The Size Distribution of Firms and Industrial Water Pollution
A Quantitative Analysis of China

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We argue that misallocation across firms amplifies industrial water pollution by distorting the firm size distribution in China. Firm-level data indicate that larger firms are more likely to use clean technology but face higher distortions. In a heterogeneous firms model with an endogenous choice of pollution treatment technologies, we show that distortions that increase with firm-level TFP lower the adoption of clean technology, amplify aggregate pollution intensity, and lower aggregate output. Quantitatively, eliminating these correlated distortions would increase output by 30% and decrease pollution by 20%. Meanwhile, environmental regulations have sizable impact on pollution but limited effects on aggregate output.

JEL: E23, L11, O44, Q52, Q53, Q56
Keywords: Firm Size Distribution, Correlated Distortion, Industrial Water Pollution

Severe pollution has accompanied China's remarkable economic growth in the past few decades, causing environmental degradation, public health damage, and millions of premature deaths.[1] Understanding the driving forces behind this severe pollution is crucial for designing and evaluating anti-pollution policies. An outstanding question is whether the severe pollution is an inevitable consequence of rapid economic growth, or largely exacerbated by policy distortions and/or market inefficiencies.

In this paper, we argue that misallocation across firms amplifies industrial water pollution by distorting the firm size distribution in China. Using nationally representative data on firm production, pollution emissions and pollution treatment technologies, we document two facts on firm size and pollution intensity (pollutants per output) in China. First, large firms have lower pollution intensity than small firms, with an elasticity of -0.37 between firm size and pollution intensity for the top five polluting industries in China. Not only do large firms generate fewer pollutants per output during production, but they are also more likely to use advanced end-of-pipe treatment technologies that require large installation costs. Second, large firms account for a smaller proportion of the total employment in China than in the United States. For the top five polluting industries in China, firms with more than 400 employees account for 20% of the total employment compared with close to 70% for the U.S. counterparts. If we take the U.S. economy as a less distorted benchmark, these facts suggest that distortions to firm size distribution could have important environmental consequences in China.

Motivated by these facts, we build a two-sector (polluting and non-polluting) model of heterogeneous firms based on Lucas (1978) to evaluate how firm-level distortions affect aggregate output and pollution. The model has the following key features. First, firms in each sector differ in their productivity levels. There is a stand-in household with a continuum of members, and each member draws a sector specific managerial ability from a given distribution. When members choose to be manager, they enter the sector their managerial ability belongs to and the ability determines the TFP of the firm they run. Second, firms differ in pollution intensity as well. As in Copeland and Taylor (1994), pollution emissions are a by-product of production for firms in the polluting sector. Following Fact 1 above, we assume that generated emissions per output during production is decreasing in firm output. Third, firms have the option to adopt an advanced end-of-pipe treatment technology, which we refer to as clean technology later, after paying a fixed adoption cost. Firms in the polluting sector are subject to environmental regulations such that firms that do not use the clean technology face a penalty that takes away a fraction of their profits. As a result, our model predicts that only firms above a certain size threshold would choose the clean technology, which is qualitatively consistent with the fact that large firms are more likely to use clean technology.

Finally, we assume that firms face distortions that increase with their TFP, or *correlated* distortions following the terminology of Restuccia and Rogerson (2008), which are modeled as progressive output taxes.[2] We find in our data that

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[2]Guner, Ventura and Xu (2008) study a closely related type of distortions, namely *size-dependent* distortions, which increase with firm or establishment size (the amount of capital, labor or output). An example is the many labor regulations in Italy, Spain and France that start to bind on firms with more than a certain number of employees. Both correlated and size-dependent distortions lead to a reallocation of resources from high TFP to low TFP firms, and they are essentially equivalent when distortions do not lead to rank reversal defined as more productive firms being smaller [Hopenhayn (2014*b*)].

firms with higher TFP exhibit substantially higher average products of capital and labor in both the polluting and non-polluting sectors, indicating that more productive firms do face higher distortions [Hsieh and Klenow (2009, 2014)]. As emphasized by previous studies, such as Restuccia and Rogerson (2008) and Hsieh and Klenow (2014), correlated distortions lower both mean firm size and aggregate output, since they limit the operations of productive firms and allow too many unproductive firms to survive and expand. A new insight from our model is that correlated distortions also increase aggregate pollution intensity for two reasons. First, they lower the profits of productive firms which reduces their incentives to adopt the clean technology. Therefore fewer firms adopt clean technology in the equilibrium with correlated distortions. Second, they take production factors away from large productive firms that also have lower pollution intensity, which drives up the average pollution intensity further.

Conventionally, environmental economics decomposes changes in aggregate pollution into three effects: scale (change in aggregate output), composition (reallocation across sectors), and technique (change in sectoral pollution intensity) effects.[3] Our model allows us to assess structurally how distortions affect aggregate pollution through the three channels. Moreover, our model and data allow us to further decompose the technique effect into within sector reallocation across firms and the change in the adoption of clean technology.

To discipline our quantitative analysis, we calibrate our model to match the observed pollution intensity, the adoption of clean technology, the correlation between distortions and firm TFP, and the firm size distribution in the Chinese data. We map the polluting sector to the top five polluting industries mentioned above, which account for 80% of the total emissions in the data, and map the non-polluting sector to the rest of the manufacturing sector. We then use the calibrated model to evaluate the effects of two policies: removing correlated distortions from both sectors and strengthening environmental regulations.

Our quantitative results show that removing correlated distortions from both sectors would increase mean firm size by 130% and aggregate output by 30%, while lowering the pollution intensity of the polluting sector by 42%. Since both sectors share similar distributions of firm TFP and distortions in the data, eliminating correlated distortions induces little reallocation across sectors. Therefore, aggregate pollution drops by 23% because the technique effect dominates the scale effect. About half of the technique effect comes from the reallocation of factors across firms, while the rest is from a 27 percentage points increase in the adoption rate of clean technology. To isolate the importance of the progressivity of distortions, we solve a version of the model in which the progressive taxes are replaced with flat taxes while holding the total tax revenue constant. Our results show that while the progressivity of distortions does not imply large output losses, it plays a central role in amplifying aggregate pollution: replacing the progressive

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[3]See for example, Grossman and Krueger (1993), Copeland and Taylor (2004) and Levinson (2009) among others.

taxes with flat taxes would cause a 7% increase in aggregate output, and a 30% decrease in aggregate pollution.

In the second experiment, we increase the regulations to the extent that the fraction of firms adopting clean technology is the same as in the first experiment. We find that the policy reduces aggregate pollution by 14%, but has very limited effect on aggregate output in the long run. Comparing this with the first experiment, we find that tightening environmental regulations improves the resource allocation on the extensive margin of resource allocation by driving small unproductive firms out of the economy. However, the allocation worsens on the intensive margin in the sense that among the remaining active firms, the production of medium-sized firms expands more at the expense of that of large firms. In contrast, the removal of correlated distortions improves the allocation on both margins.

Overall, the key message from our findings is that policy distortions and market inefficiencies force developing countries such as China to operate within the production possibility frontier between economic output and environmental quality. Therefore, contrary to a popular belief among some policymakers in developing countries, there is not necessarily a trade-off between economic growth and pollution. Both an increase in output and a reduction in pollution can be attained by reducing policy distortions and market inefficiencies.

Related Literature.—This paper is closely related to studies on the aggregate consequences of misallocation across heterogeneous firms. Pioneers in this field such as Guner, Ventura and Xu (2008), Restuccia and Rogerson (2008), and Hsieh and Klenow (2009) establish that firm-level distortions can potentially cause large losses to aggregate output, especially those that correlate with firm size or productivity.[4] Our study shows that the impact of firm-level distortions goes beyond aggregate economic output: they cause not only a large decrease in aggregate output, but also a large increase in aggregate pollution. By considering both aggregate output and pollution, our study thus provides a more complete understanding of the consequences of policy distortions given the large costs of pollution in developing countries such as China.

This paper also contributes to a growing literature in environmental economics that emphasizes the importance of firm-level heterogeneity. One line of research uses the heterogeneous firms framework to evaluate specific environmental policies. Li and Sun (2015) and Li and Shi (2017) compare the welfare effects of emission taxes and standards. Their theoretical findings imply that emission taxes will generally be favored when the dispersion of productivity across firms is large. Relatedly, Tombe and Winter (2015) use a heterogeneous firms model calibrated to the U.S. to evaluate the productivity losses from output-based intensity standards. They find that these losses can be large. We differ from this line of

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[4]For more recent works, see Bartelsman, Haltiwanger and Scarpetta (2013), Bhattacharya, Guner and Ventura (2013), Hsieh and Klenow (2014), Bento and Restuccia (2017), and Guner, Parkhomenko and Ventura (2018), among others. Restuccia and Rogerson (2013), Hopenhayn (2014a), and Restuccia and Rogerson (2017) summarize important contributions to the literature.

research by our focus on how correlated distortions affect aggregate pollution and how they interact with environmental policies.

Another line of research focuses on how international trade affects the environment through resource reallocation across heterogeneous firms. Cherniwchan, Copeland and Taylor (2017) provide a comprehensive survey of recent contributions.[5] Our paper differs from this literature in two aspects. First, most studies in this literature, with the exception of Shapiro and Walker (2018), are empirical where models are used primarily to motivate the empirical specification indirectly. Instead, we discipline our general equilibrium model using firm-level data directly in an internally consistent way, and quantitatively assess counterfactual policies using the model. Our quantitative analysis explicitly takes into account the general equilibrium effects of alternative policies, allowing us to investigate novel policy implications that are absent from reduced-form or partial equilibrium analyses. Second, using the firm-level data on treatment technologies, we provide an explanation for the negative correlation between firm size and pollution intensity. It allows us to not only investigate the overall effects of distortions, but also quantify an important channel through which the distortions affect emissions, namely, the adoption of clean technology.

Finally, our results highlight the role of firm size and misallocation for technology adoption in developing countries. As emphasized by Parente and Prescott (1994, 1999), an important task in development economics is to understand the slow rate of technology adoption in developing countries. Recent contributions include Bustos (2011), who study the role of tariff reductions, and Acemoglu et al. (2012), who focus on taxes and subsidies. Using direct observations on the adoption of pollution treatment technologies by firms, we establish empirically that firm size plays a key role in technology adoption. Our quantitative framework thus allows us to quantify how policy distortions and market inefficiencies that distort firm size would impede technology adoption.

The remainder of the paper proceeds as follows. Section I describes the data sources and empirical findings of this paper. Section II introduces the model and analyzes its property. We calibrate the model in Section III, and perform counterfactual experiments to study the effects of correlated distortions and environmental policies in Section IV. Section V concludes the paper.

I. Empirical Analysis

In this section, we describe our empirical findings on the firm-level pollution intensity and the firm size distribution in China. We start by introducing the data sources we use, and then document the key facts on pollution intensity, pollution treatment technologies and employment shares across firms of different sizes. In

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[5]Notable examples include Martin (2013), Cherniwchan (2017), Barrows and Ollivier (2018), Forslid, Okubo and Ulltveit-Moe (2018), and Shapiro and Walker (2018). A more detailed comparison between our paper and this line of research can be found in Online Appendix H.

the end, using a simple accounting exercise, we show that firm size distribution plays a quantitatively important role in aggregate pollution intensity.

A. Data Sources

This study draws upon three major data sources: the 2007 National General Survey of Pollution Sources (NGSPS) in China, the 2004 China National Economic Census (CNEC), and the Statistics of U.S. Businesses (SUSB).

