

<sup>1</sup> The effects of multi-dimensional drought on land cover change and  
<sup>2</sup> vegetation productivity in continental Chile

<sup>3</sup> Francisco Zambrano<sup>a,1,\*</sup>, Anton Vrieling<sup>b</sup>, Francisco Meza, Iongel Duran-Llacer, Francisco Fernández<sup>c</sup>,  
<sup>4</sup> Alejandro Venegas-González, Nicolas Raab, Dylan Craven

<sup>a</sup> Universidad Mayor, Hémera Centro de Observación de la Tierra, Facultad de Ciencias, Escuela de Ingeniería en Medio Ambiente y Sustentabilidad, Santiago, Chile, 7500994

<sup>b</sup> University of Twente, Faculty of Geo-Information Science and Earth Observation (ITC), Enschede, The Netherlands,

<sup>c</sup> Universidad San Sebastián,

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<sup>5</sup> **Abstract**

The north-central region of Chile has been the focus of research studies due to the persistent decrease in water supply, which is impacting the hydrological system and vegetation development. This persistent period of water scarcity has been defined as a megadrought. The aim of our study is to evaluate the land cover change over continental Chile and to examine how this is connected to drought indices of water supply, atmospheric evaporative demand (AED), soil moisture, and their effects on vegetation productivity. The drought indices were derived using monthly ERA5-Land reanalysis data spanning from 1981 to 2023. The Moderate-Resolution Imaging Spectroradiometer (MODIS) datasets were utilized to obtain information on annual land cover and monthly vegetation productivity. We analyzed short- (1, 3, 6 months) to long-term (12, 24, 36 months) time scales of drought. Our results showed that land cover change was highest in the south-central part of the country, reaching changes as high as 36% in the surface type. The water demand has increased for the whole country, with a major increase in the north. The AED and soil moisture evidence a decreasing trend, which decreases at longer time scales and from north to south. The extreme south part of the country shows an increase in supply. Vegetation productivity has a negative trend in the north-central region for all land cover types. On the other hand, forests seem to be the most resistant type to drought. The types that show to be most affected by variation in climate conditions are shrublands, savannas, and croplands. The drought indices that have the capability of explaining to a major degree the variance in vegetation productivity are the ones that consider soil moisture for twelve months and the combined effect of precipitation and AED for 24 and 12 months. The results indicate that the north-central region is the most sensitive to water supply deficits lasting longer than a year.

<sup>6</sup> **Keywords:** drought, land cover change, vegetation productivity, satellite

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<sup>7</sup> **1. Introduction**

<sup>8</sup> Drought is often classified as meteorological when there is a decrease in precipitation below the mean  
<sup>9</sup> average of several years (more than 30 years), hydrological when these anomalies last for long periods (months  
<sup>10</sup> to years) and affect water systems, and agricultural when the deficit impacts plant health anomalies and

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\*Corresponding author

Email addresses: francisco.zambrano@umayor.cl (Francisco Zambrano), a.vrieling@utwente.nl (Anton Vrieling), francisco.fernandez@uss.cl (Francisco Fernández)

<sup>1</sup>This is the first author footnote.

leads to decreased productivity (Wilhite and Glantz, 1985). However, it is important to note that drought is also influenced by human activities, which were not considered in the definitions. Thus, Van Loon et al. (2016) and AghaKouchak et al. (2021) have given an updated definition of drought for the Anthropocene, suggesting that it should be considered the feedback of humans' decisions and activities that drives the anthropogenic drought. Simultaneously, drought leads to heightened tree mortality and induces alterations in land cover and land use, ultimately affecting ecosystems (Crausbay et al., 2017). Even though many ecological studies have misinterpreted how to characterize drought, for example, sometimes considering "dry" conditions as "drought" (Slette et al., 2019). Then, Crausbay et al. (2017) proposed the ecological drought definition as "an episodic deficit in water availability that drives ecosystems beyond thresholds of vulnerability, impacts ecosystem services, and triggers feedback in natural and/or human systems." In light of current global warming, it is crucial to study the interaction between drought and ecosystems in order to understand their feedback and impact on water security. (Bakker, 2012)

Human-induced greenhouse gas emissions have increased the frequency and/or intensity of drought as a result of global warming, according to the sixth assessment report (AR6) of the Intergovernmental Panel on Climate Change (IPCC) (Calvin et al., 2023). The evidence supporting this claim has been strengthened since AR5 (IPCC, 2013). Recent studies, however, have produced contrasting findings, suggesting that drought has not exhibited a significant trend over the past forty years. (Vicente-Serrano et al., 2022; Kogan et al., 2020). Vicente-Serrano et al. (2022) analyzed the meteorological drought trend on a global scale, finding that only in a few regions has there been an increase in the severity of drought. Moreover, they attribute the increase in droughts over the past forty years solely to an increase in atmospheric evaporative demand (AED), which in turn enhances vegetation water demand, with important implications for agricultural and ecological droughts. Also, they state that "the increase in hydrological droughts has been primarily observed in regions with high water demand and land cover change". Similarly, Kogan et al. (2020) analyzed the drought trend using vegetation health methods, finding that for the globe, hemispheres, and main grain-producing countries, drought has not expanded or intensified for the last 38 years. Further, the Masson-Delmotte (2021) suggests that there is a high degree of confidence that rising temperatures will increase the extent, frequency, and severity of droughts. Also, AR6 (Calvin et al., 2023) predicts that many regions of the world will experience more severe agricultural and ecological droughts even if global warming stabilizes at 1.5°–2°C. To better evaluate the impact of drought trends on ecosystems, assessments are needed that relate meteorological and soil moisture variables to their effects on vegetation.

From 1960 to 2019, land use change has impacted around one-third of the Earth's surface, which is four times more than previously thought (Winkler et al., 2021). Multiple studies aim to analyze and forecast changes in land cover globally (Winkler et al., 2021; Song et al., 2018) and regionally (Chamling and Bera, 2020; Homer et al., 2020; Yang and Huang, 2021). Some others seek to analyze the impact of land cover change on climate conditions such as temperature and precipitation (Luyssaert et al., 2014; Pitman et al., 2012). There is less research on the interaction between drought and land cover change (Chen et al., 2022; Akinyemi, 2021; Peng et al., 2017). Peng et al. (2017) conducted a worldwide investigation utilizing net primary production to examine the spatial and temporal variations in vegetation productivity at global level. The study aimed to assess the influence of drought by comparing the twelve-month Standardized Precipitation Evapotranspiration Index (SPEI) and land cover change. According to their findings, drought is responsible for 37% of the decline in vegetation productivity, while water availability accounts for 55% of the variation. Chen et al. (2022) studied the trend of vegetation greenness and productivity and its relation to meteorological drought (SPEI of twelve months in December) and soil moisture at the global level. The results showed lower correlations (<0.2) for both variables. Akinyemi (2021) evaluates drought trends and land cover change using vegetation indices in Botswana in a semi-arid climate. These studies mostly looked at how changes in land cover and vegetation productivity are related to a single drought index (SPEI) over a single time period of 12 months. SPEI takes into account the combined effect of precipitation and AED as a water balance, but it does not allow us to know the contribution of each variable on its own. Some things worth investigating in terms of land cover change and vegetation productivity are: i) How do they respond to short- to long-term meteorological and soil moisture droughts? ii) How is the drought impacting land cover changes? And iii) How do they behave in humid and arid climatic zones regarding drought? Likewise,

62 there is a lack of understanding of how the alteration in water supply and demand is affecting land cover  
63 transformations.

64 For monitoring drought, the World Meteorological Organization recommends the SPI (Standardized Pre-  
65 cipitation Index) (WMO et al., 2012). The SPI is a multi-scalar drought index that only uses precipitation  
66 to assess short- to long-term droughts. The primary cause of drought is precipitation anomalies, and tem-  
67 perature usually makes it worse (Luo et al., 2017). Nowadays, there is an increase in attention toward  
68 using AED separately to monitor droughts (Vicente-Serrano et al., 2020). One reason is due to its attri-  
69 bution to increasing flash droughts in water-limited regions (Noguera et al., 2022). Vicente-Serrano et al.  
70 (2010) proposed the Standardized Precipitation Evapotranspiration Index (SPEI), which incorporated the  
71 temperature effect by subtracting AED from precipitation. SPEI allows for analysis of the combined effect  
72 of precipitation and AED. Since its formulation, it has been used worldwide for the study and monitoring  
73 of drought (Gebrechorkos et al., 2023; Liu et al., 2024). Hobbins et al. (2016) and McEvoy et al. (2016)  
74 developed the Evaporative Demand Drought Index (EDDI) to monitor droughts solely using the AED, and  
75 it has proven effective in monitoring flash droughts (Li et al., 2024; Ford et al., 2023). For soil moisture,  
76 several drought indices exist, such as the Soil Moisture Deficit Index (SDMI) (Narasimhan and Srinivasan,  
77 2005) and the Soil Moisture Agricultural Drought Index (SMADI) (Souza et al., 2021). Hao and AghaK-  
78 ouchak (2013) and AghaKouchak (2014) proposed the Standardized Soil Moisture Index (SSI), which has a  
79 similar formulation as the SPI, SPEI, and EDDI. Thus, there are plenty of drought indices that allow for  
80 a comprehensive assessment of drought on short- to long-term scales and that allow for the use of single  
81 variables from the earth's water balance (e.g., precipitation, AED, soil moisture). The variation in climate  
82 variables impacts vegetation development, and unfavorable conditions such as low precipitation and high  
83 temperatures usually generate a decrease in vegetation productivity. To monitor the response of vegetation,  
84 the common practice is to use satellite data. The Normalized Difference Vegetation Index (NDVI) has been  
85 widely used as a proxy for biomass production (Camps-Valls et al., 2021; Paruelo et al., 2016; Helman et al.,  
86 2014). For Chile's cultivated land, Zambrano et al. (2018) introduced the zcNDVI for assessing seasonal  
87 biomass production in response to drought. Using this information, we can advance our understanding of  
88 the impact of drought on ecosystems.

