Title: What drives study-dependent differences in distance-decay relationships of microbial communities?

Running title: Meta-Analysis of Microbial Distance-Decay Relationships

Keywords: Bacteria, Archaea, Eukarya, Mantel test, macroecology, biogeography, dispersal limitation, community dissimilarity

Abstract

Aim: Ecological communities that exist closer together in space are generally more compositionally similar than those far apart, as defined by the distance-decay of similarity relationship. However, recent research has revealed substantial variability in the distance-decay relationships of microbial communities between studies of different taxonomic groups, ecosystems, spatial scales, as well as between those using different molecular methodologies (e.g. high-throughput sequencing versus molecular fingerprinting). Here, we test how these factors influence the effect size of microbial distance-decay relationships, to draw generalisations about how microbial β-diversity scales with space.

Location: Global.

Time period: Studies published between 2005-2019 (inclusive).

Major taxa studied: Bacteria, Archaea, and microbial Eukarya.

Methods: We conducted a meta-analysis of microbial distance-decay relationships, using the Mantel correlation coefficient as a measure of effect size. We assembled 452 data points, varying in environmental/ecological context or methodological approaches, and used linear models to test the effects of each variable.

Results: Both ecological and methodological factors had significant impacts on the strength of microbial distance-decay relationships. Specifically, larger spatial extents increased the strength of these relationships, whilst differences also emerged between environments and habitats, with soils showing significantly weaker distance-decay relationships than other habitats. Methodological factors such as sequencing depth were positively related to the strength of distance-decay relationships, and choice of dissimilarity metric was also

important, with phylogenetic metrics generally giving weaker distance-decay relationships than binary or abundance-based indices.

Main conclusions: We conclude that widely studied microbial biogeographic patterns, such as the distance-decay relationship, vary by ecological context but are primarily distorted by methodological choices. Consequently, we suggest that by linking methodological approaches appropriately to the ecological context of a study, we can progress towards generalisable biogeographic relationships in microbial ecology.

Introduction

The distance-decay of community similarity is one of the most widely studied relationships in macroecology (Nekola & White, 1999; Soininen *et al.*, 2007). This relationship quantifies the decrease in compositional similarity (β-diversity) between communities with increasing geographic distance separating them, and demonstrates that nearby communities are more similar to each other than distantly-separated communities. Distance-decay relationships can arise through several different, but often interacting ecological and evolutionary processes, and consequently ecologists have extensively debated the underlying mechanisms that generate such patterns (Nekola & White, 1999; Soininen *et al.*, 2007; Hanson *et al.*, 2012). Spatial structuring of the environment can lead to distance-decay relationships, as communities close together in space are likely to experience more similar environmental conditions, and thus contain more similar communities than those situated in very different environmental conditions. Dispersal limitation can also lead to distance-decay relationships by limiting the connectivity between communities, meaning that communities closer together in space will share more species through localised dispersal than those further apart.

Distance-decay relationships are well documented in a multitude of plant and animal communities. Yet, they are of particular interest to microbial ecologists because microorganisms were typically assumed to have ubiquitous distributions for several reasons. Firstly, their small size facilitates passive dispersal over large geographic distances by vectors such as wind, bio-aerosolization, ocean currents or migrating animals (Bisson *et al.*, 2007; Favet *et al.*, 2013; Joung *et al.*, 2017; Vašutová *et al.*, 2019), thus potentially overcoming dispersal limitation as a contributing factor to microbial community composition. Additionally, microorganisms often maintain high population densities in the environment leading to dispersal by “mass effects”, whereby high dispersal rates from areas of increased

population density maintain populations in less optimal environments (Shmida & Wilson, 1985), helping them to overcome the constraints of spatially-structured environmental gradients. Finally, some microorganisms are able to enter dormant states, whether as vegetative cells or as cysts or spores (Locey *et al.*, 2019), allowing them to survive and disperse through suboptimal environments, simultaneously enhancing their dispersive abilities, and reducing the influence of spatially-structured environmental gradients (Low-Décarie *et al.*, 2016). Combined, these traits theoretically lower microbial β-diversity by increasing the amount of shared species between distant communities, in turn leading to weaker distance-decay relationships compared to macroorganisms. However, empirical tests of microbial distance-decay relationships have yielded mixed results. Many studies have detected little or no evidence of distance-decay relationships in microbial communities (Hazard *et al.*, 2013; Kivlin *et al.*, 2014), whilst others report relationships of varying steepness, across a range of spatial extents, study systems, and taxa (Dumbrell *et al.*, 2010; Martiny *et al.*, 2011; Clark *et al.*, 2017). Thus, despite hundreds of empirical studies, the generality of spatial patterns in microbial communities remains unclear, and we are no closer to understanding whether variability in the spatial scaling relationships of microbial β-diversity originates from ecological or methodological sources.

Variation in microbial distance-decay relationships could be due to different environmental or ecological contexts in studies. Here, we consider environmental context as the variability in the physico-chemical environment (e.g. temperature, pH, topology), and ecological context as the total suite of species present and their interactions. The study systems commonly of interest to microbial ecologists vary in terms of connectivity, which may facilitate or hinder dispersal between communities, thus leading to flatter or steeper distance-decay relationships, respectively. In well connected systems where dispersal is more feasible, such as oceanic waters, distance-decay relationships should be weaker than systems in which dispersal is limited, such as host-associated systems or soil systems, where distance-decay

relationships are weaker in deeper soil horizons (Li *et al.*, 2020). Moreover, study systems differ in the spatially structured environmental gradients they support. Topsoil for example, supports strong environmental gradients over distances of a few meters (e.g. Dumbrell *et al.*, 2010), leading to steep distance-decay relationships due to spatially structured environmental gradients. Additionally, different study taxa are likely to yield variable distance-decay relationships because they may differ in traits that are linked to dispersal efficacy. For example, small cell sizes lead to more efficient long distance dispersal (Wilkinson, 2001; Wilkinson *et al.*, 2012; Norros *et al.*, 2014), thus organisms with larger cell sizes, such as microbial Eukarya, should be more strongly dispersal limited than those with small cell sizes, such as Bacteria (although this may not be true for all taxa e.g. see Kivlin, 2020). Finally, it is known that spatial extent can influence our perception of ecological relationships, which may contribute to variable distance-decay relationships (Steinbauer *et al.*, 2012). Studies incorporating larger spatial extents may find stronger distance-decay relationships as they are more likely to incorporate spatial scales at which the taxa are dispersal limited and/or at which environmental conditions become spatially structured (Martiny *et al.*, 2011).

