

How self-regulation, the storage effect and their interaction contribute to coexistence in stochastic and seasonal environments

Coralie Picoche^{*,1}, Frédéric Barraquand^{1,2}

November 9, 2018

¹ University of Bordeaux, Integrative and Theoretical Ecology, LabEx COTE, Bt. B2 - Alle Geoffroy St-Hilaire, 33615 Pessac, France

² CNRS, Institute of Mathematics of Bordeaux, 351 Cours de la Libération, 33405 Talence, France

* Corresponding author. Email: coralie.picoche@u-bordeaux.fr

Abstract

Explaining coexistence in species-rich communities of primary producers remains a challenge for ecologists because of the likely competition for shared resources. Following Hutchinson's seminal suggestion, many theoreticians have tried to create diversity through a fluctuating environment, which impairs or slows down competitive exclusion. There are now several fluctuating-environment models allowing coexistence, but they often produce only a dozen of coexisting species at best. Here, we investigate how to create even richer communities in fluctuating environments, using an empirically parameterized model. Building on the forced Lotka-Volterra model of Scranton and Vasseur (2016) inspired by phytoplankton communities, we have investigated the effect of two coexistence mechanisms, namely the storage effect and higher intra- than interspecific competition strengths (i.e., strong self-regulation). We tuned the competition ratio based on empirical analyses, in which self-regulation usually dominates interspecific interactions. Although a strong self-regulation maintained more species (50%) than the storage effect (25%), we show that none of the two coexistence mechanisms considered could, by itself, ensure the coexistence of all species present at the beginning of our simulations. Realistic seasonal environments only aggravated that picture, as they decreased persistence relative to a random environment. Our results suggest that combining different mechanisms for biodiversity maintenance into community models might be more fruitful than trying to find which mechanism explains best the observed diversity levels. We additionally highlight that while biomass-trait distributions provide some clues regarding coexistence mechanisms, they cannot indicate by themselves which coexistence mechanisms are at play.

Number of words: 242

Keywords: coexistence; seasonality; competition; phytoplankton; Lotka-Volterra; storage effect

27 Introduction

28 There has been a rich debate in theoretical ecology on how to reconcile niche and neutral
29 perspectives on species coexistence (Gravel et al., 2006; Carmel et al., 2017). Under the neutral
30 perspective, all species have equal birth and death rates and compete equally (since space is
31 limited) whilst under the niche perspective, birth and death rates can vary between species
32 and various mechanisms contribute to increasing intraspecific over interspecific competition
33 (Hubbell, 2001). However, as it has been pointed out repeatedly, niche and neutral processes are
34 not mutually exclusive: they may actually act together to produce observed species coexistence
35 (Gravel et al., 2006; Mutshinda et al., 2009; Götzenberger et al., 2012).

36 An intriguing offshoot of the niche vs neutrality debate is the concept of ‘clumpy coexis-
37 tence’ (Scheffer and van Nes, 2006), whereby simultaneous influences of both niche and neutral
38 processes create several clumps of similar species along a single trait axis. Classical stabiliz-
39 ing niche differences (SNDs), such as a net intraspecific competition stronger than interspecific
40 competition, enable coexistence of multiple clumps (Chesson, 2000), while within-clump coexis-
41 tence occurs through neutral processes (Hubbell, 2001; Munoz and Huneman, 2016), as species
42 that differ too little in their fitnesses cannot exclude themselves out on reasonable timeframes.
43 Indeed, clumps on the trait axis eventually thin out in absence of immigration, but transient
44 coexistence can last for extended periods of time. This ‘emergent neutrality’ within groups
45 (Holt, 2006) has been proposed as a unifying concept for the niche and neutral theories. The
46 findings of Scheffer and van Nes (2006) have been disputed due to hidden niches in the original
47 model (Barabás et al., 2013). Hidden niches emerge through stronger intraspecific competition
48 mediated by an additional predation-like term (Barabás et al., 2013). This makes coexistence
49 in the Scheffer and van Nes model more similar to that of the classical Lotka-Volterra model
50 (Barabás et al., 2016), so that coexistence within clumps is not exactly neutral. Still, the idea
51 that niche and neutral assembly can mould communities stays potent (Haegeman and Loreau,
52 2011; Vergnon et al., 2013). Since then, several studies have suggested that ‘clumpy coexis-
53 tence’ can occur in theoretical models, most notably models incorporating temporal variation
54 (Scranton and Vasseur, 2016; Sakavara et al., 2018). In these temporal-variation models, equal
55 competitive strengths are combined with other mechanisms like the storage effect (or tempo-
56 ral niche partitioning, that is an equivalent concept for forced Lotka-Volterra models, Barabás

et al., 2012; Scranton and Vasseur, 2016). The storage effect increases the possibility of coexistence by making the interaction strength covary positively with a fluctuating environmental quality (see also Barabás et al., 2012).

Here, we build on the work of Scranton and Vasseur (2016) to investigate the possibility of coexistence through species response to fluctuating environments. Our enthusiasm for the Scranton and Vasseur (2016) model stems from our interest in phytoplankton communities, that inspired their thermal preference curves modeling intrinsic growth rates. However, Scranton and Vasseur (2016) described daily temperature as a random noise, i.e., independent and identically distributed Gaussian random variates over time. This appeared to us a key assumption to relax. Under most latitudes, temperature is indeed a seasonal signal, and seasonal forcing can strongly affect the dynamics of the community considered (Vesipa and Ridolfi, 2017). Hence, over short timescales, random temporal variations often only add noise to a largely deterministic seasonal trend. Our present work can therefore be seen as an attempt to blend Scranton and Vasseur (2016)’s stochastic framework with the periodic environments of Barabás et al. (2012), to better represent the mixture of stochastic and deterministic environmental forces affecting phytoplankton community dynamics.

Because many phytoplankton species or genera respond in similar ways to temperature despite having different optimas (Moisan et al., 2002), we hypothesized that a large seasonal variation might not necessarily foster species coexistence. In fact, similar responses to seasonal fluctuations should lead to an increased synchrony of species abundances which, in turn, should theoretically mitigate the expected temporal partitioning. How seasonality affects coexistence (as opposed to a purely randomly fluctuating environment) is therefore a key feature of this paper. In particular, we contrast cases where the storage effect is present vs absent, which conveniently maps to two different parameterizations of the forced Lotka-Volterra model. Moreover, the overall diversity obtained at the end of the simulations with Scranton and Vasseur (2016) model was relatively low compared to what we usually observe in phytoplankton communities (several dozens to hundreds of species). We have therefore sought out which mechanisms would foster a truly species-rich community for extended periods of time.

In an empirical study combining phytoplankton community-level time series and multivariate autoregressive models (Barraquand et al., 2018), we found that despite a large influence of

the environment (including temperature, irradiance, and other factors), a strong intraspecific (or intragenus) competition, when compared to interspecific interaction coefficients, was most likely the key driver of species coexistence. In other words, strong self-regulation had a large role to play in maintaining species diversity in coastal phytoplankton (Barraquand et al., 2018). These high intraspecific interaction strengths mirror those found in a number of terrestrial plant communities (Adler et al., 2018) and in animal communities (Mutshinda et al., 2009).

Here, we therefore try to establish what are the relative contributions to coexistence of the storage effect vs strong self-regulation, in a phytoplankton-like theoretical community model. This led us to cross different combinations of seasonality in the forcing signal, presence of the storage effect or not, and intra- vs interspecific competition intensity, in order to disentangle the contributions of all these factors to biodiversity maintenance.

