have two main features. First, we observed a strong self-regulation for all sites (Fig. 1, diagonal elements of all matrices), a feature that we had previously highlighted in a more detailed analysis on one of the considered study regions (Barraquand *et al.*, 2018). The ratio of mean intragenus to intergenus interaction coefficients varied between 6 and 10, not counting coefficients set to 0 before the estimation process. When we included the zeroes in the interaction matrix in the computation of the intra/inter mean interaction strength (see the Supporting Information for details of that computation), the ratio rose to 21-43. Therefore, intragenus interactions were on average one order of magnitude stronger than intergenus interactions.

Second, although the percentage of facilitative interactions varied among sites (between 40% and 71% of interactions in the selected models), facilitation remained predominant in 9 sites out of 10 (only Lazaret, in the Mediterranean Sea, has 60% negative interactions). Our observational setup being nested, with sites within regions, we could examine whether locally positive interactions remain positive in a regional context: the percentage of consistently positive interactions at the regional level varied between 30% and 53%, higher than the percentage of similarly defined negative interactions (between 15% and 40%), except for sites in the Mediterranean Sea.

We found that the percentage of true mutualism (+/+) was substantial: averaged over all sites, 32% of all interactions were (+/+) while only 12% of them were (-/-), see also Fig. S5. The sign correspondence was not always maintained between regions: the only interaction that was non-zero in the 10 sites (CHA/SKE) was mutualistic in Men er Roue only (Brittany) and mixed (+/-) in all other sites. Within the same region, however, interactions measured in different sites tended to keep the same sign. In the 3 sites of Oléron, for instance, there were 4 interactions which remained

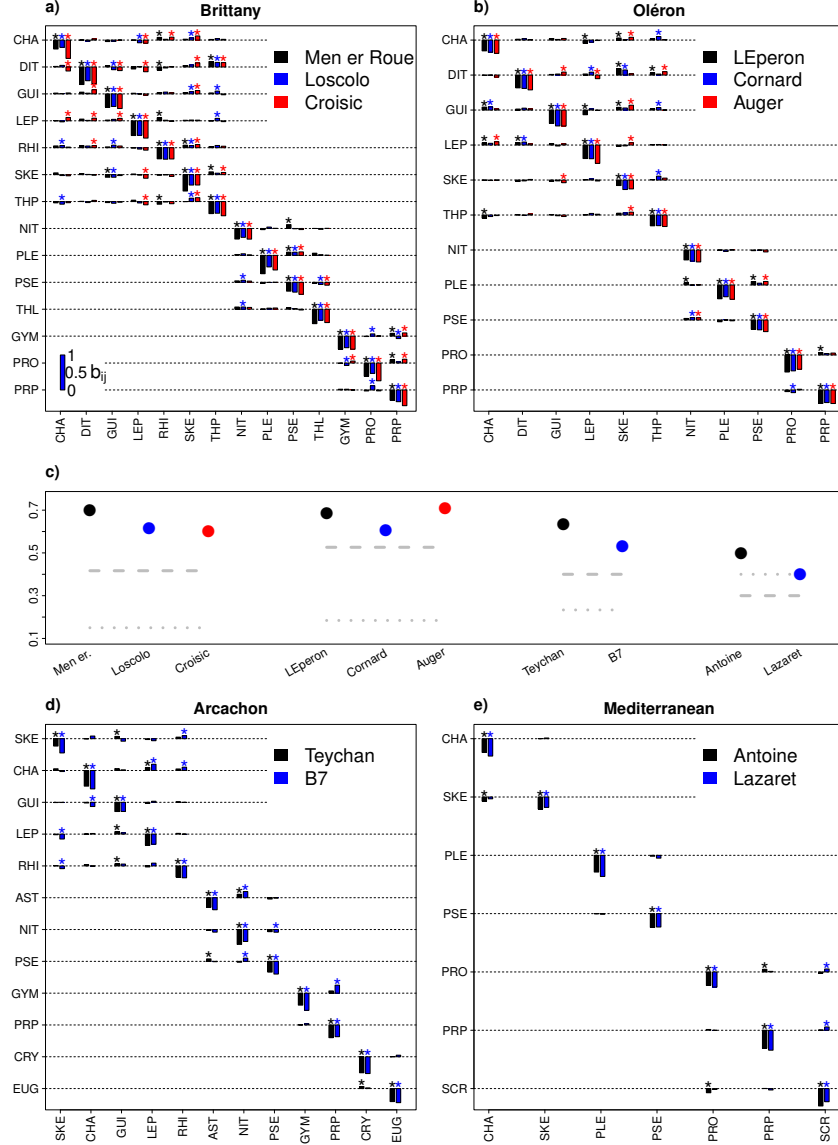


Figure 1: **Interaction matrices estimated in 10 sites along the French coastline.** Taxon  $j$  (in columns) has an effect on taxon  $i$ 's growth rate (in rows) proportional to the bar height, which corresponds to the  $\mathbf{B} - \mathbf{I}$  matrix (community composition in Table S2, most parsimonious interaction scenario presented). The scale for the coefficient values is given at the bottom left of panel a). Coefficients significantly different from 0 ( $\alpha = 5\%$ ) are marked by asterisks (\*). The fraction of positive interactions in each matrix is given by points in c) while the dashed (resp., dotted) line represents the ratio of interactions remaining positive (resp., negative) for all sites of a given region.

186 positive for both taxa involved (CHA/GUI, DIT/GUI, LEP/THP, SKE/THP), 3 of them being also  
187 mutualistic in some of the Brittany sites. This contradicts previous observations that mutualistic  
188 interactions tend to be more context-dependent than competitive interactions (Chamberlain *et al.*,  
189 2014).

