

Ambient temperature affects mammalian predator-prey interactions in an African savanna

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23 **Abstract**

- 24 1. Climate warming may alter predator-prey dynamics by changing predator attack speed and
25 prey escape speed, and by impacting the daily activity rhythms and habitat use of both
26 predators and prey. As there is no direct link between metabolic rates and temperature in
27 endotherms, it is widely assumed that climate warming will not affect interactions between
28 endothermic predators and prey.
- 29 2. This assumption ignores evidence that many endotherm species change their habitat use and
30 behaviour at high temperatures. Moreover, predator-prey chases generate metabolic heat,
31 which larger-bodied endotherms may be less able to dissipate, potentially altering chase
32 outcomes at high ambient temperatures.
- 33 3. We investigated how ambient temperature impacted predator-prey interactions in a three-
34 species system comprising one predator species, the African wild dog (*Lycaon pictus*), and
35 two prey species, dikdik (*Madoqua guentheri*) and impala (*Aepyceros melampus*).
- 36 4. We predicted that African wild dogs would kill more impala at high ambient temperatures via
37 three mechanisms. Firstly, we predicted that African wild dog activity would be reduced in the
38 day, but increase at night, enabling them to target impala, which are predictably located in
39 glades during this time period. Secondly, we predicted that all three species would seek shade
40 at high temperatures, using closed habitat where impala face higher predation rates. Third, we
41 predicted that difficulties of dissipating metabolic heat during chases at high ambient
42 temperatures would affect impala (40kg) more than dikdiks (5kg).
- 43 5. Although prey habitat selection and wild dog activity patterns changed with temperature as
44 predicted, we found that fewer, not more, wild dog scats contained impala remains when
45 ambient temperatures were high. Wild dogs spent less time hunting on hot days and, consistent
46 with models of optimal foraging under time constraints, appeared to select abundant, lower-
47 value dikdiks over rarer but higher-value impala.
- 48 6. Our findings show that weather can influence predator-prey interactions among endotherms,
49 and suggest that studies assuming otherwise should be interpreted with caution.

50
51 **Keywords:** African wild dog; antipredator behaviour; climate change; dikdik; habitat
52 selection; impala; predation; thermal tolerance
53

54 ***Introduction***

55 Climate change has far-reaching impacts on both species and ecosystems (Parmesan &
56 Yohe 2003). Meta-analyses suggest that many climate impacts on individual species operate
57 indirectly, through their effects on species interactions such as competition, herbivory,
58 predation, and parasitism (Cahill *et al.* 2012; Ockendon *et al.* 2014). As weather patterns will
59 shift under novel climate regimes, understanding how ambient temperature and other abiotic
60 processes influence species interactions will become critical to wildlife conservation in the
61 coming decades.

62 Climate is particularly critical for ectothermic species, which rely on environmental
63 temperatures to regulate their body temperature, meaning that high ambient temperatures
64 directly lead to changes in physiology, movement and behaviour. These individual
65 behavioural shifts can drive climate impacts on species interactions. For example, high
66 ambient temperatures allows ectotherms to move more rapidly, potentially making them
67 more efficient predators and more challenging prey, with cascading consequences for
68 community structure (Dell, Pawar & Savage 2014). It is usually assumed that ambient
69 temperature has no similar impacts on predation involving endotherms (e.g., Dell, Pawar &
70 Savage 2014; Harfoot *et al.* 2014); however, birds and mammals may respond to high
71 ambient temperatures by curtailing their foraging time (Ricklefs & Hainsworth 1968;
72 Quaglietta, Mira & Boitani 2018), switching from diurnal to nocturnal foraging (Levy *et al.*
73 2018), selecting different habitats (Austin 1976; Pigeon *et al.* 2016), and choosing different
74 foods (Doolan & Macdonald 1996; Garcia-Heras *et al.* 2017), all of which are likely to
75 influence their impacts on the species that they consume, or that consume them.

76 Predator-prey interactions involving large-bodied endotherms might be especially
77 sensitive to rising environmental temperatures, because the low surface-area-to-volume ratios
78 of large animals may make it difficult to dissipate heat generated by pursuing or avoiding

79 predation (Speakman & Krol 2010; Creel *et al.* 2016). Hence, with respect to the outcome of
80 predator-prey interactions, high ambient temperatures might favour endothermic prey that are
81 smaller than their endothermic predators, as the predators' large body size would mean that
82 they overheat faster than their smaller prey, leading to decreased predation rates on smaller
83 prey species. Similarly, high ambient temperatures may also benefit endothermic predators
84 that are smaller than their endothermic prey, as the large bodied prey would overheat faster
85 than the smaller bodied predator, leading to shorter chase times and higher hunting success
86 (Creel *et al.* 2016). Predation by and on large-bodied endotherms can have cascading effects
87 on community structure (e.g. Hebblewhite et al 2005), hence any effect of temperature on
88 endotherm predation might have broader impacts on ecological communities.

89 Using a three-species predator-prey system in an African savanna, we measured the
90 impact of ambient temperature on predator-prey interactions. Specifically, we examined how
91 ambient temperature affected the habitat use and activity patterns of a small-bodied prey
92 species (the 5kg Guenther's dikdik *Madoqua guentheri*), a larger-bodied prey species (the
93 40kg impala *Aepyceros melampus*), and their shared predator (the 23kg African wild dog
94 *Lycaon pictus*). At our study site in Kenya, impala and dikdiks are the two most abundant
95 ungulates, and their browsing shapes plant communities (Ford *et al.* 2014; Ford *et al.* 2015).
96 Wild dogs are crepuscular, and hunt by running prey down (Creel & Creel 1995); dikdiks and
97 impala together comprise 82% of the prey biomass that wild dogs consume at our study site
98 (Woodroffe et al. 2007).

99 We evaluated five hypotheses about the impacts of ambient temperature on predation by
100 wild dogs on impala and dikdik (Table 1). First, we hypothesised that predator-prey
101 interactions would be impacted by reduced predator activity on hot days (Table 1).
102 Previously, we have shown that wild dog activity declines as ambient temperatures increase
103 (Woodroffe et al 2017, Rabaiotti & Woodroffe 2019), which may reflect shorter hunting

periods (Woodroffe 2011b). Optimal foraging theory suggests that, when foraging time is limited, individuals should accept lower-value prey, rather than continuing to search for higher-value prey (Lucas 1983). Therefore, we predicted that, if wild dogs had reduced foraging time, they would increase predation on dikdik (which are encountered more frequently but are too small to feed a whole pack) over impala, which are larger but encountered less frequently (Woodroffe *et al.* 2007). We term this, Hypothesis 1, the ‘reduced daytime foraging hypothesis’.

Second, we hypothesised that predator-prey interactions would be impacted by greater predator nocturnality in hot weather (Table 1). We have shown previously that wild dogs are more nocturnal when daytime temperatures are high (Rabaiotti & Woodroffe 2019). While impala herds are scattered during daylight hours, at night they aggregate in clearings (“glades”) as a form of anti-predator behaviour (Augustine 2004; Otieno *et al.* 2019), and are therefore predictably located. We predicted that wild dogs would target glades at night to hunt impala, and termed this, Hypothesis 2, the ‘nocturnal prey concentration hypothesis’. We would expect that, if Hypothesis 2 was upheld, increased nocturnal wild dog activity at high ambient temperatures would lead to increased targeting of predictably located impala over less predictably located dikdik, increasing predation risk on impala (Table 1).

Third, we quantified whether habitat selection by both predators and prey might be affected by ambient temperatures, with consequences for predation. Many large mammals select shade in hotter weather (Mole *et al.* 2016; Pigeon *et al.* 2016), and we predicted that wild dogs, impala, and dikdik would do the same (Table 1). Further, we predicted that such a change would increase predation by wild dogs on impala, since impala face higher *per capita* predation rates in habitats characterised by higher woody cover (and thus shade; Ford *et al.* 2014). We term this, Hypothesis 3, the ‘shade-seeking hypothesis’.

Finally, we explored the potential consequences of both predators and prey overheating during chases. Creel *et al.* (2016) suggested that, during chases, wild dogs would overheat more slowly than their larger-bodied prey, leading to shorter and more successful chases at higher ambient temperatures. Larger species have more difficulty dissipating metabolic heat due to their lower surface-area-to-volume ratio (Peters 1986), potentially making such species more vulnerable to coursing predators. By extension, we predicted that wild dogs would overheat less rapidly than impala but more rapidly than dikdiks, leading to increased predation on impala relative to dikdiks at higher ambient temperatures (Table 1). We termed this, Hypothesis 4, the ‘prey overheating hypothesis’.

These hypothesised behavioural responses to high ambient temperatures generated contrasting sets of predictions about trophic interactions between wild dogs, impala, and dikdik (summarised in Table 1). By testing these predictions, we explored the behavioural mechanisms underlying whether, and how, ambient temperature influenced interactions between predators and prey, in an attempt to evaluate how climate change might affect this ecosystem through predator-prey interactions.

Methods

Study area

The focal area for this study was the Mpala Conservancy, Kenya (0°17' N, 36°53' E), a 200 km² area of semi-arid savanna managed jointly for livestock production and wildlife conservation. Mpala experiences little seasonal variation in ambient temperature; daily maximum temperature ranges from 20-37°C, and mean annual rainfall is 590mm, varying substantially both within and between years (Caylor, Gitonga & Martins 2017).

Mpala hosts 22 species of wild ungulate, of which dikdiks and impala are the two most abundant (Ford *et al.* 2015), at 139 individuals per km² and 20.3 individuals per km²

respectively (Augustine et al 2004). Mpala also supports five species of large carnivore in addition to wild dogs: lion (*Panthera leo*), leopard (*P. pardus*), cheetah (*Acinonyx jubatus*), spotted hyaena (*Crocuta crocuta*), and striped hyaena (*Hyaena hyaena*), all of which have dietary overlap with wild dogs (Frank, Woodroffe & Ogada 2005).

Data collection

We used Global Positioning System (GPS) collars to measure daytime and night-time activity, movement, and habitat use. GPS-collars (Savannah Tracking Ltd, Nairobi, Kenya) were fitted to 20 adult female impala (each from a different herd) between May and June 2011, as described in Ford *et al.* (2014). Each impala GPS-collar recorded a location every 20 minutes, for an average of 245 days. Likewise, GPS-collars (Savannah Tracking Ltd, Nairobi, Kenya) were fitted to 15 adult female dikdik between July 2010 and September 2011, as described in Ford & Goheen (2015). Each dikdik GPS-collar recorded a location every 10 minutes, for 18 days on average. Additionally, GPS collars (Vectronic Aerospace GmbH, Berlin, Germany) were fitted to 18 wild dogs in 13 packs between 2011 and 2016, using capture methods described in Woodroffe (2011a); all of these packs had home ranges overlapping Mpala, although none remained on Mpala year-round. Data were collected from one collar per pack at any one time. Each wild dog GPS-collar recorded locations at 01:00, 06:30, 07:00, 07:30, 08:00, 13:00, 18:00, 18:30, 19:30, and also recorded average acceleration in two planes (on a scale of 0 to 255 units) every five minutes, for 218 days on average.

