

1 **Title:** What drives study-dependent differences in distance-decay relationships of microbial
2 communities?

3 **Running title:** Meta-Analysis of Microbial Distance-Decay Relationships

4 **Keywords:** Bacteria, Archaea, Eukarya, Mantel test, macroecology, biogeography, dispersal
5 limitation, community dissimilarity

6 Abstract

7 **Aim:** Ecological communities that exist closer together in space are generally more
8 compositionally similar than those far apart, as defined by the distance-decay of similarity
9 relationship. However, recent research has revealed substantial variability in the
10 distance-decay relationships of microbial communities between studies of different
11 taxonomic groups, ecosystems, spatial scales, as well as between those using different
12 molecular methodologies (e.g. high-throughput sequencing versus molecular fingerprinting).
13 Here, we test how these factors influence the strength of microbial distance-decay
14 relationships, to draw generalisations about how microbial β -diversity scales with space.

15 **Location:** Global.

16 **Time period:** Studies published between 2005-2019 (inclusive).

17 **Major taxa studied:** Bacteria, Archaea, and microbial Eukarya.

18 **Methods:** We conducted a meta-analysis of microbial distance-decay relationships, using
19 the Mantel correlation coefficient as a measure of the strength of distance-decay
20 relationships. Our final dataset consisted of 452 data points, varying in
21 environmental/ecological context or methodological approaches, and used linear models to
22 test the effects of each variable.

23 **Results:** Both ecological and methodological factors had significant impacts on the strength
24 of microbial distance-decay relationships. Specifically, the strength of these relationships
25 varied between environments and habitats, with soils showing significantly weaker
26 distance-decay relationships than other habitats, whilst increasing spatial extents had no
27 effect. Methodological factors such as sequencing depth were positively related to the
28 strength of distance-decay relationships, and choice of dissimilarity metric was also

29 important, with phylogenetic metrics generally giving weaker distance-decay relationships
30 than binary or abundance-based indices.

31 **Main conclusions:** We conclude that widely studied microbial biogeographic patterns, such
32 as the distance-decay relationship, vary by ecological context but are primarily distorted by
33 methodological choices. Consequently, we suggest that by linking methodological
34 approaches appropriately to the ecological context of a study, we can progress towards
35 generalisable biogeographic relationships in microbial ecology.

36 Introduction

37 The distance-decay of community similarity is one of the most widely studied relationships in
38 macroecology (Nekola & White, 1999; Soininen *et al.*, 2007). This relationship quantifies the
39 decrease in compositional similarity (β -diversity) between communities with increasing
40 geographic distance separating them, and demonstrates that nearby communities are more
41 similar to each other than distantly-separated communities. Distance-decay relationships
42 arise through several different, but often interacting ecological and evolutionary processes,
43 and consequently ecologists have extensively debated the underlying mechanisms that
44 generate such patterns (Nekola & White, 1999; Soininen *et al.*, 2007; Hanson *et al.*, 2012).
45 Spatial structuring of the environment can lead to distance-decay relationships, as
46 communities close together in space are likely to experience more similar environmental
47 conditions, and thus contain more similar communities than those situated in different
48 environmental conditions. Dispersal limitation can also lead to distance-decay relationships
49 by limiting the connectivity between communities, meaning that communities closer together
50 in space will share more species through localised dispersal than those further apart.

51 Distance-decay relationships are well documented in a multitude of plant and animal
52 communities (e.g. multiple taxa - Soininen *et al.*, 2007; urban plants - Sorte *et al.*, 2008;
53 multiple aquatic taxa - Astorga *et al.*, 2012; tropical amphibians - Basham *et al.*, 2019). Yet,
54 these relationships are of particular interest to microbial ecologists as microorganisms were
55 assumed to have ubiquitous distributions for several reasons. Firstly, their small size
56 facilitates passive dispersal over large geographic distances by vectors such as wind,
57 bio-aerosolization, ocean currents or migrating animals (Bisson *et al.*, 2007; Favet *et al.*,
58 2013; Joung *et al.*, 2017; Vašutová *et al.*, 2019), thus potentially overcoming dispersal
59 limitation as a contributing factor to microbial community composition. Additionally,
60 microorganisms often maintain high population densities in the environment leading to

61 dispersal by “mass effects”, whereby high dispersal rates from areas of increased population
62 density maintain populations in less optimal environments (Shmida & Wilson, 1985), helping
63 them to overcome the constraints of spatially-structured environmental gradients. Finally,
64 some microorganisms are able to enter dormant states, whether as vegetative cells or as
65 cysts or spores (Locey *et al.*, 2020), allowing them to survive and disperse through
66 suboptimal environments, simultaneously enhancing their dispersive abilities, and reducing
67 the influence of spatially-structured environmental gradients (Low-Décarie *et al.*, 2016).
68 Combined, these traits theoretically lower microbial β -diversity by increasing the amount of
69 shared species between distant communities, in turn leading to weaker distance-decay
70 relationships compared to macroorganisms. However, empirical tests of microbial
71 distance-decay relationships have yielded mixed results. Many studies have detected little or
72 no evidence of distance-decay relationships in microbial communities (Hazard *et al.*, 2013;
73 Kivlin *et al.*, 2014), whilst others report relationships of varying strengths, across a range of
74 spatial extents, study systems, and taxa (Dumbrell *et al.*, 2010; Martiny *et al.*, 2011; Clark *et*
75 *al.*, 2017). Thus, despite hundreds of empirical studies, the generality of spatial patterns in
76 microbial communities remains unclear, and we are no closer to understanding whether
77 variability in the spatial scaling relationships of microbial β -diversity originates from
78 ecological or methodological sources.

79 Variation in microbial distance-decay relationships could be due to different environmental or
80 ecological contexts in studies. Here, we consider environmental context as the variability in
81 the physico-chemical environment (e.g. temperature, pH, topology), and ecological context
82 as the total suite of species present and their interactions. The study systems commonly of
83 interest to microbial ecologists vary in terms of connectivity, which may facilitate or hinder
84 dispersal between communities, thus leading to weaker or stronger distance-decay
85 relationships, respectively. In well connected systems where dispersal is more feasible, such
86 as oceanic waters, distance-decay relationships should be weaker than systems in which

87 dispersal is limited, such as host-associated systems or soil systems, where distance-decay
88 relationships are weaker in deeper soil horizons (Li *et al.*, 2020). Moreover, study systems
89 differ in the spatially structured environmental gradients and heterogeneity they support.
90 Sediments and soils for example, can support strong environmental gradients over distances
91 of a few meters (Dumbrell *et al.*, 2010), and can be highly heterogeneous at the millimeter
92 scale (Vos *et al.*, 2013), strengthening distance-decay relationships. Additionally, different
93 study taxa are likely to yield variable distance-decay relationships because they differ in
94 traits that are linked to dispersal efficacy. For example, small cells disperse more efficiently
95 over long distances (Wilkinson, 2001; Wilkinson *et al.*, 2012; Norros *et al.*, 2014), thus
96 organisms with larger cell sizes, such as microbial Eukarya, should be more strongly
97 dispersal limited than those with small cell sizes, such as Bacteria (although this may not be
98 true for all taxa e.g. see Kivlin, 2020). Finally, it is known that spatial extent can influence our
99 perception of ecological relationships, which may contribute to variable distance-decay
100 relationships (Steinbauer *et al.*, 2012). Studies incorporating larger spatial extents may find
101 stronger distance-decay relationships as they are more likely to incorporate spatial scales at
102 which taxa are dispersal limited and/or at which environmental conditions become spatially
103 structured (Martiny *et al.*, 2011).

104 Whilst the context in which a study was undertaken may contribute to variability in microbial
105 distance-decay relationships, so too could different methodologies. Technological advances
106 have yielded new insight into the structure and functioning of development of environmental
107 microbial communities (Clark *et al.*, 2018). However, rapid turnover in molecular
108 methodologies means that our perception of microbial β -diversity patterns integrates
109 methods that vary substantially in both coverage (ability to detect a greater proportion of the
110 community in a given sample) and resolution (ability to resolve closely related taxa) (Muyzer,
111 1999; Glenn, 2011). Early methods such as clone library sequencing and community
112 fingerprinting methods (e.g. denaturing gradient gel electrophoresis (DGGE), terminal

113 restriction fragment length polymorphism (TRFLP), or phospholipid fatty acid (PLFA)
114 analysis) are limited in their ability to detect rare taxa (Bartram *et al.*, 2011), undoubtedly
115 missing rare taxa (Low-Décarie *et al.*, 2016). In turn, this could reduce the detected
116 β -diversity, inflating estimated community similarity and weakening distance-decay
117 relationships (Hanson *et al.*, 2012). In contrast, high-throughput sequencing (HTS) platforms
118 (also frequently referred to as next-generation sequencing (NGS)) can deliver sequencing
119 depths of tens or even hundreds of thousands of sequences per sample (Caporaso *et al.*,
120 2012), thus improving both community coverage (the detected proportion of a given
121 community), and allowing more samples to be examined in a single study (sample
122 coverage). Consequently, variation in the ability of molecular methods to resolve closely
123 related taxa, and to detect rare taxa can be an additional source of variability in microbial
124 β -diversity, which by extension can either weaken or strengthen microbial distance-decay
125 relationships.

126 In addition to the molecular methods, the choice of analytical methods, such as similarity
127 metric, can influence distance-decay relationships. The similarity of communities varies
128 according to the identity and abundance of the species present, their phylogenetic
129 relationships, and by external factors such as varying sample sizes. Thus, similarity metrics
130 that vary by one or more of these characteristics would likely result in contrasting
131 distance-decay relationships (Chao *et al.*, 2005; Barwell *et al.*, 2015). For example,
132 phylogenetic indices would be expected to yield weaker distance-decay relationships than
133 other metrics, because communities that have no species in common can still be highly
134 phylogenetically similar if the species share many branches of a phylogenetic tree, thus
135 reducing the decay of similarity over geographic distance (Bryant *et al.*, 2008). On the other
136 hand, quantitative indices compare not only the composition of species present, but also
137 their abundance in each community, reflecting finer-scale changes in community structure,

138 and thus should result in stronger distance-decay relationships by providing an additional
139 axis (species abundances) by which communities can differ.

