## Climate Policy and Labor Market Inequality

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#### Abstract

This paper investigates the effects of carbon taxation on labor market inequality in Sweden. Using matched employer-employee data from the Swedish registers for the years 2004-2018, I estimate the effects of a reform that increased the stringency of the tax for a subset of firms in the manufacturing sector. Using a difference-in-difference framework, I find that the reform significantly reduced emissions among treated firms. However, it also reduced the employment of workers without a high school degree. Further results suggest that this effect is driven by a reduction in the hiring rate of this group. In addition, I find that negative employment impacts are concentrated among emission-intensive firms, which face the largest cost increases when carbon tax rates rise. The results show that carbon taxation, while effective at reducing emissions, may have strongly heterogeneous employment impacts, and that complementary policies might be needed to address labor market inequalities when implementing climate policy.

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### 1 Introduction

More countries are adopting climate policy strategies to comply with their national targets as part of the Paris Agreement. Carbon taxation remains the most prominent climate policy instrument (World Bank, 2024). However, fear of negative economic impacts and increasing labor market inequality has challenged the political support for climate policy (Vona, 2019). Despite the urgency of the climate crisis, there is still a lack of empirical evidence on the distributional impacts of climate policy and its effects on labor market inequality (Gray et al., 2023; Vrolijk and Sato, 2023).

This paper studies the effects of climate policy on firm performance and labor market inequality, combining changes in the Swedish carbon tax with administrative data on firms and workers. The tax, which is imposed on fossil fuels used for heating and engine operation, is one of the highest in the world (World Bank, 2024). I exploit a reform in the policy to isolate exogeneous variation in firm-level effective carbon tax rates. The reform increased the stringency of the carbon tax for manufacturing firms by gradually removing the possibility to apply for tax refunds over the years 2011-2018. I develop an empirical framework based on firms' differential exposure to the reform, which stems from variation in pre-reform uptake of carbon tax refunds.

I find that the increase in effective carbon tax rates led to a substantial reduction in firms' total emissions as well as emission *intensity*, by up to 40% over the time period. However, I also observe significantly negative effects on firm performance. Heterogeneity analyses reveal that, while emission reductions are found in firms with low and high emission intensity, negative effects on revenue, value added and employment are concentrated in emission-intensive firms. Further, I find that the negative effect on employment is skill-biased against low-educated workers, as it is entirely driven by a reduction in the number of workers without a high school degree. Within this group, I find stronger effects for older workers and males.

The result has relevance for a number of issues related to contemporary policy developments and the green transition. The heterogeneous impacts across workers raise distributional concerns regarding the incidence of climate policy and its potentially exacerbating effect on labor market inequality. The fact that low-educated workers suffer from substantially higher unemployment rates (Statistics Sweden, 2024), suggests a need for complement-

tary policies to ensure a 'just' transition (UNDP, 2022). While the largest manufacturing firms (in terms of employment and emissions) are regulated by the European Union Emissions Trading System (EU ETS), and therefore exempted from the domestic tax, the on average small and medium-sized firms covered by the Swedish carbon tax represent a substantial share of manufacturing employment (Eurostat, 2024). The result in this paper is therefore informative to the increasing number of countries that are adopting carbon taxation (World Bank, 2024), and for the firms to be covered by the EU's second carbon market (ETS2). The ETS2, which will be launched in 2027, will cover smaller industrial firms by upstream regulation, thus sharing key features in regulatory design with the Swedish carbon tax (European Commission, nda).<sup>1</sup>

The main contribution of this paper is that it provides the first empirical evidence on heterogeneous effects of carbon pricing on employment using linked employer-employee data. The emerging literature in this field has, to a large extent, been restricted to estimating firm-level effects of carbon pricing policies, such as carbon taxation and emissions trading. Prior studies have found no impact on average employment, despite reductions in emissions (Colmer et al., 2024; Dechezleprêtre et al., 2023; Marin et al., 2018; Martin et al., 2014). Few existing studies analyze employment impacts of climate policies using individual-level data to allow for heterogeneous effects across workers. Yamazaki (2017, 2019) analyze the implementation of a revenue-neutral carbon tax in British Columbia, Canada. Using other Canadian provinces as control groups, the author finds indications of a sectoral reallocation of production workers (wage-paid) towards cleaner sectors. Studying the same reform, Yip (2018) uses individual data from labor force surveys. The study finds that the carbon tax significantly increased unemployment, with stronger effects for low-educated males. However, without firm-level data, the author cannot distinguish between changes in firms' labor demand and changes in composition of the local economy. Walker (2013) uses linked worker-firm data to evaluate the effects on workers' earnings of the US Clean Air Act Amendments (CAAA) in 1990, which made the regulation of local air pollutants more stringent. The study finds that the reform had a long-lasting negative effect on earnings, which is entirely driven by workers that were separated from their firm, and larger for workers who were reallocated to another

<sup>&</sup>lt;sup>1</sup>As explained in Section 2, the Swedish carbon tax is imposed on upstream energy retailers. Similarly, ETS2 will implement a cap-and-trade system for fuel suppliers (European Commission, nda).

industry and county. While being an important contribution to the literature on the impacts of environmental regulation on individuals' labor market outcomes, Walker (2013) analyzes a different policy context compared to this paper. By analyzing the effects of the Swedish carbon tax reform, I study the most prominent policy tool world-wide to mitigate climate change in a recent time period. In contrast to the regulations underlying the 1990 CAAA, a carbon tax is a market-based policy instrument with the potential to incentivize emission reductions at least cost, reducing the economic burden of regulation (Colmer et al., 2024).

A second strand makes use of pre-existing definitions of 'green' tasks and occupations, predominantly based on measures available in the US Bureau of Labor Statistics' O\*NET database (Apostel and Barslund, 2024). This approach has been used to study, for example, the prevalence of green jobs and skills in the economy (Bowen et al., 2018; Curtis and Marinescu, 2022; Popp et al., 2021; Saussay et al., 2022; Vona et al., 2019), the characteristics of skills associated with green tasks (Vona et al., 2018), and employment impacts of innovation (Elliott et al., 2024). Exploiting regional variation in the stringency of environmental regulation in the US, Vona et al. (2018) points to an increasing demand for technical and engineering skills in more strictly regulated areas. Popp et al. (2021) also investigates the employment impacts of the green investments part of the 2009 American Recovery and Reinvestment Act (ARRA). The study finds that, at the regional level, newly created jobs are mainly manual labor positions, which allows for a reallocation of workers previously displaced due to environmental regulation. I contribute to this strand by studying within-firm employment changes for different workers, thus providing evidence of distributional impacts of climate policy on firms' labor demand without relying on green definitions.

In addition, this paper complements previous research on the effectiveness of carbon pricing (Ahmadi et al., 2022; Andersson, 2019; Brännlund and Lundgren, 2010; Brännlund et al., 2014; Jaraite et al., 2014; Jaraite and Maria, 2016; Leroutier, 2022; Martinsson et al., 2024;

<sup>&</sup>lt;sup>2</sup>A related strand of literature exploits variation in energy prices as a proxy for climate policy stringency. Within this framework, Marin and Vona (2019, 2021) find that climate policy is potentially skill-biased at the occupational level, leading to a higher demand for technicians and a lower demand for manual workers. However, as noted by Andersson (2019) and Brännlund et al. (2014), firms and households respond differently to variations in tax levels than prices, potentially through different expectations of future changes, leading to stronger taxation effects.

Pretis, 2022), by developing a novel empirical methodology using an unexploited register on excise tax refunds to observe firm-level treatment variation. Last, I contribute to the broader literature on the economic impacts of environmental regulation (Berman and Bui, 2001; Cole and Elliott, 2007; Curtis, 2018; Ferris et al., 2014; Gray et al., 2014; Greenstone, 2002; Liu et al., 2017, 2021; Morgenstern et al., 2002; Sheriff et al., 2019; Walker, 2011).

