**Ambient temperature affects mammalian predator-prey interactions in an African savanna**

D. Rabaiotti1,2, Adam T. Ford3, Ben Chapple2, Sophie Morrill2,

Jacob R. Goheen4 and Rosie Woodroffe1

1Institute of Zoology, Regents Park, London, UK, NW1 4RY

2Department of Genetics, Evolution and Environment, University College London, Gower Street, London, UK, WC1H 0AG

3Departent of Biology, University of British Columbia, 3187 University Way, ASC 413, Kelowna, BC Canada V1V 1V7

4Departent of Zoology and Physiology, University of Wyoming, 1000 E. University Ave., Laramie, USA, WY 82071

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***Corresponding author:*** D. Rabaiotti, Daniella.Rabaiotti@ioz.ac.uk, +447968018087

# Data Accessibility Statement

Should the article be accepted, the data will be deposited on Dryad and the DOI included at the end of the article

# Statement of Authorship

D. Rabaiotti, Rosie Woodroffe and Adam Ford, designed the research questions. D. Rabaiotti organised the datasets, designed and carried out the final analyses and wrote the paper. Rosie Woodroffe oversaw data collection on wild dogs and contributed to the writing of the paper. Adam Ford and Jacob Goheen collected the dikdik and impala data and contributed to the writing of the paper. Ben Chapple helped design and carry out the analyses on the impala and dikdik data and African wild dog habitat use. Sophie Morrill assisted in designing and carrying out the analyses on wild dog hunt times.

# Abstract

1. Climate warming may alter predator-prey dynamics by changing predator attack speed and prey escape speed, as well as the daily activity rhythms and habitat use of both predators and prey. Due to the fact there is no direct link between metabolic rates in endotherms, it is widely assumed that climate change will not affect interactions between endothermic predators and prey and temperature.
2. A variety of endotherms change their habitat use and behaviour at high temperatures, however, and chases generate metabolic heat, which larger-bodied endotherms may be less able to dissipate, causing them to move more slowly at high ambient temperatures.
3. We aimed to investigate how ambient temperature impacts predator-prey interactions in a three species system made up of one predator species, the African wild dog (*Lycaon pictus*), and two prey species, dikdik (*Madoqua guentheri*) and impala, (*Aepyceros melampus*).
4. We explored how activity rhythms, habitat use and chase speed impact predation risk from wild dogs impala and dikdik.
5. We predicted that at high ambient temperatures African wild dog predation on impala would increase through a number of mechanisms. Firstly, that at high temperatures all three species would seek shade, leading to increased use of closed habitat, where impala have higher rates of predation by wild dogs. Secondly, that wild dogs would change their activity rhythm, hunting more at night when impala are predictably located. Finally, we predicted that 40kg impala would be more vulnerable than 5kg dikdik to predation by 23kg African wild dogs.
6. In contrast with these predictions, we found fewer wild dog scats containing impala remains when ambient temperatures were high. Wild dogs spent less time hunting on hot days and, consistent with models of optimal foraging under time constraints, appeared to select abundant, lower-value dikdiks over rarer but higher-value impala.
7. Our findings show that weather can influence predator-prey interactions among endotherms.

**Keywords:** African wild dog;antipredator behaviour; climate change; dikdik; habitat selection; impala; predation thermal tolerance

# Introduction

Climate change has far-reaching impacts on both species and ecosystems (Parmesan & Yohe 2003). Meta-analyses suggest that many climate impacts on individual species operate indirectly, through their effects on species interactions such as competition, herbivory, predation, and parasitism (Cahill *et al.* 2012; Ockendon *et al.* 2014). As weather patterns will shift under novel climate regimes, understanding how ambient temperature and other abiotic processes influence species interactions will become critical to wildlife conservation in the coming decades.

Climate is particularly critical for endothermic species, which rely on environmental temperatures to regulate their body temperature, meaning that high environmental temperatures directly lead to changes in physiology, movement and behaviour. These individual behavioural shifts can drive climate impacts on species interactions. For example, rising ambient temperatures allows ectotherms to move more rapidly, potentially making them more efficient predators and more challenging prey, with cascading consequences for community structure (Dell, Pawar & Savage 2014). It is usually assumed that environmental temperature has no similar impacts on predation involving endotherms (e.g., Dell, Pawar & Savage 2014; Harfoot *et al.* 2014); however, birds and mammals may respond to high ambient temperatures by curtailing their foraging time (Ricklefs & Hainsworth 1968; Quaglietta, Mira & Boitani 2018), switching from diurnal to nocturnal foraging (Levy *et al.* 2018), selecting different habitats (Austin 1976; Pigeon *et al.* 2016), and choosing different foods (Doolan & Macdonald 1996; Garcia-Heras *et al.* 2017), all of which are likely to influence their impacts on the species that they consume, or that consume them.

Predator-prey interactions involving large-bodied endotherms might be especially sensitive to rising environmental temperatures, because the low surface-area-to-volume ratios of large animals may make it difficult to dissipate heat generated by pursuing or avoiding predation (Speakman & Krol 2010; Creel *et al.* 2016). Hence, with respect to the outcome of predator-prey interactions, high ambient temperatures might favour endothermic prey that are smaller than their endothermic predators, as the predators large body size would mean it overheats faster than its smaller prey, leading to decreased predation rates on smaller prey species. Similarly, high ambient temperatures may also benefit endothermic predators that are smaller than their endothermic prey, as the large bodied prey would overheat faster than the smaller bodied predator, leading to shorter chase times and higher hunting success on large bodied prey species (Creel *et al.* 2016). Predation by and on large-bodied endotherms can have cascading effects on community structure (e.g. Hebblewhite et al 2005), and these effects might therefore be sensitive to changes in ambient temperature.

Using a three-species predator-prey system in an African savanna, we quantified the relative importance of ambient temperature and body size as mechanisms that shape the outcome of species interactions. Specifically, we examined how ambient temperature affects the habitat use and activity patterns of small-bodied (~5kg) Guenther’s dikdik (*Madoqua guentheri*), the larger-bodied(~40kg) impala (*Aepyceros melampus*), and their shared predator, the African wild dog (*Lycaon pictus*). At our study site, impala and dikdik are the two most abundant ungulates, and their browsing shapes plant communities (Ford *et al.* 2014; Ford *et al.* 2015). Wild dogs are crepuscular, and hunt by running prey down (Creel & Creel 1995); dikdik and impala together comprise 82% of the prey biomass that wild dogs consume at our study site (Woodroffe et al. 2007). Dikdik make up a higher proportion of prey consumed by wild dogs than at other sites and are not typically considered to be within the species’ preferred prey range (Clements et al. 2014).