Whilst the context in which a study was undertaken may contribute to variability in microbial distance-decay relationships, so too could different methodologies. Technological advances have yielded new insight into the structure and functioning of development of environmental microbial communities (Clark *et al.*, 2018). However, rapid turnover in molecular methodologies means that our perception of microbial β-diversity patterns is based on methods that vary substantially in both coverage (ability to detect a greater proportion of the community in a given sample) and resolution (ability to resolve closely related taxa) (Muyzer, 1999; Glenn, 2011). Early methods such as clone library sequencing and community fingerprinting methods (e.g. denaturing gradient gel electrophoresis (DGGE), terminal restriction fragment length polymorphism (TRFLP), or phospholipid fatty acid (PLFA)

analysis) are limited in their ability to detect rare taxa (Bartram *et al.*, 2011), undoubtedly missing taxa that are extremely rare (Low-Décarie *et al.*, 2016). In turn, this could reduce the detected β-diversity, inflating estimated community similarity and weakening distance-decay relationships (Hanson *et al.*, 2012). In contrast, high-throughput sequencing (HTS) platforms (also frequently referred to as next-generation sequencing (NGS)) can deliver sequencing depths upwards of tens or even hundreds of thousands of sequences per sample (Caporaso *et al.*, 2012), thus improving both community coverage (the detected proportion of a given community), and allowing more samples to be examined in a single study (sample coverage). Consequently, variation between molecular methods in their ability to resolve closely related taxa, and to detect rare taxa can be an additional source of variability in microbial beta diversity, which by extension can either weaken or strengthen microbial distance-decay relationships.

In addition to the molecular methods, the choice of analytical methods, such as similarity metric, can influence distance-decay relationships. The similarity of communities can vary according to the identity and abundance of the species present, their phylogenetic relationships, and by external factors such as varying sample sizes. Thus, similarity metrics that vary by one or more of these characteristics would likely result in contrasting distance-decay relationships (Chao *et al.*, 2005; Barwell *et al.*, 2015). For example, phylogenetic indices would be expected to yield weaker distance-decay relationships than other metrics, because communities that have no species in common can still be highly phylogenetically similar if the species share many branches of a phylogenetic tree, thus reducing the decay of similarity over geographic distance (Bryant *et al.*, 2008). On the other hand, quantitative indices compare not only the composition of species present, but also their abundance in each community, reflecting finer-scale changes in community structure, and thus should result in stronger distance-decay relationships by providing an additional axis (species abundances) by which communities can differ.

Here, to disentangle the effects of both contextual (e.g. spatial extent, taxon, or ecosystem) and methodological (e.g. means of identifying/differentiating taxa, or similarity metric) variables on microbial distance-decay relationships, we undertook a meta-analysis to test the following specific hypotheses:

- H_1 Bacteria and Archaea will show weaker (lower effect size) distance-decay relationships than micro-eukaryotic taxa due to their smaller size and higher population densities in most environments.
- H_2 Ecosystems that contain steep physicochemical gradients will have stronger distance-decay relationships due to spatially-structured niche partitioning of communities.
- H_3 The spatial extent of a study will be positively related to the strength of any resulting distance-decay relationships, as larger extent studies incorporate greater environmental heterogeneity and lower dispersal rates between communities.
- H_4 High-throughput sequencing methods will yield stronger distance-decay relationships due to: a) their ability to resolve closely related taxa, b) their greater community coverage (e.g. number of sequences per sample, or number of individuals counted per sample), and/or c) their greater sample coverage.
- H_5 Phylogenetic similarity metrics (e.g. Unifrac, beta nearest taxon index) will result in weaker distance-decay relationships than other metrics as communities can be phylogenetically similar, yet different at fine taxonomic resolutions, whilst quantitative metrics (e.g. Bray-Curtis, Hellinger, Euclidean) will yield the strongest as they reflect changes in both species composition and abundance.

Methods

Meta-Analysis

In order to test our hypotheses, we first gathered available data on microbial distance-decay relationships via a systematic literature search. To do this, five search terms were selected to detect relevant studies (Table 1). All literature searches were conducted using the Web of Science search portal on 18/04/2020, and all results published between 1900-2019 (inclusive) were retained. To further filter the dataset to studies suitable for testing our hypotheses, search results were downloaded and manually screened using the “metagear” (Lajeunesse, 2016) package in R (version 3.4.1; R Core Team, 2019). Here, suitable studies were those that tested the relationship between community similarity and geographic distance in microbial communities, and not studies of “macroorganisms”, or studies of strain-level genetic distance (e.g. using multi-locus sequence typing). Furthermore, studies that did not test distance-decay relationships using Mantel correlation, or that used only partial Mantel tests, were also discarded. We did not identify any potentially suitable studies that were published prior to 1967, the year the Mantel test was described (Mantel, 1967), and the earliest suitable study was published in 2005.

Table 1. Details of Web of Science search terms, and the number of results for each search.

Search	Search Term	Number of results
1	TS = (biogeograph*) AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	2907
2	TS = (macroecolog*) AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	136
3	TS = ("everything is everywhere") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	66
4	TS = ("geographic distance") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	220
5	TS = ("distance decay") AND TS = (bacteria* OR archaea* OR microb* OR microorganism*)	186

From these studies, we extracted Mantel correlation coefficients (*r*) as an effect-size measure for each distance-decay relationship. The Mantel test is a permutation-based method used to test for correlation between two distance matrices, or in the context of this study, community (dis)similarity and geographic distance. The Mantel test statistic is an ideal measure of effect size for use in meta-analytical frameworks for several reasons. Firstly, the Mantel correlation test is the most frequently used method for testing distance-decay relationships in microbial ecology (Franklin & Mills, 2007; Ramette, 2007). Secondly, as the Mantel coefficient is a standardised correlation coefficient (i.e. is bound by -1 and 1), it provides an easily interpretable and comparable measure of effect size (Harrison, 2012).

We ensured all Mantel correlation coefficients reflected correlations between geographic distance and community dissimilarity, rather than similarity, by multiplying correlation coefficients by -1 where necessary (so that positive values indicate a typical distance-decay relationship). Partial Mantel statistics (which test for correlation between two matrices whilst controlling for a third) were excluded as they are influenced by other variables included in the test, and are therefore not easily comparable between studies. All Mantel correlation coefficients were transformed to *z*-scores using Fisher's *z* transformation, as recommended by Koricheva *et al.*, (2013). All subsequent statistical analyses were conducted on the transformed *z*-scores, whilst original Mantel correlation coefficients were used to make figures, for ease of interpretation.

