

Evolutionary history as important as phenological cues in the temporal assembly of woody plant communities

Deirdre Loughnan¹ and E M Wolkovich¹

October 18, 2023

¹ Department of Forest and Conservation, Faculty of Forestry, University of British Columbia, 2424 Main Mall Vancouver, BC, Canada, V6T 1Z4.

Corresponding Author: Deirdre Loughnan, deirdre.loughnan@ubc.ca

Introduction

Anthropogenic climate change is having many profound effects on plant communities globally. A well studied example of climate changes impacts is in the timing of phenological events, particularly spring events such as budburst or leafout (Fitter2002, Parmesan2006, Menzel2006, Vitasse2018). Globally, the effects of climate change have not been uniform, with some regions experiencing greater warming than others (IPCC, Schwartz2006). This has produced high variability in phenological shifts, with unique responses observed for different species and across habitats (Fitter2002, Menzel2006, Yu2010, Fridley2012). Between forest communities, observed differences in the timing of budburst has been attributed to species-level differences and geographic variability in environmental cues (Vitasse2009, Wolkovich2014, Zohner2014, Gerst2017, Vitasse2018). Identifying the relative importance of different drivers of budburst is necessary to understand and predict future changes in spring phenology, and their ultimate impacts on growing season length, carbon cycles, and species interactions (Gotelli1996, Cleland2007, Richardson2009, Richardson2013, Keenan2014).

For woody plants, there are three important environmental cues shaping the timing of budburst (Chuine2010, Polgar2011, Cook2012, Basler2014, Laube2014). Plants respond first to chilling cues, which are the low temperatures and duration of cold conditions required to break bud's winter dormancy. This is followed by forcing cues, which are the temperatures that initiates growth in the spring, and finally photoperiod cues, or a species' responsiveness to daylength. While each of these cues vary in their relative importance for a given species, they can also interact (Heide, 1993; Chuine, 2000; Caffarra and Donnelly, 2011; Flynn and Wolkovich, 2018). Previous work has shown that forcing cues can offset the effects of low chilling under warm winter conditions, while photoperiod responses can offsets weak forcing cues in cool springs (Heide, 1993; Chuine, 2000; Caffarra and Donnelly, 2011; Flynn and Wolkovich, 2018). The highly variable nature of both species cue requirements and the magnitude of their responses, however, has made it challenging to draw generalizable trends across species with large distributions.

The environmental cues that shape species budburst can vary geographically and temporally. Under future climates this could create differences in community dynamics and species composition. Increasing winter and spring temperatures will lead to the faster accumulation of chilling and forcing cues (Guy, 2014), which is associated with earlier budburst in many temperate plant species. Changes in temperature cues with climate change could therefore create novel niche space early in growing season

and facilitate species invasions or alter community assemblages (Willis2010,Wolkovich2013). Species with strong photoperiod cues may experience a reduction in their fitness, as their ability to adapt will be constrained by constant daylengths(Way2015). Differences in species cue requirements will therefore shape species adaptive potential to future change, altering the competitive landscape experienced across the growing season, and reshaping the diversity and persistence of species across forest communities.

The timing of individual species' budburst in a forest community can be highly variable, with the time from the earliest to latest budbursting species spanning several weeks in some temperate forests (Lechowicz1984). This breadth in the period of when individual species budburst in the spring allows species to fill different temporal niche (Gotelli1996). For example, understory shrub species often budburst earlier than taller canopy species, a niche difference predicted to infer overarching differences in species growth strategies. In addition to these within community differences, variation in budburst responses is likely to also exist across species' ranges (cite). At present we lack a strong understanding of the relative effects of population versus species-level variation in shaping budburst of individual species and forest communities overall.

While climate change has caused temperatures to change over recent decades, it is also important to account for the longer timescales over which forest communities have assembled (Davies2013). It is local, historic conditions that have shaped traits like phenology, the results of which we observe today. Across a species' spatial distribution we expect local cues to have differed, driving differences in their budburst phenotype and cue responses. Species with large latitudinal distributions, for example, experience differences in their photoperiod cues across their spatial range and may differ in their responsiveness to photoperiods across populations. But to date, few studies have explored how cue use differs across spatial gradients using the same forest communities, testing the relative role of local environments and biotic communities in shaping budburst. To better predict how forest communities will respond and assemble under continued climate change we need to identify the cues that drive budburst in our dominant forest species and functional groups, and how these responses can change across populations.

In this study we combined results from two growth chamber studies of woody species budburst cues. We used data from four populations, from eastern to western North America, with pairs of populations on each coast spanning 4-6° latitude. Our approach allowed us to detect general trends in how budburst of North American deciduous forest communities respond to forcing, chilling, and photoperiod cues. We were also able to explore community specific responses and detect differences between western and eastern forest communities, and across different latitudes. By including diverse assemblages of species, we also tested for trends across different functional groups, exploring differences between the dominant shrubs and tree species that characterize the understory and canopy of our forest communities.

