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TECHNICAL MEMORANDUM

## Upper Klamath Lake 2012 Data Summary Report



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## INTRODUCTION

The Klamath Tribes have been monitoring water quality in Upper Klamath Lake (UKL) since 1990. These data have been described and summarized to varying degrees in a series of reports and manuscripts (e.g., Kann 1998; Kann and Smith 1999; Kann and Welch 2005; Kann 2007 through Kann 2012). The UKL electronic water quality database was previously updated with 2012 data and appropriate quality assurance analyses (see *Excel spreadsheet: Klamath Tribes UKL Water Quality Data 1990-2012\_ver\_4-15-13.xls*). A recent report provides additional detail and more comprehensive analysis of the 1990-2009 database (Jassby and Kann 2010). The current 2012 data report is intended to serve as an annual update to the UKL water quality database, including a summary of 2012 data (basic summary statistics and graphical analysis), and limited comparison of inter-annual trends of UKL data collected for the 23 year period between 1990 and 2012.

## METHODS

Methods followed the Klamath Tribes established procedures for field collection and laboratory analysis of water quality parameters (see Klamath Tribes 2013a,b for a complete description of these methods). Beginning in 2008 for nutrient parameters and 2009 for Chlorophyll-a (CHL), laboratory analyses transitioned from Aquatic Research, INC. in Seattle WA to the Sprague River Water Quality Laboratory in Chiloquin OR. During the transition period duplicate samples were analyzed by both laboratories to confirm parameter reproducibility. During the 2012 sampling season limnological data (Table 1) were collected biweekly from the end of April through October at 10 standardized stations in UKL and Agency Lake (Figure 1; Figure 2).

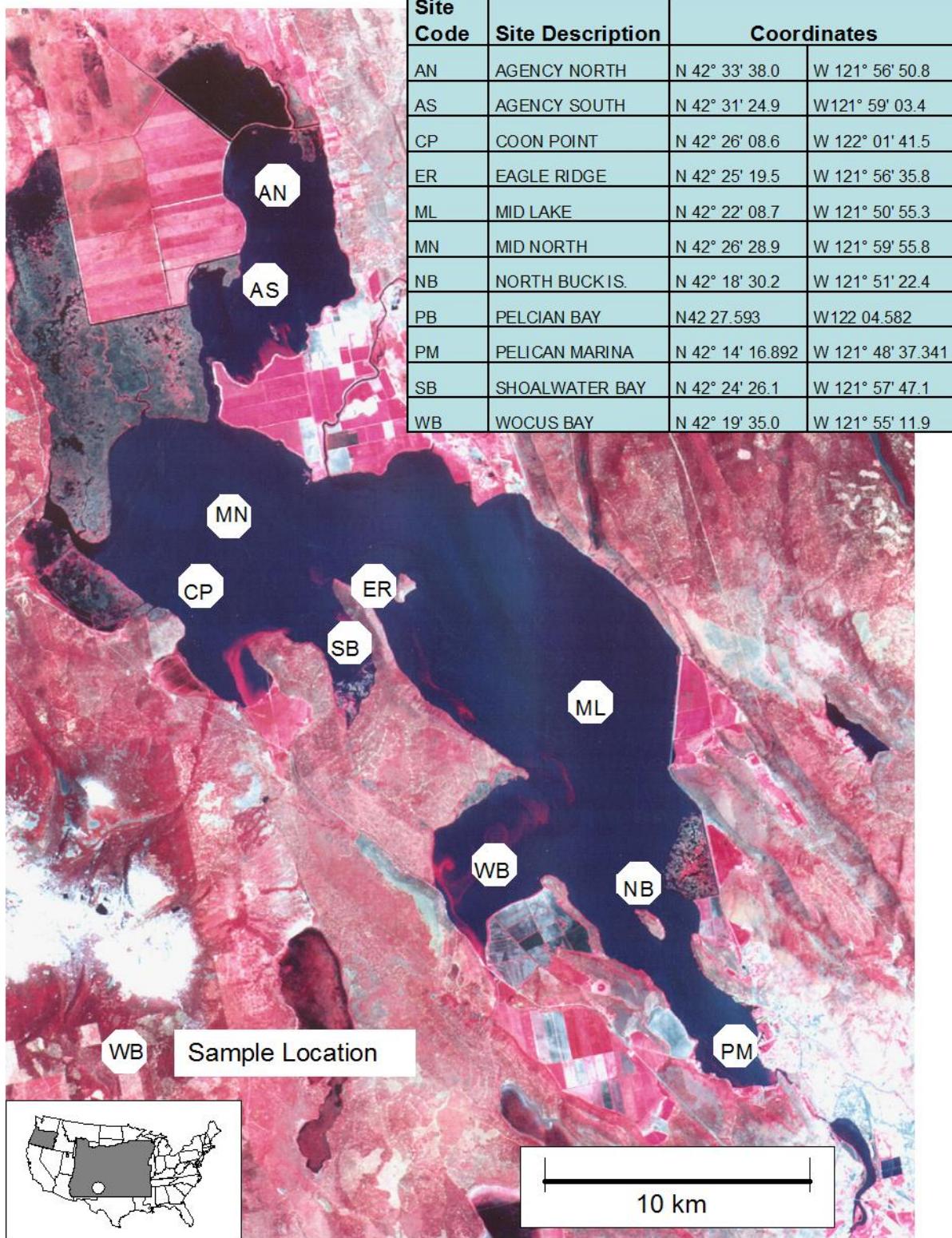
**Table 1. Limnological parameters sampled in Upper Klamath Lake, 2011.**

Parameter	Abbreviation/ Unit	Profile <sup>a</sup>	Grab <sup>b</sup>
Temperature	T (°C)	X	
Dissolved Oxygen	DO (mg/L)	X	
pH	pH	X	
Specific Conductivity	(µSiemens/cm)	X	
Secchi Transparency	Secchi (m)		
Light (Photosynthetically Active Radiation)	PAR (uEm <sup>-2</sup> s <sup>-1</sup> )	X	
Total Phosphorus	TP (µg/L)		X
Soluble Reactive phosphorus	SRP (µg/L)		X
Total Nitrogen	TN (µg/L)		X
Ammonia Nitrogen	NH <sub>4</sub> -N (µg/L)		X
Nitrate-Nitrite Nitrogen	NO <sub>3</sub> + NO <sub>2</sub> -N (µg/L)		X
Silica	SiO <sub>2</sub> (µg/L)		
Chlorophyll <i>a</i>	CHL (µg/L)		X
Phytoplankton Species Composition and Biomass <sup>c</sup>	(mm <sup>3</sup> /L)		X
Zooplankton Species Composition and Biomass <sup>c</sup>	(mg/L)		X

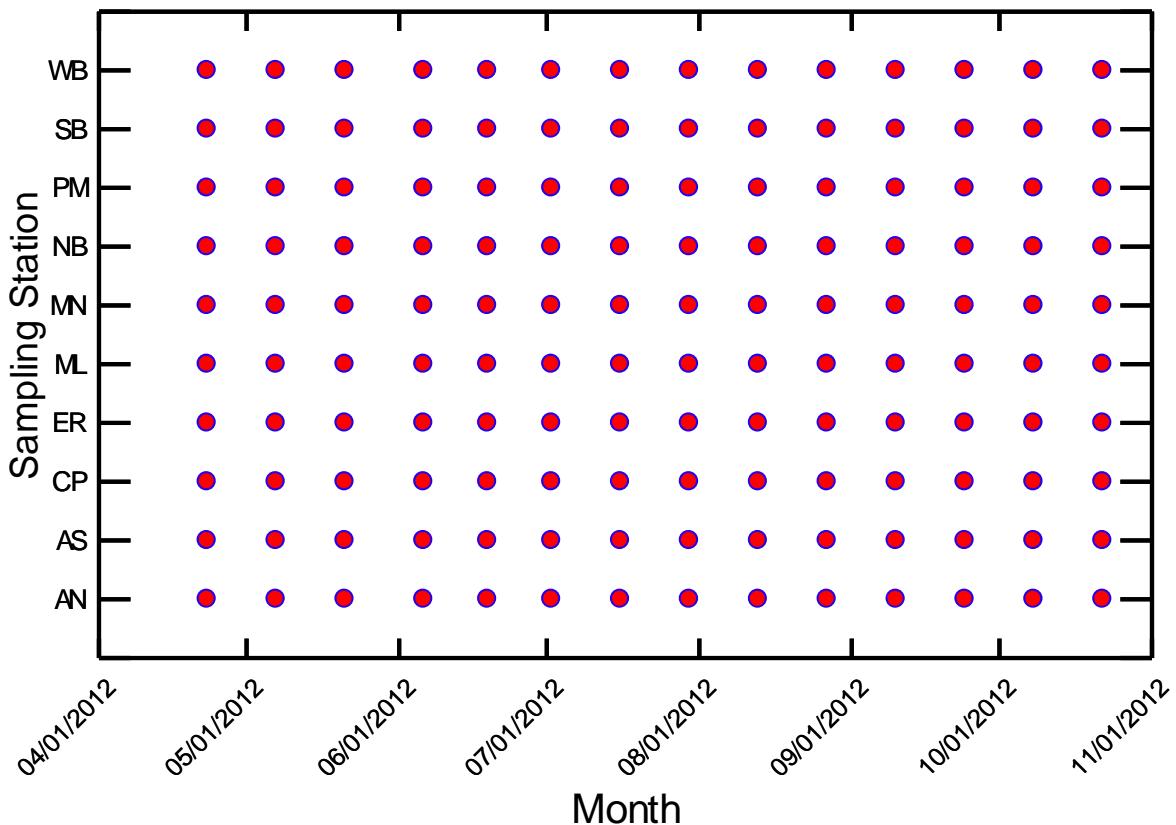
a Profile = collected with multi-parameter WQ probe at multiple depths in water column

b Grab = integrated water column sample collected with “tube sampler” except for zooplankton which was collected with a Schindler-Patalis Trap

c. Phytoplankton and zooplankton data are compiled in spreadsheets provided separately and are not analyzed herein.



**Figure 1. Location of Upper Klamath Lake sampling stations, 2012.**



**Figure 2. Spatial-temporal sampling matrix for Upper Klamath Lake, 2012.**

Data reduction consisted of computation of both lake-wide means on a given sample date and of growing season (June-September sample dates) means. Because of bloom timing differences between Upper Klamath and Agency Lake (e.g., see Kann 1998), lake-wide means and analyses are shown separately for Upper Klamath Lake only and Agency Lake only. Chlorophyll and nutrient data tended to be either normally or log-normally distributed both within a date and seasonally. Based on a comparison of both log transformed ( $\log_{10}$  or  $\log_{10}(x+1)$ ) and non-transformed data with the normal distribution using Kolmogorov-Smirnov one-sample tests or the Shapiro-Wilk standard test for normality (cf. Systat® 2004), the geometric mean tended to provide the best estimate of lake-wide or seasonal central tendency. Lake-wide variability is shown via boxplots which convey the median, interquartile range and outliers. In addition to median and interquartile values, lake-wide central tendency may be portrayed as a mean and standard error.

Nutrient quality assurance/quality control analyses are shown in the accompanying data spreadsheet (*Klamath Tribes UKL Water Quality Data 1990-2012\_ver\_4-15-13.xls*)

## RESULTS/DISCUSSION

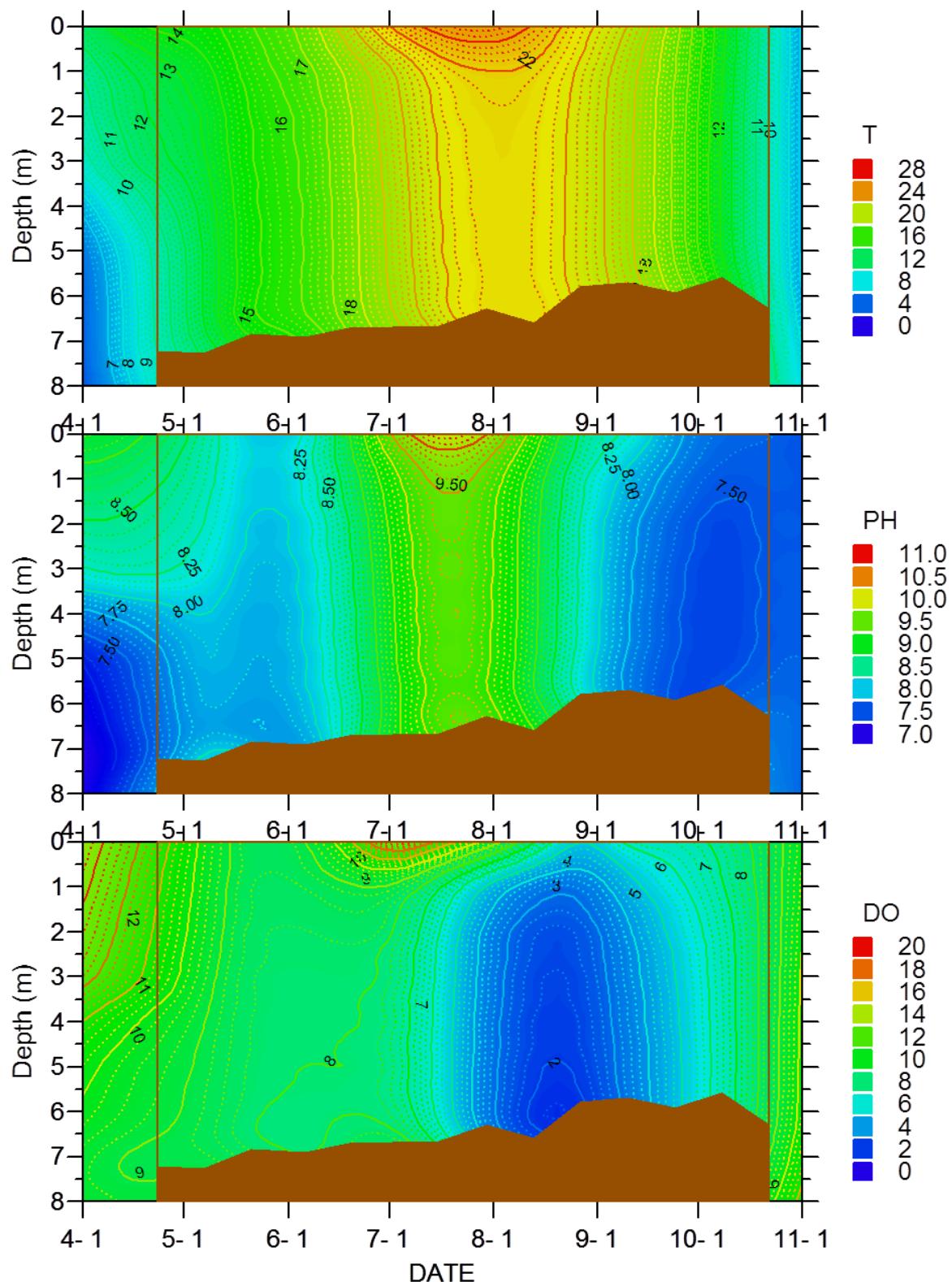
### *Seasonal and Water Column Trends in Profile Water Quality Data (T, DO, and pH)*

Water column and seasonal trends in T, DO, and pH are important aspects of water quality dynamics and fish habitat in UKL. Depth-time plots of isotherms and isopleths for these parameters allows both seasonal and depth distribution to be evaluated simultaneously. These are plotted below for two representative stations, ER located in the deep trench area, and MN located in an open-water area in the northern part of the lake (Figure 3; Figure 4). At both stations temperature ranged between 11-14 °C during late-April and early-May, with warming occurring in mid-May (~16 °C), with little additional warming into early-June (in fact Figure 7, below indicates an approximate 2 °C cooling in early-June). Overall this is in contrast to 2011 when temperatures generally remained below 12 °C into early-June. Warming continued to occur during the second half of June, although temperatures were still only ~18 °C, and only rose slightly to <20 °C through the 1<sup>st</sup> week in July. Water temperature then increased again in mid-to late-July, before peaking during the mid-August period (Figure 3; Figure 4). Maximum surface and water column temperatures occurred during the late-July to early- September period, with seasonal cooling beginning in late-August when a temperature drop of 2 °C occurred (from ~22 to <20 °C).

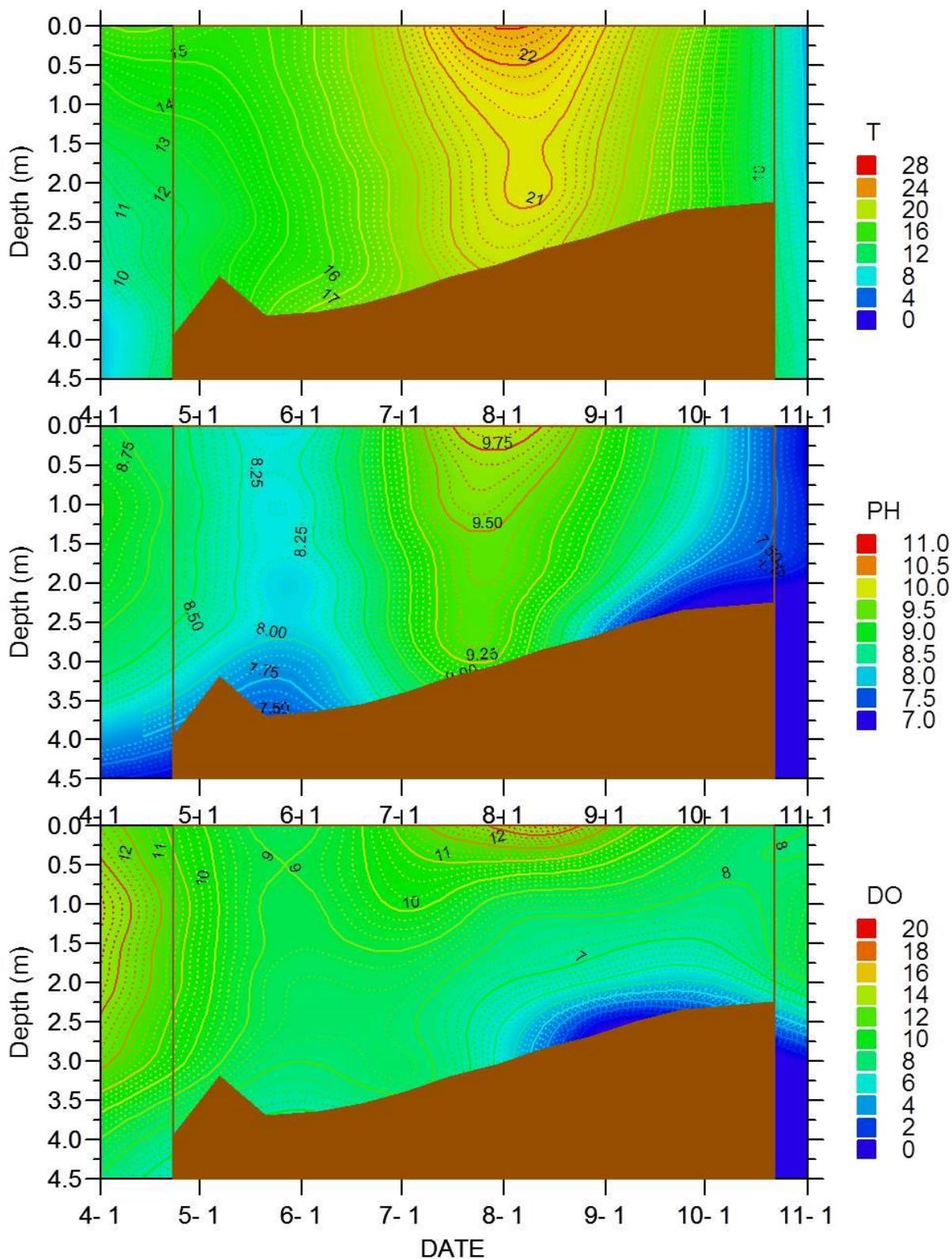
Unlike 2010 when water column pH initially increased (>9.0) in late-April and early-May (lake observations at that time indicated a massive diatom bloom and further confirmation showed very high biomass of the diatom *Asterionella formosa*), pH in 2012 was similar to 2011 and remained relatively low (<8.75) until late-June. A sharp increase to seasonal peak values then occurred by mid- to late-July with values ranging between 9.5-10.0 (Figure 3; Figure 4). UKL stations generally declined through the remainder of the season. Dissimilar to 2009-2011, pH maxima tended to be in more in sync with the period of maximum water column temperature.

