COMMUNITY ECOLOGY - ORIGINAL RESEARCH



Drivers of taxonomic, functional and phylogenetic diversities in dominant ground-dwelling arthropods of coastal heathlands

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Abstract

Although functional and phylogenetic diversities are increasingly used in ecology for a variety of purposes, their relationship remains unclear, and this relationship likely differs among taxa, yet most recent studies focused on plants. We hypothesize that communities may be diverse in functional traits due to presence of: many phylogenetic lineages, trait divergence within lineages, many species and random functional variation among species, weak filtering of traits in favorable environments, or strong trait divergence in unfavorable environments. We tested these predictions for taxa showing higher (ants), or lower (spiders, ground beetles) degrees of competition and niche construction, both of which might decouple functional traits from phylogenetic position or from the environment. Studying > 11,000 individuals and 216 species from coastal heathlands, we estimated functional as minimum spanning trees using traits related to the morphology, feeding habits and dispersal, respectively. Relationships between functional and phylogenetic diversities were overall positive and strong. In ants, this relationship disappeared after accounting for taxonomic diversities and environments, whereas in beetles and spiders taxonomic diversity is related to functional diversity only via increasing phylogenetic diversity. Environmental constraints reduced functional diversity in ants, but affected functional diversity only indirectly via phylogenetic diversity (ground beetles) and taxonomic and then phylogenetic diversity (spiders and ground beetles). Results are consistent with phylogenetic conservatism in traits in spiders and ground beetles. In ants, in contrast, traits appear more phylogenetically neutral with any new species potentially representing a new trait state, tentatively suggesting that competition or niche construction might decouple phylogenetics from trait diversity.

Keywords Neutral theory \cdot Phylogenetic conservatism \cdot Functional trait \cdot Spiders \cdot Carabids \cdot Ants \cdot Maritime clifftops \cdot Habitat filtering

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Introduction

Species communities are considered particularly valuable if they are biologically diverse. Biological diversity has for long been quantified in terms of richness of species, i.e. "taxonomic diversity" (TD). However, a community that is diverse taxonomically is not necessarily diverse in terms of phylogenetic lineages or functional traits represented by the community, quantified as minimum spanning tree connecting all species across a phylogeny or in a trait-space: the phylogenetic or functional diversities (PD, FD), sensu Faith (1992). Growing in popularity during the last twenty years (Webb et al. 2002; Campbell et al. 2010), FD and PD have been very useful in a large variety of contexts. In applied ecology for instance, phylogenetic diversity (PD) was successfully used to establish conservation prioritization (Rodrigues et al. 2005; Magura 2016; Tucker et al. 2018; Wong



et al. 2019). PD was also proven relevant for understanding community assembly rules and underlying interactions (Cavender-Bares et al. 2009). The same is true for functional diversity (FD), with cases where functional diversity led to a better understanding of ecological processes that classical taxonomic diversity alone did not properly assess (Leroy et al. 2014; Campbell et al. 2010; Wong et al. 2019).

