# EuPPollNet: A European database of plant-pollinator networks

 Jose B. Lanuza $^{1,2,3}\vert$  Tiffany M. Knight $^{3,2,4}\vert$ Nerea Montes-Perez $^1\vert$  Will Glenny $^{3,4}\vert$  Paola Acuña <sup>5</sup> | Matthias Albrecht <sup>6</sup> | Maddi Artamendi <sup>7,8</sup> | Isabelle Badenhausser <sup>9,10,11</sup> | Joanne M. Bennett <sup>12</sup> | Paolo Biella <sup>13</sup> | Ricardo Bommarco <sup>14</sup> | Andree Cappellari <sup>15</sup> | Sílvia Castro <sup>16</sup> | Yann Clough  $^{17}\vert$  Pau Colom $^{18,19}\vert$  Joana Costa $^{16,20}\vert$  Nathan Cyrille $^{21,22}\vert$  Natasha de Manincor $^{23,24}\vert$ Paula Dominguez-Lapido <sup>7</sup> | Christophe Dominik <sup>4,3</sup> | Yoko L. Dupont <sup>25</sup> | Reinart Feldmann Emeline Felten  $^{22}$ | Victoria Ferrero  $^{27}$ | William Fiordaliso  $^{28}$ | Alessandro Fisogni  $^{23}$ | Úna Fitzpatrick <sup>29</sup> | Marta Galloni <sup>30</sup> | Hugo Gaspar <sup>16</sup> | Elena Gazzea <sup>15</sup> | Irina Goia <sup>31,32</sup> | Carmelo Gómez-Martínez <sup>33</sup> | Miguel A. González-Estévez <sup>33</sup> | Juan Pedro González-Varo <sup>34</sup> | Ingo Grass <sup>35</sup> Jiří Hadrava <sup>36</sup> Nina Hautekèete <sup>23</sup> Veronica Hederström <sup>17</sup> Ruben Heleno <sup>16</sup> Sandra Hervias-Parejo <sup>33</sup> Jonna M. Heuschele <sup>3,4</sup> Bernhard Hoiss <sup>37</sup> Andrea Holzschuh <sup>37</sup> Sebastian Hopfenmüller <sup>38</sup> José M. Iriondo <sup>39</sup> Birgit Jauker <sup>40</sup> Frank Jauker <sup>41</sup> Jana Jersáková <sup>42</sup> Katharina Kallnik <sup>37</sup> Reet Karise <sup>43</sup> David Kleijn <sup>44</sup> Stefan Klotz <sup>4</sup> Theresia Krausl <sup>17</sup> Elisabeth Kühn  $^{45}$ | Carlos Lara-Romero  $^{39}$ | Michelle Larkin  $^{46}$ | Emilien Laurent  $^{22}$ | Amparo Lázaro <sup>33</sup> Felipe Librán-Embid <sup>47,48</sup> Yicong Liu <sup>4,2</sup> Sara Lopes <sup>16</sup> Francisco López-Núñez  $^{16,49}|$  João Loureiro $^{16}|$  Ainhoa Magrach $^{7,50}|$  Marika Mänd $^{43}|$  Lorenzo Marini $^{15}|$  Rafel Beltran Mas <sup>33</sup> | François Massol <sup>51</sup> | Corina Maurer <sup>6</sup> | Denis Michez <sup>24</sup> | Francisco P. Molina <sup>1</sup> | Javier Morente-López  $^{52}$ | Sarah Mullen  $^{53}$ | Georgios Nakas  $^{54}$ | Lena Neuenkamp  $^{55,56}$ | Arkadiusz Nowak <sup>57,58</sup> | Catherine J. O'Connor <sup>16,59</sup> | Aoife O'Rourke <sup>53</sup> | Erik Öckinger <sup>14</sup> | Jens M. Olesen  $^{60}$ | Øystein H. Opedal  $^{61}$ | Theodora Petanidou  $^{54}$ | Yves Piquot  $^{23}$ | Simon G. Potts  $^{62}$ | Eileen F. Power <sup>63</sup> Willem Proesmans <sup>24,22</sup> Demetra Rakosy <sup>4,3,64</sup> Sara Reverté <sup>24</sup> Stuart P. M. Roberts  $^{62}$ l Maj Rundlöf  $^{65}$ l Laura Russo  $^{66,53}$ l Bertrand Schatz  $^{67}$ l Jeroen Scheper  $^{44}$ l Oliver Schweiger  $^{4,3}|$  Pau Enric Serra $^{33}|$  Catarina Siopa $^{16}|$  Henrik G. Smith  $^{65,17}|$  Dara Stanley  $^{68}|$  Valentin Ştefan <sup>4,3</sup> | Ingolf Steffan-Dewenter <sup>37</sup> | Jane C. Stout <sup>63</sup> | Louis Sutter <sup>69</sup> | Elena Motivans Švara <sup>3,4,2</sup>| Sebastian Świerszcz <sup>57,70</sup>| Amibeth Thompson <sup>2,3,71</sup>| Anna Traveset <sup>33</sup>| Annette Trefflich <sup>72</sup> Robert Tropek <sup>73,74</sup> Teja Tscharntke <sup>48</sup> Adam J. Vanbergen <sup>22</sup> Montserrat Vilà <sup>1,75</sup> Ante Vujić <sup>76</sup> | Cian White <sup>53</sup> | Jennifer B. Wickens <sup>62</sup> | Victoria B. Wickens <sup>62</sup> | Marie Winsa <sup>14</sup> | Leana Zoller <sup>2,3,77</sup> Ignasi Bartomeus <sup>1</sup>

# 1 Corresponding author= barragansljose@gmail.com

- All authors excluding the first four and last are ordered alphabetically
- <sup>1</sup> Doñana Biological Station (EBD-CSIC), Seville, Spain, <sup>2</sup> Martin Luther University Halle-Wittenberg,
- Institute of Biology, Halle, Germany, <sup>3</sup> German Centre for Integrative Biodiversity Research (iDiv)
- 35 Halle-Jena-Leipzig, <sup>4</sup> Department of Community Ecology, Helmholtz Centre for Environmental

Research - UFZ, Halle, Germany, <sup>5</sup> Department of Plant Biology, Faculty of Science, University of Vigo, Vigo, Spain, <sup>6</sup> Agroecology and Environment, Agroscope, Zürich, Switzerland, <sup>7</sup> Basque Centre for Climate Change-BC3, Leioa, Spain, <sup>8</sup> University of the Basque Country, EuskalHerriko Unibertsitatea (UPV-EHU), Leioa, Spain, <sup>9</sup> Centre of Biological Studies of Chizé, La Rochelle 39 University, Villiers en Bois, France, <sup>10</sup> LTSER "ZA Plaine & Val de Sèvre", CNRS, Villiers en Bois, France, <sup>11</sup> Multidisciplinary Research Unit for Grasslands and Forage Crops, INRAE, Lusignan, France, 41 <sup>12</sup> Fenner School of Environment & Society, The Australian National University, Canberra, Australia, <sup>13</sup> ZooPlantLab, Department of Biotechnology and Biosciences, University of Milano-Bicocca, Milan, 43 Italy, <sup>14</sup> Department of Ecology, Swedish University of Agricultural Sciences, Uppsala, Sweden, <sup>15</sup> Department of Agronomy, Food, Natural Resources, Animals and Environment, University of Padua, 45 Padua, Italy, <sup>16</sup> Department of Life Sciences, Centre for Functional Ecology, University of Coimbra, Coimbra, Portugal, <sup>17</sup> Centre for Environmental and Climate Science, Lund University, Lund, 47 Sweden, <sup>18</sup> Department of Evolutionary Biology, Ecology, and Environmental Sciences, University of Barcelona, Barcelona, Spain, <sup>19</sup> Biodiversity Research Institute (IRBio), Barcelona, Spain, <sup>20</sup> Linking 49 Landscape, Environment, Agriculture and Food, School of Agriculture, University of Lisbon, Portugal, 50 <sup>21</sup> Biogéosciences, UMR 6282 CNRS, University of Burgundy, Dijon, France, <sup>22</sup> Agroecology, INRAE, 51 Institut Agro, University of Burgundy, University of Burgundy Franche-Comté, Dijon, France, <sup>23</sup> Univ. Lille, CNRS, UMR 8198 - Evo-Eco-Paleo, F-59000 Lille, France, <sup>24</sup> Laboratory of Zoology, Research 53 Institute of Biosciences, University of Mons, Mons, Belgium, <sup>25</sup> Department of Agroecology, Aarhus 54 University, Denmark, <sup>26</sup> Helmholtz Centre for Environmental Research - UFZ, Leipzig, Germany, 55 <sup>27</sup> Department of Biodiversity and Environmental Management, University of León, León, Spain, <sup>28</sup> 56 Ecology of Interactions and Global Change, Research Institute in Biosciences, University of Mons, Mons, Belgium, <sup>29</sup> National Biodiversity Data Centre, County Waterford, Ireland, <sup>30</sup> Department of 58 Biological, Geological and Environmental Sciences (BiGeA), University of Bologna, Bologna, Italy, 59 Faculty of Biology and Geology, Babes-Bolyai University, Cluj-Napoca, Romania, <sup>32</sup> Centre 60 for Systems Biology, Biodiversity and Bioresources (3B), Babeş-Bolyai University, Cluj-Napoca, Romania, <sup>33</sup> Mediterranean Institute for Advanced Studies (IMEDEA, UIB-CSIC), Esporles, Spain, 62 <sup>34</sup> Department of Biology, Institute of Marine Research (INMAR), University of Cádiz, Puerto Real, Spain, <sup>35</sup> Department of Ecology of Tropical Agricultural Systems, University of Hohenheim, Stuttgart, 64 Germany, <sup>36</sup> Department of Zoology, Faculty of Science, Charles University, Prague, Czechia, <sup>37</sup> Department of Animal Ecology and Tropical Biology, Biocenter, University of Würzburg, Würzburg, Germany, <sup>38</sup> Cultural Landscape Günztal Foundation, Ottobeuren, Germany, <sup>39</sup> Global Change Research Institute (IICG-URJC), Rey Juan Carlos University, Madrid, Spain, 40 Justus Liebig University Giessen, Giessen, Germany, 41 Institute of Landscape Ecology and Resource Management, 69 Justus Liebig University Giessen, Giessen, Germany, 42 Department of Ecosystems Biology, Faculty 70 of Science, University of South Bohemia, České Budějovice, Czechia, <sup>43</sup> Institute of Agricultural and 71 Environmental Sciences, Estonian University of Life Sciences, Tartu, Estonia, 44 Plant Ecology and 72 Nature Conservation Group, Wageningen University, Wageningen, The Netherlands, <sup>45</sup> Department 73 of Conservation Biology & Social-Ecological Systems, Helmholtz Centre for Environmental Research UFZ, Halle, Germany, <sup>46</sup> Botany and Plant Science, School of Natural Sciences and Ryan Institute, 75 University of Galway, Galway, Ireland, <sup>47</sup> Institute of Animal Ecology and Systematics, Justus Liebig 76 University of Gießen, Germany, 48 Agroecology, University of Göttingen, Göttingen, Germany, 49 77 Research Centre for Natural Resources Environment and Society (CERNAS), Polytechnic Institute of Coimbra, Coimbra Agriculture School, Coimbra, Portugal, <sup>50</sup> IKERBASQUE, Basque Foundation 79 forScience, Bilbao, Spain, <sup>51</sup> Univ. Lille, CNRS, Inserm, CHU Lille, Institut Pasteur de Lille, U1019 UMR 9017 - CIIL - Center for Infection and Immunity of Lille, F-59000 Lille, France, <sup>52</sup> Plant 81 Evolutionary Ecology, Institute of Ecology, Evolution and Diversity, Faculty of Biological Sciences, Goethe University Frankfurt, Max-von-Laue-Str. 13, 60438, Frankfurt am Main, Germany. 53 Botany