Water column DO values were initially elevated in late-April and early-May (10-11 mg/L), and declined in mid-May (7-8 mg/L) at which time they continued to remain stable until late-June (7-9 mg/L). Surface DO then increased in late-June (10-12 mg/L; preceding the pH increase that occurred later in July), and DO at lower depths remained near 7 mg/L (Figure 3 and Figure 4). DO then declined during late-July (at ER) and early-August (MN) at deeper depths (<2-4 mg/L), especially at ER (Figure 3). Low DO at ER extended through much of the water column (Figure 3), as compared to MN where low DO only occurred near bottom (Figure 4). Lower DO conditions were present until early- to mid-September. As noted previously (e.g., Kann 2012), trends in pH and DO can be influenced by temperature and algal dynamics (cool late-spring and early-summer conditions were associated with low algal productivity, a delayed bloom, and moderate bloom decline in 2011). However, algal productivity remained low in May and June despite water temperatures that were substantially warmer than 2011 (see below), indicating that factors other than water temperature also influence algal productivity and subsequent DO and pH dynamics.

Similar depth-time plots were constructed for these stations for all years of data (1990-2011) and are shown in Appendix I. Although a comprehensive inter-annual analysis will not be performed here, unlike 2011 when water column temperatures showed a later peak than most years (early August as opposed to late-July), 2012 peak temperature timing was more typical of other years. However, DO tended to be low on a water column-wide basis at ER, showing more extreme lows



**Figure 3. Depth-time distributions of isotherms of T ( $^{\circ}\text{C}$ ) and isopleths of D.O (mg/L) and pH at UKL station Eagle Ridge (ER), 2012. Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).**



**Figure 4. Depth-time distributions of isotherms of T (°C) and isopleths of D.O (mg/L) and pH at UKL station Mid North (MN), 2012. Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).**

than many previous years (indicative of a severe bloom decline). Low off-bottom DO was also apparent at MN relative to other years. The pH seasonal peak tended to be higher than the previous several years at MN, especially deeper in the water column (Appendix I). As shown below and in earlier data and analytical reports (e.g., Kann 2011; Jassby and Kann 2010), differences in pH and dissolved oxygen can be explained in part by the interaction of both climate and bloom dynamics, which can also be influenced by lake level.

### *2012 Station Distributions*

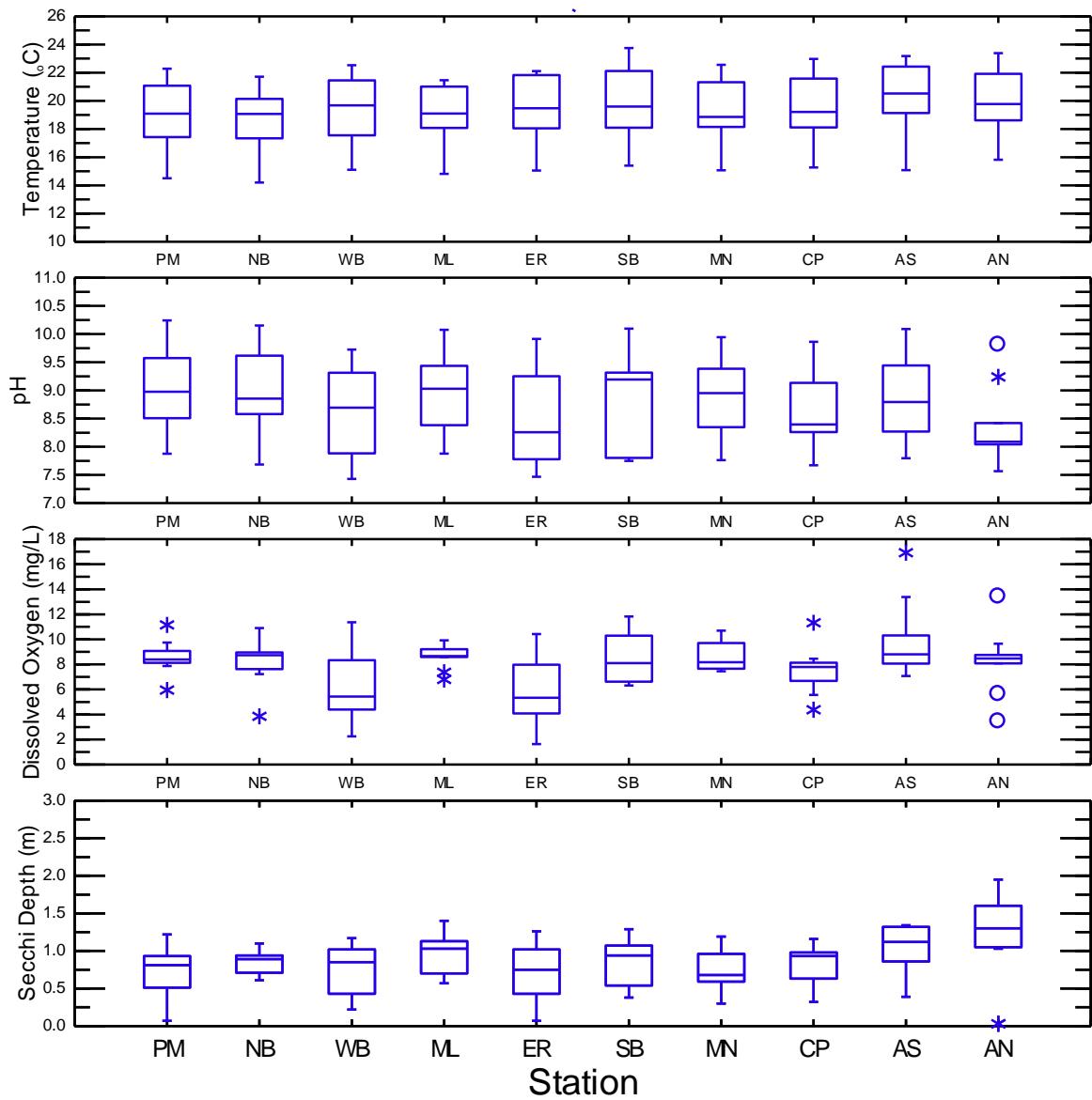
The distribution of parameter values for each station for the June-September period (chosen here to encompass the major algal growing season in UKL) are shown in Figure 5 and Figure 6. Although the seasonal timing of water quality has been shown to vary among stations (see below analyses comparing individual stations by date), the season-wide distributions as indicated by the interquartile range (25<sup>th</sup>-75<sup>th</sup> percentiles or box hinges in the plots below) tend to overlap for most parameters. In addition, although the timing of sample collection can affect the distribution of these variables (particularly temperature, pH and dissolved oxygen—see Jassby and Kann 2010), the below plots reflect water column means which are less sensitive to the effect of sample timing than are surface values.

Nonetheless, as with previous years, certain stations tended to stand out on a seasonal basis. For example, the pH distribution (as indicated by the upper or lower quartile) was skewed higher for PM and NB (and to a lesser degree AS), and skewed lower for WB, ER, SM, and AN (most notable for AN; Figure 5). ER showed the lowest overall DO, closely followed by WB, while PM, NB, ML, AS, and AN tended to show higher overall distributions. Secchi depth (transparency) was somewhat lower at ER and WB and higher at AS and AN. These among-station comparisons were not necessarily consistent with either 2010 or 2011 patterns (see Kann 2011; 2012), and among-station patterns are not always consistent from year-to-year.

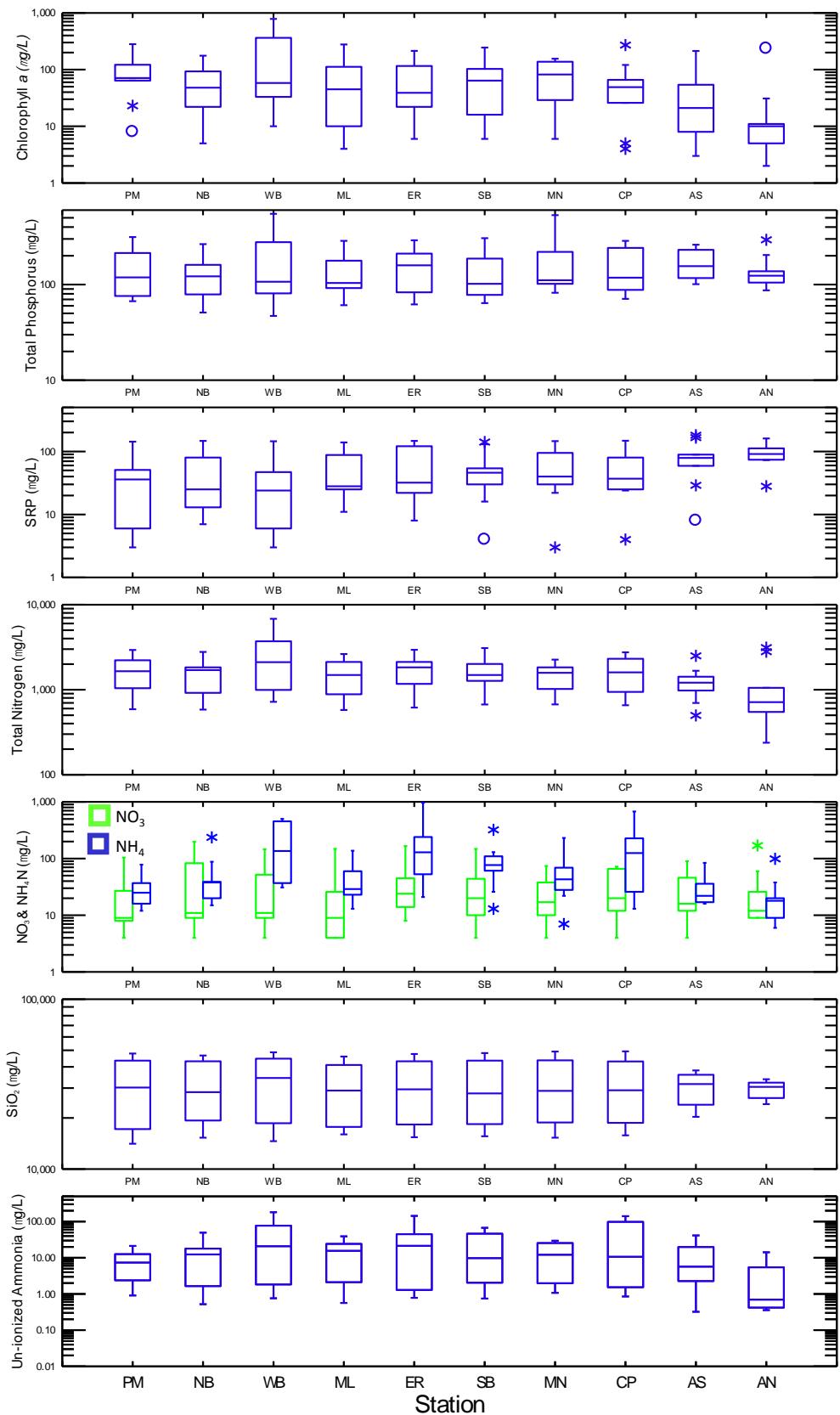
Stations PM, SB, and MN were among the highest with respect to median CHL, while WB showed significantly higher upper quartile CHL. The distribution if CHL at stations ML, AS, and AN were among the lowest (Figure 6). However, the inter-quartile CHL range was similar among many other stations. Both AS and AN showed noticeably lower CHL in 2012, especially compared to previous years (Kann 2012). In contrast to 2010 and 2011 when the AS and AN stations stood out with noticeably higher upper quartile and median values for TP, values were similar to or somewhat lower (see AN) than other stations in 2012 (Figure 6). However, SRP values for AS and AN were skewed high relative to other stations. Both PM and WB were skewed low for SRP relative to other stations.

Similar to previous years, Agency Lake stations were among the lowest for nitrogen, particularly for NH<sub>4</sub>-N, but also for NO<sub>3</sub>-N, and TN (Figure 6; Table 2). The upper quartile value and interquartile range for TN were highest at WB. Similar to 2010 and 2011, ER, SB, and CP were among the highest for ammonia, but WB was also high in 2012 (NH<sub>4</sub>-N; Figure 6; Table 2). Un-ionized ammonia also tended to be highest at WB, ER, SB, and CP in 2012 (Figure 6). Upper quartile NO<sub>3</sub>-N was higher overall at NB.

Median silica values (~30,000 µg/L) were similar among stations, although the Agency Lake stations showed a narrower interquartile range (Figure 6). See below for a description of seasonal silica dynamics.



**Figure 5.** Station distributions of T (°C), pH, D.O (mg/L), and Secchi depth, June-September, 2012.



**Figure 6. Station distributions of CHL, TP, SRP, TN, NO<sub>3</sub>+ NO<sub>2</sub>-N, NH<sub>4</sub>-N, SiO<sub>2</sub> and un-ionized ammonia, June-September, 2012.**



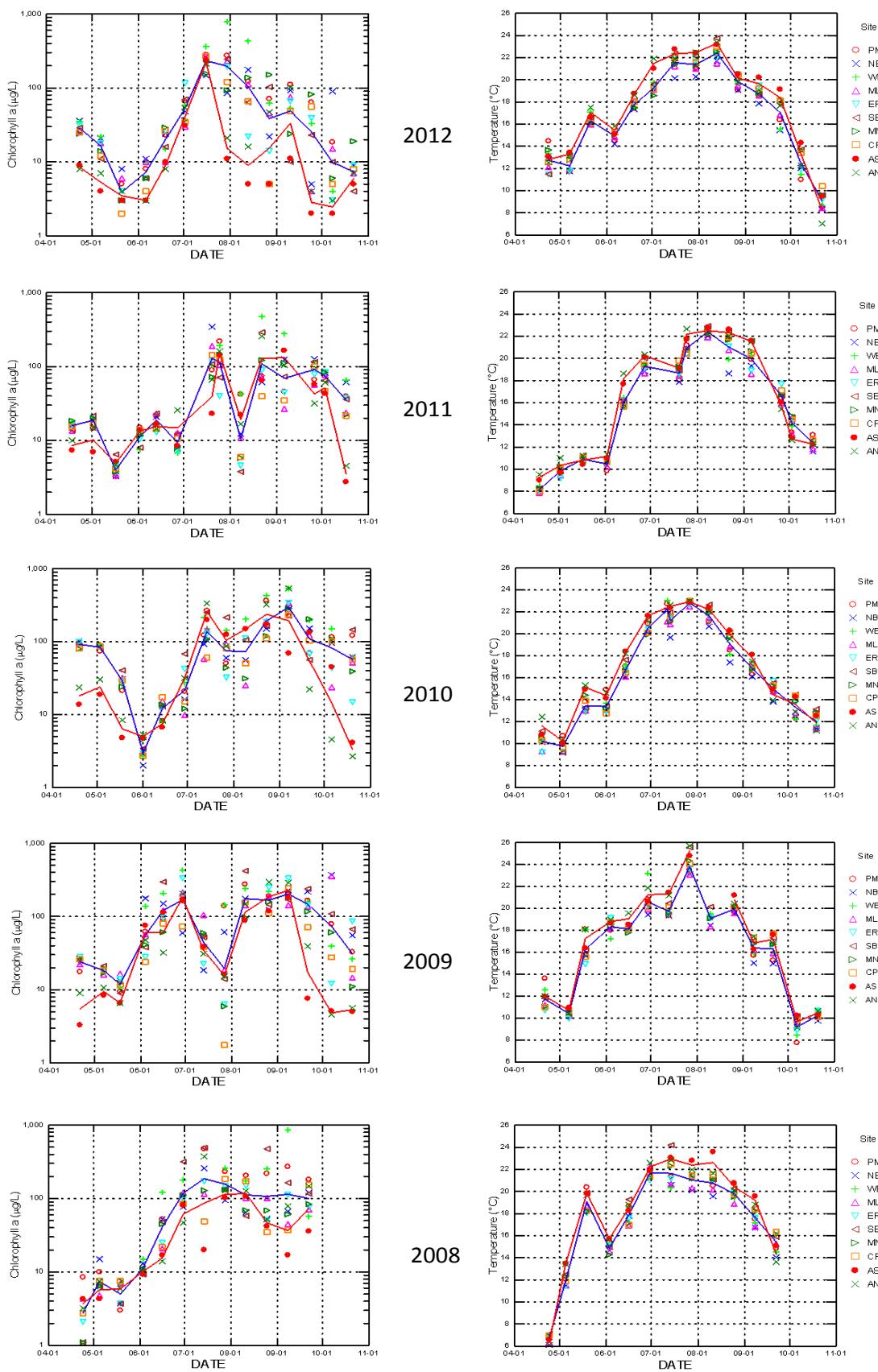
## *Seasonal Chlorophyll Pattern and Climate Interaction*

Seasonal differences in algal biomass (CHL) among stations in 2012 show that, similar to the previous four years (2008-2011), but unlike 2006 (Kann 2011) and 2007 when AS and AN increased earlier and declined earlier in the season relative to UKL stations; early season CHL in Agency Lake was low in late-April and early-May and was then similar to UKL through the initial bloom peak (Figure 7). However, the Agency Lake bloom decline occurred much earlier than UKL in 2012 (and to a much greater degree than it did in 2011). The similarity between Agency and UKL Lakes in terms of the June algal biomass increase and seasonal maxima and decline in the later years may reflect greater connectivity between the two lakes due wetland restoration activities on the Williamson Delta Preserve (e.g., Wong et al. 2010; 2011). CHL concentration at the more southerly stations vs. northerly stations did not tend to show as distinct a seasonal pattern as previous years when relatively higher CHL occurred later in the bloom cycle (e.g., August) at the southerly stations.

As noted in previous annual data reports (Kann 2008 to 2012), water temperature partially explained the early season CHL patterns among the years. For example, low temperatures coincided with a depressed early-June bloom in 2006, and in 2008 much cooler lake-wide water temperature (median value  $<7^{\circ}\text{C}$ ) in late April and early-May also coincided with low CHL levels. However, it was clear that factors other than temperature were also affecting bloom dynamics in those years (Figure 7).

For example, in 2010, late-April and early-May CHL was noticeably higher than the previous four years (generally  $>80\text{ }\mu\text{g/L}$ ) due to an unusually large diatom bloom (*Asterionella formosa*) occurring at that time—despite temperatures in a range similar to many of the previous years (Kann 2011). The large 2010 diatom bloom then declined rapidly beginning in mid-May and by early-June chlorophyll levels were less than  $10\text{ }\mu\text{g/L}$ . In contrast, CHL levels in 2011 were only slightly elevated in late-April and early-May (generally  $<20\text{ }\mu\text{g/L}$ ), and except for a decline in mid-May ( $<7\text{ }\mu\text{g/L}$ ), they remained generally less than  $20\text{ }\mu\text{g/L}$  (often less than  $10\text{ }\mu\text{g/L}$  at many stations) through the end of June (Figure 7). During this same period water column temperature remained very cool ( $<11^{\circ}\text{C}$  through early June) and although mid-June temperature increased to  $\sim 16^{\circ}\text{C}$  in UKL (they were 1-2 deg. warmer in Agency Lake), they only rose slightly, remaining  $<20^{\circ}\text{C}$  through most of July (Figure 7). In contrast, water temperatures during the previous five years generally exceeded  $20^{\circ}\text{C}$  by early-July, if not sooner (Figure 7).

In 2012, the CHL pattern was more similar to 2010, although the spring levels  $\sim 30\text{ }\mu\text{g/L}$  were still substantially lower than the  $\sim 100\text{ }\mu\text{g/L}$  achieved in 2010. May-June levels were similar, as was the peak which occurred mid to late-July of both 2001 and 2012. Water temperature warmed more rapidly than 2011, and CHL also increased to levels  $>50\text{ }\mu\text{g/L}$  by early-July. CHL did not undergo a lake-wide decline in August as it did in 2011.



**Figure 7. Seasonal CHL and temperature trends for UKL stations, 2008-2012 (blue line shows the median value for UKL-only, red line shows the median value for Agency Lake-only).**

Because water temperature in the above plots is measured biweekly, and due to UKL's shallow depth a short lag-time is generally observed with respect to equilibrium with ambient air temperatures (e.g., Wood et al. 2006), it is also instructive to evaluate daily air temperatures as another indicator of water column warming.

