

EuPPollNet: A European database of plant-pollinator networks

Jose B. Lanuza ^{1,2,3} | Tiffany M. Knight ^{3,2,4} | Nerea Montes-Perez ¹ | Will Glenny ^{3,4} | Paola Acuña ⁵ | Matthias Albrecht ⁶ | Maddi Artamendi ^{7,8} | Isabelle Badenhäusser ^{9,10,11} | Joanne M. Bennett ¹² | Paolo Biella ¹³ | Ricardo Bommarco ¹⁴ | Andree Cappellari ¹⁵ | Sílvia Castro ¹⁶ | Yann Clough ¹⁷ | Pau Colom ^{18,19} | Joana Costa ^{16,20} | Nathan Cyrille ^{21,22} | Natasha de Manincor ^{23,24} | Paula Dominguez-Lapido ⁷ | Christophe Dominik ^{4,3} | Yoko L. Dupont ²⁵ | Reinart Feldmann ²⁶ | Emeline Felten ²² | Victoria Ferrero ²⁷ | William Fiordaliso ²⁸ | Alessandro Fisogni ²³ | Úna Fitzpatrick ²⁹ | Marta Galloni ³⁰ | Hugo Gaspar ¹⁶ | Elena Gazzea ¹⁵ | Irina Goia ^{31,32} | Carmelo Gómez-Martínez ³³ | Miguel A. González-Estévez ³³ | Juan Pedro González-Varo ³⁴ | Ingo Grass ³⁵ | Jiří Hadrava ³⁶ | Nina Hautekèete ²³ | Veronica Hederström ¹⁷ | Ruben Heleno ¹⁶ | Sandra Hervias-Parejo ³³ | Jonna M. Heuschele ^{3,4} | Bernhard Hoiss ³⁷ | Andrea Holzschuh ³⁷ | Sebastian Hopfenmüller ³⁸ | José M. Iriondo ³⁹ | Birgit Jauker ⁴⁰ | Frank Jauker ⁴¹ | Jana Jersáková ⁴² | Katharina Kallnik ³⁷ | Reet Karise ⁴³ | David Kleijn ⁴⁴ | Stefan Klotz ⁴ | Theresia Krausl ¹⁷ | Elisabeth Kühn ⁴⁵ | Carlos Lara-Romero ³⁹ | Michelle Larkin ⁴⁶ | Emilien Laurent ²² | Amparo Lázaro ³³ | Felipe Librán-Embida ^{47,48} | Yicong Liu ^{4,2} | Sara Lopes ¹⁶ | Francisco López-Núñez ^{16,49} | João Loureiro ¹⁶ | Ainhoa Magrach ^{7,50} | Marika Mänd ⁴³ | Lorenzo Marini ¹⁵ | Rafel Beltran Mas ³³ | François Massol ⁵¹ | Corina Maurer ⁶ | Denis Michez ²⁴ | Francisco P. Molina ¹ | Javier Morente-López ⁵² | Sarah Mullen ⁵³ | Georgios Nakas ⁵⁴ | Lena Neuenkamp ^{55,56} | Arkadiusz Nowak ^{57,58} | Catherine J. O'Connor ^{16,59} | Aoife O'Rourke ⁵³ | Erik Öckinger ¹⁴ | Jens M. Olesen ⁶⁰ | Øystein H. Opedal ⁶¹ | Theodora Petanidou ⁵⁴ | Yves Piquot ²³ | Simon G. Potts ⁶² | Eileen F. Power ⁶³ | Willem Proesmans ^{24,22} | Demetra Rakosy ^{4,3,64} | Sara Reverté ²⁴ | Stuart P. M. Roberts ⁶² | Maj Rundlöf ⁶⁵ | Laura Russo ^{66,53} | Bertrand Schatz ⁶⁷ | Jeroen Scheper ⁴⁴ | Oliver Schweiger ^{4,3} | Pau Enric Serra ³³ | Catarina Siopa ¹⁶ | Henrik G. Smith ^{65,17} | Dara Stanley ⁶⁸ | Valentin Ștefan ^{4,3} | Ingolf Steffan-Dewenter ³⁷ | Jane C. Stout ⁶³ | Louis Sutter ⁶⁹ | Elena Motivans Švara ^{3,4,2} | Sebastian Świerszcz ^{57,70} | Amibeth Thompson ^{2,3,71} | Anna Traveset ³³ | Annette Trefflich ⁷² | Robert Tropek ^{73,74} | Teja Tšcharntke ⁴⁸ | Adam J. Vanbergen ²² | Montserrat Vilà ^{1,75} | Ante Vujić ⁷⁶ | Cian White ⁵³ | Jennifer B. Wickens ⁶² | Victoria B. Wickens ⁶² | Marie Winsa ¹⁴ | Leana Zoller ^{2,3,77} | Ignasi Bartomeus ¹

Corresponding author= barragansljose@gmail.com

All authors excluding the first four and last are ordered alphabetically

¹ Doñana Biological Station (EBD-CSIC), Seville, Spain, ² Martin Luther University Halle-Wittenberg, Institute of Biology, Halle, Germany, ³ German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, ⁴ Department of Community Ecology, Helmholtz Centre for Environmental

36 Research - UFZ, Halle, Germany, ⁵ Department of Plant Biology, Faculty of Science, University
 37 of Vigo, Vigo, Spain, ⁶ Agroecology and Environment, Agroscope, Zürich, Switzerland, ⁷ Basque
 38 Centre for Climate Change-BC3, Leioa, Spain, ⁸ University of the Basque Country, EuskalHerriko
 39 Unibertsitatea (UPV-EHU), Leioa, Spain, ⁹ Centre of Biological Studies of Chizé, La Rochelle
 40 University, Villiers en Bois, France, ¹⁰ LTSER "ZA Plaine & Val de Sèvre", CNRS, Villiers en Bois,
 41 France, ¹¹ Multidisciplinary Research Unit for Grasslands and Forage Crops, INRAE, Lusignan, France,
 42 ¹² Fenner School of Environment & Society, The Australian National University, Canberra, Australia,
 43 ¹³ ZooPlantLab, Department of Biotechnology and Biosciences, University of Milano-Bicocca, Milan,
 44 Italy, ¹⁴ Department of Ecology, Swedish University of Agricultural Sciences, Uppsala, Sweden, ¹⁵
 45 Department of Agronomy, Food, Natural Resources, Animals and Environment, University of Padua,
 46 Padua, Italy, ¹⁶ Department of Life Sciences, Centre for Functional Ecology, University of Coimbra,
 47 Coimbra, Portugal, ¹⁷ Centre for Environmental and Climate Science, Lund University, Lund,
 48 Sweden, ¹⁸ Department of Evolutionary Biology, Ecology, and Environmental Sciences, University of
 49 Barcelona, Barcelona, Spain, ¹⁹ Biodiversity Research Institute (IRBio), Barcelona, Spain, ²⁰ Linking
 50 Landscape, Environment, Agriculture and Food, School of Agriculture, University of Lisbon, Portugal,
 51 ²¹ Biogéosciences, UMR 6282 CNRS, University of Burgundy, Dijon, France, ²² Agroecology, INRAE,
 52 Institut Agro, University of Burgundy, University of Burgundy Franche-Comté, Dijon, France, ²³ Univ.
 53 Lille, CNRS, UMR 8198 - Evo-Eco-Paleo, F-59000 Lille, France, ²⁴ Laboratory of Zoology, Research
 54 Institute of Biosciences, University of Mons, Mons, Belgium, ²⁵ Department of Agroecology, Aarhus
 55 University, Denmark, ²⁶ Helmholtz Centre for Environmental Research - UFZ, Leipzig, Germany,
 56 ²⁷ Department of Biodiversity and Environmental Management, University of León, León, Spain, ²⁸
 57 Ecology of Interactions and Global Change, Research Institute in Biosciences, University of Mons,
 58 Mons, Belgium, ²⁹ National Biodiversity Data Centre, County Waterford, Ireland, ³⁰ Department of
 59 Biological, Geological and Environmental Sciences (BiGeA), University of Bologna, Bologna, Italy,
 60 ³¹ Faculty of Biology and Geology, Babeş-Bolyai University, Cluj-Napoca, Romania, ³² Centre
 61 for Systems Biology, Biodiversity and Bioresources (3B), Babeş-Bolyai University, Cluj-Napoca,
 62 Romania, ³³ Mediterranean Institute for Advanced Studies (IMEDEA, UIB-CSIC), Esporles, Spain,
 63 ³⁴ Department of Biology, Institute of Marine Research (INMAR), University of Cádiz, Puerto Real,
 64 Spain, ³⁵ Department of Ecology of Tropical Agricultural Systems, University of Hohenheim, Stuttgart,
 65 Germany, ³⁶ Department of Zoology, Faculty of Science, Charles University, Prague, Czechia, ³⁷
 66 Department of Animal Ecology and Tropical Biology, Biocenter, University of Würzburg, Würzburg,
 67 Germany, ³⁸ Cultural Landscape Günztal Foundation, Ottobeuren, Germany, ³⁹ Global Change
 68 Research Institute (ICG-URJC), Rey Juan Carlos University, Madrid, Spain, ⁴⁰ Justus Liebig
 69 University Giessen, Giessen, Germany, ⁴¹ Institute of Landscape Ecology and Resource Management,
 70 Justus Liebig University Giessen, Giessen, Germany, ⁴² Department of Ecosystems Biology, Faculty
 71 of Science, University of South Bohemia, České Budějovice, Czechia, ⁴³ Institute of Agricultural and
 72 Environmental Sciences, Estonian University of Life Sciences, Tartu, Estonia, ⁴⁴ Plant Ecology and
 73 Nature Conservation Group, Wageningen University, Wageningen, The Netherlands, ⁴⁵ Department
 74 of Conservation Biology & Social-Ecological Systems, Helmholtz Centre for Environmental Research
 75 - UFZ, Halle, Germany, ⁴⁶ Botany and Plant Science, School of Natural Sciences and Ryan Institute,
 76 University of Galway, Galway, Ireland, ⁴⁷ Institute of Animal Ecology and Systematics, Justus Liebig
 77 University of Gießen, Germany, ⁴⁸ Agroecology, University of Göttingen, Göttingen, Germany, ⁴⁹
 78 Research Centre for Natural Resources Environment and Society (CERNAS), Polytechnic Institute
 79 of Coimbra, Coimbra Agriculture School, Coimbra, Portugal, ⁵⁰ IKERBASQUE, Basque Foundation
 80 for Science, Bilbao, Spain, ⁵¹ Univ. Lille, CNRS, Inserm, CHU Lille, Institut Pasteur de Lille, U1019
 81 - UMR 9017 - CIIL - Center for Infection and Immunity of Lille, F-59000 Lille, France, ⁵² Plant
 82 Evolutionary Ecology, Institute of Ecology, Evolution and Diversity, Faculty of Biological Sciences,
 83 Goethe University Frankfurt, Max-von-Laue-Str. 13, 60438, Frankfurt am Main, Germany, ⁵³ Botany

84 Department, Trinity College Dublin, Dublin, Ireland, ⁵⁴ Laboratory of Biogeography & Ecology,
 85 Department of Geography, University of the Aegean, Mytilene, Greece, ⁵⁵ Department of Botany,
 86 Institute of Ecology and Earth Sciences, University of Tartu, Tartu, Estonia, ⁵⁶ Institute of Landscape
 87 Ecology, Münster University, Münster, Germany, ⁵⁷ Polish Academy of Sciences Botanical Garden,
 88 Center for Biological Diversity Conservation in Powsin, Warsaw, Poland, ⁵⁸ Botanical Garden of the
 89 Wrocław University, Wrocław, Poland, ⁵⁹ Cardiff School of Biosciences, Cardiff University, Cardiff, UK,
 90 ⁶⁰ Department of Biology, University of Aarhus, Aarhus, Denmark, ⁶¹ Division of Biodiversity and
 91 Evolution, Department of Biology, Lund University, Lund, Sweden, ⁶² Centre for Agri-Environmental
 92 Research, School of Agriculture, Policy and Development, University of Reading, Reading, UK, ⁶³
 93 Botany, School of Natural Sciences, Trinity College Dublin, Dublin, Ireland, ⁶⁴ Thünen-Institute of
 94 Biodiversity, Braunschweig, Germany, ⁶⁵ Department of Biology, Lund University, Lund, Sweden,
 95 ⁶⁶ Department of Ecology and Evolutionary Biology, University of Tennessee, Knoxville, TN,
 96 USA, ⁶⁷ CEFEE, CNRS, University of Montpellier, EPHE, IRD, Montpellier, France, ⁶⁸ School of
 97 Agriculture and Food Science, University College Dublin, Dublin, Ireland, ⁶⁹ Plant Productions
 98 Systems, Agroscope, Zürich, Switzerland, ⁷⁰ Institute of Agroecology and Plant Production, Wrocław
 99 University of Environmental and Life Sciences, Wrocław, Poland, ⁷¹ University of Freiburg, Chair of
 100 Nature Conservation and Landscape Ecology, Freiburg, Germany, ⁷² State Institute of Agriculture
 101 and Horticulture Saxony-Anhalt, Bernburg, Germany, ⁷³ Institute of Entomology, Biology Centre,
 102 Czech Academy of Sciences, České Budějovice, Czechia, ⁷⁴ Department of Ecology, Faculty of Science,
 103 Charles University, Prague, Czechia, ⁷⁵ Department of Plant Biology and Ecology, University of
 104 Seville, Seville, Spain, ⁷⁶ Department of Biology and Ecology, Faculty of Sciences, University of Novi
 105 Sad, Novi Sad, Serbia, ⁷⁷ Department of Ecology & Evolutionary Biology, University of Colorado,
 106 Boulder, CO, USA
 107

Abstract

Motivation: Pollinators play a crucial role in maintaining Earth’s terrestrial biodiversity and human food production by mediating sexual reproduction for most flowering plants. Indeed, the network of interactions formed by plants and pollinators constitutes the backbone of plant-pollinator community stability and functioning. However, rapid human-induced environmental changes are compromising the long-term persistence of plant-pollinator interaction networks. One of the major challenges for pollinator conservation is the lack of robust generalisable data capturing how plant-pollinator communities are structured across space and time. Here, we present the EuPPollNet (European Plant-Pollinator Networks) database, a fully open and reproducible European-level database containing harmonized taxonomic data on plant-pollinator interactions referenced in both space and time, along with other ecological variables of interest. This database offers an open workflow that allows researchers to track data-curation decisions and edit them according to their preferences. We present the taxonomic and sampling coverage of EuPPollNet, and summarize key structural properties in plant-pollinator networks. We hope EuPPollNet will stimulate future research that fills the taxonomic, ecological, and geographical data gaps on plant-pollinator interactions that we have identified. Further, the variation in the structure of the networks in EuPPollNet provides a strong basis for future studies aimed at quantifying drivers of plant-pollinator network change and guiding future conservation planning for plants and pollinators.

Main Types of Variables Included: EuPPollNet contains 1,162,913 interactions between plants and pollinators from 1,864 distinct networks (i.e., distinct sampling event in space or time), which belong to 54 different studies distributed across 23 European countries. In addition, information about sampling methodology, habitat type, bio-climatic region, and further taxonomic rank information for both plant and pollinator species are also provided (i.e., genus, family and order).