Materials and Methods

Field sampling

We combined data from two growth chamber studies using branch clippings of North American deciduous woody plants. Such cutting experiments are a commonly used and powerful approach to understanding the responses of adult trees to known environmental conditions (Cite Vitasse). Our first study used samples taken from two eastern populations—Harvard Forest, Massachusetts, USA (42.55°N, 72.20°W) and St. Hippolyte, Quebec, Canada (45.98°N, 74.01°W), which we sampled from the 20-28 January, 2015. The second growth chamber study spanned two western populations—Smithers (54.78°N, 127.17°W) and E.C. Manning Park (49.06°N, 120.78°W), British Columbia, Canada, respectively, sampled from the 19-28 of October, 2019.

We observed budburst for 47 species spanning our two transects, with 28 species at our eastern transect and 22 species at our western transect, 3 of which occur at both transects. We selected the dominant deciduous species in the forest communities at each population and to maximize the number of species occurring across sites. Of the species we sampled in our eastern transect, 13 were shrubs and 15 were trees, while in our western forest community we sampled 18 shrub and 4 tree species. At all sites, we tagged 15-20 healthy, mature, individuals for each species in the summer prior to the growth chamber study. Depending on the size of the individual, we collected 1-16 cuttings from 6-20 tagged plants, with samples being taken using a pole pruner from the ground. We kept samples cold during sampling and immediately placed them in water upon returning from the field. Our eastern study was conducted at the Arnold Arboretum, Boston, MA, USA, and the western study at the University of British Columbia, BC, Canada. We used daily weather data from local weather stations to calculate the field chilling prior to collection (see Table S2).

Growth chamber study

To estimate the effects of chilling, forcing and photoperiod, we used growth chambers with two levels of chilling—with either no additional chilling or 30 days at 4°C for our eastern study, and 30 days or 70 days of chilling at 4°C for our western study (all dark)—after which plants were moved to one of two levels of forcing—a cool regime of 15:5°C and a warm regime of 20:10°C—crossed with two photoperiods of either 8 or 12 hours—for a total of 8 distinct treatments in each study. Our design was similar for both eastern and western species, except for the thermoperiodicity in forcing treatments. In our eastern study, we set the duration of forcing temperatures to correspond with the photoperiod treatments. While this is a widespread and common approach in phenology chamber studies, it leads to a confounding effect in the thermoperiodicity (Buonaiuto et al., 2023). Thus we applied forcing temperatures for 12 hours independent of the duration of the photoperiod treatments in our western study. We had seven to ten replicates per treatment in our eastern study and eight replicates in our western study, with the exception of *Sorbus scopulina*, for which we only had six replicates per treatment. Further details regarding the re-cutting of samples, changing water in containers, and rotation through chambers are presented in Flynn and Wolkovich (2018).

We assessed phenological observations of budburst were assessed using the BBCH scale, adapted for our specific species (Finn et al., 2007). Every 1-3 days, we observed each sample and recorded all phenological stages up to full leaf expansion (defined as code 17 by Finn et al., 2007). Here, we present the response in budburst date, defined as the day of budbreak or shoot elongation (defined as code 07 by Finn et al. (2007)). In total, our eastern study spanned 82 days, with over 19320 phenological observations. The western study in turn spanned 113 and we made 47844 phenological observations across all samples.

Statistical Analysis

To test for differences in cue responses across species and populations, we used a phylogenetic mixed effect model with partial pooling across species. This approach allowed us to account for both the evolutionary relatedness of our species, and estimate the species level cue responses and differences across sites. Species phylogenetic relatedness was obtained by pruning the Smith and Brown (2018) megatree of angiosperms.

In our model, we included each of our three cues as continuous variables. To account for the greater amounts of field chilling experienced in our eastern study, we converted our chilling treatments to chill portions for each individual population using local weather station data and the chillR package (Luedeling, 2020). Chill portions were calculated for both the chilling experienced in the field prior to

sampling and during the experiment in our 4°C chilling chambers. We also converted forcing temperatures to mean daily temperatures in each treatment to account for differences in thermoperiodicity between the two studies (Buonaiuto et al., 2023). Finally, we z-scored each cue and site using two standard deviations to allow results across parameters to be directly compared (Gelman, 2008). In addition to the overall predictors, we included all two-way interactions between cues and between cues and sites:

$$\begin{aligned} \hat{y}_i \sim & \text{normal}(\alpha_{sp_i} + \beta_{site2} + \beta_{site3} + \beta_{site4} + \\ & \beta_{chill_{sp_i}} + \beta_{force_{sp_i}} + \beta_{photo_{sp_i}} + \\ & \beta_{force \times chill_{sp_i}} + \beta_{chill \times photo_{sp_i}} + \beta_{photoperiod \times chill_{sp_i}} + \\ & \beta_{force \times site2_{sp_i}} + \beta_{force \times site3_{sp_i}} + \beta_{force \times site4_{sp_i}} + \\ & \beta_{chill \times site2_{sp_i}} + \beta_{chill \times site3_{sp_i}} + \beta_{chill \times site4_{sp_i}} + \\ & \beta_{photoperiod \times site2_{sp_i}} + \beta_{photoperiod \times site3_{sp_i}} + \beta_{photoperiod \times site4_{sp_i}}) \end{aligned}$$