The remainder of this paper proceeds as follows. Section 2 describes the relevant features of the Swedish carbon tax and the reform studied in this paper. Section 3 describes the dataset and presents descriptive statistics. Section 4 presents the empirical framework, and Section 5 presents the results. Section 6 provides a discussion of the results, and concludes.

## 2 Institutional Background

The Swedish carbon tax was implemented in 1991, and established a price on emissions from fossil fuels consumed for heating or engine operation (SFS 1994:1776, nd). The tax is measured in Swedish Krona (SEK) per volume of fuel and varies across fuels based on their carbon content, such that the SEK/ton CO<sub>2</sub> tax rate is constant. To facilitate administration, the regulation has adopted a tax suspension regime in which, in principle, upstream firms that import, produce or sell energy products are tax liable and must register as taxpayers. The tax is levied when a fuel is sold by a registered taxpayer to a consumer. Between registered taxpayers, however, taxation is suspended (Hammar and Åkerfeldt, 2011). Most industrial firms are not registered taxpayers, but are instead affected by the carbon tax through higher prices on fossil fuels as energy retailers pass on their tax payments to consumers. The regulatory design thus relies on sufficient cost pass-through from the energy sector to incentivize emission reductions in the overall economy.

Figure 1a plots the carbon tax in SEK per ton CO<sub>2</sub> between 2004-2018.<sup>3</sup> However, for the industrial sector, the regulation has featured generous rebates over time. Between 2004-2010, the government offered a 79% refund of the tax paid on industrial fuel consumption fulfilling certain criteria. First, tax refunds were only granted for fuel consumed in the manufacturing process for uses other than motorized vehicles.<sup>4</sup> Second, the manufacturing

<sup>&</sup>lt;sup>3</sup>The average exchange rate over the period was 9.39 SEK/EUR.

<sup>&</sup>lt;sup>4</sup>Examples of industrial motorized vehicles are excavators and wheel loaders. This condition also excludes

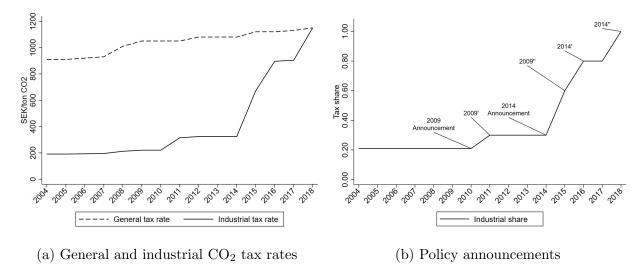


Figure 1: Timing of policy implementation. Figure 1a shows evolution of the Swedish carbon tax with and without industrial rebates, in SEK/ton CO<sub>2</sub>. Figure 1b shows the timing of announcements of the two policy changes (2009 and 2014), and their corresponding increases in industrial tax shares.

process in which fuel has been used must be the *main activity* of the firm. This implies that refunds were largely given to firms for heating in the manufacturing process.<sup>5</sup> Firms received the tax refund through application up to three years after fuel purchase, assuming a 100% pass-through of the tax to fuel prices. The resulting net tax rate is represented by the solid line in Figure 1a.<sup>6</sup> For firms regulated by the EU ETS, the carbon tax was completely removed in 2011 to avoid double carbon pricing (Ryner, 2022).

The manufacturing rebates were gradually reduced from 2011, and completely removed in 2018. The phase-out of the rebates was communicated in two steps, which are shown in fuels used for transportation of goods on roads.

<sup>&</sup>lt;sup>5</sup>The rebate was also given to utilities delivering heating to manufacturing firms for this purpose, such that total tax burden along the supply chain was independent of whether heating was delivered or generated on-site by the firm.

<sup>&</sup>lt;sup>6</sup>Additional rebates have been directed to specific sectors. Fuels used in the production of energy products and in some metallurgical and mineralogical processes are completely exempt from carbon taxation, as was fuels used for special vehicles in manufacturing in the mining industry until 2020. In addition, tax payments were capped at 0.8% of sales until 2015.

Figure 1b. In 2009, the government released a new plan to achieve its medium-term climate targets. The plan included an increase in the share of the carbon tax paid by industry from 21% to 30% in 2011, with an additional increase to 60% in 2015. An assessment was planned to be made in 2015 to evaluate the effectiveness and socioeconomic costs of the policy (Government Bill 2008/09:162, 2009). After elections in 2014, the new government already presented an updated climate plan which mandated further emission reductions domestically (Government Bill 2014/15:1, 2014). The new plan featured an increase in the industrial tax share in 2016, and a complete phase-out of the rebates in 2018.

The removal of carbon tax rebates constitutes a suitable setting to study the environmental and economic impacts of climate policy, for three reasons. First, cross-sectional variation in uptake of tax rebates before announcement means that firms were differentially exposed to the reform. Second, the fact that rebates were not based on industry classification, but rather fuel usage, allows for a comparison of firms within the same industries, reducing the risk of confounding factors. Third, the effective tax rate increased by up to five times over the reform period for the most affected firms, meaning that the reform substantially raised the incentives for emission reduction. Hence, the removal of rebates induced meaningful and plausibly (conditionally) exogenous variation in climate policy stringency across firms over time.

### 3 Data

Data sources The sample is constructed from the Energy Use in Manufacturing survey (ISEN). It is a mandatory annual survey for all manufacturing firms with more than 9 employees, and collects information on the cost and quantity of energy consumption by fuel type. The dataset used in this paper covers the years 2004-2018, and is linked to administrative tax records, which includes information about firms' accounting. By combining fuel consumption with fuel-specific emission factors from the Swedish Environmental Protection Agency, I obtain firm-level annual emissions. The dataset is further linked to individuals in working age (16-64) by the organizational number of a worker's employer in November in a given year, which provides information on worker characteristics such as age, gender,

educational history and income.

Importantly, the dataset is linked to a register containing information about firms' excise duty refunds, which covers the energy and carbon taxes on fuels. This data is available from 2008, which, combined with fuel consumption data, allows me to calculate the implied net carbon tax rate for each firm in a given year. Since the majority of these firms do not pay the tax directly to the authority, I calculate indirect gross tax payments based on fuel consumption and official tax rates (in SEK per volume). From this I obtain firm-level net carbon tax rates by subtracting any deduction or refund observed in the tax register. In some cases, the resulting tax rates are negative. One potential reason for this is measurement error in the fuel consumption survey or when matching fuels to tax rates in the regulatory text, which categorizes fuels differently. A second potential reason is the possibility for firms to apply for refunds retrospectively up to three years after purchasing a fuel, which could result in an accumulation of refunds exceeding gross tax payments in some years. I approach this issue by setting all negative tax rates to zero, as these firms are likely to have had some tax rebate these years.

Sample restriction As outlined in the section below, treatment is defined as having a tax rebate in 2008. In the following analysis I make use of two samples with different selection criteria. Regardless of sample, however, I remove firms that were ever regulated by the EU ETS. This is done to avoid endogenous selection in and out of regulation of the domestic carbon tax, since EU ETS firms have been subject to different rebates (and a complete exemption from the tax since 2011). In the main analysis, I restrict the sample to a balanced panel of firms with observations in all years between 2004-2018. This makes it possible to evaluate differential trends between treated and control firms in the relevant outcomes before the implementation of the reform, and removes any compositional effects over time. It also allows me to further restrict the sample to firms with positive emissions in all pre-reform years 2004-2008, which increases the comparability between firms and therefore internal validity. The secondary sample is characterized by a less restrictive selection criteria.

