Title: What drives study-dependent differences in distance-decay relationships of microbial communities?

Running title: Meta-Analysis of Microbial Distance-Decay Relationships

Keywords: Bacteria, Archaea, Eukarya, Mantel test, macroecology, biogeography, dispersal limitation, community dissimilarity

Abstract

Aim: Ecological communities that exist closer together in space are generally more compositionally similar than those far apart, as defined by the distance-decay of similarity relationship. However, recent research has revealed substantial variability in the distance-decay relationships of microbial communities between studies of different taxonomic groups, ecosystems, spatial scales, as well as between those using different molecular methodologies (e.g. high-throughput sequencing versus molecular fingerprinting). Here, we test how these factors influence the strength of microbial distance-decay relationships, to draw generalisations about how microbial β-diversity scales with space.

Location: Global.

Time period: Studies published between 2005-2019 (inclusive).

Major taxa studied: Bacteria, Archaea, and microbial Eukarya.

Methods: We conducted a meta-analysis of microbial distance-decay relationships, using the Mantel correlation coefficient as a measure of the strength of distance-decay relationships. Our final dataset consisted of 452 data points, varying in environmental/ecological context or methodological approaches, and used linear models to test the effects of each variable.

Results: Both ecological and methodological factors had significant impacts on the strength of microbial distance-decay relationships. Specifically, the strength of these relationships varied between environments and habitats, with soils showing significantly weaker distance-decay relationships than other habitats, whilst increasing spatial extents had no effect. Methodological factors such as sequencing depth were positively related to the strength of distance-decay relationships, and choice of dissimilarity metric was also

important, with phylogenetic metrics generally giving weaker distance-decay relationships than binary or abundance-based indices.

Main conclusions: We conclude that widely studied microbial biogeographic patterns, such as the distance-decay relationship, vary by ecological context but are primarily distorted by methodological choices. Consequently, we suggest that by linking methodological approaches appropriately to the ecological context of a study, we can progress towards generalisable biogeographic relationships in microbial ecology.

Introduction

The distance-decay of community similarity is one of the most widely studied relationships in macroecology (Nekola & White, 1999; Soininen *et al.*, 2007). This relationship quantifies the decrease in compositional similarity (β-diversity) between communities with increasing geographic distance separating them, and demonstrates that nearby communities are more similar to each other than distantly-separated communities. Distance-decay relationships arise through several different, but often interacting ecological and evolutionary processes, and consequently ecologists have extensively debated the underlying mechanisms that generate such patterns (Nekola & White, 1999; Soininen *et al.*, 2007; Hanson *et al.*, 2012). Spatial structuring of the environment can lead to distance-decay relationships, as communities close together in space are likely to experience more similar environmental conditions, and thus contain more similar communities than those situated in different environmental conditions. Dispersal limitation can also lead to distance-decay relationships by limiting the connectivity between communities, meaning that communities closer together in space will share more species through localised dispersal than those further apart.

Distance-decay relationships are well documented in a multitude of plant and animal communities (e.g. multiple taxa - Soininen *et al.*, 2007; urban plants - Sorte *et al.*, 2008; multiple aquatic taxa - Astorga *et al.*, 2012; tropical amphibians - Basham *et al.*, 2019). Yet, these relationships are of particular interest to microbial ecologists as microorganisms were assumed to have ubiquitous distributions for several reasons. Firstly, their small size facilitates passive dispersal over large geographic distances by vectors such as wind, bio-aerosolization, ocean currents or migrating animals (Bisson *et al.*, 2007; Favet *et al.*, 2013; Joung *et al.*, 2017; Vašutová *et al.*, 2019), thus potentially overcoming dispersal limitation as a contributing factor to microbial community composition. Additionally, microorganisms often maintain high population densities in the environment leading to

dispersal by “mass effects”, whereby high dispersal rates from areas of increased population density maintain populations in less optimal environments (Shmida & Wilson, 1985), helping them to overcome the constraints of spatially-structured environmental gradients. Finally, some microorganisms are able to enter dormant states, whether as vegetative cells or as cysts or spores (Locey *et al.*, 2020), allowing them to survive and disperse through suboptimal environments, simultaneously enhancing their dispersive abilities, and reducing the influence of spatially-structured environmental gradients (Low-Décarie *et al.*, 2016). Combined, these traits theoretically lower microbial β-diversity by increasing the proportion of shared species between distant communities, in turn leading to weaker distance-decay relationships compared to macroorganisms. However, empirical studies have yielded mixed results on the strength of microbial distance-decay relationships, where strength is defined as the degree to which geographic distance and community dissimilarity are correlated. Many studies have detected little or no evidence of distance-decay relationships in microbial communities (Hazard *et al.*, 2013; Kivlin *et al.*, 2014), whilst others report relationships of varying strengths, across a range of spatial extents, study systems, and taxa (Dumbrell *et al.*, 2010; Martiny *et al.*, 2011; Clark *et al.*, 2017). Thus, despite hundreds of empirical studies, the generality of spatial patterns in microbial communities remains unclear, and we are no closer to understanding whether variability in the spatial scaling relationships of microbial β-diversity originates from ecological or methodological sources.

Variation in microbial distance-decay relationships could be due to different environmental or ecological contexts in studies. Here, we consider environmental context as the variability in the physico-chemical environment (e.g. temperature, pH, topology), and ecological context as the total suite of species present and their interactions. The study systems commonly of interest to microbial ecologists vary in terms of connectivity, which may facilitate or hinder dispersal between communities, thus leading to weaker or stronger distance-decay relationships, respectively. In well connected systems where dispersal is more feasible, such

as oceanic waters, distance-decay relationships should be weaker than systems in which dispersal is limited, such as host-associated systems or soil systems, where distance-decay relationships are weaker in deeper soil horizons (Li *et al.*, 2020). Moreover, study systems differ in the spatially structured environmental gradients and heterogeneity they support. Sediments and soils for example, can support strong environmental gradients over distances of a few meters (Dumbrell *et al.*, 2010), and can be highly heterogeneous at the millimeter scale (Vos *et al.*, 2013), strengthening the correlation between distance and community dissimilarity. Additionally, different study taxa are likely to yield variable distance-decay relationships because they differ in traits that are linked to dispersal efficacy. For example, small cells disperse more efficiently over long distances (Wilkinson, 2001; Wilkinson *et al.*, 2012; Norros *et al.*, 2014), thus organisms with larger cell sizes, such as microbial Eukarya, should be more strongly dispersal limited than those with small cell sizes, such as Bacteria (although this may not be true for all taxa e.g. see Kivlin, 2020). Finally, it is known that spatial extent can influence our perception of ecological relationships, which may contribute to variable distance-decay relationships (Steinbauer *et al.*, 2012). Studies incorporating larger spatial extents would be expected to show exponential decay of similarity, as communities are more likely to originate from distinct species pools, with high dispersal limitation. In contrast, studies with smaller spatial extents are generally expected to follow power-law decay, although the spatial scales at which the distance-decay relationship follows either of these forms may also depend on the size of the study organisms (Martiny *et al.*, 2011; Nekola & McGill, 2014; Luan *et al.*, 2020).

Whilst the context in which a study was undertaken may contribute to variability in microbial distance-decay relationships, so too could different methodologies. Technological advances have yielded new insight into the structure and functioning of development of environmental microbial communities (Clark *et al.*, 2018). However, rapid turnover in molecular methodologies means that our perception of microbial β-diversity patterns integrates

methods that vary substantially in both coverage (ability to detect a greater proportion of the community in a given sample) and resolution (ability to resolve closely related taxa) (Muyzer, 1999; Glenn, 2011). Early methods such as clone library sequencing and community fingerprinting methods (e.g. denaturing gradient gel electrophoresis (DGGE), terminal restriction fragment length polymorphism (TRFLP), or phospholipid fatty acid (PLFA) analysis) are limited in their ability to detect rare taxa (Bartram *et al.*, 2011), and often miss them completely (Low-Décarie *et al.*, 2016). In turn, this could reduce the detected β-diversity, inflating estimated community similarity and weakening distance-decay relationships (Hanson *et al.*, 2012). In contrast, high-throughput sequencing (HTS) platforms (also frequently referred to as next-generation sequencing (NGS)) can deliver sequencing depths of tens or even hundreds of thousands of sequences per sample (Caporaso *et al.*, 2012), thus improving both community coverage (the detected proportion of a given community), and allowing more samples to be examined in a single study (sample coverage). Consequently, variation in the ability of molecular methods to resolve closely related taxa, and to detect rare taxa can be an additional source of variability in microbial β-diversity, which by extension can either weaken or strengthen microbial distance-decay relationships.

In addition to the molecular methods, the choice of analytical methods, such as similarity metric, can influence distance-decay relationships. The similarity of communities varies according to the identity and abundance of the species present, their phylogenetic relationships, and by external factors such as varying sample sizes. Thus, similarity metrics that vary by one or more of these characteristics would likely result in contrasting distance-decay relationships (Chao *et al.*, 2005; Barwell *et al.*, 2015). For example, phylogenetic indices would be expected to yield weaker distance-decay relationships than other metrics, because communities that have no species in common can still be highly phylogenetically similar if the species share many branches of a phylogenetic tree, thus

reducing the decay of similarity over geographic distance (Bryant *et al.*, 2008). On the other hand, quantitative indices compare not only the composition of species present, but also their abundance in each community, reflecting finer-scale changes in community structure, and thus should result in stronger distance-decay relationships by providing an additional axis (species abundances) by which communities can differ.