In order to test our hypotheses, several variables relating to the context and methodology of each distance-decay relationship were recorded. Details of these variables are described in Box 1.

Box 1. Details of the explanatory variables extracted from each study.

Resolution

Each distance-decay relationship was categorised into either high-resolution (high-throughput or Sanger sequencing), low resolution (molecular e.g. ARISA, TRFLP, DGGE, PhyloChip, PLFA), or low resolution (morphological), based on the method's ability to distinguish between closely related organisms.

Community Coverage
This refers to the sequencing depth in sequencing-based studies, or number of individuals counted in morphology-based studies, per sample. For sequencing studies, we recorded the number of sequences after rarefaction, or if this was not given, the average number of sequences per sample. As there is no comparable measure of coverage for fingerprinting studies, we excluded them from analyses of community coverage.

Sample Coverage
Sample coverage refers to the sample size (e.g. number of communities/samples) of each distance-decay relationship.

Dissimilarity Index
The dissimilarity index used to calculate each distance-decay relationship. Recorded dissimilarity indices were then categorised as quantitative (Bray-Curtis, Horn-Morisita, Euclidean, Hellinger, Theta), qualitative (Jaccard, Raup-Crick, Sørensen, Simpson, βsim), or phylogenetic (weighted or unweighted Unifrac, Rao, β-mean nearest taxon distance, β-mean pairwise distance).

Study Taxon
Each distance-decay relationship was binned into the following broad taxonomic categories based on the taxonomy of the focal organisms (Archaea, Bacteria, Fungi, or other microbial Eukarya), or combination of these categories if a relationship was based on multiple taxa (for example due to using sequencing primers that detect both Archaea and Bacteria). Fungi grouped separately from other micro-Eukaryotes due to their distinct reproductive strategy (e.g. spore-production) and the fact the they are frequently targeted using distinct molecular approaches (e.g. via taxon-specific primer sets), in contrast to most other studies of micro-Eukarya.

Spatial Extent
This is the maximum distance separating communities in km. If this was not stated in text or provided in supplementary material (e.g. in a geographic distance matrix), it was calculated from given geographic coordinates, estimated from a plot of the distance-decay relationship, or estimated from scaled maps.

Environment
We broadly categorised distance-decay relationships based on the type of environment (agriculture, air, aquifer, coastal wetlands/intertidal, desert, dune, forest, glacier, grassland, lake, marine, coastal marshes, mine, river, snow, urban) within which they were sampled. Whilst these categories are not mutually exclusive, we categorised each study based on which environment best represented the environmental context in which each study was undertaken.

Habitat

The type of environmental material that the sampled communities occupied. We categorised distance-decay relationships as: air, host-associated, sediment, snow, soil, water.

Statistical Analyses

In order to determine whether distance-decay relationships varied between categorical variables (as in hypotheses 1, 2, 4, and 5), we used ANOVA tests. In tests where significant differences between groups were found, Tukey's Honest Significant Difference (HSD) tests were used to determine which groups were different. Linear models were used to test relationships between effect sizes and continuous variables such as spatial extent and community coverage. The variables spatial extent and community coverage were initially log transformed to aid model fitting, as they spanned several orders of magnitude. To compare the overall influence of ecological vs methodological factors on microbial distance-decay relationships, we compared two full models (including all relevant variables) using AIC scores, on a subset of the data for which all variables were successfully recorded. We then sought to identify a smaller number of variables that adequately predicted the effect size of microbial distance-decay relationships by using a drop-term likelihood ratio procedure.

Results

Our Web of Science searches resulted in 2,982 unique search results. Manual screening of the abstracts yielded 951 studies that were deemed to be potentially suitable for use in this analysis. A total of 452 Mantel correlation coefficients were successfully obtained from 187 studies represented in 61 journals (Fig. S1). Reported Mantel correlation coefficients ranged from -0.33 to 0.95, with a mean of 0.27 (std. error = 0.011).

Influence of Context on the Distance-Decay Relationship

In order to determine whether contextual factors can influence the strength of distance-decay relationships, the influence of ecological factors including study taxa, study system, and spatial scale were tested. Within the dataset, the most commonly studied taxa were Bacteria (n = 238), followed by Fungi (n = 93), other microbial Eukaryotes (n = 67), and Archaea (n = 26). We found no clear differences in the effect sizes of distance-decay relationships between these taxa ($F_{5,\ 441}$ = 0.97, P = 0.43), although distance-decay relationships incorporating bacterial and fungal communities showed the smallest effect sizes, albeit only from six studies (Fig. 1).

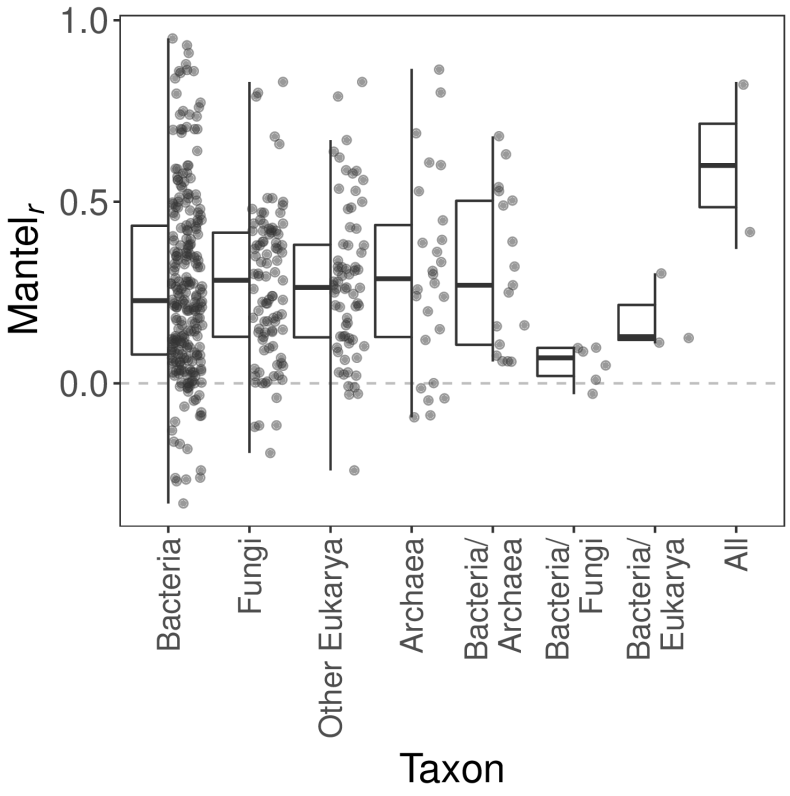


Figure 1. Effect sizes (Mantel$_r$) of distance-decay relationships based on different study taxa. A larger effect size indicates stronger positive correlation between community dissimilarity and geographic distance. The “All” category consists of studies that incorporated all microbial taxonomic groups, whilst combined categories (e.g. Bacteria/Archaea) incorporate communities from multiple taxonomic groups (e.g. bacterial and archaeal communities).