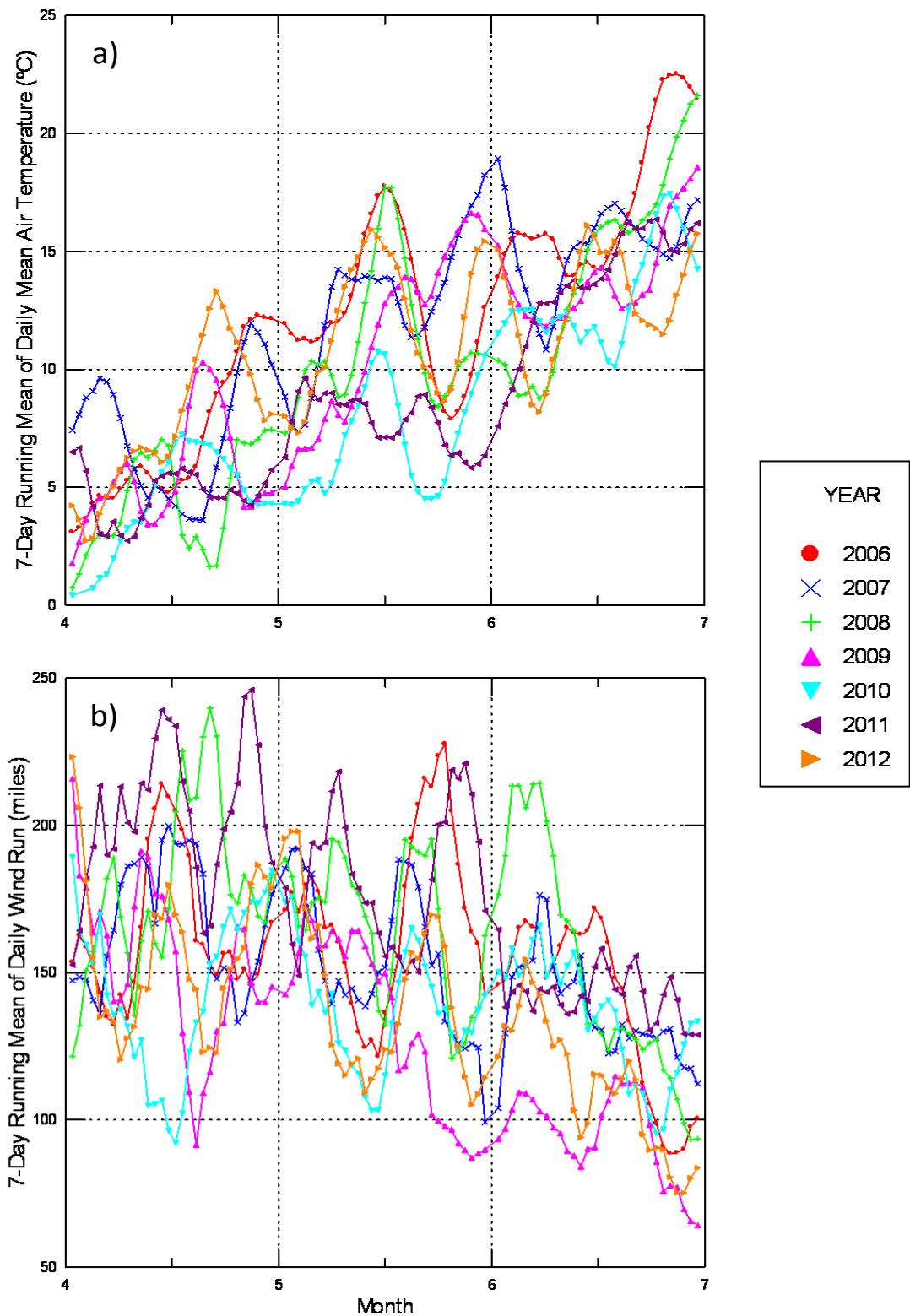
Hourly data obtained from the USBR AgriMet station located near Agency Lake (Figure 13a) indicated at least partial tracking of May air temperature and CHL levels (Kann 2011; 2012). For example, temperature declines in mid-May of 2006 and 2008 that remained near or below 15 °C through mid-June were associated with suppressed CHL levels in early-June (Kann 2011). In 2007 and 2009, air temperatures warmed between mid- and late-May and were associated with elevated CHL levels in early June, and in 2010, when temperatures cooled substantially in mid-May and portions of June, CHL also remained suppressed during early and mid-June (Figure 7; Figure 8a)

Likewise, air temperature in 2011 was similar to other years in the beginning of May, but unlike other years it remained relatively constant through May (it was also among the lowest when compared to previous years and showed further cooling towards the end of May) before increasing in June to levels similar to other years (Figure 8a). A significant departure from other years then occurred in July of 2011 when the upper quartile, median, and lower quartile values of daily mean air temperature were substantially lower than corresponding values for the previous 5 years (Figure 9).

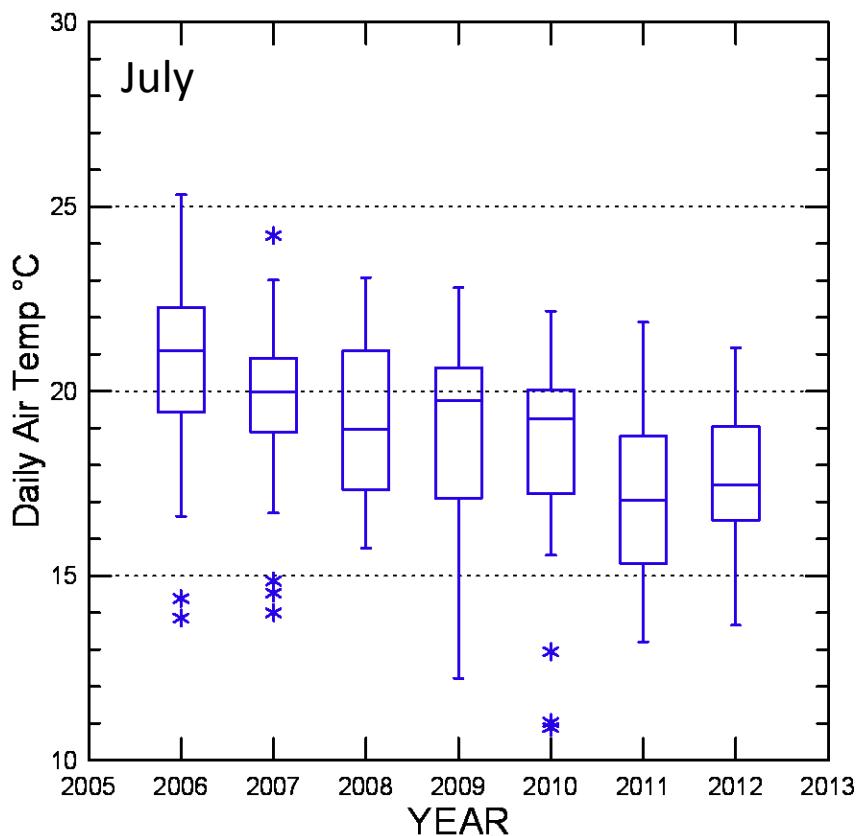
In 2012, temperatures were intermediate to warmer in late-April, intermediate in early-May, with cyclical warming and cooling occurring through June, with mid-to late June temperatures tending to be lower than previous years (Figure 8a). July of 2012 was slightly higher than 2011, but like 2011 was lower than the 2006-2010 period. (Figure 9).

In general, previous analyses indicating a threshold temperature of ~15 for *Aphanizomenon* bloom development in Upper Klamath Lake (Kann 1998; Kann 2011) continue to be supported. However, as noted previously (Kann 2011) high CHL levels due to spring diatom blooms can be achieved even at temperatures much cooler than 15 °C. Furthermore, similar to 2011, based on 2012 patterns it appears that once the 15 °C threshold was reached, cool temperatures towards the end of June and into July also had an apparent effect on continuing algal biomass development.

Analysis of wind speed as an indicator of the extent of water column mixing showed that the periods directly preceding and during the typical period of June bloom development in 2006 and 2008 (which had relatively low CHL during that period) were characterized by generally higher wind speed relative to 2007 (Figure 8b), which had relatively higher CHL (Kann 2011). Wind speed during 2009 was more similar to 2007 (which had lower wind preceding and during development of the early June bloom), and also tended to have relatively higher CHL compared to 2006 or 2008, which were suppressed. The pattern for 2010 was less clear and may have been confounded by the massive diatom bloom which crashed immediately preceding the June period when *Aphanizomenon* typically begins to increase. Although wind was somewhat low to intermediate during the typical bloom initiation period in 2010 (late-May to early-June; Figure 8b), CHL still remained suppressed, possibly reflecting the unusually cool period occurring during late-May (Figure 8a). Likewise, in 2011 relatively high wind speeds in April and May were associated with relatively low algal biomass (Figure 8b).



**Figure 8.** Time series of the 7-day running mean of daily air temperature (a) and 7-day running mean of the daily wind run in miles (b), April-June, 2006-2012. Data are from the Bureau of Reclamation AgriMet station located at Agency Lake (AGKO).

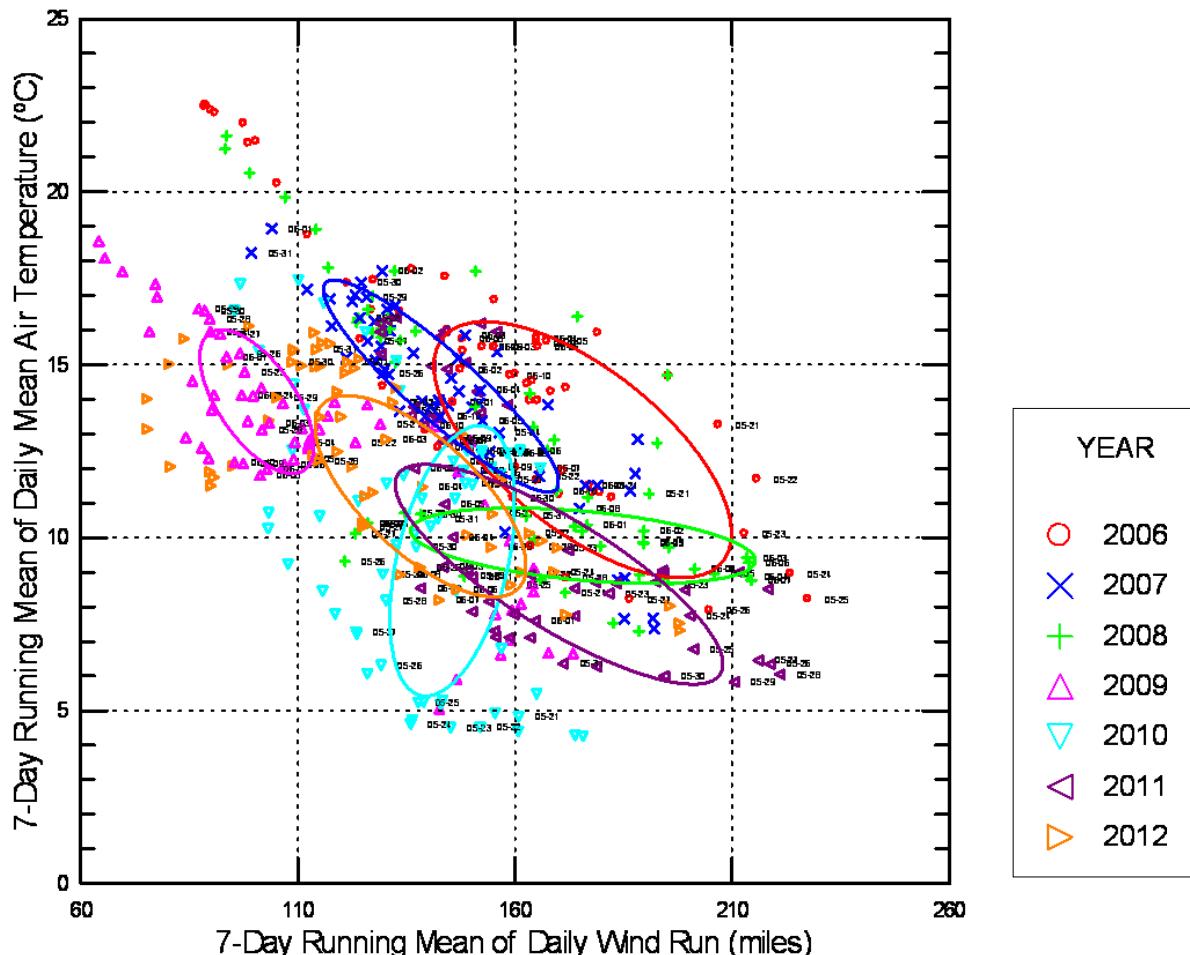


**Figure 9. Annual distribution of Agency Lake AgriMet (AGKO) daily air temperatures during July, 2006-2012.**

Wind speeds during the 2012 bloom development period (late-May to June) were on the low side relative to other years, which did not necessarily lead to enhanced CHL levels as expected, again indicating that factors other than climate also dictate bloom development.

Also similar to previous 2006-2011 analysis of air temperature and wind speed data that showed wind and temperature to be related such that warm/calm conditions co-occur and that cool/windy conditions co-occur (Kann 2012), these parameters also tended to co-occur in 2012 (Figure 10). Supporting observations from the above time-series graphs, confidence ellipses computed for the period encompassing 10 days prior to and subsequent to June 1<sup>st</sup> (the typical historical period of initial *Aphanizomenon* increase) show that both 2006 and 2008 (red and green ellipses in Figure 10) tended to be cooler and windier than during the same periods in 2007 and 2009 (blue and pink ellipses). 2009 showed the lowest wind speed of the four years (Figure 8b and Figure 10) and was associated with higher early- and mid-June CHL than the other years (Figure 7) (Kann 2010). Temperatures in 2010 were cooler overall than the other years and wind speeds were only intermediate (Figure 10). During 2011 the late-May to early-June period was among the coolest and windiest of the six years portrayed (Figure 10), and as noted above also showed relatively low algal biomass levels. 2012 wind speeds were lower, and temperatures intermediate relative to other years.

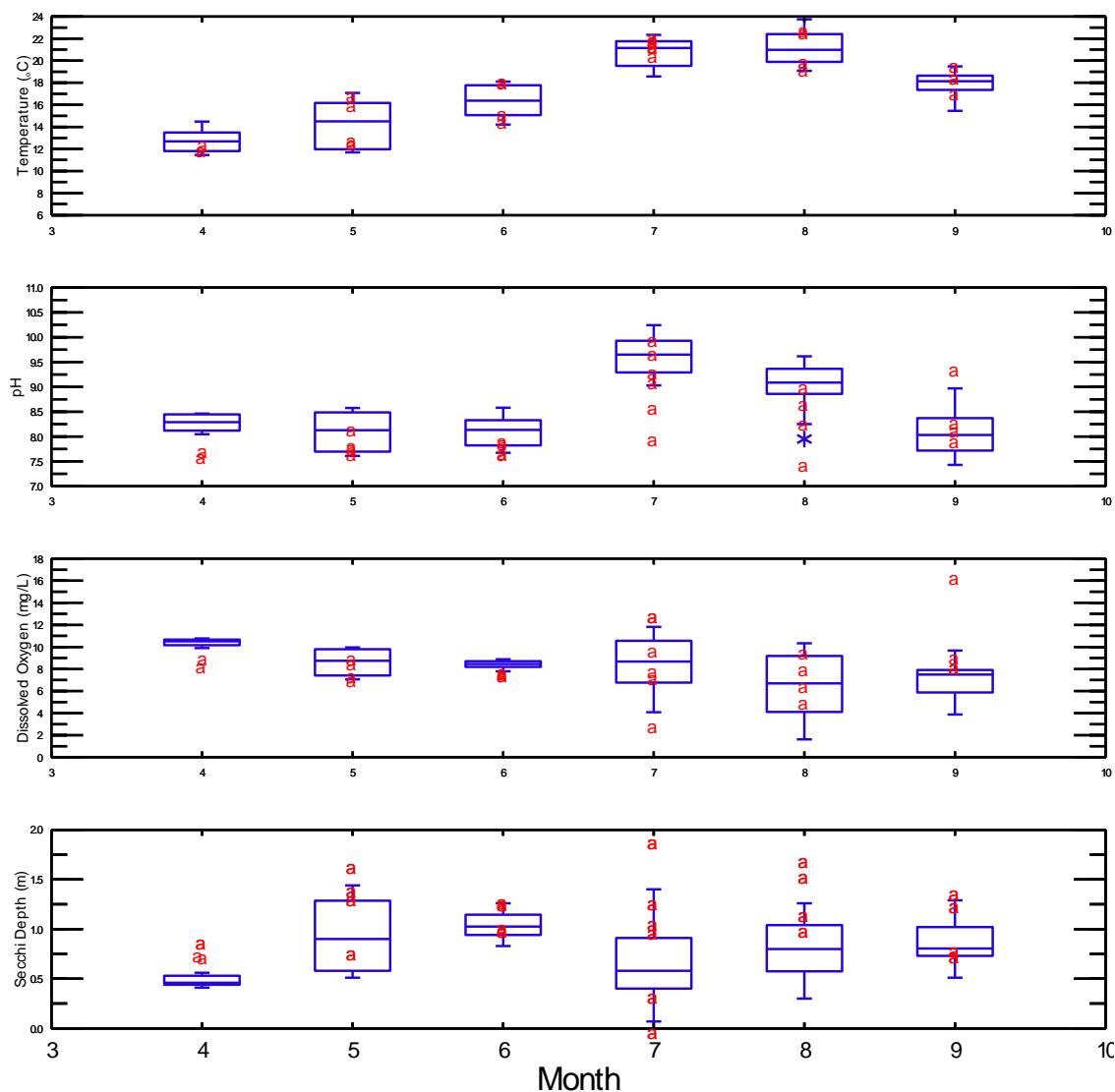
These climate data indicate that cooler and well mixed conditions during the usual early season bloom development period (e.g., Kann and Welch 2005) contribute to variability in year-to-year bloom development. Multivariate analyses performed on the longer 1990-2009 data set also showed that wind and temperature, along with lake elevation were determinants of CHL levels in UKL (Jassby and Kann 2010). As noted below these factors also interact with varying year-to-year variability in nutrient concentrations.



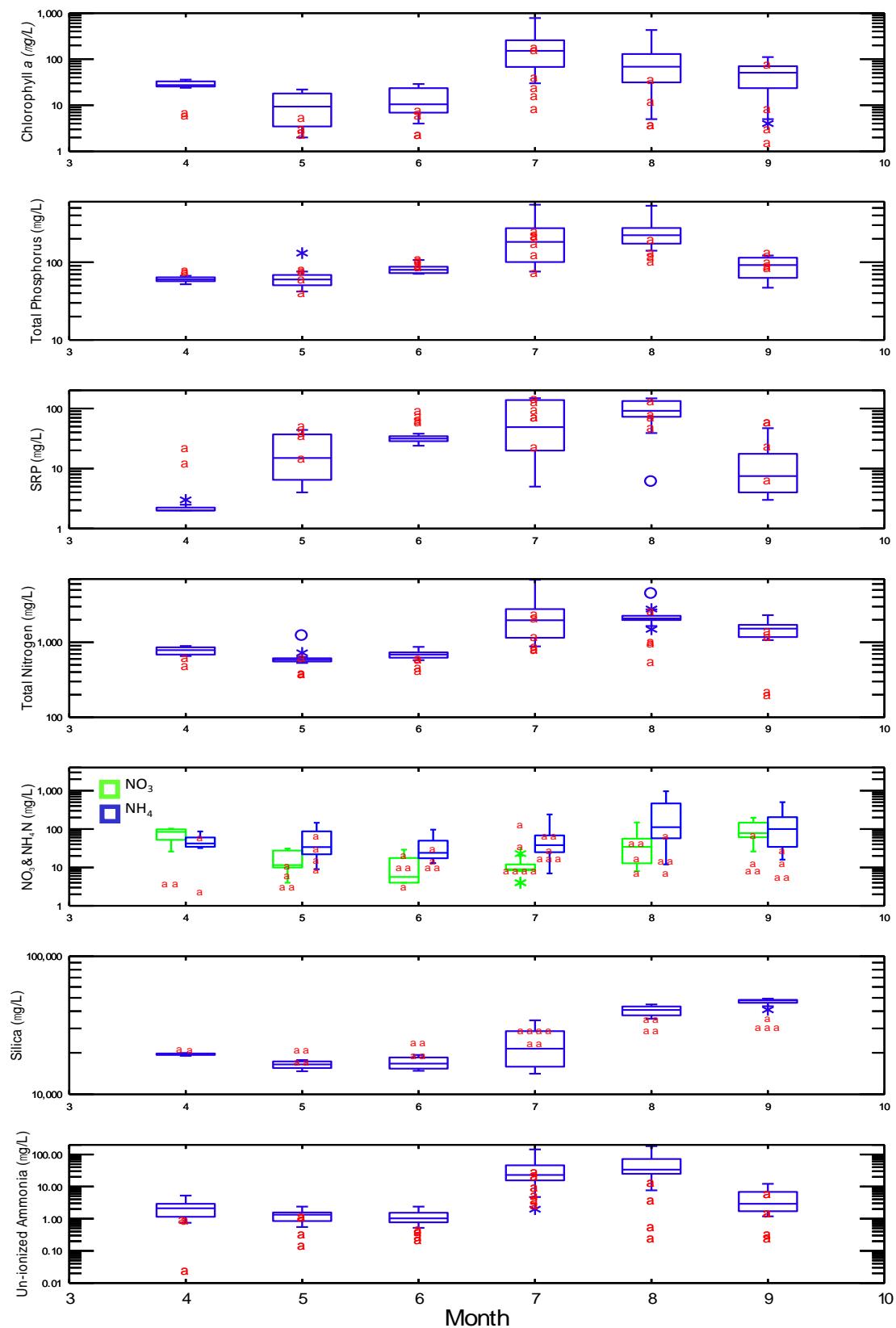
**Figure 10.** Scatter plot of the 7-day running mean of the daily wind run (miles) vs. 7-day running mean of daily air temperature ( $^{\circ}\text{C}$ ) during May and June. Data are from the Bureau of Reclamation AgriMet station located at Agency Lake (AGKO). Data labels are day of the month. Confidence ellipses are drawn for dates occurring during the last 10 days of May and first 10 days of June; confidence ellipses are centered on the sample means of the x and y variables where the unbiased sample standard deviations of x and y determine its major axes and the sample covariance between x and y, its orientation (Systat 2013).