Here, to disentangle the effects of both contextual (e.g. spatial extent, taxon, or ecosystem) and methodological (e.g. means of identifying/differentiating taxa, or similarity metric) variables on microbial distance-decay relationships, we undertook a meta-analysis to test the following specific hypotheses:

- H_1 Bacteria and Archaea will show weaker (lower correlation between geographic distance and community dissimilarity) distance-decay relationships than micro-eukaryotic taxa due to their smaller size and higher population densities in most environments.
- H_2 Environments that are able to maintain steep physicochemical gradients, such as sediments and soils, will have stronger (higher correlation between geographic distance and community dissimilarity) distance-decay relationships than those such as seawater or air, where environmental gradients are more diffuse.
- H_3 Spatial extent will be positively related to the strength of the distance-decay relationship as, at large spatial scales, increased dispersal limitation and environmental heterogeneity will decrease the variance in community similarity at a given spatial distance, resulting in stronger distance-decay relationships.
- H_4 High-throughput sequencing methods will yield stronger distance-decay relationships due to: a) their ability to resolve closely related taxa, b) their greater community coverage (e.g. number of sequences per sample, or number of individuals counted per sample), and/or c) their greater sample coverage.

- H_5 Phylogenetic similarity metrics (e.g. Unifrac, beta nearest taxon index) will result in weaker distance-decay relationships than other metrics as communities can be phylogenetically similar, yet different at fine taxonomic resolutions, whilst quantitative metrics (e.g. Bray-Curtis, Hellinger, Euclidean) will yield the strongest as they reflect changes in both species composition and abundance.

Methods

Meta-Analysis

In order to test our hypotheses, we first gathered available data on microbial distance-decay relationships via a systematic literature search. To do this, five search terms were selected to detect relevant studies (Table 1). All literature searches were conducted using the Web of Science search portal on 18/04/2020, and all results published between 1900-2019 (inclusive) were retained. To further filter the dataset to studies suitable for testing our hypotheses, search results were downloaded and manually screened using the "metagear" (Lajeunesse, 2016) package in R (version 3.4.1; R Core Team, 2019). Here, suitable studies were those that tested the relationship between community similarity and geographic distance in microbial communities, and not studies of "macroorganisms", or studies of strain-level genetic distance (e.g. using multi-locus sequence typing). Furthermore, studies that did not test distance-decay relationships using Mantel correlation, or that used only partial Mantel tests, were also discarded. We did not identify any potentially suitable studies that were published prior to 1967, the year the Mantel test was described (Mantel, 1967), and the earliest suitable study was published in 2005.

Table 1. Details of Web of Science search terms, and the number of results for each search.

Search	Search Term	Number of results
1	TS = (biogeograph*) AND TS = (bacteria* OR archaea* OR	2907

	microb* OR microorganism*)	
2	TS = (macroecolog*) AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	136
3	TS = ("everything is everywhere") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	66
4	TS = ("geographic distance") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	220
5	TS = ("distance decay") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	186

From these studies, we extracted Mantel correlation coefficients (*r*) as an effect-size measure for each distance-decay relationship, which we refer to throughout as distance-decay strength. The Mantel test is a permutation-based method used to test for correlation between two distance matrices, or in the context of this study, community (dis)similarity and geographic distance. The Mantel test statistic is an ideal measure of effect size for use in meta-analytical frameworks for several reasons. Firstly, the Mantel correlation test is the most frequently used method for testing distance-decay relationships in microbial ecology (Franklin & Mills, 2007; Ramette, 2007). Secondly, as the Mantel coefficient is a standardised correlation coefficient (i.e. is bound by -1 and 1), it provides an easily interpretable and comparable measure of effect size (Harrison, 2012).

We ensured all Mantel correlation coefficients reflected correlations between geographic distance and community dissimilarity, rather than similarity, by multiplying correlation coefficients by -1 where necessary (so that positive values indicate a typical distance-decay relationship). Partial Mantel statistics (which test for correlation between two matrices whilst controlling for a third) were excluded as they are influenced by other variables included in the test, and are therefore not easily comparable between studies. All Mantel correlation coefficients were transformed to *z*-scores using Fisher's *z* transformation, as recommended by Rosenberg *et al.* (2013). All subsequent statistical analyses were conducted on the

204 transformed *z*-scores, whilst original Mantel correlation coefficients were used to make
205 figures, for ease of interpretation.

206 In order to test our hypotheses, several variables relating to the context and methodology of
207 each distance-decay relationship were recorded. Details of these variables are described in
208 Box 1.

209 Box 1. Details of the explanatory variables extracted from each study.

Resolution
Each distance-decay relationship was categorised into either high-resolution (high-throughput or Sanger sequencing), low resolution (molecular e.g. ARISA, TRFLP, DGGE, PhyloChip, PLFA), or low resolution (morphological), based on the method's ability to distinguish between closely related organisms.

Community Coverage
This refers to the sequencing depth in sequencing-based studies, or number of individuals counted in morphology-based studies, per sample. For sequencing studies, we recorded the number of sequences after rarefaction, or if this was not given, the average number of sequences per sample. As there is no comparable measure of coverage for fingerprinting studies, we excluded them from analyses of community coverage.

Sample Coverage
Sample coverage refers to the sample size (e.g. number of communities/samples) of each distance-decay relationship.

Dissimilarity Index
The dissimilarity index used to calculate each distance-decay relationship. Recorded dissimilarity indices were then categorised as quantitative (Bray-Curtis, Horn-Morisita, Euclidean, Hellinger, Theta), qualitative (Jaccard, Raup-Crick, Sørensen, Simpson, βsim), or phylogenetic (weighted or unweighted Unifrac, Rao, β-mean nearest taxon distance, β-mean pairwise distance).

Correlation Type
Studies were categorised according to the type of correlation coefficient used in the analysis distance-decay relationship (e.g. Spearman's or Pearson's correlation coefficient). The correlation type was only recorded if the type of correlation coefficient was explicitly mentioned.

Study Taxon
Each distance-decay relationship was binned into the following broad taxonomic categories based on the taxonomy of the focal organisms (Archaea, Bacteria, Fungi, or other microbial Eukarya), or combination of these categories if a relationship was based on multiple taxa (for example due to using sequencing primers that detect both Archaea and Bacteria). Fungi grouped separately from other micro-Eukaryotes due to their distinct reproductive strategy (e.g. spore-production) and the fact the they are frequently targeted

using distinct molecular approaches (e.g. via taxon-specific primer sets), in contrast to most other studies of micro-Eukarya.

Spatial Extent
This is the maximum distance separating communities in km. If this was not stated in text or provided in supplementary material (e.g. in a geographic distance matrix), it was calculated from given geographic coordinates, estimated from a plot of the distance-decay relationship, or estimated from scaled maps.

Environment
We broadly categorised distance-decay relationships based on the type of environment (agriculture, air, aquifer, coastal wetlands/intertidal, desert, dune, forest, glacier, grassland, lake, marine, coastal marshes, mine, river, snow, urban) within which they were sampled. Whilst these categories are not mutually exclusive, we categorised each study based on which environment best represented the environmental context in which each study was undertaken. For studies on lakes, we also recorded whether relationships originated from a single lake, or across multiple lakes.

Habitat
The type of environmental material that the sampled communities occupied. We categorised distance-decay relationships as: air, host-associated, sediment, snow, soil, water.

Statistical Analyses

In order to determine whether distance-decay relationships varied between categorical variables (as in hypotheses 1, 2, 4, and 5), we used ANOVA tests. In tests where significant differences between groups were found, Tukey's Honest Significant Difference (HSD) tests were used to determine which groups were different. Linear mixed-effect models were used to separately test for relationships between the strength (correlation between geographic distance and community dissimilarity, expressed as the Mantel correlation coefficient) of distance-decay relationships and single continuous variables such as spatial extent and community coverage, using a random intercept to account for heteroscedasticity due to some studies contributing multiple relationships in each model. P-values and R^2 values were calculated for each term in these models using the approach described by Nakagawa & Schielzeth (2013). The variables spatial extent and community coverage were initially $\log_{10}$ transformed to aid model fitting, as they spanned several orders of magnitude. To compare

the overall influence of ecological vs methodological factors on microbial distance-decay relationships, we compared two full models (including all relevant variables) using AIC scores, on a subset of the data for which all variables were successfully recorded. We report the results of all null hypothesis tests in terms of statistical “clarity” rather than “significance”, in line with recommendations from Dushoff *et al.* (2019).

Results

Our Web of Science searches resulted in 2,982 unique search results. Manual screening of the abstracts yielded 951 studies that were deemed to be potentially suitable for use in this analysis. A total of 452 Mantel correlation coefficients were successfully obtained from 187 studies represented in 61 journals (Fig. S1). Reported Mantel correlation coefficients ranged from -0.33 to 0.95, with a mean of 0.27 (std. error = 0.011), whilst a summary of the variables collected is shown in Table 2.

Table 2. Summary of collected data. For categorical variables, the number of individual distance-decay relationships in each category are shown, whereas minima, maxima, median and mean values are shown for continuous variables. Detailed descriptions of each variable are found in Box 1, and raw data can be found in Table S1.

Ecological variables		Methodological variables	
Variable	Summary	Variable	Summary
[a] Study taxon	Archaea: *n* = 26 Bacteria: *n* = 238 Eukarya: *n* = 67 Fungi: *n* = 93 Archaea + Bacteria: *n* = 17 Bacteria + Eukarya: *n* = 3 Bacteria + Fungi: *n* = 6 All: *n* = 2	Resolution	High: *n* = 345 Intermediate: *n* = 84 Low: *n* = 23
Spatial extent (km)	Min = 0.0001 Mean = 1,543 Median = 220 Max = 18,700	Community coverage (number of individuals/	Min = 8 Mean = 217,357 Median = 1,257 Max = 34,192,561

	NA = 15	sequences)	NA = 115
Environment type	Agriculture: *n* = 16 Air: *n* = 13 Aquifer: *n* = 1 Coastal: *n* = 8 Desert: *n* = 4 Dune: *n* = 1 Forest: *n* = 76 Glacier: *n* = 5 Grassland: *n* = 96 Lake: *n* = 76 Marine: *n* = 88 Marsh: *n* = 3 Mine: *n* = 1 River: *n* = 57 Snow: *n* = 3 Urban: *n* = 4	Dissimilarity index	β-MNTD: *n* = 13 β-MPD: *n* = 1 β-sim: *n* = 4 Bray-Curtis: *n* = 218 Bray-Curtis$_{Sim}$: *n* = 3 Bray-Curtis$_{Nes}$: *n* = 1 Canberra: *n* = 1 Euclidean: *n* = 9 Hellinger: *n* = 4 Jaccard: *n* = 49 Mash: *n* = 2 Morisita-Horn: *n* = 4 Rao: *n* = 2 Raup-Crick: *n* = 19 Simpson: *n* = 2 Sorensen: *n* = 42 Theta: *n* = 1 Unweighted Unifrac: *n* = 17 Weighted Unifrac: *n* = 59 NA: *n* = 1
Habitat type	Air: *n* = 16 Host: *n* = 75 Sediment: *n* = 78 Snow: *n* = 3 Soil: *n* = 141 Water: *n* = 137 NA: *n* = 2	Correlation type	Pearson: *n* = 62 Spearman: *n* = 86 NA: *n* = 304
		Sample coverage (Number of samples)	Min = 4 Mean = 52.88 Median = 25 Max = 1,010 NA = 1

[a] The “All” category consists of studies that incorporated all microbial taxonomic groups, whilst combined categories (e.g. Archaea + Bacteria) incorporate communities from multiple taxonomic groups (e.g. bacterial and archaeal communities).