Methods

Model description

The model described in Scranton and Vasseur (2016) is based on the Lotka-Volterra competition model. Fluctuations in the environment are introduced in the model by temperature-dependent intrinsic growth rates (see Eq. 1-2, all coefficients are defined in Table 1) so that species growth rates can be expressed as:

$$\frac{dN_i}{dt} = r_i(\tau)N_i \left(1 - \sum_{j=1}^S \alpha_{ij}N_j \right) - mN_i \quad (1)$$

$$r_i(\tau) = a_r(\tau_0) e^{E_r \frac{(\tau - \tau_0)}{k\tau\tau_0}} f_i(\tau) \quad (2)$$

$$\text{where } f_i(\tau) = \begin{cases} e^{-|\tau - \tau_i^{opt}|^3/b_i}, & \tau \leq \tau_i^{opt} \\ e^{-5|\tau - \tau_i^{opt}|^3/b_i}, & \tau > \tau_i^{opt} \end{cases} \quad (3)$$

$$\text{and } b_i \text{ such as } \int r_i(\tau) d\tau = A \quad (4)$$

Model parameters are detailed in Table 1, and we set their values to match the features of phytoplankton communities as in Scranton and Vasseur's work (2016). The niche of each species i is defined by its thermal optimum τ_i^{opt} . Thermal performance curves defined in Eq. 3

are parameterized so that all species share the same niche area (Eq. 4), which sets a trade-off between maximum growth rates and niche width.

Table 1: Parameter definitions and values for the model described in Eqs. 1-4. Parameter values are not specified when they vary with time and/or the species considered.

Name	Definition	Value (unit)
S	Initial number of species	60 (NA)
N_i	Biomass density of the i^{th} species	(kg/area)
τ	Temperature	(K)
$r_i(\tau)$	Growth rate of species i as a function of temperature	$(\frac{\text{kg}}{\text{kg} \times \text{year}})$
α_{ij}	Strength of competition of species $j \rightarrow i$	0.001 (area/kg)
b_i	Normalization constant for the thermal decay rate	(K^3)
m	Mortality rate	$15(\frac{\text{kg}}{\text{kg} \times \text{year}})$
τ_0	Reference temperature	293 (K) / 20 ($^{\circ}\text{C}$)
$a_r(\tau_0)$	Growth rate at reference temperature	$386(\frac{\text{kg}}{\text{kg} \times \text{year}})$
E_r	Activation energy	0.467 (eV)
k	Boltzmann's constant	$8.6173324 \cdot 10^{-5} (\text{eV} \cdot \text{K}^{-1})$
$f_i(\tau)$	Fraction of the maximum rate achieved for the i^{th} species	(NA)
μ_τ	Mean temperature	293 (K)
σ_τ	Standard deviation for temperature	5 (K)
τ_{\min}	Minimum thermal optimum	288 (K)
τ_{\max}	Maximum thermal optimum	298 (K)
A	Niche breadth	$10^{3.1}(\frac{\text{kg}}{\text{kg} \times \text{year}})$
τ_i^{opt}	Thermal optimum for growth of the i^{th} species	(K)
θ	Scaling between random and seasonal noise	(0;1.3) (NA)
κ	Ratio of intra-to-interspecific competition strength	(1;10) (NA)

The original environmental forcing is a normally distributed variable centered on 293 K (20 $^{\circ}\text{C}$), with a 5 K dispersion. Temperature varies from one day to the next, but is kept constant throughout the day. At the monthly or annual temporal scale usually used in ecological studies, temperature could therefore be considered as a white noise (Vasseur and Yodzis, 2004). However, from a mathematical viewpoint, the noise is slightly autocorrelated as the integration process goes below the daily time step. We therefore use the expression ‘random noise’ to describe this forcing, as opposed to the ‘seasonal noise’ described hereafter. To construct the seasonal noise, we add to the random forcing signal a lower-frequency component, using a sinusoidal function with a period of 365 days (Eq. 5). We tune the ratio of low-to-high frequency with the variable θ so as to keep the same energy content - i.e., equal total variance - in the forcing signal.

$$\tau(t) = \mu_\tau + \theta \sigma_\tau \sin(2\pi t) + \epsilon_t, \text{ where } \epsilon_t \sim \mathcal{N}\left(0, \sigma_\tau \sqrt{1 - \frac{\theta^2}{2}}\right) \quad (5)$$

Note that the upper limit for θ , $\sqrt{2}$, corresponds to a completely deterministic model which we do not explore here (but see Zhao (1991) proving bounded coexistence). We choose to keep the stochasticity in the signal and to model a plausible temperature signal with $\theta = 1.3$ (illustrated in Fig. 1b) when considering a seasonal forcing of the dynamics.

The formulation of the forced Lotka-Volterra model of Scranton and Vasseur (2016) implies a storage effect, as the net effect of competition exerted by species j on i is the product of the temperature-related growth rate $r_i(\tau)$ and the competitive strength α_{ij} exerted by species j multiplied by its abundance N_j . Therefore, total net competition ($\sum_{j=1}^S r_i(\tau)\alpha_{ij}N_j$) covaries positively with the growth rate values $r_i(\tau)$, which defines the storage effect (Chesson, 1994; Fox, 2013; Ellner et al., 2016). To remove the assumption of an explicit storage effect, we created another version of the model using the mean value of a species' growth rate (\bar{r}_i) to weight the interaction coefficients (see Eq. 6). The mean growth rate value was computed by first generating the temperature time series and then averaging all r_i over the corresponding sequences of τ values.

$$\frac{dN_i}{dt} = N_i \left(r_i(\tau) - \sum_{j=1}^S \bar{r}_i \alpha_{ij} N_j \right) - mN_i \quad (6)$$

Following Eq. 6, net competition remains unaffected by the environmental conditions, in contrast to intrinsic growth rates, while preserving the same average magnitude as in Eq. 1.

Strong self-regulation is ensured by the addition of the coefficient κ , which is the ratio of intra-to-interspecific competition strength. We can therefore re-write the interaction coefficients α_{ij} in Eq. 7

$$\alpha_{ij} = \alpha (1 + (\kappa - 1)\delta_{ij}) \quad (7)$$

where δ_{ij} is the Kronecker symbol, equal to 1 if $i = j$ and to 0 otherwise. The value of the parameter $\kappa = 10$ was chosen from analyses of phytoplanktonic data (Barraquand et al., 2018)¹. Hereafter, the expression “strong self-regulation” characterizes dynamics where the intraspecific competition strength is 10 times higher than the interspecific competition strength, as opposed to “equal competitive strengths” where intra- and interspecific competition strengths are equal.

¹The values of non-zero interspecific competition coefficients reported in Fig. 5 of Barraquand et al. (2018) are somewhat higher than here (and κ lower, closer to 4) because the best-fitting model actually set to zero some intergroup competition coefficients; using a model where all α_{ij} are non-zero leads in contrast to $\kappa = 10$.

In addition to two types of environmental forcings (random noise with $\theta = 0$, and seasonal noise with $\theta = 1.3$), we compare the results for four formulations of the model: with and without an explicit storage effect (Eq. 1 and Eq. 6, respectively); with strong self-regulation or equal intra- and interspecific competition strength ($\kappa = 10$ or $\kappa = 1$, respectively). These are summed up in Table 2.

Table 2: Growth rate of species i in the four formulations of the model we present.

$\frac{1}{N_i} \frac{dN_i}{dt} + m_i$	Storage effect	No storage effect
Strong self-regulation ($\kappa = 10$)	$r_i(\tau) \left(1 - \sum_{j=1}^S \alpha (1 + 9\delta_{ij}) N_j \right)$	$r_i(\tau) - \sum_{j=1}^S \bar{r}_i \alpha (1 + 9\delta_{ij}) N_j$
Equal competitive strengths ($\kappa = 1$)	$r_i(\tau) \left(1 - \sum_{j=1}^S \alpha N_j \right)$	$r_i(\tau) - \sum_{j=1}^S \bar{r}_i \alpha N_j$

Set-up

We replicate the ‘Species sorting’ experiment of Scranton and Vasseur (2016) to investigate how the structure of synthetic phytoplankton communities varies under the different scenarios we described above. We focused on the dynamics of a community initialized with 60 species with thermal optima uniformly spaced along the interval $[15^\circ\text{C}, 25^\circ\text{C}]$, and with the same initial density ($\frac{1}{\alpha S}$). Each simulation was run for 5000 years in 1-day intervals. When the density of a species dropped below 10^{-6} kg/area, it was considered extinct. We ran 100 simulations for each combination of parameters (type of environmental signal, storage effect and value of κ).

