

Bound within Boundaries: How Well Do Protected Areas Match Movement Corridors of Their Most Mobile Protected Species?

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Abstract

1. Conserving and managing large portions of land to connect wildlife reserves is an increasingly used strategy to maintain and restore connectivity among wildlife populations. Boundaries of such conservation areas are often determined based on expert opinion and socio-political constraints, yet the extent to which they match species' movement corridors is rarely examined. This is mainly due to a lack of data, particularly on wide-ranging movement behavior such as dispersal. Nevertheless, empirically assessing the adequacy of protected areas is key for the implementation of targeted management actions and efficient use of limited conservation funds.
2. Between 2011 and 2019, we collected high-resolution GPS movement data on 16 dispersing African wild dog (*Lycaon pictus*) coalitions from a free-ranging population in the Kavango-Zambezi Transfrontier Conservation Area (KAZA-TFCA). Spanning five countries and 520'000 km² the KAZA-TFCA is the world's largest transboundary conservation area and a prime example for international conservation efforts. We used integrated step selection analysis to estimate habitat selection of dispersers and to create a permeability surface for the entire KAZA-TFCA. We compared landscape permeability across different regions within the KAZA-TFCA as well as outside its boundaries. Lastly, we calculated least-cost paths and corridors to verify that major movement routes were adequately encompassed within the KAZA-TFCA.
3. Permeability within the boundaries of the KAZA-TFCA was more than double compared to areas outside it. Furthermore, we observed a five-fold permeability difference among the five KAZA-TFCA countries. We also showed that major movement corridors of wild dogs ran within the KAZA-TFCA, although some minor routes remained outside formally protected areas.
4. Differences in permeability were mainly related to different degrees of human activities across regions, and to the presence or absence of rivers, swamps or open water. The relation between permeability and other landscape features was less evident.
5. *Synthesis and Applications:* In this study, we showed how pertinent dispersal data of a highly mobile species can be used to empirically evaluate the adequacy of already-existing or planned protected areas. Furthermore, we observed regional differences in landscape permeability that highlight the need for a coordinated effort towards maintaining or restoring connectivity, especially where transboundary dispersal occurs.

1 Introduction

2 Connectivity among subpopulations is a crucial pre-requisite for many species to thrive
3 and persist (Fahrig, 2003). Accordingly, preserving and protecting movement corridors be-
4 tween wildlife reserves has become a task of utmost importance (Doerr et al., 2011; Rudnick
5 et al., 2012), resulting in an ever-growing number of large and often transboundary pro-
6 tected areas. While boundaries of such areas are often drawn according to expert opinion
7 and socio-political needs, subjective assessments have revealed deficiencies in the past (Cle-
8 venger et al., 2002; Pullinger and Johnson, 2010). Thus, an empirical evaluation of the
9 adequacy of already-existing or planned protected areas using pertinent animal movement
10 data is paramount for targeted use of valuable and scarce conservation funds (Pullinger and
11 Johnson, 2010).

12 In recent years, a growing body of research has used animal relocation data to iden-
13 tify movement corridors and assess connectivity at large scales (e.g. Chetkiewicz et al.,
14 2006; Squires et al., 2013; Elliot et al., 2014). Identification of potential movement corri-
15 dors typically relies on the estimation of permeability surfaces, which return the ease or
16 willingness at which the focal species traverses a specific landscape (Sawyer et al., 2011).
17 Such surfaces are created based on species' relative selection strengths (Avgar et al., 2017),
18 which can be quantified using a suite of selection functions (Zeller et al., 2012). Specifically,
19 selection strengths are estimated by comparing spatial covariates (e.g. environmental and
20 anthropogenic) at locations visited by the animal to the same spatial covariates at locations
21 available to the animal (Zeller et al., 2012). Importantly, selection functions require adequate
22 landscape and relocation data that are representative of the process being studied (Diniz
23 et al., 2019). Altough selection during residence and dispersal may coincide (Fattebert et al.,
24 2015), it appears that relocation data collected on dispersing individuals outperforms data
25 collected on resident individuals in the detection of large-scale movement corridors (Elliot
26 et al., 2014; Abrahms et al., 2017; Diniz et al., 2019). Nevertheless, dispersal data is inher-
27 ently difficult to collect and remains scarce in the connectivity literature (Vasudev et al.,
28 2015). As such, most permeability surfaces upon which movement corridors are identified
29 are created using relocation data collected on resident individuals. This has likely limited
30 our ability to meaningfully assess the effectiveness of protected areas in securing functional
31 connectivity for their protected species.

32 One initiative that aims at restoring and enhancing connectivity across large scales is the
33 Kavango-Zambezi Transfrontier Conservation Area (KAZA-TFCA), which constitutes the
34 world's largest transfrontier conservation area, spanning over 520'000 km² and five coun-

tries (www.kavangozambezi.org). While the KAZA-TFCA was originally set to facilitate movements of African elephants (*Loxodonta africana*; Tshipa, 2017), it is also key to the conservation of other wide-ranging species such as African wild dogs (*Lycaon pictus*; Woodroffe and Sillero-Zubiri, 2012; Cozzi et al., 2020), lions (*Panthera leo*; Elliot et al., 2014; Cushman et al., 2018), and cheetahs (*Acinonyx jubatus*; Weise et al., 2017). To date, however, few studies have attempted to assess the adequacy of the KAZA-TFCA using global positioning system (GPS) relocation data of its protected species at large spatial scales (Elliot et al., 2014; Tshipa, 2017; Brennan et al., 2020). Thus, how well the boundaries of the KAZA-TFCA reflect natural movement patterns and dispersal corridors of its most mobile protected species is virtually unknown.