We evaluated five hypotheses about the impacts of ambient temperature on predation by wild dogs on impala and dikdik (Table 1). First, we explored the potential consequences for a scenario in which predators reduced their activity during daytime, when ambient temperatures are higher (Table 1). Previously, we have shown that wild dogs are less active on hot days (Rabaiotti & Woodroffe 2019), which may reflect shorter hunting periods (Woodroffe 2011b). Optimal foraging theory suggests that, when foraging time is limited, individuals should accept lower-value prey, rather than continuing to search for higher-value prey (Lucas 1983). Therefore, we expected that, if wild dogs had reduced foraging time, they would increase predation on dikdik (which are encountered more frequently but are too small to feed a whole pack) over impala, which are larger but encountered less frequently (Woodroffe *et al.* 2007). We term this, Hypothesis 1, the ‘reduced daytime foraging hypothesis’.

Secondly, we examined how prey behaviours, coupled with changes in predator behaviours, can increase predation risk on one prey species when compared to another. Impala have been shown to aggregate at night in small clearings (“glades”), as a form of anti-predator behaviour (Augustine 2004; Otieno et al. 2019), and are therefore predictably located. We predicted that Impala would be more predictably located at night, and termed this, Hypothesis 2, the ‘nocturnal prey concentration hypothesis’.

We have previously shown that following hot days, wild dogs are more active at night, which may reflect increased nocturnal hunting (Rabaiotti & Woodroffe 2019), and we hypothesised that this would be the case with this dataset, terming this, Hypothesis 3a, the ‘increased night-time foraging’ hypothesis. We also predicted that wild dogs should therefore select areas in and around glades when hunting at night (Table 1) – terming this, Hypothesis 3b, the ‘nocturnal prey targeting hypothesis’. We would expect, if Hypothesis 2, 3a and 3b were upheld, that increased nocturnal wild dog activity at high ambient temperatures would lead to increased targeting of predictably located impala over less predictably located dikdik, increasing predation risk on impala (Table 1).

Fourth, we quantified whether habitat selection by both predators and prey might be affected by ambient temperatures, with consequences for predation. Many large mammals select shade in hotter weather (Mole *et al.* 2016; Pigeon *et al.* 2016), and we predicted that wild dogs, impala, and dikdik would do the same (Table 1). Further, we predicted that such a change would increase predation by wild dogs on impala, since impala face higher *per capita* predation rates in habitats characterised by higher woody cover (and thus shade; Ford *et al.* 2014). We term this, Hypothesis 4, the ‘shade-seeking hypothesis’.

Finally, we explored the potential consequences of both predators and prey overheating during chases. Creel *et al.* (2016) suggested that reduced daytime activity was not, in fact, due to reduced time spent hunting, but instead a result of decreased chase time at high ambient temperatures. Creel *et al.* (2016) suggests that, during chases, wild dogs would overheat more slowly than larger-bodied prey, leading to shorter and more successful chases at high ambient temperatures. Larger prey species have more difficulty dissipating heat due to their lower surface-area-to-volume ratio (Peters 1986), potentially making such species more vulnerable to coursing predators. By extension, we predicted that wild dogs would overheat less rapidly than impala but more rapidly than dikdik, leading to increased predation on impala relative to dikdik (Table 1). We termed this, Hypothesis 5, the ‘prey overheating hypothesis’.

These four hypothesised behavioural responses to high ambient temperatures generated four different scenarios, comprising contrasting sets of predictions about trophic interactions between wild dogs, impala, and dikdik (summarised in Table 1). By testing these predictions, we explored the behavioural mechanisms underlying whether, and how, ambient temperature influenced interactions between these three species, in an attempt to evaluate how climate change might affect this ecosystem through predator-prey interactions.

# Methods

## Study area

The focal area for this study was the Mpala Conservancy, Kenya, a 200 km2 area of semi-arid savanna managed jointly for livestock production and wildlife conservation (0°17’ N, 36°53’ E). Mpala experiences little seasonal variation in ambient temperature; daily maximum temperature ranges from 20-37°C, and mean annual rainfall is 590mm, varying substantially both within and between years (Caylor, Gitonga & Martins 2017).

Mpala hosts 22 species of wild ungulate, of which dikdik and impala are the two most abundant (Ford *et al.* 2015), at 299 individuals per km2 and 161 individuals per km2 respectively (Shorrocks et al 2008). Mpala also supports six species of large carnivore: lion (*Panthera leo*), leopard (*P. pardus*), cheetah (*Acinonyx jubatus*), spotted hyaena (*Crocuta crocuta*), striped hyaena (*Hyaena hyaena*) and African wild dog (Frank, Woodroffe & Ogada 2005).

## Data collection

We used Global Positioning System (GPS) collars to measure daytime and night-time activity, movement, and habitat use. GPS-collars (Savannah Tracking Ltd, Nairobi, Kenya) were fitted to 20 adult female impala (each from a different herd) between May and June 2011, as described in Ford *et al.* (2014). Each impala GPS-collar recorded a location every 20 minutes, for an average of 245 days. Likewise, GPS-collars (Savannah Tracking Ltd, Nairobi, Kenya) were fitted to 15 adult female dikdik between July 2010 and September 2011, as described in Ford & Goheen (2015). Each dikdik GPS-collar recorded a location every 10 minutes, for 18 days on average. Additionally, GPS collars (Vectronic Aerospace GmbH, Berlin, Germany) were fitted to 18 wild dogs in 13 packs between 2011 and 2016, using capture methods described in Woodroffe (2011a); all of these packs had home ranges overlapping Mpala, although none remained on Mpala year-round. Data were collected from one collar per pack at any one time. Each wild dog GPS-collar recorded locations at 01:00, 06:30, 07:00, 07:30, 08:00, 13:00, 18:00, 18:30, 19:30, and also recorded average acceleration in two planes (on a scale of 0 to 255) every five minutes, for 218 days on average.

We used faecal analysis to quantify the relative frequency of predation by wild dogs on impala and dikdik. Wild dog scats were collected during 2001-2004 across a 5,700km2 study area which included Mpala Conservancy, and analysed as described in Woodroffe *et al.* (2007). We avoided pseudoreplication by including only one randomly-selected scat collected from each pack on each occasion (Woodroffe *et al.* 2007). Most scat samples came from different individuals to those from which the GPS data was obtained due to the average life-span of African wild dogs being 7 years, and also the difficulty of determining which individual each scat belonged to.

Daily rainfall and daily maximum ambient temperature were collected from a weather station at Mpala Research Centre (Caylor, Gitonga & Martins 2017). Finally, we used GPS-collar data to identify periods when wild dog packs were denning (raising pups in a den, recognisable from the movement path which shows a “starburst” pattern of repeat visits to the same location, Woodroffe, Groom & McNutt 2017) as activity levels have been shown to increase during this period (Rabaiotti 2019).