$$\begin{aligned} \beta_{force_{sp}} & \sim \text{normal}(\mu_{force_{sp}}, \sigma_{force}^2) \\ \dots & \\ \beta_{photoperiod \times site4} & \sim \text{normal}(\mu_{photoperiod \times site4}, \sigma_{photoperiod \times site4}^2) \end{aligned}$$

We included the phylogenetic effect as a variance covariance matrix in the parameterization of the normal random vector:

$$\boldsymbol{\alpha} = [\alpha_1, \dots, \alpha_n]^T \text{ such that } \boldsymbol{\alpha} \sim \text{multi-normal}(\mu_{\boldsymbol{\alpha}}, \mathbf{V}), \quad (1)$$

$$(2)$$

The timing of budburst prior to evolution is represented as root values of the means of the multivariate normal distributions. For more detail on this method of phylogenetic modeling, see Morales-Castilla et al (XXXX).

We used semi-informative priors for each of our model parameters, validating the model code using test data prior to analysis. Our model produced \hat{R} values close to 1, well mixed chains, and n_{eff} values that exceeded 10% of the model iterations. Our model was fitted using the Stan language (Stan Development Team, 2018) using the rstan package in R (R Development Core Team, 2017) (version).

We present model estimates, relative to baseline conditions of low chilling, low forcing, and short photoperiod treatments, and—when relevant—for the Smithers population. We present results as means and 90% posterior uncertainty intervals from our Bayesian phylogenetic models.

Results

On average, we observed species budburst 28.1 days (uncertainty interval: 45, 15.2) after the start of our forcing and photoperiod treatments (all estimates are given as mean \pm 90 uncertainty intervals,

henceforth ‘UI,’ and budburst dates given as relative to baseline conditions, see methods for more details). Some species, such as *Aronia melanocarpa*, budburst as early as day 13.6, and as late as day 52.1 for *Quercus velutina*. The difference in timing of our earliest and latest species spanned a period of 38.5 and 30.7 days for our eastern and western species respectively. Overall we found species budburst to be strongly phylogenetically structured (λ of 0.8, UI: 0.9, 0.6) and that all cues led to an advance in budburst date. Of the three cues, chilling had the strongest effect (-15.2, UI: -17.3, -13.2) and photoperiod the weakest (-3.6, UI: -4.3, -3.0, Fig. 1). But we found a strong, positive interaction between forcing and chilling (9.1, UI: 7.6, 10.5). This results in subadditive effects between chilling and forcing cues, with low chilling being offset by high forcing conditions (Fig. 2 and see Table 1 for model output).

Overall we found population effects were small compared to differences between cues. Across all species, there was considerable overlap in the responses of our four populations to each cue (Fig. 1). As such, low forcing, chilling, and shorter photoperiods all produced later estimated budburst dates compared to higher temperature and longer photoperiod treatments across all populations (Fig. 3 a-c, shown for baseline conditions) and across transects (Fig. 3, see Table 1 for model output). Populations were similar in their overall budburst dates, but eastern populations were marginally earlier (34.1, UI: 40.7, 27.1) compared to our western populations (46.5, UI: 52.7, 39.4, see also Table 1 for model output).

Budburst of individual species showed distinct differences in their timing and the relative importance of cues. Both chilling and forcing responses varied with budburst, with later budbursting species having slightly stronger responses to each cue respectively (Fig. 4). Despite our expectations that species responses should differ across the canopy layer, with shrubs being less sensitive to cues than tree species, we did not find strong differences across these two functional groups (Fig. S3). Many of our earliest species were shrubs with weaker responses to chilling and forcing, like *Cornus stolonifera*. But 36.7% of our shrubs species, including *Menziesia ferruginea* and *Symphoricarpos alba*, exhibited the opposite response and budburst relatively late (Fig. S3). Similarly for tree species, some matched our predictions for their cue responses, but 23.5% budburst earlier than expected. Specific tree species, such as *Quercus velutina*, did have stronger chilling and photoperiod responses as predicted, and of all our focal species, such as *Fagus grandifolia* produced the strongest photoperiod response. But overall there were no clear differences between the cue responses of trees and shrubs across our four forest communities.

Our results depict the relative importance of phenological cues to variation across species, but we found the contribution of cues to budburst order and timing to be dependent of the magnitude of the cue, and small compared to overall species differences. The rank order of species estimated budburst changed more under high cue conditions than under low cues (Fig S4). The budburst ranking for the three species that occurred in both transects remained relatively consistent across the two communities (Fig. 4), with only *Alnus incana* in our western community experiencing a large change in rank with cues (Fig. S5).

