

Genotypic variation in a foundation tree results in heritable ecological network structure

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ABSTRACT

Elucidating the genetic basis to ecological network structure is fundamental to understanding evolution in complex ecosystems. Although previous work has demonstrated that genetic variation can influence food webs and trophic chains, we are unaware of a study that has quantified the heritability of network structure of a foundation species associated community. To examine this, in a 20+ year old common garden we observed epiphytic lichens associated with narrowleaf cottonwood (*Populus angustifolia*), a riparian ecosystem foundation tree species. We constructed and conducted genetic analyses of signed, weighted, directed lichen interaction networks on individual trees. We found three primary results. First, genotype identity significantly predicted lichen network similarity; i.e., replicates of the same genotype supported more similar lichen networks than different genotypes. Second, broad sense heritability estimates showed that plant genotype explained network similarity ($H^2 = 0.41$), degree ($H^2 = 0.32$) and centralization ($H^2 = 0.33$). Third, of several tree phenotypic traits examined, bark roughness was both heritable ($H^2 = 0.32$) and significantly predicted lichen network similarity ($R^2 = 0.26$). These results support a mechanistic, genetic pathway from variation in a heritable tree trait to ecological network structure and demonstrate that evolution has the potential to act at the community level to influence not only abundances of organisms but also interactions at the scale of entire networks.

Keywords: networks | heritability | community | genetics | lichen | cottonwood | *Populus* | common garden

INTRODUCTION

Evolution occurs in the context of complex ecological networks. Community genetics studies have shown that genetic variation in foundation species, which have large effects on ecosystems by modulating and stabilizing local conditions (Ellison et al., 2005), plays a significant role in defining distinct communities of interacting organisms: such as endophytes, pathogens, lichens, arthropods, and soil microbes (Busby et al., 2015; Barbour et al., 2009; Lamit et al., 2015a). Multiple studies have now demonstrated that genetic variation influences numerous functional traits (e.g., phytochemical, phenological, morphological)

29 that in combination result in a multivariate functional trait phenotype (Holeski et al., 2012) in which
30 individual plant genotypes support different communities and ecosystem processes (Bailey et al., 2009;
31 Whitham et al., 2012). Recently, the importance of genetic variation in structuring ecological systems
32 was reviewed, and not only were many instances of strong genetic effects found in many ecosystems but
33 the effect of intraspecific variation was at times greater than inter-specific variation (Des Roches et al.,
34 2018). There is now evidence to support that selection occurs among groups of species (Wade, 2007)
35 and that genetic variation and phylogenetic relatedness contribute to variation in community assembly
36 (Crutsinger, 2016) and species interactions (Whitham et al., 2006; Bailey et al., 2009; Moya-Laraño,
37 2011). These evolutionary dynamics have the potential to shape the structure of ecological interaction
38 networks (Rezende et al., 2007; Guimarães et al., 2007; Gómez et al., 2009).

39 Empirical and theoretical work in network ecology and evolutionary biology point to the need for
40 examinations of the genetic basis of ecological network structure. Analyses of ecological networks have
41 demonstrated that indirect effects can lead to self-organization, producing sign-changing, amplifying
42 and/or dampening effects (Fath and Patten, 1998; Newman, 2006; Sole and Bascompte, 2006) and other
43 studies have demonstrated that indirect effects of interactions among species can lead to network structures
44 that amplify or dampen the effects of selection, such as the formation of star-like structures in which there
45 is a “central” species or core group of species (Lieberman et al., 2005). Also, work by Toju et al. (2014,
46 2016, 2017) observed consistent patterns of centralized interactions of species modules (i.e., groups of
47 species that interact more strongly within their group than with other species) focused around hubs of
48 plant-fungal interactions. In other words, a small number of plant and fungal symbionts tended to have
49 disproportionate numbers of interactions with other species and likely are the drivers in determining
50 community assembly, structure and dynamics. Interspecific indirect genetic effects (IIGE) theory (*sensu*
51 Shuster et al. (2006)) in evolutionary biology also points to the importance of studying the genetics of
52 interaction network structure. Genetically based differences in network structure among individuals
53 can be acted upon by natural selection when there are fitness consequences of different networks of
54 IIGEs, leading to community evolution per Whitham et al. (2020) and, by extension, interaction network
55 evolution. For example, although the analysis was of abundances rather than interaction networks, Gehring
56 et al. (2014, 2017) found that the mycorrhizal communities on the roots of drought tolerant and intolerant
57 trees are dominated by different orders of ectomycorrhizal fungal mutualists that also differ in the benefits
58 they provide that enhance tree performance. Because drought tolerant genotypes are three times more
59 likely to survive record droughts, selection acts both on the tree and its fungal community and with
60 increased drought the community phenotype has changed over time. Also, in an antagonistic interaction
61 context, Busby et al. (2015) found that with the addition of a damaging leaf pathogen to cottonwoods in a
62 common garden, the impacts of these strong interactors results in a different and diminished community
63 of arthropods relative to control trees. These examples collectively support the possibility that selection
64 acting on the tree may alter the network structure of associated communities in which different networks
65 are more likely to survive drought and pathogen outbreaks, respectively. Regardless of whether the IIGE
66 is unilateral (i.e., tree affects the community) or reciprocal (i.e., the community also affects the relative
67 fitness of the tree), selection at the level of the tree population or its community, or both, can change
68 network structure and alter community dynamics (Whitham et al., 2020).

69 In this context, the “genetic similarity rule” of community genetics provides a useful framework
70 we can apply to interaction networks at the nexus of ecological and evolutionary dynamics. In a study
71 combining experimental common gardens and landscape-scale observations of interactions between
72 *Populus* spp. (cottonwoods) and arthropods, Bangert et al. (2006) observed that individual genotypes
73 that are more genetically similar will tend to have similar phytochemical traits and thus tend to have
74 similar interactions with other species. Although this is likely to have consequences for interactions and
75 network structure, studies in the network ecology literature generally do not include a genetic component
76 (Lau et al., 2017) and community genetics studies have primarily focused on community composition
77 in terms of the abundance of species (Des Roches et al., 2018). Some studies have examined the effects
78 of genetic variation on trophic chains in plant-associated communities (including *Populus*, *Solidago*,
79 *Oenothera*, *Salix*) (Bailey et al., 2005; Johnson, 2008; Smith et al., 2011, 2015; Barbour et al., 2016) and
80 generally found that increasing genotypic diversity leads to increased trophic complexity. We are aware
81 of only two studies that explicitly examined the effect of genotypic variation on interaction networks
82 between tree individuals and associated herbivores using ecological network metrics (Lau et al., 2016a;
83 Keith et al., 2017). Both found that genotypic diversity generates increased network modularity (i.e.,

84 compartmentalization); however, both were examining networks at the scale of forest stands, rather than
85 networks associated with individual trees; therefore, neither was able to observe replicated networks in
86 order to statistically test for genetic effects on network structure and quantify the genetic component (i.e.,
87 heritable variation) in network structure.