89 Chile's diverse climatic and ecosystem types (Beck et al., 2023; Luebert and Pliscoff, 2022) make it an ideal  
90 natural laboratory for studying climate and ecosystems. Additionally, the country has experienced severe  
91 drought conditions that have had significant effects on vegetation and water storage. North-central Chile has  
92 faced a persistent precipitation deficit since 2010, defined as a mega drought. (Garreaud et al., 2017), which  
93 has impacted the Chilean ecosystem. This megadrought was defined by the Standardized Precipitation  
94 Index (SPI) of twelve months in December having values below one standard deviation. Some studies have  
95 addressed how this drought affects single ecosystems in terms of forest development (Miranda et al., 2020;  
96 Venegas-González et al., 2018), forest fire occurrence (Urrutia-Jalabert et al., 2018), and crop productivity  
97 (Zambrano, 2023; Zambrano et al., 2018, 2016). We found one study regarding land cover and drought in  
98 Chile. The study by Fuentes et al. (2021) evaluates water scarcity and land cover change in Chile between  
99 29° and 39° of south latitude. Fuentes et al. (2021) used the SPEI of one month for evaluating drought,  
100 which led to misleading results. For example, they did not find a temporal trend in the SPEI but found a  
101 decreasing trend in water availability and an increase trend on AED, which in turn should have been capable  
102 of being captured with longer time scales of the SPEI. The term "megadrought" in Chile is used to describe  
103 a prolonged water shortage that lasts for several years, resulting in a permanent deficit that impacts the  
104 hydrological system (Boisier et al., 2018). Therefore, it is crucial to evaluate temporal scales that consider  
105 the cumulative impact over a period of several years. The association between drought and the environment  
106 in Chile is not well comprehended. Hence, it is imperative to acquire a more profound comprehension of the  
107 manner in which climatic and soil moisture droughts influence environmental dynamics, in order to make  
108 well-informed decisions on adaptation strategies.

109 Here, we analyze the multi-dimensional impacts of drought across ecosystems in continental Chile. More  
110 specifically, we aim to assess: i) short- to long-term temporal trends in multi-scalar drought indices; ii)  
111 temporal changes in land-use cover and the direction and magnitude of their relationships with trends in

<sup>112</sup> drought indices; and iii) the trend in vegetation productivity and its relationship with drought indices across  
<sup>113</sup> Chilean ecosystems.

## <sup>114</sup> 2. Study area

<sup>115</sup> Continental Chile has diverse climate conditions with strong gradients from north to south and east to west  
<sup>116</sup> ([Aceituno et al., 2021](#)) (Figure 1 a), which determines its great ecosystem diversity ([Luebert and Pliscoff,](#)  
<sup>117</sup> [2022](#)) (Figure 1 c). The Andes Mountains are a main factor in climate latitudinal variation ([Garreaud, 2009](#)).  
<sup>118</sup> “Norte Grande” and “Norte Chico” predominate in an arid desert climate with hot (Bwh) and cold (Bwk)  
<sup>119</sup> temperatures. At the south of “Norte Chico,” the climate changes to an arid steppe with cold temperatures  
<sup>120</sup> (Bsk). In these two northern regions, the land is mostly bare, with a small surface of vegetation types  
<sup>121</sup> such as shrubland and grassland. In the zones “Centro” and the north half of “Sur,” the main climate is  
<sup>122</sup> Mediterranean, with warm to hot summers (Csa and Csb). Land cover in “Centro” comprises a significant  
<sup>123</sup> amount of shrubland and savanna (50%), grassland (16%), forest (8%), and croplands (5%). An oceanic  
<sup>124</sup> climate (Cfb) predominates in the south of “Sur” and the north of “Austral.” Those zones are high in forest  
<sup>125</sup> and grassland. The southern part of the country has a tundra climate, and in “Austral,” it is a cold semi-arid  
<sup>126</sup> area with an extended surface of grassland, forest, and, to a lesser extent, savanna.

## <sup>127</sup> 3. Materials and Methods

### <sup>128</sup> 3.1. Data

#### <sup>129</sup> 3.1.1. Gridded meteorological and vegetation data

<sup>130</sup> To analyze land cover change, we use the classification scheme by the IGBP (International Geosphere-  
<sup>131</sup> Biosphere Programme) from the product MCD12Q1 collection 6.1 from MODIS. The MCD12Q1 has a  
<sup>132</sup> yearly frequency from 2001 to 2022 and defines 17 classes. To derive a proxy for vegetation productivity, we  
<sup>133</sup> used the Normalized Difference Vegetation Index (NDVI) from the product MOD13A3 collection 6.1 from  
<sup>134</sup> MODIS ([Didan, 2015](#)). MOD13A3 provides vegetation indices at 1km of spatial resolution and monthly  
<sup>135</sup> frequency. The NASA EOSDIS Land Processes Distributed Active Archive Center (LP DAAC), USGS Earth  
<sup>136</sup> Resources Observation and Science (EROS) Center, Sioux Falls, South Dakota, provided the MOD13A3 and  
<sup>137</sup> MCD12Q1 from the online Data Pool, accessible at <https://lpdaac.usgs.gov/tools/data-pool/>.

Table 1: Description of the satellite and reanalysis data used

| Product | Sub-product | Variable                               | Spatial Resolution | Period    | Units | Short Name |  |
|---------|-------------|--|--------------------|-----------|-------|------------|--|
| ERA5L   |             | Precipitation                          | 0.1°               | 1981-2023 | mm    | P          |  |
|         |             | Maximum temperature                    |                    |           | °C    | $T_{max}$  |  |
|         |             | Minimum temperature                    |                    |           | °C    | $T_{min}$  |  |
|         |             | Volumetric Soil Water Content at 1m    |                    |           | m3/m3 | SM         |  |
| ERA5L*  | MOD13A3.061 | Atmospheric Evaporative Demand         | 0.1°               | 1981-2023 | mm    | AED        |  |
| MODIS   |             | Normalized Difference Vegetation Index | 1 km               | 2000-2023 |       | NDVI       |  |
|         |             | land cover IGBP scheme                 |                    | 2001-2022 |       | land cover |  |

\*Calculated from maximum and minimum temperatures derived from ERA5L with Eq. 1.

<sup>138</sup> For soil moisture, water supply, and water demand variables, we used ERA5L (ECMWF Reanalysis version  
<sup>139</sup> 5 over land) ([Muñoz-Sabater et al., 2021](#)), a reanalysis dataset that provides the evolution of atmospheric and  
<sup>140</sup> land variables since 1950. It has a spatial resolution of 0.1° (9 km), hourly frequency, and global coverage.

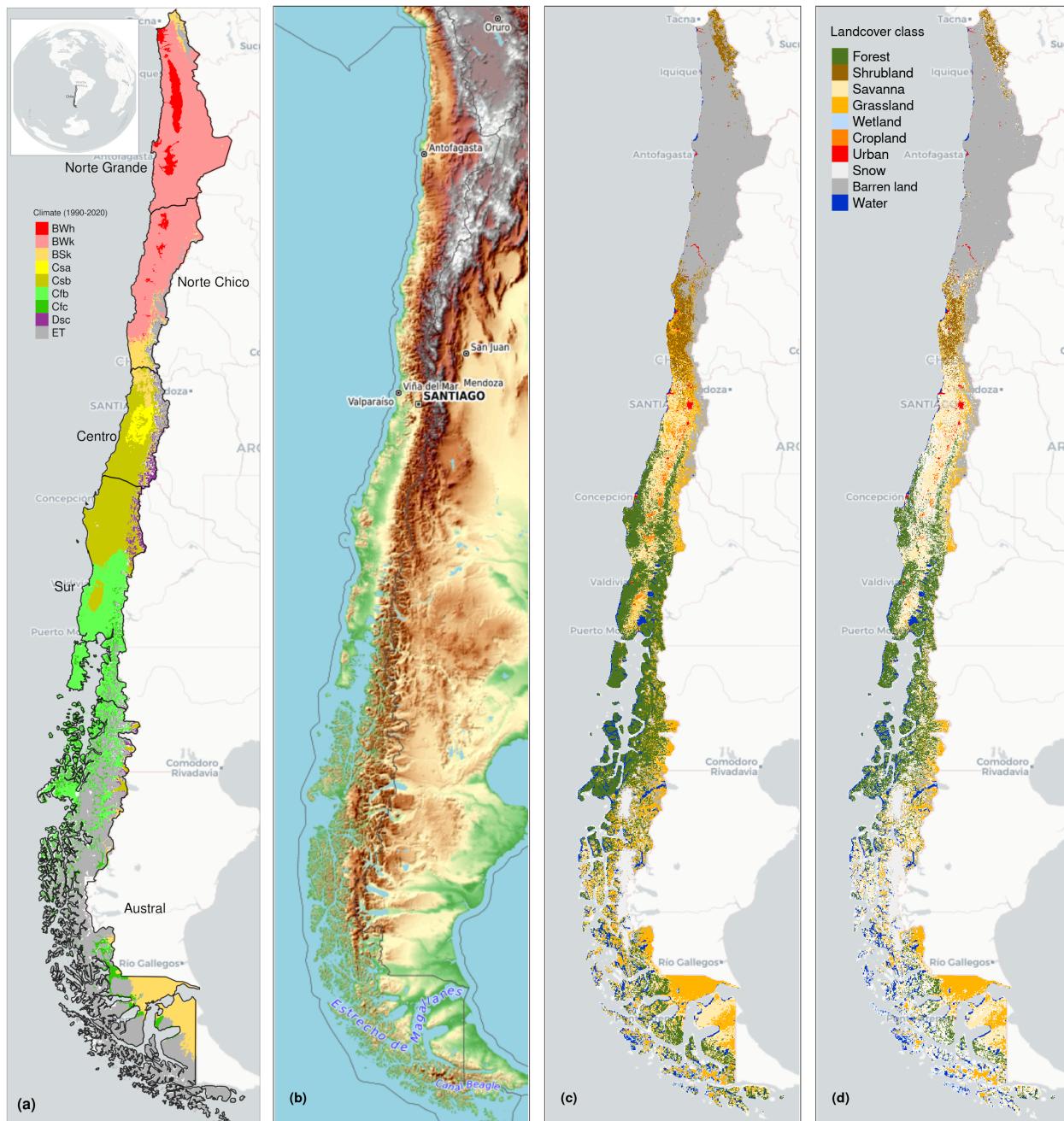


Figure 1: (a) Chile with the Koppen-Geiger climate classes and the five macrozones “Norte Grande”, “Norte Chico”, “Centro”, “Sur”, and “Austral”. (b) Topography reference map. (c) land cover classes for 2022. (d) Persistent land cover classes (> 80%) for 2001-2022

141 We selected the variables for total precipitation, maximum and minimum temperature at 2 meters, and  
 142 volumetric soil water layers between 0 and 100cm of depth (layer 1 to layer 3). Table 1 shows a summary  
 143 of the data and its main characteristics.