A positive relationship between different aspects of diversities has often been taken for granted, for instance when using PD as a proxy for FD (Webb et al. 2002). However, existing theory on ecological communities makes different hypotheses and predictions on relationships between diversities. Specifically, we may hypothesize that communities may be diverse in functional traits due to the presence of many phylogenetic lineages and trait conservatism within lineages (sensu Wiens et al. 2010). Phylogenetic conservatism is often considered to be strong (Darwin 1859 in Webb et al. 2002; Wiens 2010), but has rarely been demonstrated locally (Pavoine et al. 2010) and hence the links between local PD and FD are not fully elucidated (Gerhold et al. 2015; Cadotte et al. 2019). Only if local phylogenetic conservatism is perfect, does an increase in PD inevitably increase FD (Webb et al. 2002). However, locally, phylogenetic conservatism may not be perfect or even be absent (Eterovick et al. 2010; Haak et al. 2014; Grundler et al. 2018; Baulechner et al. 2020). In that case, we may hypothesize that community assembly is driven by the most simple, neutral processes (Hubbell 2001): traits are unimportant so that any species arriving or going extinct carries a random chance of representing a unique trait state within a given community (Purves and Turnbull 2010; Novack-Gottshall 2016). FD would hence increase with taxonomic diversity (TD), not with PD. Alternatively, we may hypothesize that under low PD, character displacement occurs: the many related species outcompete each other if too similar, leading to a negative relationship between PD and FD (Prinzing et al. 2008) or at least a disappearance of positive relationship. Finally, FD may be driven by the environment rather than PD or TD of the community. We may hypothesize that severe environmental constraints select against certain functional trait states and hence reduce FD (Webb et al. 2002), or that such constraints impose character displacement among species and reduce niche packing (Safi et al. 2011; Prinzing 2016) and thereby increase FD. FD would hence decrease or increase, respectively, with environmental constraints. Different hypotheses might apply to different taxa. Taxa where competition is important are least likely to show a positive relationship between PD and FD and most likely to show a negative. Also, in taxa capable of niche construction (i.e. enduringly manipulating their selective environment: Odling-Smee et al. 2013), traits might become emancipated from the environment, resulting in species diversity being little constrained by environmental filters.

Despite this diversity of theoretically possible relationships between PD and FD, several authors argued that PD is a proxy of FD, and that the more traits are used to calculate FD the more it will resemble PD (Flynn et al. 2011; Khalil et al. 2018; Tucker et al. 2018). As recently highlighted by Cadotte et al. (2019), most studies showed a positive correlation between phylogenetic and functional diversity (PFC). Yet the spectrum of correlation strength is wide, with a few studies even reporting negative correlations (Prinzing et al. 2008; Bernard-Verdier et al. 2013). This gray area led to debate about the relationships between PD and FD (Pavoine et al. 2013; Tucker et al. 2018), and highlights the need for more empirical studies. Yet, most of previous work was done on plant communities (Cavender-Bares et al. 2009; Campbell et al. 2010), this redundancy of study models being recurrent in ecology, and easily fixed by multiplying model taxa (Coelho et al. 2009; Gerlach et al. 2013; Wong et al. 2019).

Our study therefore proposes to compare the PFC between several dominant taxa of arthropods using a standardized sampling protocol taking advantage of a largescale restoration design on maritime clifftops, and to assess whether taxonomic diversities and/or abiotic forces drive both functional and phylogenetic diversities in each taxa. We studied three groups of arthropods, spiders, ground beetles and ants. These arthropods are known to complement vegetation survey (Pétillon et al. 2014; Hacala et al. 2020), and thereby offer an interesting opportunity to better understand the link between PD and FD (Wong et al. 2019). Spiders, ants and ground beetles are dominant ground-active macroarthropods in several temperate habitats, and their traits and phylogeny are relatively well known (Bond et al. 2014; Pedley et al. 2014; Schirmel et al. 2016; Magura 2016; Parr et al. 2017). To our knowledge, one publication has already treated PD and FD for a single group of ground-dwelling arthropods (Arnan et al. 2015), and some have conducted simple bivariate correlations between FD to PD (Corbelli et al. 2015; Liu et al. 2016), all having observed a positive PFC (but see Ridel et al. 2021). The novelty of our study resides in the fact that we look for more than just bivariate correlation but investigate the complex driving forces behind patterns of taxonomic functional and phylogenetic diversities of spiders, carabids and ants. Moreover, we use our above framework of hypotheses to predict differences between spiders, carabids and ants in what drives FD: The three groups differ in the two characters that may influence the drivers of FD. First, competition has been frequently demonstrated among ants (Cerda et al. 2013) but less so in ground beetles (Niemelä 1993) and least in spiders (Wise 2006). Given our above hypotheses, we hence expect negative relationships between PD and FD to be most likely or positive relationships between PD and FD least likely in ants, and the inverse in spiders. Second, niche construction