In comparing our earliest and latest budbursting species, we found the differences in the cue responses were relatively small (Fig. 4). This is contrary to our prediction of stronger cue requirements—especially chilling cues—for later budbursting species. For example, an early (shrub) species *Lyonia ligustrina* had chilling and forcing cues of -15.4 and -10.4 respectively, which are comparable to the estimates of -16.6 and -10.4 for the much later tree species, *Quercus alba*. Yet our model predicts these species to have a 15.9 day difference in day of budburst. This highlights how the timing of budburst across species was mostly attributed to species level differences outside of cue responses (Fig. 5). Across the range of cue values, cues explained only between 38.6 and 67.0% of variation for eastern species and between 48.3 and 60.6% for western species, with the intercept explaining the rest (Fig. 5).

Discussion

Quick overview

1. In our study of woody plant budburst phenology across North America, we found the assembly of species' temporal niche to only be partially explained by environmental cues and population level variation.
2. But within a community, there was considerable species level variation.
3. Evidence of a strong phylogenetic structure to the timing of species bb

Temporal assembly across North America

1. While our western and eastern forest communities differ in both their species composition and local environments—only small differences in budburst phenology or cue responses across populations
 - (a) Expected local habitat selection = differing selective pressures across populations = leading to differences in cues responses across populations
 - (b) But we only observed slightly earlier budburst in our eastern populations when compared to western populations (Fig 1).
 - (c) Also found no latitudinal gradients across populations despite differences in local photoperiods (Fig 1)—contrasts previous work in which poleward populations were later (Lieth1974, Zettlemyer2021)
 - (d) Lack of population-level trends suggests the drivers that determine the timing of budburst are not shaped by current geography, at least at our continental scale

Community composition and interspecific variation in phenology

1. We found high variation in species cue responses
 - (a) Species varied in the timing of budburst—early to late budbursting spp span a similar period as natural communities (Maycock1961)—suggesting that our experiment captures a realistic breadth in phenology within our forest communities
 - (b) Generally—all species showed some cue response
 - (c) Cues consistent with previous studies—ie.chilling being strongest, photoperiod weakest, but complex interactions between cues—eg (Heide, 1993; Caffarra and Donnelly, 2011; Flynn and Wolkovich, 2018)
 - (d) May be an important adaptation under future warmer winters when chilling might be insufficient and higher forcing needed to offset it
 - (e) While spp do differ in cue responses, identifying general trends cue responses across assemblages of species remains challenging.
2. Shrub and tree species differ greatly in their physiology—filling different ecological niche—but we found little variation in their cue responses on a whole.
 - (a) Most tree species budburst later than shrubs species—in line with previous work—earlier bb in shrubs compared to trees (Panchen2014, Yu2015)
 - (b) But about a quarter of tree species = earlier budburst dates, more similar to the timing of shrub species, and a third of shrubs budburst at similar times as trees —suggesting more nuance to this than previously found—advantage of us having such a large assemblages of species

- (c) As trees advance phenologically—with earlier canopy closure and reduced light predicted (Donnelly2019)—some less responsive shrubs will experience reduced fitness—but our finding suggests many have similar cue responses and are likely to also advance with warming with the potential to maintain their relative temporal niche space.
- (d) Indicates that cues we think are important for bb may also be selecting for other functional traits—such as traits related to light capture and photosynthesis or nutrient uptake for example
- (e) Suggests we need a more detailed examination of the the underlying mechanism—the fitness of species across different temporal niche = determined by species overall suite of traits—incorporating phenology into a broader trait framework would provide greater insight into the considerable variation observed within each group.

Community assembly in responses to cues versus evolutionary history

1. With continued warming, our ability to predict future spring growth will depend on our understanding of cue responses at the species and community level.
 - (a) Within a community, species varied in their timing of bb by several weeks— but the similarity across populations suggests strong niche conservatism and stabilizing selection with little change in response to local habitats
 - (b) This finding illustrates the complexity of predicting how changing environmental cues will impact individual species within communities—suggests population level variation in environmental cue will have a negligible impact.
 - (c) But can't simply focus on current cues—species current phenotypes are the result of multiple interacting and complex environmental cues that have shaped species over evolutionary time scales (Ackerly2009).
2. In accounting for the effects of species' phylogeny, we found a high level of phylogenetic relatedness among clades of species in the timing of budburst.
 - (a) A high proportion of variation in budburst is explained by species' level effects, with cues only explaining about two-thirds of variation in budburst (Fig 4)
 - (b) Our study applied extreme differences in cues to test for responses—useful for longer-term predictions/parameterizing process based models—but variation due to phylogenetic structure likely underestimated
 - (c) Surprising evolutionary history has such a large effect—given phenology is thought of as a highly plastic trait, with local conditions expected to drive variation in the type and magnitude of cue responses
 - (d) But it indicates we are not accounting for other key drivers of spring phenology—additional cues or functional traits—limiting our predictive abilities.
3. The evolutionary history and ancestral phenotype of a species has the potential to effect their ability to adapt to future climates.
 - (a) Could prevent adaptation to new environmental conditions, if spp are physiologically constrained by their ancestral, optimum phenotype
 - (b) Constraints could skew species assemblages, favouring spp with traits adapted for warmer climates—lower chilling and photoperiod requirements—better able to track and benefit from future climate conditions
 - (c) But could also indicate limited evolutionary time—some temperate species, like *Nyssa* and *Fagus* have had a shorter geologic time in which to adapt to temperatures