## Identifying hunting periods of wild dogs

To estimate the time that wild dogs spent hunting, and to compare prey and predator behaviour at times when predation risk was highest, we used accelerometry data to identify wild dog hunting bouts. First, we summed the two accelerometer measurements for each 5-minute period, to give an overall measure of activity (from 0 to 510). The scale was from 0 to 510 as the accelerometry data is produced on a scale of 0 to 255 on each axis due to 255 being the maximum representable number on an 8-bit byte, which is used by these collars for data storage, and these axes were summed to give a maximum value of 510. We then defined hunting bouts based on three criteria: (i) activity >0 for >20 minutes; (ii) total activity during the activity bout >500; (iii) followed by three or more consecutive records of 0 activity. These criteria excluded activity bouts which were too short to represent hunting, or which related to less energetic behaviours, such as socialising. Criteria were derived using visual observations of activity plots and examining the characteristics of activity bouts that occurred in time periods during which it is common to observe the dogs hunting. For each hunting bout, we recorded start time, end time, duration (in minutes), and intensity (total activity divided by duration). The distributions of start and stop times are shown in Figure S1. We classified hunting bouts as “morning”, “evening”, “night” or “midday”, based on their start and stop times, as detailed in Table S1 and Figure S1. Bouts that spanned multiple time periods were excluded from the analysis. After using accelerometry data to delineate hunting bouts, we categorised each wild dog location as falling inside or outside a hunting bout.

We categorised dikdik and impala GPS collar locations into four time periods. We defined “morning” as the time period between sunrise (taken from the *R* package *suncalc*, Agafonkin & Thieurmel 2017) and the third quartile of end times for hunting bouts (approximately 3.5h after sunrise; Figure S1). We classified impala and dikdik “evening” GPS-locations as those recorded between the first quartile of start times for hunting bouts (approximately 2hr10min before sunset; Figure S1) and sunset (which fell between 1841h and 1912h). We categorised impala and dikdik GPS-locations as “midday” if they were recorded between the “morning” and “evening” periods, and “night” if they were recorded between sunset and sunrise. Prey activity was categorised using the timing of predator hunting bouts as these times were when the location of prey species would impact predation risk by African wild dogs.

## Habitat use

We analysed habitat use from a woody cover map of Mpala created from a 2011 Quickbird satellite image (Digital Globe, Longmont, CO, USA) by Ford *et al.* (2014). We estimated the woody cover (a measure of shadiness) associated with each impala, dikdik, and wild dog GPS-collar location as the proportion of woody cover within a circular area of radius 40m, centred on the collar location.

We used the same habitat map to classify each GPS-collar location in relation to glades. For impala and dikdik, we recorded whether each location fell inside or outside a glade. As <1% of wild dog hunting locations fell inside glades, we recorded the distance to the nearest glade for each wild dog location, using the plugin “*NNJoin*” in *QGIS* (QGIS Development Team 2018).

## Statistical analyses

We used multi-model inference to evaluate associations between each outcome variable and a range of explanatory variables (Table S2, Table S3). For each outcome variable, we built a series of statistical models using explanatory variables and biologically-meaningful interactions between explanatory variables, with individual identity as a random variable. We then used Akaike’s Information Criterion (AIC) to compare models using the *R* package *MuMIn* (Bartoń 2017). We considered all models with AIC scores within 7 units of the best (lowest AIC) model (*i.e.,* ΔAIC<7) to have some level of support (Burnham et al. 2002), referring to this array of models as the “top set”. We used model averaging (Burnham et al. 2002) to estimate the effect on the outcome variable of each explanatory variable in the top set.

The reduced daytime foraging time hypothesis (Hypothesis 1) predicted that wild dogs would spend less time hunting on hot days. To test this hypothesis, we split the activity data into morning, midday and evening. We analysed continuous outcome variables describing hunt duration, start time, stop time, hunt intensity and a binary outcome variable describing whether or not a hunt was recorded for each time period. Only 10% of hunting bouts occurred during midday (accounting for 2% of daily activity), so these were not subjected to statistical analyses. Morning and evening hunt periods were analysed separately as a number of the covariates (Table S2) were predicted to impact morning, evening in different ways, for example high ambient temperature is predicted to decrease the duration of morning hunts and increase the duration of evening hunts, and moonlight is predicted to increase the duration of evening hunts, but have no impact on morning hunts. For each outcome variable, we constructed Generalised Linear Mixed Models (GLMMs) using the package *nlme* (Pinheiro *et al.* 2015) in *R* (R Core Team 2015), with Gaussian error distribution for the continuous outcome variables and binomial error distribution for the binary outcome variable. Each model included the identity of individual animals as a random effect.

Although our hypotheses concerned the potential effects of ambient temperature, we included three other explanatory variables known to influence activity patterns of wild dogs, namely denning status, rainfall, moonlight and, in models of night-time activity, time of moonrise and moonset. For further details see Table S2.

In the nocturnal prey concentration hypothesis (Hypothesis 2), we predicted that impala would use glades more at night. To test this we calculated the proportion of each impala’s locations falling within glades during the morning, midday, evening, and night periods for each 24-h period. We analysed these outcome variables using GLMMs with Gaussian error distribution, using time of day, maximum daily ambient temperature the previous day, daily rainfall, and rainfall phase as candidate explanatory variables. Following Ford *et al.* (2014), we considered days to fall within “wet phases” if >50mm of rain had fallen in the previous four weeks, and classed all other days as falling in “dry phases”. In constructing the array of models for multi-model inference, this rainfall phase variable was never included in the same model as daily rainfall, because the two were correlated.

In the increased night time foraging hypothesis (Hypothesis 3a) wild dogs would be more likely to hunt, and increase their hunting activity at night. To investigate this, hunting bouts that fell between sunset and sunrise were classified as night-time, and the same continuous outcome variables and explanatory variables were modelled as in the investigation of Hypothesis 1, however, as nocturnal periods span dates, the maximum ambient temperature for the proceeding daytime period was used in these models. As in the investigation of Hypothesis 1, we constructed Generalised Linear Mixed Models (GLMMs), with Gaussian error distribution for the continuous outcome variables and binomial error distribution for the binary outcome variable, for each outcome variable using the package *nlme* (Pinheiro *et al.* 2015) in *R* (R Core Team 2015), and each model included the identity of individual animals as a random effect.

In the nocturnal prey targeting hypothesis (Hypothesis 3b) wild dogs would be expected to preferentially target glades at night. To test the hypothesis that wild dogs hunted in glades more often at night, we calculated the mean distance to the nearest glade for each night-time hunt period. As when investigating Hypothesis 2, the nocturnal prey concentration hypothesis, we analysed these outcome variables using GLMMs with Gaussian error distribution, using the same explanatory variables.

In the shade-seeking hypothesis (Hypothesis 4) we hypothesised that wild dogs, impala, and dikdik would increase their use of shaded habitat at high ambient temperatures. To test this hypothesis, we constructed a series of models with use of woody cover as the outcome variable. To avoid pseudoreplication, we averaged the woody cover values for each individual across each morning, midday, evening or night-time period. For wild dogs, only locations from hunting bouts were included. We analysed these outcome variables using GLMMs with individual identity as a random effect, building a separate array of models for each time of day, and for all times of day together. Candidate explanatory variables were the same as for the analyses of wild dog activity (i.e., temperature, rainfall, moonlight, and, for wild dogs, denning), but also included a variable describing rainfall phase (Table S3) as this is predicted to impact herbivore behaviour and location (Goheen et al 2013).