is very strong in ants through building of formicaries and farming fungi or aphids (Odling-Smee et al. 2013), it exists to a lower degree in many spiders due to web-building, and it is probably inexistent or weak in ground beetles. Given our above hypotheses, ants might hence depend least on abiotic factors that constrain individual survival in other taxa, such as salinity or humidity (Petillon et al. 2008 for ground beetles and spiders) and ground beetles might depend most on these factors. However, we acknowledge that ants might depend more on factors that constrain the construction of formicaries such as soil depth (Schmidt et al 2017).

Methods

Sampling and identification

Sampling took place in three coastal sites of Brittany, Western France in 2017: L'Apothicairerie (47° 21′ 44.0″ N, 3° 15′ 34.9″ W), La Pointe de l'Enfer (47° 37′ 18.3″ N 3° 27′ 46.9″ W) and La Pointe de Pen-Hir, located on the mainland (48° 15′ 03″ N, 4° 37′ 25″ W). These sites were selected for they are all comparable with similar dominant vegetation (a short and dry heathland dominated by *Erica* spp. and *Ulex* spp.), and under ongoing ecological passive restoration. The degradation on each sites was subdivided into 3 degradation states. This division was defined by vegetation states, bare-ground surface and age from restoration actions. This subdivision was only used to spacialize the sampling for bare ground and was used as a proxy of degradation in our models (see full description of study sites and pictures in Hacala et al. 2020 and landscape context in Online Resource 1).

Two 400 m² plots of homogeneous vegetation were set for each degradation state, and four pitfall traps (80 mm in diameters and 100 mm deep) were set at each plot. Each plot were situated at least 100 m from the nearest differing habitats (see Online Resource 1). Traps were half-filled with a salted solution (250 g L⁻¹) with a drop of odorless soap and settled 10 m apart to avoid interference and local pseudoreplication (Topping and Sunderland 1992). This resulted in 71 traps (in one station, the sampling area was too restricted to set 4 traps spaced of 10 m apart, so 1 was removed) active between mid-March to mid-June 2017, and emptied every 2 weeks. We hence focus on the spring community as pooling with the autumn community would pool different body sizes (Lövei and Sunderland 1996) across species that de facto do not co-occur in space and time. Total plant cover was estimated in a 5 m radius circle around each trap, and all species were identified and their percentage cover estimated. Environmental variables where inferred from the vegetation communities using Ellenberg's index values (1992) extracted from Hill et al. (2004) and corrected for the British Isles, following Lafage et al. (2015, 2019) and Diekmann (2003). Anthropogenic degradation was assessed using bare ground as a proxy of degradation intensity (Hacala et al. 2020). Pitfall samples were sorted in laboratory, arthropods transferred to ethanol 70%, and stored at the University of Rennes 1. Spiders, grounds beetles and ants were identified to species level. Spiders were identified using Roberts (1985) and Nentwig et al. (2019). Ground beetles were identified using Luff (2007) and Jeannel (1941). Ants were identified to species levels using Blatrix et al. (2013).

Trait gathering

Three criteria were set to select the traits for calculating FD of all three taxa. (1) They exist under a form or can be transformed in a way to be comparable between taxa (e.g. numerical into categorical). (2) They must be reported as drivers of species assemblages for each taxon. (3) They must be available in the literature for the three taxa. The functional trait definition provided by Violle et al. (2007) and followed by Wong et al. (2019) was finally used, with the exclusion of environmental preferences as they are not strictly speaking a functional trait. Three traits were selected and their values gathered from existing literature: dispersal abilities, body size and trophic guild (Table 1), all these traits being known to be affected by environmental responses in spider, carabid and ant assemblages (Schirmel et al. 2012; Lafage et al. 2015). Dispersal trait was chosen to express dispersal at a landscape scale to account for the colonization ability of the studied species. The spider capacity to disperse by the air (ballooning) was the criterion used to discriminate dispersal ability. For ground beetles, the ability to fly based on wing development was used for dispersal ability. For the ants, the queen's way of locomotion for founding a new colony (by flight or by foot after a colony division) was used as comparable landscape level of dispersal. Body size was divided into three discreet classes to facilitate the comparison with the other variables. The division into size classes was done for each class to contain a comparable number of species. The guild trait was acquired through the feeding regime for ants and ground beetles. For spiders, the foraging strategy was used to distinguish guilds inside a predator-only taxon. We stress that traits were never inferred from phylogenetic position (see Table 1 for details and literature used).