## 2012 Monthly and Seasonal Water Quality, Chlorophyll, and Nutrient Patterns

Basic statistics for monthly distributions for all sampling years are shown in Appendix 1. Peak water temperatures occurred in July and August of 2012, with the July median higher than August (this is in contrast to 2011 when the August median was higher) (Figure 11). Monthly distributions for pH in 2012 were similar to 2006-2008 and 2011 which showed a progressive seasonal increase with seasonal maxima occurring in July that coincided with the lowest Secchi depth (indicating reduced transparency) and highest CHL distributions (note that in 2010 high pH values occurred in April, declined in May and June and showed a bimodal peak in July and September). In contrast to 2011 CHL did not show a bimodal peak in 2012 (Figure 7 and Figure 12), and pH declined into September (Figure 11). Lower DO occurred during August in 2012, and although the timing of low DO was similar to other years, DO was relatively low compared to 2011 and other years despite the lack of a strong lake-wide bloom decline. As noted above, 2012 CHL peaked in July, and did not exhibit a second peak in September (Figure 12).



**Figure 11. Monthly distributions of T ( $^{\circ}\text{C}$ ), pH, D.O (mg/L), and Secchi depth, 2012 (symbol “a” denotes values for Agency Lake plotted separately from the box plot distribution).**



**Figure 12. Monthly distributions of CHL, TP, SRP, TN,  $\text{NO}_3 + \text{NO}_2\text{-N}$ ,  $\text{NH}_4\text{-N}$ ,  $\text{SiO}_2$  and un-ionized ammonia, 2012 (symbol “a” denotes values for Agency Lake plotted separately from the box plot distribution).**

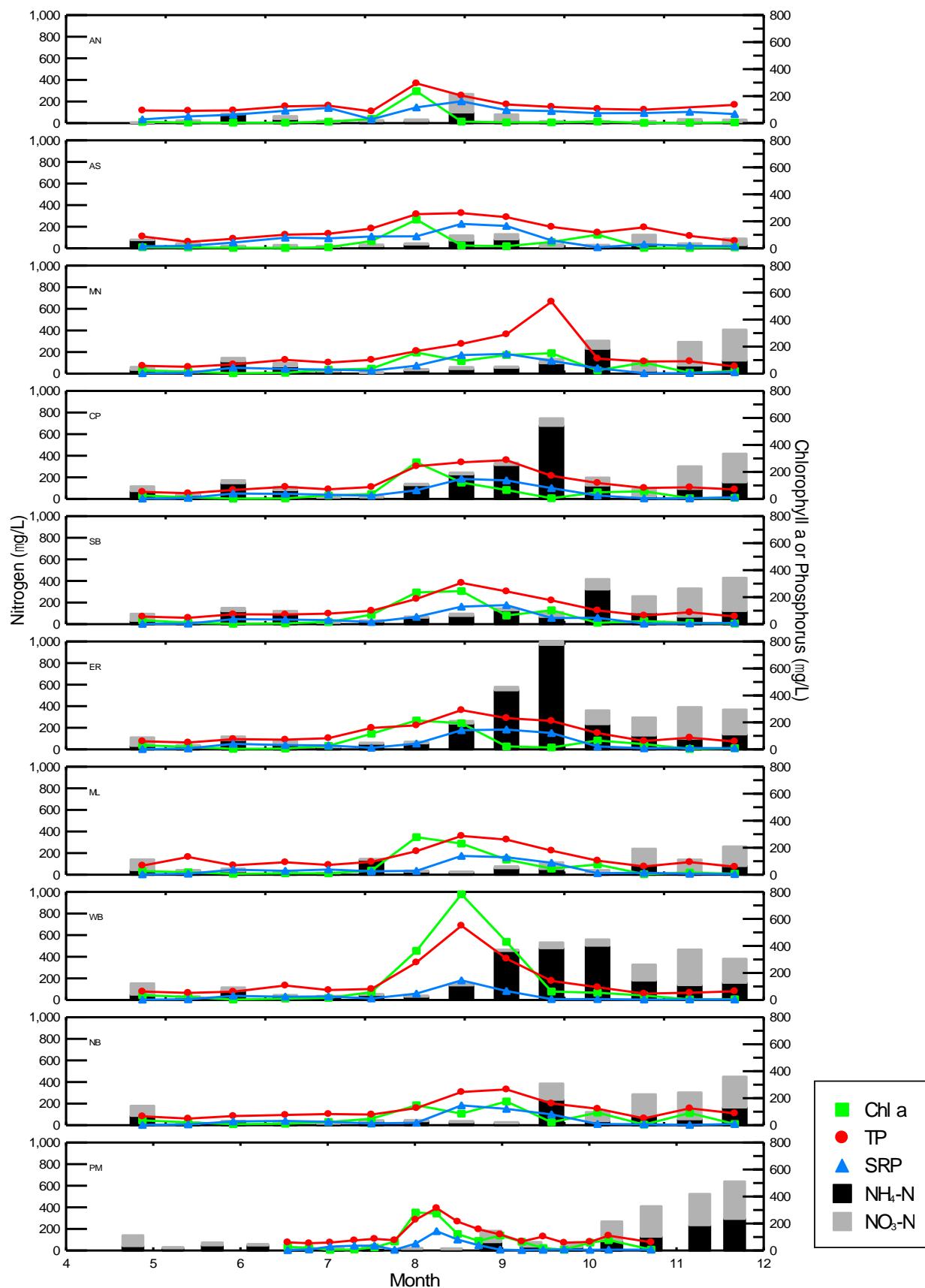
TP in 2012 remained low and relatively constant through June before increasing in July during the overall CHL increase (Figure 12). Unlike 2011 when SRP remained low and was depressed well into July, SRP in 2012 was elevated in May and June. Although TP and TN typically increase during June, values remained low for both parameters during June of 2012, with values then increasing during July when CHL increased with the initiation of the *Aphanizomenon* bloom. In contrast to 2011 when about a month delay was observed (when compared to the typical seasonal pattern), 2012 SRP increased in May and June, while 2012 NH<sub>4</sub><sup>+</sup> increased in July (Figure 12).

A further look at the 2012 time-series with respect to CHL and dissolved nutrients shows that, as in other years, SRP at the UKL stations generally remained low through the initial July CHL peak before increasing during the algal biomass decline in early August (Figure 13). Also as in previous years (e.g., 2009-2011), this trend did not apply to the Agency Lake stations which showed elevated SRP in April (AN) and May (AN and AS), with levels then increasing and remaining high during the July CHL increase (Figure 13). As noted previously, there is an indication that SRP is limiting the early season bloom in UKL (as with 2011, the 2012 early season bloom also occurred in July), especially since internal sources of phosphorus are also increasing during that time period (Kann 2010; Walker et al. 2012).

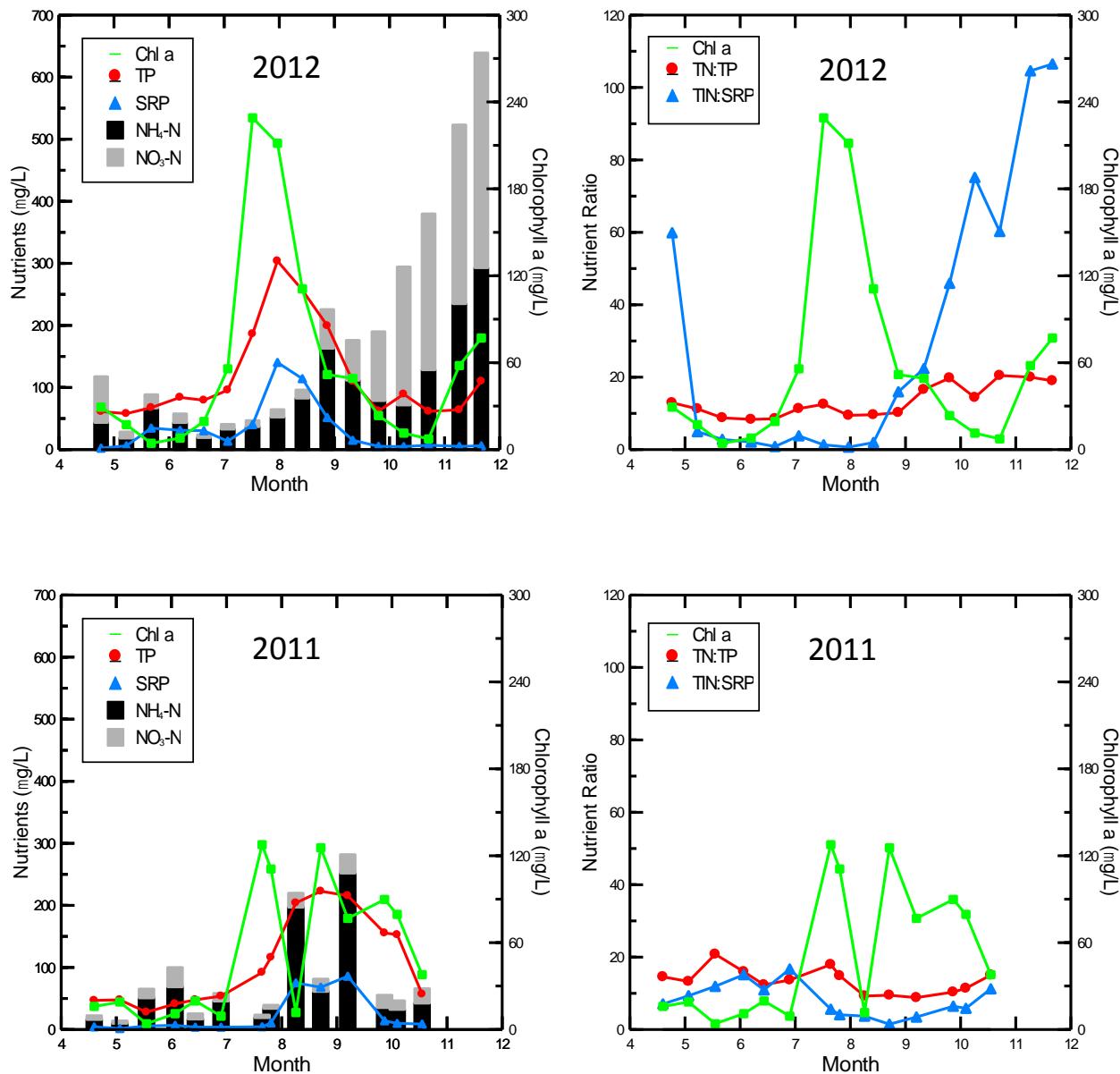
In 2012 and most previous years TIN (the sum of NH<sub>4</sub>-N and NO<sub>3</sub>-N) levels were relatively low during the spring period, and because levels in 2011 tended to be somewhat elevated when compared to previous years, it was speculated that these relatively higher TIN levels (as well as climate) may play have played a role in suppressing the June to early-July *Aphanizomenon* bloom in 2011(Kann 2012). Elevated spring TIN levels in 2012 levels were also associated with a somewhat delayed bloom (although not as delayed as 2011) (Figure 14). Levels were notably lower on a lake-wide basis during the period leading up to the 2010 bloom (Kann 2012). Similar to SRP, TIN then increased substantially during the August CHL decline (Figure 13). As in earlier years, SRP in 2012 tended to decline into the fall months. However, unlike 2011 when TIN also declined in the fall, values continued to increase in 2012 (Figure 14). Spring and fall TIN tended to show an increased proportion of NO<sub>3</sub>-N, while summer TIN was comprised predominantly of NH<sub>4</sub>-N (Figure 13; Figure 14).

Both 2010 and 2011 showed ratios of TN:TP during April to be ~15, which in general would tend to favor the type of diatom blooms observed in spring (Kann 2012). However, unlike 2010 when the TN:TP and TIN:SRP ratios declined in May and June during the period preceding the rise of nitrogen-fixing *Aphanizomenon* in UKL (TN:TP ratios were generally lower than 12 and TIN:SRP<2.5), ratios increased substantially during Figure 14). This increase appears to be driven by the higher TIN values during his period, and as noted above may provide a partial explanation for the late onset of the nitrogen-fixing *Aphanizomenon* bloom. This is also supported by the relatively low April-June TN:TP and TIN:SRP ratios in 2012 (TN:TP ~10; TIN:SRP<2), which were associated with an earlier *Aphanizomenon*-associated algal biomass rise (Figure 14). The much lower 2012 TIN:SRP ratios are partially caused by higher SRP than occurred in 2011.

Similar to 2011, an apparent geographical grouping of stations occurred in 2012 (Figure 13); Agency Lake stations (higher early-season SRP and TP and higher SRP overall); northerly stations MN, CP, SB, ER (higher TIN during the August bloom decline and extending into October); and southerly stations ML, NB, PM (somewhat lower TIN during the August bloom decline, and increase in September). As noted above, differences in magnitude and timing of



**Figure 13. Chlorophyll, SRP, and TIN time-series for UKL and Agency Lake Stations, 2012.**



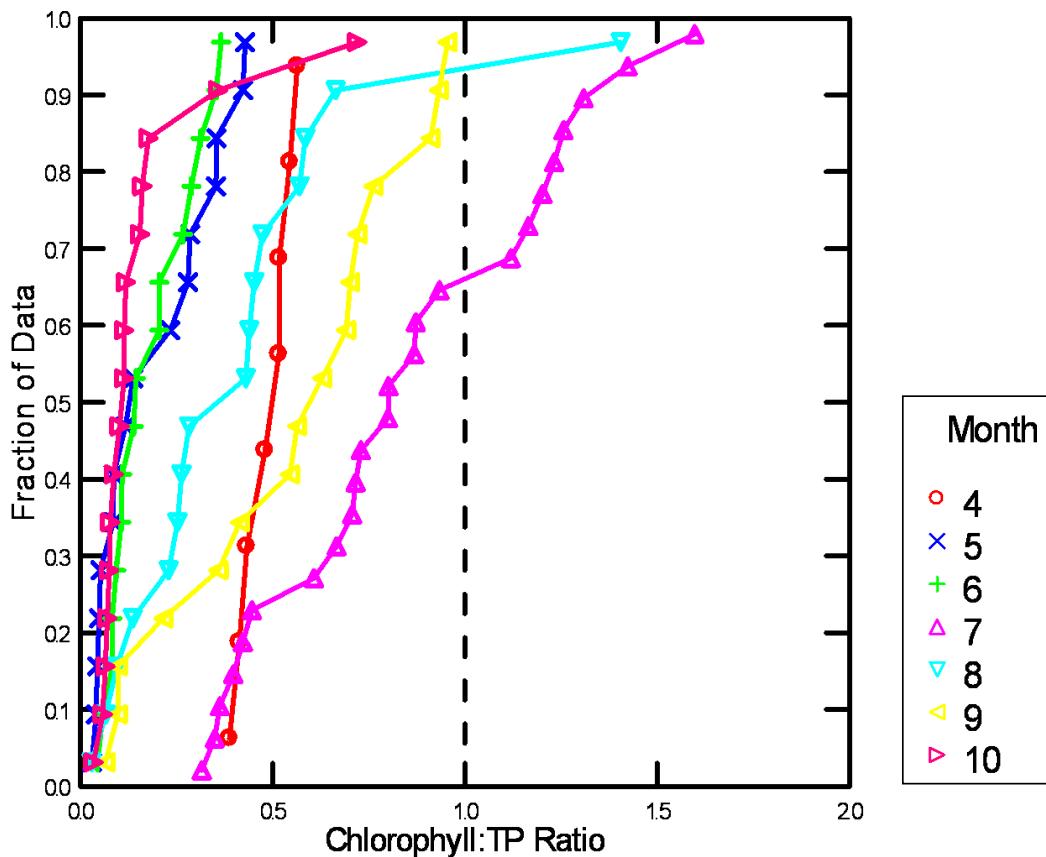
**Figure 14. Lake-wide mean Chlorophyll, SRP, TIN, and nutrient ratio time-series for UKL Stations, 2011 and 2012.**

CHL at the Agency Lake stations are not as apparent as in earlier years, possibly due to dike breaching and greater connectivity between the two lakes (Figure 12).

Silica (a parameter not previously reported in annual data reports) showed declining and lowest seasonal values in April and May, stable values in June, and a substantial increase beginning in July, with elevated values continuing through September (Figure 12). These trends are likely tied to silica uptake during spring diatom blooms, and subsequent summer sediment recycling and lack of uptake due to diatom decline during periods of *Aphanizomenon* dominance. More detailed silica trends along with algal biomass and nutrients are presented in Appendix III. These graphs indicate that the silica increase is concomitant with initial large CHL and TP increases in July, and that silica concentration increases (>40,500 µg/L) continue into September before gradually declining in the fall, and continue to decline to seasonal lows in the spring. The Agency Lake sites showed a more muted pattern with somewhat higher values in the spring

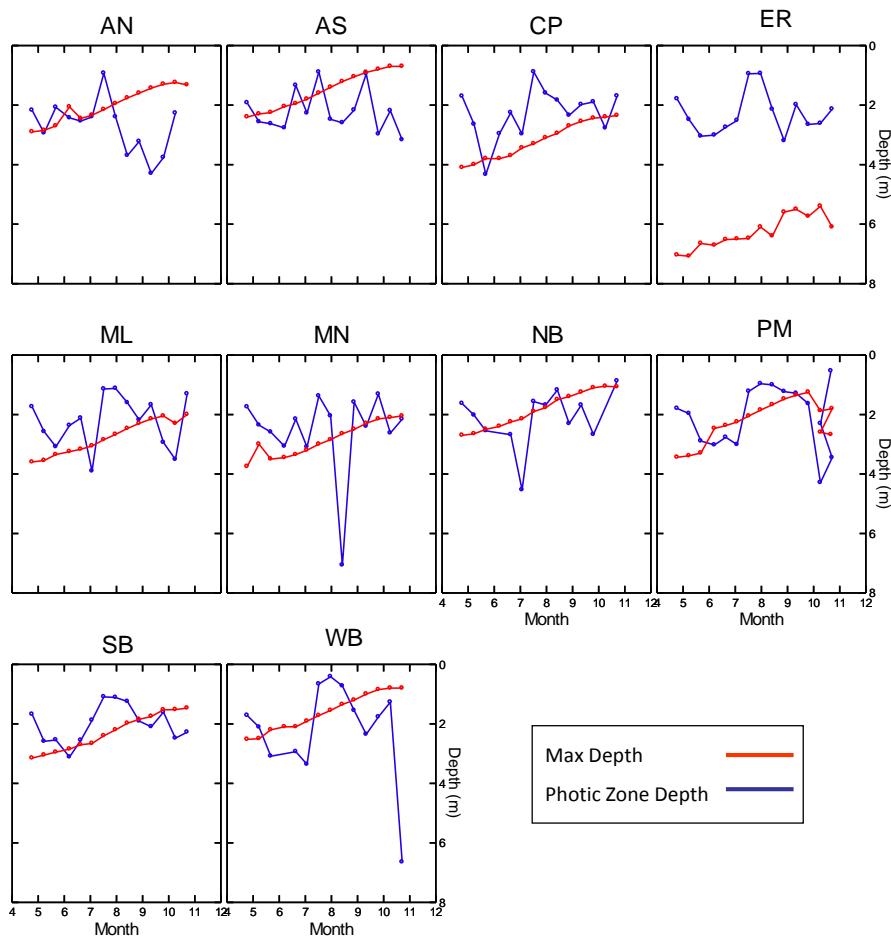
compared to other stations, and the magnitude of summer increases were less pronounced, especially at AN (Appendix III).