144 3.2. Short- to long-term drought

145 3.2.1. Atmospheric Evaporative Demand (AED)

146 In order to compute the drought indices that use water demand, it is necessary to first calculate the  
147 AED. To do this, we employed the Hargreaves method ([Hargreaves, 1994; Hargreaves and Samani, 1985](#)) by  
148 applying the following equation:

$$AED = 0.0023 \cdot Ra \cdot (T + 17.8) \cdot (T_{max} - T_{min})^{0.5} \quad (1)$$

149 where  $Ra$  ( $MJ\ m^2\ day^{-1}$ ) is extraterrestrial radiation;  $T$ ,  $T_{max}$ , and  $T_{min}$  are mean, maximum, and  
150 minimum temperature ( $^{\circ}C$ ) at 2m. For calculating  $Ra$  we used the coordinate of the latitud of the centroid  
151 of each pixel. We chose the method of Hargreaves to estimate AED because of its simplicity, which only  
152 requires temperatures and extrarrestrial radiation. Also, it has been recommended over other methods (e.g.,  
153 Penman-Monteith) when the access to climatic variables is limited ([Vicente-Serrano et al., 2014](#)).

154 3.2.2. Non-parametric calculation of drought indices

155 To derive the drought indices of water supply and demand, soil moisture, and vegetation (i.e., the proxy  
156 of productivity), we used the ERA5L dataset and the MODIS product, with a monthly frequency for 1981–  
157 2023 and 2000–2023, respectively. The drought indices correspond to a historical anomaly with regard to  
158 a variable (e.g., meteorological, vegetation, or soil moisture). To account for the anomaly, the common  
159 practice is to derive it following a statistical parametric methodology in which it is assumed that the  
160 statistical distribution of the data is known ([Heim, 2002](#)). A wrong decision is usually the highest source of  
161 uncertainty ([Laimighofer and Laaha, 2022](#)). In the case of Chile, due to its high degree of climatic variability,  
162 it is complex to choose a proper distribution without previous research. Here, we follow a non-parametric  
163 methodology for the calculation of the drought indices, in a similar manner as the framework proposed by  
164 [Farahmand and AghaKouchak \(2015\); Hobbins et al. \(2016\); McEvoy et al. \(2016\)](#).

165 For the purpose of monitoring water supply drought, we used the well-known Standardized Precipitation  
166 Index (SPI), which relies on precipitation data. To evaluate water demand, we chose the Evaporative  
167 Demand Drought Index (EDDI), developed by [Hobbins et al. \(2016\)](#) and [McEvoy et al. \(2016\)](#), which is based  
168 on the AED. The United States currently monitors drought using the EDDI (<https://psl.noaa.gov/eddi/>) as  
169 an experimental index. To consider the combined effect of water supply and demand, we selected the SPEI  
170 ([Vicente-Serrano et al., 2010](#)). For SPEI, an auxiliary variable  $D = P - AED$  is calculated. Soil moisture  
171 is the main driver of vegetation productivity, particularly in semi-arid regions ([Li et al., 2022](#)). Hence,  
172 for soil water drought, we used the SSI (Standardized Soil Moisture Index) ([Hao and AghaKouchak, 2013;](#)  
173 [AghaKouchak, 2014](#)). In our case, for the SSI, we used the average soil moisture from ERA5L at 1m depth.  
174 Finally, for the proxy of productivity, we used the zcNDVI proposed by [Zambrano et al. \(2018\)](#), which was  
175 derived from the monthly time series of NDVI retrieved from MOD13A1. All the indices are multi-scalar  
176 and can be used for the analysis of short- to long-term droughts.

177 To derive the drought indices, first we must calculate the sum of the variables with regard to the time scale  
178 (s). In this case, for generalization purposes, we will use  $V$ , referring to variables  $P$ ,  $AED$ ,  $D$ ,  $NDVI$ , and  
179  $SM$  (Table 1). We cumulated each  $V$  over the time series of  $n$  values (months), and for the time scales  $s$ :

$$A_{si} = \sum_{i=n-s-i+2}^{n-i+1} V_i \quad \forall i \geq n - s + 1 \quad (2)$$

180 The  $A_{si}$  corresponds to a moving window (convolution) that sums the variable for time scales  $s$  from  
181 the last month, month by month, until the first month in which it could sum for  $s$  months. An inverse  
182 normal approximation ([Abramowitz and Stegun, 1968](#)) obtains the empirically derived probabilities once

<sup>183</sup> the variable cumulates over time for the scale  $s$ . Then, we used the empirical Tukey plotting position (Wilks,  
<sup>184</sup> 2011) over  $A_i$  to derive the  $P(a_i)$  probabilities across a period of interest:

$$P(A_i) = \frac{i - 0.33}{n + 0.33} \quad (3)$$

<sup>185</sup> The drought indices *SPI*, *SPEI*, *EDDI*, *SSI*, and *zcNDVI* are obtained following the inverse normal  
<sup>186</sup> approximation:

$$DI(A_i) = W - \frac{C_0 + C_1 \cdot W + c_2 \cdot W^2}{1 + d_1 \cdot W + d_2 \cdot W^2 + d_3 \cdot W^3} \quad (4)$$

<sup>187</sup>  $DI$  is referring to the drought index calculated for the variable  $V$  (i.e., SPI, SPEI, EDDI, SSI, and zcNDVI).  
<sup>188</sup> The values for the constats are:  $C_0 = 2.515517$ ,  $C_1 = 0.802853$ ,  $C_2 = 0.010328$ ,  $d_1 = 1.432788$ ,  $d_2 =$   
<sup>189</sup>  $0.189269$ , and  $d_3 = 0.001308$ . For  $P(A) \leq 0.5$ ,  $W = \sqrt{-2 \cdot \ln(P(A_i))}$ , and for  $P(A_i) > 0.5$ , replace  $P(A_i)$   
<sup>190</sup> with  $1 - P(A_i)$  and reverse the sign of  $DI(A_i)$ .

<sup>191</sup> The drought indices were calculated for time scales of 1, 3, 6, 12, 24, and 36 months at a monthly frequency  
<sup>192</sup> for 1981–2023 in order to be used for short- to long-term evaluation of drought. In the case of the proxy of  
<sup>193</sup> vegetation productivity (zcNDVI) it was calculated for a time scale of six months at monthly frequency for  
<sup>194</sup> 2000–2023. For zcNDVI, we test time scales of 1, 3, 6, and 12 months; we choose to use six months because  
<sup>195</sup> that shows a more stable representation of vegetation productivity.

### <sup>196</sup> 3.2.3. Trend of drought indices

<sup>197</sup> To estimate if there are significant positive or negative trends for the drought indices, we used the non-  
<sup>198</sup> parametric test of Mann-Kendall (Kendall, 1975). To determine the magnitude of the trend, we used Sen's  
<sup>199</sup> slope (Sen, 1968). Some of the advantages of applying this methodology are that the Sen's slope is not  
<sup>200</sup> affected by outliers as regular regression does, and it is a non-parametric method that is not influenced by  
<sup>201</sup> the distribution of the data. We applied Mann-Kendall to see if the trend was significant and Sen's slope  
<sup>202</sup> to estimate the magnitude of the trend. We did this to the six time scales from 1981 to 2023 (monthly  
<sup>203</sup> frequency) and the indices SPI, EDDI, SPEI, and SSI. Thus, we have trends per index and time scale (24 in  
<sup>204</sup> total). Then, we extracted the trend aggregated by macrozone and per land cover persistent macroclasses.

### <sup>205</sup> 3.3. Interaction of land cover and drought

#### <sup>206</sup> 3.3.1. Land cover change

<sup>207</sup> To analyze the land cover change, we use the IGBP scheme from the MCD12Q1 collection 6.1 from MODIS.  
<sup>208</sup> This product has been previously used for studies of drought and land cover in Chile (Fuentes et al., 2021;  
<sup>209</sup> Zambrano et al., 2018). From the 17 classes, we regrouped into ten macroclasses, as follows: classes 1–4 to  
<sup>210</sup> forest, 5–7 to shrublands, 8–9 to savannas, 10 as grasslands, 11 as wetlands, 12 and 14 to croplands, 13 as  
<sup>211</sup> urban, 15 as snow and ice, 16 as barren, and 17 to water bodies. Thus, we have a land cover raster time series  
<sup>212</sup> with the ten macroclasses for 2001 and 2023. We validate the land cover macroclasses regarding a highly  
<sup>213</sup> detailed (30 m of spatial resolution) land cover map made for Chile by Zhao et al. (2016) for 2013–2014.  
<sup>214</sup> Our results showed a global accuracy of ~0.82 and a F1 score of ~0.66. Section S2 in the Supplementary  
<sup>215</sup> Material shows the procedure for validation.

<sup>216</sup> We calculated the surface occupied per land cover class into the five macrozones (“Norte Grande” to  
<sup>217</sup> “Austral”) per year for 2001–2023. After that, we calculated the trend's change in surface per land cover  
<sup>218</sup> type and macroclass. We used Mann-Kendall for the significance of the trend (Kendall, 1975) and Sen's  
<sup>219</sup> slope to calculate the magnitude (Sen, 1968).

