

# The Spatial and Temporal Domains of Modern Ecology

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## Abstract

**To properly understand ecological phenomena, it is necessary to observe them across a range of spatial and temporal scales. Ecologists first raised this point in the 1980s, and since then the ability to collect multi-scale observations has grown rapidly. To assess modern ecology's progress in addressing scale, we analyzed the resolution, extent, interval, and duration of observations in 348 studies published between 2004-2014. We found that the scale domains of observations are fairly narrow, and are collected primarily with conventional field techniques. In the spatial domain, most observations have resolutions of  $\leq 1$  m<sup>2</sup> and extents of  $\leq 10,000$  ha. In the temporal domain, most observations were either unreplicated or of low**

frequency ( $\geq 1$  month interval), and were made over relatively short durations ( $\leq 1$  year). Compared to prior meta-analyses from the 1980s and early 2000s, observational durations and resolutions remain largely unchanged, but intervals have become finer and extents larger. Despite such gains, a large gulf exists between the scales at which phenomena are actually observed, and the scales those observations ostensibly represent, raising concerns about observational representativeness. Adding to these concerns, scales were not clearly reported in most studies, suggesting that it is a minor consideration. Journals can help mitigate this problem by implementing scale reporting standards, which can spur ecologists to more rapidly adopt new observational technologies, and thereby close key gaps in current observational domains.

1 The scales at which ecosystems are observed plays a critical role in shaping our understand-  
2 ing of their structure and function (1–3). Ecological patterns emerge from temporal and spatial  
3 domains that may be coarser or finer than the processes that shape them, which means that in-  
4 vestigation across multiple scales is essential for understanding ecological phenomena (1). This  
5 awareness has grown rapidly since the 1980s, accelerated by the need to understand how changes  
6 in the global climate, ocean, and land systems are affecting everything from individual popula-  
7 tions (4) to entire biomes (5), while technological advances in areas such as remote sensing and  
8 genetics are making it ever-easier to quantify ecological features across a broad and increasing  
9 range of scales (2, 6).

10 Given the growing awareness of scale, expanding data gathering capabilities, and the fact that  
11 the most comprehensive (and arguably best-known) meta-analyses of ecological research scales  
12 were published nearly 30 years ago (7, 8), it is both timely and important to assess the scales of  
13 contemporary ecological investigation. To address this need, we quantified the spatial and temporal  
14 domains of empirical observations that were reported within recently (2004–2014) published eco-

logical studies (here domain means the distribution of observations within the spectrum of one or more scale dimensions<sup>1</sup>). Empirical observations are critical for developing and testing the models that explain why ecological patterns vary in time and space (1, 7), therefore the spatio-temporal domains of observations provide an important indicator of the field's progress towards achieving a holistic, predictive understanding of ecosystems (1, 2).

Our analysis focused on two dimensions of spatial scale, resolution (grain) and extent, and two of temporal scale, interval and duration (Table 1, and see SI for full definitions). Resolution is the area of an individual spatial replicate within which a complete measurement (as opposed to a sub-sample) of the feature of interest was made. Extent is the area enclosed by the outer-most spatial replicates, or, if the system or habitat being sampled was distinct from its surrounding matrix (e.g. forest patches in grassland habitats), the summed area of sampled patches. Interval refers to the average time elapsed between individual temporal replicates. Duration measures the time elapsed between the first and last temporal replicates, or, in the case of temporally unreplicated observations, the estimated time spent collecting the observation. We also assessed observational scales within two additional dimensions, *actual* extent (the integrated area of spatial replicates) and *actual* duration (the summed observational time of temporal replicates). We evaluated these additional dimensions to gain insight into how much the actual scales of observation (i.e. how much space and time is covered by the measurement) differ from those they are intended to represent.

Our analysis was based on a review of 348 papers randomly selected from 42,918 published between 2004-2014 in the top 30 (based on 2012 impact factor) ecology-themed journals. We extracted scale data from 378 observations of “natural” (i.e. non-experimentally manipulated) ecological features that were reported within 133 of the reviewed papers (plus an additional 62 that these cited as the source of observations). We excluded experiments because they tend to be of limited extent, duration, and resolution due to their higher logistical costs (7, 8), and would

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<sup>1</sup>This definition differs slightly from Wiens' (3), who defined “domain of scale” as “a portion of the scale spectrum within which process-pattern relationships are consistent regardless of scale.”

Table 1: The scale dimensions of ecological observations assessed in this meta-analysis.

	Component	Units	Description
Spatial	Resolution	m <sup>2</sup>	Area of an individual spatial replicate (plot)
	Extent	ha	Area encompassed by all spatial replicates
	Actual extent	ha	Summed area of all spatial replicates
Temporal	Interval	days	Time elapsed between successive temporal replicates
	Duration	days	Time elapsed between first and last temporal replicates
	Actual duration	days	Summed observational time of all temporal replicates

therefore likely bias our findings towards finer scales, while minimizing the impact that new observing methods (e.g. satellite imaging, wireless sensing) may have had in expanding the scales of ecological investigation (9–11).