<sup>&</sup>lt;sup>7</sup>The dataset currently misses observations for 2013.

<sup>&</sup>lt;sup>8</sup>Since the empirical framework of the paper is based on a binary definition of treatment (having a tax rebate or not), the exact level will not be important for the result.

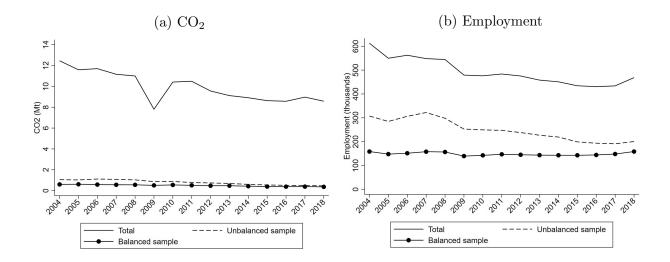


Figure 2: Aggregate emissions and employment over time

This sample consists of firms that are observed, with positive emissions, at least in 2007 and 2008. This unbalanced panel is used to evaluate the sensitivity of the result to compositional changes (i.e. firm exit), and to conduct heterogeneity analysis across industries with higher statistical power. A comparison of the balanced sample and the unbalanced one will also be informative regarding the external validity of the result.

Figure 2 shows the coverage of the two samples in terms of emissions and employment, in relation to all manufacturing in ISEN. Both samples constitutes a small share of manufacturing emissions. This is due to the selection criteria excluding firms which were *ever* regulated by the EU ETS. Manufacturing emissions are characterized by a heavily skewed distribution, with a strong selection of energy- and emission-intensive firms falling under EU ETS. However, this is not the case for the number of employed workers, where the analyzed firms make up a substantial share of manufacturing employment.

<sup>&</sup>lt;sup>9</sup>Observations in 2007 are used to construct control variables related to exposure to the Great Recession in the sensitivity analysis.

**Descriptive statistics** The above restrictions result in a dataset of 3,222 unique firms, of which 1,163 belongs to the balanced sample. Table 1 presents summary statistics for a selection of variables for the control and treatment group, respectively, for the two samples. Treated firms (i.e. those with a carbon tax rebate in 2008) are on average larger in terms of employment, revenue and capital (fixed assets). The difference is, however, most pronounced when comparing CO<sub>2</sub>. Treated firms use more fossil fuels in relation to their total energy consumption, and causes substantially more emissions. To some degree, these differences might reflect the higher incentives for emission-intensive firms to apply for tax refunds, and more knowledge about refund application possibilities among larger firms. Still, Figure 9 in the Appendix shows that there is considerable overlap in outcomes between the control and treatment group when log-transformed. Figure 3 shows the distribution of treatment across industries. Reassuringly, there is within-industry variation in treatment, as both treated and control firms are found in the majority of manufacturing industries. The dataset also covers a repeated cross-section of 298,213 unique individuals employed at the firms at some point in the sample period. The typical worker is a high school educated male, and there are small differences in worker composition between treated and control firms in the main (balanced) sample.

Figure 4 plots the raw trends in average outcomes for the treatment and control group in relation to the announcement of the reform (2009) and the year of implementation (2011). A salient feature is the impacts of the Great Recession in 2009, which caused a sudden fall in firm performance. Despite differences in levels, average outcomes for the two groups run parallel over all pre-reform years, with the exception of capital. Firms in the control group seems to have a steeper increase in the leading years, which warrants extra caution when analyzing this outcome. A general sensitivity analysis with respect to influence from the Great Recession is carried out in the empirical section.

Table 1: Descriptive Statistics

	Mean 2008							
		Unbalanc	Balanced					
	All	Control	Treated	Control	Treated			
	(1)	(2)	(3)	(4)	(5)			
Firms	3,222	1,628	1,594	464	699			
Employment	92.64	64.71	121.15	107.43	152.91			
Revenue (mSEK)	239.07	141.66	338.57	255.75	410.68			
Fixed assets (mSEK)	85.73	56.93	115.14	128.83	159.91			
$CO_2$ emissions (ton)	320.55	102.47	543.29	136.64	711.96			
Fossil energy share	0.30	0.22	0.39	0.18	0.35			
Workers	298,213	105,218	192,995	49,848	106,885			
No high school	0.19	0.19	0.19	0.18	0.20			
High school	0.56	0.61	0.54	0.60	0.57			
Above high school	0.25	0.20	0.27	0.22	0.23			
STEM	0.17	0.14	0.19	0.15	0.16			
Female	0.24	0.22	0.24	0.24	0.23			
Age	42.82	42.87	41.80	41.89	41.74			

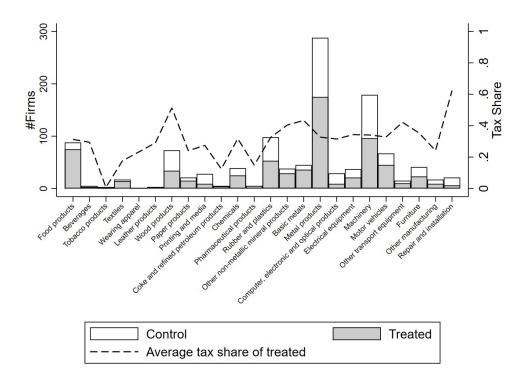


Figure 3: Treatment by industry

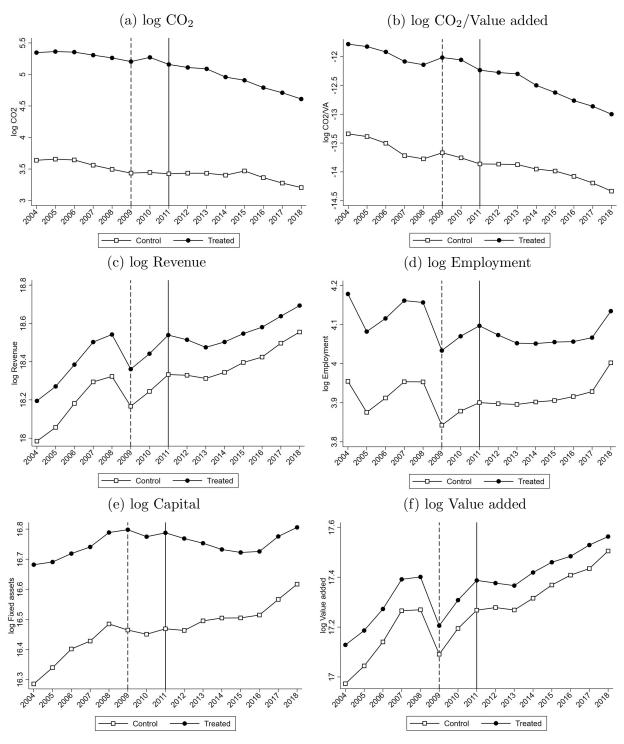


Figure 4: Trends in average outcomes

## 4 Empirical Framework

The cross-sectional variation in tax rebate uptake creates firm-level variation in exposure to the reform. Treated firms are those whose carbon tax rebates were removed over the treatment period. I define the first year of treatment as the year of the first announcement to decrease the tax rebates for industrial firms, which happened in 2009. The control group consists of firms that already paid the full carbon tax rate in 2008, before the rebates were phased out. These firms are arguably unaffected by the policy change.