Later in this study, we will examine the variation in vegetation productivity across various land cover types and how water demand and supply, and soil moisture affect it. In order to avoid variations due to a change in the land cover type from year-to-year that will wrongly impact NDVI, we developed a persistence mask for land cover for 2001–2022. Thereby, we reduce an important source of variation on a regional scale. Therefore, we generated a raster mask for IGBP MODIS per pixel using macroclasses that remained unchanged for at least 80% of the years (2001–2022). This enabled us to identify regions where the land cover macroclasses are persistent.

### 3.3.2. Relationship between land cover and drought trends

We wanted to explore the relationship between the trend in land cover classes and the trend in the drought indices. For this purpose, in order to have more representative results, we conducted the analysis over sub-basins within continental Chile. We use 469 basins, which have a surface area between 0.0746 and 24,000 ( $km^2$ ), and a median area of 1,249 ( $km^2$ ). For each basin, we calculate the relative trend per land cover type, considering the proportion of the type relative to the total surface of the basin. Then, we extracted per basin the average trend of the drought indices SPI, SPEI, EDDI, SSI, and all their time scales 1, 3, 6, 12, 24, and 36. Also, we extracted the average trend in the proxy of vegetation productivity (zcNDVI). We wanted to analyze which drought indices and time scales have a major impact on changes in land cover type.

We have 25 predictors, including drought indices and vegetation productivity. We analyzed the 25 predictors per type of landcover, thus running six random forest models. Random forest uses multiple decision trees and allows for classification and regression. Some advantages are that it allows to find no linear relationship, reduces overfitting, and allows to derive the variable importance. We used random forests for regression and trained 1000 forests. To obtain more reliable results, we resampled by creating ten folds, running a random forest per fold, and calculating the r-squared (rsq), root mean square error (RMSE), and variable importance.

The variable importance helps for a better understanding of the relationships by finding which variable has a higher contribution to the model. We calculate the variable's importance by permuting out-of-bag (OOB) data per tree and computing the mean standard error in the OOB. After permuting each predictor variable, we repeat the process for the resting variable. We repeated this process ten times (per fold) to obtain the performance metrics (rsq, RMSE, and variable importance).

### 3.4. Drought impacts on vegetation productivity

We analyzed the trend of vegetation productivity over the unchanged land cover macroclasses. To achieve this, we used the persistent mask of land cover macroclasses. This way, we tried to reduce the noise in the vegetation due to a change in land cover from year to year. We used the zcNDVI as a proxy of vegetation productivity. In Chile's cultivated land, [Zambrano et al. \(2018\)](#) introduced the zcNDVI for assessing seasonal biomass production in relation to climate.

We examine the drought indices of water demand, water supply, and soil moisture and their correlation with vegetation productivity. The objective is to determine the impact of soil moisture and water demand and supply on vegetation productivity. We want to address three main questions: Which of the drought variables—supply, demand, or soil moisture—most helps to explain the changes in plant productivity? How do the short- to long-term time scales of the drought variable affect vegetation productivity in Chile? And finally, how strong is the relationship between the variables and the drought index? Thus, we will be able to advance in understanding how climate is affecting vegetation, considering the impact on the five land cover types: forest, cropland, grassland, savanna, and shrubland.

We conducted an analysis on the linear correlation between the indices SPI, SPEI, EDDI, and SSI over time periods of 1, 3, 6, 12, 24, and 36 months with zcNDVI. We used a method similar to that used by [Meroni et al. \(2017\)](#) which compared the SPI time-scales with the cumulative FAPAR (Fraction of Absorbed Photosynthetically Active Radiation). We performed a pixel-to-pixel linear correlation analysis for each

266 index within the persistent mask of land cover macroclasses. We first compute the Pearson coefficient of  
267 correlation for each of the six time scales. A time scale is identified as the one that attains the highest  
268 correlation ( $p < 0.05$ ). We then extracted the Pearson correlation coefficient corresponding to the time  
269 scales where the value peaked. As a result, for each index, we generated two raster maps: 1) containing  
270 the raster with values of the time scales and drought index that reached the maximum correlation, and 2)  
271 having the magnitude of the correlation obtained by the drought index at the time scales.

272 *3.5. Software*

273 For the downloading, processing, and analysis of the spatio-temporal data, we used the open source software  
274 for statistical computing and graphics, R (R Core Team, 2023). For downloading ERA5L, we used the  
275 `{ecmwfr}` package (Hufkens et al., 2019). For processing raster data, we used `{terra}` (Hijmans, 2023) and  
276 `{stars}` (Pebesma and Bivand, 2023). For managing vectorial data, we used `{sf}` (Pebesma, 2018). For the  
277 calculation of AED, we used `{SPEI}` (Beguería and Vicente-Serrano, 2023). For mapping, we use `{tmap}`  
278 (Tennekes, 2018). For data analysis and visualization, the suite `{tidyverse}` (Wickham et al., 2019) was used.  
279 For the random forest modeling, we used the `{tidymodels}`(Kuhn and Wickham, 2020) and `{ranger}`(Wright  
280 and Ziegler, 2017) packages.

281 **4. Results**

282 *4.1. Short- to long-term drought*

283 Figure 2 shows the spatial variation of the trend for the drought indices from short- to long-term scales.  
284 SPI and SPEI have a decreasing trend from “Norte Chico” to “Sur.” However, there is an increasing trend  
285 in “Austral.” The degree of the trend is stronger at higher time scales. The SSI indicates that in “Norte  
286 Grande,” there are surfaces that have increased in the southwest part and in the northeast have decreased,  
287 and is shown for all time scales. Similar to SPI and SPEI, SSI decreases at higher time scales. EDDI showed  
288 a positive trend for the whole of continental Chile, with a higher trend toward the north and a descending  
289 gradient toward the south. The degree of trend increases at higher time scales.

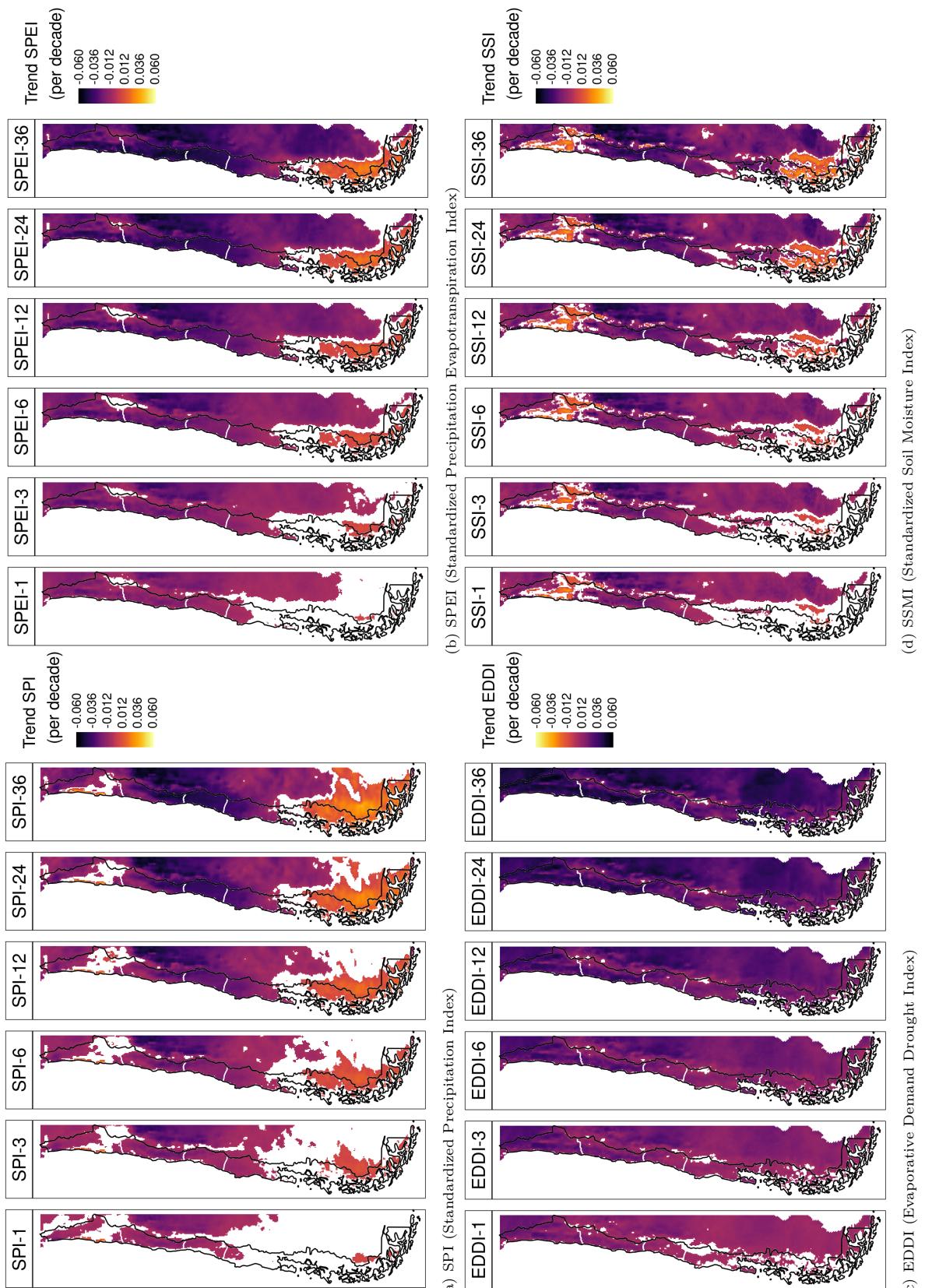


Figure 2: Linear trend of the drought index (\*) at time scales of 1, 3, 6, 12, 24, and 36 months for 1981-2023

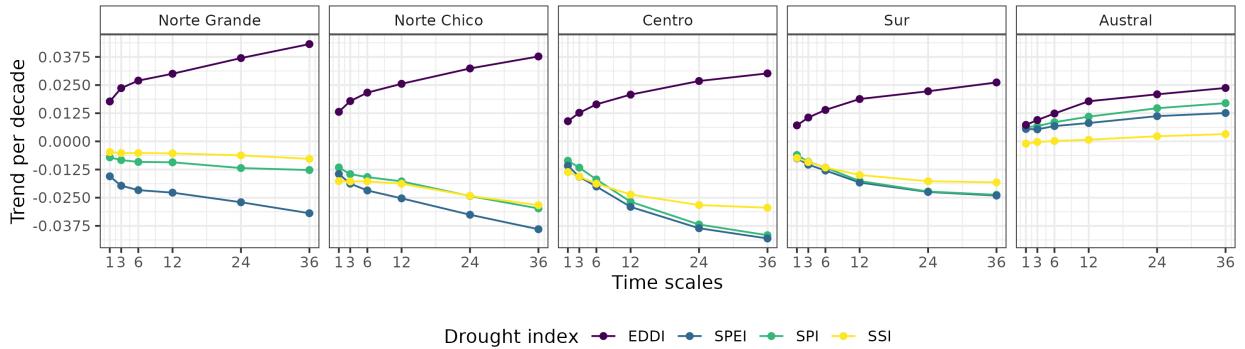


Figure 3: Trend per decade for the drought indices SPI, EDDI, SPEI, and SSI aggregated by macrozone.