To account for uncertainty in the estimation of observational dimensions due to 1) unclear methodological description in the reviewed papers, and 2) observer interpretation, we conducted a resampling analysis (n=1000) in which scale values were randomly perturbed within the bounds of estimated inter-observer variation (SI). We constructed histograms for each dimension from the mean of the perturbed ensembles, and estimated 95% confidence intervals for each histogram bin (Fig. 1). We constructed kernel density estimates from the full resampled ensemble in order to assess observational distributions within different juxtapositions of the four primary (resolution, extent, interval, duration) space-time dimensions (Fig. 2).

## Results

### *Domains*

In terms of resolution, the majority (67%) of observations were collected in plots of <1 m<sup>2</sup> resolution, 24% were collected within plots of 1 m<sup>2</sup> up to 1 ha, and the remaining 9% in plots of ≥1 ha (Fig. 1A). The extent of 19% of observations was <10 ha, 23% covered 10–1,000 ha, 12% 1,000–10,000 ha, 19% 10,000–100,000 ha, 12% 100,000–1,000,000 ha, and 15% >1,000,000 ha (Fig. 1B).

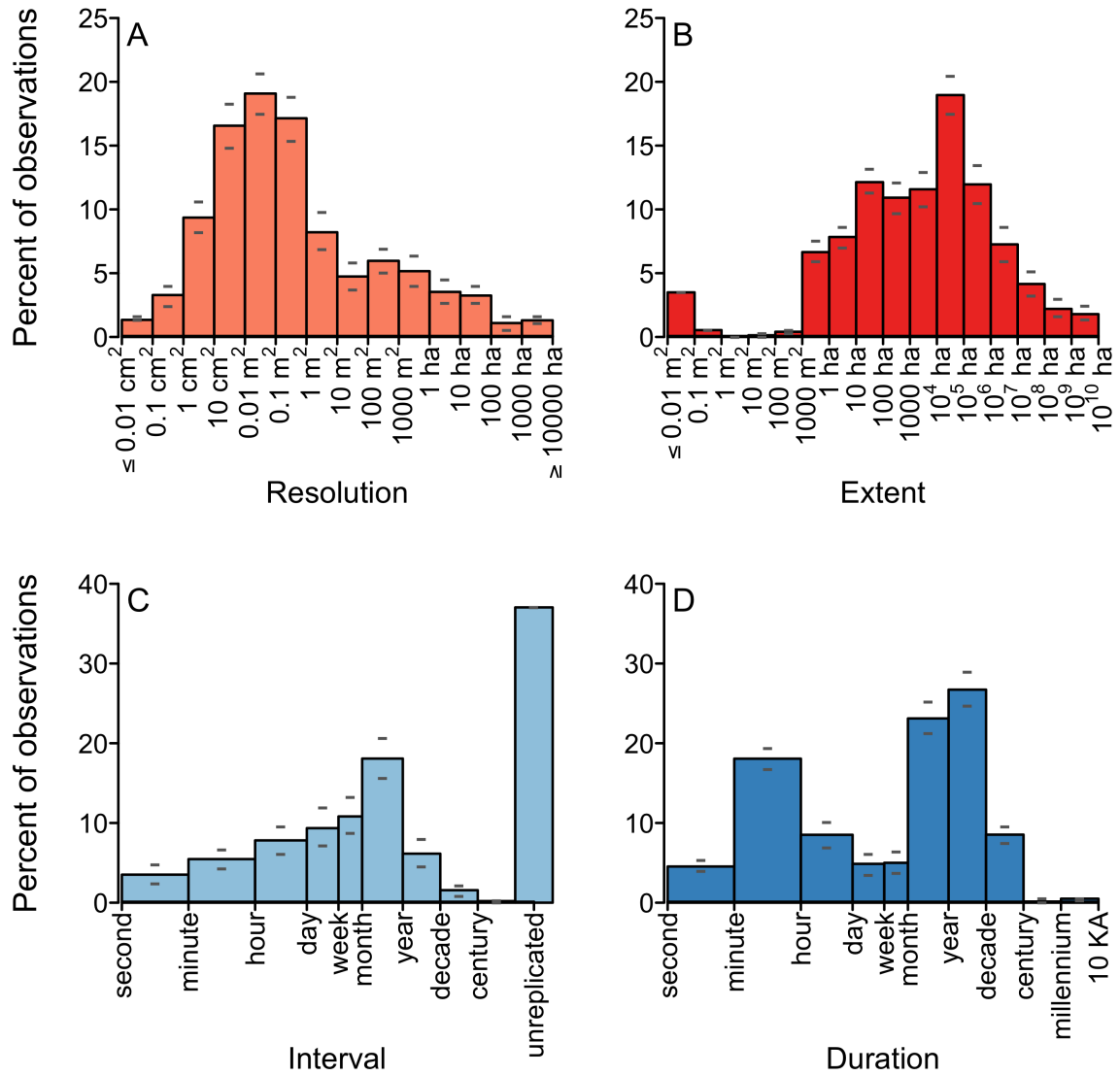


Figure 1: Histograms of the resolution (A), extent (B), interval (C), and duration (D) of observations collected from the surveyed ecological studies. Bars represent the average percentages for each bin realized after 1000 perturbed resamples, while grey bars indicate the 95% confidence interval.

In the temporal dimensions, 37% of observations were not repeated (Fig. 1C), 17% were repeated at short intervals (sub-second to daily), 20% at daily to monthly intervals, 18% at monthly to yearly intervals, 6% at yearly to decadal intervals, and 2% at decadal or greater intervals. Duration was one day or less for 31% of sampled observations, while 10% covered one day to one month, 23% lasted one month to one year, 27% covered 1-10 years, and 9% spanned a decade or more (including several paleoecological studies covering centuries to millennia; Fig. 1D).