All simulations were run with Matlab’s ode45 algorithm, an explicit Runge-Kutta (4,5) integration scheme with an absolute error tolerance of 10^{-8} , and relative error tolerance of 10^{-3} . The code is available in a GitHub repository².

Results

Typical dynamics of the community following Eq. 1 (the model of Scranton and Vasseur, 2016), with both a purely Gaussian noise (original choice of Scranton and Vasseur, 2016; Fig. 1a) and a seasonal noise described in Eq. 5 (our variant, Fig. 1b), are shown in Fig. 1c and

²<https://github.com/CoraliePicoche/Seasonality>, will be made public upon acceptance or at the reviewer’s request and stored in Zenodo.

d, respectively. A sinusoidal forcing produces the strongly seasonally structured dynamics that are typical of phytoplankton. Even though only 5 species can be seen in Fig. 1c, there were 14 species still present at the end of the simulation forced by a random noise, with large disparities in the range of their biomasses. A third of the species kept a biomass above 10 kg/area (setting area = 1 ha, with a depth of a few meters, produces a realistic standing biomass; Reynolds, 2006) while 6 out of the 14 species biomasses remained below the unit. All persisting species in the random noise simulation were clustered within a 3.2°C-range of thermal optima (see the biomass distribution as a function of the thermal optimum in Supplementary Material, Fig. A1). No obvious temporal patterns (e.g., cycles) could be seen in the community forced by random noise. On the contrary, seasonal cycles were clear in the seasonally-forced case of Fig. 1d. Only 4 species coexisted at the end of the simulation with seasonal noise, gathered in two groups with large thermal optimum differences (5.7°C between the maximum thermal optimum of the first group and the minimum thermal optimum of the second group). When temperatures were high, the group with higher thermal optima reached its maximum biomass, then as temperature decreases through the season, these species leave room for the growth of the low-temperature group.

The decrease in persistence due to seasonality observed in Fig. 1 was confirmed in all our simulations (Fig. 2). In cases where final species richness varied from one simulation to another (namely, the two middle cases in Fig. 2: with storage effect but without strong self-regulation, or without storage effect but with strong self-regulation), seasonality reduced the number of extant species to, on average, 27% and 48% of their original values, respectively (Fig. 2). A seasonal signal therefore led to a much smaller average persistence. There was also less variance in persistence between seasonally forced simulations compared to random noise simulations.

Both a strong self-regulation and the storage effect markedly increased persistence. Without any of these coexistence mechanisms, only one species persisted at the end of the simulations. When only the storage effect was present, the number of extant species varied between 8 and 20 (14.8 ± 2.4) with random noise, or 2 and 6 (4.1 ± 0.7) with a seasonal signal. On the other hand, when only a strong self-regulation was present, the number of extant species nearly doubled, varying between 20 and 32 (27.5 ± 2.4), or 12 and 15 (13.3 ± 0.6), with a random or a seasonal noise, respectively. Remarkably, when the storage effect and a strong self-regulation

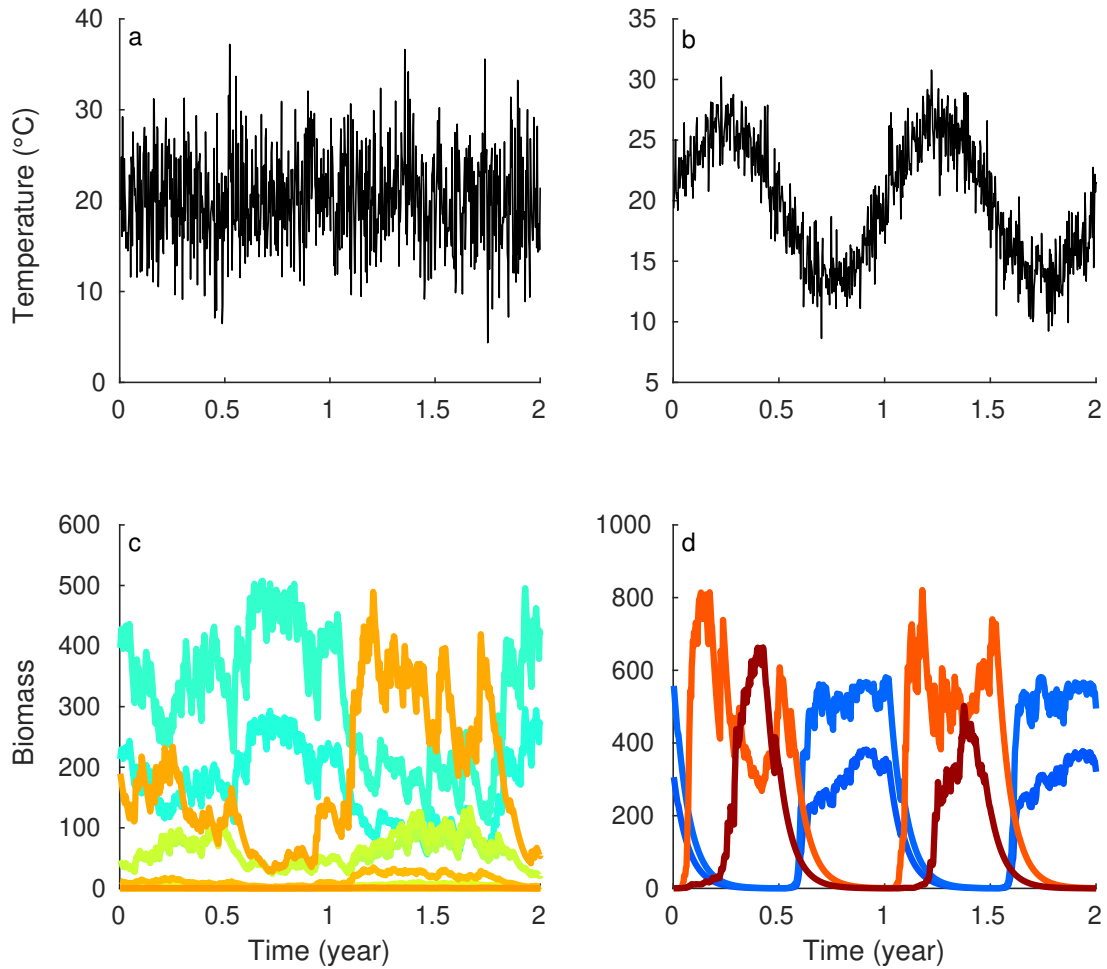


Figure 1: Time series of temperature (top; a-b) and extant species (bottom; c-d) for the 2 last years of a 5000-year simulation, with storage effect but no differences between intraspecific and interspecific competition strengths. The forcing temperature is either a random noise (a) or a seasonal noise (b), leading to community dynamics with more erratic fluctuations (c) vs seasonally structured fluctuations (d). Line colors of species biomasses correspond to their thermal optimum (from blue, corresponding to low thermal optimum, to red, corresponding to high thermal optimum).

both affected the community dynamics, all species persisted in the community, while neither of these mechanisms was able to produce that result alone, for either random or seasonal noise.