The Figure 3 displays the averaged aggregation per macrozone, the drought index, and the timescale. The macrozones that reached the lowest trend for SPI, SPEI, and SSI are “Norte Chico” and “Centro,” where the indices also decrease at longer time scales. Potentially explained due to the prolonged reduction in precipitation that has affected the hydrological system in Chile. At 36 months, it reaches trends between -0.03 and -0.04 (z-score) per decade for SPI, SPEI, and SSI. For “Sur,” the behavior is similar, decreasing at longer scales and having between -0.016 and -0.025 per decade for SPI, SPEI, and SSI. “Norte Grande” has the highest trend at 36 months for EDDI (0.042 per decade), and “Centro” has the lowest for SPI and SPEI. In “Norte Grande” and “Norte Chico,” which are in a semi-arid climate, it is evident that the EDDI has an effect on the difference between the SPI and SPEI index, which is not seen in the other macrozones. Contrary to the other macrozones, “Austral” showed an increase in all indices, being the highest for EDDI at 36 months (0.025) and the lowest for SSI, which shows only a minor increase in the trend.

#### 4.2. Interaction of land cover and drought

##### 4.2.1. Land cover change

Table 2: Surface of the land cover class that persist during 2001-2022

| macrozone    | Surface [km <sup>2</sup> ] |          |           |         |           |             |
|--------------|----------------------------|----------|-----------|---------|-----------|-------------|
|              | Forest                     | Cropland | Grassland | Savanna | Shrubland | Barren land |
| Norte Grande |                            |          | 886       |         | 7,910     | 171,720     |
| Norte Chico  |                            | 90       | 4,283     | 589     | 16,321    | 84,274      |
| Centro       | 3,739                      | 1,904    | 7,584     | 19,705  | 844       | 12,484      |
| Sur          | 72,995                     | 1,151    | 7,198     | 15,906  |           | 2,175       |
| Austral      | 60,351                     |          | 54,297    | 19,007  | 249       | 7,218       |
| Total        | —                          | 137,085  | 3,145     | 74,247  | 55,206    | 25,324      |
|              |                            |          |           |         |           | 277,870     |

For vegetation, we obtained and use hereafter five macroclasses of land cover from IGBP MODIS: forest, shrubland, savanna, grassland, and croplands. Figure 1c shows the spatial distribution of the macroclasses through Chile for the year 2022. Figure 1d shows the macroclasses of land cover persistance (80%) during 2021–2022, respectively (Table 2). Within continental Chile, barren land is the land cover class with the highest surface area ( $277,870 \text{ km}^2$ ). The largest type of vegetation, with  $137,085 \text{ km}^2$ , is forest. Grassland has  $74,247 \text{ km}^2$ , savanna  $55,206 \text{ km}^2$ , shrubland  $25,341 \text{ km}^2$ , and cropland  $3,146 \text{ km}^2$  (Table 2). The macrozones with major changes for 2001–2022 were “Centro,” “Sur,” and “Austral,” with 36%, 31%, and 34% of their surface changing the type of land cover, respectively (Figure 1 and Table 3). Figure 4 shows the summary of the proportion of surface per land cover class and macrozone, derived from the persistance mask over continental Chile.

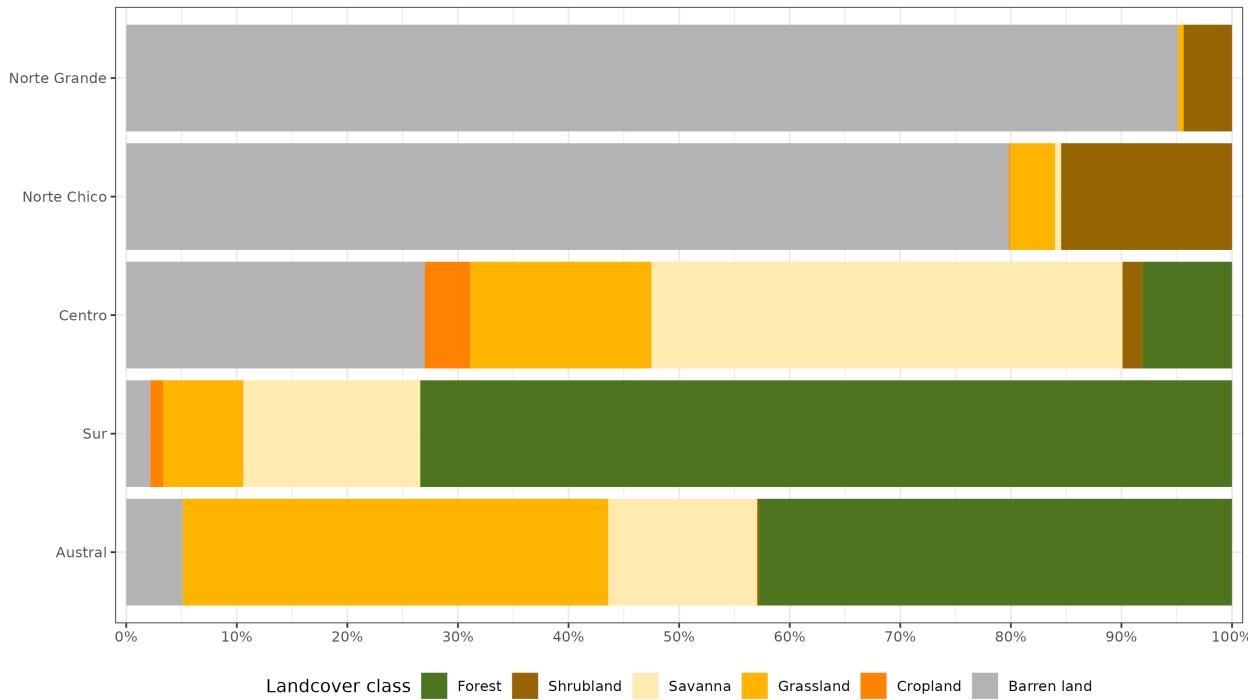
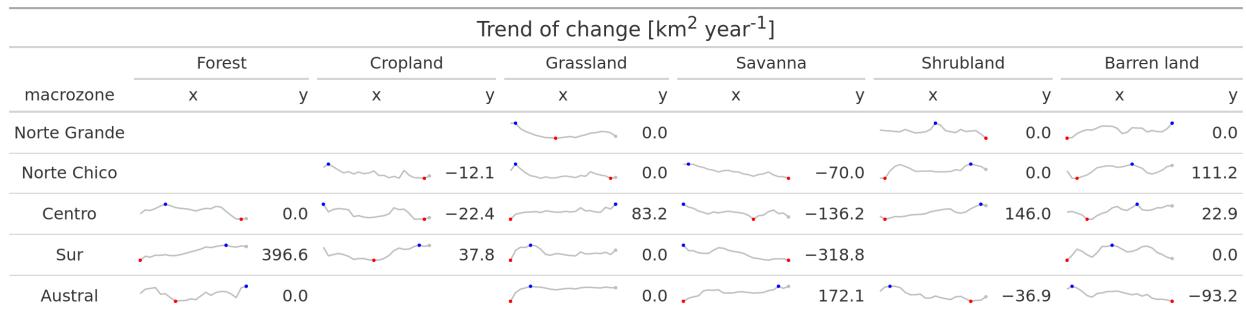


Figure 4: Proportion of land cover class from the persistent land cover for 2001-2022 (>80%) per macrozone

Table 3: The value of Sen's slope trend next to the time-series plot of surface per land cover class (IGBP MCD12Q1.016) for 2001–2022 through Central Chile. Values of zero indicate that there was not a significant trend. Red dots on the plots indicate the maximum and minimum values of surface.



From the trend analysis in Table 3, we can indicate that the “Norte Chico” shows an increase in barren land of  $111 \text{ km}^2 \text{ yr}^{-1}$  and a reduction in the class savanna of  $70 \text{ km}^2 \text{ yr}^{-1}$ . In the “Centro” and “Sur,” there are changes with an important reduction in savanna ( $136$  to  $318 \text{ km}^2 \text{ yr}^{-1}$ ), and an increase in shrubland and grassland. Showing a change for more dense vegetation types. The area under cultivation (croplands) appears to be shifting from the “Centro” to the “Sur.” Also, there is a high increase in forest ( $397 \text{ km}^2 \text{ yr}^{-1}$ ) in the “Sur,” seemingly replacing the savanna lost (Table 3).