NGSPS.—The NGSPS was a joint effort of multiple national ministries in China. It covers all of the pollution units from the agricultural, industrial, and domestic sources, as well as facilities for the centralized treatment of pollution. In this study, we focus on the industrial pollution sources, which consist of businesses generating pollutants from the 39 manufacturing industries. The survey is organized directly by the State Council with data collection and quality control done by specially trained field staffs. According to the *Regulation on the National General Survey of Pollution Sources [Decree of the State Council of the People's Republic of China (No. 508)]*, faking data would lead to financial penalties or even lawsuits. Therefore, the dataset is administrative in nature and is subject to less concerns of systematic under-reporting.[6] While we are not aware of other economic studies that use NGSPS, a number of studies in environmental sciences have used this dataset. See, for example, Niu et al. (2016), Qi et al. (2017) and the references therein. A more detailed description of the dataset can be found in Online Appendix A.

The NGSPS collects information on the emissions of multiple air, water, and solid waste pollutants. In this paper, we focus on water pollution mainly because the data are more accurately measured than air pollution.[7] Among all the water pollutants, we focus on *chemical oxygen demand* (COD), a major water pollutant that is commonly used to measure water quality. It measures the overall quantity of contaminants that will eventually cause oxygen loss and thus the death of living creatures. We choose COD because a much larger share of firms report positive emissions of COD compared to other pollutants. If we focus on other pollutants, such as hexavalent chromium, we would limit ourselves to a much smaller share of firms that engage in special production processes like leather tanning or pigment

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[6]In fact, uniform under-reporting of emissions by small and large firms will not change the basic conclusions of this paper. Our results will be contaminated only if under-reporting by large firms is more severe than that by small firms, which is less likely in most polluting industries since larger firms are usually inspected more carefully. For instance, many large firms in these industries are identified as *key sources*, and therefore are inspected and monitored by the government routinely and carefully, making it harder for them to fake data during this one-time survey.

[7]After our interview with the field staff, we learned that the differences in data quality was mainly caused by the differences in the sampling environment. For water pollution, firms rarely discharge wastewater directly. Instead, the wastewater is first stored in a sedimentation tank to settle heavier solids to the bottom and let lighter materials to float. This process creates a quiescent environment for sampling. For air pollutants, the sampling tube has to be placed in the chimneys, where the environment is much more volatile. Moreover, if firms were to manipulate their instant emissions when the field staff is on-site sampling, the sedimentation process makes it more difficult, since firms would have to clean the tank prior to the on-site visit of the field staff.

TABLE 1—STATISTICS OF THE TOP 5 POLLUTING INDUSTRIES BY COD

	Paper	Food Processing	Textiles	Chemical Materials	Beverages
% in Total COD Emission	33.4	15.1	14.0	10.4	4.27
% of Emissions by Key Sources	99.6	91.8	91.1	99.7	65.1
% of Production by Key Sources	87.2	69.3	48.3	98.6	88.1

Sources: NGSPS. The acronyms respectively refer to (with the two-digit GB/T4547-2002 classification code in parentheses): Paper and Paper Products (C22), Processing of Food from Agricultural Products (C13), Textiles (C17), Raw Chemical Materials and Chemical Products (C26), and Beverages (C15).

manufacturing.[8] Moreover, COD emissions are typically highly correlated with the emissions of other pollutants when both are positive.[9]

The data reveal that industrial water pollutants are typically concentrated in a handful of industries, including COD. In our empirical analysis, we focus on the top five polluting industries in terms of COD emissions, which account for nearly 80% of all COD emissions from industrial sources. Within narrowly defined manufacturing industries, the NGSPS divides firms into two large groups—*key sources* and *regular sources.* Firms identified as key sources are deemed as more polluting, and therefore are inspected and monitored by the government more carefully. During this census, the field staff also collected extra information on these firms, and the data from these firms were inspected with greater care. Importantly, the key sources include both small and large firms, and the empirical densities of the log output of the key sources resemble the bell shape of those of all firms.[10] As a result, when empirically analyzing the relationship of firm size with pollution intensity and treatment technology adoption, we focus on the key sources, which account for more than 90% of total COD emissions and 80% of total output from the top five polluting industries (Table 1). The variables we use are the firm's output, the amount of major pollutants generated and emitted, installation costs of pollutant treatment technologies, industry (up to four-digit GB/T4574-2002), ownership classification, and province.

CNEC.—The CNEC was conducted by the National Bureau of Statistics in China. It was designed to cover all businesses that undertake economic activities in the secondary and tertiary industries in China. We use observations that belong to the manufacturing sector. The variables we use are the firm's output, labor compensation, book value of capital stock, number of employees, industry (up to four-digit GB/T4574-2002), ownership classification, and province.[11]

[8]See Table A.1 in Online Appendix A for the share of firms that report positive emissions for different water pollutants. We would like to stress that many firms do not report emissions of specific pollutants such as cyanidium or chromium simply because they do not use related materials in production. It is not because of under-reporting or other issues related to data quality. In fact, since these pollutants are more hazardous than others, firms that discharge these pollutants are usually monitored more closely.

[9]In the Paper and Paper Products industry, for example, the correlation between the emissions of COD and NH_4^+ is corr(COD, NH_4^+) = 0.82 and that between COD and BOD is corr(COD, BOD) = 0.94.

[10]Please see Online Appendix A.1 for the formal definition of the key sources, and Figure B.1 in Online Appendix B for the empirical densities of the log output of the key sources.

[11]A widely used firm-level dataset in China is the Annual Surveys of Industrial Firms (ASIF). We use

FIGURE 1. FIRM SIZE AND POLLUTION INTENSITY
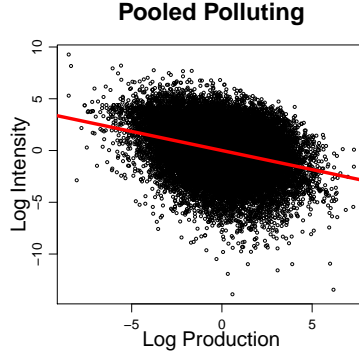
Source: NGSPS.

Note: The line is the least square fit. Residuals after controlling for the covariates are plotted.

SUSB.—The SUSB from the U.S. Census Bureau, is an annual series that provides national and subnational data on the distribution of economic activities by firm size and industry. We use the data on total employment by firm size and industry from 2004.

B.   Firm Size and Pollution Intensity

Firm Size and Pollution Intensity.—To examine the statistical relationship between pollution intensity and firm size for the polluting sectors in China, we run the following regression using data on firms identified as key sources in the top five polluting industries:

$$(1) \qquad \log(\text{COD}_i) = \underset{(0.17)}{-3.75} + \underset{(0.01)}{0.63} \times \log(\text{Output}_i) + \mathbf{X}\gamma + \varepsilon_i,$$

where $\mathbf{X} = \{\mathbf{X_s}, \mathbf{X_p}, \mathbf{X_o}\}$ refers to the dummy variables for two-digit industry, province, and ownership respectively. The estimates are all statistically significant at the 0.1% level, with the standard errors reported in the parentheses below the estimates.

These estimates imply that as a firm's output increases by 1%, its COD emissions would on average increase by only 0.63%. We define pollution intensity as COD emissions per unit of output. Then the above estimates imply an elasticity of -0.37 between firm-level pollution intensity and output. That is, other things

CNEC instead of ASIF because that ASIF does not have information on non-state firms with sales below 5 million CNY, which are crucial for our analysis. If we used the 2004 ASIP, we would have missed 29% of the employment and 80% of the firms.

Table 2—Firm Size and Treatment Technologies

Technology	Mean Efficiency	Adoption Rate	Median Costs	Median Output
Physical	63.37%	25.89%	100	100
Chemical	74.96%	33.64%	308	219
Biological	80.90%	38.60%	923	500

Source: NGSPS.
Note: The numbers reported are for the Paper and Paper Products (C22) industry. Abatement efficiency is defined as $1 -$ COD Emitted/COD Generated. The median installation costs of physical technologies and median output of firms that adopt them are normalized to 100.

equal, a 1% increase in firm output is associated with a 0.37% decrease in pollution intensity. Figure 1 visualizes this negative relationship, where the residuals from regressing intensity and output on the dummies are plotted to remove the influences of other covariates. The estimation has an R^2 of 0.41, suggesting that nearly half of the variations in COD emissions can be explained by the variations in firm output and in the three sets of dummies.[12] We would also like to point out that the coefficient of the dummy on whether a firm is state-owned or not, is statistically insignificant. As a result, we do not distinguish between whether a firm is state-owned or not in our later analysis.

The Role of End-of-Pipe Treatment Technologies.—An important question is, why do larger firms have lower pollution intensity? To answer this question, we exploit the detailed information on the end-of-pipe wastewater treatment technologies that firms use in the NGSPS. The NGSPS groups wastewater treatment technologies into five categories: physical, chemical, physio-chemical, biological, and combined technologies. In the subsequent analysis, we drop physio-chemical technologies because fewer than 0.5% of the firms adopt this type of technologies. In addition, because combined technologies look very similar to biological technologies for all the technical aspects we are interested in, we group them with biological technologies.[13] In sum, we are left with three broad categories of technologies.

To illustrate the difference in technology adoption by firms of different sizes, we use the Paper and Paper Products industry as an example. By focusing on one specific industry, we abstract from the potential heterogeneity in production processes across industries. Table 2 shows the mean abatement efficiency, median installation costs, and median output of firms that adopt these technologies, where the abatement efficiency is defined as one minus the ratio of emitted COD to generated COD. We find that for each 100 tons of COD generated, firms with biological technologies on average emit 19.1 tons of COD, compared to 36.6 tons

[12]We also tried other econometric specifications, and found similar negative correlations between pollution intensity and firm size. For instance, we estimated versions of Equation (1) for each industry and with robust standard errors clustered in provinces. Please see Online Appendix C for more information.

[13]Examples of the technologies attributed to these three base categories include filtering, centrifuging, precipitation, and separation (physical); oxidation reduction and neutralization (chemical); and aerobic biological treatment and activated sludge processes (biological). Broadly speaking, physical technology corresponds to the *Preliminary* and *Primary* treatment in the U.S., while chemical and biological technologies correspond to the *Secondary* and *Tertiary* treatment.

for firms with physical technologies. Meanwhile, the median installation costs of biological technologies are 8 times higher than those of physical technologies. In addition, the median output of the firms using biological technologies are 4 times higher than those of firms with physical technologies. These point to increasing returns to scale in the adoption of cleaner technologies. As a result, small firms lack the profit margins needed to take advantage of these increasing returns to scale, while large firms are more likely to adopt cleaner technologies.

To see the statistical relationship between firm size and technology adoption, we pool all five polluting industries together and estimate a linear probability model of treatment technology adoption with the same set of covariates as in Equation (1):

$$y_i = \underset{(0.04)}{0.23} + \underset{(0.001)}{0.05} \times \log(\text{Output}_i) + \mathbf{X}\gamma + \epsilon_i,$$

where y_i is a dummy variable which equals to 1 when a firm uses biological technologies. The estimates show that doubling a firm's output leads to a 5 percentage points increase in the adoption rate of biological technologies. This confirms our previous finding with the paper industry that larger firms are more likely to use cleaner technologies.

