

Evolutionary history explains as much variation in the temporal assembly of woody plant communities as temperature and photoperiod

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Abstract

Species life history events are advancing with climate change, but these phenological responses are highly variable. The impacts of climate change are spatially implicit, varying across populations, in addition to high species-level variation. But understanding the causes of this variability is critical if we are to accurately predict future changes and the cascading impacts phenology has on communities. We took a more holistic approach to account the diversity in drivers of phenological variability, using spring budburst to test the relative importance of population-level variation, species differences, and the potentially confounding effects of phylogenetic structure on current day phenology. We combined budburst data from two growth chamber studies and used a Bayesian phylogenetic mixed effects model to test cue responses of 47 woody plant species across four populations in eastern and western North America. While we found weak population-level differences in budburst and cues, budburst showed a strong phylogenetic structure. Cue responses, however, only explained 38.6 to 60.6% of variation, with considerable species-level differences. Our findings suggest phenology is largely the result of environmental cues and evolutionary history, and not geographic differences. Understanding the unidentified drivers of species-level differences, whether they be physiological constraints or other key traits, will allow us to predict future community responses to climate change and conserve ecosystem services.

Introduction

Climate change is altering the timing of species life history events—phenology—with average advances of 2.6-2.8 days per decade (Parmesan, 2007; Thackeray et al., 2016; Cohen et al., 2018). Phenological events and their responses to climate change, however, are highly variable. Events, such as leafout, can span a period of weeks in most communities and are shifting in response to rising temperatures at different rates (Parmesan, 2007; Fitter and Fitter, 2002; Menzel et al., 2006; Yu et al., 2010; Fridley, 2012).

Understanding and explaining this variability is an important goal that remains extremely challenging. Some degree of variability is likely due to differences in climate change itself across space—as some areas warm faster than others, potentially generating larger phenological shifts (Hoegh-Guldberg et al., 2018). Climate change alone, however, explains a limited amount of the total variation. In

contrast, substantial variation appears related to species-level differences in phenology (Vitasse et al., 2009; Wolkovich and Cleland, 2014; Zohner and Renner, 2014; Vitasse et al., 2018), but we still have a limited understanding of how these species differences compare to other potential sources of variation, including those at the population-level. Understanding the role and scale of species-level variation, however, is necessary to predict future changes in phenology, and ultimately, its impacts on community dynamics and ecosystem services, including carbon cycles and pollination (Gotelli and Graves, 1996; Cleland et al., 2007; Richardson, A.D., O’Keefe, 2009).

Different species often have unique phenologies—even within the same location and climate conditions. In varying the timing of their phenological events within a community, species can limit the overlap in their resource use and thus, the competition they experience. For example, in forest communities, understory species often budburst earlier than canopy trees, when light and soil nutrients are most available. This niche partitioning allows species to differ in their temporal resource use as environmental conditions change across the growing season, ultimately filling different temporal niche (Gotelli and Graves, 1996).

While temporal niche differences theoretically should allow more species to coexist in a community, there is considerable evidence that they are also shaped by population differences. Phenology can be highly variable across populations to select for locally optimum trait phenotypes (ADDCITES). This has been shown across many ecosystems, including for bird breeding phenology in montane meadows in Sierra Nevada, that differs by 12 days across high and low elevation populations (Saracco et al., 2019). How much of this phenological variability is due to local adaptation or plasticity combined with interannual variation in climate is unclear. Across populations, we may expect the presence of strong population-level differences in phenology to indicate that local adaptation is an important driver of variation.

Identifying if phenological differences across populations is due to local adaptation requires evidence that populations have differing underlying cue systems. Both animals and plants often cue to variation in temperature and photoperiod (Renner and Zohner, 2018; Bonamour et al., 2019), with expectations that responses may shift with latitude. Because both temperature and photoperiod vary across latitudinal gradients, populations may use different temperature thresholds and photoperiods, driving observed latitudinal gradients in phenology (ADDCITES).

To predict how communities will respond to continued climate change requires a holistic approach that accounts for these multiple drivers of phenological variability. For a given community, we must account for differences in temperature and photoperiod cues across species that shape their temporal niches. Species differences, however, may not be predicted perfectly by current temporal niches, as past climate variability during the long evolutionary timescales over which communities assembled could effect species-level variation today (Davies et al., 2013)—a potential effect that recent phylogenetic methods can help test. But this cannot be done in isolation of population-level variation due to plasticity, local adaptation, or both. Conducting experiments in controlled environments can help remove the relative effects of interannual climate variability that drives plastic variation in phenology, allowing us to identify species primary cues and test for evidence that populations vary in these cues.

Spring budburst offers an excellent system to test for species-level and latitudinal (population-level) patterns in phenology and cue responses. Budburst of temperate woody plants is known to respond to temperature cues in both the winter and spring, referred to as chilling and forcing, respectively, as well as daylength (photoperiod) (Chaine et al., 2010; Polgar and Primack, 2011; Cooke et al., 2012; Basler and Körner, 2014; Laube et al., 2014). These three cues interact to shape the start of spring growth, with variation in the relative importance of individual cues across species. Phenotypic differences between species—such as functional groups that represent varying growth and resource-use strategies—may further promote phenological differences and ultimately optimize species temporal

niche within a community.

Here we combined results from two growth chamber studies of woody species budburst cues. We used data from four populations, from eastern to western North America, with pairs of populations on each coast spanning 4-6° latitude respectively. Using a phylogenetic approach allowed us to detect general trends in budburst cue responses in North American deciduous forest communities. We also explored community specific responses to detect differences between western and eastern forest communities, and across latitudes. By including diverse assemblages of species, we tested for differences between functional groups, comparing the dominant shrub and tree species that characterize our forest understories and canopy.