The analysis is based on two empirical models. The first approach is an event-study capturing the dynamics of the estimated treatment effect between 2004 - 2018. It is represented by the following equation

$$\ln Y_{jt} = \eta_j + \alpha_{It} + \sum_{k=2004}^{2018} \beta^k \times \mathbf{1}(t=k) \times D_j + \epsilon_{jt}$$
(1)

where 2008 is the omitted year of reference.<sup>10</sup>  $Y_{jt}$  is the outcome of firm j in year t. I control for firm fixed effects  $\eta_j$  and year-by-industry fixed effects  $\alpha_{It}$  to accommodate shocks specific to industry I. Treatment  $D_j$  equals one if firm j had a carbon tax rebate in 2008.  $\beta^k$  captures the marginal effect of higher carbon tax stringency (through lower rebates) in year k. Treatment adoption occurs simultaneously for all firms, and the binary definition of treatment overcomes potential issues related to negative weights and heterogeneous treatment effects discussed in the recent econometrics literature (Callaway et al., 2024).  $\epsilon_{jt}$  is an error term allowed to correlate over time within firms. The second empirical model is a long difference approach which estimates the following two-period equation

$$\ln Y_{it} = \eta_i + Post_t + \Gamma_I \times Post_t + \beta D_i \times Post_t + \epsilon_{it}$$
 (2)

where t is either 2008 or 2018, and  $Post_t = \mathbf{1}(t=2018)$  is an indicator for the final year of the reform.  $\Gamma_I \times Post_t$  is an interaction of industry indicators and the year indicator

<sup>&</sup>lt;sup>10</sup>This means that treatment effects will be compared to differences in the year before the observed impacts of the Great Recession (see Figure 4).

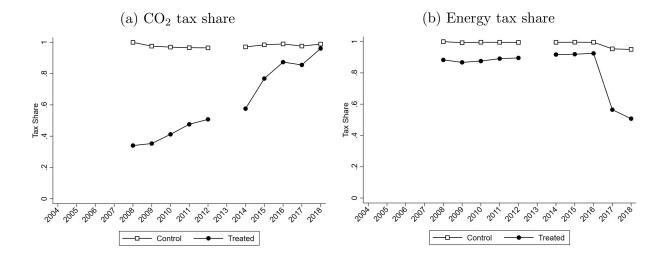


Figure 5: Firm-level fuel tax shares

to control for industry shocks.  $\beta$  represents the long-difference estimate of the complete phase-out of the tax rebates.

Both approaches rely on the assumption that pre-announcement tax rebate status  $D_j$  is exogenous to unobserved, within-industry changes in outcomes  $\epsilon_{jt}$ . Figure 5 provides information of a potential source of bias, namely concurring changes in other fuel policies. Using the actual data, Figure 5a shows that average calculated  $CO_2$  tax shares follow the pattern of the reform for treated firms, with average shares close to 1 for control firms, validating the treatment assignment procedure. Figure 5b shows the respective average energy tax share, which is also imposed on fuels. The figure shows no evidence of diverging trends between the groups for most of the time period. However, treated firms' energy tax share falls in 2017 and 2018, which could influence outcomes in these years. Event-study estimates makes it possible to compare treatment effects across years, and therefore to assess the importance of these changes. If the assumption holds, the empirical model will identify the average treatment effect on the treated of increases in the stringency of climate policy.

### 5 Result

#### 5.1 Main result

Figure 6 presents the estimated effects of increasing climate policy stringency on firms' environmental and economic performance. Reassuringly, I do not find significant differential trends in outcomes before the reform except for capital, which was already observed in the raw data. Figure 6a and 6b show large, significant reductions in treated firms' total emissions and emission intensity. Starting in 2011, firms reduce emissions continuously until 2015, after which the negative effects stabilize around -40%, without signs of recovery. The reform also seems to have caused a negative impact on firm performance, although not in the proportion of the emission reductions. Both revenue and employment falls after the implementation of the reform, with significantly estimated point estimates around -7%. Similar patterns are observed for capital and value added, although differential trends before the reform and large confidence intervals precludes conclusions regarding the effects on these outcomes. Treated firms are seemingly more affected by the Great Recession, as seen by the downward jump in 2009 in revenue and employment. However, this decline is fully recovered during the announcement period, and has an opposite direction of the estimated treatment effect, suggesting that differential recovery paths after the crisis are not likely to drive the result. Further analysis on the sensitivity to this shock is carried out below.

### 5.2 Sensitivity of the main result

Business cycle exposure This section provides more evidence on the robustness of the main result to different specifications, starting with controlling for observables related to exposure to business cycles. The announcement of the reform coincides with the negative impacts of the Great Recession in 2009. In addition to the industry-year fixed effects already included in the baseline regressions, I construct variables related to three dimensions of firms' exposure to business cycles, namely export share of sales  $(EX'_j)$ , employment size  $(L'_j)$ , and capital size  $(K_j)$ .  $EX'_j$  is a vector of two indicator variables, which equal to 1 if the firm's exports as a share of total sales in 2007 is in the range (0%, 50%) or >= 50%, respectively.  $L'_j$  is also a vector of two indicator variables, which equal 1 for firms whose number of employees

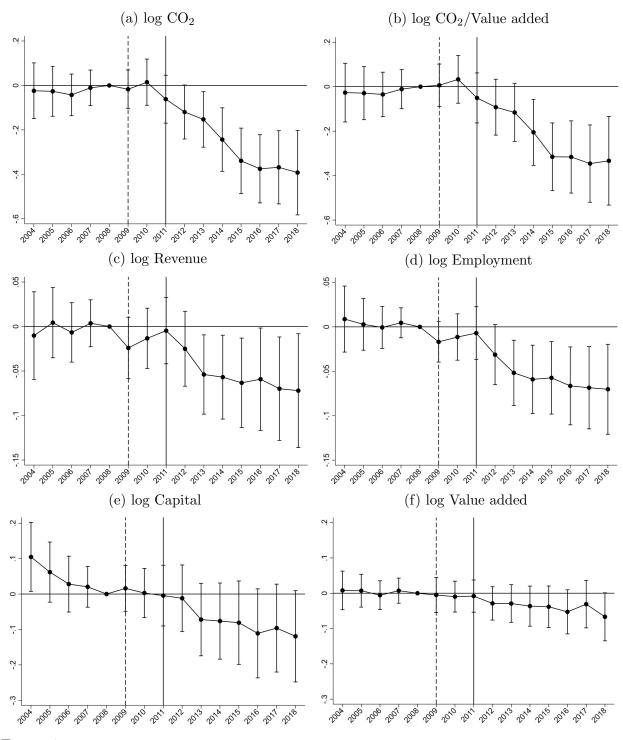


Figure 6: Event study results using a balanced panel of firms. Regressions includes industry-by-year fixed effects, and standard errors are clustered by firm. Capped spikes show 95% confidence intervals.

in 2007 is in the range (49,250) or >= 250, respectively.  $K_j$  is an indicator variable equal to 1 if the firm's fixed assets in 2007 exceed the 2-digit industry median for that year. These variables are interacted with year fixed effects (or  $Post_t$  in the long difference specification), to allow for separate, non-parametric time trends along these dimensions. This will, for example, capture different exposures to exchange rate fluctuations for exporting versus non-exporting firms, or different impacts relating to firm size.