Phylogenetic tree building

Phylogenetic trees were constructed by combining phylogenetic and taxonomic data from literature, assuming identical branch length between genus (1) and species (0.5) as real branch lengths because sequences were not available for all identified species.

The spider phylogenetic tree was adapted from Wheeler et al. (2017). Genera that were not present in Wheeler's tree were placed using Arnedo et al. (2009), Frick et al. (2010)



Table 1 Details of functional trait for each taxon. Each trait is decomposed in its modalities for each of the three taxon

Trait	Spider	Refs.	Ground beetle	Refs.	Ant	Refs.
Dispersal	Ballooning 0/1	[1, 2]	Apterous/ polymorphic/ macropterous	[5]	Colony division/ mixte/independ- ent	[9]
Size	Small: <3 Medium: [3; 5] Large: >5	[3]	Small: < 5 Medium: [5; 10] Large: > 10	[5]	Small: <3 Medium: [3; 4] Large: >4	[8–9]
Guild	Other hunters Ground hunters Sheet web weavers Space web weavers Sensing web weavers Specialists Orb web weavers Ambush hunters	[4]	Predator Omnivore Herbivore	[5–7]	Omnivor Predator Nectarivor Granivore Parasite	[8–9]

The bibliographic source from which the traits were gathered are specified (Hacala et al. 2020)

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and Wang et al. (2015) for Linyphiidae and Agnarsson (2004), Azevedo et al. (2018), Maddison (2015), Millidge (1978), Piacentini and Ramírez (2019) and Scharff et al. (2019) for other families. The carabid phylogenetic tree was adapted from López-López and Vogler (2017), Martínez-Navarro et al. (2005), Ober and Maddison (2008), Ruiz et al. (2009) and Sasakawa & Kubota (2007). The ant phylogenetic tree was adapted from Moreau et al. (2006).

Diversity calculating

The two diversity metrics, PD and FD, were calculated using the BAT package (Cardoso et al. 2015). The diversities values were calculated at the trap scale with a data pooling from the season of sampling for each taxon. Distance matrices were computed with Gower distance from the FD package (Laliberté et al. 2014) for functional distances and with as phylo from the ape() package (Paradis et al. 2019)

for phylogenetic distances. PD and FD were calculated by meaning the Jackknifes estimates from apha-estimated () function from BAT to account for sampling variability.

Statistical analysis

Correlations between PD and FD were estimated in a Bayesian framework with a Student's t distribution with the brms package (Bürkner 2018). We used 2000 iterations on 4 chains. Model convergence was checked by visually inspecting diagnostic plots.

To select environmental variables affecting PD and FD, models were built within a Bayesian framework using brms (Bürkner, 2018) with two chains and default priors. All environmental variables were standardized and centered. The models included % bare-ground, litter depth, soil depth, and vegetation height. It also included community-weighted means of nitrogen level, light, salinity, pH and humidity



based on Elenberg indicator values (modified by Hill et al. 1999). Model convergence was checked by visually inspecting diagnostic plots and using Rhat value. Parameter selection was based on "HDI+ROPE decision rule" (Kruschke and Liddell 2018) with a range value determined as -0.1 * sd(y), 0.1 * sd(y) (Kruschke and Liddell 2018) and was performed using bayestestR (Makowski et al., 2019). We also provide we the probability of direction (pd), which is the probability that the posterior distribution of a parameter is strictly positive or negative. The thresholds beyond which the variable was considered as significant (i.e. non negligible) were pd > 95% and ROPE < 2.5%.