The distance-decay relationships in our dataset originated from 16 different environments. Of these, five were represented by three, or fewer, distance-decay relationships, and so were excluded from further analyses (marsh; n = 3, snow; n = 3, dune, mine, aquifer; n = 1). The most frequently studied environments were grasslands (n = 96), marine (n = 88), and lakes and forests (n = 76 for both). We found clear differences in the effect sizes of distance-decay relationships between environments (Fig. 2A; $F_{10,\ 432}$ = 3.187, P < 0.001). Specifically, and perhaps counter-intuitively, grassland-based studies tended to have weaker distance-decay relationships than those from aquatic environments such as lakes, rivers, or the marine environment (|coef| > 0.17, P < 0.05 for all comparisons). Urban environments, which included built environments such as sewers and indoor air, also produced low effect sizes, although with only four data points, this difference was not statistically clear (P > 0.43 for all comparisons). A more detailed analysis of the interaction between environment type and habitat revealed that, whilst environments ($F_{9,\ 420}$ = 3.29, P < 0.001) and habitat ($F_{3,\ 420}$ = 6.65, P < 0.001) differ from each other, their interaction was not statistically significant ($F_{4,\ 420}$ = 1.93, P = 0.10). In fact, within environments, only marine host-associated and marine water-based distance-decay relationships were clearly different from each other (Fig. 2B), with host-associated communities showing significantly larger effect sizes (coef = 0.35, P < 0.001).

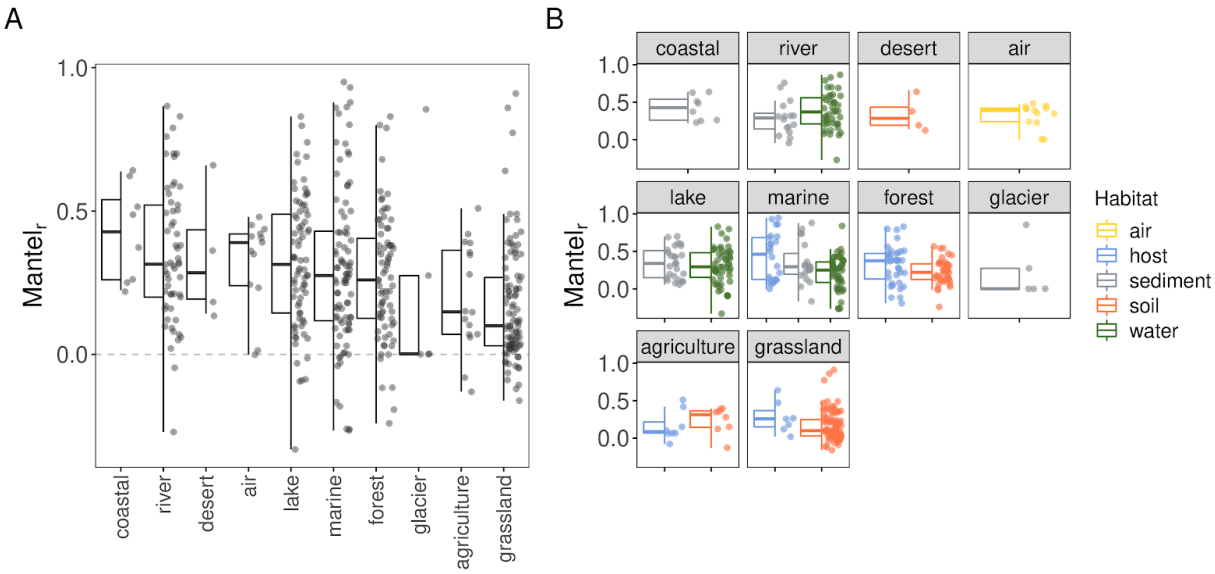


Figure 2. Variation in Mantel correlation coefficients (effect size) of distance-decay relationships between different environments (A) and habitat types (B). Environment categories are arranged from highest to lowest mean effect size.

Finally, we found a positive relationship between the (log) spatial extent and the effect size of microbial distance-decay relationships (Fig. 3; coef = 0.03, t = 4.66, R^2 = 0.05, P < 0.001), such that studies incorporating large spatial scales tend to have stronger distance-decay relationships. As larger spatial scale studies might also incorporate greater sampling coverage, we also tested for collinearity between the spatial scale of a study and the sampling coverage, but found no correlation between these variables (ρ = 0.06, P = 0.19).

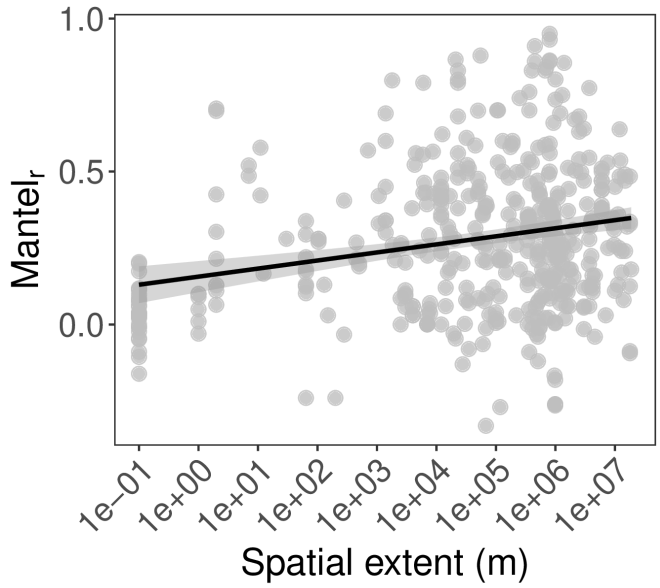


Figure 3. The relationship between spatial extent and the Mantel correlation coefficient of microbial distance-decay relationships. The best fit line represents the fit of a linear regression between the log of spatial extent and Mantel correlation coefficient, and the grey shaded region shows 95% confidence intervals.