In the prey overheating hypothesis (Hypothesis 5), we predicted that chase distances of impala would be shorter on hotter days, while chase distances of dikdik would be longer, leading to greater predation on impala on hotter days (Table 1). We could not measure chase distance, as our GPS-collar locations were recorded too infrequently, so our evaluation of this scenario relied on testing the hypothesis that wild dogs killed impala more frequently on hot days. This outcome was also predicted if the nocturnal prey concentration, increased night-time foraging and nocturnal prey targeting (Hypotheses 2 and 3) were all upheld, and under the shade seeking hypothesis (Hypothesis 4). We tested this hypothesis by using a GLM with binomial error distribution to analyse whether or not wild dog scats contained impala remains. In this model, candidate explanatory variables were ambient temperature during the previous day (as prey hairs have been found in wild dog scat between 6 and 80 hours after prey consumption (Davies-Mostert et al 2010)), and land use (community vs private land, to reflect variation in impala abundance). Pack or individual identity were often unknown for wild dog scats, so these models did not include random effects.

# Results

## Daily movement patterns

African wild dogs showed a strongly crepuscular activity pattern, with hunts and daily activity concentrated in the morning and evening time periods (Figure S2). Sixteen percent of wild dog hunting periods, and 17% of daily activity, occurred at night (Figure S3).

Impala had the highest levels of displacement between GPS locations at night (1.24km ±2.11) and lower displacement distances in the morning (0.90km ±1.4), daytime (0.97km ±1.7) and evening (0.89km ±1.3) time periods.

Dikdik had similar levels of displacement between GPS locations throughout the day, with the higher levels of displacement in the morning (0.27km ±0.3) and at night (0.26km ±0.27) and slightly lower displacement distances in the daytime (0.20km ±0.27) and evening (0.19km ±0.23) time periods.

## Effects of ambient temperature on wild dog hunting patterns

Consistent with the reduced daytime foraging time hypothesis (Hypothesis 1), wild dogs’ daytime hunting periods were shorter in both the morning and evening time periods at high ambient temperatures (Table 2, Figure 1d). These shorter hunting periods reflected earlier start and stop times in the morning (Table S5) and later start times in the evening (Table S6). In addition to being shorter, both morning and evening hunts entailed less intense activity on hotter days (Table S5, Table S6). Evening hunts were less likely to occur at all on days with higher ambient temperatures (Table S6), though there was no such effect on morning hunts (Table S5). Rainfall may have mitigated the effects of high ambient temperatures, with rainfall:temperature interactions included in some of the top models for hunt duration and intensity (Table S5, Table S6). Packs were consistently more active during daytime when they were denning (Table S5, Table S6).

Similarly, and as predicted under the increased night time foraging hypothesis (Hypothesis 3a), wild dogs were more likely to hunt at night following daytime periods with high ambient temperatures (Table 2). There was also some evidence that ambient temperature increased the duration and intensity of night-time hunts (Table 2). Nocturnal hunting activity was increased at higher levels of moonlight (Table 2, Table S6), with corresponding reductions in daytime activity on dates with high moonlight indices (Table 2, Table S5, Table S6).

## Habitat selection

The three species differed in their use of woody cover, with impala using the most open areas and wild dogs using the areas with the highest woody cover (Figure 1a). Impala used the least wooded areas at night, and during the morning period, whereas wild dogs used the most wooded areas at night (Figure 1a). The use of woody cover by dikdik was relatively consistent throughout the day (Figure 1a).

As predicted under the shade-seeking hypothesis (Hypothesis 4), impala selected areas with denser woody cover on hotter days, during morning, midday, and evening periods (Table 3). In contrast, there was weak and inconsistent evidence for dikdik selecting woody cover based on ambient temperature, and no evidence of such selection by hunting wild dog packs (Table 3).

The three species also varied in their use of glades. Consistent with the nocturnal prey concentration scenario (Hypothesis 2), impala were more likely to be located in glades at night than at other times (Figure 1c), although there was no such pattern for dikdik (Table 4, Figure 1c). In contrast with predictions under Hypothesis 3b, the nocturnal prey targeting scenario, however, wild dogs were not located closer to glades at night than at other times of day, and were no more likely to be found close to glades at night when daytime temperatures had been high (Table 4, Figure 1b). Impala were less likely to be located in glades on moonlit nights, and following days with high ambient temperatures, whereas dikdik appeared more likely to use glades on moonlit nights (Table 4). Pack size was the most consistent predictor of wild dog proximity to glades, with larger packs found in closer proximity (Table 4).

## Prey selection

Among 795 wild dog scats, 71 (9%) contained impala remains and 609 (77%) contained dikdik remains. As predicted under the reduced daytime foraging scenario (Hypothesis 1) but not the other three scenarios, wild dog scats were less likely to contain impala remains when ambient temperatures had been higher on the previous day (Table 5, Figure 1e).

# Discussion

Our analyses revealed clear associations between ambient temperature and the behaviour of both predator and prey species, which appeared to influence trophic interactions. Our findings were most consistent with the reduced daytime foraging hypothesis (Hypothesis 1), under which we predicted that, on hot days, wild dogs would spend less of the daylight period hunting, and therefore prefer abundant small prey over larger but less abundant prey. Consistent with these predictions, we found that wild dogs spent less time hunting during daytime hours (Figure 1d, Table S5, Table S6). Previously we posited that, on average, dikdik and impala would yield comparable returns, because the greater energy intake achievable by hunting impala was offset by the shorter travel distances associated with hunting dikdik (Woodroffe *et al.* 2007). However, optimal foraging theory predicts that a predator with limited time should select a more abundant but lower-value prey (such as dikdik) rather than waiting to locate rarer, higher value prey (such as impala, Lucas 1983). High ambient temperatures would therefore be expected to favour wild dogs eating dikdik more than impala, and our observations were consistent with this prediction (Table 5).

Under the nocturnal prey concentration, increased night time foraging and nocturnal prey targeting scenarios hypotheses (Hypotheses 2-3b), we predicted that, when ambient temperatures were high, wild dogs would hunt at night, targeting impala which are predictably located in glades. Although wild dogs hunted more often at night in hot weather (Table 2), and impala were located in glades at night (Table 4, Figure 1c), there was no evidence that wild dogs targeted impala at night. Wild dogs were no closer to glades at night than at other times of day, were no closer to glades on nights when ambient daytime temperatures had been high (Table 4), and were less likely to consume impala in hot weather (Table 5, Figure 1e). Hence, Scenario 2 did not generate the predicted consequences for either predator behaviour or predation risk.