88 Here, we investigate how genetic variation in a foundation tree species determines the structure of a
89 network of interactions among a community of tree associated lichens. We used a long-term (20+ years),
90 common garden experiment with clonally replicated *Populus angustifolia* individuals of known genetic
91 identity (Martinsen et al., 2001). We focused on a community of epiphytic lichen species, as previous
92 research has demonstrated significant compositional effects of genotypic variation on lichen in this system
93 (Lamit et al., 2011, 2015b,a) and epiphytic organisms in other systems (Winfrey et al., 2011; Zytynska
94 et al., 2011). Applying a probability-theory based network modeling approach (Araújo et al., 2011), we
95 constructed a set of interaction network models for the lichens associated with individual trees. Using
96 these models, we then examined the genetic basis of the structure of these ecological networks via several
97 network metrics that measure different aspects of network structure at the scale of individual species (i.e.,
98 nodes) or the entire network observed on each tree genotype. Given the potential importance of focal or
99 “central” nodes (e.g., species) for determining network dynamics (Lieberman et al., 2005), we focused on
100 network metrics that measure centrality for individual species and centralization for whole networks. Both
101 of these metrics measure how much a species is connected in the network relative to other species. As
102 there is a preponderance of evidence that in natural systems evolution occurs in communities comprised of
103 networks of interacting species (Lau et al., 2016a; Keith et al., 2017; Thompson, 2013; Bascompte et al.,
104 2006), we set out to test two hypotheses. First, per the community similarity rule (Bangert et al., 2006)
105 and IIGE theory (Whitham et al., 2020), we hypothesize that trees of the same genotype (i.e., clones) will
106 support more similar lichen interaction networks relative to individuals of other genotypes. In other words,
107 epiphytic lichen network structure is heritable, which can be calculated via comparisons of within and
108 among group variation in network structure. Second, heritability of lichen network structure is the result
109 of underlying phenotypic covariation in tree traits important to interactions between trees and lichens and
110 among lichens. Evidence that such trait covariance generates variation in interactions among community
111 members provides an intermediate genetics-based mechanism for the underlying factors determining
112 lichen distribution and abundance. In combination, evaluating these two hypotheses is fundamental to
113 understanding variation and dynamics of network structure and evolution.

114 MATERIALS AND METHODS

115 Study System

116 The study was conducted along the Weber River, UT (USA), which is a cottonwood (*Populus* spp.)
117 dominated riparian ecosystem. Although two native species, *Populus angustifolia* (James) and *Populus*
118 *fremontii* (S. Watson), occur here and are known to hybridize, in order to focus on intra-specific genetic
119 variation we only sampled pure or advanced generation back-crosses of *P. angustifolia*. Bark lichens
120 have been intensively sampled in this system and provide an ideal community in which to observe and
121 model interaction networks, as their sessile nature permits accurate identification of individuals and their
122 highly localized, direct contact interactions and slow population turnover rates facilitate the assessment of
123 interactions among lichen species on individual trees (Lamit et al., 2015b).

124 A long-term, common garden experiment was used to isolate the effect of tree genotype from the
125 effect of the localized microenvironment associated with each individual and spatial autocorrelation.
126 Established in 1992, asexually propagated clones of genotyped *P. angustifolia* individuals were obtained
127 from wild collections and planted in a fully randomized design at the Ogden Nature Center, Ogden, UT.
128 From the population of established individuals in the common garden, we sampled a total of ten genotypes,
129 replicated between 3 and 8 times each. These individuals comprised a set of tree genotypes with lichen
130 communities that have been well studied by previous investigations (Lamit et al., 2011, 2015b,a).

131 Bark Lichens and Trait Observations

132 We conducted a modified sampling procedure originally developed by Lamit et al. (2015b). On each
133 tree, presence or absence of each lichen species was assessed in a total of 50 1 cm² cells arrayed in
134 a 10 cm² checkerboard pattern. Given the small size and sessile nature of lichens, we were able to
135 rapidly assess lichen interactions by quantifying thalli of different species occurring in close proximity.
136 Sampling was restricted to the northern aspect of the trunk to maximize the abundance of lichens and

control for the effect of trunk aspect. Two adjacent 100 cm^2 quadrats centered at 50 cm and 95 cm from ground level were sampled (Fig 1). The observed lichen community included: *Athallia holocarpa*, *Candelariella subdeflexa*, *Myriolecis hagenii*, *Rinodina freyi*, *Physcia adscendens*, *Physciella melanachra*, *Physcia undulata*, *Xanthomendoza galericulata*, *Xanthomendoza montana*. Several other species were not observed in the present study but are known to occur in this region: *Melanohalea elegantula*, *Melanohalea sublivacea*, *Phaeophyscia ciliata* and *Phaeophyscia orbicularis*.

The cell size and checkerboard sampling pattern was chosen to isolate the individuals in each cell. In a survey of *Xanthomendoza galericulata* in the common garden, we had observed a median thallus size of $0.12 \pm 0.001\text{ cm}^2$ (1 S.E.) (Supporting Information, Fig. 1). This expected thallus size formed the basis for our sampling design, such that lichen observations were spatially independent of thalli present in other cells but exposed to similar micro-environmental conditions created by the bark and the location of the sampling area on an individual tree. Therefore, we were confident in treating the cell-wise observations in quadrats as independent with respect to lichen-lichen interactions.

We quantified tree traits inside or in close proximity to the lichen quadrats. We assessed bark texture/structure in the quadrat as the percent of 1 cm^2 cells with “rough” bark, i.e., bark with fractured surface. In addition, we also examined several bark chemistry traits by taking bark samples immediately adjacent to each quadrat. We used previously collected phytochemical data from Lamit et al. (2011), including the concentration of condensed tannins, carbon and nitrogen. Additionally, we quantified bark pH for each tree. Bark samples were collected by excavating adjacent to the quadrat down to a depth of 2 mm. Bark pieces were air dried for storage and later processing. Samples were prepped for pH measurements by crushing with a mortar and pestle until all pieces were approximately 0.5 cm in diameter, creating equivalent surface areas among samples. 0.5 g of crushed bark was placed in a 15 ml Falcon collection tube with 5 ml of deionized water. Tubes were capped and let sit for 24 hrs prior to pH measurement with a SevenGoTM pH meter (Mettler Toledo).

Lichen Network Modeling

For each tree, the repeated observations of lichens were used to construct replicated interaction networks, i.e. one for each individual tree. Unipartite networks were generated using the conditional probabilities of each species pair, i.e., the probability of observing one species given an observation of another species $P(S_i|S_j)$, based on the method developed by Araújo et al. (2011). To calculate conditional probabilities, we quantified the individual probabilities of species occurrences $P(S_i)$ and the joint probability of co-occurrences $P(S_i, S_j)$ using the frequencies of each species and their co-occurrences. Using the axioms of probability, we can calculate the conditional probabilities of each species pair as $P(S_i|S_j) = \frac{P(S_i, S_j)}{P(S_j)}$. This yields a matrix that could possibly be asymmetrical, i.e., $P(S_i|S_j)$ does not have to be equal to $P(S_j|S_i)$. Also, the diagonal, $P(S_i|S_i)$, is equal to one for all species present and zero for species that were not observed in any cell.