Department, Trinity College Dublin, Dublin, Ireland,  $^{54}$  Laboratory of Biogeography & Ecology, Department of Geography, University of the Aegean, Mytilene, Greece, <sup>55</sup> Department of Botany, Institute of Ecology and Earth Sciences, University of Tartu, Tartu, Estonia, <sup>56</sup> Institute of Landscape Ecology, Münster University, Münster, Germany, <sup>57</sup> Polish Academy of Sciences Botanical Garden, Center for Biological Diversity Conservation in Powsin, Warsaw, Poland, <sup>58</sup> Botanical Garden of the Wrocław University, Wrocław, Poland, <sup>59</sup> Cardif School of Biosciences, Cardif University, Cardif, UK, 89 <sup>60</sup> Department of Biology, University of Aarhus, Aarhus, Denmark, <sup>61</sup> Division of Biodiversity and Evolution, Department of Biology, Lund University, Lund, Sweden, <sup>62</sup> Centre for Agri-Environmental 91 Research, School of Agriculture, Policy and Development, University of Reading, Reading, UK, <sup>63</sup> Botany, School of Natural Sciences, Trinity College Dublin, Dublin, Ireland, <sup>64</sup> Thünen-Institute of 93 Biodiversity, Braunschweig, Germany, <sup>65</sup> Department of Biology, Lund University, Lund, Sweden, <sup>66</sup> Department of Ecology and Evolutionary Biology, University of Tennessee, Knoxville, TN, 95 USA, <sup>67</sup> CEFE, CNRS, University of Montpellier, EPHE, IRD, Montpellier, France, <sup>68</sup> School of Agriculture and Food Science, University College Dublin, Dublin, Ireland, <sup>69</sup> Plant Productions Systems, Agroscope, Zürich, Switzerland, <sup>70</sup> Institute of Agroecology and Plant Production, Wrocław University of Environmental and Life Sciences, Wrocław, Poland, <sup>71</sup> University of Freiburg, Chair of Nature Conservation and Landscape Ecology, Freiburg, Germany, <sup>72</sup> State Institute of Agriculture and Horticulture Saxony-Anhalt, Bernburg, Germany, <sup>73</sup> Institute of Entomology, Biology Centre, Czech Academy of Sciences, České Budějovice, Czechia, <sup>74</sup> Department of Ecology, Faculty of Science, 101 102 Charles University, Prague, Czechia, 75 Department of Plant Biology and Ecology, University of 103 Seville, Seville, Spain, <sup>76</sup> Department of Biology and Ecology, Faculty of Sciences, University of Novi 104 Sad, Novi Sad, Serbia, 77 Department of Ecology & Evolutionary Biology, University of Colorado, 105 Boulder, CO, USA 106 107

# **Abstract**

127

128

129

130

131

132

**Motivation:** Pollinators play a crucial role in maintaining Earth's terrestrial biodiversity and human food production by mediating sexual reproduction for most flowering plants. In-110 deed, the network of interactions formed by plants and pollinators constitutes the backbone 111 of plant-pollinator community stability and functioning. However, rapid human-induced envi-112 ronmental changes are compromising the long-term persistence of plant-pollinator interaction 113 networks. One of the major challenges for pollinator conservation is the lack of robust gen-114 eralisable data capturing how plant-pollinator communities are structured across space and 115 time. Here, we present the EuPPollNet (European Plant-Pollinator Networks) database, a 116 fully open and reproducible European-level database containing harmonized taxonomic data 117 on plant-pollinator interactions referenced in both space and time, along with other ecolog-118 ical variables of interest. This database offers an open workflow that allows researchers to 119 track data-curation decisions and edit them according to their preferences. We present the 120 taxonomic and sampling coverage of EuPPollNet, and summarize key structural properties 121 in plant-pollinator networks. We hope EuPPollNet will stimulate future research that fills 122 the taxonomic, ecological, and geographical data gaps on plant-pollinator interactions that we 123 have identified. Further, the variation in the structure of the networks in EuPPollNet pro-124 vides a strong basis for future studies aimed at quantifying drivers of plant-pollinator network 125 change and guiding future conservation planning for plants and pollinators. 126

Main Types of Variables Included: EuPPollNet contains 1,162,913 interactions between plants and pollinators from 1,864 distinct networks (i.e., distinct sampling event in space or time), which belong to 54 different studies distributed across 23 European countries. In addition, information about sampling methodology, habitat type, bio-climatic region, and further taxonomic rank information for both plant and pollinator species are also provided (i.e., genus, family and order).

Spatial location and grain: The database contains 1,214 different sampling locations from 134 different natural and anthropogenic habitats that fall in 7 different bio-climatic regions. All 135 records are geo-referenced and presented in the World Geodetic System 1984 (WGS84).

Time period and grain: Species interaction data was collected between 2004 and 2021. All records are time-referenced and most of the studies documented interactions within a single flowering season (68.52%).

Major taxa and level of measurement: The database contains interaction data at the species level for 94.39% of the records, including a total of 1,411 plant and 2,223 pollinator species. The database covers 5.56% of the European species of flowering plants, 34.38% of bees, 26.21% of butterflies, and 33.63% of syrphid species at the European level.

Software format: The database was built with the R programming language and is stored as ".rds" and ".csv" formats. The construction of the database is fully reproducible and can be accessed at the following link: https://github.com/JoseBSL/EuPPollNet.git.

# 146 KEYWORDS

Angiosperms, connectance, flowering plants, nestedness, plant-pollinator networks, pollinators, species

148 interactions

# **1 | INTRODUCTION**

161

162

163

164

165

166

167

168

169

170

171

172

173

174

175

176

177

178

179

180

181

182

183

184

185

186

187

188

Plant-pollinator interactions involve a great diversity of species, largely attributed to their co-150 evolutionary history (Ollerton, 2017), and are critically important for terrestrial biodiversity 151 and economic productivity. The synergistic effects of climate change with other global change 152 pressures are threatening worldwide biodiversity (Bellard et al., 2014; Sala et al., 2000), includ-153 ing plant and pollinator species as well as their interactions (Eichenberg et al., 2021; Goulson 154 et al., 2015; Settele et al., 2016). Under this scenario, the increasing availability of biodiversity 155 data plays a major role in our ecological understanding of species status, trends, and con-156 servation (Heberling et al., 2021; Zattara & Aizen, 2021). However, our knowledge of plant 157 and pollinator species and their network of interactions still exhibits major temporal, spatial 158 and taxonomic biases (Archer et al., 2014; Marshall et al., 2024; Poisot et al., 2021; Troia & 159 McManamay, 2016), limiting our ability to effectively protect their biodiversity. 160

The interactions between different plant and pollinator species within a community form complex networks. Macro-ecological analyses of the topology of these networks have revealed common properties, such as truncated power-law degree distributions (Jordano et al., 2003) or nestedness (Bascompte et al., 2003). Large-scale analyses across multiple studies can quantify patterns across geographic regions (Olesen & Jordano, 2002; Traveset et al., 2016) or environmental gradients (Ramos-Jiliberto et al., 2010; Rech et al., 2016; Saunders et al., 2023) that cannot be examined in a single study. Although macro-ecological approaches that use ecological interactions make significant contributions to knowledge (Windsor et al., 2023), such synthesis work must consider variation across studies in the spatio-temporal nature of the data (Burkle & Alarcón, 2011; García et al., 2024). For instance, plant-pollinator studies tend to differ in sampling effort and methodology which affect the structure of the resulting plantpollinator networks (Gibson et al., 2011; Jordano, 2016; Schwarz et al., 2020). Most plantpollinator networks have unobserved interactions (Chacoff et al., 2012; Olesen et al., 2011), and thus research that attempts to synthesize across published studies must have access to raw data on interactions in order to statistically account for sampling effort and completeness. This emphasizes the importance of providing data in its rawest possible form in datasets that will be utilized for synthesis and macro-ecological studies.

Europe is one of the continents with a larger amount of available biodiversity data (Proença et al., 2017), yet still exhibits major gaps (Bennett et al., 2018; Wetzel et al., 2018). While species checklists need to be treated carefully, especially at a macro-ecological scale (Grenié et al., 2023), the growing number of European plant and pollinator checklists (Reverté et al., 2023), along with occurrence data (Zattara & Aizen, 2021), is setting a foundation for the conservation of its flora and their pollinators. However, species richness is just one component of biodiversity and documenting the interaction between plants and pollinators is essential for understanding biodiversity change (Jordano, 2016). Numerous works have studied plant-pollinator interactions in the last decades, generating thousands of plant-pollinator interaction networks worldwide. Several initiatives have tried to integrate plant-pollinator interaction data into databases such as Mangal (Poisot et al., 2016) or GloBI (Poelen et al., 2014), resulting

in numerous large scale comparative analyses that have enhanced our understanding of the ecology of plants and pollinators (e.g., European wild bee data trends; Marshall et al., 2024). Despite all these resources, Europe lacks accessible harmonized plant-pollinator interaction data that allow researchers to evaluate plant-pollinator interactions at a European level. For example, only over a dozen of European plant-pollinator networks are included in Mangal, while GloBI focuses on pairwise interactions disconnected from the community context. Assembling and curating the existing information on EU plant-pollinator networks will guide research efforts, conservation planning and will set a foundation for future global change research. 

Here, we present the European Plant-Pollinator Networks database (EuPPollNet), which con-tains harmonized information on plant-pollinator interactions at the European level. The pollinator taxonomic groups include the main orders of entomofauna that visit and polli-nate flowering plants in Europe. These comprise insect species from the orders Hymenoptera, Diptera, Lepidoptera, and Coleoptera, accounting for almost the totality of recorded interac-tions in EuPPollNet (99.86%). To understand the scope of the database, we examined the taxonomic and sampling coverage of the different plant and pollinator species at the European level with the help of the most up-to-date species checklists and rarefaction analyses. In addi-tion, for bees and plants, we evaluated if there is a phylogenetic signal in the presence-absence of interaction data. 

EuPPollNet contains one of the largest sets of plant-pollinator networks collated to date, providing a unique opportunity to examine the prevalence of key structural metrics across networks. For example, despite the large theoretical literature on the meaning of a nested structure in plant-pollinator networks (Bascompte & Jordano, 2007; Guimaraes, 2020), where specialists species interact only with subsets of generalists species, this pattern has only been empirically evaluated with a relatively small number of networks (Bascompte et al., 2003; Payrató-Borras et al., 2019; Staniczenko et al., 2013), and is still debated how structural metrics such as connectance and nestedness change across latitudes and bioclimatic regions (Olesen & Jordano, 2002; Song et al., 2017; Trøjelsgaard & Olesen, 2013).

Overall, EuPPollNet aims to cover a wide range of taxonomic groups and habitats, while also providing other variables of interest that define the ecological context and sampling methods of the study. In addition, EuPPollNet offers a transparent and accessible workflow of its data management and species harmonization that allows the database to be reused and to expand over time. This database provides a large number of community-level networks with curated and harmonized data, distinguishing it from other currently available resources that contain plant-pollinator interactions. We expect that EuPPollNet can be used to evaluate macro-ecological processes in plant-pollinator networks, guide conservation planning, and set a baseline for global change research.

# 25 2 | METHODS

# Data acquisition

The EuPPollNet database includes published and unpublished studies compiled initially by a 227 wide number of researchers and institutions within the European continent as defined by the European Environment Agency (Stanners & Bourdeau, 1995). As this database is the result 229 of one of the work packages of the European project Safeguard (Safeguarding European wild 230 pollinators; https://doi.org/10.3030/101003476), first, data was directly requested from mem-231 bers of the Safeguard project in May 2022. Second, the request was extended to data owners 232 outside of the project, with data collection concluding in August 2024. These other data own-233 ers were identified by direct communication with colleagues suggested by Safeguard members 234 and by directly searching for studies on Google Scholar of under-represented regions within 235 the database. While Google Scholar lacks reproducibility (Gusenbauer & Haddaway, 2020), 236 it still remains the most comprehensive search engine to date (Gusenbauer, 2019). This ap-237 proach maximized the potential number of studies that could be incorporated in this database. 238 The search strings used were 'plant-pollinator interactions' and 'plant-pollinator networks'. To 239 maintain high quality standards that will support robust future ecological research, we only 240 included studies that met the following criteria: 1) studies containing time- and geo-referenced 241 records of plant-pollinator interactions; and 2) studies that quantify interactions by document-242 ing the contact between a floral visitor - referred to as a 'pollinator' throughout the manuscript, even though pollination efficiency is not evaluated - and the reproductive structure of a specific 244 sampled plant (i.e., phyto-centric networks). 245

## 246 Dataset description

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

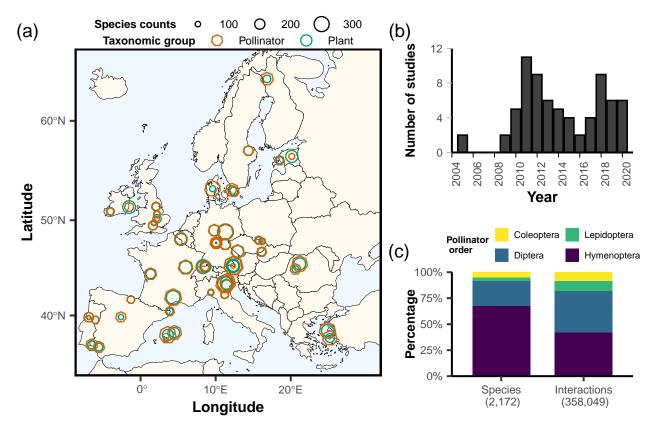
262

The database contains 54 independent published and unpublished studies conducted during the time period 2004 - 2021 in 23 different countries (Figure 1a and Figure 1b; see Figure S1 for exact locations). The studies differ in sampling effort and methodology, and thus documenting sampling methods and sampling effort is an important feature of EuPPollNet. Most studies took place within a single flowering season (68.52%), sampled a given location for an average of 6.99 days, and exclusively sampled diurnal plant-pollinator interactions, with transects being the most common sampling method (64.81%). All the studies documented interactions with Hymenopterans (with 50.00% considering all Hymenopterans, 46.15% only wild bees and 3.85% only bumblebees), 92.31% documented interactions with Dipterans (with 46.15% considering all Dipterans, 46.15% only syrphids and 5.77% recorded syrphids plus bombylids or tachinid flies), 63.46% with Lepidopterans, and 32.69% with Coleopterans. The database includes a total of 1,162,913 distinct interactions. Most of the pollinator species belong to the orders Hymenoptera, Diptera, Lepidoptera and Coleoptera (89.11%), which account for nearly all interactions in this database (99.86%). Species that belong to other orders (10.89%) are not explored in this study as they account for a minor fraction of the total interactions (0.14%). Hymenoptera and Diptera contain the highest number of species,

with each comprising approximately 1,000 species in the database. However, the majority of plant-pollinator interactions are from Hymenoptera species (90.15%; **Figure 1c**). Notably, the western honey bee, *Apis mellifera*, represents 69.89% of the total interaction records from the database and an average of 30.74% of the total interactions per network.