Juxtaposing these observational dimensions provides further insight into the spatio-temporal domains of observations (Fig. 2). Contrasting resolution with interval reveals that the majority of temporally replicated observations (unrepeated observations were excluded because they lack interval values) had resolutions of 10 cm<sup>2</sup>-1 m<sup>2</sup> and were revisited at daily to yearly intervals (Fig. 2A). A less dense, oblong concentration of observations bounded on the lower right by monthly to yearly observations at 100 m<sup>2</sup> and on the upper left by near-daily to monthly observations with 1-10 ha resolution is also evident. This lower right to upper left orientation reflects the tradeoff between resolution and interval that is typical of satellite imaging (12), and stands in contrast to the upper right to lower left line that stretches between this concentration and the high frequency (minute-hour intervals), high spatial resolution (0.1-100 cm<sup>2</sup>) observations. This line demonstrates the opposite tradeoff that occurs with field-based observations, where larger plot sizes demand greater effort that in turn reduces sampling frequency (8).

Contrasting duration and extent (for all observations) reveals two primary domains of observational concentration. The first consists of observations spanning one month to one decade in time and 10-1000 ha in space, while the second is defined by observations of one year to several decades that cover 10,000 to 1,000,000 ha (Fig. 2B). Three other notable, but lesser areas of concentration are also evident, including small area observations (0.1-1 ha) covering one month to decade, and short duration, temporally unreplicated observations (<1 day) of either 1-10 ha or 10,000-1,000,000 ha.