- 299 (d) Identifying how and to what extent latent traits contribute to the phylogenetic structur-
300 ing of temperate forest communities = unknown, but important if we are to predict how
301 communities will respond to continued changes in climate.

302 **Predicting budburst phenology under future climates**

- 303 1. Integrating the relative effects of climate, evolutionary history, and ecological process is critical
304 to forecasting future phenologies under climate change.
- 305 2. Requires community wide approaches such as ours—one of the first to include these three types
306 of drivers at a large geographic scale and across forest communities.
- 307 3. Critical bc provides greater diversity of spp and insights into how communities as a whole will
308 respond
- 309 4. Our approach address many of the challenges to predicting species responses to further climate
310 change—complexity of regional and increasingly variable seasonality
- 311 5. And yet, our findings highlight our limited understanding of the factors underlying phenological
312 variation within forest communities and groups of species.
- 313 6. to forecast future changes = we need to identify the key underlying ancestral traits that link
314 species' evolutionary histories and are driving the high spp variability observed within temperate
315 forest communities.

1 Tables and figures

Table 1: Summary output from a phylogenetic Bayesian model in which species are partially pooled and phylogeny is included on the intercept. The model includes photoperiod and site as dummy variables, while the forcing and chilling effect is included as continuous. Parameter estimates were used to estimate day of budburst for sites under low cue conditions and for Smithers only when relevant.

	mean	sd	5%	95%
Intercept	12.51	3.14	7.40	17.60
Phylogenetic effect	0.79	0.12	0.60	0.90
Forcing	-9.55	0.74	-10.70	-8.30
Photoperiod	-3.62	0.41	-4.30	-3.00
Chilling	-15.21	1.25	-17.30	-13.20
Manning Park	2.09	0.36	1.50	2.70
Harvard Forest	-6.04	1.03	-7.80	-4.40
St. Hippolyte	-8.71	0.97	-10.30	-7.10
Forcing x photoperiod	0.23	0.71	-1.00	1.40
Forcing x chilling	9.06	0.90	7.60	10.50
Photoperiod x chilling	-0.67	0.90	-2.20	0.80
Forcing x Manning Park	-1.76	0.77	-3.00	-0.50
Photoperiod x Manning Park	0.58	0.79	-0.70	1.90
Chilling x Manning Park	-0.36	1.60	-3.00	2.20
Forcing x Harvard Forest	3.81	1.22	1.80	5.80
Photoperiod x Harvard Forest	-1.96	0.86	-3.30	-0.60
Chilling x Harvard Forest	9.97	2.03	6.60	13.40
Forcing x St. Hippolyte	5.25	1.19	3.20	7.20
Photoperiod x St. Hippolyte	-2.13	0.84	-3.50	-0.70
Chilling x St. Hippolyte	8.65	1.70	5.90	11.50

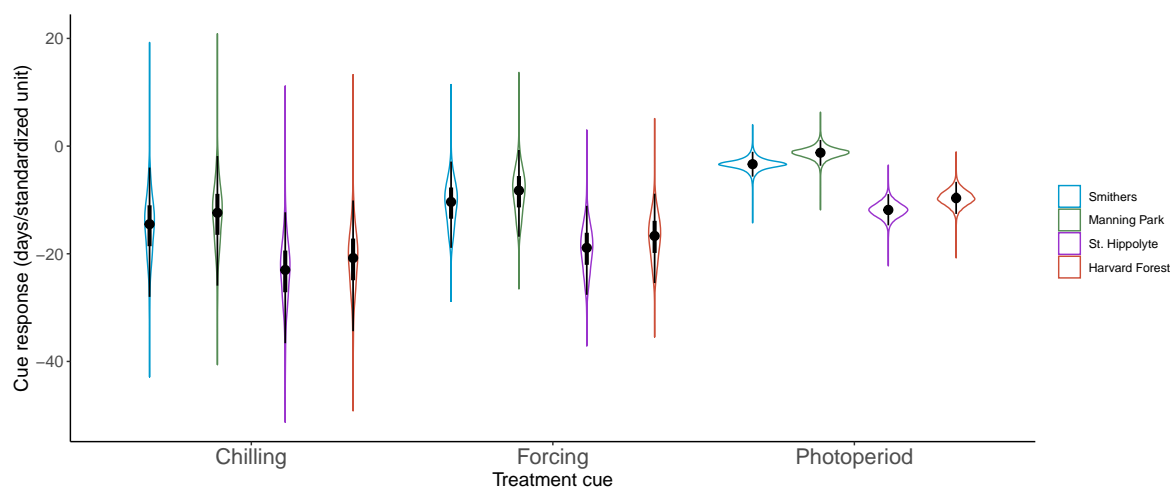


Figure 1: Posterior distributions of estimated chilling, forcing, and photoperiod responses with site level effects for individual populations. Black circles represent the median cue response, thicker black lines the 50% uncertainty interval, while the thinner black line the 90% uncertainty interval. The y-axis spans the entire range of the data. Cues were z-scored using two standard deviations, see methods.