140 Here, to disentangle the effects of both contextual (e.g. spatial extent, taxon, or ecosystem)
141 and methodological (e.g. means of identifying/differentiating taxa, or similarity metric)
142 variables on microbial distance-decay relationships, we undertook a meta-analysis to test
143 the following specific hypotheses:

- 144 • H₁ Bacteria and Archaea will show weaker distance-decay relationships than
145 micro-eukaryotic taxa due to their smaller size and higher population densities in
146 most environments.
- 147 • H₂ Environments that are able to maintain steep physicochemical gradients, such as
148 sediments and soils, will have stronger distance-decay relationships than those such
149 as seawater or air, where environmental gradients are more diffuse.
- 150 • H₃ Spatial extent will be positively related to the strength of the distance-decay
151 relationship as, at large spatial scales, increased dispersal limitation and
152 environmental heterogeneity will decrease the variance in community similarity at a
153 given spatial distance, resulting in stronger distance-decay relationships.
- 154 • H₄ High-throughput sequencing methods will yield stronger distance-decay
155 relationships due to: a) their ability to resolve closely related taxa, b) their greater
156 community coverage (e.g. number of sequences per sample, or number of
157 individuals counted per sample), and/or c) their greater sample coverage.
- 158 • H₅ Phylogenetic similarity metrics (e.g. Unifrac, beta nearest taxon index) will result
159 in weaker distance-decay relationships than other metrics as communities can be
160 phylogenetically similar, yet different at fine taxonomic resolutions, whilst quantitative
161 metrics (e.g. Bray-Curtis, Hellinger, Euclidean) will yield the strongest as they reflect
162 changes in both species composition and abundance.

163 **Methods**

164 *Meta-Analysis*

165 In order to test our hypotheses, we first gathered available data on microbial distance-decay
166 relationships via a systematic literature search. To do this, five search terms were selected to
167 detect relevant studies (Table 1). All literature searches were conducted using the Web of
168 Science search portal on 18/04/2020, and all results published between 1900-2019
169 (inclusive) were retained. To further filter the dataset to studies suitable for testing our
170 hypotheses, search results were downloaded and manually screened using the “metagear”
171 (Lajeunesse, 2016) package in R (version 3.4.1; R Core Team, 2019). Here, suitable studies
172 were those that tested the relationship between community similarity and geographic
173 distance in microbial communities, and not studies of “macroorganisms”, or studies of
174 strain-level genetic distance (e.g. using multi-locus sequence typing). Furthermore, studies
175 that did not test distance-decay relationships using Mantel correlation, or that used only
176 partial Mantel tests, were also discarded. We did not identify any potentially suitable studies
177 that were published prior to 1967, the year the Mantel test was described (Mantel, 1967),
178 and the earliest suitable study was published in 2005.

179 Table 1. Details of Web of Science search terms, and the number of results for each search.

Search	Search Term	Number of results
1	TS = (biogeograph*) AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	2907
2	TS = (macroecolog*) AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	136
3	TS = ("everything is everywhere") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	66
4	TS = ("geographic distance") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	220

5	TS = ("distance decay") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	186
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180 From these studies, we extracted Mantel correlation coefficients (r) as an effect-size
181 measure for each distance-decay relationship. The Mantel test is a permutation-based
182 method used to test for correlation between two distance matrices, or in the context of this
183 study, community (dis)similarity and geographic distance. The Mantel test statistic is an ideal
184 measure of effect size for use in meta-analytical frameworks for several reasons. Firstly, the
185 Mantel correlation test is the most frequently used method for testing distance-decay
186 relationships in microbial ecology (Franklin & Mills, 2007; Ramette, 2007). Secondly, as the
187 Mantel coefficient is a standardised correlation coefficient (i.e. is bound by -1 and 1), it
188 provides an easily interpretable and comparable measure of effect size (Harrison, 2012).

189 We ensured all Mantel correlation coefficients reflected correlations between geographic
190 distance and community dissimilarity, rather than similarity, by multiplying correlation
191 coefficients by -1 where necessary (so that positive values indicate a typical distance-decay
192 relationship). Partial Mantel statistics (which test for correlation between two matrices whilst
193 controlling for a third) were excluded as they are influenced by other variables included in
194 the test, and are therefore not easily comparable between studies. All Mantel correlation
195 coefficients were transformed to z-scores using Fisher's z transformation, as recommended
196 by Rosenberg *et al.*, (2013). All subsequent statistical analyses were conducted on the
197 transformed z-scores, whilst original Mantel correlation coefficients were used to make
198 figures, for ease of interpretation.

199 In order to test our hypotheses, several variables relating to the context and methodology of
200 each distance-decay relationship were recorded. Details of these variables are described in
201 Box 1.

202 Box 1. Details of the explanatory variables extracted from each study.

Resolution

Each distance-decay relationship was categorised into either high-resolution (high-throughput or Sanger sequencing), low resolution (molecular e.g. ARISA, TRFLP, DGGE, PhyloChip, PLFA), or low resolution (morphological), based on the method's ability to distinguish between closely related organisms.

Community Coverage

This refers to the sequencing depth in sequencing-based studies, or number of individuals counted in morphology-based studies, per sample. For sequencing studies, we recorded the number of sequences after rarefaction, or if this was not given, the average number of sequences per sample. As there is no comparable measure of coverage for fingerprinting studies, we excluded them from analyses of community coverage.

Sample Coverage

Sample coverage refers to the sample size (e.g. number of communities/samples) of each distance-decay relationship.

Dissimilarity Index

The dissimilarity index used to calculate each distance-decay relationship. Recorded dissimilarity indices were then categorised as quantitative (Bray-Curtis, Horn-Morisita, Euclidean, Hellinger, Theta), qualitative (Jaccard, Raup-Crick, Sørensen, Simpson, β sim), or phylogenetic (weighted or unweighted Unifrac, Rao, β -mean nearest taxon distance, β -mean pairwise distance).

Correlation Type

Studies were categorised according to the type of correlation coefficient used in the analysis distance-decay relationship (e.g. Spearman's or Pearson's correlation coefficient). The correlation type was only recorded if the type of correlation coefficient was explicitly mentioned.

Study Taxon

Each distance-decay relationship was binned into the following broad taxonomic categories based on the taxonomy of the focal organisms (Archaea, Bacteria, Fungi, or other microbial Eukarya), or combination of these categories if a relationship was based on multiple taxa (for example due to using sequencing primers that detect both Archaea and Bacteria). Fungi grouped separately from other micro-Eukaryotes due to their distinct reproductive strategy (e.g. spore-production) and the fact they are frequently targeted using distinct molecular approaches (e.g. via taxon-specific primer sets), in contrast to most other studies of micro-Eukarya.

Spatial Extent

This is the maximum distance separating communities in km. If this was not stated in text or provided in supplementary material (e.g. in a geographic distance matrix), it was calculated from given geographic coordinates, estimated from a plot of the distance-decay relationship, or estimated from scaled maps.

Environment

We broadly categorised distance-decay relationships based on the type of environment (agriculture, air, aquifer, coastal wetlands/intertidal, desert, dune, forest, glacier, grassland, lake, marine, coastal marshes, mine, river, snow, urban) within which they were sampled.

Whilst these categories are not mutually exclusive, we categorised each study based on which environment best represented the environmental context in which each study was undertaken. For studies on lakes, we also recorded whether relationships originated from a single lake, or across multiple lakes.

Habitat

The type of environmental material that the sampled communities occupied. We categorised distance-decay relationships as: air, host-associated, sediment, snow, soil, water.

203 *Statistical Analyses*

204 In order to determine whether distance-decay relationships varied between categorical
205 variables (as in hypotheses 1, 2, 4, and 5), we used ANOVA tests. In tests where significant
206 differences between groups were found, Tukey's Honest Significant Difference (HSD) tests
207 were used to determine which groups were different. Linear mixed-effect models were used
208 to test relationships between the strength of distance-decay relationships and continuous
209 variables such as spatial extent and community coverage, using a random intercept to
210 account for heteroscedasticity due to some studies contributing multiple relationships. The
211 variables spatial extent and community coverage were initially \log_{10} transformed to aid model
212 fitting, as they spanned several orders of magnitude. To compare the overall influence of
213 ecological vs methodological factors on microbial distance-decay relationships, we
214 compared two full models (including all relevant variables) using AIC scores, on a subset of
215 the data for which all variables were successfully recorded. We report the results of all null
216 hypothesis tests in terms of statistical "clarity" rather than "significance", in line with
217 recommendations from Dushoff *et al.*, (2019)

218 **Results**

219 Our Web of Science searches resulted in 2,982 unique search results. Manual screening of
220 the abstracts yielded 951 studies that were deemed to be potentially suitable for use in this
221 analysis. A total of 452 Mantel correlation coefficients were successfully obtained from 187

studies represented in 61 journals (Fig. S1). Reported Mantel correlation coefficients ranged from -0.33 to 0.95, with a mean of 0.27 (std. error = 0.011), whilst a summary of the variables collected is shown in Table 2.

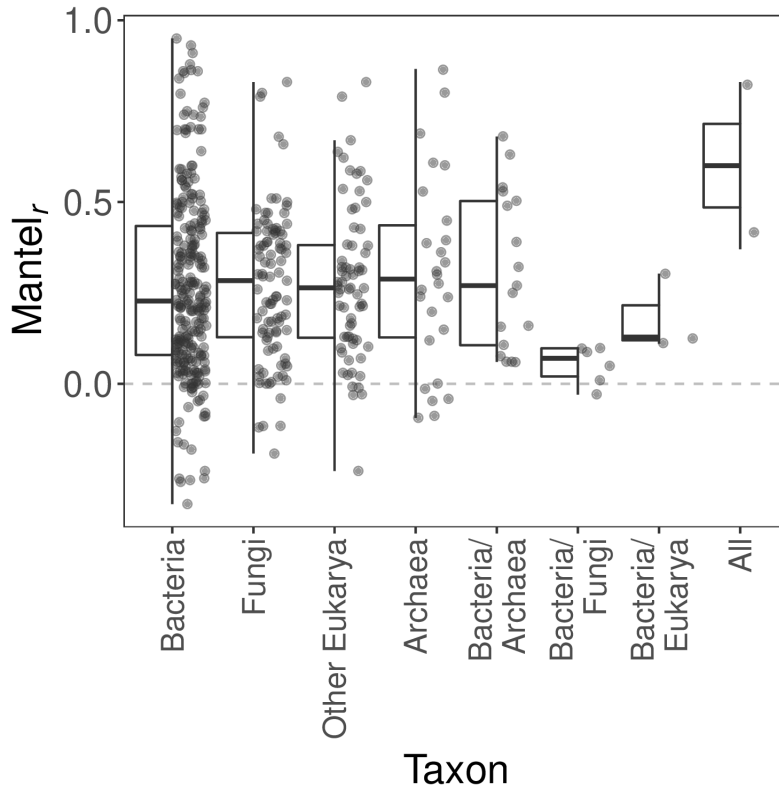
Table 2. Summary of collected data. For categorical variables, the number of individual distance-decay relationships in each category are shown, whereas minima, maxima, median and mean values are shown for continuous variables. Detailed descriptions of each variable are found in Box 1, and raw data can be found in Table S1.