Influence of Methodological Factors on the Distance-Decay Relationship

To determine whether the microbial distance-decay relationship may be influenced by methodological factors, we tested for relationships between the method of community characterisation, sampling depth, or choice of community similarity index and the effect size of microbial distance-decay relationships. We grouped community characterisation methods according to their ability to distinguish between closely related taxa. There were no clear differences in the distance-decay effect sizes between methods of differing resolutions ($F_{2, 449}$ = 0.562, P = 0.57), nor were there clear differences between different molecular methods (Fig. S2, $F_{7, 437}$ = 1.97, P = 0.06), considering only those methods that had >4 distance-decay relationships (excluding Ion Torrent; n = 4, phylo-chip; n = 2, and Pac-Bio; n = 1).

Whilst we observed no differences in distance-decay relationships between different resolution methods, we observed a positive relationship between ($\log_{10}$) community coverage and the strength of microbial distance-decay relationships (Fig. 4A; n = 337, coef = 0.04, t = 2.39, P < 0.01). However, this relationship was weak (R^2 = 0.01), and when two distance-decay relationships with extremely high community coverage were removed, the slope was indistinguishable from 0 (coef = 0.03, t = 1.78, R^2 = 0.01, P = 0.08).

The logistics of multiplexing samples on high-throughput sequencing runs means that there is often a trade-off between the community coverage and sampling coverage of a study. However, we found no evidence of negative correlation between these two factors (Pearson's ρ = -0.03, P = 0.54). Neither did we detect any clear relationship between the number of samples ($\log_{10}$ sample coverage) and the distance-decay effect size (Fig. 4B; n = 451, coef = -0.04, t = -1.03, P = 0.30).

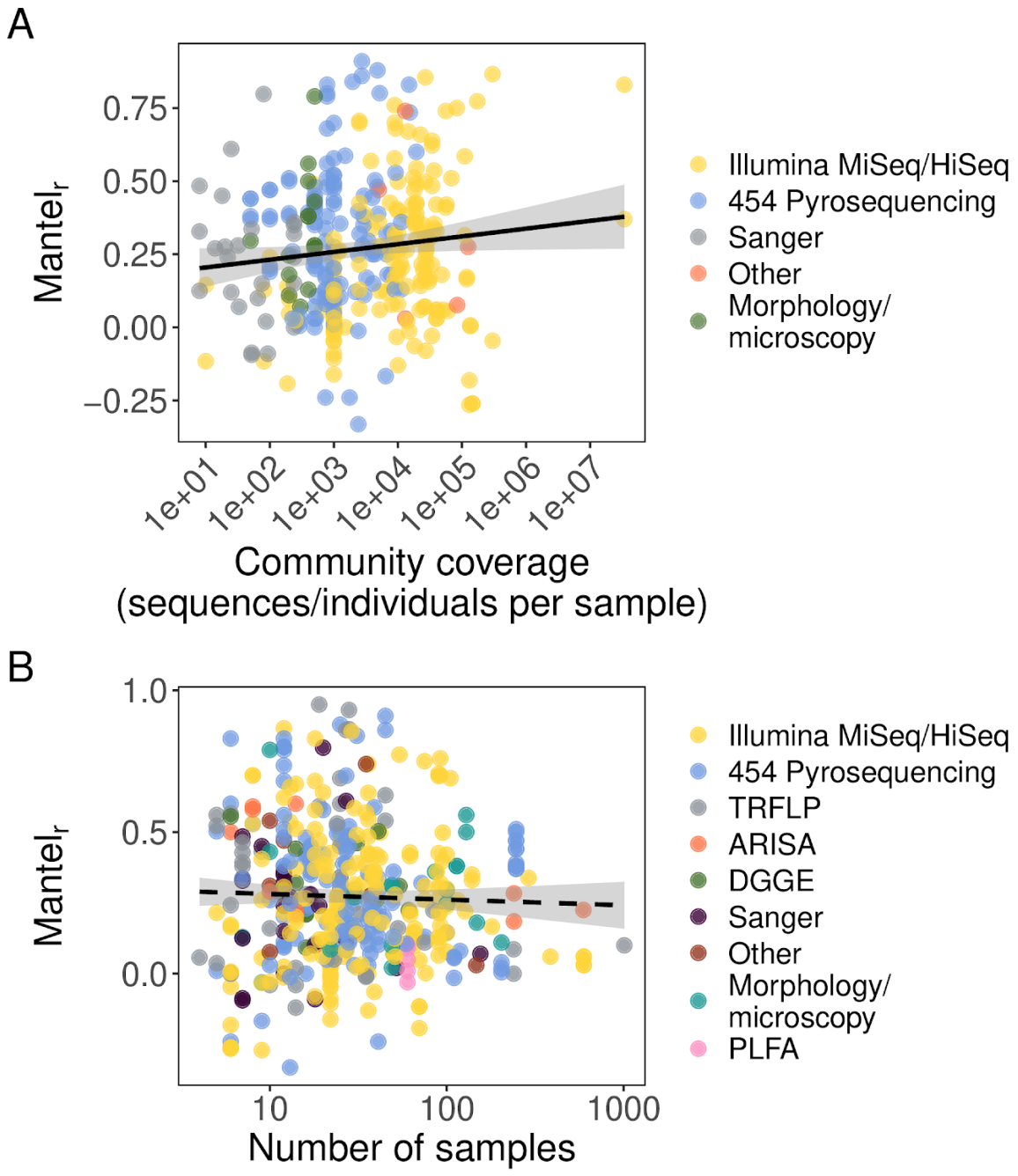
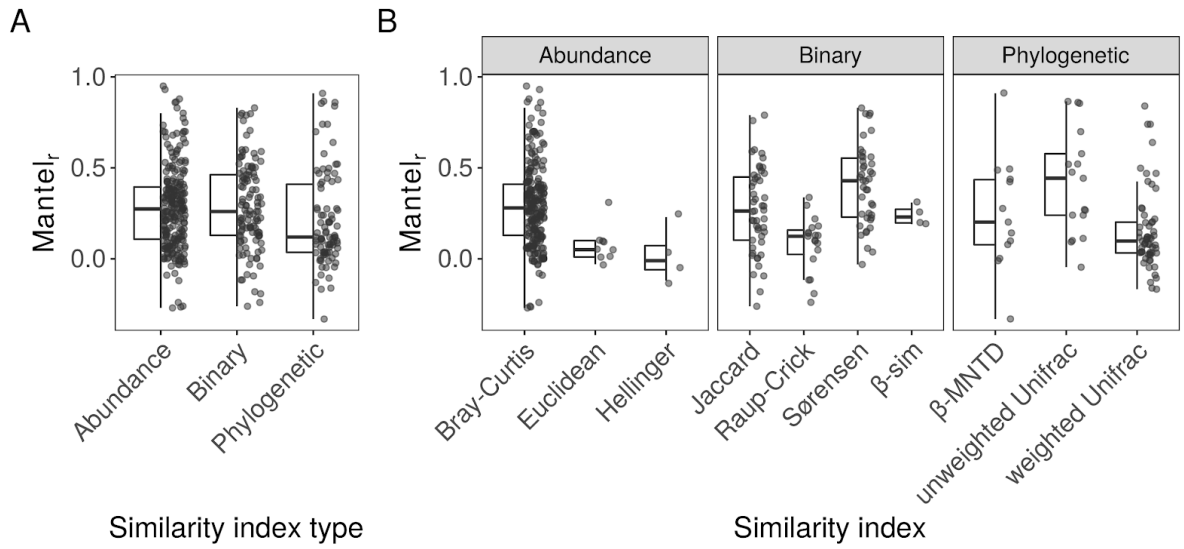