We then applied an analytical procedure to remove non-significant links between species. This procedure determines if the joint probability of a species pair (i.e., $P(S_i, S_j)$) is different from zero (Fig. 2). Here, a confidence interval $CI_{95\%}$ is calculated as $CI_{95\%} = E[S_i S_j] * Z_{95\%} * \sqrt{V(S_i S_j)}$, where the expected frequency of co-occurrences $E(S_i S_j)$ is the total number of cells surveyed (N) times the independent probabilities of each species $P(S_i) * P(S_j)$, $Z_{95\%}$ is the Z-score for 95% from a Z-distribution and the expected variance of $E(S_i S_j)$ is the total number of cells times the expected probability of $S_i S_j$ and its compliment (i.e., $V(S_i S_j) = N * E[P(S_i, S_j)] * (1 - E[P(S_i, S_j)])$). If the observed number of co-occurrence falls within the confidence interval, the joint probability $P(S_i, S_j)$ is concluded to be equal to the product of the individual probabilities (i.e., $P(S_i)P(S_j)$), and the conditional probability reduces to the individual probability of that species $P(S_i)$. Therefore, unless the co-occurrence of a species pair falls outside the confidence interval, the probability that the observation of one species given the other is no different than simply observing that species alone. This enables us to remove links from a given network by re-scaling the resulting conditional probabilities through subtraction of the individual probabilities from the conditional probabilities (i.e., how different the conditional probability is from the independent probability), which makes any species with a non-significant conditional probability zero.

The resulting matrix ($\mathbf{D} = D_{ij}$) can be interpreted as one species’ impact on another with zero being no effect and values less than or greater than zero being negative and positive effects, respectively. We will refer to \mathbf{D} as a signed, weighted interaction matrix. As such, \mathbf{D} has the properties that it can be asymmetric (i.e., D_{ij} does not necessarily equal D_{ji}) and that it scales between -1 and 1, and, therefore, does not have

191 the mathematical properties of a probabilistic network (Poisot et al., 2016). Also, as the method does not
192 track individuals within species; therefore, the “intra-specific” observations are the same species being
193 counted across the cells of the grid and a resulting $D_{ii} = 0$. In the context of these networks, asymmetry
194 and positive/negative valued connections are distinct quantities. In-coming and out-going connections can
195 be interpreted as “influenced by” and “influenced”, respectively; while positive and negative are within
196 this study interpreted as one species increasing or decreasing, respectively, the probability of another
197 species’ occurrence.

198 **Analyses, Software and Data**

199 To quantify the structural variation of lichen networks we calculated several metrics at both the level of
200 node and whole networks. Although there are many other network metrics, for the sake of simplicity
201 we focus on a subset that represent the primary interesting features of network structure, see Lau et al.
202 (2017). We calculated the number of interactions or “links” in each network (degree), which provides
203 a measure of the size of the network (Lau et al., 2016a; Borrett and Lau, 2014). We also calculated the
204 centralization of each network using Freeman’s centrality, which measures the evenness of the distribution
205 of interactions among the species in the network, using the `sna` package (Butts, 2019). In a network
206 with low centralization species have similar strengths and numbers of interactions. A network with high
207 centralization tends to have one or a small number of species that interact with other species. We used a
208 related function to calculate the centrality of each species (i.e., node level centrality) in each network as
209 well. To calculate separate metrics for positive and negative links, as the networks contained not only
210 positive and negative connections but also directional connections (both in-coming and out-going), we
211 calculated the same network metrics for all combinations of these types of connections using recently
212 developed methods for signed, weighted and directed networks (Everett and Borgatti, 2014) using the
213 `signnet` package (Schoch, 2020).

214 We used a combination of parametric and non-parametric, permutation based frequentist statistical
215 analyses to test for the effects of genetic variation on lichen communities and their interaction networks.
216 To assess the effect of genotype on traits as univariate response variables (including the metrics of
217 network structure), we used additive, random effects models with Restricted Maximum Likelihood
218 (REML) conducted in R via the `lme4` and `RLRsim` packages (Bates et al., 2015; Scheipl et al., 2008).
219 To conform to test assumptions, traits were root transformed with the exception of condensed tannin
220 concentration and carbon-nitrogen ratio, which were rank and \log_{10} transformed, respectivley. Differences
221 in node level centrality among species was tested using ANOVA and Tukey-HSD multiple comparison
222 tests. Correlations among trait variables and network metrics were quantified and tested using linear
223 correlations of Pearson’s r . For multivariate response variables, such as lichen community composition
224 and network structure, we used distance based multivariate statistical approaches. To quantify the
225 similarity of lichen networks among individual trees, we calculated the pairwise Euclidean distance of
226 the \mathbf{D} interaction matrices among all trees (Newman, 2010). For community composition we applied
227 Bray-Curtis similarity to a matrix of species abundances obtained by aggregating the gridded observations
228 by summing over the binary cell-wise species presence-absences. To test for the effects of genotype and
229 other predictor variables on community and network similarity we conducted Permutational Analysis
230 of Variance (PERMANOVA) with `vegan` (Oksanen et al., 2019) using 100000 permutations. For
231 visualization of multivariate patterns, we used Non-metric Multi-Dimensional Scaling (NMDS) (Goslee
232 and Urban, 2007) to produce dimensionally reduced ordinations of these multivariate responses and
233 fitted vectors for continuous predictor variables to the ordinated values (Oksanen et al., 2019), using
234 100 random initial configurations with a maximum of 1000 iterations and a change in stress threshold
235 of less than 10^{-12} . This was repeated for one to four dimension configurations, and the configuration
236 with the lowest dimensionality and an unexplained variation less than 10% was selected. For all tests
237 where genotype was used as a predictor, we quantified the heritability of the response variable. Because
238 the trees in the garden were clonal replicates of each genotype, we calculated broad-sense heritability,
239 which is the genotypic variance divided by the total phenotypic variance (Conner and Hartl, 2004),
240 which can be interpreted as a measure of the phenotypic variance due to genotypic variation. All
241 analyses were conducted using R version 4.0.2 (R Core Team, 2020). Code and data for the project
242 are openly available as a reproducible workflow using `drake` (Landau, 2018) archived via Zenodo
243 <https://doi.org/10.5281/zenodo.4581639>.

244 **RESULTS**

245 In support of our first hypotheses, we found that tree genotype influenced lichen network structure and
 246 that multiple lichen network metrics were heritable. Tree genotype significantly predicted the structural
 247 similarity of lichen networks and, overall, network-level metrics responded significantly to tree genotype,
 248 including network degree and centralization including both in-coming and out-going links or when
 249 separated into in-coming only or out-going only (Table 1, Fig. 4). Metrics including only positive
 250 links also showed a significant effect of tree genotype, including positive degree and positive in-going
 251 centralization. Metrics calculated with negative links were not significant, including degree (negative)
 252 and both in-coming (negative) and out-going centralization (negative). Interestingly, although network
 253 similarity and multiple network metrics were significantly predicted by tree genotype, we did not observe
 254 a significant genotypic effect for community composition ($\text{Pseudo-}F_{9,27} = 0.751$, $R^2 = 0.20$, $p\text{-value} =$
 255 0.888).