Across the KAZA-TFCA, the African wild dog (*Lycaon pictus*) represents a highly mobile and endangered flagship species for conservation efforts. Once widespread across the entire Sub-Saharan continent, wild dogs have been widely extirpated through human persecution, habitat destruction, and disease outbreaks (Woodroffe and Sillero-Zubiri, 2012). As a result, the species has become one of Africa's most endangered large carnivores, and currently only survives in small, spatially scattered subpopulations (Woodroffe and Sillero-Zubiri, 2012). Within these subpopulations, wild dogs form cooperative breeding packs of up to thirty individuals (Creel and Creel, 2002), whose social structure is strongly governed by the process of dispersal (McNutt, 1996; Behr et al., 2020). Both males and females disperse from their natal pack, either alone or in same-sex dispersing coalitions, and search for unrelated mates and a suitable territory to settle (McNutt, 1996; Cozzi et al., 2020; Behr et al., 2020). During dispersal, wild dogs can cover several hundred kilometers and cross international borders (Masenga et al., 2016; Woodroffe et al., 2019; Cozzi et al., 2020). Despite the importance of dispersal for the long-term viability of this species, little empirical information is available on habitat selection and potential movement barriers during dispersal. The few studies that have collected dispersal data have shown that dispersers quickly move over large distances, avoid human-dominated landscapes and areas densely covered by trees, but prefer proximity to water (Masenga et al., 2016; Woodroffe et al., 2019; O'Neill et al., 2020; Cozzi et al., 2020).

Here, we collected and analyzed GPS relocation data on 16 dispersing wild dogs in as many dispersing coalitions from a free-ranging population in northern Botswana to assess the adequacy of the KAZA-TFCA in securing connectivity. We estimated relative selection strengths towards environmental and anthropogenic landscape features, and used the obtained coefficients to predict a permeability surface spanning the entire KAZA-TFCA.

69 We then investigated how landscape permeability varied regionally and internationally, and
70 we compared permeability within and outside the KAZA-TFCA boundaries. Finally, we
71 calculated least-cost paths and corridors to identify major movement routes and to verify
72 that these were successfully covered by the KAZA-TFCA.

73 **2 Methods**

74 **2.1 Study Area**

75 The study area (centered at -17°13'9"S, 23°56'4"E; Figure 1a) was outlined by a rectangu-
76 lar bounding box stretching over 1.3 Mio km² and encompassing the entire KAZA-TFCA
77 (Figure 1b). The KAZA-TFCA lies in the basins of the Okavango and Zambezi rivers and
78 includes parts of Angola, Botswana, Namibia, Zimbabwe, and Zambia. With a total area of
79 over 520'000 km² it constitutes the earth's largest transboundary conservation area and is
80 characterized by diverse landscapes, including savanna, grassland, and dry or moist wood-
81 land habitats. Rainfall in the study area is seasonal and lasts from November to March.
82 The KAZA-TFCA also comprises the Okavango Delta, which represents a highly dynamic
83 hydrological flood-pulsing system (McNutt, 1996; Wolski et al., 2017). The extent of the
84 flood in the delta greatly changes within and between years, depending on the amount
85 of rain that descends from the catchment areas in Angola and reaches the distal ends of
86 the delta between July and August (Figure S4). The flood drastically affects surrounding
87 landscapes, so that during maximum extent (ca. 12'000 km²) the delta becomes a patchy
88 conglomerate of swamps, open water, and islands, whereas these structures run dry when
89 the flood retracts to its minimum extent (ca. 5'000 km²; Wolski et al., 2017). Despite 36
90 national parks (NPs) and other protected areas, there is considerable human influence in
91 some regions of the KAZA-TFCA, mainly originating from farms, human density, and road
92 traffic.

93 **2.2 GPS Relocation Data**

94 We used a population of free-ranging African wild dogs inhabiting the Okavango Delta in
95 northern Botswana as a source population for dispersing individuals. This population has
96 been extensively studied since 1989 (McNutt, 1996; Cozzi et al., 2013, 2020; Behr et al.,
97 2020). Between 2011 and 2019, we systematically collected GPS relocation data on 16
98 coalitions of dispersing African wild dogs (7 female and 9 male coalitions). Candidate
99 dispersing individuals were identified based on age, number of same-sex siblings, pack size,

and presence of unrelated individuals of the opposite sex in their pack (McNutt, 1996; Behr et al., 2020). Selected individuals were immobilized according to protocols described in Osofsky et al. (1996), and fitted with GPS/Satellite radio collars (*Vertex Lite*; *Vectronic Aerospace GmbH, Berlin, Germany*) while still with their natal pack. Collars weighed 330g, accounting for about 1.5% of a wild dog's body weight. A 5cm long decomposable cotton piece was added to the collar belt to guarantee collar drop-off after about 12-18 months. All required procedures were undertaken and supervised by a Botswana-registered wildlife veterinarian. During dispersal, GPS collars were programmed to record a GPS relocation every 4 hours and to regularly transmit data via iridium satellite system to a base station.

Because we were interested in dispersal behavior only, we discarded any GPS data collected while individuals were still with their natal packs and after settlement in a new territory (Cozzi et al., 2020). We identified the exact time of emigration and settlement based on direct field observations and through visual inspection of the net squared displacement (NSD) metric. NSD quantifies the squared Euclidean distance of a relocation to a reference point (Börger and Fryxell, 2012), which in our case was the center of the dispersing coalition's natal home range. Thus, dispersal was deemed to have started when a coalition had left its natal home range and continued until the NSD metric remained stationary, indicating that the coalition had successfully settled (Figure S1). In our analysis, we did not differentiate between male and female dispersing coalitions, for previous research found little differences between sexes during dispersal (Woodroffe et al., 2019; Cozzi et al., 2020).