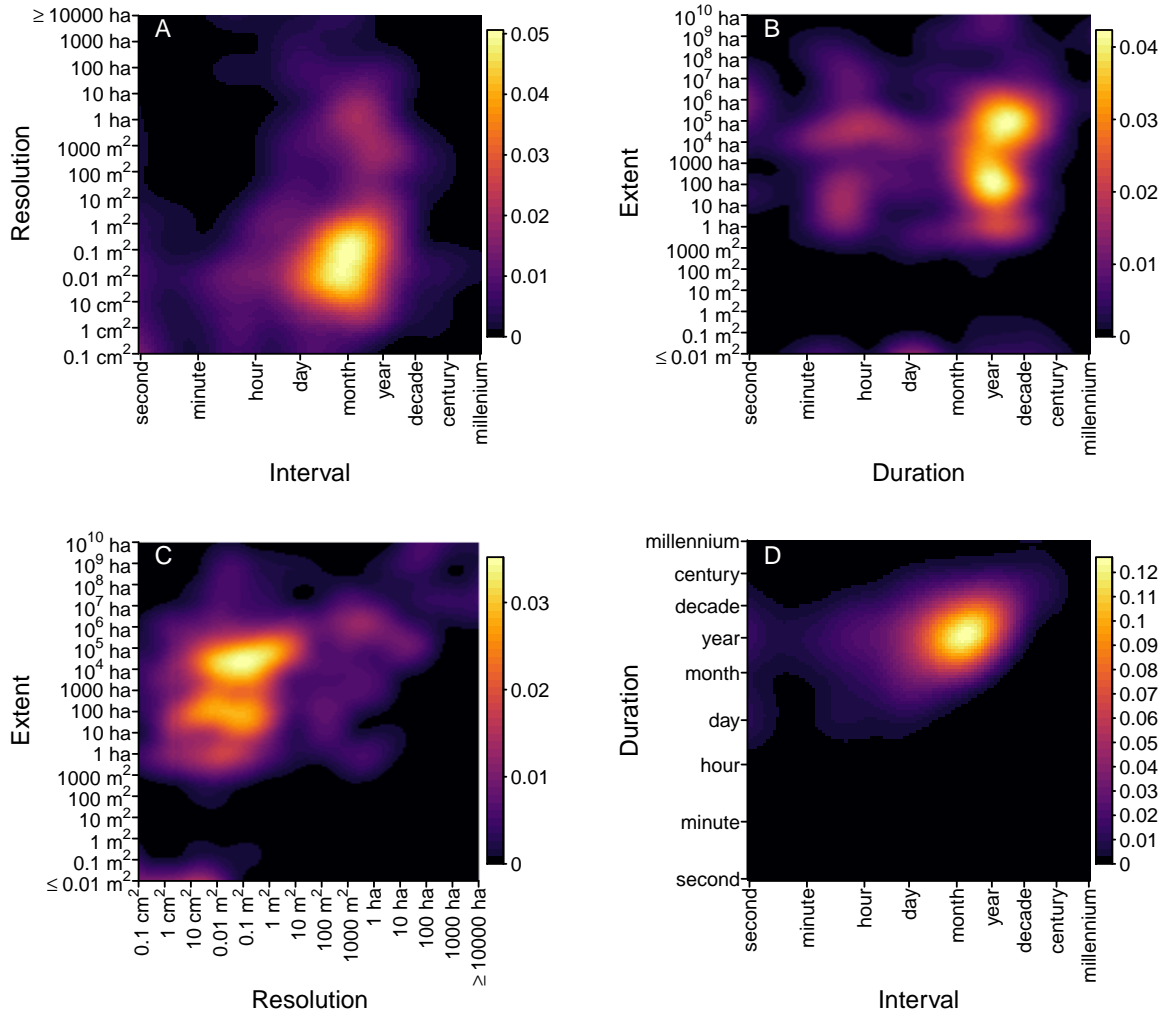


Figure 2: Kernel density estimates of observational densities within the domains defined by A) interval and resolution (of temporally replicated observations only), B) duration and extent, C) resolution and extent, and D) interval and duration (of temporally replicated observations). Density estimates were applied to the log-transformed values of each observational dimension, and density estimates are rescaled to represent percentages.

82 Comparing the two spatial dimensions against one another (for all observations) shows a pri-  
 83 mary concentration of observations with  $10 \text{ cm}^2$  to  $100 \text{ m}^2$  resolution that have extents ranging  
 84 between slightly over 1,000 to nearly 1,000,000 ha (Fig. 2C). The second-most prominent con-

centration consists of higher resolution ( $1 \text{ cm}^2$ - $1 \text{ m}^2$ ), smaller extent (10-1,000 ha) observations, beneath which lies a third and fainter concentration of  $1$ - $1,000 \text{ cm}^2$  resolution,  $1000 \text{ m}^2$  to  $<10 \text{ ha}$ . These three concentrations suggest a tendency for observational extent to increase with resolution, which is a relationship that becomes more pronounced in the less densely observed portion of the domain where resolutions  $\geq 100 \text{ m}^2$ .

Most temporally replicated observations have daily to decadal intervals and durations of  $\geq 1$  month to 1 decade. The orientation of this concentration shows that interval increases with duration; observations lasting one month to one year tend to have daily to monthly intervals, while those lasting one year to one decade tend to have yearly to decadal intervals. The low densities of observations having sub-daily intervals shows that relatively few high frequency, long duration ecological measurements are undertaken.

For our final analysis of observational domains, we evaluated the differences between the scales represented by the extent and duration dimensions and the scales that ecological observations actually cover. To make this assessment, we  $\log_{10}$  transformed and then subtracted the values of i) actual extent from extent and ii) actual duration from duration, yielding the magnitude of difference between each pair of dimensions for each observation. We plotted these values (summarized in box plots) against their corresponding actual extent/duration values to evaluate whether these differences vary with scale (Fig. 3). This analysis showed that most (81%) of the assessed observations had actual extents of  $\leq 1 \text{ ha}$ , which on average were 4 to nearly 8 orders of magnitude smaller than the primary measure of extent (Fig 3A). These differences declined linearly with actual extent, becoming approximately equal for the 5% of observations having  $\geq 100,000 \text{ ha}$  actual extent. The actual duration of 64% of observations was  $<1$  day, which on average was 3 to nearly 9 orders of magnitude shorter than the time span covered by the conventional metric of duration (Fig. 3B). As with extent, the magnitude of difference declined steadily with actual duration, effectively disappearing for the 16% of observations with  $\geq 1$  month of actual duration.