Ecological variables		Methodological variables	
Variable	Summary	Variable	Summary
Study taxon	Archaea: $n = 26$ Bacteria: $n = 238$ Eukarya: $n = 67$ Fungi: $n = 93$ Archaea + Bacteria: $n = 17$ Bacteria + Eukarya: $n = 3$ Bacteria + Fungi: $n = 6$ All: $n = 2$	Resolution	High: $n = 345$ Intermediate: $n = 84$ Low: $n = 23$
Spatial extent (km)	Min = 0.0001 Mean = 1,543 Median = 220 Max = 18,700 NA = 15	Community coverage (number of individuals/sequences)	Min = 8 Mean = 217,357 Median = 1,257 Max = 34,192,561 NA = 115
Environment type	Agriculture: $n = 16$ Air: $n = 13$ Aquifer: $n = 1$ Coastal: $n = 8$ Desert: $n = 4$ Dune: $n = 1$ Forest: $n = 76$ Glacier: $n = 5$ Grassland: $n = 96$ Lake: $n = 76$ Marine: $n = 88$ Marsh: $n = 3$ Mine: $n = 1$ River: $n = 57$ Snow: $n = 3$ Urban: $n = 4$	Dissimilarity index	β -MNTD: $n = 13$ β -MPD: $n = 1$ β -sim: $n = 4$ Bray-Curtis: $n = 218$ Bray-Curtis _{Sim} : $n = 3$ Bray-Curtis _{Nes} : $n = 1$ Canberra: $n = 1$ Euclidean: $n = 9$ Hellinger: $n = 4$ Jaccard: $n = 49$ Mash: $n = 2$ Morisita-Horn: $n = 4$ Rao: $n = 2$ Raup-Crick: $n = 19$ Simpson: $n = 2$ Sorensen: $n = 42$ Theta: $n = 1$

			Unweighted Unifrac: $n = 17$ Weighted Unifrac: $n = 59$ NA: $n = 1$
Habitat type	Air: $n = 16$ Host: $n = 75$ Sediment: $n = 78$ Snow: $n = 3$ Soil: $n = 141$ Water: $n = 137$ NA: $n = 2$	Correlation type	Pearson: $n = 62$ Spearman: $n = 86$ NA: $n = 304$
		Sample coverage (Number of samples)	Min = 4 Mean = 52.88 Median = 25 Max = 1,010 NA = 1

229 *Influence of Context on the Distance-Decay Relationship*

230 In order to determine whether contextual factors can influence the strength of
231 distance-decay relationships, the influence of ecological factors including study taxa, study
232 system, and spatial scale were tested. Within the dataset, the most commonly studied taxa
233 were Bacteria ($n = 238$), followed by Fungi ($n = 93$), other microbial Eukaryotes ($n = 67$),
234 and Archaea ($n = 26$). We found no clear differences in the strength of distance-decay
235 relationships between these taxa ($F_{5, 441} = 0.99$, $P = 0.43$), although distance-decay
236 relationships incorporating bacterial and fungal communities showed the weakest
237 relationships, albeit only from six studies (Fig. 1).

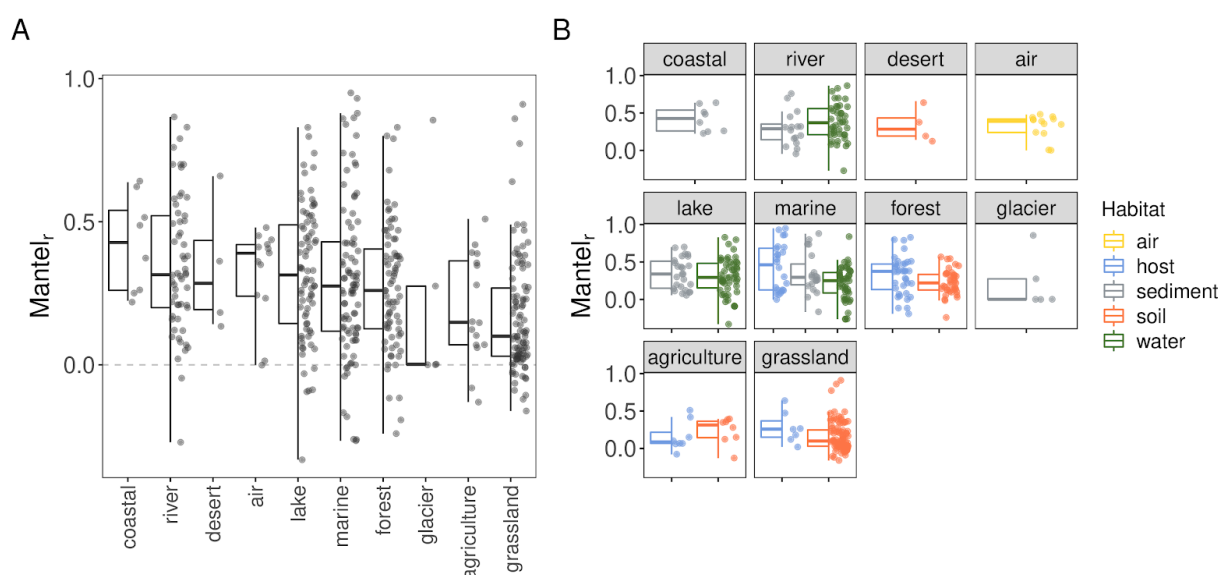


238 Figure 1. The strength (Mantel_r) of distance-decay relationships based on different study
 239 taxa. A larger Mantel_r value indicates a stronger distance-decay relationship. The “All”
 240 category consists of studies that incorporated all microbial taxonomic groups, whilst
 241 combined categories (e.g. Bacteria/Archaea) incorporate communities from multiple
 242 taxonomic groups (e.g. bacterial and archaeal communities).

243 The distance-decay relationships in our dataset originated from 16 different environments.
 244 Of these, five were represented by three, or fewer, distance-decay relationships, and so
 245 were excluded from further analyses (marsh; $n = 3$, snow; $n = 3$, dune, mine, aquifer; $n = 1$).
 246 The most frequently studied environments were grasslands ($n = 96$), marine ($n = 88$), and
 247 lakes and forests ($n = 76$ for both). We found clear differences in the strength of
 248 distance-decay relationships between environments (Fig. 2A; $F_{10, 432} = 3.187$, $P < 0.001$).
 249 Specifically, and perhaps counter-intuitively, grassland-based studies had weaker
 250 distance-decay relationships than those from aquatic environments such as lakes, rivers, or

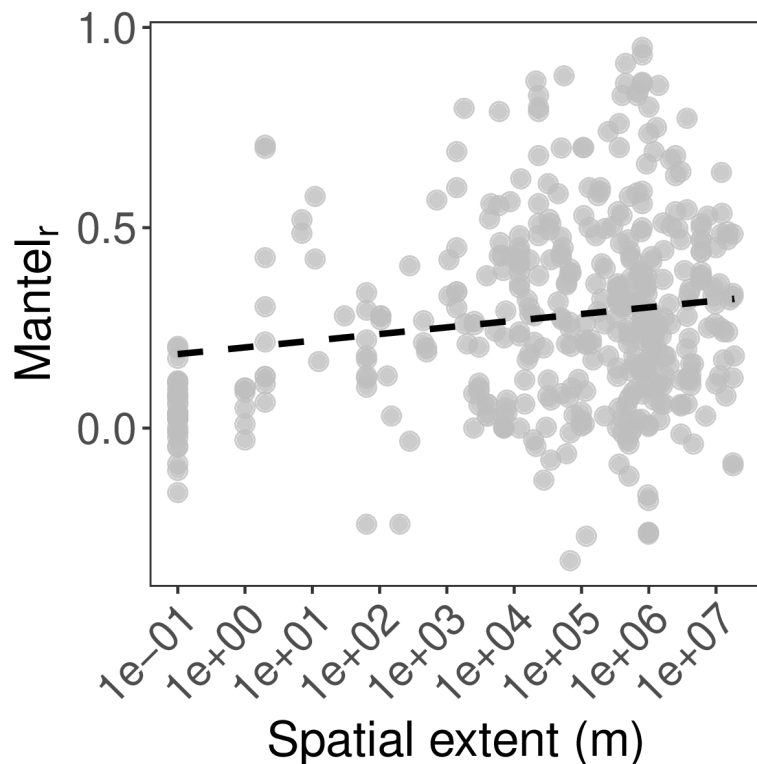
251 the marine environment ($|\text{coef}| > 0.17$, $P < 0.05$ for all comparisons). Urban environments,
 252 which included built environments such as sewers and indoor air, also produced weak
 253 distance-decay relationships, although with only four data points, this difference was not
 254 statistically clear ($P > 0.43$ for all comparisons). We also found no difference in the strength
 255 of distance-decay relationships between studies conducted in single lakes compared to
 256 those incorporating multiple lakes ($F_{1, 74} = 0.11$, $P = 0.74$), despite the average spatial extent
 257 of multiple-lake studies being approximately 32-fold greater than that of single-lake studies
 258 (Fig. S2).

259 A more detailed analysis of the interaction between environment type and habitat revealed
 260 that, whilst environments ($F_{9, 420} = 3.29$, $P < 0.001$) and habitat ($F_{3, 420} = 6.65$, $P < 0.001$)
 261 differ from each other, their interaction was not statistically significant ($F_{4, 420} = 1.93$, $P =$
 262 0.10). In fact, within environments, only marine host-associated and marine water-based
 263 distance-decay relationships were clearly different from each other (Fig. 2B), with
 264 host-associated communities showing significantly stronger distance-decay relationships
 265 ($\text{coef} = 0.35$, $P < 0.001$).



266 Figure 2. Variation in Mantel correlation coefficients of distance-decay relationships between
 267 different environments (A) and habitat types (B). Environment categories are arranged from
 268 strongest to weakest mean distance-decay relationship.