Spatial location and grain: The database contains 1,214 different sampling locations from 13 different natural and anthropogenic habitats that fall in 7 different bio-climatic regions. All records are geo-referenced and presented in the World Geodetic System 1984 (WGS84).

Time period and grain: Species interaction data was collected between 2004 and 2021. All records are time-referenced and most of the studies documented interactions within a single flowering season (68.52%).

Major taxa and level of measurement: The database contains interaction data at the species level for 94.39% of the records, including a total of 1,411 plant and 2,223 pollinator species. The database covers 5.56% of the European species of flowering plants, 34.38% of bees, 26.21% of butterflies, and 33.63% of syrphid species at the European level.

Software format: The database was built with the R programming language and is stored as “.rds” and “.csv” formats. The construction of the database is fully reproducible and can be accessed at the following link: <https://github.com/JoseBSL/EuPPollNet.git>.

146 **KEYWORDS**

147 Angiosperms, connectance, flowering plants, nestedness, plant-pollinator networks, pollinators, species
148 interactions

1 | INTRODUCTION

Plant-pollinator interactions involve a great diversity of species, largely attributed to their co-evolutionary history (Ollerton, 2017), and are critically important for terrestrial biodiversity and economic productivity. The synergistic effects of climate change with other global change pressures are threatening worldwide biodiversity (Bellard et al., 2014; Sala et al., 2000), including plant and pollinator species as well as their interactions (Eichenberg et al., 2021; Goulson et al., 2015; Settele et al., 2016). Under this scenario, the increasing availability of biodiversity data plays a major role in our ecological understanding of species status, trends, and conservation (Heberling et al., 2021; Zattara & Aizen, 2021). However, our knowledge of plant and pollinator species and their network of interactions still exhibits major temporal, spatial and taxonomic biases (Archer et al., 2014; Marshall et al., 2024; Poisot et al., 2021; Troia & McManamay, 2016), limiting our ability to effectively protect their biodiversity.

The interactions between different plant and pollinator species within a community form complex networks. Macro-ecological analyses of the topology of these networks have revealed common properties, such as truncated power-law degree distributions (Jordano et al., 2003) or nestedness (Bascompte et al., 2003). Large-scale analyses across multiple studies can quantify patterns across geographic regions (Olesen & Jordano, 2002; Traveset et al., 2016) or environmental gradients (Ramos-Jiliberto et al., 2010; Rech et al., 2016; Saunders et al., 2023) that cannot be examined in a single study. Although macro-ecological approaches that use ecological interactions make significant contributions to knowledge (Windsor et al., 2023), such synthesis work must consider variation across studies in the spatio-temporal nature of the data (Burkle & Alarcón, 2011; García et al., 2024). For instance, plant-pollinator studies tend to differ in sampling effort and methodology which affect the structure of the resulting plant-pollinator networks (Gibson et al., 2011; Jordano, 2016; Schwarz et al., 2020). Most plant-pollinator networks have unobserved interactions (Chacoff et al., 2012; Olesen et al., 2011), and thus research that attempts to synthesize across published studies must have access to raw data on interactions in order to statistically account for sampling effort and completeness. This emphasizes the importance of providing data in its rawest possible form in datasets that will be utilized for synthesis and macro-ecological studies.

Europe is one of the continents with a larger amount of available biodiversity data (Proença et al., 2017), yet still exhibits major gaps (Bennett et al., 2018; Wetzel et al., 2018). While species checklists need to be treated carefully, especially at a macro-ecological scale (Grenié et al., 2023), the growing number of European plant and pollinator checklists (Reverté et al., 2023), along with occurrence data (Zattara & Aizen, 2021), is setting a foundation for the conservation of its flora and their pollinators. However, species richness is just one component of biodiversity and documenting the interaction between plants and pollinators is essential for understanding biodiversity change (Jordano, 2016). Numerous works have studied plant-pollinator interactions in the last decades, generating thousands of plant-pollinator interaction networks worldwide. Several initiatives have tried to integrate plant-pollinator interaction data into databases such as *Mangal* (Poisot et al., 2016) or *GloBI* (Poelen et al., 2014), resulting

in numerous large scale comparative analyses that have enhanced our understanding of the ecology of plants and pollinators (e.g., European wild bee data trends; [Marshall et al., 2024](#)). Despite all these resources, Europe lacks accessible harmonized plant-pollinator interaction data that allow researchers to evaluate plant-pollinator interactions at a European level. For example, only over a dozen of European plant-pollinator networks are included in *Mangal*, while *GloBI* focuses on pairwise interactions disconnected from the community context. Assembling and curating the existing information on EU plant-pollinator networks will guide research efforts, conservation planning and will set a foundation for future global change research.

Here, we present the European Plant-Pollinator Networks database (EuPPollNet), which contains harmonized information on plant-pollinator interactions at the European level. The pollinator taxonomic groups include the main orders of entomofauna that visit and pollinate flowering plants in Europe. These comprise insect species from the orders Hymenoptera, Diptera, Lepidoptera, and Coleoptera, accounting for almost the totality of recorded interactions in EuPPollNet (99.86%). To understand the scope of the database, we examined the taxonomic and sampling coverage of the different plant and pollinator species at the European level with the help of the most up-to-date species checklists and rarefaction analyses. In addition, for bees and plants, we evaluated if there is a phylogenetic signal in the presence-absence of interaction data.

EuPPollNet contains one of the largest sets of plant-pollinator networks collated to date, providing a unique opportunity to examine the prevalence of key structural metrics across networks. For example, despite the large theoretical literature on the meaning of a nested structure in plant-pollinator networks ([Bascompte & Jordano, 2007](#); [Guimaraes, 2020](#)), where specialists species interact only with subsets of generalists species, this pattern has only been empirically evaluated with a relatively small number of networks ([Bascompte et al., 2003](#); [Payrató-Borras et al., 2019](#); [Staniczenko et al., 2013](#)), and is still debated how structural metrics such as connectance and nestedness change across latitudes and bioclimatic regions ([Olesen & Jordano, 2002](#); [Song et al., 2017](#); [Trøjelsgaard & Olesen, 2013](#)).

Overall, EuPPollNet aims to cover a wide range of taxonomic groups and habitats, while also providing other variables of interest that define the ecological context and sampling methods of the study. In addition, EuPPollNet offers a transparent and accessible workflow of its data management and species harmonization that allows the database to be reused and to expand over time. This database provides a large number of community-level networks with curated and harmonized data, distinguishing it from other currently available resources that contain plant-pollinator interactions. We expect that EuPPollNet can be used to evaluate macro-ecological processes in plant-pollinator networks, guide conservation planning, and set a baseline for global change research.

2 | METHODS

Data acquisition

The EuPPollNet database includes published and unpublished studies compiled initially by a wide number of researchers and institutions within the European continent as defined by the European Environment Agency (Stanners & Bourdeau, 1995). As this database is the result of one of the work packages of the European project Safeguard (Safeguarding European wild pollinators; <https://doi.org/10.3030/101003476>), first, data was directly requested from members of the Safeguard project in May 2022. Second, the request was extended to data owners outside of the project, with data collection concluding in August 2024. These other data owners were identified by direct communication with colleagues suggested by Safeguard members and by directly searching for studies on Google Scholar of under-represented regions within the database. While Google Scholar lacks reproducibility (Gusenbauer & Haddaway, 2020), it still remains the most comprehensive search engine to date (Gusenbauer, 2019). This approach maximized the potential number of studies that could be incorporated in this database. The search strings used were ‘*plant-pollinator interactions*’ and ‘*plant-pollinator networks*’. To maintain high quality standards that will support robust future ecological research, we only included studies that met the following criteria: 1) studies containing time- and geo-referenced records of plant-pollinator interactions; and 2) studies that quantify interactions by documenting the contact between a floral visitor - referred to as a ‘pollinator’ throughout the manuscript, even though pollination efficiency is not evaluated - and the reproductive structure of a specific sampled plant (i.e., phyto-centric networks).

Dataset description

The database contains 54 independent published and unpublished studies conducted during the time period 2004 - 2021 in 23 different countries (Figure 1a and Figure 1b; see Figure S1 for exact locations). The studies differ in sampling effort and methodology, and thus documenting sampling methods and sampling effort is an important feature of EuPPollNet. Most studies took place within a single flowering season (68.52%), sampled a given location for an average of 6.99 days, and exclusively sampled diurnal plant-pollinator interactions, with transects being the most common sampling method (64.81%). All the studies documented interactions with Hymenopterans (with 50.00% considering all Hymenopterans, 46.15% only wild bees and 3.85% only bumblebees), 92.31% documented interactions with Dipterans (with 46.15% considering all Dipterans, 46.15% only syrphids and 5.77% recorded syrphids plus bombylids or tachinid flies), 63.46% with Lepidopterans, and 32.69% with Coleopterans. The database includes a total of 1,162,913 distinct interactions. Most of the pollinator species belong to the orders Hymenoptera, Diptera, Lepidoptera and Coleoptera (89.11%), which account for nearly all interactions in this database (99.86%). Species that belong to other orders (10.89%) are not explored in this study as they account for a minor fraction of the total interactions (0.14%). Hymenoptera and Diptera contain the highest number of species,

with each comprising approximately 1,000 species in the database. However, the majority of plant-pollinator interactions are from Hymenoptera species (90.15%; **Figure 1c**). Notably, the western honey bee, *Apis mellifera*, represents 69.89% of the total interaction records from the database and an average of 30.74% of the total interactions per network.

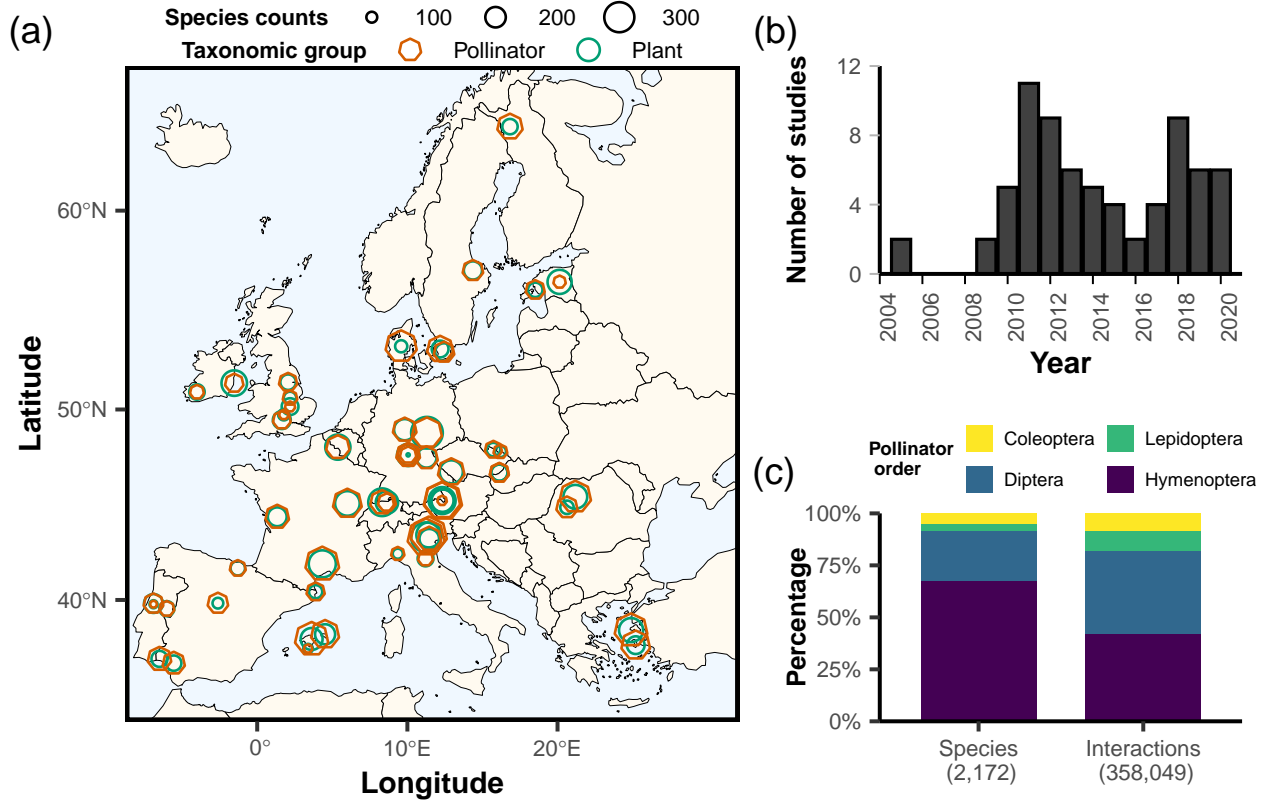


Figure 1. (a) Locations of the studies in EuPollNet across the European continent, showing the total number of pollinator (i.e., orange heptagon) and plant (i.e., green circles) species per study. The sizes of these shapes are proportional to the respective species counts. For visualization purposes, we have selected only a single location per study. (b) Number of studies by year in EuPollNet. (c) Proportion of species and interactions across the four main pollinator orders in EuPollNet, excluding interactions from *Apis mellifera*. The total number of species and interactions is indicated in parentheses at the bottom.

Data structure

The EuPPollNet database is available in both .csv and .rds formats and contains a total of 30 columns (**Table 1**), where each row represents a single interaction between a plant and a pollinator species. These columns include information about the study and network identifiers (columns 1 and 2), sampling method (3), habitat type as described by the author, and a unified habitat classification across studies (4 and 5), bioregion where the network is located (6), country, locality, and latitude-longitude coordinates (7 to 10), date of the interaction (11), number of interactions (12), taxonomic information about plants (13 to 20), taxonomic information about pollinators (21 to 28), and information about the availability of floral count data (29). The flower count data is provided in a separate file (.csv or .rds) and can be merged with the interaction data through the “Flower_data_merger” column (30). Note that although two-thirds of studies include information on floral abundance, the methods and units vary greatly across studies. To construct a plant-pollinator network matrix within a single flowering season at the site level, users should group interactions by plant and pollinator species, site, study, and year. Finally, metadata at the study level is provided in a separate file, including information about the authors, digital object identifier (if available), sampling time, and taxonomic coverage of the main pollinator groups for each study.

284 Taxonomic harmonization

285 All plant and pollinator species names were checked and harmonized in R using **rgbif** (Cham-
286 berlin et al., 2022). The protocol for plants and pollinators is similar but slightly different
287 given the availability of the different taxonomic resources. For transparency, we have included
288 in the database the original species name, the new assigned name, and, if the name of the
289 species is uncertain (e.g., species complex or species alike). In addition, taxonomic informa-
290 tion at genus, family and order level was downloaded for each species.

291 For plants: (i) we initially verified the exact matches against the GBIF species checklist; (ii)
292 we selected unmatched cases and fixed orthographic errors; (iii) we retrieved again taxonomic
293 information for those unmatched cases, evaluated accuracy of fuzzy matching and manually
294 fixed records that are still not found; (iv) finally, we used the World Flora Taxonomic Backbone
295 (Govaerts et al., 2021; WFO, July 7, 2022) as the ultimate filter for taxonomic information as
296 we used it to calculate the plant taxonomic coverage of our database.