In contrast to 2009 when chlorophyll to TP ratios greater than 1 (which indicate potential P limitation- see Kann 2010) were observed at a high frequency in June during the initial bloom increase, and to 2010 when CHL: TP ratios >1 occurred in April, part of May, July, and part of September (Kann 2011), the 2012 frequency of CHL: TP ratios >1 was similar to 2011, occurring at a high frequency only in July (Figure 15).



**Figure 15. Quantile plot (cumulative frequency) of April-October chlorophyll to TP ratios in Upper Klamath Lake, 2012.**

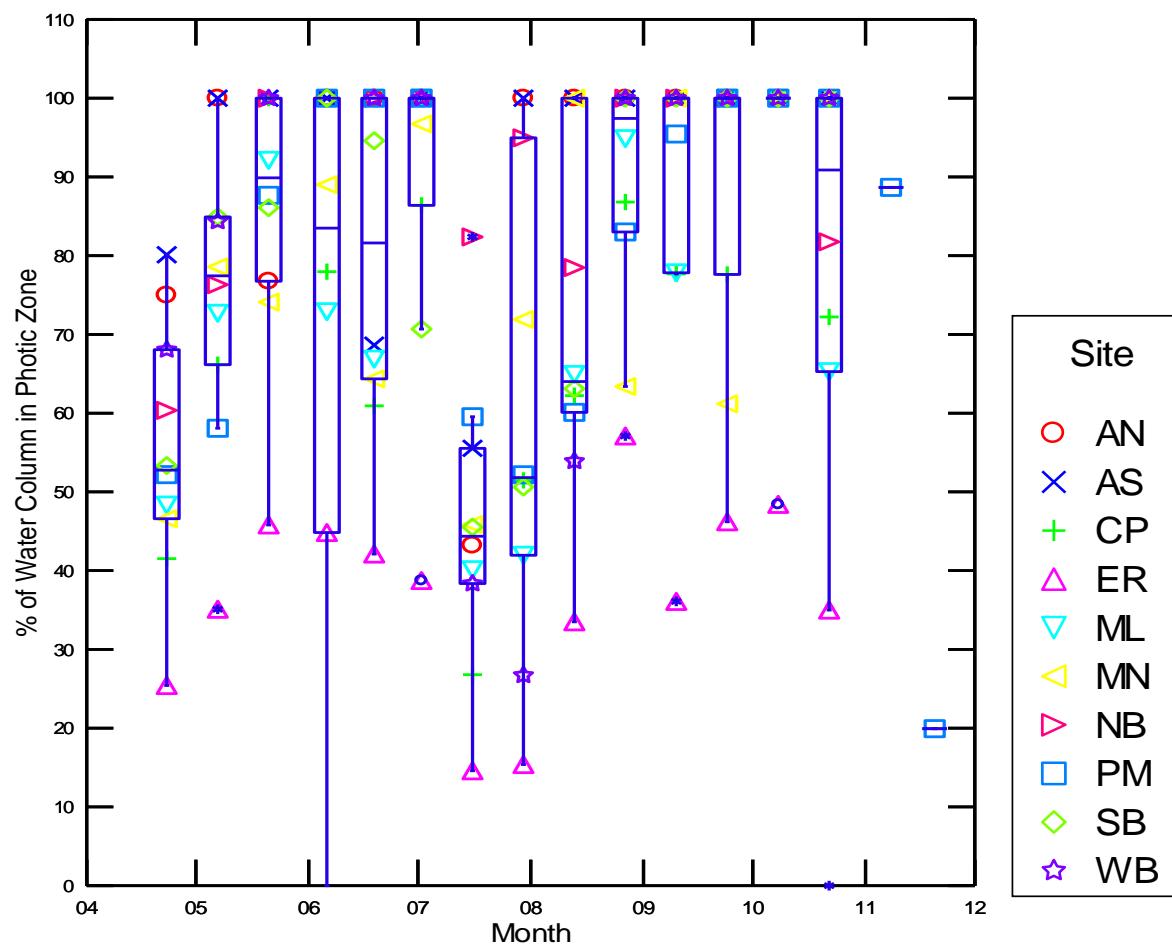
The underwater light environment is another factor that can influence both bloom dynamics and other water quality parameters, especially those that are photosynthetically driven. Although not discussed in detail here, a plot of photic zone depth (defined as the depth where 99% of incident light is absorbed as computed from extinction coefficients) relative to the maximum depth at UKL and Agency Lake stations shows that, as in other years, despite the shallow nature of the system that the photic zone depth was often shallower than maximum depth in 2012 (Figure 16; occurring when the blue line is above red line). The typical UKL pattern shows a relatively shallow photic zone during the spring diatom bloom, a deeper photic zone that extends the depth of the water column during much of May and early-June, a shallower photic zone during late-June to mid-July algal blooms, a decline (i.e., deeper photic zone- although not as extreme as the May decline) during August bloom declines, and finally another shallow photic zone period



**Figure 16. P**hotic zone depth and maximum depth at UKL and Agency Lake stations in 2012 (periods when the blue line is shallower than the red line indicate that a portion of the water column is not within the photic zone).

during bloom rebound in late-August and September (Kann 2010-2112). The 2012 pattern (Figure 16) was similar to other years, showing that at times a percentage of the water column is outside of the photic zone (e.g., does not have sufficient light for photosynthesis; Figure 17). Light limitation is more apparent at the deeper ER station which showed a greater percentage of the water column to be light limited.

To the extent that underwater light is influenced by seasonal algal dynamics (in concert with ambient light and the interaction with lake depth), decreases in available light during the early spring of 2010 (as they were in 2009) were likely influenced by diatom blooms (Kann 2011). However, unlike 2009 when a “clear water” phase occurred in May as the diatoms declined, in 2010 the “clear water” phase did not occur until later in May and extended to the end of June. A decline in available light did not occur until mid-July of 2010 (this occurred during June of 2009) as the *Aphanizomenon* bloom increased. In 2011 there was no apparent “clear water” phase until late-June, after which the algal bloom increased and water transparency declined (Kann 2012). In 2012, a reduced photic zone occurred in early-spring, with a “clear water” phase occurring in late-May, and although the photic zone was again reduced in early June, the lake was then relatively clear again in late-June prior to a sharp decrease in transparency in mid-July (Figure 17).

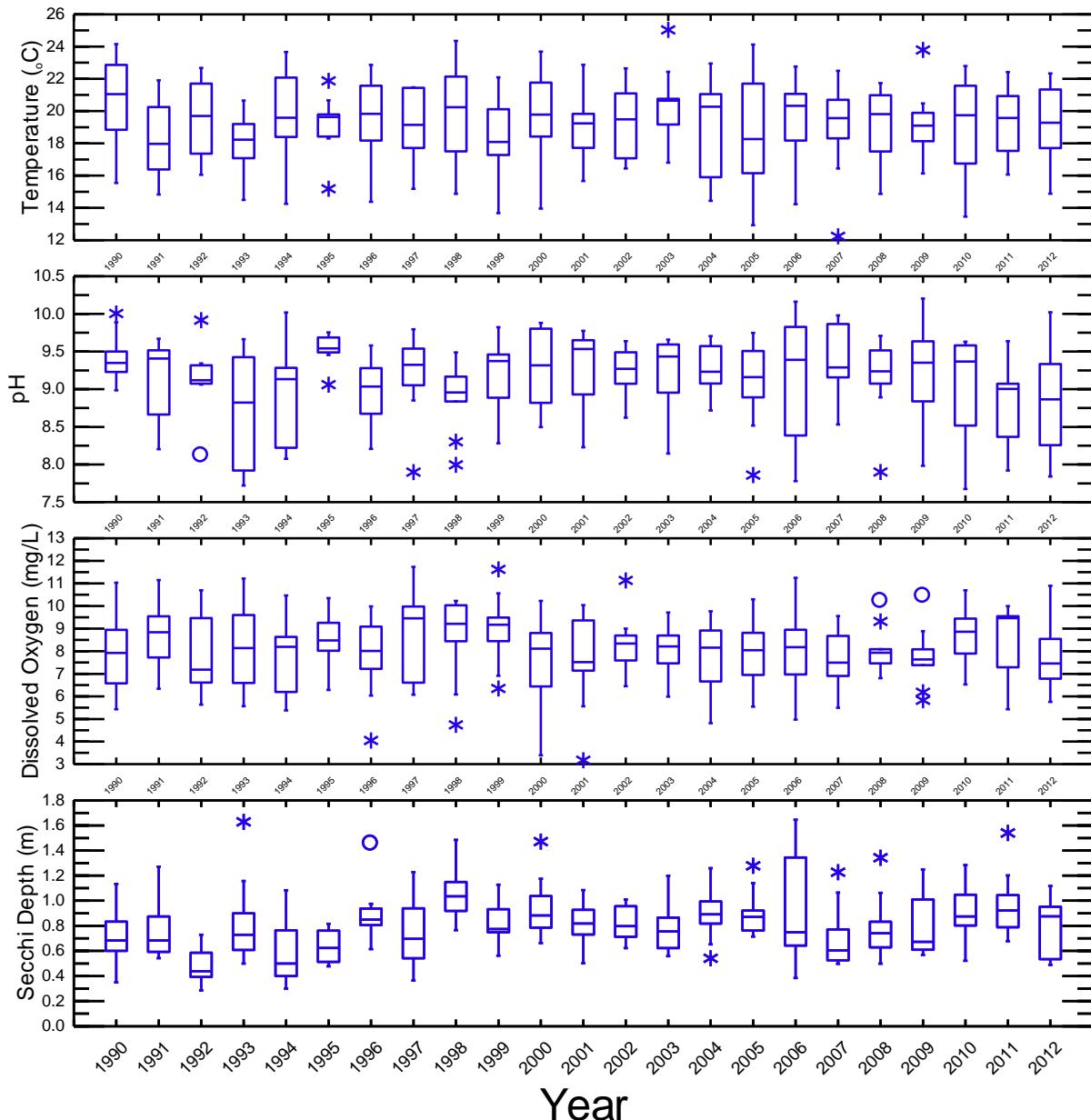


**Figure 17. Percent of the water column in the photic zone for UKL and Agency Lake Stations, 2012.**

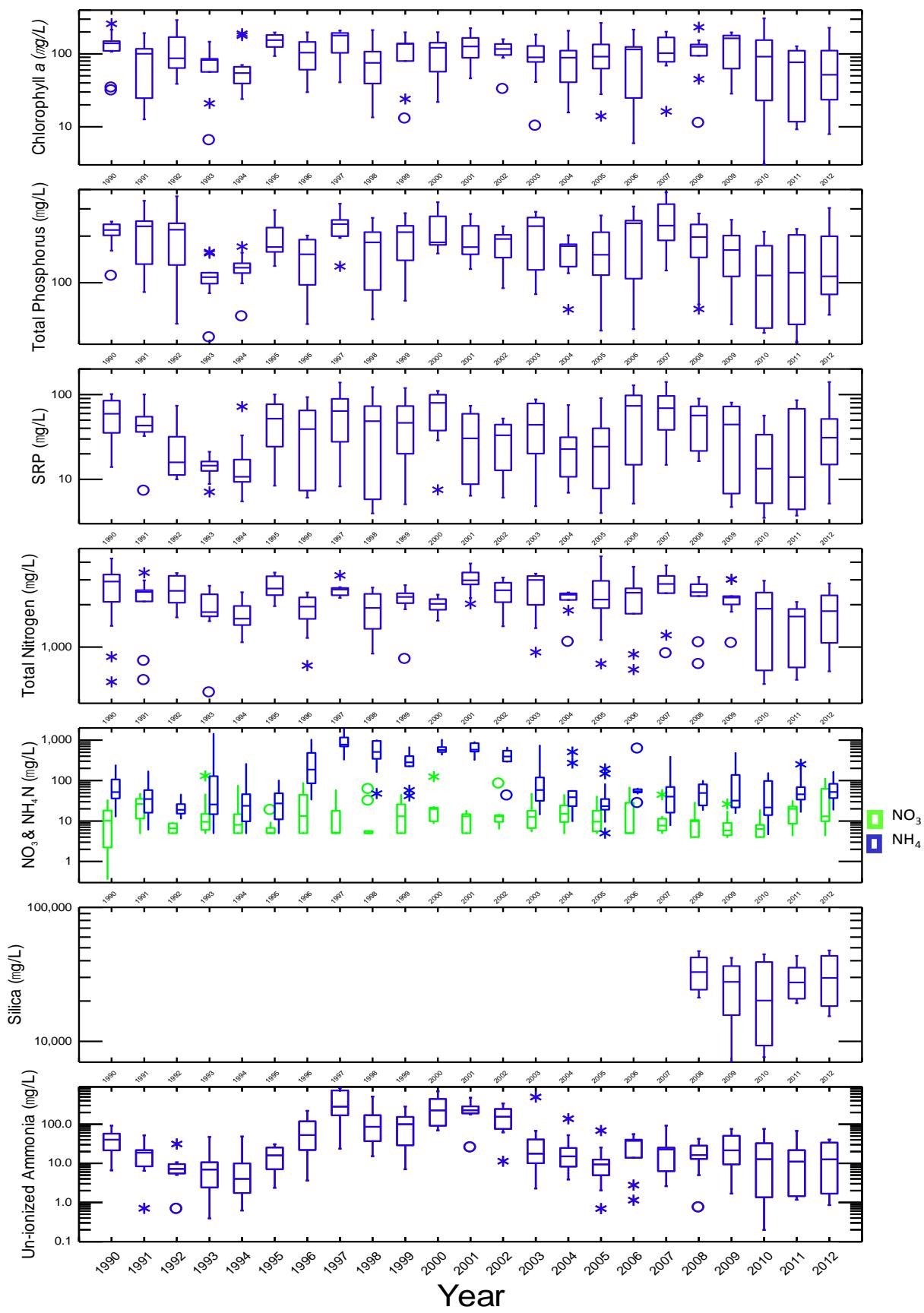
As is typical for many shallow lake ecosystems, the concentration of nutrients, their ratios, the underwater light climate, and climatic variables (e.g., temperature and wind speed) are important determinants of annual bloom dynamics of *Aphanizomenon* in UKL. During the 2010 and 2009 growing seasons (see Kann 2010; 2011) it appears that the late-spring decline in TN:TP (indicating more nitrogen limiting conditions), a later (June as opposed to May) “clear water” phase (nitrogen fixation generally has a high energy/light requirement) and cooler May-June temperatures were important determinants of *Aphanizomenon* bloom timing. Likewise, relatively high TIN concentrations and high TIN:SRP ratios, a late “clear water” phase, generally cooler and windier conditions during late-May and early-June, and cool temperatures in July apparently influenced bloom dynamics in 2011. The bloom pattern in 2012 fell somewhere in between the 2009-2010 and 2011 pattern, with declining TIN:SRP ratios and a “clear water” phase also preceding the summer *Aphanizomenon* increase. The interaction of these variables as well as other controlling factors such as lake level and hydrodynamic patterns could be explored further with additional multivariate statistical analyses on the entire dataset. As shown by Jassby and Kann (2010), lake level and climate interact to determine bloom magnitude during the early season.

## Comparison of 2012 to Previous 1990-2011 Data

To facilitate inter-annual comparisons of the major water quality variables, lake-wide means and medians were computed for UKL-only and Agency Lake-only. The distributions for the June-September period are shown in Figures 18-21 and summary statistics in Tables 3 and 4. Similar to 2010 and 2011, the June-Sep UKL-only pH distribution for 2012 was among the lowest for the period of record, with the median and lower quartile values similar to 1993 (Figure 18; Table 3).



**Figure 18. June-September distribution of UKL-only lake-wide means for T ( $^{\circ}\text{C}$ ), pH, D.O (mg/L), and Secchi depth, 1990-2012.**



**Figure 19. June-September distribution of UKL-only lake-wide means for CHL, TP, SRP, TN, NO<sub>3</sub>+NO<sub>2</sub>-N and NH<sub>4</sub>-N, 1990-2012.**







In contrast to 2011 when median DO concentration was higher than all other years for the period of record, median DO in 2012 was among the lowest despite the lack of a large bloom decline. As expected due to its controlling effect on pH, median CHL in 2012 tended to be among the lowest for the period of record (Figure 19). Although 2010 was also low (Kann 2011), overall values were noticeably lower than 2010 (Figure 19).

A similar pattern was shown for 2012 TP, although SRP was somewhat higher and more similar to values occurring in the early 2000's (Figure 19). However, similar to TP and Chl, TN was also among the lowest for the period of record. For the 23 years of record, the ammonia distribution was similar from 1990-1995, was elevated from 1996-2002, and then decreased to pre-1996 levels during the past 10 years (2003-2012). However, both 2009 and 2010 showed somewhat higher ammonia concentrations than the previous 5 years, and the lower-quartiles were higher in 2011 and 2012 than the previous 4 years (Figure 19). Inter-annual silica variability is indicated, with 2009 and 2010 showing reduced lower quartile values, possibly due to enhanced diatom blooms in spring of those years (Figure 19).

For Agency Lake, both pH and DO were also noticeably lower compared to previous years (Figure 20; Table 4). Lower quartile values of CHL, TP, and TN in Agency Lake were also among the lowest for the period of record (Figure 21; Table 4). In addition, the 2012 NO<sub>3</sub>-NO<sub>2</sub>-N distribution increased slightly from the previous 2 years, and NH<sub>4</sub>-N tended to follow the overall 23 year cyclical pattern described above, and had similar median but high lower quartile value compared to 2011 (Figure 21). Both Agency and UKL Lakes continued to show several periods of apparent sub-decadal cyclical increases and decreases for nutrient parameters over the period of record (Figure 19 and Figure 21). Inter-annual silica variability in Agency Lake is low relative to UKL.

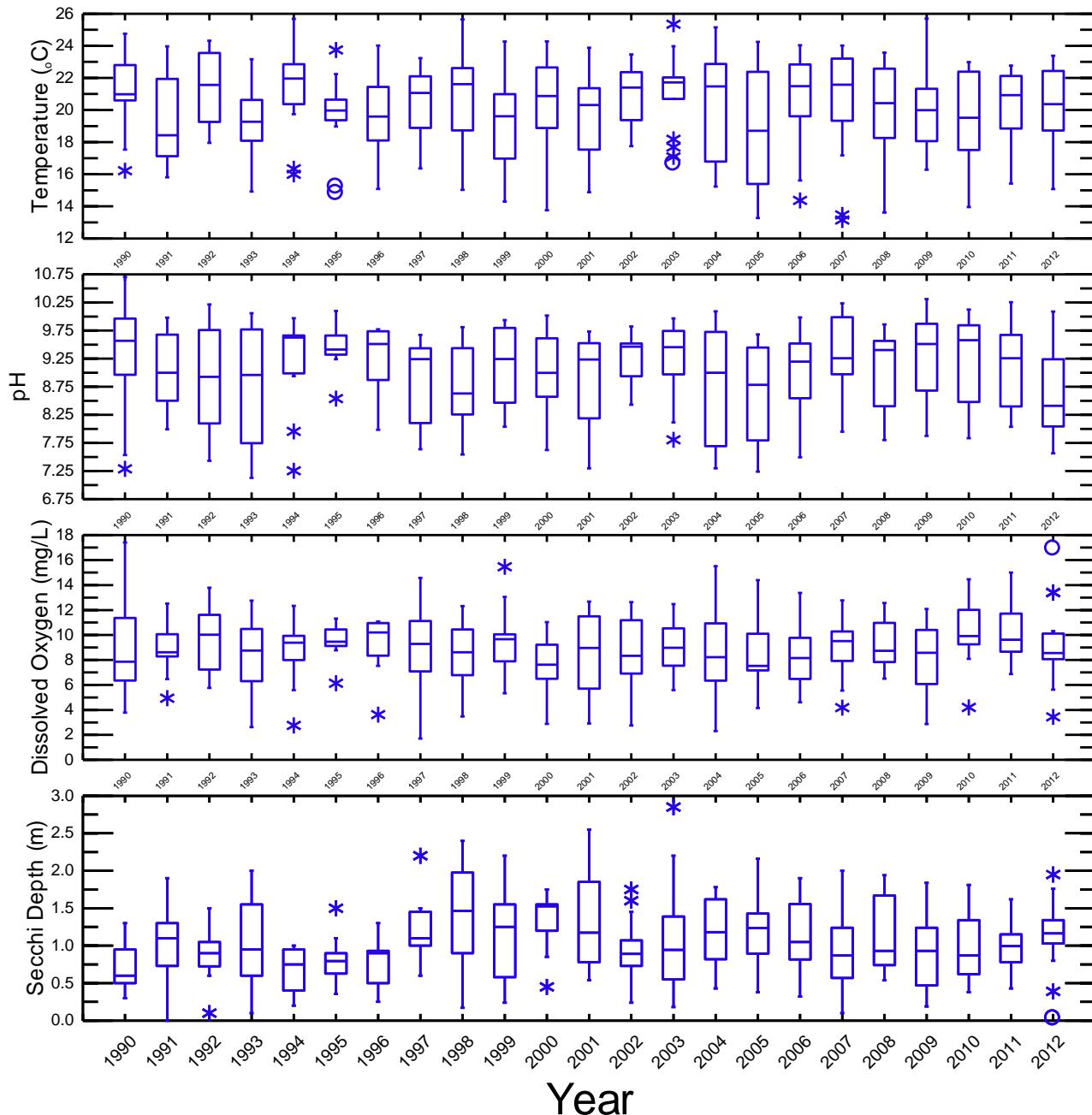
## SUMMARY

With the addition of 2012 data, the UKL water quality/limnological database now includes 23 years of data and includes the years 1990-2012. Given the dynamic and variable nature of shallow, high productivity lakes such as UKL, a long-term monitoring program is essential for assessing change relative to management programs, as well as for understanding lake dynamics.