## 190 Interaction network analysis

191 The stability (*sensu* resilience, Ives & Carpenter 2007) of all interaction matrices was not strongly  
192 affected by the percentage of positive interactions or the mean and variance of the intergenus  
193 interactions (Fig. 2). There was a slight increase in stability with weighted connectance, with a drop  
194 in eigenvalue modulus for weighted connectances between 0.09 and 0.1. The maximum modulus of  
195 the interaction matrix eigenvalues remained between 0.65 and 0.80.

196 Given that a direct complexity-stability (*sensu* resilience) link was not obvious, we investigated  
197 whether the matrix coefficients had some particular structure that could help theoretical ecology to  
198 make better null models of joint community dynamics and interactions (James *et al.*, 2015). We  
199 defined two scores, vulnerability (summed effect of others on the focal taxon growth rate, eq. S5)  
200 and impact (summed effect of the focal taxon onto other taxa's growth rates, eq. S6). Relations  
201 between inter- and intra-genus interactions emerged (Fig. 3): genera that were more self-regulating  
202 also had also a higher vulnerability score and a lower impact score. Those two influences are likely  
203 to trade-off: a high degree of self-regulation somehow buffers the effect of outside influences on  
204 population dynamics. Taxa that were less self-regulating were also more likely to have a stronger  
205 effect onto other taxa. As these genera tended to be more abundant (Fig. S7), this could be mediated  
206 by the average density of a genus. It is important to note, however, that these trends are weak and  
207 there is therefore a considerable amount of randomness dominating the interaction matrix: many  
208 scenarios of self-regulation vs limitation by others are therefore possible.

209 Aside from the trade-offs of Fig. 3, we found no remarkable patterns of covariation between  
210 matrix elements (Fig. S5) other than a mean-variance scaling of interaction coefficients (Fig. S6).

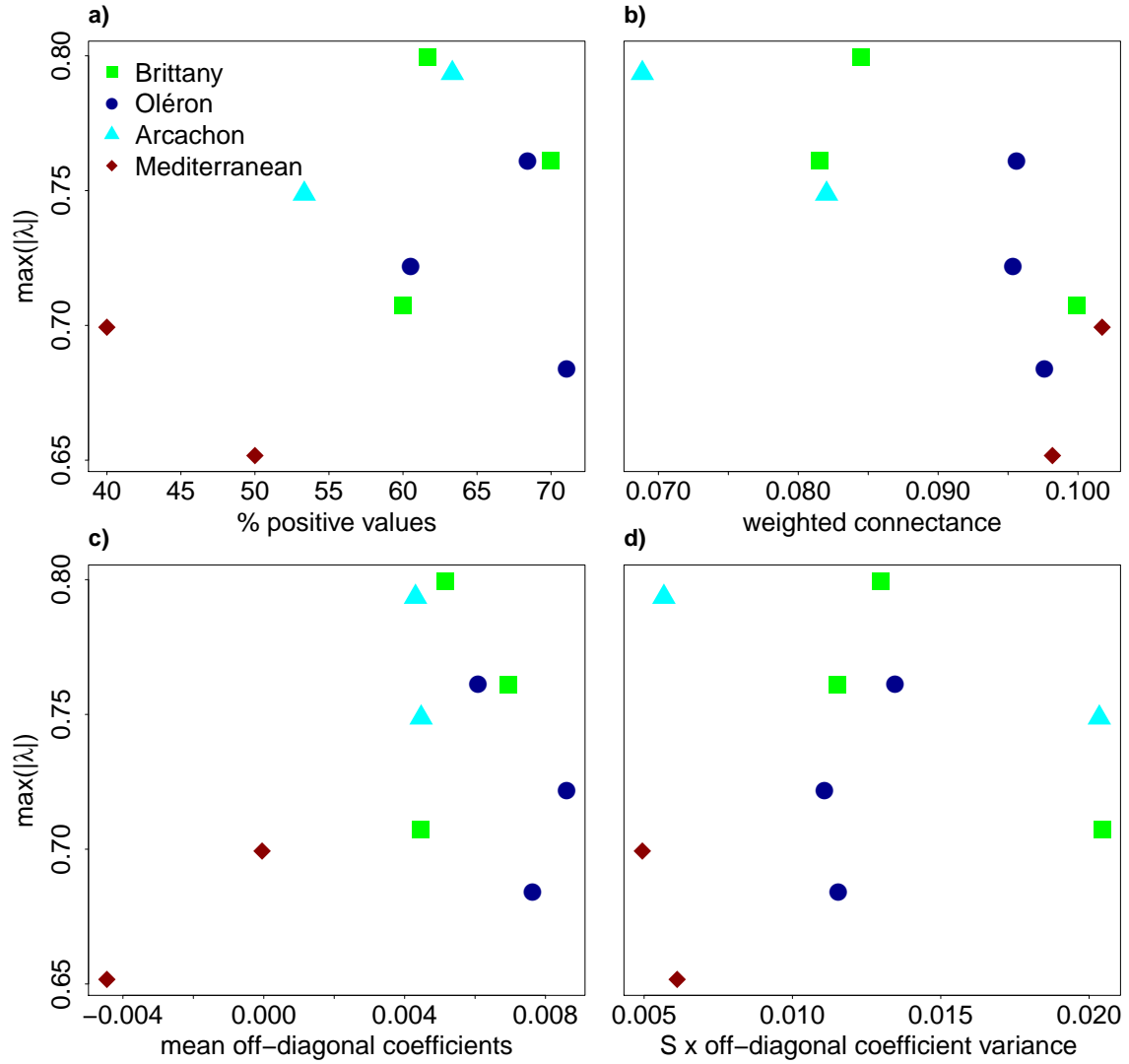


Figure 2: **Relation between stability and complexity of the interaction networks.** The maximum modulus of the eigenvalues of the interaction matrix  $\mathbf{B}$  indicates stability *sensu* resilience. Off-diagonal coefficient variance is multiplied by the dimension of the network, that is the number of species in the region. Each color or shape corresponds to a given region. The formula for weighted connectance is given in the Supporting Information.