297 For pollinators: (i) we first created a checklist of species names for the most representative
298 pollinator groups at the European level by combining the most up to date published checklists
299 of bees and syrphids (Reverté et al., 2023), and butterflies (Wiemers et al., 2018); (ii) we
300 compared pollinator species names against the checklist and recovered some unmatched cases
301 with restrictive fuzzy matching by using **stringdist** package (Van der Loo et al., 2014); (iii)
302 we fixed unmatched records when necessary and retrieved the taxonomic information for all
303 species from GBIF; (iv) we fixed the non-found cases in the GBIF checklist and made sure that
304 all species names from bees, syrphids and butterflies were named according to their respective
305 species checklists. Coleoptera species names were only checked against the GBIF checklist.

306 Taxonomic coverage

307 To assess the completeness of plant and pollinator species in the EuPPollNet database at the
308 European level, we used the aforementioned checklists for plants and pollinators. Specifically
309 for plants, we refined the checklist to include only plants occurring in Europe and excluded
310 taxonomic groups not associated with biotic pollination. We did this by first excluding the fam-
311 ilies considered to have exclusively a wind pollination mode (see Culley et al., 2002), and then
312 by filtering out the genera with wind or non-biotic pollination from families that exhibit both
313 biotic and non-biotic pollination modes. Additionally, we manually included exotic species and
314 added unresolved species names that were not present in the accepted names of the checklist
315 at the current version of usage. For pollinators, we only evaluated the taxonomic coverage of
316 groups with species checklists available in Europe (i.e., bees, syrphids and butterflies). While
317 there is not a good understanding of pollinator diversity in other taxonomic groups (e.g., bee-
318 tles), it was assumed that their coverage within the database is equal to the average coverage
319 of bees, syrphids, and butterflies (mean coverage = 31.4; sd = 4.51). Therefore, the total
320 number of of flower visiting species from other taxonomic groups (i.e., non-bee, non syrphid
321 and non-butterfly flower visitors) at European level was extrapolated by assuming that their

Table 1. Column names and their descriptions within the EuPPollNet database.

Variable	Description
1. Study_id	Identifier of the study
2. Network_id	Identifier of a site sampled within a study
3. Sampling_method	Type of plant-pollinator sampling
4. Authors_habitat	Type of habitat as described by the authors
5. EuPPollNet_habitat	Type of habitat homogenized across studies
6. Bioregion	European biogeographical regions
7. Country	Country where the plant-pollinator interaction was observed
8. Locality	Locality where the plant-pollinator interaction was observed
9. Latitude	North-south position of the observed interaction in decimal degrees
10. Longitude	East-west position of the observed interaction in decimal degrees
11. Date	Year, month and day when the observation took place
12. Interaction	Number of interactions. By default is 1 as interactions are provided ungrouped
13. Plant_original_name	Plant species name given by the authors
14. Plant_accepted_name	Harmonized plant species name in the database
15. Plant_rank	Taxonomic rank of the observation
16. Plant_order	Order taxonomic rank of the observed plant species
17. Plant_family	Family taxonomic rank of the observed plant species
18. Plant_genus	Genus taxonomic rank of the observed plant species
19. Plant_unsure_id	Category to indicate if the plant species name is unsure (Yes) or not (No)
20. Plant_uncertainty_type	If the name is unsure, type of species uncertainty is provided
21. Pollinator_original_name	Pollinator species name given by the authors
22. Pollinator_accepted_name	Harmonized pollinator species name in the database
23. Pollinator_rank	Taxonomic rank of the observation
24. Pollinator_order	Order taxonomic rank of the observed pollinator species
25. Pollinator_family	Family taxonomic rank of the observed pollinator species
26. Pollinator_genus	Genus taxonomic rank of the observed pollinator species
27. Pollinator_unsure_id	Category to indicate if the pollinator species name is unsure (Yes) or not (No)
28. Pollinator_uncertainty_type	If the name is unsure, type of species uncertainty is provided
29. Flower_data	Floral data availability (Yes) or (No)
30. Flower_data_merger	Column to merge floral data with the interaction dataset

coverage is equal to the mean coverage of bees, syrphids, and butterflies. Consequently, we provide an estimate for the total number of pollinators across the European continent.

Finally, to evaluate if the presence-absence of interaction records for bees and flowering plants follows a phylogenetic pattern within the database, we calculated its phylogenetic signal at genus and family level, respectively. The phylogenetic signal was calculated by using the *phylosig* function from the **phytools** package (Revell, 2012). We extracted the phylogenetic information for bees from a genus level phylogeny (Hedtke et al., 2013) and processed it using the packages **ape** (Paradis et al., 2019), **MCMCglmm** (Hadfield, 2010) and **phytools**. For plants, the phylogenetic tree was obtained from a species level plant phylogeny (Smith & Brown, 2018) with the help of the **rtree** package (Li, 2023).

Sampling coverage

The completeness of the EuPPollNet database was evaluated by exploring the rarefied accumulation curves of plant and pollinator species and their interactions across the different networks. In addition, we computed the accumulation curve of pollinator species with an increasing number of plant species as an indicator for how many pollinator species are likely responsible for the pollination of flowering plants (e.g., Kleijn et al., 2015 for crops). The rarefied and extrapolated sampling curves were obtained using the **iNEXT** package (Hsieh & Chao, 2016). The different rarefied curves were complemented with 100 bootstrapped accumulation curves.

Habitat type and bioclimatic region

We describe the habitat type for each site using information from Corine Land Cover (CLC, version 2018) extracted using the Terra package (Hijmans et al., 2022), visual inspection of Google Earth imagery and the habitat classification from the authors. These different habitat categories (see definitions in supplementary text) allow a quick comparison and understanding of the habitat types from the database. Moreover, Europe is characterized by a great variety of environmental conditions that harbor different biota. Thus, to allow authors to explore the set of studies that share similar environmental conditions and species, we assigned a biogeographical region to each site. The biogeographical regions were downloaded from the European Environment Agency (version 2016) and were matched to the different sites using a spatial join from the **sf** package (Pebesma et al., 2018).

Network analyses

To provide a general overview of the structure of plant-pollinator networks in EuPPollNet, we quantified connectance and nestedness for each network and examined how these network metrics change across different latitudes and bioclimatic regions in Europe. We selected these two network metrics as they are commonly evaluated in plant-pollinator network studies and

capture structural properties with a straightforward interpretation. We implemented “standardised” versions of connectance and nestedness to account for the effect of sampling effort on network metrics. As connectance is negatively associated with network size (Jordano, 1987), we evaluated how network connectance was associated with the number of species (i.e., log of geometric mean of plants and pollinators) and extracted the residuals from this association (i.e., residual connectance) as a measurement of corrected connectance. The relationship between residual connectance and species richness was investigated using a beta regression. This approach was chosen because connectance displays a non-normal distribution with continuous values bounded between 0 and 1. The model was implemented using the `betareg` package (Cribari-Neto & Zeileis, 2010). We used NODFc to compare nestedness across networks, as it corrects by connectance and the number of species in comparison to other nestedness metrics that change with network size (Song et al., 2017). This metric was calculated using the **maxn-odf** package (Hoeppke & Simmons, 2021). Both residual connectance and NODFc were used as dependent variables to evaluate their association with latitude. In addition, to quantify how connectance and nestedness change with network size, we determined their association with the number of species per network using the Kendall rank correlation coefficient.

Finally, to compare if networks are more or less nested than expected by chance, we employed the traditional z-score approach with the widely used nestedness metric (NODF) from Almeida-Neto et al. (2008). The z-score approach only compares each unique network against their randomized versions, avoiding the influence of network size. As NODF is a metric computed from binary matrices, we calculated 100 null models for the binarized version of each network using the ‘curveball’ algorithm (Strona et al., 2014). This algorithm implements the configuration model (i.e. random rewiring of all links, without self-links or double links) and thus keeps the exact number of connections per species (i.e., realised degree). Since null model selection can significantly influence statistical results (Kaiser, 2015), we also implemented a null model that reorganizes the quantitative networks before binarising them. To that end, we used the ‘quasiswap_count’ algorithm, which is a non-sequential algorithm for quantitative networks that maintains constant connectance and the number of connections per species. Each empirical network was randomized 100 times using each method. These null models were implemented with the **vegan** (Oksanen et al., 2013) package. Both connectance and nestedness (NODF) were estimated for each network using the function *networklevel* from **bipartite** (Dormann et al., 2008).

3 | RESULTS

Taxonomic coverage

Europe hosts approximately over 5,000 species of pollinators, including 2,138 bee species, 913 syrphid species, 496 butterflies species, and about 1,700 species from other taxonomic groups, as estimated based on EuPPollNet average taxonomic coverage. Additionally, there are around 25,000 species of plants that benefit from animal pollination when excluding non-biotic pollinated species. EuPPollNet contains a total of 2,223 pollinators and 1,411 plant species. The coverage of the main pollinator groups occurring in Europe is 34.38% for bees, 33.63% for syrphids and 26.21% for butterflies (see **Figure S2** for coverage at the family level for bees and butterflies, and at the subfamily level for syrphids). Bees (i.e., Anthophila) constitute 89.65% of the interactions in EuPPollNet, and 77.95% of the interactions when excluding honey bees. Within the database, 83.82% of bee genera have at least one species with interaction records, and the average coverage of species at the bee genus level is 36.99% (**Figure 2**). The presence or absence of interaction records for bees does not follow a phylogenetic pattern ($\lambda = 0.07$; $P = 0.65$). The database coverage of all flowering plant species occurring in Europe is 5.56% (**Figure 3**), with an average coverage of 9.04% at the plant family level. Approximately, half of the plant families have at least one species with interaction records (52.56%), and the presence or absence of interaction data for the different plant species also does not follow a statistically relevant phylogenetic pattern ($\lambda = 0.26$; $P = 0.07$).

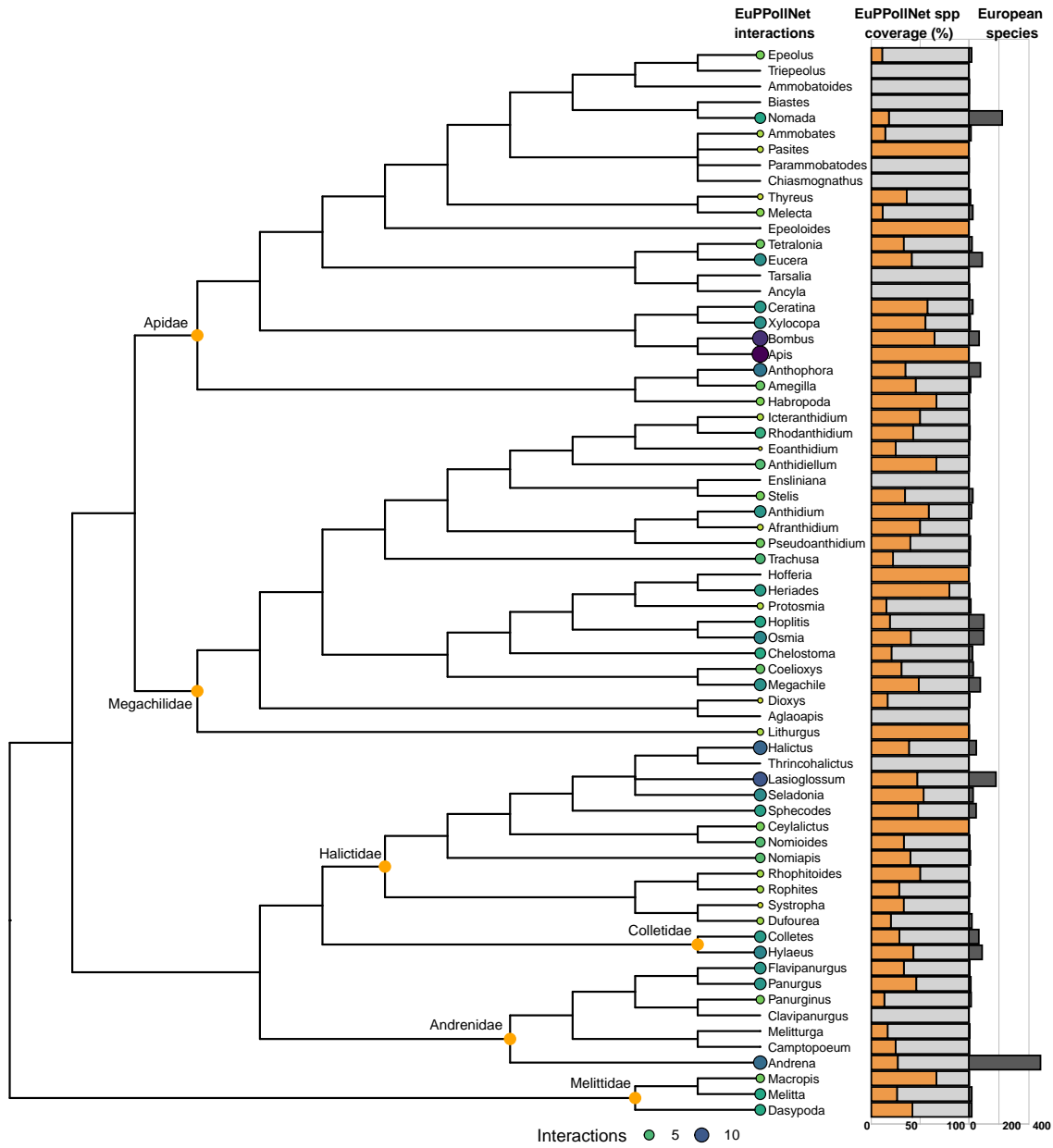


Figure 2. Phylogenetic and taxonomic coverage of bee genera at European level. The number of interactions recorded per genus in the database is illustrated using circles, with their sizes proportional to the number of interactions on a logarithmic scale, complemented by a gradient of colors ranging from yellow to dark purple. Additionally, the coverage of species recorded in EuPPollNet for each genus is depicted using two types of bars: orange bars representing the percentage of species included in the database and light grey bars indicating the percentage of species not included in the database, out of the total number of bee species in Europe. Dark grey bars represent the total number of species per genus at European level.



Figure 3. Phylogenetic and taxonomic coverage of the plant families at European level. The number of interactions recorded per family in the database is illustrated using circles, with their sizes proportional to the number of interactions on a logarithmic scale, complemented by a gradient of colors ranging from yellow to dark purple. Additionally, the coverage of species recorded in EuPPollNet for each family is depicted using two types of bars: orange bars representing the percentage of species included in the database and light grey bars indicating the percentage of species not included in the database out of the total number of bee species per family in Europe. Dark grey bars represent the total number of species per family at European level.

Sampling coverage

The estimated sampling coverage of plant and pollinator species within EuPPollNet across the different networks is approximately 97% for both taxonomic groups. This suggests that the rarefied accumulation curves of both plant and pollinator species exhibit already a “quasi-asymptotic” growth of species richness by considering the current number of networks (**Figure 4a-4b**). The predicted observed species richness by doubling the sampling effort on the already sampled habitat types within the database will only increase pollinator richness by 23.79% and plant richness by 21.53%. However, the sampling coverage of interactions is 74.35%, and by doubling the sampling effort the predicted number of unique interactions recorded will have approximately a twofold increase (53.68%; **Figure 4c**). When we consider the accumulated pollinator richness across sampled plant species, this curve also shows a “quasi-asymptotic” growth with a sampling coverage value of 96.54%. The predicted recorded pollinator species by doubling the number of plants sampled is expected to increase by 22.22% (**Figure 4d**). We find that a small portion of plant species and pollinator species are shared across a broad range of networks and that most plant (85.68%) and pollinator (87.72%) species are exclusively found in less than 1% of networks (**Figures 4e-4f**). The most common plant (*Trifolium pratense*) and pollinator (*Bombus pascuorum* when excluding *Apis mellifera*) species are found in 36.07% and 62.70% of networks, respectively.