Materials and Methods

Field sampling

We combined data from two growth chamber studies using branch clippings of North American deciduous woody plants. Such cutting experiments are a commonly used and powerful approach to understanding the responses of adult trees to known environmental conditions (Vitasse et al., 2014). Our first study used samples taken from two eastern populations—Harvard Forest, Massachusetts, USA (42.55°N, 72.20°W) and St. Hippolyte, Quebec, Canada (45.98°N, 74.01°W), which we sampled from the 20-28 January, 2015. The second growth chamber study spanned two western populations—Smithers (54.78°N, 127.17°W) and E.C. Manning Park (49.06°N, 120.78°W), British Columbia, Canada, respectively, sampled from the 19-28 of October, 2019.

We observed budburst for 47 species spanning our two transects, with 28 species at our eastern transect and 22 species at our western transect, 3 of which occur at both transects. We selected the dominant deciduous species in the forest communities at each population and maximized the number of species occurring across sites. Of the species in our eastern transect, there were 13 shrubs and 15 trees, while our western transect included 18 shrub and 4 tree species. At all sites, we tagged 15-20 healthy, mature, individuals for each species in the summer prior to sampling. Depending on the size of the individual, we collected 1-16 cuttings from 6-20 tagged plants, using a pole pruner from the ground. We kept samples cold and immediately placed them in water upon returning from the field. We conducted the eastern study at the Arnold Arboretum, Boston, MA, USA, and the western study at the University of British Columbia, BC, Canada. We used daily weather data from local weather stations to calculate the field chilling prior to collection (see Table ??).

Growth chamber study

Drawing on decades of work identifying the primary budburst cues, we tested the effects of chilling, forcing, and photoperiod cues. We used growth chambers with two levels of chilling—with either no additional chilling or 30 days at 4°C for our eastern study, and 30 days or 70 days of chilling at 4°C for our western study (all dark)—after which we moved plants to one of two levels of forcing—a cool regime of 15:5°C and a warm regime of 20:10°C—crossed with two photoperiods of either 8 or 12 hours—for a total of 8 distinct treatments in each study. Our design was similar for both eastern and western species, except for differences in the timing of sample collection and the thermoperiodicity in forcing treatments. We collected samples in our eastern study in late January, experiencing considerable field chilling. But we collected the western samples in late October, with most chilling occurring under controlled conditions. In our eastern study, we also set the duration of forcing temperatures to correspond with the photoperiod treatments. While this is a widespread and common approach in phenology chamber studies, it leads to a confounding effect in the thermoperiodicity (Buonaiuto et al., 2023). Thus we applied forcing temperatures for 12 hours independent of the duration of the

photoperiod treatments in our western study. We had seven to ten replicates per treatment in our eastern study and eight replicates in our western study, with the exception of *Sorbus scopulina*, for which we only had six replicates per treatment. Further details regarding the re-cutting of samples, water changes, and rotation through chambers are presented in Flynn and Wolkovich (2018).

We assessed phenological stages of budburst using the BBCH scale, adapted for our specific species (Finn et al., 2007). Every 1-3 days, we observed each sample and recorded all phenological stages up to full leaf expansion (defined as code 17 by Finn et al., 2007). Here, we present the response in budburst date, defined as the day of budbreak or shoot elongation (code 07 by Finn et al. (2007)). Our eastern study spanned 82 days, with over 19320 phenological observations, while the western study spanned 113 days and we made 47844 phenological observations in total.

Statistical Analysis

To test for differences in cue responses across species and populations, we used a phylogenetic mixed effects model with partial pooling across species. This approach accounts for both the evolutionary relatedness of our species, and estimates the species level cue responses and differences across sites. We obtained species phylogenetic relatedness by pruning the Smith and Brown (2018) megatree of angiosperms.

In our model, we included each of our three cues as continuous variables. To account for the greater amounts of field chilling experienced in our eastern study, we converted our chilling treatments to chill portions for each individual population using local weather station data and the chillR package (Luedeling, 2020). We calculated chill portions for both the chilling experienced in the field prior to sampling and during the experiment in our 4°C chilling chambers. We also converted forcing temperatures to mean daily temperatures in each treatment to account for differences in thermoperiodicity between the two studies (Buonaiuto et al., 2023). Finally, we z-scored each cue and site using two standard deviations to allow direct comparisons between results across parameters (Gelman, 2008). In addition to the overall predictors, we included all two-way interactions between cues and between cues and sites.

Our modeling approach allowed us to combine observations of budburst (y_i) across species (sp , to estimate both a species-level intercept (α) and slope estimates for individual cues, sites, and all possible interactions (β) to estimate the day of budburst (y) relative to the first day of forcing conditions.

$$\begin{aligned}
 y_i = & \alpha_{sp_i} + \beta_{site2} + \beta_{site3} + \beta_{site4} + \\
 & \beta_{chill_{sp_i}} + \beta_{force_{sp_i}} + \beta_{photo_{sp_i}} + \\
 & \beta_{force \times chilling_{sp_i}} + \beta_{chilling \times photo_{sp_i}} + \beta_{photoperiod \times chilling_{sp_i}} + \\
 & \beta_{force \times site2_{sp_i}} + \beta_{force \times site3_{sp_i}} + \beta_{force \times site4_{sp_i}} + \\
 & \beta_{chilling \times site2_{sp_i}} + \beta_{chilling \times site3_{sp_i}} + \beta_{chilling \times site4_{sp_i}} + \\
 & \beta_{photoperiod \times site2_{sp_i}} + \beta_{photoperiod \times site3_{sp_i}} + \beta_{photoperiod \times site4_{sp_i}} + \epsilon_i
 \end{aligned}$$

$$\epsilon_i \sim \text{normal}(0, \sigma_y^2)$$

The intercept and slopes were modeled with partial pooling at the species-level:

$$\begin{aligned}
\beta_{force_{sp}} &\sim \text{normal}(\mu_{force_{sp}}, \sigma_{force}^2) \\
&\dots \\
\beta_{photoperiod \times site4} &\sim \text{normal}(\mu_{photoperiod \times site4}, \sigma_{photoperiod \times site4}^2)
\end{aligned}$$

171 We included the phylogenetic effect as a variance covariance matrix (\mathbf{V}) in the parameterization of the
 172 normal random vector:

$$\boldsymbol{\alpha} = \begin{bmatrix} \alpha_1 \\ \alpha_2 \\ \dots \\ \alpha_n \end{bmatrix} \text{ such that } \boldsymbol{\alpha} \sim \text{multi-normal}(\mu_{\alpha}, \mathbf{V}), \quad (1)$$

$$(2)$$

173 The timing of budburst prior to evolution is represented as root values of the means of the mul-
 174 tivariate normal distributions. The For more detail on this method of phylogenetic modeling, see
 175 Morales-Castilla et al (2024).