	Response	df	RLRT	H2	p-value
Lichen Network Similarity	9	3.5821	0.41	0.0537	
Degree	9	3.5175	0.32	0.0255	
Degree (positive)	9	3.6925	0.32	0.0229	
Degree (negative)	9	0.0327	0.03	0.3859	
Centralization	9	4.0444	0.33	0.0184	
Centralization In-Degree	9	4.4812	0.35	0.0142	
Centralization In-Degree (positive)	9	3.9852	0.33	0.0190	
Centralization In-Degree (negative)	9	0.3304	0.11	0.2508	
Centralization Out-Degree	9	3.8615	0.32	0.0205	
Centralization Out-Degree (positive)	9	3.5585	0.31	0.0248	
Centralization Out-Degree (negative)	9	0.0862	0.05	0.3446	

Table 1. Genotypic effects on the associated lichen network structure. *RLRT* is the statistic from the restricted likelihood ratio tests.

256 The genetic response of network centralization was driven by variation in *Athallia holocarpa*. Cen-
 257 tralinity varied significantly among species ($F_{8,324} = 7.99$, $R^2 = 0.16$, $p\text{-value} < 0.0001$). The node-level
 258 metrics for *A. holocarpa* displayed the strongest response to tree genotype with high levels of heritability
 259 of positive centrality for both the in-coming ($RLRT = 3.61$, $H^2 = 0.32$, $p\text{-value} = 0.0240$) and out-going
 260 ($RLRT = 3.13$, $H^2 = 0.30$, $p\text{-value} = 0.0327$) perspectives but not for either negative centrality metrics
 261 in-coming ($RLRT = 0$, $H^2 = 0$, $p\text{-value} = 1$) or out-going ($RLRT = 0$, $H^2 = 0$, $p\text{-value} = 0.4543$). None
 262 of the other species' centralities showed a genotypic response (Supporting Information, Fig. 2) with
 263 the exception of *X. montana* ($RLRT = 2.92$, $H^2 = 0.32$, $p\text{-value} = 0.0375$); however, the centrality of *X.*
 264 *montana* was much lower overall relative to *A. holocarpa* and the variation in *X. montana* centrality was
 265 restricted to two genotypes (Fig. 5).

266 In support of our second hypothesis, analysis of trait covariation revealed that genotype indirectly
 267 influenced lichen network centralization via genetically based variation in bark roughness. The percent
 268 cover of rough bark ($RLRT = 4.8526$, $H^2 = 0.3221$, $p\text{-value} = 0.0113$) and condensed tannins ($RLRT =$
 269 3.0522 , $H^2 = 0.3205$, $p\text{-value} = 0.0343$) both displayed significant responses to tree genotype. None of
 270 the other bark traits, pH ($RLRT = 0.00$, $H^2 = 0.00$, $p\text{-value} = 1.0000$) or carbon-nitrogen ratio ($RLRT$
 271 $= 0.0000$, $H^2 = 0.0000$, $p\text{-value} = 1.0000$), showed a significant response to tree genotype and none
 272 other than bark roughness was correlated with network similarity (Table 2); therefore, we focused our
 273 subsequent analyses on the indirect effect of genotype on lichen network structure via bark roughness. We
 274 found that bark roughness was significantly correlated with network similarity and other lichen network
 275 metrics, including negative correlations with overall network degree ($df = 35$, $t = -2.13$, $r = -0.34$, $p\text{-value}$
 276 $= 0.04$) and centralization ($df = 35$, $t = -2.52$, $r = -0.39$, $p\text{-value} = 0.02$). In other words, trees with more
 277 similar levels of bark roughness tended to have lichen interaction networks with similar structure (Fig. 6).
 278 To quantify the genetic bases of this effect of bark roughness on network structure, we used the residual
 279 values from regressions of the network metrics and bark roughness in subsequent tests of the effect of
 280 tree genotype and found no significant effect of tree genotype for either degree ($RLRT = 0.00$, $H^2 = 0.00$,
 281 $p\text{-value} = 1.0000$) or centralization ($RLRT = 0.00$, $H^2 = 0.00$, $p\text{-value} = 1.0000$), and, thus, the bulk of the

282 genetically based variation in the network metrics can be explained by bark roughness.

	df	SS	R2	Pseudo-F	p-value
Bark Roughness	1	20850.09	0.26	12.9234	0.0101
Condensed Tannins	1	5993.66	0.07	3.7150	0.0813
pH	1	1273.19	0.02	0.7892	0.3712
Carbon:Nitrogen Ratio	1	3896.18	0.05	2.4150	0.1890
Residual	32	51627.33	0.64		
Total	36	80993.59	1.00		

Table 2. PERMANOVA Pseudo-F Table of lichen network similarity response to bark traits.

283 DISCUSSION

284 Ecological and Evolutionary Importance of Network Heritability

285 Although previous studies have examined aspects of networks, such as trophic complexity (Barbour et al.,
286 2016) and forest stand level interaction network structure (Lau et al., 2016b; Keith et al., 2017), this
287 is the first study that we are aware of to examine the heritability of network structure with replicated
288 networks at the genotype scale. Previous work in the evolution of ecological networks have primarily
289 focused on macro-evolutionary dynamics (Rezende et al., 2007; Weber et al., 2017; Valverde et al., 2018;
290 Harmon et al., 2019) or have been simulation based individual-level models that integrate intraspecific
291 variation to the species level (Maliet et al., 2020), even though recent syntheses have pointed to the
292 importance of processes operating across scales of organization (Guimarães, 2020). There are two
293 important functional ramifications of genetically based variation in network structure. First, heritability
294 of network structure suggests that some amount of interaction network complexity is determined and
295 therefore could be predicted by genetic identity. Variation in space and time create variation in ecological
296 networks that influences evolutionary dynamics via shifts in ecological dynamics, such as population
297 demographics (Guimarães, 2020). Given that ecosystems are comprised of hundreds and thousands of
298 species, each having a multitude of interactions, the potential to find traction for making predictions in the
299 context of ecological, let alone evolutionary, dynamics seems daunting. The promise of predictability
300 lies in the presence of asymmetries in ecosystems, such as hierarchy created by foundation species via
301 differences in body size and/or life-history strategies (Ellison et al., 2005). Second, heritability (i.e.,
302 genetic determination) means that there is structure in the spatial or temporal variation that is created
303 by individuals of foundation species whose traits are in part determined by underlying trait differences.
304 Although this variation is inherently a function of both genetic and environmental effects (Conner and
305 Hartl, 2004), the community and network-level effects are also a function of the scale of the interaction
306 (Shuster et al., 2006; Lau et al., 2017).

