

1 **Machine Learning-Guided Synthetic Microbial Communities Enable Functional and**
2 **Sustainable Degradation of Persistent Environmental Pollutants**

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12

13 **Abstract**

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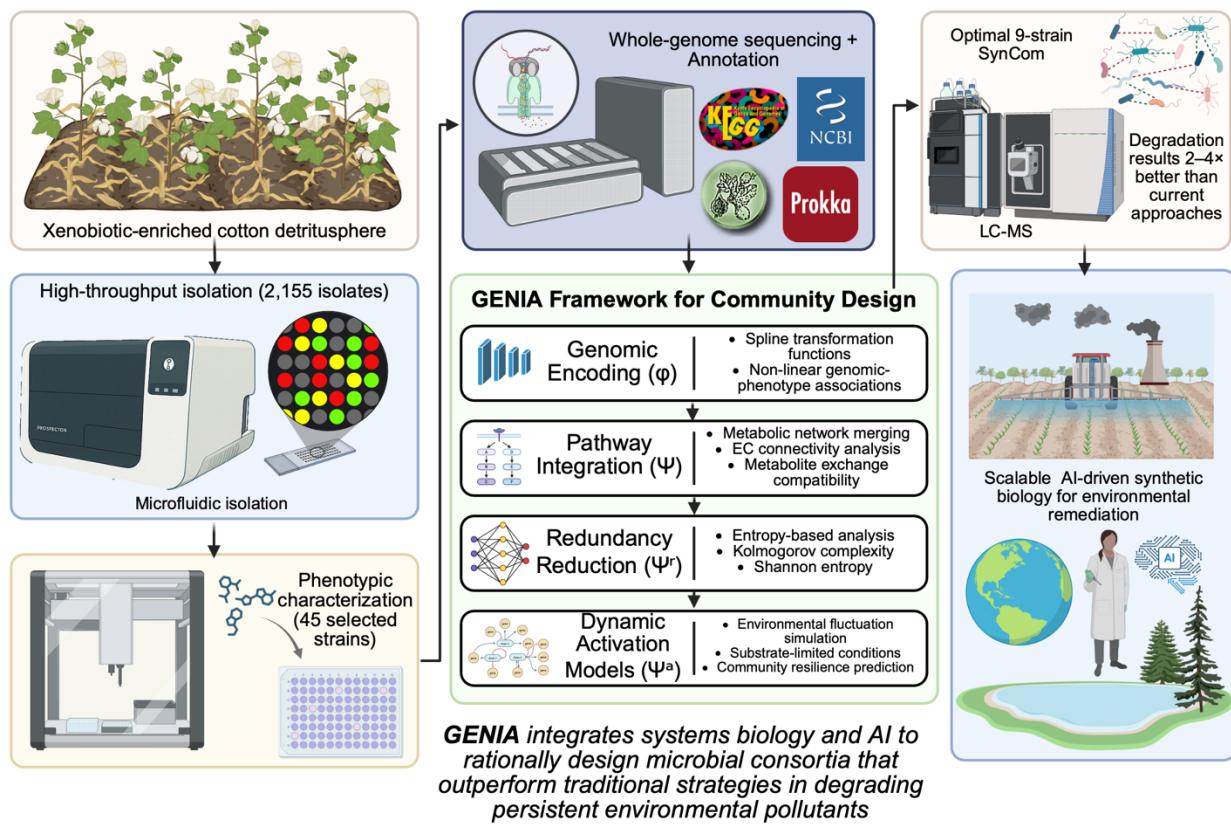
15 Persistent environmental pollutants demand the use of diverse microbial metabolic capabilities for
16 effective degradation. While naturally occurring consortia or single strains often fall short in
17 efficiency, synthetic microbial communities (SynComs) hold greater promise for enhanced
18 degradation. To address this challenge, we developed GENIA (Genomically and Environmentally
19 Networked Intelligent Assemblies), a genome-informed and machine learning-guided framework
20 for the rational design of SynComs capable of multi-pollutant degradation under simulated
21 environmental conditions. Using a microfluidic high-throughput cultivation platform, 2,155
22 bacterial strains were isolated from xenobiotic-enriched environments and screened for pollutant-
23 specific growth. Whole-genome sequencing and functional annotation of 45 prioritized strains
24 revealed metabolic traits associated with the potential degradation of challenging persistent
25 environmental pollutants as proof of concept, i.e., lignin oxidation, atrazine dechlorination, and
26 PFAS defluorination. These genomic profiles were encoded into spline-based graph
27 representations and integrated within the GENIA pipeline, which combines graph neural networks,
28 pathway complementarity modeling, and functional redundancy minimization to predict optimal
29 community assemblies. The resulting nine-member community—comprising *Pantoea dispersa*,
30 *Atlantibacter hermannii*, *Pseudomonas fulva*, *Paenibacillus polymyxa*, *Bacillus cabrialesii*,
31 *Micrococcus luteus*, *Bacillus pseudomycoides*, *Bacillus licheniformis*, and *Pseudomonas*

32 *pergaminensis*—was predicted to exhibit broad catabolic capacity and minimal intra-community
33 competition. Kinetic experiments in minimal medium demonstrated simultaneous multi-pollutant
34 degradation: lignin (91.6% removal by day 5), atrazine (91.4% removal by day 3), and PFOS
35 (93.1% removal within seven days), representing a 2-4-fold improvement over existing
36 approaches. GENIA establishes a scalable and generalizable framework that integrates systems-
37 level genomics, phenotypic screening, and predictive modeling to engineer ecologically coherent
38 microbial consortia with application to complex environmental bioremediation.

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Graphical Abstract



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43 Keywords

44 PFAS biodegradation; synthetic microbial communities (SynComs); machine learning; PFOA;
45 PFOS; atrazine, lignin, metabolic complementarity; microbial bioremediation; whole genome
46 sequencing

47

48

49 **INTRODUCTION**

50

51 Microbial communities have shaped life on Earth not through competition, but through
52 cooperation, a principle captured by Lynn Margulis: "*Life did not take over the globe by combat,*
53 *but by networking*" (1). This ecological interdependence inspires synthetic microbial communities
54 (SynComs), intentionally assembled to perform functions beyond the reach of individual strains
55 or natural consortia. Yet, despite the conceptual appeal, few studies have validated SynCom as a
56 practical strategy for solving real-world environmental problems. Persistent pollutants such as
57 lignin, atrazine, and per- and polyfluoroalkyl substances (PFAS) exemplify global environmental
58 challenges that demand innovative remediation approaches (2–5). These pollutants pose persistent
59 threats to both terrestrial and aquatic ecosystems due to their structural stability and widespread
60 use in agriculture and industry (6–8). Atrazine, a widely used herbicide, persists in groundwater
61 and has been associated with endocrine disruption and ecotoxicity (9–11). Lignin, although
62 natural, is a complex aromatic polymer that hinders biomass valorization and requires microbial
63 degradation for effective bioconversion (12,13). PFAS, including perfluorooctanoic acid (PFOA)
64 and perfluorooctane sulfonate (PFOS), are highly fluorinated compounds used in coatings and
65 surfactants that resist degradation, leading to global contamination and bioaccumulation (14–16).
66 Microbial bioremediation has been explored as a sustainable strategy to degrade these pollutants,
67 leveraging specific enzymes such as peroxidases, dehalogenases, and monooxygenases (17–19).
68 However, most natural microbial consortia exhibit functional redundancy and competitive
69 interactions that limit degradation efficiency (20). Furthermore, the isolation and characterization
70 of novel strains with xenobiotic-degrading capabilities remain a bottleneck, particularly for
71 pollutants with limited bioavailability or transport across membranes (21).

72

73 To address these limitations, high-throughput microfluidic platforms such as the Prospector®
74 system allow cultivation of challenging or slow-growing environmental bacteria by enabling
75 nanowell-scale isolation, minimizing interspecies competition, and promoting unbiased recovery
76 of environmental taxa (22,23). Following isolation, genome-resolved metabolic modeling can
77 guide rational community design by integrating genomic potential with functional performance.
78 Recent advances in whole-genome sequencing, particularly using long-read platforms such as

79 Oxford Nanopore Technologies, have enhanced the resolution of biosynthetic and degradative
80 pathways in environmental strains (24–26).

81

82 In parallel, machine learning (ML) and network-based modeling frameworks have shown promise
83 in designing SynComs for targeted functions, including biodegradation and nutrient cycling
84 (27,28). Graph Neural Networks (GNNs) and attention-based architectures enable integration of
85 heterogeneous biological data, allowing predictions of emergent properties such as community
86 stability and pollutant degradation capacity (29–31). However, these computational approaches
87 are often developed independently of experimental validation, limiting their practical utility.

88

89 Here, we present an integrated experimental-computational framework, GENIA (Genomically and
90 Environmentally Networked Intelligent Assemblies), that couples high-throughput strain isolation,
91 functional screening, genome sequencing, and network modeling to design synthetic communities
92 for the degradation of lignin, atrazine, and PFAS. We isolated 2,155 bacterial strains from the
93 cotton detritusphere—a complex, microbially active niche shaped by intensive agrochemical use
94 and enriched in recalcitrant compounds such as lignin (from plant biomass), atrazine (a legacy
95 herbicide), and per- and polyfluoroalkyl substances (PFAS) from irrigation and soil
96 amendments—and performed systematic phenotypic screening on defined pollutant substrates.
97 Genomes of the most effective degraders ($n = 45$) were sequenced, annotated, and analyzed for
98 key catabolic enzymes and transporters.

99

100 Using spline-based genome encodings and metabolic network integration, we constructed
101 predictive models of community-level degradation. Functional redundancy was minimized
102 through Kolmogorov complexity analysis, and dynamic activation models simulated pathway
103 engagement under stress. Community performance was predicted using a hybrid GNN architecture
104 incorporating Graph Attention Networks and Node2Vec embeddings, trained on empirical growth
105 data.

106

107 This study presents the first demonstration of machine learning-guided synthetic community
108 assembly for simultaneous degradation of structurally diverse persistent pollutants, establishing a

109 comprehensive workflow that bridges high-throughput microbiology, systems genomics, and
110 predictive AI modeling.