#### 4.2.2. Relationship between drought indices and land cover change

According to Table 4, the random forest models for estimating the landcover trend from the trends in drought indices reach an r-squared between 0.32 and 0.39 for the types of forest, grassland, savanna, shrubland, and barren land. It is more likely that short- and medium-term increases in AED (EDDI-6 and

Table 4: The five most important trends of drought indices in estimating the landcover trend per land cover type and the r-squared (rsq) reached by each random forest model.

| Position | Forest<br>(rsq=0.32) | Cropland<br>(rsq=0.06) | Grassland<br>(rsq=0.39) | Savanna<br>(rsq=0.23) | Shrubland<br>(rsq=0.23) | Barren_land<br>(rsq=0.32) |
|----------|----------------------|------------------------|-------------------------|-----------------------|-------------------------|---------------------------|
| 1        | EDDI-36              | EDDI-36                | EDDI-6                  | EDDI-6                | EDDI-6                  | EDDI-12                   |
| 2        | EDDI-24              | SSI-36                 | EDDI-12                 | EDDI-12               | SPI-36                  | EDDI-6                    |
| 3        | EDDI-12              | EDDI-24                | EDDI-24                 | SPI-36                | SPEI-36                 | SPI-6                     |
| 4        | SSI-36               | EDDI-12                | SPEI-6                  | EDDI-36               | EDDI-3                  | SPEI-6                    |
| 5        | SSI-6                | SSI-24                 | SPI-6                   | EDDI-24               | SPI-24                  | EDDI-24                   |

323 EDDI-12) and short-term precipitation deficits (SPI-6 and SPEI-6) contributed to changes in grassland and  
 324 bare land. The short-term increase of AED (EDDI-3 and EDDI-6) and the longer duration of the precipitation  
 325 deficit (SPI-24, SPI-36, and SPEI-36) most likely contribute to the changes in shrubland. The changes  
 326 in savanna are associated with a short- and long-term increase in AED and a three-year precipitation deficit  
 327 (SPI-36). The increase in cumulative AED from 12 to 36 months is the strongest associated variable that  
 328 contributes to changes in forests, followed by the reduction of soil moisture over six and 36 months. The  
 329 supplementary material in Section S3 provides further details about the variable's importance.

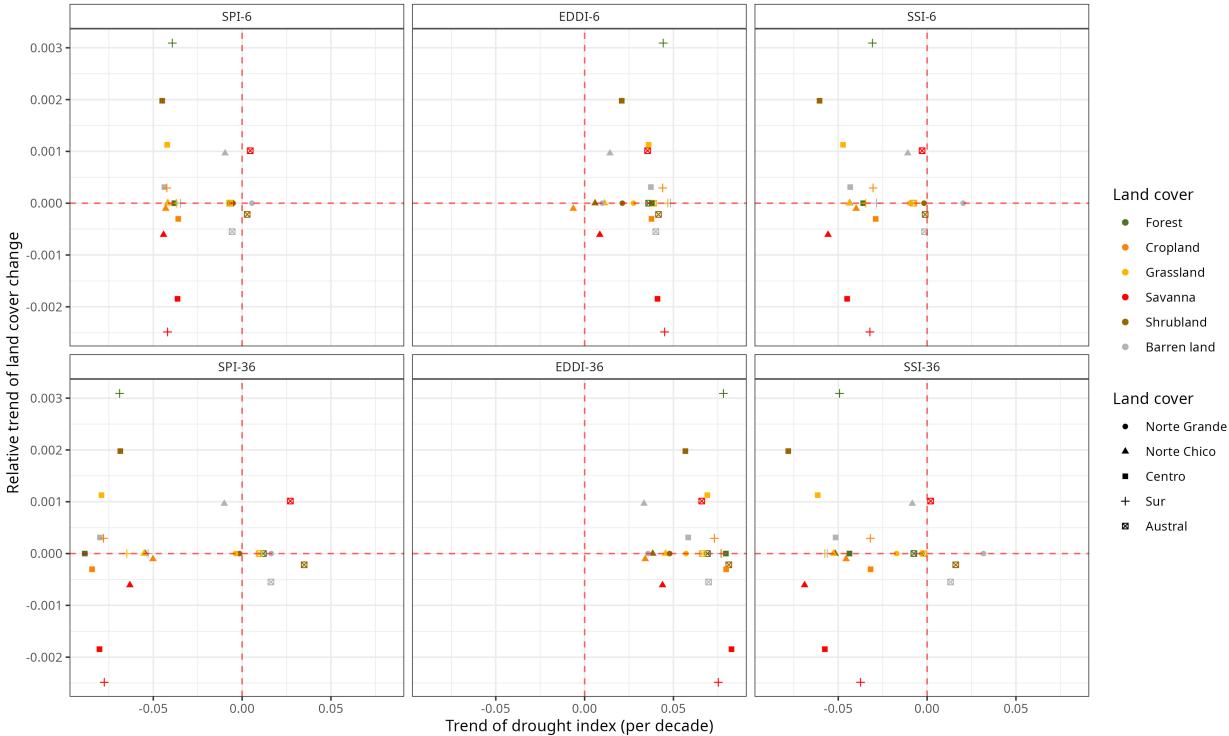


Figure 5: Relationship between the trend in land cover change (y-axis) and the trend in drought indices (x-axis) for the five macrozones. Vertical panels correspond to 1, 3, 6, 12, 24, and 36 months of the time scale by drought index. Horizontal panels show each drought index

330 We study the connection between the SPI, EDDI, and SSI drought indices and changes in land cover in  
 331 Figure 5. To do this, we compare the relative changes in land cover (in terms of the total surface area per

332 land cover type and macrozone) over six and thirty-six months. Figure 5 shows that the forest in the “Sur,”  
 333 shrubland and grassland in “Centro,” barren land in “Norte Chico,” and savanna in “Austral” showed an  
 334 increase in the surface of landcover associated with an increase in EDDI. Savanna in “Centro,” “Sur,” and  
 335 “Norte Chico” decreases with the increase in EDDI. The SPI and SSI showed similar behavior regarding  
 336 the trend in land cover type. A decrease in SPI and SSI is associated with an increase in the surface in  
 337 shrubland and grassland in “Centro,” forest in “Sur,” and barren land in “Norte Chico,” as well as a  
 338 decrease trend in savanna in “Norte Chico,” “Centro,” and “Sur.”

339 *4.3. Drought impacts on vegetation productivity within land cover*

340 *4.3.1. Trends in vegetation productivity*



Figure 6: (a) Map of the linear trend of the index zcNDVI for 2000–2023. Greener colors indicate a positive trend; redder colors correspond to a negative trend and a decrease in vegetation productivity. Grey colors indicate either no vegetation or a change in land cover type for 2001–2022. (b) Temporal variation of zcNDVI-6 aggregated at macrozone level within continental Chile. Each horizontal panel corresponds to a macrozone from ‘Norte Grande’ to ‘Austral’.

341 The temporal variation within the macrozones is shown in Figure 6b). There is a negative trend in “Norte  
 342 Chico” with -0.035 and “Centro” with -0.02 per decade. Vegetation reached its lowest values for 2019-2022,  
 343 with an extreme condition in early 2020 and 2022 in the “Norte Chico” and “Centro”. The “Sur” and  
 344 “Austral” show a positive trend of around 0.012 and 0.016, respectively, per decade (Figure 6).

345 In Figure 6 it is showed the spatial map of trends in zcNDVI (Figure 6a). In “Norte Grande,” vegetation  
 346 productivity, as per the z-index, exhibits a yearly increase of 0.027 for grassland and 0.032 for shrubland. In  
 347 the “Norte Chico,” savanna has the lowest trend of -0.062, cropland -0.047, shrubland -0.042, and grassland  
 348 -0.037. In “Centro,” shrubland reaches -0.07, savanna -0.031, cropland -0.024, forest -0.017, and grassland  
 349 -0.005 per decade. This decrease in productivity could be associated either with a reduction in vegetation  
 350 surface, a decrease in biomass, or browning.

351 *4.3.2. Correlation between vegetation productivity and drought indices*

352 Figure 7 shows the highest coefficient of determination ( $r^2$ , or rsq) found in the regression analysis  
 353 between different drought indicators and zcNDVI over time scales (1, 3, 6, 12, 24, and 36 months). The  
 354 spatial variation of time scales reached per index is mostly for time scales above 12 months. In the case of  
 355 SSI, the predominant scales are 6 and 12 months. For all indices, to the north, the time scales are higher

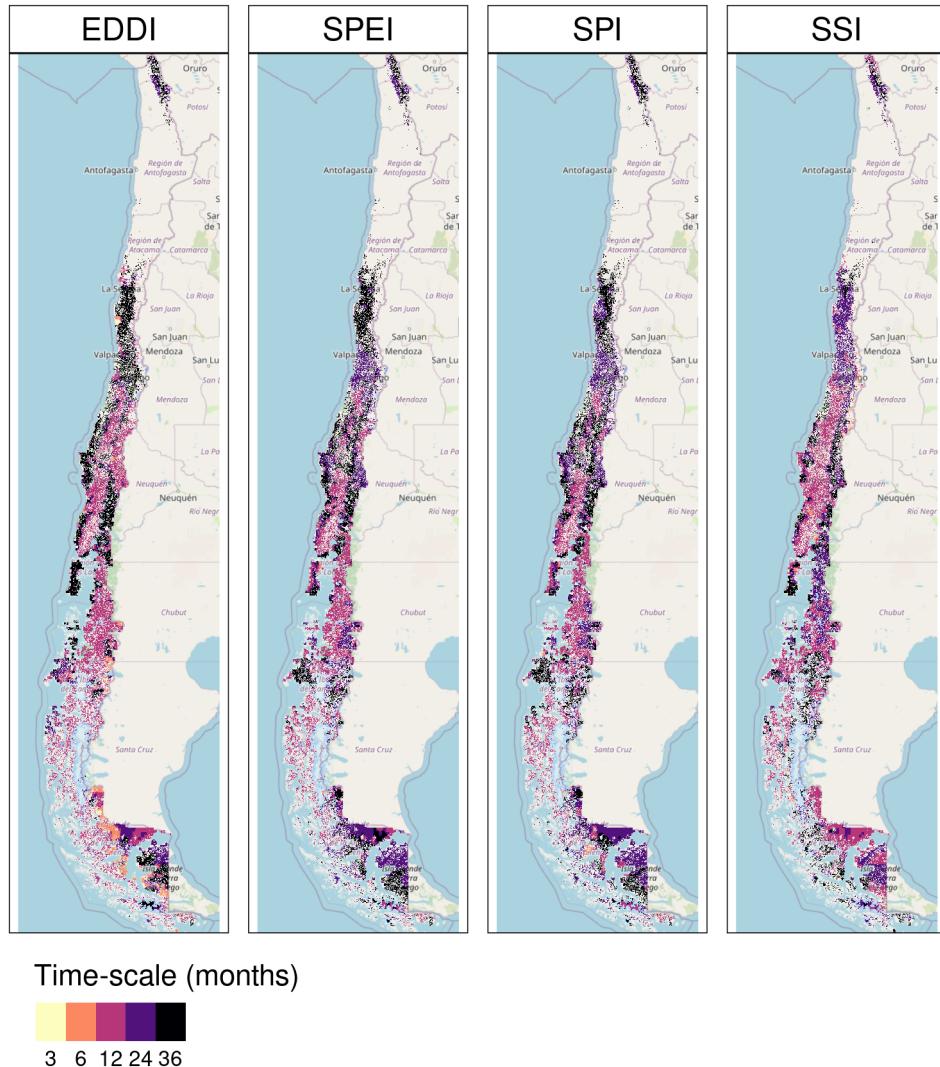


Figure 7: Time scales per drought index that reach the maximum coefficient of determination