269 The spatial extents of recorded distance-decay relationships ranged from 10 cm to more
 270 than 18,000 km, and minimal spatial extents varied notably across environments and
 271 habitats, with terrestrial and soil-based studies often conducted over smaller spatial scales
 272 (Fig. S3). After accounting for differences between studies, we found no evidence of a
 273 statistically clear relationship between the spatial extent of a study and the strength of the
 274 observed distance-decay relationship (coef = 0.02, marginal $R^2 = 0.020$, $t = 1.58$, $P = 0.11$).
 275 Finally, as larger spatial scale studies might also incorporate greater sampling coverage, we
 276 tested for collinearity between the spatial scale of a study and the sampling coverage, but
 277 found no correlation between these variables ($\rho = 0.06$, $P = 0.19$).



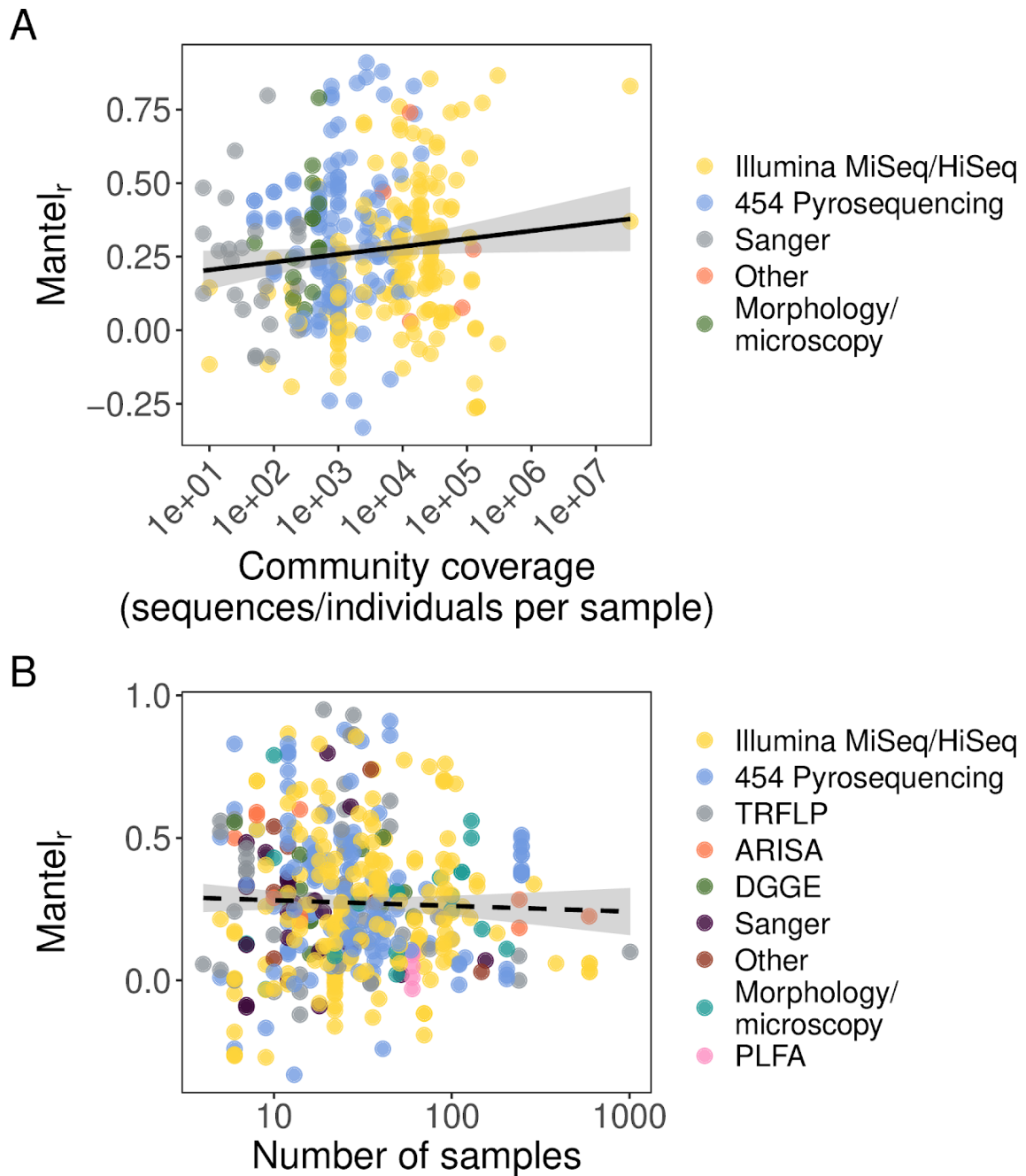
278 Figure 3. The relationship between spatial extent and the Mantel correlation coefficient of
279 microbial distance-decay relationships. The dashed line represents the fit of a mixed-effects
280 model between the \log_{10} of spatial extent and Mantel correlation coefficient, with a
281 study-dependent random intercept.

282 *Influence of Methodological Factors on the Distance-Decay Relationship*

283 We grouped community characterisation methods according to their ability to distinguish
284 between closely related taxa. There were no clear differences in the strength of
285 distance-decay relationships between different resolution methods ($F_{2, 449} = 0.562$, $P = 0.57$),
286 nor were there clear differences between different molecular methods (Fig. S4, $F_{7, 437} = 1.97$,
287 $P = 0.06$), considering only those methods that had >4 distance-decay relationships across
288 the entire dataset (excluding Ion Torrent; $n = 4$, phylo-chip; $n = 2$, and Pac-Bio; $n = 1$).

289 Whilst we observed no differences in distance-decay relationships between different
290 resolution methods, after accounting for study-dependent differences, we found a positive
291 relationship between (\log_{10}) community coverage and the strength of microbial
292 distance-decay relationships (Fig. 4A; $n = 337$, conditional $R^2 = 0.57$, coef = 0.06, $t = 2.73$, P
293 < 0.01), although the marginal effect of community coverage was weak (marginal $R^2 = 0.04$).

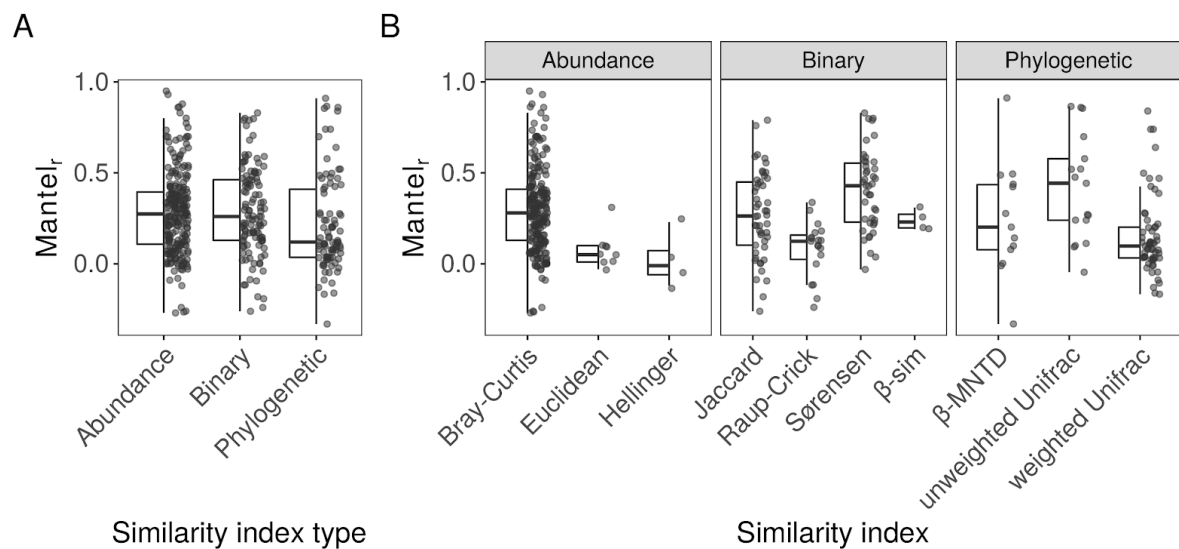
294 The logistics of multiplexing samples on high-throughput sequencing runs means that there
295 is often a trade-off between the community coverage and sampling coverage of a study.
296 However, we found no evidence of negative correlation between these two factors
297 (Pearson's $\rho = -0.03$, $P = 0.54$). Nor did we detect any clear relationship between the
298 number of samples (\log_{10} sample coverage) and the strength of distance-decay relationships,
299 even after accounting for study-specific differences with a mixed effects model (Fig. 4B; $n =$
300 451, coef = -0.06, marginal $R^2 = 0.01$, $t = -1.40$, $P = 0.16$).



301 Figure 4. The relationship between the strength of microbial distance-decay relationships
 302 (Mantel_r) and A) community coverage, quantified as the number of sequences or individuals
 303 counted per sample, and B) sample coverage, quantified as the number of individual
 304 samples used to construct distance-decay relationships. Points are individual Mantel
 305 correlation coefficients, coloured by the molecular technique used to characterise the

306 microbial community. Solid lines indicate statistically significant relationships ($P < 0.05$),
307 whilst dashed lines indicate non-significant relationships ($P > 0.05$), and shaded grey ribbons
308 represent 95% confidence intervals. Abbreviated molecular methods in the legend are
309 defined as follows (TRFLP = Terminal Restriction Fragment Length Polymorphism; ARISA =
310 Automated Ribosomal Intergenic Spacer Analysis; DGGE = Denaturing Gradient Gel
311 Electrophoresis; PLFA = Phospholipid Fatty Acid analysis).

312 Choice of similarity index also had a clear impact on the strength of microbial distance-decay
313 relationships. As well as recording the specific similarity index used, we categorised indices
314 into types (binary, abundance, or phylogenetic) to test for broad differences in
315 distance-decay relationships. We analysed the nested interaction between similarity index
316 and index type, and found no clear differences between different index types (Fig. 5A; $F_{2, 424}$
317 = 1.48, $P = 0.23$). However, the interaction between index type and similarity index was
318 significant ($F_{7, 424} = 7.20$, $P = 0.001$). Post-hoc analysis revealed differences between similarity
319 indices within and between index types (Fig. 5B). Distance-decay relationships based on the
320 Raup-Crick index were weaker than those based on either Sørensen (coef = -0.38, $P < 0.01$)
321 or unweighted Unifrac indices (coef = -0.44, $P < 0.01$), whilst those based on weighted
322 Unifrac were weaker than both Sørensen (coef = -0.29, $P < 0.001$) and unweighted Unifrac
323 (coef = -0.35 $P < 0.05$). Finally, most studies did not explicitly state the correlation type used
324 to generate each Mantel test ($n = 304$), but of those that did, Spearman's correlation
325 coefficient was more frequently used ($n = 86$) than Pearson's ($n = 62$). We found no clear
326 difference in the strength of microbial distance-decay relationships using these two methods
327 ($F_{1, 146} = 2.47$, $P = 0.12$).