Under the shade-seeking hypothesis (Hypothesis 4), we predicted that all three species would increase their daytime use of woody cover at high ambient temperatures. However, we found that only impala did so (Table 3). Of the three species, impala used the most open habitat (Figure 1a, b and c), which may have resulted in a greater need to seek shade at high ambient temperatures. Physiological studies suggest that dikdik are dependent upon shade to thermoregulate (Kamau & Maloiy 1985) but, being small-bodied, they may be able to use small patches of shade without moving into denser habitat. There was no evidence that wild dogs hunted in denser cover on hot days (Table 3), perhaps because hunting periods occurred before and after the hottest times of day (Figure 1a). The tendency of impala to move into denser cover at high ambient temperatures would be expected to increase their risk of being killed, since wild dogs typically occupied denser cover (Figure 1a), and impala are more likely to be killed in denser cover (Ford *et al.* 2014). Nevertheless, we found that impala remains were less likely to be found in wild dog scats following periods of high ambient temperature (Table 5). Hence, although impala behaviour changed in line with the predictions of the shade-seeking hypothesis (Hypothesis 4), this change did not generate the predicted impact on predation risk.

Under the prey overheating (Hypothesis 5) we predicted that high ambient temperatures would reduce chase distances for impala and increase them for dikdik, as larger-bodied animals would be slowed down by the heat to a greater extent than smaller-bodied animals (Creel *et al.* 2016). We were not able to test these hypotheses directly because our monitoring methods did not allow us to measure chase distance or speed. However, under this scenario we also predicted that predation upon impala would increase, relative to predation on dikdik, when ambient temperatures were high. Our observations showed the opposite pattern (Table 5, Figure 1e); hence, a key prediction of the chase overheating scenario was not upheld by our analyses.

Our findings favour the reduced daytime foraging hypothesis (Hypothesis 1) over the other scenarios. Under Hypothesis 1, we would expect energy intake by wild dogs to be reduced at high ambient temperatures, due to constrained foraging time and a consequent acceptance of lower-energy prey. This prediction is consistent with our observations that this study population experiences higher mortality and lower reproductive success at high ambient temperatures (Woodroffe, Groom & McNutt 2017; Rabaiotti *et al.* in review). In contrast, under the other three scenarios, wild dogs’ food intake (and potentially survival and reproductive success), would be expected to improve at high temperatures, because impala (higher-value prey) would be more accessible due to being predictably located (Hypotheses 2 nocturnal prey concentration), in dense cover where they are vulnerable to predators (Hypothesis 4, shade-seeking), or more easily captured due to their tendency to overheat during high speed chases (Hypothesis 5, prey overheating). Hence, while demographic patterns cannot confirm the reduced foraging time hypothesis (Hypothesis 1) as the most likely mechanism whereby ambient temperature influences predator-prey interactions in this system, they do contribute to refuting Hypotheses 2-5.

Our findings suggest that climate change might have two wider impacts through its effect on predation by wild dogs. First, we have shown previously that wild dogs suppressed dikdik numbers, and dikdik browsing influenced tree abundance (Ford *et al.* 2015). Although we found that wild dog predation on dikdik did not trigger cascading effects on vegetation (Ford *et al.* 2015), rising environmental temperatures would be expected to intensify wild dog predation on dikdik, which might generate wider impacts on community structure.

Second, our findings suggest that wild dog populations’ resilience in the face of climate change might be affected by the abundance of small, abundant prey. Our study site is unusual both in the density of dikdik it supports (Augustine 2010; Ford et al. 2015) and the degree to which wild dogs rely on dikdik as primary prey (Woodroffe *et al.* 2007). Across much of the geographic range of wild dogs, dikdik are altogether absent (Kingdon & Hoffman 2013), and impala comprise the bulk of wild dog diets (Creel, Mills & McNutt 2004; Mbizah, Marino & Groom 2012). Without abundant small prey to hunt on hot days, we would expect wild dogs to more frequently make no kill at all, with potentially greater impacts on energy intake than may occur at our study site. Low reproductive success and high mortality at high ambient temperatures have been reported for wild dogs from two sites where impala are wild dogs’ principal prey (Woodroffe, Groom & McNutt 2017; Rabaiotti *et al.* In Review), although site-specific conditions complicate any comparison of the magnitude of temperature effects at the different sites.

Our results help to explain the negative impacts of high ambient temperatures on the survival and reproductive success of the African wild dog, an endangered species (Woodroffe & Sillero-Zubiri 2013). Our findings suggest that prey diversity may help to buffer wild dog populations against the effects of climate change, because abundant prey are readily located even when hunting time is constrained. Measures which maintain prey densities and diversity (such as limiting offtake by people) may help to conserve wild dogs in a warming climate. However, interventions which artificially raise prey densities (such as installation of waterholes) may risk increasing the densities of wild dogs’ predators (Creel & Creel 1996; Mills & Gorman 1997), and might have harmful consequences for wild dog conservation.

More generally, we have shown that ambient temperature can influence patterns of predation, even when both predator and prey are tropical endotherms. This finding contrasts with the assumptions of several general ecosystem and community ecology models (e.g., Dell, Pawar & Savage 2014 , Harfoot et al 2014), and suggests that such models may not accurately represent the responses of ecological communities in which endotherms play important roles. However, our findings also highlight the difficulty of accurately predicting exactly how environmental temperature would be expected to influence predation involving endotherms: all four of the scenarios that we investigated were plausible, but they generated conflicting hypotheses about how predation on impala and dikdik might vary in response to weather conditions, and only empirical testing indicated the true direction of the impact. Further investigations of how temperature-induced changes in the behaviour of predator and prey species together influence trophic interactions would help to build a more general picture of the relationship between ambient temperature and endotherm predation within ecological communities.

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|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| Table 1: Hypotheses and predicted changes in predation risk at high ambient temperatures | | | | |
| **Hypothesis** | **Species** | **Explanation** | **Implication for predator** | **Expected dietary change at high temperatures** |
| Hypothesis 1: Reduced daytime foraging | Wild dogs | Wild dogs are expected to spend less time hunting in daytime when temperatures are high1. | Dikdik live at higher densities so are predicted to be selected by a predator with limited time2. | Increased ratio of dikdik to impala |
| Hypothesis 2: Nocturnal prey concentration | Impala | Impala gather in glades at night. | Impala are predictably located at night making them easier to locate3. | Decreased ratio of dikdik to impala |
| Hypothesis 3a: Increased night time foraging | Wild dogs | Wild dogs are expected to increase activity at night, because daytime foraging is constrained1. | At high temperatures wild dogs are more active at night, when impala are easier to locate as they gather in glades3. Impala are therefore at higher risk when wild dogs hunt more at night. | Decreased ratio of dikdik to impala |
| Hypothesis 3b: Nocturnal prey targeting | Wild dogs | Wild dogs’ nocturnal activity is concentrated around glades because this is where their prey are predictably located. |
| Hypothesis 4: Shade seeking | All species | Wild dogs, impala and dikdik move into denser vegetation at high temperatures to avoid overheating. | Predation risk on impala increases as predation by wild dogs on impala is higher in denser vegetation4. | Decreased ratio of dikdik to impala |
| Hypothesis 5: Prey overheating | Impala | At high temperatures larger bodied animals overheat more quickly, and therefore impala overheat faster than dikdik when temperatures are hot. | Chase times for impala fall as they overheat more quickly than the smaller bodied wild dogs5. | Decreased ratio of dikdik to impala. |
| 1(Rabaiotti 2019), 4(Lucas 1983), 3(Augustine 2004); 3 (Ford *et al.* 2014); 4(Ford *et al.* 2014),5(Creel *et al.* 2016). | | | | |