Figure 4. The relationship between the effect size of microbial distance-decay relationships (Mantel correlation coefficient) and A) community coverage, quantified as the number of sequences or individuals counted per sample, and B) sample coverage, quantified as the number of individual samples used to construct distance-decay relationships. Points are individual Mantel correlation coefficients, coloured by the molecular technique used in

characterising the microbial community. Solid lines indicate statistically significant relationships ($P < 0.05$), whilst dashed lines indicate non-significant relationships ($P > 0.05$), and grey ribbons represent 95% confidence intervals. Abbreviated molecular methods in the legend are defined as follows (TRFLP = Terminal Restriction Fragment Length Polymorphism; ARISA = Automated Ribosomal Intergenic Spacer Analysis; DGGE = Denaturing Gradient Gel Electrophoresis; PLFA = Phospholipid Fatty Acid analysis).

Choice of similarity index also had a clear impact on the effect size of microbial distance-decay relationships. As well as recording the specific similarity index used, we categorised these indices into types (binary, abundance, or phylogenetic) to look for broad differences in distance-decay relationships between them. We analysed the nested interaction between similarity index and index type, and found no clear differences between different index types (Fig. 5A; $F_{2, 424} = 1.48$, P = 0.23). However, the interaction between index type and similarity index was significant ($F_{7, 424} = 7.20$, P 0.001). Post-hoc analysis revealed differences between similarity indices within and between index types (Fig. 5B). Distance-decay effect sizes based on the Raup-Crick index were weaker than those based on either Sørensen (coef = -0.38, P < 0.01) or unweighted Unifrac indices (coef = -0.44, P < 0.01), whilst those based on weighted Unifrac were weaker than both Sørensen (coef = -0.29, P < 0.001) and unweighted Unifrac (coef = -0.35 P < 0.05).



309 Figure 5. Variation in effect sizes of microbial distance-decay relationships calculated with
310 different similarity index types (A), or individual indices (B). Only indices with four or more
311 effect sizes were plotted for clarity.

312 *Comparison of Contextual and Methodological Variables*

313 In order to determine whether eco-environmental context or methodological factors better
314 explain the effect size of microbial distance decay relationship, we specified two models,
315 with variables from these two categories, using a subset of the original data for which values
316 were obtained for all variables (n = 323). The two models each had four variables, and used
317 similar degrees of freedom (context model df = 26, methodological model df = 27). The
318 methodological model outperformed the contextual model in terms of both AIC (Akaike
319 Information Criterion) and R^2 measures of model performance (Table 2). Notably, neither
320 model explained a high proportion of the variance, although both AIC and likelihood ratio
321 tests supported both models over a null (intercept only) model.

Table 2. Comparison of models specified using either contextual, or methodological variables. Akaike Information Criterion (AIC) and adjusted R^2 quantify the likelihood and fit of a model relative to the number of predictor variables, respectively.

Model	AIC	Adj-R^2	Likelihood ratio comparison to null (intercept only) model			
			ΔAIC	Sum of squares	F (df)	*P* value
Contextual	146.89	0.11	-13.69	5.34	2.61	< 0.001
Methodological	134.11	0.14	-26.46	6.47	3.17 (25)	< 0.001

Discussion

Previous research into the spatial ecology of microbial communities has not yielded a consistent distance-decay relationship. Our meta-analysis of 452 microbial distance-decay relationships suggests that the reasons for this lack of consistency are two-fold. Firstly, the differing contexts within which studies are conducted contribute variability to reported distance-decay relationships. In particular, we found that differing study systems and spatial extents were associated with variation in microbial distance-decay relationships. Secondly, methodological differences between studies, including dissimilarity index, data resolution, and sample coverage, all significantly affected observed distance-decay relationships. A central tenet of macroecology is the search for universal patterns and relationships; our results suggest generalisable relationships may only emerge when methodological approaches are appropriately coupled to ecological context.

Our comparison of distance-decay relationships between different study systems revealed that agricultural and especially grassland-based studies had weaker relationships than studies of other environments. Within these environments, soils were by far the most

frequently studied habitat, and we initially expected that, as soils maintain strong physicochemical gradients over small vertical and horizontal spatial scales (e.g. Dumbrell *et al.*, 2010), that these distance-decay relationships would be stronger than in other environments or habitats. It is possible that the environmental gradients present in soils do not change linearly over geographic distance, for example if similar environmental conditions are patchily distributed. Alternatively, soil microorganisms may be able to disperse more effectively than previously thought, perhaps via association with other soil organisms (e.g. bacterial migration along fungal hyphae; Warmink *et al.*, 2011), migratory species such as birds (Bisson *et al.*, 2007), wind blown soil particles (Favet *et al.*, 2013), or via bioaerosols (Joung *et al.*, 2017). The depth profile over which soil samples integrate may also play a role in obscuring distance-decay relationships, as surface soils show stronger distance-decay relationships than deeper ones, likely due to the greater intensity of dispersing propagules entering the surface (Li *et al.*, 2020). Furthermore, soils harbour extensive microbial “seed banks” of dormant organisms and/or relic DNA that could weaken the distance-decay relationship (Lennon & Jones, 2011; Carini *et al.*, 2016; Lennon *et al.*, 2018). Dormant cells and relic DNA are not subject to environmental selection yet, are routinely detected in molecular community assays, and thus may diminish the perceived effects of spatially-structured environmental selection on microbial communities (Locey *et al.*, 2019). Thus, in habitats such as soils, distinguishing dormant from active cells could result in stronger distance-decay relationships than those recorded previously, though the extent to which this phenomenon plays a role in other environments is less clear .