264

265



**Figure 1**. (a) Locations of the studies in EuPPollNet across the European continent, showing the total number of pollinator (i.e., orange heptagon) and plant (i.e., green circles) species per study. The sizes of these shapes are proportional to the respective species counts. For visualization purposes, we have selected only a single location per study. (b) Number of studies by year in EuPPollNet. (c) Proportion of species and interactions across the four main pollinator orders in EuPPollNet, excluding interactions from *Apis mellifera*. The total number of species and interactions is indicated in parentheses at the bottom.

#### 267 Data structure

The EuPPollNet database is available in both .csv and .rds formats and contains a total of 30 columns (Table 1), where each row represents a single interaction between a plant and a 269 pollinator species. These columns include information about the study and network identifiers 270 (columns 1 and 2), sampling method (3), habitat type as described by the author, and a 271 unified habitat classification across studies (4 and 5), bioregion where the network is located 272 (6), country, locality, and latitude-longitude coordinates (7 to 10), date of the interaction 273 (11), number of interactions (12), taxonomic information about plants (13 to 20), taxonomic 274 information about pollinators (21 to 28), and information about the availability of floral count 275 data (29). The flower count data is provided in a separate file (.csv or .rds) and can be 276 merged with the interaction data through the "Flower data merger" column (30). Note 277 that although two-thirds of studies include information on floral abundance, the methods and 278 units vary greatly across studies. To construct a plant-pollinator network matrix within a 279 single flowering season at the site level, users should group interactions by plant and pollinator 280 species, site, study, and year. Finally, metadata at the study level is provided in a separate 281 file, including information about the authors, digital object identifier (if available), sampling 282 time, and taxonomic coverage of the main pollinator groups for each study. 283

#### **Taxonomic harmonization**

285

286

287

288

289

290

291

292

293

294

295

296

297

298

299

300

301

302

303

304

305

306

307

310

311

312

313

316

317

318

319

320

321

All plant and pollinator species names were checked and harmonized in R using rgbif (Chamberlain et al., 2022). The protocol for plants and pollinators is similar but slightly different given the availability of the different taxonomic resources. For transparency, we have included in the database the original species name, the new assigned name, and, if the name of the species is uncertain (e.g., species complex or species alike). In addition, taxonomic information at genus, family and order level was downloaded for each species.

For plants: (i) we initially verified the exact matches against the GBIF species checklist; (ii) we selected unmatched cases and fixed orthographic errors; (iii) we retrieved again taxonomic information for those unmatched cases, evaluated accuracy of fuzzy matching and manually fixed records that are still not found; (iv) finally, we used the World Flora Taxonomic Backbone (Govaerts et al., 2021; WFO, July 7, 2022) as the ultimate filter for taxonomic information as we used it to calculate the plant taxonomic coverage of our database.

For pollinators: (i) we first created a checklist of species names for the most representative pollinator groups at the European level by combining the most up to date published checklists of bees and syrphids (Reverté et al., 2023), and butterflies (Wiemers et al., 2018); (ii) we compared pollinator species names against the checklist and recovered some unmatched cases with restrictive fuzzy matching by using **stringdist** package (Van der Loo et al., 2014); (iii) we fixed unmatched records when necessary and retrieved the taxonomic information for all species from GBIF; (iv) we fixed the non-found cases in the GBIF checklist and made sure that all species names from bees, syrphids and butterflies were named according to their respective species checklists. Coleoptera species names were only checked against the GBIF checklist.

#### Taxonomic coverage

To assess the completeness of plant and pollinator species in the EuPPollNet database at the European level, we used the aforementioned checklists for plants and pollinators. Specifically 308 for plants, we refined the checklist to include only plants occurring in Europe and excluded taxonomic groups not associated with biotic pollination. We did this by first excluding the families considered to have exclusively a wind pollination mode (see Culley et al., 2002), and then by filtering out the genera with wind or non-biotic pollination from families that exhibit both biotic and non-biotic pollination modes. Additionally, we manually included exotic species and added unresolved species names that were not present in the accepted names of 314 the checklist at the current version of usage. For pollinators, we only evaluated the taxo-315 nomic coverage of groups with species checklists available in Europe (i.e., bees, syrphids and butterflies). While there is not a good understanding of pollinator diversity in other taxonomic groups (e.g., beetles), it was assumed that their coverage within the database is equal to the average coverage of bees, syrphids, and butterflies (mean coverage = 31.4; sd = 4.51). Therefore, the total number of of flower visiting species from other taxonomic groups (i.e., non-bee, non syrphid and non-butterfly flower visitors) at European level was extrapolated by

Table 1. Column names and their descriptions within the EuPPollNet database.

Variable	Description
1. Study_id	Identifier of the study
2. Network_id	Identifier of a site sampled within a study
3. Sampling_method	Type of plant-pollinator sampling
4. Authors_habitat	Type of habitat as described by the authors
5. EuPPollNet_habitat	Type of habitat homogenized across studies
6. Bioregion	European biogiographical regions
7. Country	Country where the plant-pollinator interaction was observed
8. Locality	Locality where the plant-pollinator interaction was observed
9. Latitude	North-south position of the observed interaction in decimal degrees
10. Longitude	East-west position of the observed interaction in decimal degrees
11. Date	Year, month and day when the observation took place
12. Interaction	Number of interactions. By default is 1 as interactions are provided ungrouped
13. Plant_original_name	Plant species name given by the authors
14. Plant_accepted_name	Harmonized plant species name in the database
15. Plant_rank	Taxonomic rank of the observation
16. Plant_order	Order taxonomic rank of the observed plant species
17. Plant_family	Family taxonomic rank of the observed plant species
18. Plant_genus	Genus taxonomic rank of the observed plant species
19. Plant_unsure_id	Category to indicate if the plant species name is unsure (Yes) or not (No)
20. Plant_uncertainty_type	If the name is unsure, type of species uncertainty is provided
21. Pollinator_original_name	Pollinator species name given by the authors
$22.\ Pollinator\_accepted\_name$	Harmonized pollinator species name in the database
23. Pollinator_rank	Taxonomic rank of the observation
24. Pollinator_order	Order taxonomic rank of the observed pollinator species
25. Pollinator_family	Family taxonomic rank of the observed pollinator species
26. Pollinator_genus	Genus taxonomic rank of the observed pollinator species
27. Pollinator_unsure_id	Category to indicate if the pollinator species name is unsure (Yes) or not (No)
28. Pollinator_uncertainty_type	If the name is unsure, type of species uncertainty is provided
29. Flower_data	Floral data availability (Yes) or (No)
30. Flower_data_merger	Column to merge floral data with the interaction dataset

assuming that their coverage is equal to the mean coverage of bees, syrphids, and butterflies.
Consequently, we provide an estimate for the total number of potential pollinators across the
European continent.

Finally, to evaluate if the presence-absence of interaction records for bees and flowering plants 325 follows a phylogenetic pattern within the database, we calculated its phylogenetic signal at 326 genus and family level, respectively. The phylogenetic signal was calculated by using the 327 phylosiq function from the phytools package (Revell, 2012). We extracted the phylogenetic 328 information for bees from a genus level phylogeny (Hedtke et al., 2013) and processed it using 329 the packages ape (Paradis et al., 2019), MCMCglmm (Hadfield, 2010) and phytools. For 330 plants, the phylogenetic tree was obtained from a species level plant phylogeny (Smith & 331 Brown, 2018) with the help of the **rtree** package (Li, 2023). 332

## 333 Sampling coverage

The completeness of the EuPPollNet database was evaluated by exploring the rarefied accumulation curves of plant and pollinator species and their interactions across the different networks.

In addition, we computed the accumulation curve of pollinator species with an increasing number of plant species as an indicator for how many pollinator species are likely responsible for the pollination of flowering plants (e.g., Kleijn et al., 2015 for crops). The rarefied and extrapolated sampling curves were obtained using the iNEXT package (Hsieh & Chao, 2016). The different rarefied curves were complemented with 100 bootstrapped accumulation curves.

# 341 Habitat type and bioclimatic region

We describe the habitat type for each site using information from Corine Land Cover (CLC, 342 version 2018) extracted using the Terra package (Hijmans et al., 2022), visual inspection of 343 Google Earth imagery and the habitat classification from the authors. These different habitat 344 categories (see definitions in supplementary text) allow a quick comparison and understanding 345 of the habitat types from the database. Moreover, Europe is characterized by a great variety 346 of environmental conditions that harbor different biota. Thus, to allow authors to explore 347 the set of studies that share similar environmental conditions and species, we assigned a 348 biogeographical region to each site. The biogeographical regions were downloaded from the 349 European Environment Agency (version 2016) and were matched to the different sites using a 350 spatial joint from the sf package (Pebesma et al., 2018). 351

#### Network analyses

352

To provide a general overview of the structure of plant-pollinator networks in EuPPollNet, we quantified connectance and nestedness for each network and examined how these network metrics change across different latitudes and bioclimatic regions in Europe. We selected these

two network metrics as they are commonly evaluated in plant-pollinator network studies and capture structural properties with a straightforward interpretation. We implemented "standardised" versions of connectance and nestedness to account for the effect of sampling effort on network metrics. As connectance is negatively associated with network size (Jordano, 1987), we evaluated how network connectance was associated with the number of species (i.e., log of geometric mean of plants and pollinators) and extracted the residuals from this association (i.e., residual connectance) as a measurement of corrected connectance. The relationship between residual connectance and species richness was investigated using a beta regression. This approach was chosen because connectance displays a non-normal distribution with continuous values bounded between 0 and 1. The model was implemented using the betareg package (Cribari-Neto & Zeileis, 2010). We used NODFc to compare nestedness across networks, as it corrects by connectance and the number of species in comparison to other nestedness metrics that change with network size (Song et al., 2017). This metric was calculated using the maxnodf package (Hoeppke & Simmons, 2021). Both residual connectance and NODFc were used as dependent variables to evaluate their association with latitude. In addition, to quantify how connectance and nestedness change with network size, we determined their association with the number of species per network using the Kendall rank correlation coefficient.

356

357

358

359

360

361

362

363

364

365

366

367

368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

Finally, to compare if networks are more or less nested than expected by chance, we employed the traditional z-score approach with the widely used nestedness metric (NODF) from Almeida-Neto et al. (2008). The z-score approach only compares each unique network against their randomized versions, avoiding the influence of network size. As NODF is a metric computed from binary matrices, we calculated 100 null models for the binarized version of each network using the 'curveball' algorithm (Strona et al., 2014). This algorithm implements the configuration model (i.e. random rewiring of all links, without self-links or double links) and thus keeps the exact number of connections per species (i.e., realised degree). Since null model selection can significantly influence statistical results (Kaiser, 2015), we also implemented a null model that reorganizes the quantitative networks before binarising them. To that end, we used the 'quasiswap count' algorithm, which is a non-sequential algorithm for quantitative networks that maintains constant connectance and the number of connections per species. Each empirical network was randomized 100 times using each method. These null models were implemented with the **vegan** (Oksanen et al., 2013) package. Both connectance and nestedness (NODF) were estimated for each network using the function networklevel from bipartite (Dormann et al., 2008).