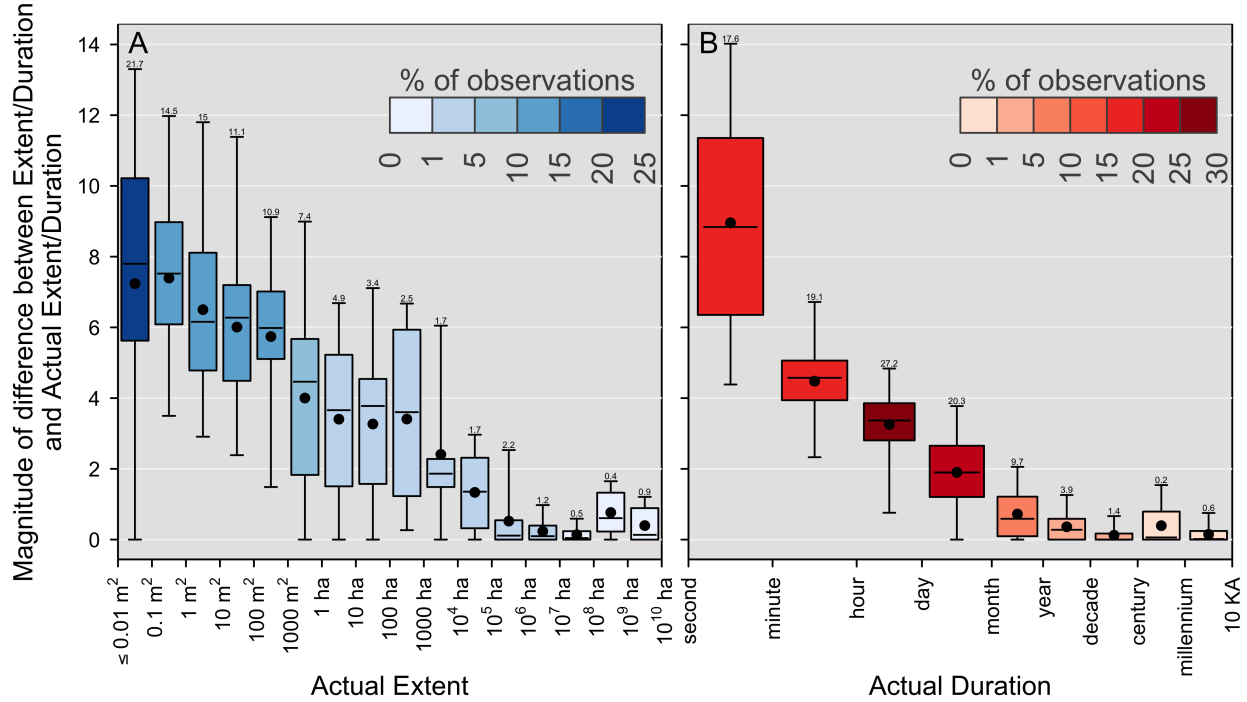


Figure 3: The difference between extent and *actual* extent (the summed area of spatial replicates) (A) and duration and *actual* duration (the summed sampling duration across temporal replicates) (B). Difference values are expressed in terms of how many orders of magnitude larger (or longer) extent (duration) is than actual extent (actual duration), and are summarized (as box plots, with circle in box representing the mean and line the median) in bins representing increasing scales of actual extent/duration. The percentages of observations falling within each bin are indicated by the color of the inter-quartile and the numeric value above the upper whisker.

### Observational methods

We classified the method used to collect each observation into several broad categories, which were field methods (manual *in situ* data collection), automated (*in situ*) sensing, remote sensing, other geographic data, and paleo-reconstruction approaches. Field methods were used for 80% of observations, automated sensing for 12.4%, remote sensing for 6.4%, and paleo-reconstruction and other geographic data each for less than 1% each. Using linear regression (weighted by the number of observations per publication year) to assess whether the relative frequency of observing methods changed during the 10 year study period, the use of remote sensing appeared to increase

by 1.3% per year from 2004-2014 ( $R^2 = 0.25$ ,  $p < 0.12$ ), and field methods declined by the same percentage ( $R^2 = 0.1$ ,  $p < 0.18$ ), although both slopes failed to meet the customary threshold for statistical significance. Automated sensing methods showed no trend over time (SI).

#### ***Potential biases and uncertainties in quantifying scales***

There were several potential methodological aspects that could have influenced our assessment of ecology's spatial and temporal domains. The first stems from our finding that many studies did not precisely report observational scales, which meant that we had to estimate, rather than simply record, these values for most observations (specifically, in 63%, 60%, and 69% of cases for resolution, extent, and actual extent, and 36%, 64%, and 83% of cases for interval, duration, and effective duration, respectively). The inevitable estimation errors may have biased our overall findings. However, we attempted to quantify and account for this error by assessing inter-rater disagreement and incorporating this uncertainty into our resampling methodology. The resulting confidence intervals (Fig. 1) suggest that it was unlikely that estimation errors unduly influenced our findings.