and diminish toward the south until the south part of “Austral,” where they increase. In Figure 8, the map of Pearson correlation values ( $r$ ) is shown. The EDDI reached correlations above 0.5 between “Norte Chico” and “Sur.” The correlation changes from negative to positive toward the Andes Mountains and to the sea, just as in the northern part of “Austral.” The SPI and SPEI have similar results, with the higher values in “Norte Chico” and “Centro” being higher than 0.6. Following a similar spatial pattern as EDDI but with an opposite sign. The SSI showed to be the index that has a major spatial extension with a higher correlation. It has a similar correlation to SPI and SPEI for “Norte Chico” and “Sur,” but has improvements for “Sur.”

In Table 5, we aggregate per macrozone and landcover the correlation analysis presented in Figure 7 and Figure 8. According to what is shown, forests seem to be the most resistant to drought. Showing that only “Centro” is slightly ( $rsq = 0.25$ ) impacted by a 12-month soil moisture deficit (SSI-12). In the “Norte Chico” and to a lesser extent in the “Norte Grande,” it is evident that a SSI-12 with a  $rsq = 0.45$  and a decrease in water supply (SPI-36 and SPEI-24 with  $rsq = 0.28$  and 0.34, respectively) have an impact on grasslands. However, this type was unaffected by soil moisture, water supply, or demand in macrozones further south. The types that show to be most affected by variation in climate conditions are shrublands,

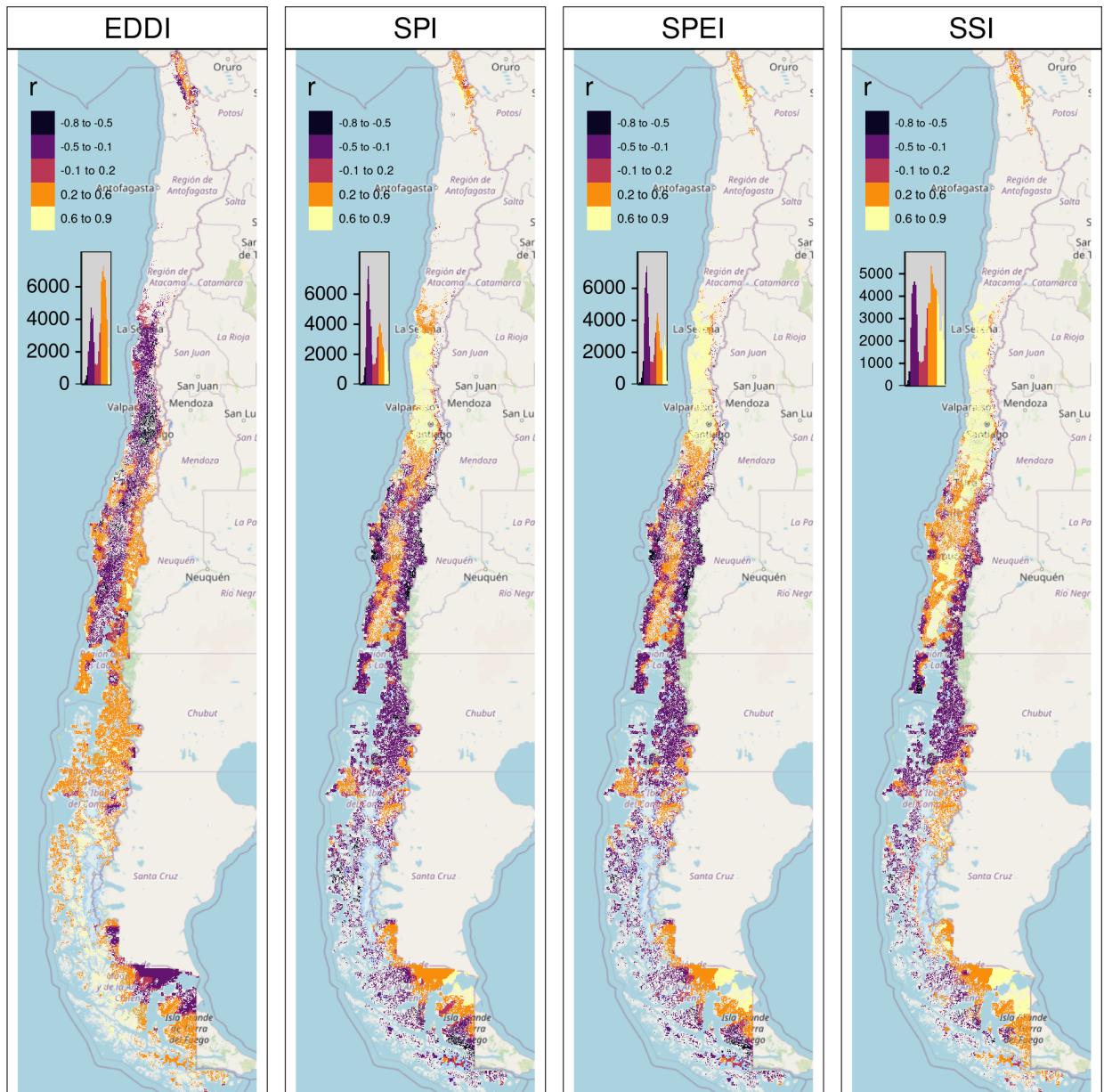
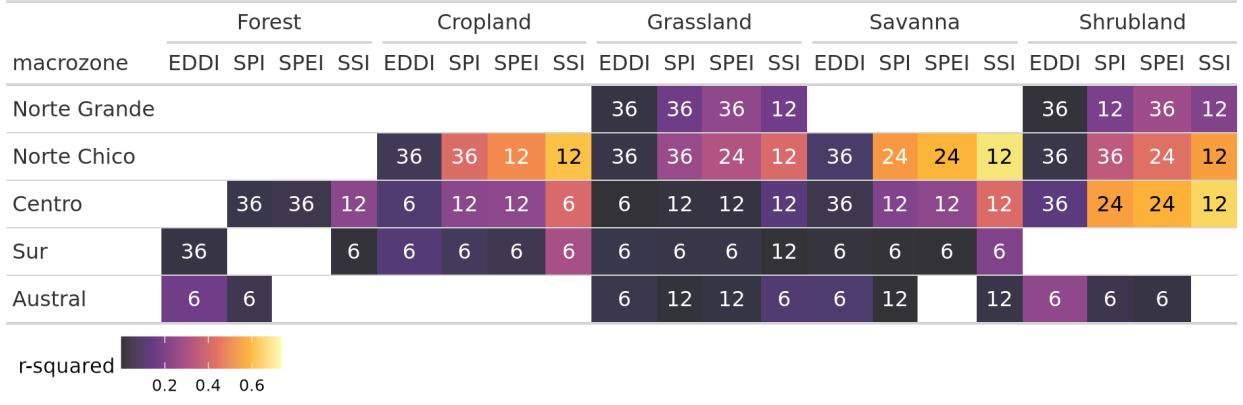


Figure 8: Pearson correlation value for the time scales and drought index that reach the maximum coefficient of determination

370 savannas, and croplands. For savannas in “Norte Chico,” the SSI-12 and SPI-24 reached an rsq of 0.74  
 371 and 0.58, respectively. This value decreases to the south, but the SSI-12 is still the variable explaining  
 372 more of the variation in vegetation productivity (rsq = 0.45 in “Centro” and 0.2 in “Sur”). In the case  
 373 of croplands, the SPEI-12, SPI-36, and SSI-12 explain between 45% and 66% of the variability in “Norte  
 374 Chico.” The type of land most impacted by climatic variation was shrubland, where soil moisture explained  
 375 59% and precipitation, 37%, in “Norte Chico” and “Centro,” with SSI-12 being the most relevant variable,  
 376 then SPI-36 in “Norte Chico” and SPI-24 in “Sur.”

Table 5: Summary per land cover macroclass and macrozone regarding the correlation between zcNDVI with the drought indices EDDI, SPI, SPEI, and SSI for time scales of 1, 3, 6, 12, 24, and 36. The numbers in each cell indicate the time scale that reached the maximum correlation for the land cover and macrozone, and the color indicates the strength of the r-squared obtained with the index and the time scale.