2.3 Spatial Covariates

To investigate relative selection strengths of dispersing wild dogs, we used a set of georeferenced covariates (Figure 2) that we aggregated in the categories *land cover*, *protection status*, and *anthropogenic*. *Land cover* comprised the covariates water cover (binary), distance to water (continuous), percentage cover by shrubs/grassland (continuous), and percentage cover by trees (continuous). To capture the pulsing behavior of the Okavango Delta, we classified satellite imagery and frequently updated layers for water cover and corresponding layers depicting distance to water. *Protection status* contained a binary covariate, indicating whether an area was protected or not. *Anthropogenic* included covariates rendering the presence of roads (binary), the distance to roads (continuous), and a proxy for human influence (continuous) that took into account human density, farming, and roads. We prepared all covariates as spatial raster layers from freely available online services and

133 from remotely sensed satellite imagery. To ensure a consistent resolution (i.e. cell-size or
134 grain) across covariates, we coarsened or interpolated all layers to a resolution of 250m x
135 250m. For further details on the preparation and source of each covariate, see Appendix
136 A.3. We performed processing and manipulation of data as well as all spatial and statistical
137 analyses using R, version 3.6.1 (R Core Team, 2019).

138 **2.4 Habitat Selection Model**

139 We used an integrated step selection function (iSSF; Avgar et al., 2016) to investigate
140 dispersers' relative selection strengths towards the above-mentioned spatial covariates. That
141 is, we paired each realized step (i.e. the connecting line between two consecutive GPS
142 relocations; Turchin, 1998) with 24 random steps that were generated by sampling turning
143 angles from a uniform distribution $U(-\pi, +\pi)$ and step lengths from a gamma distribution
144 fitted to realized steps (Avgar et al., 2016). A realized step and its 24 associated random
145 steps formed a stratum and received a unique identifier. Along each step, we extracted
146 the above-mentioned covariates (Table S3), standardized extracted values using a z-score
147 transformation, and checked for correlation using Pearson's Correlation Coefficient r . None
148 of the covariates were overly correlated ($|r| > 0.6$; Latham et al., 2011) and we retained
149 all of them for modeling. Our habitat selection model then assumed that dispersing wild
150 dogs assigned a selection score $w(x)$ of the following exponential form to each realized and
151 random step (Fortin et al., 2005):

$$w(x) = \exp(\beta_1 x_1 + \beta_2 x_2 + \dots + \beta_n x_n) \quad (\text{Equation 1})$$

152 The selection score $w(x)$ of a step depended on its associated covariates (x_1, x_2, \dots, x_n) , as
153 well as on the animal's relative selection strengths towards these covariates $(\beta_1, \beta_2, \dots, \beta_n)$.
154 To estimate relative selection strengths (i.e. the β 's) for each covariate, we used mixed effects
155 conditional logistic regression analysis as suggested by Muff et al. (2020). We implemented
156 their method using the R-package *glmmTMB* (Brooks et al., 2017) and used dispersing
157 coalition ID to model random slopes. We also modelled random intercepts with an arbitrary
158 high variance of 10^6 to make use of the poisson trick (see Muff et al., 2020). We defined three
159 movement metrics, namely the cosine of the turning angle ($\cos(ta)$), the step length (sl) and
160 the logarithm of the step length ($\log(sl)$), as core covariates and ran stepwise forward model
161 selection based on Akaike's Information Criterion (AIC; Burnham and Anderson, 2002) for
162 all other covariates. The inclusion of movement metrics served to reduce biases in estimated
163 habitat selection coefficients that may have arisen due to movement behavior (Avgar et al.,

164 2016). To validate the predictive power of the most parsimonious habitat selection model,
165 we ran k-fold cross-validation for case-control studies as described in Fortin et al. (2009)
166 (details in Appendix A.5).

167 2.5 Permeability Surface

168 Using the most parsimonious habitat selection model, we predicted a permeability surface
169 spanning the entire extent of the KAZA-TFCA. That is, we applied Equation 1 to our
170 spatial covariates and calculated the selection score $w(x)$ for each raster cell. Because our
171 representation of water was dynamic we collapsed all dynamic water maps into a single
172 map using areas that were covered by water in at least 10% of the cases. We used the
173 resulting map to also calculate a layer returning the distance to water. To reduce the
174 influence of outliers in predicted permeability scores, we followed Squires et al. (2013) and
175 curtailed predicted scores between the 1st and 99th percentile of their original values. To
176 compare permeability across different regions, we normalized the permeability surface to a
177 range between 0 (most impermeable) and 1 (most permeable), and we determined median
178 permeability within and outside the KAZA-TFCA, within and outside formally protected
179 areas, and within each of the five KAZA-TFCA countries.

180 2.6 Least-Cost Paths and Corridors

181 To identify movement corridors of dispersing wild dogs, we specified source points and
182 calculated factorial least-cost paths (LCPs) as well as factorial least-cost corridors (LCCs)
183 among them (Elliot et al., 2014). We generated source points by overlaying the study area
184 with a regular grid of points spaced at 100 km. We only considered those points that
185 fell within protected areas $> 700 \text{ km}^2$, which conforms with home-range requirements of
186 African wild dogs (Pomilia et al., 2015). Finally, we defined centroids as source points
187 for those protected areas $> 700 \text{ km}^2$ that were not assigned any source points from the
188 regular grid. Because wild dogs residing outside of protected areas are rare and unviable
189 (Van der Meer et al., 2014), we consider our selection of source points within protected areas
190 to be appropriate. In total, we generated 68 source points, which resulted in 2'278 unique
191 pairwise combinations and therefore 2'278 unique LCPs and LCCs. We computed factorial
192 LCPs and LCCs between source points using the R-package *gdistance* (further details in
193 Appendix A.7). After computation, we tallied overlapping LCPs and LCCs, respectively,
194 into single connectivity maps. Because LCPs return discrete paths, whereas LCCs return
195 continuous corridors, we present both methods.

196 **3 Results**

197 **3.1 Dispersal Events**

198 In total, we collected 4'169 GPS relocations during dispersal (Figure S2 & Table S1), re-
199 sulting in an average of 261 ($SD = 207$) locations per dispersing coalition. Coalitions on
200 average dispersed for 48 days ($SD = 44$), covered a mean Euclidean distance of 54 km (SD
201 = 71) and a cumulative distance of 597 km ($SD = 508$).