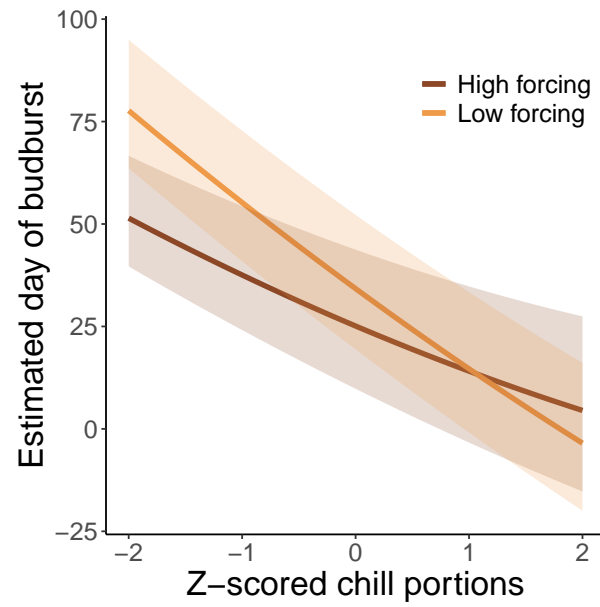


Figure 2: Estimated day of budburst of first bud in response to chill portions and forcing, estimated for our defined baseline conditions. We found a positive interaction between chilling and forcing cues, with high forcing compensating for low chilling to produce earlier budburst, while budburst occurred earlier under low forcing following high chilling.

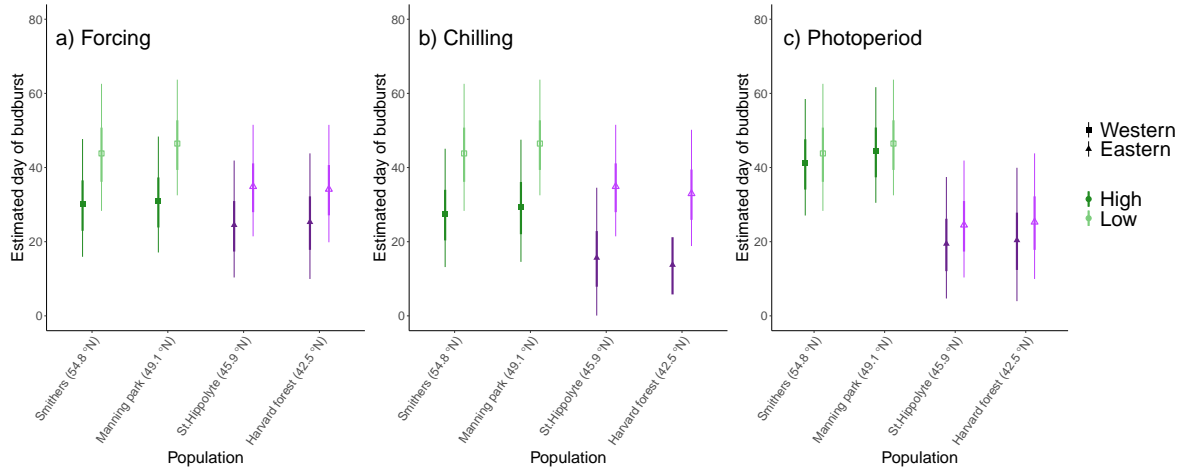


Figure 3: Estimated day of budburst of first bud in response to (a) forcing across populations under low chilling conditions and short photoperiods, (b) chilling across populations under low forcing and short photoperiods, and (c) across photoperiods under our baseline forcing and chilling conditions for species sampled from our four populations. The thin error bars represent the 90% uncertainty interval, while the thicker error bars represent the 50% uncertainty interval.