Influence of Context on the Distance-Decay Relationship

In order to determine whether contextual factors can influence the strength of distance-decay relationships, the influence of ecological factors including study taxa, study

system, and spatial scale were tested. Within the dataset, the most commonly studied taxa were Bacteria (n = 238), followed by Fungi (n = 93), other microbial Eukaryotes (n = 67), and Archaea (n = 26). We found no clear differences in the strength of distance-decay relationships between these taxa ($F_{5,\ 441}$ = 0.99, P = 0.43), although distance-decay relationships incorporating bacterial and fungal communities showed the weakest relationships, albeit only from six studies (Fig. 1).

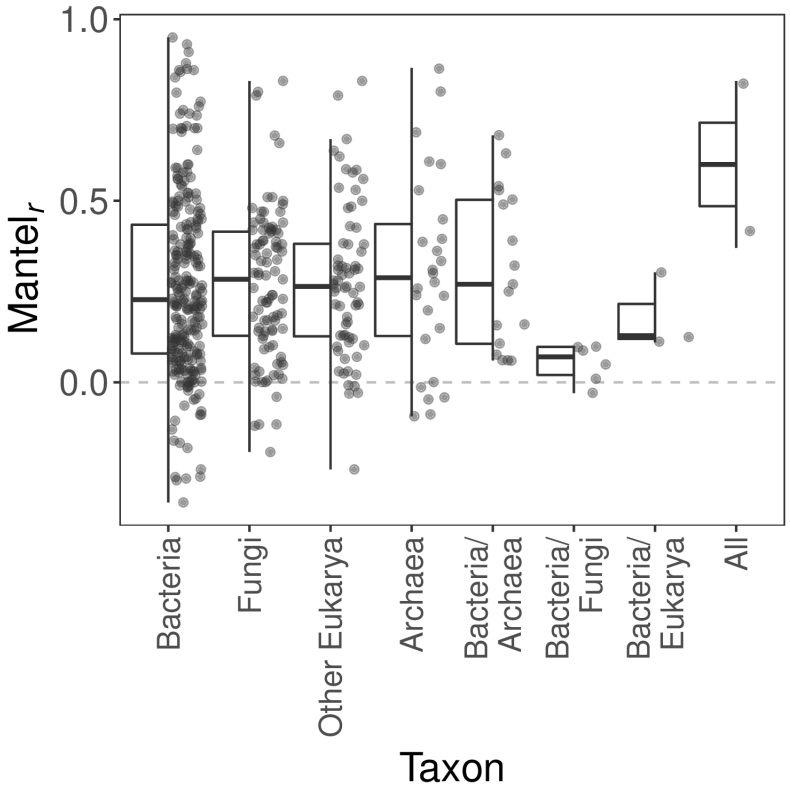
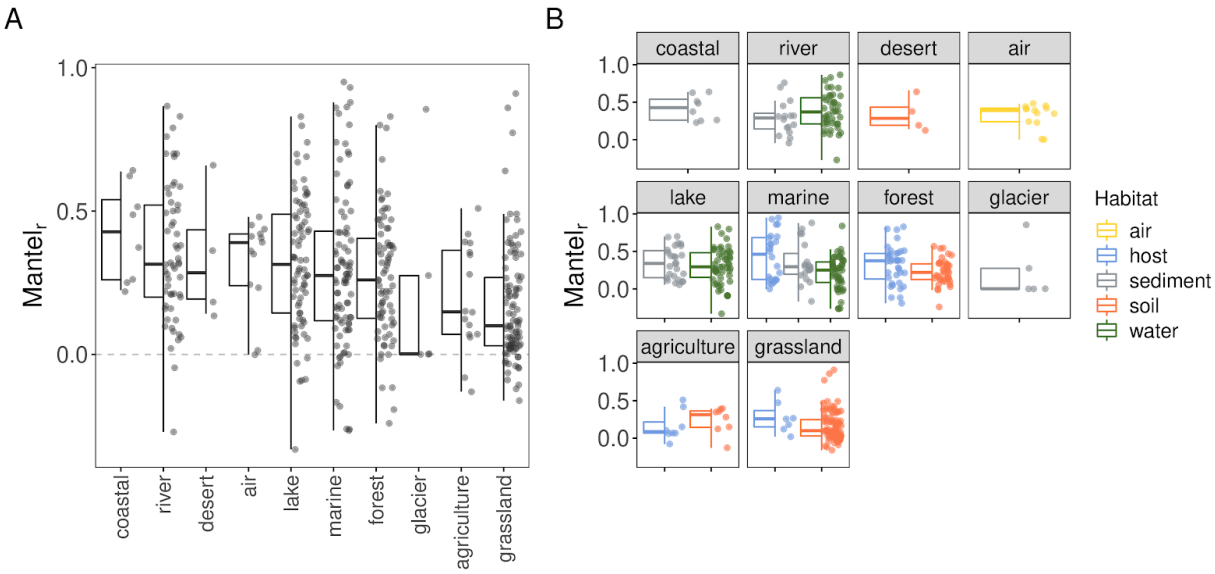


Figure 1. The strength (Mantel$_r$) of distance-decay relationships based on different study taxa. A larger Mantel$_r$ value indicates a stronger distance-decay relationship. The "All" category consists of studies that incorporated all microbial taxonomic groups, whilst combined categories (e.g. Bacteria/Archaea) incorporate communities from multiple taxonomic groups (e.g. bacterial and archaeal communities).

The distance-decay relationships in our dataset originated from 16 different environments. Of these, five were represented by three, or fewer, distance-decay relationships, and so

were excluded from further analyses (marsh; n = 3, snow; n = 3, dune, mine, aquifer; n = 1). The most frequently studied environments were grasslands (n = 96), marine (n = 88), and lakes and forests (n = 76 for both). We found clear differences in the strength of distance-decay relationships between environments (Fig. 2A; $F_{10, 432}$ = 3.187, P < 0.001). Specifically, and perhaps counter-intuitively, grassland-based studies had weaker distance-decay relationships than those from aquatic environments such as lakes, rivers, or the marine environment (|coef| > 0.17, P < 0.05 for all comparisons). Urban environments, which included built environments such as sewers and indoor air, also produced weak distance-decay relationships, although with only four data points, this difference was not statistically clear (P > 0.43 for all comparisons). We also found no difference in the strength of distance-decay relationships between studies conducted in single lakes compared to those incorporating multiple lakes ($F_{1, 74}$ = 0.11, P = 0.74), despite the average spatial extent of multiple-lake studies being approximately 32-fold greater than that of single-lake studies (Fig. S2).

A more detailed analysis of the interaction between environment type and habitat revealed that, whilst environments ($F_{9, 420}$ = 3.29, P < 0.001) and habitat ($F_{3, 420}$ = 6.65, P < 0.001) differ from each other, their interaction was not statistically significant ($F_{4, 420}$ = 1.93, P = 0.10). In fact, within environments, only marine host-associated and marine water-based distance-decay relationships were clearly different from each other (Fig. 2B), with host-associated communities showing significantly stronger distance-decay relationships (coef = 0.35, P < 0.001).



276 Figure 2. Variation in Mantel correlation coefficients of distance-decay relationships between
277 different environments (A) and habitat types (B). Environment categories are arranged from
278 strongest to weakest mean distance-decay relationship.

279 The spatial extents of recorded distance-decay relationships ranged from 10 cm to more
280 than 18,000 km, and minimal spatial extents varied notably across environments and
281 habitats, with terrestrial and soil-based studies often conducted over smaller spatial scales
282 (Fig. S3). After accounting for differences between studies, we found no evidence of a
283 statistically clear relationship between the spatial extent of a study and the strength of the
284 observed distance-decay relationship (coef = 0.02, marginal R^2 = 0.020, t = 1.58, P = 0.11).
285 Finally, as larger spatial scale studies might also incorporate greater sampling coverage, we
286 tested for collinearity between the spatial scale of a study and the sampling coverage, but
287 found no correlation between these variables (ρ = 0.06, P = 0.19).