# 3 | RESULTS

### Taxonomic coverage

Europe hosts approximately over 5,000 species of pollinators, including 2,138 bee species 913 391 syrphid species, 496 butterflies species, and about 1,700 species from other taxonomic groups, 392 as estimated based on EuPPollNet average taxonomic coverage. Additionally, there are around 393 25,000 species of plants that benefit from animal pollination when excluding non-biotic polli-394 nated species. EuPPollNet contains a total of 2,223 pollinators and 1,411 plant species. The 395 coverage of the main pollinator groups occurring in Europe is 34.38% for bees, 33.63% for syr-396 phids and 26.21% for butterflies (see Figure S2 for coverage at the family level for bees and 397 butterflies, and at the subfamily level for syrphids). Bees (i.e., Anthophila) constitute 89.65% 398 of the interactions in EuPPollNet, and 77.95% of the interactions when excluding honey bees. 399 Within the database, 83.82% of bee genera have at least one species with interaction records, 400 and the average coverage of species at the bee genus level is 36.99% (Figure 2). The presence 401 or absence of interaction records for bees does not follow a phylogenetic pattern ( $\lambda = 0.07$ ; 402 P = 0.65). The database coverage of all flowering plant species occurring in Europe is 5.56% 403 (**Figure 3**), with an average coverage of 9.04% at the plant family level. Approximately, 404 half of the plant families have at least one species with interaction records (52.56%), and the 405 presence or absence of interaction data for the different plant species also does not follow a 406 statistically relevant phylogenetic pattern ( $\lambda = 0.26$ ; P = 0.07).

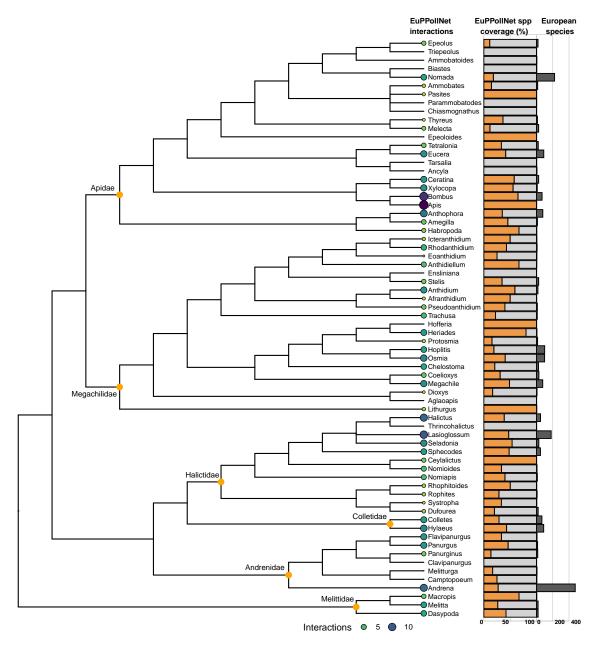


Figure 2. Phylogenetic and taxonomic coverage of bee genera at European level. The number of interactions recorded per genus in the database is illustrated using circles, with their sizes proportional to the number of interactions on a logarithmic scale, complemented by a gradient of colors ranging from yellow to dark purple. Additionally, the coverage of species recorded in EuPPollNet for each genus is depicted using two types of bars: orange bars representing the percentage of species included in the database and light grey bars indicating the percentage of species not included in the database, out of the total number of bee species in Europe. Dark grey bars represent the total number of species per genus at European level.

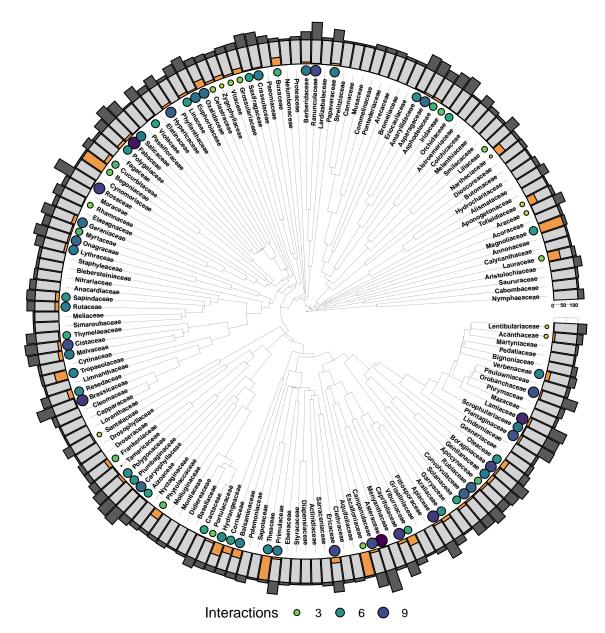


Figure 3. Phylogenetic and taxonomic coverage of the plant families at European level. The number of interactions recorded per family in the database is illustrated using circles, with their sizes proportional to the number of interactions on a logarithmic scale, complemented by a gradient of colors ranging from yellow to dark purple. Additionally, the coverage of species recorded in EuPPollNet for each family is depicted using two types of bars: orange bars representing the percentage of species included in the database and light grey bars indicating the percentage of species not included in the database out of the total number of bee species per family in Europe. Dark grey bars represent the total number of species per family at European level.

# 408 Sampling coverage

The estimated sampling coverage of plant and pollinator species within EuPPollNet across the different networks is approximately 97% for both taxonomic groups. This suggests that 410 the rarefied accumulation curves of both plant and pollinator species exhibit already a "quasi-411 asymptotic" growth of species richness by considering the current number of networks (Figure 412 4a-4b). The predicted observed species richness by doubling the sampling effort on the already 413 sampled habitat types within the database will only increase pollinator richness by 23.79% and 414 plant richness by 21.53%. However, the sampling coverage of interactions is 74.35%, and by 415 doubling the sampling effort, the predicted number of unique interactions recorded will have 416 approximately a twofold increase (53.68%; **Figure 4c**). When we consider the accumulated 417 pollinator richness across sampled plant species, this curve also shows a "quasi-asymptotic" 418 growth with a sampling coverage value of 96.54%. The predicted recorded pollinator species 419 by doubling the number of plants sampled is expected to increase by 22.22% (Figure 4d). We 420 find that a small portion of plant species and pollinator species are shared across a broad range 421 of networks and that most plant (85.68%) and pollinator (87.72%) species are exclusively found 422 in less than 1% of networks (Figures 4e-4f). The most common plant (Trifolium pratense) 423 and pollinator (Bombus pascuorum when excluding Apis mellifera) species are found in 36.07% 424 and 62.70% of networks, respectively.

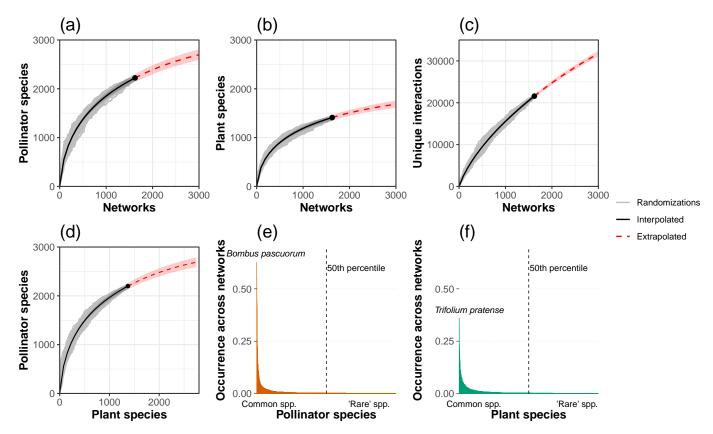


Figure 4. Graphs (a-b-c) indicate the accumulation curves for pollinators, plants, and the number of unique pairwise interactions across networks. Grey solid lines represent 100 randomized accumulation curves, the black solid lines represent the interpolated curve (i.e., the mean across curves), and the red dashed lines illustrate the extrapolated curve for approximately 3000 networks. The solid black points indicate the number of species and interactions contained in the database. Graph (d) shows the accumulation curve of pollinator species across an increasing number of plant species. This last graph uses the same color and shape structure as the ones in the top panel. Graphs (e-f) indicate the percentage of occurrence (i.e., incidence) of plant and pollinator species across networks. Species on the left (i.e., common) are found in many networks, while species on the right (i.e., rare) are found in few or only a single network. Note that *Apis mellifera* is the most common pollinator but was excluded from this visualization.

# 426 Habitat type and bioclimatic region

The proportion of species from the major pollinator orders within the database differed across 427 habitats and bioclimatic regions (Figure 5). Hymenoptera was the main taxonomic order in 428 the majority of habitats, exceeded only by Diptera for the habitat categories of riparian vege-429 tation and moors and heathland. Overall, the proportion of flower visitors from Lepidoptera 430 and Coleoptera were low across all habitats but Coleopteran flower visitors were notably more 431 abundant in sclerophyllous vegetation and beaches, dunes and sands habitat categories. Sim-432 ilar patterns were observed when exploring the pollinator proportions by bioclimatic region. 433 Hymenopterans were abundant across all bioclimatic regions and Dipterans were particularly 434 abundant in the Boreal, Alpine and Atlantic regions. Lepidopterans had low proportions 435 across all bioclimatic regions and Coleopterans were only relevant in the Mediterranean region 436 at European level. Notably, the number of studies (Figure 5) and sampling sites (Figure 437 S3) also differed across habitats and bioclimatic regions. The habitats sampled by a higher 438 number of studies in the database were intensive grasslands (26), semi-natural grasslands (15) 439 and sclerophyllous vegetation (10). However, the habitats that contain a higher number of 440 sampling sites were intensive grasslands (601), agricultural margins (432) and agricultural land 441 (141). The bioclimatic regions with a higher number of studies were Continental (24), Atlantic 442 (13) and Mediterranean (13); and those that contain a higher number of sampling sites were Continental (490), Atlantic (459) and Boreal (439). 444

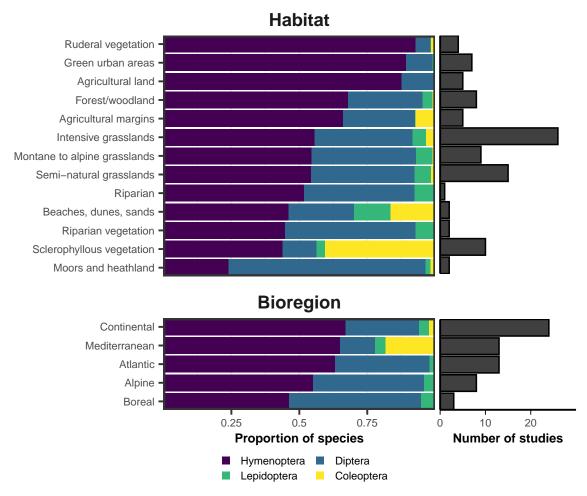


Figure 5. Proportion of species from the major pollinator orders by habitat types and bioclimatic regions in the EuPPollNet database. The orders, from left to right, include Hymenoptera, Diptera, Lepidoptera and Coleoptera. The horizontal barplot on the right indicates the number of studies that were conducted on each habitat type or bioclimatic region. Note that a single study can contribute to more than one habitat or bioclimatic region. Areas with a greater number of studies are more likely to depict accurate proportions of the different pollinator orders in those systems. The Pannonian and Steppic bioregions were excluded from this visualization because they contain only few networks from a single sampling day and site.

#### 445 Network properties

Connectance values ranged between 0.03 to 0.4 ( $\bar{x} = 0.14$ ) and followed a negative exponential relationship with the number of species per network (Kendall  $\tau = -0.75$ , P < 0.01; Figure 6a). 447 Nestedness values (NODFc) ranged between 1.34 to 8.63 ( $\bar{x} = 2.87$ ), and, as expected, were 448 not independent of the mean number of species (Kendall  $\tau = -0.05$ , P = 0.08; Figure S4). 449 Although latitude has a statistically significant impact on residual connectance, it explains 450 only a small portion of the observed variability in residual connectance and nestedness across 451 networks. (connectance:  $R^2 = 0.02$ , P < 0.01, Figure 6c; NODFc:  $R^2 \sim 0$ , P = 0.83, Figure 452 6d). Overall, networks towards higher latitudes showed lower residual connectance but similar 453 nestedness compared to networks located at lower latitudes. Note that residual connectance 454 and normalised nestedness showed a moderate significant negative correlation (Kendall  $\tau =$ 455 -0.43, P < 0.01). Empirical networks did not show statistically different nestedness (NODF) to 456 the simulated ones (**Figure 6b** and **Figure S5**). The 'curveball' method for binary networks 457 resulted in 12.02% of networks statistically less nested than null expectations, 85.71% showing 458 no difference, and 1.57% being more nested. The 'quasiswap' count' algorithm for quantitative 459 networks resulted in 11.4% of networks being less nested, 88.07% showing no difference and 460 0.53% being more nested. Note that in both cases, NODF is calculated on binarised matrices. 461

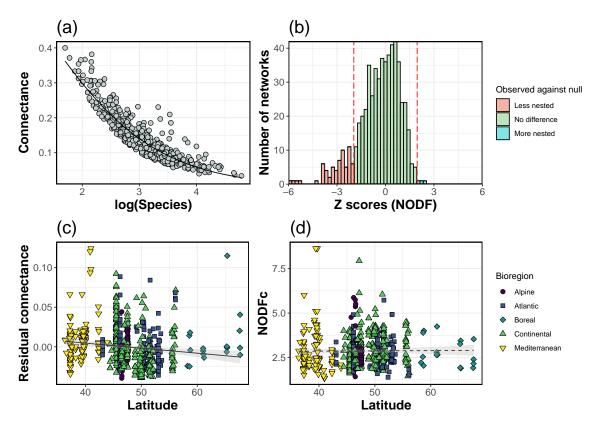


Figure 6. Graph (a) shows the association between network connectance and the geometric mean of plant and pollinator species per network on a log-scale with the respective fitted line from a Beta regression. Graph (b) shows the distribution of z-scores when comparing the nestedness from the empirical networks with their randomised counterparts (100 null models for each network with the curveball algorithm). The vertical red dashed lines represent the z critical value for a two tailed test with alpha = 0.05. Z-scores to the left of the first vertical red dashed line indicate that networks are less nested than expected by chance (red), those between the two dashed lines indicate no statistical difference from random expectations (green), and those to the right indicate that networks are more nested than expected by chance (blue). Graphs (c-d) show the fitted regression of residual connectance and nestedness across the latitudinal range of the studies from the database. The solid fitted line indicates a significant association, while the dashed lines indicate a non-significant association. The bioclimatic region of each network is indicated with points of different shapes and colours.