The result is presented in Table 2. Odd columns show the baseline result of the long difference model, which repeats the observed result of Figure 6. Even columns controls for business cycles proxies. Adding these control variables have minimal impact on the estimated effects, suggesting that biases from overlapping changes in economic conditions are not a large concern for the empirical research design.

	log	$CO_2$	log Co	$O_2/VA$	log R	evenue	log Emp	loyment	log (	Capital log		y VA
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)	(10)	(11)	(12)
$D \times Post$	-0.403***	-0.396***	-0.347***	-0.343***	-0.080**	-0.082**	-0.070***	-0.065**	-0.114*	-0.129**	-0.067*	-0.065*
	(0.098)	(0.098)	(0.103)	(0.103)	(0.032)	(0.032)	(0.025)	(0.025)	(0.066)	(0.065)	(0.034)	(0.034)
Observations	1,990	1,990	1,974	1,974	2,320	2,320	2,326	2,326	2,304	2,304	2,306	2,306
$\eta_j$	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>	<b>√</b>
$Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	✓	✓	$\checkmark$	✓	$\checkmark$	✓	✓
$\Gamma_I \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	✓	✓	$\checkmark$	✓	$\checkmark$	✓	✓
$EX'_i \times Post_t$		$\checkmark$		$\checkmark$		✓		✓		$\checkmark$		✓
$L'_i \times Post_t$		$\checkmark$		$\checkmark$		✓		$\checkmark$		$\checkmark$		$\checkmark$
$K_j \times Post_t$		$\checkmark$		$\checkmark$		✓		$\checkmark$		$\checkmark$		$\checkmark$

Table 2: Long difference results using a balanced panel of firms. Standard errors in parenthesis are clustered by firm. \*\*\* p<0.01, \*\* p<0.05, \* p<0.1

Firm turnover I also investigate the sensitivity of the baseline result to sample restriction. In Figure 7 I re-estimate the event-study model using the unbalanced sample. As before, the result shows significant reductions in emissions and emission intensity, with similar magnitudes as with the balanced sample. Likewise, I find a negative effect of the reform on revenue and employment. Regarding emissions, I observe differential pre-trends, which is likely due to the firms which are now entering the sample between 2004-2007. Nevertheless, the estimation suggests that firm exit does not play a role in previous findings, and that the validity of the baseline result is not limited to the subset of surviving firms over the entire time period.

### 5.3 Heterogeneity Analysis

**Firm heterogeneity** This section explores heterogeneity in the previous result, starting with firm characteristics. In order to investigate the extent to which estimated treatment effects vary across firms, I focus on two dimensions. First, I construct the variable  $CO_2$  int. which equals 1 for firms with a CO<sub>2</sub> intensity (in terms of value added) above the 2-digit industry median in 2007. Second, I construct the variable Capital int., which equals to 1 for firms with a capital intensity, measured as the ratio of fixed assets to employees, above the 2-digit industry median in 2007. Emission-intensive firms face higher incentives to reduce emissions as effective tax rates rise, due to the higher costs of compliance. It is also possible that capital-intensive firms have different modes of compliance than labor-intensive firms (e.g. varying energy-labor substitution possibilities), leading to heterogeneous responses in outcomes. Table 3 shows the result along these dimensions, where the new variables are interacted with the previous treatment term. Column (2) reports the effects on total emissions. The point estimates suggest that both low and high emission-intensive firms reduce emissions, with high emission-intensive firms responding more strongly (although not significantly). Point estimates in column (4), which shows result for emission intensity, has a similar pattern as total emissions. Interestingly, the negative employment effects from previous estimations seem to be concentrated among emission-intensive firms. This is in line with emission-intensive firms facing higher cost increases due to the reform. Capital-intensive firms have in general less negative impacts, although statistical precision for this interaction

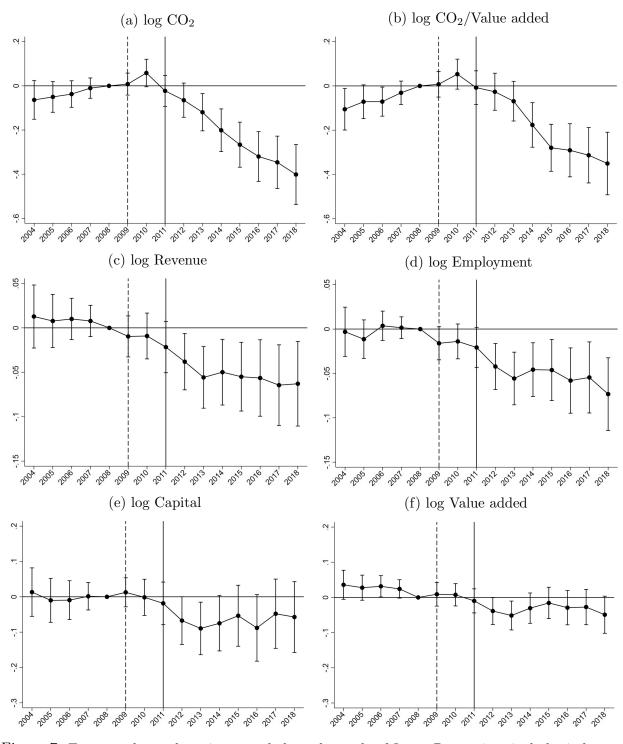


Figure 7: Event study results using an unbalanced sample of firms. Regressions includes industry-by-year fixed effects, and standard errors are clustered by firm. Capped spikes show 95% confidence intervals.

term is again low. Table 4 summarizes the result of Table 3 by presenting the estimated net treatment effects for the four subgroups, 'Low emission intensity & Labor intense', 'Low emission intensity & Capital intense', 'High emission intensity & Labor intense,' and 'High emission intensity & Capital intense'.

	$\log \mathrm{CO}_2$		$\log \mathrm{CO_2/VA}$		log Revenue		log Employment		log Capital		log VA	
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)	(10)	(11)	(12)
$D \times Post$	-0.396***	-0.336**	-0.343***	-0.368**	-0.082**	-0.040	-0.065**	0.009	-0.129**	-0.145	-0.065*	-0.013
	(0.099)	(0.168)	(0.103)	(0.179)	(0.032)	(0.050)	(0.025)	(0.042)	(0.065)	(0.112)	(0.034)	(0.058)
$D \times Post \times CO_2$ int.		-0.252		-0.098		-0.085		-0.134**		0.042		-0.110
		(0.179)		(0.182)		(0.061)		(0.053)		(0.146)		(0.069)
$D \times Post \times Capital \ int.$		0.205		0.285		0.016		0.020		0.077		-0.018
		(0.179)		(0.187)		(0.063)		(0.049)		(0.127)		(0.067)
Observations	1,990	1,990	1,974	1,974	2,320	2,320	2,326	2,326	2,304	2,304	2,306	2,306
$CO_2 \ intj \times Post_t$		<b>√</b>		<b>√</b>		<b>√</b>		✓		<b>√</b>		<b>√</b>
Capital $intj \times Post_t$		$\checkmark$		$\checkmark$		$\checkmark$		✓		✓		$\checkmark$

Table 3: Long difference results using a balanced panel of firms.  $CO_2$  int.<sub>j</sub> is an indicator equal to one for firms with a  $CO_2$  intensity (defined as ton  $CO_2$ /value added) above the 2-digit industry median (balanced sample) in 2007. Capital int.<sub>j</sub> is an indicator equal to one for firms with a capital ratio (fixed assets/employees) above the 2-digit industry median (balanced sample) in 2007. All regressions include firm FE, industry-year FE, and the business cycle exposure controls in Table 2. Standard errors in parenthesis are clustered by firm. \*\*\* p<0.01, \*\* p<0.05, \* p<0.1

	$\log \mathrm{CO}_2$	$\log \mathrm{CO_2/VA}$	log Revenue	log Employment	log Capital	log VA
	(1)	(2)	(3)	(4)	(5)	(6)
Low emission & Labor int	-0.336**	-0.368**	-0.040	0.009	-0.145	-0.013
	(0.168)	(0.179)	(0.050)	(0.042)	(0.112)	(0.058)
Low emission & Capital int	-0.130	-0.083	-0.023	0.029	-0.068	-0.031
	(0.164)	(0.171)	(0.057)	(0.042)	(0.097)	(0.055)
High emission & Labor int	-0.588***	-0.466***	-0.125**	-0.125**	-0.102	-0.124**
	(0.145)	(0.151)	(0.053)	(0.050)	(0.142)	(0.062)
High emission & Capital int	-0.382**	-0.181	-0.109*	-0.104**	-0.025	-0.142**
	(0.163)	(0.162)	(0.059)	(0.048)	(0.132)	(0.065)