202 **3.2 Habitat Selection Model**

203 Our most parsimonious habitat selection model ($\Delta AIC > 2$ than any alternative model;
204 Table S4) retained the covariates *water*, *distance to water*, *trees*, *shrubs/grassland*, and
205 *human influence*, beside the fixed covariates $\cos(ta)$, *sl*, and $\log(sl)$ (Figure 3a). With respect
206 to environmental covariates, dispersers avoided moving through water ($\beta = -0.53$, 95%
207 CI -0.79 to -0.27) but selected for locations in its vicinity, although the latter effect was
208 not significant ($\beta = -0.33$, 95% CI = -0.73 to 0.08). Dispersers avoided areas that were
209 densely covered by trees ($\beta = -0.31$, CI = -0.47 to -0.15) and preferred areas covered by
210 shrubs/grassland ($\beta = 0.25$, 95% CI = 0.07 to 0.42). Finally, dispersers avoided areas that
211 were influenced by humans ($\beta = -0.45$, 95% CI = -0.82 to -0.08). With the exception
212 of *distance to water*, random effects revealed little variability between dispersal coalitions
213 (Appendix A6).

214 Results from the k-fold cross-validation suggested that our prediction was significant and
215 robust, as highlighted by the fact that the 95%-CIs intervals of $\bar{r}_{s,realized}$ and $\bar{r}_{s,random}$ did
216 not overlap (Figure 3b). Likewise, the significant correlation between ranks and correspond-
217 ing frequencies for realized steps suggested a good fit between predictions and observations
218 (Figure 3b).

219 **3.3 Permeability Surface**

220 Our prediction of landscape permeability revealed substantial differences across regions in
221 the study area (Figure 4). Comparisons of median permeability values (Table 1) showed
222 that permeability inside the KAZA-TFCA was more than two times as high as permeability
223 outside it. Permeability varied by country, with a five-fold permeability difference among
224 them. Angola and Botswana were characterized by comparably highly permeable landscapes,
225 Zimbabwe and Zambia were relatively impermeable, and Namibia ranged in between the two
226 extremes (Table 1). Visual inspection of our covariate layers indicated that high permeability

in Angola and Botswana was mainly related to a combination of low human influence, low tree cover, high shrubs/grassland cover, and a close distance to water. Although swamps, wetlands, and permanent water themselves provided little permeability, their surroundings acted as strong attractants to dispersers. The low permeability that characterized Zambia and Zimbabwe, on the other hand, was mainly caused by substantial human influence. Albeit the KAZA-TFCA covered most permeability hot-spots, several highly permeable regions remained uncovered by its borders. Across all countries, protected areas provided roughly double the permeability of unprotected landscapes (Table 1).

3.4 Least-Cost Paths & Least-Cost Corridors

Our least-cost analysis revealed three major movement corridors of which all were well-contained within the KAZA-TFCA boundaries (Figure 5). One major corridor ran SE-NW and connected the Okavango-Linyanti ecosystem in Botswana with Luengue-Luiana NP in Angola. A second corridor ran W-E between Chobe NP in Botswana and Zimbabwe's Hwange NP. A third major corridor ran NE-SW, completely across unprotected areas, and connected Kafue NP in Zambia with more central regions of the KAZA-TFCA. Several minor corridors branched off from these three major corridors; these included a southward connection between the Okavango-Linyanti and the Central Kalahari Game Reserve, a southwesterly corridor connecting Luengue-Luiana NP with Namibia's Khaudum NP, and a northeasterly extension of the Hwange corridor into Zimbabwe's Matusadona NP. According to our predictions, the landscapes in the Okavango-Linyanti region were the highest frequented dispersal routes within the KAZA-TFCA (Figure 5b). Our model did not detect any significant direct corridors between Zimbabwe and Zambia or Zambia and Angola, and only a very limited W-E direct connection between the Okavango region and Namibia's Khaudum NP. Except for the corridor into the Central Kalahari National Park, our model did not detect any significant connectivity outside the boundaries of the KAZA-TFCA. Furthermore, we found little to no direct connectivity between peripheral points; that is, most paths and corridors connecting two adjacent peripheral points ran through more central regions before heading towards their destination at the periphery (Figure 5).

4 Discussion

We used GPS relocation data collected on dispersing African wild dogs to investigate whether their main movement corridors are contained within the boundaries of the world's largest transboundary conservation area, namely the KAZA-TFCA. Our analysis suggests that the

259 KAZA-TFCA indeed encompasses all major corridors of African wild dogs, demonstrating
260 the potential value of such an initiative. We thus exemplified how pertinent dispersal data
261 of a highly mobile species can be used to assess the adequacy of already existing or planned
262 protected areas. Our approach is neither limited to the African wild dog, nor to our study
263 area and thus applicable to any study system. All covariates used throughout this study
264 are readily available on a global scale and many of them are likely to be important deter-
265 minants of movement behavior, landscape permeability, and connectivity for other species
266 (Zeller et al., 2012; Thurfjell et al., 2014). Interestingly, our predicted network of least
267 cost-paths and corridors for African wild dogs shows surprising similarities to corridors of
268 dispersing lions inhabiting the same ecosystem (Elliot et al., 2014; Cushman et al., 2018).
269 This not only reinforces confidence in our own predictions but also suggests potential syner-
270 gies for the conservation of these two, and possibly more, species. Expanding our analytical
271 framework to additional species will likely yield important insights on the consistency of
272 inter-specific movement corridors, thus highlighting areas that are exceptionally valuable for
273 the conservation of several species (e.g. Brennan et al., 2020).