# 3 | DISCUSSION

481

482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

497

498

499

500

501

EuPPollNet offers the largest set of plant-pollinator studies and networks compiled to date 463 at European level. The database contains 1,411 plant and 2,223 pollinator species with over 464 a million interaction records. While the overall sampling coverage of species and interactions 465 is relatively high across the sampled sites, the taxonomic coverage of plants and the main 466 pollinator groups at the European level is still relatively low (i.e., 5.6% for flowering plants 467 and 34.4% for bee species). This likely reflects that most plant and pollinator species are rare 468 and geographically restricted. For example, given that most of the plant-pollinator networks 469 from the database are sampled on intensive grasslands, and habitat heterogeneity is a crucial 470 factor in understanding pollinator diversity at European level (Hass et al., 2018; Kleijn et al., 471 2015; Martínez-Núñez et al., 2022), adding studies on other habitat types is likely to result in 472 a rapid increase of the coverage of plant and pollinator species and their interactions. Indeed, 473 plant and pollinator species were rarely shared across multiple sites, indicating that there are 474 few "common" species and many "rare" ones at the metaweb or continental level. This high 475 number of "rare" species results in an upward slope of the species or interaction accumulation 476 curves (Thompson & Withers, 2003). In other words, minimal sampling efforts are capturing 477 a substantial number of species and interactions, but achieving a comprehensive inventory will 478 require numerous sampling events within and across habitats, particularly for plant-pollinator 479 interactions. 480

Bees are responsible for the majority of the sampled interactions at the metaweb level. As not all surveys included all pollinator groups, this result may partly be influenced by the taxonomic groups sampled across studies. However, the relevance of bees and other pollinator orders for network topology changed across habitats and bioclimatic regions in accordance to the literature. For instance, plant-pollinator communities in the Mediterranean were dominated by bees, while communities in Alpine or Boreal regions were fly species rich or fly-dominated. These patterns are consistent with our current understanding of bee diversity, which peaks in dry or temperate areas (Leclercq et al., 2023; Orr et al., 2021); and with the fact that colder environments (i.e., altitude and latitude wise) harbor a larger fraction of fly pollinators compared to other taxa (Elberling & Olesen, 1999; Lefebvre et al., 2018). In addition, beetles were only commonly documented as floral visitors in the Mediterranean region. This study cannot determine whether pollination ecologists traditionally document flower-beetle interactions only in the Mediterranean, or if there are fewer flower visitations by beetles outside this region. Nevertheless, the high proportion of beetles as floral visitors provides further support for their potential role as pollinators in the Mediterranean (Herrera, 2019; León-Osper & Narbona, 2022). The number of butterfly species and interactions were relatively low compared to the other taxa. While Europe contains fewer butterfly species than other regions of the world (Ollerton, 2017), their relevance as pollinators is likely underestimated within this database. This is because a large fraction of studies ( $\sim 40\%$ ) did not sample butterflies, and conventional sampling methods for monitoring other insect pollinators (e.g., bees or flies) may be inadequate for sampling plant-butterfly interactions (Isaac et al., 2011). Honey bees were present in 87% of networks and conducted on average a third of the total interactions per network. The proportion of honey bees in networks across Europe is higher than in natural communities (i.e., large unmanaged assemblages of plant species) across the world (~13%; Hung et al., 2018). This potentially reflects the dominance of intensive grassland habitats in EuPPollNet and their widespread distribution across European landscapes (Isselstein et al., 2005), the highly generalised nature of honey bees, their native status and above all, the widespread practice of beekeeping in Europe (Herrera, 2020; Magrach et al., 2017; Steffan-Dewenter & Tscharntke, 2000).

Although Europe contains a much larger number of flowering plants than pollinator species (~5 to 1 ratio according to our extrapolation from checklists), the observed number of pollinator species in the database was almost double that of the plants. This could be explained by the fact that all networks are phytocentric, resulting in sampling bias towards pollinator species (Jordano, 2016; Vizentin-Bugoni et al., 2018). While animal-centered sampling is likely to increase the plant-pollinator species ratio (e.g., Encinas-Viso et al., 2023), the spatial scale and environmental context of the sampled communities will also influence their observed diversity, especially given the ability to move of pollinators and the sessile nature of plants. In addition, we found that the accumulation curve of pollinators per plant species does not saturate, which indicates low redundancy of pollinators and that many are regionally "rare". Rare pollinators can be functionally important for plant species at the landscape level (Simpson et al., 2022; Winfree et al., 2018), highlighting the need to conduct further sampling events to identify these rare species across different regions and to effectively understand and protect plant-pollinator biodiversity.

Consistent with Olesen & Jordano (2002), we found that residual connectance (i.e., the deviation from the expected connectance for a given network size) was lower at higher latitudes, while normalised nestedness increased towards higher latitudes. Networks at lower latitudes in Europe are exposed to higher temperatures, which can result in higher visitation rates (Arroyo et al., 1985; Classen et al., 2015; Herrera, 2019), and the overall level of pollinator generalization is known to be higher at lower latitudes (Schleuning et al., 2012). These factors should increase the number of possible connections that can be established between plants and pollinators for a given network size, resulting in more connected networks at lower latitudes in Europe. However, our results cannot be extrapolated to lower latitudes outside Europe, as tropical systems might behave differently. Finally, a result that can be surprising is that most empirical networks (~85%) showed a non-nested structure. This result contrasts with earlier work suggesting that plant-pollinator networks are typically nested (Bascompte et al., 2003) and it aligns with recent evidence showing that nestedness is not a common feature in plant-pollinator networks when evaluated against restrictive null models that conserve the observed species degree (Payrató-Borras et al., 2019; Figure S6). Note that while species degree distributions are sufficient to explain the emergence of nestedness, this does not preclude nestedness from being a useful metric for comparison across networks. For example, plantpollinator networks are more nested than plant-herbivore networks precisely because these two network types differ in their composition of generalist and specialist species (Thébault & Fontaine, 2010).

Although this database covers a wide range of habitats across 23 countries, it contains temporal and geographical biases that can impact our understanding of plant-pollinator communities 545 (Hughes et al., 2021). For instance, none of the studies in this database sampled nocturnal 546 pollinators, which can impact our view of network structure (García et al., 2024), and most 547 studies were conducted during a single flowering season, limiting our ability to evaluate tempo-548 ral trends of plant-pollinator communities in the face of environmental changes (Alarcón et al., 549 2008; Chacoff et al., 2018). In addition, most plant-pollinator networks are sampled from cen-550 tral Europe, while Eastern Europe and the Mediterranean region are underrepresented. This 551 is consistent with previous studies which also report lack of plant-pollinator data for those 552 regions (Bennett et al., 2018; Marshall et al., 2024), highlighting that this database shows ex-553 isting patterns in data availability despite the absence of a systematic search for studies. The 554 lack of data for Eastern Europe, which contains vast landscapes with semi-natural grasslands 555 experiencing rapid land use change (Sutcliffe et al., 2015), and for the Mediterranean region, 556 which is severely impacted by climate change (Duchenne et al., 2020; Jaworski et al., 2022; 557 Pareja-Bonilla et al., 2023), is particularly concerning. These areas are well known for their rich 558 pollinator diversity (Miličić et al., 2018; Reverté et al., 2023), and their under-representation 559 is likely contributing to the low taxonomic coverage of this database at the European level. 560 Although some of the most well studied countries in Europe (e.g., Belgium, The Netherlands) 561 have already experienced land use change and biodiversity loss at the end of the 20th century 562 (Carvalheiro et al., 2013), plant-pollinator communities in Europe and across the globe still 563 face current and future threats from climate change (Bartomeus et al., 2011; Duchenne et al., 564 2020), land use change (Batáry et al., 2015; Reidsma et al., 2006), and the introduction of 565 alien species (Vanbergen et al., 2018; Vilà et al., 2009). Therefore, continuous monitoring pro-566 grams are needed in order to evaluate spatio-temporal changes of species and their interactions 567 across different European habitats and regions. This will allow local and large scale analyses 568 of the status and trends of plant-pollinator communities, effectively informing management 569 and conservation actions. 570

In conclusion, the EuPPollNet database enables researchers to explore spatial, taxonomic and structural properties of plant-pollinator networks within Europe. In contrast to previous databases, EuPPollNet provides interaction data along with sampling information that could help researchers to better control for sampling effort and completeness and to select the most suitable networks for their research questions. Here, we have shown how connectance and nestedness change across their latitudinal range, and that plant-pollinator networks are as nested as expected given plant and pollinator generalist levels. These analyses aim to highlight the variability present across Europe in the structure of plant-pollinator networks and illustrate the opportunities available to develop and test questions about spatio-temporal network change using EuPPollNet. The reproducible workflow allows researchers to adapt and reuse this database, enabling the continuous addition of new networks to better evaluate the status and trends of plant-pollinator communities. Finally, we hope this database becomes an iterative resource that keeps growing and improving over time to better understand and conserve European biodiversity.

571

572

573

574

575

576

577

578

579

580

581

582

583

584

# REFERENCES

- Alarcón, R., Waser, N. M., & Ollerton, J. (2008). Year-to-year variation in the topology of a plant–pollinator interaction network. *Oikos*, 117(12), 1796–1807.
- Almeida-Neto, M., Guimaraes, P., Guimaraes Jr, P. R., Loyola, R. D., & Ulrich, W. (2008).

  A consistent metric for nestedness analysis in ecological systems: Reconciling concept and measurement. *Oikos*, 117(8), 1227–1239.
- Archer, C. R., Pirk, C. W. W., Carvalheiro, L. G., & Nicolson, S. W. (2014). Economic and ecological implications of geographic bias in pollinator ecology in the light of pollinator declines. *Oikos*, 123(4), 401–407.
- Arroyo, M. T. K., Armesto, J. J., & Primack, R. B. (1985). Community studies in pollination ecology in the high temperate andes of central chile II. Effect of temperature on visitation rates and pollination possibilities. *Plant Systematics and Evolution*, 149(3–4), 187–203.
- Bartomeus, I., Ascher, J. S., Wagner, D., Danforth, B. N., Colla, S., Kornbluth, S., & Winfree, R. (2011). Climate-associated phenological advances in bee pollinators and bee-pollinated plants. *Proceedings of the National Academy of Sciences*, 108(51), 20645–20649.
- Bascompte, J., & Jordano, P. (2007). Plant-animal mutualistic networks: The architecture of
   biodiversity. Annu. Rev. Ecol. Evol. Syst., 38, 567–593.
- Bascompte, J., Jordano, P., Melián, C. J., & Olesen, J. M. (2003). The nested assembly
   of plant–animal mutualistic networks. *Proceedings of the National Academy of Sciences*,
   100(16), 9383–9387.
- Batáry, P., Dicks, L. V., Kleijn, D., & Sutherland, W. J. (2015). The role of agri-environment schemes in conservation and environmental management. *Conservation Biology*, 29(4), 1006–1016.
- Bellard, C., Leclerc, C., Leroy, B., Bakkenes, M., Veloz, S., Thuiller, W., & Courchamp,
   F. (2014). Vulnerability of biodiversity hotspots to global change. Global Ecology and
   Biogeography, 23(12), 1376–1386.
- Bennett, J. M., Thompson, A., Goia, I., Feldmann, R., Ştefan, V., Bogdan, A., Rakosy, D.,
  Beloiu, M., Biro, I.-B., Bluemel, S., et al. (2018). A review of european studies on
  pollination networks and pollen limitation, and a case study designed to fill in a gap. *AoB*Plants, 10(6), ply068.
- Burkle, L. A., & Alarcón, R. (2011). The future of plant–pollinator diversity: Understanding
   interaction networks across time, space, and global change. American Journal of Botany,
   98(3), 528–538.
- Carvalheiro, L. G., Kunin, W. E., Keil, P., Aguirre-Gutiérrez, J., Ellis, W. N., Fox, R., Groom,
   Q., Hennekens, S., Van Landuyt, W., Maes, D., et al. (2013). Species richness declines and
   biotic homogenisation have slowed down for NW-european pollinators and plants. *Ecology Letters*, 16(7), 870–878.
- Chacoff, N. P., Resasco, J., & Vázquez, D. P. (2018). Interaction frequency, network position,
   and the temporal persistence of interactions in a plant–pollinator network. *Ecology*, 99(1),
   21–28.
- Chacoff, N. P., Vázquez, D. P., Lomáscolo, S. B., Stevani, E. L., Dorado, J., & Padrón, B.
   (2012). Evaluating sampling completeness in a desert plant–pollinator network. *Journal*