111

112 MATERIALS AND METHODS

113

114 **Overall study design.** The strategy used encompassed field, lab, informatics, and *in vitro*
115 evaluation to develop and validate a minimal synthetic community for degradation of targeted
116 environmental pollutants: PFAS, atrazine, and lignin. First, we recovered large number of bacterial
117 isolates using the Prospector high throughput culturomics method, allowing us to screen for
118 bacteria that can utilize the pollutants of interest as a nutritional substrate, and potentially degrade
119 them. The bacteria identified as potential degraders were sequenced for their genomes, and our
120 novel GENIA strategy (Fig 1) was applied towards genome-scale models. Using a set of machine
121 learning strategies, we identified low-redundancy and complementary communities that can
122 efficiently degrade all three pollutants of interest. Finally, we validated the strategy *in vitro* to test
123 the removal of the pollutant molecules from the lab enrichments. Detailed descriptions are
124 presented below.

125

126 **High-throughput Culturomics in Strain Isolation Using the Prospector® System.** Bacteria
127 were isolated from cotton stalks-associated detritusphere using the Prospector® Microbial
128 Isolation and Cultivation System (Isolation Bio, San Carlos, CA). The detritusphere was selected
129 as an isolation source due to its enrichment in lignocellulosic residues and legacy of agricultural
130 and industrial contaminants, including atrazine and perfluoroalkyl substances (PFAS) (32,33). The
131 cotton stalks were collected from Texas A&M Brazos Bottom Experimental Farm Station
132 (30.551157826632647, -96.43135829477686). These stalks had been left behind after harvest in
133 November 2024 and were sampled for our study in April 2025. For microbial extraction, 0.5 g of
134 lignocellulosic material was weighed and diluted with 4.5 mL of PBS, followed by vortexing for
135 1 minute to detach large soil particles. Samples were then sonicated for 5 minutes at 40 kHz
136 frequency to dislodge microorganisms from the detritusphere. Serial dilutions of this suspension
137 were prepared in R2A medium using the Most Probable Number (MPN) technique with OzBlue
138 as a redox indicator for growth detection, MPN enabled quantification of the original microbial
139 concentration, which was subsequently used to calculate the appropriate dilution factor required

140 to achieve a target density of approximately one cell per 3 nL (34). Diluted suspensions (1.5 mL)
141 were mixed with 1.5 mL R2A medium containing the Prospector® viability dye (final
142 concentration: 200 $\mu\text{mol L}^{-1}$) and loaded into arrays of >6000 nanowells (3 nL per well) using the
143 system's vacuum-loading chamber. Arrays were sealed and incubated aerobically at 27 °C for 48
144 hours. Viable isolates were identified based on fluorescence changes (excitation: 530 nm;
145 emission: 590 nm) and aseptically transferred to 96-well plates containing R2A broth using the
146 automated Prospector® transfer system (35). A total of 2,155 pure isolates were obtained.
147

148 **Functional Screening of Pollutant Degradation Potential.** To assess the potential for pollutant
149 degradation, isolates were screened for growth using lignin, atrazine, or PFAS as primary carbon
150 sources in minimal medium. Analytical-grade compounds were used throughout: alkali lignin
151 (Sigma-Aldrich, ≥97% purity), atrazine (Tokyo Chemical Industry Co., >97% purity),
152 perfluorooctanesulfonic acid potassium salt (PFOS-K, Agilent Technologies, ≥98% purity), and
153 perfluorooctanoic acid (PFOA, Agilent Technologies, ≥95% purity). PFAS compounds were
154 supplied as stock solutions in methanol and added to achieve final concentrations of 20 mg L^{-1}
155 each, with methanol content maintained below 0.04% (v/v), and control treatments with equivalent
156 methanol concentrations without PFAS were included to account for potential solvent utilization.
157 Before inoculation, cells were washed twice with phosphate-buffered saline (PBS, pH 7.0),
158 resuspended, and normalized to an optical density of $\text{OD}_{600} = 0.5$. Automated inoculation into 96-
159 well plates was performed using the Opentrons OT-2 liquid handling platform (36). Each well
160 received 20 μL of normalized inoculum, 180 μL of M9 minimal medium (Na_2HPO_4 , 6.78 g L^{-1} ;
161 KH_2PO_4 , 3.0 g L^{-1} ; NH_4Cl , 1.0 g L^{-1} ; NaCl , 0.5 g L^{-1} ; MgSO_4 , 1 mM; CaCl_2 , 0.1 mM),
162 supplemented with alkali lignin (500 mg L^{-1}), atrazine (30 mg L^{-1}), or PFAS (20 mg L^{-1} each of
163 PFOA and PFOS) (13, 37, 38) and 5 μL OzBlue redox indicator. Cultures were incubated at 27 °C
164 for 10 days with orbital shaking at 120 rpm. Growth was assessed daily by measuring relative
165 fluorescence units (RFU) at 560 nm excitation and 590 nm emission wavelengths using a
166 microplate reader (SpectraMax, Molecular Devices). Isolates were considered positive for
167 pollutant utilization when fluorescence readings exceeded the mean fluorescence of negative
168 controls by at least 2 standard deviations and showed sustained growth over a minimum of 3
169 consecutive days. Forty-five strains demonstrating consistent growth on at least one pollutant were
170 selected for further analysis.

171

172 **Whole-Genome Sequencing and Assembly.** Genomic DNA was extracted from each isolate
173 using the ZymoBIOMICS DNA Miniprep Kit (Zymo Research, USA). High molecular weight
174 DNA was used to construct sequencing libraries with Oxford Nanopore standard protocols.
175 Sequencing was performed on the PromethION platform (Plasmidsaurus, Eugene, OR). Genome
176 assemblies were generated using Flye v2.9 (39), followed by three rounds of polishing with Racon
177 v1.4.21 (40) and one round with Medaka v1.7.2 (41). Assembly quality was assessed using
178 QUAST v5.2.0 (42), and completeness and contamination were evaluated with CheckM v1.2.2
179 (43), ensuring >98% completeness and <1.5% contamination.

180

181 **Genome Annotation and Pathway Characterization.** Genome annotation was performed using
182 PGAP (44), Prokka v1.14.6 (45), and RAST (46). Prodigal v2.6.3 was used for gene prediction
183 (47). Functional annotation employed KEGG (48), COG (49), and BiGG Models database (50).
184 BLASTx and BLASTp were used to identify orthologs relevant to biodegradation with thresholds
185 of $\geq 70\%$ identity, $\geq 70\%$ query coverage, and E-value $\leq 1e^{-5}$ (51). Specific enzymes were identified
186 and classified based on their EC numbers from the functional annotation databases. Lignin-
187 degrading enzymes included lignin peroxidases (EC 1.11.1.14), manganese peroxidases (EC
188 1.11.1.13), aryl-alcohol oxidases (EC 1.1.3.7), dye-decolorizing peroxidases (EC 1.11.1.19), and
189 vanillate demethylases (EC 3.1.1.68). Atrazine degradation pathways included atrazine
190 chlorohydrolase (EC 3.8.1.8), hydroxyatrazine ethylaminohydrolase (EC 3.5.4.43), and cyanuric
191 acid hydrolase (EC 3.5.2.15); the dealkylation pathway involving N-dealkylases and cytochrome
192 P450 enzymes (EC 1.14.14.-); and the oxidative pathway with atrazine deaminase and various
193 oxidoreductases (EC 1.4.99.-). Additional enzymes included melamine deaminase (EC 3.5.4.45),
194 ammeline aminohydrolase (EC 3.5.1.84), and ammelide aminohydrolase (EC 3.5.1.102) for
195 complete triazine ring mineralization. For PFAS, enzymes such as haloalkane dehalogenase (EC
196 3.8.1.5), alkane monooxygenases (EC 1.14.15.3), NAD(P)H oxidoreductases (EC 1.6.99.-), and
197 cytochrome P450s (EC 1.14.14.-) were detected. Genes encoding membrane transporters,
198 including ABC transporters, MFS, RND efflux pumps, and TRAP systems, were frequently
199 associated with atrazine and PFAS metabolism.

200

201 **Community Design and Network Integration Modeling.** The GENIA framework shown in Fig.
202 1 guided the synthetic community design. Annotated genomes were encoded using spline
203 transformation functions (ϕ) to capture non-linear associations between genomic content and
204 phenotype (52). Encoded profiles were processed by the pathway integration module (Ψ), which
205 generated composite networks by merging individual strain metabolic maps based on EC
206 connectivity and metabolite exchange compatibility (53–55). Functional redundancy was reduced
207 using entropy-based redundancy analysis (Ψ^r) via Kolmogorov complexity and Shannon entropy
208 (56,57). Community resilience and conditional functionality were predicted using dynamic
209 activation models (Ψ^a) that simulated environmental fluctuations and substrate-limited conditions
210 (58,59). Graph neural network models integrating Graph Attention Networks and Node2Vec
211 embeddings were trained on substrate utilization data to predict community-level degradation
212 potential (60,61).

213

214 **Cross-Validation with INaP 2.0.** To assess the predictive robustness of the GENIA-derived
215 community design, we conducted cross-validation using INaP 2.0 (Integrated Network and
216 Pathway prediction pipeline, v2.0), a complementary genome-informed platform for microbial
217 interaction modeling (62). INaP 2.0 integrates taxonomic, metabolic, and pathway co-occurrence
218 matrices to simulate cross-feeding and syntrophic dependencies among microbial pairs and
219 consortia. The identical 45 genomes were annotated using the INaP 2.0 pipeline, and functional
220 overlap with GENIA predictions was quantified by computing Jaccard similarity indices between
221 predicted metabolic capabilities and pollutant degradation modules.