274 Our results emphasize that human influences constitute some of the main barriers to
275 connectivity among wild dog populations. This conforms to findings on dispersing wild
276 dogs from eastern Africa (Masenga et al., 2016; O'Neill et al., 2020) but conflicts with
277 findings from South Africa by Davies-Mostert et al. (2012), who reported a high willingness
278 of dispersers to cross human-dominated landscapes. Such differences may arise from the fact
279 that our model infers preferences by comparing *used* and *available* habitats, whereas Davies-
280 Mostert et al. (2012) only recorded net dispersal distances, thereby precluding such an
281 analysis. Thus, we believe that differences to Davies-Mostert et al. (2012) may be explained
282 by the unavailability of alternative routes through natural landscapes, which may have forced
283 dispersers in South Africa to cross human dominated landscapes despite a strong aversion to
284 do so. In this regard, our representation of dispersal corridors and the resulting connectivity
285 appear conservative, as dispersers may be able to make the best out of a bad situation and
286 cross landscapes characterized by considerably unfavorable conditions (Palomares et al.,
287 2000; Elliot et al., 2014). Nevertheless, successful conservation of this species relies on
288 policymakers' and local authorities' willingness and ability to provide and conserve natural
289 areas that remain free from anthropogenic pressures. This is not only paramount in light
290 of increasing connectivity and facilitating dispersal, but also in terms of reducing human-
291 caused mortality during dispersal. In fact, previous studies have shown that human-caused
292 mortality represents a major threat to wild dogs' ability to disperse (Woodroffe et al., 2019;

²⁹³ Cozzi et al., 2020).

²⁹⁴ Besides human influence, we identified water as additional obstacle to dispersal. This
²⁹⁵ corroborates earlier studies showing that water bodies are almost impenetrable to resident
²⁹⁶ packs (Abrahms et al., 2017) and only infrequently crossed by dispersing individuals (Cozzi
²⁹⁷ et al., 2020). An accurate and dynamic representation of water is thus imperative and
²⁹⁸ particularly relevant in seasonal or flood-pulsing ecosystems such as the Okavango Delta.

²⁹⁹ Although dispersers avoided moving through water, they selected locations in its vicinity.
³⁰⁰ This behavior may be caused by the occurrence of prey close to water (Bonyongo, 2005). For
³⁰¹ the same reason, however, competitors such as lions, spotted hyenas, and resident wild
³⁰² dogs may also use areas close to water (Valeix et al., 2010), thereby occasionally forcing
³⁰³ dispersing wild dogs to switch behavior and move into prey-poorer areas away from water
³⁰⁴ (Creel and Creel, 2002; Mills and Gorman, 1997). This reasoning may also explain the
³⁰⁵ large confidence intervals surrounding the corresponding β -estimate. Given the influence
³⁰⁶ that resident conspecifics, competitors, and prey can have on dispersers (Cozzi et al., 2018;
³⁰⁷ Armansin et al., 2019) future studies should strive to collect and incorporate intra- and
³⁰⁸ interspecific relationships into analyses of landscape connectivity.

³⁰⁹ Overall, our findings on habitat selection during dispersal coincide with findings from
³¹⁰ dispersing wild dogs in Kenya (O'Neill et al., 2020), suggesting that there are strong com-
³¹¹ monalities between dispersers from these very different ecosystems. Thus, despite wild dogs'
³¹² ability to cope with diverse habitats and adapt to changing conditions (Woodroffe, 2011),
³¹³ the fundamental factors included in our study appear to influence wild dogs from other
³¹⁴ ecosystems alike. Nevertheless, expanding our analysis to dispersers emigrating from other
³¹⁵ source populations would invaluable contribute to our understanding of dispersal.

³¹⁶ Locally, we identified the Okavango-Linyanti region as a potential dispersal hub through
³¹⁷ which dispersing wild dogs gain access to more peripheral regions of the KAZA-TFCA.
³¹⁸ It appears that the absence of human activities, the central position within the KAZA-
³¹⁹ TFCA, and the presence of relatively impermeable water bodies (e.g. Okavango Delta,
³²⁰ Linyanti Swamp) funnel dispersal movements, resulting in a highly frequented corridor.
³²¹ Furthermore, the lack of permeable areas between peripheral source points often resulted
³²² in corridors stretching longer Euclidean distances, in an arc-shaped route via a stretch of
³²³ suitable habitat through the Okavango-Linyanti ecosystem. This is an expected outcome
³²⁴ in case structural and functional connectivity coincide, i.e. when dispersers follow suitable
³²⁵ habitats to disperse (Fattebert et al., 2015; Hauenstein et al., 2019). The key role of the
³²⁶ Okavango-Linyanti region for overall connectivity within the KAZA-TFCA thus calls for

actions to secure its protection status in the future. While the region is currently a Wildlife Management Area, it has neither the status of a National Park nor that of a Game Reserve. A similar case of non-formally protected but key dispersal landscape is represented by the area south of Kafue NP in Zambia, for which a disruption of its main and narrow dispersal corridor would result in considerable isolation of its subpopulations. We also revealed a potential southwards corridor between the Okavango-Linyanti ecosystem and the Central Kalahari National Park. Elliot et al. (2014) and Cushman et al. (2018) identified a similar corridor for dispersing lions, suggesting that upholding and protecting a link between those ecosystems is pivotal. Some areas through which the corridor runs are neither part of the KAZA-TFCA nor profit from any form of protection status. Human presence and activities along the national road that longitudinally traverses this corridor may limit functional connectivity (Cozzi et al., 2020).

Our approach of identifying movement corridors based on pre-defined start and end points implicitly assumes that individuals know the end point of their dispersal journey and that they have almost complete knowledge of associated movement costs (Panzacchi et al., 2016). Since dispersers often move into unknown territory, this may not necessarily be the case (Abrahms et al., 2017; Cozzi et al., 2020). However, specification of pre-defined end points might not be necessary, as the parametrized iSSF model can be used as mechanistic movement model to simulate dispersal events from known source points, yet without restricting the domain of potential end points (Signer et al., 2017). Consequently, movement corridors would emerge more naturally as the result of a myriad of simulated dispersal events (Zeller et al., 2020).

Besides estimating corridors, individual based simulations may be used to generate permeability surfaces (Avgar et al., 2016; Signer et al., 2017). Such surfaces have been shown to reduce the risk of overestimating $w(x)$ and consequently connectivity, particularly in areas that lie far from suitable habitats (Signer et al., 2017). While a simulation-based approach is conceptually straightforward, computational requirements are high, especially for such a large extent as constituted by the KAZA-TFCA.