## 5. Discussion

### 5.1. Drought trend and attribution to land cover change

Vicente-Serrano et al. (2022), in a study at the global scale of drought trends, indicates that there have not been significant trends in meteorological drought since 1950. Also, state that the increase in hydrological trend in some parts of the globe (northeast Brazil and the Mediterranean region) is related to changes in land cover and specifically to the rapidly increasing irrigated area, which consequently increases water extraction. Kogan et al. (2020) analyzed the agricultural drought impact globally and in the main grain producer countries, finding that “since 1980, the Earth warming has not changed the drought area or intensity.” In our study, we took into account the variation in vegetation productivity in Chile, specifically in areas without any changes in land cover, to prevent any misleading conclusions about the increase in water demand due to land cover change. Our results show a contrasting perspective. The SPI, SPEI, and SSI (water supply) showed a decrease in trends, except for the southern part, and an increase in EDDI (water demand). The trend, positive or negative, was stronger as the time scales increased. Trends in the long term (e.g., 36 months) are evidence of how human-induced climate change is affecting Chile, which seems to be due to an intense hydrological drought resulting from the persistence of the precipitation deficit. We found that there has been a significant trend in the decline of vegetation productivity (zcNDVI) since 2000 for the north-central part of the country, which has been extreme between 2020 and 2022 and has impacted natural and cultivated land. Additionally, we demonstrated that the drought, primarily due to an increase in AED, accounts for about 30% of the changes in land cover types (excluding croplands). These changes are associated with a decrease in water demand from vegetation. Moreover, the most water-demanding type, cropland, showed a decrease in the north-central region, while barren land showed an increase. The north-central part of the country primarily experienced these changes due to a higher increase in AED. Thus, we have evidence of a significant decline in water supply and an increase in AED for the north-central part of Chile, which show to be the most relevant variables for drought conditions. Some questions arise regarding what is occurring with the cultivated land. We used the unchanged land cover to ensure that an increase in surface area is not considered in the trend analysis. For croplands, it could happen that some areas have changed the types of crops for others with higher water demand, which consequently increases water demand. However, this effect should be minor compared to the decrease in water supply and increase in water demand at this scale of analysis.

This shows that the main cause of the hydrological drought in Chile was a steady drop in water supply made worse by an increase in AED, but it seems that in zones most affected by drought, the main cause is

408 not an increase in water demand due to an intensification of cultivated land (e.g., an increase in irrigated  
409 crops). North-central Chile has experienced a decline in vegetation productivity across land cover types,  
410 which is primarily attributable to variations in water supply and soil moisture. An increase in water demand,  
411 such as an increase in the surface area of irrigated crops, could strengthen this trend.

412 *5.2. Impact of short- to long-term droughts on land cover types*

413 We found that the 12-month soil moisture deficit most affects the productivity of vegetation in all land  
414 cover types along Chile. In a study in the Yangtze River Basin in China, [Jiang et al. \(2020\)](#) analyzed the  
415 impact of drought on vegetation using the SPEI and the Enhanced Vegetation Index (EVI). They found  
416 that cropland was more sensitive to drought than grassland, showing that cropland responds strongly to  
417 short- and medium-term drought (< SPEI-6). In our case, the SPEI-12 was the one that most impacted the  
418 croplands in “Norte Chico” and “Centro.” In general, most studies show that croplands are most sensitive  
419 to short-term drought (< SPI-6) ([Zambrano et al., 2016](#); [Potopová et al., 2015](#); [Dai et al., 2020](#); [Rhee et al.,  
420 2010](#)). Short-term precipitation deficits impact soil water, and thus less water is available for plant growth.  
421 However, we found that in “Norte Chico,” an SPI-36 and SPEI-12 had a higher impact, which are associated  
422 with hydrological drought (long-term), and in “Centro,” an SPI-12 and SPEI-12. Thus, we attribute this  
423 impact to the hydrological drought that has decreased groundwater storage ([Tauçare et al., 2024](#)), which in  
424 turn is impacted by long-term deficits, and consequently, the vegetation is more dependent on groundwater.  
425 In “Sur” and “Austral,” the correlations between drought indices and vegetation productivity decrease, as  
426 do the time scales that reach the maximum r-squared. The possible reason for this is that the most resistant  
427 types, forest and grassland, predominate south of “Centro.” Also, drought episodes have been less frequent  
428 and intense. The drought episodes have had a lower impact on water availability for vegetation.

429 According to [Senf et al. \(2020\)](#), severe drought conditions in Europe are a significant cause of tree mortality.  
430 However, we discovered that forests, as the most resilient land cover class to drought, experience less variation  
431 in drought indices. Supporting this is [Fathi-Taperasht et al. \(2022\)](#), who asserts that Indian forests are the  
432 most drought-resistant and recover rapidly. Similarly, the work of [Wu et al. \(2024\)](#), who analyzed vegetation  
433 loss and recovery in response to meteorological drought in the humid subtropical Pearl River basin in China,  
434 indicates that forests showed higher drought resistance. Using Vegetation Optical Depth (VOD), kNDVI,  
435 and EVI, [Xiao et al. \(2023\)](#) tests the resistance of ecosystems and finds that ecosystems with more forests  
436 are better able to handle severe droughts than croplands. They attribute the difference to a deeper rooting  
437 depth for trees, a higher water storage capacity, and different water use strategies between forest and  
438 cropland ([Xiao et al., 2023](#)). In contrast, [Venegas-González et al. \(2023\)](#), who studied Cryptocarya alba and  
439 Beilschmiedia miersii (both from the Lauraceae family) that live in sclerophyllous forests in Chile, found  
440 that the trees’ overall growth had slowed down. This could mean that the natural dynamics of their forests  
441 have changed. They attributed it to the cumulative effects of the unprecedented drought (i.e., hydrological  
442 drought).

443 Thus, we attribute that forest to being the most resistant to drought, due to the fact that most of the species  
444 comprising it are highly resilient to water scarcity compared to the other land cover classes. Nonetheless, if  
445 we want to go deep in our analysis, we should use earth observation data that is able to capture a higher  
446 level of detail. For example, when we used MOD13A3 with a 1km spatial resolution to measure vegetation  
447 condition, it took the average condition of 1 square kilometer. Then, to use remote sensing to look at how a  
448 certain type of forest (like sclerophyllous forest) changes in response to drought on a local level, we should  
449 use operational products with higher spatial resolutions, like those from Landsat or Sentinel. This will let  
450 us do a more thorough analysis.

451 *5.3. Vegetation productivity and drought.*

452 The main external factors that affect biomass production by vegetation are actual evapotranspiration and  
453 soil moisture, and the rate of ET in turn depends on the availability of water storage in the root zone.

454 Thus, soil moisture plays a key role in land carbon uptake and, consequently, in the production of biomass  
455 (Humphrey et al., 2021). Moreover, Zhang et al. (2022) indicate there is a bidirectional causality between  
456 soil moisture and vegetation productivity. Lastly, some studies have redefined agricultural drought as soil  
457 moisture drought from a hydrological perspective (Van Loon et al., 2016; Samaniego et al., 2018). Even  
458 though soil moisture is the external factor most determinant of vegetation biomass, there are multiple internal  
459 factors, such as species, physiological characteristics, and plant hydraulics, that would affect vegetation  
460 productivity. Because of that, we believe that agricultural drought, referring to the drought that impacts  
461 vegetation productivity, is the most proper term, as originally defined by Wilhite and Glantz (1985).

462 The study results showed that the soil moisture-based drought index (SSI) was better at explaining vegeta-  
463 tion productivity across land cover macroclasses than meteorological drought indices like SPI, SPEI, and  
464 EDDI. In the early growing season and especially in irrigated rather than rainfed croplands, soil moisture  
465 has better skills than SPI and SPEI for estimating gross primary production (GPP). This according to  
466 Chatterjee et al. (2022) evaluation of the SPI and SPEI and their correlation with GPP in the CONUS.  
467 Also, Zhou et al. (2021) indicate that the monthly scaled Standardized Water Deficit Index (SWDI) can  
468 accurately show the effects of agricultural drought in most of China. Nicolai-Shaw et al. (2017) also looked  
469 at the time-lag between the SWDI and the Vegetation Condition Index (VCI). They found that there was  
470 little to no time-lag in croplands but a greater time-lag in forests.

471 In our case, there is strong spatial variability throughout Chile and between classes, mainly attributable to  
472 climate heterogeneity, hydrological status, or vegetation resistance to water scarcity. The semi-arid “Norte  
473 Chico” and the Mediterranean “Centro” were where SSI had the best performance. In Chile, medium-term  
474 deficits of 12 months are more relevant in the response of vegetation, which decreases to the south, and in the  
475 case of croplands, they seem to react in a shorter time, with six months (SSI-6) in “Centro.” This variation  
476 for croplands could be related to the fact that in “Norte Chico,” the majority of crops are irrigated, but  
477 to the south there is a higher proportion of rainfed agriculture, which is most dependent on the short-term  
478 availability of water. Rather, in the “Norte Chico,” the orchards are more dependent on the storage of water  
479 in dams of groundwater reservoirs, which are affected by long-term drought (e.g., SPI-36).

#### 480 5.4. Future outlook (to complete)

## 481 6. Conclusion

482 There is a trend toward decreasing water supply in most parts of Chile, particularly in the “Centro” and  
483 “Norte Chico” regions. The whole country showed an increase in AED. Vegetation productivity only showed  
484 a decrease in the “Norte Chico” and “Centro,” being highest for shrubland and croplands. Forest is the land  
485 cover most resistant to drought, as shown along Chile, and shrubland and cropland are the most sensitive.

486 A soil moisture deficit of 12 months (SSI-12) is highly correlated with vegetation productivity for the land  
487 cover classes of shrubland, savannas, croplands, and forest in “Norte Chico” and “Centro.” For the southern  
488 part of the country with humid conditions, the correlation with SSI decreases. Soil moisture overcomes  
489 the capacity to explain vegetation productivity over the supply and demand drought indices in the entire  
490 territory.

491 The variation in vegetation productivity appears to be associated with climate variation rather than an-  
492 thropogenic factors (e.g., an increase in water demand by irrigated crops). Even though switching to more  
493 demanding crops on the land could increase the impact of drought on vegetation, this would need to be  
494 more thoroughly investigated, for instance at the watershed level.

495 The results of this study could help in the development of a robust forecasting system for land cover classes  
496 in Chile, helping to improve preparedness for climate change impacts on vegetation.

497    **Supplementary material**

498    **References**

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