176
 177 We used semi-informative priors for each of our model parameters, validating the model code using
 178 test data prior to analysis. Our model produced \hat{R} values close to 1, well mixed chains, and n_{eff}
 179 values that exceeded 10% of the model iterations. Our model was fitted using the Stan language (Stan
 180 Development Team, 2018) using the rstan package in R (R Development Core Team, 2017) (version).

181
 182 We present model estimates, relative to baseline conditions of low chilling, low forcing, and short pho-
 183 toperiod treatments, and—when relevant—for the Smithers population. We present results as means
 184 and 90% posterior uncertainty intervals from our Bayesian phylogenetic models.

185

186 Results

187 On average, we observed species budburst 28.1 days (uncertainty interval: 45, 15.2) after the start
 188 of forcing and photoperiod treatments (all estimates are given as mean \pm 90% uncertainty intervals,
 189 henceforth ‘UI,’ and budburst dates given as relative to baseline conditions, see methods for more de-
 190 tails). Some species, such as *Aronia melanocarpa*, budburst as early as day 13.6, and as late as day 52.1
 191 for *Quercus velutina*. The difference in the timing of the earliest and latest species spanned a period of
 192 38.5 and 30.7 days for the eastern and western species respectively. Overall we found species budburst
 193 was strongly phylogenetically structured (λ of 0.8, UI: 0.9, 0.6) and that all cues led to an advance in
 194 budburst date. Of the three cues, chilling had the strongest effect (-15.2, UI: -17.3, -13.2) and pho-
 195 toperiod the weakest (-3.6, UI: -4.3, -3.0, Fig. 1). But we found a strong, positive interaction between
 196 forcing and chilling (9.1, UI: 7.6, 10.5). This results in subadditive effects between chilling and forcing
 197 cues, with low chilling being offset by high forcing conditions (Fig. 2 and see Table 1 for model output).

198

199 Overall we found small population-level effects compared to differences between cues. Across all species,
 200 there was considerable overlap in the responses of the four populations to each cue (Fig. 1). As such,
 201 low forcing, chilling, and shorter photoperiods all produced later estimated budburst dates compared
 202 to higher temperature and longer photoperiod treatments across all populations (Fig. 3 a-c, shown
 203 for baseline conditions) and across transects (Fig. 3, see Table 1 for model output). Populations
 204 overall budburst dates did not differ, but eastern populations budburst marginally earlier (34.1, UI:

40.7, 27.1) compared to the western populations (46.5, UI: 52.7, 39.4, see also Table 1 for model output).

Budburst of individual species showed distinct differences in their timing and the relative importance of cues. Both chilling and forcing responses varied with budburst, with later budbursting species having slightly stronger responses to each cue respectively (Fig. 4). Despite our expectations that species responses should differ across the canopy layer, with shrubs being less sensitive to cues than tree species, we did not find strong differences across these two functional groups (Fig. 6). Shrubs, like *Cornus stolonifera*, showed weak responses to chilling and forcing. But 36.7% of the shrubs species, including *Menziesia ferruginea* and *Symphoricarpos alba*, exhibited the opposite response and budburst relatively late (Fig. 6). Similarly for tree species, some matched our predictions for their cue responses, but 23.5% budburst earlier than expected. Specific tree species, such as *Quercus velutina*, did have stronger chilling and photoperiod responses as predicted, such as *Fagus grandifolia* produced the strongest photoperiod response. But overall we did not find clear differences between the cue responses of trees and shrubs across the four forest communities.

Our results depict the relative importance of phenological cues to variation across species, but we found the contribution of cues to budburst order and timing to be dependent of the magnitude of the cue, and small compared to overall species differences unexplained by cues (represented in the modeling framework as the intercept). The rank order of species estimated budburst changed more under high cue conditions than under low cues (Fig 7). The budburst ranking for the three species that occurred in both transects remained relatively consistent across the two communities (Fig. 4), with only *Alnus incana* in the western community experiencing a large change in rank with cues (Fig. 7).

In comparing the earliest and latest budbursting species, we found relatively small differences in cues (Fig. 4). This is contrary to our prediction of stronger cue requirements—especially chilling cues—for later budbursting species. For example, an early shrub species, *Lyonia ligustrina*, had chilling and forcing cues of -15.4 and -10.4 respectively, which are comparable to the cue estimates of -16.6 and -10.4 for the much later tree species, *Quercus alba*. Yet the model predicts these species to have a 15.9 day difference in day of budburst. This highlights how the timing of budburst across species was mostly attributed to species level differences outside of cue responses (intercept values, see Fig. 5). Across the range of cue values, cues explained only between 38.6 and 67.0% of variation for eastern species and between 48.3 and 60.6% for western species, with the intercept explaining the rest (Fig. 5).

Discussion

In our study the assembly of species temporal niche of woody plant budburst across North America was partly explained by species differences in environmental cues, with little variation explained by differences between populations. Despite populations differing in their species assemblages, and ratio of shrub to tree species, they had similar budburst cue responses and timing. But in comparing species within communities, we found budburst to span several weeks. We also found slight gradients in environmental cue responses across each region, with the strongest responses to chilling and forcing cues. Our findings support that species phenology is largely shaped by temperature and light, but there is also strong phylogenetic structuring in budburst.

Temporal assembly

Despite differences in the species composition and local conditions of our forest communities, under controlled conditions, populations differed only slightly in their phenology and cue responses. We found that our Eastern populations budburst slightly earlier than our western populations (Fig 1). This con-

trasts our expectation that local habitats impose unique selective pressures, driving differences in cues responses. This negligible variation could be due to the earlier collection date of our western species, which reduced the field chilling they received relative to our eastern samples.