## Literature comparison

Finally, we sought to put these results in a broader context by compiling the intra vs inter group estimates of previous MAR(1) studies of long-term observational count data (listed in Table S4).



Figure 3: **Relation between vulnerability/impact and self-regulation.** Average vulnerability (effects of others on the focal taxon growth rate, a-b) and impact (effects of the focal taxon on others' growth rates, c-d), as well as self-regulation, are computed for untransformed (a-c) or absolute (b-d) values of the coefficients of the interaction matrix ( $\mathbf{B} - \mathbf{I}$ ) for the 10 study sites. Each color corresponds to a given region (Fig S1). Linear regressions are shown as black lines.

214 We found that the order of magnitude of intra/inter interaction strengths considered here is not  
 215 particularly above those found for most planktonic systems to which MAR(1) models have been

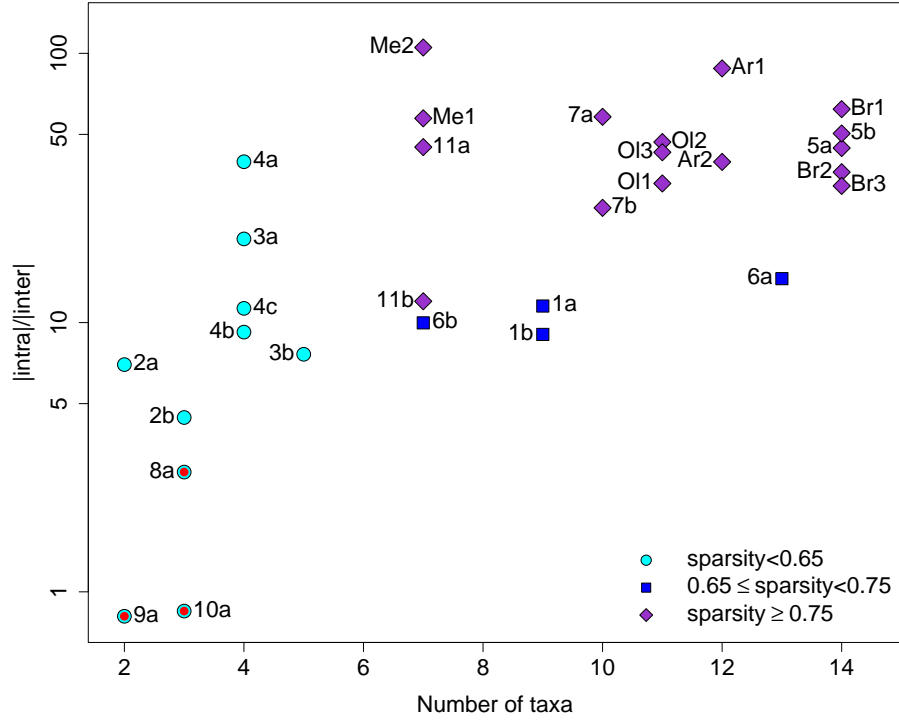


Figure 4: **Ratio of intra- to inter-group interaction strength in Multivariate AutoRegressive (MAR) models.** The reference for each study is given in Table S4. Codes beginning with letters correspond to the present study (Ar: Arcachon; Ol: Oléron; Br: Brittany; Me: Mediterranean Sea). The symbol color and shape correspond to the sparsity of the interaction matrix (e.g., the proportion of null interactions in the matrix). Red dots correspond to terrestrial and/or low dimension predator-prey systems, giving a lower bound for the intra/inter ratio. Intergroup interactions were set to 0 when they were not specified in the articles (in most cases, authors removed non-significant interactions at the 5% level; Fig. S9 is the same figure taking into account only significant interactions)

216 fitted, considering that our systems are relatively high-dimensional and that the higher the number  
 217 of taxa, the larger the intraspecific regulation (Barabás *et al.*, 2017). We included in Fig. 4 not only  
 218 plankton studies but also a couple of vertebrate or insect studies on less diverse communities, where  
 219 interactions are stronger, in order to provide lower bounds for the intra/inter ratio. The conclusion  
 220 from this comparison seems to be that, unlike small communities that can be tight-knit, any diverse  
 221 field system of competitors and facilitators has evolved large niche differences making intragroup  
 222 competition much larger in magnitude than intergroup interactions.

## Discussion

### Strong self-regulation and facilitation

We found very large niche differences between genera, translating into much higher intragenus than intergenus effects on growth rates (i.e., strong self-regulation), together with a high degree of facilitative net interactions.

The intra/intertaxa interaction strength ratio (Levine & HilleRisLambers, 2009) that we found, from 6–10 to above 20, depending on whether one includes interactions set to zero before the estimation process, could appear very high in light of previous intra/interspecific competition strength estimates of 4 to 5 by Adler *et al.* (2018b). Additional estimates using the unconstrained interaction matrix yielded ratios between 8 and 11 depending on the site (Table S3 and Fig. S8 in the Supporting Information), but weak intertaxa effects are likely to be inflated in the full model. Therefore, a intra/inter ratio of 10 seems like a conservative estimate, twice that of Adler *et al.* (2018b) who use a different model, i.e., a Lotka-Volterra competition model. We outline how to relate a MAR(1) model to a discrete-time Lotka-Volterra equivalent in the Supporting Information; even though there is a relationship between intra/inter ratios in both models, the relationship is not trivial when abundances vary greatly between species. Hence, to some degree, intra/inter ratios can differ between model frameworks or ways of measuring density-dependencies (e.g., a high measurement error due to using proxies of densities for plants can result in bias in interaction coefficient estimates, Detto *et al.*, 2019). However, a ratio intra/inter at least twice larger than the ones previously found may call for other explanations. One could also argue that our high intra/inter ratio arises because we consider the genus as our baseline taxonomic unit, rather than the species. It is logical that niche differentiation increases as one gets up the phylogenetic tree, and that getting down to the species level could slightly decrease that ratio (but see Narwani *et al.*, 2017, in which phylogenetic closeness decreases competition strength). However, taxonomic resolution is unlikely to be the sole explanation for the high intra/inter ratio of interaction strength found here, for two reasons. First, phytoplankton species belonging to different genera are often found to compete in experiments (Titman, 1976; Tilman *et al.*, 1982; Descamps-Julien & Gonzalez, 2005). In the field-based dataset