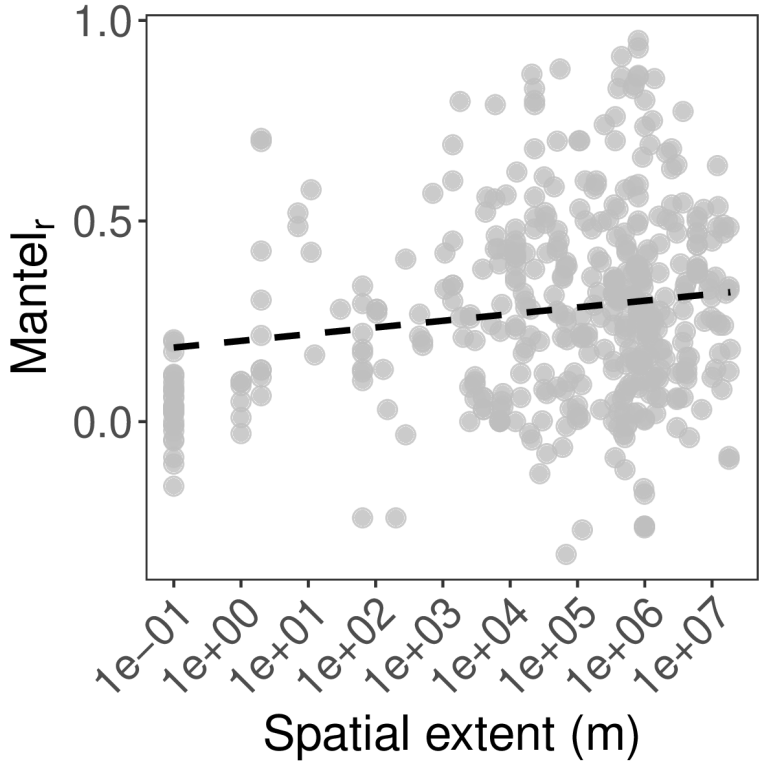


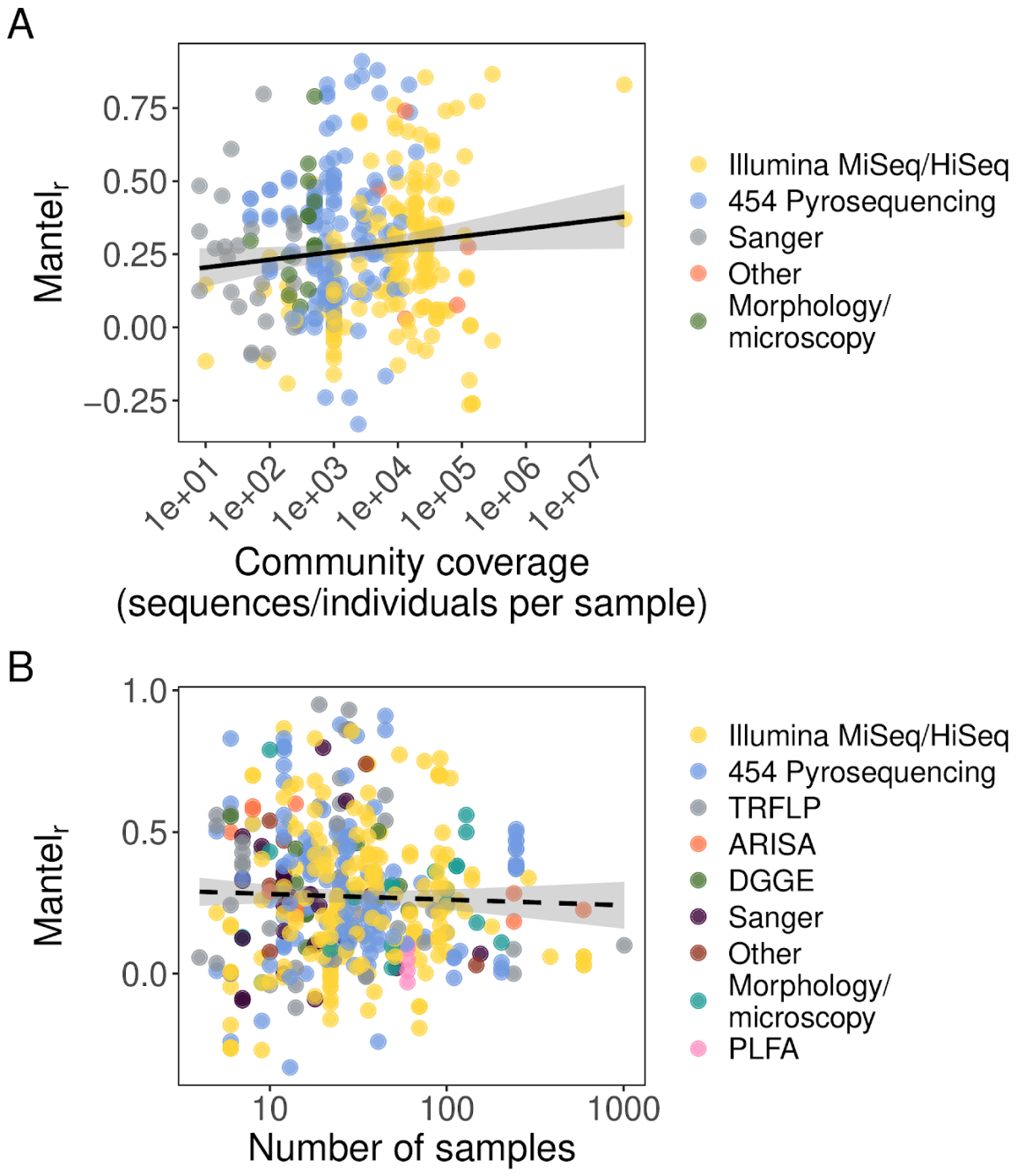
Figure 3. The relationship between spatial extent and the Mantel correlation coefficient of microbial distance-decay relationships. The dashed line represents the fit of a mixed-effects model between the $\log_{10}$ of spatial extent and Mantel correlation coefficient, with a study-dependent random intercept.

Influence of Methodological Factors on the Distance-Decay Relationship

We grouped community characterisation methods according to their ability to distinguish between closely related taxa. There were no clear differences in the strength of distance-decay relationships between different resolution methods ($F_{2, 449} = 0.562$, $P = 0.57$), nor were there clear differences between different molecular methods (Fig. S4, $F_{7, 437} = 1.97$, $P = 0.06$), considering only those methods that had >4 distance-decay relationships across the entire dataset (excluding Ion Torrent; n = 4, phylo-chip; n = 2, and Pac-Bio; n = 1).

Whilst we observed no differences in distance-decay relationships between different resolution methods, after accounting for study-dependent differences, we found a positive relationship between ($\log_{10}$) community coverage and the strength of microbial distance-decay relationships (Fig. 4A; n = 337, conditional R^2 = 0.57, coef = 0.06, t = 2.73, P < 0.01), although the marginal effect of community coverage was weak (marginal R^2 = 0.04).

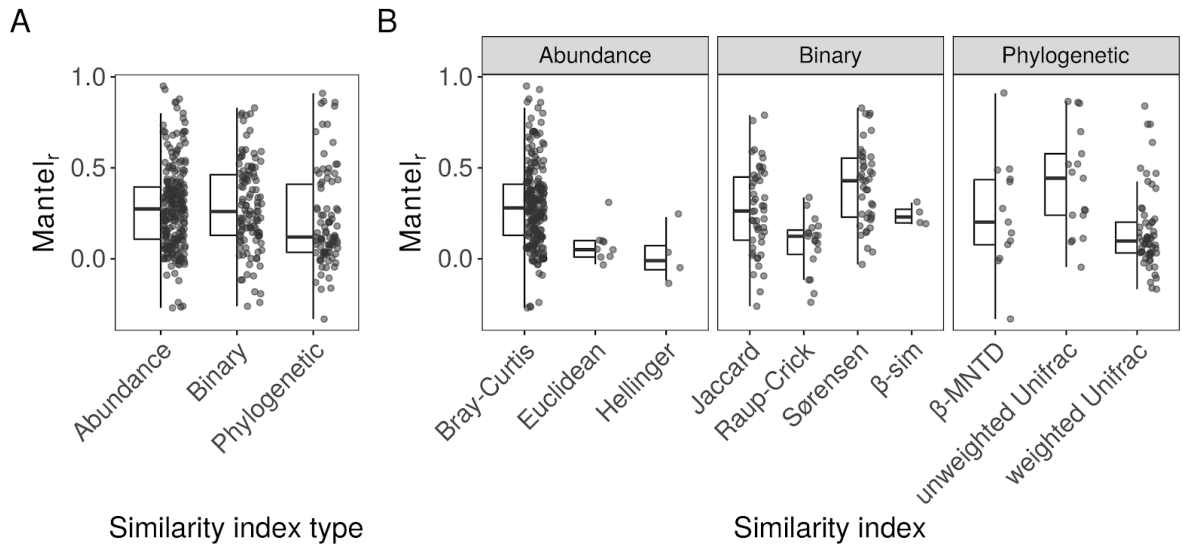
The logistics of multiplexing samples on high-throughput sequencing runs means that there is often a trade-off between the community coverage and sampling coverage of a study. However, we found no evidence of negative correlation between these two factors (Pearson's ρ = -0.03, P = 0.54). Nor did we detect any clear relationship between the number of samples ($\log_{10}$ sample coverage) and the strength of distance-decay relationships, even after accounting for study-specific differences with a mixed effects model (Fig. 4B; n = 451, coef = -0.06, marginal R^2 = 0.01, t = -1.40, P = 0.16).



311 Figure 4. The relationship between the strength of microbial distance-decay relationships
312 (Mantel$_r$) and A) community coverage, quantified as the number of sequences or individuals
313 counted per sample, and B) sample coverage, quantified as the number of individual
314 samples used to construct distance-decay relationships. Points are individual Mantel
315 correlation coefficients, coloured by the molecular technique used to characterise the

316 microbial community. Solid lines indicate statistically significant relationships ($P < 0.05$),
317 whilst dashed lines indicate non-significant relationships ($P > 0.05$), and shaded grey ribbons
318 represent 95% confidence intervals. Abbreviated molecular methods in the legend are
319 defined as follows (TRFLP = Terminal Restriction Fragment Length Polymorphism; ARISA =
320 Automated Ribosomal Intergenic Spacer Analysis; DGGE = Denaturing Gradient Gel
321 Electrophoresis; PLFA = Phospholipid Fatty Acid analysis; Sanger = Sanger sequencing of
322 cloned phylogenetically informative genes).