222

223 **Experimental Validation of Synthetic Communities.** Nine strains predicted by GENIA to form
224 an optimal community were systematically validated through controlled co-culture experiments.
225 Individual strains were first cultivated separately to mid-exponential phase ($OD_{600} = 0.8\text{--}1.0$) in
226 M9 minimal medium supplemented with 0.2% glucose. Cell densities were determined by serial
227 dilution plating and adjusted to $9 \times 10^6 \text{ CFU mL}^{-1}$ for each strain. The synthetic community was
228 assembled by mixing equal volumes of each strain suspension in an equal ratio, resulting in a final
229 theoretical density of 10^6 CFU mL^{-1} per strain. The assembled synthetic communities were
230 inoculated into fresh M9 minimal medium (Na_2HPO_4 , 6.78 g L^{-1} ; KH_2PO_4 , 3.0 g L^{-1} ; NH_4Cl , 1.0 g L^{-1} ; NaCl , 0.5 g L^{-1} ; MgSO_4 , 1 mM ; CaCl_2 , 0.1 mM) containing one of the three target pollutants

232 as the sole carbon source. Treatment conditions included kraft lignin (500 mg L^{-1}), atrazine (30 mg L^{-1}), or PFAS mixture (20 mg L^{-1} each of PFOA and PFOS). Control treatments without
233 pollutants and with individual strain monocultures were included for comparative analysis.
234 Incubations were performed in 250 mL baffled Erlenmeyer flasks with 50 mL working volume in
235 6 biological replicates per treatment at 27°C with orbital shaking at 150 rpm for 7 days. Samples
236 were collected every day at consistent time points (09:00 h). At each sampling point, 2 mL aliquots
237 were collected under aseptic conditions, where 1 mL was immediately processed for chemical
238 analysis and the reminder 1 mL preserved in 20% glycerol at -80°C for molecular analysis.
239 Atrazine and PFAS degradation were quantified by LC-MS/MS after solid-phase extraction using
240 C18 cartridges, employing compound-specific multiple reaction monitoring (MRM) transitions
241 with deuterated internal standards to ensure analytical accuracy (63,64). Additionally, fluoride
242 concentrations were determined using the Hach SPADNS 2 Method 10225 with Accuvac®
243 ampules, following EPA-compliant procedures equivalent to EPA reference method SM 4500-F
244 D for drinking water and wastewater analysis (65). Lignin degradation was monitored by UV-Vis
245 spectrophotometry, measuring the decrease in aromatic content at 280 nm and 310 nm, with
246 vanillin and syringaldehyde formation quantified at their characteristic absorption maxima of 308
247 nm and 274 nm, respectively (66).

249

250 **Statistical Analysis and Computational Framework.** Statistical analyses were conducted using
251 a comprehensive computational framework implemented in Python 3.11 with GPU acceleration
252 via CUDA 12.1. The primary machine learning framework utilized TensorFlow 2.15 with Keras
253 API for deep learning implementations, complemented by PyTorch 2.1 for graph neural network
254 architectures (67,68). Data preprocessing and traditional statistical analyses employed scikit-learn
255 1.4, NumPy 1.25, and SciPy 1.11 for robust numerical computations (69,70).

256

257 Community composition differences were assessed using permutational multivariate analysis of
258 variance (PERMANOVA) with 10,000 permutations, implemented through the PERMANOVA-S
259 framework to accommodate multiple distance metrics simultaneously (71,72). Non-parametric
260 tests included Analysis of Similarities (ANOSIM), Multi-Response Permutation Procedures
261 (MRPP), and the microbiome higher criticism analysis (MiHC) for sparse association detection

262 (73,74). Effect sizes were calculated using Cohen's d for pairwise comparisons and eta-squared
263 (η^2) for multivariate effects.

264

265 Model performance was evaluated using stratified k-fold cross-validation (k=10) with temporal
266 blocking to prevent data leakage in time-series analyses. Hyperparameter optimization employed
267 Bayesian optimization using Gaussian Process regression implemented in the Optuna framework
268 with Tree-structured Parzen Estimator (TPE) sampling (75). Model selection criteria included area
269 under the precision-recall curve (AUPRC), F1-score, Matthew's correlation coefficient (MCC),
270 and balanced accuracy to handle class imbalance in degradation outcomes.

271

272 Neural network architectures were optimized using Neural Architecture Search (NAS) with
273 evolutionary algorithms and early stopping based on validation loss plateauing (patience=20
274 epochs). Regularization techniques included dropout ($p=0.3$), batch normalization, and L2 weight
275 decay ($\lambda=0.001$). Learning rate scheduling employed cosine annealing with warm restarts, and
276 gradient clipping prevented exploding gradients in recurrent components (76,77).

277

278 Post-hoc power analyses were conducted using Monte Carlo simulations with 5,000 bootstrap
279 resamples to ensure adequate statistical power ($1-\beta \geq 0.80$) for detecting biologically meaningful
280 effects. False discovery rate (FDR) correction was applied using the Benjamini-Hochberg
281 procedure for multiple comparisons, with a significance threshold set at $\alpha = 0.05$. Effect size
282 interpretation followed Cohen's conventions for ecological data: small ($\eta^2 = 0.01$), medium ($\eta^2 =$
283 0.06), and large ($\eta^2 = 0.14$).

284

285 Predictive uncertainty was quantified using Monte Carlo dropout and ensemble methods with 100
286 forward passes. Model interpretability employed SHAP (SHapley Additive exPlanations) values
287 and integrated gradients to identify feature importance in community assembly predictions (78,79).
288 Confidence intervals for degradation rates were calculated using bias-corrected and accelerated
289 (BCa) bootstrap with 2,000 resamples.

290

291 **Data Availability.** Genome sequences and annotations are deposited in GenBank under accession
292 numbers listed in the Supporting Information. Complete analytical code, preprocessed datasets,

293 and hyperparameter configurations are available at
294 https://github.com/EsauDelaVega/GENIA_Framework_Biodegradation.git. Interactive
295 visualizations and supplementary analyses are accessible through Jupyter notebooks in the project
296 repository.

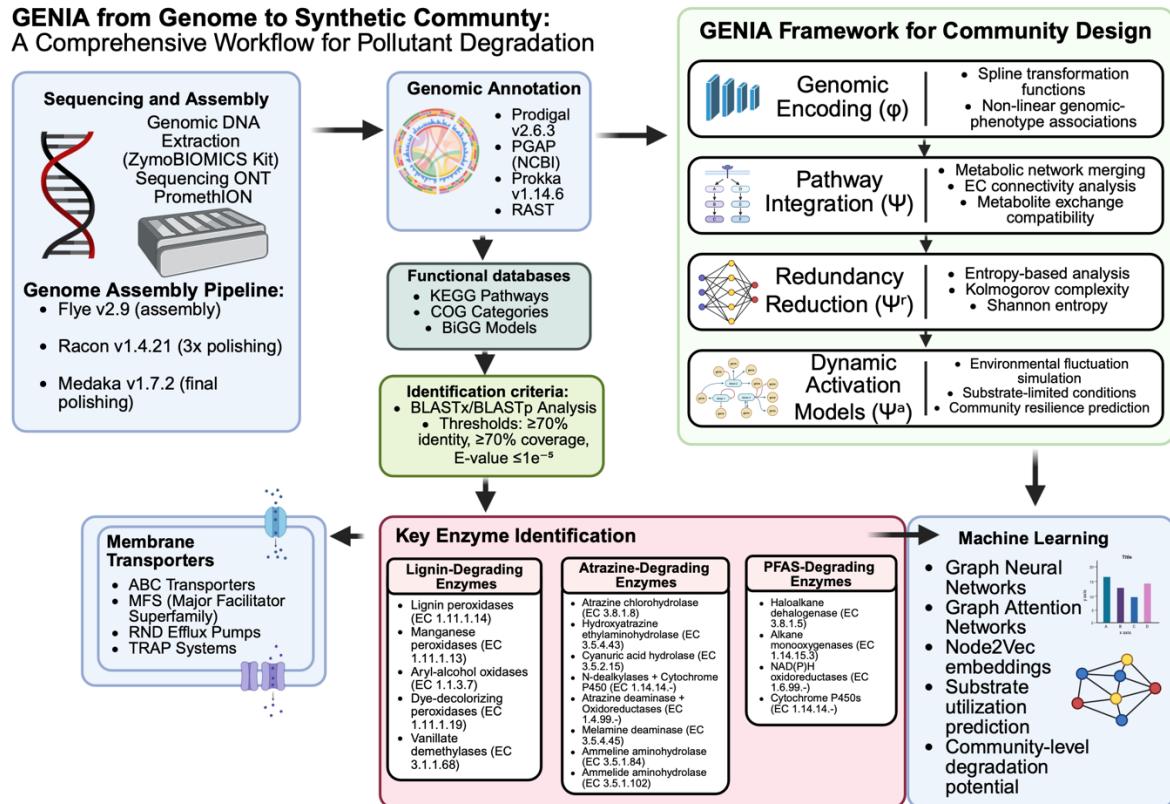
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298 RESULTS

299

300 The GENIA framework was developed as an integrated computational pipeline for rational design
301 of synthetic microbial communities capable of multi-pollutant degradation (Figure 1). The
302 workflow integrates high-throughput strain isolation through microfluidic cultivation, whole-
303 genome sequencing and assembly, comprehensive functional annotation through multiple
304 pipelines, and machine learning-guided community optimization. The framework incorporates
305 four sequential computational modules: genomic encoding (ϕ) using spline transformation
306 functions, pathway integration (Ψ) for metabolic network merging, redundancy reduction (Ψ^r)
307 through entropy-based analysis, and dynamic activation models (Ψ^a) for environmental simulation.
308 Machine learning components integrate Graph Neural Networks, Graph Attention Networks, and
309 Node2Vec embeddings to predict optimal community assemblies.

310



311

312 **Figure 1. GENIA framework for machine learning-guided design of pollutant-degrading synthetic microbial**
313 **communities.** The integrated workflow combines high-throughput isolation, Oxford Nanopore sequencing, and
314 comprehensive annotation (Prodigal, PGAP, Prokka, RAST) with KEGG, COG, and BiGG databases. GENIA
315 processes genomic data through four sequential modules: (1) Genomic Encoding (φ) using spline transformations,
316 (2) Pathway Integration (Ψ) for metabolic network merging, (3) Redundancy Reduction (Ψ^r) employing Kolmogorov
317 complexity analysis, and (4) Dynamic Activation Models (Ψ^a) for environmental simulation. Machine learning
318 integration uses Graph Neural Networks and Node2Vec embeddings to predict optimal synthetic communities
319 targeting lignin, atrazine, and PFAS degradation pathways with enhanced capacity and minimal competition.