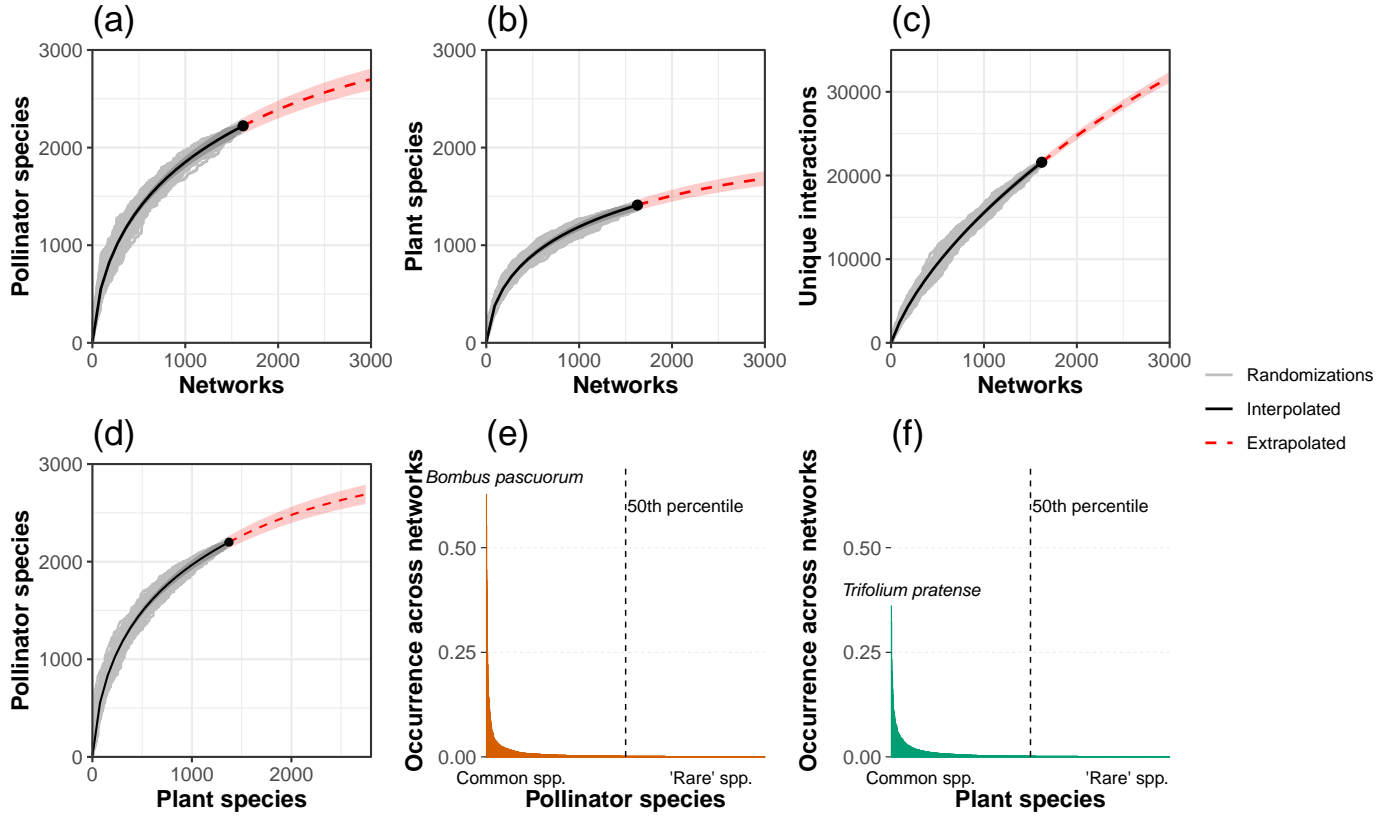


Figure 4. Graphs (a-b-c) indicate the accumulation curves for pollinators, plants, and the number of unique pairwise interactions across networks. Grey solid lines represent 100 randomized accumulation curves, the black solid lines represent the interpolated curve (i.e., the mean across curves), and the red dashed lines illustrate the extrapolated curve for approximately 3000 networks. The solid black points indicate the number of species and interactions contained in the database. Graph (d) shows the accumulation curve of pollinator species across an increasing number of plant species. This last graph uses the same color and shape structure as the ones in the top panel. Graphs (e-f) indicate the percentage of occurrence (i.e., incidence) of plant and pollinator species across networks. Species on the left (i.e., common) are found in many networks, while species on the right (i.e., rare) are found in few or only a single network. Note that *Apis mellifera* is the most common pollinator but was excluded from this visualization.

Habitat type and bioclimatic region

The proportion of species from the major pollinator orders within the database differed across habitats and bioclimatic regions (**Figure 5**). Hymenoptera was the main taxonomic order in the majority of habitats, exceeded only by Diptera for the habitat categories of riparian vegetation and moors and heathland. Overall, the proportion of flower visitors from Lepidoptera and Coleoptera were low across all habitats but Coleopteran flower visitors were notably more abundant in sclerophyllous vegetation and beaches, dunes and sands habitat categories. Similar patterns were observed when exploring the pollinator proportions by bioclimatic region. Hymenopterans were abundant across all bioclimatic regions and Dipterans were particularly abundant in the Boreal, Alpine and Atlantic regions. Lepidopterans had low proportions across all bioclimatic regions and Coleopterans were only relevant in the Mediterranean region at European level. Notably, the number of studies (**Figure 5**) and sampling sites (**Figure S3**) also differed across habitats and bioclimatic regions. The habitats sampled by a higher number of studies in the database were intensive grasslands (26), semi-natural grasslands (15) and sclerophyllous vegetation (10). However, the habitats that contain a higher number of sampling sites were intensive grasslands (601), agricultural margins (432) and agricultural land (141). The bioclimatic regions with a higher number of studies were Continental (24), Atlantic (13) and Mediterranean (13); and those that contain a higher number of sampling sites were Continental (490), Atlantic (459) and Boreal (439).

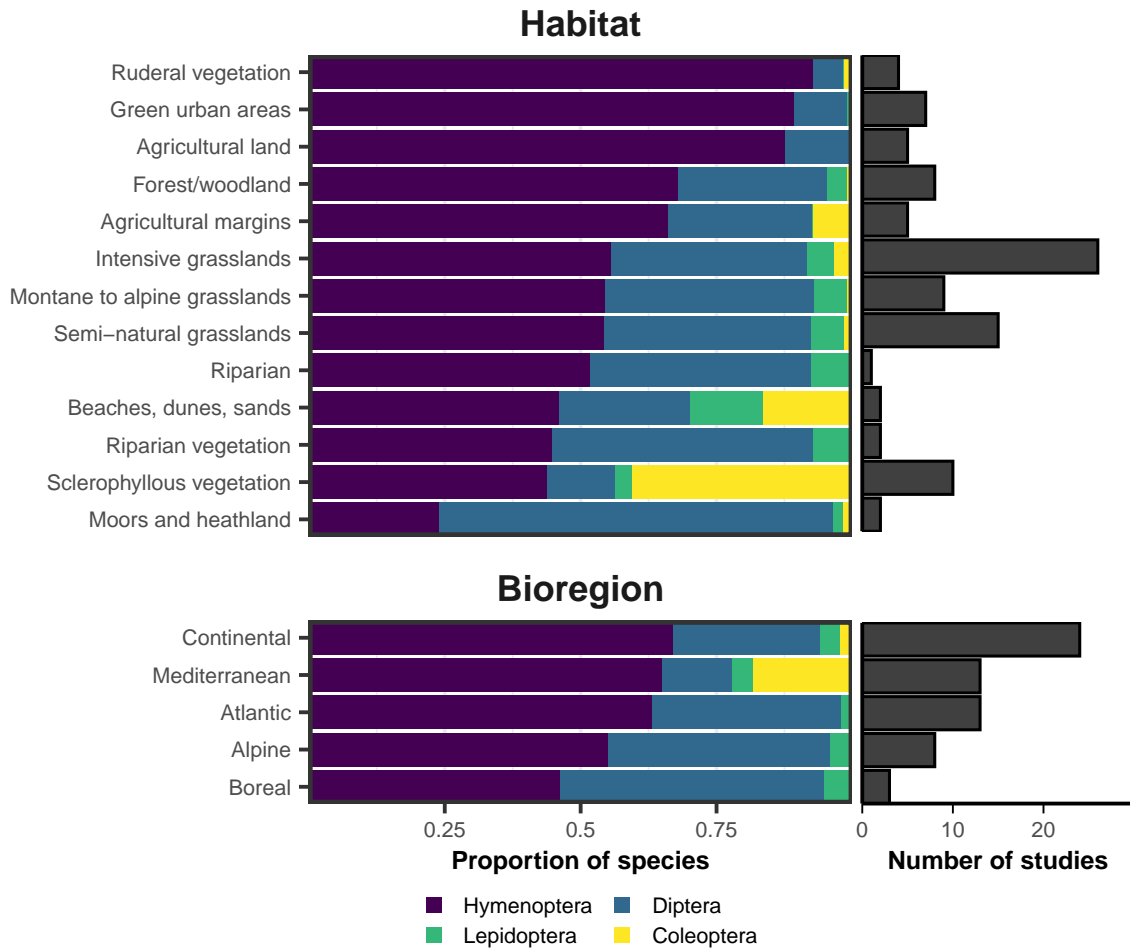


Figure 5. Proportion of species from the major pollinator orders by habitat types and bioclimatic regions in the EuPPollNet database. The orders, from left to right, include Hymenoptera, Diptera, Lepidoptera and Coleoptera. The horizontal barplot on the right indicates the number of studies that were conducted on each habitat type or bioclimatic region. Note that a single study can contribute to more than one habitat or bioclimatic region. Areas with a greater number of studies are more likely to depict accurate proportions of the different pollinator orders in those systems. The Pannonian and Steppic bioregions were excluded from this visualization because they contain only few networks from a single sampling day and site.

444 Network properties

445 Connectance values ranged between 0.03 to 0.4 ($\bar{x} = 0.14$) and followed a negative exponential
446 relationship with the number of species per network (Kendall $\tau = -0.75$, $P < 0.01$; **Figure 6a**).
447 Nestedness values (NODFc) ranged between 1.34 to 8.63 ($\bar{x} = 2.87$), and, as expected, were
448 not independent of the mean number of species (Kendall $\tau = -0.05$, $P = 0.08$; **Figure S4**).
449 Although latitude has a statistically significant impact on residual connectance, it explains
450 only a small portion of the observed variability in residual connectance and nestedness across
451 networks. (connectance: $R^2 = 0.02$, $P < 0.01$, **Figure 6c**; NODFc: $R^2 \sim 0$, $P = 0.83$, **Figure**
452 **6d**). Overall, networks towards higher latitudes showed lower residual connectance but similar
453 nestedness compared to networks located at lower latitudes. Note that residual connectance
454 and normalised nestedness showed a moderate significant negative correlation (Kendall $\tau =$
455 -0.43 , $P < 0.01$). Empirical networks did not show statistically different nestedness to the
456 simulated ones (**Figure 6b** and **Figure S5**). The ‘curveball’ method for binary networks
457 resulted in 12.02% of networks statistically less nested than null expectations, 85.71% showing
458 no difference, and 1.57% being more nested. The ‘quasiswap_count’ algorithm for quantita-
459 tive networks resulted in 11.4% of networks being less nested, 88.07% showing no difference
460 and 0.53% being more nested. Note that in both cases, NODFc is calculated on binarised
461 matrices.

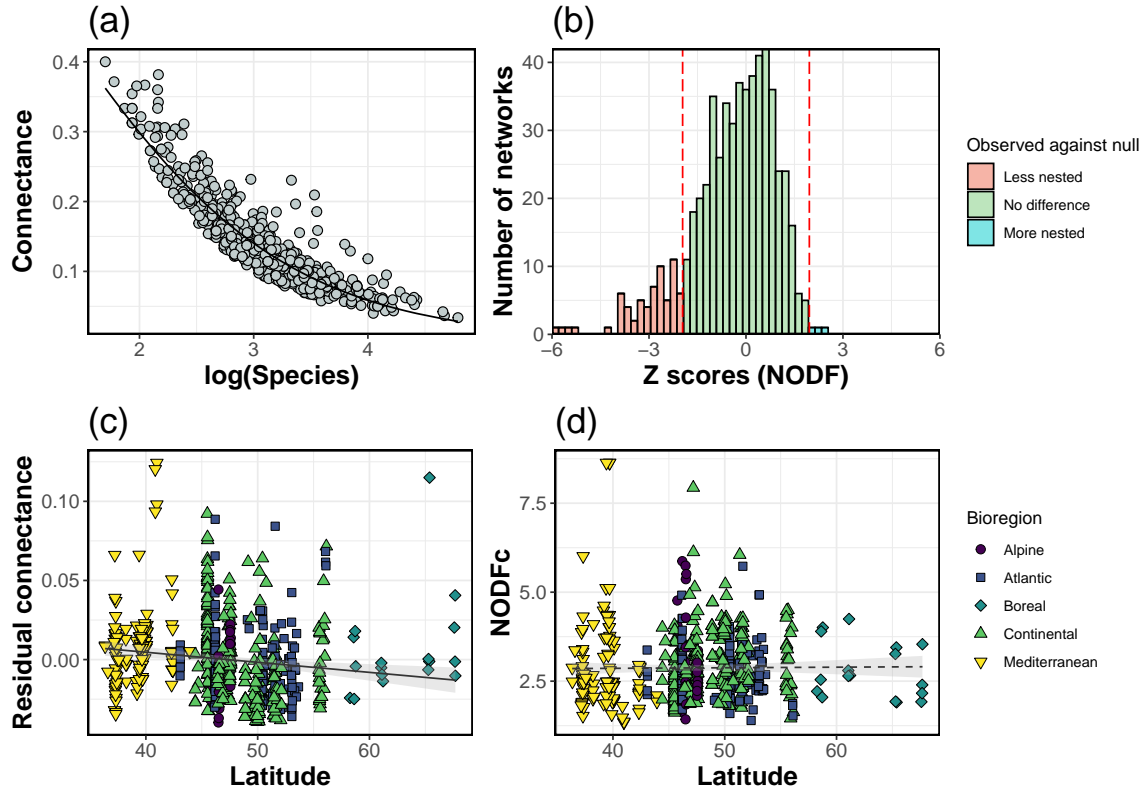


Figure 6. Graph (a) shows the association between network connectance and the geometric mean of plant and pollinator species per network on a log-scale with the respective fitted line from a Beta regression. Graph (b) shows the distribution of z-scores when comparing the nestedness from the empirical networks with their randomised counterparts (100 null models for each network with the curveball algorithm). The vertical red dashed lines represent the z critical value for a two tailed test with $\alpha = 0.05$. Z-scores to the left of the first vertical red dashed line indicate that networks are less nested than expected by chance (red), those between the two dashed lines indicate no statistical difference from random expectations (green), and those to the right indicate that networks are more nested than expected by chance (blue). Graphs (c-d) show the fitted regression of residual connectance and nestedness across the latitudinal range of the studies from the database. The solid fitted line indicates a significant association, while the dashed lines indicate a non-significant association. The bioclimatic region of each network is indicated with points of different shapes and colours.