- of Animal Ecology, 81(1), 190–200.
- Chamberlain, S., Oldoni, D., & Waller, J. (2022). Rgbif: Interface to the global biodiversity information facility API. http://www-cs-faculty.stanford.edu/~uno/abcde.html
- Classen, A., Peters, M. K., Kindeketa, W. J., Appelhans, T., Eardley, C. D., Gikungu, M. W., Hemp, A., Nauss, T., & Steffan-Dewenter, I. (2015). Temperature versus resource constraints: Which factors determine bee diversity on mount kilimanjaro, tanzania? *Global Ecology and Biogeography*, 24(6), 642–652.
- Cribari-Neto, F., & Zeileis, A. (2010). Beta regression in r. *Journal of Statistical Software*, 34, 1–24.
- Culley, T. M., Weller, S. G., & Sakai, A. K. (2002). The evolution of wind pollination in angiosperms. *Trends in Ecology & Evolution*, 17(8), 361–369.
- Dormann, C. F., Gruber, B., & Fründ, J. (2008). Introducing the bipartite package: Analysing ecological networks. *Interaction*, 1(0.2413793), 8–11.
- Duchenne, F., Thébault, E., Michez, D., Elias, M., Drake, M., Persson, M., Rousseau-Piot,
  J., Pollet, M., Vanormelingen, P., & Fontaine, C. (2020). Phenological shifts alter the
  seasonal structure of pollinator assemblages in europe. Nature Ecology & Evolution, 4(1),
  115–121.
- Eichenberg, D., Bowler, D. E., Bonn, A., Bruelheide, H., Grescho, V., Harter, D., Jandt, U.,
   May, R., Winter, M., & Jansen, F. (2021). Widespread decline in central european plant
   diversity across six decades. Global Change Biology, 27(5), 1097–1110.
- Elberling, H., & Olesen, J. M. (1999). The structure of a high latitude plant-flower visitor system: The dominance of flies. *Ecography*, 22(3), 314–323.
- Encinas-Viso, F., Bovill, J., Albrecht, D. E., Florez-Fernandez, J., Lessard, B., Lumbers, J.,
   Rodriguez, J., Schmidt-Lebuhn, A., Zwick, A., & Milla, L. (2023). Pollen DNA metabarcod ing reveals cryptic diversity and high spatial turnover in alpine plant-pollinator networks.
   Molecular Ecology, 32(23), 6377-6393.
- García, Y., Giménez-Benavides, L., Iriondo, J. M., Lara-Romero, C., Méndez, M., Morente López, J., & Santamaría, S. (2024). Addition of nocturnal pollinators modifies the structure
   of pollination networks. Scientific Reports, 14(1), 1226.
- Gibson, R. H., Knott, B., Eberlein, T., & Memmott, J. (2011). Sampling method influences
   the structure of plant–pollinator networks. Oikos, 120(6), 822–831.
- Goulson, D., Nicholls, E., Botías, C., & Rotheray, E. L. (2015). Bee declines driven by combined stress from parasites, pesticides, and lack of flowers. *Science*, 347(6229), 1255957.
- Govaerts, R., Nic Lughadha, E., Black, N., Turner, R., & Paton, A. (2021). The world
   checklist of vascular plants, a continuously updated resource for exploring global plant
   diversity. Scientific Data, 8(1), 215.
- Grenié, M., Berti, E., Carvajal-Quintero, J., Dädlow, G. M. L., Sagouis, A., & Winter, M.
   (2023). Harmonizing taxon names in biodiversity data: A review of tools, databases and
   best practices. Methods in Ecology and Evolution, 14(1), 12–25.
- Guimaraes, P. R. (2020). The structure of ecological networks across levels of organization.
   Annual Review of Ecology, Evolution, and Systematics, 51, 433–460.
- Gusenbauer, M. (2019). Google scholar to overshadow them all? Comparing the sizes of 12 academic search engines and bibliographic databases. *Scientometrics*, 118(1), 177–214.

- Gusenbauer, M., & Haddaway, N. R. (2020). Which academic search systems are suitable for systematic reviews or meta-analyses? Evaluating retrieval qualities of google scholar, PubMed, and 26 other resources. Research Synthesis Methods, 11(2), 181–217.
- Hadfield, J. D. (2010). MCMC methods for multi-response generalized linear mixed models:
  The MCMCglmm r package. *Journal of Statistical Software*, 33, 1–22.
- Hass, A. L., Kormann, U. G., Tscharntke, T., Clough, Y., Baillod, A. B., Sirami, C., Fahrig,
   L., Martin, J.-L., Baudry, J., Bertrand, C., et al. (2018). Landscape configurational
   heterogeneity by small-scale agriculture, not crop diversity, maintains pollinators and plant
   reproduction in western europe. Proceedings of the Royal Society B: Biological Sciences,
   285(1872), 20172242.
- Heberling, J. M., Miller, J. T., Noesgaard, D., Weingart, S. B., & Schigel, D. (2021). Data
   integration enables global biodiversity synthesis. Proceedings of the National Academy of
   Sciences, 118(6), e2018093118.
- Hedtke, S. M., Patiny, S., & Danforth, B. N. (2013). The bee tree of life: A supermatrix approach to apoid phylogeny and biogeography. *BMC Evolutionary Biology*, 13, 1–13.
- Herrera, C. M. (2019). Complex long-term dynamics of pollinator abundance in undisturbed mediterranean montane habitats over two decades. *Ecological Monographs*, 89(1), e01338.
- Herrera, C. M. (2020). Gradual replacement of wild bees by honeybees in flowers of the mediterranean basin over the last 50 years. *Proceedings of the Royal Society B*, 287(1921), 20192657.
- Hijmans, R. J., Bivand, R., Forner, K., Ooms, J., Pebesma, E., & Sumner, M. D. (2022).
   Package "terra".
- Hoeppke, C., & Simmons, B. I. (2021). Maxnodf: An r package for fair and fast comparisons of nestedness between networks. *Methods in Ecology and Evolution*, 12(4), 580–585.
- Hsieh, T., & Chao, A. (2016). iNEXT: An r package for rarefaction and extrapolation of
   species diversity (hill numbers). Methods in Ecology and Evolution, 7(12), 1451–1456.
- Hughes, A. C., Orr, M. C., Ma, K., Costello, M. J., Waller, J., Provoost, P., Yang, Q., Zhu,
  C., & Qiao, H. (2021). Sampling biases shape our view of the natural world. *Ecography*,
  44(9), 1259–1269.
- Hung, K.-L. J., Kingston, J. M., Albrecht, M., Holway, D. A., & Kohn, J. R. (2018). The worldwide importance of honey bees as pollinators in natural habitats. *Proceedings of the Royal Society B: Biological Sciences*, 285(1870), 20172140.
- Isaac, N. J., Cruickshanks, K. L., Weddle, A. M., Marcus Rowcliffe, J., Brereton, T. M., Dennis,
   R. L., Shuker, D. M., & Thomas, C. D. (2011). Distance sampling and the challenge of
   monitoring butterfly populations. *Methods in Ecology and Evolution*, 2(6), 585–594.
- Isselstein, J., Jeangros, B., & Pavlu, V. (2005). Agronomic aspects of biodiversity targeted
   management of temperate grasslands in europe—a review. Agronomy Research, 3(2), 139–
   151.
- Jaworski, C. C., Geslin, B., Zakardjian, M., Lecareux, C., Caillault, P., Nève, G., Meunier, J.-Y., Dupouyet, S., Sweeney, A. C., Lewis, O. T., et al. (2022). Long-term experimental drought alters floral scent and pollinator visits in a mediterranean plant community despite overall limited impacts on plant phenotype and reproduction. *Journal of Ecology*, 110(11),

- 2628-2648.
- Jordano, P. (1987). Patterns of mutualistic interactions in pollination and seed dispersal:
  Connectance, dependence asymmetries, and coevolution. *The American Naturalist*, 129(5),
  657–677.
- Jordano, P. (2016). Sampling networks of ecological interactions. *Functional Ecology*, 30(12), 1883–1893.
- Jordano, P., Bascompte, J., & Olesen, J. M. (2003). Invariant properties in coevolutionary networks of plant–animal interactions. *Ecology Letters*, 6(1), 69–81.
- Kaiser, L. (2015). Nestedness and null models in ecology. Bachelor's Thesis, Freiburg, Germany: Albert-Ludwigs-University Freiburg.
- Kleijn, D., Winfree, R., Bartomeus, I., Carvalheiro, L. G., Henry, M., Isaacs, R., Klein, A.-M., Kremen, C., M'gonigle, L. K., Rader, R., et al. (2015). Delivery of crop pollination services is an insufficient argument for wild pollinator conservation. *Nature Communications*, 6(1), 7414.
- Leclercq, N., Marshall, L., Caruso, G., Schiel, K., Weekers, T., Carvalheiro, L. G., Dathe, H. H., Kuhlmann, M., Michez, D., Potts, S. G., et al. (2023). European bee diversity: Taxonomic and phylogenetic patterns. *Journal of Biogeography*, 50(7), 1244–1256.
- Lefebvre, V., Villemant, C., Fontaine, C., & Daugeron, C. (2018). Altitudinal, temporal and trophic partitioning of flower-visitors in alpine communities. *Scientific Reports*, 8(1), 4706.
- León-Osper, M., & Narbona, E. (2022). Unravelling the mystery of red flowers in the mediterranean basin: How to be conspicuous in a place dominated by hymenopteran pollinators. Functional Ecology, 36(11), 2774–2790.
- Li, D. (2023). Rtrees: An r package to assemble phylogenetic trees from megatrees. *Ecography*, 2023(7), e06643.
- Magrach, A., González-Varo, J. P., Boiffier, M., Vilà, M., & Bartomeus, I. (2017). Honeybee
   spillover reshuffles pollinator diets and affects plant reproductive success. Nature Ecology
   Evolution, 1(9), 1299–1307.
- Marshall, L., Leclercq, N., Carvalheiro, L. G., Dathe, H. H., Jacobi, B., Kuhlmann, M., Potts,
   S. G., Rasmont, P., Roberts, S. P., & Vereecken, N. J. (2024). Understanding and addressing shortfalls in european wild bee data. *Biological Conservation*, 290, 110455.
- Martínez-Núñez, C., Kleijn, D., Ganuza, C., Heupink, D., Raemakers, I., Vertommen, W., &
   Fijen, T. P. (2022). Temporal and spatial heterogeneity of semi-natural habitat, but not crop diversity, is correlated with landscape pollinator richness. *Journal of Applied Ecology*, 59(5), 1258–1267.
- Miličić, M., Vujić, A., & Cardoso, P. (2018). Effects of climate change on the distribution of hoverfly species (diptera: Syrphidae) in southeast europe. *Biodiversity and Conservation*, 27, 1173–1187.
- Oksanen, J., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., O'hara, R., Simpson, G. L., Solymos, P., Stevens, M. H. H., Wagner, H., et al. (2013). Package "vegan." *Community Ecology Package, Version*, 2(9), 1–295.
- Olesen, J. M., Bascompte, J., Dupont, Y. L., Elberling, H., Rasmussen, C., & Jordano, P. (2011). Missing and forbidden links in mutualistic networks. *Proceedings of the Royal*

- 756 Society B: Biological Sciences, 278(1706), 725–732.
- Olesen, J. M., & Jordano, P. (2002). Geographic patterns in plant–pollinator mutualistic networks. *Ecology*, 83(9), 2416–2424.
- Ollerton, J. (2017). Pollinator diversity: Distribution, ecological function, and conservation.