Across latitudinal gradients, we expected to find evidence of population differences, as latitude correlates with reduced photoperiod and temperatures. Greater shifts in phenology across latitudinal gradients have been found by studies using *in situ* phenological data (Post et al., 2018; Alecrim et al.). However, this work is based on long-term observations in the field, with confounding differences in study duration and start dates, and use variable methodologies and geographic extent (Post et al., 2018; Alecrim et al.). The lack of population-level effects that we observed suggests budburst is not shaped strongly by current geography or local adaptation, at least at our continental scale.

Community composition and interspecific variation in phenology

Our species varied in their responses to environmental cues, creating large potential differences in their temporal niche and ecological roles. Species ranged from early to late budburst dates and spanned a similar period as observed in natural communities (Lechowicz, 1984; Richardson, A.D., O’Keefe, 2009). This suggests that with our experiment captured a realistic breadth in temperate forest budburst phenology. Furthermore, all our focal species responded to each environmental cue, with strong responses to chilling and weak responses to photoperiod, which is consistent with previous studies. Cues, however, had complex interactions that appear potentially advantageous under warming climates (Heide, 1993; Caffarra and Donnelly, 2011; Flynn and Wolkovich, 2018). For example, the interaction between chilling and forcing ensures species still budburst if warmer winters cause insufficient chilling, but will require additional forcing when chilling is low. Despite these strong differences in species overall cue responses, we did not find the clear, generalizable trends across species with similar growth strategies.

Shrub and tree species differ greatly in their physiology, filling different ecological niche space. But we found little overall variation in cue responses. Most of our tree species budburst later than our shrubs species, a relative order also found by previous studies (Gill et al., 1998; Panchen et al., 2014). About a quarter of our tree species also budburst early, with timing more similar to that of shrubs. Similarly, we found a third of shrub species to budburst at similar times as the majority of our trees. This deviation from our expectations for how shrub and tree species partition their budburst timing suggests there is more nuance to these patterns than previously found. As trees species advance phenologically—resulting in earlier canopy closure and reduced light availability (Donnelly and Yu, 2019)—shrub species with weak cue responses may have reduced fitness. But our finding suggests many shrubs will also advance phenologically and have the potential to maintain their relative temporal niche and mitigate potential effects on community dynamics.

The number of shrub and tree species exhibiting later than expected budburst phenologies suggests the cues we think are important for budburst could also be important for other traits. Earlier budbursting species may exhibit a suite of traits that infers a greater ability to capture light and photosynthesize prior to canopy closure or improve nutrient uptake later in the season. While we associate these traits with shrubs that dominate the forest understory, it is possible that temporal niche partitioning among tree species have selected for earlier budburst in some tree species as well.

Phenology has not been well integrated with our broader current understanding of how traits shape species niche and growth. But, plant strategy theory predicts phenology to correlate with other key traits that facilitate species ability to take advantage of available resources and grow early in the spring (CITEGRIME), or infer the greater competitive abilities needed when budbursting later. Differences in species temporal niche are also likely shaped by differences in other key plant traits. Currently, we do not understand the mechanisms that underlie these trait relationships, but incorporating phenology into a broader trait framework could provide insights into the drivers selecting for species phenotypes.

Community assembly in responses to cues versus evolutionary history

Our ability to accurately forecast phenology is predicated on our understanding of cue responses at the population, species and community levels. In North America, springs conditions show high inter-annual variability, with forest communities experiencing a breadth of environmental cues in a given year (Schwartz and Reiter, 2000; Zohner et al., 2017). It may be advantageous for species to have consistent cue responses to optimize budburst across years and space.

Population-level variation in cue responses were negligible at the species-level, but we did find important differences (In what? This reads like you mean population level difference) (Fig. 5). As climate change leads to stronger temperature cues, we are already observing advances in species budburst, but the strongest effects will occur in high latitude populations that are warming faster and to a greater extent than others (Hoegh-Guldberg et al., 2018) (Fig. 5). In future, budburst across populations may become increasingly different, as the compounding effects of warming temperatures impose varying selection across species distributions.

At a finer scale, our findings suggest the order of species budburst in a community will change with variable warming. Community dynamics are most likely to change in populations that experience greater warming. As the order in which species budburst shifts (Fig S4), species that previously exhibited distinct temporal niche will increasingly interact. This may have cascading effects on species coexistence, changing their competitive landscape, and synchrony with pollinators or herbivores. But we cannot focus solely on species cues to understand this temporal community assembly. Species phenotypes are the result of multiple interacting and complex environmental cues that act over evolutionary time scales (Ackerly, 2009).

Our study included the three cues—chilling, forcing and photoperiod—mostly commonly attributed to species spring phenology, but about a third of total variation was not explained by these cues (Fig 4). This suggests our model is incomplete. Contrary to our expectations, this variation was not explained by differences in geography or latitudinal gradients in cues. Our results, instead, point to a large effect of species evolutionary history on budburst.

Still missing from our understanding of budburst phenology are the unidentified latent traits that underlies the observed phylogenetic structure. Species differ in many physiological traits that also shape the temporal niche of temperate woody species. For example, we might predict shrub species to possess photosynthetic machinery that allows them to better use the light available prior to canopy closure in early spring. By not accounting for these additional cues or traits that drive phenology, we have a limited ability to predict how communities will respond to future changes.

The evolutionary history and ancestral phenotype of a species will also have a strong effect on their adaptive potential to future climates. Species traits are shaped over 'deep' timescales, with previous evolution influencing responses to the present climate. The conservation of ancestral phenotypes could impose physiologically constraints, preventing species from adapting to new environments. In this way, phylogenetic trends could be an indication of limited evolutionary time to adapt. Some temperate species, like *Nyssa* and *Fagus*, have had a shorter geologic time to adapt to their local temperatures (Lechowicz, 1984). These phylogenetic effects could skew species assemblages, favouring species with traits and cue requirements adapted for historical climates.