3 | DISCUSSION

EuPPollNet offers the largest set of plant-pollinator studies and networks compiled to date at European level. The database contains 1,411 plant and 2,223 pollinator species with over a million interaction records. While the overall sampling coverage of species and interactions is relatively high across the sampled sites, the taxonomic coverage of plants and the main pollinator groups at the European level is still relatively low (i.e., 5.6% for flowering plants and 34.4% for bee species). This likely reflects that most plant and pollinator species are rare and geographically restricted. For example, given that most of the plant-pollinator networks from the database are sampled on intensive grasslands, and habitat heterogeneity is a crucial factor in understanding pollinator diversity at European level (Hass et al., 2018; Kleijn et al., 2015; Martínez-Núñez et al., 2022), adding studies on other habitat types is likely to result in a rapid increase of the coverage of plant and pollinator species and their interactions. Indeed, plant and pollinator species were rarely shared across multiple sites, indicating that there are few “common” species and many “rare” ones at the metaweb or continental level. This high number of “rare” species results in an upward slope of the species or interaction accumulation curves (Thompson & Withers, 2003). In other words, minimal sampling efforts are capturing a substantial number of species and interactions, but achieving a comprehensive inventory will require numerous sampling events within and across habitats, particularly for plant-pollinator interactions.

Bees are responsible for the majority of the sampled interactions at the metaweb level. As not all surveys included all pollinator groups, this result may partly be influenced by the taxonomic groups sampled across studies. However, the relevance of bees and other pollinator orders for network topology changed across habitats and bioclimatic regions in accordance to the literature. For instance, plant-pollinator communities in the Mediterranean were dominated by bees, while communities in Alpine or Boreal regions were fly species rich or fly-dominated. These patterns are consistent with our current understanding of bee diversity, which peaks in dry or temperate areas (Leclercq et al., 2023; Orr et al., 2021); and with the fact that colder environments (i.e., altitude and latitude wise) harbor a larger fraction of fly pollinators compared to other taxa (Elberling & Olesen, 1999; Lefebvre et al., 2018). In addition, beetles were only commonly documented as floral visitors in the Mediterranean region. This study cannot determine whether pollination ecologists traditionally document flower-beetle interactions only in the Mediterranean, or if there are fewer flower visitations by beetles outside this region. Nevertheless, the high proportion of beetles as floral visitors provides further support for their potential role as pollinators in the Mediterranean (Herrera, 2019; León-Osper & Narbona, 2022). The number of butterfly species and interactions were relatively low compared to the other taxa. While Europe contains fewer butterfly species than other regions of the world (Ollerton, 2017), their relevance as pollinators is likely underestimated within this database. This is because a large fraction of studies (~40%) did not sample butterflies, and conventional sampling methods for monitoring other insect pollinators (e.g., bees or flies) may be inadequate for sampling plant-butterfly interactions (Isaac et al., 2011). Honey bees were present in 87% of networks and conducted on average a third of the total interactions per network.

The proportion of honey bees in networks across Europe is higher than in natural communities (i.e., large unmanaged assemblages of plant species) across the world (~13%; [Hung et al., 2018](#)). This potentially reflects the dominance of intensive grassland habitats in EuPPollNet and their widespread distribution across European landscapes ([Isselstein et al., 2005](#)), the highly generalised nature of honey bees, their native status and above all, the widespread practice of beekeeping in Europe ([Herrera, 2020](#); [Magrath et al., 2017](#); [Steffan-Dewenter & Tschardt, 2000](#)).

Although Europe contains a much larger number of flowering plants than pollinator species (~5 to 1 ratio according to our extrapolation from checklists), the observed number of pollinator species in the database was almost double that of the plants. This could be explained by the fact that all networks are phytocentric, resulting in sampling bias towards pollinator species ([Jordano, 2016](#); [Vizentin-Bugoni et al., 2018](#)). While animal-centered sampling is likely to increase the plant-pollinator species ratio (e.g., [Encinas-Viso et al., 2023](#)), the spatial scale and environmental context of the sampled communities will also influence their observed diversity, especially given the ability to move of pollinators and the sessile nature of plants. In addition, we found that the accumulation curve of pollinators per plant species does not saturate, which indicates low redundancy of pollinators and that many are regionally “rare”. Rare pollinators can be functionally important for plant species at the landscape level ([Simpson et al., 2022](#); [Winfree et al., 2018](#)), highlighting the need to conduct further sampling events to identify these rare species across different regions and to effectively understand and protect plant-pollinator biodiversity.

Consistent with [Olesen & Jordano \(2002\)](#), we found that residual connectance (i.e., the deviation from the expected connectance for a given network size) was lower at higher latitudes, while normalised nestedness increased towards higher latitudes. Networks at lower latitudes in Europe are exposed to higher temperatures, which can result in higher visitation rates ([Arroyo et al., 1985](#); [Classen et al., 2015](#); [Herrera, 2019](#)), and the overall level of pollinator generalization is known to be higher at lower latitudes ([Schleuning et al., 2012](#)). These factors should increase the number of possible connections that can be established between plants and pollinators for a given network size, resulting in more connected at lower latitudes in Europe. However, our results cannot be extrapolated to lower latitudes outside Europe, as tropical systems might behave differently. Finally, a result that can be surprising is that most empirical networks (~85%) showed a non-nested structure. This result contrasts with earlier work suggesting that plant-pollinator networks are typically nested ([Bascompte et al., 2003](#)) and it aligns with recent evidence showing that nestedness is not a common feature in plant-pollinator networks when evaluated against restrictive null models that conserve the observed species degree ([Payrató-Borrás et al., 2019](#); Figure S6). Note that while species degree distributions are sufficient to explain the emergence of nestedness, this does not preclude nestedness from being a useful metric for comparison across networks. For example, plant-pollinator networks are more nested than plant-herbivore networks precisely because these two network types differ in their composition of generalist and specialist species ([Thébault & Fontaine, 2010](#)).

Although this database covers a wide range of habitats across 23 countries, it contains tempo-

ral and geographical biases that can impact our understanding of plant-pollinator communities (Hughes et al., 2021). For instance, none of the studies in this database sampled nocturnal pollinators, which can impact our view of network structure (García et al., 2024), and most studies were conducted during a single flowering season, limiting our ability to evaluate temporal trends of plant-pollinator communities in the face of environmental changes (Alarcón et al., 2008; Chacoff et al., 2018). In addition, most plant-pollinator networks are sampled from central Europe, while Eastern Europe and the Mediterranean region are underrepresented. This is consistent with previous studies which also report lack of plant-pollinator data for those regions (Bennett et al., 2018; Marshall et al., 2024), highlighting that this database shows existing patterns in data availability despite the absence of a systematic search for studies. The lack of data for Eastern Europe, which contains vast landscapes with semi-natural grasslands experiencing rapid land use change (Sutcliffe et al., 2015), and for the Mediterranean region, which is severely impacted by climate change (Duchenne et al., 2020; Jaworski et al., 2022; Pareja-Bonilla et al., 2023), is particularly concerning. These areas are well known for their rich pollinator diversity (Miličić et al., 2018; Reverté et al., 2023), and their under-representation is likely contributing to the low taxonomic coverage of this database at the European level. Although some of the most well studied countries in Europe (e.g., Belgium, The Netherlands) have already experienced land use change and biodiversity loss at the end of the 20th century (Carvalho et al., 2013), plant-pollinator communities in Europe and across the globe still face current and future threats from climate change (Bartomeus et al., 2011; Duchenne et al., 2020), land use change (Batáry et al., 2015; Reidsma et al., 2006), and the introduction of alien species (Vanbergen et al., 2018; Vilà et al., 2009). Therefore, continuous monitoring programs are needed in order to evaluate spatio-temporal changes of species and their interactions across different European habitats and regions. This will allow local and large scale analyses of the status and trends of plant-pollinator communities, effectively informing management and conservation actions.

In conclusion, the EuPPollNet database enables researchers to explore spatial, taxonomic and structural properties of plant-pollinator networks within Europe. In contrast to previous databases, EuPPollNet provides interaction data along with sampling information that could help researchers to better control for sampling effort and completeness and to select the most suitable networks for their research questions. Here, we have shown how connectance and nestedness change across their latitudinal range, and that plant-pollinator networks are as nested as expected given plant and pollinator generalist levels. These analyses aim to highlight the variability present across Europe in the structure of plant-pollinator networks and illustrate the opportunities available to develop and test questions about spatio-temporal network change using EuPPollNet. The reproducible workflow allows researchers to adapt and reuse this database, enabling the continuous addition of new networks to better evaluate the status and trends of plant-pollinator communities. Finally, we hope this database becomes an iterative resource that keeps growing and improving over time to better understand and conserve European biodiversity.

REFERENCES

- Alarcón, R., Waser, N. M., & Ollerton, J. (2008). Year-to-year variation in the topology of a plant–pollinator interaction network. *Oikos*, *117*(12), 1796–1807.
- Almeida-Neto, M., Guimaraes, P., Guimaraes Jr, P. R., Loyola, R. D., & Ulrich, W. (2008). A consistent metric for nestedness analysis in ecological systems: Reconciling concept and measurement. *Oikos*, *117*(8), 1227–1239.
- Archer, C. R., Pirk, C. W. W., Carnevalheiro, L. G., & Nicolson, S. W. (2014). Economic and ecological implications of geographic bias in pollinator ecology in the light of pollinator declines. *Oikos*, *123*(4), 401–407.
- Arroyo, M. T. K., Armesto, J. J., & Primack, R. B. (1985). Community studies in pollination ecology in the high temperate andes of central Chile II. Effect of temperature on visitation rates and pollination possibilities. *Plant Systematics and Evolution*, *149*, 187–203.
- Bartomeus, I., Ascher, J. S., Wagner, D., Danforth, B. N., Colla, S., Kornbluth, S., & Winfree, R. (2011). Climate-associated phenological advances in bee pollinators and bee-pollinated plants. *Proceedings of the National Academy of Sciences*, *108*(51), 20645–20649.
- Bascompte, J., & Jordano, P. (2007). Plant-animal mutualistic networks: The architecture of biodiversity. *Annu. Rev. Ecol. Evol. Syst.*, *38*, 567–593.
- Bascompte, J., Jordano, P., Melián, C. J., & Olesen, J. M. (2003). The nested assembly of plant–animal mutualistic networks. *Proceedings of the National Academy of Sciences*, *100*(16), 9383–9387.
- Batáry, P., Dicks, L. V., Kleijn, D., & Sutherland, W. J. (2015). The role of agri-environment schemes in conservation and environmental management. *Conservation Biology*, *29*(4), 1006–1016.
- Bellard, C., Leclerc, C., Leroy, B., Bakkenes, M., Veloz, S., Thuiller, W., & Courchamp, F. (2014). Vulnerability of biodiversity hotspots to global change. *Global Ecology and Biogeography*, *23*(12), 1376–1386.
- Bennett, J. M., Thompson, A., Goia, I., Feldmann, R., Ștefan, V., Bogdan, A., Rakosy, D., Beloiu, M., Biro, I.-B., Bluemel, S., et al. (2018). A review of European studies on pollination networks and pollen limitation, and a case study designed to fill in a gap. *AoB Plants*, *10*(6), ply068.
- Burkle, L. A., & Alarcón, R. (2011). The future of plant–pollinator diversity: Understanding interaction networks across time, space, and global change. *American Journal of Botany*, *98*(3), 528–538.
- Carnevalheiro, L. G., Kunin, W. E., Keil, P., Aguirre-Gutiérrez, J., Ellis, W. N., Fox, R., Groom, Q., Hennekens, S., Van Landuyt, W., Maes, D., et al. (2013). Species richness declines and biotic homogenisation have slowed down for NW-European pollinators and plants. *Ecology Letters*, *16*(7), 870–878.
- Chacoff, N. P., Resasco, J., & Vázquez, D. P. (2018). *Interaction frequency, network position, and the temporal persistence of interactions in a plant–pollinator network*.
- Chacoff, N. P., Vázquez, D. P., Lomáscolo, S. B., Stevani, E. L., Dorado, J., & Padrón, B. (2012). Evaluating sampling completeness in a desert plant–pollinator network. *Journal of Animal Ecology*, *81*(1), 190–200.

- Chamberlain, S., Oldoni, D., & Waller, J. (2022). *Rgbif: Interface to the global biodiversity information facility API*.
- Classen, A., Peters, M. K., Kindeketa, W. J., Appelhans, T., Eardley, C. D., Gikungu, M. W., Hemp, A., Nauss, T., & Steffan-Dewenter, I. (2015). Temperature versus resource constraints: Which factors determine bee diversity on mount Kilimanjaro, Tanzania? *Global Ecology and Biogeography*, 24(6), 642–652.
- Cribari-Neto, F., & Zeileis, A. (2010). Beta regression in R. *Journal of Statistical Software*, 34, 1–24.
- Culley, T. M., Weller, S. G., & Sakai, A. K. (2002). The evolution of wind pollination in angiosperms. *Trends in Ecology & Evolution*, 17(8), 361–369.
- Dormann, C. F., Gruber, B., & Fründ, J. (2008). Introducing the bipartite package: Analysing ecological networks. *Interaction*, 1(0.2413793), 8–11.
- Duchenne, F., Thébault, E., Michez, D., Elias, M., Drake, M., Persson, M., Rousseau-Piot, J., Pollet, M., Vanormelingen, P., & Fontaine, C. (2020). Phenological shifts alter the seasonal structure of pollinator assemblages in Europe. *Nature Ecology & Evolution*, 4(1), 115–121.
- Eichenberg, D., Bowler, D. E., Bonn, A., Bruehlheide, H., Grescho, V., Harter, D., Jandt, U., May, R., Winter, M., & Jansen, F. (2021). Widespread decline in central European plant diversity across six decades. *Global Change Biology*, 27(5), 1097–1110.
- Elberling, H., & Olesen, J. M. (1999). The structure of a high latitude plant-flower visitor system: The dominance of flies. *Ecography*, 22(3), 314–323.
- Encinas-Viso, F., Bovill, J., Albrecht, D. E., Florez-Fernandez, J., Lessard, B., Lumbers, J., Rodriguez, J., Schmidt-Lebuhn, A., Zwick, A., & Milla, L. (2023). Pollen DNA metabarcoding reveals cryptic diversity and high spatial turnover in alpine plant–pollinator networks. *Molecular Ecology*, 32(23), 6377–6393.
- García, Y., Giménez-Benavides, L., Iriondo, J. M., Lara-Romero, C., Méndez, M., Morente-López, J., & Santamaría, S. (2024). Addition of nocturnal pollinators modifies the structure of pollination networks. *Scientific Reports*, 14(1), 1226.
- Gibson, R. H., Knott, B., Eberlein, T., & Memmott, J. (2011). Sampling method influences the structure of plant–pollinator networks. *Oikos*, 120(6), 822–831.
- Goulson, D., Nicholls, E., Botías, C., & Rotheray, E. L. (2015). Bee declines driven by combined stress from parasites, pesticides, and lack of flowers. *Science*, 347(6229), 1255957.
- Govaerts, R., Nic Lughadha, E., Black, N., Turner, R., & Paton, A. (2021). The world checklist of vascular plants, a continuously updated resource for exploring global plant diversity. *Scientific Data*, 8(1), 215.
- Grenié, M., Berti, E., Carvajal-Quintero, J., Dädlow, G. M. L., Sagouis, A., & Winter, M. (2023). Harmonizing taxon names in biodiversity data: A review of tools, databases and best practices. *Methods in Ecology and Evolution*, 14(1), 12–25.
- Guimaraes, P. R. (2020). The structure of ecological networks across levels of organization. *Annual Review of Ecology, Evolution, and Systematics*, 51, 433–460.
- Gusenbauer, M. (2019). Google scholar to overshadow them all? Comparing the sizes of 12 academic search engines and bibliographic databases. *Scientometrics*, 118(1), 177–214.
- Gusenbauer, M., & Haddaway, N. R. (2020). Which academic search systems are suitable