Another potential source of bias lies within our scale-estimation protocols, chiefly with respect to our rule for estimating resolution (the smallest areal unit of *complete* measurement). We selected this definition for the sake of consistency, but some papers reported resolution as a larger area in which sub-samples were taken. For these, our estimates are finer than what the studies' authors apparently considered to be plot resolution. Additionally, our domain estimates would presumably be somewhat different if we had included experimentally manipulated observations. For example, average resolution and duration would likely be finer (7, 8).

Finally, because our review did not include papers beyond 2014, the omission of studies from the most recent years could have introduced bias into our domain estimates. Indeed, if the trend towards increasing use of remote sensing between 2004-2014 was not spurious, we can project that a repeated study applied to papers published between 2004-2017 would find remote sensing used

for 7.7% of observations (a 22% increase), which would increase mean extent by 17.4% (95% CI = -1.3-67%; or 0.07 orders of magnitude) above the 2004-2014 average (see SI for details of calculation). Further evidence for this trend lies within the extent values themselves, which increased 0.25 orders of magnitude per year between 2004-2014 ( $R^2 = 0.25$ ,  $p < 0.07$ ). This somewhat clearer trend also suggests that including more recent studies would have increased mean extent, but by a more modest 5.5% (0.02 orders of magnitude).

## Discussion

### *Insights into the scale domains of modern ecology*

Our results suggest that the scale domains of modern ecological observations are fairly narrow, and are collected primarily with conventional field techniques. Spatially, the majority of observations have grains of  $\leq 1 \text{ m}^2$  and extents of  $\leq 10,000 \text{ ha}$  (Fig 1A;B). In the temporal domains, most observations are either un-replicated snapshots, or of low frequency ( $\geq 1$  month interval; Fig. 1C) and relatively short duration ( $\leq 1$  year; Fig. 1D; 2D). Contrasting observational dimensions reveals that larger extents are associated with larger plot sizes (Fig. 2C), while longer durations are associated with longer intervals (Fig. 2D). The latter association presumably reflects a cost-imposed tradeoff between sampling frequency and temporal duration that is characteristic of traditional field-based observation. The same tradeoff is also responsible for the inverse relationship between resolution and interval that dominates that domain space (Fig. 2A). As a result of these tradeoffs, there are notable “holes” in the domains defined by high frequency (daily to sub-daily intervals) observations having 1) high to moderate resolutions ( $\geq 1 \text{ m}^2$  up to 100 ha; Fig. 2A) and 2) decadal or longer durations (Fig. 2B).

Have these domains changed since the seminal papers on scale first began to appear in the late 1980s (1, 3, 7)? A comprehensive answer to this question would require a similar study focused on earlier literature, but an analysis of results presented in three prior studies provides partial insight. The first and most comprehensive dataset consists of duration values extracted by Tilman (7) from

623 studies published between 1977-1987 in the journal *Ecology*. The average duration of the most comparable subset of those values ( $n=419$ ; see SI) was 3.6 years, compared to 3.3 years for observations in our sample (or 5.1 if temporally un-replicated observations are excluded). The second dataset is found in Kareiva and Anderson (8), who present the resolutions of 97 community ecology experiments published in *Ecology* between 1980-1986. The average of those ( $12,657 \text{ m}^2$ ) was substantially smaller than the mean of our sample ( $1,496,069 \text{ m}^2$ ), but comparing the 80th percentile value ( $197 \text{ m}^2$ ) of Kareiva and Anderson's (8) to ours ( $115 \text{ m}^2$ ) shows that the majority of contemporary observations are finer-grained than most 1980s-era experiments. The third dataset is provided by Porter et al (11), who compared the extent and interval of 25 studies published in 2003 and 2004 (also in *Ecology*). The mean interval was 178 days, compared to 684 days in our sample, but the 80th percentile value in our study was 169 days compared to 329 days in theirs. Extent in our sample was substantially larger according to multiple summary statistics, including the mean ( $114,965,072 \text{ ha}$  in our study versus  $368,403 \text{ ha}$ ), median ( $5,051 \text{ ha}$  versus  $9 \text{ ha}$ ), and 95th percentile ( $46,424,808 \text{ ha}$  versus  $136,000 \text{ ha}$ ).

Although limited due to methodological differences (e.g. a focus on experiments versus un-manipulated systems), these comparisons suggest that the duration and, less clearly, resolution of ecological observations have changed little in the past 30 years, but observational frequency and particularly extent have increased. The growth in observational extent is also reflected by the positive year-on-year trend within our own dataset, which corresponds to the increasing use of remote sensing (see figures in SI).