We assessed the relative contribution of environmental variables selected by Bayesian models using structural equation modeling (SEM). The SEM approach also allowed us to assess the links between TD, PD and FD taking environment into account. A significant correlated error between the two variables would indicate the existence of an unknown parameter influencing both variables. We used the piecewiseSEM package (Lefcheck 2016) as it allows using mixed models in association with nlme package (Pinheiro et al. 2020). Our initial model included the following links: (1) PD is affected by TD and selected environmental variables, (2) FD is affected by TD, selected environmental variables and PD, (3) TD is affected by selected environmental variables and (4) there is correlated error between PD and FD. Site was used as a random factor in every link modeled using nlme (Pinheiro-Silva et al., 2020). After the specification of the initial model, we re-defined our model excluding nonsignificant links (p < 0.05) using a stepwise approach until Δ AICc < 2 between two subsequent models. Finally, we assessed model fit using Fisher's C statistic.

Results

Simple correlations

Correlations between phylogenetic and functional diversities were 0.59 (95% CI 0.41–0.73), 0.65 (95% CI 0.46–0.79) and 0.72 (95% CI 0.57–0.82) for spiders, carabid beetles and ants respectively (Fig. 1), and overall increased with decreasing species richness of each taxa (153, 40 and 23, respectively).

Variable pre-selection

Humidity and salinity were the best explanatory variables for spider PD (Table 2). None of the environmental variables included in the model explained spider FD (Table 2). As for spider PD, spider TD was best explained by humidity and salinity. Percent bare-ground and salinity were the best explanatory variables for carabid beetle PD Carabid beetle FD and TD were best explained by salinity (Table 2). Soil

depth was the best explanatory variable for ant PD, FD and TD (Table 2).

Structural equation models

When testing the relationship between diversity metrics and environmental variables, our final SEMs indicated good fit with the data both for spiders (Fisher's C = 17.19, p = 0.07; Fig. 2a) and carabid beetles (Fisher's C = 9.77, p = 0.64, Fig. 2b). The fit of the ant model could not be estimated as the best model was fully saturated (Fig. 2c).

Environmental variables (here humidity and salinity indexes) were only linked to spider TD. Direction of these links were opposite with similar sizes. Spider PD was strongly and positively related to TD while spider FD was strongly and positively linked to PD.

Similar (but weaker) relationships were found between carabid beetle α -, phylogenetic and functional diversities (Fig. 2). Environmental variables affecting carabid diversities were salinity and % bare-ground. Salinity positively influenced TD and PD. % bare-ground had a negative effect on PD.

Ant phylogenetic and functional diversities were strongly positively linked to TD. Soil depth was the only environmental variable selected by the model. It had a weak and positive effect on FD. Selected, but not significant, paths were found between functional and phylogenetic diversities (p = 0.07), between soil depth and TD (p = 0.08) and between soil depth and PD (p = 0.22).

Discussion

Correlations were overall positive and high between FD and PD, following general theory (Webb et al. 2002). These simple correlations between PD and FD may or may not represent mathematical artifacts from an increase of both, FD and PD with TD. We will hence focus the discussion on the results that account for TD and other variables in parallel. Accounting for TD and environments made disappear this PD/FD relationship in ants, but maintained them in spiders and ground beetles, where TD related to FD only via increasing PD. Environmental constraints reduced FD in ants, but affected FD only indirectly via PD (ground beetles) and via TD and then PD (spiders and ground beetles).