Emissions Generated During Production.—Another reason behind the negative correlation between pollution intensity and firm size is that large firms also generate less COD per output during the production stage. In particular, if we divide our sample into two groups according to whether a firm is using biological technologies or not and repeat regression (1) again, we find significantly negative correlations between COD per output and firm size within each group:

(2) Biological: $\log(\text{COD}_i) = \underset{(0.20)}{-4.37} + \underset{(0.01)}{0.67} \times \log(\text{Output}_i) + \mathbf{X}\gamma + \varepsilon_i,$

(3)   Others: $\log(\text{COD}_i) = \underset{(0.31)}{-3.41} + \underset{(0.01)}{0.64} \times \log(\text{Output}_i) + \mathbf{X}\gamma + X_e\gamma_e + \varepsilon_i,$

where X_e is the dummy for chemical equipment.[14]

Such within group negative correlations could be because that larger firms use cleaner or fewer inputs to produce the same amount of output. For one example, the paper pulp industry uses two different types of inputs: the bagasse and the wood. According to the *Handbook of Emission Coefficients* by the Chinese Academy of Sciences, the bagasse, which is mostly used by firms with an annual production of less than 100 kilotons, generates 140–180 kg of COD per ton of output, and the wood, which is mostly used by firms with an annual production of more than 100 kilotons, generates 30–55 kg of COD per ton of output. As another example, Bloom et al. (2010) find that better managed firms use less energy

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[14]In our model and quantitative analysis below, we assume two types of treatment technologies. This is why we divide firms into two groups here. Running regression (1) separately for firms using physical and chemical equipments yield similarly significantly negative correlations between COD per output and firm size for each technology. Online Appendix C.3 reports the results in detail.

to produce the same amount of output in the U.K. Unfortunately, we cannot test these hypotheses directly using our data. Instead, we focus on firms' decisions to adopt treatment technologies, and model the emissions reduction during the production stage in a reduced-form way based on Equations (2) and (3).

C. *Firm Size Distribution and Aggregate Pollution Intensity*

The negative correlation between pollution intensity and firm size suggests that firm size distribution plays a potentially important role in aggregate pollution intensity. To illustrate the point, we document facts on the firm size distribution in China and in the U.S., and compute the change in the aggregate pollution intensity if the firm size distribution for the top five polluting industries in China were replaced with that in the U.S. We choose the U.S. as the benchmark because it is a relatively less distorted economy, and China and the U.S. are both large economies with a complete set of industries so that we can find comparable industries in both economies.[15]

Since the SUSB does not report information on output, we focus on the shares of employment accounted for by firms of different sizes, since firm-level employment is strongly correlated with output. We use the *International Standard Industrial Classification of All Economic Activities, Revision 3.1* (ISIC Rev. 3.1) published by the United Nations to bridge the different industrial classification systems adopted by China (GB/T4574-2002) and the United States (NAICS 2002). Specifically, we use the crosswalks of GB/2002 at the four-digit level and those of NAICS/2002 at the six-digit level to the ISIC Rev. 3.1 issued respectively by the National Bureau of Statistics of China and the U.S. Census Bureau, respectively.

The left panel of Figure 2 depicts the firm size distribution for the top five polluting industries pooled together. We find that approximately 70% of employment in the United States is accounted for by firms with 400+ employees, while in China the number is only 20%.[16] Compared with the U.S., a much larger share of production is carried out by small and medium-sized firms in China. From the right panel of Figure 2, we find a similar pattern for the whole manufacturing sector.

An Accounting Exercise.—To see the potential importance of firm size distribution, we conduct the following accounting exercise. For the top five polluting industries in China, we fix the total output and the estimated firm size–pollution intensity relationship, and replace the firm size distributions with those in the U.S.

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[15]It is a commonly used strategy in the macro-development literature to take the U.S. economy as an undistorted benchmark. See Guner, Ventura and Xu (2008), Hsieh and Klenow (2009, 2014), Bartelsman, Haltiwanger and Scarpetta (2013), and Adamopoulos and Restuccia (2014), among others.

[16]In Online Appendix E, we compare the firm size distributions between China and the U.S. for each individual polluting industry. We find that while there is substantial heterogeneity across individual industries, large firms in China account for a much smaller employment shares than those in the U.S. for each industry.
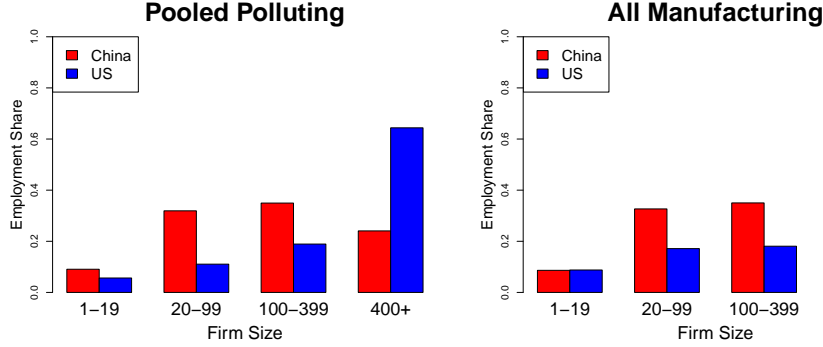
Figure 2. Employment Distributions

Sources: CNEC and SUSB.

Then we compute the resulting average pollution intensity for these polluting industries pooled together. This simple exercise is complicated by the fact that the NGSPS reports firm-level output instead of number of employees. Therefore, we construct an employment–output relationship from a linear regression using the CNEC data.[17]

We find that the average pollution intensity for the pooled top five polluting industries drops by 33%, when the firm size distributions in these industries are changed to those in the U.S. This would cause a 25% decrease in the average pollution intensity for the whole manufacturing sector if we leave other manufacturing industries untouched. Admittedly a crude approximation, this exercise nevertheless shows that distortions to firm size distribution could potentially have a large impact on aggregate pollution intensity.

II. The Model

The above accounting exercise is illuminating but has several limitations. First, the firm size distribution is mechanically changed to that in the U.S. Although the U.S. economy is often taken as a benchmark in the development literature, it is unclear whether the differences in firm size distribution between China and the U.S. are mainly caused by distortions or by some other factors. Second, the accounting exercise captures the technique channel only, while distortions may operate through the scale and composition effects as well. Third, the relationship between firm size and emissions intensity is taken as exogenous in the accounting exercise. However, distortions to firm size may also influence firm's incentive to

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[17]There are different methods to construct the employment–output relationship using the CNEC data, and all of them provide similar results. See Online Appendix F for details.

adopt different treatment technologies, which endogenizes the firm size-intensity relationship.

To address these issues, we build a two-sector (polluting and non-polluting) model of heterogeneous firms based on the classic Lucas (1978) span-of-control model.[18] The model features heterogeneity in firm productivities in both sectors, heterogeneity in pollution intensity and endogenous choice of pollution treatment technologies in the polluting sector. Motivated by the data, firms in both sectors are subject to correlated distortions [Restuccia and Rogerson (2008)] in the form of progressive output taxes. In the context of our model, the polluting sector is mapped to the top five polluting industries and the non-polluting sector to the rest of the manufacturing sector, while clean and dirty technologies are interpreted as biological and other (physical plus chemical) equipments as specified in Section I.B.

A. Setup

We assume that the products of the two sectors in the model economy are perfect substitutes to each other.[19] The polluting sector generates emissions as a by-product of production as in Copeland and Taylor (1994), while the non-polluting sector does not. The top five polluting industries (hence the polluting sector) account for nearly 80% of total industrial COD emissions and 20% of total industrial output, with the rest accounted for by the non-polluting sector. Firms in both sectors behave competitively.

Household.—There is a representative household with a continuum of members that maximizes the life-time utility $\sum_{t=0}^{\infty} \beta^t U(C_t)$, where C_t is the consumption at time t and β is the discount factor.

A fraction μ of the members are endowed with managerial talent that can be used in the polluting sector, drawn from a distribution with cumulative distribution function (CDF) $G^d(z)$, where the superscript d refers to the polluting sector. Meanwhile, a fraction $1 - \mu$ of the members are endowed with managerial talent for the non-polluting sector with CDF $G^c(z)$, where the superscript c refers to the non-polluting sector. The support for both distributions are $Z \triangleq [0, \bar{z}]$. We assume that z is fixed once drawn. Household members face an occupational choice between worker and manager. A worker supplies one unit of labor inelastically in exchange for wage income, and is perfectly mobile between sectors, while a manager in each sector runs a neoclassical firm and earns profits. As we see in

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[18]A widely used alternative in the misallocation literature is the closed-economy version of Melitz (2003). Hsieh and Klenow (2009) establish that these two frameworks are isomorphic for the effects of misallocation on aggregate output when the number of firms is fixed. In Appendix G.3, we show the qualitative equivalence between the two families of models.

[19]In Online Appendix J, we consider the case in which the final good is a CES aggregation of the goods produced in two sectors, and conduct counterfactual experiments under different values of the elasticity of substitution in the CES function. For the counterfactual experiments we discuss in the main text, the quantitative results in the CES cases are similar to those in the perfect substitute case, largely because the experiments induce very limited reallocation between sectors.

Figure 2, the firm size distributions in the polluting sector and the whole manufacturing sector are almost identical. Based on this observation, we assume that $G^d(z) = G^c(z) \triangleq G(z)$ from now on.

In addition, we assume that the household owns all the firms and capital in this economy. The goods produced in the non-polluting sector are taken as the numeraire. With perfect substitution between goods, the price of the goods produced in the polluting sector is also 1. R and W are the capital rental price and wage rate, respectively.

Firms.—The production functions in the two sectors take the same form. Production requires managerial talent z, capital k, and labor l. For a firm run by a manager with managerial talent z, its output is

$$y = F(z, k, l) = z^{1-\gamma}(k^\alpha l^{1-\alpha})^\gamma,$$

where $0 < \gamma < 1$ is the span-of-control parameter, which governs the returns to scale at the firm level. Due to diminishing returns to scale at the firm level, a non-degenerate distribution of firm size can be supported, and firms earn positive profits as the payment to managerial talent z. As is typically the case with studies that follow Lucas (1978), managerial talent z is the only source of firm-level TFP in this model. In what follows, we also refer to z as firm-level TFP.

In the non-polluting sector, firms do not discharge pollutants and are not subject to environmental regulations. In the polluting sector, pollutants e are generated as a by-product of production. Total emissions $e = \mathcal{E}(j, y)$ depend on the amount of output y, and the pollution treatment technology firms use, with $j = 1$ refers to the clean technology and $j = 0$ the dirty technology. Based on the empirical findings from Section I, we assume that

$$(4) \qquad\qquad \log\left(\frac{e}{y}\right) = \psi_0^{(j)} + \psi_1^{(j)} \log y, \quad j = 0, 1,$$

where $\psi_1^{(j)} < 0$. This formulation indicates that larger firms emit fewer pollutants per unit of output, which is consistent with our empirical findings in Section I.B.

Adopting the clean technology incurs fixed cost Rk_E. We assume that the environmental authority punishes firms for using the dirty technology. Similar in spirit to Shapiro and Walker (2018), we model the environmental regulations in a parsimonious way as a penalty which equals to a proportion ξ of the firm's profits. The penalty is later redistributed to the household as lump-sum transfers.[20] Therefore, the central trade-off in a firm's technology choice is weighing

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[20]In practice, China employs both economic incentives and command-and-control instruments in its current industrial pollution management and control system. A powerful tool used by the Chinese government is to randomly audit and then shut down firms that do not comply with environmental regulations for a certain period of time, which would cost these firms a fraction of their profits. See Trevor Nace, "China shuts down tens of thousands of factories in widespread pollution crackdown," Forbes, October 24th 2017, for an anecdotal example. We present a short summary of China's current

the adoption cost Rk_E against the profit losses from the environmental penalties. To the extent that firms' profits increase with firm size, we implicitly capture the size-dependency of the environmental regulations, in the sense that larger firms are usually monitored and inspected more closely. A key prediction from these assumptions is that, only firms above a certain size threshold would choose the clean technology, which is qualitatively consistent with our empirical findings on technology adoption in Section I.B.[21]

Correlated Distortions.—Finally, we introduce the correlated distortions at the firm level to both sectors as implicit progressive output taxes which increase with a firm's productivity levels. The functional form of the tax is:

$$(5) \qquad \tau_z^i = \max\left\{0, 1 - \phi_0^i z^{\phi_1^i}\right\}, i = c, d,$$

where the parameter ϕ_0^i determines the mean level of the taxes, and the parameter ϕ_1^i determines the progressivity of the taxes. When $\phi_1^i < 0$, the tax rates are increasing in firm TFP z, and the progressivity of the taxes is decreasing in the value of ϕ_1^i.[22]

The tax function (5) is consistent with the positive relationship between firm TFP and average product of factors we find in the CNEC data. This can be seen from Figure 3, which depicts the geometric average of the average product of labor and capital $(y/k)^\alpha (y/l)^{1-\alpha}$ against firm TFP z for both the polluting and non-polluting sectors.[23] It can also be seen that the relationships between the average product of factors and firm TFP in these two sectors share very similar slopes. Based on this observation, we assume $\tau_z^c = \tau_z^d \triangleq \tau_z$ from now on.

