

# STATISTICAL ECO(-TOXICO)LOGY

IMPROVING THE UTILIZATION OF DATA FOR  
ECOLOGICAL RISK ASSESSMENT

by

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## INTRODUCTION AND OBJECTIVES

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### 1.1 PESTICIDES IN FRESHWATER ECOSYSTEMS

### 1.2 ECOLOGICAL RISK ASSESSMENT

Ecological risk assessment (ERA) tries to estimate risks to animals, populations or ecosystems and is used as a tool for decision making under uncertainty (Newman, 2015). The decision to be made is, whether a (new) pesticide can be approved for usage and a potential release in the environment without a risk to the environment. Ecological risk is defined as a combination of the severity and the probability of occurrence of a potential adverse effect (Suter, 2007). Therefore, ERA is based on two components: Effect- and exposure assessment. A combination of both is needed to characterise ecological risks.

Effect assessment characterizes the strength of effects using laboratory and semi-field experiments. It establishes relationships between the concentration of a compound and the observed ecological effects. In the European Union a tiered approach with increasing complexity and realism. Lower tier assessment is based on highly standardized single species laboratory experiments, whereas higher tier assessment is refined by testing additional species, extended laboratory experiments or model ecosystem experiments. To address the various uncertainties in effect assessment (e.g. experimental variation, variation between species, variation in environmental conditions etc) the retrieved toxicity values are multiplied by a assessment factor between 0.01 (lower tier assessment) and 0.5 (higher tier assessment) depending on data quality, which yields to a regulatory acceptable concentration (RAC) (EFSA, 2013).

Exposure Assessment for freshwaters aims to characterise the probability of a adverse effect by deriving a predicted environmental concentration (PEC) in surface waters and sediments (Newman, 2015). It is mainly based on modeling the fate of chemicals in the environment using computer simulations. In the European Union, the FOCUS models are used (FOCUS, 2001; EFSA, 2013). To calculate PECs these models need many compound specific input parameters like the molecular weight, water solubility, partitioning coefficients and dissipation time. Additionally, information on the application regime and crop type is needed. FOCUS models the concentration within edge-of-field streams of 1 meter width and 30 cm depth (Er-lacher and Wang, 2011). Such a stream width corresponds to a catchment size of 7 km<sup>2</sup> (See figure ??) [ref to small streams supplement](#) which is considered as small stream (Lorenz et al., 2016). Nevertheless, recent research showed that FOCUS models fail predict measured field concentrations of pesticides (Knäbel et al., 2012; Knäbel et al., 2014).

The final step in ERA is risk characterisation. It puts together the information gained from effect and exposure assessment. Risk can expressed in several ways, a quantitative way being the risk quotient approach: A PEC / RAC ratio greater than one indicating potential risks (Suter, 2007; EFSA, 2013; Amiard-Triquet, 2015). Substances with a ration lower than one could be approved for usage and potential release to the environment.

### 1.3 ENVIRONMENTAL MONITORING

### 1.4 STATISTICAL ECOTOXICOLOGY

Ecological effect assessment generates data on ecological effects using experiments. The produced datasets range from small univariate datasets (lower tier assessment) to medium sized multivariate datasets (higher tier assessment). These datasets are analyzed using statistical techniques in order to deliver usable information for assessment. Therefore, effect assessment relies heavily on statistical methods. Statistical ecotoxicology combines statistics with the specific needs and constraints of ecotoxicology. It aims to provide solutions to statistical challenges in ecotoxicology (Fox and Landis, 2016a), guidance on experimental designs (Johnson et al., 2015)

and tools to integrate big data (Van den Brink et al., 2016) to improve accuracy of ERA.

The relationships between the concentration of a compound and the observed effects are usually analysed using dose-response models (Ritz, 2010), which can be used to derive an effective concentration for x% effect ( $EC_x$ ). Nevertheless, such relationships cannot always be established from experimental data. For example, model ecosystem experiments are conducted to characterize effects on whole biological communities. However, because of multivariate responses and potential indirect effects, there is no clear dose-response relationship and no models for this kind of data available. There are also other examples where fitting dose-response models is problematic (Green, 2016). In such cases there is usually a no-observed-effect concentration (NOEC) computed. The NOEC is the highest tested concentration that does not lead to significant deviation from the control response and therefore relies on null hypothesis significance testing (NHST). However, the use of NOEC as toxicity measure in ecological effect assessment has been heavily in the past (Laskowski, 1995; Chapman et al., 1996; Warne and Dam, 2008; Fox et al., 2012; Jager, 2012; Fox and Landis, 2016b).

Instead of conducting experiments, toxicity could be also predicted from molecular structures using quantitative structure-activity relationships (QSAR), which are calculated using machine-learning techniques (Murrell et al., 2015; Cortes-Ciriano, 2016). However, improving these models to give sufficient prediction accuracy needs to integrate more input data (Kühne et al., 2013).