250 studied here, the same genera that are considered in experiments are found not to compete (or only  
 251 weakly), hence there must be some niche differentiation occurring in the field but not in the lab.  
 252 Second, the only other study that managed to provide MAR(1) estimates down to the species level  
 253 for phytoplankton, that of [Huber & Gaedke \(2006\)](#), provides an intra/interspecific strength ratio  
 254 similar to ours (point 7a in Fig. 4). Strong self-regulation seems therefore a genuine feature of field  
 255 phytoplanktonic communities. We discuss below possible mechanistic interpretations.

256 Another main finding of our study is the large frequency of positive interactions, with 30% truly  
 257 mutualistic (+/+) interactions and between 40 and 70% facilitative effects. Although a seasonal  
 258 environment can generate some positive covariation between taxa, those effects have already been  
 259 filtered out by the inclusion of our 2 abiotic covariates (Fig. S4). The facilitative effects shown  
 260 here are therefore residual effects, once abiotic trends are accounted for. Between 40 and 70%  
 261 facilitation can be compared to the meta-analysis by [Adler \*et al.\* \(2018b\)](#) who also found facilitative  
 262 interactions, but less than here ( $\approx 30\%$ ). However, [Adler \*et al.\* \(2018b\)](#)'s review contains many  
 263 experiments while the plant literature is replete with field examples of facilitation ([Brooker \*et al.\*,  
 264 2008](#); [McIntire & Fajardo, 2014](#)), so that plant facilitation could be higher in the field. At the  
 265 moment, it is therefore unknown how the predominance of facilitative interactions that we found  
 266 in phytoplankton compares to facilitation in terrestrial plants. We note that several authors using  
 267 MAR(1) models previously forbade positive interactions within the same trophic level, so that the  
 268 fraction of facilitative interactions in plankton cannot be computed from literature-derived MAR(1)  
 269 estimates.

270 The large niche differences and facilitative interactions that arise when considering a single  
 271 trophic level are an emergent property, resulting from hidden effects of resource or predator  
 272 partitioning/sharing ([Chesson, 2018](#)). In our previous publication investigating in detail the Arcachon  
 273 study sites ([Barraquand \*et al.\*, 2018](#)), we have argued that for phytoplankton, the strong intragroup  
 274 density-dependence could arise from effects of natural enemies ([Haydon, 1994](#)). Natural enemies  
 275 could also very well create apparent mutualism between prey species ([Abrams \*et al.\*, 1998](#); [de Ruiter  
 276 & Gaedke, 2017](#)). We believe this to be likely for the present study, given that the study regions  
 277 (Arcachon, Oléron, Brittany, Mediterranean) have similar predators (zooplankton, e.g., [Jamet \*et al.\*,](#)

278 2001; Modéran *et al.*, 2010; Tortajada *et al.*, 2012) and parasites (viruses, e.g., Ory *et al.*, 2010;  
 279 fungi). Though natural enemies are good candidates to explain the observed niche differences and  
 280 emerging facilitation, one must bear in mind that other known drivers of phytoplankton dynamics  
 281 such as allelopathy (Felpeto *et al.*, 2018), auxotrophy (Tang *et al.*, 2010) or hydrodynamics (Lévy  
 282 *et al.*, 2018) can all, in theory, help create different niches and an emerging facilitation (see last  
 283 subsection of the Discussion). Finally, resources that are usually considered limiting for all species  
 284 might in fact not always be: Burson *et al.* (2018) show that phytoplanktonic taxa specialize on  
 285 different components of the light spectrum. This constitutes an example of fine-scale resource  
 286 partitioning of one resource, light, that all species and genera are usually thought to compete for.

## 287 **No complexity-stability relationship but connections between self-regulation** 288 **and intergroup interactions**

289 There was no relation between the complexity of the communities (measured as either the weighted  
 290 connectance or the interaction coefficient variance) and their stability (measured by the largest  
 291 modulus of the eigenvalues, which quantifies the return time to a point equilibrium, i.e., resilience).  
 292 This result is conditional upon our model being a good approximate description of the system  
 293 (i.e., no multiyear limit cycles or chaotic attractors as the mapping between eigenvalues and actual  
 294 stability is distorted in that case, Certain *et al.*, 2018). However, we already showed on a subset of  
 295 this data that a fixed point in a MAR(1) model, perturbed by seasonality and abiotic variables, is an  
 296 accurate description of the system (Barraquand *et al.*, 2018). Therefore, we are confident that the  
 297 absence of complexity-resilience relationship found here is not a mere artefact of an inadequate model.  
 298 This absence of direct link between complexity and stability could be an actual feature of empirical  
 299 systems, as shown previously by Jacquet *et al.* (2016) using a different technique. This result seems to  
 300 contradict theory based on random matrices, especially for competitive and/or mutualistic networks  
 301 (Allesina & Tang, 2012). However, one must bear in mind that such result could also be generated  
 302 by the limited size of our networks, as random matrix theory relies on asymptotics (Allesina &  
 303 Tang, 2015). We should also mention that our interaction matrices (based on a discrete-time model)  
 304 are not strictly analogous to the ones used most frequently in theoretical ecology (continuous-time

model), though the spectral radius (largest modulus) is here tightly related to the real part of the lead eigenvalue in equivalent continuous-time models (see Supporting Information). Thus while the jury is still out regarding the absence of complexity-resilience relation found here, it may well be a genuine absence. In addition to complexity metrics, we also found that the percentage of mutualistic interactions, that is thought to affect the stability of a network, either positively or negatively (Mougi & Kondoh, 2012; Coyte *et al.*, 2015; García-Callejas *et al.*, 2018), does not in fact have a major impact on our networks' resilience.