- for systematic reviews or meta-analyses? Evaluating retrieval qualities of google scholar, PubMed, and 26 other resources. *Research Synthesis Methods*, 11(2), 181–217.
- Hadfield, J. D. (2010). MCMC methods for multi-response generalized linear mixed models: The MCMCglmm r package. *Journal of Statistical Software*, 33, 1–22.
- Hass, A. L., Kormann, U. G., Tschardtke, T., Clough, Y., Baillod, A. B., Sirami, C., Fahrig, L., Martin, J.-L., Baudry, J., Bertrand, C., et al. (2018). Landscape configurational heterogeneity by small-scale agriculture, not crop diversity, maintains pollinators and plant reproduction in western europe. *Proceedings of the Royal Society B: Biological Sciences*, 285(1872), 20172242.
- Heberling, J. M., Miller, J. T., Noesgaard, D., Weingart, S. B., & Schigel, D. (2021). Data integration enables global biodiversity synthesis. *Proceedings of the National Academy of Sciences*, 118(6), e2018093118.
- Hedtke, S. M., Patiny, S., & Danforth, B. N. (2013). The bee tree of life: A supermatrix approach to apoid phylogeny and biogeography. *BMC Evolutionary Biology*, 13, 1–13.
- Herrera, C. M. (2019). Complex long-term dynamics of pollinator abundance in undisturbed mediterranean montane habitats over two decades. *Ecological Monographs*, 89(1), e01338.
- Herrera, C. M. (2020). Gradual replacement of wild bees by honeybees in flowers of the mediterranean basin over the last 50 years. *Proceedings of the Royal Society B*, 287(1921), 20192657.
- Hijmans, R. J., Bivand, R., Forner, K., Ooms, J., Pebesma, E., & Sumner, M. D. (2022). *Package “terra”*.
- Hoeppke, C., & Simmons, B. I. (2021). Maxnodf: An r package for fair and fast comparisons of nestedness between networks. *Methods in Ecology and Evolution*, 12(4), 580–585.
- Hsieh, T., & Chao, A. (2016). iNEXT: An r package for rarefaction and extrapolation of species diversity (h ill numbers). *Methods in Ecology and Evolution*, 7(12), 1451–1456.
- Hughes, A. C., Orr, M. C., Ma, K., Costello, M. J., Waller, J., Provoost, P., Yang, Q., Zhu, C., & Qiao, H. (2021). Sampling biases shape our view of the natural world. *Ecography*, 44(9), 1259–1269.
- Hung, K.-L. J., Kingston, J. M., Albrecht, M., Holway, D. A., & Kohn, J. R. (2018). The worldwide importance of honey bees as pollinators in natural habitats. *Proceedings of the Royal Society B: Biological Sciences*, 285(1870), 20172140.
- Isaac, N. J., Cruickshanks, K. L., Weddle, A. M., Marcus Rowcliffe, J., Brereton, T. M., Dennis, R. L., Shuker, D. M., & Thomas, C. D. (2011). Distance sampling and the challenge of monitoring butterfly populations. *Methods in Ecology and Evolution*, 2(6), 585–594.
- Isselstein, J., Jeangros, B., & Pavlu, V. (2005). Agronomic aspects of biodiversity targeted management of temperate grasslands in europe—a review. *Agronomy Research*, 3(2), 139–151.
- Jaworski, C. C., Geslin, B., Zakardjian, M., Lecareux, C., Caillault, P., Nève, G., Meunier, J.-Y., Dupouyet, S., Sweeney, A. C., Lewis, O. T., et al. (2022). Long-term experimental drought alters floral scent and pollinator visits in a mediterranean plant community despite overall limited impacts on plant phenotype and reproduction. *Journal of Ecology*, 110(11), 2628–2648.

- Jordano, P. (1987). Patterns of mutualistic interactions in pollination and seed dispersal: Connectance, dependence asymmetries, and coevolution. *The American Naturalist*, 129(5), 657–677.
- Jordano, P. (2016). Sampling networks of ecological interactions. *Functional Ecology*, 30(12), 1883–1893.
- Jordano, P., Bascompte, J., & Olesen, J. M. (2003). Invariant properties in coevolutionary networks of plant–animal interactions. *Ecology Letters*, 6(1), 69–81.
- Kaiser, L. (2015). Nestedness and null models in ecology. *Freiburg, Germany: Albert-Ludwigs-University Freiburg*.
- Kleijn, D., Winfree, R., Bartomeus, I., Carvalheiro, L. G., Henry, M., Isaacs, R., Klein, A.-M., Kremen, C., M’gonigle, L. K., Rader, R., et al. (2015). Delivery of crop pollination services is an insufficient argument for wild pollinator conservation. *Nature Communications*, 6(1), 7414.
- Leclercq, N., Marshall, L., Caruso, G., Schiel, K., Weekers, T., Carvalheiro, L. G., Dathe, H. H., Kuhlmann, M., Michez, D., Potts, S. G., et al. (2023). European bee diversity: Taxonomic and phylogenetic patterns. *Journal of Biogeography*, 50(7), 1244–1256.
- Lefebvre, V., Villemant, C., Fontaine, C., & Daugeron, C. (2018). Altitudinal, temporal and trophic partitioning of flower-visitors in alpine communities. *Scientific Reports*, 8(1), 4706.
- León-Osper, M., & Narbona, E. (2022). Unravelling the mystery of red flowers in the mediterranean basin: How to be conspicuous in a place dominated by hymenopteran pollinators. *Functional Ecology*, 36(11), 2774–2790.
- Li, D. (2023). Rtrees: An r package to assemble phylogenetic trees from megatrees. *Ecography*, 2023(7), e06643.
- Magrath, A., González-Varo, J. P., Boiffier, M., Vilà, M., & Bartomeus, I. (2017). Honeybee spillover reshuffles pollinator diets and affects plant reproductive success. *Nature Ecology & Evolution*, 1(9), 1299–1307.
- Marshall, L., Leclercq, N., Carvalheiro, L. G., Dathe, H. H., Jacobi, B., Kuhlmann, M., Potts, S. G., Rasmont, P., Roberts, S. P., & Vereecken, N. J. (2024). Understanding and addressing shortfalls in european wild bee data. *Biological Conservation*, 290, 110455.
- Martínez-Núñez, C., Kleijn, D., Ganuza, C., Heupink, D., Raemakers, I., Vertommen, W., & Fijen, T. P. (2022). Temporal and spatial heterogeneity of semi-natural habitat, but not crop diversity, is correlated with landscape pollinator richness. *Journal of Applied Ecology*, 59(5), 1258–1267.
- Miličić, M., Vujić, A., & Cardoso, P. (2018). Effects of climate change on the distribution of hoverfly species (diptera: Syrphidae) in southeast europe. *Biodiversity and Conservation*, 27, 1173–1187.
- Oksanen, J., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., O’hara, R., Simpson, G. L., Solymos, P., Stevens, M. H. H., Wagner, H., et al. (2013). Package “vegan.” *Community Ecology Package, Version*, 2(9), 1–295.
- Olesen, J. M., Bascompte, J., Dupont, Y. L., Elberling, H., Rasmussen, C., & Jordano, P. (2011). Missing and forbidden links in mutualistic networks. *Proceedings of the Royal Society B: Biological Sciences*, 278(1706), 725–732.

- Olesen, J. M., & Jordano, P. (2002). Geographic patterns in plant–pollinator mutualistic networks. *Ecology*, 83(9), 2416–2424.
- Ollerton, J. (2017). Pollinator diversity: Distribution, ecological function, and conservation. *Annual Review of Ecology, Evolution, and Systematics*, 48, 353–376.
- Orr, M. C., Hughes, A. C., Chesters, D., Pickering, J., Zhu, C.-D., & Ascher, J. S. (2021). Global patterns and drivers of bee distribution. *Current Biology*, 31(3), 451–458.
- Paradis, E., Blomberg, S., Bolker, B., Brown, J., Claude, J., Cuong, H. S., Desper, R., & Didier, G. (2019). Package “ape.” *Analyses of Phylogenetics and Evolution, Version*, 2(4), 47.
- Pareja-Bonilla, D., Arista, M., Morellato, L. P. C., & Ortiz, P. L. (2023). Better soon than never: Climate change induces strong phenological reassembly in the flowering of a mediterranean shrub community. *Annals of Botany*, mcad193.
- Payrató-Borras, C., Hernández, L., & Moreno, Y. (2019). Breaking the spell of nestedness: The entropic origin of nestedness in mutualistic systems. *Physical Review X*, 9(3), 031024.
- Pebesma, E. J. et al. (2018). Simple features for r: Standardized support for spatial vector data. *R J.*, 10(1), 439.
- Poelen, J. H., Simons, J. D., & Mungall, C. J. (2014). Global biotic interactions: An open infrastructure to share and analyze species-interaction datasets. *Ecological Informatics*, 24, 148–159.
- Poisot, T., Baiser, B., Dunne, J. A., Kéfi, S., Massol, F., Mouquet, N., Romanuk, T. N., Stouffer, D. B., Wood, S. A., & Gravel, D. (2016). Mangal–making ecological network analysis simple. *Ecography*, 39(4), 384–390.
- Poisot, T., Bergeron, G., Cazelles, K., Dallas, T., Gravel, D., MacDonald, A., Mercier, B., Violet, C., & Vissault, S. (2021). Global knowledge gaps in species interaction networks data. *Journal of Biogeography*, 48(7), 1552–1563.
- Proença, V., Martin, L. J., Pereira, H. M., Fernandez, M., McRae, L., Belnap, J., Böhm, M., Brummitt, N., García-Moreno, J., Gregory, R. D., et al. (2017). Global biodiversity monitoring: From data sources to essential biodiversity variables. *Biological Conservation*, 213, 256–263.
- Ramos-Jiliberto, R., Domínguez, D., Espinoza, C., Lopez, G., Valdovinos, F. S., Bustamante, R. O., & Medel, R. (2010). Topological change of andean plant–pollinator networks along an altitudinal gradient. *Ecological Complexity*, 7(1), 86–90.
- Rech, A. R., Dalsgaard, B., Sandel, B., Sonne, J., Svenning, J.-C., Holmes, N., & Ollerton, J. (2016). The macroecology of animal versus wind pollination: Ecological factors are more important than historical climate stability. *Plant Ecology & Diversity*, 9(3), 253–262.
- Reidsma, P., Tekelenburg, T., Van den Berg, M., & Alkemade, R. (2006). Impacts of land-use change on biodiversity: An assessment of agricultural biodiversity in the european union. *Agriculture, Ecosystems & Environment*, 114(1), 86–102.
- Revell, L. J. (2012). Phytools: An r package for phylogenetic comparative biology (and other things). *Methods in Ecology and Evolution*, 2, 217–223.
- Reverté, S., Miličić, M., Ačanski, J., Andrić, A., Aracil, A., Aubert, M., Balzan, M. V., Bartomeus, I., Bogusch, P., Bosch, J., et al. (2023). National records of 3000 european bee and hoverfly species: A contribution to pollinator conservation. *Insect Conservation and*

798 *Diversity*, 16(6), 758–775.

799 Sala, O. E., Stuart Chapin, F., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-
800 Sanwald, E., Huenneke, L. F., Jackson, R. B., Kinzig, A., et al. (2000). Global biodiversity
801 scenarios for the year 2100. *Science*, 287(5459), 1770–1774.

802 Saunders, M. E., Kendall, L. K., Lanuza, J. B., Hall, M. A., Rader, R., & Stavert, J. R. (2023).
803 Climate mediates roles of pollinator species in plant–pollinator networks. *Global Ecology*
804 *and Biogeography*, 32(4), 511–518.

805 Schleuning, M., Fründ, J., Klein, A.-M., Abrahamczyk, S., Alarcón, R., Albrecht, M., Ander-
806 sson, G. K., Bazarian, S., Böhning-Gaese, K., Bommarco, R., et al. (2012). Specialization
807 of mutualistic interaction networks decreases toward tropical latitudes. *Current Biology*,
808 22(20), 1925–1931.

809 Schwarz, B., Vázquez, D. P., CaraDonna, P. J., Knight, T. M., Benadi, G., Dormann, C.
810 F., Gauzens, B., Motivans, E., Resasco, J., Blüthgen, N., et al. (2020). Temporal scale-
811 dependence of plant–pollinator networks. *Oikos*, 129(9), 1289–1302.

812 Settele, J., Bishop, J., & Potts, S. G. (2016). Climate change impacts on pollination. *Nature*
813 *Plants*, 2(7), 1–3.

814 Simpson, D. T., Weinman, L. R., Genung, M. A., Roswell, M., MacLeod, M., & Winfree,
815 R. (2022). Many bee species, including rare species, are important for function of entire
816 plant–pollinator networks. *Proceedings of the Royal Society B*, 289(1972), 20212689.

817 Smith, S. A., & Brown, J. W. (2018). Constructing a broadly inclusive seed plant phylogeny.
818 *American Journal of Botany*, 105(3), 302–314.

819 Song, C., Rohr, R. P., & Saavedra, S. (2017). Why are some plant–pollinator networks more
820 nested than others? *Journal of Animal Ecology*, 86(6), 1417–1424.

821 Staniczenko, P., Kopp, J. C., & Allesina, S. (2013). The ghost of nestedness in ecological
822 networks. *Nature Communications*, 4(1), 1–6.

823 Stanners, D., & Bourdeau, P. (1995). *Europe’s environment: The dobríř assessment*.

824 Steffan-Dewenter, I., & Tschardtke, T. (2000). Resource overlap and possible competition
825 between honey bees and wild bees in central europe. *Oecologia*, 122, 288–296.

826 Strona, G., Nappo, D., Boccacci, F., Fattorini, S., & San-Miguel-Ayanz, J. (2014). A fast and
827 unbiased procedure to randomize ecological binary matrices with fixed row and column
828 totals. *Nature Communications*, 5(1), 4114.

829 Sutcliffe, L. M., Batáry, P., Kormann, U., Báldi, A., Dicks, L. V., Herzon, I., Kleijn, D.,
830 Tryjanowski, P., Apostolova, I., Arlettaz, R., et al. (2015). Harnessing the biodiversity
831 value of central and eastern european farmland. *Diversity and Distributions*, 21(6), 722–
832 730.