328 Figure 5. Variation in the strength of microbial distance-decay relationships (*Mantel_r*)
 329 calculated with different similarity index types (A), or individual indices (B). Only indices with
 330 four or more distance-decay relationships were plotted for clarity.

331 *Comparison of Contextual and Methodological Variables*

332 In order to determine whether eco-environmental context or methodological factors better
 333 explain the strength of microbial distance decay relationship, we specified two models, with
 334 variables from these two categories, using a subset of the original data for which values
 335 were obtained for all variables ($n = 323$). Each model had four variables, and used similar
 336 degrees of freedom (context model $df = 26$, methodological model $df = 27$). The
 337 methodological model outperformed the contextual model in terms of both AIC (Akaike
 338 Information Criterion) and R^2 measures of model performance (Table 3). Notably, neither
 339 model explained a high proportion of the variance, although both AIC and likelihood ratio
 340 tests supported both models over a null (intercept only) model.

Table 3. Comparison of models specified using either contextual, or methodological variables. Akaike Information Criterion (AIC) and adjusted R^2 quantify the likelihood and fit of a model relative to the number of predictor variables, respectively.

Model	AIC	Adj- R^2	Likelihood ratio comparison to null (intercept only) model			
			Δ AIC	Sum of squares	F (df)	P value
Contextual	146.89	0.11	-13.69	5.34	2.61	< 0.001
Methodological	134.11	0.14	-26.46	6.47	3.17 (25)	< 0.001

Discussion

Previous research into the spatial ecology of microbial communities has not yielded a consistent distance-decay relationship. Our meta-analysis of 452 microbial distance-decay relationships suggests that the reasons for this lack of consistency are two-fold. Firstly, the differing contexts within which studies are conducted contribute variability to reported distance-decay relationships. In particular, we found that differing study systems were associated with variation in microbial distance-decay relationships. Secondly, methodological differences between studies, including dissimilarity index, data resolution, and sample coverage, all significantly affected observed distance-decay relationships. A central tenet of macroecology is the search for universal patterns and relationships; our results suggest generalisable relationships may only emerge when methodological approaches are appropriately coupled to ecological context.

Our comparison of distance-decay relationships between different study systems revealed that agricultural and especially grassland-based studies had weaker relationships than studies of other environments. Within these environments, soils were by far the most

359 frequently studied habitat, and we initially expected that, as soils maintain strong
360 physicochemical gradients over small vertical and horizontal spatial scales (e.g. Dumbrell *et al.*,
361 2010), that these distance-decay relationships would be stronger than in other
362 environments or habitats. It is possible that the environmental gradients present in soils do
363 not change linearly over geographic distance, for example if similar environmental conditions
364 are patchily distributed. Alternatively, soil microorganisms may be able to disperse more
365 effectively than previously thought, perhaps via association with other soil organisms (e.g.
366 bacterial migration along fungal hyphae; Warmink *et al.*, 2011), migratory species such as
367 birds (Bisson *et al.*, 2007), wind blown soil particles (Favet *et al.*, 2013), or via bioaerosols
368 (Joung *et al.*, 2017). The depth profile over which soil samples integrate may also play a role
369 in obscuring distance-decay relationships, as surface soils show stronger distance-decay
370 relationships than deeper ones, likely due to the greater intensity of dispersing propagules
371 entering the surface (Li *et al.*, 2020). Furthermore, soils harbour extensive microbial “seed
372 banks” of dormant organisms and/or relic DNA that could weaken the distance-decay
373 relationship (Lennon & Jones, 2011; Carini *et al.*, 2016; Lennon *et al.*, 2018). Dormant cells
374 and relic DNA are not subject to environmental selection yet, are routinely detected in
375 molecular community assays, likely diminishing the perceived effects of spatially-structured
376 environmental selection on microbial communities (Locey *et al.*, 2020). Thus, in habitats
377 such as soils, distinguishing dormant from active cells could result in stronger
378 distance-decay relationships than those recorded previously, although evidence of the same
379 effect on distance-decay slopes is mixed (Meyer *et al.*, 2018; Locey *et al.*, 2020). The extent
380 to which this phenomenon plays a role in other environments is also unclear.

381 Originally, we expected the weakest distance-decay relationships to occur in connected
382 aquatic environments such as rivers, oceans, or within single lakes, as the movement of
383 water may provide an effective dispersal mechanism, homogenising microbial communities
384 over larger spatial and environmental distances. In contrast, we found that aquatic

385 communities actually showed stronger distance-decay relationships. Soininen *et al.* (2007)
386 recorded similar distance-decay rates between terrestrial, marine and aquatic ecosystems,
387 showing that context-dependent distance-decay relationships may be a feature of microbial
388 communities. We also found that the strength of distance-decay relationships was not
389 different in studies based on single, or multiple, lakes, despite the difference in spatial
390 extents of these studies. Lakes act as habitat islands within a terrestrial matrix and so
391 dispersal limitation and environmental heterogeneity should be greater across multiple lakes
392 than within a single lake, resulting in stronger distance-decay relationships in multi-lake
393 studies. One explanation is that catchment-scale environmental parameters such as geology
394 may homogenise environmental conditions across multiple lakes, meaning that
395 environmental distances are similar within and between lakes. Alternatively, other
396 biogeographic processes such as mass effects may homogenise communities between
397 hydrologically connected lakes (Lindström & Bergström, 2004), especially where lakes are of
398 different sizes (Reche *et al.*, 2005). Host-associated communities showed relatively strong,
399 but variable distance-decay relationships. We suggest that this is caused jointly by the
400 ecology of the host species, in combination with the degree of host-specificity with the
401 associated microbiome. For example, if the host is not dispersal limited, and associates with
402 a large variety of microorganisms, then the distance-decay relationship may be relatively
403 weaker than those of either dispersal limited hosts, or highly specific associated
404 microbiomes.

405 The scale-dependence of various biogeographical relationships is well studied
406 (Hillebrand, 2004; Bissett *et al.*, 2010; Martiny *et al.*, 2011; Soininen *et al.*, 2011), albeit with
407 contrasting results. Soininen *et al.* (2011) reported that distance-decay relationships of
408 various microbial communities were generally steeper over greater spatial extents, whilst our
409 results suggest that increasing spatial extent does not significantly increase the strength of
410 distance-decay relationships. As we analysed distance-decay strength rather than

411 steepness, our results are not necessarily contradictory. A strong distance-decay
412 relationship occurs when, at a given spatial distance, all pairs of communities are equally
413 dissimilar to one another, whereas a steep distance-decay occurs when communities
414 separated by different distances are highly dissimilar to each other. We initially expected that
415 spatial extent might alter the strength of distance-decay relationships as, at greater
416 distances, decreased dispersal and increased environmental heterogeneity should reduce
417 the variance in compositional similarity between pairs of communities (at a given distance).
418 Instead, it could be that the spatial configuration or connectivity of the communities could be
419 more important than spatial extent *per se*. For example, at a given spatial distance, some
420 pairs of communities could be linked by dispersal and others not, increasing the variation in
421 community similarity at each distance, and weakening the distance-decay relationship. In
422 practice, this could occur in lake systems where at a certain geographic distance, some pairs
423 of communities fall within the same lake and some in different lakes or when long-distance
424 dispersal vectors link some pairs of communities separated by large distances, but not
425 others, as has been proposed for halophilic microbial communities dispersing on migratory
426 birds for example (Clark *et al.*, 2017; Kemp *et al.*, 2018). Furthermore, we observed that the
427 minimum spatial extents differed according to the environment they were conducted in.
428 Studies from terrestrial environments (e.g. grasslands and forests) or those based on soils
429 generally incorporated smaller spatial extents than those based on aquatic systems (with the
430 exception of some host-associated marine studies) or on habitats such as water or air. This
431 could be due to the logistics of sampling at small scales. For example, sampling planktonic
432 microbial communities at small (centimeters to meters) scales could be confounded by
433 mixing caused by the sampling process or by tidal movements of water. Additionally, since
434 many studies analysing microbial distance-decay relationships aimed to discern between
435 environmental and spatial effects on microbial communities, it may be widely assumed that
436 aquatic environments are more homogenous and/or that microorganisms are not dispersal

437 limited at these scales compared to more physically stable environments such as soils or
438 sediments.

439 Distance-decay relationships are frequently interpreted as evidence for neutral community
440 assembly processes such as dispersal limitation, in the microbial literature. Across microbial
441 taxa, cell size is a trait thought to influence dispersal efficacy (Wilkinson, 2001; Wilkinson *et al.*,
442 2012; Zinger *et al.*, 2019), and so larger microorganisms such as micro-Eukarya should
443 show stronger distance-decay relationships than smaller microorganisms such as Bacteria
444 or Archaea. However, we found no evidence for this, suggesting that phylogenetically
445 structured traits such as cell size may be less important compared to other contextual and
446 methodological factors, or that the broad domain-level classification used here does not
447 sufficiently capture different microbial cell sizes. As discussed previously, distance-decay
448 relationships can arise from spatially autocorrelated environmental gradients as well as
449 dispersal limitation (Nekola & White, 1999). Therefore, the lack of differences in
450 biogeographic patterns observed at the domain level may be the result of a trade-off
451 between dispersal-related processes and environmental filtering. For instance, bacterial
452 distance-decay relationships may be less strongly influenced by dispersal than
453 environmental filtering, and vice versa for Eukarya. Consequently, these influences may
454 balance out at broad taxonomic levels, resulting in similar biogeographic patterns at the
455 domain level.

456 In comparison to contextual factors, methodological factors were found to have a greater
457 influence on microbial distance-decay relationships. The development of molecular methods,
458 including high-throughput sequencing platforms, has vastly improved our ability to
459 characterise microbial communities (Roesch *et al.*, 2007; Caporaso *et al.*, 2012). However,
460 these methods differ in their resolution, community coverage, and ability to multiplex large
461 numbers of samples, all of which we hypothesised could strengthen or weaken

462 distance-decay relationships by altering our estimation of microbial β -diversity. In contrast,
463 we observed only a weak relationship between the strength of distance-decay relationships
464 and community coverage, and no clear effects of different resolution methods, or the number
465 of samples, suggesting that molecular methodology may not play as large a role in
466 determining microbial biogeographic patterns as previously thought.