We used faecal analysis to quantify the relative frequency of predation by wild dogs on impala and dikdik. Wild dog scats were collected during 2001-2004 across a 5,700km² study area which included Mpala Conservancy, and analysed as described in Woodroffe *et al.* (2007). We avoided pseudoreplication by including only one randomly-selected scat

collected from each pack on each occasion (Woodroffe *et al.* 2007). Because the GPS-collar data were collected ≥ 7 years after the scat samples, they came from the descendants of the individuals which deposited the scats. There were no major shifts in temperature and rainfall between the two time periods, however (Table S1).

Daily rainfall and daily maximum ambient temperature were collected from a weather station at Mpala Research Centre (Caylor, Gitonga & Martins 2017). Data on the timing of sunrise and sunset were taken from the *R* package *suncalc* (Agafonkin & Thieurmél 2017). Moonlight was calculated by multiplying percentage illumination by the number of hours the moon was in the sky between sunset and sunrise. Finally, we used GPS-collar data to identify periods when wild dog packs were denning (raising pups in a den, recognisable from the movement path which shows a “starburst” pattern of repeat visits to the same location, Woodroffe *et al.* 2017) as activity levels have been shown to increase during this period (Woodroffe *et al.* 2017). Mornings and evenings where individuals were babysitting (guarding the pups and therefore not hunting) were defined as days where the individual did not move from the den site during the morning or evening – these were removed from the analyses.

Identifying hunting periods of wild dogs

To estimate the time that wild dogs spent hunting, and to compare prey and predator behaviour at times when predation risk was highest, we used accelerometry data collected by GPS-collars to identify wild dog hunting bouts. First, we summed the two accelerometer measurements for each 5-minute period, to give an overall measure of activity. We then defined hunting bouts based on three criteria: (i) activity > 0 units for > 20 minutes; (ii) total activity during the activity bout > 500 units; (iii) followed by three or more consecutive records of 0 unit activity. These criteria excluded activity bouts which were too short to

represent hunting, or which related to less energetic behaviours, such as socialising. Criteria were derived using visual inspection of activity plots and examining the characteristics of activity bouts in time periods during which wild dogs are typically observed hunting. For each hunting bout identified in this way, we recorded start time, end time, duration (in minutes), and intensity (total activity divided by duration). The distributions of start and stop times are shown in Fig. S1. We classified hunting bouts as “morning”, “evening”, “night” or “midday”, based on their start and stop times, as detailed in Table S2 and Fig. S1. Bouts that spanned multiple time periods were excluded from the analysis. After using accelerometry data to identify specific hunting bouts, we categorised each wild dog location as falling inside or outside a hunting bout.

Categorising activity periods of prey species

We categorised dikdik and impala GPS collar locations into four time periods, reflecting diel variation in wild dog hunting activity and hence predation risk. We defined “morning” as the time period between sunrise (taken from the *R* package *suncalc*, Agafonkin & Thieurmel 2017) and the third quartile of end times for wild dog hunting bouts (approximately 3.5h after sunrise; Fig. S1). We classified impala and dikdik “evening” GPS-locations as those recorded between the first quartile of start times for hunting bouts (approximately 2hr10min before sunset; Fig. S1) and sunset (which fell between 1841h and 1912h). We categorised impala and dikdik GPS-locations as “midday” if they were recorded between the “morning” and “evening” periods, and “night” if they were recorded between sunset and sunrise.

Habitat use

We analysed habitat use from a vegetation map of Mpala created from a 2011 Quickbird satellite image (Digital Globe, Longmont, CO, USA) by Ford *et al.* (2014). We estimated the

woody cover (a measure of shadiness) associated with each impala, dikdik, and wild dog GPS-collar location as the proportion of woody cover within a circular area of radius 40m, centred on the collar location, calculated in *QGIS* (QGIS Development Team 2018).

We used the same habitat map to classify each GPS-collar location in relation to glades. For impala and dikdik, we recorded whether each location fell inside or outside a glade. As <1% of wild dog hunting locations fell inside glades, we recorded the distance to the nearest glade for each wild dog location, using the plugin “*NNJoin*” in *QGIS* (QGIS Development Team 2018).

Statistical analyses

We used multi-model inference to evaluate associations between wild dog hunting outcome variables (start time, stop time, duration, probability of occurrence, and intensity of morning, evening and night time hunts) and a range of explanatory variables (Table S3, Table S4). The reduced daytime foraging time hypothesis (Table 1, Hypothesis 1) predicted that wild dogs would spend less time hunting on hot days. To test this hypothesis, we split the activity data into morning, midday and evening. We analysed continuous outcome variables describing hunt duration, start time, stop time, and hunt intensity, as well as a binary outcome variable describing whether or not a hunt was recorded for each time period. Only 10% of hunting bouts occurred during the midday period (accounting for 2% of daily activity), so we considered these abnormal events and did not subject them to further statistical analyses. Morning and evening hunt periods were analysed separately as a number of the covariates (Table S3) were predicted to impact morning and evening hunts in different ways. For each outcome variable, we constructed Generalised Linear Mixed Models (GLMMs) using the package *nlme* (Pinheiro *et al.* 2015) in *R*, with Gaussian error distribution for the continuous outcome variables and binomial error distribution for the binary outcome variable. Each

model included the identity of individual animals and the date as random effects. Although our hypotheses concerned the potential effects of ambient temperature, we included several other explanatory variables known to influence activity patterns of wild dogs, namely denning status, rainfall, moonlight and, in models of night-time activity, time of moonrise and moonset. Detailed justifications for including each explanatory variable are shown in Table S3.

For each outcome variable, we built a series of statistical models using explanatory variables and biologically-meaningful interactions between explanatory variables, with individual identity as a random variable. We then used Akaike's Information Criterion (AIC) to compare models using the *R version 3.3.2* (version 3.3.2, R Core Team 2017) package *MuMIn* (Bartoń 2017). We considered all models with delta AIC scores within 2 units of the lowest AIC model (*i.e.*, $\Delta\text{AIC} < 2$) to have a high level of support (Burnham et al. 2002), referring to this array of models as the "top set". We used model averaging (Burnham et al. 2002) to estimate the effect on the outcome variable of each explanatory variable in the top set.

In the nocturnal prey selection hypothesis (Table 1, Hypothesis 2), we predicted that impala would use glades more at night (Table 1, 2a: nocturnal prey concentration). To test this hypothesis, we calculated the proportion of each impala's locations falling within glades during the morning, midday, evening, and night periods for each 24-h period. We used mixed effects models with in glade (1) or not (0) as the independent variable and time period (day (morning, evening and midday combined) or night) as the explanatory variable and individual identity and date as random variables to test whether prey species were significantly more likely to be found in glades at night than during the day.

In the nocturnal prey selection hypothesis, we also predicted that wild dogs would be more likely to hunt at night following hot days (Table 1, 2b: increased night-time hunting). To test this hypothesis, we analysed continuous outcome variables describing the duration, start time, stop time, and intensity of hunts occurring between sunset and sunrise, as well as a binary outcome variable describing whether or not a hunt was recorded each night. These outcome variables were compared with the same explanatory variables used in Hypothesis 1 except that, as nocturnal periods span dates, the maximum ambient temperature for the proceeding daytime period was used. As in the models of daytime hunting behaviour (Hypothesis 1), we constructed Generalised Linear Mixed Models (GLMMs), with Gaussian error distribution for the continuous outcome variables and binomial error distribution for the binary outcome variable, for each outcome variable using the package *nlme* (Pinheiro *et al.* 2015) in *R* (R Core Team 2015), and each model included the identity of individual animals as a random effect. Model selection was performed using model averaging in the same way as when investigating daytime hunts (Hypothesis 1).

In the nocturnal prey selection hypothesis, we also predicted that wild dogs would preferentially target glades at night (Table 1, 2c: night-time habitat selection). To test this hypothesis, we calculated the mean distance to the nearest glade for each night-time hunt period. As when investigating prey use of glades, we analysed this outcome variable using a mixed effects model with time period (day (morning, evening and midday combined) or night) as the explanatory variable and individual identity and date as random variables, in order to test whether wild dogs were significantly closer to glades at night than during the day.

In the shade-seeking hypothesis (Table 1, Hypothesis 3) we predicted that wild dogs, impala, and dikdiks would increase their use of shaded habitat at high ambient temperatures. To test this hypothesis, we constructed a series of models with use of woody cover as the

outcome variable. To avoid pseudoreplication, we averaged the woody cover values for each individual across each morning, midday, evening or night-time period. For wild dogs, we only included measurements of woody cover at locations taken during hunting bouts. We analysed these outcome variables using GLMMs with individual identity as a random effect, building a separate array of models for each time of day, and for all times of day together. Candidate explanatory variables were the same as for the analyses of wild dog activity (i.e., temperature, rainfall, moonlight, and, for wild dogs, denning), but also included a variable describing rainfall phase (Table S4), which was predicted to impact herbivore behaviour and location (Goheen et al 2013). Following Ford *et al.* (2014), we considered days to fall within “wet phases” if >50mm of rain had fallen in the previous four weeks, and classed all other days as falling in “dry phases”. In constructing the array of models for multi-model inference, this rainfall phase variable was never included in the same model as daily rainfall, because the two were correlated. Model selection was performed using model averaging in the same way as when investigating wild dog hunting (Hypothesis 1 and 2b).

The prey overheating hypothesis (Table 1, Hypothesis 4) could not be tested directly from our datasets, because our GPS-collar locations were recorded too infrequently to estimate chase speeds, so our evaluation of this scenario relied on testing the prediction that wild dogs killed impala more frequently on hot days (Table 1). This outcome was also predicted under Hypotheses 2 and 3, however, and could not be used to specifically support Hypothesis 4. We tested this prediction by using a GLM with binomial error distribution to analyse whether or not wild dog scats contained impala remains. In this model, candidate explanatory variables were ambient temperature during the previous seven days (to account for delays between a scat being deposited and collected), and land use (as a previous study showed that impala were consumed far less frequently on community land, relative to private land, Woodroffe *et al.* 2007). Pack or individual identity were often unknown for wild dog scats, so these models

did not include random effects. Model selection was performed using model averaging in the same way as when investigating wild dog hunting (Hypothesis 1 and 2b) and woody cover selection (Hypothesis 3).