T**able 2** Variables associated with the **duration of hunting periods** of African wild dogs during daytime, and the **occurrence** of night-time hunts. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc < 5) for the duration of hunting periods in the morning and evening (in minutes), and the occurrence (or not) of hunts at night. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

|  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | Estimate | | | Lower  95% CI | | Upper  95% CI | | | Variable  Importance (n) |
| Morning hunt duration (minutes) | Intercept | | | 220.06 | 198.78 | | 241.33 | | | — (4) |
| Denning (Yes) | | | -57.48 | -137.38 | | 22.41 | | | 1.00 (4) |
| Temperature (°C) | | | -1.33 | -2.007 | | -0.66 | | | 0.88 (3) |
| Moonlight night before | | | 0.054 | -0.24 | | 0.35 | | | 0.68 (2) |
| Rainfall (mm) | | | -2.97 | -6.31 | | 0.36 | | | 0.60 (1) |
| Rainfall:Temperature | | | 0.13 | 0.004 | | 0.25 | | | 0.60 (1) |
| Denning:Temperature | | | 5.015 | 2.83 | | 7.19 | | | 0.58 (1) |
| Evening hunt duration (minutes) | Intercept | | | 226.49 | 213.95 | | 239.04 | | | — (2) |
| Temperature (°C) | | | -3.01 | -3.45 | | -2.57 | | | 1.00 (2) |
| Moonlight | | | -1.84 | -2.063 | | -1.62 | | | 1.00 (2) |
| Moonrise | | | -0.006 | -0.12 | | 0.11 | | | 1.00 (2) |
| Denning (Yes) | | | 7.43 | -18.14 | | 33.003 | | | 1.00 (2) |
| Rainfall (mm) | | | -0.14 | -2.30 | | 2.16 | | | 0.20 (1) |
| Rainfall:Temperature | | | 0.013 | -0.071 | | 0.098 | | | 0.20 (1) |
| Denning:Temperature | | | 2.04 | 0.50 | | 3.58 | | | 0.15 (1) |
| Night-time hunt occurrence | Intercept | | -0.41 | | | -0.52 | | -0.30 | — (2) | | |
| Temperature | | 0.019 | | | 0.016 | | 0.023 | 1.00 (2) | | |
| Moonlight | | 0.031 | | | 0.028 | | 0.033 | 1.00 (2) | | |
| Denning | | -0.069 | | | -0.094 | | -0.044 | 0.52 (1) | | |

|  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | Estimate | Lower  95% CI | Upper  95% CI | Variable  importance (n) |
| impala - morning | Intercept | 0.05 | 0.03 | 0.06 | (1) |
| Temperature | 0.002 | 0.001 | 0.002 | 0.98 (1) |
| impala - midday | Intercept | 0.08 | 0.06 | 0.09 | (1) |
| Temperature | 0.001 | 0.001 | 0.002 | 1.00 (1) |
| impala - evening | Intercept | 0.09 | 0.09 | 0.1 | (2) |
| Wet *vs* dry phase | -0.002 | -0.004 | -0.0006 | 0.73 (1) |
| Temperature | 0.0005 | 0.0001 | 0.0008 | 0.2 (1) |
| impala - night | Intercept | 0.06 | 0.05 | 0.06 | (1) |
| Rainfall | 0.0002 | -0.0001 | 0.0006 | 0.86 (1) |
| dikdik - crepuscular | Intercept | 0.1 | 0.1 | 0.2 | (2) |
| Wet *vs* dry phase | 0.0055 | -0.0004 | 0.01 | 0.78 (1) |
| Temperature | -0.0008 | -0.001 | 0.000003 | 0.11 (1) |
| dikdik - midday | Intercept | 0.1 | 0.1 | 0.2 | (4) |
| Wet *vs* dry phase | -0.0002 | -0.006 | 0.006 | 0.59 (1) |
| Moonlight | 0.0009 | 0.0003 | 0.001 | 0.21 (1) |
| Rain (mm) | 0.0005 | 0.0002 | 0.0009 | 0.10 (1) |
| Temperature | 0.0004 | -0.0004 | 0.001 | 0.09 (1) |
| dikdik - night | Intercept | 0.1 | 0.1 | 0.2 | (2) |
| Moonlight | -0.02 | -0.03 | -0.01 | 0.71 (1) |
| Wet *vs* dry phase | 0.02 | 0.02 | 0.03 | 0.32 (1) |
| wild dog - crepuscular | Intercept | 0.2 | 0.1 | 0.2 | - (1) |
| Denning vs not | -0.04 | -0.05 | -0.02 | 1.00 (1) |
| wild dog - night | Intercept | 0.2 | 0.1 | 0.2 | - (1) |
| Denning *vs* not | -0.1 | -0.1 | -0.09 | 1.00 (1) |

**Table 3** Variables associated with **selection of woody cover** by African wild dogs, impala, and dikdik. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc < 5). The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

**Table 4** Variables associated with **use of glades** by wild dogs, impala, and dikdik. The table presents Estimated effects of explanatory variables included in the top model sets (ΔAICc < 5. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | Estimate | | Lower  95% CI | Upper  95% CI | Variable  importance (n) |
| impala in glade – 24h | Intercept | | 0.04 | 0.02 | 0.06 | (1) |
| Morning *vs* midday | | 0.06 | 0.06 | 0.07 | 1.00 (1) |
| Evening *vs* midday | | -0.009 | -0.01 | -0.005 | 1.00 (1) |
| Night *vs* midday | | 0.2 | 0.2 | 0.2 | 1.00 (1) |
| impala in glade - night | Intercept | | 0.3 | -0.007 | 0.03 | (1) |
| Moonlight | | -0.07 | -0.08 | -0.06 | 1.00 (1) |
| Rainfall | | -0.003 | -0.003 | -0.002 | 0.94 (1) |
| Temperature | | 0.006 | 0.004 | 0.008 | 0.06 (1) |
| dikdik in glade – 24h | Intercept | | 0.03 | 0.02 | 0.04 | — (1) |
| Wet *vs* dry phase | | -0.02 | -0.02 | -0.006 | 0.96 (1) |
| dikdik in glade - night | Intercept | | 0.01 | -0.007 | 0.03 | (1) |
| Moonlight | | 0.04 | 0.03 | 0.05 | 0.97(1) |
| wild dog distance to glade – 24h | Intercept | | 0.4 | 0.1 | 0.7 | — (4) |
| Denning *vs* not | | 0.04 | 0.02 | 0.06 | 0.74 (1) |
| Wet *vs* dry phase | | -0.01 | -0.03 | 0.01 | 0.18 (1) |
| Rainfall | | 0.002 | 0.0007 | 0.004 | 0.04 (1) |
| Temperature | | -0.007 | -0.005 | 0.004 | 0.04 (1) |
| wild dog distance to glade – night | Intercept | | 0.4 | 0.4 | 0.7 | (6) |
| Moonlight | | 0.06 | 0.02 | 0.1 | 0.69(3) |
| Wet *vs* dry phase | | -0.04 | -0.07 | -0.01 | 0.27 (2) |
| Rainfall | | 0.005 | 0.002 | 0.008 | 0.04 (1) |
| Denning *vs* not | | -0.04 | -0.08 | 0.005 | 0.04 (1) |
| Temperature | | -0.003 | -0.01 | 0.004 | 0.04 (1) |