Originally, we expected that studies of aquatic microbial communities may show the weakest distance-decay relationships as riverine or oceanic hydrology may provide an effective dispersal mechanism, homogenising microbial communities over larger spatial and environmental gradients over larger spatial scales. In contrast, we found that aquatic communities actually showed stronger distance-decay relationships. Soininen *et al.* (2007)

recorded similar distance-decay rates between terrestrial, marine and aquatic ecosystems, showing that context-dependent distance-decay relationships may be a feature of microbial communities. Host-associated communities showed relatively strong, but variable distance-decay relationships. We suggest that this is caused jointly by the ecology of the host species, in combination with the degree of host-specificity with the associated microbial community. For example, if the host is not dispersal limited, and associates with a large variety of microorganisms, then the distance-decay relationship may be relatively weaker than those of either dispersal limited hosts, or highly specific associated microbiomes.

Scale-dependent relationships have been reported previously (Bissett *et al.*, 2010; Martiny *et al.*, 2011; Soininen *et al.*, 2011), albeit with contrasting results. Our results are comparable to those of Soininen *et al.* (2011), who reported that distance-decay relationships of various microbial communities were generally steeper as greater spatial scales were incorporated. The scale-dependence of this relationship may be explained by greater environmental heterogeneity in large-scale studies, thus communities are subjected to different environmental filters, resulting in more dissimilar communities. In combination with this, communities separated by very large geographic distances should have minimal dispersal between them, assuming connectivity is negatively related to geographic distance. Alternatively, this result may be a statistical artefact, caused by studies with large spatial extents incorporating many zero similarity community comparisons (i.e. communities with no species in common), therefore biasing measured distance-decay relationships (Millar *et al.*, 2011; Steinbauer *et al.*, 2012).

Distance-decay relationships are frequently interpreted as evidence for neutral community assembly processes such as dispersal limitation, in the microbial literature. Across microbial taxa, cell size is a trait thought to influence dispersal efficacy (Wilkinson, 2001; Wilkinson *et al.*, 2012; Zinger *et al.*, 2019), and so larger microorganisms such as micro-Eukarya should

show stronger distance-decay relationships than smaller microorganisms such as Bacteria or Archaea. However, we found no evidence for this, suggesting that phylogenetically structured traits such as cell size may be less important compared to other contextual and methodological factors, or that the broad domain-level classification used here does not sufficiently capture different microbial cell sizes. As discussed previously, distance-decay relationships can arise from spatially autocorrelated environmental gradients as well as dispersal limitation (Nekola & White, 1999). Therefore, the lack of differences in biogeographic patterns observed at the domain level may be the result of a trade-off between dispersal-related processes and environmental filtering. For instance, bacterial distance-decay relationships may be less strongly influenced by dispersal than environmental filtering, and vice versa for Eukarya. Consequently, these influences may balance out at broad taxonomic levels, resulting in similar biogeographic patterns at the domain level.

In comparison to contextual factors, methodological factors were found to have a greater influence on microbial distance-decay relationships. The development of molecular methods, including high-throughput sequencing platforms, has vastly improved our ability to characterise microbial communities (Roesch *et al.*, 2007; Caporaso *et al.*, 2012). However, these methods differ in their resolution, community coverage, and ability to multiplex large numbers of samples, all of which we hypothesised could strengthen or weaken distance-decay relationships by altering our estimation of microbial β-diversity. In contrast, we observed only a weak relationship between distance-decay effect sizes and community coverage, and no clear relationships with different resolution methods, or with the number of samples, suggesting that molecular methodology may not play as large a role in determining microbial biogeographic patterns as previously thought.

The ability to resolve closely related taxa has previously been found to be an important determinant of our ability to detect biogeographical patterns, as such patterns may only emerge when taxa are defined at sufficiently high resolution (Hanson *et al.*, 2012). Yet, other studies show that bioinformatically altering taxonomic resolution frequently has little effect on microbial biogeographic patterns. For example, increasing the similarity threshold at which operational taxonomic units are defined is thought to be equivalent to increasing the taxonomic resolution (Callahan *et al.*, 2017). Yet, empirical biogeographic relationships often appear robust to such manipulation, in a variety of taxa and ecosystems (Clark *et al.*, 2017; Glassman & Martiny, 2018; Meyer *et al.*, 2018), supporting our finding that resolution may not be important. Perhaps most molecular methodologies operate above resolutions at which biogeographic patterns begin to change, or more worryingly, perhaps we are still studying microbial biogeography at too low a resolution.

Aside from resolution, another important variable related to molecular methodology is community coverage. One of the few universal patterns that appears to hold true for most microbial communities is the “long-tailed” species abundance-distributions (Dumbrell *et al.*, 2010; Shoemaker *et al.*, 2017; Maček *et al.*, 2019), which is caused by the majority of microorganisms in a community being rare. The rarer taxa in microbial communities also tend to be the least widespread (Clark *et al.*, 2017; Lindh *et al.*, 2017; Meyer *et al.*, 2018; Shade & Stopnisek, 2019) and thus, only detecting the more abundant, widespread organisms would overestimate compositional similarity across communities, and consequently, weaken distance-decay relationships due to the lower rate of turnover (Meyer *et al.*, 2018). Perhaps of more concern is that even with existing sequencing platforms, our surveys of environmental microbial communities still miss taxa that are vanishingly rare in the environment, such as extremophiles that persist in non-extreme habitats (Low-Décarie *et al.*, 2016). The ability of common species to reflect ecological patterns of the wider

community is debated (Galand *et al.*, 2009; Heino & Soininen, 2010; van Dorst *et al.*, 2014) and is linked to a wider debate on the ecological importance of rare species that is far beyond the scope of this work (e.g. Gaston, 2012). However, rare microorganisms are well known to be of critical importance in the context of environmental perturbations (Shade *et al.*, 2014; Low-Décarie *et al.*, 2016) and in providing ecosystem processes (e.g. sulfate-reduction in peat soils, Hausmann *et al.*, 2016; and anaerobic ammonia-oxidation in river sediments Lansdown *et al.*, 2016) and as a result, ignoring them may further distance biogeographic patterns from ecosystem-level processes.