The firm-level distortions τ_z are meant to capture a variety of policies and institutions that reallocate production factors from large productive to small unproductive firms, following the *indirect* approach in the misallocation literature [Restuccia and Rogerson (2013)]. These policies and institutions include size-dependent taxes and subsidies that favor small firms, imperfect tax enforcement that induces tax evasion among small firms, financial constraints and contrac-

environmental regulations based on Lin (2013) in Online Appendix D.3.

[21]In Online Appendix D.2, we provide further evidence that the ratio of a firm's expenditures on treatment technology to firm output decreases with firm size, which supports the fixed cost assumption.

[22]This tax function has a long history in macro-public finance literature, which dates back to at least Bénabou (2002). We thank a referee for pointing this out. Recent studies that use this function include Hsieh and Klenow (2014), Bento and Restuccia (2017) and Guner, Parkhomenko and Ventura (2018). In Online Appendix D.1, we discuss distortions that take more general forms, and present empirical evidence in support of this specific form. See Figure D.2 in Online Appendix D for an illustration of this tax function.

[23]To estimate firm TFP and average product of factors, we need to assign the values to the parameter that determines the income share of capital, α, and the span-of-control parameter, γ. We take these values from our calibration in Section III, which implies that $\alpha = 0.54$ and $\gamma = 0.93$. The positive relationship between firm TFP and average product of factors, however, hold for a wide range of values for α and γ, in addition to the values used here. Moreover, similar patterns are also observed in other countries. For instance, Hsieh and Klenow (2014) find a strong and positive relationship between firm TFP and average product of factors in India and Mexico, while Bento and Restuccia (2017) document similar patterns for dozens of developing countries.
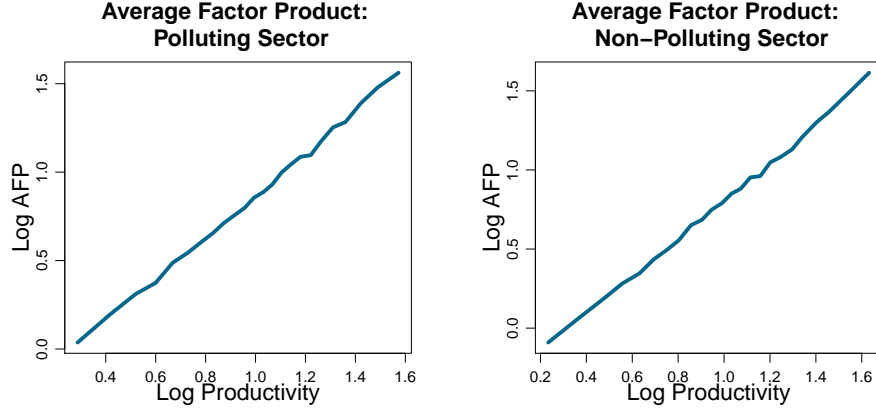
Average Factor Product: Polluting Sector

Average Factor Product: Non–Polluting Sector

FIGURE 3. AVERAGE PRODUCT OF FACTORS VERSUS FIRM PRODUCTIVITY IN THE CROSS-SECTION

Source: CNEC.

Note: Both variables have been normalized to reflect the relative values to the sectoral averages, which are similar across sectors.

tual frictions that impede the growth of productive firms [Cooley and Quadrini (2001), Clementi and Hopenhayn (2006) and Akcigit, Alp and Peters (2018)], and local protectionism and internal trade barriers that disproportionately affect large productive firms [Eberhardt, Wang and Yu (2016) and Tombe and Zhu (2019)]. While it is of great policy interest to study specific policies and institutions underlying τ_z, we leave this important task to future work.

B. *Firm's Profit Maximization Problem*

Firms in the non-polluting sector solve the standard profit maximization problem:

$$\pi^c(z) = \max_{k,l} \left\{ (1 - \tau_z)z^{1-\gamma}(k^{\alpha}l^{1-\alpha})^{\gamma} - Wl - Rk \right\},$$

where the superscript c refers to the non-polluting sector.

Firms in the polluting sector choose first the optimal treatment technology and then the optimal production plan. The profit function $\pi^d(z)$ is the envelope of the profits when using the dirty technology $\pi_0^d(z)$ and clean technology, $\pi_1^d(z)$:

$$\pi^d(z) = \max \left\{ \pi_0^d(z), \pi_1^d(z) \right\},$$

where the superscript d refers to the polluting sector, and the subscripts denotes the treatment technologies. Firms using the clean technology are not subject to

environmental penalties, hence the profit $\pi_1^d(z)$ is:

$$\pi_1^d(z) = \max_{k,l} \left\{ (1 - \tau_z)z^{1-\gamma}(k^\alpha l^{1-\alpha})^\gamma - Wl - R(k + k_E) \right\}.$$

However, when a firm chooses the dirty technology, it will be punished by the environmental authority, and lose a fraction ξ of its gross profit. Its net profit $\pi_0^d(z)$ is thus:

$$\pi_0^d(z) = \max_{k,l} \left\{ (1 - \xi) \left[(1 - \tau_z)z^{1-\gamma}(k^\alpha l^{1-\alpha})^\gamma - Wl - Rk \right] \right\}.$$

The firm's problem implies that there exists a firm size threshold $\tilde{l}^d$ above which firm will choose the clean technology, which is consistent with our empirical finding that larger firms are more likely to use the clean technology. This can be seen by comparing the potential profits from using clean versus dirty technology:

$$\pi_1^d(z) - \pi_0^d(z) = \frac{\xi(1-\gamma)}{(1-\alpha)\gamma}Wl^d(z) - Rk_E.$$

In addition, when $1 - \gamma + \phi_1 > 0$, which is satisfied given our calibrated values for γ and ϕ_1, $\pi_i^d(z)$ and $l^d(z)$ are both increasing z. Then, as we show in Lemma 2 of Online Appendix G.1, there exists a threshold $\tilde{z}_d$ such that only firms with TFP higher than $\tilde{z}_d$ would choose clean technology. There also exist two thresholds $\hat{z}_i, i = c, d$, such that in sector i, only members with $z > \hat{z}_i$ choose to be managers.

C. Stationary Equilibrium

In this section, we close the model by specifying the household's optimization problem, and then define the stationary equilibrium.

The household solves the standard consumption-saving problem:

(6)
$$\max_{C_t, K_{t+1}} \sum_{t=0}^{\infty} \beta^t U(C_t)$$

$$s.t.$$

$$C_t + K_{t+1} - (1 - \delta)K_t = I_t,$$

where K_t is aggregate capital, δ is the depreciation rate, and I_t is household income, which we specify in detail shortly.[24] The solution to (6) is the standard

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[24]Here we do not explicitly model the household's preference over water quality. Adding water quality to the household's utility function will not change the firm's choices in a competitive equilibrium, as no individual household member has any control over the aggregate environmental quality. However, if we study the social planner's problem instead, the household's preference over water quality will become a major modeling choice.

intertemporal Euler equation:

$$U'(C_t) = \beta U'(C_{t+1})(1 + R_{t+1} - \delta),$$

which pins down the equilibrium interest rate.

Household income I_t comes from three sources: wage and interest income, firms' profits, and lump-sum transfers from taxes τ_z and environmental penalties ξ. If we let T denote the transfers, then household income is[25]

$$I = RK + W[\mu G(\hat{z}_d) + (1 - \mu)G(\hat{z}_c)] + T$$
$$+ \mu \left[\int_{\hat{z}_d}^{\tilde{z}_d} \pi_0^d(z)dG(z) + \int_{\tilde{z}_d}^{\overline{z}} \pi_1^d(z)dG(z) \right] + (1 - \mu) \int_{\hat{z}_c}^{\overline{z}} \pi^c(z)dG(z),$$

where the transfer T is given by

$$T = \mu \xi \int_{\hat{z}_d}^{\tilde{z}_d} \pi^d(z)dG(z) + \tau_z \left[\mu \int_{\hat{z}_d}^{\overline{z}} \pi^d(z)dG(z) + (1 - \mu) \int_{\hat{z}_c}^{\overline{z}} \pi^c(z)dG(z) \right].$$

Let Y be aggregate output and E be aggregate pollution, then the stationary equilibrium of the model is defined as follows.

DEFINITION 1:   *A stationary equilibrium of the model consists of prices* $\{W, R\}$, *allocations* $\{C, K, Y\}$, *firms' policy functions* $\{k^j(z), l^j(z), y^j(z), \pi^j(z)\}, j = c, d$, *thresholds for household's occupational choices* $\{\hat{z}_c, \hat{z}_d\}$ *and polluting firms' technology adoption* $\tilde{z}_d$, *as well as aggregate pollution* E, *such that:*

(i) *Given factor prices* $\{W, R\}$, $\{C, K, \hat{z}_c, \hat{z}_d\}$ *solve the household's optimization problem;*

(ii) *Given factor prices* $\{W, R\}$, $\{k^j(z), l^j(z), y^j(z), \pi^j(z)\}, j = c, d$, *and* $\tilde{z}_d$ *solve the firms' optimization problems;*

(iii) *Factor prices* $\{W, R\}$ *clear all markets:*

- *Labor Market:*

$$\mu G(\hat{z}_d) + (1 - \mu)G(\hat{z}_c) = \mu \int_{\hat{z}_d}^{\overline{z}} l^d(z)dG(z) + (1 - \mu) \int_{\hat{z}_c}^{\overline{z}} l^c(z)dG(z),$$

- *Capital Market:*

$$K = \mu \left[\int_{\hat{z}_d}^{\overline{z}} k^d(z)dG(z) + k_E[G(\overline{z}) - G(\tilde{z}_d)] \right] + (1 - \mu) \int_{\hat{z}_c}^{\overline{z}} k^c(z)dG(z),$$

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[25]We have implicitly assumed that firms using clean and dirty technologies coexist in the equilibrium, meaning that $\hat{z}_d < \tilde{z}_d$. This is also the case in our quantitative exercises later.

- *Product Market (Walras' Law):*

$$C + K - (1 - \delta)K = \mu \int_{\hat{z}_d}^{\overline{z}} y^d(z) dG(z) + (1 - \mu) \int_{\hat{z}_c}^{\overline{z}} y^c(z) dG(z);$$

(iv) *Aggregate Pollution:*

$$E = \mu \left[\int_{\hat{z}_d}^{\tilde{z}_d} e\left(0, y^d(z)\right) dG(z) + \int_{\tilde{z}_d}^{\overline{z}} e\left(1, y^d(z)\right) dG(z) \right].$$

The algorithm to compute the stationary equilibrium is sketched in Online Appendix I.