833 Thébault, E., & Fontaine, C. (2010). Stability of ecological communities and the architecture
834 of mutualistic and trophic networks. *Science*, 329(5993), 853–856.

835 Thompson, G. G., & Withers, P. C. (2003). Effect of species richness and relative abundance
836 on the shape of the species accumulation curve. *Austral Ecology*, 28(4), 355–360.

837 Traveset, A., Tur, C., Trøjelsgaard, K., Heleno, R., Castro-Urgal, R., & Olesen, J. M. (2016).
838 Global patterns of mainland and insular pollination networks. *Global Ecology and Biogeog-*
839 *raphy*, 25(7), 880–890.

840 Troia, M. J., & McManamay, R. A. (2016). Filling in the GAPS: Evaluating completeness and

- 841 coverage of open-access biodiversity databases in the united states. *Ecology and Evolution*,
842 6(14), 4654–4669.
- 843 Trøjelsgaard, K., & Olesen, J. M. (2013). Macroecology of pollination networks. *Global Ecology*
844 *and Biogeography*, 22(2), 149–162.
- 845 Van der Loo, M. P. et al. (2014). The stringdist package for approximate string matching. *R*
846 *J.*, 6(1), 111.
- 847 Vanbergen, A. J., Espíndola, A., & Aizen, M. A. (2018). Risks to pollinators and pollination
848 from invasive alien species. *Nature Ecology & Evolution*, 2(1), 16–25.
- 849 Vilà, M., Bartomeus, I., Dietzsch, A. C., Petanidou, T., Steffan-Dewenter, I., Stout, J. C., &
850 Tscheulin, T. (2009). Invasive plant integration into native plant–pollinator networks across
851 europe. *Proceedings of the Royal Society B: Biological Sciences*, 276(1674), 3887–3893.
- 852 Vizenin-Bugoni, J., Maruyama, P. K., Souza, C. S. de, Ollerton, J., Rech, A. R., & Sazima,
853 M. (2018). Plant-pollinator networks in the tropics: A review. *Ecological Networks in the*
854 *Tropics: An Integrative Overview of Species Interactions from Some of the Most Species-*
855 *Rich Habitats on Earth*, 73–91.
- 856 Wetzell, F. T., Bingham, H. C., Groom, Q., Haase, P., Kõljalg, U., Kuhlmann, M., Martin, C.
857 S., Penev, L., Robertson, T., Saarenmaa, H., et al. (2018). Unlocking biodiversity data:
858 Prioritization and filling the gaps in biodiversity observation data in europe. *Biological*
859 *Conservation*, 221, 78–85.
- 860 Wiemers, M., Balletto, E., Dincă, V., Fric, Z. F., Lamas, G., Lukhtanov, V., Munguira, M. L.,
861 Swaay, C. A. van, Vila, R., Vliegenthart, A., et al. (2018). An updated checklist of the
862 european butterflies (lepidoptera, papilionoidea). *ZooKeys*, 811, 9.
- 863 Windsor, F. M., Hoogen, J. van den, Crowther, T. W., & Evans, D. M. (2023). Using ecological
864 networks to answer questions in global biogeography and ecology. *Journal of Biogeography*,
865 50(1), 57–69.
- 866 Winfree, R., Reilly, J. R., Bartomeus, I., Cariveau, D. P., Williams, N. M., & Gibbs, J. (2018).
867 Species turnover promotes the importance of bee diversity for crop pollination at regional
868 scales. *Science*, 359(6377), 791–793.
- 869 Zattara, E. E., & Aizen, M. A. (2021). Worldwide occurrence records suggest a global decline
870 in bee species richness. *One Earth*, 4(1), 114–123.

871 ACKNOWLEDGEMENTS

872 We thank all the taxonomists and ecologists that has made this database possible by contribut-
873 ing with their fieldwork data.

874 FUNDING INFORMATION

875 This research was funded by the H2020 European project Safeguard (101003476) and by
876 the Federal State of Saxony-Anhalt (MLU-BioDivFund). RT was supported by the Czech

877 Science Foundation (Project No. 21-24186M). NDM, NH, YP and FM were financially sup-
878 ported by the ANR ARSENIC project (grant no. 14-CE02-0012), the ANR NGB project
879 (grant no. 17-CE32- 011), the Region Nord-Pas-de-Calais, the CNRS, the French Ministère de
880 l'Enseignement Supérieur et de la Recherche, the Hauts-de-France Region and the European
881 Regional Funds.

882 **CONFLICT OF INTEREST**

883 None.

884 **DATA AVAILABILITY**

885 All data and code to produce of this database and manuscript are available at Zenodo (LINK)
886 and Github (<https://github.com/JoseBSL/EuPPollNet.git>).

SUPPORTING INFORMATION

Title: EuPPollNet: A European database of plant-pollinator networks

Authors: Jose B. Lanuza ^{1,2,3} | Tiffany M. Knight ^{3,2,4} | Nerea Montes-Perez ¹ | Will Glenny ^{3,4} | Paola Acuña ⁵ | Matthias Albrecht ⁶ | Maddi Artamendi ^{7,8} | Isabelle Badenhäusser ^{9,10,11} | Joanne M. Bennett ¹² | Paolo Biella ¹³ | Ricardo Bommarco ¹⁴ | Andree Cappellari ¹⁵ | Sílvia Castro ¹⁶ | Yann Clough ¹⁷ | Pau Colom ^{18,19} | Joana Costa ^{16,20} | Nathan Cyrille ^{21,22} | Natasha de Manincor ^{23,24} | Paula Dominguez-Lapido ⁷ | Christophe Dominik ^{4,3} | Yoko L. Dupont ²⁵ | Reinart Feldmann ²⁶ | Emeline Felten ²² | Victoria Ferrero ²⁷ | William Fiordaliso ²⁸ | Alessandro Fisogni ²³ | Úna Fitzpatrick ²⁹ | Marta Galloni ³⁰ | Hugo Gaspar ¹⁶ | Elena Gazzea ¹⁵ | Irina Goia ^{31,32} | Carmelo Gómez-Martínez ³³ | Miguel A. González-Estévez ³³ | Juan Pedro González-Varo ³⁴ | Ingo Grass ³⁵ | Jiří Hadrava ³⁶ | Nina Hautekèete ²³ | Veronica Hederström ¹⁷ | Ruben Heleno ¹⁶ | Sandra Hervias-Parejo ³³ | Jonna M. Heuschele ^{3,4} | Bernhard Hoiss ³⁷ | Andrea Holzschuh ³⁷ | Sebastian Hopfenmüller ³⁸ | José M. Iriondo ³⁹ | Birgit Jauker ⁴⁰ | Frank Jauker ⁴¹ | Jana Jersáková ⁴² | Katharina Kallnik ³⁷ | Reet Karise ⁴³ | David Kleijn ⁴⁴ | Stefan Klotz ⁴ | Theresia Krausl ¹⁷ | Elisabeth Kühn ⁴⁵ | Carlos Lara-Romero ³⁹ | Michelle Larkin ⁴⁶ | Emilien Laurent ²² | Amparo Lázaro ³³ | Felipe Librán-Embida ^{47,48} | Yicong Liu ^{4,2} | Sara Lopes ¹⁶ | Francisco López-Núñez ^{16,49} | João Loureiro ¹⁶ | Ainhoa Magrach ^{7,50} | Marika Mänd ⁴³ | Lorenzo Marini ¹⁵ | Rafel Beltran Mas ³³ | François Massol ⁵¹ | Corina Maurer ⁶ | Denis Michez ²⁴ | Francisco P. Molina ¹ | Javier Morente-López ⁵² | Sarah Mullen ⁵³ | Georgios Nakas ⁵⁴ | Lena Neuenkamp ^{55,56} | Arkadiusz Nowak ^{57,58} | Catherine J. O'Connor ^{16,59} | Aoife O'Rourke ⁵³ | Erik Öckinger ¹⁴ | Jens M. Olesen ⁶⁰ | Øystein H. Opedal ⁶¹ | Theodora Petanidou ⁵⁴ | Yves Piquot ²³ | Simon G. Potts ⁶² | Eileen F. Power ⁶³ | Willem Proesmans ^{24,22} | Demetra Rakosy ^{4,3,64} | Sara Reverté ²⁴ | Stuart P. M. Roberts ⁶² | Maj Rundlöf ⁶⁵ | Laura Russo ^{66,53} | Bertrand Schatz ⁶⁷ | Jeroen Scheper ⁴⁴ | Oliver Schweiger ^{4,3} | Pau Enric Serra ³³ | Catarina Siopa ¹⁶ | Henrik G. Smith ^{65,17} | Dara Stanley ⁶⁸ | Valentin Ștefan ^{4,3} | Ingolf Steffan-Dewenter ³⁷ | Jane C. Stout ⁶³ | Louis Sutter ⁶⁹ | Elena Motivans Švara ^{3,4,2} | Sebastian Świerszcz ^{57,70} | Amibeth Thompson ^{2,3,71} | Anna Traveset ³³ | Annette Trefflich ⁷² | Robert Tropek ^{73,74} | Teja Tschardt ⁴⁸ | Adam J. Vanbergen ²² | Montserrat Vilà ^{1,75} | Ante Vujić ⁷⁶ | Cian White ⁵³ | Jennifer B. Wickens ⁶² | Victoria B. Wickens ⁶² | Marie Winsa ¹⁴ | Leana Zoller ^{2,3,77} | Ignasi Bartomeus ¹

Contains:

- Supplementary text 1
- Figure S1
- Figure S2
- Figure S3
- Figure S4
- Figure S5

Supplementary text 1

EuPPollNet habitat definitions adapted from authors and CORINE Land Cover (CLC). To guide these definitions, they are connected to habitat types defined in the European Nature Information System (EUNIS 2021; <https://eunis.eea.europa.eu/habitats-code-browser-revised.jsp>). Specific examples from EUNIS habitats are provided when they can be precisely linked to EuPPollNet habitat types.

- 1) **Ruderal vegetation:** Plants growing on highly disturbed sites such as road sides or mineral extraction sites. This category partially matches the category of ‘vegetated man made habitats’ from EUNIS (code **V**; e.g., category of ‘dry perennial anthropogenic herbaceous vegetation’, code **V38**).
- 2) **Agricultural margins:** Sides of crops that can include any type of vegetation from low growing plants to trees. This category partially matches the category of ‘vegetated man made habitats’ from EUNIS (code **V**; e.g., category of ‘hedgerows’, code **V4**).
- 3) **Green urban areas:** Parks, private gardens or small pastures within an urban setting. Botanical gardens are included in this category. This category partially matches the category of ‘vegetated man made habitats’ from EUNIS (code **V**; e.g., category of ‘cultivated areas of gardens and parks’, code **V2**).
- 4) **Agricultural land:** Includes any type of crop and any type of vegetation growing within them. This category partially matches the category of ‘vegetated man made habitats’ from EUNIS (code **V**; e.g., category of ‘intensive unmixed crops’, code **V11**).
- 5) **Forest/woodland understory:** Any plant community sampled within a wooded group of plants. The forest may be situated in an agricultural setting or in a fully natural scenario. We have included agroforestry areas as well as open to dense forests in this category but excluded forest that contains sclerophyllous vegetation. This category is similar to the **EUNIS habitat type** ‘forest and other wooded land’ (code **T**).
- 6) **Semi-natural grassland:** Acidic, neutral, or calcareous unimproved grasslands located at elevations lower than 1000 m that are managed extensively through mowing or grazing but have not been improved, i.e., they have never received significant fertilizer or herbicide inputs. This category partially matches the category of ‘grasslands and lands dominated by forbs, mosses or lichens’ from EUNIS (code **R**).
- 7) **Intensive grassland:** Communities dominated by non-woody, low-growing plants that are heavily influenced by human disturbance, such as agriculture, mowing, moderate to high grazing, or urban environments. The plant composition is generally dominated by graminoid species. This category partially matches the EUNIS categories of ‘vegetated man made habitats’ (code **V**; e.g., category of ‘artificial grasslands and herb dominated habitats’, code **V4**) and ‘grasslands and lands dominated by forbs, mosses or lichens’ (code **R**).

- 960 8) **Sclerophyllous vegetation:** Any type of system with a dominant shrub community
961 adapted to drought. Typical of the Mediterranean region. Note, that we have included
962 in this category also woodlands (open coniferous forest) where the shrub community was
963 the main focus of the study. This category partially matches the **EUNIS habitat type**
964 ‘forest and other wooded land’ (code **T**).
- 965 9) **Beaches, dunes, sands:** Plant communities growing on sandy soil. This category
966 partially matches the **EUNIS habitat type** ‘coastal habitats’ (code **N**).
- 967 10) **Riparian vegetation:** Plant communities growing on river margins. This category
968 partially matches the **EUNIS habitat type** ‘forest and other wooded land’ (code **T**)
969 but also the category of ‘inland waters’ which is currently under review.
- 970 11) **Montane to alpine grasslands:** Communities dominated by non-woody, low-growing
971 plants that experience little or no human disturbance and are found in high-elevation
972 areas (>1,000 m). This category partially matches the category of ‘grasslands and lands
973 dominated by forbs, mosses or lichens’ from EUNIS (code **R**; e.g., alpine and subalpine
974 grasslands with code **R4**).
- 975 12) **Moors and heathland:** Low-growing woody vegetation (i.e., typically < 5 m) that is
976 characteristic of poorly fertile soils where environmental conditions prevent the natural
977 formation of forests. This category is similar to the **EUNIS habitat type** ‘heathland,
978 scrub and tundra’ (code **S**).