323 Choice of similarity index also had a clear impact on the strength of microbial distance-decay
324 relationships. As well as recording the specific similarity index used, we categorised indices
325 into types (binary, abundance, or phylogenetic) to test for broad differences in
326 distance-decay relationships. We analysed the nested interaction between similarity index
327 and index type, and found no clear differences between different index types (Fig. 5A; $F_{2, 424}$
328 = 1.48, P = 0.23). However, the interaction between index type and similarity index was
329 significant ($F_{7, 424}$ = 7.20, P 0.001). Post-hoc analysis revealed differences between similarity
330 indices within and between index types (Fig. 5B). Distance-decay relationships based on the
331 Raup-Crick index were weaker than those based on either Sørensen (coef = -0.38, P < 0.01)
332 or unweighted Unifrac indices (coef = -0.44, P < 0.01), whilst those based on weighted
333 Unifrac were weaker than both Sørensen (coef = -0.29, P < 0.001) and unweighted Unifrac
334 (coef = -0.35 P < 0.05). Finally, most studies did not explicitly state the correlation type used
335 to generate each Mantel test (n = 304), but of those that did, Spearman's correlation
336 coefficient was more frequently used (n = 86) than Pearson's (n = 62). We found no clear
337 difference in the strength of microbial distance-decay relationships using these two methods
338 ($F_{1, 146}$ = 2.47, P = 0.12).



339 Figure 5. Variation in the strength of microbial distance-decay relationships (*Mantel*$_r$)
340 calculated with different similarity index types (A), or individual indices (B). Only indices with
341 four or more distance-decay relationships were plotted for clarity.

342 *Comparison of Contextual and Methodological Variables*

343 In order to determine whether eco-environmental context or methodological factors better
344 explain the strength of microbial distance decay relationship, we specified two models, with
345 variables from these two categories, using a subset of the original data for which values
346 were obtained for all variables (n = 323). Each model had four variables, and used similar
347 degrees of freedom (context model df = 26, methodological model df = 27). The
348 methodological model outperformed the contextual model in terms of both AIC (Akaike
349 Information Criterion) and R^2 measures of model performance (Table 3). Notably, neither
350 model explained a high proportion of the variance, although both AIC and likelihood ratio
351 tests supported both models over a null (intercept only) model.

Table 3. Comparison of models specified using either contextual, or methodological variables. Akaike Information Criterion (AIC) and adjusted R^2 quantify the likelihood and fit of a model relative to the number of predictor variables, respectively.

Model	AIC	Adj-R^2	Likelihood ratio comparison to null (intercept only) model			
			ΔAIC	Sum of squares	F (df)	*P* value
Contextual	146.89	0.11	-13.69	5.34	2.61	< 0.001
Methodological	134.11	0.14	-26.46	6.47	3.17 (25)	< 0.001

Discussion

Previous research into the spatial ecology of microbial communities has not yielded a consistent distance-decay relationship. Our meta-analysis of 452 microbial distance-decay relationships suggests that the reasons for this lack of consistency are two-fold. Firstly, the differing contexts within which studies are conducted contribute variability to reported distance-decay relationships. In particular, we found that differing study systems were associated with variation in microbial distance-decay relationships. Secondly, methodological differences between studies, including dissimilarity index, data resolution, and sample coverage, all significantly affected observed distance-decay relationships. A central tenet of macroecology is the search for universal patterns and relationships; our results suggest generalisable relationships may only emerge when methodological approaches are appropriately coupled to ecological context.

Our comparison of distance-decay relationships between different study systems revealed that agricultural and especially grassland-based studies had weaker relationships than studies of other environments. Within these environments, soils were by far the most

frequently studied habitat, and we initially expected that, as soils maintain strong physicochemical gradients over small vertical and horizontal spatial scales (e.g. Dumbrell *et al.*, 2010), that these distance-decay relationships would be stronger than in other environments or habitats. It is possible that the environmental gradients present in soils do not change linearly over geographic distance, for example if similar environmental conditions are patchily distributed. Alternatively, soil microorganisms may be able to disperse more effectively than previously thought, perhaps via association with other soil organisms (e.g. bacterial migration along fungal hyphae; Warmink *et al.*, 2011), migratory species such as birds (Bisson *et al.*, 2007), wind blown soil particles (Favet *et al.*, 2013), or via bioaerosols (Joung *et al.*, 2017). The depth profile over which soil samples integrate may also play a role in obscuring distance-decay relationships, as surface soils show stronger distance-decay relationships than deeper ones, likely due to the greater intensity of dispersing propagules entering the surface (Li *et al.*, 2020). Furthermore, soils harbour extensive microbial “seed banks” of dormant organisms and/or relic DNA that could weaken the distance-decay relationship (Lennon & Jones, 2011; Carini *et al.*, 2016; Lennon *et al.*, 2018). Dormant cells and relic DNA are not subject to environmental selection, yet they are routinely detected in molecular community assays, likely diminishing the perceived effects of spatially-structured environmental selection on microbial communities (Locey *et al.*, 2020). Thus, in habitats such as soils, distinguishing dormant from active cells could result in stronger distance-decay relationships than those recorded previously, although evidence of the same effect on distance-decay slopes is mixed (Meyer *et al.*, 2018; Locey *et al.*, 2020). The extent to which this phenomenon plays a role in other environments is also unclear.

Originally, we expected the weakest distance-decay relationships to occur in connected aquatic environments such as rivers, oceans, or within single lakes, as the movement of water may provide an effective dispersal mechanism, homogenising microbial communities over larger spatial and environmental distances. In contrast, we found that aquatic

communities actually showed stronger distance-decay relationships than terrestrial systems. Soininen *et al.* (2007) recorded similar distance-decay rates between terrestrial, marine and aquatic ecosystems, showing that context-dependent distance-decay relationships may be a feature of microbial communities. We also found that the strength of distance-decay relationships was not different in studies based on single, or multiple, lakes, despite the difference in spatial extents of these studies. Lakes act as habitat islands within a terrestrial matrix and so dispersal limitation and environmental heterogeneity should be greater across multiple lakes than within a single lake, resulting in stronger distance-decay relationships in multi-lake studies. One explanation is that catchment-scale environmental parameters such as geology may homogenise environmental conditions across multiple lakes, meaning that environmental distances are similar within and between lakes. Alternatively, other biogeographic processes such as mass effects may homogenise communities between hydrologically connected lakes (Lindström & Bergström, 2004), especially where lakes are of different sizes (Reche *et al.*, 2005). Host-associated communities showed relatively strong, but variable distance-decay relationships. We suggest that this is caused jointly by the ecology of the host species, in combination with the degree of host-specificity with the associated microbiome. For example, if the host is not dispersal limited, and associates with a large variety of microorganisms, then the distance-decay relationship may be relatively weaker than those of either dispersal-limited hosts, or highly specific associated microbiomes.

The scale-dependence of various biogeographical relationships is well studied (Hillebrand, 2004; Bissett *et al.*, 2010; Martiny *et al.*, 2011; Soininen *et al.*, 2011), albeit with contrasting results. Soininen *et al.* (2011) reported that distance-decay relationships of various microbial communities were generally steeper over greater spatial extents, whilst our results suggest that increasing spatial extent does not significantly increase the strength of distance-decay relationships. As we analysed distance-decay strength rather than

steepness, our results are not necessarily contradictory. A strong distance-decay relationship occurs when, at a given spatial distance, all pairs of communities are equally dissimilar to one another, whereas a steep distance-decay occurs when communities separated by different distances are highly dissimilar to each other. We initially expected that spatial extent might alter the strength of distance-decay relationships as, at greater distances, decreased dispersal and increased environmental heterogeneity should reduce the variance in compositional similarity between pairs of communities (at a given distance). Instead, it could be that the spatial configuration or connectivity of the communities could be more important than spatial extent *per se*. For example, at a given spatial distance, some pairs of communities could be linked by dispersal and others not, increasing the variation in community similarity at each distance, and weakening the distance-decay relationship. In practice, this could occur in lake systems where at a certain geographic distance, some pairs of communities fall within the same lake and some in different lakes or when long-distance dispersal vectors link some pairs of communities separated by large distances, but not others, as has been proposed for halophilic microbial communities dispersing on migratory birds for example (Clark *et al.*, 2017; Kemp *et al.*, 2018). Furthermore, we observed that the minimum spatial extents differed according to the environment they were conducted in. Studies from terrestrial environments (e.g. grasslands and forests) or those based on soils generally incorporated smaller spatial extents than those based on aquatic systems (with the exception of some host-associated marine studies) or on habitats such as water or air. This could be due to the logistics of sampling at small scales. For example, sampling planktonic microbial communities at small (centimeters to meters) scales could be confounded by mixing caused by the sampling process or by tidal movements of water. Additionally, since many studies analysing microbial distance-decay relationships aimed to discern between environmental and spatial effects on microbial communities, it may be widely assumed that aquatic environments are more homogenous and/or that microorganisms are not dispersal

limited at these scales compared to more physically stable environments such as soils or sediments.

Distance-decay relationships are frequently interpreted as evidence for neutral community assembly processes such as dispersal limitation, in the microbial literature. Across microbial taxa, cell size is a trait thought to influence dispersal efficacy (Wilkinson, 2001; Wilkinson *et al.*, 2012; Zinger *et al.*, 2019), and so larger microorganisms such as micro-Eukarya should show stronger distance-decay relationships than smaller microorganisms such as Bacteria or Archaea. However, we found no evidence for this, suggesting that phylogenetically structured traits such as cell size may be less important compared to other contextual and methodological factors, or that the broad domain-level classification used here does not sufficiently capture different microbial cell sizes. As discussed previously, distance-decay relationships can arise from spatially autocorrelated environmental gradients as well as dispersal limitation (Nekola & White, 1999). Therefore, the lack of differences in biogeographic patterns observed at the domain level may be the result of a trade-off between dispersal-related processes and environmental filtering. For instance, bacterial distance-decay relationships may be less strongly influenced by dispersal than environmental filtering, and vice versa for Eukarya. Consequently, these influences may balance out at broad taxonomic levels, resulting in similar biogeographic patterns at the domain level.