Predicting budburst phenology under future climates

Our results provide new insights into the factors and ecological processes critical to forecasting future phenologies. The consistent trends we observed across populations facilitate forecasting across similar regions with limited local phenological data but similar species assemblages. The high phylogenetic structure in budburst timing could enable us to forecast trends in phylogenetically related species for well resolved and well sampled genera (Molina-Venegas et al., 2018).

Our analytical approach and findings have broad applications across plant communities and in forecasting species responses to climate change. To this end, future efforts should invest in identifying latent traits contributing to the high phylogenetic structure of budburst. Combining community wide approaches with phylogenetic relationships, such as we have done, allow us to combine the effects of species evolutionary history with current ecological processes. This holistic approach will enable us to predict changes in spring phenology and mitigate their cascading effects on forest communities and ecosystem services.

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Tables and figures

Table 1: Summary output from a phylogenetic Bayesian model in which species are partially pooled and phylogeny is included on the intercept. The model includes photoperiod and site as dummy variables, while the forcing and chilling effect is included as continuous. Parameter estimates were used to estimate day of budburst for sites under low cue conditions and for Smithers only when relevant.

	mean	sd	5%	95%
Intercept	12.51	3.14	7.40	17.60
Phylogenetic effect	0.79	0.12	0.60	0.90
Forcing	-9.55	0.74	-10.70	-8.30
Photoperiod	-3.62	0.41	-4.30	-3.00
Chilling	-15.21	1.25	-17.30	-13.20
Manning Park	2.09	0.36	1.50	2.70
Harvard Forest	-6.04	1.03	-7.80	-4.40
St. Hippolyte	-8.71	0.97	-10.30	-7.10
Forcing x photoperiod	0.23	0.71	-1.00	1.40
Forcing x chilling	9.06	0.90	7.60	10.50
Photoperiod x chilling	-0.67	0.90	-2.20	0.80
Forcing x Manning Park	-1.76	0.77	-3.00	-0.50
Photoperiod x Manning Park	0.58	0.79	-0.70	1.90
Chilling x Manning Park	-0.36	1.60	-3.00	2.20
Forcing x Harvard Forest	3.81	1.22	1.80	5.80
Photoperiod x Harvard Forest	-1.96	0.86	-3.30	-0.60
Chilling x Harvard Forest	9.97	2.03	6.60	13.40
Forcing x St. Hippolyte	5.25	1.19	3.20	7.20
Photoperiod x St. Hippolyte	-2.13	0.84	-3.50	-0.70
Chilling x St. Hippolyte	8.65	1.70	5.90	11.50

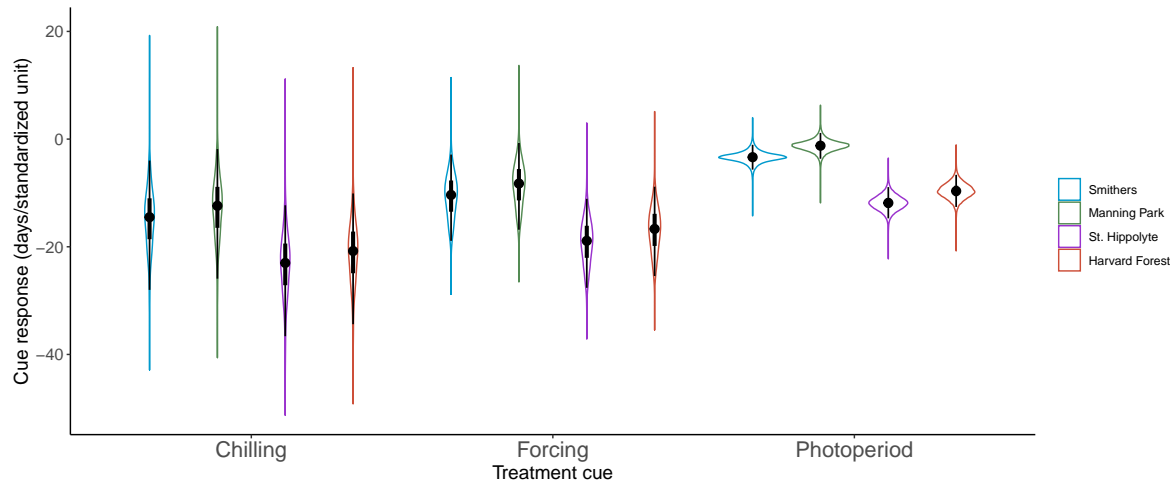


Figure 1: Posterior distributions of estimated chilling, forcing, and photoperiod responses with site level effects for individual populations. Black circles represent the median cue response, thicker black lines the 50% uncertainty interval, while the thinner black line the 90% uncertainty interval. The y-axis spans the entire range of the data. We z-scored cues using two standard deviations, see methods.

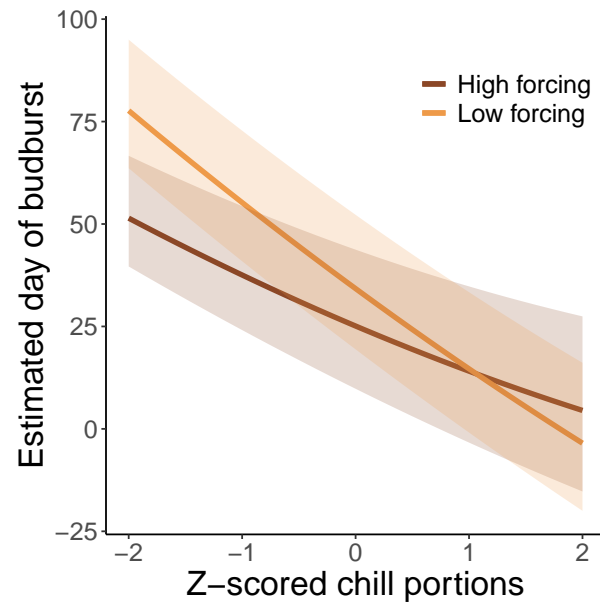


Figure 2: Estimated day of budburst of first bud in response to chill portions and forcing, estimated for our defined baseline conditions. We found a positive interaction between chilling and forcing cues, with high forcing compensating for low chilling to produce earlier budburst, while budburst occurred earlier under low forcing following high chilling.