The biomass-trait distribution of the community was affected by the type of forcing even when the richness of the community was stable (Fig. 3). Without storage effect nor strong self-regulation, there was only one species left at the end of the simulations. A random noise favored species with intermediate thermal optima: the final species had a thermal optimum between 18.9°C and 21.4°C (only a fourth of the initial range of thermal optima) for two simulations out of three and the maximum final biomasses over 100 simulations was reached in this range (Fig. 3a). This distribution may indicate a selection for the highest long-term growth rates, averaged over time (see scaled growth rates in Fig. 3). Seasonality with no coexistence mechanisms also

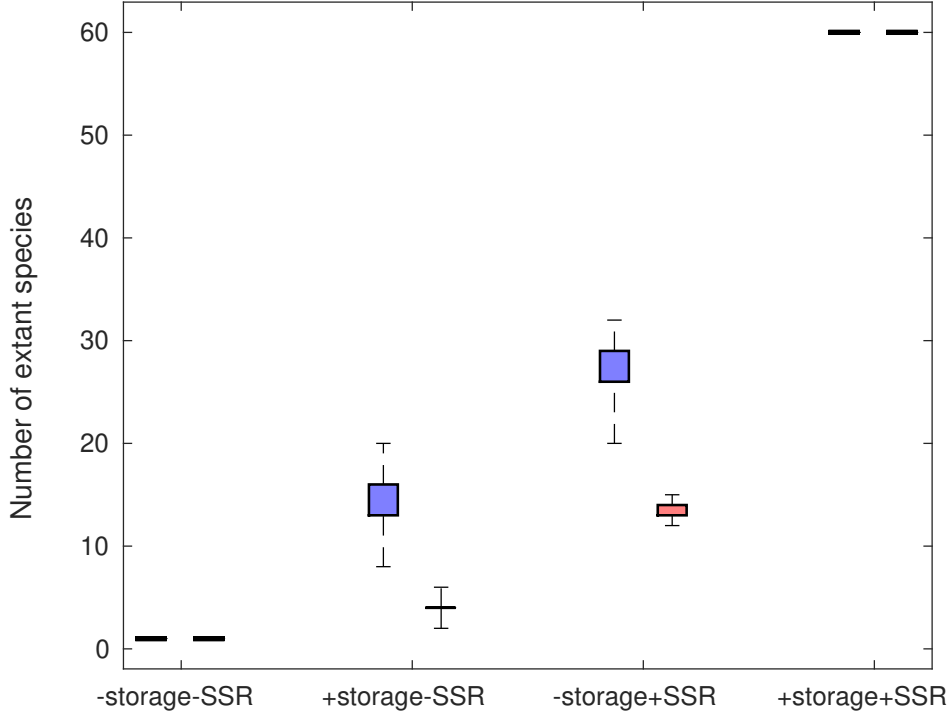


Figure 2: Number of species still present at the end of 100 simulations (5000 years each), initialized with 60 species, with a random forcing signal (blue) or a seasonal noise (red). The signs + or -storage refer to presence or absence of the storage effect, respectively; + / - SSR, presence or absence of Strong Self-Regulation, respectively. Community compositions are stable in the cases -storage-SSR and +storage+SSR, for which 1 or 60 species are still present at the end of all simulations, respectively. Due to low variance, the whiskers here represent min and max rather than 1.5 interquartile range.

led to a single final species but, in this case, the species always had a higher maximum growth rate (thermal optimum above 22°C). Species with a higher thermal optimum were more likely to persist and to reach a higher biomass at the end of the simulation. 38% of the simulations therefore ended with the species having the highest temperature optimum, 25°C. The shift in trait distribution towards higher maximum growth rates with a seasonal noise vs higher average growth rates with a random noise was consistent for all model types considered.

When both storage effect and strong self-regulation were present, the 60 initial species co-existed with small variations in biomasses for each species over the 100 simulations (mean CV=0.008 across simulations with either a random or a seasonal noise, Fig. 3b and d). The forcing signal modified only the distribution of biomasses resulting in contrasted community structures despite equal richness in both simulation types. With a random noise, the distribution was unimodal with a maximum biomass reached for the second highest long-term average growth rate (corresponding to a thermal optimum of 20.2°C). On the contrary, a seasonal signal

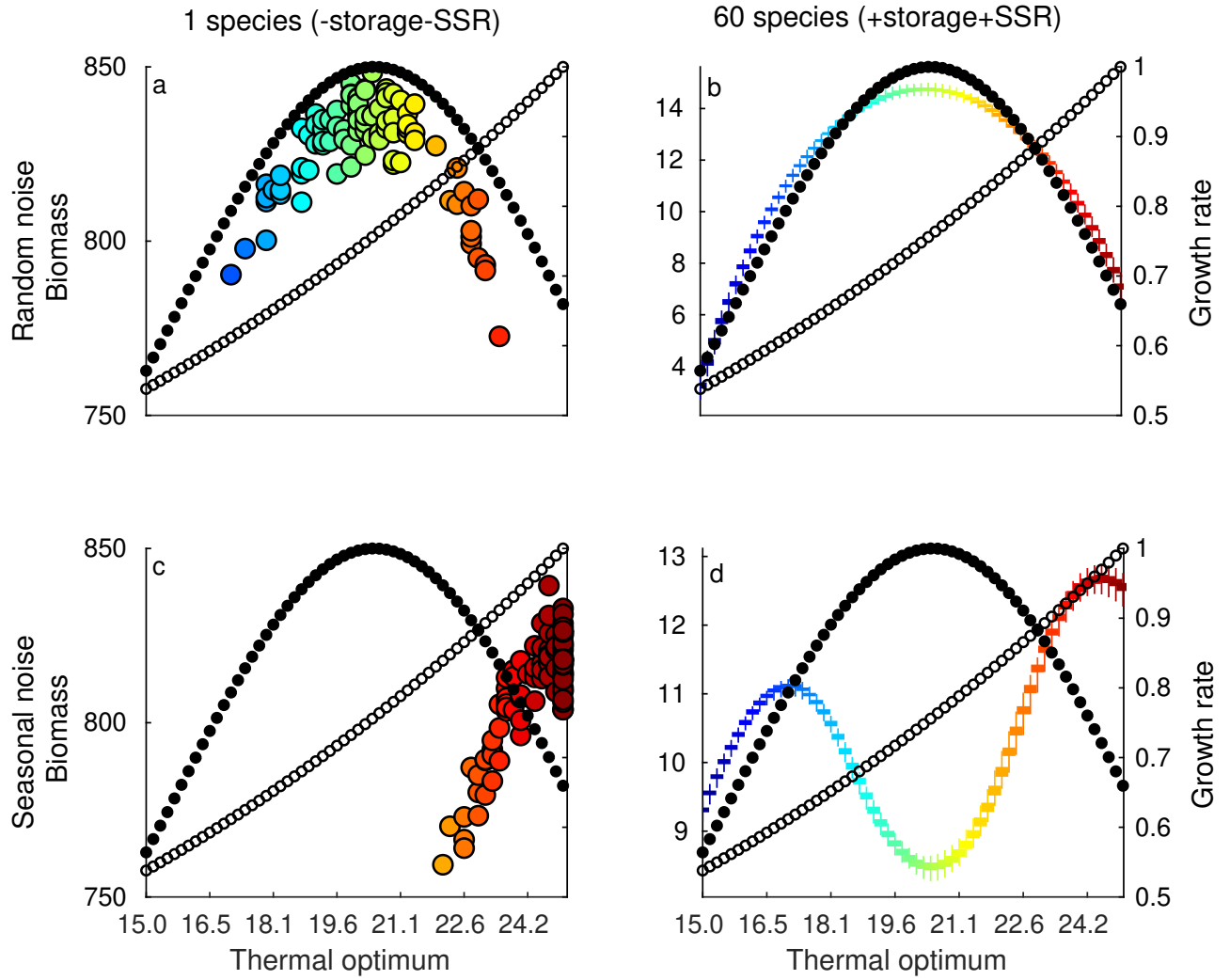


Figure 3: Mean biomass distribution over the last 200 years for 100 simulations, as a function of the species thermal optima. Here we consider the two stable-composition cases and two types of forcing signal. On the left column, simulations without storage effect nor strong self-regulation are presented. Only one species is present at the end of the simulations and its mean value is represented by one large colored circle per simulation. There can be several circles for the same species, corresponding to multiple simulations ending with this species alone. On the right column, simulations with storage effect and strong self-regulation are represented. All species are present at the end of the simulations and small boxplots present the variation in the temporal average of biomass with a given trait, for 100 simulations. The forcing signal is either a random (top) or a seasonal noise (bottom). Each species is identified by its thermal optimum through its color code. Scaled (divided by maximum) average and maximum growth rates are shown as small filled and open circles, respectively, and are indexed on the right y-axis.

led to a bimodal distribution (centered on 17.0°C and 24.4°C), each corresponding to one season, with highest biomasses for higher thermal optima (Fig. 3d). The minimum biomass was reached for the highest long-term average growth rate at an intermediate temperature (20.4°C).