Table 4: Linear combinations of estimated treatment effects from Table 3. Standard errors in parentheses are calculated using the Delta method. \*\*\* p<0.01, \*\* p<0.05, \* p<0.1

Worker heterogeneity In this part, I explore heterogeneous impacts of the employment effects for different types of workers. Table 5 presents the effects on employment for different educational categories, which are 'No high school', 'High school', and 'Above high school'. The latter category is divided further into STEM (Science, Technology, Engineering and Mathematics) and non-STEM degrees. Panel A shows the result for all workers in the sample. I find that the negative effect on average employment from previous sections is entirely driven by a reduction in the number of workers without a high school degree, with a significant point estimate of -0.165. Workers in the other educational groups are not significantly affected. I neither observe any difference among highly educated workers between STEM and non-STEM degrees.

To investigate to the extent the negative effect on low education workers is driven by part-time workers, I make further sample restrictions in Panel B. As I do not observe worked hours, I use the notion that, due to the relatively compressed wage distribution, few workers

<sup>&</sup>lt;sup>11</sup>The workers are categorized by their highest obtained degree.

in Sweden that earns less than 60% of the national median income are full-time workers. I therefore focus on workers with annual incomes above this threshold, and run the same regressions.<sup>12</sup> The result shows a slightly smaller point estimate of -0.139, which is still significant at 1%. Part-time workers are hence unlikely to be driving the negative impacts on low-educated workers.

Next, I present further results on the heterogeneity with respect to gender and age. In Table 6, I disaggregate the result on employment for males and females, and four different age brackets, within each educational category. The result again shows that workers with a high school degree or above are unaffected by the policy. The disaggregated heterogeneity analysis reveals that the negative employment effects for low-educated workers are largely driven by a reduction in the number of male workers, and workers between 40-64 years old.

Finally, I investigate the margins of adjustments behind the negative effects on employment, by looking separately at changes in hiring and separation rates. The result from the event-study estimation is presented in Figure 8, which shows the effects of the policy change on hiring and separation rates for each educational category at the firm. The negative point estimates for hiring rates after 2011 and close-to-zero estimates for separation rates for workers without a high school degree suggests that firms are adjusting their labor force by reducing their hiring rate. However, the statistical uncertainty in this analysis is very large.

### 6 Discussion and Conclusion

In this paper, I provide new empirical evidence on the impacts of climate policy on firm performance and labor market inequality, using a novel methodology and rich administrative datasets. I find that more stringent climate policy, induced by the removal of carbon tax refunds, significantly reduced emissions among Swedish manufacturing firms. I also find negative effects on revenue, value added, and employment among emission-intensive firms. While the negative impacts on firms' economic performance are consistent with a higher marginal cost of production due to a higher net-of-tax price of fuels (e.g. through higher fos-

<sup>&</sup>lt;sup>12</sup>As median incomes are not published for 2008 by the statistical authorities, I use the mean income, which over-excludes workers.

	log Employment								
	Above high								
	No high school	High school	All	STEM					
	(1)	(2)	(3)	(4)					
		Panel A: All we	orkers						
$D \times Post$	-0.165***	-0.024	-0.050	-0.045					
	(0.041)	(0.027)	(0.039)	(0.041)					
Observations	2,264	2,326	2,134	1,934					
		Panel B: Full-time	$workers^{\dagger}$						
$D \times Post$	-0.139***	-0.032	-0.041	-0.053					
	(0.040)	(0.027)	(0.039)	(0.042)					
Observations	2,176	2,324	2,078	1,866					
$\eta_j$	<b>√</b>	✓	✓	<b>√</b>					
$Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$					
$\Gamma_I \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$					
$EX'_j \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$					
$L'_j \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$					
$K_j \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$					

Table 5: Long difference results on employment by education using a balanced panel of firms. 'STEM' includes workers with a higher education (above high school) in Science, Technology, Engineering and Mathematics. Standard errors in parenthesis are clustered by firm. \*\*\* p<0.01, \*\* p<0.05, \* p<0.1.

 $<sup>^{\</sup>dagger}$ Excluding workers with incomes below 60% of mean income.

	log Employment								
					Age				
	Male	Female	16 - 29	30 - 39	40 - 49	50 - 64			
	(1)	(2)	(3)	(4)	(5)	(6)			
			Panel A: I	No high school					
$D \times Post$	-0.172***	-0.071	-0.063	0.013	-0.146**	-0.127***			
	(0.042)	(0.060)	(0.077)	(0.073)	(0.065)	(0.045)			
Observations	2,216	1,380	1,344	1,150	1,358	2,032			
			Panel B:	High school					
$D \times Post$	-0.027	-0.040	-0.085	-0.012	-0.010	-0.041			
	(0.028)	(0.043)	(0.057)	(0.052)	(0.046)	(0.041)			
Observations	2,324	2,062	2,036	2,104	2,232	2,242			
			Panel C: Ab	oove high schoo	ol				
$D \times Post$	-0.036	-0.040	-0.080	0.047	-0.085	-0.047			
	(0.041)	(0.053)	(0.074)	(0.061)	(0.057)	(0.053)			
Observations	2,062	1,490	1,034	1,510	1,524	1,576			
$\eta_j$	✓	✓	✓	<b>√</b>	<b>√</b>	<b>√</b>			
$Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$			
$\Gamma_I \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$			
$EX_j' \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$			
$L'_j \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$			
$K_j \times Post_t$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$	$\checkmark$			

Table 6: Long difference results on employment using a balanced panel of firms. Standard errors in parenthesis are clustered by firm. \*\*\* p<0.01, \*\* p<0.05, \* p<0.1

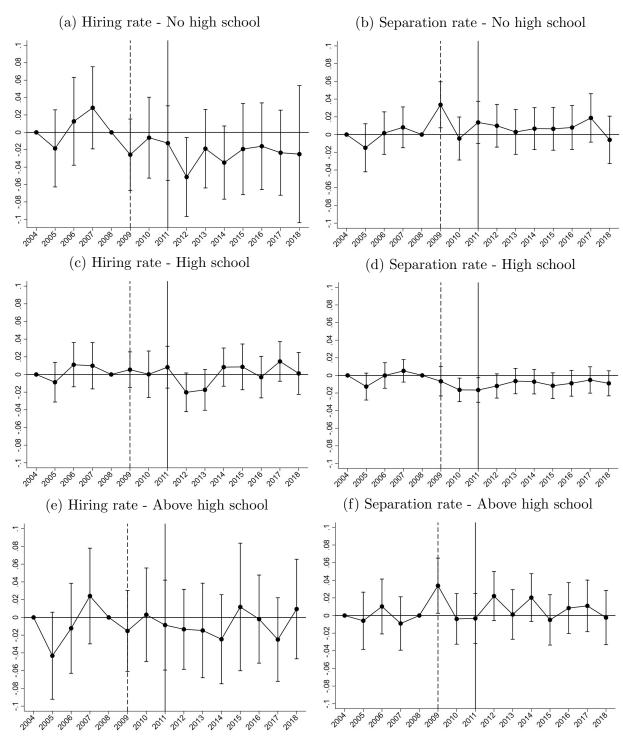


Figure 8: Event study results on labor turnover using a balanced panel of firms. Regressions includes industry-by-year fixed effects, and standard errors are clustered by firm. Capped spikes show 95% confidence intervals.