320

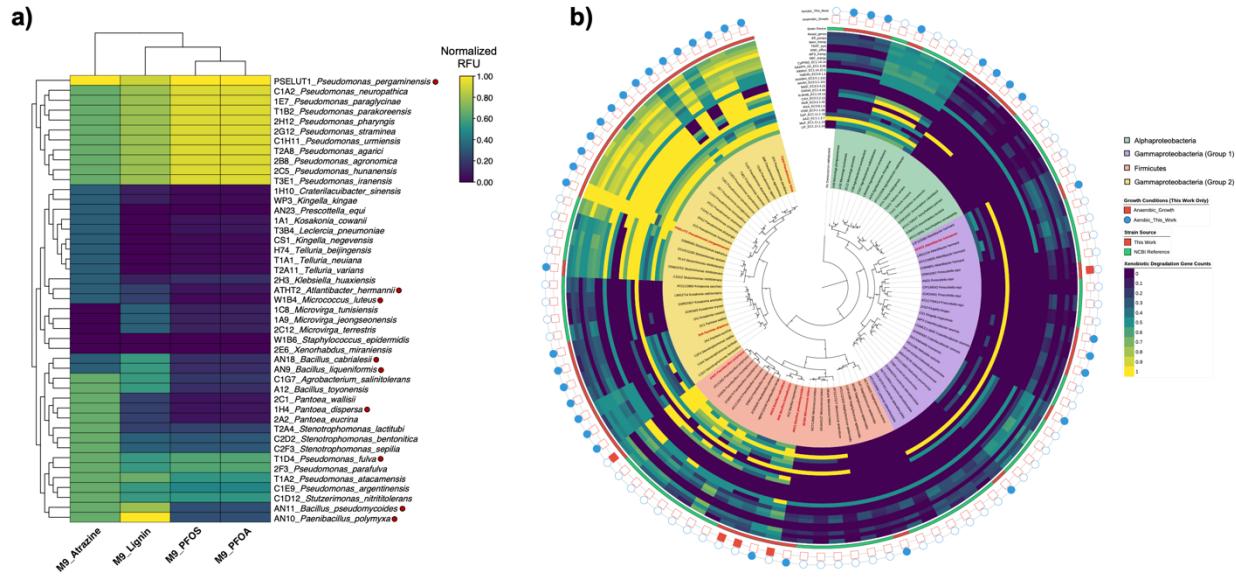
321 **Functional Diversity and Phylogenomic Distribution of Biodegradation Capabilities**

322 Comprehensive functional screening of the 45 selected strains on minimal medium containing
323 individual pollutants showed distinct patterns of substrate utilization and degradation potential
324 (Figure 2a). The hierarchical clustering analysis grouped strains based on their growth
325 performance across the four test conditions (M9 with addition of atrazine, M9 + lignin, M9 +
326 PFOS, or PFOA), showing functional clusters that transcended taxonomic boundaries. Lignin
327 utilization showed the broadest representation across strains, with notable performers including
328 *Paenibacillus polymyxa* (AN10, 8.80 ± 0.56 RFU), *Pseudomonas pergaminensis* (PSELUT1, 8.00
329 ± 0.52 RFU), and multiple *Pseudomonas* species achieving 7.20 ± 0.45 RFU after 72 hours of

330 incubation. Atrazine degradation capabilities were more selective, with *Pseudomonas*
331 *pergaminensis* demonstrating the highest activity (3.60 ± 0.29 RFU), followed by a cluster of
332 strains including *Bacillus pseudomycoides* (AN11), *Pantoea* species, and most *Pseudomonas*
333 isolates showing moderate but consistent growth (2.40 ± 0.17 RFU). PFAS degradation proved
334 most challenging, with *Pseudomonas pergaminensis* again leading performance on both PFOS
335 (4.20 ± 0.36 RFU) and PFOA (2.52 ± 0.23 RFU). In comparison, most other *Pseudomonas* strains
336 achieved moderate PFAS utilization (3.90 ± 0.31 RFU for PFOS, 2.34 ± 0.19 RFU for PFOA).
337 Detailed genomic analysis showed extensive enzymatic repertoires underlying the observed
338 growth phenotypes (Figure 2b). The xenobiotic degradation gene count heatmap demonstrated that
339 diverse enzyme families supported biodegradation potential distributed heterogeneously across the
340 strain collection. *Pseudomonas pergaminensis* (PSELUT1) exhibited the most comprehensive
341 enzymatic profile, harboring multiple copies of lignin peroxidases (3 genes), PFAS-degrading
342 haloalkane dehalogenases (2 genes), and extensive transporter systems (26 total membrane
343 transporters). *Paenibacillus polymyxa* (AN10) showed exceptional lignin degradation capacity
344 with 5 dye-decolorizing peroxidase genes and 4 vanillate demethylase genes, correlating with its
345 superior lignin utilization performance. The enzyme distribution analysis revealed clear
346 specialization patterns across taxonomic groups. *Pseudomonas* species consistently possessed the
347 broadest enzymatic capabilities, with comprehensive PFAS degradation pathways including
348 multiple copies of cytochrome P450 enzymes (5-6 genes) and extensive efflux pump systems (4-
349 5 genes). *Bacillus* strains showed strong lignin and atrazine degradation potential but notably
350 lacked PFAS-specific enzymes, with AN11 (*B. pseudomycoides*) displaying 4 dye-decolorizing
351 peroxidase genes but only 2 haloalkane dehalogenase genes. *Microvirga* species exhibited an
352 inverse specialization, possessing robust lignin degradation enzymes (2 lignin peroxidases, 2
353 manganese peroxidases) but completely lacking atrazine chlorohydrolase and hydroxyatrazine
354 ethylaminohydrolase genes, explaining their poor atrazine utilization phenotypes.
355
356 Phylogenetic analysis demonstrated that biodegradation capabilities were widely distributed
357 across bacterial families rather than being phylogenetically constrained, supporting horizontal
358 gene transfer as a significant evolutionary mechanism. The integration of quantitative growth data
359 with genomic enzyme profiles validated the GENIA framework's predictive capacity. It provided

360 the foundation for rational community selection based on complementary metabolic functions
361 rather than functional redundancy.

362



363

364 **Figure 2. Functional diversity and phylogenetic distribution of biodegradation capabilities across bacterial**
365 **isolates.** (a) Hierarchical clustering heatmap showing normalized growth performance of 45 selected strains on
366 minimal medium (M9) supplemented with individual pollutants. Rows represent bacterial strains, columns represent
367 growth conditions (Atrazine, Lignin, PFOS, PFOA). Color intensity indicates normalized delta relative fluorescence
368 units (RFU) over 3 consecutive days of sustained growth (scale: 0.00-1.00, dark purple = low growth, yellow = high
369 growth)—clustering based on Euclidean distance with Ward linkage. Red dots indicate strains selected for synthetic
370 community assembly. (b) Maximum likelihood phylogenomic tree constructed from 312 orthologous gene families
371 using GTR+Gamma+I substitution model with 1,000 bootstrap replicates (bootstrap values >95% for major clades).
372 Concentric rings display respiratory metabolism (blue circles = aerobic, red squares = anaerobic), strain genomes
373 source (red = strains isolated in this study, green = NCBI reference strains), and normalized xenobiotic degradation
374 enzyme counts (blue = low abundance, yellow = high abundance). The analysis reveals phylogenetically independent
375 distribution of biodegradation capabilities, demonstrating that functional potential transcends taxonomic boundaries
376 and supporting horizontal gene transfer as a mechanism for xenobiotic metabolism evolution.

377

378 **Quantitative Network Analysis of Machine Learning-Guided Bacterial Consortium Design**

379 The GENIA framework successfully integrated whole-genome sequencing data from 45 bacterial
380 isolates with phenotypic growth profiles to predict optimal synthetic communities for multi-
381 contaminant degradation. Machine learning analysis identified nine bacterial strains with

382 complementary metabolic capabilities, resulting in two distinct network representations that
383 quantitatively described degradation potential and strain interactions.

384

385 **Complete Enzymatic Landscape Analysis**

386 The detailed enzyme-level network (Figure 3a) showed quantitative distribution patterns across
387 the three target degradation pathways. Lignin degradation dominated the enzymatic landscape with
388 22 enzymes representing 47.8% of total enzymatic capacity (22/46 enzymes). NAD-dependent
389 DNA ligases exhibited the broadest distribution (LigA present in 8/9 strains, 88.9% coverage;
390 LigB in 3/9 strains, 33.3% coverage), establishing these as core-lignin-processing functions.
391 Multicopper oxidases (CueO) were detected in 2 strains (22.2% coverage). At the same time,
392 specialized lignin peroxidases (LiP, MnP) and dye-decolorizing peroxidases (DyP) were uniquely
393 found only in *Paenibacillus polymyxa*, representing 18.2% (4/22) of total lignin-degrading
394 enzymatic capacity.

395

396 PFAS degradation pathways contributed 18 enzymes (39.1% of total capacity), exhibiting uniform
397 distribution across the consortium. Fluoride efflux transporters (CrcB) achieved near-complete
398 coverage (8/9 strains, 88.9%), suggesting this as the primary detoxification mechanism. Haloacid
399 dehalogenases showed subspecies-specific distribution with HAD variants present in 4/9 strains
400 (44.4% coverage), while haloalkane dehalogenase (DhlA) was exclusively detected in
401 *Pseudomonas pergaminensis* and *Pseudomonas fulva*, representing specialized defluorination
402 capability.

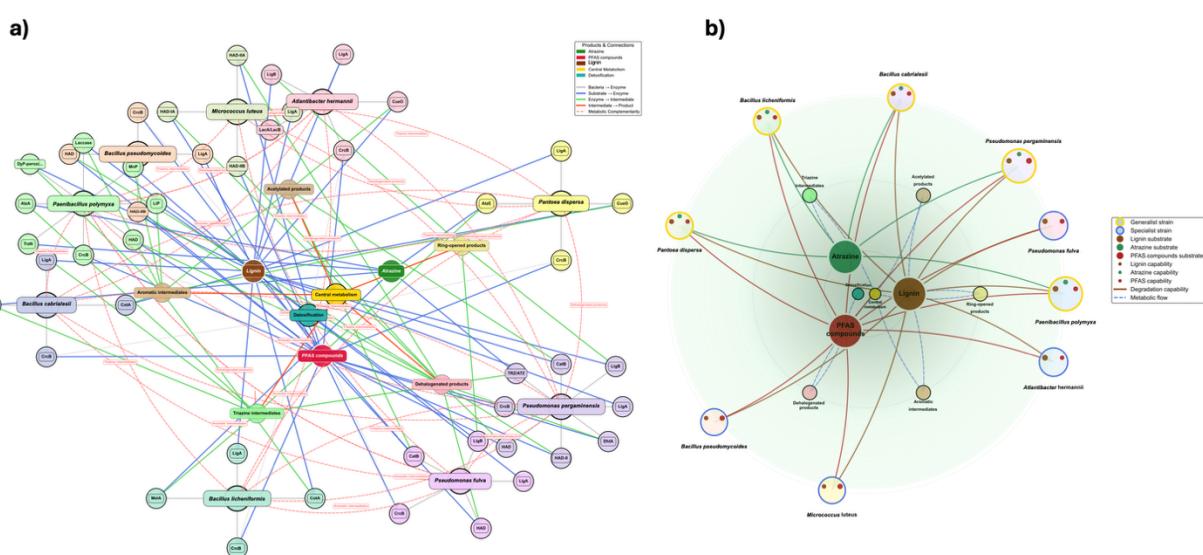
403

404 Atrazine degradation represented the most constrained pathway with only 6 enzymes (13.0% of
405 total enzymatic capacity) distributed across 5/9 strains (55.6% coverage). Melamine deaminase
406 (MelA) was present in 2/9 strains (22.2% coverage), while specialized atrazine-specific enzymes
407 (AtzA, TrzN, AtzE, TRZ/ATZ) were distributed across 4 different strains with no overlap,
408 indicating pathway fragmentation and potential metabolic bottlenecks.