  Annual Review of Ecology, Evolution, and Systematics, 48, 353–376.
- Orr, M. C., Hughes, A. C., Chesters, D., Pickering, J., Zhu, C.-D., & Ascher, J. S. (2021). Global patterns and drivers of bee distribution. *Current Biology*, 31(3), 451–458.
- Paradis, E., Blomberg, S., Bolker, B., Brown, J., Claude, J., Cuong, H. S., Desper, R., &
   Didier, G. (2019). Package "ape." Analyses of Phylogenetics and Evolution, Version, 2(4),
   47.
- Pareja-Bonilla, D., Arista, M., Morellato, L. P. C., & Ortiz, P. L. (2023). Better soon than
   never: Climate change induces strong phenological reassembly in the flowering of a mediter ranean shrub community. Annals of Botany, mcad193.
- Payrató-Borras, C., Hernández, L., & Moreno, Y. (2019). Breaking the spell of nestedness: The entropic origin of nestedness in mutualistic systems. *Physical Review X*, 9(3), 031024.
- Pebesma, E. J. et al. (2018). Simple features for r: Standardized support for spatial vector data. R J., 10(1), 439.
- Poelen, J. H., Simons, J. D., & Mungall, C. J. (2014). Global biotic interactions: An open
   infrastructure to share and analyze species-interaction datasets. *Ecological Informatics*, 24,
   148–159.
- Poisot, T., Baiser, B., Dunne, J. A., Kéfi, S., Massol, F., Mouquet, N., Romanuk, T. N.,
   Stouffer, D. B., Wood, S. A., & Gravel, D. (2016). Mangal–making ecological network
   analysis simple. *Ecography*, 39(4), 384–390.
- Poisot, T., Bergeron, G., Cazelles, K., Dallas, T., Gravel, D., MacDonald, A., Mercier, B.,
   Violet, C., & Vissault, S. (2021). Global knowledge gaps in species interaction networks
   data. Journal of Biogeography, 48(7), 1552–1563.
- Proença, V., Martin, L. J., Pereira, H. M., Fernandez, M., McRae, L., Belnap, J., Böhm,
   M., Brummitt, N., García-Moreno, J., Gregory, R. D., et al. (2017). Global biodiversity
   monitoring: From data sources to essential biodiversity variables. *Biological Conservation*,
   213, 256–263.
- Ramos-Jiliberto, R., Domínguez, D., Espinoza, C., Lopez, G., Valdovinos, F. S., Bustamante, R. O., & Medel, R. (2010). Topological change of andean plant-pollinator networks along an altitudinal gradient. *Ecological Complexity*, 7(1), 86–90.
- Rech, A. R., Dalsgaard, B., Sandel, B., Sonne, J., Svenning, J.-C., Holmes, N., & Ollerton, J. (2016). The macroecology of animal versus wind pollination: Ecological factors are more important than historical climate stability. *Plant Ecology & Diversity*, 9(3), 253–262.
- Reidsma, P., Tekelenburg, T., Van den Berg, M., & Alkemade, R. (2006). Impacts of land-use change on biodiversity: An assessment of agricultural biodiversity in the european union.

  Agriculture, Ecosystems & Environment, 114(1), 86–102.
- Revell, L. J. (2012). Phytools: An r package for phylogenetic comparative biology (and other things). *Methods in Ecology and Evolution*, 2, 217–223.
- Reverté, S., Miličić, M., Ačanski, J., Andrić, A., Aracil, A., Aubert, M., Balzan, M. V.,
  Bartomeus, I., Bogusch, P., Bosch, J., et al. (2023). National records of 3000 european bee

- and hoverfly species: A contribution to pollinator conservation. *Insect Conservation and Diversity*, 16(6), 758–775.
- Sala, O. E., Stuart Chapin, F., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-Sanwald, E., Huenneke, L. F., Jackson, R. B., Kinzig, A., et al. (2000). Global biodiversity scenarios for the year 2100. *Science*, 287(5459), 1770–1774.
- Saunders, M. E., Kendall, L. K., Lanuza, J. B., Hall, M. A., Rader, R., & Stavert, J. R. (2023).
   Climate mediates roles of pollinator species in plant–pollinator networks. Global Ecology
   and Biogeography, 32(4), 511–518.
- Schleuning, M., Fründ, J., Klein, A.-M., Abrahamczyk, S., Alarcón, R., Albrecht, M., Andersson, G. K., Bazarian, S., Böhning-Gaese, K., Bommarco, R., et al. (2012). Specialization of mutualistic interaction networks decreases toward tropical latitudes. *Current Biology*, 22(20), 1925–1931.
- Schwarz, B., Vázquez, D. P., CaraDonna, P. J., Knight, T. M., Benadi, G., Dormann, C. F., Gauzens, B., Motivans, E., Resasco, J., Blüthgen, N., et al. (2020). Temporal scale-dependence of plant-pollinator networks. *Oikos*, *129*(9), 1289–1302.
- Settele, J., Bishop, J., & Potts, S. G. (2016). Climate change impacts on pollination. *Nature Plants*, 2(7), 1–3.
- Simpson, D. T., Weinman, L. R., Genung, M. A., Roswell, M., MacLeod, M., & Winfree, R. (2022). Many bee species, including rare species, are important for function of entire plant–pollinator networks. *Proceedings of the Royal Society B*, 289(1972), 20212689.
- Smith, S. A., & Brown, J. W. (2018). Constructing a broadly inclusive seed plant phylogeny.

  \*\*American Journal of Botany, 105(3), 302–314.
- Song, C., Rohr, R. P., & Saavedra, S. (2017). Why are some plant–pollinator networks more nested than others? *Journal of Animal Ecology*, 86(6), 1417–1424.
- Staniczenko, P., Kopp, J. C., & Allesina, S. (2013). The ghost of nestedness in ecological networks. *Nature Communications*, 4(1), 1–6.
- Stanners, D., & Bourdeau, P. (1995). Europe's environment: The dobříš assessment. Office for Official Publications of the European Community.
- Steffan-Dewenter, I., & Tscharntke, T. (2000). Resource overlap and possible competition between honey bees and wild bees in central europe. *Oecologia*, 122, 288–296.
- Strona, G., Nappo, D., Boccacci, F., Fattorini, S., & San-Miguel-Ayanz, J. (2014). A fast and unbiased procedure to randomize ecological binary matrices with fixed row and column totals. *Nature Communications*, 5(1), 4114.
- Sutcliffe, L. M., Batáry, P., Kormann, U., Báldi, A., Dicks, L. V., Herzon, I., Kleijn, D.,
  Tryjanowski, P., Apostolova, I., Arlettaz, R., et al. (2015). Harnessing the biodiversity
  value of central and eastern european farmland. *Diversity and Distributions*, 21(6), 722–
  730.
- Thébault, E., & Fontaine, C. (2010). Stability of ecological communities and the architecture of mutualistic and trophic networks. *Science*, 329(5993), 853–856.
- Thompson, G. G., & Withers, P. C. (2003). Effect of species richness and relative abundance on the shape of the species accumulation curve. *Austral Ecology*, 28(4), 355–360.
- Traveset, A., Tur, C., Trøjelsgaard, K., Heleno, R., Castro-Urgal, R., & Olesen, J. M. (2016).

  Global patterns of mainland and insular pollination networks. Global Ecology and Biogeog-

- raphy, 25(7), 880-890.
- Troia, M. J., & McManamay, R. A. (2016). Filling in the GAPS: Evaluating completeness and coverage of open-access biodiversity databases in the united states. *Ecology and Evolution*, 6(14), 4654–4669.
- Trøjelsgaard, K., & Olesen, J. M. (2013). Macroecology of pollination networks. *Global Ecology* and *Biogeography*, 22(2), 149–162.
- Van der Loo, M. P. et al. (2014). The stringdist package for approximate string matching.

  The R Journal, 6(1), 111.
- Vanbergen, A. J., Espíndola, A., & Aizen, M. A. (2018). Risks to pollinators and pollination from invasive alien species. *Nature Ecology & Evolution*, 2(1), 16–25.
- Vilà, M., Bartomeus, I., Dietzsch, A. C., Petanidou, T., Steffan-Dewenter, I., Stout, J. C., & Tscheulin, T. (2009). Invasive plant integration into native plant—pollinator networks across europe. *Proceedings of the Royal Society B: Biological Sciences*, 276(1674), 3887–3893.
- Vizentin-Bugoni, J., Maruyama, P. K., Souza, C. S. de, Ollerton, J., Rech, A. R., & Sazima, M.
   (2018). Plant-pollinator networks in the tropics: A review. In W. Dáttilo & V. Rico-Gray
   (Eds.), Ecological networks in the tropics: An integrative overview of species interactions
   from some of the most species-rich habitats on earth (pp. 73–91). Springer.
- Wetzel, F. T., Bingham, H. C., Groom, Q., Haase, P., Kõljalg, U., Kuhlmann, M., Martin, C.
   S., Penev, L., Robertson, T., Saarenmaa, H., et al. (2018). Unlocking biodiversity data:
   Prioritization and filling the gaps in biodiversity observation data in europe. Biological
   Conservation, 221, 78–85.
- Wiemers, M., Balletto, E., Dincă, V., Fric, Z. F., Lamas, G., Lukhtanov, V., Munguira, M. L.,
   Swaay, C. A. van, Vila, R., Vliegenthart, A., et al. (2018). An updated checklist of the
   european butterflies (lepidoptera, papilionoidea). ZooKeys, 811, 9.
- Windsor, F. M., Hoogen, J. van den, Crowther, T. W., & Evans, D. M. (2023). Using ecological networks to answer questions in global biogeography and ecology. *Journal of Biogeography*, 50(1), 57-69.
- Winfree, R., Reilly, J. R., Bartomeus, I., Cariveau, D. P., Williams, N. M., & Gibbs, J. (2018).

  Species turnover promotes the importance of bee diversity for crop pollination at regional scales. *Science*, 359(6377), 791–793.
- Zattara, E. E., & Aizen, M. A. (2021). Worldwide occurrence records suggest a global decline in bee species richness. *One Earth*, 4(1), 114–123.

## 874 ACKNOWLEDGEMENTS

We thank all the taxonomists and ecologists that has made this database possible by contributing with their fieldwork data.

# FUNDING INFORMATION

This research was funded by the H2020 European project Safeguard (101003476) and by
the Federal State of Saxony-Anhalt (MLU-BioDivFund). RT was supported by the Czech
Science Foundation (Project No. 21-24186M). NDM, NH, YP and FM were financially supported by the ANR ARSENIC project (grant no. 14-CE02-0012), the ANR NGB project
(grant no. 17-CE32- 011), the Region Nord-Pas-de-Calais, the CNRS, the French Ministère de
l'Enseignement Supérieur et de la Recherche, the Hauts-de-France Region and the European
Regional Funds.

# **SET OF INTEREST**

886 None.

## 887 DATA AVAILABILITY

All data and code to produce of this database and manuscript are available at Zenodo (LINK) and Github (https://github.com/JoseBSL/EuPPollNet.git).