467 The ability to resolve closely related taxa has previously been found to be an important
468 determinant of our ability to detect biogeographical patterns, as such patterns may only
469 emerge when taxa are defined at sufficiently high resolution (Hanson *et al.*, 2012). Yet, other
470 studies show that bioinformatically altering taxonomic resolution frequently has little effect on
471 microbial biogeographic patterns. For example, increasing the similarity threshold at which
472 operational taxonomic units are defined is thought to be equivalent to increasing the
473 taxonomic resolution (Callahan *et al.*, 2017). Yet, empirical biogeographic relationships often
474 appear robust to such manipulation, in a variety of taxa and ecosystems (Clark *et al.*, 2017;
475 Glassman & Martiny, 2018; Meyer *et al.*, 2018), supporting our finding that resolution may
476 not be important. Perhaps most molecular methodologies operate above resolutions at
477 which biogeographic patterns begin to change, or more worryingly, perhaps we are still
478 studying microbial biogeography at too low a resolution.

479 Aside from resolution, another important variable related to molecular methodology is
480 community coverage. One of the few universal patterns that appears to hold true for most
481 microbial communities is the “long-tailed” species abundance-distributions (Dumbrell *et al.*,
482 2010; Shoemaker *et al.*, 2017; Maček *et al.*, 2019), which is caused by the majority of
483 microorganisms in a community being rare. The rarer taxa in microbial communities also
484 tend to be the least widespread (Clark *et al.*, 2017; Lindh *et al.*, 2017; Meyer *et al.*, 2018;
485 Shade & Stopnisek, 2019) and thus, only detecting the more abundant, widespread
486 organisms would overestimate compositional similarity across communities, and

487 consequently, weaken distance-decay relationships due to the lower rate of turnover (Meyer
488 *et al.*, 2018). Perhaps of more concern is that even with existing sequencing platforms, our
489 surveys of environmental microbial communities still miss taxa that are vanishingly rare in
490 the environment, such as extremophiles that persist in non-extreme habitats (Low-Décarie *et*
491 *al.*, 2016). The ability of common species to reflect ecological patterns of the wider
492 community is debated (Galand *et al.*, 2009; Heino & Soininen, 2010; van Dorst *et al.*, 2014)
493 and is linked to a wider debate on the ecological importance of rare species that is far
494 beyond the scope of this work (e.g. Gaston, 2012). However, rare microorganisms are well
495 known to be of critical importance in the context of environmental perturbations (Shade *et*
496 *al.*, 2014; Low-Décarie *et al.*, 2016) and in providing ecosystem processes (e.g.
497 sulfate-reduction in peat soils, Hausmann *et al.*, 2016; and anaerobic ammonia-oxidation in
498 river sediments Lansdown *et al.*, 2016) and as a result, ignoring them may further distance
499 biogeographic patterns from ecosystem-level processes.

500 Against expectation, we observed no clear differences in distance-decay relationships using
501 different similarity metric types, and differences between specific metrics were minimal.
502 Distance-decay relationships based on the weighted Unifrac distance and the Raup-Crick
503 index were weaker than those based on other metrics. The Raup-Crick index is less
504 influenced by concurrent changes in species richness between communities, and as such is
505 a more pure reflection of shifts in β -diversity (Chase *et al.*, 2011). Consequently, by removing
506 the potentially confounding effects of richness differences, the Raup-Crick index will likely
507 result in more variable estimates of similarity between communities, which would lead to
508 weaker distance-decay relationships.

509 Phylogenetic metrics, such as Unifrac, cluster communities at a lower resolution, as two
510 communities can be closely genetically related, yet distinct at fine taxonomic resolutions
511 (e.g. species or strain-level). For example, Bryant *et al.* (2008) found that Unifrac similarity

512 was approximately three times higher than the compositional similarity of the same set of
513 bacterial communities. Further, phylogenetic metrics may be inappropriate in less
514 phylogenetically diverse environments (e.g. extreme systems) where phylogenetic diversity
515 can be largely constrained to one taxon (e.g. the Halobacteria in hypersaline environments),
516 leaving few “phylogenetic degrees of freedom” left to separate communities (Fukuyama,
517 2019). However, this does not account for the observed difference between weighted and
518 unweighted versions of the Unifrac index, the former of which accounts for species’ relative
519 abundance data, whilst the latter is binary (presence/absence based). A criticism of the
520 weighted Unifrac index is that too much weight is placed on abundant taxa (Chen *et al.*,
521 2012). As abundant species are generally more widespread, placing too much weight on
522 abundant taxa would have the effect of making communities appear artificially similar,
523 exacerbating the effects of using a phylogenetic metric. As we observed no difference
524 between binary and abundance-based compositional indices, the differences observed with
525 weighted Unifrac appear to be the result of combining phylogenetic and weighted indices.
526 We therefore suggest that weighted phylogenetic metrics may underestimate microbial
527 biogeographic patterns, unless appropriate weight is given to rare and abundant taxa (Chen
528 *et al.*, 2012).

529 Our analysis of 452 microbial distance-decay relationships also revealed the overwhelming
530 preference of microbial ecologists to use classic dissimilarity indices such as the Bray-Curtis
531 ($n = 218$), Jaccard ($n = 49$), Sørensen ($n = 42$) indices. These choices no doubt reflect a
532 wider trend in ecology as a whole, however, it is pertinent to draw attention to more recently
533 developed metrics that may be more appropriate given the properties of microbial datasets
534 and the hypotheses being tested. Biotic interactions are drivers of microbial β -diversity
535 (Hanson *et al.*, 2012), yet classic dissimilarity metrics do not account for co-occurrence
536 information in communities. To this end, a new family of metrics described by Schmidt *et al.*,
537 (2017) include information on the average interactions of the taxa present, thus providing a

538 novel approach to integrating co-occurrence data into distance-decay relationships.
539 Microbiome sequencing data also have several characteristics that may be problematic in
540 the analysis of community (dis)similarities. For example, the non-biological variance of
541 sample sizes in sequence datasets can result in statistical artefacts that confound
542 biogeographic relationships (Baselga, 2007). Here, modifications made to some classic
543 indices by Chao *et al.* (2005) reduce the sensitivity of these indices to variable sample sizes
544 by accounting for unobserved species, thus reducing the need for post-sequencing
545 normalisation of sample sizes (McMurdie & Holmes, 2014). Furthermore, “fuzzy logic”-based
546 similarity indices are able to reduce the impact of false-absences or -presences on estimates
547 of β -diversity, which is beneficial for microbial ecology studies where rarefaction may induce
548 false-absences, and taxonomic assignment errors or contamination may lead to
549 false-presences. Additionally, most high-throughput sequence datasets are compositional.
550 Compositionality occurs as the arbitrary total number of sequences per sample imposed by
551 the sequencing machine causes species counts (abundances) to be dependent on each
552 other (e.g. if species A increases in abundance, species B and C will appear relatively less
553 abundant, even if their absolute abundance hasn’t changed). Binary indices should be
554 suitable as occurrences (presence/absences) are not affected by compositionality, unless
555 increases in the abundance of one or more species cause others to drop below the detection
556 limit, in which case fuzzy indices may be appropriate. Alternatively, metrics such as the
557 Aitchison distance perform well when appropriate (centered log-ratio) transformations are
558 applied to counts (Gloor *et al.*, 2017), or recently developed metrics such as the Rank Bias
559 Overlap index show promise for analysing similarity between communities based on species
560 abundance ranks (Webber *et al.*, 2010). Finally, many similarity metrics have been shown to
561 merge compositional turnover (replacement of species) and nestedness (whereby
562 communities are subsets of one another), thereby blurring the contribution of distinct
563 ecological processes to total community (dis)similarity. To combat this, modified versions of

564 classic indices such as Jaccard, Sorensen, and Bray-Curtis have been developed, allowing
565 the partitioning of community similarity metrics into their turnover and nestedness
566 components (Baselga, 2010; Podani & Schmera, 2011). We echo the call of Green and
567 Bohannan (2006) for microbial ecologists to exercise more care in their choice of
568 dissimilarity metrics, especially as many of these new metrics are implemented in popular
569 and freely accessible software, such as R (e.g. Baselga and Orme, 2012).

570 Overall, our analyses revealed that methodological factors explain more variation in
571 microbial distance-decay relationships than ecological context, but that both sets of factors
572 alter our perception of this biogeographic pattern. Given the importance of methodological
573 factors in determining the strength of microbial biogeographic patterns, it is intuitive to
574 recommend standardising approaches across studies in order to minimise the statistical
575 signals associated with methodological variance. However, our results show variance due to
576 differing ecological contexts would still hinder drawing generalisable relationships across
577 studies. Instead, we suggest that tailoring methodological choices towards specific
578 ecological contexts may enhance generalisable relationships in microbial ecology. For
579 instance, in searching for consistent relationships between ocean waters and terrestrial soils,
580 it would be unrealistic to sample both at the same spatial grain and extent, as the
581 heterogeneity in the physicochemical environment, and dispersal processes of their
582 microbial communities, are fundamentally different. Similarly, we should not necessarily
583 expect the relationships between soils and river sediments to be comparable, as
584 microorganisms in soils can feasibly disperse in any direction, whereas in rivers or streams
585 dispersal would be largely constrained by flow direction. Consequently, tailoring
586 methodological approaches, such as the sampling design and/or (geographical) distance
587 measure, to better reflect the environmental heterogeneity and dispersal dynamics between
588 contrasting ecological contexts may enable us to negotiate the hierarchy of interacting
589 factors that obscure macroecological patterns in microbial communities.

590 **Conclusions**

591 Our meta-analysis of >450 microbial distance-decay relationships revealed that factors
592 related to the eco-environmental context within which a study was conducted, as well as the
593 methodology of the study, jointly influence quantification of this classic biogeographic
594 pattern. Against expectation, factors related to molecular methodology had relatively little
595 effect on distance-decay relationships, whilst choice of dissimilarity metric was more
596 important, highlighting that even after using robust, modern molecular methods, analytical
597 choices have the power to obscure or enhance biogeographic patterns. We detected clear
598 relationships between microbial distance-decay relationships and various contextual and
599 methodological variables, yet combining these variables explained only a modest amount of
600 variation in our dataset. This lack of explanatory power indicates that microbial
601 biogeographic patterns depend on a number of contextual variables beyond those analysed
602 here. Moving forward, we suggest that microbial ecologists should place greater emphasis
603 on quantifying habitat connectivity as a means of understanding the dispersal process that
604 lead to spatial patterns such as the distance-decay relationship. Additionally, we recommend
605 that experiment designs/data collection strategies should be replicated spatially,
606 taxonomically, temporally, or any combination therein where possible (e.g. Meyer *et al.*,
607 2018; Alzarhani *et al.*, 2019; Zinger *et al.*, 2019), facilitating a more generalised
608 understanding of the variation in spatial microbial community patterns. The question of
609 whether microbial communities show spatial patterns such as distance-decay relationships
610 should be laid to rest; disentangling the web of ecological and environmental drivers that
611 shape these patterns is the next challenge in microbial biogeography.