In comparison to contextual factors, methodological factors were found to have a greater influence on microbial distance-decay relationships. The development of molecular methods, including high-throughput sequencing platforms, has vastly improved our ability to characterise microbial communities (Roesch *et al.*, 2007; Caporaso *et al.*, 2012). However, these methods differ in their resolution, community coverage, and ability to multiplex large numbers of samples, all of which we hypothesised could strengthen or weaken

distance-decay relationships by altering our estimation of microbial β-diversity. In contrast, we observed only a weak relationship between the strength of distance-decay relationships and community coverage, and no clear effects of different resolution methods, or the number of samples, suggesting that molecular methodology may not play as large a role in determining microbial biogeographic patterns as previously thought.

The ability to resolve closely related taxa has previously been found to be an important determinant of our ability to detect biogeographical patterns, as such patterns may only emerge when taxa are defined at sufficiently high resolution (Hanson *et al.*, 2012). Yet, other studies show that bioinformatically altering taxonomic resolution frequently has little effect on microbial biogeographic patterns. For example, increasing the similarity threshold at which operational taxonomic units are defined is thought to be equivalent to increasing the taxonomic resolution (Callahan *et al.*, 2017). Yet, empirical biogeographic relationships often appear robust to such manipulation, in a variety of taxa and ecosystems (Clark *et al.*, 2017; Glassman & Martiny, 2018; Meyer *et al.*, 2018), supporting our finding that resolution may not be important. Perhaps most molecular methodologies operate above resolutions at which biogeographic patterns begin to change, or more worryingly, perhaps we are still studying microbial biogeography at too low a resolution.

Aside from resolution, another important variable related to molecular methodology is community coverage. One of the few universal patterns that appears to hold true for most microbial communities is the “long-tailed” species abundance-distributions (Dumbrell *et al.*, 2010; Shoemaker *et al.*, 2017; Maček *et al.*, 2019), which is caused by the majority of microorganisms in a community being rare. The rarer taxa in microbial communities also tend to be the least widespread (Clark *et al.*, 2017; Lindh *et al.*, 2017; Meyer *et al.*, 2018; Shade & Stopnisek, 2019) and thus, only detecting the more abundant, widespread organisms would overestimate compositional similarity across communities, and

consequently, weaken distance-decay relationships due to the lower rate of turnover (Meyer *et al.*, 2018). Perhaps of more concern is that even with existing sequencing platforms, our surveys of environmental microbial communities still miss taxa that are vanishingly rare in the environment, such as extremophiles that persist in non-extreme habitats (Low-Décarie *et al.*, 2016). The ability of common species to reflect ecological patterns of the wider community is debated (Galand *et al.*, 2009; Heino & Soininen, 2010; van Dorst *et al.*, 2014) and is linked to a wider debate on the ecological importance of rare species that is far beyond the scope of this work (e.g. Gaston, 2012). However, rare microorganisms are well known to be of critical importance in the context of environmental perturbations (Shade *et al.*, 2014; Low-Décarie *et al.*, 2016) and in providing ecosystem processes (e.g. sulfate-reduction in peat soils, Hausmann *et al.*, 2016; and anaerobic ammonia-oxidation in river sediments Lansdown *et al.*, 2016) and as a result, ignoring them may further distance biogeographic patterns from ecosystem-level processes.

Against expectation, we observed no clear differences in distance-decay relationships using different similarity metric types, and differences between specific metrics were minimal. Distance-decay relationships based on the weighted Unifrac distance and the Raup-Crick index were weaker than those based on other metrics. The Raup-Crick index is less influenced by concurrent changes in species richness between communities, and as such is a more pure reflection of shifts in β-diversity (Chase *et al.*, 2011). Consequently, by removing the potentially confounding effects of richness differences, the Raup-Crick index will likely result in more variable estimates of similarity between communities, which would lead to weaker distance-decay relationships.

Phylogenetic metrics, such as Unifrac, cluster communities at a lower resolution, as two communities can be closely genetically related, yet distinct at fine taxonomic resolutions (e.g. species or strain-level). For example, Bryant *et al.* (2008) found that Unifrac similarity

was approximately three times higher than the compositional similarity of the same set of bacterial communities. Further, phylogenetic metrics may be inappropriate in less phylogenetically diverse environments (e.g. extreme systems) where phylogenetic diversity can be largely constrained to one taxon (e.g. the haloarchaea in hypersaline environments), leaving few “phylogenetic degrees of freedom” left to separate communities (Fukuyama, 2019). However, this does not account for the observed difference between weighted and unweighted versions of the Unifrac index, the former of which accounts for species’ relative abundance data, whilst the latter is binary (presence/absence based). A criticism of the weighted Unifrac index is that too much weight is placed on abundant taxa (Chen *et al.*, 2012). As abundant species are generally more widespread, placing too much weight on abundant taxa would have the effect of making communities appear artificially similar, exacerbating the effects of using a phylogenetic metric. As we observed no difference between binary and abundance-based compositional indices, the differences observed with weighted Unifrac appear to be the result of combining phylogenetic and weighted indices. We therefore suggest that weighted phylogenetic metrics may underestimate microbial biogeographic patterns, unless appropriate weight is given to rare and abundant taxa (Chen *et al.*, 2012).

Our analysis of 452 microbial distance-decay relationships also revealed the overwhelming preference of microbial ecologists to use classic dissimilarity indices such as the Bray-Curtis (n = 218), Jaccard (n = 49), Sørensen (n = 42) indices. These choices undoubtedly reflect a wider trend in ecology as a whole, however, it is pertinent to draw attention to more recently developed metrics that may be more appropriate given the properties of microbial datasets and the hypotheses being tested. Biotic interactions are drivers of microbial β-diversity (Hanson *et al.*, 2012), yet classic dissimilarity metrics do not account for co-occurrence information in communities. To this end, a new family of metrics described by Schmidt *et al.* (2017) include information on the average interactions of the taxa present, thus providing a

549 novel approach to integrating co-occurrence data into distance-decay relationships.
550 Microbiome sequencing data also have several characteristics that may be problematic in
551 the analysis of community (dis)similarities. For example, the non-biological variance of
552 sample sizes in sequence datasets can result in statistical artefacts that confound
553 biogeographic relationships (Baselga, 2007). Here, modifications made to some classic
554 indices by Chao *et al.* (2005) reduce the sensitivity of these indices to variable sample sizes
555 by accounting for unobserved species, thus reducing the need for post-sequencing
556 normalisation of sample sizes (McMurdie & Holmes, 2014). Furthermore, “fuzzy logic”-based
557 similarity indices are able to reduce the impact of false-absences or -presences on estimates
558 of β-diversity, which is beneficial for microbial ecology studies where rarefaction may induce
559 false-absences, and taxonomic assignment errors or contamination may lead to
560 false-presences. Additionally, most high-throughput sequence datasets are compositional.
561 Compositionality occurs as the arbitrary total number of sequences per sample imposed by
562 the sequencing machine causes species counts (abundances) to be dependent on each
563 other (e.g. if species A increases in abundance, species B and C will appear relatively less
564 abundant, even if their absolute abundance hasn’t changed). Binary indices should be
565 suitable as occurrences (presence/absences) are not affected by compositionality, unless
566 increases in the abundance of one or more species cause others to drop below the detection
567 limit, in which case fuzzy indices may be appropriate. Alternatively, metrics such as the
568 Aitchison distance perform well when appropriate (centered log-ratio) transformations are
569 applied to counts (Gloor *et al.*, 2017), or recently developed metrics such as the Rank Bias
570 Overlap index show promise for analysing similarity between communities based on species
571 abundance ranks (Webber *et al.*, 2010). Finally, many similarity metrics have been shown to
572 merge compositional turnover (replacement of species) and nestedness (whereby
573 communities are subsets of one another), thereby blurring the contribution of distinct
574 ecological processes to total community (dis)similarity. To combat this, modified versions of

classic indices such as Jaccard, Sorensen, and Bray-Curtis have been developed, allowing the partitioning of community similarity metrics into their turnover and nestedness components (Baselga, 2010; Podani & Schmera, 2011). We echo the call of Green and Bohannan (2006) for microbial ecologists to exercise more care in their choice of dissimilarity metrics, especially as many of these new metrics are implemented in popular and freely accessible software, such as R (e.g. Baselga and Orme, 2012).

Overall, our analyses revealed that methodological factors explain more variation in microbial distance-decay relationships than ecological context, but that both sets of factors alter our perception of this biogeographic pattern. Given the importance of methodological factors in determining the strength of microbial biogeographic patterns, it is intuitive to recommend standardising approaches across studies in order to minimise the statistical signals associated with methodological variance. However, our results show variance due to differing ecological contexts would still hinder drawing generalisable relationships across studies. Instead, we suggest that tailoring methodological choices towards specific ecological contexts may enhance generalisable relationships in microbial ecology. For instance, in searching for consistent relationships between ocean waters and terrestrial soils, it would be unrealistic to sample both at the same spatial grain and extent, as the heterogeneity in the physicochemical environment, and dispersal processes of their microbial communities, are fundamentally different. Similarly, we should not necessarily expect the relationships between soils and river sediments to be comparable, as microorganisms in soils can feasibly disperse in any direction, whereas in rivers or streams dispersal would be largely constrained by flow direction. Consequently, tailoring methodological approaches, such as the sampling design and/or (geographical) distance measure, to better reflect the environmental heterogeneity and dispersal dynamics between contrasting ecological contexts may enable us to negotiate the hierarchy of interacting factors that obscure macroecological patterns in microbial communities.