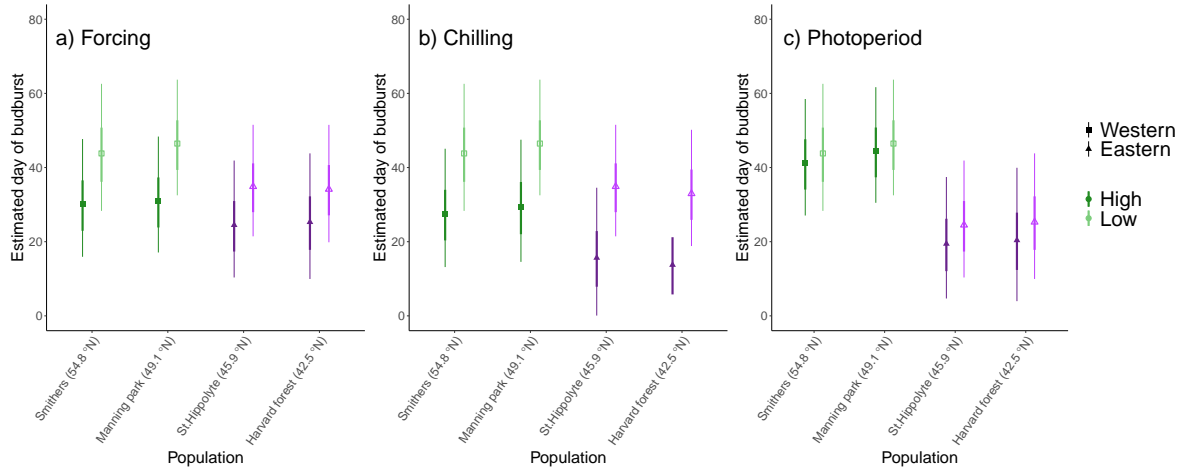


Figure 3: Estimated day of budburst of first bud in response to (a) forcing across populations under low chilling conditions and short photoperiods, (b) chilling across populations under low forcing and short photoperiods, and (c) across photoperiods under our baseline forcing and chilling conditions for species sampled from our four populations. The thin error bars represent the 90% uncertainty interval, while the thicker error bars represent the 50% uncertainty interval.

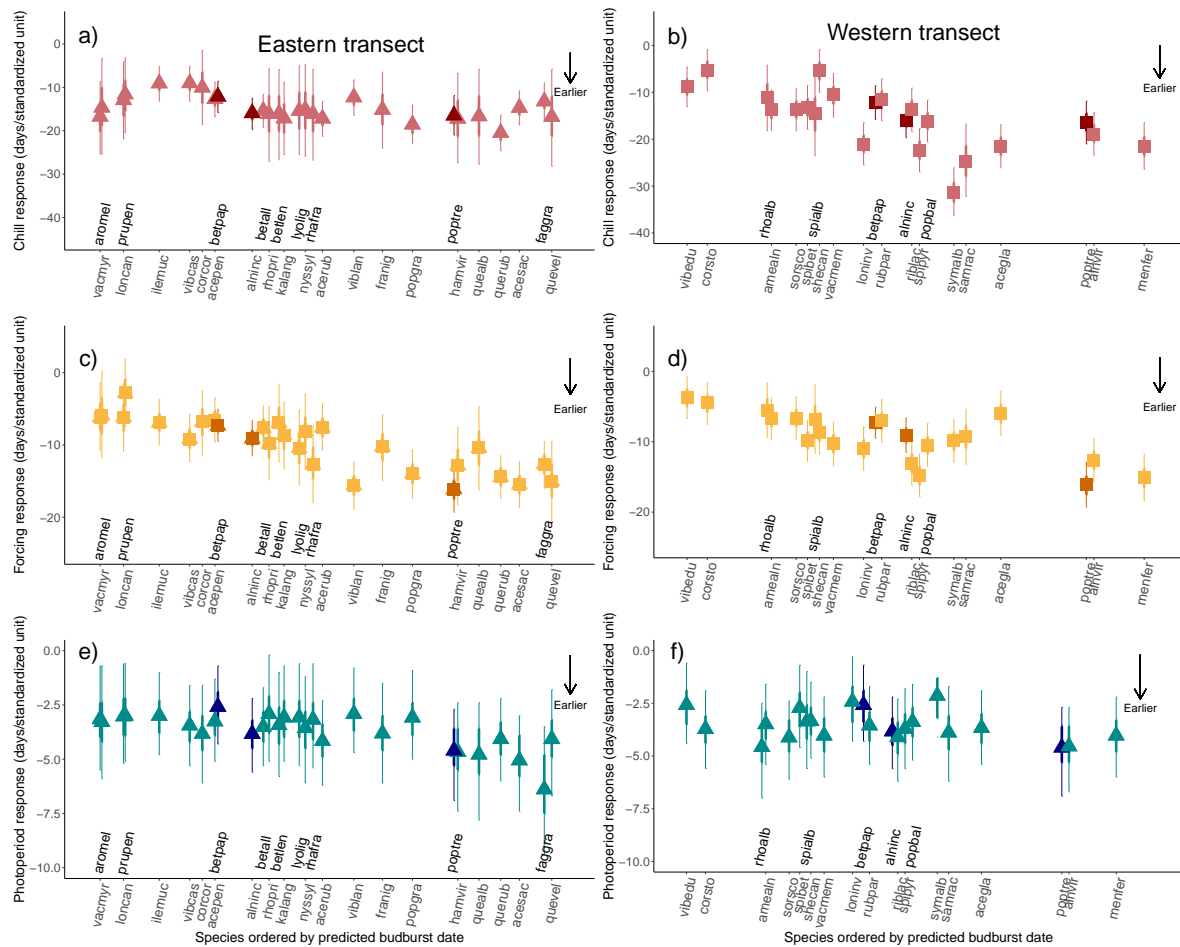


Figure 4: Estimated species chilling (a,b), forcing (c,d) and photoperiod (e,f) responses ranked by increasing estimated budburst dates for both the eastern (a, c, e) and western (b, d, f) populations. Cues are plotted on differing y-axis to better depict species differences. For each species, the thinner error bars represent the 90% uncertainty interval, and thicker bars represent the 50% uncertainty interval. Species that occur in both the eastern and western transects are depicted using darker colours. To make the results across cues directly comparable, we standardized predictors (here, cues) via z -scores using two standard deviations, see methods for further details.