In addition to weighted connectance and interaction variance, indices at the genus level (vulnerability and impact) approximate the average effects exerted and sustained by any given taxa in the different study sites. While, at the network level, network structure (either complexity measures or the percentage of mutualistic interactions) did not affect resilience, a relation emerged between self-regulation, necessary for coexistence, and genus-level indices. We found that the more a genus is self-regulated, the more it tends to be vulnerable to other genera's impacts and the less it impacts other genera. We examined whether vulnerability and impact could be affected by phylogenetic correlations; they were not, as on Fig. 3, points were not clustered according to genus, family or phylum. High self-regulation usually indicates large niche differences with the rest of the community, and it makes therefore sense that a species/genus whose needs strongly differ from the others only marginally impacts the resources of the other coexisting species. This is what we expect under strong niche partitioning. A low self-regulation was also correlated with high average abundance, which echoes findings by Yenni *et al.* (2017) who demonstrated that rare species usually show stronger self-regulation. This correlation between relative rarity and self-regulation could explain the lesser impact of highly self-regulated species/genus: a taxon which dominates the community composition can have a major effect on the others, especially as they usually cover more space, while it is harder for the less common taxa to have large impacts. In contrast, it was more difficult to explain the relationship between self-regulation and vulnerability: a genus that is more self-regulated and less common was found here to be on average more vulnerable to other genera's increases in densities. Such relation implies greater stability (*sensu* resilience, Ives *et al.* 2003, and also invariability, Arnoldi *et al.* 2019) for the network as a whole, because the taxa that are the more vulnerable

333 to other taxa’s impacts are also those whose dynamics are intrinsically more buffered. By which  
 334 mechanisms this could happen is so far unclear and open to speculation. It could just be a “mass  
 335 effect”: common taxa are in high enough numbers to deplete resources or change the environment in  
 336 ways that affect the less common ones, but the reverse is not true. As a final note on relationships  
 337 between interaction matrix coefficients, we caution that the trends evidenced are all relatively weak:  
 338 considerable stochasticity still dominates the distribution of interaction matrix coefficients.

## 339 Ghosts of competition past and present

340 Overall, the dominance of niche differentiation in observational plankton studies – based on our  
 341 analysis of the REPHY dataset and re-analysis of the MAR(1) literature – is similar to what has been  
 342 recently found in plant community studies (Volkov *et al.*, 2007; Adler *et al.*, 2018b) or empirically  
 343 parameterized food webs including horizontal diversity (Barabás *et al.*, 2017). Large niche differences  
 344 might be due to the ghost of competition past, i.e., competition has occurred in the past, leading to  
 345 strong selection and subsequent evolution, and then to progressive niche separation. In this scenario,  
 346 species have evolved niches that allow them not to compete or to interact only weakly (very strong  
 347 facilitative effects might be likewise destabilizing, Coyte *et al.*, 2015). The likely predator effects that  
 348 we highlighted above could be comprised within such niche differentiation *sensu largo*: specialized  
 349 predators can make strong conspecific density-dependence emerge (Bagchi *et al.*, 2014; Comita *et al.*,  
 350 2014), while switching generalists can also promote diversity (Vallina *et al.*, 2014). Both predators  
 351 and resources have often symmetrical effects and can therefore contribute almost equally to such  
 352 past niche differentiation (Chesson, 2018).

353 An intriguing new possibility, dubbed the “ghost of competition present” (Tuck *et al.*, 2018),  
 354 suggests by contrast that spatial distributions in relation to abiotic factors might have a large  
 355 impact on the interaction strengths inferred from temporal interaction models such as ours. Recent  
 356 combinations of model fitting and removal experiments have shown that model fitting usually  
 357 underestimates the effect of competitors that are uncovered by removal experiments (Tuck *et al.*,  
 358 2018; Adler *et al.*, 2018a). This could occur for instance if species are spatially segregated (at a small  
 359 scale) because each species only exists within a domain where it is relatively competitive (Pacala’s

spatial segregation hypothesis, chapter 15 in [Pacala & Levin 1997](#)), while a focal species could spread out if competitors were removed. This means that a species can be limited by competitors, but act so as to minimize competition (a little like avoidance behaviour in animals) and maximize opportunities for positive interactions, which implies that competition is in effect hard to detect when all species are present. This mechanism would require some spatial segregation between phytoplankton species at the scale of interactions, i.e., at the microscale. At the moment, while it is known that fine-scale hydrodynamics generate inhomogeneities at the microscale ([Barton \*et al.\*, 2014](#); [Breier \*et al.\*, 2018](#)) it is yet quite unclear how they might affect multivariate spatial patterns of species distributions (*sensu* [Bolker & Pacala 1999](#) or [Murrell & Law 2003](#)). Moreover, even with some microscale spatial segregation between species, a “ghost of competition present” mechanism might not work in phytoplankton as in terrestrial plants, because the turbulent, ever-changing aquatic environment imposes additional constraints on the spatial distribution of organisms.