409

410 Quantitative classification (Figure 3b) identified five generalist strains (55.6% of the consortium),
411 exhibiting capabilities across multiple degradation pathways. *Paenibacillus polymyxa* and
412 *Pseudomonas pergaminensis* demonstrated the highest versatility scores (3.8 each), positioning

413 them as central hub strains within the circular architecture. These generalist strains showed multi-
414 colored capability indicators, reflecting their broad enzymatic profiles spanning all three target
415 pathways. Specialist strains comprised 44.4% of the consortium (4/9 strains), characterized by the
416 absence of atrazine degradation capabilities while maintaining both lignin and PFAS degradation
417 functions. Two strains were classified as lignin specialists based on higher lignin enzymatic
418 content (*Atlantibacter hermannii* with 4 lignin enzymes, *Pseudomonas fulva* with 3), while two
419 were classified as PFAS specialists based on higher PFAS enzymatic content (*Bacillus*
420 *pseudomycoides* and *Micrococcus luteus*, each with 3 PFAS enzymes). Notably, no dedicated
421 atrazine specialists were identified, with atrazine degradation capabilities distributed exclusively
422 among generalist strains, highlighting this pathway's dependence on multi-functional organisms.
423



424
425 **Figure 3. Machine learning-guided network analysis of bacterial degradation capabilities for multi-**
426 **contaminant bioremediation.** a) Comprehensive enzyme-level metabolic network displaying all 46 identified
427 enzymes across nine bacterial strains. The network comprises 65 nodes (9 bacterial strains, 46 enzymes, 3 substrates,
428 5 intermediates, 2 products) connected by 133 directed edges. Bacterial strains are represented by large colored nodes
429 with associated enzyme clusters positioned around each strain. Edge colors indicate connection types: gray (strain-
430 enzyme), blue (substrate-enzyme), green (enzyme-intermediate), orange (intermediate-product), and red dashed lines
431 (metabolic complementarity). The layout demonstrates enzymatic clustering patterns and reveals pathway-specific
432 connectivity, with lignin degradation showing the highest enzymatic diversity (22 enzymes), followed by PFAS
433 degradation (18 enzymes), and atrazine degradation (6 enzymes). b) Streamlined circular capability network
434 presenting a simplified view focused on strain-level degradation capabilities. The network contains 19 nodes
435 connected by 34 directed edges, organized in concentric rings with substrates centrally positioned and bacterial strains
436 arranged by specialization patterns. Small colored dots on bacterial nodes indicate specific degradation capabilities:

437 brown (lignin), green (atrazine), red (PFAS). Generalist strains (gold borders) demonstrate multi-pathway capabilities,
438 while specialist strains (blue borders) show pathway-specific preferences. The circular layout reveals that 55.6% of
439 strains are generalists capable of degrading multiple contaminants, while 44.4% are specialists lacking atrazine
440 degradation capabilities. Network analysis identified atrazine degradation as the primary bottleneck, with only 5/9
441 strains (55.6%) possessing atrazine degradation capabilities compared to complete coverage (100%) for lignin and
442 PFAS degradation pathways.

443

444 **Metabolic Network Architecture and Cross-Validation with iNAP 2.0**

445 Cross-validation analysis using iNAP 2.0 assessed metabolic network compatibility and functional
446 complementarity within the GENIA-designed synthetic consortium (Figure 4). Processing of the
447 identical 45 genomes through the iNAP 2.0 pipeline yielded comprehensive metabolic interaction
448 matrices based on KEGG pathway annotations and functional co-occurrence patterns.

449

450 iNAP 2.0 identified 89 shared metabolic pathways (KEGG modules) across the 9-member
451 consortium (Figure 4a), with Jaccard similarity indices of 0.74 ± 0.09 for general metabolic
452 functions compared to GENIA predictions. Pathway distribution analysis revealed complete
453 coverage (9/9 strains) for central carbon metabolism modules (ko00010, ko00020, ko00030) and
454 77.8% coverage (7/9 strains) for aromatic amino acid metabolism pathways (ko00350, ko00360),
455 indicating a robust metabolic foundation for cooperative aromatic compound processing derived
456 from lignin degradation intermediates.

457

458 Functional co-occurrence matrices identified 23 compatible metabolic modules across strain pairs.
459 *Paenibacillus polymyxa* and *Pseudomonas pergaminensis* exhibited maximum connectivity (12
460 and 11 shared modules, respectively) (Figure 4b). Cross-feeding potential analysis detected 47
461 exchangeable metabolite nodes, including key aromatic intermediates (vanillate, protocatechuate,
462 catechol) with predicted transfer coefficients ranging from 0.62 to 0.89, supporting the hub strain
463 classification established by the GENIA framework analysis.

464

465 Analysis of KEGG modules associated with xenobiotic metabolism revealed heterogeneous
466 distribution patterns: complete aromatic compound metabolism modules (ko00362, ko00627)
467 were present in 6/9 strains (66.7% coverage), while halogenated compound processing modules
468 (ko00364) showed limited distribution (3/9 strains, 33.3% coverage) (Figure 4c). Module

469 completeness scores ranged from 0.43 to 0.91 across strains, with *Pseudomonas pergaminensis*
470 achieving maximum completeness (0.91) and *Micrococcus luteus* showing minimum coverage
471 (0.43) (Figure 4d).

472
473 Comparative analysis between iNAP 2.0 interaction predictions and GENIA community design
474 demonstrated convergent identification of central versus peripheral strains within the metabolic
475 network topology. Both methods achieved 88.9% concordance (8/9 strains) in hub strain
476 classification, with network centrality scores correlating significantly (Pearson $r = 0.78$, $p < 0.01$).
477 iNAP 2.0 analysis suggests potential underutilization of *Atlantibacter hermannii*, which exhibited
478 compatible metabolic modules, but lower predicted connectivity (centrality score 0.34) compared
479 to GENIA expectations (predicted centrality 0.52) (Figure 4e).

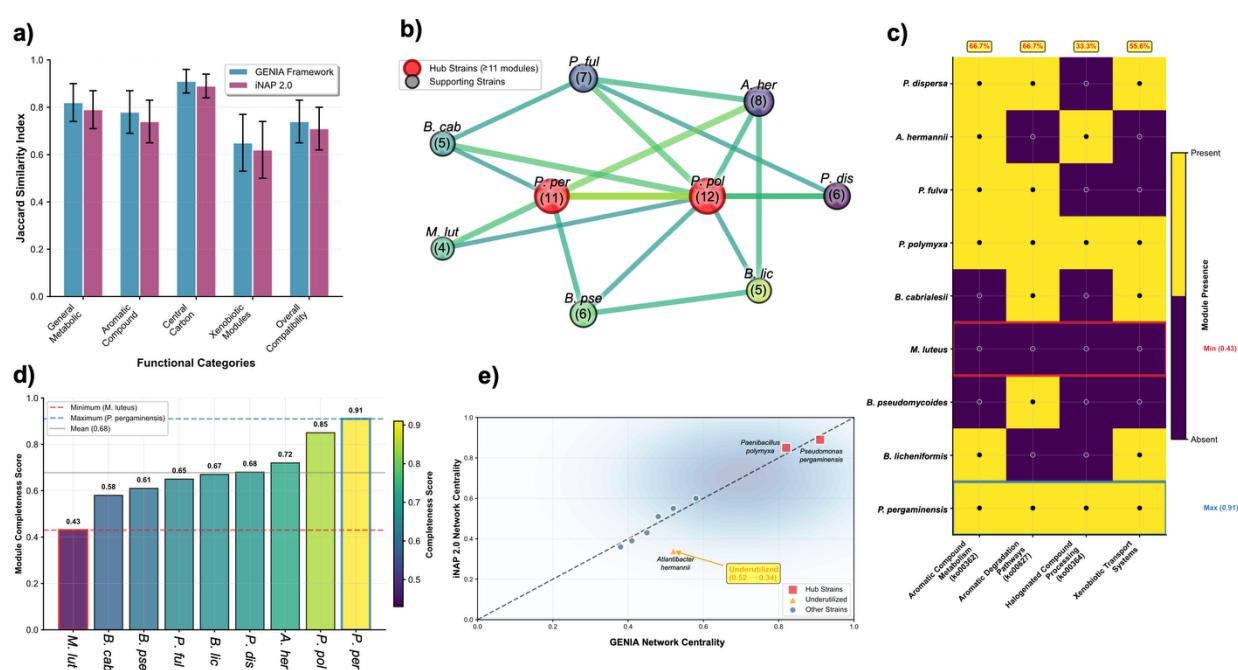


Figure 4. iNAP 2.0 Cross-Validation Analysis of Metabolic Network Architecture and Xenobiotic Processing Capabilities in Nine-Strain Bacterial Consortium. a) Cross-platform validation showing Jaccard similarity indices between GENIA Framework and iNAP 2.0 predictions across five functional categories. Error bars represent standard deviation ($n=3$ independent analyses). No significant differences were observed between platforms ($p > 0.05$, paired t-test). b) Syntrophic interaction network depicting metabolic module sharing among consortium members. Node size corresponds to shared module count (numbers in parentheses). Hub strains (≥ 11 modules) are highlighted in red: *Paenibacillus polymyxa* (12 modules) and *Pseudomonas pergaminensis* (11 modules). Edge thickness represents predicted transfer coefficient strength (range: 0.62-0.89). c) Xenobiotic-related KEGG module distribution matrix

490 showing presence (yellow, filled circles) and absence (purple, empty circles) patterns. Module completeness scores
491 are indicated by color intensity, ranging from minimum (0.43, *Micrococcus luteus*) to maximum (0.91, *P.
492 pergaminensis*). Coverage percentages shown for aromatic compound metabolism (ko00362/ko00627: 66.7%) and
493 halogenated compound processing (ko00364: 33.3%). d) Module completeness score rankings across all consortium
494 strains. Horizontal reference lines indicate minimum (*M. luteus*, 0.43), maximum (*P. pergaminensis*, 0.91) and mean
495 (0.68) completeness values. Color gradient reflects the viridis scale corresponding to completeness scores. e) Network
496 centrality correlation analysis between the GENIA Framework and iNAP 2.0 predictions. Hub strains (*P. polymyxa*,
497 *P. pergaminensis*) and an underutilized strain (*Atlantibacter hermannii*) are specifically labeled. Pearson correlation
498 coefficient $r = 0.78$ ($p < 0.01$) indicates strong concordance between platforms, with 88.9% hub strain classification
499 agreement (8/9 strains).