Table 2: Chill units from our two western sites, E.C. Manning Park (MP) and Smithers B.C.(SM) Canada, and our two eastern sites, Harvard Forest (HF) USA and St. Hippolyte(SH) Canada.

Population	Chilling.treatment	Chilling.Hours	Utah.Model	Chill.Portions
Harvard forest	Field chilling	892	814.50	56.62
Harvard forest	Field chilling + 30 d at 4 degree C	2140	2062.50	94.06
St. Hippolyte	Field chilling	682	599.50	44.63
St. Hippolyte	Field chilling + 30 d at 4 degree C	1930	1847.50	82.06
Smithers	Field chilling + 30 d at 4 degree C	1317	1368.00	54.95
Smithers	Field chilling + 70 d at 4 degree C	1965	2016.00	74.67
Manning Park	Field chilling + 30 d at 4 degree C	1213	1377.00	55.09
Manning Park	Field chilling + 70 d at 4 degree C	1861	2025.00	75.33

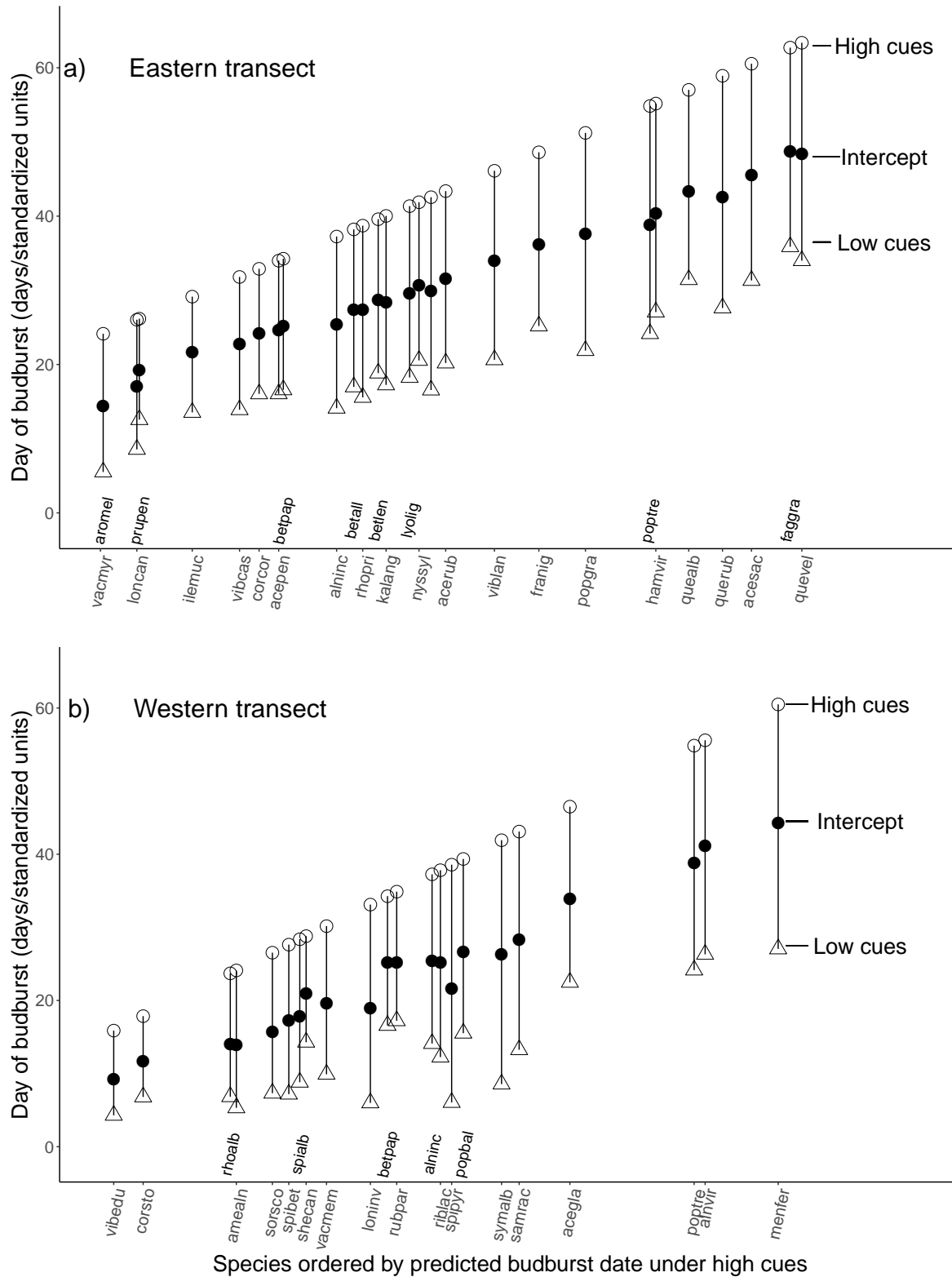


Figure 5: Comparisons of our estimated budburst based on the full model (intercept plus all cues), shown in white, versus the intercepts only (without cues), shown in black, ordered according to differences in species estimated budburst dates for both the eastern (a) and western (b) populations under high cue conditions, depicted as circles, and low cue conditions, depicted as triangles.