## Acknowledgements

This study was only made possible by the dedication of all members of the REPHY program by Ifremer ([REPHY, 2017](#)), providing invaluable data through years of fieldwork. We are grateful to David Murrell for his careful reading and suggestions, and to Peter Adler for helpful exchanges. We also want to thank Bérangère Péquin for her constructive comments and Gyuri Barabás for thoughtful suggestions that improved the connection made to continuous-time stability theory. This study was supported by the French ANR through LabEx COTE (ANR-10-LABX-45).

## Supporting Information

This article contains supporting information.

## Authors’ contributions

CP and FB contributed equally to the project design and the methodology. The computer code was written by CP. The authors jointly interpreted the results and co-wrote the manuscript after an early draft by FB.

## 386 Data accessibility

387 The REPHY dataset has already been published (REPHY, 2017) and all scripts for MAR models  
 388 and subsequent network analyses are available online in a GitHub repository ([https://github.](https://github.com/CoraliePicoche/REPHY-littoral)  
 389 [com/CoraliePicoche/REPHY-littoral](https://github.com/CoraliePicoche/REPHY-littoral)).

## References

- Abrams, P., Holt, R. & Roth, J. (1998). Apparent competition or apparent mutualism? Shared predation when populations cycle. *Ecology*, **79**, 201–212.
- Adler, P., Ellner, S. & Levine, J. (2010). Coexistence of perennial plants: an embarrassment of niches. *Ecology Letters*, **13**, 1019–1029.
- Adler, P., Kleinhesselink, A., Hooker, G., Taylor, J., Teller, B. & Ellner, S. (2018a). Weak interspecific interactions in a sagebrush steppe? Conflicting evidence from observations and experiments. *Ecology*, **99**, 1621–1632.
- Adler, P., Smull, D., Beard, K., Choi, R., Furniss, T., Kulmatiski, A., Meiners, J., Tredennick, A. & Veblen, K. (2018b). Competition and coexistence in plant communities: intraspecific competition is stronger than interspecific competition. *Ecology Letters*, **21**, 1319–1329.
- Allesina, S. & Tang, S. (2012). Stability criteria for complex ecosystems. *Nature*, **483**, 205–208.
- Allesina, S. & Tang, S. (2015). The stability-complexity relationship at age 40: a random matrix perspective. *Population Ecology*, **57**, 63–75.
- Armstrong, R. & McGehee, R. (1976). Coexistence of species competing for shared resources. *Theoretical Population Biology*, **9**, 317–328.
- Armstrong, R. & McGehee, R. (1980). Competitive exclusion. *The American Naturalist*, **115**, 151–170.

- Arnoldi, J.F., Loreau, M. & Haegeman, B. (2019). The inherent multidimensionality of temporal variability: how common and rare species shape stability patterns. *Ecology Letters*, **22**, 1557–1567.
- Bagchi, R., Gallery, R., Gripenberg, S., Gurr, S., Narayan, L., Addis, C., Freckleton, R. & Lewis, O. (2014). Pathogens and insect herbivores drive rainforest plant diversity and composition. *Nature*, **506**, 85–88.
- Barabás, G., Michalska-Smith, M. & Allesina, S. (2017). Self-regulation and the stability of large ecological networks. *Nature Ecology & Evolution*, **1**, 1870–1875.
- Barraquand, F., Picoche, C., Maurer, D., Carassou, L. & Auby, I. (2018). Coastal phytoplankton community dynamics and coexistence driven by intragroup density-dependence, light and hydrodynamics. *Oikos*, **127**, 1834–1852.
- Barton, A., Ward, B., Williams, R. & Follows, M. (2014). The impact of fine-scale turbulence on phytoplankton community structure: Phytoplankton and turbulence. *Limnology and Oceanography: Fluids and Environments*, **4**, 34–49.
- Bersier, L.F., Banasek-Richter, C. & Cattin, M.F. (2002). Quantitative descriptors of food-web matrices. *Ecology*, **83**, 2394.
- Bolker, B. & Pacala, S. (1999). Spatial moment equations for plant competition: understanding spatial strategies and the advantages of short dispersal. *The American Naturalist*, **153**, 575–602.
- Breier, R., Lalescu, C., Waas, D., Wilczek, M. & Mazza, M. (2018). Emergence of phytoplankton patchiness at small scales in mild turbulence. *Proceedings of the National Academy of Sciences*, **115**, 12112–12117.
- Brooker, R., Maestre, F., Callaway, R., Lortie, C., Cavieres, L., Kunstler, G., Liancourt, P., Tielbörger, K., Travis, J., Anthelme, F., Armas, C., Coll, L., Corcket, E., Delzon, S., Forey, E., Kikvidze, Z., Olofsson, J., Pugnaire, F., Quiroz, C., Saccone, P., Schiffrers, K., Seifan, M., Touzard, B. & Michalet, R. (2008). Facilitation in plant communities: the past, the present, and the future. *Journal of Ecology*, **96**, 18–34.

