# The Spatial and Temporal Domains of Modern Ecology

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### **Abstract**

To properly understand ecological phenomena, it is necessary to observe them across a range of spatial and temporal scales. Ecologists first raised this point in the 1980s, and since then the ability to collect multi-scale observations has grown rapidly. To assess modern ecology's progress in addressing scale, we analyzed the resolution, extent, interval, and duration of observations in 348 observational studies published between 2004-2014. We found that the scale domains of observations are fairly narrow, and are collected primarily with conventional field techniques. In the spatial domain, most observations have resolutions of  $\leq 1$  m2 and extents of  $\leq 10,000$  ha. In the temporal domain, most observations were either unrepli-

cated or of low frequency ( $\geq 1$  month interval), and were made over relatively short durations ( $\leq 1$  year). Compared to prior meta-analyses from the 1980s and early 2000s, observational durations and resolutions remain largely unchanged, but intervals have become finer and extents larger. Despite such gains, a large gulf exists between the scales at which phenomena are actually observed, and the scales those observations ostensibly represent, revealing portions of space and time not truly captured by replicates and raising concerns about observational comprehensiveness. Adding to these concerns, scales were not clearly reported in most studies, suggesting that it is a minor consideration. Journals can help mitigate this problem by implementing scale reporting standards, which can spur ecologists to more rapidly adopt new observational technologies, and thereby close key gaps in current observational domains.

The scales at which ecosystems are observed plays a critical role in shaping our understanding of their structure and function (I-3). Ecological patterns emerge from temporal and spatial
domains that may be coarser or finer than the processes that shape them, which means that investigation across multiple scales is essential for understanding ecological phenomena (I). This
awareness has grown rapidly since the 1980s, accelerated by the need to understand how changes
in the global climate, ocean, and land systems are affecting everything from individual populations (I) to entire biomes (I), while technological advances in areas such as remote sensing and
genetics are making it ever-easier to quantify ecological features across a broad and increasing
range of scales (I).

Given the growing awareness of scale, expanding data gathering capabilities, and the fact that
the most comprehensive (and arguably best-known) meta-analyses of ecological research scales
were published nearly 30 years ago (I), it is both timely and important to assess the scales of
contemporary ecological investigation. To address this need, we quantified the spatial and temporal

domains (here domain means the distribution of observations within the spectrum of one or more scale dimensions<sup>1</sup>) of empirical observations (defined here as ecological observations collected under un-controlled or non-manipulated conditions) that were reported within recently (2004-2014) published ecological studies. Empirical observations are critical for developing and testing the models that explain why ecological patterns vary in time and space (1, 7), therefore the spatio-temporal domains of observations provide an important indicator of the field's progress towards achieving a holistic, predictive understanding of ecosystems (1, 2).

Our analysis focused on two dimensions of spatial scale, resolution (grain) and extent, and two 21 of temporal scale, interval and duration (Table 1, and see SI for full definitions). Resolution is the area of an individual spatial replicate within which a complete measurement (as opposed to a subsample) of the feature of interest was made. Extent is the area enclosed by the outer-most spatial replicates, or, if the system or habitat being sampled was distinct from its surrounding matrix (e.g. forest patches in grassland habitats), the summed area of sampled patches. Interval refers to the average time elapsed between individual temporal replicates. Duration measures the time elapsed between the first and last temporal replicates, or, in the case of temporally unreplicated observations, the estimated time spent collecting the observation. We also assessed observational scales within two additional dimensions, actual extent (the summed area of spatial replicates) and actual duration (the summed observational time of temporal replicates). We evaluated these additional dimensions to gain insight into how much the actual scales of observation (i.e. how much space and time is covered by the measurement) differ from the scales that the observations are explicitly or implicitly intended to represent. This difference may contain important information about how effectively ecological observations characterize ecological phenomena. First, an increasing gap between actual and intended observational scale inherently implies greater interpolation or extrapolation of observed measurements, raising the odds of over-leveraging data. Second (and re-

<sup>&</sup>lt;sup>1</sup>This definition differs slightly from Wiens' (3), who defined "domain of scale" as "a portion of the scale spectrum within which process-pattern relationships are consistent regardless of scale."

latedly), since natural systems are frequently complex, non-linear, and non-random (9-11), a larger gap may increase the likelihood of encountering unanticipated data challenges such as censored data (sensu(12)), as phenomena may resolve themselves in the space or time between observations.

Table 1: The scale dimensions of ecological observations assessed in this meta-analysis.