D. *Correlated Distortions, Technology Adoption, and Aggregate Pollution*

In this section, we show qualitatively how correlated distortions modeled as Equation (5) affect the adoption of clean technology across firms and the aggregate pollution. First, we show that correlated distortions have a negative impact on firms' incentives to adopt the clean technology. Second, using a simplified one-sector model with a fixed number of firms and exogenous firm-level pollution intensity, we characterize the conditions under which correlated distortions lead to higher aggregate pollution. The core trade-off here is between the scale and technique effects, because the correlated distortions generally cause a decrease in average pollution intensity but an increase in aggregate output. All the proofs and related analytical results can be found in Appendices G.1 and G.2.

Technology Adoption.—Proposition 1 shows that correlated distortions discourage firms from adopting of clean technology, especially those with medium levels of TFP.

PROPOSITION 1:   *When $1 - \gamma + \phi_1 > 0$, a firm chooses to install clean technology only when its productivity is higher than a threshold. Moreover, the threshold when the firm faces distortions $\tilde{z}_f$ is higher than that when the firm is not subject to any distortions $\tilde{z}_n$.*

Figure 4 illustrates the intuition behind Proposition 1. Recall that while the adoption cost Rk_E is independent of z, the penalty for not using the clean technology $\xi\pi_0(z)$ is increasing in z. Therefore, the adoption of the clean technology pays off for firms with higher z. On the other hand, the elasticity of profits to productivity z is $1 - \gamma$ when there are no distortions, and is $1 - \gamma + \phi_1$ when distortions exist. Since $\phi_1 < 0$, the slope of the profit functions are flatter, and the threshold where firms switch technologies is higher when they face distortions. As a result, while firms with TFP in the range $[\tilde{z}_n, \tilde{z}_f]$ would choose the clean technology when they are not subject to distortions and choose the dirty technology when they are.[26]

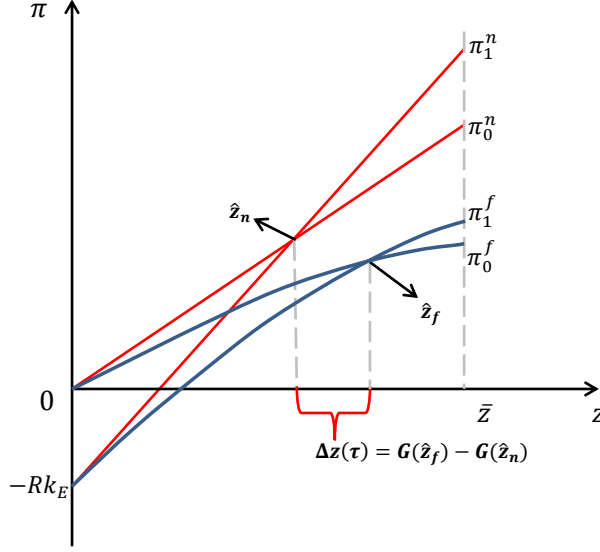[26]Of course, when the correlated distortions apply to all firms in the economy, they would affect the

FIGURE 4. CORRELATED DISTORTIONS AND TREATMENT TECHNOLOGIES

Note: There are four profit functions in the figure, $\pi_0^n, \pi_0^f, \pi_1^n$, and π_1^f. The superscripts n and f denote whether firms face correlated distortions in the economy or not, and the subscripts 0 and 1 denote whether firms choose the dirty or clean technology, respectively.

Aggregate Pollution.—To derive a closed-form expression for the relationship between correlated distortions and aggregate pollution, we build a simplified one-sector model that abstracts from the endogenous occupation choice and treatment technology choice. This complements our analysis in Proposition 1, which focuses on how distortions affect firms' decision in that the general equilibrium effects are captured. Proposition 2 summarizes our main findings.

PROPOSITION 2:   *In the simplified model, there exists a threshold*

$$\hat{\phi}_1 = \frac{\psi_1(1-\gamma)}{1-\gamma(1+\psi_1)},$$

such that E is increasing in ϕ_1 when $\phi_1 < \hat{\phi}_1$, and decreasing when $\phi_1 \geq \hat{\phi}_1$.

Proposition 2 says that the relationship between E and the progressivity of the distortions $|\phi_1|$ is inverted-U shaped with the turning point $|\hat{\phi}_1|$ increasing in the elasticity of emission intensity to output $|\psi_1|$.[27] Because correlated distortions generally decreases the total output while increasing the average pollution inten-

equilibrium wage W and hence $\tilde{z}_f$ in equilibrium. We find later in our quantitative analysis that $\tilde{z}_f > \tilde{z}_n$ even when general equilibrium effects are considered.

[27]Since ϕ_1 and ψ_1 are both negative, the progressivity and elasticity are stronger when $|\phi_1|$ and $|\psi_1|$ are larger, meaning that the raw values ϕ_1 and ψ_1 are smaller.

sity, the intuition behind Proposition 2 is a race between the scale and technique effects. In particular, the scale effect is larger when the progressivity of the distortions is stronger (a smaller negative ϕ_1), while the technique effect is larger when the elasticity of emission intensity to output is larger (a smaller negative $\psi_1 < 0$). As a result, when $|\phi_1|$ is not too large relative to $|\psi_1|$, the technique effect dominates. Hence an increase in ϕ_1 (which means a decrease in the progressivity) causes higher aggregate pollution E. Otherwise, the scale effect dominates, and an increase in ϕ_1 reduces E. The average distortions ϕ_0 (which is positive) always reduces E, however, since it does not affect the firm size distribution, and hence only has the scale effect. The overall effect of the correlated distortions is thus a combination of those from ϕ_0 and ϕ_1.

III. Calibration

In this section, we calibrate our model to the Chinese data. We first describe our key calibration strategy, and then explain how our choice of the model parameters matches the key aspects of the Chinese economy at both the aggregate and firm levels.

A. Calibration Strategy

The main task of our quantitative analysis is to study how correlated distortions affect aggregate pollution by changing the firm size distribution. Therefore, it is important that our model matches well the pollution intensity by firm size, the relative distortions faced by large versus small firms, and the firm size distribution in China. As we mentioned in the previous section, the polluting sector is mapped to the top five polluting industries and the non-polluting sector to the rest of the manufacturing sector, while clean and dirty technologies are interpreted as biological and other (physical plus chemical) technologies.

Following the evidence in Section I.B, the relationship between pollution intensity and firm size is given by Equation (4), and the emissions by firms using different treatment technologies are

$$e = \mathcal{E}(j, y) = \mathrm{e}^{\psi_0^{(j)}} y^{1 + \psi_1^{(j)}}, j = 0, 1,$$

where $\{\psi_0^{(j)}, \psi_1^{(j)}\}$ are based on the reduced-form estimates from Equations (2) and (3).

The firm size distribution in the model is determined by both the talent distribution $G(\cdot)$ and the progressivity of distortions $-\phi_1$. This can be seen from Equation (7), derived from firms' first order conditions:

$$(7) \qquad\qquad (1 - \tau_z) z^{1 - \gamma} = \Phi l^{1 - \gamma},$$

where Φ is a constant. As a result, the progressivity of the distortions ϕ_1 has to be calibrated together with the talent distribution $G(\cdot)$. To do this, if we let $\theta_l = y/l$ be the average product of labor, the firm's first order conditions imply

$$\theta_l = \frac{y}{l} = \frac{W}{(1-\alpha)\gamma(1-\tau_z)}.$$

Substituting the tax function (5) into the above equation, we have

$$(8) \qquad \left(\frac{z_i}{z_j}\right)^{\phi_1} = \frac{\theta_{l,j}}{\theta_{l,i}}.$$

Equation (8) can then be used to identify ϕ_1 using the variations in average product of labor of firms with different TFP.

To calibrate $G(\cdot)$, we cannot use Equation (7) and the firm-level data to back out the distribution semi-parametrically. The reason is that the endogenous occupation choice in our model makes the distribution of firm TFP in equilibrium also endogenous, making the semi-parametric way of calibration untractable in practice. This is unlike other studies that assume a fixed distribution of firm TFP in equilibrium, for instance Bento and Restuccia (2017). To keep the calibration process tractable, motivated by Equation (7), we assume that the distribution of the "after-tax" TFP $z' = (1 - \tau_z)z^{1-\gamma}$ takes the following form: the majority on the bottom side characterized by a truncated log-normal distribution with log mean μ and standard deviation σ, and a top value z'_{max} with a total mass g_{max} representing the large firms. This mass of top managers allows our model to match the observed concentration of employment among large firms, which have lower emission intensities than other firms.[28] We then choose the values of $(\mu, \sigma, z'_{max}, g_{max})$ together with γ and ϕ_1 to match the observed firm size distribution and Equation (8). The true firm-level TFP z and its distribution $G(z)$ can be backed out by

$$z = \left[z'\phi_0^{1/(\gamma-1)}\right]^{\frac{1-\gamma}{1-\gamma+\phi_1}},$$

given a choice of ϕ_0.

B.   *The Choice of Parameter Values*

In total, we need to choose values for 17 parameters: the discount factor β, depreciation rate δ, production technology parameters $\{\alpha, \gamma\}$, the relative size of the polluting sector μ, treatment technology parameters $\{\psi_0^{(0)}, \psi_1^{(0)}, \psi_0^{(1)}, \psi_1^{(1)}, k_E, \xi\}$, correlated distortions $\{\phi_0, \phi_1\}$, and distributional parameters $\{\mu_z, \sigma_z, z'_{max}, g_{max}\}$.

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[28]This strategy is popular among macroeconomic studies on distributions with fat tails. See for example, Castañeda, Díaz-Giménez and Ríos-Rull (2003) and Guner, Ventura and Xu (2008).

Six of the seventeen parameters can be determined exogenously. We set the depreciation rate δ to 10% [Song, Storesletten and Zilibotti (2011)].
$\{\psi_0^{(0)}, \psi_1^{(0)}, \psi_0^{(1)}, \psi_1^{(1)}\}$ are taken from the reduced-form estimates from regressions (2) and (3) in Section I.B, giving us $\psi_0^{(0)} = -3.41$, $\psi_1^{(0)} = -0.36$, $\psi_0^{(1)} = -4.37$, and $\psi_1^{(1)} = -0.33$. In the context of our model, these estimates suggest that, on average, for two firms with the same level of output but different treatment technologies, the firm using the clean technology emits 40% to 60% less pollutants than the one with the dirty technology. We set $\mu = 0.20$, which is the share of manufacturing firms accounted for by the five polluting industries in the CNEC data. The share of manufacturing output accounted for by these five industries is also quite close, which is 18%.

The remaining parameters have to be calibrated jointly due to general equilibrium feedback. Still, some parameters can be calibrated more or less independently from others, which substantially improves the tractability of the calibration process. In particular, we choose β and α to match the capital to output ratio of 1.65, and capital share of 0.5 in China respectively [Bai, Hsieh and Qian (2006)]. We set ξ and k_E such that the model reflects the average cost to install clean technology in the data. For the five polluting industries, we find in the NGSPS that 57% of the firms use biological technology, and the average adoption cost of biological equipments is approximately 2.5% of the average firm output for these firms. These two targets help us pin down the values of ξ and k_E.

We then calibrate $\{\gamma, \mu_z, \sigma_z, z'_{max}, g_{max}, \phi_1\}$ such that the model generates a firm size distribution comparable to the data and Equation (8) holds. To link Equation (8) to the data, we set $\{z_i, z_j\}$ and $\{\theta_{l,i}, \theta_{l,j}\}$ to their average levels in the top and bottom quintiles respectively. Because the calculation of z depends on γ, Equation (8) also defines a relationship between ϕ_1 and γ. For the size distribution, we focus on the number of firms and the share of employment in each of the five size groups (Figure 5). In practice, we adjust $\{\gamma, \mu_z, \sigma_z, z'_{max}, g_{max}\}$. For each γ, we compute the corresponding ϕ_1 using Equation (8). We then compute the size distributions in the equilibrium and repeat the process for different combinations of the five parameters until the size distributions in the equilibrium are close to those in the data.