Results

Daily movement patterns

African wild dogs showed a strongly crepuscular activity pattern, with hunts and daily activity concentrated in the morning and evening time periods (Fig. 1). Sixteen percent of wild dog hunting periods, and 17% of total daily activity, occurred at night (Fig. 2).

Impala had the highest levels of displacement between GPS locations at night ($1.24\text{km} \pm 2.11(\text{s.d.})$) and lower displacement distances in the morning ($0.90\text{km} \pm 1.4$), daytime ($0.97\text{km} \pm 1.7$) and evening ($0.89\text{km} \pm 1.3$) time periods.

Dikdik had similar levels of displacement between GPS locations throughout the day, with the higher levels of displacement in the morning ($0.27\text{km} \pm 0.3$) and at night ($0.26\text{km} \pm 0.27$) and slightly lower displacement distances in the daytime ($0.20\text{km} \pm 0.27$) and evening ($0.19\text{km} \pm 0.23$) time periods.

Associations between ambient temperature and wild dog hunting

Consistent with the reduced daytime foraging time hypothesis (Hypothesis 1), wild dogs' daytime hunting periods were shorter in both the morning and evening time periods at high ambient temperatures (Table 2, Fig. 3d). These shorter hunting periods reflected earlier start and stop times in the morning (Table S5) and later start times in the evening (Table S6). In addition to being shorter, both morning and evening hunts entailed less intense activity on hotter days (Table S5, Table S6). Evening hunts were less likely to occur at all on days with higher ambient temperatures (Table S6), though there was no such effect on morning hunts

(Table S5). Rainfall may have mitigated the effects of high ambient temperatures, with rainfall:temperature interactions included in some of the top models for hunt duration (Table S5, Table S6). Packs were consistently more active during daytime when they were denning (Table S5, Table S6).

Similarly, and as predicted under the increased nocturnal prey selection hypothesis (Hypothesis 2), wild dogs were more likely to hunt at night following daytime periods with high ambient temperatures (Table 2). There was also some evidence that the duration and intensity of night-time hunts increased following hot days (Table 2). Nocturnal hunting activity was increased at higher levels of moonlight (Table 2, Table S6), with corresponding reductions in daytime activity on dates with high moonlight indices (Table 2, Table S5, Table S6).

Habitat selection

The three species differed in their use of woody cover, with impala using the most open areas and wild dogs using the areas with the highest woody cover (Fig. 3a). Impala use of open areas was most pronounced at night, and during the morning period, whereas wild dogs used the wooded areas to a greater extent at night (Fig. 3a). The use of woody cover by dikdiks was relatively consistent throughout the day (Fig. 3a).

As predicted under the shade-seeking hypothesis (Hypothesis 3), impala selected areas with denser woody cover on hotter days, during morning and midday periods, although not during the evening (Table 3). In contrast, there was no evidence for dikdiks selecting woody cover based on ambient temperature, and also no evidence of such selection by hunting wild dog packs (Table 3).

The three species also varied in their use of glades. Consistent with the nocturnal prey selection hypothesis (Hypothesis 2a), impala were significantly more likely to be located in

glades at night than in the day (Table 4, Fig. 3c), although there was no such pattern for dikdiks (Table 4, Fig. 3c). In contrast with predictions under Hypothesis 2c, however, wild dogs were significantly further from glades when hunting at night than when hunting in the day (Table 4, Fig. 3b).

Prey selection

Among 795 wild dog scats, 71 (9%) contained impala remains and 609 (77%) contained dikdik remains. As predicted under the restricted foraging time hypothesis (Hypothesis 1) but not the other three hypotheses, wild dog scats were less likely to contain impala remains when ambient temperatures had been higher in the previous seven days (Table 5, Fig. 3e).

Discussion

Our analyses revealed clear associations between ambient temperature and the behaviour of both predator and prey species, which appeared to influence trophic interactions. Our findings were most consistent with the restricted foraging time hypothesis (Hypothesis 1), under which we predicted that, on hot days, wild dogs would spend less of the daylight period hunting, and therefore prefer abundant small prey over larger but less abundant prey. Consistent with these predictions, we found that wild dogs spent less time hunting during daytime hours (Fig. 3d, Table S5, Table S6). Previously we posited that, on average, dikdik and impala would yield comparable returns, because the greater energy intake achievable by hunting impala was offset by the shorter travel distances associated with hunting dikdik (Woodroffe *et al.* 2007). However, optimal foraging theory predicts that a predator with limited time should select more abundant but lower-value prey (such as dikdiks) rather than waiting to locate rarer, higher value prey (such as impala, Lucas 1983).

High ambient temperatures would therefore be expected to favour wild dogs eating dikdiks more than impala, and our observations were consistent with this prediction (Table 5).

Under the nocturnal prey selection hypothesis (Hypothesis 2), we predicted that, when ambient temperatures were high, wild dogs would hunt at night, targeting impala which are predictably located in glades. Although wild dogs hunted more often at night in hot weather (Table 2), and impala were located in glades at night (Table 4, Fig. 3c), there was no evidence that wild dogs targeted impala at night. Wild dogs were no closer to glades at night than at other times of day, were no closer to glades on nights when ambient daytime temperatures had been high (Table 4), and were less likely to consume impala in hot weather (Table 5, Fig. 3e). Hence, patterns of predator behaviour and prey selection were not consistent with Hypothesis 2.

Under the shade-seeking hypothesis (Hypothesis 3), we predicted that all three species would increase their daytime use of woody cover at high ambient temperatures. However, we found that only impala did so (Table 3). Of the three species, impala used the most open habitat (Fig. 3a, b and c), which may have resulted in a greater need to seek shade at high ambient temperatures. Physiological studies suggest that dikdiks are dependent upon shade to thermoregulate (Kamau & Maloiy 1985) but, being small-bodied and less gregarious than impala, they may be able to use small patches of shade without moving into denser habitat. There was no evidence that wild dogs hunted in denser cover on hot days (Table 3), perhaps because hunting periods occurred before and after the hottest times of day (Fig. 3a). The tendency of impala to move into denser cover at high ambient temperatures would be expected to increase their risk of being killed, since wild dogs typically occupied denser cover (Fig. 3a), and impala are more likely to be killed in denser cover (Ford *et al.* 2014). Nevertheless, we found that impala remains were less likely to be found in wild dog scats following periods of high ambient temperature (Table 5). Hence, although impala behaviour

427 changed in line with the predictions of the shade-seeking hypothesis (Hypothesis 3), this
428 change did not generate the predicted impact on predation.

429 Under the prey overheating hypothesis (Hypothesis 4) we predicted that high ambient
430 temperatures would reduce chase distances for impala and increase them for dikdiks, as
431 larger-bodied animals would be slowed down by the heat to a greater extent than smaller-
432 bodied animals (Creel *et al.* 2016). We were not able to test this hypothesis directly because
433 our monitoring methods did not allow us to measure chase distance or speed. However, under
434 this scenario we also predicted that predation upon impala would increase, relative to
435 predation on dikdiks, when ambient temperatures were high. Our observations showed the
436 opposite pattern (Table 5, Fig. 3e); hence, a key prediction of the prey overheating hypothesis
437 was not upheld by our observations. Dikdiks make up a higher proportion of prey consumed
438 by wild dogs than at other sites and are not typically considered to be within the species'
439 range of preferred prey sizes (Clements *et al.* 2014). The prey overheating hypothesis may be
440 more relevant in areas that have greater fluctuations in temperature (e.g. as seen in other parts
441 of wild dogs species range) or where larger prey are more abundant (e.g., wildebeest).

442 Other evidence also favours the reduced foraging time hypothesis (Hypothesis 1) over
443 the other hypotheses. Under Hypothesis 1, we would expect energy intake by wild dogs to be
444 reduced at high ambient temperatures, due to constrained foraging time and a consequent
445 acceptance of lower-energy prey. This prediction is consistent with our observations that this
446 study population experiences higher mortality and lower reproductive success at high
447 ambient temperatures (Woodroffe, Groom & McNutt 2017; Rabaiotti *et al.* in review). In
448 contrast, under the other three scenarios, wild dogs' food intake (and thus potentially their
449 survival and reproductive success), would be expected to improve at high temperatures,
450 because impala (higher-value prey) would be more accessible due to being predictably
451 located (Hypothesis 2), in dense cover where they are vulnerable to predators (Hypothesis 3),

or more easily captured due to their tendency to overheat during high speed chases (Hypothesis 4). Hence, while demographic patterns cannot confirm the reduced foraging time hypothesis (Hypothesis 1) as the most likely mechanism whereby ambient temperature influences predator-prey interactions in this system, they do contribute to refuting Hypotheses 2-4.

Our findings suggest that climate change might have two wider impacts through its effect on predation by wild dogs. First, we have shown previously that wild dogs suppressed dikdik numbers, and dikdik browsing influenced tree abundance (Ford *et al.* 2015). Although we found that wild dog predation on dikdiks did not trigger cascading effects on vegetation (Ford *et al.* 2015), rising environmental temperatures would be expected to intensify wild dog predation on dikdiks, which might generate wider impacts on community structure.

Second, our findings suggest that wild dog populations' resilience in the face of climate change might be affected by the abundance of small, abundant prey. Our study site is unusual both in the density of dikdiks it supports (Augustine 2010; Ford *et al.* 2015) and the degree to which wild dogs rely on dikdiks as prey (Woodroffe *et al.* 2007). Across much of the geographic range of wild dogs, dikdiks are altogether absent (Kingdon & Hoffman 2013), and impala comprise the bulk of wild dog diets (Creel, Mills & McNutt 2004; Mbizah, Marino & Groom 2012). Without abundant small prey to hunt on hot days, we would expect wild dogs to more frequently make no kill at all, with potentially greater impacts on energy intake than may occur at our study site. Low reproductive success and high mortality at high ambient temperatures have been reported for wild dogs from two sites where impala are wild dogs' principal prey (Woodroffe, Groom & McNutt 2017; Rabaiotti *et al.* In Review), although site-specific conditions complicate any comparison of the magnitude of temperature effects at the different sites.

Our results help to explain the negative impacts of high ambient temperatures on the survival and reproductive success of the African wild dog, an endangered species (Woodroffe & Sillero-Zubiri 2013). Our findings suggest that prey diversity may help to buffer wild dog populations against the effects of climate change, because abundant prey are readily located even when hunting time is constrained. Measures which maintain prey densities and diversity (such as limiting offtake by people) may help to conserve wild dogs in a warming climate. However, interventions which artificially raise prey densities (such as installation of waterholes) may risk increasing the densities of wild dogs' predators (Creel & Creel 1996; Mills & Gorman 1997), and might have harmful consequences for wild dog conservation.