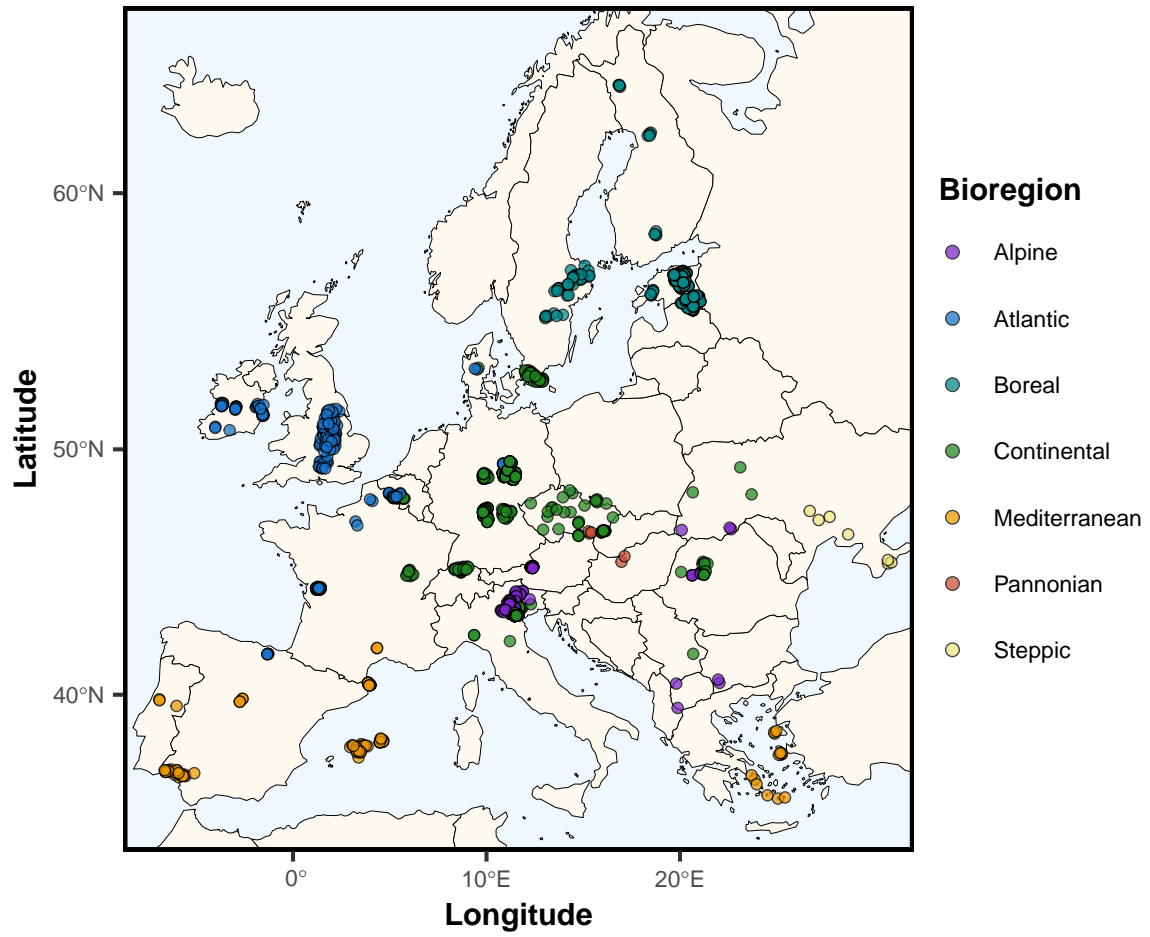


Figure S1. Geographical location of all networks in the EuPPollNet database coloured by bioregion.

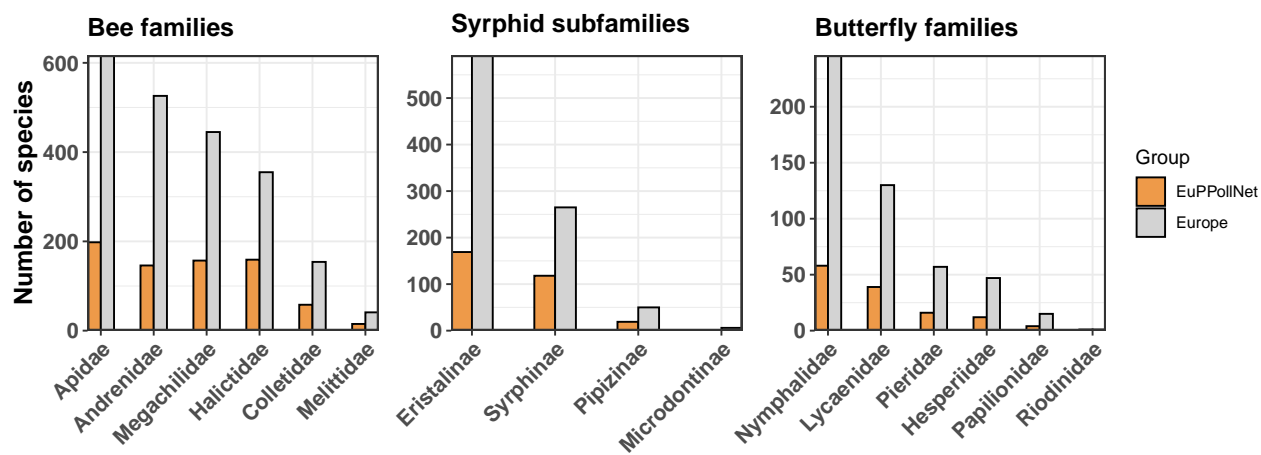


Figure S2. Coverage of the EuPPollNet species for bees (family level), syrphids (subfamily level) and butterflies (family level) in relation to the total number of European species within these taxonomic groups.

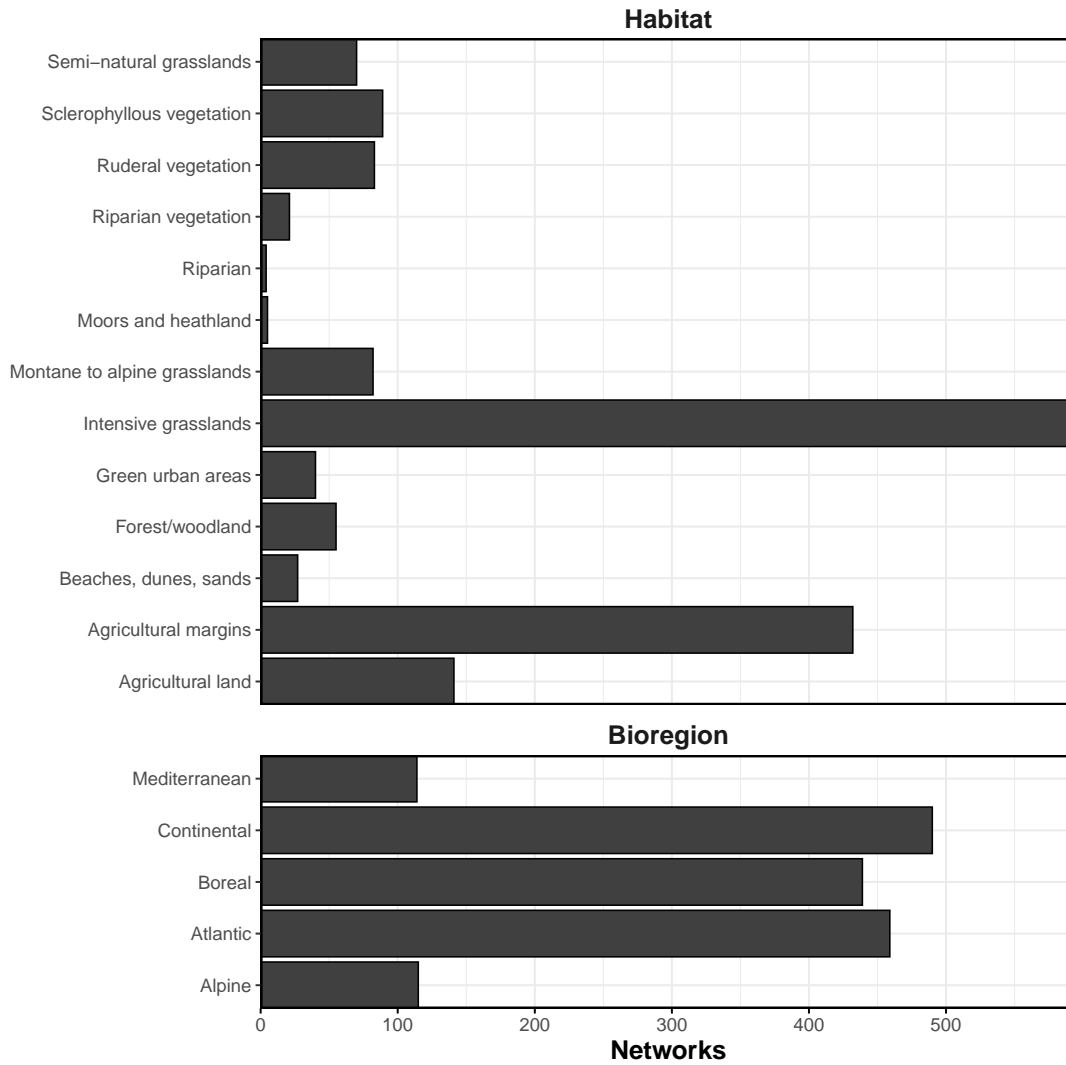


Figure S3. Barplot indicating the number of networks by habitat and bioclimatic region within the database.

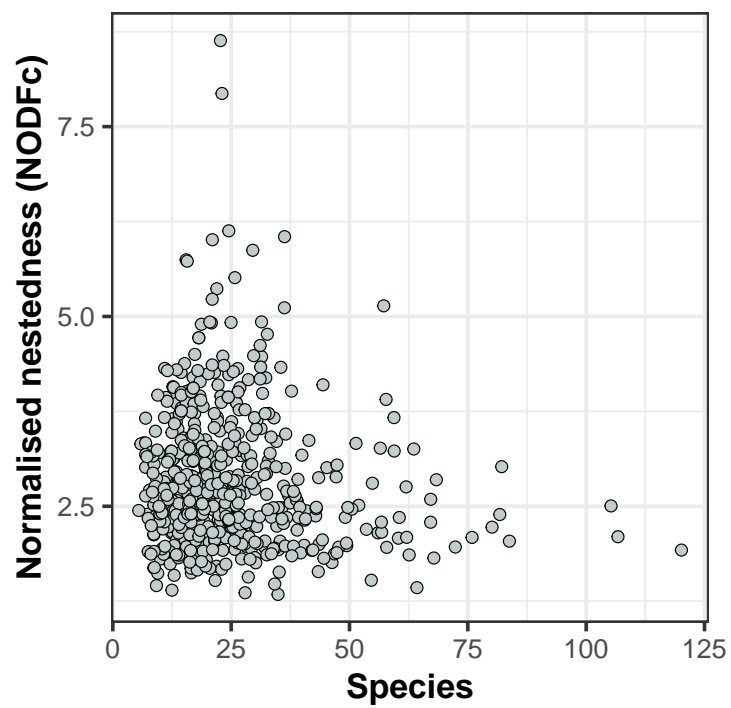


Figure S4. Association between nestedness (NODFc) and the geometric mean of plant and pollinator species per network

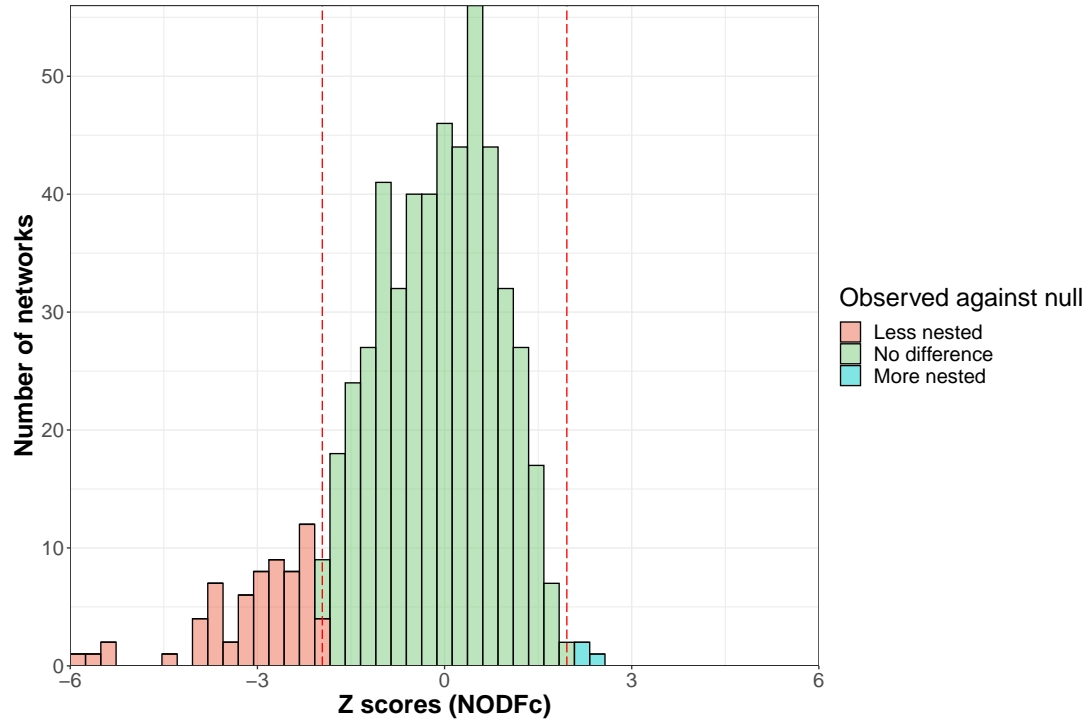


Figure S5. Distribution of z-scores comparing the nestedness of empirical networks with their randomized counterparts with the ‘quasiswap_count’ algorithm (100 null models for each network). The quasiswap_count algorithm maintains row and column sums with fixed connectance.

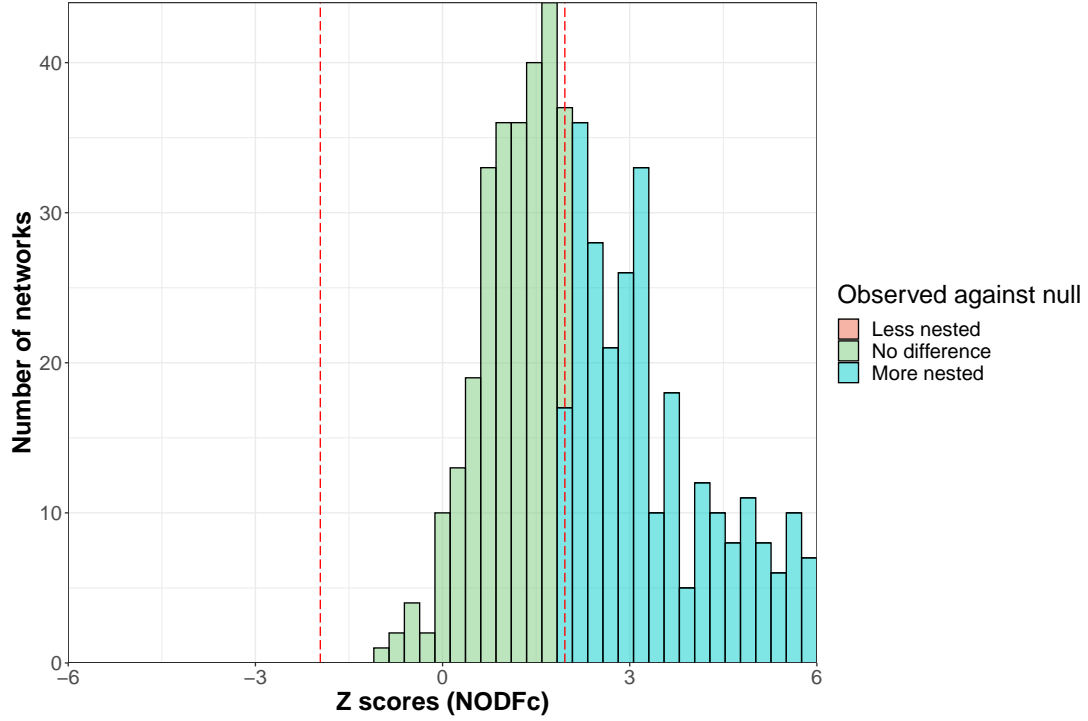


Figure S6. Distribution of z-scores comparing the nestedness of empirical networks with their randomized counterparts with a probabilistic null model from Bascompte et al., 2003 (100 null models for each network). This is the less restrictive approach, as it does not constrain the realised species degree. It first calculates the probability for each cell, and then if a random number between 0 and 1 exceeds this probability, an observed interaction is assigned to the cell. This shows that when the realised degree is not constrained, the number of networks that are statistically more nested is notably higher, highlighting that the nested structure is likely a byproduct of the species degree distribution.