Component		Units	Description
Spatial	Resolution	$m^2$	Area of an individual spatial replicate (e.g. plot)
	Extent	ha	Area encompassed by all spatial replicates
	Actual extent	ha	Summed area of all spatial replicates
Temporal	Interval	days	Time elapsed between successive temporal replicates
	Duration	days	Time elapsed between first and last temporal replicates
	Actual duration	days	Summed observational time of all temporal replicates

Our analysis was based on a review of 348 papers randomly selected from 42,918 published between 2004-2014 in the top 30 (based on 2012 impact factor) ecology-themed journals. We extracted scale data from 378 observations of "natural" (i.e. non-experimentally manipulated) ecological features that were reported within 133 of the reviewed papers (plus an additional 62 that these cited as the source of observations). We excluded experiments because they tend to be of limited extent, duration, and resolution due to their higher logistical costs (7, 8), and would therefore likely bias our findings towards finer scales, while minimizing the impact that new observing methods (e.g. satellite imaging, wireless sensing) may have had in expanding the scales of ecological investigation (13–15).

To account for uncertainty in the estimation of observational dimensions due to 1) unclear methodological description in the reviewed papers, and 2) observer interpretation, we conducted a resampling analysis (n=1000) in which scale values were randomly perturbed within the bounds of estimated inter-observer variation (SI). We constructed histograms for each dimension from the mean of the perturbed ensembles, and estimated 95% confidence intervals for each histogram bin (Fig. 1). We constructed kernel density estimates from the full resampled ensemble in order to

<sup>56</sup> assess observational distributions within different juxtapositions of the four primary (resolution,

extent, interval, duration) space-time dimensions (Fig. 2).

#### 8 Results

than 0.8%.

#### 59 Observational methods

To account for potential differences in scales related to methodology, we classified each observation according to the following broad categories, which were field methods (manual *in situ* data collection), automated (*in situ*) sensing, remote sensing/other geographic data (hereafter remote observations), and paleo-reconstruction approaches. Field methods were used for 80% of observations, automated sensing for 12.4%, remote sensing for 6.9%, and paleo-reconstruction for less

#### 66 Distributions within the four primary dimensions

In terms of resolution, the majority (67%) of observations (across all methods) were collected in plots of <1 m<sup>2</sup> resolution, 24% were collected within plots of 1 m<sup>2</sup> up to 1 ha, and the remaining 9% in plots of  $\ge 1$  ha (Fig. 1A). These distributions primarily reflect those of field observations, the dominant observational methodology. Examining the distributions for each observational method (Fig. S1 in SI) shows that automated sensing and paleo-reconstruction had resolutions that were generally finer (85% or more <0.1 m<sup>2</sup>) than field observations (47% <0.1 m<sup>2</sup>), while the majority of remote observations were much coarser (70% >100 m<sup>2</sup>).

The extent of 19% of observations was <10 ha, 23% covered 10-1,000 ha, 12% 1,000-10,000 ha, 19% 10,000-100,000 ha, 12% 100,000-1,000,000 ha, and 15% >1,000,000 ha (Fig. 1B). As with resolution, the extent covered by automated sensing methods tended to be smaller (52% <100 ha) than those of field observations (31% <100 ha), while all but 4% of remote observations covered areas  $\ge 10,000$  ha (as did the small number of paleo-reconstructions).

In the temporal dimensions, 37% of observations were not repeated (Fig. 1C), 17% were repeated at short intervals (sub-second to daily), 20% at daily to monthly intervals, 18% at monthly

to yearly intervals, 6% at yearly to decadal intervals, and 2% at decadal or greater intervals. With respect to temporally replicated observation (Fig. S1 in SI), automated sensing techniques had the finest intervals ( $61\% \le 1$  day;  $100\% \le 1$  year), followed by remote observation ( $37\% \le 1$  day;  $78\% \le 1$  year), field observations ( $17\% \le 1$  day;  $86\% \le 1$  year), and paleo-reconstructions ( $21\% \le 1$  decade).

Duration was one day or less for 31% of sampled observations (due to lack of temporal replication), while 10% covered one day to one month, 23% lasted one month to one year, 27% covered 1-10 years, and 9% spanned a decade or more (including several paleoecological studies covering centuries to millennia; Fig. 1D). Paleo-reconstructions naturally had the longest duration (67% > 1 decade), while just  $\sim$ 40% of field, automated, and remote observations had durations of 1 year or longer.