More generally, we have shown that ambient temperature can influence patterns of predation, even when both predator and prey are tropical endotherms. This finding contrasts with the assumptions of several general ecosystem and community ecology models (e.g., Dell, Pawar & Savage 2014, Harfoot et al 2014), and suggests that such models may not accurately represent the responses of ecological communities in which endotherms play important roles. However, our findings also highlight the difficulty of accurately predicting exactly how environmental temperature would be expected to influence predation involving endotherms: all four of the hypotheses that we investigated were plausible, but they generated conflicting predictions about how predation on impala and dikdiks might vary in response to weather conditions, and only empirical testing indicated the true direction of the impact. Further investigations of how temperature-induced changes in the behaviour of predator and prey species together influence trophic interactions would help to build a more general picture of the relationship between ambient temperature and endotherm predation within ecological communities.

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Author Contributions

D. Rabaiotti, Rosie Woodroffe and Adam Ford, designed the research questions. D. Rabaiotti organised the datasets, designed and carried out the final analyses and wrote the paper. Rosie Woodroffe oversaw data collection on wild dogs and contributed to the writing of the paper. Adam Ford and Jacob Goheen collected the dikdik and impala data and contributed to the writing of the paper. Ben Chapple helped design and carry out the analyses on the impala and dikdik data and African wild dog habitat use. Sophie Morrill assisted in designing and carrying out the analyses on wild dog hunt times.

Data Availability Statement

Should the article be accepted, the data will be deposited on Dryad and the DOI included at the end of the article

References

- Agafonkin, V. & Thieurmél, B. (2017) *suncalc: compute sun position, sunlight phases, moon position and lunar phase*. <https://CRAN.R-project.org/package=suncalc>.
- Augustine, D.J. (2004) Influence of cattle management on habitat selection by impala on central Kenyan rangeland. *Journal of Wildlife Management*, 68(4), 916-923.
- Augustine, D.J. and McNaughton, S.J., (2004). Regulation of shrub dynamics by native browsing ungulates on East African rangeland. *Journal of Applied Ecology*, 41(1), pp.45-58.
- Augustine, D.J., 2010. Response of native ungulates to drought in semi-arid Kenyan rangeland. *African Journal of Ecology*, 48(4), pp.1009–1020. Available at: <http://doi.wiley.com/10.1111/j.1365-2028.2010.01207.x>.
- Austin, G.T. (1976) Behavioral adaptations of verdin to desert. *Auk*, 93(2), 245-262.
- Bartoń, K. (2017) *MuMIn: Multi-Model Inference*. <https://CRAN.R-project.org/package=MuumIn>.
- Burnham, K.P., Anderson, D.R. & Burnham, K.P., 2002. *Model selection and multimodel inference : a practical information-theoretic approach*, Springer.
- Cahill, A.E., Aiello-Lammens, M.E., Fisher-Reid, M.C., Hua, X., Karanewsky, C.J., Ryu, H.Y., Sbeglia, G.C., Spagnolo, F., Waldron, J.B., Warsi, O. & Wiens, J.J. (2012) How does climate change cause extinction? *Proceedings of the Royal Society B-Biological Sciences*, 280(1750), 20121890.
- Caylor, K.K., Gitonga, J. & Martins, D.J. (2017) Mpala Research Centre Meteorological and Hydrological Dataset Mpala Research Centre, Kenya.
- Clements, H.S., Tambling, C.J., Hayward, M.W., Kerley, G.I.H., 2014. An objective approach to determining the weight ranges of prey preferred by and accessible to the five large African carnivores. *PLoS ONE* 9(7), e101054.

547 Cozzi, G., Broekhuis, F., McNutt, J.W., Turnbull, L, MacDonald, D and Stable, B., 2012. Fear
 548 of the dark or dinner by moonlight ? Reduced temporal partitioning among Africa ' s large
 549 carnivores *Ecology*, 93(12), pp.2590–2599.

550 Creel, S. & Creel, N.M. (1995) Communal hunting and pack size in African wild dogs,
 551 *Lycaon pictus*. *Animal Behaviour*, 50(5), 1325-1339.

552 Creel, S., Creel, N.M., Creel, A.M. & Creel, B.M. (2016) Hunting on a hot day: effects of
 553 temperature on interactions between African wild dogs and their prey. *Ecology*. 97(11),
 554 2910-2916.

555 Creel, S., Mills, M.G.L. & McNutt, J.W. (2004) Demography and population dynamics of
 556 African wild dogs in three critical populations. *The biology & conservation of wild canids* (eds
 557 D.W. Macdonald & C. Sillero-Zubiri), pp. 337-350. Oxford University Press, Oxford.

558 Creel, S.R. & Creel, N.M. (1996) Limitation of African wild dogs by competition with
 559 larger carnivores. *Conservation Biology*, **10(2)**, 1-15.

560 Davies-Mostert, H.T., Mills, M.G., Kent, V. and Macdonald, D.W., 2010. Reducing potential
 561 sources of sampling bias when quantifying the diet of the African wild dog through scat
 562 analysis. *African Journal of Wildlife Research*, 40(2), pp.105-113.

563 Dell, A.I., Pawar, S. & Savage, V. (2014) Temperature dependence of trophic interactions
 564 are driven by asymmetry of species responses and foraging strategy. *Journal of Animal*
 565 *Ecology*, **83(1)**, 70-84.

566 Doolan, S.P. & Macdonald, D.W. (1996) Diet and foraging behaviour of group-living
 567 meerkats, *Suricata suricatta*, in the southern Kalahari. *Journal of Zoology*, **239(4)**, 697-716.

568 Ford, A.T. & Goheen, J.R. (2015) An experimental study on risk effects in a dwarf
 569 antelope, *Madoqua guentheri*. *Journal of Mammalogy*, **96(5)**, 918-926.

570 Ford, A.T., Goheen, J.R., Augustine, D.J., Kinnaird, M.F., O'Brien, T.G., Palmer, T.M., Pringle,
 571 R.M. & Woodroffe, R. (2015) Recovery of African wild dogs suppresses prey but does not
 572 trigger a trophic cascade. *Ecology*, **96(10)**, 2705-2714.

573 Ford, A.T., Goheen, J.R., Otieno, T.O., Bidner, L., Isbell, L.A., Palmer, T.M., Ward, D.,
574 Woodroffe, R. & Pringle, R.M. (2014) Large carnivores make savanna tree communities less
575 thorny. *Science*, **346**(6207), 346-349.

576 Frank, L.G., Woodroffe, R. & Ogada, M.O. (2005) People and predators in Laikipia District,
577 Kenya. *People and wildlife - Conflict or coexistence?* (eds R. Woodroffe, S. Thirgood & A.R.
578 Rabinowitz), pp. 286-304. Cambridge University Press, Cambridge.

579 Funston, P.J., Mills, M.G.L. & Biggs, H.C. (2001) Factors affecting the hunting success of
580 male and female lions in the Kruger National Park. *Journal of Zoology*, **253**(4), 419-431.

581 Garcia-Heras, M.S., Mougeot, F., Simmons, R.E. & Arroyo, B. (2017) Regional and temporal
582 variation in diet and provisioning rates suggest weather limits prey availability for an
583 endangered raptor. *Ibis*, **159**(3), 567-579.

584 Goheen, J.R., Palmer, T.M., Charles, G.K., Helgen, K.M., Kinyua, S.N., Maclean, J.E., Turner,
585 B.L., Young, H.S. and Pringle, R.M., 2013. Piecewise disassembly of a large-herbivore
586 community across a rainfall gradient: the UHURU experiment. *PloS one*, 8(2).

587 Harfoot, M.B.J., Newbold, T., Tittensor, D.P., Emmott, S., Hutton, J., Lyutsarev, V., Smith,
588 M.J., Scharlemann, J.P.W. & Purves, D.W. (2014) Emergent global patterns of ecosystem
589 structure and function from a mechanistic general ecosystem model. *PLOS Biology*, **12**(4),
590 24.

591 Hebblewhite, M., White, C.A., Nietvelt, C.G., McKenzie, J.A., Hurd, T.E., Fryxell, J.M., Bayley,
592 S.E. and Paquet, P.C., 2005. Human activity mediates a trophic cascade caused by wolves.
593 *Ecology*, 86(8), pp.2135-2144.

594 Jackson, C.R., Power, R.J., Groom, R.J., Masenga, E.H., Mjinga, E.E., Fyumagwa, R.D.,
595 Røskoft, E. and Davies-Mostert, H., 2014. Heading for the hills: risk avoidance drives den site
596 selection in African wild dogs. *PLoS One*, 9(6).

597 Kamau, J.M.Z. & Maloiy, G.M.O. (1985) Thermoregulation and heat balance in the dikdik
598 antelope (*Rhynchotragus kirki*) - a field and laboratory study. *Comparative Biochemistry and*
599 *Physiology a-Physiology*, **81(2)**, 335-340.

600 Kingdon, J. & Hoffman, M. (2013) *Mammals of Africa: Volume VI, Hippopotamuses, Pigs,*
601 *Deer, Giraffe and Bovids*. Bloomsbury, London.

602 Levy, O., Dayan, T., Porter, W.P. & Kronfeld-Schor, N. (2018) Time and ecological
603 resilience: can diurnal animals compensate for climate change by shifting to nocturnal
604 activity? *Ecological Monographs*, 89(1), 01334, doi:10.1002/ecm.1334.

605 Lucas, J.R. (1983) The role of foraging time constraints and variable prey encounter in
606 optimal diet choice. *American Naturalist*, **122(2)**, 191-209.

607 Mbizah, M.M., Marino, J. & Groom, R.J. (2012) Diet of four sympatric carnivores in Savé
608 Valley Conservancy, Zimbabwe: implications for conservation of the African wild dog
609 (*Lycaon pictus*). *South African Journal of Wildlife Research*, **42(2)**, 94-103.

610 Mills, M.G.L. & Gorman, M.L. (1997) Factors affecting the density and distribution of wild
611 dogs in the Kruger National Park. *Conservation Biology*, **11(6)**, 1397-1406.

612 Mole, M.A., Rodrigues DÁraujo, S., van Aarde, R.J., Mitchell, D. & Fuller, A. (2016) Coping
613 with heat: behavioural and physiological responses of savanna elephants in their natural
614 habitat. *Conservation Physiology*, **4(1)**, cow044.

615 Ockendon, N., Baker, D.J., Carr, J.A., White, E.C., Almond, R.E.A., Amano, T., Bertram, E.,
616 Bradbury, R.B., Bradley, C., Butchart, S.H.M., Doswald, N., Foden, W., Gill, D.J.C., Green, R.E.,
617 Sutherland, W.J., Tanner, E.V.J. & Pearce-Higgins, J.W. (2014) Mechanisms underpinning
618 climatic impacts on natural populations: altered species interactions are more important
619 than direct effects. *Global Change Biology*, **20(7)**, 2221-2229.