Our work shows how dispersal data of a highly mobile species can be used to identify movement corridors and to assess the adequacy of protected areas. In our case, the predicted movement corridors of African wild dogs were well contained within the boundaries of the world's largest transboundary conservation area, namely the KAZA-TFCA, suggesting that it will significantly contribute to the long-term viability of this species. Moreover, our connectivity network allowed revealing potential dispersal hubs through which dispersers

361 gain access to more remote regions of the study area. Finally, our investigations showed
362 that human influence constitutes one of the main barriers to dispersal and substantially
363 reduces landscape connectivity. Successful conservation of wide-ranging species, such as
364 exemplified by the African wild dog, will therefore be contingent on the willingness of local
365 authorities, policymakers, and land managers to preserve areas that remain free from human
366 strains. Ultimately, our work contributes to the growing field of connectivity studies and
367 provides and easily expandable framework for assessing the adequacy of already-existing or
368 planned protected areas.

369 **5 Authors' Contributions**

370 D.D.H., D.M.B., A.O. and G.C. conceived the study and designed methodology; D.M.B.,
371 G.C., and J.W.M. collected the data; D.D.H. and D.M.B. analysed the data; G.C. and A.O.
372 assisted with modelling; D.D.H., D.M.B., and G.C. wrote the first draft of the manuscript
373 and all authors contributed to the drafts at several stages and gave final approval for pub-
374 lication.

375 **6 Data Availability**

376 GPS movement data of dispersing coalitions will be made available on dryad at the time of
377 publication. We will also provide example R-code on Github to demonstrate our least-cost
378 modelling approach (<https://github.com/DavidDHofmann>).

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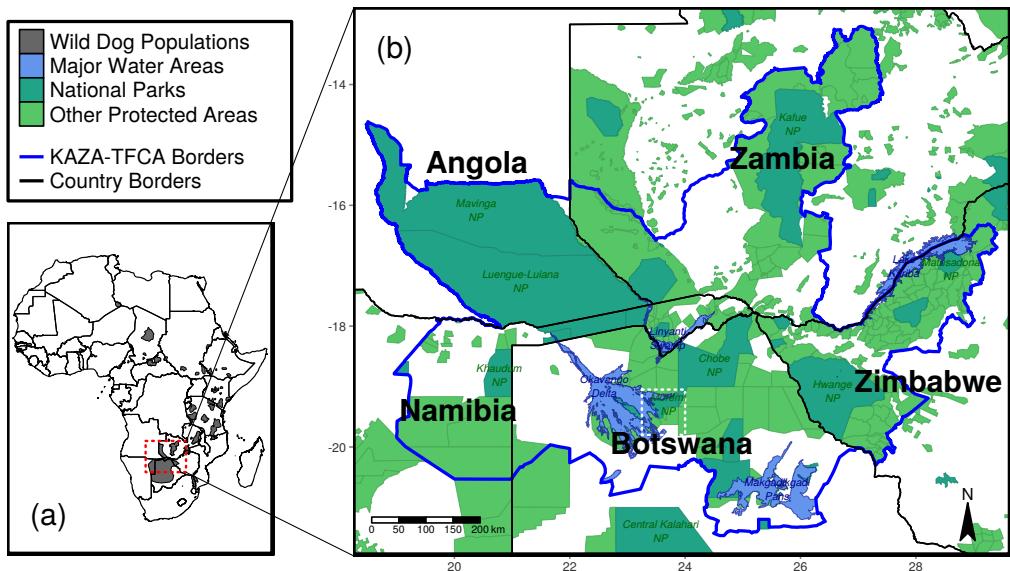


Figure 1: Overview of our study area. (a) The red dotted rectangle depicts the study area, which was confined by a bounding box encompassing the entire KAZA-TFCA. Gray areas indicate remaining wild dog populations according to the IUCN (Woodroffe and Sillero-Zubiri, 2012). (b) The white rectangle illustrates the area within which dispersing coalitions were collared. Since Game Reserves in Botswana virtually serve the same purpose as National Parks, we use the terms interchangeably for this region.