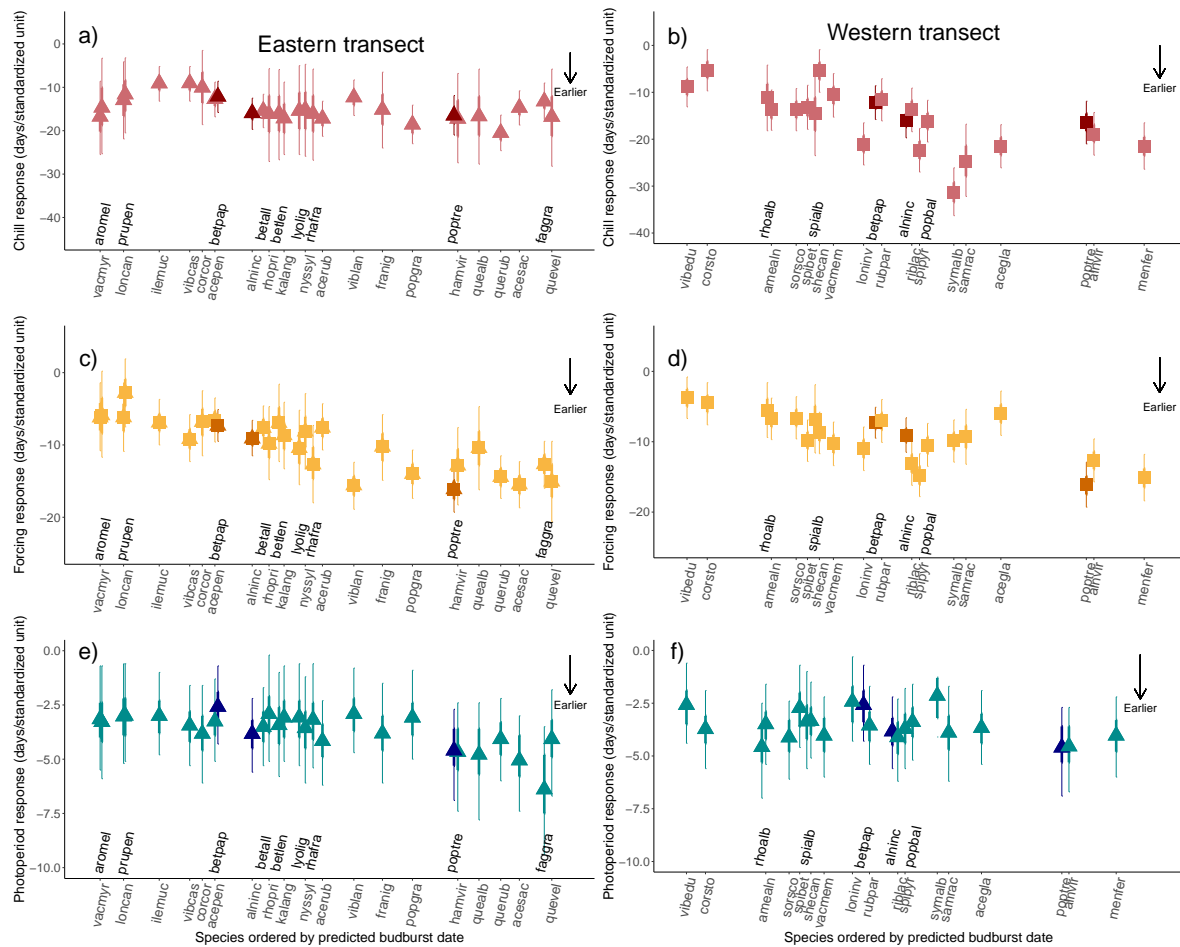


Figure 4: Estimated species' chilling (a,b), forcing (c,d) and photoperiod (e,f) responses ranked by increasing estimated budburst dates for both the eastern (a, c, e) and western (b, d, f) populations. Cues are plotted on differing y-axis to better depict species differences. For each species, the thinner error bars represent the 90% uncertainty interval, and thicker bars represent the 50% uncertainty interval. Species that occur in both the eastern and western transects are depicted using darker colours. Each cue was z-scored, using two standard deviations, to make the results across cues directly comparable, see methods.

References

- Buonaiuto, D. M., E. M. Wolkovich, and M. J. Donahue. 2023. Experimental designs for testing the interactive effects of temperature and light in ecology : The problem of periodicity. *Functional Ecology* 37:1747–1756.
- Caffarra, A., and A. Donnelly. 2011. The ecological significance of phenology in four different tree species: Effects of light and temperature on bud burst. *International Journal of Biometeorology* 55:711–721.
- Chuine, I. 2000. A unified model for budburst of trees. *Journal of Theoretical Biology* 207:337–347.
- Finn, G. A., A. E. Straszewski, and V. Peterson. 2007. A general growth stage key for describing trees and woody plants. *Annals of Applied Biology* 151:127–131.
- Flynn, D. F. B., and E. M. Wolkovich. 2018. Temperature and photoperiod drive spring phenology across all species in a temperate forest community. *New Phytologist* 219:1353–1362.
- Gelman, A. 2008. Scaling regression inputs by dividing by two standard deviations. *Statistics in Medicine* 27:2865–2873.
- Guy, R. D. 2014. The early bud gets to warm. *New Phytologist* 202:7–9.
- Heide, O. M. 1993. Daylength and thermal time responses of budburst during dormancy release in some northern deciduous trees. *Physiologia Plantarum* 88:531–540.
- Körner, C., and D. Basler. 2010. Phenology Under Global Warming. *Science* 327:1461–1463.
- Luedeling, E. 2020. chillR: Statistical Methods for Phenology Analysis in Temperate Fruit Trees. <https://CRAN.R-project.org/package=chillR>.
- R Development Core Team. 2017. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria .
- Smith, S. A., and J. W. Brown. 2018. Constructing a broadly inclusive seed plant phylogeny. *American journal of botany* 105:302–314.
- Stan Development Team. 2018. RStan: the R interface to Stan. R package version 2.17.3.