# SUPPORTING INFORMATION

Title: EuPPollNet: A European database of plant-pollinator networks

Authors: Jose B. Lanuza <sup>1,2,3</sup> Tiffany M. Knight <sup>3,2,4</sup> Nerea Montes-Perez <sup>1</sup> Will Glenny 892 <sup>3,4</sup>| Paola Acuña <sup>5</sup>| Matthias Albrecht <sup>6</sup>| Maddi Artamendi <sup>7,8</sup>| Isabelle Badenhausser <sup>9,10,11</sup>| 893 Joanne M. Bennett  $^{12}\vert$  Paolo Biella  $^{13}\vert$  Ricardo Bommarco  $^{14}\vert$  Andree Cappellari  $^{15}\vert$  Sílvia 894 Castro <sup>16</sup> | Yann Clough <sup>17</sup> | Pau Colom <sup>18,19</sup> | Joana Costa <sup>16,20</sup> | Nathan Cyrille <sup>21,22</sup> | Natasha 895 de Manincor <sup>23,24</sup> | Paula Dominguez-Lapido <sup>7</sup> | Christophe Dominik <sup>4,3</sup> | Yoko L. Dupont <sup>25</sup> | 896 Reinart Feldmann <sup>26</sup> Emeline Felten <sup>22</sup> Victoria Ferrero <sup>27</sup> William Fiordaliso <sup>28</sup> Alessandro Fisogni <sup>23</sup> | Úna Fitzpatrick <sup>29</sup> | Marta Galloni <sup>30</sup> | Hugo Gaspar <sup>16</sup> | Elena Gazzea <sup>15</sup> | Irina Goia <sup>31,32</sup> Carmelo Gómez-Martínez <sup>33</sup> Miguel A. González-Estévez <sup>33</sup> Juan Pedro González-Varo 899 <sup>34</sup> Ingo Grass <sup>35</sup> Jiří Hadrava <sup>36</sup> Nina Hautekèete <sup>23</sup> Veronica Hederström <sup>17</sup> Ruben Heleno 900 <sup>16</sup> Sandra Hervias-Parejo <sup>33</sup> Jonna M. Heuschele <sup>3,4</sup> Bernhard Hoiss <sup>37</sup> Andrea Holzschuh 901 <sup>37</sup> Sebastian Hopfenmüller <sup>38</sup> José M. Iriondo <sup>39</sup> Birgit Jauker <sup>40</sup> Frank Jauker <sup>41</sup> Jana 902 Jersáková <sup>42</sup> | Katharina Kallnik <sup>37</sup> | Reet Karise <sup>43</sup> | David Kleijn <sup>44</sup> | Stefan Klotz <sup>4</sup> | Theresia 903 Krausl <sup>17</sup> | Elisabeth Kühn <sup>45</sup> | Carlos Lara-Romero <sup>39</sup> | Michelle Larkin <sup>46</sup> | Emilien Laurent <sup>22</sup> | 904 Amparo Lázaro <sup>33</sup> | Felipe Librán-Embid <sup>47,48</sup> | Yicong Liu <sup>4,2</sup> | Sara Lopes <sup>16</sup> | Francisco López-905 Núñez <sup>16,49</sup> João Loureiro <sup>16</sup> Ainhoa Magrach <sup>7,50</sup> Marika Mänd <sup>43</sup> Lorenzo Marini <sup>15</sup> Rafel Beltran Mas <sup>33</sup> François Massol <sup>51</sup> Corina Maurer <sup>6</sup> Denis Michez <sup>24</sup> Francisco P. Molina <sup>1</sup> Javier Morente-López  $^{52}$ | Sarah Mullen  $^{53}$ | Georgios Nakas  $^{54}$ | Lena Neuenkamp  $^{55,56}$ | Arkadiusz 908 Nowak <sup>57,58</sup> | Catherine J. O'Connor <sup>16,59</sup> | Aoife O'Rourke <sup>53</sup> | Erik Öckinger <sup>14</sup> | Jens M. Olesen 909 <sup>60</sup> Øystein H. Opedal <sup>61</sup> Theodora Petanidou <sup>54</sup> Yves Piquot <sup>23</sup> Simon G. Potts <sup>62</sup> Eileen F. 910 Power <sup>63</sup> Willem Proesmans <sup>24,22</sup> Demetra Rakosy <sup>4,3,64</sup> Sara Reverté <sup>24</sup> Stuart P. M. Roberts 911 <sup>62</sup> Maj Rundlöf <sup>65</sup> Laura Russo <sup>66,53</sup> Bertrand Schatz <sup>67</sup> Jeroen Scheper <sup>44</sup> Oliver Schweiger 912 <sup>4,3</sup>| Pau Enric Serra <sup>33</sup>| Catarina Siopa <sup>16</sup>| Henrik G. Smith <sup>65,17</sup>| Dara Stanley <sup>68</sup>| Valentin 913 Ştefan <sup>4,3</sup> | Ingolf Steffan-Dewenter <sup>37</sup> | Jane C. Stout <sup>63</sup> | Louis Sutter <sup>69</sup> | Elena Motivans Švara <sup>3,4,2</sup>| Sebastian Świerszcz <sup>57,70</sup>| Amibeth Thompson <sup>2,3,71</sup>| Anna Traveset <sup>33</sup>| Annette Trefflich 915 <sup>72</sup> Robert Tropek <sup>73,74</sup> Teja Tscharntke <sup>48</sup> Adam J. Vanbergen <sup>22</sup> Montserrat Vilà <sup>1,75</sup> Ante 916 Vujić <sup>76</sup> | Cian White <sup>53</sup> | Jennifer B. Wickens <sup>62</sup> | Victoria B. Wickens <sup>62</sup> | Marie Winsa <sup>14</sup> | Leana 917 Zoller <sup>2,3,77</sup> Ignasi Bartomeus <sup>1</sup> 918

# 919 Contains:

891

- Supplementary text 1
- Figure S1
- Figure S2
- Figure S3
- Figure S4
- Figure S5
- Figure S6

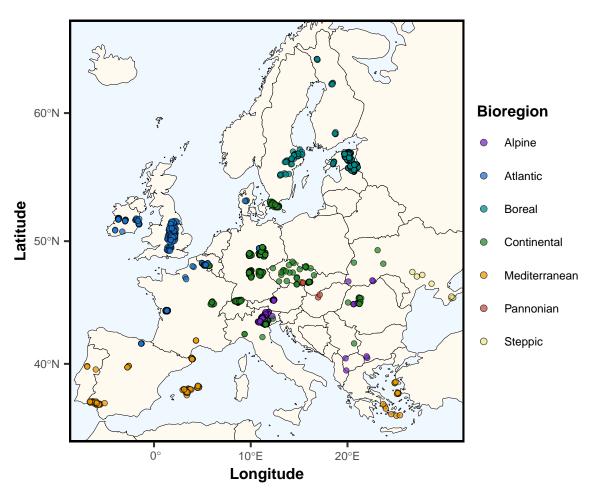
#### Supplementary text 1

EuPPollNet habitat definitions adapted from authors and CORINE Land Cover (CLC). To guide these definitions, they are connected to habitat types defined in the European Nature Information System (EUNIS 2021; https://eunis.eea.europa.eu/habitats-code-browser-revised.jsp). Specific examples from EUNIS habitats are provided when they can be precisely linked to EuPPollNet habitat types.

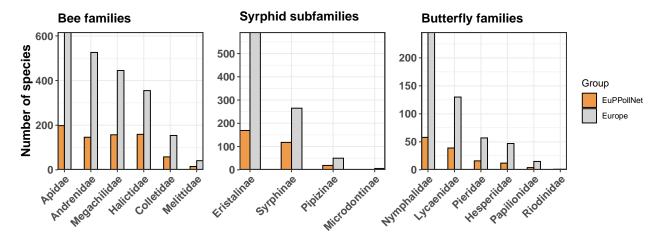
- 1) Ruderal vegetation: Plants growing on highly disturbed sites such as road sides or mineral extraction sites. This category partially matches the category of 'vegetated man made habitats' from EUNIS (code V; e.g., category of 'dry perennial anthropogenic herbaceous vegetation', code V38).
- 2) **Agricultural margins**: Sides of crops that can include any type of vegetation from low growing plants to trees. This category partially matches the category of 'vegetated man made habitats' from EUNIS (code V; e.g., category of 'hedgerows', code V4).
- 3) Green urban areas: Parks, private gardens or small pastures within an urban setting. Botanical gardens are included in this category. This category partially matches the category of 'vegetated man made habitats' from EUNIS (code  $\mathbf{V}$ ; e.g., category of 'cultivated areas of gardens and parks', code  $\mathbf{V2}$ ).
- 4) **Agricultural land**: Includes any type of crop and any type of vegetation growing within them. This category partially matches the category of 'vegetated man made habitats' from EUNIS (code V; e.g., category of 'intensive unmixed crops', code V11).
- 5) Forest/woodland understory: Any plant community sampled within a wooded group of plants. The forest may be situated in an agricultural setting or in a fully natural scenario. We have included agroforestry areas as well as open to dense forests in this category but excluded forest that contains sclerophyllous vegetation. This category is similar to the EUNIS habitat type 'forest and other wooded land' (code T).
- 6) Semi-natural grassland: Acidic, neutral, or calcareous unimproved grasslands located at elevations lower than 1000 m that are managed extensively through mowing or grazing but have not been improved, i.e., they have never received significant fertilizer or herbicide inputs. This category partially matches the category of 'grasslands and lands dominated by forbs, mosses or lichens' from EUNIS (code  $\bf R$ ).
- 7) Intensive grassland: Communities dominated by non-woody, low-growing plants that are heavily influenced by human disturbance, such as agriculture, mowing, moderate to high grazing, or urban environments. The plant composition is generally dominated by graminoid species. This category partially matches the EUNIS categories of 'vegetated man made habitats' (code V; e.g., category of 'artificial grasslands and herb dominated habitats', code V4) and 'grasslands and lands dominated by forbs, mosses or lichens' (code R).

- 8) Sclerophyllous vegetation: Any type of system with a dominant shrub community 964 adapted to drought. Typical of the Mediterranean region. Note, that we have included 965 in this category also woodlands (open coniferous forest) where the shrub community was 966 the main focus of the study. This category partially matches the **EUNIS** habitat type 967 'forest and other wooded land' (code  $\mathbf{T}$ ). 968
- 9) Beaches, dunes, sands: Plant communities growing on sandy soil. This category 969 partially matches the **EUNIS** habitat type 'coastal habitats' (code N). 970
- 10) Riparian vegetation: Plant communities growing on river margins. This category 971 partially matches the EUNIS habitat type 'forest and other wooded land' (code T) 972 but also the category of 'inland waters' which is currently under review. 973
- 11) Montane to alpine grasslands: Communities dominated by non-woody, low-growing 974 plants that experience little or no human disturbance and are found in high-elevation 975 areas (>1,000 m). This category partially matches the category of 'grasslands and lands 976 dominated by forbs, mosses or lichens' from EUNIS (code R; e.g., alpine and subalpine 977 grasslands with code  $\mathbf{R4}$ ). 978
- 12) Moors and heathland: Low-growing woody vegetation (i.e., typically < 5 m) that is 979 characteristic of poorly fertile soils where environmental conditions prevent the natural formation of forests. This category is similar to the **EUNIS** habitat type 'heathland, 981 scrub and tundra' (code S). 982

980



 $\textbf{Figure S1}. \ \ \text{Geographical location of all networks in the EuPPollNet database coloured by bioregion}. \\$ 



**Figure S2**. Coverage of the EuPPollNet species for bees (family level), syrphids (subfamily level) and butterflies (family level) in relation to the total number of European species within these taxonomic groups.

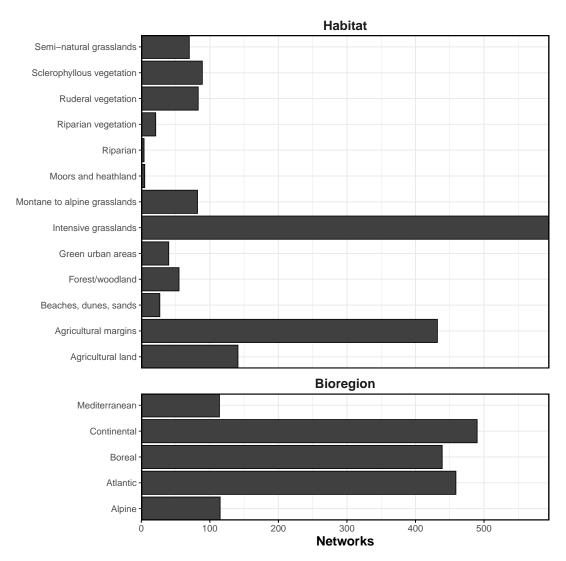
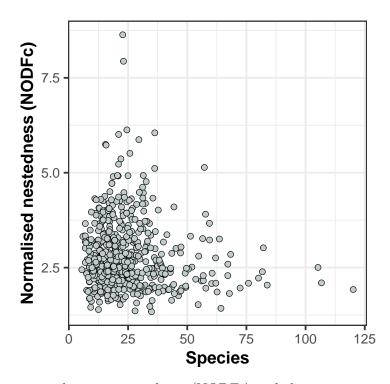
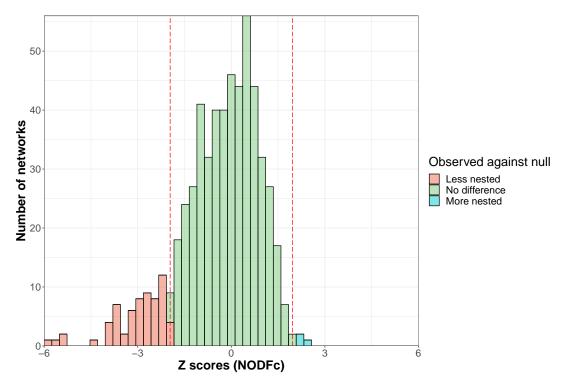


Figure S3. Barplot indicating the number of networks by habitat and bioclimatic region within the database.



 $\textbf{Figure S4}. \ \, \text{Association between nestedness (NODFc) and the geometric mean of plant and pollinator species per network } \\$ 



**Figure S5**. Distribution of z-scores comparing the nestedness of empirical networks with their randomized counterparts with the 'quasiswap\_count' algorithm (100 null models for each network). The quasiswap\_count algorithm maintains row and column sums with fixed connectance.

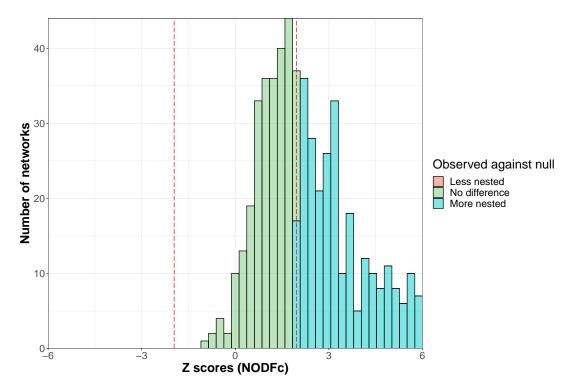


Figure S6. Distribution of z-scores comparing the nestedness of empirical networks with their randomized counterparts with a probabilistic null model from Bascompte et al., 2003 (100 null models for each network). This is the less restrictive approach, as it does not constrain the realised species degree. It first calculates the probability for each cell, and then if a random number between 0 and 1 exceeds this probability, an observed interaction is assigned to the cell. This shows that when the realised degree is not constrained, the number of networks that are statistically more nested is notably higher, highlighting that the nested structure is likely a byproduct of the species degree distribution.