Finally, following Bento and Restuccia (2017), we calibrate ϕ_0 such that the average tax burden in the economy matches the value added tax rate Chinese manufacturing firms face, which is 13%. This gives us the value of $\phi_0 = 1.15$.[29] Table 3 lists the model parameters, their targets, and the calibrated values.

Discussions.—Given the calibrated values of the parameters, our model matches well the key aspects of the Chinese economy at both the aggregate and firm levels.

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[29]In general, it is difficult to estimate the average distortions in an economy, because many factors that lead to resource misallocation are either poorly measured or not measured at all. Reassuringly, a change in ϕ_0 causes a uniform change in firm-level distortions, and therefore has very limited impact on firm size distribution and aggregate pollution intensity. As a result, a different calibration target for ϕ_0 will not affect the main insight of our paper.

Table 3—Calibration

Parameter		Value	Target
Production	δ	0.1000	Depreciation Rate
	α	0.5376	Capital Share (0.5)
	γ	0.9300	Size Distribution†
Treatment	$\psi_0^{(0)}$	-3.4144	Estimates from Regression (2)
	$\psi_1^{(0)}$	-0.3636	
	$\psi_0^{(1)}$	-4.3748	Estimates from Regression (3)
	$\psi_1^{(1)}$	-0.3288	
	k_E	4.6000	Clean Firms: Fixed Cost/Output Ratio (2.5%)
	ξ	0.2300	Clean Technology Adoption Rate
Distortions	ϕ_0	1.15	Average Value Added Tax Rate(13%)
	ϕ_1	-0.03	Elasticity of Average Product to TFP Eq. (8)†
Preference	β	0.8750	Capital/output Ratio (1.65)
Productivity	μ_z	-2.4567	Size Distribution†
	σ_z	4.0020	
	z'_{max}	24855	
	g_{max}	0.00048	
	μ	0.20	Fraction of Firms in the Polluting Sector

Note: †The parameters are jointly calibrated.

Figure 5 plots the distributions of the fractions (left panel) and employment shares (right panel) of firms with different sizes for the whole manufacturing sector in the model and in the data.[30] Overall, the model matches the two distributions well. The mean and median of the firm size distribution 53.74 and 19.56, which are not directly targeted in the calibration, match their empirical counterparts of 52.19 and 20 respectively, suggesting that the top managers captures well the upper tail of the firm size distribution.

The calibrated value of the span-of-control parameter γ, which determines the firm-level returns to scale, lies close to the upper end of the range identified by the literature.[31] A relatively high value of γ is needed to match the employment share of large firms (in particular those with 400+ employees) given the existence of correlated distortions, because we need high returns to scale so that the right amount of resources will be concentrated among firms with high TFP. However, if γ is too large, the high demand for factors from the large firms would push up the factor prices, and therefore drive many small firms with low TFP out of the market through endogenous occupation choices. This would jeopardize our model's fit of the fraction of small firms to the data. Our choice of $\gamma = 0.93$ is a compromise of the two countervailing forces.

[30]Since the distributions of managerial talent are the same in the polluting and non-polluting sectors, and the calibrated value of k_E is not too large, the firm size distributions for the two sectors in the model are close to each other, which is consistent with data patterns in Figure 2.

[31]The values of firm-level returns to scale used in the macro literature range from 0.80 [Guner, Ventura and Xu (2008)] to 0.95 [Bartelsman, Haltiwanger and Scarpetta (2013)]. Estimates from micro studies differ across industries. For example, Olley and Pakes (1996) estimate the value to be between 0.8 and 0.9 for the U.S. telecommunications equipment industry, while the estimates from Baily et al. (1992) point towards close to constant returns to scale for many manufacturing industries in the U.S.. See also the discussions in Atkeson and Kehoe (2005).
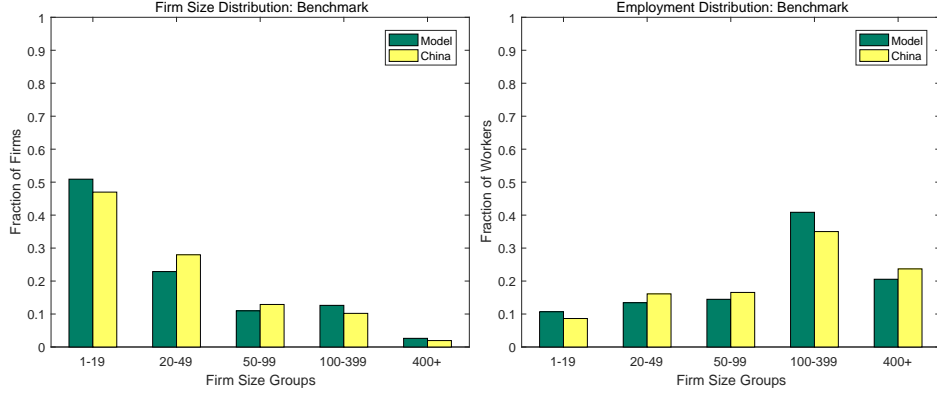
FIGURE 5. FIRM SIZE DISTRIBUTION: MODEL VERSUS DATA

Note: The firm size and employment distributions of the whole manufacturing sector are reported.

IV. Quantitative Exercises

In the section, we use our calibrated model to evaluate the effects of alternative policies on aggregate output and pollution. We conduct two sets of counter-factual experiments. In Experiment (i), we follow the *indirect* approach in the misallocation literature and remove correlated distortions from both sectors by changing τ_z to 0.[32] This is meant to capture the overall effects of a variety of policies and institutions that reallocate resources from high to low TFP firms. To isolate the importance of the progressivity of distortions, later we replace the progressive taxes with flat taxes by setting $\phi_1 = 0$, and change ϕ_0 to keep the total implicit tax revenue constant. In Experiment (ii), we tighten the environ-mental regulations by increasing ξ to the extent that the fraction of firms using the clean technology is the same as in Experiment (i). We then compare the results from these two experiments, and illustrate the key mechanisms that drive these different results.

A. Counterfactual Experiments

Removing the Correlated Distortions.—In Experiment (i), we remove correlated distortions from both sectors by setting $\tau_z = 0$ for all z. Column (i) of Table 4 summarizes the results. Overall, removing correlated distortions from both sec-

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[32]In Online Appendix I, we build and calibrate a similar model in which the final consumption is a CES aggregation of the products from two sectors, and use the model to evaluate the impact of removing the correlated distortions from the polluting sector only. The main difference between this experiment and Experiment (i) in the main text is that this one induces large reallocation of factors from the non-polluting to the polluting sector. As a result, the reduction in aggregate pollution from removing distortions will be smaller.

TABLE 4—THE EFFECTS OF CORRELATED DISTORTIONS AND ENVIRONMENTAL REGULATIONS

Statistics	Polluting			Non-polluting		
	Benchmark	(i)	(ii)	Benchmark	(i)	(ii)
Output	100.00	131.16	98.99	100.00	129.62	100.28
Capital	100.00	163.06	99.00	100.00	161.26	100.28
Consumption	100.00	123.63	100.02	100.00	123.63	100.02
Wage	100.00	160.00	99.96	100.00	160.00	99.96
Output per Worker	100.00	128.25	100.15	100.00	128.60	99.95
Output per Firm	100.00	297.62	110.17	100.00	311.85	100.28
Average Productivity	100.00	221.24	111.11	100.00	235.39	100.00
% in Output	19.73	19.91	19.52	80.27	80.09	80.48
# of Firms	100.00	44.07	89.86	100.00	41.56	100.00
Mean Size	59.98	139.19	65.98	52.18	126.53	52.35
Median Size	23.67	43.09	27.88	18.61	34.31	18.67
Pollution	100	76.67	85.82			
Intensity	100	58.46	86.73			
Clean Share	57.77	85.61	85.10			
Regulation	23.00	23.00	35.50			

Note: All values are percentages except for mean size and median size, which are the numbers of workers. Consumption and wage are not sector specific.

TABLE 5—RESOURCE ALLOCATION ON THE INTENSIVE MARGIN

Economy	QU_1	QU_2	QU_3	QU_4	QU_5
Polluting Sector:					
Benchmark	2.79	4.26	7.45	16.81	68.70
Case (i)	1.53	2.89	6.46	18.25	70.87
Case (ii)	3.02	4.63	8.01	17.79	66.55
Non-polluting Sector:					
Benchmark	2.42	3.78	6.65	15.39	71.75
Case (i)	1.30	2.49	5.68	16.97	73.57
Case (ii)	2.42	3.78	6.65	15.39	71.75

Note: QU_1 to QU_5 represent the first to fifth quintiles, respectively.

tors increases aggregate output by about 30% in both sectors (scale effect), while lowering the pollution intensity of the polluting sector by 42% (technique effect). Since both sectors share the same distributions of firm-level productivity and distortions, removing correlated distortions induces little reallocation across sectors (composition effect). Therefore, the technique effect from eliminating correlated distortions dominates the scale and composition effects, and aggregate pollution drops by 23%. The decrease in aggregate pollution is consistent with Proposition 2, given the calibrated values for (ϕ_1, γ), and the estimated relationship between pollution intensity and firm size.

Removing correlated distortions causes large reallocation of production factors across firms, and leads to improved allocation along both the extensive and intensive margins. The *extensive* margin is defined as the selection of firms into production, and the *intensive* margin as the production distribution among the active firms. Specifically, since the distortions τ_z are higher for high TFP firms in the benchmark economy, a large share of the production factors would flow
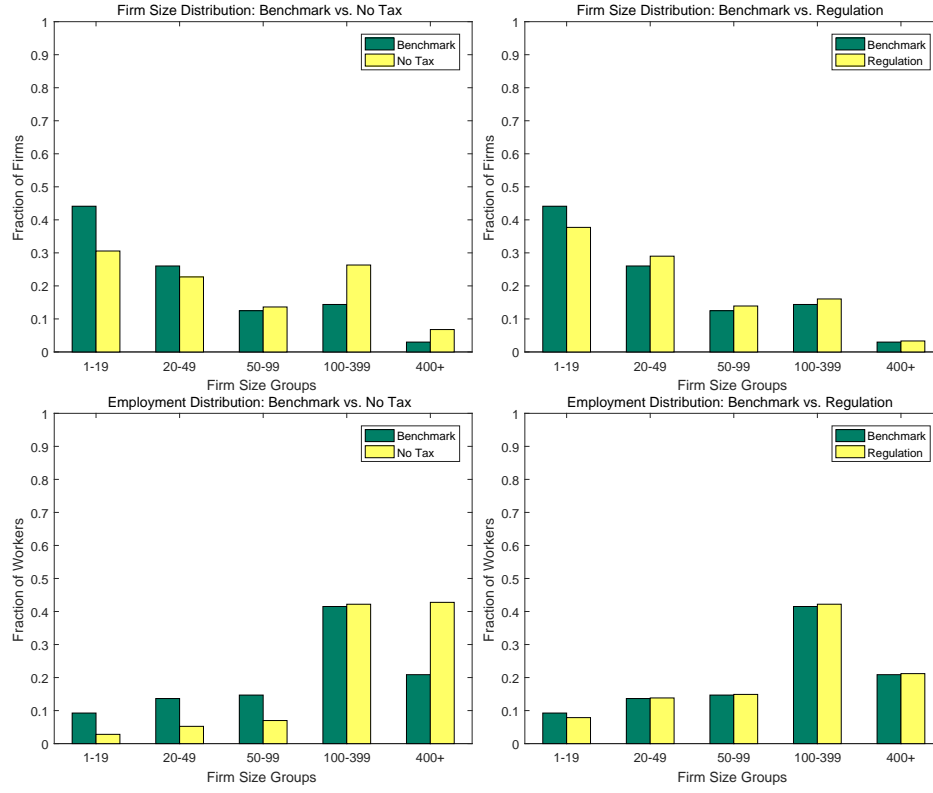
FIGURE 6. THE EFFECTS OF DISTORTIONS AND REGULATIONS ON FIRM SIZE DISTRIBUTIONS

Note: The distributions of the polluting sector are reported.

to these firms after the removal of distortions (intensive margin). The increased labor demand from these firms would drive up the equilibrium wage by 60%, which would in turn drive many unproductive firms out of the market through the endogenous occupation choices (extensive margin). Due to the changes along the extensive margin, the total number of firms is cut by more than half, and the mean firm size more than doubles. Along the intensive margin, as we see from Table 5, surviving firms with TFP in the top two quintiles expand at the expense of those with lower TFP. The two panels on the left of Figure 6 show the changes in firm size distributions following the removal of the correlated distortions.