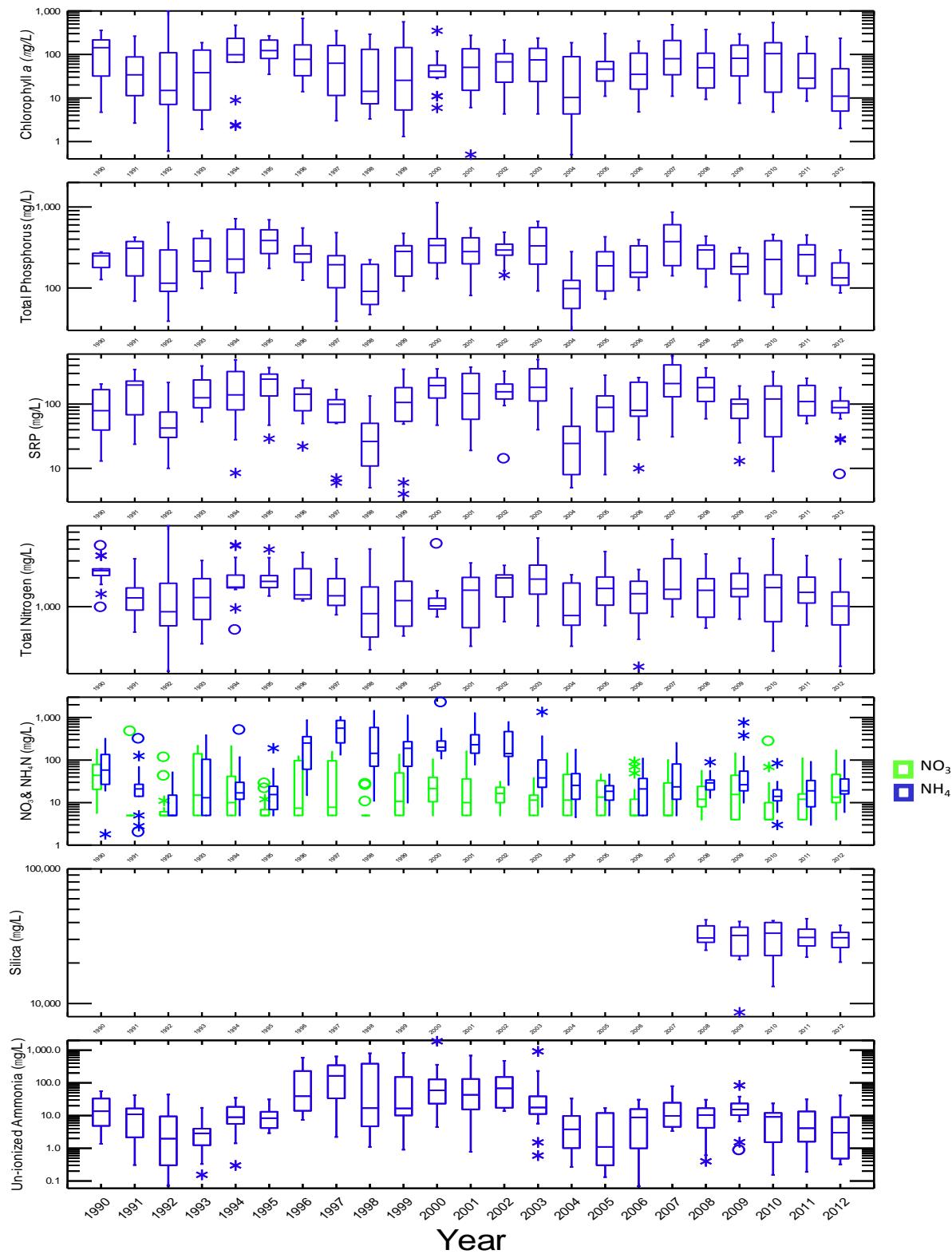
For example, as noted in earlier reports, ongoing wetland restoration is occurring in vast areas of the periphery of UKL (e.g., Wong et al. 2010; 2011), riparian and nutrient management plans (e.g., Oregon 1010 and TMDL plans) have been developed, and water use plans have been implemented (e.g., KBRT Wood River Valley programs).

Continued monitoring is recommended to accommodate the restoration time-frame (restoration of ecological function can be a multi-decade process) for Klamath Basin activities and to increase statistical power (sample size) for multi-variable analyses. Such a long-term database allows for statistical time series or trend analysis, as well as multi-variable assessment of the relationship between controlling variables (e.g., climate) and important water quality parameters (e.g., see Jassby and Kann 2010).

Further analysis (beyond the scope of the current data summary report) of the noticeable differences in CHL, as well as other water quality parameters among years will provide an opportunity to gain further insight into annual controlling factors of bloom dynamics. Additional multivariate analyses, time-series and trend analyses such as Seasonal Kendal Tests, as well as integration with current lake literature on shallow lakes and *Aphanizomenon* bloom dynamics are recommended. The analysis of the long-term Upper Klamath Lake phytoplankton and zooplankton datasets will also significantly aid in understanding annual water quality variability.



**Figure 20. June-September distribution of Agency Lake means for T (°C), pH, D.O (mg/L), and Secchi depth, 1990-2012.**



**Figure 21. June-September distribution of Agency Lake means for CHL, TP, SRP, TN, NO<sub>3</sub>+ NO<sub>2</sub>-N, SiO<sub>2</sub> and NH<sub>4</sub>-N, 1990-2012.**







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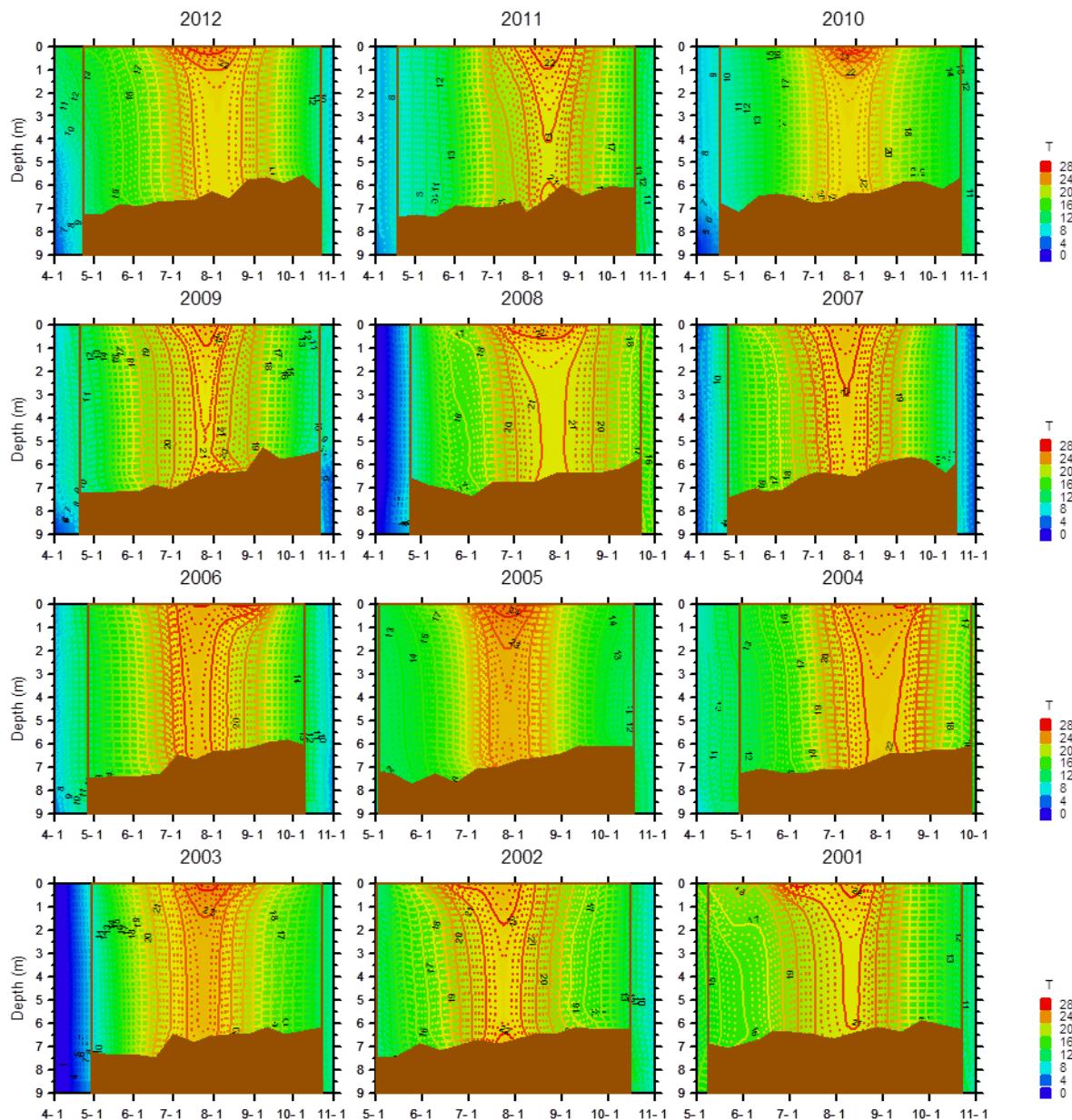
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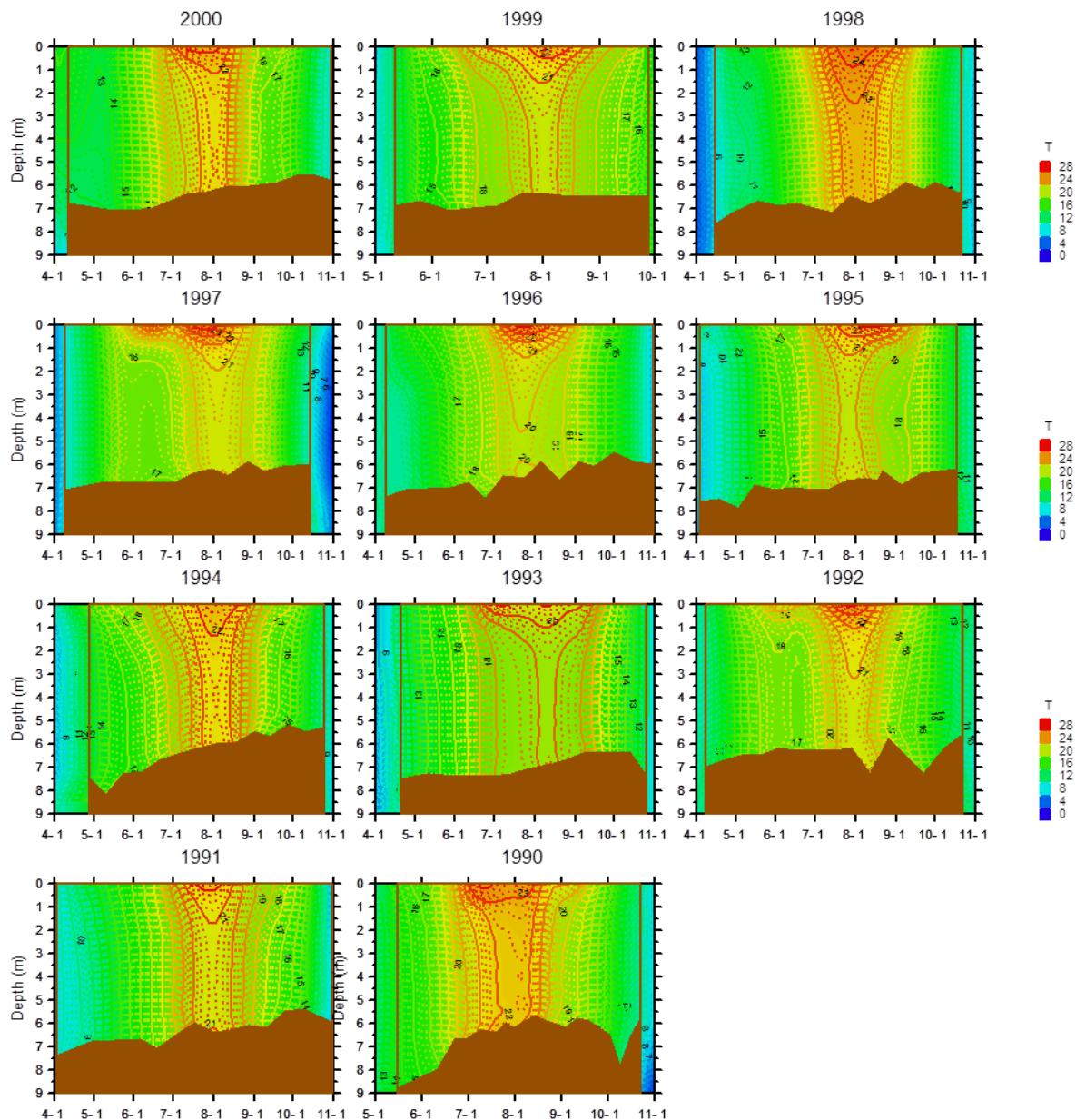
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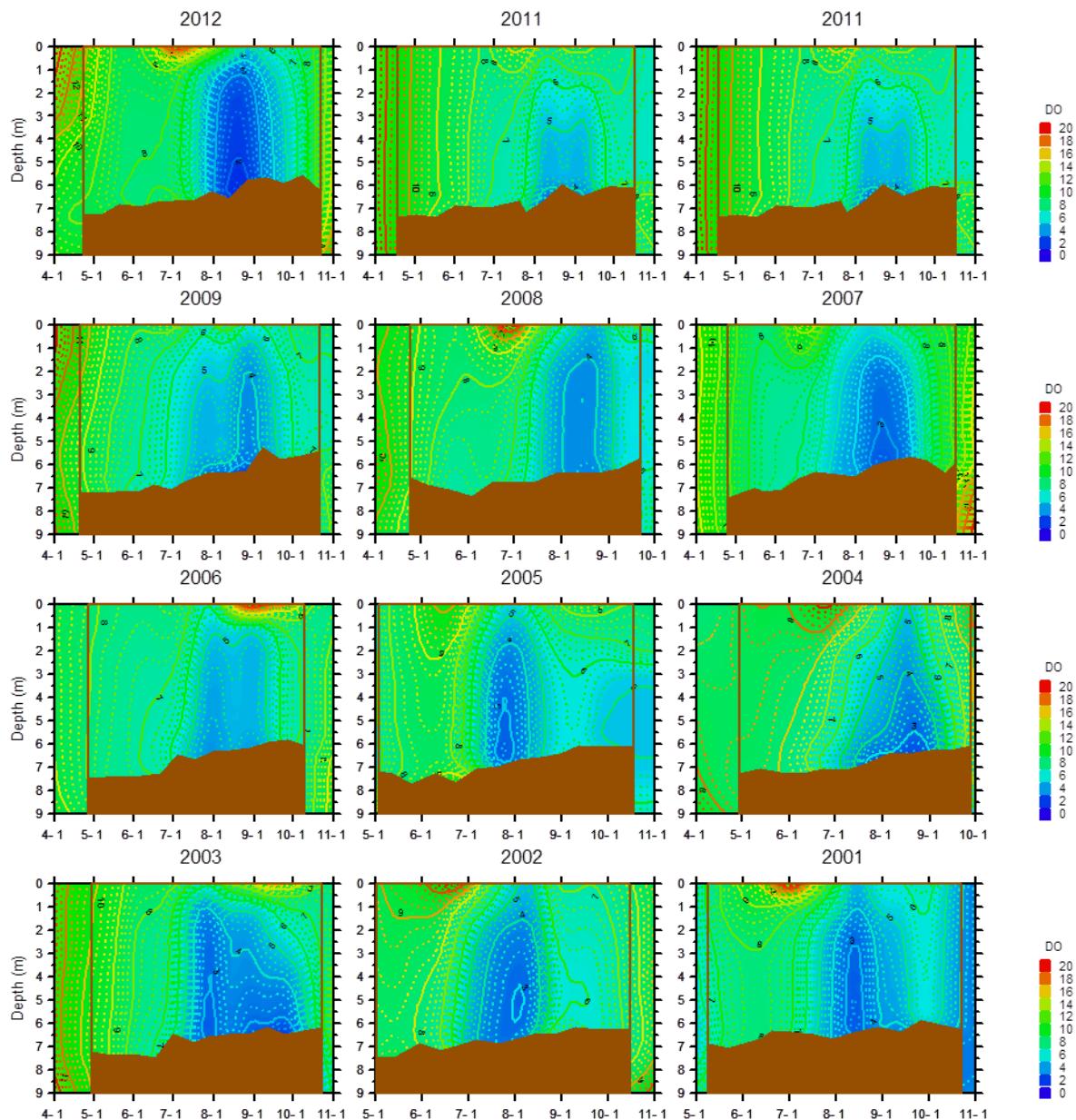
## APPENDIX I: Seasonal and water column trends in water quality profile data (T, DO, and pH)



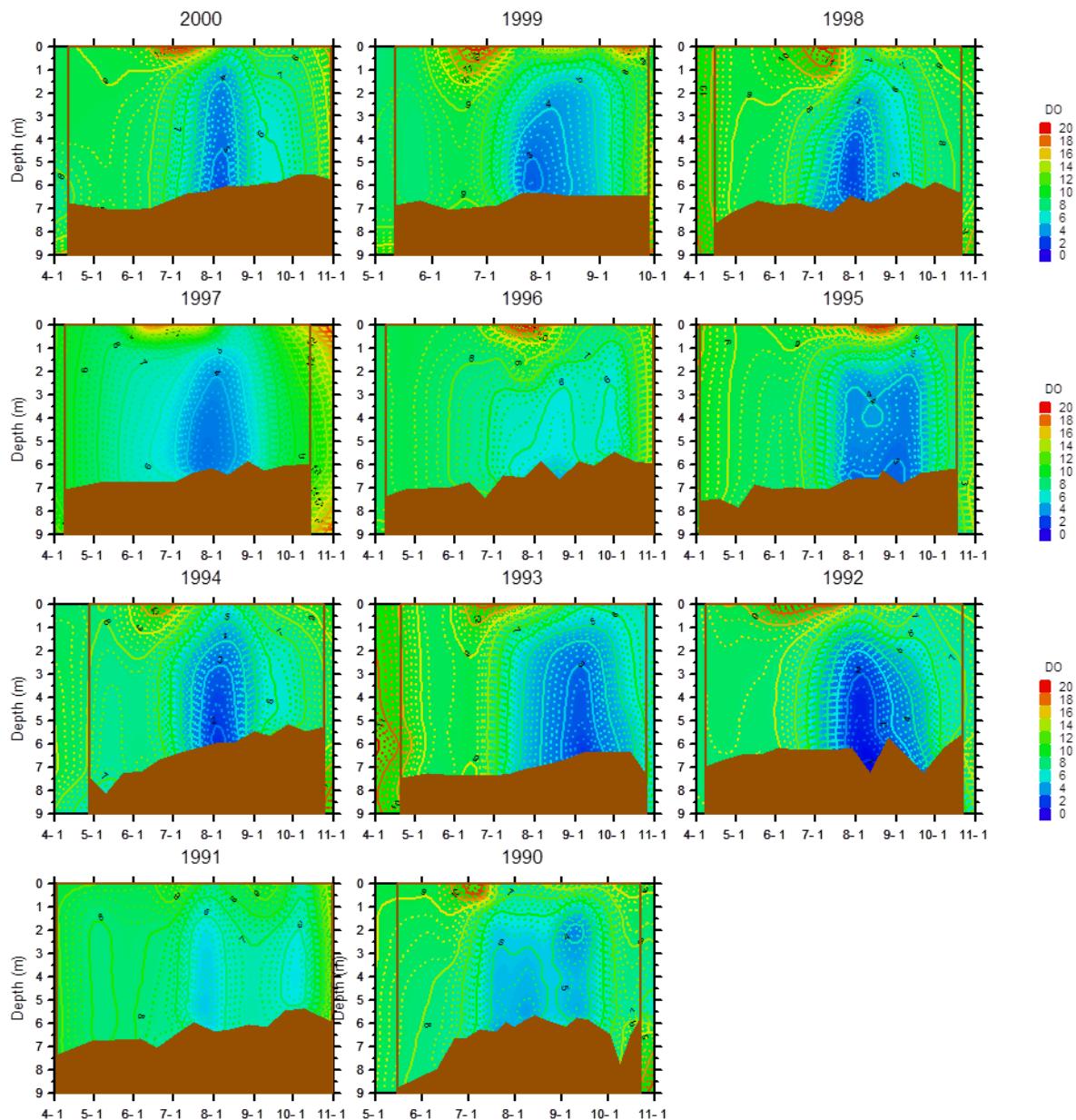
Depth-time distributions of isotherms of temperature ( $^{\circ}\text{C}$ ) at UKL station Eagle Ridge (ER), 2001-2012. Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