#### 92 Spatial and temporal domains

Juxtaposing these observational dimensions provides further insight into the spatial and temporal domains of observations (Fig. 2). Contrasting resolution with interval reveals that the majority of temporally replicated observations (unrepeated observations were excluded because they lack interval values) had resolutions of 10 cm<sup>2</sup>-1 m<sup>2</sup> and were revisited at daily to yearly intervals (Fig. 2A). A less dense, oblong concentration of observations bounded on the upper left by monthly to yearly observations at 100 m<sup>2</sup> resolution and on the lower right by near-daily to monthly observations with 1-10 ha resolution is also evident. The four observational methods occupied substantially different portions of the domain space, as indicated by the locations of their median values (and illustrated further in Fig. S2 in the SI): the median domain of field observations was between 0.1-1 m<sup>2</sup> of resolution with a monthly interval, whereas remote observations had coarser median 102 resolutions (1000 m<sup>2</sup>) but finer median intervals ( $\sim$ 1 day). Paleo-reconstructions and automated 103 sensing techniques were both finely resolved (medians from 10 cm<sup>2</sup> to 0.01 m<sup>2</sup>), but automated ap-104 proaches had hourly-daily median intervals compared to multi-decadal for paleo-reconstructions.

Comparing the interval and duration of temporally replicated observations showed most obser-106 vations had daily to decadal intervals and durations of  $\geq 1$  month up to 1 decade (Fig. 2B). The orientation of this concentration shows that interval increases with duration; observations lasting 108 one month to one year tend to have daily to monthly intervals, while those lasting 1 year to 1 109 decade tend to have yearly to decadal intervals. This tendency is reflected in the median domain 110 locations of the primary observational methods: automated sensing had the finest median inter-111 val (hour-day) and shortest duration (month-year), followed by remote sensing (median interval 112 slightly greater than one day and median duration of 1 year), field observations (median monthly 113 interval and duration just over 1 year), and finally paleo-reconstructions (median interval 1 decade and millennial duration). The low densities of observations having sub-daily intervals shows that 115 relatively few high frequency, long duration ecological measurements are undertaken; this position 116 Contrasting the two spatial dimensions against one another (for all observations) shows a pri-117 mary concentration of observations with 10 cm<sup>2</sup> to 100 m<sup>2</sup> resolution with extents ranging from 118 just over 1,000 to nearly 1,000,000 ha (Fig. 2C). The second-most prominent concentration con-119 sists of higher resolution (1 cm<sup>2</sup>-1 m<sup>2</sup>), smaller extent (10-1,000 ha) observations, beneath which 120 lies a third and fainter concentration of 1-1,000 cm<sup>2</sup> resolution, 1000 m<sup>2</sup> to <10 ha. These three 121 concentrations suggest a tendency for observational extent to increase with resolution, which is 122 further evident in the median domain values (and kernel densities; Fig. S2 in SI) of automated (0.01 m<sup>2</sup> resolution, 100 ha extent), field (0.1-1 m<sup>2</sup> resolution, 1,000-10,000 ha extent), and remote (1,000 m<sup>2</sup> resolution, 1-10 million ha extent) observations. Paleo-reconstructions were an outlier from this relationship, having very fine median resolution (0.01 m<sup>2</sup>) but large extent (1 million ha), a result that likely reflects the very low sample size for this observational type. 127

Two primary domains of observational concentration are revealed by juxtaposing duration and extent (across all observations). The first consists of observations spanning one month to one decade in time and 10-1,000 ha in space, while the second is defined by observations of one year

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to several decades that cover 10,000 to 1,000,000 ha (Fig. 2D). Three other notable, but lesser areas of concentration are also evident, including small area observations (0.1-1 ha) covering one month to decade, and short duration, temporally unreplicated observations (<1 day) of either 1-10 ha or 10,000-1,000,000 ha. The median observation (1 year duration, 100 ha extent) from automated sensing lies near the center of the second major concentration, while the median extents of remote (1-10 million ha) and field observations bound the primary concentration at its upper and lower extents, while the median duration of both observational types falls between 1 month to 1 year.

#### 138 Differences between actual and ostensible scales

For our final analysis of observational domains, we evaluated the differences between the scales 139 represented by the extent and duration dimensions and the scales that ecological observations ac-140 tually cover. To make this assessment, we  $log_{10}$  transformed and then subtracted the values of i) 141 actual extent from extent and ii) actual duration from duration, yielding the magnitude of differ-142 ence between each pair of dimensions for each observation. We plotted these values (summarized 143 in box plots) against their corresponding extent/duration values to evaluate whether these differ-144 ences varied with scale (Fig. 3). Extent was on average 5.6 orders of magnitude larger than actual 145 extent, and this difference increased with extent, reaching a maximum of 8 orders of magnitude between 100 million and 1 billion ha of extent (Fig. 3A). The difference fell to 3 orders of magni-147 tude at 10 billion ha, a domain that was covered by the <2% of observations that were primarily collected with remote sensing. Remote observations had the smallest mean difference magnitude (1.9), compared to 5.7 or larger for the other three methods (see Fig. S3 in SI for method-specific plots). 151