**Table 5** Variables associated with **African wild dog consumption of impala**. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc<5) for African wild dog consumption of impala. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| Explanatory variable | Estimate | Lower 95% CI | Upper 95% CI | Variable importance (n) |
| Intercept | 0.74 | 0.5 | 1.01 | — (2) |
| Daily maximum temperature | -0.021 | -0.03 | -0.01 | 1.00 (1) |
| Community *vs* private land | -0.63 | -0.1 | -0.02 | 0.10 (1) |

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Figure 1 a) Mean woody cover use by impala, dikdik and wild dogs across time periods. Bars represent standard deviation. b) Mean wild dog distance to glades, in km, across time periods. Bars represent standard deviation. c) Probability of glade use by impala and dikdik across time periods. d) Duration of African wild dog morning and evening hunts at different daily maximum temperatures (°C). Points represent the means, lines indicate the conditional mean, and shaded areas represent the 95% confidence interval. e) Probability of the presence of impala in wild dog scat at different daily maximum temperatures (°C). The line represents the conditional mean, and shaded areas represent the 95% confidence interval.

**Temperature affects predator-prey interactions in an African savanna**

D. Rabaiotti1,2, Adam T. Ford3, Ben Chapple2, Sophie Morrill2,

Jacob Goheen4 and Rosie Woodroffe1

**Supporting Information**

**Figure S1** Frequency distribution of (A) start times and (B) stop times of African wild dog hunting periods, identified using collar-mounted accelerometers.



A

B

Time

**Figure S2** Daily movement patterns of African wild dogs (A) activity measured by collar-mounted accelerometers fitted to 18 wild dogs; blue bands denote the start and stop times used to classify morning and evening hunts (B) Mean maximum temperature throughout the day at the study site



**Figure S3** Percentage of hunts and total activity which fall in Morning, Midday, Evening and Night periods, and the percentage of bouts and total activity which spanned multiple time periods, and were therefore determined not to be hunts.

A screenshot of a cell phone

Description automatically generated

A close up of a map

Description automatically generatedF**igure S4** Distance to glades for wild dogs, impala and dikdik during morning, day, evening and night time periods.

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| **Table S1** Start and stop times of hunting periods allocated to each time period within the analysis. | | | | |
| Type of Bout | Earliest start time | Latest start time | Earliest stop time | Latest stop time |
| Morning | 04:55:39 | 06:56:04 | 07:30:41 | 12:51:33 |
| Midday | 06:56:05 | 16:01:54 | 12:51:34 | 18:47:02 |
| Evening | 16:01:55 | 18:26:52 | 18:47:03 | 21:06:56 |
| Night | 18:26:53 | 04:55:38 | 21:06:57 | 07:30:40 |

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| **Table S2** Temperature variance and difference in means between the time period the scats were collected (2001-2004) and when the GPS data was collected (2011-2016) | | |
| Variance 2001-2004 | Variance 2011-2016 | Difference in means |
| 3.61 | 4.01 | 0.73 |

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| **Table S2** Explanatory variables, other than temperature, included in models of wild dog hunts, and justification for their inclusion. | | | |
| Explanatory variable | Included in models for which time of day  (Morning, Evening, Night) | Justification | Citation |
| Denning status (yes/no) | All | Wild dogs are more active during the denning period | Woodroffe, Groom and McNutt 2017, Rabaiotti and Woodroffe 2019 |
| Total 24h rainfall (mm) | Both | Rainfall mitigates the impact of high temperatures on activity levels | Rabaiotti and Woodroffe 2019 |
| Moonlight (proportion illumination\*hours of moonlight) | Both (Moonlight the previous night was used in morning models, and the following night in evening models) | Wild dogs are more active on moonlit nights | Cozzi et al. 2012, Rabaiotti and Woodroffe 2019 |
| Moonrise (time) | Night | Wild dogs are more active on moonlit nights | Cozzi et al. 2012, Rabaiotti and Woodroffe 2019 |
| Moonset (time) | Night | Wild dogs are more active on moonlit nights | Cozzi et al. 2012, Rabaiotti and Woodroffe 2019 |

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| **Table S4** Explanatory variables, other than temperature, included in models of wild do, impala and dikdik habitat use, glade use and distance to glades and justification for their inclusion. | | | | |
| Explanatory variable | Included in models for which time of day  (Morning, Evening, Night) | Species | Justification | Citation |
| Denning status (yes/no) | All | Wild dogs | Wild dogs use more rugged areas with less vegetation during the denning period | Jackson et al 2014 |
| Total 24h rainfall (mm) | All | All | Rainfall mitigates the impact of high temperatures on activity levels and influences herbivore movement | Rabaiotti and Woodroffe 2019, Ford et al 2014 |
| Rainfall phase | All | All | Rainfall phase impacts herbivore movement and assemblages | Goheen et al 2013 |
| Moonlight (proportion illumination\*hours of moonlight) | Both (Moonlight the previous night was used in morning models, and the following night in evening models) | All | Wild dogs are more active on moonlit nights | Cozzi et al. 2012, Rabaiotti and Woodroffe 2019 |

**Table S5** Variables associated with the characteristics of African wild dog hunting periods in the **morning**. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc < 5) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

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| --- | --- | --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | | Estimate | Lower  95% CI | Upper  95% CI | Variable  importance (n) | |
| Occurrence | Intercept | | 0.80 | 0.77 | 0.83 | | — (1) |
| Moonlight before | | -0.0097 | -0.011 | -0.0079 | | 1.00 (1) |
| Duration  (minutes) | Intercept | | 220.06 | 198.78 | 241.33 | | — (4) |
| Denning *vs* not | | 57.48 | -137.38 | 22.41 | | 1.00 (4) |
| Temperature (°C) | | -1.33 | -2.007 | -0.66 | | 0.88 (3) |
| Moonlight before | | 0.054 | -0.24 | 0.35 | | 0.68 (2) |
| Rainfall (mm) | | -2.97 | -6.31 | 0.36 | | 0.60 (1) |
| Rainfall:Temperature | | 0.13 | 0.004 | 0.25 | | 0.60 (1) |
| Denning:Temperature | | 5.015 | 2.83 | 7.19 | | 0.58 (1) |
| Intensity | Intercept | | 51.43 | 46.78 | 56.076 | | — (3) |
| Denning *vs* not | | 1.56 | 0.14 | 0.16 | | 0.89 (2) |
| Temperature (°C) | | -0.26 | -0.38 | -0.14 | | 0.57 (2) |
| Start time | Intercept | 06:31:41 | | 06:24:07 | 06:39:14 | | (1) |
| Denning *vs* not | -00:07:46 | | -00:06:14 | -00:09:18 | | 1(1) |
| Temperature (°C) | -00:01:15 | | -00:01:01 | -00:01:29 | | 0.99(1) |
| Stop time | Intercept | | 09:54:52 | 09:20:48 | 10:28:56 | | (3) |
| Temperature (°C) | -00:02:28 | | -00:01:53 | -00:03:03 | | 0.72(2) |
| Denning *vs* not | 00:13:25 | | 00:09:37 | 00:17:14 | | 0.57(2) |