- Burson, A., Stomp, M., Greenwell, E., Grosse, J. & Huisman, J. (2018). Competition for nutrients and light: testing advances in resource competition with a natural phytoplankton community. *Ecology*, **99**, 1108–1118.
- Cavieres, L. & Badano, E. (2009). Do facilitative interactions increase species richness at the entire community level? *Journal of Ecology*, **97**, 1181–1191.
- Certain, G., Barraquand, F. & Gårdmark, A. (2018). How do MAR(1) models cope with hidden nonlinearities in ecological dynamics? *Methods in Ecology and Evolution*, **9**, 1975–1995.
- Chamberlain, S., Bronstein, J. & Rudgers, J. (2014). How context dependent are species interactions? *Ecology Letters*, **17**, 881–890.
- Chesson, P. (2018). Updates on mechanisms of maintenance of species diversity. *Journal of Ecology*, **106**, 1773–1794.
- Chesson, P. & Huntly, N. (1997). The roles of harsh and fluctuating conditions in the dynamics of ecological communities. *The American Naturalist*, **150**, 519–553.
- Comita, L., Queenborough, S., Murphy, S., Eck, J., Xu, K., Krishnadas, M., Beckman, N. & Zhu, Y. (2014). Testing predictions of the Janzen-Connell hypothesis: a meta-analysis of experimental evidence for distance- and density-dependent seed and seedling survival. *Journal of Ecology*, **102**, 845–856.
- Coyte, K., Schluter, J. & Foster, K. (2015). The ecology of the microbiome: Networks, competition, and stability. *Science*, **350**, 663–666.
- Cushing, J., Levarge, S., Chitnis, N. & Henson, S.M. (2004). Some discrete competition models and the competitive exclusion principle. *Journal of Difference Equations and Applications*, **10**, 1139–1151.
- Descamps-Julien, B. & Gonzalez, A. (2005). Stable coexistence in a fluctuating environment: an experimental demonstration. *Ecology*, **86**, 2815–2824.



- Detto, M., Visser, M., Wright, S. & Pacala, S. (2019). Bias in the detection of negative density dependence in plant communities. *Ecology Letters*, **22**, 1923–1939.
- Felpeto, A., Roy, S. & Vasconcelos, V. (2018). Allelopathy prevents competitive exclusion and promotes phytoplankton biodiversity. *Oikos*, **127**, 85–98.
- García-Callejas, D., Molowny-Horas, R. & Araújo, M. (2018). The effect of multiple biotic interaction types on species persistence. *Ecology*, **99**, 2327–2337.
- García-Callejas, D., Molowny-Horas, R. & Araújo, M. (2018). Multiple interactions networks: towards more realistic descriptions of the web of life. *Oikos*, **127**, 5–22.
- Griffiths, J., Hajdu, S., Downing, A., Hjerne, O., Larsson, U. & Winder, M. (2015). Phytoplankton community interactions and environmental sensitivity in coastal and offshore habitats. *Oikos*, **125**, 1134–1143.
- Gross, K. (2008). Positive interactions among competitors can produce species-rich communities. *Ecology Letters*, **11**, 929–936.
- Hampton, S., Holmes, E., Scheef, L., Scheuerell, M., Katz, S., Pendleton, D. & Ward, E. (2013). Quantifying effects of abiotic and biotic drivers on community dynamics with multivariate autoregressive (MAR) models. *Ecology*, **94**, 2663–2669.
- Hampton, S., Izmet’Eva, L., Moore, M., Katz, S., Dennis, B. & Silow, E. (2008). Sixty years of environmental change in the world’s largest freshwater lake - Lake Baikal, Siberia. *Global Change Biology*, **14**, 1947–1958.
- Hampton, S., Scheuerell, M. & Schindler, D. (2006). Coalescence in the lake washington story: Interaction strengths in a planktonic food web. *Limnology and Oceanography*, **51**, 2042–2051.
- Haydon, D. (1994). Pivotal assumptions determining the relationship between stability and complexity: an analytical synthesis of the stability-complexity debate. *The American Naturalist*, **144**, 14–29.

- Hernández Fariñas, T., Bacher, C., Soudant, D., Belin, C. & Barillé, L. (2015). Assessing phytoplankton realized niches using a French national phytoplankton monitoring network. *Estuarine, Coastal and Shelf Science*, **159**, 15–27.
- Holmes, E.E., Ward, E.J. & Scheuerell, M.D. (2014). Analysis of multivariate time-series using the MARSS package. *User guide: <http://cran.r-project.org/web/packages/MARSS/vignettes/UserGuide.pdf>*.
- Hubbell, S. (2001). *The unified neutral theory of biodiversity and biogeography*. Princeton University Press.
- Huber, V. & Gaedke, U. (2006). The role of predation for seasonal variability patterns among phytoplankton and ciliates. *Oikos*, **114**, 265–276.
- Huisman, J. & Weissing, F. (2001). Biological conditions for oscillations and chaos generated by multispecies competition. *Ecology*, **82**, 2682–2695.
- Hutchinson, G. (1961). The paradox of the plankton. *The American Naturalist*, **95**, 137–145.
- Ives, A., Gross, K. & Klug, J. (1999). Stability and Variability in Competitive Communities. *Science*, **286**, 542–544.
- Ives, A.R. & Carpenter, S.R. (2007). Stability and diversity of ecosystems. *Science*, **317**, 58–62.
- Ives, A.R., Dennis, B., Cottingham, K.L. & Carpenter, S.R. (2003). Estimating community stability and ecological interactions from time-series data. *Ecological Monographs*, **73**, 301–330.
- Jacquet, C., Moritz, C., Morissette, L., Legagneux, P., Massol, F., Archambault, P. & Gravel, D. (2016). No complexity-stability relationship in empirical ecosystems. *Nature Communications*, **7**.
- James, A., Plank, M., Rossberg, A., Beecham, J., Emmerson, M. & Pitchford, J. (2015). Constructing random matrices to represent real ecosystems. *The American Naturalist*, **185**, 680–692.
- Jamet, J.L., Boge, G., Richard, S., Geneys, C. & Jamet, D. (2001). The zooplankton community in bays of Toulon area (northwest Mediterranean Sea, France). *Hydrobiologia*, **557**, 155–165.