The large increase in aggregate output is caused by both an improved allocation of production factors across firms, and a large increase in capital accumulation by about 60% due to decreased average tax rate from 13% to 0%. The increase in capital accumulation plays a much larger role than improved allocation of factors in the increase in aggregate output. Our quantitative exercises in Section IV.B below illustrate this point further. The large decline in average pollution intensity

is also caused by two factors: the improved allocation of production factors, and increased adoption of the clean technology. After the removal of correlated distortions, more production factors move to large productive firms which also have lower pollution intensity. This reallocation lowers the average pollution intensity. Meanwhile, as we discuss in Section III.D, correlated distortions lower the elasticity of firm profits to firm TFP z, and therefore reduce their incentives to adopt the clean technology. Removing correlated distortions would do exactly the opposite and encourage the adoption of clean technology.

To evaluate the relative contributions of the improved allocation of factors and improved adoption of clean technologies, we consider a hypothetical economy that is identical to our benchmark economy, except that the clean technology shows no improvement in abatement efficiency over the dirty technology, that is, $\psi_0^{(1)} = \psi_0^{(0)}$ and $\psi_1^{(1)} = \psi_1^{(0)}$. First, we remove correlated distortions from this hypothetical economy, which reveals the importance of the improved allocation of factors. We then re-introduce the differences between clean and dirty technologies to the hypothetical economy, so that the resulting equilibrium is exactly the same as that in Experiment (i). This further reveals the importance of the improved adoption of clean technologies. We find that the improved adoption of clean technologies account for 48% of the decrease in average intensity.

Finally, the polluting and non-polluting sectors respond in similar ways to the removal of correlated distortions, and little reallocation is induced across sectors. Recall that the distributions of managerial talent and distortions are the same across these two sectors, and the only material difference between these two sectors is the presence of the fixed adoption cost of clean technology k_E and environmental regulations. Since the calibrated value of k_E is not too large, the differences in the reaction of these two sectors to the removal of distortions are rather limited.

One notable difference is that the mean firm size is somewhat larger in the polluting sector, because the environmental regulations work as an additional selection mechanism. The interaction between the correlated distortions and environmental regulations leads to further efficiency losses in the polluting sector, because average firm size and hence the effective distortions increase. As a result, compared to the non-polluting sector, the removal of correlated distortions leads to a 1.5 percentage points larger increase in output in the polluting sector.

Tightening Environmental Regulations.—In Experiment (ii), we increase environmental penalties ξ to the extent that the fraction of firms using the clean technology increases to the same 56% as in Experiment (i). The results are shown in Column (ii) of Table 4. Overall, in the polluting sector, the increase in ξ leads to small decreases in aggregate output, capital stock and equilibrium wage, a sizable change in firm size distribution, and substantial decreases in aggregate pollution and pollution intenisty. On the other hand, the increase in ξ has very limited spillover on both the aggregate outcomes and firm size distribution to the non-polluting sector.

The increase in ξ affects the resource allocation in the polluting sector along both the extensive and intensive margins. The most notable changes come along the extensive margin. In our model as well as in the data, it is the small unproductive firms that choose the dirty technology. Therefore, an increase in environmental penalties would mostly fall on small unproductive firms and lower their profits. This would drive a substantial fraction of small firms out of the market through the endogenous occupation choices, and lead to a sizable increase in average firm size. Meanwhile, labor supply increases as the managers of these exiting firms now choose to be workers, which suppresses the equilibrium wage. These changes along the extensive margin are efficiency-improving in the presence of correlated distortions, since the direction of the factor reallocation is opposite to that induced by the correlated distortions.

While small firms account for a large fraction of firms, they account for only a small fraction of production factors, which implies that the total amount of factors affected by the extensive changes is small. In addition, the polluting sector only account for 20% of production factors in the benchmark economy. Therefore, the induced decline in equilibrium wage is very small (0.04%). In addition, our calibrated value of the adoption cost of clean technology k_E is not too large. Therefore, the impact on surviving firms in the polluting sector is also limited, which can be seen in Table 5. Despite this, the allocation of factors along the intensive margin worsens, in the sense that the share of factors used by the most productive group decreases, because more distorted productive firms benefit less from the wage decline. This offsets the improvement in efficiency from the reallocation along the extensive margin.

Since the total amount of factors involved in reallocation along the two margins is small, most of the decrease in average pollution intensity comes from the increased adoption of the clean technology. We repeat the same decomposition exercise as we did in Experiment (i), and find that 94% and 96% of the reductions in aggregate pollution and average intensity, respectively, stem from the increased adoption of the clean technology. Since the output of the polluting sector only decreases only slightly in this experiment, the decreases in aggregate pollution and intensity are quantitatively similar at around 14%.

The non-polluting sector is indirectly affected in this experiment by the general equilibrium response in wage. Since the induced change in equilibrium wage is very small, the increase in ξ has very small effects on both the aggregate outcomes and firm size distribution, even when the products of the two sectors are perfect substitutes.

Comparing the Two Experiments.—Both of the above experiments cause the reallocation of factors across firms, but in different ways. Removing the correlated distortions from both sectors improve resource allocation along both the extensive and intensive margins, which push the economy closer to the production possibility frontiers. In contrast, while tightening environmental regulations leads to tougher selection of firms, which pushes the extensive margin closer to

TABLE 6—THE EFFECTS OF THE PROGRESSIVITY OF DISTORTIONS

Statistics	Polluting			Non-polluting		
	Benchmark	(i')	(i)	Benchmark	(i')	(i)
Output	100.00	108.15	131.16	100.00	106.87	129.62
Capital	100.00	110.91	163.06	100.00	109.63	161.26
Consumption	100.00	106.57	123.63	100.00	106.57	123.63
Wage	100.00	108.76	160.00	100.00	108.76	160.00
Output per Worker	100.00	105.74	128.25	100.00	106.03	128.60
# of Firms	100.00	44.07	44.07	100.00	41.56	41.56
Mean Size	59.98	139.19	139.19	52.18	126.53	126.53
Pollution	100	70.04	76.67			
Intensity	100	64.78	58.46			
Clean Share	57.77	73.33	85.61			

Note: All of the values are percentages except for mean firm size, which are the numbers of workers. Consumption and wage are not sector specific.

its first-best level, the resource allocation across firms along the intensive margin worsens.[33] As a result, although the fraction of firms using clean technology is set to be the same in the two experiments, the reduction of aggregate pollution intensity is larger in Experiment (i), because the larger (and cleaner) firms take a much larger market share compared to Experiment (ii).

B. Effects of the Progressivity of the Distortions

In Experiment (i), by setting $\tau_z = 0$, we change both the progressivity and and the average level of distortions. To isolate the importance of the progressivity of τ_z, we solve a version of the model where the progressive taxes are replaced with uniform taxes by setting $\phi_1 = 0$, and change ϕ_0 to keep the total implicit tax revenue constant.[34] Not surprisingly, the implied uniform tax rate $\tau_z = 18\%$ is higher than the average tax rate of 13% under the progressive tax system. Column (i') in Table 6 summarizes the results of this experiment and compares them to those from Experiment (i). A comparison of Column Benchmark and Column (i') reveals the effects of a change in progressivity of τ_z, while a comparison of Column (i') and (i) highlights the effects of a uniform change in tax rates across firms.

As we can see from Table 6, removing the progressivity of τ_z increases aggregate output, capital, consumption and output per worker. However, the 6% to 11% increases here are rather moderate compared to the 20% to 60% increases when all the distortions are eliminated. These differences are mainly caused by the relatively high level of the uniform tax rate of 18% in Experiment (i'), which dis-

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[33]In reality, tightening environmental regulations may cause other undesirable outcomes. For example, extra resources are needed to enforce tightening regulations, and more rent-seeking activities may possibly be induced by the regulations. Due to data limitations, we leave a complete welfare analysis of tightening environmental regulations to future research.

[34]This strategy has a long tradition in the public finance and development economics literature. See for example, Ventura (1999), Conesa and Krueger (2006), Bhattacharya, Guner and Ventura (2013) and Bento and Restuccia (2017), among others.

courages capital accumulation and results in much lower capital stock compared to the case of $\tau_z = 0$.[35]

However, removing the progressivity of τ_z causes a large change in firm size distribution, and therefore large declines in aggregate pollution and pollution intensity. In fact, the resulting number of firms and mean firm size are the same between Experiment (i) and (i'), since the only difference between them is a uniform increase in tax rates across firms. The average pollution intensity decreases by 35% when the progressivity of τ_z is removed, compared to a 42% decline when all the correlated distortions are eliminated. Again, the decline in average pollution intensity is caused by both an improved resource allocation and improved adoption of the clean technology among firms. However, the second channel is not as strong as in Experiment (i) (16 versus 28 percentage points), since the increase in the slope of the firm's profit functions is not as large as in Experiment (i). Moreover, since the output of the polluting sector only increases by 8% in this experiment compared to 30% in Experiment (i), the 30% decline in aggregate pollution here is larger than the 23% in Experiment (i).

In sum, our findings imply that it is the progressivity of firm-level distortions, rather than the average level of distortions, that plays a dominate role in amplifying aggregate pollution and pollution intensity.

V. Conclusions

This paper proposes a novel explanation for the industrial water pollution problem in China. In our theory, large productive firms are more likely to use the clean technology and have lower pollution intensity, largely due to the fixed adoption cost of clean technology and imperfect environmental regulations. Distortions that increase with firm TFP reallocate production factors away from large productive firms, therefore reduce aggregate output while increasing aggregate pollution intensity. We support our theory by firm-level data on production, emissions, and pollution treatment technologies from China. Quantitatively, we find that distortions correlated with firm TFP reduce aggregate output by about 30% and increase aggregate pollution by about 20%.

Our findings have novel implications for both resource misallocation and environmental pollution. They imply that the welfare losses from misallocation include not only lower economic output, which is the focus of previous studies, and but also more pollution. This could have important welfare consequences given the severity of pollution in China and other developing countries. Our findings also challenge a popular view that economic growth necessarily comes at the cost of more pollution in developing countries. They suggest that a growth path with

[35]Our results on aggregate output also appear rather moderate compared to those in Restuccia and Rogerson (2008), who control for the capital accumulation effects in their quantitative exercises. This is mainly because that the distortions in their paper cause large rank reversals between firm TFP and size, while the distortions in our paper preserve ranks. Please see the related discussions in Hopenhayn (2014b).

both higher output and lower pollution is attainable, if policy and institutional distortions can be removed in a cost effective way.