Conclusions

Our meta-analysis of >450 microbial distance-decay relationships revealed that factors related to the eco-environmental context within which a study was conducted, as well as the methodology of the study, jointly influence quantification of this classic biogeographic pattern. Against expectation, factors related to molecular methodology had relatively little effect on distance-decay relationships, whilst choice of dissimilarity metric was more important, highlighting that even after using robust, modern molecular methods, analytical choices have the power to obscure or enhance biogeographic patterns. We detected clear relationships between microbial distance-decay relationships and various contextual and methodological variables, yet combining these variables explained only a modest amount of variation in our dataset. This lack of explanatory power indicates that microbial biogeographic patterns depend on a number of contextual variables beyond those analysed here. In future, we suggest that microbial ecologists should place greater emphasis on quantifying habitat connectivity to better understand the dispersal processes that lead to spatial patterns such as the distance-decay relationship. Additionally, we recommend that experimental designs and data-collection strategies should be replicated spatially, taxonomically, temporally, or any combination therein where possible (e.g. Meyer *et al.*, 2018; Alzarhani *et al.*, 2019; Zinger *et al.*, 2019), facilitating a more generalised understanding of the variation in spatial microbial community patterns. The question of whether microbial communities show spatial patterns such as distance-decay relationships should be laid to rest; disentangling the web of ecological and environmental drivers that shape these patterns is the next challenge in microbial biogeography.

Data Availability Statement

Full raw data analysed in this manuscript are provided in Table S1. Full raw data and R code used in this manuscript will be uploaded to the Dryad data repository upon acceptance of this article.

Acknowledgements

This work was funded by a UK Natural Environment Research Council (NERC) quota award PhD studentship (471757).

References

Alzarhani, A.K., Clark, D.R., Underwood, G.J.C., Ford, H., Cotton, T.E.A. & Dumbrell, A.J. (2019) Are drivers of root-associated fungal community structure context specific? *The ISME Journal*, **13**, 1330.

Astorga, A., Oksanen, J., Luoto, M., Soininen, J., Virtanen, R. & Muotka, T. (2012) Distance decay of similarity in freshwater communities: do macro- and microorganisms follow the same rules? *Global Ecology and Biogeography*, **21**, 365–375.

Bartram, A.K., Lynch, M.D.J., Stearns, J.C., Moreno-Hagelsieb, G. & Neufeld, J.D. (2011) Generation of Multimillion-Sequence 16S rRNA Gene Libraries from Complex Microbial Communities by Assembling Paired-End Illumina Reads. *Applied and Environmental Microbiology*, **77**, 3846–3852.

Barwell, L.J., Isaac, N.J.B. & Kunin, W.E. (2015) Measuring β-diversity with species abundance data. *The Journal of Animal Ecology*, **84**, 1112–1122.

Baselga, A. (2010) Partitioning the turnover and nestedness components of beta diversity. *Global Ecology and Biogeography*, **19**, 134–143.

Baselga, A. & Orme, C.D.L. (2012) betapart: an R package for the study of beta diversity. *Methods in Ecology and Evolution*, **3**, 808–812.

Basham, E.W., Seidl, C.M., Andriamahohatra, L.R., Oliveira, B.F. & Scheffers, B.R. (2019) Distance–decay differs among vertical strata in a tropical rainforest. *Journal of Animal Ecology*, **88**, 114–124.

Bissett, A., Richardson, A.E., Baker, G., Wakelin, S. & Thrall, P.H. (2010) Life history determines biogeographical patterns of soil bacterial communities over multiple spatial scales. *Molecular Ecology*, **19**, 4315–4327.

Bisson, I.-A., Marra, P.P., Burtt, E.H., Sikaroodi, M. & Gillevet, P.M. (2007) A Molecular Comparison of Plumage and Soil Bacteria Across Biogeographic, Ecological, and Taxonomic Scales. *Microbial Ecology*, **54**, 65–81.

Bryant, J.A., Lamanna, C., Morlon, H., Kerkhoff, A.J., Enquist, B.J. & Green, J.L. (2008) Microbes on mountainsides: Contrasting elevational patterns of bacterial and plant diversity. *Proceedings of the National Academy of Sciences*, **105**, 11505–11511.

Callahan, B.J., McMurdie, P.J. & Holmes, S.P. (2017) Exact sequence variants should replace operational taxonomic units in marker-gene data analysis. *The ISME Journal*, **11**, 2639–2643.

Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Huntley, J., Fierer, N., Owens, S.M., Betley, J., Fraser, L., Bauer, M., Gormley, N., Gilbert, J.A., Smith, G. & Knight, R. (2012) Ultra-high-throughput microbial community analysis on the Illumina HiSeq and MiSeq platforms. *The ISME Journal*, **6**, 1621–1624.

Carini, P., Marsden, P.J., Leff, J.W., Morgan, E.E., Strickland, M.S. & Fierer, N. (2016) Relic DNA is abundant in soil and obscures estimates of soil microbial diversity. *Nature*

Microbiology, **2**, 1–6.
Chao, A., Chazdon, R.L., Colwell, R.K. & Shen, T.-J. (2005) A new statistical approach for assessing similarity of species composition with incidence and abundance data. *Ecology Letters*, **8**, 148–159.
Chase, J.M., Kraft, N.J.B., Smith, K.G., Vellend, M. & Inouye, B.D. (2011) Using null models to disentangle variation in community dissimilarity from variation in α-diversity. *Ecosphere*, **2**, art24.
Chen, J., Bittinger, K., Charlson, E.S., Hoffmann, C., Lewis, J., Wu, G.D., Collman, R.G., Bushman, F.D. & Li, H. (2012) Associating microbiome composition with environmental covariates using generalized UniFrac distances. *Bioinformatics*, **28**, 2106–2113.
Clark, D.R., Ferguson, R.M.W., Harris, D.N., Nicholass, K.J.M., Prentice, H.J., Randall, K.C., Randell, L., Warren, S.L. & Dumbrell, A.J. (2018) Streams of data from drops of water: 21st century molecular microbial ecology. *Wiley Interdisciplinary Reviews: Water*, **5**, e1280.
Clark, D.R., Mathieu, M., Mourot, L., Dufossé, L., Underwood, G.J.C., Dumbrell, A.J. & McGenity, T.J. (2017) Biogeography at the limits of life: Do extremophilic microbial communities show biogeographical regionalization? *Global Ecology and Biogeography*, **26**, 1435–1446.
van Dorst, J., Bissett, A., Palmer, A.S., Brown, M., Snape, I., Stark, J.S., Raymond, B., McKinlay, J., Ji, M., Winsley, T. & Ferrari, B.C. (2014) Community fingerprinting in a sequencing world. *FEMS microbiology ecology*, **89**, 316–330.
Dumbrell, A.J., Nelson, M., Helgason, T., Dytham, C. & Fitter, A.H. (2010) Relative roles of niche and neutral processes in structuring a soil microbial community. *The ISME Journal*, **4**, 337–345.
Dushoff, J., Kain, M.P. & Bolker, B.M. (2019) I can see clearly now: Reinterpreting statistical significance. *Methods in Ecology and Evolution*, **10**, 756–759.
Favet, J., Lapanje, A., Giongo, A., Kennedy, S., Aung, Y.-Y., Cattaneo, A., Davis-Richardson, A.G., Brown, C.T., Kort, R., Brumsack, H.-J., Schnetger, B., Chappell, A., Kroijenga, J., Beck, A., Schwibbert, K., Mohamed, A.H., Kirchner, T., de Quadros, P.D., Triplett, E.W., Broughton, W.J. & Gorbushina, A.A. (2013) Microbial hitchhikers on intercontinental dust: catching a lift in Chad. *The ISME Journal*, **7**, 850–867.
Franklin, R.B. & Mills, A.L. (2007) *Statistical Analysis Of Spatial Structure In Microbial Communities*. *The Spatial Distribution of Microbes in the Environment* (ed. by R.B. Franklin) and A.L. Mills), pp. 31–60. Springer Netherlands, Dordrecht.
Fukuyama, J. (2019) Emphasis on the deep or shallow parts of the tree provides a new characterization of phylogenetic distances. *Genome Biology*, **20**, 131.
Galand, P.E., Casamayor, E.O., Kirchman, D.L. & Lovejoy, C. (2009) Ecology of the rare microbial biosphere of the Arctic Ocean. *Proceedings of the National Academy of Sciences*, **106**, 22427–22432.
Gaston, K.J. (2012) The importance of being rare. *Nature*, **487**, 46–47.
Glassman, S.I. & Martiny, J.B.H. (2018) Broadscale Ecological Patterns Are Robust to Use of Exact Sequence Variants versus Operational Taxonomic Units. *mSphere*, **3**.
Glenn, T.C. (2011) Field guide to next-generation DNA sequencers. *Molecular Ecology Resources*, **11**, 759–769.
Gloor, G.B., Macklaim, J.M., Pawlowsky-Glahn, V. & Egozcue, J.J. (2017) Microbiome Datasets Are Compositional: And This Is Not Optional. *Frontiers in Microbiology*, **8**.
Green, J. & Bohannan, B.J.M. (2006) Spatial scaling of microbial biodiversity. *Trends in Ecology & Evolution*, **21**, 501–507.
Hanson, C.A., Fuhrman, J.A., Horner-Devine, M.C. & Martiny, J.B.H. (2012) Beyond biogeographic patterns: processes shaping the microbial landscape. *Nature Reviews*