620 Otieno, T.O. et al., 2019. Human- and risk-mediated browsing pressure by sympatric
621 antelope in an African savanna. *Biological Conservation*, 232, pp.59–65. Available at:
622 <https://www.sciencedirect.com/science/article/pii/S0006320718312771>

Parmesan, C. & Yohe, G. (2003) A globally coherent fingerprint of climate change impacts across natural systems. *Nature*, **421(6918)**, 37-42.

Peters, R.H., 1986. *The ecological implications of body size*, Cambridge University Press.

Pigeon, K.E., Cardinal, E., Stenhouse, G.B. & Cote, S.D. (2016) Staying cool in a changing landscape: the influence of maximum daily ambient temperature on grizzly bear habitat selection. *Oecologia*, **181(4)**, 1101-1116.

Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D. & R Core Team (2015) *nlme: Linear and Nonlinear Mixed Effects Models. R package version 3.1-119*. <http://CRAN.R-project.org/package=nlme>.

QGIS Development Team (2018) *QGIS Geographic Information System*. Open Source Geospatial Foundation Project, <http://qgis.osgeo.org>.

Quaglietta, L., Mira, A. & Boitani, L. (2018) Extrinsic and intrinsic factors affecting the daily rhythms of a semiaquatic carnivore in a mediterranean environment. *Hystrix-Italian Journal of Mammalogy*, **29(4)**, 128-136.

R Core Team (2015) *R: A language and environment for statistical computing*. R Foundation for Statistical Computing <http://www.R-project.org>, Vienna, Austria.

Rabaiotti, D., Groom, R., McNutt, J.W., Watermeyer, J. & Woodroffe, R. (In Review) High temperatures and human pressures interact to influence mortality in an African carnivore.

Rabaiotti, D. & Woodroffe, R. (2019) Coping with climate change: limited behavioural responses to hot weather in a tropical carnivore. *Oecologia*, **189(3)**, 587-599.

Ricklefs, R.E. & Hainsworth, F.R. (1968) Temperature dependent behavior of cactus wren. *Ecology*, **49(2)**, 227-233.

Shorrocks, B., Cristescu, B. and Magane, S., 2008. Estimating density of Kirk's dik-dik (*Madoqua kirkii* Günther), impala (*Aepyceros melampus* Lichtenstein) and common zebra (*Equus burchelli* Gray) at Mpala, Laikipia District, Kenya. *African Journal of Ecology*, **46(4)**, pp.612-619.

649 Speakman, J.R. & Krol, E. (2010) Maximal heat dissipation capacity and hyperthermia
650 risk: neglected key factors in the ecology of endotherms. *Journal of Animal Ecology*, **79(4)**,
651 726-746.

652 Woodroffe, R. (2011a) Demography of a recovering African wild dog (*Lycaon pictus*)
653 population. *Journal of Mammalogy*, **92(2)**, 305-315.

654 Woodroffe, R. (2011b) Ranging behaviour of African wild dog packs in a human-
655 dominated landscape. *Journal of Zoology*, **283(2)**, 88-97.

656 Woodroffe, R., Groom, R. & McNutt, J.W. (2017) Hot dogs: high ambient temperatures
657 influence reproductive success in a tropical mammal. *Journal of Animal Ecology*, **86(6)**
658 1329-1338.

659 Woodroffe, R., Lindsey, P.A., Romañach, S.S. & ole Ranah, S.M.K. (2007) African wild dogs
660 (*Lycaon pictus*) can subsist on small prey: implications for conservation. *Journal of*
661 *Mammalogy*, **88(1)**, 181-193.

662 Woodroffe, R. & Sillero-Zubiri, C. (2013) *African wild dog Red List Assessment*. IUCN.
663

Table 1 Hypotheses describing how predator behaviour, prey behaviour, and predation risk might vary with ambient temperatures

Hypothesis	Species	Explanation	Implication for predator	Expected dietary change at high temperatures
Hypothesis 1: Restricted foraging time	Wild dogs	Wild dogs are expected to spend less time hunting in daytime when temperatures are high ¹ .	Predators with limited foraging time are predicted to select more abundant prey even if it contains less energy ² (in this case dikdiks).	Increased ratio of dikdik to impala
Hypothesis 2: Nocturnal prey selection				
<i>2a: Nocturnal prey concentration</i>	Impala	Impala are predicted to gather in glades at night, as described prior to wild dog recolonisation ³ .	Predators find impala more easily at night as they are predictably located in glades ³ .	Decreased ratio of dikdik to impala
<i>2b: Increased night time hunting</i>	Wild dogs	Wild dogs are expected to spend more time hunting at night, because daytime foraging is constrained ¹ .	As a visual hunter active at night, wild dogs may rely on prey which are easily located.	
<i>2c: Night-time habitat selection</i>	Wild dogs	Wild dogs are predicted to hunt in or near glades at night, as impala are predictably located there.	Following hot days, wild dogs spend more time hunting at night, targeting glades where they can easily locate impala.	Decreased ratio of dikdik to impala
Hypothesis 3: Shade seeking	All species	Wild dogs, impala and dikdik are all predicted to move into denser vegetation at high ambient temperatures to avoid overheating.	Wild dogs are able to hunt impala more successfully, as impala are more vulnerable to predation in denser vegetation ⁴ .	Decreased ratio of dikdik to impala
Hypothesis 4: Prey overheating	Impala	When running at high ambient temperatures, larger bodied animals overheat more quickly, and therefore impala slow down more than dikdiks.	Wild dogs are able to capture impala more easily because impala overheat more quickly when running than the smaller-bodied wild dogs ⁵ .	Decreased ratio of dikdik to impala.

¹(Rabaiotti 2019), ⁴(Lucas 1983), ³(Augustine 2004); ³(Ford *et al.* 2014); ⁴(Ford *et al.* 2014), ⁵(Creel *et al.* 2016).

Table 2 Variables associated with the duration of hunting periods of African wild dogs in the morning and evening, and the occurrence of night-time hunts. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AIC_c < 2$) for the duration of hunting periods in the morning and evening (in minutes), and the occurrence (or not) of hunts at night. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Outcome variable	Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable Importance (n)
Morning hunt duration (minutes)	Intercept	225.038	206.146	245.931	— (2)
	Denning (Yes)	-76.723	-154.789	1.343	1.00 (2)
	Temperature (°C)	-1.359	-2.035	-0.684	1.00 (2)
	Moonlight night before	0.058	-0.241	0.356	0.72 (1)
	Rainfall (mm)	-2.962	-6.296	0.372	0.72 (1)
	Rainfall:Temperature	0.128	0.004	0.252	0.72 (1)
	Denning:Temperature	5.014	2.834	7.195	0.72 (1)
Evening hunt duration (minutes)	Intercept	226.427	214.182	238.627	— (1)
	Temperature (°C)	-3.005	-3.438	-2.572	1.00 (1)
	Moonlight	-1.846	-2.065	-1.627	1.00 (1)
	Moonrise	-0.005	-0.120	0.110	1.00 (1)
	Denning (Yes)	15.823	13.113	18.534	1.00 (1)
Night-time hunt occurrence	Intercept	-0.414	-0.522	-0.305	— (2)
	Temperature	0.020	0.016	0.023	1.00 (2)
	Moonlight	0.031	0.029	0.033	1.00 (2)
	Denning	-0.069	-0.094	-0.044	0.52 (1)

Table 3 Variables associated with selection of woody cover by African wild dogs, impala, and dikdiks. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AICc < 2$). The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Outcome variable	Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable importance (n)
impala - morning	Intercept	0.0465	0.0352	0.0578	(1)
	Temperature	0.0019	0.0003	0.0022	1.00 (1)
impala - midday	Intercept	0.0762	0.0640	0.0884	(1)
	Temperature	0.0014	0.0011	0.0018	1.00 (1)
impala - evening	Intercept	0.0948	0.0878	0.1018	(1)
	Wet vs dry phase	-0.0021	-0.0035	-0.0006	1.00 (1)
impala - night	Intercept	0.0574	0.0531	0.0616	(1)
	Rainfall	0.0003	0.0002	0.0004	1.00 (1)
dikdik – morning and evening	Intercept	0.1211	0.1011	0.1412	(1)
	Wet vs dry phase	0.0055	-0.0003	0.0114	1.00 (1)
dikdik - midday	Intercept	0.1288	0.1102	0.1474	(1)
	Wet vs dry phase	-0.0002	-0.0063	0.0059	1.00 (1)
dikdik - night	Intercept	0.1313	0.1075	0.1551	(2)
	Moonlight	-0.0225	-0.0283	-0.0168	0.70 (1)
	Wet vs dry phase	0.0234	0.0170	0.0298	0.30 (1)
wild dog – morning and evening	Intercept	0.1666	0.1506	0.1827	- (1)
	Denning vs not	-0.0356	-0.0465	-0.0247	1.00 (1)
wild dog - night	Intercept	0.1480	0.1216	0.1745	- (1)
	Denning vs not	-0.1205	-0.1481	-0.0993	1.00 (1)

Table 4 Relationship between time of day (day or night) and use of glades by wild dogs, impala, and dikdiks.