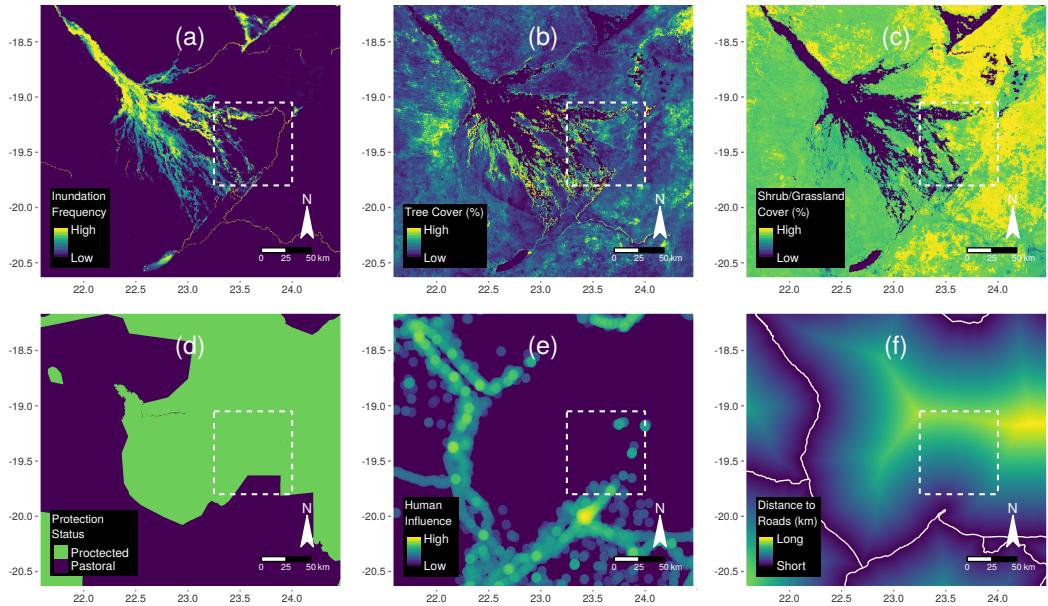


Figure 2: Overview of spatial covariates that we included in our models. We prepared all covariates for the entire study area but for better visibility we only plot them for the surroundings of the Okavango Delta. The white rectangle in each plot depicts the area within which dispersing coalitions were collared. (a) Averaged layer of all dynamic (binary) water maps. (b) Percentage cover of trees. (c) Percentage cover of shrubs/grassland. Anything that was not covered by trees or shrubs/grassland was deemed to be bare land. (d) Protection status of the area. (e) Human influence proxy composed of human density, farms, and roads. (f) Distance to nearest road (white lines depict actual roads).

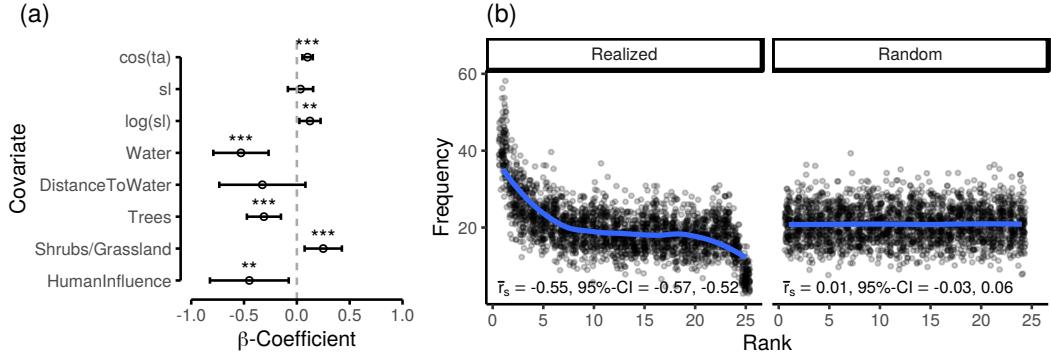


Figure 3: (a) Estimated selection coefficients from the most parsimonious habitat selection model. Negative coefficients indicate avoidance of a covariate, positive coefficients selection of a covariate. Whiskers delineate the 95%-CIs for estimated parameters. Significance codes: ** $p < 0.05$, *** $p < 0.01$. (b) Results from the k-fold cross validation for case-control studies. The left graph shows rank frequencies of *realized* steps according to predictions, whereas the right graph shows rank frequencies of *randomly selected* steps according to predictions. \bar{r}_s indicates the mean correlation coefficient resulting from 100 repetitions of the k-fold cross validation. The blue smoothing line was fitted using a locally weighted polynomial regression and serves to aid the eye in detecting the trends. Correlation coefficients suggest that our prediction was significant and robust, evidenced by the fact that the confidence intervals of $\bar{r}_{s,realized}$ and $\bar{r}_{s,random}$ did overlap and by the fact that there was strong and significant correlation between ranks and associated frequency for realized steps.

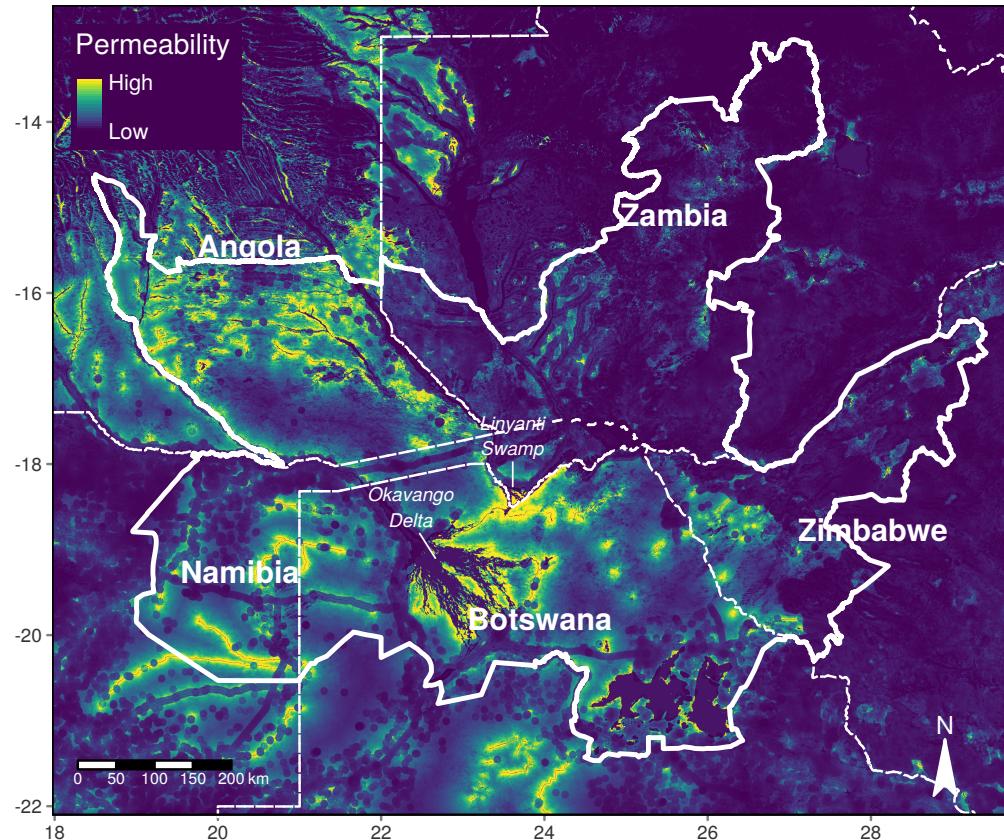


Figure 4: Predicted permeability surface for the extent of the KAZA-TFCA. Permeability was predicted by calculating selection scores $w(x) = \exp(\beta_1 x_1 + \beta_2 x_2 + \dots + \beta_n x_n)$ for each raster cell based on the raster cell's underlying covariates (x_i) and estimated selection strength (β_i). Areas that dispersers find easy to traverse are depicted in bright colors. Bold white lines delineate the borders of the KAZA-TFCA, whereas dashed white lines show country borders.