The magnitude of difference between the duration and actual duration of observations was somewhat smaller, averaging 3.4 across all observations, ranging from just under 2 for the shortest durations (hour-day) to over 4 for observations lasting 1 decade to 1 century (Fig. 3B). As with extent, the difference fell substantially for the longest durations (century to 10,000 years), as these

domains were covered by paleo-reconstructions (Fig. S3 in SI), which have effectively no difference between duration and actual duration because coring techniques capture continuous temporal records. The mean difference magnitudes for the other three observing methods ranged from just over 3 (field observations and automated sensing) to nearly 6 (remote observations).

#### 160 Potential biases and uncertainties in quantifying scales

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There were several potential methodological aspects that could have influenced our assessment 161 of ecology's spatial and temporal domains. The first stems from our finding that many studies 162 did not precisely report observational scales, which meant that we had to estimate, rather than 163 simply record, these values for most observations (specifically, in 63%, 60%, and 69% of cases 164 for resolution, extent, and actual extent, and 36%, 64%, and 83% of cases for interval, duration, 165 and effective duration, respectively). The inevitable estimation errors may have biased our overall 166 findings. However, we attempted to quantify and account for this error by assessing inter-rater 167 disagreement and incorporating this uncertainty into our resampling methodology. The resulting 168 confidence intervals (Fig. 1) suggest that it was unlikely that estimation errors unduly influenced 169 our findings. 170

Another potential source of bias lies within our scale-estimation protocols, chiefly with respect to our rule for estimating resolution (the smallest areal unit of *complete* measurement). We selected this definition for the sake of consistency, but some papers reported resolution as a larger area in which sub-samples were taken. For these, our estimates are finer than what the studies' authors apparently considered to be plot resolution. Additionally, our domain estimates would presumably be somewhat different if we had included experimentally manipulated observations. For example, average resolution and duration would likely be finer (7, 8).

The relatively small sample size of our study may also bias these findings. Although we used a randomized title draw to obtain a representative sample, we reviewed just 0.8% of papers published during the 10 year period. Our sample may therefore under- or over-represent observational cover-

age in certain scale domains, particularly for analyses broken down by observational method. The case where this is of most concern is for paleo-reconstructions, the small sample size of which resulted in a likely over-estimate of the typical extent of such observations (e.g. Fig. 2B, C; however, the interval and duration findings are probably more representative).

Finally, because our review did not include papers beyond 2014, the omission of studies from 185 the most recent years could have introduced bias into our domain estimates. To evaluate this possi-186 bility, we used linear regressions to examine whether the relative frequency of observing methods 187 changed during the 10 year study period, and whether our findings regarding scale domains might 188 be different if we included more recent studies. Although our sample size was too small to assign 189 statistical significance, we found a possible positive trend in the use of remote observations and 190 a corresponding decline in field observations over the 10 year period. If these trends were not 191 spurious, they suggest that including studies from 2015-2017 would result in a somewhat larger 192 relative sample of remote observations, which would in turn increase slightly the mean extent of 193 ecological observations (see SI for details of trend analyses). 194

#### 195 Discussion

#### 196 Insights into the scale domains of modern ecology

Our results suggest that the scale domains of modern ecological observations are fairly narrow, and are collected primarily with conventional field techniques. Spatially, the majority of observations have grains of  $\leq 1$  m<sup>2</sup> and extents of  $\leq 10,000$  ha (Fig 1A;B). In the temporal domains, most observations are either un-replicated snapshots, or of low frequency ( $\geq 1$  month interval; Fig. 1C) and relatively short duration ( $\leq 1$  year; Fig. 1D).

Contrasting observational dimensions reveals that larger extents are associated with larger plot sizes (Fig. 2C), while longer durations tend to be associated with longer intervals (Fig. 2B).