**Table S6** Variables associated with the characteristics of African wild dog hunting periods in the **evening**. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc < 5) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

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| --- | --- | --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | | Estimate | Lower 95% CI | Upper 95% CI | Variable  importance (n) | |
| Occurrence | Intercept | | 1.21 | 1.08 | 1.35 | | — (2) |
| Moonlight | | -0.010 | -0.012 | -0.008 | | 1.00 (2) |
| Temperature (°C) | | -0.017 | -0.021 | -0.013 | | 0.97 (2) |
|  | Denning *vs* not | | 0.042 | 0.016 | 0.069 | | 0.08 (1) |
| Duration  (minutes) | Intercept | | 226.49 | 213.95 | 239.04 | | — (2) |
| Temperature (°C) | | -3.01 | -3.45 | -2.57 | | 1.00 (2) |
| Moonlight | | -1.84 | -2.06 | -1.62 | | 1.00 (2) |
| Moonrise | | -0.006 | -0.12 | 0.11 | | 1.00 (2) |
| Denning *vs* not | | 7.43 | -18.14 | 33.00 | | 1.00 (2) |
| Rainfall (mm) | | -0.14 | -2.30 | 2.16 | | 0.20 (1) |
| Rainfall:Temperature | | 0.013 | -0.071 | 0.098 | | 0.20 (1) |
| Denning:Temperature | | 2.04 | 0.50 | 3.58 | | 0.15 (1) |
| Intensity | Intercept | | 62.98 | 46.78 | 56.08 | | — (4) |
| Temperature (°C) | | -0.83 | -0.96 | -0.70 | | 1.00 (4) |
| Denning *vs* not | | 7.50 | 2.55 | 12.45 | | 1.00 (4) |
| Moonlight | | -0.21 | -0.27 | -0.15 | | 0.45 (3) |
| Moonrise | | -0.006 | -0.038 | 0.027 | | 0.45 (3) |
| Rainfall (mm) | | -1.14 | -2.04 | -0.78 | | 0.14 (1) |
| Rainfall:Temperature | | 0.058 | 0.034 | 0.082 | | 0.14 (1) |
| Denning:Temperature | | -0.58 | -1.01 | -0.13 | | 0.07 (1) |
| Start time | Intercept | 15:43:55 | | 15:34:49 | 15:53:01 | | (2) |
| Temperature (°C) | 00:03:30 | | 00:03:11 | 00:03:49 | | 1.00 (2) |
| Denning *vs* not | -00:10:49 | | -00:08:52 | -00:09:18 | | 1.00 (2) |
| Moonlight | 00:01:01 | | 00:00:52 | 00:01:10 | | 0.90 (1) |
| Moonrise | -00:00:01 | | -00:00:05 | 00:00:05 | | 0.90 (1) |
| Stop time | Intercept | 19:42:14 | | 19:38:50 | 19:45:39 | | (2) |
| Denning *vs* not | 00:13:25 | | 00:09:37 | 00:17:14 | | 0.55(1) |
| Moonlight | 00:00:52 | | 00:00:43 | 00:01:02 | | 0.44(1) |
| Moonrise | 00:00:01 | | -00:00:04 | 00:00:06 | | 0.44(1) |

**Table S7** Variables associated with the characteristics of African wild dog hunting periods **at night**. The table presents estimated effects of explanatory variables included in the top model sets (ΔAICc < 5) for each outcome variable. The relative importance of each variable is shown along with the number of models in the top model set in which it was included (n). Bold highlighting indicates estimates for which the 95% confidence interval excluded zero.

|  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- |
| Outcome variable | Explanatory variable | | Estimate | Lower 95% CI | Upper 95% CI | Variable importance (n) | |
| Occurrence | Intercept | | -0.41 | -0.52 | -0.30 | | — (2) |
| Temperature | | 0.019 | 0.016 | 0.023 | | 1.00 (2) |
| Moonlight | | 0.031 | 0.028 | 0.033 | | 1.00 (2) |
| Denning | | -0.069 | -0.094 | -0.044 | | 0.52 (1) |
| Duration  (minutes) | Intercept | | 80.69 | 47.72 | 113.66 | | — (4) |
| Moonlight | | 5.59 | 5.00 | 6.18 | | 1.00 (4) |
| Moonrise | | 0.83 | 0.42 | 1.25 | | 1.00 (4) |
| Denning *vs* not | | 102.12 | -52.87 | 257.11 | | 0.94 (4) |
| Temperature (°C) | | 0.71 | -0.55 | 1.97 | | 0.79 (3) |
| Rainfall (mm) | | -0.54 | -7.66 | 6.57 | | 0.53 (2) |
| Rainfall:Temperature | | 0.004 | -0.27 | 0.27 | | 0.51 (1) |
| Denning:Temperature | | -7.42 | -13.76 | -1.10 | | 0.47 (2) |
| Intensity | Intercept | | 28.51 | 20.67 | 36.36 | | — (4) |
| Moonlight | | 0.86 | 0.75 | 0.98 | | 1.00 (4) |
| Moonrise | | -0.011 | -0.10 | -0.072 | | 1.00 (4) |
| Denning *vs* not | | 1.82 | -3.17 | 6.84 | | 0.75 (2) |
| Temperature (°C) | | 0.42 | 0.18 | 0.66 | | 0.52 (2) |
| Start time | Intercept | 15:52:55 | | 09:34:32 | 22:11:08 | | (2) |
| Moonrise | 00:24:22 | | 00:20:30 | 00:24:22 | | 1.00 (2) |
| Moonlight | -00:05:15 | | -00:10:38 | 00:00:08 | | 1.00 (2) |
| Temperature (°C) | -00:29:38 | | -00:18:57 | -00:40:20 | | 0.22 (1) |
| Stop time | Intercept | 09:14:33 | | 06:28:31 | 12:00:35 | | (3) |
| Moonlight | -00:24:17 | | -00:20:20 | -00:28:14 | | 1.00 (3) |
| Moonrise | 00:03:22 | | 00:00:31 | 00:06:12 | | 1.00 (3) |
| Denning *vs* not | 02:22:25 | | 01:11:11 | 03:33:39 | | 0.25 (1) |
|  | Temperature | 00:19:04 | | 00:11:10 | 00:26:57 | | 0.08 (1) |