307 Notably, even if the composition of the communities is the same among individuals and genotypes,
308 interactions may not be. We didn't observe compositional differences using the same data from which
309 the lichen networks were derived. If we only had our composition dataset from this study, we would
310 have concluded no response of the lichen community to tree genotype, even though the underlying
311 interactions among lichen species does vary among genotypes. As such differences in network structure
312 could occur without observable differences in species richness or community composition, which have
313 been the primary focus of the majority of community genetics studies (Whitham et al., 2020). Community
314 composition of lichens has previously been observed to be different among tree genotypes in the same
315 experimental garden (Lamit et al., 2011, 2015b). The different results observed in the present study is
316 likely a result of differences in lichen quantification and the tree genotypes observed leading to overall
317 higher abundances of observed lichens to assure the possibility of observing lichen interactions. The
318 previous study used a visual percent cover estimation, unlike the current study, which observed lichens at
319 the scale of 1 cm² cells, which could over-estimate cover depending on the frequency at which actual
320 thallus size was less than 1 cm², as well as both the northern and southern aspects of each tree. These
321 differences do not negate the findings of either study.

322 The present study's finding of differences in network structure without significant compositional
323 differences points to the importance of quantifying how network structure changes in response to genetic
324 variation. Having not observed a compositional effect of tree genotype without measuring the network

325 structure could lead to the conclusion of no genetic effect on the community, even though differences in
326 network structure are leading to altered, local evolutionary dynamics. It is possible that these underlying
327 differences in interactions among lichens could lead to differences in community composition at a future
328 point in time via their effects on species abundances (Shuster et al., 2006); however, this is not needed
329 for evolutionary dynamics to occur via selection that leads to shifts in trait distributions without shifting
330 species abundance distributions, which is possible under stabilizing, disruptive and directional selection
331 (Conner and Hartl, 2004), so long as the relative abundances of each species is imperceptibly changed.
332 Thus, it is imperative that further community genetics research assess or at least be aware of the potential
333 effects of variation in interactions and not just observe species abundances, otherwise community level
334 genetic effects may be underestimated, especially when cumulative interaction effects are taken into
335 account (Borrett et al., 2007, 2010).

336 Furthermore, the demonstration of the heritability of interaction networks, without significant dif-
337 ferences in community composition, provides clear empirical evidence that IIGEs need to encompass
338 the structure of interaction networks. Although IIGE theory provides a quantitative framework within
339 which to approach evolutionary theory at higher levels of biological organization (from populations to
340 communities and ecosystems), this theory has focused on modeling the strong effects of foundation
341 species (Shuster et al., 2006; Whitham et al., 2012, 2020) and has not yet integrated developments from
342 the ecological or evolutionary network theory literature. Thus, it has not developed a way to examine
343 complex interactions among species; however, previous studies have demonstrated this network context is
344 likely to be important, as altering the structure of interaction networks provides a means for genetic effects
345 to be dampened or magnified within the system of interacting species (Smith et al., 2011; Keith et al.,
346 2017). Although such a synthesis necessitates a much greater effort than can be afforded in this paper, it
347 is possible to point to several productive pathways forward. In terms of interaction networks, foundation
348 species are relatively central within the system of interactions, that is their direct and/or indirect effects
349 are greater than other species. So, when the more centralized (foundation) species have genetically based
350 interactions, genetic effects will tend to be propagated and possibly magnified in the community. Here, we
351 found that even though species with relatively higher abundances (e.g., *X. galericulata*) were present in
352 the community, the response of the network was not predominately due to their interactions. In addition,
353 although there were species that tended to be central within the lichen networks, the similarity of networks
354 showed the highest heritability compared to any of the centrality metrics, which suggests that there are
355 aspects to network structure that not solely due to the central species. Taken together, these results point to
356 the importance of considering the impact of network structure and that it is a potentially productive path
357 forward for the development and application of IIGE theory.

358 **Network Structure and Levels of Selection**

359 The demonstration of evolution at any scale requires demonstrating three key elements (Conner and Hartl,
360 2004), which multilevel selection theory posits can occur simultaneously at multiple levels of ecological
361 organization (Whitham et al., 2003, 2020). First, there must be variation in the structure (composition,
362 abundance, species interactions, diversity, interaction network structure) of communities. Second, these
363 differences must be genetically based and heritable in which community structure is passed from one
364 generation to the next. For example, numerous studies show that related individuals tend to support the
365 same communities of insects and microbes, and ecosystem processes of biodiversity, nutrient cycling and
366 stability, whereas unrelated individuals support more different communities and ecosystem processes
367 (Bangert et al., 2006, 2008; Barbour et al., 2009; Whitham et al., 2020). Importantly, the current study
368 shows that networks are also heritable traits that greatly increases its utility as a community phenotype
369 that selection can act upon. Third, selection must act on these differences to favor some communities over
370 others leading to change over time (i.e., community evolution). The differential survival and performance
371 of individual tree genotypes will simultaneously result in selection occurring on the lichen community
372 and network structure that it supports. Since our findings show that networks are heritable, another metric
373 of community evolution would be to quantify how networks change over time in response to an invasive
374 species, climate change, or some other agent of selection, which represents a frontier for future ecological
375 network studies.

376 The current study shows the utility of considering the spatial and temporal scales of ecological and
377 evolutionary dynamics. In line with previous empirical studies (Bangert et al., 2006; Zook et al., 2010;
378 Zytynska et al., 2012), we found that the assembly of ecological networks can have a measurable genetic

379 basis depending on the spatial scale of interactions, due in part, to asymmetries in size and longevity of
380 organisms. Intra-specific, genotypic diversity among cottonwood trees appears to be a major factor in the
381 creation of meta-communities of lichens on individual trees that form interaction modules with different
382 dynamics. When communities are comprised of individuals whose habitat is primarily determined by
383 another organism, these communities inherently form modules within the larger ecosystem, as they tend to
384 interact more with each other than with other individuals (Lau et al., 2017). Our study demonstrates that
385 the localized environmental differences determined by the genetic variation within a single tree species
386 can not only impact community composition, as repeatedly demonstrated in other community genetics
387 studies (Whitham et al., 2006; Des Roches et al., 2018), but also shape the structure of interactions among
388 individuals. Some network structures are likely to be more stable, either in response to disturbance or via
389 self-organized dynamics. For example, centralized networks, although more efficient, are theorized to be
390 more susceptible to targeted “attacks” in the terminology of defense networks. As mentioned previously,
391 one class of networks that are theorized to have amplifying effects on networks have centralized “star”
392 shapes with one or a few species at the center and radiating interactions out from the central core
393 (Lieberman et al., 2005). This is structurally what we have observed with the networks that tend to occur
394 on some of the genotypes in our study, i.e., the more centralized networks. It is likely that these networks
395 could function as hot-spots of evolutionary dynamics resulting from the amplifying effect the centralized
396 network structure found on that tree genotype, as multiple studies have found significant impacts of the
397 removal of foundation species in different systems (Keith et al., 2017; Des Roches et al., 2018).