In cases where the richness of the community varied, the overall shape (multimodal vs unimodal) of the marginal distribution of extant species with respect to the trait axis were similar for both types of environmental forcings (Fig. 4). By contrast, the type of coexistence

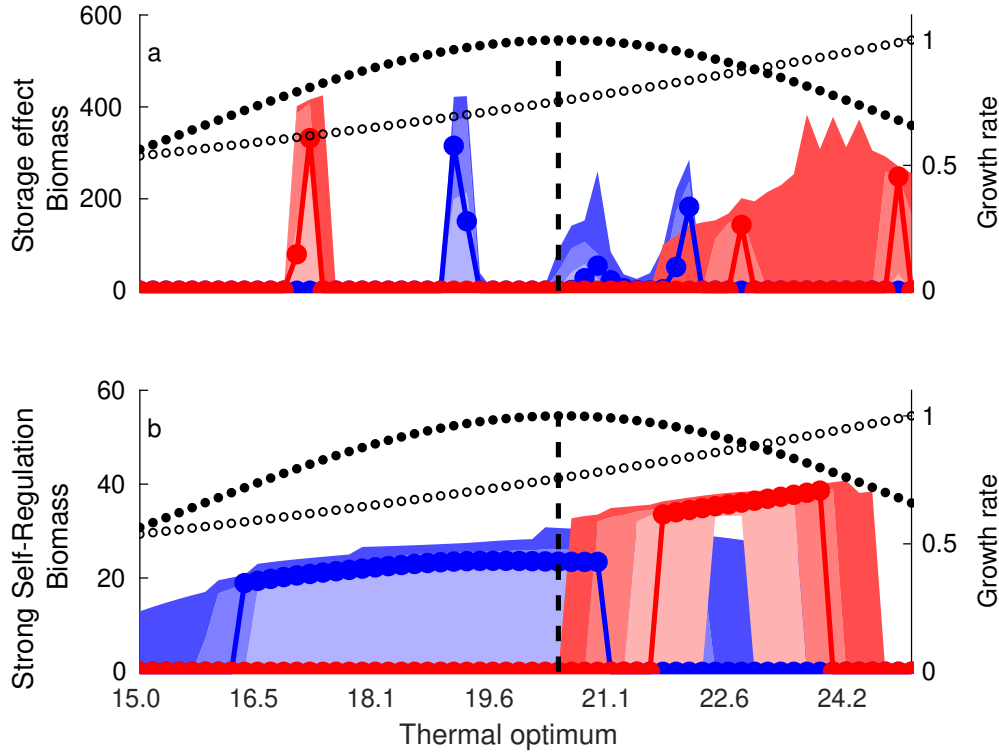


Figure 4: Mean biomass distribution over the last 200 years for 100 simulations, as a function of thermal optima, with storage effect and equal competitive strengths (a) and without storage effect, with strong self-regulation (b). The forcing signal is either a random (in blue) or a seasonal noise (in red). Shades of the same color correspond to the 50th, 90th and 100th percentiles of the distributions while colored lines correspond to one representative simulation. Scaled (divided by maximum) average (whose maximum is indicated by the dashed line) and maximum growth rates are shown as filled and open circles, respectively, and indexed on the right y-axis.

mechanism generated different shapes. Indeed, the storage effect (when acting alone) led to a multimodal biomass distribution with respect to thermal optima. We always observed 3 modes with a random noise and 3 modes in 95% of the seasonal simulations (Fig. 4a). With a random noise, extant species were grouped in rather similar clumps regarding species thermal optima (between 18.8°C and 22.2°C) whereas species tended to be further apart in the seasonal case, covering a total range of 7.7°C, with species grouping in the higher part of the thermal range, above 22°C. On the other hand, strong self-regulation led to a quasi-uniform biomass distribution (Fig. 4 b). Species in communities forced by a random noise stayed in the lower range of thermal optima (in 96% of the simulations, the highest thermal optimum was 22.4°C, see Fig. A2 in the Supplementary Material) while they were filtered out in communities subjected to a seasonal fluctuation of their environment, for which species with thermal optima above 20.5°C

235 persisted. As before (Fig. 3), seasonality promoted species with a higher maximum growth
236 rate since the autocorrelated temperatures enabled them to achieve this highest growth rate
237 for a longer period of time than a random noise would have.

238 Discussion

239 We have simulated competitive Lotka-Volterra dynamics forced by a fluctuating environment
240 (e.g., temperature fluctuations) under a range of scenarios allowing more or less coexistence.
241 Two coexistence mechanisms, the storage effect and strong self-regulation (i.e., intraspecific
242 competition much stronger than interspecific competition), could be either present or absent,
243 which led to four scenarios. These four scenarios were crossed with two possibilities for the forc-
244 ing signal, a random noise (mostly white) and a stochastic yet seasonal signal, both with equal
245 temporal variance. Our investigation therefore built on the model of Scranton and Vasseur
246 (2016), which included a random forcing and a storage effect, but considered seven additional
247 combinations of mechanisms. This was motivated by our wish to include two observed fea-
248 tures of phytoplankton dynamics: seasonal cycles (Winder and Cloern, 2010) and strong self-
249 regulation (Chesson, 2000; Adler et al., 2010; Barraquand et al., 2018). Many mechanisms can
250 lead to intraspecific competition being stronger than interspecific competition: nonlinearities in
251 the functional forms of competition or mutualism that contribute to increasing self-regulation
252 (Kawatsu and Kondoh, 2018), or predation as well as parasitism (see e.g., the generalist preda-
253 tors in Haydon, 1994). Strong self-regulation seems nonetheless an ubiquitous feature in com-
254 petition networks of primary producers (Adler et al., 2018), and perhaps even more general
255 networks (Barabás et al., 2017).

256 Before discussing the ecological interpretation of our results, we first recall some technical
257 assumptions made in this study. All our simulations lasted for a fixed duration (5000 time
258 steps) as in Scranton and Vasseur (2016). This means that short- and medium-term transients
259 (a few years to hundreds of years) were completely negligible at the end of the time series,
260 but very long transients could remain. We realized that convergence could be incomplete after
261 5000 years in some cases (e.g., random noise + storage effect + equal competitive strength).
262 Such simulations would take up to 15 000 years to converge and the rate of convergence would
263 slow over time, as can also be observed for similar models (Scheffer and van Nes, 2006). We

264 kept a fixed time integration window rather than waiting for convergence for both technical and
265 ecological reasons. From a technical standpoint, adding 10 000 years of numerical integration (or
266 more) for the sake of reaching equilibrium would have been very challenging computationally,
267 and comparison with the values reported by Scranton and Vasseur (2016) would have been
268 compromised. From an ecological standpoint, waiting for full convergence when there are
269 extremely long transients (Hastings et al., 2018) is also quite artificial: there is no reason to
270 believe that very long transients (i.e., transients that are maintained for thousands of years)
271 have any less ecological reality than an attractor that is deemed stable. Speed of convergence
272 is therefore an issue to judge whether transients should be considered or excluded, and a very
273 long yet fixed time window for integration allows advantageously to compare all mechanisms.

274 Another assumption pertains to competition coefficients. To allow for comparison with
275 Scranton and Vasseur (2016), we did not introduce variability in intraspecific competition
276 strength or interspecific competition strength. By contrast, data-based coefficients vary between
277 species (Barraquand et al., 2018), with a majority of weak interactions (as suggested in Wootton
278 and Emmerson, 2005) and more variance in intraspecific coefficients. Stump (2017) recently
279 considered the potential effects of competition coefficient variability (also called non-diffuse
280 competition), as did Kokkoris et al. (2002); more variance in interspecific competition strength
281 is usually detrimental to coexistence (see Stump (2017) for a classification of the various effects).
282 Setting the competition coefficients using a multidimensional trait-based framework, like that
283 of Ashby et al. (2017), would provide a natural development to the work presented here; it is in
284 our opinion difficult to speculate on those variance effects because both intra- and interspecific
285 competition coefficient variances may matter to community persistence.