Against expectation, we observed no clear differences in distance-decay relationships using different similarity metric types, and differences between specific metrics were minimal. Distance-decay relationships based on the weighted Unifrac distance were weaker than those based on other metrics. Phylogenetic metrics, such as Unifrac, cluster communities at a lower resolution, as two communities can be closely genetically related, yet distinct at fine taxonomic resolutions (e.g. species or strain-level). For example, Bryant *et al.* (2008) found that Unifrac similarity was approximately three times higher than the compositional similarity of the same set of bacterial communities. Further, phylogenetic metrics may be inappropriate in less phylogenetically diverse environments (e.g. extreme systems) where phylogenetic diversity can be largely constrained to one taxon (e.g. the Halobacteria in hypersaline environments), leaving few “phylogenetic degrees of freedom” left to separate communities (Fukuyama, 2019). However, this does not account for the observed difference between weighted and unweighted versions of the Unifrac index, the former of which accounts for species’ relative abundance data, whilst the latter is binary (presence/absence based). A criticism of the weighted Unifrac index is that too much weight is placed on abundant taxa (Chen *et al.*, 2012). As abundant species are generally more widespread, placing too much weight on abundant taxa would have the effect of making communities appear artificially similar, exacerbating the effects of using a phylogenetic metric. As we observed no

difference between binary and abundance-based compositional indices, the differences observed with weighted Unifrac appear to be the result of combining phylogenetic and weighted indices. We therefore suggest that weighted phylogenetic metrics may underestimate microbial biogeographic patterns, unless appropriate weight is given to rare and abundant taxa (Chen *et al.*, 2012).

Our analysis of 452 microbial distance-decay relationships also revealed the overwhelming preference of microbial ecologists to use classic dissimilarity indices such as the Bray-Curtis (n = 218), Jaccard (n = 49), Sørensen (n = 42) indices. These choices no doubt reflect a wider trend in ecology as a whole, however, it is pertinent to draw attention to more recently developed metrics that may be more appropriate given the properties of microbial datasets and the hypotheses being tested. Biotic interactions are drivers of microbial β-diversity (Hanson *et al.*, 2012), yet classic dissimilarity metrics do not account for co-occurrence information in communities. To this end, a new family of metrics described by Schmidt *et al.*, (2017) include information on the average interactions of the taxa present, thus providing a novel approach to integrating co-occurrence data into distance-decay relationships. One problematic characteristic of high-throughput sequence datasets is the non-biological variance of sample sizes, which can result in statistical artefacts that confound biogeographic relationships (Baselga, 2007). Here, modifications made to some classic indices by Chao *et al.* (2005) reduce the sensitivity of these indices to variable sample sizes by accounting for unobserved species, thus reducing the need for post-sequencing normalisation of sample sizes (McMurdie & Holmes, 2014). Furthermore, “fuzzy logic”-based similarity indices are able to reduce the impact of false-absences or -presences on estimates of β-diversity, which is beneficial for microbial ecology studies where rarefaction may induce false-absences, and taxonomic assignment errors or contamination may lead to false-presences. Finally, many similarity metrics have been shown to merge compositional turnover (replacement of species) and nestedness (whereby communities are subsets of one

another), thereby blurring the contribution of distinct ecological processes to total community (dis)similarity. To combat this, modified versions of classic indices such as Jaccard, Sorensen, and Bray-Curtis have been developed, allowing the partitioning of community similarity metrics into their turnover and nestedness components (Baselga, 2010; Podani & Schmera, 2011). We echo the call of Green and Bohannan (2006) for microbial ecologists to exercise more care in their choice of dissimilarity metrics, especially as many of these new metrics are implemented in popular and freely accessible software, such as R (e.g. Baselga and Orme, 2012).

Overall, our analyses revealed that methodological factors explain more variation in microbial distance-decay relationships than ecological context, but that both sets of factors alter our perception of this biogeographic pattern. Given the importance of methodological factors in determining the strength of microbial biogeographic patterns, it is intuitive to recommend standardising approaches across studies in order to minimise the statistical signals associated with methodological variance. However, our results show variance due to differing ecological contexts would still hinder drawing generalisable relationships across studies. Instead, we suggest that tailoring methodological choices towards specific ecological contexts may enhance generalisable relationships in microbial ecology. For instance, in searching for consistent relationships between ocean waters and terrestrial soils, it would be unrealistic to sample both at the same spatial grain and extent, as the heterogeneity in the physicochemical environment, and dispersal processes of their microbial communities, are fundamentally different. Similarly, we should not necessarily expect the relationships between soils and river sediments to be comparable, as microorganisms in soils can feasibly disperse in any direction, whereas in rivers or streams dispersal would be largely constrained by flow direction. Consequently, tailoring methodological approaches, such as the sampling design and/or (geographical) distance measure, to better reflect the environmental heterogeneity and dispersal dynamics between

contrasting ecological contexts may enable us to negotiate the hierarchy of interacting factors that obscure macroecological patterns in microbial communities.

Conclusions

Our meta-analysis of >450 microbial distance-decay relationships revealed that factors related to the eco-environmental context within which a study was conducted, as well as the methodology of the study, jointly influence quantification of this classic biogeographic pattern. Against expectation, factors related to molecular methodology had relatively little effect on distance-decay relationships, whilst choice of dissimilarity metric was more important, highlighting that even after using robust, modern molecular methods, analytical choices have the power to obscure or enhance biogeographic patterns. Whilst we were able to detect clear relationships between microbial distance-decay relationships and various contextual and methodological variables, combining these variables explained only a modest amount of variation in our dataset. This lack of explanatory power highlights the fact that microbial biogeographic patterns may depend on a great number of contextual variables beyond those analysed here, and that understanding the environmental, or methodological, factors that drive this context-dependence may enable us to unify the seemingly disparate patterns observed by microbial ecologists over the past few decades.

Data Availability Statement

Full raw data analysed in this manuscript are provided in Table S1. Full raw data and code used in this manuscript will be uploaded to the Dryad data repository upon acceptance of this article.

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