However, even though observational extent is increasing, there remains for most observations a large gulf between the area that is actually sampled and that which the spatial replicates purportedly represent (Fig. 3A). A similarly large discrepancy is also evident between the time spanned by repeated observations and the time that is spent actually observing a phenomenon (Fig. 3B). These differences between the actual and *representative* scales of observation have implications for eco-

logical understanding, as the unobserved portions of space and time may contain important patterns and processes that are not captured by replicates, due to phenomenon-dependent factors such as autocorrelation and representativeness of the sampling scheme (13–17). Brief, infrequent snapshots, or fine-grained, spatially sparse replicates, may be sufficient to characterize many phenomena (as an example with respect to temporal replication, annual changes in tree cover are well-represented by low frequency satellite imaging (18)), but may be inadequate for more dynamic phenomena. For example, wildfire extent and duration can be mapped by daily return satellites (19, 20), but the instantaneous nature of the imaging means that it cannot be used to observe fire behavior (21). To capture such behavior, long periods of continuous observation may be more important for understanding dynamics than frequent repeats.

It is therefore important to examine whether the scales of the phenomena being observed are adequately captured by the design of replicates. Our methods suggest one possible procedure for assessing the *scale representativeness* of replicates: 1) measure auto-correlation (spatial or serial) in the replicates, 2) add the autocorrelation length to the replicate area/duration, 3) calculate an autocorrelation-adjusted actual extent/duration, and 4) plot where it falls between actual extent/duration and extent/duration. The distance between the adjusted actual value and the ostensible value can provide a measure of how well the replicates represent the intended scale of observation. If the gap remains large, then alternative sampling methods may be used to close it. For example, remote sensing provides wall-to-wall spatial coverage of a study area, erasing the difference between actual extent and extent. Furthermore, the interval of high-resolution imaging (higher resolution is preferred in images as it allows individual features to be better discerned (22, 23)) is now approaching daily to sub-daily scales (24, 25), allowing improved representation of spatial and temporal dynamics. For phenomena that can't be measured from space, either because they are not visible or because they require continuous observation, new approaches for collecting *in situ* or near-surface observations (e.g. low-cost wireless sensors (11, 26, 27), citizen observers (28),

and autonomous vehicles (29)) can be used to increase the spatial and temporal coverage of observations.

The aforementioned insights regarding modern observational domains must be tempered by the uncertainty within our own scale estimates, as detailed in the preceding section. However, most of this uncertainty is attributable to unclear reporting of scale values in the majority of papers we reviewed (a problem also noted in geography studies (30)). This tendency towards vague documentation offers one final insight, which is that, despite decades of accumulated knowledge regarding the importance of scale in ecology (1–3, 31), scale appears to remain a low priority throughout much of the discipline. Beyond contributing to the broader problem of scientific reproducibility (32), inattentiveness to scale increases the risk that observations inadequately represent the phenomenon of interest, thereby limiting the generalizability of any derived ecological knowledge (3, 30, 31). To mitigate this problem, we recommend that ecological journals require authors to quantify and clearly report the values of resolution, extent, interval, and duration.

### ***Looking forward***

Our study suggests that the concept of scale has yet to fully permeate the discipline of ecology. Evidence for this assertion lies in the continued narrowness of ecology’s observational scale domains and the poor documentation of scale dimensions in the literature. However, the increasing extent of ecological observations, enabled by remote sensing and presumably motivated by many ecologists’ appreciation of scale-related issues, suggests that ecology’s scale domains are gradually changing. In the coming years, the accelerating gains in technology and analytical methods will allow researchers new and unprecedented capabilities to peer into, and thus close, the prominent holes in observational scale domains. A renewed, discipline-wide focus on scale’s importance, including the adoption of stricter scale-reporting standards by journals, will help spur ecologists to address these gaps, while fostering the improved transferability of knowledge within the discipline.

## References

1. Levin, S. A. The problem of pattern and scale in ecology. *Ecology* **73**, 1943–1967 (1992).
2. Chave, J. The problem of pattern and scale in ecology: What have we learned in 20 years? *Ecology Letters* **16**, 4–16 (2013).
3. Wiens, J. A. Spatial scaling in ecology. *Functional Ecology* **3**, 385–397 (1989). PT: J.
4. Tingley, M. W., Koo, M. S., Moritz, C., Rush, A. C. & Beissinger, S. R. The push and pull of climate change causes heterogeneous shifts in avian elevational ranges. *Global Change Biology* **18**, 3279–3290 (2012).
5. Xiao, J. F. & Moody, A. Photosynthetic activity of US biomes: Responses to the spatial variability and seasonality of precipitation and temperature. *Global Change Biology* **10**, 437–451 (2004).
6. Schneider, D. C. The rise of the concept of scale in ecology. *BioScience* **51**, 545–553 (2001).
7. Tilman, D. Ecological experimentation: Strengths and conceptual problems. In *Long-Term Studies in Ecology*, 136–157 (Springer, 1989).
8. Kareiva, P. & Andersen, M. Spatial aspects of species interactions: The wedding of models and experiments. In *Community Ecology*, 35–50 (Springer, 1988).
9. Turner, W. *et al.* Remote sensing for biodiversity science and conservation. *Trends in Ecology & Evolution* **18**, 306–314 (2003).
10. Pettorelli, N. *et al.* Satellite remote sensing for applied ecologists: Opportunities and challenges. *Journal of Applied Ecology* **51**, 839–848 (2014).
11. Porter, J. *et al.* Wireless Sensor Networks for Ecology. *BioScience* **55**, 561–572 (2005).