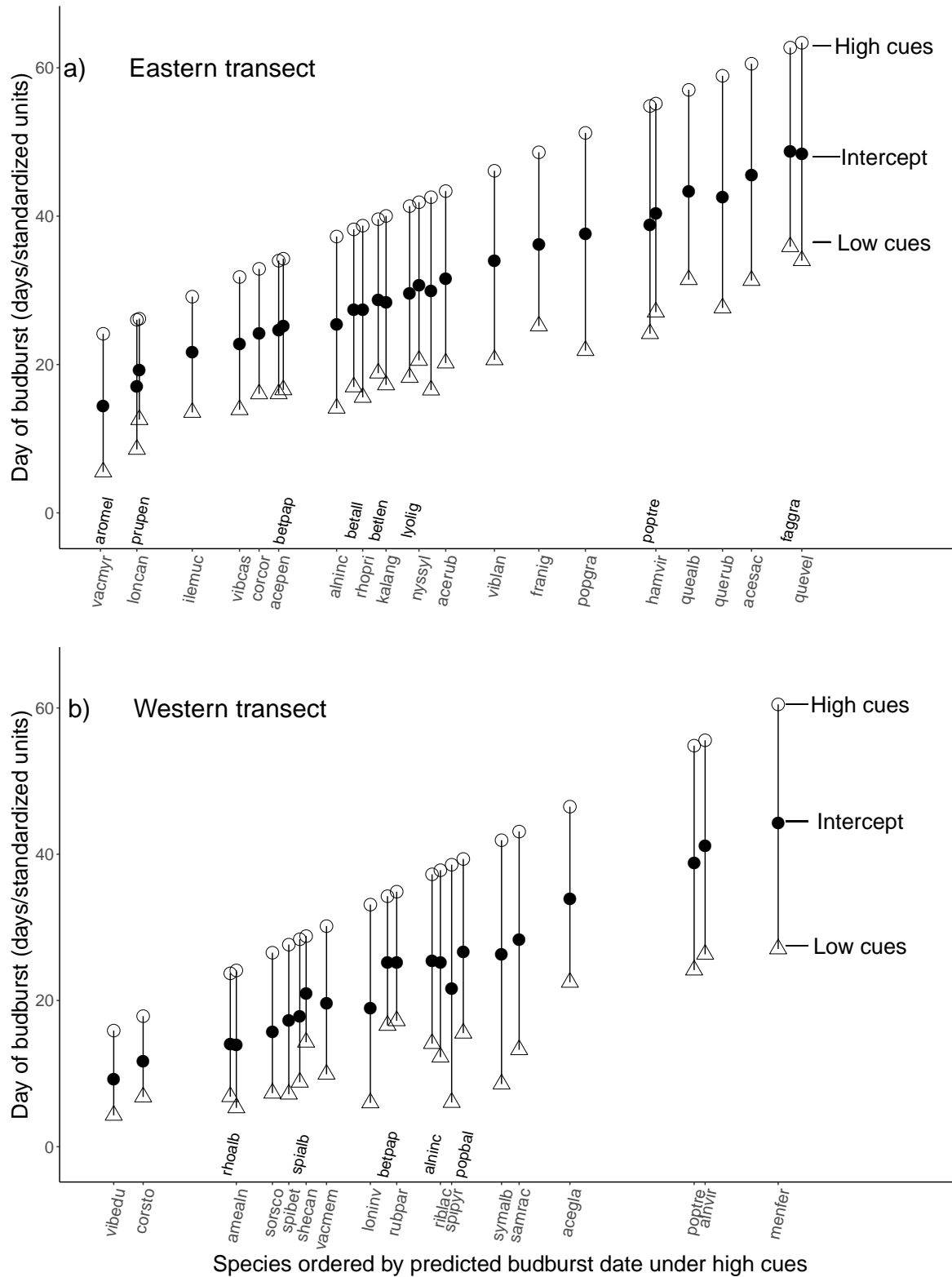


Figure 5: Comparisons of our estimated budburst based on the full model (intercept plus all cues), shown in white, versus the intercepts only (without cues), shown in black, ordered according to differences in species estimated budburst dates for both the eastern (a) and western (b) populations under high cue conditions, depicted as circles, and low cue conditions, depicted as triangles.