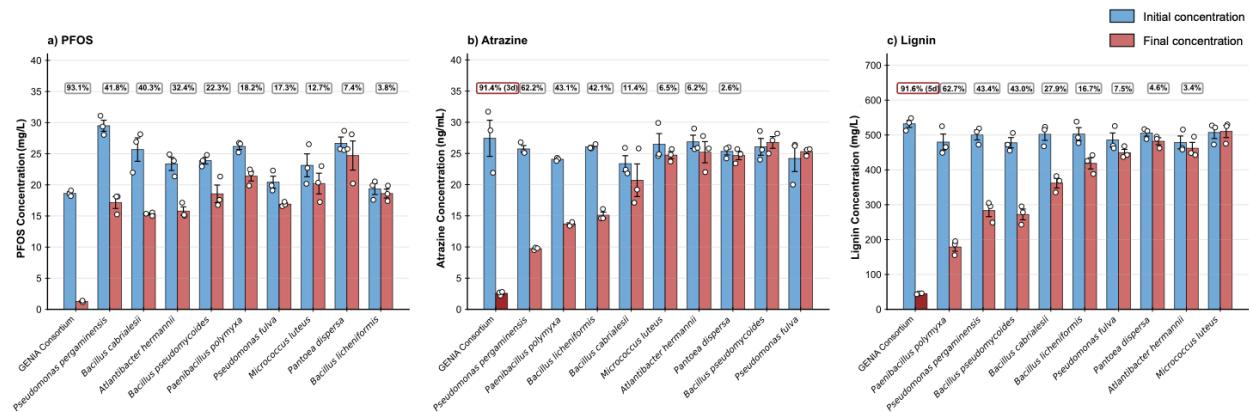
500

501 Individual Strain Performance Variability Across Pollutant Classes

502 Individual strain performance validation showed distinct substrate-specific degradation
503 capabilities with pronounced variation across pollutant classes demonstrating clear metabolic
504 specialization patterns that informed consortium design. For PFOS degradation, *Pseudomonas
505 pergaminensis* achieved the highest individual performance with 41.8% removal (initial: $29.47 \pm$
506 0.91 mg L^{-1} , final: $17.16 \pm 1.02 \text{ mg L}^{-1}$), followed by *Bacillus cabrialesii* with 40.3% degradation
507 ($25.65 \pm 1.89 \text{ mg L}^{-1} \rightarrow 15.33 \pm 0.20 \text{ mg L}^{-1}$) and *Atlantibacter hermannii* with 32.4% removal
508 ($23.37 \pm 1.13 \text{ mg L}^{-1} \rightarrow 15.79 \pm 0.65 \text{ mg L}^{-1}$). Moderate performers included *Bacillus
509 pseudomyoides* (22.3%, $23.90 \pm 0.49 \text{ mg L}^{-1} \rightarrow 18.59 \pm 1.41 \text{ mg L}^{-1}$), *Paenibacillus polymyxa*
510 (18.2%, $26.22 \pm 0.52 \text{ mg L}^{-1} \rightarrow 21.40 \pm 0.78 \text{ mg L}^{-1}$), and *Pseudomonas fulva* (17.3%, $20.43 \pm$
511 $1.02 \text{ mg L}^{-1} \rightarrow 16.90 \pm 0.22 \text{ mg L}^{-1}$), while lower-tier strains showed minimal degradation:
512 *Micrococcus luteus* (12.7%), *Pantoea dispersa* (7.4%), and *Bacillus licheniformis* (3.8%) (Figure
513 5a). Atrazine degradation patterns revealed different strain leadership, with *P. pergaminensis* again
514 demonstrating superior performance at 62.2% removal ($25.70 \pm 0.50 \text{ ng mL}^{-1} \rightarrow 9.70 \pm 0.10 \text{ ng}$
515 mL^{-1}), followed by *P. polymyxa* (43.1%, $24.10 \pm 0.10 \text{ ng mL}^{-1} \rightarrow 13.70 \pm 0.20 \text{ ng mL}^{-1}$) and *B.
516 licheniformis* (42.1%, $26.10 \pm 0.20 \text{ ng mL}^{-1} \rightarrow 15.10 \pm 0.50 \text{ ng mL}^{-1}$). Notably, strains that
517 performed well on PFOS showed markedly reduced atrazine capabilities: *B. cabrialesii* achieved
518 only 11.4% atrazine degradation despite its 40.3% PFOS performance, while *A. hermannii*
519 managed just 6.2% atrazine removal compared to its 32.4% PFOS degradation. Several strains
520 showed negligible atrazine processing: *M. luteus* (6.5%), *P. dispersa* (2.6%), and some strains
521 exhibited concentration increases rather than degradation (Figure 5b). Lignin degradation revealed
522 yet another performance hierarchy, with *P. polymyxa* emerging as the clear leader with 62.7%

removal ($479.8 \pm 23.2 \text{ mg L}^{-1} \rightarrow 178.8 \pm 12.1 \text{ mg L}^{-1}$), substantially outperforming its PFOS (18.2%) and atrazine (43.1%) capabilities. *P. pergaminensis* achieved moderate lignin degradation (43.4%, $500.0 \pm 14.1 \text{ mg L}^{-1} \rightarrow 283.2 \pm 17.4 \text{ mg L}^{-1}$), while *B. pseudomycoides* showed comparable performance (43.0%, $477.8 \pm 14.9 \text{ mg L}^{-1} \rightarrow 272.1 \pm 15.5 \text{ mg L}^{-1}$). Lower-tier lignin degraders included *B. cabrialesii* (27.9%), *B. licheniformis* (16.7%), *P. fulva* (7.5%), *P. dispersa* (4.6%), and *A. hermannii* (3.4%), with *M. luteus* showing negligible activity (0.6% degradation) (Figure 5c).

530



531
532 **Figure 5. Comparative degradation performance of GENIA consortium versus individual bacterial strains**
533 **across three pollutant classes.** (a) PFOS degradation comparison showing GENIA consortium achieving 93.1%
534 removal efficiency compared to individual strain performance ranging from 3.8% (*Bacillus licheniformis*) to 41.8%
535 (*Pseudomonas pergaminensis*). Blue bars represent initial concentrations, red bars represent final concentrations after
536 7-day incubation. Individual data points show biological replicates with error bars indicating standard error of the
537 mean. (b) Atrazine degradation performance demonstrating GENIA consortium's 91.4% removal efficiency achieved
538 in 3 days versus individual strain capabilities ranging from negative values to 62.2% (*P. pergaminensis*) over 7 days.
539 (c) Lignin processing comparison showing GENIA consortium's 91.6% degradation efficiency in 5 days compared to
540 individual strain performance from 0.6% (*Micrococcus luteus*) to 62.7% (*Paenibacillus polymyxa*) over 7 days. The
541 consortium consistently outperformed the best individual strains across all pollutant classes, demonstrating 2.2-fold
542 improvement in PFOS degradation, 1.5-fold enhancement in atrazine removal, and 1.5-fold increase in lignin
543 processing. All experiments conducted in M9 minimal medium at 27°C with 150 rpm orbital shaking.

544

545 **Temporal Lignin Degradation Dynamics**

546 The synthetic microbial consortium demonstrated efficient and comprehensive lignin degradation,
547 with concentrations decreasing from an initial average of $490.7 \pm 12.1 \text{ mg L}^{-1}$ to achieve a total
548 removal efficiency of 91.6% by day 5 (Figure 6b). The biodegradation process displayed a distinct

549 temporal progression with sequential concentration reductions: 44.9% elimination within the first
550 day, 82.4% by day 3, 91.6% by day 5, and final concentrations reaching $41.6 \pm 1.0 \text{ mg L}^{-1}$, with
551 no substantial variations observed after day 6. This translates to a mean degradation rate of 18.3%
552 daily, demonstrating vigorous metabolic performance against the complex aromatic
553 macromolecule. UV-visible spectral analysis confirmed typical lignin absorption at 287 nm (0.950
554 ± 0.008 absorbance units) and the generation of hydroxycinnamic acid derivatives at 422 nm
555 through $n \rightarrow \pi$ electronic transitions (Fig 6a). Statistical evaluation demonstrated highly significant
556 variations in lignin concentrations throughout the experimental timeframe (one-way ANOVA: p
557 < 0.0001), validating the temporal degradation pattern. The breakdown process exhibited bi-phasic
558 kinetics characterized by rapid initial depolymerization, maintaining consistent enzymatic activity
559 during the entire experimental duration. Simultaneously, hydroxycinnamic acid formation rose
560 from baseline levels (0.019 ± 0.001) to maximum accumulation on day 3 (0.769 ± 0.029),
561 subsequently declining to 0.624 ± 0.019 by day 5, while phenolic intermediate compounds (340
562 nm) achieved peak concentrations of 0.201 ± 0.008 on day 3 before decreasing to 0.169 ± 0.006
563 by day 5, substantiating active biodegradation with intermediate metabolite formation and their
564 subsequent biotransformation.