sil fuel prices or a switch to more expensive biofuels), they are in contrast to previous studies on carbon pricing (Colmer et al., 2024; Dechezleprêtre et al., 2023; Marin et al., 2018; Martin et al., 2014). There are several potential explanations behind this result. First, the reform studied in this paper led to a considerable increase in effective tax rates, by approximately 88 euros per ton CO<sub>2</sub> between 2010 and 2018. This increase can be compared to permit price fluctuations within the EU ETS, which is the most studied policy in the literature, that varied between 0 and 30 euros between 2005 and 2015 (the implementation year and the latest year included in related research, respectively) (Dechezleprêtre et al., 2023). Second, the initial phases of the EU ETS were characterized by free allocation of emission allowances (instead of auctioning), leading to potentially large windfall profits among over-allocated firms (Ellerman et al., 2016), thus mitigating negative output effects. Third, the selection of firms into different forms of regulation raises the point of heterogeneous treatment effects across firm characteristics. The industrial firms covered by the EU ETS are substantially larger and more energy-intensive then the firms regulated by the Swedish carbon tax. These firms may have different financial and technological constraints, and therefore respond differently to carbon pricing.

The skill-biased effect on employment against low-educated workers is in line with the (scant) previous literature analyzing heterogeneous climate policy impacts (Yamazaki, 2017, 2019; Yip, 2018), and lends support to the notion that the impacts of the green transition share similarities with those of general technological change (Marin and Vona, 2019). Previous research has linked automation and technology upgrading with increasing inequality between high- and low-skill workers (Akerman et al., 2015; Autor, 2019; Graetz and Michaels, 2018). It is, however, important to note the potentially different mechanisms behind the skill-bias in, for example, automation and carbon taxation, where the former is characterized by market-driven, productivity-enhancing (at the firm-level) task displacement and a potentially negative substitution effect between new technology and low-skill workers (Acemoglu and Restrepo, 2022). It is possible that the estimated skill-bias in this paper is channeled by a combination of a negative output effect and ambiguous substitution effects, depending on the induced behavior among regulated firms and their possibilities to adopt new technologies. The existence of a dominating output effect would suggest different relative impacts on high- versus low-educated workers under, for example, green subsidies (Popp et al., 2021).

The lower labor demand for workers without a high school degree, whose unemployment rates are exceptionally high (Statistics Sweden, 2024), highlights the importance of re-skilling the workforce to mitigate undesired distributional impacts (European Commission, ndb). However, to fully understand the transitional costs of the green transition, one must be able to observe the impacts on individuals' career trajectories. These costs will depend on the extent to which individuals are reallocated to new sectors in which their skills are less compatible (Walker, 2013), and their ability to move to expanding, green firms (Curtis et al., 2024; Weber, 2020). These issues warrant further research.

# 7 Appendix

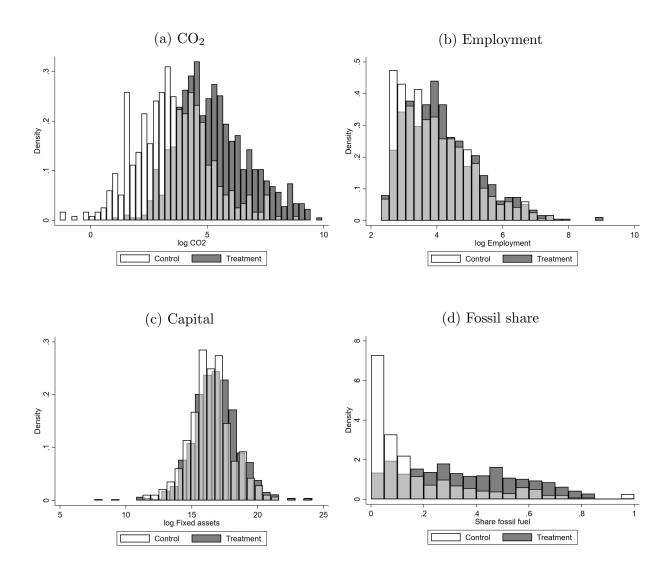


Figure 9: 2008 distribution by treatment status

### References

- Acemoglu, D. and Restrepo, P. (2022). Tasks, automation, and the rise in us wage inequality. *Econometrica*, 90(5):1973–2016.
- Ahmadi, Y., Yamazaki, A., and Kabore, P. (2022). How do carbon taxes affect emissions? plant-level evidence from manufacturing. *Environmental and Resource Economics*, 82(2):285–325.
- Akerman, A., Gaarder, I., and Mogstad, M. (2015). The skill complementarity of broadband internet. The Quarterly Journal of Economics, 130(4):1781–1824.
- Andersson, J. J. (2019). Carbon taxes and CO<sub>2</sub> emissions: Sweden as a case study. *American Economic Journal: Economic Policy*, 11(4):1–30.
- Apostel, A. and Barslund, M. (2024). Measuring and characterising green jobs: A literature review. *Energy Research & Social Science*, 111:103477.
- Autor, D. H. (2019). Work of the past, work of the future. In *AEA Papers and Proceedings*, volume 109, pages 1–32. American Economic Association 2014 Broadway, Suite 305, Nashville, TN 37203.
- Berman, E. and Bui, L. T. (2001). Environmental Regulation and Productivity: Evidence from Oil Refineries. *Review of Economics and Statistics*, 83(3):498–510.
- Bowen, A., Kuralbayeva, K., and Tipoe, E. L. (2018). Characterising green employment: The impacts of 'greening'on workforce composition. *Energy Economics*, 72:263–275.
- Brännlund, R. and Lundgren, T. (2010). Environmental policy and profitability: evidence from swedish industry. *Environmental Economics and Policy Studies*, 12:59–78.
- Brännlund, R., Lundgren, T., and Marklund, P.-O. (2014). Carbon intensity in production and the effects of climate policy—Evidence from Swedish industry. *Energy Policy*, 67:844–857.

- Callaway, B., Goodman-Bacon, A., and Sant'Anna, P. H. (2024). Difference-in-differences with a continuous treatment. Technical report, National Bureau of Economic Research.
- Cole, M. A. and Elliott, R. J. (2007). Do environmental regulations cost jobs? an industry-level analysis of the uk. The BE Journal of Economic Analysis & Policy, 7(1).
- Colmer, J., Martin, R., Muûls, M., and Wagner, U. J. (2024). Does pricing carbon mitigate climate change? firm-level evidence from the european union emissions trading system. *Review of Economic Studies*, page rdae055.
- Curtis, E. M. (2018). Who Loses Under Cap-and-Trade Programs? The Labor Market Effects of the NOx Budget Trading Program. Review of Economics and Statistics, 100(1):151–166.
- Curtis, E. M. and Marinescu, I. (2022). Green energy jobs in the US: What are they, and where are they? Technical report, National Bureau of Economic Research.
- Curtis, E. M., O'Kane, L., and Park, R. J. (2024). Workers and the green-energy transition: Evidence from 300 million job transitions. *Environmental and Energy Policy and the Economy*, 5(1):127–161.
- Dechezleprêtre, A., Nachtigall, D., and Venmans, F. (2023). The joint impact of the European Union emissions trading system on carbon emissions and economic performance. Journal of Environmental Economics and Management, 118:102758.
- Ellerman, A. D., Marcantonini, C., and Zaklan, A. (2016). The european union emissions trading system: ten years and counting. *Review of Environmental Economics and Policy*.
- Elliott, R. J., Kuai, W., Maddison, D., and Ozgen, C. (2024). Eco-innovation and (green) employment: A task-based approach to measuring the composition of work in firms. *Journal of Environmental Economics and Management*, page 103015.
- European Commission (n.d.a). ETS2: buildings, road transport and additional sectors. https://climate.ec.europa.eu/eu-action/eu-emissions-trading-system-eu-ets/ets2-buildings-road-transport-and-additional-sectors\_en.