Outcome variable		Estimate	Lower 95% CI	Upper 95% CI	p
impala – probability of being in a glade	Intercept	0.0586	0.0401	0.0772	
	Night <i>vs</i> day	0.1878	0.1832	0.1923	<0.0001
dikdik – probability of being in a glade	Intercept	0.0126	0.0016	0.0237	
	Night <i>vs</i> day	0.0219	0.0091	0.0349	0.09
wild dog - distance to glade	Intercept	0.0041	0.0037	0.0043	
	Night <i>vs</i> day	0.0002	0.0001	0.0003	0.001

Table 5 Variables associated with **African wild dog consumption of impala**. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AIC_c < 2$) for African wild dog consumption of impala. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable importance (n)
Intercept	0.355	0.181	0.530	— (1)
Daily maximum temperature	-0.007	-0.013	-0.001	1.00 (1)

Figure 1 Daily movement patterns of African wild dogs (A) activity measured by collar-mounted accelerometers fitted to 18 wild dogs; blue bands denote the start and stop times used to classify morning and evening hunts (B) Mean maximum temperature throughout the day at the study site

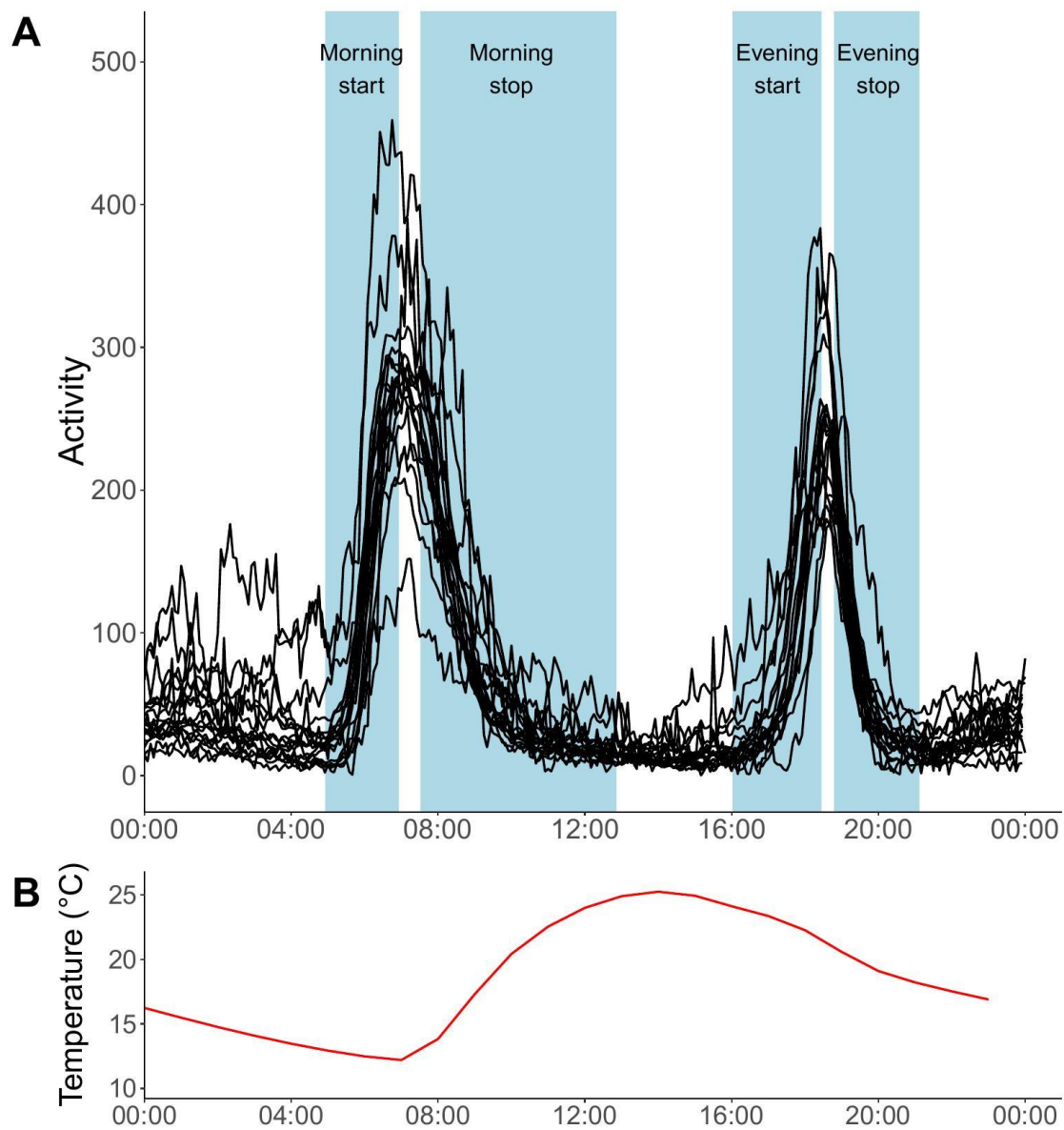


Figure 2 Percentage of hunts and total activity which fall in Morning, Midday, Evening and Night periods, and the percentage of bouts and total activity which spanned multiple time periods, and were therefore determined not to be hunts.

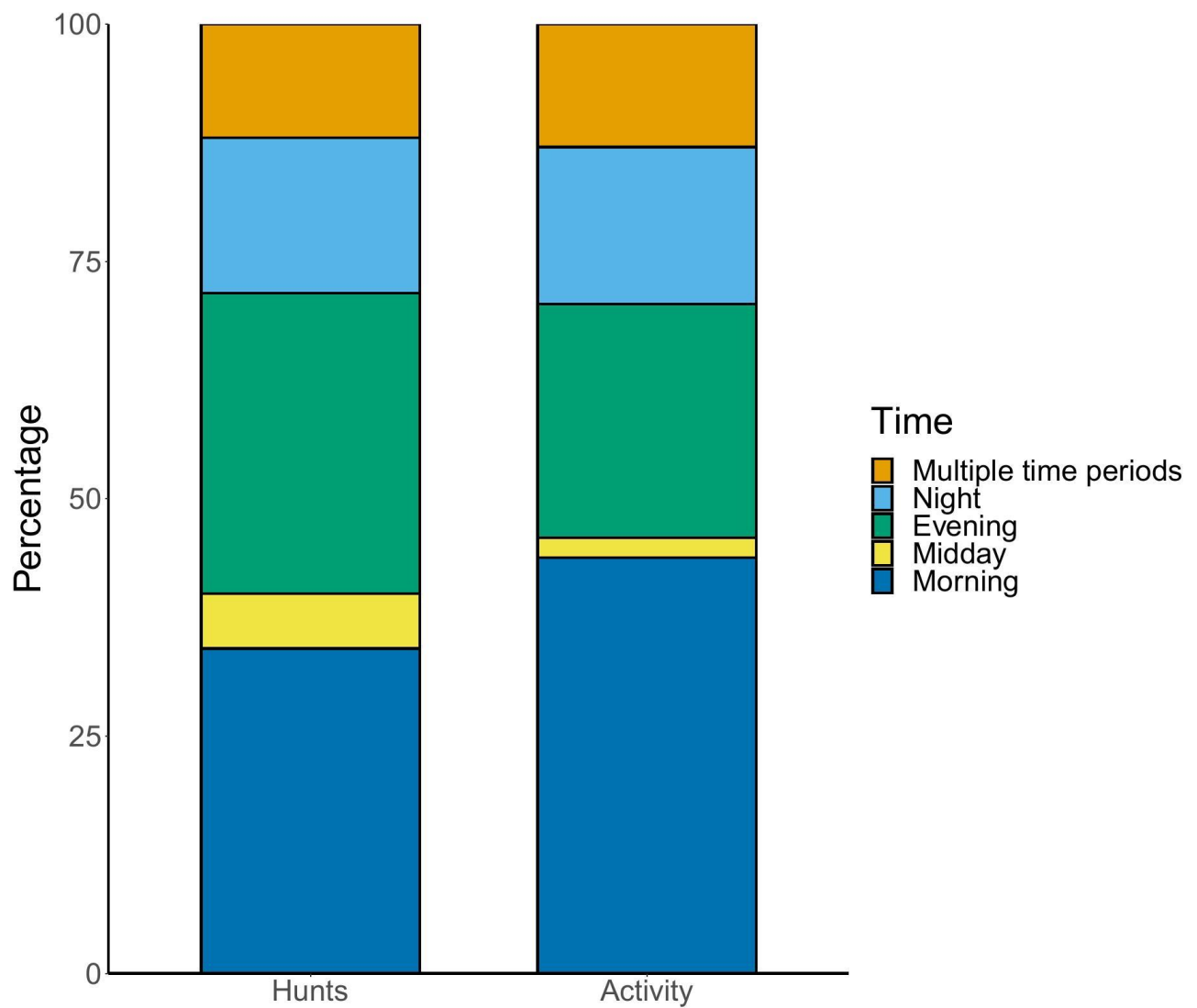


Figure 3 Diel variation in habitat selection by African wild dogs, dikdik, and impala a)

Mean woody cover use by impala, dikdik and wild dogs across time periods. b) Mean wild dog distance to glades, in km, across time periods. c) Probability of glade use by impala and dikdik across time periods. Bars represent standard error.