612 **Data Availability Statement**

613 Full raw data analysed in this manuscript are provided in Table S1. Full raw data and R code
614 used in this manuscript will be uploaded to the Dryad data repository upon acceptance of
615 this article.

616 References

- 617 .
- 618 Alzarhany, A.K., Clark, D.R., Underwood, G.J.C., Ford, H., Cotton, T.E.A. & Dumbrell, A.J.
619 (2019) Are drivers of root-associated fungal community structure context specific?
620 *The ISME Journal*, **13**, 1330.
- 621 Astorga, A., Oksanen, J., Luoto, M., Soininen, J., Virtanen, R. & Muotka, T. (2012) Distance
622 decay of similarity in freshwater communities: do macro- and microorganisms follow
623 the same rules? *Global Ecology and Biogeography*, **21**, 365–375.
- 624 Bartram, A.K., Lynch, M.D.J., Stearns, J.C., Moreno-Hagelsieb, G. & Neufeld, J.D. (2011)
625 Generation of Multimillion-Sequence 16S rRNA Gene Libraries from Complex
626 Microbial Communities by Assembling Paired-End Illumina Reads. *Applied and
627 Environmental Microbiology*, **77**, 3846–3852.
- 628 Barwell, L.J., Isaac, N.J.B. & Kunin, W.E. (2015) Measuring β -diversity with species
629 abundance data. *The Journal of Animal Ecology*, **84**, 1112–1122.
- 630 Baselga, A. (2010) Partitioning the turnover and nestedness components of beta diversity.
631 *Global Ecology and Biogeography*, **19**, 134–143.
- 632 Baselga, A. & Orme, C.D.L. (2012) betapart: an R package for the study of beta diversity.
633 *Methods in Ecology and Evolution*, **3**, 808–812.
- 634 Basham, E.W., Seidl, C.M., Andriamahohatra, L.R., Oliveira, B.F. & Scheffers, B.R. (2019)
635 Distance–decay differs among vertical strata in a tropical rainforest. *Journal of
636 Animal Ecology*, **88**, 114–124.
- 637 Bissett, A., Richardson, A.E., Baker, G., Wakelin, S. & Thrall, P.H. (2010) Life history
638 determines biogeographical patterns of soil bacterial communities over multiple
639 spatial scales. *Molecular Ecology*, **19**, 4315–4327.
- 640 Bisson, I.-A., Marra, P.P., Burt, E.H., Sikaroodi, M. & Gillevet, P.M. (2007) A Molecular
641 Comparison of Plumage and Soil Bacteria Across Biogeographic, Ecological, and
642 Taxonomic Scales. *Microbial Ecology*, **54**, 65–81.
- 643 Bryant, J.A., Lamanna, C., Morlon, H., Kerkhoff, A.J., Enquist, B.J. & Green, J.L. (2008)
644 Microbes on mountainsides: Contrasting elevational patterns of bacterial and plant
645 diversity. *Proceedings of the National Academy of Sciences*, **105**, 11505–11511.
- 646 Callahan, B.J., McMurdie, P.J. & Holmes, S.P. (2017) Exact sequence variants should
647 replace operational taxonomic units in marker-gene data analysis. *The ISME Journal*,
648 **11**, 2639–2643.
- 649 Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Huntley, J., Fierer, N., Owens,
650 S.M., Betley, J., Fraser, L., Bauer, M., Gormley, N., Gilbert, J.A., Smith, G. & Knight,
651 R. (2012) Ultra-high-throughput microbial community analysis on the Illumina HiSeq
652 and MiSeq platforms. *The ISME Journal*, **6**, 1621–1624.
- 653 Carini, P., Marsden, P.J., Leff, J.W., Morgan, E.E., Strickland, M.S. & Fierer, N. (2016) Relic
654 DNA is abundant in soil and obscures estimates of soil microbial diversity. *Nature
655 Microbiology*, **2**, 1–6.
- 656 Chao, A., Chazdon, R.L., Colwell, R.K. & Shen, T.-J. (2005) A new statistical approach for
657 assessing similarity of species composition with incidence and abundance data.
658 *Ecology Letters*, **8**, 148–159.

659 Chase, J.M., Kraft, N.J.B., Smith, K.G., Vellend, M. & Inouye, B.D. (2011) Using null models
660 to disentangle variation in community dissimilarity from variation in α -diversity.
661 *Ecosphere*, **2**, art24.

662 Chen, J., Bittinger, K., Charlson, E.S., Hoffmann, C., Lewis, J., Wu, G.D., Collman, R.G.,
663 Bushman, F.D. & Li, H. (2012) Associating microbiome composition with
664 environmental covariates using generalized UniFrac distances. *Bioinformatics*, **28**,
665 2106–2113.

666 Clark, D.R., Ferguson, R.M.W., Harris, D.N., Nicholass, K.J.M., Prentice, H.J., Randall, K.C.,
667 Randell, L., Warren, S.L. & Dumbrell, A.J. (2018) Streams of data from drops of
668 water: 21st century molecular microbial ecology. *Wiley Interdisciplinary Reviews:*
669 *Water*, **5**, e1280.

670 Clark, D.R., Mathieu, M., Mourot, L., Dufossé, L., Underwood, G.J.C., Dumbrell, A.J. &
671 McGenity, T.J. (2017) Biogeography at the limits of life: Do extremophilic microbial
672 communities show biogeographical regionalization? *Global Ecology and*
673 *Biogeography*, **26**, 1435–1446.

674 van Dorst, J., Bissett, A., Palmer, A.S., Brown, M., Snape, I., Stark, J.S., Raymond, B.,
675 McKinlay, J., Ji, M., Winsley, T. & Ferrari, B.C. (2014) Community fingerprinting in a
676 sequencing world. *FEMS microbiology ecology*, **89**, 316–330.

677 Dumbrell, A.J., Nelson, M., Helgason, T., Dytham, C. & Fitter, A.H. (2010) Relative roles of
678 niche and neutral processes in structuring a soil microbial community. *The ISME*
679 *Journal*, **4**, 337–345.

680 Dushoff, J., Kain, M.P. & Bolker, B.M. (2019) I can see clearly now: Reinterpreting statistical
681 significance. *Methods in Ecology and Evolution*, **10**, 756–759.

682 Favet, J., Lapanje, A., Giongo, A., Kennedy, S., Aung, Y.-Y., Cattaneo, A., Davis-Richardson,
683 A.G., Brown, C.T., Kort, R., Brumsack, H.-J., Schnetger, B., Chappell, A., Kroijenga,
684 J., Beck, A., Schwibbert, K., Mohamed, A.H., Kirchner, T., de Quadros, P.D., Triplett,
685 E.W., Broughton, W.J. & Gorbushina, A.A. (2013) Microbial hitchhikers on
686 intercontinental dust: catching a lift in Chad. *The ISME Journal*, **7**, 850–867.

687 Franklin, R.B. & Mills, A.L. (2007) *Statistical Analysis Of Spatial Structure In Microbial*
688 *Communities. The Spatial Distribution of Microbes in the Environment* (ed. by R.B.
689 Franklin) and A.L. Mills), pp. 31–60. Springer Netherlands, Dordrecht.

690 Fukuyama, J. (2019) Emphasis on the deep or shallow parts of the tree provides a new
691 characterization of phylogenetic distances. *Genome Biology*, **20**, 131.

692 Galand, P.E., Casamayor, E.O., Kirchman, D.L. & Lovejoy, C. (2009) Ecology of the rare
693 microbial biosphere of the Arctic Ocean. *Proceedings of the National Academy of*
694 *Sciences*, **106**, 22427–22432.

695 Gaston, K.J. (2012) The importance of being rare. *Nature*, **487**, 46–47.

696 Glassman, S.I. & Martiny, J.B.H. (2018) Broad-scale Ecological Patterns Are Robust to Use
697 of Exact Sequence Variants versus Operational Taxonomic Units. *mSphere*, **3**.

698 Glenn, T.C. (2011) Field guide to next-generation DNA sequencers. *Molecular Ecology*
699 *Resources*, **11**, 759–769.

700 Gloor, G.B., Macklaim, J.M., Pawlowsky-Glahn, V. & Egozcue, J.J. (2017) Microbiome
701 Datasets Are Compositional: And This Is Not Optional. *Frontiers in Microbiology*, **8**.

702 Green, J. & Bohannan, B.J.M. (2006) Spatial scaling of microbial biodiversity. *Trends in*
703 *Ecology & Evolution*, **21**, 501–507.

704 Hanson, C.A., Fuhrman, J.A., Horner-Devine, M.C. & Martiny, J.B.H. (2012) Beyond
705 biogeographic patterns: processes shaping the microbial landscape. *Nature Reviews*
706 *Microbiology*, **10**, 497–506.

707 Harrison, F. (2012) Getting started with meta-analysis. *Journal of Applied Ecology*, 1–10.

708 Hausmann, B., Knorr, K.-H., Schreck, K., Tringe, S.G., Glavina del Rio, T., Loy, A. & Pester,
709 M. (2016) Consortia of low-abundance bacteria drive sulfate reduction-dependent

710 degradation of fermentation products in peat soil microcosms. *The ISME Journal*, **10**,
 711 2365–2375.

712 Hazard, C., Gosling, P., Gast, C.J. van der, Mitchell, D.T., Doohan, F.M. & Bending, G.D.
 713 (2013) The role of local environment and geographical distance in determining
 714 community composition of arbuscular mycorrhizal fungi at the landscape scale. *The*
 715 *ISME Journal*, **7**, 498–508.

716 Heino, J. & Soininen, J. (2010) *Are common species sufficient in describing turnover in*
 717 *aquatic metacommunities along environmental and spatial gradients*.

718 Hillebrand, H. (2004) On the Generality of the Latitudinal Diversity Gradient. *The American*
 719 *Naturalist*, **163**, 192–211.