Large amount of data are available that could be use for effect and exposure assessment. For example, the US EPA ECOTOX database (U.S. EPA, 2016), the Pesticides Properties Database (Lewis et al., 2016) and ETOX (Umweltbundesamt, 2016) provide toxicity data that could be used for effect assessment. Databases like Physprop (Howard and Meylan, 2016) and Pubchem (Kim et al., 2016) provide chemical properties that are needed as input for exposure models. Monitoring data provides information on realised concentrations, could be used for validation of models and retrospective risk assessment. This big data can provide new information and opportunities for ERA (Dafforn et al., 2015). However, it needs to be linked in order to be used effectively in ERA.

## 1.5 OBJECTIVES AND OUTLINE OF THE THESIS

The overall goal of this thesis was to contribute to the emerging field of statistical ecotoxicology and ERA. The main objectives were (i) to scrutinize new methods in statistical ecotoxicology, (ii) explore available monitoring data and (iii) provide tools to deal with big data. Figure 1.1 provides a conceptual overview on the parts of this thesis and its relations to ERA as outlined in the previous sections.

The thesis starts with a comparison of statistical methods to analyse ecotoxicological experiments in effect assessment (Chapter ??). Specific questions addressed were:

- Are newer statistical methods more powerful than currently used methods for NHST?
- How much statistical power do current experimental designs in ecotoxicology exhibit?

Exposure assessment aims at predicting concentrations in small streams. Chapter 2 focuses on large-scale realised environmental concentrations and the drivers thereof. Specific goals were:

- Compile all available monitoring data on pesticides in small streams in Germany
- Explore the relationship between agricultural land use and streams size and measured pesticide concentrations.
- Study annual dynamics of pesticide exposure, as well as influence of precipitation on measured pesticide concentrations.
- Assess the current pollution in German streams and identify responsible pesticides.

The compilation of monitoring data from different data sources, lead to a big inhomogeneous amount of data that first needs to be harmonized. Chapters ?? (chemical data) and ?? (biological data) describe software solutions to simplify and accelerate the workflow of:

- validating and harmonizing chemical and taxonomic data



- linking datasets
- retrieving properties and identifiers

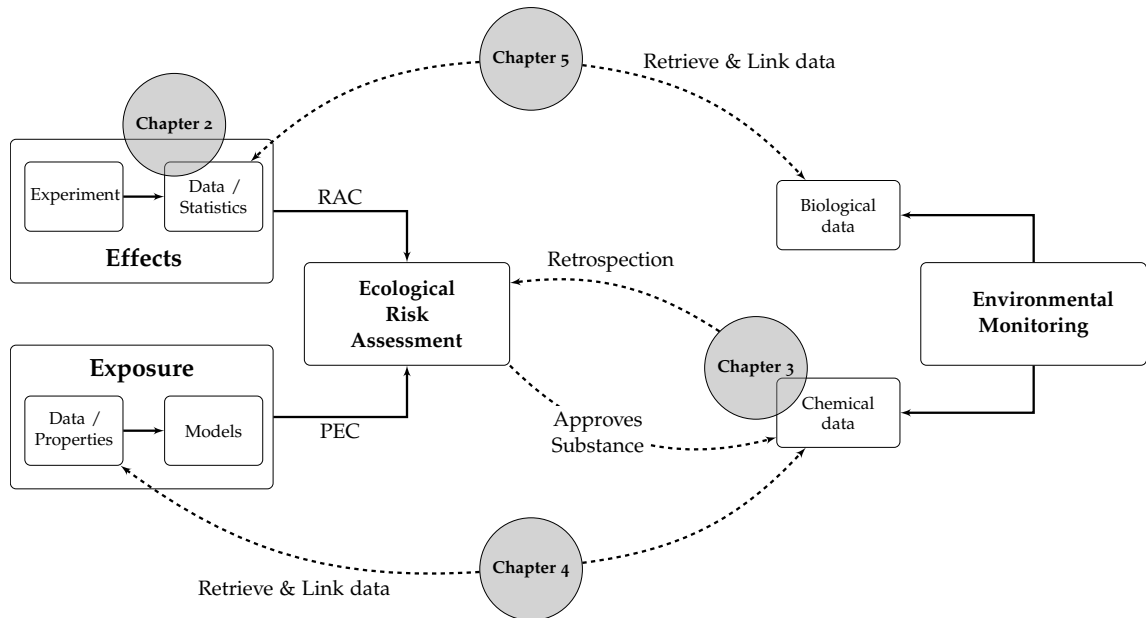


Figure 1.1: Conceptual overview on data in ecological risk assessment, environmental monitoring and the parts addressed by this thesis.

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## LARGE SCALE RISKS FROM PESTICIDES IN SMALL STREAMS

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2.1 ABSTRACT

2.2 REFERENCES



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## DISCUSSION

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### 3.1 STATISTICAL ECOTOXICOLOGY

### 3.2 LEVERAGING MONITORING DATA FOR ECOLOGICAL RISK ASSESSMENT

### 3.3 CHALLENGES TO UTILIZE 'BIG DATA' IN ERA

### 3.4 CONCLUSIONS AND OUTLOOK

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