The latter association primarily reflects a cost-imposed tradeoff between sampling frequency and temporal duration that is characteristic of traditional field-based observation, the dominant obser-

vational mode, but is also reflected by the relative domain locations of the four main observational methods, suggesting that high observational frequency poses a durational cost across observational 207 methods. A similar cost tradeoff is also evident in for the inverse relationship between resolution 208 and interval that dominates that domain space (Fig. 2A), which again primarily relates to charac-209 teristics of field-based observations, in which larger plot sizes demand greater effort that in turn 210 reduces sampling frequency (8). Less obvious is the opposite tradeoff that occurs with remote 211 observation (but shown in Fig. S2 in SI), where finer resolution is desired to increase detail, but 212 typically comes at the cost of longer intervals (?). 213

As a result of these aforementioned tradeoffs, there are several notable gaps in the spatiotemporal domains of ecological observations, which line in the domains defined by high frequency 215 (daily to sub-daily intervals) observations having 1) high to moderate resolutions ( $\geq 1 \text{ m}^2 \text{ up to } 100 \text{ m}^2 \text{ m}$ 216 ha; Fig. 2A) and 2) decadal or longer durations (Fig. 2B). Another hole is evident in the high to moderate resolution, large extent (1 million-10 billion ha) domain (Fig. 2C).

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Have these domains changed since the seminal papers on scale first began to appear in the late 219 1980s (1, 3, 7)? A comprehensive answer to this question would require a similar study focused on 220 earlier literature, but an analysis of results presented in three prior studies provides partial insight. 221 The first and most comprehensive dataset consists of duration values extracted by Tilman (7) from 222 623 studies published between 1977-1987 in the journal *Ecology*. The average duration of the 223 most comparable subset of those values (n=419; see SI) was 3.6 years, compared to 3.3 years for 224 observations in our sample (or 5.1 if temporally un-replicated observations are excluded). The second dataset is found in Kareiva and Anderson (8), who present the resolutions of 97 community 226 ecology experiments published in *Ecology* between 1980-1986. The average of those (12,657 m<sup>2</sup>) 227 was substantially smaller than the mean of our sample (1,496,070 m<sup>2</sup>), but comparing the 80th 228 percentile value (197 m<sup>2</sup>) of Kareiva and Anderson's (8) to ours (115 m<sup>2</sup>) shows that the majority 229 of contemporary observations are finer-grained than most 1980s-era experiments. The third dataset is provided by Porter et al (*15*), who compared the extent and interval of 25 studies published in 2003 and 2004 (also in *Ecology*). The mean interval was 178 days, compared to 684 days in our sample, but the 80th percentile value in our study was 169 days compared to 329 days in theirs.

Extent in our sample was substantially larger according to multiple summary statistics, including the mean (114,965,072 ha in our study versus 368,403 ha), median (5,051 ha versus 9 ha), and 95th percentile (46,424,808 ha versus 136,000 ha).

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Although limited due to methodological differences (e.g. a focus on experiments versus unmanipulated systems), these comparisons suggest that the duration and, less clearly, resolution of ecological observations have changed little in the past 30 years, but observational frequency and particularly extent have increased. A weak positive trend in our own data suggests that average extent of ecological observations is steadily increasing (Fig. S5 in SI), and likely corresponds to the increasing use of remote observation technologies (Fig. S4 in SI).

However, even though observational extent appears to be increasing, there remains for most 243 observations a large gulf between the area that is actually sampled and that which the spatial repli-244 cates purportedly represent (Fig. 3A). A similarly large discrepancy is also evident between the 245 time spans covered by repeated observations and the time that is spent actually observing a phe-246 nomenon (Fig. 3B). These differences between the actual and representative scales of observation 247 have implications for ecological understanding, as the unobserved portions of space and time may 248 contain important patterns and processes that are not captured by replicates, due to phenomenondependent factors such as autocorrelation and representativeness of the sampling scheme (16–20). Brief, infrequent snapshots, or fine-grained, spatially sparse replicates, may be sufficient to char-251 acterize many phenomena (as an example with respect to temporal replication, annual changes in 252 tree cover are well-represented by low frequency satellite imaging (21)), but may be inadequate 253 for more dynamic phenomena. For example, wildfire extent and duration can be mapped by daily 254 return satellites (22, 23), but the instantaneous nature of the imaging means that it cannot be used to observe fire behavior (24). To capture such behavior, long periods of continuous observation
may be more important for understanding dynamics than frequent repeats.