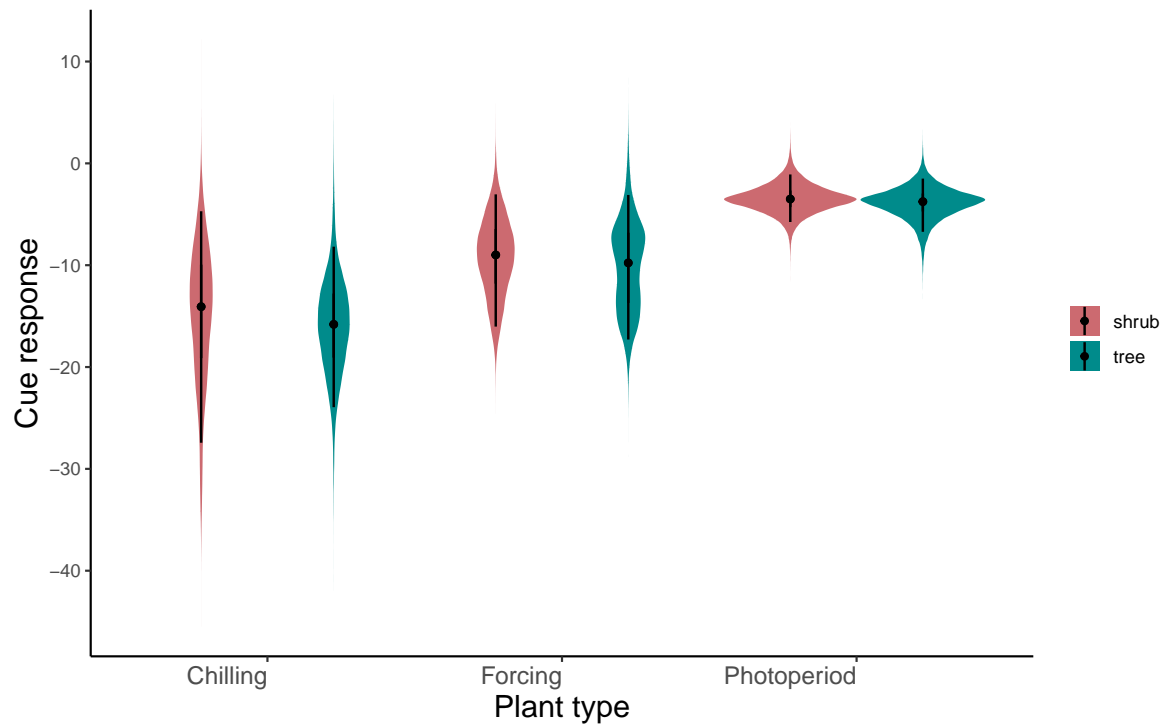


Figure 6: Comparisons of posterior distributions for cues estimates between shrub and tree species. Black circles represent the median cue response, while the thinner black line the 90% quantile interval. The coloured distribution is the the posterior density of the posteriors of the cue responses for all species within a given architectural type. The y-axis spans the entire range of the data.

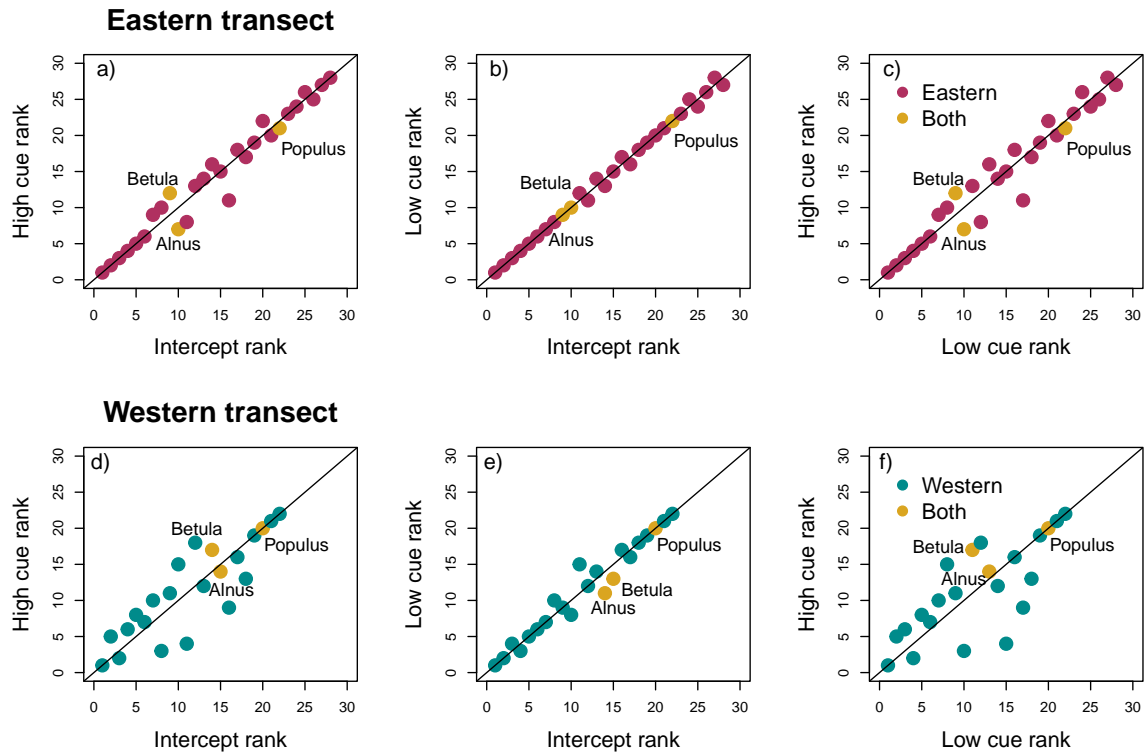


Figure 7: Estimated changes in species ranked budburst order, compared between species level effects (species intercept) and under high cue conditions (a and d), species level effects and under low cue condition (b and e), and ranked order under low and high cue conditions (c and f) for our eastern in red (a-c) and western species in blue (d-f). For the three species that occur in both transect, shown in yellow, *Alnus incana* exhibited the greatest rank change with a difference of three, while *Betula papyrifera* had a rank difference of two and *Populus tremuloides* a rank difference of one.