Differences of drivers between taxonomic, functional and phylogenetic diversities

The anthropogenic gradient of degradation, represented by bare ground, affected PD only in ground beetles, while this variable is known to affect TD of ground beetles in other habitats (see e.g. Pétillon et al. 2008 in salt marshes). Other



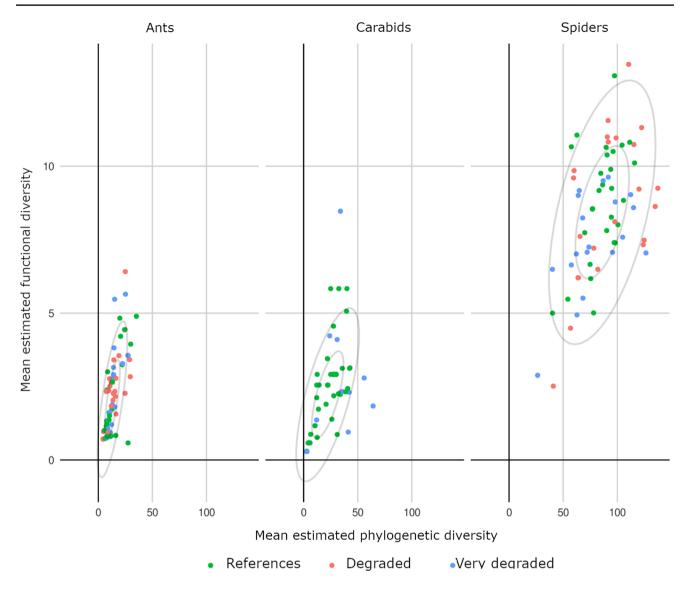


Fig. 1 Plot of mean estimated functional diversity as a function of phylogenetic diversity. Ellipses correspond to 95% and 5% normal confidence ellipses. Each dot correspond to a trap. $R^2 = 0.38$; 0.42,

and 0.52. See Fig. 2 for a broader picture accounting also for taxonomic diversity and the environment (Hacala et al. 2020)

environmental variables on the other hand have interacted with PD and FD in several ways, positively (humidity for spiders, salinity for ground beetles and soil depth for ants) or negatively (salinity for spiders), but most of their effects were indirect: With two exceptions (bare ground on PD of ground beetles and soil depth on FD of ants), environmental variables acted only on TD which then cascaded its effects onto PD which in turn affected FD. Both, PD and FD are estimated using minimum spanning trees, but possibly a functional minimum spanning tree contains more polytomies, and hence adding a species without adding a new trait adds little branch length (see Online Resource 2 for a simulation supporting this idea). Moreover, environments selecting for species may do so based on traits that we

did not consider in the calculus of FD, and that are highly phylogenetically convergent, leading to a major increase in PD. The fact that PD often increases with TD might reflect either the effect of an unmeasured environmental variable (e.g. weather conditions as they are known to be harsh in the studied habitat see Sawtschuck 2010) affecting both TD and PD, or a coexistence facilitated among phylogenetically distant species, which would result in a positive relationship between TD and PD. Traits involved in stress response, like drought tolerance, indeed appear to have evolved convergently (Dunn et al. 1976). Increase in the diversity of the functional traits that we did consider would only be a side effect of such selection for phylogenetically convergent tolerances or resistances. Other traits more directly related



Table 2 Results Bayesian models for variable selection to be used in the SEM models (Hacala et al. 2020)

		R^2	Variable selected	Pd (%)	Median	89% CI	% in ROPE
Spiders	PD	0.59	Humidity	99.6	16.37	[7.45; 26.31]	0
			Salinity	97.8	- 21	[-38.68; -4.90]	0
	FD	0.26	_				
	TD	0.66	Humidity	99.4	0.56	[0.27; 0.92]	0
			Salinity	99.4	-0.98	[-1.56; -0.34]	0
Carabid beetles	PD	0.44	% Bare ground	99.2	-6.9	[-11.73; -2.67]	0
			Salinity	99.9	21.01	[8.11; 32.19]	0
	FD	0.24	Salinity	99.4	2.61	[0.78; 4.09]	0
	TD	0.44	Salinity	89.2	0.64	[-0.29; 1.45]	0
Ants	PD	0.31	Soil depth	97.5	2.84	[0.68; 5.11]	0.4
	FD	0.4	Soil depth	99.1	0.59	[0.18; 0.99]	0
	TD	0.37	Soil depth	98.1	0.36	[0.09; 0.66]	1.1