- Kinlock, N.L. (2019). A Meta-analysis of Plant Interaction Networks Reveals Competitive Hierarchies as well as Facilitation and Intransitivity. *The American Naturalist*, **194**, 640–653.
- Levine, J. & HilleRisLambers, J. (2009). The importance of niches for the maintenance of species diversity. *Nature*, **461**, 254–257.
- Lévy, M., Franks, P. & Smith, K. (2018). The role of submesoscale currents in structuring marine ecosystems. *Nature Communications*, **9**, 4758.
- Li, L. & Chesson, P. (2016). The effects of dynamical rates on species coexistence in a variable environment: the paradox of the plankton revisited. *The American Naturalist*, **188**, E46–E58.
- McIntire, E. & Fajardo, A. (2014). Facilitation as a ubiquitous driver of biodiversity. *New Phytologist*, **201**, 403–416.
- Modéran, J., Bouvais, P., David, V., Le Noc, S., Simon-Bouhet, B., Niquil, N., Miramand, P. & Fichet, D. (2010). Zooplankton community structure in a highly turbid environment (Charente estuary, France): Spatio-temporal patterns and environmental control. *Estuarine, Coastal and Shelf Science*, **88**, 219–232.
- Mougi, A. & Kondoh, M. (2012). Diversity of interaction types and ecological community stability. *Science*, **337**, 349–351.
- Murrell, D. & Law, R. (2003). Heteromyopia and the spatial coexistence of similar competitors. *Ecology Letters*, **6**, 48–59.
- Mutshinda, C.M., O’Hara, R.B. & Woiwod, I.P. (2009). What drives community dynamics? *Proceedings of the Royal Society B: Biological Sciences*, **276**, 2923–2929.
- Narwani, A., Bentlage, B., Alexandrou, M., Fritschie, K., Delwiche, C., Oakley, T. & Cardinale, B. (2017). Ecological interactions and coexistence are predicted by gene expression similarity in freshwater green algae. *Journal of Ecology*, **105**, 580–591.
- Ory, P., Hartmann, H., Jude, F., Dupuy, C., Del Amo, Y., Catala, P., Mornet, F., Huet, V., Jan, B., Vincent, D., Sautour, B. & Montanié, H. (2010). Pelagic food web patterns: do they modulate

- virus and nanoflagellate effects on picoplankton during the phytoplankton spring bloom? Effects of viruses and nanoflagellates on picoplankton. *Environmental Microbiology*, pp. 2755–2772.
- Pacala, S. & Levin, S. (1997). *Spatial ecology: the role of space in population dynamics and interspecific interactions*. Princeton University Press, Princeton, NJ.
- REPHY (2017). *REPHY dataset - French Observation and Monitoring program for Phytoplankton and Hydrology in coastal waters. 1987-2016 Metropolitan data*. <https://www.seanoe.org/data/00361/47248/>.
- Rosindell, J., Hubbell, S. & Etienne, R. (2011). The unified neutral theory of biodiversity and biogeography at age ten. *Trends in Ecology & Evolution*, **26**, 340.
- de Ruiter, P. & Gaedke, U. (2017). Emergent facilitation promotes biological diversity in pelagic food webs. *Food Webs*, **10**, 15–21.
- Scheef, L., Hampton, S. & Izmet'Eva, L. (2013). Inferring plankton community structure from marine and freshwater long-term data using multivariate autoregressive models. *Limnology and Oceanography: Methods*, **11**, 475–484.
- Tang, Y.Z., Koch, F. & Gobler, C.J. (2010). Most harmful algal bloom species are vitamin B1 and B12 auxotrophs. *Proceedings of the National Academy of Sciences*, **107**, 20756–20761.
- Tilman, D., Kilham, S. & Kilham, P. (1982). Phytoplankton community ecology: the role of limiting nutrients. *Annual Review of Ecology and Systematics*, **13**, 349–372.
- Titman, D. (1976). Ecological competition between algae: experimental confirmation of resource-based competition theory. *Science*, **192**, 463–465.
- Tortajada, S., Niquil, N., Blanchet, H., Grami, B., Montanié, H., David, V., Glé, C., Saint-Béat, B., Johnson, G., Marquis, E., Del Amo, Y., Dubois, S., Vincent, D., Dupuy, C., Jude, F., Hartmann, H. & Sautour, B. (2012). Network analysis of the planktonic food web during the spring bloom in a semi enclosed lagoon (Arcachon, SW France). *Acta Oecologica*, **40**, 40–50.

- Tuck, S., Porter, J., Rees, M. & Turnbull, L. (2018). Strong responses from weakly interacting species. *Ecology Letters*, **21**, 1845–1852.
- Utermöhl, H. (1958). Zur Vervollkommnung der quantitativen Phytoplankton-Methodik. *Mitt. int. Ver. theor. angew. Limnol.*, **9**.
- Vallina, S., Ward, B., Dutkiewicz, S. & Follows, M. (2014). Maximal feeding with active prey-switching: A kill-the-winner functional response and its effect on global diversity and biogeography. *Progress in Oceanography*, **120**, 93–109.
- Venables, W.N. & Smith, D.M. (2013). the R Core Team. *An Introduction to R. Notes on R: A Programming Environment for Data Analysis and Graphics*. URL: [ran. r-project. org/doc/manuals/r-release/R-intro. pdf](http://ran.r-project.org/doc/manuals/r-release/R-intro.pdf).
- Violle, C., Nemergut, D., Pu, Z. & Jiang, L. (2011). Phylogenetic limiting similarity and competitive exclusion. *Ecology Letters*, **14**, 782–787.
- Volkov, I., Banavar, J., Hubbell, S. & Maritan, A. (2003). Neutral theory and relative species abundance in ecology. *Nature*, **424**, 1035–1037.
- Volkov, I., Banavar, J., Hubbell, S. & Maritan, A. (2007). Patterns of relative species abundance in rainforests and coral reefs. *Nature*, **450**, 45–49.
- Yenni, G., Adler, P. & Ernest, S.K.M. (2017). Do persistent rare species experience stronger negative frequency dependence than common species? Asymmetric NFD and persistent rare species. *Global Ecology and Biogeography*, **26**, 513–523.