- European Commission (n.d.b). The Just Transition Mechanism: making sure no one is left behind. https://commission.europa.eu/strategy-and-policy/priorities-2019-2024/european-green-deal/finance-and-green-deal/just-transition-mechanism\_en.
- Eurostat (2024). Businesses in the manufacturing sector. https://ec.europa.eu/eurostat/statistics-explained/index.php?title=Manufacturing\_statistics\_-\_ NACE\_Rev.\_2&oldid=502915.
- Ferris, A. E., Shadbegian, R. J., and Wolverton, A. (2014). The Effect of Environmental Regulation on Power Sector Employment: Phase I of the Title IV SO2 Trading Program. Journal of the Association of Environmental and Resource Economists, 1(4):521–553.
- Government Bill 2008/09:162 (2009). En sammanhållen klimat- och energipolitik. https://www.regeringen.se/contentassets/cf41d449d2a047049d7a34f0e23539ee/ensammanhallen-klimat-och-energipolitik—klimat-prop.-200809162.
- Government Bill 2014/15:1 (2014). Budgetpropositionen för 2015. https://www.regeringen.se/contentassets/f479a257aa694bf097a3806bbdf6ff19/forslagtill-statens-budget-for-2015-finansplan-och-skattefragor-kapitel-1-7/.
- Graetz, G. and Michaels, G. (2018). Robots at work. Review of Economics and Statistics, 100(5):753–768.
- Gray, W. B., Shadbegian, R., and Wolverton, A. (2023). Environmental regulation and labor demand: What does the evidence tell us? *Annual Review of Resource Economics*, 15(1):177–197.
- Gray, W. B., Shadbegian, R. J., Wang, C., and Meral, M. (2014). Do epa regulations affect labor demand? evidence from the pulp and paper industry. *Journal of Environmental Economics and Management*, 68(1):188–202.
- Greenstone, M. (2002). The Impacts of Environmental Regulations on Industrial Activity: Evidence from the 1970 and 1977 Clean Air Act Amendments and the Census of Manufactures. *Journal of Political Economy*, 110(6):1175–1219.

- Hammar, H. and Åkerfeldt, S. (2011). CO<sub>2</sub> Taxation in Sweden 20 Years of Experience and Looking Ahead.
- Jaraite, J., Kazukauskas, A., and Lundgren, T. (2014). The effects of climate policy on environmental expenditure and investment: Evidence from Sweden. *Journal of Environmental Economics and Policy*, 3(2):148–166.
- Jaraitė, J. and Maria, C. D. (2016). Did the eu ets make a difference? an empirical assessment using lithuanian firm-level data. *The Energy Journal*, 37(2):68–92.
- Leroutier, M. (2022). Carbon pricing and power sector decarbonization: Evidence from the UK. *Journal of Environmental Economics and Management*, 111:102580.
- Liu, M., Shadbegian, R., and Zhang, B. (2017). Does environmental regulation affect labor demand in china? evidence from the textile printing and dyeing industry. *Journal of Environmental Economics and Management*, 86:277–294.
- Liu, M., Tan, R., and Zhang, B. (2021). The Costs of "Blue Sky": Environmental Regulation, Technology Upgrading, and Labor Demand in China. *Journal of Development Economics*, 150:102610.
- Marin, G., Marino, M., and Pellegrin, C. (2018). The impact of the European Emission Trading Scheme on multiple measures of economic performance. *Environmental and Resource Economics*, 71(2):551–582.
- Marin, G. and Vona, F. (2019). Climate policies and skill-biased employment dynamics: Evidence from EU countries. *Journal of Environmental Economics and Management*, 98:102253.
- Marin, G. and Vona, F. (2021). The impact of energy prices on socioeconomic and environmental performance: Evidence from french manufacturing establishments, 1997–2015. European Economic Review, 135:103739.
- Martin, R., De Preux, L. B., and Wagner, U. J. (2014). The Impact of a Carbon Tax on Manufacturing: Evidence from Microdata. *Journal of Public Economics*, 117:1–14.

- Martinsson, G., Sajtos, L., Strömberg, P., and Thomann, C. (2024). The effect of carbon pricing on firm emissions: Evidence from the swedish co2 tax. *The Review of Financial Studies*, 37(6):1848–1886.
- Morgenstern, R. D., Pizer, W. A., and Shih, J.-S. (2002). Jobs versus the environment: an industry-level perspective. *Journal of environmental economics and management*, 43(3):412–436.
- Popp, D., Marin, G., Vona, F., and Chen, Z. (2021). The employment impact of a green fiscal push: Evidence from the american recovery and reinvestment act. *Brookings Papers on Economic Activity*, 2021:1–69.
- Pretis, F. (2022). Does a carbon tax reduce co2 emissions? evidence from british columbia. Environmental and Resource Economics, 83(1):115–144.
- Ryner, E. (2022). KI-kommentar: Energi- och miljöskatter i Sverige och internationellt. Konjunkturinstitutet.
- Saussay, A., Sato, M., Vona, F., and O'Kane, L. (2022). Who's fit for the low-carbon transition? emerging skills and wage gaps in job and data.
- SFS 1994:1776 (n.d.). Act on Excise Duties on Energy (Lag (1994:1776) om skatt på energi). [Accessed: 2024-09-11].
- Sheriff, G., Ferris, A. E., and Shadbegian, R. J. (2019). How Did Air Quality Standards Affect Employment at US Power Plants? The Importance of Timing, Geography, and Stringency. Journal of the Association of Environmental and Resource Economists, 6(1):111–149.
- Statistics Sweden (2024). Population aged 15-74 (LFS) by labour status, level of education and sex. Year 2005 2023.
- UNDP (2022). What is just transition? And why is it important? .
- Vona, F. (2019). Job losses and political acceptability of climate policies: why the 'job-killing'argument is so persistent and how to overturn it. *Climate Policy*, 19(4):524–532.

- Vona, F., Marin, G., and Consoli, D. (2019). Measures, drivers and effects of green employment: evidence from us local labor markets, 2006–2014. Journal of Economic Geography, 19(5):1021–1048.
- Vona, F., Marin, G., Consoli, D., and Popp, D. (2018). Environmental regulation and green skills: An empirical exploration. *Journal of the Association of Environmental and Resource Economists*, 5(4):713–753.
- Vrolijk, K. and Sato, M. (2023). Quasi-experimental evidence on carbon pricing. *The World Bank Research Observer*, 38(2):213–248.
- Walker, W. R. (2011). Environmental Regulation and Labor Reallocation: Evidence from the Clean Air Act. *American Economic Review*, 101(3):442–47.
- Walker, W. R. (2013). The Transitional Costs of Sectoral Reallocation: Evidence From the Clean Air Act and the Workforce. *The Quarterly Journal of Economics*, 128(4):1787–1835.
- Weber, J. G. (2020). How should we think about environmental policy and jobs? an analogy with trade policy and an illustration from us coal mining. Review of Environmental Economics and Policy.
- World Bank (2024). State and Trends of Carbon Pricing.
- Yamazaki, A. (2017). Jobs and Climate Policy: Evidence from British Columbia's Revenue-Neutral Carbon Tax. *Journal of Environmental Economics and Management*, 83:197–216.
- Yamazaki, A. (2019). Who bears more burdens of carbon taxes? heterogeneous employment effects within manufacturing plants.
- Yip, C. M. (2018). On the Labor Market Consequences of Environmental Taxes. *Journal of Environmental Economics and Management*, 89:136–152.