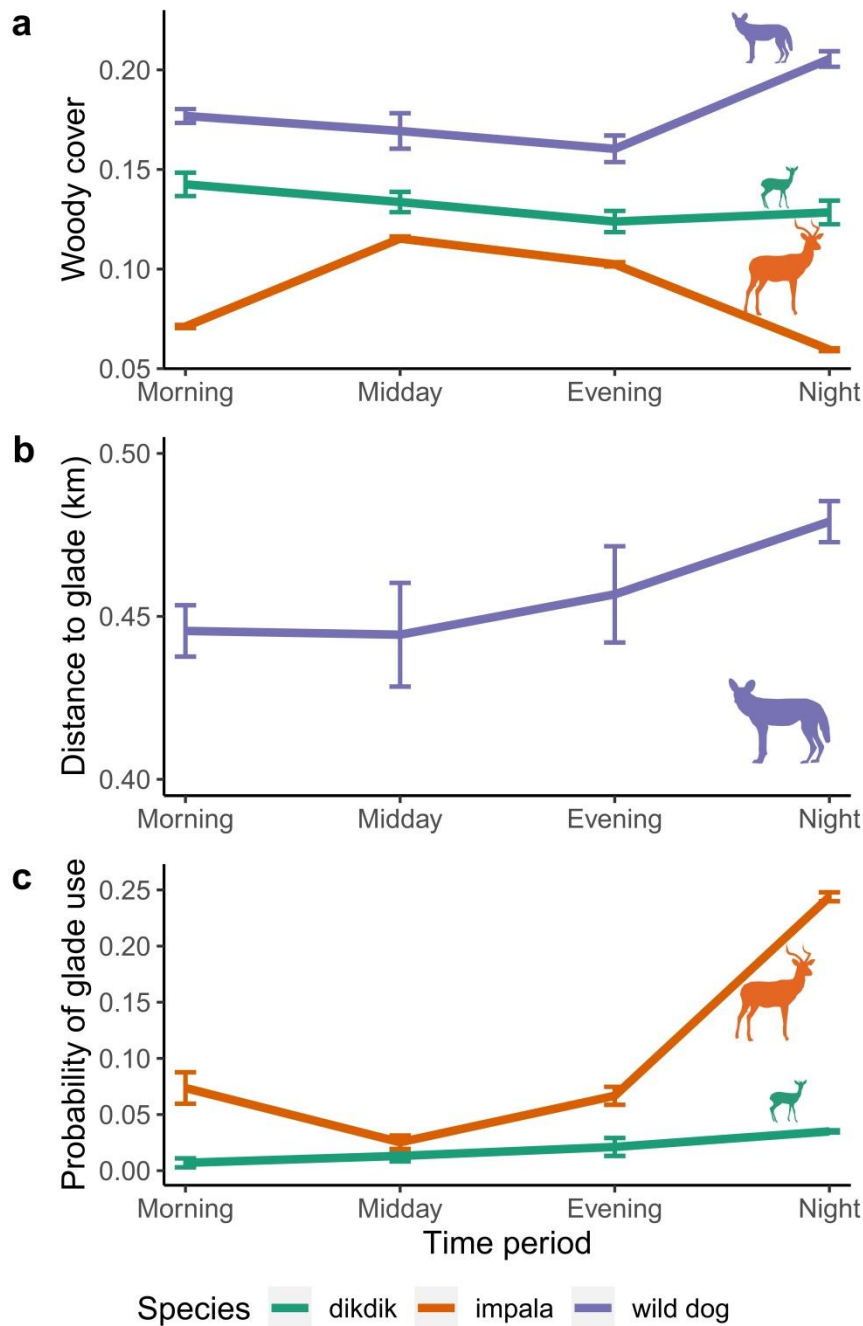
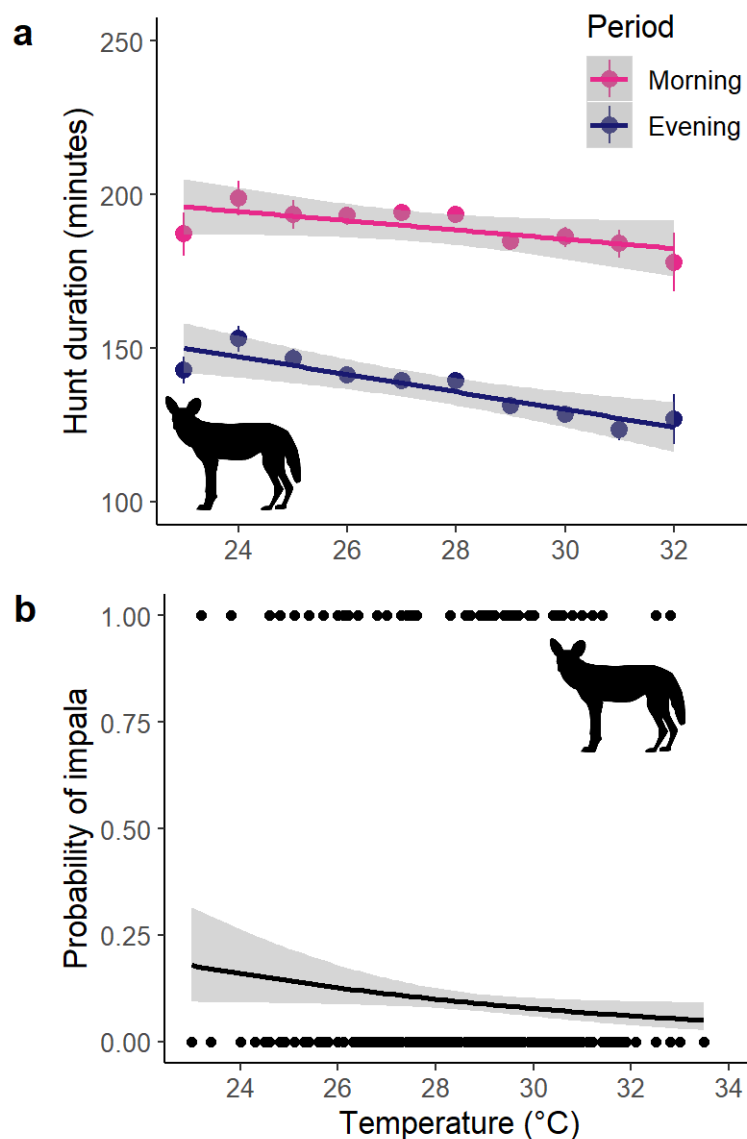


Figure 4 Associations between ambient temperature and African wild dog hunting behaviour a) Duration of African wild dog morning and evening hunts at different daily maximum temperatures ($^{\circ}\text{C}$). Points represent the means for each 1°C interval, vertical bars indicate the standard error b) Probability of the presence of impala in wild dog scat at different daily maximum temperatures ($^{\circ}\text{C}$). Lines represent conditional mean, and shaded areas represent the 95% confidence interval.



Temperature affects predator-prey interactions in an African savanna

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Jacob Goheen⁴ and Rosie Woodroffe¹

Supplementary Material

Figure S1 Frequency distribution of (A) start times and (B) stop times of African wild dog hunting periods, identified using collar-mounted accelerometers.

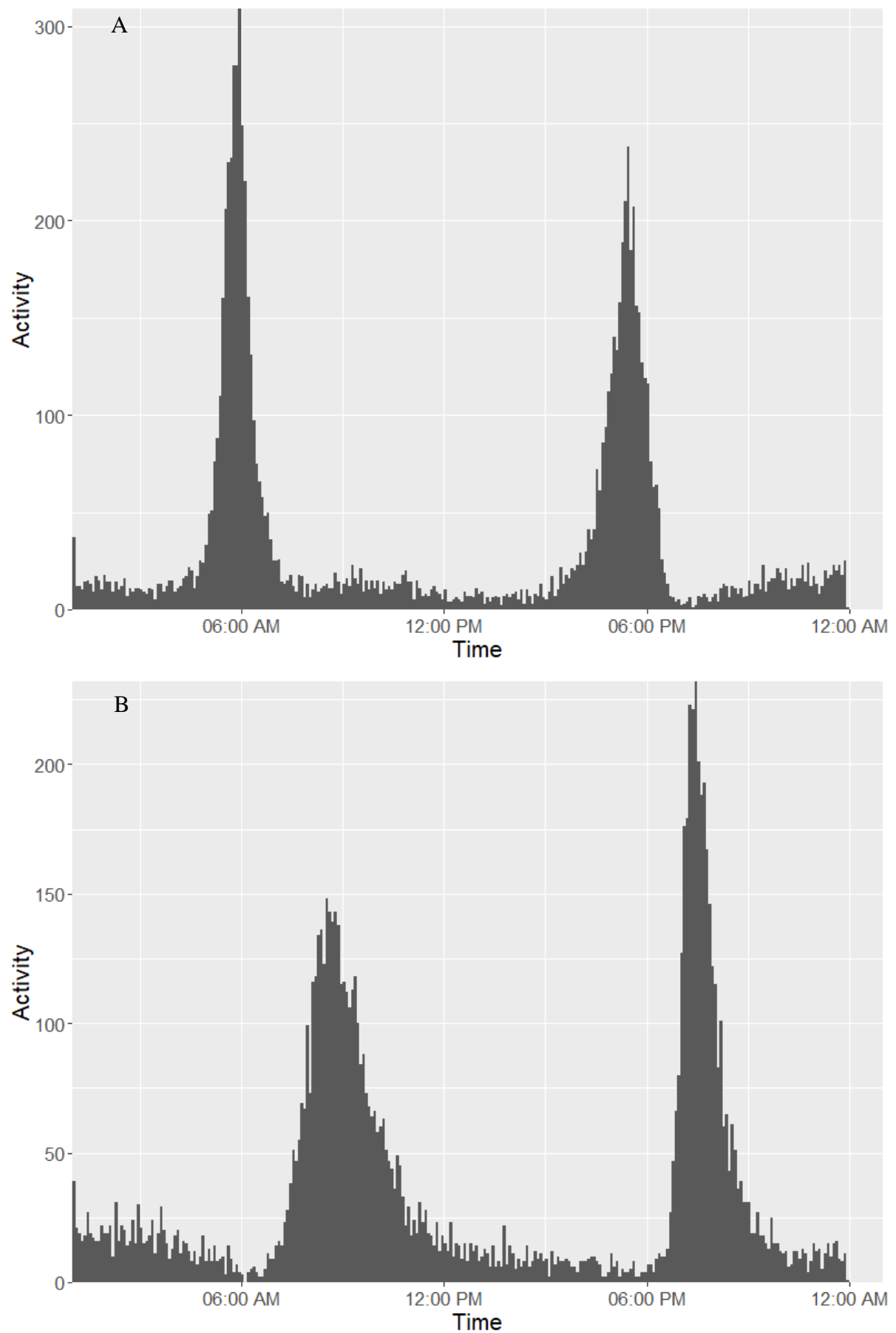


Figure S2 Distance to glades for wild dogs, impala and dikdiks during morning, day, evening and night time periods. Note that, as dikdik ranges are very small, their distance to glades will be influenced by the locations of home ranges where dikdiks were captured.

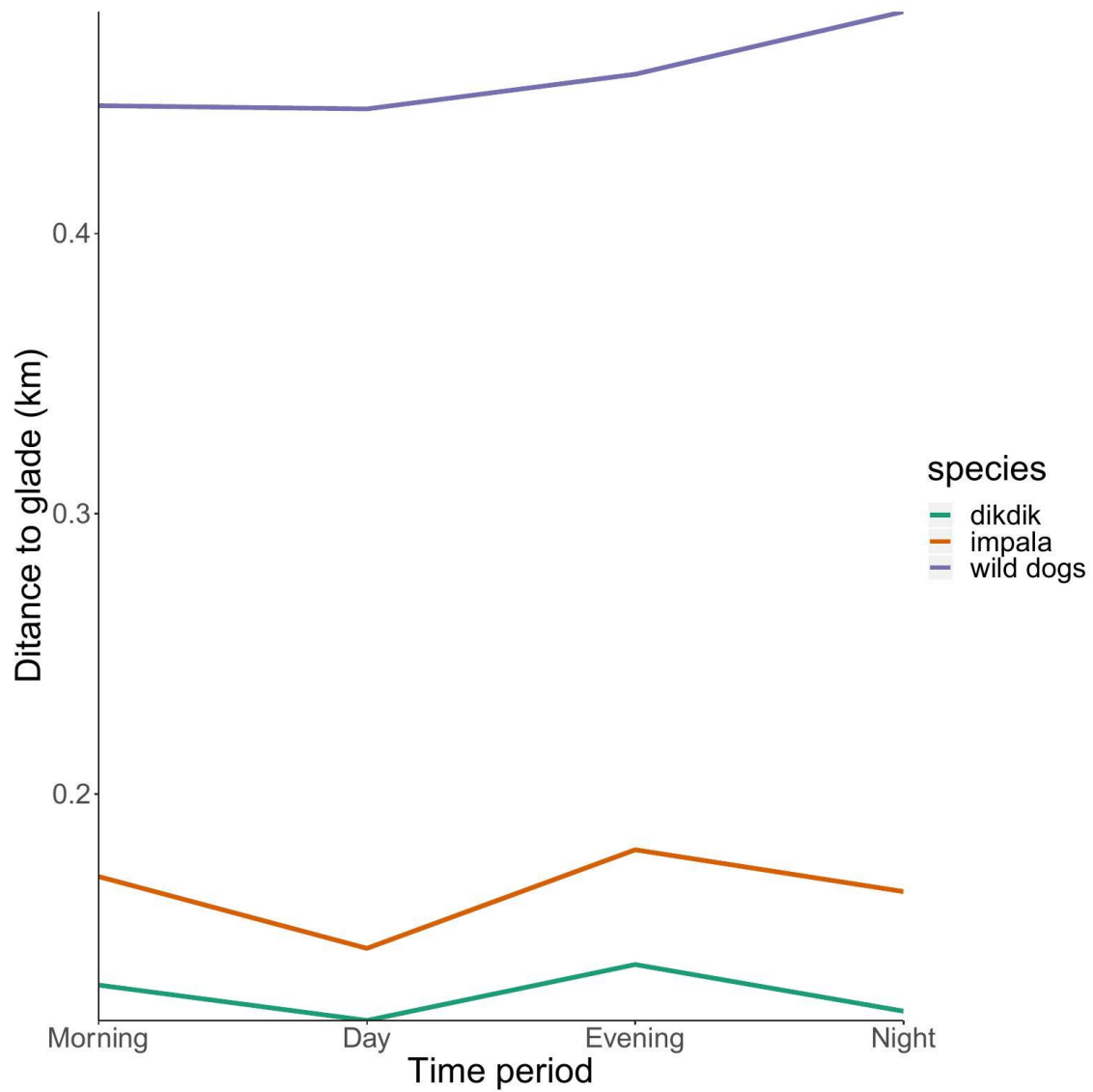


Table S1 Annual temperature and rainfall values at Mpala Research Centre throughout the period of the study.