398 The importance of the scale of network organization to create hierarchical structure (Guimarães, 2020)
399 and the potential for foundation species to create this structure in the vast majority of ecosystems (Ellison
400 et al., 2005; Whitham et al., 2006) suggests that future work would be aided by determining species with
401 large differences in body-size and longevity, which could be contributing to interaction network structure
402 (e.g., modules and centralized species). As heritable variation is the raw material for natural selection to
403 act upon, a genetic basis for interaction network structure indicates that conserving genetic variation is
404 important to consider in efforts to restore or preserve complex species interactions and their associated
405 ecosystem functions (Whitham et al., 2012; Evans et al., 2013; Whitham et al., 2020). Many ecological
406 network studies that have an evolutionary context have focused on asymmetry and the quantification
407 of its structure in communities (Bascompte et al., 2006; Díaz-Castelazo et al., 2010; Guimarães et al.,
408 2011; Thompson, 2013). The findings of these studies have generally supported the hypothesis that
409 nestedness (the degree to which species tend to interact with similar subsets of the community) tends to
410 promote stability in mutualistic, primarily bipartite (i.e., two-mode), networks and modularity contributes
411 to the stabilization of antagonistic networks (Elias et al., 2013; Grilli et al., 2016). Although there is
412 growing evidence that the nestedness of mutualistic networks is not necessarily the result of selection
413 for systems-level properties that promote stability but could be either a product of asymptotic abundance
414 distributions leading to uneven interaction frequencies (Staniczenko et al., 2013) and/or a by-product of
415 selection and divergence creating network “spandrels” in ecosystems (Valverde et al., 2018), this does not
416 preclude the functional consequences of network structure but rather the developmental or evolutionary
417 processes that have produced the structure. We were not able to examine nestedness or modularity of
418 the lichen networks as metrics for analyzing weighted, directed *and* signed networks could not be found.
419 Hopefully future network theoretic developments will make the appropriate metrics available to conduct
420 these analyses.

421 Game theoretic applications to ecological networks provide another theoretical basis that relates more
422 directly to the heritability of network centralization found in the present study. Although developed at the
423 population scale, the network theory could be applied at the community scale as well (Lieberman et al.,
424 2005). One seemingly useful direction is the classification of networks into two general categories, rooted
425 and cyclic, in which rooted networks have interactions in which evolutionary effects emanate from one or
426 multiple origins but these effects do not have feedbacks to the origin, whereas cyclic networks contain
427 feedbacks to one or more origins. This is equivalent to “unidirectional” and “reciprocal” genetic effects
428 in the context of IIGE theory (Whitham et al., 2020). As we do not have an estimate of the effect of the
429 lichen at the larger scale of the fitness of the tree they occur on, we can not determine whether the lichen
430 networks in this system are cyclic or not. In terrestrial ecosystems, lichens play important ecological
431 roles, such as substrate stabilization (Root et al., 2011) and nitrogen fixation (Nelson et al., 2018). Some
432 epiphytic lichens can have demonstrable effects on the availability of nutrients for the trees that they are
433 associated with (Norby and Sigal, 1989). Although none of the lichens the present study’s system is

known to fix nitrogen, it is possible that they might add micro-nutrients or provide some other unobserved benefit to their host trees. Elucidating the presence of and quantifying such feedbacks would permit the determination of the cyclic nature evolutionary dynamics. If there are positive effects of lichens on host trees that might increase their ability to respond to environmental stress, then selection could enhance tree performance and trees with superior communities are more likely to survive. Gehring et al. (2014, 2017) showed this with ectomycorrhizal communities in which trees with superior mutualist communities were more likely to survive drought and community evolution occurred (Whitham et al., 2020). However, such feedbacks to the higher level of the foundation species are not a requirement for evolution, and, regardless of whether networks are rooted (no feedbacks) or cyclic (feedbacks present), selection at the community level leading to evolution can still occur. Specifically within the context of the current study, even if lichens and their interactions do not feed back to affect the performance and ultimate fitness of the tree on which they reside, non-random death of trees, such as those observed for drought in arid systems Sthultz et al. (2009); Gehring et al. (2017), can still result in selection at the community level and evolution. For example, when a tree dies from some event (e.g., a drought, fire, storm, etc.), its lichen network is selected against while intact networks persist on other trees that survive this selection event.

Conclusion

We found support for both of our hypotheses. First, tree genotype influenced the network structure of lichen communities associated with narrowleaf cottonwoods in a riparian forest ecosystem. Network similarity and metrics of network structure tended to be more similar on trees of the same genotype. Generally, this genetic effect was manifested in positive interactions and largely driven by *A. holocarpa*. Second, the genetically based trait, bark roughness, was observed to affect network variation, largely via shifts in positive in-coming and out-going interactions. Chemistry traits, whether genetically based (e.g., tannin concentration) or not, were not significantly correlated with lichen network structure. Bark roughness has been demonstrated previously to be under strong genetic control in cottonwoods (Bdeir et al., 2017) and other foundation tree species, such as *Eucalyptus* (Nantongo et al., 2020). Bark roughness has also been shown to be an important tree trait influencing bark lichens (Lamit et al., 2015b); however this is the first demonstration of a link from genetics to lichen network structure. As such, these results have important implications for the influence of genetically based variation in ecosystems with networks of interacting species. Going forward, future work could extend the many previous community genetics studies that have focused on sessile organisms, such as gallin insects (Bailey et al., 2005; Whitham et al., 2006; Crutsinger et al., 2014; Smith et al., 2011; Keith et al., 2017), to quantify the frequency of these interactions in the context of the larger community. Network modeling and analysis will provide useful tools for the identification of species within network modules that are most important to study in systems where little is known about the natural history of organisms in an ecosystem. Such investigations will bring us closer to understanding the evolutionary drivers of Darwin's entangled bank and the interconnectedness of species in complex communities (Darwin, 1859; Dátilo et al., 2016).

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Figure 1. The communities of bark lichens were observed in a common garden of replicated genotypes of narrowleaf cottonwood trees (*P. angustifolia*) at the Ogden Nature Center (Ogden, UT). (A) Lichens were sampled within a fixed area (100 cm^2) on individual trees at two heights, 50cm and 95cm from the ground. (B-I) Photos showing lichen species observed, respectively: *Xanthomendoza montana*, *Candelariella subdeflexa*, *Rinodina freyi*, *Athallia holocarpa*, *Physcia adscendens*, *Physciella melanachra*, *Physcia undulata* and *Myriolecis hagenii*. Photo Credits: L.J. Lamit (B-D) and R. Reese Næsborg (E-I).