Microbiology, **10**, 497–506.
Harrison, F. (2012) Getting started with meta-analysis. *Journal of Applied Ecology*, 1–10.
Hausmann, B., Knorr, K.-H., Schreck, K., Tringe, S.G., Glavina del Rio, T., Loy, A. & Pester, M. (2016) Consortia of low-abundance bacteria drive sulfate reduction-dependent degradation of fermentation products in peat soil microcosms. *The ISME Journal*, **10**, 2365–2375.
Hazard, C., Gosling, P., Gast, C.J. van der, Mitchell, D.T., Doohan, F.M. & Bending, G.D. (2013) The role of local environment and geographical distance in determining community composition of arbuscular mycorrhizal fungi at the landscape scale. *The ISME Journal*, **7**, 498–508.
Heino, J. & Soininen, J. (2010) *Are common species sufficient in describing turnover in aquatic metacommunities along environmental and spatial gradients*.
Hillebrand, H. (2004) On the Generality of the Latitudinal Diversity Gradient. *The American Naturalist*, **163**, 192–211.
Joung, Y.S., Ge, Z. & Buie, C.R. (2017) Bioaerosol generation by raindrops on soil. *Nature Communications*, **8**, 1–10.
Kemp, B.L., Tabish, E.M., Wolford, A.J., Jones, D.L., Butler, J.K. & Baxter, B.K. (2018) The Biogeography of Great Salt Lake Halophilic Archaea: Testing the Hypothesis of Avian Mechanical Carriers. *Diversity*, **10**, 124.
Kivlin, S.N. Global mycorrhizal fungal range sizes vary within and among mycorrhizal guilds but are not correlated with dispersal traits. *Journal of Biogeography*, **n/a**.
Kivlin, S.N., Winston, G.C., Goulden, M.L. & Treseder, K.K. (2014) Environmental filtering affects soil fungal community composition more than dispersal limitation at regional scales. *Fungal Ecology*, **12**, 14–25.
Lajeunesse, M.J. (2016) Facilitating systematic reviews, data extraction and meta-analysis with the metagear package for r. *Methods in Ecology and Evolution*, **7**, 323–330.
Lansdown, K., McKew, B.A., Whitby, C., Heppell, C.M., Dumbrell, A.J., Binley, A., Olde, L. & Trimmer, M. (2016) Importance and controls of anaerobic ammonium oxidation influenced by riverbed geology. *Nature Geoscience*, **9**, 357–360.
Lennon, J.T. & Jones, S.E. (2011) Microbial seed banks: the ecological and evolutionary implications of dormancy. *Nature Reviews. Microbiology*, **9**, 119–130.
Lennon, J.T., Muscarella, M.E., Placella, S.A. & Lehmkuhl, B.K. (2018) How, When, and Where Relic DNA Affects Microbial Diversity. *mBio*, **9**.
Li, P., Li, W., Dumbrell, A.J., Liu, M., Li, G., Wu, M., Jiang, C. & Li, Z. (2020) Spatial Variation in Soil Fungal Communities across Paddy Fields in Subtropical China. *mSystems*, **5**.
Lindh, M.V., Sjöstedt, J., Ekstam, B., Casini, M., Lundin, D., Hugerth, L.W., Hu, Y.O.O., Andersson, A.F., Andersson, A., Legrand, C. & Pinhassi, J. (2017) Metapopulation theory identifies biogeographical patterns among core and satellite marine bacteria scaling from tens to thousands of kilometers. *Environmental Microbiology*, **19**, 1222–1236.
Lindström, E.S. & Bergström, A.-K. (2004) Influence of inlet bacteria on bacterioplankton assemblage composition in lakes of different hydraulic retention time. *Limnology and Oceanography*, **49**, 125–136.
Locey, K.J., Muscarella, M.E., Larsen, M.L., Bray, S.R., Jones, S.E. & Lennon, J.T. (2020) Dormancy dampens the microbial distance–decay relationship. *Philosophical Transactions of the Royal Society B: Biological Sciences*, **375**, 20190243.
Low-Décarie, E., Fussmann, G.F., Dumbrell, A.J. & Bell, G. (2016) Communities that thrive in extreme conditions captured from a freshwater lake. *Biology Letters*, **12**, 20160562.
Luan, L., Jiang, Y., Cheng, M., Dini-Andreote, F., Sui, Y., Xu, Q., Geisen, S. & Sun, B. (2020) Organism body size structures the soil microbial and nematode community assembly

at a continental and global scale. *Nature Communications*, **11**, 6406.
Maček, I., Clark, D.R., Šibanc, N., Moser, G., Vodnik, D., Müller, C. & Dumbrell, A.J. (2019) Impacts of long-term elevated atmospheric CO2 concentrations on communities of arbuscular mycorrhizal fungi. *Molecular Ecology*, **28**, 3445–3458.
Mantel, N. (1967) The Detection of Disease Clustering and a Generalized Regression Approach. *Cancer Research*, **27**, 209–220.
Martiny, J.B.H., Eisen, J.A., Penn, K., Allison, S.D. & Horner-Devine, M.C. (2011) Drivers of bacterial β-diversity depend on spatial scale. *Proceedings of the National Academy of Sciences*, **108**, 7850–7854.
McMurdie, P.J. & Holmes, S. (2014) Waste Not, Want Not: Why Rarefying Microbiome Data Is Inadmissible. *PLOS Computational Biology*, **10**, e1003531.
Meyer, K.M., Memiaghe, H., Korte, L., Kenfack, D., Alonso, A. & Bohannan, B.J.M. (2018) Why do microbes exhibit weak biogeographic patterns? *The ISME Journal*, **12**, 1404–1413.
Muyzer, G. (1999) DGGE/TGGE a method for identifying genes from natural ecosystems. *Current Opinion in Microbiology*, **2**, 317–322.
Nakagawa, S. & Schielzeth, H. (2013) A general and simple method for obtaining R2 from generalized linear mixed-effects models. *Methods in Ecology and Evolution*, **4**, 133–142.
Nekola, J.C. & McGill, B.J. (2014) Scale dependency in the functional form of the distance decay relationship. *Ecography*, **37**, 309–320.
Nekola, J.C. & White, P.S. (1999) The distance decay of similarity in biogeography and ecology. *Journal of Biogeography*, **26**, 867–878.
Norros, V., Rannik, Ü., Hussein, T., Petäjä, T., Vesala, T. & Ovaskainen, O. (2014) Do small spores disperse further than large spores? *Ecology*, **95**, 1612–1621.
Podani, J. & Schmera, D. (2011) A new conceptual and methodological framework for exploring and explaining pattern in presence – absence data. *Oikos*, **120**, 1625–1638.
R Core Team (2019) *R: A Language and Environment for Statistical Computing*, R Foundation for Statistical Computing, Vienna, Austria.
Ramette, A. (2007) Multivariate analyses in microbial ecology. *FEMS Microbiology Ecology*, **62**, 142–160.
Reche, I., Pulido-Villena, E., Morales-Baquero, R. & Casamayor, E.O. (2005) Does Ecosystem Size Determine Aquatic Bacterial Richness? *Ecology*, **86**, 1715–1722.
Roesch, L.F.W., Fulthorpe, R.R., Riva, A., Casella, G., Hadwin, A.K.M., Kent, A.D., Daroub, S.H., Camargo, F.A.O., Farmerie, W.G. & Triplett, E.W. (2007) Pyrosequencing enumerates and contrasts soil microbial diversity. *The ISME journal*, **1**, 283–290.
Rosenberg, M.S., Rothstein, H.R. & Gurevitch, J. (2013) Effect sizes: Conventional choices and calculations. *Handbook of Meta-analysis in Ecology and Evolution*, 61–71.
Shade, A., Jones, S.E., Caporaso, J.G., Handelsman, J., Knight, R., Fierer, N. & Gilbert, J.A. (2014) Conditionally Rare Taxa Disproportionately Contribute to Temporal Changes in Microbial Diversity. *mBio*, **5**.
Shade, A. & Stopnisek, N. (2019) Abundance-occupancy distributions to prioritize plant core microbiome membership. *Current Opinion in Microbiology*, **49**, 50–58.
Shmida, A. & Wilson, M.V. (1985) Biological Determinants of Species Diversity. *Journal of Biogeography*, **12**, 1–20.
Shoemaker, W.R., Locey, K.J. & Lennon, J.T. (2017) A macroecological theory of microbial biodiversity. *Nature Ecology & Evolution*, **1**, 1–6.
Soininen, J., Korhonen, J.J., Karhu, J. & Vetterli, A. (2011) Disentangling the spatial patterns in community composition of prokaryotic and eukaryotic lake plankton. *Limnology and Oceanography*, **56**, 508–520.

Soininen, J., McDonald, R. & Hillebrand, H. (2007) The distance decay of similarity in ecological communities. *Ecography*, **30**, 3–12.
Sorte, F.A.L., McKinney, M.L., Pyšek, P., Klotz, S., Rapson, G.L., Celesti-Grapow, L. & Thompson, K. (2008) Distance decay of similarity among European urban floras: the impact of anthropogenic activities on β diversity. *Global Ecology and Biogeography*, **17**, 363–371.
Steinbauer, M.J., Dolos, K., Reineking, B. & Beierkuhnlein, C. (2012) Current measures for distance decay in similarity of species composition are influenced by study extent and grain size. *Global Ecology and Biogeography*, **21**, 1203–1212.
Vašutová, M., Mleczko, P., López-García, A., Maček, I., Boros, G., Ševčík, J., Fujii, S., Hackenberger, D., Tuf, I.H., Hornung, E., Páll-Gergely, B. & Kjøller, R. (2019) Taxi drivers: the role of animals in transporting mycorrhizal fungi. *Mycorrhiza*, **29**, 413–434.
Vos, M., Wolf, A.B., Jennings, S.J. & Kowalchuk, G.A. (2013) Micro-scale determinants of bacterial diversity in soil. *FEMS Microbiology Reviews*, **37**, 936–954.
Warmink, J.A., Nazir, R., Corten, B. & van Elsas, J.D. (2011) Hitchhikers on the fungal highway: The helper effect for bacterial migration via fungal hyphae. *Soil Biology and Biochemistry*, **43**, 760–765.
Webber, W., Moffat, A. & Zobel, J. (2010) A similarity measure for indefinite rankings. *ACM Transactions on Information Systems*, **28**, 1–38.
Wilkinson, D.M. (2001) What is the upper size limit for cosmopolitan distribution in free-living microorganisms? *Journal of Biogeography*, **28**, 285–291.
Wilkinson, D.M., Koumoutsaris, S., Mitchell, E.A.D. & Bey, I. (2012) Modelling the effect of size on the aerial dispersal of microorganisms. *Journal of Biogeography*, **39**, 89–97.
Zinger, L., Taberlet, P., Schimann, H., Bonin, A., Boyer, F., Barba, M.D., Gaucher, P., Gielly, L., Giguet-Covex, C., Iribar, A., Réjou-Méchain, M., Rayé, G., Rioux, D., Schilling, V., Tymen, B., Viers, J., Zouiten, C., Thuiller, W., Coissac, E. & Chave, J. (2019) Body size determines soil community assembly in a tropical forest. *Molecular Ecology*, **28**, 528–543.