It is therefore important to examine whether the scales of the phenomena being observed are 258 adequately captured by the design of replicates. Our methods suggest one possible procedure 259 for assessing the scale representativeness of replicates: 1) measure auto-correlation (spatial or 260 serial) in the replicates, 2) add the autocorrelation length to the replicate area/duration, 3) calcu-261 late an autocorrelation-adjusted actual extent/duration, and 4) plot where it falls between actual 262 extent/duration and extent/duration. The distance between the adjusted actual value and the osten-263 sible value can provide a measure of how well the replicates represent the intended scale of obser-264 vation. Though increasing spatial or temporal coverage may not always be the goal of a study (e.g., 265 when spatial or temporal autocorrelation is a measure of interest), if the gap between actual and 266 ostensible values remains large, then alternative sampling methods may be used to close it. For ex-267 ample, remote sensing provides wall-to-wall spatial coverage of a study area, erasing the difference 268 between actual extent and extent. Furthermore, the interval of high-resolution imaging (higher res-269 olution is preferred in images as it allows individual features to be better discerned (25, 26)) is 270 now approaching daily to sub-daily scales (27, 28), allowing improved representation of spatial 271 and temporal dynamics. For phenomena that can't be measured from space, either because they 272 are not visible or because they require continuous observation, new approaches for collecting in 273 situ or near-surface observations (e.g. low-cost wireless sensors (15, 29, 30), citizen observers (31), 274 and autonomous vehicles (32)) can be used to increase the spatial and temporal coverage of observations.

The aforementioned insights regarding modern observational domains must be tempered by the uncertainty within our own scale estimates, as detailed in the preceding section. However, most of this uncertainty is attributable to unclear reporting of scale values in the majority of papers we reviewed (a problem also noted in geography studies (33)). This tendency towards vague

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documentation offers one final insight, which is that, despite decades of accumulated knowledge regarding the importance of scale in ecology (1-3, 34), scale appears to remain a low priority throughout much of the discipline. Beyond contributing to the broader problem of scientific reproducibility (35), inattentiveness to scale increases the risk that observations inadequately represent the phenomenon of interest, thereby limiting the generalizability of any derived ecological knowledge (3, 33, 34). To mitigate this problem, we recommend that ecological journals require authors to quantify and clearly report the values of resolution, extent, interval, and duration.

#### 288 Looking forward

Our study suggests that the concept of scale has yet to fully permeate the discipline of ecology. Ev-289 idence for this assertion lies in the continued narrowness of ecology's observational scale domains 290 and the poor documentation of scale dimensions in the literature. However, the increasing extent of 291 ecological observations, enabled by remote sensing and presumably motivated by many ecologists' 292 appreciation of scale-related issues, suggests that ecology's scale domains are gradually changing. 293 In the coming years, the accelerating gains in technology and analytical methods will allow re-294 searchers new and unprecedented capabilities to peer into, and thus close, the prominent holes in 295 observational scale domains. A renewed, discipline-wide focus on scale's importance, including 296 the adoption of stricter scale-reporting standards by journals, will help spur ecologists to address 297 these gaps, while fostering the improved transferability of knowledge within the discipline. 298

## References

- 1. Levin, S. A. The problem of pattern and scale in ecology. *Ecology* 73, 1943–1967 (1992).
- 2. Chave, J. The problem of pattern and scale in ecology: what have we learned in 20years?

  Ecology Letters 16, 4–16 (2013).
- 30. Wiens, J. A. Spatial scaling in ecology. Functional Ecology 3, 385–397 (1989). PT: J.

- 4. Tingley, M. W., Koo, M. S., Moritz, C., Rush, A. C. & Beissinger, S. R. The push and pull of climate change causes heterogeneous shifts in avian elevational ranges. *Global Change Biology* **18**, 3279–3290 (2012).
- 5. Xiao, J. F. & Moody, A. Photosynthetic activity of US biomes: responses to the spatial variability and seasonality of precipitation and temperature. *Global Change Biology* **10**, 437–451 (2004).
- 6. Schneider, D. C. The rise of the concept of scale in ecology. *BioScience* **51**, 545–553 (2001).
- 7. Tilman, D. Ecological experimentation: strengths and conceptual problems. In *Long-term* studies in ecology, 136–157 (Springer, 1989).
- 8. Kareiva, P. & Andersen, M. Spatial aspects of species interactions: The wedding of models and experiments. In *Community ecology*, 35–50 (Springer, 1988).
- 9. Levin, S. A. Ecosystems and the Biosphere as Complex Adaptive Systems. *Ecosystems* 1, 431–436 (1998).
- 10. Pringle, R. M. & Tarnita, C. E. Spatial Self-Organization of Ecosystems: Integrating Multiple Mechanisms of Regular-Pattern Formation. *Annual Review of Entomology* **62**, 359–377 (2017).
- 11. Rietkerk, M. & van de Koppel, J. Regular pattern formation in real ecosystems. *Trends in Ecology & Evolution* **23**, 169–175 (2008).
- 12. Efron, B. The Efficiency of Cox's Likelihood Function for Censored Data. *Journal of the*American Statistical Association **72**, 557–565 (1977).
- 13. Turner, W. *et al.* Remote sensing for biodiversity science and conservation. *Trends in Ecology*& Evolution **18**, 306–314 (2003).