286 Finally, our study is limited to communities whose species have fast population dynamics
287 relative to the yearly timescale, like phytoplankton and likely other fast-living organisms, so
288 that many generations can occur in a year. Different effects of seasonality may occur in species
289 that have slower life histories or with generations that extend over multiple years (e.g., multiyear
290 cycles and chaotic attractors, Rinaldi et al. 1993; Taylor et al. 2013; Tyson and Lutscher 2016).
291 Persistence may be affected differently by seasonality in such cases with slower community
292 dynamics.

293 With these assumptions in mind, we have found that first, temporally forced Lotka-Volterra

294 dynamics cannot sustain any diversity with our phytoplankton-based set of parameters, unless
 295 the structure is geared to include either the storage effect or a strong self-regulation. Although
 296 this absence of diversity-enhancing effect of “pure” environmental variation has already been
 297 stated by other authors (Chesson and Huntly, 1997; Barabás et al., 2012; Fox, 2013; Scranton
 298 and Vasseur, 2016), this is not always intuitive (Fox, 2013), so we feel compelled to stress it once
 299 more: temporal variation in growth rate alone cannot ensure coexistence within competitive
 300 communities. A nice point made by Scranton and Vasseur (2016) was that a built-in storage
 301 effect in a forced Lotka-Volterra model, parameterized for phytoplankton communities, could
 302 lead to some degree of coexistence. Our investigation reproduced these results, using the
 303 random noise considered by Scranton and Vasseur (2016). However, an arguably more lifelike
 304 noisy and seasonal temperature forcing considerably lessened the richness of the community
 305 after 5000 years, decreasing from 15 to 4 species on average. Even imagining that groups
 306 represented here are genera or classes rather than species, this is a fairly low diversity for a
 307 phytoplankton-like community (see e.g., Chapter 1 in Reynolds, 2006). This suggests that
 308 the storage effect may not, on its own, be sufficient to maintain species-rich communities
 309 (e.g., dozens to hundreds of species). We have therefore sought out whether a stronger self-
 310 regulation could maintain a higher diversity, using field-based intra- vs intergroup (species or
 311 genera) competition strength ratio (Barraquand et al., 2018), where the intragroup density-
 312 dependence was estimated 10 times stronger. Implementing such strong self-regulation, in the
 313 forced Lotka-Volterra models that we considered, produced a higher level of diversity than the
 314 storage effect (almost double). Of course, the result is somehow contingent upon the strength
 315 of self-regulation. Our estimates are a little stronger than what was found in perennial plants
 316 (Adler et al., 2010), where interspecific competition was suggested 4 or 5 times stronger than
 317 intraspecific. Still, the widespread effects of natural enemies in phytoplankton (zooplankton,
 318 parasites) may contribute to increasing the strength of self-regulation (Barraquand et al., 2018;
 319 Chesson, 2018) relative to other systems, hence we believe that 10 times stronger intraspecific
 320 competition constitutes a reasonable order of magnitude.

321 However, such strong self-regulation was still insufficient to maintain the whole community
 322 diversity (60 species) by itself, especially when the seasonal forcing (always decreasing species
 323 richness) was considered. The diversity within clumps of similar values of thermal optima

324 was considerably decreased once seasonality was implemented. This diversity reduction occurs
325 because within a season, the signal autocorrelation gives long, contiguous time intervals to the
326 best competitor to exclude its less adapted competitors. This makes the results likely to hold
327 not only for seasonal environments, but more generally for autocorrelated ones above the daily
328 scale, i.e., “red” noise. In contrast, the random noise scenario – which can be considered white
329 noise above the daily temporal scales – generates large temperature shifts more frequently, and
330 thereby forbids such competitive exclusion. In a seasonal setting, a species with the highest
331 long-term (arithmetically) averaged growth rate may not be the best competitor, and can
332 disappear as a result of a strong competition from both low- and high-temperature tolerant
333 species. This holds with or without a storage effect.

334 Our results may appear at odds with recent proposals that seasonal forcing in itself would
335 help maintain diversity (Sakavara et al., 2018). However, we compared the effect of seasonal
336 forcing to that of other forcing signals while controlling for total variance. Thus, the contrast
337 between our results and those of Sakavara et al. (2018) may be due to the role of forcing variance
338 over time (we compare scenarios under a constant total variance). Overall, while seasonality
339 may be slightly better than no forcing at all in maintaining diversity if the storage effect is
340 present, seasonal forcing of parameters does not improve coexistence when compared to white
341 noise.

342 In addition to community diversity, the biomass-trait relationship also varied from one
343 simulation to another. Some regularities did emerge across simulations though. The storage
344 effect begot several clumps along the trait space (as observed by Scranton and Vasseur, 2016).
345 The seasonality that we added to the temperature signal led to more distant clumps on the
346 trait axis, with less species per clump. Conversely, strong self-regulatory mechanisms alone led
347 to relatively uniform biomass distributions, with species forming a single large cluster, which
348 covered a fraction of the initial trait space. Therefore, the shape of the distribution was af-
349 fected by the coexistence mechanism at work while the average trait value was modified by
350 the type of environmental forcing, even though the mean value of the environmental signal
351 did not change. The biomass-trait distributions therefore constitute clues to interpret com-
352 munity dynamics (D’Andrea and Ostling, 2016; Loranger et al., 2018), although we certainly
353 recommend to interpret them with caution to avoid over-generalization. The identification

of multiple modes in biomass-trait relationships and SADs is relatively recent (Dornelas and Connolly, 2008; Matthews et al., 2014) and is a rare pattern in theoretical models (McGill et al., 2007). Barabás et al. (2013) convincingly argued that multimodality could arise from the demographic stochasticity of a single model run (with either self-regulation or neutrality, but without the clumpy coexistence emerging from the storage effect). However, our results are based on many model runs, for which either the storage effect alone or the storage effect + strong self-regulation in a seasonal context consistently produced multimodal distributions, while simulations without the storage effect always led to a single cluster along the trait axis. Our suggestion for empirical studies is as follows: if only one spatial location is observed, caution in interpreting multiple clumps on the trait axis is of course required, as Barabás et al. (2013) highlighted. However, with several locations - or in a theoretical context - one could average across locations to reproduce similar graphs to the ones produced here. Clumps in the trait axis when averaged across model runs/locations are therefore a signature of a coexistence induced by the storage effect, for the cases that we considered in the article. Of course, other mechanisms that we did not include in our models may produce similar patterns (Rael et al., 2018) or obfuscate these patterns – typically strong self-regulation weakens the clustering on the trait axis. We therefore view clustering on the trait axis (when averaged over several samples) as an interesting clue suggesting to look for a storage effect, rather than any definite proof that the storage effect is at work.

In our previous empirical study of coastal phytoplankton dynamics (Barraquand et al., 2018), we did not find any storage effect. This, however, does not mean that it could not be observed in other planktonic systems. Given the consequences of the storage effect for species richness and composition presented here, we are skeptical that the storage effect could by itself help explaining phytoplankton diversity. However, our results suggest that in phytoplankton-like seasonal environments, even though empirically-based self-regulation produce much more diversity than the storage effect when considered in isolation, the storage effect can help diversity maintenance when combined to other mechanisms. Indeed, the combination storage effect + strong self-regulation is non-additive: the cases where both self-regulation and the storage effect were present showed more diversity than generated by any mechanism on its own.