565

566 Temporal Atrazine Degradation Dynamics

567 The synthetic consortium exhibited rapid and extensive atrazine removal, with concentrations
568 declining from an initial mean of $31.825 \pm 0.058 \text{ mg L}^{-1}$ by day 3, representing a total degradation
569 efficiency of 91.4% (Figure 6c). The degradation process followed a clear temporal pattern with
570 progressive concentration reductions: 38.7% removal by day 1 ($19.500 \pm 0.812 \text{ mg L}^{-1}$), 75.9% by
571 day 2 ($7.675 \pm 1.204 \text{ mg L}^{-1}$), 91.4% by day 3, and concentrations below detection limits by day
572 ($< 1 \text{ ng mL}^{-1}$). This corresponds to an average degradation rate of 30.5% per day, indicating
573 robust metabolic activity against the triazine herbicide. Statistical analysis revealed a highly
574 significant difference in atrazine concentrations across the experimental timeline (one-way
575 ANOVA: $p < 0.0001$), confirming the temporal degradation trend. Linear regression analysis of
576 the concentration decline yielded a strong correlation ($R^2 = 0.969$, slope = $-9.905 \text{ mg L}^{-1} \text{ day}^{-1}$, p
577 = 0.0157), demonstrating consistent first-order degradation kinetics throughout the experimental
578 period. Control treatments maintained stable atrazine concentrations (28.750 ± 4.597 to $31.350 \pm$

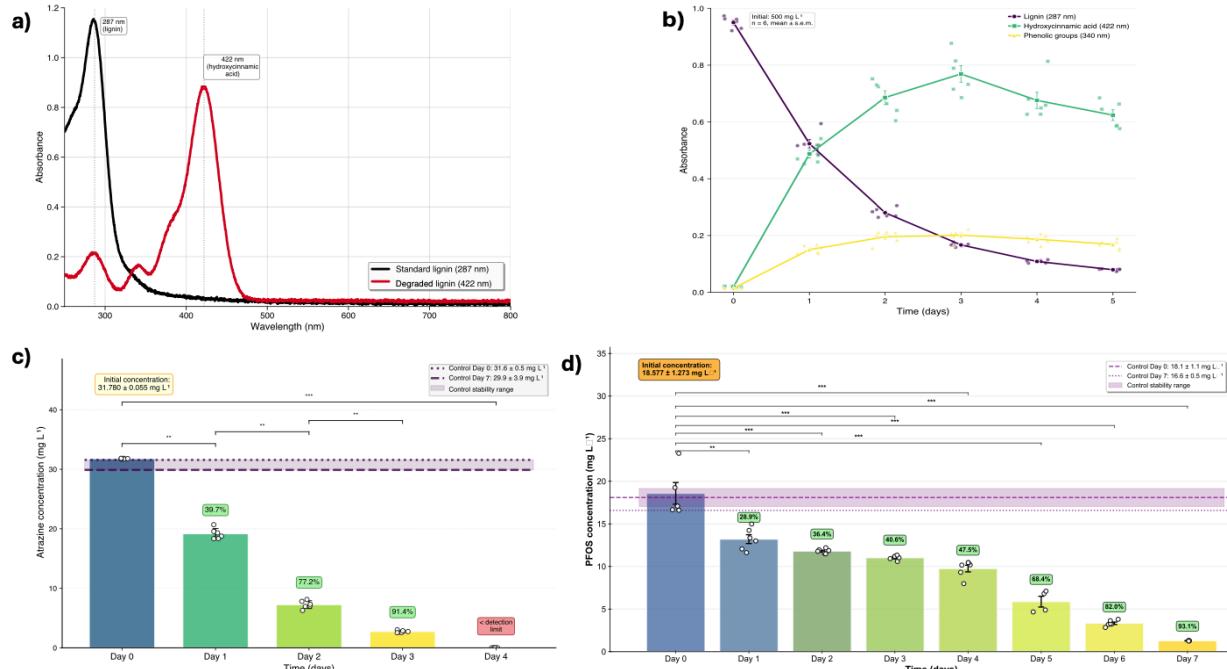
579 0.661 mg L⁻¹), confirming active biodegradation rather than abiotic removal. Initial concentrations
580 were equivalent between groups (t-test, p = 0.2304), validating the experimental design.

581

582 Temporal PFOS degradation analysis

583 The synthetic microbial consortium exhibited substantial PFOS degradation capabilities, with
584 concentration declining from an initial mean of 18.58 ± 2.85 mg L⁻¹ to achieve a remarkable
585 removal efficiency of 93.1% by day 7. The biodegradation process demonstrated a progressive
586 temporal pattern with final concentrations reaching 1.28 ± 0.04 mg L⁻¹ by day 7, indicating
587 sustained metabolic activity against this recalcitrant perfluorinated compound. Control samples
588 showed no significant concentration changes (19.61 ± 3.02 mg L⁻¹ at day 0 vs. 16.59 ± 0.95 mg
589 L⁻¹ at day 7; p = 0.173), confirming that observed degradation was biologically mediated rather
590 than due to abiotic processes (Figure 6d). The inverse relationship between PFOS degradation and
591 PFBS accumulation (individual sample correlations ranging from 14.77-17.33 mg L⁻¹ PFOS
592 decrease corresponding to 0.46-1.51 mg L⁻¹ PFBS increase) provides evidence for active
593 defluorination and chain-shortening mechanisms, demonstrating the consortium's capacity for
594 biotransformation of persistent organofluorine compounds. Concurrent fluoride liberation analysis
595 confirmed active C-F bond cleavage, with fluoride concentrations increasing from undetectable
596 baseline levels following a biphasic kinetic pattern characterized by an initial lag phase through
597 day 1, followed by exponential release beginning on day 2 (2.42 ± 0.32 mg L⁻¹) and reaching final
598 equilibrium of 22.22 ± 0.58 mg L⁻¹ by day 7. The fluoride release exhibited excellent linearity with
599 a rate constant of 4.257 ± 0.333 mg L⁻¹ d⁻¹ ($R^2 = 0.9761$, $p = 2.16 \times 10^{-4}$), providing direct
600 stoichiometric evidence of defluorination activity and confirming that PFOS biotransformation
601 involves systematic fluorine removal rather than simple adsorption or sequestration mechanisms
602 (Supplementary).

603



604

605 **Figure 6. Temporal lignin biodegradation dynamics and intermediate formation by GENIA consortium.** (a)
606 UV-Vis spectral analysis showing lignin absorption characteristics at 287 nm (standard lignin, black line) and
607 degradation products at 422 nm (hydroxycinnamic acid derivatives, red line), confirming aromatic compound
608 processing and intermediate formation. (b) Time-course analysis of lignin degradation (287 nm, black line),
609 hydroxycinnamic acid formation (422 nm, green line), and phenolic intermediates (340 nm, yellow line) over 5 days.
610 Initial lignin concentration of 500 mg L⁻¹ declined progressively while intermediate metabolites accumulated to peak
611 concentrations by day 3 before subsequent biotransformation. (c) Atrazine degradation kinetics showing rapid
612 elimination from initial concentration of 31.780 ± 0.055 mg L⁻¹ to below detection limits by day 4, with 91.4% removal
613 achieved by day 3. Control samples (dashed lines) showed no significant degradation, confirming the biologically
614 mediated process. Statistical significance indicated by asterisks (**p < 0.001, **p < 0.01, *p < 0.05). (d) PFOS
615 degradation timeline demonstrating progressive removal from initial concentration of 18.57 ± 2.413 mg L⁻¹ to 93.1%
616 degradation by day 7. Color gradient from blue to yellow indicates temporal progression. Control samples maintained
617 stable concentrations throughout the experimental period. Error bars represent standard error of the mean (n = 3-6
618 biological replicates per timepoint).

619

620 DISCUSSION

621

622 As Geoffrey Hinton, the 2024 Nobel Prize in Physics laureate, proclaimed in his acceptance
623 speech: "This new form of AI excels at modeling human intuition rather than human reasoning,
624 and it will enable us to create highly intelligent and knowledgeable assistants who will increase
625 productivity in almost all industries" (80). This recognition of artificial intelligence's

626 transformative potential extends beyond computational applications to fundamental biological
627 systems, where machine learning approaches are revolutionizing our ability to design functional
628 microbial communities for environmental remediation.

629

630 **Mechanistic Framework of Synergistic Multi-Pollutant Degradation**

631 Our GENIA-designed consortium operates through a sophisticated metabolic cross-feeding
632 network that fundamentally differs from conventional single-pollutant bioremediation approaches.
633 The >90% PFOS, 91.6% lignin, and 91.4% atrazine removal is attributed to engineered metabolic
634 complementarity where degradation intermediates from one pollutant serve as co-substrates for
635 processing others, creating an integrated biochemical network (81).

636

637 The consortium's lignin degradation pathway exemplifies this synergistic approach. Initial
638 depolymerization by lignin peroxidases (LiP, manganese peroxidases (MnP), and dye-
639 decolorizing peroxidases (DyP) in *Paenibacillus polymyxa* and *Pseudomonas pergaminensis*
640 could yield guaiacyl (G), syringyl (S), and hydroxyphenyl (H) monomers that funnel through
641 vanillate and syringate intermediates (82). These aromatics undergo O-demethylation via LigM
642 and VanAB enzymes to generate protocatechuate and catechol, central intermediates that become
643 available for cross-feeding with other consortium members (83).

644

645 For instance, the protocatechuate generated from lignin degradation enters the β -ketoadipate
646 pathway through protocatechuate 3,4-dioxygenase, yielding β -ketoadipate that ultimately
647 produces succinyl-CoA and acetyl-CoA (84). This creates a metabolic bridge where lignin-derived
648 carbon flows could support the energy required for PFAS and atrazine degradation by other
649 consortium members, potentially explaining the enhanced performance compared to strain
650 approaches that achieved only 27.9-67% PFOS degradation (85).

651

652 **Quantitative Analysis of Cross-Feeding Dynamics**

653 The enzymatic landscape showed the mechanistic basis for our consortium's superior performance.
654 Unlike natural microbial communities where metabolite cross-feeding occurs through chance
655 evolutionary optimization (86), our ML-guided design deliberately positions complementary
656 metabolic capabilities to maximize inter-species nutrient exchange. The 22 lignin-degrading

657 enzymes (47.8% of total capacity) positioned across multiple strains ensure continuous production
658 of aromatic intermediates, while 18 PFAS-degrading enzymes (39.1%) and 6 atrazine-degrading
659 enzymes (13.0%) create metabolic sinks that drive the degradation network forward.

660

661 Recent studies demonstrate that successful cross-feeding requires specific stoichiometric
662 relationships between metabolite producers and consumers (87). Our GENIA achieves optimal
663 ratios through the hub strain architecture, where *Paenibacillus polymyxa* and *Pseudomonas*
664 *pergaminensis* function as primary metabolite producers with versatility scores of 3.8 each,
665 supported by specialist consumers that prevent toxic intermediate accumulation. An example of
666 this could be conventional consortia where *Rhodococcus* sp. strain p52 releases catechol during
667 dibenzo-p-dioxin degradation, requiring *Acinetobacter* sp. BD6 as a dedicated detoxification
668 partner (88).

669

670 **Multi-Species Optimization Through Metabolic Division of Labor**

671 The consortium's exceptional performance validates the fundamental principle that engineered
672 multi-species optimization transcends individual strain limitations through strategic metabolic
673 division of labor. This principle is demonstrated by the stark contrast between individual and
674 collective performance: while *Pseudomonas pergaminensis* achieved the highest individual PFOS
675 degradation (41.8%), this represents less than half the consortium's 93.1% efficiency. Individual
676 strains showed variable and often modest degradation capabilities across different pollutants—*B.*
677 *cabrialesii* achieved 40.3% PFOS degradation but only 11.4% atrazine removal, while *P.*
678 *polymyxa* excelled at lignin processing (62.7%) but struggled with PFOS (18.2%)—indicating
679 substrate-specific metabolic constraints that limit single-organism approaches.