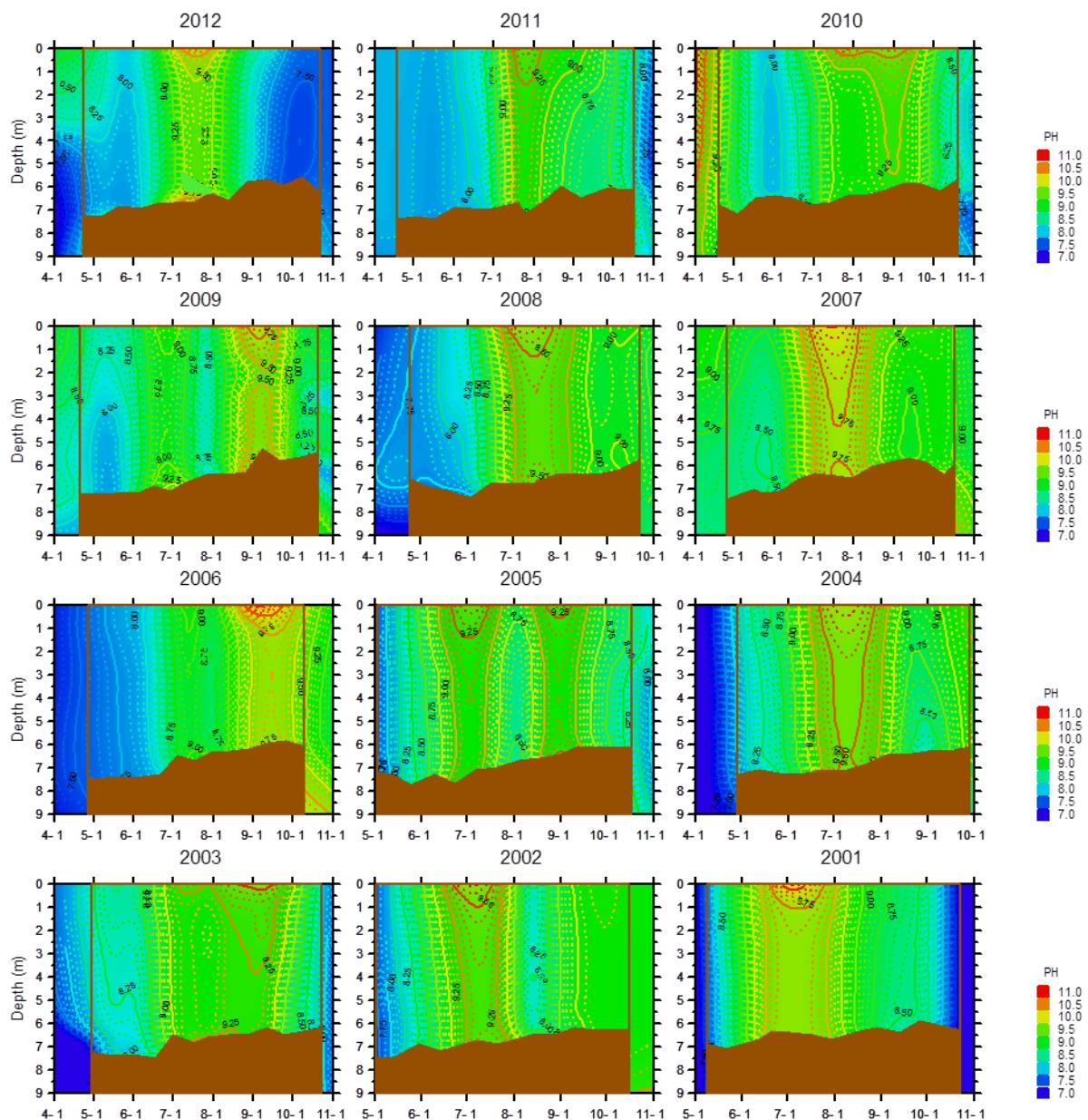
**Depth-time distributions of isotherms of temperature ( $^{\circ}\text{C}$ ) at UKL station Eagle Ridge (ER), 1990-2000. Note:**  
**1)** brown shaded area on the abscissa denotes the bottom profile depth, and **2)** contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



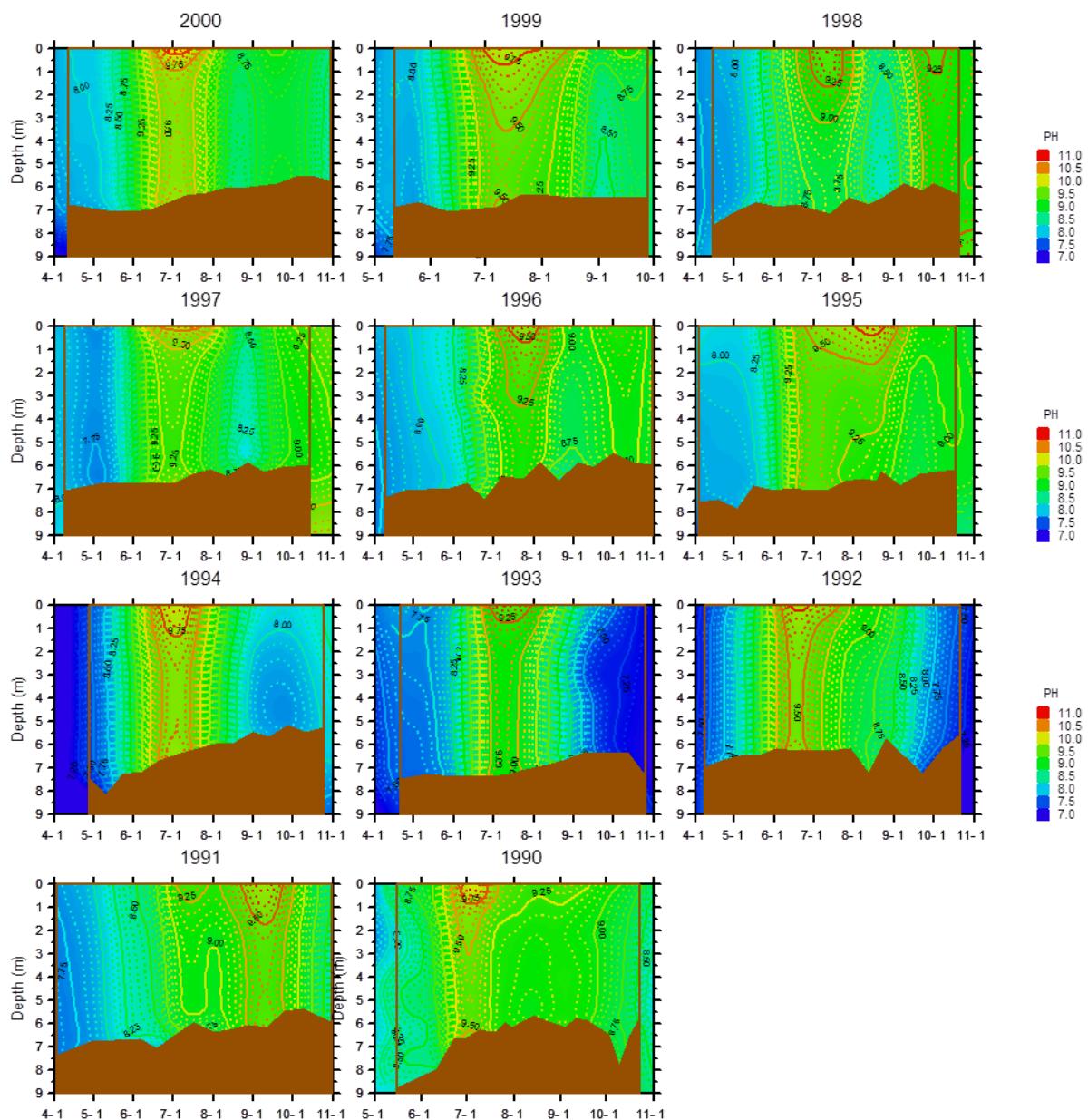
**Depth-time distributions of isopleths of dissolved oxygen (mg/L) at UKL station Eagle Ridge (ER), 2001-2012.**  
 Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



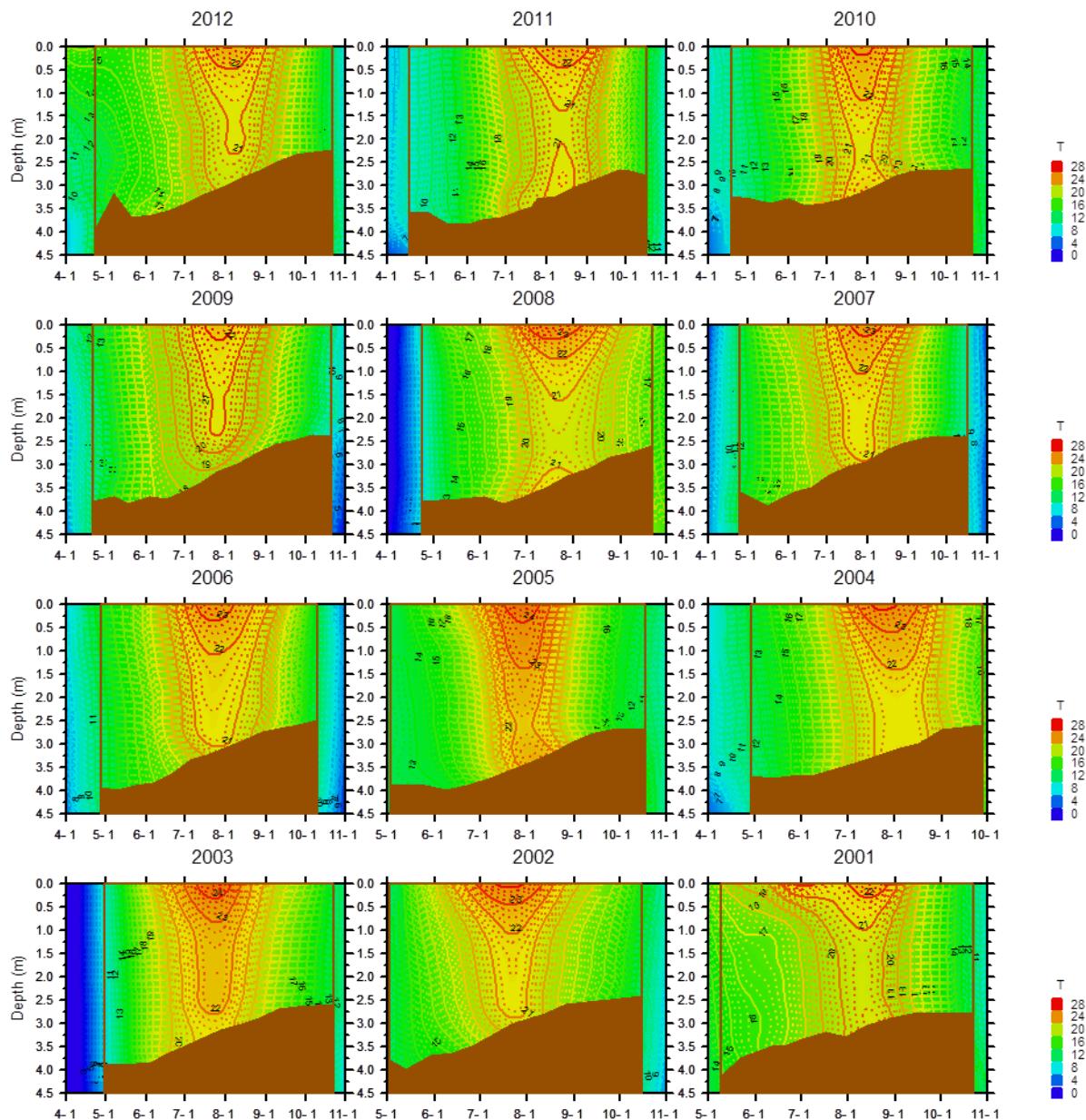
**Depth-time distributions of isopleths of dissolved oxygen (mg/L) at UKL station Eagle Ridge (ER), 1990-2000.**  
**Note:** 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



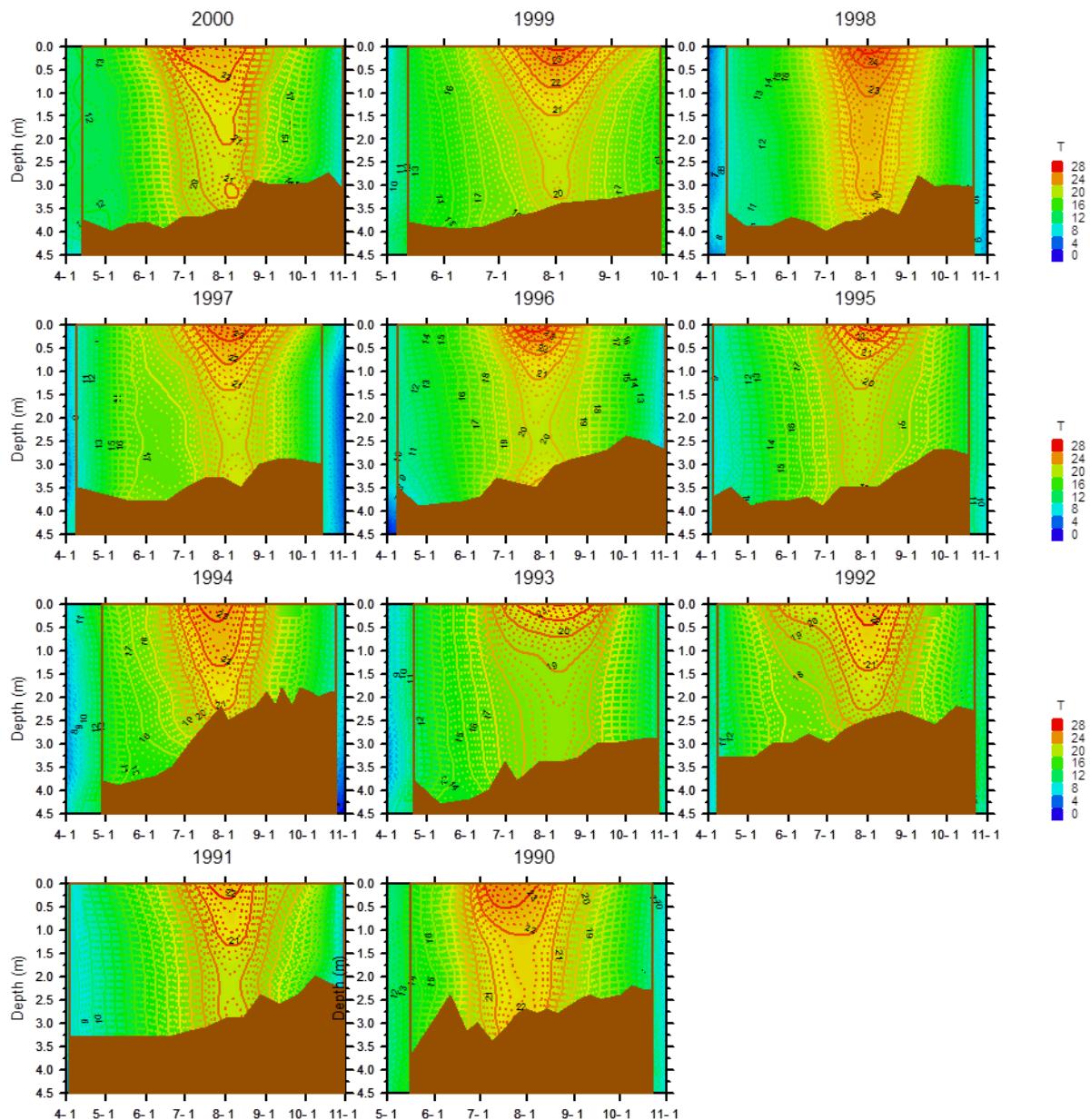
**Depth-time distributions of isopleths of pH at UKL station Eagle Ridge (ER), 2001-2012.** Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



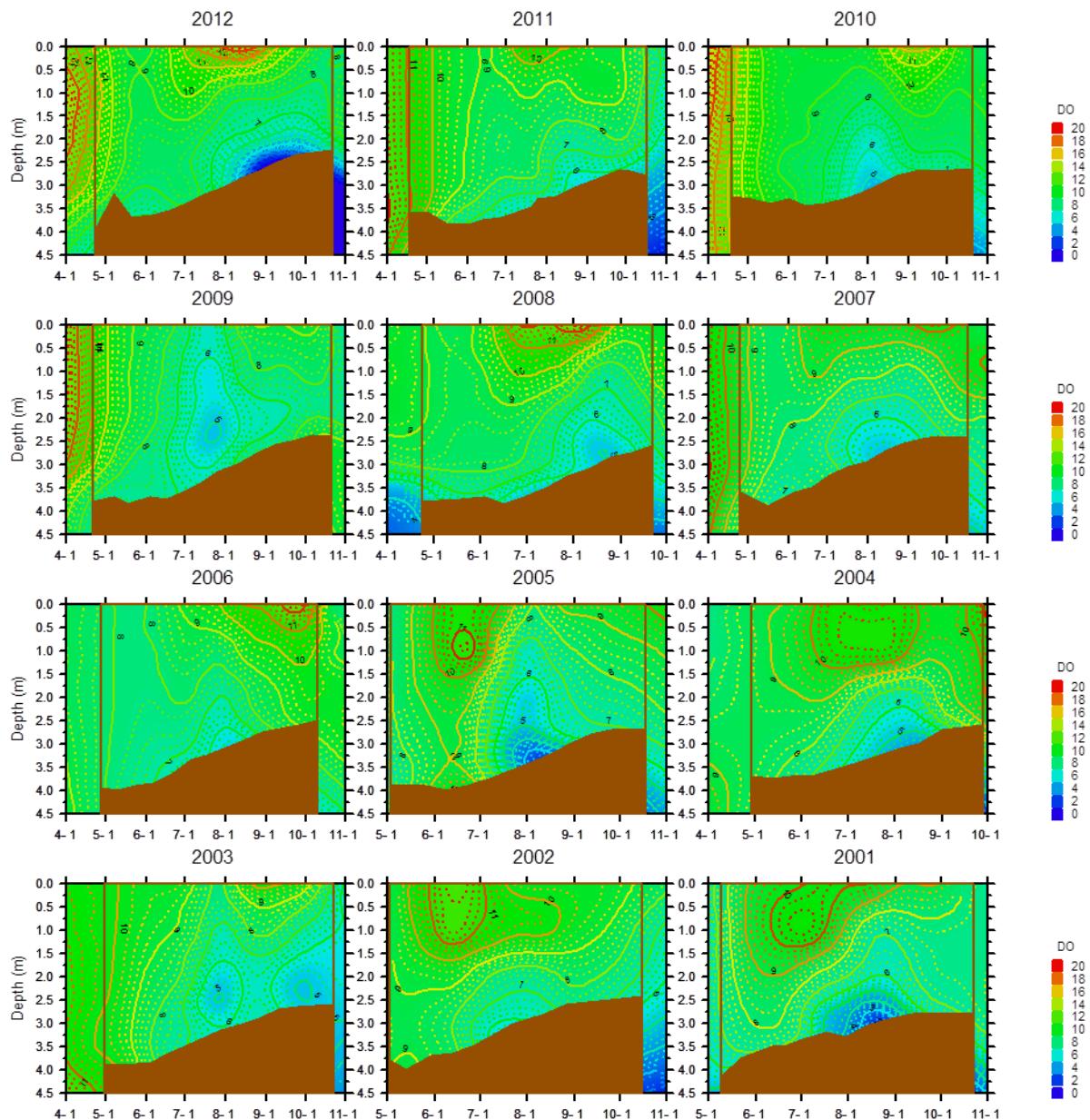
**Depth-time distributions of isopleths of pH at UKL station Eagle Ridge (ER), 1990-2000.** Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