The above results suggest the very exciting idea that multiple coexistence mechanisms might

combine superadditively, thus helping us to better understand the astounding diversity of primary producers. This logic could, in principle, be extended to mechanisms that we have not considered here (e.g., spatial structure, specialized natural enemies, that could be as important here for plankton as they are for tropical trees, see Bagchi et al., 2014; Comita et al., 2014; Barraquand et al., 2018). Previous research has however demonstrated that generalist seed predation could weaken the storage effect (Kuang and Chesson, 2009, 2010) thus different mechanisms might not always combine superadditively as we found here. That said, superadditivity has been found in some cases, i.e., pathogens could enhance the storage effect (Mordecai, 2015). Better explaining plant or microbial diversity would then not be about selecting the best unique mechanism susceptible to explain the observed diversity, but rather better combining those mechanisms together. This may obviously be an annoyance for those who like to sharpen Occam’s razor, but it clearly holds opportunities for theoreticians wishing to investigate synergies between coexistence mechanisms in highly diverse communities. Aside from the synergies between predator diversity-enhancing effects, strong self-regulation through various means and storage effects (on the temporal axis), one obvious follow-up of this research would be interactions with spatial structure. Spatial structure occurs both endogeneously, through spatially restricted movements and interactions, and exogeneously, through spatial variation in environmental covariates (Bolker, 2003). Numerous studies (e.g., Bolker and Pacala, 1999; Murrell and Law, 2002) have shown that spatially restricted movements and interactions - very small-scale spatial structure - can help coexistence, which we believe would be especially important for phytoplankton since many species form colonies (Reynolds, 2006; see discussion in Barraquand et al., 2018). Moreover, although temperature is usually relatively spatially homogeneous over space, other drivers (e.g., rainfall, pH in terrestrial ecosystems; salinity in aquatic ones) may exhibit spatial variation which is a main factor for coexistence (Snyder, 2008). The odds that different (resource) niches, natural enemies, spatial limits to competition and temporal niche partitioning all interact to promote the very high-dimensional coexistence observed in the field seem much higher than for any of those mechanisms alone. Whether the diversity-enhancing effects of these mechanisms combine subadditively (as in Kuang and Chesson, 2010) or superadditively like here is therefore worthy of further research.

413 Acknowledgements

414 We thank Alix Sauve for thoughtful comments and some bibliographic references, as well as
415 György Barabás and an anonymous referee for constructive feedback. This study was supported
416 by the French ANR through LabEx COTE (ANR-10-LABX-45).

417 References

- 418 Adler PB, Ellner SP, Levine JM (2010) Coexistence of perennial plants: an embarrassment of
419 niches. *Ecology letters* 13(8):1019–1029, DOI 10.1111/j.1461-0248.2010.01496.x
- 420 Adler PB, Smull D, Beard KH, Choi RT, Furniss T, Kulmatiski A, Meiners JM, Tredennick
421 AT, Veblen KE (2018) Competition and coexistence in plant communities: intraspecific
422 competition is stronger than interspecific competition. *Ecology Letters* 21(9):1319–1329, DOI
423 10.1111/ele.13098
- 424 Ashby B, Watkins E, Lourenço J, Gupta S, Foster KR (2017) Competing species leave
425 many potential niches unfilled. *Nature Ecology & Evolution* 1(10):1495–1501, DOI 10.1038/
426 s41559-017-0295-3
- 427 Bagchi R, Gallery RE, Gripenberg S, Gurr SJ, Narayan L, Addis CE, Freckleton RP, Lewis
428 OT (2014) Pathogens and insect herbivores drive rainforest plant diversity and composition.
429 *Nature* 506(7486):85–88, DOI 10.1038/nature12911
- 430 Barabás G, Meszéna G, Ostling A (2012) Community robustness and limiting similarity in
431 periodic environments. *Theoretical Ecology* 5(2):265–282, DOI 10.1007/s12080-011-0127-z
- 432 Barabás G, D’Andrea R, Rael R, Meszéna G, Ostling A (2013) Emergent neutrality or hidden
433 niches? *Oikos* 122(11):1565–1572, DOI 10.1111/j.1600-0706.2013.00298.x
- 434 Barabás G, Michalska-Smith MJ, Allesina S (2016) The Effect of Intra- and Interspecific Compe-
435 tition on Coexistence in Multispecies Communities. *The American Naturalist* 188(1):E1–E12,
436 DOI 10.1086/686901
- 437 Barabás G, Michalska-Smith MJ, Allesina S (2017) Self-regulation and the stability of

438 large ecological networks. *Nature Ecology & Evolution* 1(12):1870–1875, DOI 10.1038/
439 s41559-017-0357-6

440 Barraquand F, Picoche C, Maurer D, Carassou L, Auby I (2018) Coastal phytoplankton com-
441 munity dynamics and coexistence driven by intragroup density-dependence, light and hydro-
442 dynamics. *Oikos* In press, DOI 10.1101/171264

443 Bolker B, Pacala S (1999) Spatial Moment Equations for Plant Competition: Understand-
444 ing Spatial Strategies and the Advantages of Short Dispersal. *The American Naturalist*
445 153(6):575–602, DOI 10.1086/303199

446 Bolker BM (2003) Combining endogenous and exogenous spatial variability in analytical pop-
447 ulation models. *Theoretical Population Biology* 64(3):255–270, DOI 10.1016/S0040-5809(03)
448 00090-X

449 Carmel Y, Suprunenko YF, Kunin WE, Kent R, Belmaker J, Bar-Massada A, Cornell SJ
450 (2017) Using exclusion rate to unify niche and neutral perspectives on coexistence. *Oikos*
451 126(10):1451–1458, DOI 10.1111/oik.04380

452 Chesson P (1994) Multispecies Competition in Variable Environments. *Theoretical Population*
453 *Biology* 45:227–276, DOI 10.1006/tpbi.1994.1013

454 Chesson P (2000) Mechanisms of maintenance of species diversity. *Annual review of Ecology*
455 *and Systematics* 31:343–366, DOI 10.1146/annurev.ecolsys.31.1.343

456 Chesson P (2018) Updates on mechanisms of maintenance of species diversity. *Journal of Ecol-*
457 *ogy* 106(5):1773–1794, DOI 10.1111/1365-2745.13035

458 Chesson P, Huntly N (1997) The roles of harsh and fluctuating conditions in the dynamics of
459 ecological communities. *The American Naturalist* 150(5):519–553, DOI 10.1086/286080

460 Comita LS, Queenborough SA, Murphy SJ, Eck JL, Xu K, Krishnadas M, Beckman N, Zhu Y
461 (2014) Testing predictions of the Janzen-Connell hypothesis: a meta-analysis of experimental
462 evidence for distance- and density-dependent seed and seedling survival. *Journal of Ecology*
463 102(4):845–856, DOI 10.1111/1365-2745.12232

464 D’Andrea R, Ostling A (2016) Challenges in linking trait patterns to niche differentiation. *Oikos*
465 125(10):1369–1385, DOI 10.1111/oik.02979

466 Dornelas M, Connolly SR (2008) Multiple modes in a coral species abundance distribution.
467 *Ecology Letters* 11(10):1008–1016, DOI 10.1111/j.1461-0248.2008.01208.x

468 Ellner SP, Snyder RE, Adler PB (2016) How to quantify the temporal storage effect using
469 simulations instead of math. *Ecology Letters* 19(11):1333–1342, DOI 10.1111/ele.12672

470 Fox JW (2013) The intermediate disturbance hypothesis should be abandoned. *Trends in Ecol-*
471 *ogy & Evolution* 28(2):86–92, DOI 10.1016/j.tree.2012.08.014

472 Götzenberger L, de Bello F, Bråthen KA, Davison J, Dubuis A, Guisan A, Lepš J, Lindborg
473 R, Moora M, Pärtel M, Pellissier L, Pottier J, Vittoz P, Zobel K, Zobel M (2012) Ecological
474 assembly rules in plant communities-approaches, patterns and prospects. *Biological Reviews*
475 87(1):111–127, DOI 10.1111/j.1469-185X.2011.00187.x

476 Gravel D, Canham CD, Beaudet M, Messier C (2006) Reconciling niche and neutrality: the
477 continuum hypothesis: Reconciling niche and neutrality. *Ecology Letters* 9(4):399–409, DOI
478 10.1111/j.1461-0248.2006.00884.x