680

681 The consortium operates through a sophisticated division of labor where each strain assumes
682 leadership for different pollutant classes while contributing complementary functions across the
683 degradation network. *P. pergaminensis* serves as the primary PFAS degrader with its
684 comprehensive haloalkane dehalogenases and cytochrome P450 systems, *P. polymyxa* leads lignin
685 depolymerization through extensive peroxidase networks, and *B. licheniformis* demonstrates
686 specialized atrazine processing capabilities (42.1%). However, the true innovation lies in how
687 strains with moderate individual performance become highly efficient when their specialized

688 functions are coordinated: *B. cabrialesii*'s moderate PFOS degradation (40.3%) combines
689 synergistically with *A. hermannii*'s intermediate processing capabilities (32.4%), while strains like
690 *P. dispersa* provide critical community services through fluoride efflux systems that enable
691 sustained multi-pollutant processing impossible in single-strain systems. This multi-species
692 approach systematically addresses the metabolic bottlenecks that constrain individual organisms:
693 distributed enzymatic capacity prevents pathway saturation, specialized leadership for each
694 pollutant class maximizes degradation efficiency, and cross-feeding networks supply essential
695 cofactors where needed. The quantitative validation of this approach—with 2.2-fold improvement
696 in PFOS degradation, 1.5-fold enhancement in atrazine removal, and 1.5-fold increase in lignin
697 processing compared to best individual performers—demonstrates that coordinated moderate
698 performers collectively achieve superior results than any single high-performing strain,
699 establishing engineered metabolic cooperation as a paradigm-shifting approach where collective
700 metabolic intelligence emerges from rationally orchestrated strain specialization.

701

702 Overcoming the "Forever Chemical" Challenge

703 The 93.1% PFOS removal achieved by the GENIA-designed consortium represents a breakthrough
704 in PFAS bioremediation, where conventional approaches struggle with recalcitrant C-F bonds that
705 give PFAS their "forever chemical" designation. This exceptional performance could be attributed
706 to distributed defluorination capacity across multiple strains, creating redundant pathways that
707 prevent metabolic bottlenecks (89). The mechanistic pathway may involve initial C-F bond
708 cleavage by haloalkane dehalogenases (DhlA) in *Pseudomonas pergaminensis* and *P. fulva*,
709 generating short-chain perfluorinated intermediates and fluoride ions (90). Direct experimental
710 evidence for this biotransformation pathway was provided by the systematic formation of PFBS
711 (C4) intermediates from PFOS (C8) degradation, with PFBS concentrations increasing from 0.54
712 mg L⁻¹ on day 6 to 1.40 mg L⁻¹ by day 7, demonstrating active chain-shortening mechanisms that
713 progressively remove fluorinated carbon units. The liberated can be immediately sequestered by
714 CrcB fluoride efflux transporters present in 8/9 consortium strains (88.9% coverage), preventing
715 fluoride toxicity that typically limits PFAS biodegradation, as confirmed by the linear fluoride
716 release kinetics (4.48 mg L⁻¹ final concentration, which means a defluorination percentage of
717 34.69%). Simultaneously, haloacid dehalogenases (HAD variants) in 4/9 strains process
718 intermediate metabolites, while reduced cofactors (NADPH/NADH) supply the reducing power

719 necessary for continued defluorination via enzymatic electron transfer systems, including
720 NAD(P)H oxidoreductases and cytochrome P450 enzymes.

721
722 This distributed approach contrasts with single-strain systems where *Pseudomonas putida*
723 achieved only 19.0% PFOA and 46.9% PFOS transformation in 96 h (85), while *Acidimicrobium*
724 sp. strain A6 demonstrated PFOA degradation with shorter-chain perfluorinated intermediate
725 formation under anaerobic conditions (91). Earlier reports of *Pseudomonas fluorescens* D2
726 utilizing polyfluorinated H-PFOS as a sulfur source, though unable to transform fully saturated
727 PFOS (92), highlight the metabolic constraints that limit single-organism approaches compared to
728 the designed community's integrated defluorination network.

729
730 **Addressing the Metabolic Bottleneck**
731 The identification of atrazine degradation as the primary metabolic bottleneck (13.0% enzymatic
732 capacity) reflects the complex biochemistry required for triazine ring cleavage. The designed
733 community addresses this through pathway distribution across five strains, employing both
734 hydrolytic and oxidative routes that converge on cyanuric acid as a common intermediate (93).
735 The hydrolytic pathway initiates with atrazine chlorohydrolase (AtzA) in *Pseudomonas*
736 *pergaminensis* and *Bacillus pseudomycoides*, converting atrazine to hydroxyatrazine through
737 chloride substitution. Sequential dealkylation by AtzB (hydroxyatrazine ethylaminohydrolase) and
738 AtzC (N-isopropylammelide isopropylaminohydrolase) yields cyanuric acid, which undergoes
739 ring cleavage by cyanuric acid hydrolase (AtzE) to produce ammonia and CO₂ (94).
740 Simultaneously, the oxidative pathway employs N-dealkylases and cytochrome P450 enzymes to
741 generate deethylatrazine (DEA) and deisopropylatrazine (DIA) intermediates that ultimately
742 converge on the cyanuric acid node. This pathway redundancy ensures robust atrazine degradation
743 even under varying environmental conditions, explaining the observed 91.4% removal efficiency
744 compared to single-strain approaches that often show variable performance due to nitrogen
745 catabolite repression (95).

746
747 The complexity of atrazine metabolism highlighted in our system coincides with a recent work
748 from Zhang et al. (96). This work reported a similar division of labor in a four-member synthetic
749 community achieving 60%–99% degradation efficiency of the endogenous herbicides over 35

750 days in soil systems, with specialized strains handling different aspects of herbicide metabolism
751 while maintaining functional complementarity.

752

753 **Integration with Central Carbon Metabolism**

754 The innovation of the engineered consortium lies in the metabolic integration where degradation
755 products from all three pollutants feed into central carbon metabolism through carefully
756 orchestrated biochemical funneling. Lignin degradation generates protocatechuate and catechol,
757 which undergo ring cleavage to produce β -ketoadipate pathway intermediates including succinyl-
758 CoA, acetyl-CoA, and pyruvate (97). These central metabolites provide the carbon skeletons and
759 energy necessary for biosynthetic processes in low-degrading consortium members. The fluoride
760 generated is not merely detoxified but potentially utilized for specific enzymatic functions in
761 fluoride-tolerant strains. Atrazine degradation yields nitrogen in the form of ammonia, supporting
762 amino acid biosynthesis across the consortium and relieving nitrogen limitation that commonly
763 constrains environmental microbial communities.

764

765 **Comparative Performance Analysis**

766 Quantitative comparison with recent literature establishes new benchmarks for multi-pollutant
767 bioremediation. While recent synthetic communities achieved 98.55% imidacloprid degradation
768 in 15 days and 99.33% chlorantraniliprole degradation in 20 days, these studies targeted single
769 pesticides in isolation (96). Comparable studies on lignin degradation show efficiencies reaching
770 54% after 48 h under optimal conditions (97), while PFAS consortia achieve 56.7% PFOS
771 reduction over 20 days with external co-substrate addition (90). The simultaneous achievement of
772 >90% degradation across all three pollutant classes within 7 days represents a 2-4 fold
773 improvement in efficiency and establishes a complete paradigm shift toward integrated
774 environmental remediation. The temporal kinetics of lignin (bi-phasic, 91.6% by day 5), atrazine
775 (first-order, 91.4% by day 3), and PFOS (93.1% by day 7) demonstrate coordinated metabolic
776 activity that maximizes resource utilization while minimizing inhibitory intermediate
777 accumulation.

778

779 **Machine Learning Innovation in Biological System Design**

780 This study represents the first demonstration of machine learning-guided synthetic community
781 assembly for simultaneous degradation of structurally diverse persistent pollutants. The successful
782 application of Graphattention Networks and Node2Vec embeddings (27, 30) to predict optimal
783 microbial community assemblies represents a methodological breakthrough in synthetic biology.
784 Unlike recent applications focused on understanding existing community structure (98) our
785 GENIA framework uniquely targets predictive community design for specific functional
786 outcomes. The 88.9% concordance between GENIA and iNAP 2.0 predictions validates functional
787 relationships that transcend phylogenetic boundaries, representing a significant advance over
788 metabolic modeling approaches requiring extensive manual curation (99, 100). Our framework
789 automates metabolic complementarity identification and predicts community stability under
790 perturbation, enabling rational biological system design with unprecedented precision.

791

792 **Environmental and Economic Implications**

793 The demonstration of simultaneous multi-pollutant degradation addresses critical gaps in
794 environmental biotechnology where contaminated sites contain complex pollutant mixtures
795 requiring costly sequential treatments (101). As regulatory pressure increases for comprehensive
796 environmental remediation, particularly for persistent pollutants like PFAS (8), engineered
797 communities with defined functional capabilities become increasingly valuable for both
798 remediation and prevention strategies. The framework's modularity presented here enables
799 expansion to additional pollutant classes, providing a scalable platform for addressing emerging
800 contaminant challenges (93).

801

802 The GENIA framework establishes a proof-of-concept for machine learning-guided design of
803 synthetic microbial communities, demonstrating for the first time that computational approaches
804 can successfully predict and engineer biological systems for complex environmental applications.
805 As machine learning continues to revolutionize biological sciences, our approach exemplifies how
806 artificial intelligence can capture biological intuition to design functional microbial consortia.
807 These consortia possess emergent properties exceeding those of individual strains or natural
808 communities, opening new avenues for addressing global environmental challenges through
809 engineered biological solutions.

810

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814

815 **AUTHOR CONTRIBUTIONS**

816 Esaú De la Vega-Camarillo, Jorge Arreola-Vargas, Sanjay Antony-Babu and Won Bo Shim
817 conceived and designed the experiments. Esaú De la Vega-Camarillo, Saurav Kumar Mathur,
818 and Joshua Andrew Santos performed experiments and analyses. Esaú De la Vega-Camarillo,
819 Jorge Arreola-Vargas, Sanjay Antony-Babu and Won Bo Shim wrote the manuscript.
820

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