Year	Temperature (°C)			Rainfall (mm)	
	Minimum	Annual Mean	Maximum	Total	Annual Mean
2000	3.85	19.99	31.54	280.17	0.78
2001	8.23	19.76	32.19	670.56	1.84
2002	6.92	19.81	38.03	483.62	1.32
2003	4.39	19.59	32.64	684.21	1.77
2004	7.67	19.53	38.78	718.05	1.97
2005	3.61	19.18	34.02	508.00	1.40
2006	6.76	19.80	32.51	543.81	1.49
2007	5.66	19.55	33.08	505.99	1.39
2008	5.87	19.69	32.15	410.96	1.19
2009	7.53	20.66	33.31	244.09	0.67
2010	Data not available due to the weather station being broken				
2011	7.15	18.60	31.95	703.35	2.87
2012	3.07	18.73	32.12	639.85	1.76
2013	5.79	19.01	31.86	660.93	1.90
2014	6.53	19.51	31.59	395.72	1.19
2015	6.09	19.33	32.98	548.36	1.61
2016	8.18	19.98	32.89	706.12	1.94

Table S2 Start and stop times of hunting periods allocated to each time period within the analysis.

Type of Bout	Earliest start time	Latest start time	Earliest stop time	Latest stop time
Morning	04:55:39	06:56:04	07:30:41	12:51:33
Midday	06:56:05	16:01:54	12:51:34	18:47:02
Evening	16:01:55	18:26:52	18:47:03	21:06:56
Night	18:26:53	04:55:38	21:06:57	07:30:40

Table S3 Explanatory variables, other than temperature, included in models of wild dog hunts, and justification for their inclusion.

Explanatory variable	Included in models for which time of day (Morning, Evening, Night)	Justification	Citation
Denning status (yes/no)	All	We have previously shown that wild dogs are more active during the denning period	Woodroffe, Groom and McNutt 2017, Rabaiotti and Woodroffe 2019
Total 24h rainfall (mm)	All	We have previously shown that rainfall mitigates the impact of high temperatures on wild dog activity levels so we would expect a negative interaction with temperature during the day and a positive one at night.	Rabaiotti and Woodroffe 2019
Moonlight (proportion illumination*hours of moonlight)	All (Moonlight the previous night was used in morning models, and the following night in evening models)	We and others have previously shown that wild dogs are more active on moonlit nights and would therefore expect a positive relationship between moonlight and nocturnal activity, and potentially a positive relationship with morning and evening activity.	Cozzi et al. 2012,
Moonrise (time)	Night		Rabaiotti and
Moonset (time)	Night		Woodroffe 2019

Table S4 Explanatory variables, other than temperature, included in models of wild dog, impala and dikdik woody cover use and justification for their inclusion.

Explanatory variable	Included in models for which time of day (Morning, Evening, Night)	Species	Justification	Citation
Denning status (yes/no)	All	Wild dogs	Wild dogs use more rugged areas with less vegetation during the denning period	Jackson et al 2014
Total 24h rainfall (mm)	All	All	Rainfall mitigates the impact of high temperatures on activity levels and influences herbivore movement.	Rabaiotti and Woodroffe 2019, Ford et al 2014
Rainfall phase	All	All	Rainfall phase impacts herbivore movement and assemblages	Goheen et al 2013
Moonlight (proportion illumination*hours of moonlight)	All (Moonlight the previous night was used in morning models, and the following night in evening models)	All	Wild dogs are more active on moonlit nights	Cozzi et al. 2012, Rabaiotti and Woodroffe 2019

Table S5 Variables associated with the characteristics of African wild dog hunting periods in the **morning**. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AICc < 2$) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Outcome variable	Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable importance (n)
Occurrence	Intercept	0.8005	0.7698	0.8312	— (1)
	Moonlight before	-0.0097	-0.0115	-0.0079	1.00 (1)
Duration (minutes)	Intercept	225.038	206.146	245.931	— (2)
	Denning vs not	-76.723	-154.789	1.343	1.00 (2)
	Temperature (°C)	-1.359	-2.035	-0.684	1.00 (2)
	Moonlight before	0.058	-0.241	0.356	0.72 (1)
	Rainfall (mm)	-2.962	-6.296	0.372	0.72 (1)
	Rainfall:Temperature	0.128	0.004	0.252	0.72 (1)
	Denning:Temperature	5.014	2.834	7.195	0.72 (1)
Intensity	Intercept	50.9285	46.4112	55.4459	— (2)
	Denning vs not	1.5610	0.7941	2.3279	1.00 (1)
	Temperature (°C)	-0.2552	-0.3743	-0.1362	0.52 (2)
Start time	Intercept	06:31:41	06:24:07	06:39:14	(1)
	Denning vs not	-00:07:46	-00:06:14	-00:09:18	1(1)
	Temperature (°C)	-00:01:15	-00:01:01	-00:01:29	0.99(1)
Stop time	Intercept	09:54:52	09:20:48	10:28:56	(3)
	Temperature (°C)	-00:02:28	-00:01:53	-00:03:03	0.72(2)
	Denning vs not	00:13:25	00:09:37	00:17:14	0.57(2)

Table S6 Variables associated with the characteristics of African wild dog hunting periods in the **evening**. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AICc < 2$) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Outcome variable	Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable importance (n)
Occurrence	Intercept	1.2315	1.1211	1.3419	— (1)
	Moonlight	-0.0098	-0.0118	-0.0079	1.00 (1)
	Temperature (°C)	-0.0167	-0.0206	-0.0129	0.97 (1)
Duration (minutes)	Intercept	226.427	214.182	238.627	— (1)
	Temperature (°C)	-3.005	-3.438	-2.572	1.00 (1)
	Moonlight	-1.846	-2.065	-1.627	1.00 (1)
	Moonrise	-0.005	-0.120	0.110	1.00 (1)
	Denning vs not	15.823	13.113	18.534	1.00 (1)
Intensity	Intercept	62.7066	59.0173	66.3959	— (2)
	Temperature (°C)	-0.8203	-0.9451	-0.6954	1.00 (4)
	Denning vs not	6.4783	5.6993	7.2573	1.00 (2)
	Moonlight	-0.2122	-0.2744	-0.1499	0.36 (1)
	Moonrise	-0.0043	-0.0369	0.0283	0.36 (1)
Start time	Intercept	15:43:55	15:34:49	15:53:01	(1)
	Temperature (°C)	00:03:30	00:03:11	00:03:49	1.00 (1)
	Denning vs not	-00:10:49	-00:08:52	-00:09:18	1.00 (1)
	Moonlight	00:01:01	00:00:52	00:01:10	1.00 (1)
	Moonrise	-00:00:01	-00:00:05	00:00:05	1.00 (1)
Stop time	Intercept	19:42:14	19:38:50	19:45:39	(2)
	Denning vs not	00:13:25	00:09:37	00:17:14	0.55(1)
	Moonlight	00:00:52	00:00:43	00:01:02	0.44(1)
	Moonrise	00:00:01	-00:00:04	00:00:06	0.44(1)

Table S7 Variables associated with the characteristics of African wild dog hunting periods **at night**. The table presents estimated effects of explanatory variables included in the top model sets ($\Delta AIC_c < 2$) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n).

Outcome variable	Explanatory variable	Estimate	Lower 95% CI	Upper 95% CI	Variable importance (n)
Occurrence	Intercept	-0.414	-0.522	-0.305	— (2)
	Temperature	0.020	0.016	0.023	1.00 (2)
	Moonlight	0.031	0.029	0.033	1.00 (2)
	Denning	-0.069	-0.094	-0.044	0.52 (1)
Duration (minutes)	Intercept	80.2071	47.2221	113.1922	— (3)
	Moonlight	5.5885	4.9986	6.1783	1.00 (3)
	Moonrise	0.8356	0.4175	1.2537	1.00 (3)
	Denning vs not	106.6300	-50.2589	263.5188	1.00 (3)
	Temperature (°C)	0.7343	-0.5236	1.9921	0.79 (2)
	Rainfall (mm)	-0.5488	-7.7647	6.6672	0.52 (1)
	Rainfall:Temperature	0.0042	-0.2645	0.2729	0.52 (1)
Intensity	Denning:Temperature	-7.4281	-13.7490	-1.1071	0.52 (1)
	Intercept	28.2814	20.3200	36.2428	— (2)
	Moonlight	0.8649	0.7477	0.9822	1.00 (2)
	Moonrise	-0.0111	-0.0942	0.0720	1.00 (2)
	Denning vs not	2.1092	-0.0491	4.2676	1.00 (2)
Start time	Temperature (°C)	0.4325	0.1926	0.6724	0.52 (1)
	Intercept	15:52:55	09:34:32	22:11:08	(1)
	Moonrise	00:24:22	00:20:30	00:24:22	1.00 (1)
Stop time	Moonlight	-00:05:15	-00:10:38	00:00:08	1.00 (1)
	Intercept	09:14:33	06:28:31	12:00:35	(1)
	Moonlight	-00:24:17	-00:20:20	-00:28:14	1.00 (1)
	Moonrise	00:03:22	00:00:31	00:06:12	1.00 (1)