The analytical framework in this paper can be extended to other pollutants, sectors, and countries. For other pollutants, China also has a severe air pollution problem, and casual observations reveal that small and medium-sized firms in high-polluting industries such as steel contribute disproportionately to air pollution in China. For other sectors, Wu et al. (2018) find that misallocation caused by migration and land policies is an important cause of agricultural pollution in China. For other countries, Hsieh and Klenow (2014) and Bento and Restuccia (2017) find evidence of distortions that are highly correlated with firm TFP in a number of developing countries, while Dasgupta, Lucas and Wheeler (1998) find that small firms have higher air pollution intensity in Brazil and Mexico. It would be interesting to assess the impact of distortions on pollution in these countries.

REFERENCES

Acemoglu, Daron, Philippe Aghion, Leonardo Bursztyn, and David Hemous. 2012. "The Environment and Directed Technical Change." *American Economic Review*, 102(1): 131–166.

Adamopoulos, Tasso, and Diego Restuccia. 2014. "The Size Distribution of Farms and International Productivity Difference." *American Economic Review*, 104(6): 1667–1697.

Akcigit, Ufuk, Harun Alp, and Michael Peters. 2018. "Lack of Selection and Limits to Delegation: Firm Dynamics in Developing Countries." *NBER Working Paper No. 21905*.

Atkeson, Andrew, and Patrick J. Kehoe. 2005. "Modeling and Measuring Organization Capital." *Journal of Political Economy*, 113(5): 1026–1053.

Bai, Chong-En, Chang-Tai Hsieh, and Yingyi Qian. 2006. "The Return to Capital in China." *Brookings Papers on Economic Activity*, 2006(2): 61–88.

Baily, Martin Neil, Charles Hulten, David Campbell, Timothy Bresnahan, and Richard E. Caves. 1992. "Productivity Dynamics in Manufacturing Plants." *Brookings Papers on Economic Activity. Microeconomics*, 1992: 187–267.

Barrows, Geoffrey, and Hélène Ollivier. 2018. "Cleaner Firms or Cleaner Products? How Product Mix Shapes Emission Intensity from Manufacturing." *Journal of Environmental Economics and Management*, 88: 134–158.

Bartelsman, Eric, John Haltiwanger, and Stefano Scarpetta. 2013. "Cross-Country Differences in Productivity: The Role of Allocation and Selection." *American Economic Review*, 103(1): 305–334.

Bénabou, Roland. 2002. "Tax and Education Policy in a Heterogeneous-Agent Economy: What Levels of Redistribution Maximize Growth and Efficiency?" *Econometrica*, 70(2): 481–517.

Bento, Pedro, and Diego Restuccia. 2017. "Misallocation, Establishment Size, and Productivity." *American Economic Journal: Macroeconomics*, 9(3): 267–303.

Bhattacharya, Dhritiman, Nezih Guner, and Gustavo Ventura. 2013. "Distortions, Endogenous Managerial Skills and Productivity Differences." *Review of Economic Dynamics*, 16(1): 11–25.

Bloom, Nicholas, Christos Genakos, Ralf Martin, and Raffaella Sadun. 2010. "Modern Management: Good for the Environment or Just Hot Air?" *Economic Journal*, 120(544): 551–572.

Bustos, Paula. 2011. "Trade Liberalization, Exports, and Technology Upgrading: Evidence on the Impact of MERCOSUR on Argentinian Firms." *American Economic Review*, 101(1): 304–34.

Castañeda, Ana, Javier Díaz-Giménez, and José-Víctor Ríos-Rull. 2003. "Accounting for Earnings and Wealth Inequality." *Journal of Political Economy*, 111(4): 818–857.

Cherniwchan, Jevan. 2017. "Trade Liberalization and the Environment: Evidence from NAFTA and U.S. Manufacturing." *Journal of International Economics*, 105: 130–149.

Cherniwchan, Jevan, Brian R. Copeland, and M. Scott Taylor. 2017. "Trade and the Environment: New Methods, Measurements and Results." *Annual Review of Economics*, 9: 59–85.

Clementi, Gian Luca, and Hugo A. Hopenhayn. 2006. "A Theory of Financing Constraints and Firm Dynamics." *The Quarterly Journal of Economics*, 121(1): 229–265.

Conesa, Juan Carlos, and Dirk Krueger. 2006. "On the Optimal Progressivity of the Income Tax Code." *Journal of Monetary Economics*, 53(7): 1425–1450.

Cooley, Thomas F., and Vincenzo Quadrini. 2001. "Financial Markets and Firm Dynamics." *American Economic Review*, 91(5): 1286–1310.

Copeland, Brian R., and M. Scott Taylor. 1994. "North-South Trade and the Environment." *The Quarterly Journal of Economics*, 109(3): 755–787.

Copeland, Brian R., and M. Scott Taylor. 2004. "Trade, Growth, and the Environment." *Journal of Economic Literature*, 42(1): 7–71.

Dasgupta, Susmita, Robert E.B. Lucas, and David Wheeler. 1998. "Small Plants, Pollution and Poverty: New Evidence from Brazil and Mexico." *World Bank.*

Eberhardt, Markus, Zheng Wang, and Zhihong Yu. 2016. "From One to Many Central Plans: Drug Advertising Inspections and Intra-National Protectionism in China." *Journal of Comparative Economics*, 44(3): 608–622.

Forslid, Rikard, Toshihiro Okubo, and Karen Helene Ulltveit-Moe. 2018. "Why Are Firms that Export Cleaner? International Trade, Abatement, and Environmental Emissions." *Journal of Environmental Economics and Management*, 91: 166–183.

Grossman, Gene M., and Alan B. Krueger. 1993. "Environmental Impacts of A North American Free Trade Agreement." In *The U.S.-Mexico Free Trade Agreement.* , ed. Peter M. Garber. Cambridge, Massachusetts:MIT Press.

Guner, Nezih, Andrii Parkhomenko, and Gustavo Ventura. 2018. "Managers and Productivity Differences." *Review of Economic Dynamics*, 29: 256–282.

Guner, Nezih, Gustavo Ventura, and Yi Xu. 2008. "Macroeconomic Implications of Size-dependent Policies." *Review of Economic Dynamics*, 11(4): 721–744.

Hopenhayn, Hugo A. 2014*a*. "Firms, Misallocation, and Aggregate Productivity: A Review." *Annual Review of Economics*, 6(1): 735–770.

Hopenhayn, Hugo A. 2014*b*. "On the Measure of Distortions." *NBER Working Paper No. 20404.*

Hsieh, Chang-Tai, and Peter J. Klenow. 2009. "Misallocation and Manufacturing TFP in China and India." *The Quarterly Journal of Economics*, 124(4): 1403–1448.

Hsieh, Chang-Tai, and Peter J. Klenow. 2014. "The Life-cycle of Plants in India and Mexico." *The Quarterly Journal of Economics*, 129(3): 1035–1084.

Levinson, Arik. 2009. "Technology, International Trade, and Pollution from US Manufacturing." *American Economic Review*, 99(5): 2177–2192.

Lin, Liguo. 2013. "Enforcement of Pollution Levies in China." *Journal of Public Economics*, 98: 32–43.

Li, Zhe, and Jianfei Sun. 2015. "Emission Taxes and Standards in a General Equilibrium with Entry and Exit." *Journal of Economic Dynamics and Control*, 61: 34–60.

Li, Zhe, and Shouyong Shi. 2017. "Emission Taxes and Standards in a General Equilibrium with Productivity Dispersion and Abatement." *Macroeconomic Dynamics*, 21(8): 1857–1886.

Lucas, Robert E. Jr. 1978. "On the Size Distribution of Business Firms." *The Bell Journal of Economics*, 9(2): 508–523.

Martin, Leslie A. 2013. "Energy Efficiency Gains from Trade: Greenhouse Gas Emission and India's Manufacturing Firms." *Unpublished.*

Melitz, Marc J. 2003. "The Impact of Trade on Intra-Industry Reallocations and Aggregate Industry Productivity." *Econometrica*, 71(6): 1695–1725.

Niu, Kunyu, Jian Wu, Fang Yu, and Jingli Guo. 2016. "Construction and Operation Costs of Wastewater Treatment and Implications for the Paper Industry in China." *Environmental Science and Technology*, 50(22): 12339–12347.

Olley, George Steven, and Ariel Pakes. 1996. "The Dynamics of Productivity in the Telecommunications Equipment Industry." *Econometrica*, 64(6): 1263–1297.

Parente, Stephen L., and Edward C. Prescott. 1994. "Barriers to Technology Adoption and Development." *Journal of Political Economy*, 102(2): 298–321.

Parente, Stephen L., and Edward C. Prescott. 1999. "Monopoly Rights: A Barrier to Riches." *American Economic Review*, 89(5): 1216–1233.

Qi, Ji, Bo Zheng, Meng Li, Fang Yu, Chuchu Chen, Fei Liu, Xiafei Zhou, Jing Yuan, Qiang Zhang, and Kebin He. 2017. "A High-resolution Air Pollutants Emission Inventory in 2013 for the Beijing-Tianjin-Hebei Region, China." *Atmospheric Environment*, 170: 156–168.

Restuccia, Diego, and Richard Rogerson. 2008. "Policy Distortions and Aggregate Productivity with Heterogeneous Establishments." *Review of Economic Dynamics*, 11(4): 707–720.

Restuccia, Diego, and Richard Rogerson. 2013. "Misallocation and Productivity." *Review of Economic Dynamics*, 16(1): 1–10.

Restuccia, Diego, and Richard Rogerson. 2017. "The Causes and Costs of Misallocation." *Journal of Economic Perspectives*, 31(3): 151–174.

Shapiro, Joseph S., and Reed Walker. 2018. "Why is Pollution from U.S. Manufacturing Declining? The Roles of Environmental Regulation, Productivity, and Trade." *American Economic Review*, 108(12): 3814–3854.

Song, Zheng, Kjetil Storesletten, and Fabrizio Zilibotti. 2011. "Growing Like China." *American Economic Review*, 101(1): 196–233.

The World Bank. 2007. *Cost of Pollution in China: Economic Estimates of Physical Damages.* Washington D.C.:World Bank Publications.

Tombe, Trevor, and Jennifer Winter. 2015. "Environmental Policy and Misallocation: The Productivity Effect of Intensity Standards." *Journal of Environmental Economics and Management*, 72: 137–163.

Tombe, Trevor, and Xiaodong Zhu. 2019. "Trade, Migration and Productivity: A Quantitative Analysis of China." *American Economic Review*, 109(5): 1843–1872.

Vennemo, Haakon, Kristin Aunan, Henrik Lindhjem, and Hans Martin Seip. 2009. "Environmental Pollution in China: Status and Trends." *Review of Environmental Economics and Policy*, 3(2): 209–230.

Ventura, Gustavo. 1999. "Flat Tax Reform: A Quantitative Exploration." *Journal of Economic Dynamics and Control*, 23(9-10): 1425–1458.

Wu, Yiyun, Xican Xi, Xin Tang, Deming Luo, Baojing Gu, Shu Kee Lam, Peter Vitousek, and Deli Chen. 2018. "Policy Distortions, Farm Size, and the Overuse of Agricultural Chemicals in China." *Proceedings of the National Academy of Sciences of the United States of America*, 115(27): 7010–7015.

Zheng, Siqi, and Matthew E. Kahn. 2013. "Understanding China's Urban Pollution Dynamics." *Journal of Economic Literature*, 51(3): 731–772.