720 Joung, Y.S., Ge, Z. & Buie, C.R. (2017) Bioaerosol generation by raindrops on soil. *Nature*
 721 *Communications*, **8**, 1–10.

722 Kemp, B.L., Tabish, E.M., Wolford, A.J., Jones, D.L., Butler, J.K. & Baxter, B.K. (2018) The
 723 Biogeography of Great Salt Lake Halophilic Archaea: Testing the Hypothesis of Avian
 724 Mechanical Carriers. *Diversity*, **10**, 124.

725 Kivlin, S.N. Global mycorrhizal fungal range sizes vary within and among mycorrhizal guilds
 726 but are not correlated with dispersal traits. *Journal of Biogeography*, **n/a**.

727 Kivlin, S.N., Winston, G.C., Goulden, M.L. & Treseder, K.K. (2014) Environmental filtering
 728 affects soil fungal community composition more than dispersal limitation at regional
 729 scales. *Fungal Ecology*, **12**, 14–25.

730 Lajeunesse, M.J. (2016) Facilitating systematic reviews, data extraction and meta-analysis
 731 with the metagear package for r. *Methods in Ecology and Evolution*, **7**, 323–330.

732 Lansdown, K., McKew, B.A., Whitby, C., Heppell, C.M., Dumbrell, A.J., Binley, A., Olde, L. &
 733 Trimmer, M. (2016) Importance and controls of anaerobic ammonium oxidation
 734 influenced by riverbed geology. *Nature Geoscience*, **9**, 357–360.

735 Lennon, J.T. & Jones, S.E. (2011) Microbial seed banks: the ecological and evolutionary
 736 implications of dormancy. *Nature Reviews. Microbiology*, **9**, 119–130.

737 Lennon, J.T., Muscarella, M.E., Placella, S.A. & Lehmkuhl, B.K. (2018) How, When, and
 738 Where Relic DNA Affects Microbial Diversity. *mBio*, **9**.

739 Li, P., Li, W., Dumbrell, A.J., Liu, M., Li, G., Wu, M., Jiang, C. & Li, Z. (2020) Spatial Variation
 740 in Soil Fungal Communities across Paddy Fields in Subtropical China. *mSystems*, **5**.

741 Lindh, M.V., Sjöstedt, J., Ekstam, B., Casini, M., Lundin, D., Hugerth, L.W., Hu, Y.O.O.,
 742 Andersson, A.F., Andersson, A., Legrand, C. & Pinhassi, J. (2017) Metapopulation
 743 theory identifies biogeographical patterns among core and satellite marine bacteria
 744 scaling from tens to thousands of kilometers. *Environmental Microbiology*, **19**,
 745 1222–1236.

746 Lindström, E.S. & Bergström, A.-K. (2004) Influence of inlet bacteria on bacterioplankton
 747 assemblage composition in lakes of different hydraulic retention time. *Limnology and*
 748 *Oceanography*, **49**, 125–136.

749 Locey, K.J., Muscarella, M.E., Larsen, M.L., Bray, S.R., Jones, S.E. & Lennon, J.T. (2020)
 750 Dormancy dampens the microbial distance–decay relationship. *Philosophical*
 751 *Transactions of the Royal Society B: Biological Sciences*, **375**, 20190243.

752 Low-Décarie, E., Fussmann, G.F., Dumbrell, A.J. & Bell, G. (2016) Communities that thrive
 753 in extreme conditions captured from a freshwater lake. *Biology Letters*, **12**,
 754 20160562.

755 Maček, I., Clark, D.R., Šibanc, N., Moser, G., Vodnik, D., Müller, C. & Dumbrell, A.J. (2019)
 756 Impacts of long-term elevated atmospheric CO₂ concentrations on communities of
 757 arbuscular mycorrhizal fungi. *Molecular Ecology*, **28**, 3445–3458.

758 Mantel, N. (1967) The Detection of Disease Clustering and a Generalized Regression
 759 Approach. *Cancer Research*, **27**, 209–220.

760 Martiny, J.B.H., Eisen, J.A., Penn, K., Allison, S.D. & Horner-Devine, M.C. (2011) Drivers of

761 bacterial β -diversity depend on spatial scale. *Proceedings of the National Academy*
 762 *of Sciences*, **108**, 7850–7854.

763 McMurdie, P.J. & Holmes, S. (2014) Waste Not, Want Not: Why Rarefying Microbiome Data
 764 Is Inadmissible. *PLOS Computational Biology*, **10**, e1003531.

765 Meyer, K.M., Memiaghe, H., Korte, L., Kenfack, D., Alonso, A. & Bohannan, B.J.M. (2018)
 766 Why do microbes exhibit weak biogeographic patterns? *The ISME Journal*, **12**,
 767 1404–1413.

768 Muyzer, G. (1999) DGGE/TGGE a method for identifying genes from natural ecosystems.
 769 *Current Opinion in Microbiology*, **2**, 317–322.

770 Nekola, J.C. & White, P.S. (1999) The distance decay of similarity in biogeography and
 771 ecology. *Journal of Biogeography*, **26**, 867–878.

772 Norros, V., Rannik, Ü., Hussein, T., Petäjä, T., Vesala, T. & Ovaskainen, O. (2014) Do small
 773 spores disperse further than large spores? *Ecology*, **95**, 1612–1621.

774 Podani, J. & Schmera, D. (2011) A new conceptual and methodological framework for
 775 exploring and explaining pattern in presence – absence data. *Oikos*, **120**,
 776 1625–1638.

777 R Core Team (2019) *R: A Language and Environment for Statistical Computing*, R
 778 Foundation for Statistical Computing, Vienna, Austria.

779 Ramette, A. (2007) Multivariate analyses in microbial ecology. *FEMS Microbiology Ecology*,
 780 **62**, 142–160.

781 Reche, I., Pulido-Villena, E., Morales-Baquero, R. & Casamayor, E.O. (2005) Does
 782 Ecosystem Size Determine Aquatic Bacterial Richness? *Ecology*, **86**, 1715–1722.

783 Roesch, L.F.W., Fulthorpe, R.R., Riva, A., Casella, G., Hadwin, A.K.M., Kent, A.D., Daroub,
 784 S.H., Camargo, F.A.O., Farmerie, W.G. & Triplett, E.W. (2007) Pyrosequencing
 785 enumerates and contrasts soil microbial diversity. *The ISME journal*, **1**, 283–290.

786 Rosenberg, M.S., Rothstein, H.R. & Gurevitch, J. (2013) Effect sizes: Conventional choices
 787 and calculations. *Handbook of Meta-analysis in Ecology and Evolution*, 61–71.

788 Shade, A., Jones, S.E., Caporaso, J.G., Handelsman, J., Knight, R., Fierer, N. & Gilbert, J.A.
 789 (2014) Conditionally Rare Taxa Disproportionately Contribute to Temporal Changes
 790 in Microbial Diversity. *mBio*, **5**.

791 Shade, A. & Stopnisek, N. (2019) Abundance-occupancy distributions to prioritize plant core
 792 microbiome membership. *Current Opinion in Microbiology*, **49**, 50–58.

793 Shmida, A. & Wilson, M.V. (1985) Biological Determinants of Species Diversity. *Journal of*
 794 *Biogeography*, **12**, 1–20.

795 Shoemaker, W.R., Locey, K.J. & Lennon, J.T. (2017) A macroecological theory of microbial
 796 biodiversity. *Nature Ecology & Evolution*, **1**, 1–6.

797 Soininen, J., Korhonen, J.J., Karhu, J. & Vetterli, A. (2011) Disentangling the spatial patterns
 798 in community composition of prokaryotic and eukaryotic lake plankton. *Limnology*
 799 *and Oceanography*, **56**, 508–520.

800 Soininen, J., McDonald, R. & Hillebrand, H. (2007) The distance decay of similarity in
 801 ecological communities. *Ecography*, **30**, 3–12.

802 Sorte, F.A.L., McKinney, M.L., Pyšek, P., Klotz, S., Rapson, G.L., Celesti-Grapow, L. &
 803 Thompson, K. (2008) Distance decay of similarity among European urban floras: the
 804 impact of anthropogenic activities on β diversity. *Global Ecology and Biogeography*,
 805 **17**, 363–371.

806 Steinbauer, M.J., Dolos, K., Reineking, B. & Beierkuhnlein, C. (2012) Current measures for
 807 distance decay in similarity of species composition are influenced by study extent
 808 and grain size. *Global Ecology and Biogeography*, **21**, 1203–1212.

809 Vašutová, M., Mleczko, P., López-García, A., Maček, I., Boros, G., Ševčík, J., Fujii, S.,
 810 Hackenberger, D., Tuf, I.H., Hornung, E., Páll-Gergely, B. & Kjoller, R. (2019) Taxi
 811 drivers: the role of animals in transporting mycorrhizal fungi. *Mycorrhiza*, **29**,

812 413–434.

813 Vos, M., Wolf, A.B., Jennings, S.J. & Kowalchuk, G.A. (2013) Micro-scale determinants of
814 bacterial diversity in soil. *FEMS Microbiology Reviews*, **37**, 936–954.

815 Warmink, J.A., Nazir, R., Corten, B. & van Elsas, J.D. (2011) Hitchhikers on the fungal
816 highway: The helper effect for bacterial migration via fungal hyphae. *Soil Biology and*
817 *Biochemistry*, **43**, 760–765.

818 Webber, W., Moffat, A. & Zobel, J. (2010) A similarity measure for indefinite rankings. *ACM*
819 *Transactions on Information Systems*, **28**, 1–38.

820 Wilkinson, D.M. (2001) What is the upper size limit for cosmopolitan distribution in free-living
821 microorganisms? *Journal of Biogeography*, **28**, 285–291.

822 Wilkinson, D.M., Koumoutsaris, S., Mitchell, E.A.D. & Bey, I. (2012) Modelling the effect of
823 size on the aerial dispersal of microorganisms. *Journal of Biogeography*, **39**, 89–97.

824 Zinger, L., Taberlet, P., Schimann, H., Bonin, A., Boyer, F., Barba, M.D., Gaucher, P., Gielly,
825 L., Giguët-Covex, C., Iribar, A., Réjou-Méchain, M., Rayé, G., Rioux, D., Schilling, V.,
826 Tymen, B., Viers, J., Zouiten, C., Thuiller, W., Coissac, E. & Chave, J. (2019) Body
827 size determines soil community assembly in a tropical forest. *Molecular Ecology*, **28**,
828 528–543.