- 14. Pettorelli, N. *et al.* Satellite remote sensing for applied ecologists: opportunities and challenges. *Journal of Applied Ecology* **51**, 839–848 (2014).
- 15. Porter, J. et al. Wireless Sensor Networks for Ecology. BioScience 55, 561–572 (2005).
- 16. Underwood, A. J. Experiments in Ecology: Their Logical Design and Interpretation Using

  Analysis of Variance (Cambridge University Press, 1997).
- 17. Palmer, M. W. & White, P. S. Scale dependence and the species-area relationship. *American*Naturalist 717–740 (1994).
- 18. Cao, Y., Williams, D. D. & Larsen, D. P. Comparison of ecological communities: The problem of sample representativeness. *Ecological Monographs* **72**, 41–56 (2002).
- 19. Legendre, P. Spatial autocorrelation trouble or new paradigm? *Ecology* **74**, 1659–1673 (1993).
- 20. Collins, S. L., Micheli, F. & Hartt, L. A method to determine rates and patterns of variability in ecological communities. *Oikos* **91**, 285–293 (2000).
- 21. Hansen, M. C. *et al.* High-resolution global maps of 21st-century forest cover change. *Science* 342, 850–853 (2013).
- 22. Roy, D., Jin, Y., Lewis, P. & Justice, C. Prototyping a global algorithm for systematic fireaffected area mapping using MODIS time series data. *Remote Sensing of Environment* **97**, 137–162 (2005).
- <sup>344</sup> 23. Jones, B. M. *et al.* Fire behavior, weather, and burn severity of the 2007 Anaktuvuk River tundra fire, North Slope, Alaska. *Arctic, Antarctic, and Alpine Research* **41**, 309–316 (2009).

- 24. Clements, C. B. *et al.* Observing the dynamics of wildland grass fires: FireFlux-a field validation experiment. *Bulletin of the American Meteorological Society* **39**, 1369–1382 (2007).
- Dark, S. J. & Bram, D. The modifiable areal unit problem (MAUP) in physical geography.
   Progress in Physical Geography 31, 471–479 (2007).
- 26. Hay, G. J., Blaschke, T., Marceau, D. J. & Bouchard, A. A comparison of three image-object methods for the multiscale analysis of landscape structure. *ISPRS Journal of Photogrammetry*352 and Remote Sensing 57, 327–345 (2003).
- 27. Drusch, M. *et al.* Sentinel-2: ESA's Optical High-Resolution Mission for GMES Operational
   Services. *Remote Sensing of Environment* 120, 25–36 (2012).
- 28. Hand, E. Startup liftoff. *Science* **348**, 172–177 (2015).
- Wolf, A., Falusi, J., Caylor, K., Sheffield, J. & Wood, E. A GSM-based surface meteorology
   network in service of improved African hydrological data assimilation and drought forecasting
   (San Francisco, 2012).
- 359 30. Collins, S. L. *et al.* New opportunities in ecological sensing using wireless sensor networks. *Frontiers in Ecology and the Environment* **4**, 402–407 (2006).
- 31. Dickinson, J. L. *et al.* The current state of citizen science as a tool for ecological research and public engagement. *Frontiers in Ecology and the Environment* **10**, 291–297 (2012).
- 363 32. Anderson, K. & Gaston, K. J. Lightweight unmanned aerial vehicles will revolutionize spatial ecology. *Frontiers in Ecology and the Environment* **11**, 138–146 (2013).
- 33. Margulies, J. D., Magliocca, N. R., Schmill, M. D. & Ellis, E. C. Ambiguous Geographies:

  Connecting Case Study Knowledge with Global Change Science. *Annals of the American*Association of Geographers **106**, 572–596 (2016).