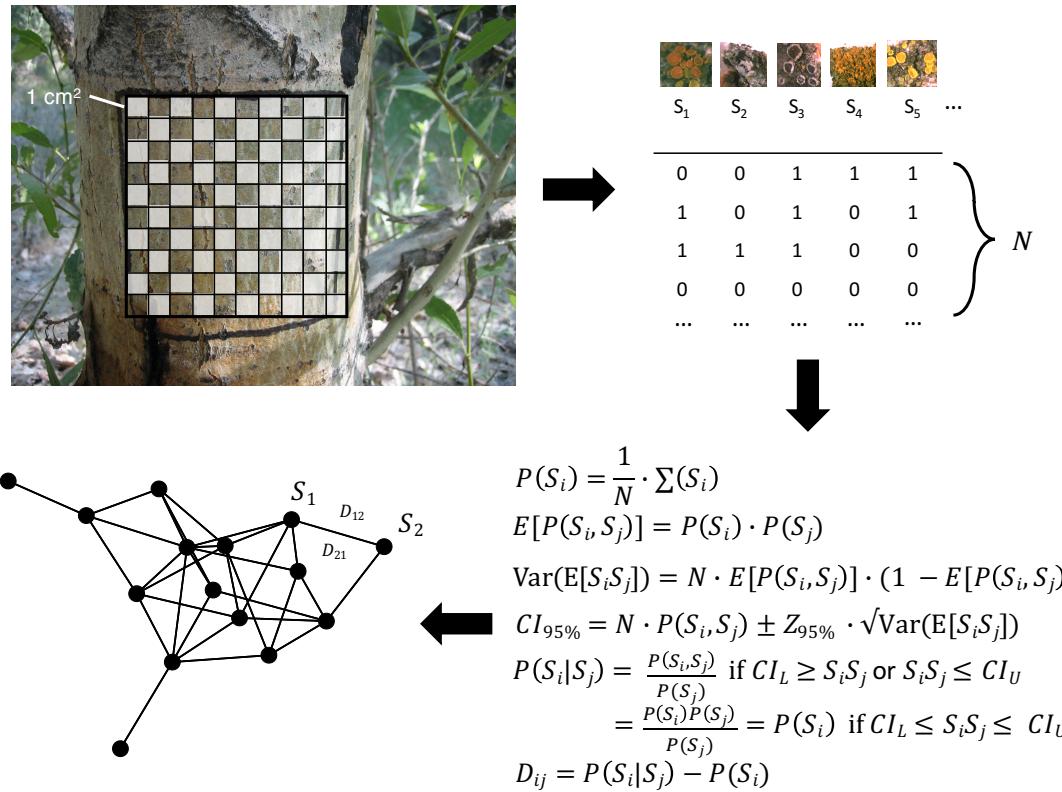


Figure 2. Lichen interaction networks were constructed by conducting field observations in 1 cm² cells within a 100 cm² grid on each tree using a checkerboard pattern (grey cells). Thus, a set of N total cell observations were recorded for each tree with the presence or absence of each species recorded for each cell. Applying the probability-based network modeling method adapted from (Araújo et al., 2011), we calculated the conditional probabilities, $P(S_i | S_j)$, for all species pairs and removed (i.e., set equal to zero) species pairs whose joint probabilities, $P(S_i S_j)$, were not significant using a confidence interval based comparison of their observed co-occurrence frequency, $S_i S_j$, to that expected due to chance alone, $E[P(S_i S_j)] = P(S_i)P(S_j)$, and $P(S_i | S_j)$ reduces to $P(S_i)$, the observed individual probability of species S_i .

