### STATISTICAL ECO(-TOXICO)LOGY

### IMPROVING THE UTILIZATION OF DATA FOR ECOLOGICAL RISK ASSESSMENT

by

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# 1 GENERAL DISCUSSION AND OUTLOOK

#### STATISTICAL ECOTOXICOLOGY

The simulation study performed in chapter ?? clearly showed that common experimental designs exhibit unacceptably low statistical power (Szöcs and Schäfer, 2016; Van Der Hoeven, 1998). This underpins the criticism accumulated over the last 30 years towards the usage of NOEC as endpoint (Fox and Landis, 2016). Nevertheless, the NOEC is still one of the standard endpoint for mesocosm experiments in higher tier risk assessment (EFSA, 2013).

Recently, a posteriori calculations of statistical power have been proposed to counteract these limitations and aid the interpretation treatment-related effects in model ecosystems (Brock et al., 2015). The "minimum detectable difference" (MDD) estimates the difference between to means that must exist in order to produce a statistically significant result (p <0.05 (Gelman and Stern, 2006)) and could be used to interpret NOEC. However, a posteriori calculations have been shown to have logical flaws when used for interpretation of non-significant results (Hoenig and Heisey, 2001; Nakagawa and Foster, 2004). However, conducting and report of a priori power calculations, as performed in chapter ??, might provide researchers important information to optimize their study designs, ensuring that their experimental designs have appropriate power (Johnson et al., 2015).

Moreover, similar simulations can not only be used to analyse data of factorial designs, but also from regression designs. Indeed, simulations could be used to determine optimal designs for dose-response models and  $EC_x$  determination, balancing precision and resources. Regression designs are generally more powerful and provide more information than factorial designs (Cottingham et al., 2005). Regression designs in mesocosm experiments, assigning the replicates to more tested concentrations, might also provide more insights. However, currently statistical tools to analyse a community dose-reponse relations, providing a  $EC_{x,community}$  are not well explored. Separate dose-response models could be fit to each species (Ritz, 2010), leading to a  $EC_x$  for each species in a mesocosm study. Subsequently, this  $EC_x$  values could be combined and summarised

using Species Sensitivity Distributions (Posthuma et al., 2002), providing a hazardous concentration ( $HC_{x,community}$ ) for x % of species affected (Maltby et al., 2005). Another possibility would be to use a logistic type of ordination (van den Brink et al., 2003). Reduced-Rank vector generalized linear models (RR-VGLM) could be used to fit such type of models (Yee, 2015; Yee and Hastie, 2003) but they have not been applied in ecotoxicology yet.

In a similar vein, community ecology is currently experiencing a shift towards new class of multivariate methods, incorporating statistical models for abundances across many taxa simultaneously (ter Braak and Šmilauer, 2015; Warton et al., 2015a; Warton et al., 2015b; Warton et al., 2012). However, this methods have not been applied frequently and their applicability to ecotoxicological data is currently unclear (Szöcs et al., 2015). All this models have in common, that the choice of statistical model is primarily based ...

### LEVERAGING MONITORING DATA FOR ECOLOGICAL RISK ASSESSMENT

## CHALLENGES TO UTILIZE 'BIG DATA' IN ECOLOGICAL RISK ASSESSMENT

#### CONCLUSIONS

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