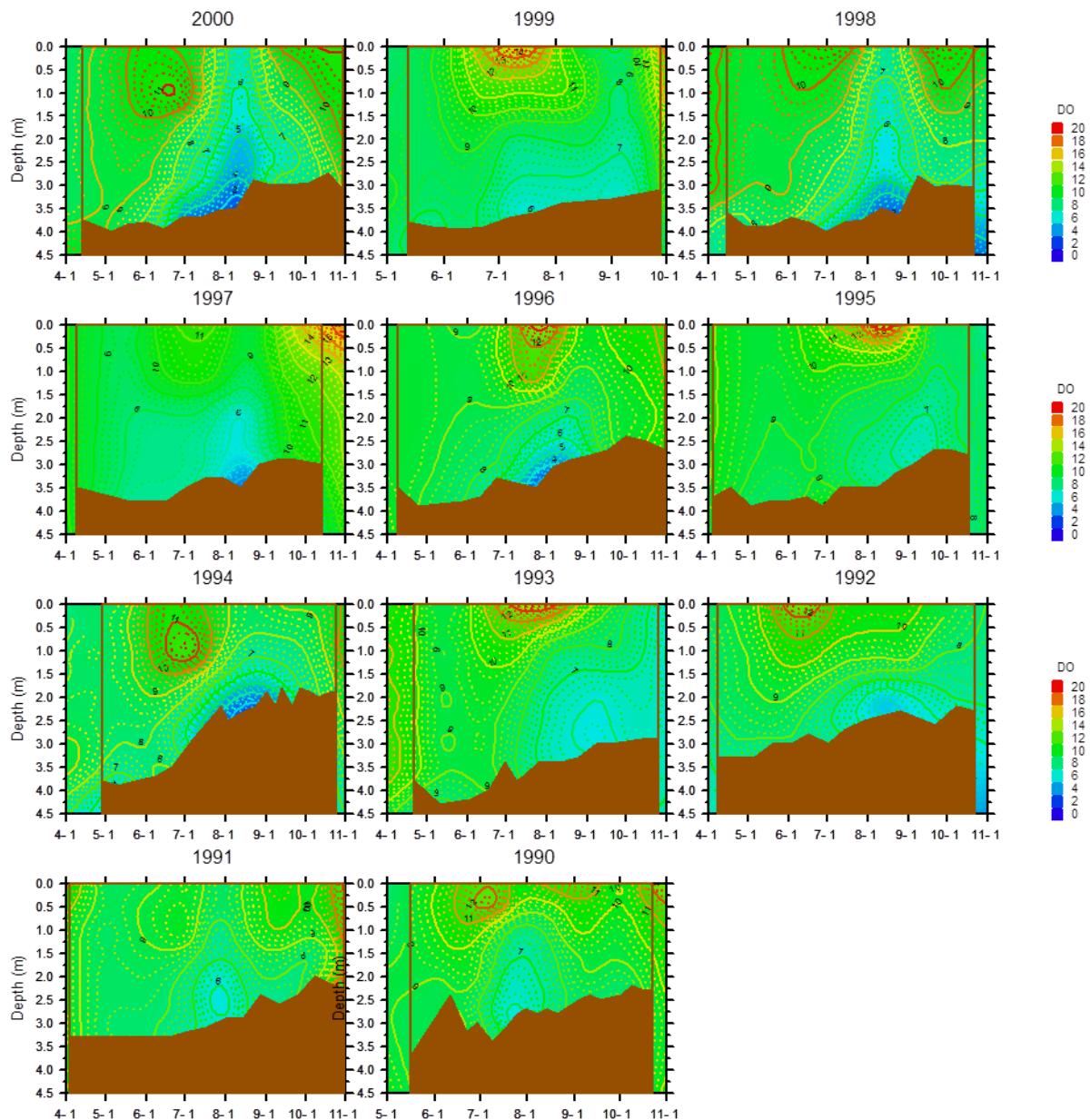
**Depth-time distributions of isotherms of temperature (°C) at UKL station Mid-North (MN), 2001-2012. Note:**  
**1)** brown shaded area on the abscissa denotes the bottom profile depth, and **2)** contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



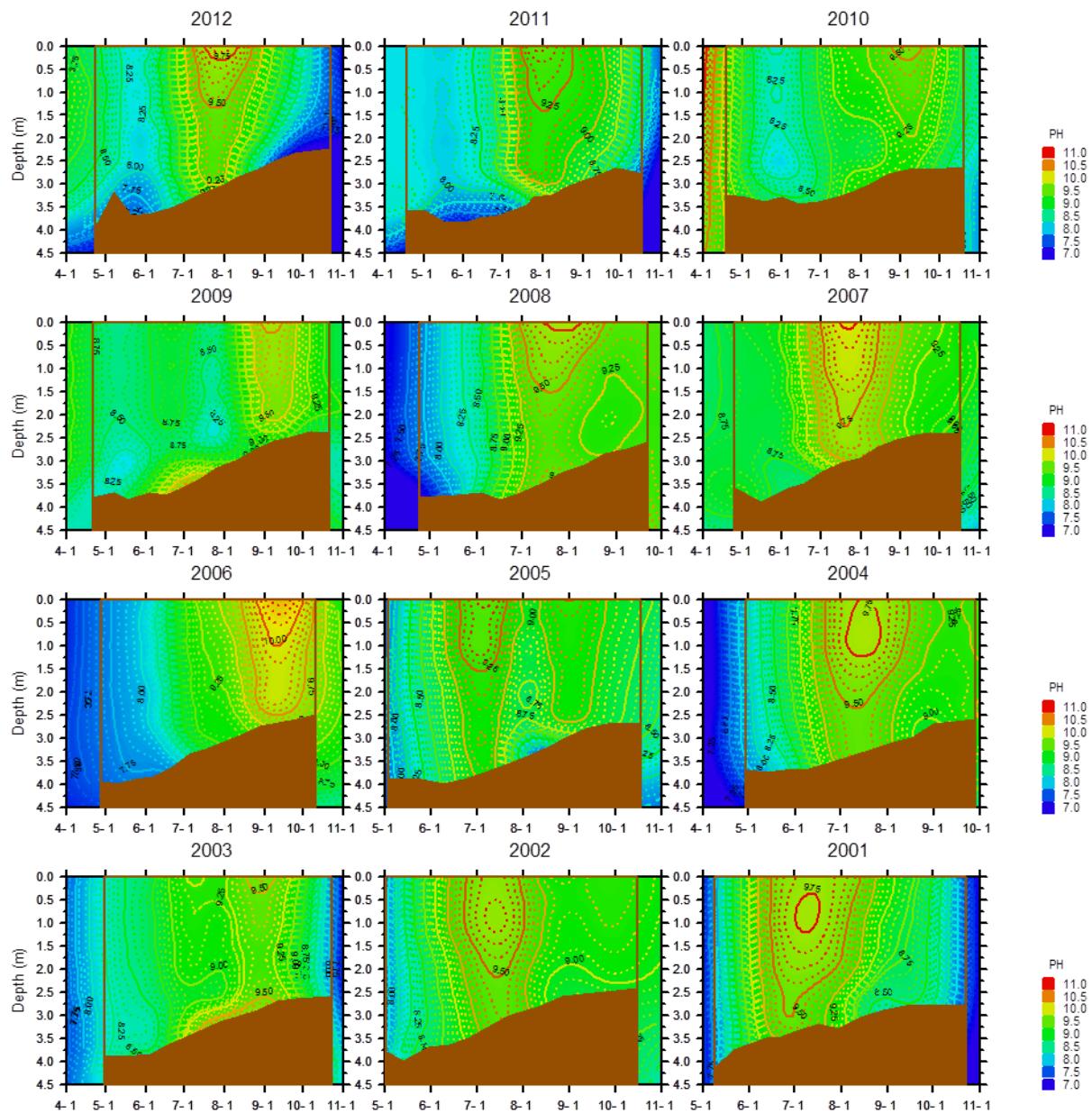
**Depth-time distributions of isotherms of temperature (°C) at UKL station Mid-North (MN), 1990-2000.** Note:  
1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



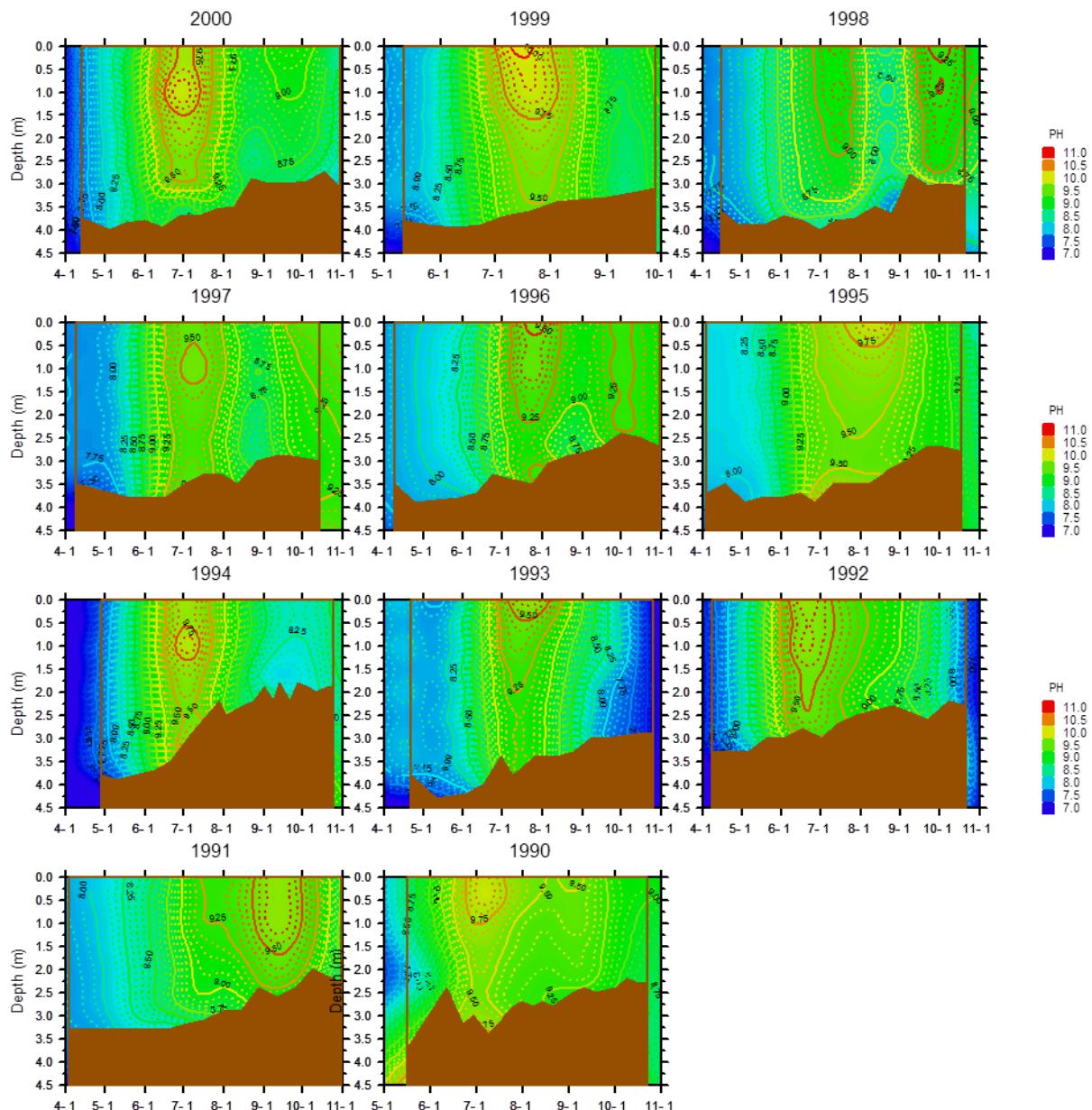
**Depth-time distributions of isopleths of dissolved oxygen (mg/L) at UKL station Mid-North (MN), 2001-2012.**  
**Note:** 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



**Depth-time distributions of isopleths of dissolved oxygen (mg/L) at UKL station Mid-North (MN), 1990-2000.**  
 Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).



**Depth-time distributions of isopleths of pH at UKL station Mid-North (MN), 2001-2012. Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).**



**Depth-time distributions of isopleths of pH at UKL station Mid-North (MN), 1990-2000.** Note: 1) brown shaded area on the abscissa denotes the bottom profile depth, and 2) contours are not valid outside of vertical brown lines (begin and end dates for seasonal sampling).













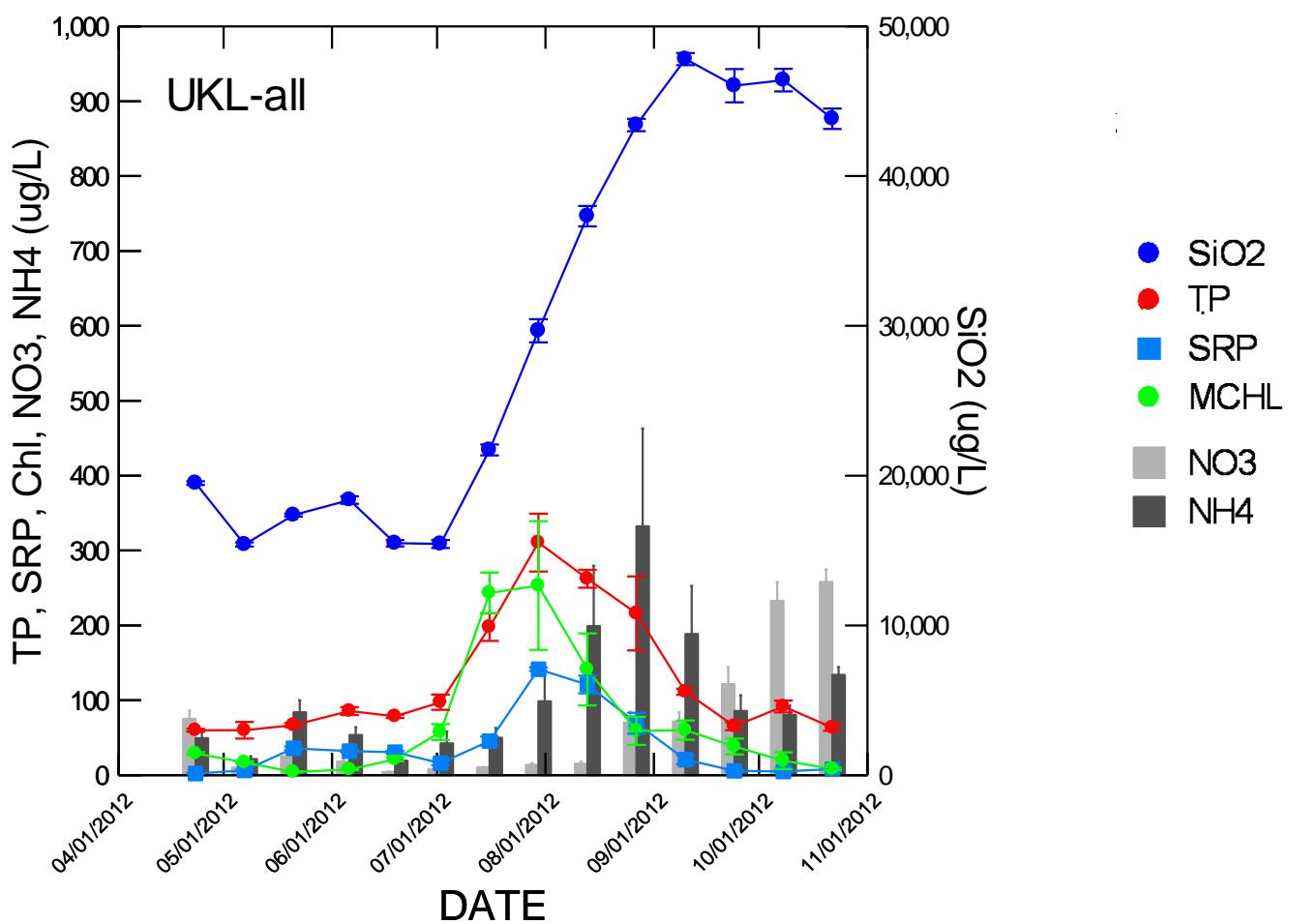






Year	Month	Parameter	Temp- erature (oC)	pH	Dissolved Oxygen (mg/L)	Secchi Depth (m)	Chloro- phyll a (µg/L)	Total Phosphorus (µg/L)	Soluble Reactive Phosphorus (µg/L)	Total Nitrogen (µg/L)	NO3+NO2 Nitrogen (µg/L)	NH4 Nitrogen (µg/L)	Un- ionized Ammonia (µg/L)
2011	7	UQ	20.8	9.73	10.5	0.83	167	122	10	1805	4	36	22
2011	8	N of Cases	16	16	16	16	16	16	16	16	16	16	16
2011	8	Median	22.0	9.05	7.4	1.03	43	206	71	1870	19	114	47
2011	8	Arithmetic Mean	21.7	9.03	7.3	1.01	85	218	75	2092	25	200	55
2011	8	LQ	21.4	8.85	5.5	0.78	8	184	64	1525	16	49	19
2011	8	UQ	22.5	9.28	9.6	1.41	81	228	83	2285	26	292	74
2011	9	N of Cases	16	16	16	16	16	16	16	16	16	16	16
2011	9	Median	18.1	8.98	8.8	0.79	87	187	35	1660	24	166	28
2011	9	Arithmetic Mean	18.3	8.91	8.4	0.89	93	186	52	1772	31	172	35
2011	9	LQ	16.7	8.72	6.4	0.65	51	145	14	1525	14	40	10
2011	9	UQ	20.1	9.11	10.3	1.01	116	211	96	1890	43	279	62
2012	6	N of Cases	16.0	16.00	16.0	16.00	16	16	16	16	16	16	16
2012	6	Median	16.4	8.13	8.5	1.03	11	80	32	687	6	24	1
2012	6	Arithmetic Mean	16.3	8.10	8.4	1.04	14	82	31	692	11	36	1
2012	6	LQ	15.1	7.82	8.2	0.94	7	73	29	622	4	18	1
2012	6	UQ	17.8	8.33	8.7	1.15	24	88	35	734	19	50	2
2012	7	N of Cases	24.0	24.00	24.0	24.00	24	24	24	24	24	24	24
2012	7	Median	21.1	9.65	8.7	0.58	152	183	49	1975	9	38	23
2012	7	Arithmetic Mean	20.7	9.62	8.7	0.63	185	202	68	2180	11	64	39
2012	7	LQ	19.5	9.29	6.8	0.40	68	101	20	1155	9	25	16
2012	7	UQ	21.8	9.93	10.6	0.91	258	275	138	2770	12	69	46
2012	8	N of Cases	16.0	16.00	16.0	16.00	16	16	16	16	16	16	16
2012	8	Median	21.0	9.09	6.7	0.80	69	223	92	2085	35	114	34
2012	8	Arithmetic Mean	21.1	9.04	6.4	0.78	100	239	95	2243	43	266	55
2012	8	LQ	19.9	8.86	4.1	0.58	34	174	73	1980	13	58	25
2012	8	UQ	22.4	9.37	9.2	1.04	130	276	133	2260	57	467	72
2012	9	N of Cases	16.0	16.00	16.0	16.00	16	16	16	16	16	16	16
2012	9	Median	18.1	8.03	7.5	0.81	51	92	8	1520	79	101	3
2012	9	Arithmetic Mean	17.9	8.08	7.0	0.85	49	88	13	1524	97	137	4
2012	9	LQ	17.3	7.72	5.9	0.73	24	63	4	1175	62	35	2
2012	9	UQ	18.6	8.37	7.9	1.02	71	115	18	1715	147	206	7

**APPENDIX III: 2012 Seasonal trends in silica and other nutrient parameters in UKL  
(lake-wide mean shown with standard error).**



## 2012 Seasonal trends in silica and other nutrient parameters by station in UKL