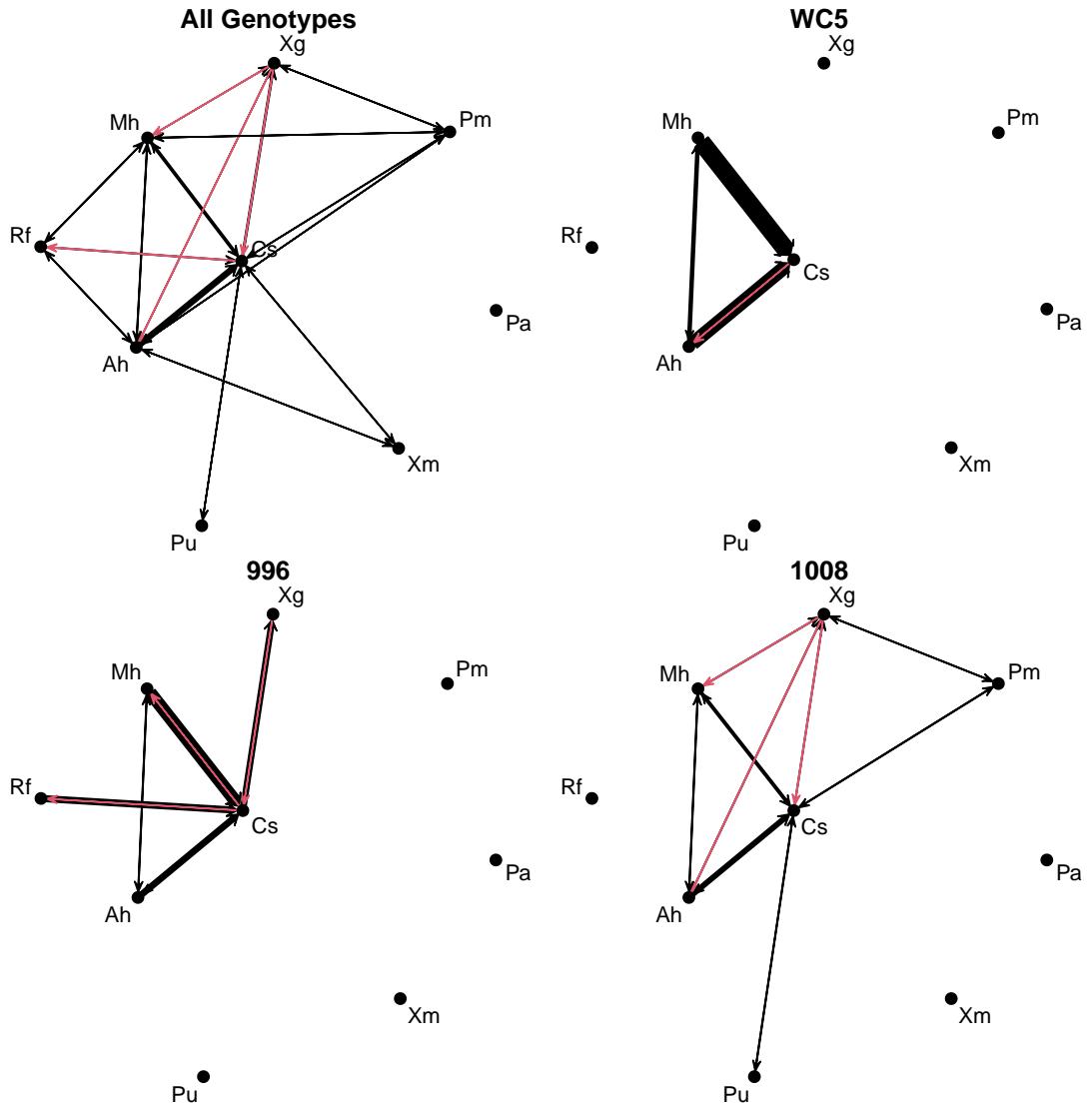


Figure 3. Lichen networks varied in structure among tree genotypes. Network diagrams of the mean lichen interaction matrices averaged for all trees and for illustrative genotypes (996, WC5 and 1008) showing a range of interaction network structure. Directionality (arrowheads) and sign (red = negative, black = positive) of interactions are shown as edges between species, which are scaled by their magnitude. Species names are abbreviated by the first letter of the genus and specific epithet: Ah = *Athallia holocarpa*, Cs = *Candelariella subdeflexa*, Mh = *Myriolecis hagenii*, Rf = *Rinodina freyi*, Pa = *Physcia adscendens*, Pm = *Physciella melanchra*, Pu = *Physcia undulata*, Xg = *Xanthomendoza galericulata*, Xm = *Xanthomendoza montana*. The sign of the interaction is indicative of greater (positive) or lesser (negative) paired occurrences than expected relative to the overall frequency of occurrence of each species. Ecologically, the links in the network are likely the product of multiple types of interactions (e.g. mutualism, parasitism, competition, facilitation) that could vary over both space and time.

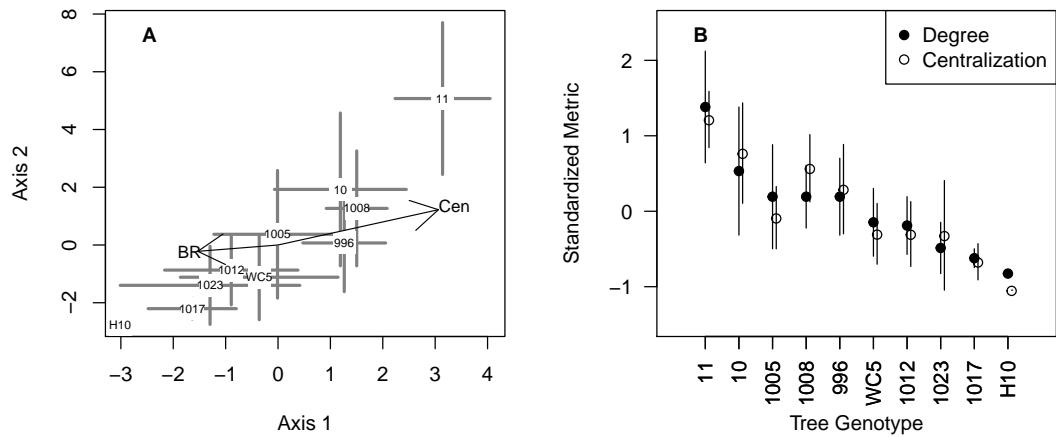


Figure 4. The similarity of lichen networks varied among tree genotypes. A. The plot shows genotype centroids of NMDS ordinated ($R^2 = 0.999$, stress = 0.008) lichen network similarity (± 1 S.E.). Genotype centroids that are closer together tend to have more similar lichen network structure. Arrows showing the direction (arrowhead) and magnitude (length) of the vectors of correlation between bark roughness (BR) and network centralization (Cen) and the ordinated networks. B. Plot showing the standardized ($\frac{x-\bar{x}}{\sigma}$) means (± 1 S.E.) for the two of the genetically based lichen network metrics: overall degree (i.e., total number of links) and centralization, which is a measure of the dominance of one species in the network.

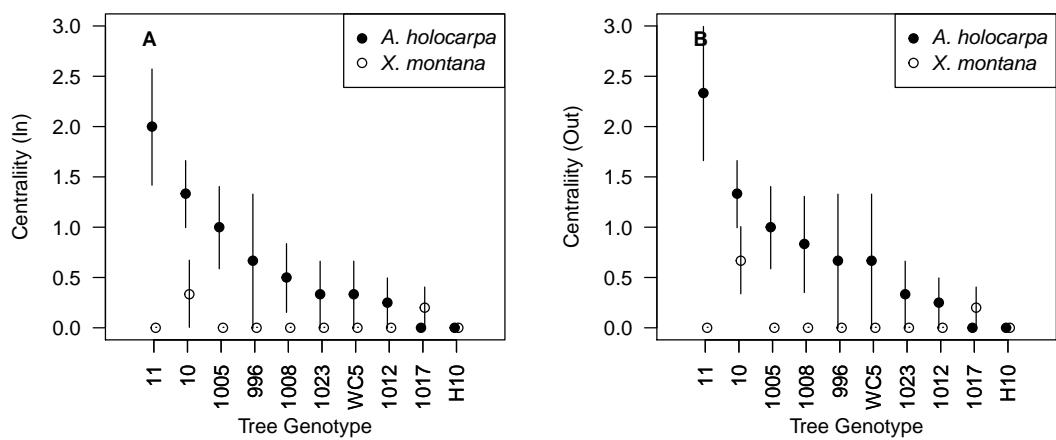


Figure 5. Dot-plots showing the mean (dot) and ± 1 SE of in-degree (A) and out-degree (B) centrality for two species, *A. holocarpa* and *X. montana*. *Athallia holocarpa* centrality was highly variable among genotypes. *Xanthomendoza montana* centrality, both in- and out-degree, was only non-zero for two genotypes, and only out-degree centrality displayed a significant response to genotype.

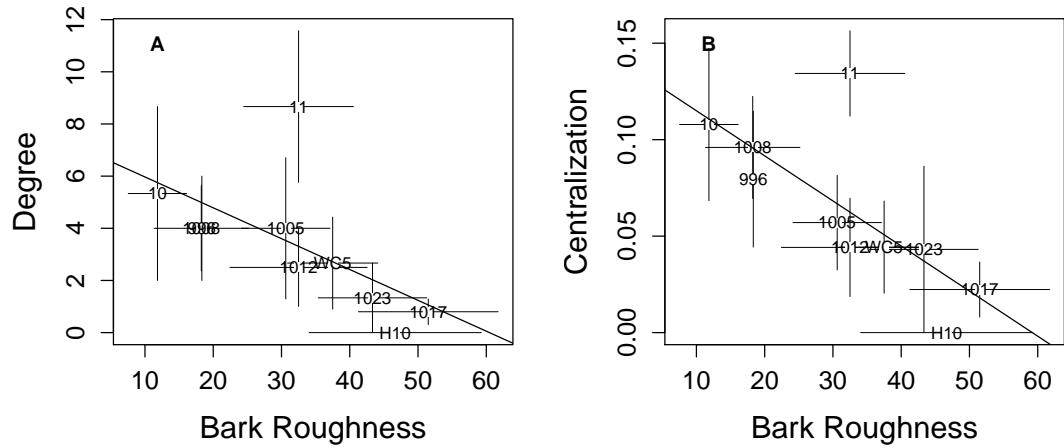


Figure 6. Bivariate plots of the negative relationship between bark roughness and two network metrics: A) degree and B) centralization. Each plot displays the genotype mean (± 1 S.E) for both variables and a least-squares regression line calculated using the genotype means. Generally, as roughness increased the number of interactions (degree) and dominance of those interactions (centralization) decreased.