- 263 12. Estes, L. D. *et al.* A platform for crowdsourcing the creation of representative, accurate land-  
264 cover maps. *Environmental Modelling & Software* **80**, 41–53 (2016).
- 265 13. Underwood, A. J. *Experiments in Ecology: Their Logical Design and Interpretation Using*  
266 *Analysis of Variance* (Cambridge University Press, 1997).
- 267 14. Palmer, M. W. & White, P. S. Scale dependence and the species-area relationship. *American*  
268 *Naturalist* 717–740 (1994).
- 269 15. Cao, Y., Williams, D. D. & Larsen, D. P. Comparison of ecological communities: The problem  
270 of sample representativeness. *Ecological Monographs* **72**, 41–56 (2002).
- 271 16. Legendre, P. Spatial autocorrelation - trouble or new paradigm? *Ecology* **74**, 1659–1673  
272 (1993).
- 273 17. Collins, S. L., Micheli, F. & Hartt, L. A method to determine rates and patterns of variability  
274 in ecological communities. *Oikos* **91**, 285–293 (2000).
- 275 18. Hansen, M. C. *et al.* High-resolution global maps of 21st-century forest cover change. *Science*  
276 **342**, 850–853 (2013).
- 277 19. Roy, D., Jin, Y., Lewis, P. & Justice, C. Prototyping a global algorithm for systematic fire-  
278 affected area mapping using MODIS time series data. *Remote Sensing of Environment* **97**,  
279 137–162 (2005).
- 280 20. Jones, B. M. *et al.* Fire behavior, weather, and burn severity of the 2007 Anaktuvuk River  
281 tundra fire, North Slope, Alaska. *Arctic, Antarctic, and Alpine Research* **41**, 309–316 (2009).
- 282 21. Clements, C. B. *et al.* Observing the dynamics of wildland grass fires: FireFlux-a field vali-  
283 dation experiment. *Bulletin of the American Meteorological Society* **39**, 1369–1382 (2007).



- 284 22. Dark, S. J. & Bram, D. The modifiable areal unit problem (MAUP) in physical geography.  
285 *Progress in Physical Geography* **31**, 471–479 (2007).
- 286 23. Hay, G. J., Blaschke, T., Marceau, D. J. & Bouchard, A. A comparison of three image-object  
287 methods for the multiscale analysis of landscape structure. *ISPRS Journal of Photogrammetry*  
288 *and Remote Sensing* **57**, 327–345 (2003).
- 289 24. Drusch, M. *et al.* Sentinel-2: ESA's Optical High-Resolution Mission for GMES Operational  
290 Services. *Remote Sensing of Environment* **120**, 25–36 (2012).
- 291 25. Hand, E. Startup liftoff. *Science* **348**, 172–177 (2015).
- 292 26. Wolf, A., Falusi, J., Caylor, K., Sheffield, J. & Wood, E. A GSM-based surface meteorology  
293 network in service of improved African hydrological data assimilation and drought forecasting  
294 (San Francisco, 2012).
- 295 27. Collins, S. L. *et al.* New opportunities in ecological sensing using wireless sensor networks.  
296 *Frontiers in Ecology and the Environment* **4**, 402–407 (2006).
- 297 28. Dickinson, J. L. *et al.* The current state of citizen science as a tool for ecological research and  
298 public engagement. *Frontiers in Ecology and the Environment* **10**, 291–297 (2012).
- 299 29. Anderson, K. & Gaston, K. J. Lightweight unmanned aerial vehicles will revolutionize spatial  
300 ecology. *Frontiers in Ecology and the Environment* **11**, 138–146 (2013).
- 301 30. Margulies, J. D., Magliocca, N. R., Schmill, M. D. & Ellis, E. C. Ambiguous Geographies:  
302 Connecting Case Study Knowledge with Global Change Science. *Annals of the American*  
303 *Association of Geographers* **106**, 572–596 (2016).
- 304 31. Wheatley, M. & Johnson, C. Factors limiting our understanding of ecological scale. *Ecological*  
305 *Complexity* **6**, 150–159 (2009).

306 32. Baker, M. 1,500 scientists lift the lid on reproducibility. *Nature News* **533**, 452 (2016).

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