479 Haegeman B, Loreau M (2011) A mathematical synthesis of niche and neutral theories in
480 community ecology. *Journal of Theoretical Biology* 269(1):150–165, DOI 10.1016/j.jtbi.2010.
481 10.006

482 Hastings A, Abbott KC, Cuddington K, Francis T, Gellner G, Lai YC, Morozov A, Petrovskii S,
483 Scranton K, Zeeman ML (2018) Transient phenomena in ecology. *Science* 361(6406):eaat6412,
484 DOI 10.1126/science.aat6412

485 Haydon D (1994) Pivotal assumptions determining the relationship between stability and com-
486 plexity: an analytical synthesis of the stability-complexity debate. *The American Naturalist*
487 144(1):14–29, DOI 10.1086/285658

488 Holt R (2006) Emergent neutrality. *Trends in Ecology & Evolution* 21(10):531–533, DOI 10.
489 1016/j.tree.2006.08.003

490 Hubbell SP (2001) The Unified Neutral Theory of Biodiversity and Biogeography (MPB-32).
 491 Princeton University Press

492 Kawatsu K, Kondoh M (2018) Density-dependent interspecific interactions and the complexity–
 493 stability relationship. *Proc R Soc B* 285(1879):20180698, DOI 10.1098/rspb.2018.0698

494 Kokkoris GD, Jansen VAA, Loreau M, Troumbis AY (2002) Variability in interaction strength
 495 and implications for biodiversity. *Journal of Animal Ecology* 71(2):362–371, DOI 10.1046/j.
 496 1365-2656.2002.00604.x

497 Kuang JJ, Chesson P (2009) Coexistence of annual plants: generalist seed predation weakens
 498 the storage effect. *Ecology* 90(1):170–182, DOI 10.1890/08-0207.1

499 Kuang JJ, Chesson P (2010) Interacting coexistence mechanisms in annual plant communi-
 500 ties: frequency-dependent predation and the storage effect. *Theoretical population biology*
 501 77(1):56–70, DOI 10.1016/j.tpb.2009.11.002

502 Loranger J, Munoz F, Shipley B, Violle C (2018) What makes trait-abundance relationships
 503 when both environmental filtering and stochastic neutral dynamics are at play? *Oikos* DOI
 504 10.1111/oik.05398

505 Matthews TJ, Borges PAV, Whittaker RJ (2014) Multimodal species abundance distributions:
 506 a deconstruction approach reveals the processes behind the pattern. *Oikos* 123(5):533–544,
 507 DOI 10.1111/j.1600-0706.2013.00829.x

508 McGill BJ, Etienne RS, Gray JS, Alonso D, Anderson MJ, Benecha HK, Dornelas M, Enquist
 509 BJ, Green JL, He F, Hurlbert AH, Magurran AE, Marquet PA, Maurer BA, Ostling A,
 510 Soykan CU, Ugland KI, White EP (2007) Species abundance distributions: moving beyond
 511 single prediction theories to integration within an ecological framework. *Ecology Letters*
 512 10(10):995–1015, DOI 10.1111/j.1461-0248.2007.01094.x

513 Moisan JR, Moisan TA, Abbott MR (2002) Modelling the effect of temperature on the maximum
 514 growth rates of phytoplankton populations. *Ecological Modelling* 153(3):197–215, DOI 10.
 515 1016/S0304-3800(02)00008-X

516 Mordecai EA (2015) Pathogen impacts on plant diversity in variable environments. *Oikos*
517 124(4):414–420, DOI 10.1111/oik.01328

518 Munoz F, Huneman P (2016) From the neutral theory to a comprehensive and multiscale
519 theory of ecological equivalence. *The Quarterly Review of Biology* 91(3):321–342, DOI 10.
520 1086/688098

521 Murrell DJ, Law R (2002) Heteromyopia and the spatial coexistence of similar competitors.
522 *Ecology Letters* 6(1):48–59, DOI 10.1046/j.1461-0248.2003.00397.x

523 Mutshinda CM, O’Hara RB, Woiwod IP (2009) What drives community dynamics? *Proceedings*
524 *of the Royal Society B: Biological Sciences* 276(1669):2923–2929, DOI 10.1098/rspb.2009.0523

525 Rael RC, D’Andrea R, Barabás G, Ostling A (2018) Emergent niche structuring leads to in-
526 creased differences from neutrality in species abundance distributions. *Ecology* 99(7):1633–
527 1643, DOI 10.1002/ecy.2238

528 Reynolds CS (2006) *The ecology of phytoplankton*. Cambridge University Press

529 Rinaldi S, Murator S, Kuznetsov Y (1993) Multiple attractors, catastrophes and chaos in sea-
530 sonally perturbed predator-prey communities. *Bulletin of Mathematical Biology* 55(1):15–35,
531 DOI 10.1007/BF02460293

532 Sakavara A, Tsirtsis G, Roelke DL, Mancy R, Spatharis S (2018) Lumpy species coexistence
533 arises robustly in fluctuating resource environments. *Proceedings of the National Academy*
534 *of Sciences* 115(4):738–743, DOI 10.1073/pnas.1705944115

535 Scheffer M, van Nes EH (2006) Self-organized similarity, the evolutionary emergence of groups
536 of similar species. *Proceedings of the National Academy of Sciences* 103(16):6230–6235, DOI
537 10.1073/pnas.0508024103

538 Scranton K, Vasseur DA (2016) Coexistence and emergent neutrality generate synchrony among
539 competitors in fluctuating environments. *Theoretical Ecology* 9(3):353–363, DOI 10.1007/
540 s12080-016-0294-z

541 Snyder RE (2008) When does environmental variation most influence species coexistence? *The-*
542 *oretical Ecology* 1(3):129–139, DOI 10.1007/s12080-008-0015-3

543 Stump SM (2017) Multispecies Coexistence without Diffuse Competition; or, Why Phylogenetic
544 Signal and Trait Clustering Weaken Coexistence. *The American Naturalist* 190(2):213–228,
545 DOI 10.1086/692470

546 Taylor RA, White A, Sherratt JA (2013) How do variations in seasonality affect population
547 cycles? *Proceedings of the Royal Society B: Biological Sciences* 280(1754), DOI 10.1098/
548 rspb.2012.2714

549 Tyson R, Lutscher F (2016) Seasonally Varying Predation Behavior and Climate Shifts Are
550 Predicted to Affect Predator-Prey Cycles. *The American Naturalist* 188(5):539–553, DOI
551 10.1086/688665

552 Vasseur DA, Yodzis P (2004) The Color Of Environmental Noise. *Ecology* 85(4):1146–1152,
553 DOI 10.1890/02-3122

554 Vergnon R, van Nes EH, Scheffer M (2013) Interpretation and predictions of the Emergent neu-
555 trality model: a reply to Barabás et al. *Oikos* 122(11):1573–1575, DOI 10.1111/j.1600-0706.
556 2013.00790.x

557 Vesipa R, Ridolfi L (2017) Impact of seasonal forcing on reactive ecological systems. *Journal of*
558 *Theoretical Biology* 419:23–35, DOI 10.1016/j.jtbi.2017.01.036

559 Winder M, Cloern JE (2010) The annual cycles of phytoplankton biomass. *Philosophi-*
560 *cal Transactions of the Royal Society B: Biological Sciences* 365(1555):3215–3226, DOI
561 10.1098/rstb.2010.0125

562 Wootton JT, Emmerson M (2005) Measurement of Interaction Strength in Nature. *Annual*
563 *Review of Ecology, Evolution, and Systematics* 36(1):419–444, DOI 10.1146/annurev.ecolsys.
564 36.091704.175535

565 Zhao XQ (1991) The qualitative analysis of n-species Lotka-Volterra periodic competition
566 systems. *Mathematical and Computer Modelling* 15(11):3–8, DOI 10.1016/0895-7177(91)
567 90100-L