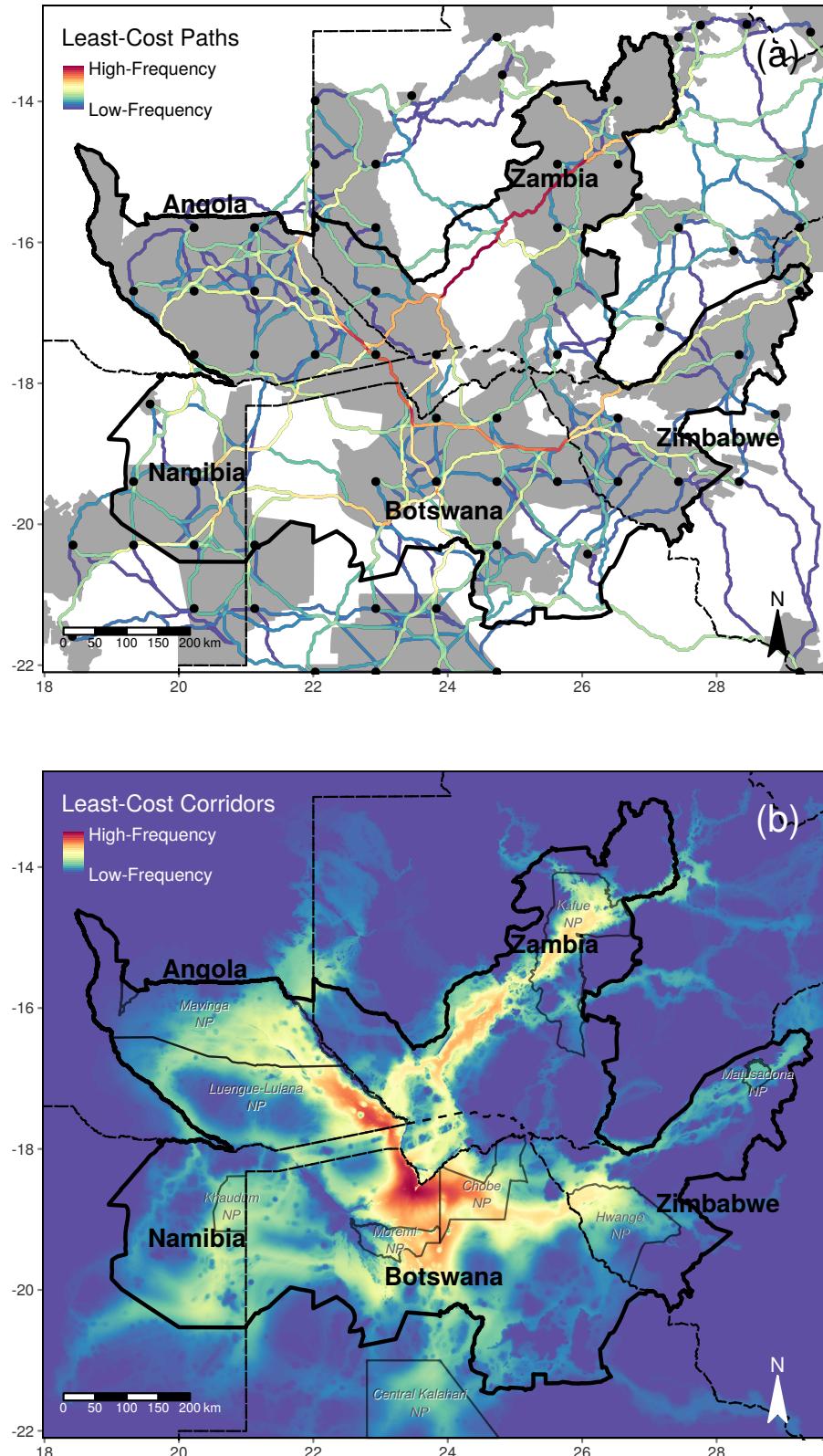


Figure 5: (a) Source points (black dots) and corresponding least-cost paths leaving from protected areas (dark grey borders). Note that only contiguous protected areas covering more than 700 km^2 are depicted. Continuous thin black lines indicate the borders of the KAZA-TFCA, whereas dashed black lines delineate country-borders. (b) Least-cost corridors between the same source points as illustrated in subfigure (a). For ease of spatial reference, we also labeled some national parks (NPs, in dark-grey).

Table 1: Comparison of median permeability (interquantile range in brackets) across countries, separated into areas within and outside the KAZA-TFCA, as well as within and outside formally protected areas. High values indicate high permeability, whereas low values correspond to low permeability.

Country	KAZA-TFCA		Protection Status		
	Inside	Outside	Protected	Pastoral	Overall
Angola	0.35 (0.41)	0.12 (0.32)	0.35 (0.41)	0.12 (0.32)	0.19 (0.38)
Botswana	0.24 (0.30)	0.14 (0.16)	0.27 (0.35)	0.14 (0.18)	0.18 (0.25)
Namibia	0.20 (0.30)	0.12 (0.17)	0.22 (0.30)	0.10 (0.14)	0.14 (0.24)
Zambia	0.05 (0.09)	0.02 (0.05)	0.04 (0.09)	0.03 (0.05)	0.03 (0.06)
Zimbabwe	0.06 (0.16)	0.05 (0.04)	0.07 (0.17)	0.04 (0.04)	0.05 (0.06)
Overall	0.15 (0.29)	0.06 (0.14)	0.14 (0.30)	0.06 (0.14)	0.08 (0.21)