- 368 34. Wheatley, M. & Johnson, C. Factors limiting our understanding of ecological scale. *Ecological Complexity* **6**, 150–159 (2009).
- 35. Baker, M. 1,500 scientists lift the lid on reproducibility. *Nature News* **533**, 452 (2016).
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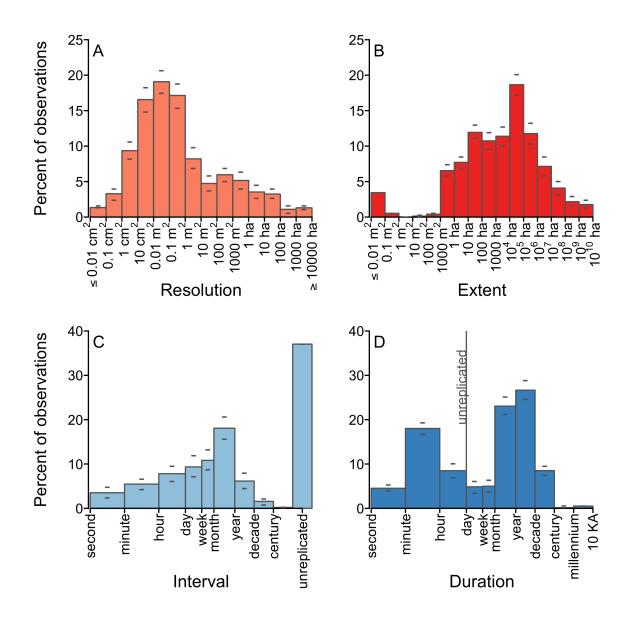


Figure 1: Histograms of the resolution (A), extent (B), interval (C), and duration (D) of observations collected from the surveyed ecological studies. Bars represent the average percentages for each bin realized after 1000 perturbed resamples, while grey bars indicate the 95% confidence interval. Bar widths in C-D indicate differences in scale between x-axis labels. The grey vertical line in D indicates that the majority (>95%) of observations of  $\leq 1$  day duration were temporally unreplicated.

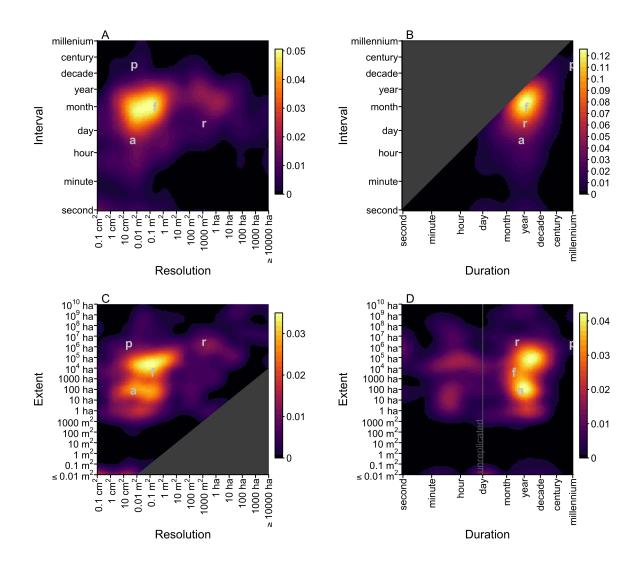


Figure 2: Kernel density estimates of observational densities within the domains defined by A) interval and resolution (of temporally replicated observations only), B) duration and extent, C) resolution and extent, and D) interval and duration (of temporally replicated observations). Density estimates were applied to the log-transformed values of each observational dimension, and density estimates are rescaled to represent percentages. Letters in the plots denote the median values of different observational methods (f=field observations; a = automated sensing; r = remote sensing; p = paleo-observations). The grey shaded areas represent physically impossible domains (intervals greater than duration and resolutions greater than extent).

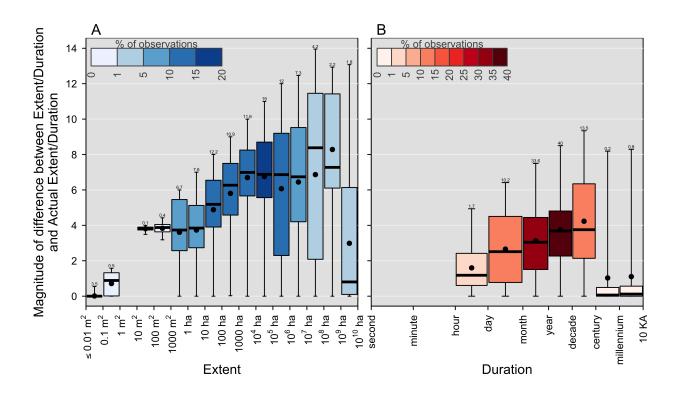


Figure 3: The difference between extent and *actual* extent (the summed area of spatial replicates) (A) and duration and *actual* duration (the summed sampling duration across temporal replicates) (B). Difference values are expressed in terms of how many orders of magnitude larger (or longer) extent (duration) is than actual extent (actual duration), and are summarized (as box plots, with circle in box representing the mean and line the median) in bins representing increasing scales of actual extent/duration. The percentages of observations falling within each bin are indicated by the color of the inter-quartile and the numeric value above the upper whisker. The grey vertical line in D indicates that the majority (>95%) of observations of  $\leq 1$  day duration were temporally unreplicated.