 R^2 conditional R^2 , Pd probability of direction (probability that the posterior distribution of a parameter is strictly positive or negative), CI confidence interval, % in ROPE percentage of the posterior distribution in the Region of Practical Equivalence

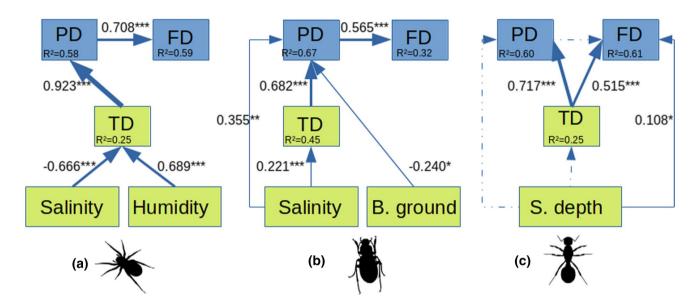


Fig. 2 Best piecewise SEMs showing links between taxonomic, phylogenetic and functional diversity and environmental variables for **a** spiders, **b** carabid beetles and **c** ants. Thickness of arrows is proportional to the standardized path coefficients (directionality and size

given within boxes). Asterisks give significance level of linkages (*<0.05, **<0.01, ***<0.001), and dashed lines correspond to paths included but not significant (p>0.05). Conditional R^2 values are given within the boxes containing variables (Hacala et al. 2020)

to environmental stresses could be worthwhile to integrate in the future, as they could be available through the development of online databases (see e.g. Lowe et al. 2020 for spiders).

Differences of drivers between taxa

Diversity patterns differing between taxa was one of our expectations (Wong et al. 2019). We had expected positive relationships between PD and FD to be least likely in ants and most likely in spiders. This expectation was confirmed

for ants, albeit not for spiders. We had also expected that ants might depend least (and ground beetles most) on abiotic factors that constrain individual survival in other taxa, such as salinity or humidity, albeit ants might depend more on factors that constrain the construction of formicaries such as soil depth. This overall expectation was confirmed for ants, but not for carabid beetles. We will below discuss the relationships individually.

Spider TD responded negatively to salinity and drought. Salinity is a known environmental filter for spiders (Desender and Maelfait 1999; Pétillon et al. 2005, 2007;



Traut 2005) hence the salinity-resisting spider species are selected, reducing the local species richness. TD directly decreased with salinity, but PD and FD were only affected through TD. Salinity resistance, a trait not included in our analysis, thus may have appeared randomly in different parts of spider phylogenic tree (see Online Resource 1): sorting for the few salinity tolerant species does not sort for phylogenetically particularly proximate or distant species. Humidity directly impacted only TD and impacted PD and FD only indirectly through TD. This could again be explained by environmental preferences and tolerances (not considered as traits here) if randomly distributed over spider phylogeny. Spider assemblage in the studied environment most likely result from a combination of humidity tolerant and mesohygrophilous species like in other ecotones (Traut 2005), which would explain the positive effect of humidity on spider TD (see also Entling et al. 2007, Weinninger & Fagan 2000 and Lafage et al. 2015; 2019). The positive effect of PD on FD in spiders suggests a phylogenetic signal of traits. This may appear contradictory with the tolerance traits considered above interpreted as being random across spider phylogeny, but these tolerance traits may not be representative of the whole spectrum of traits that characterize spiders. The majority of traits could show phylogenetic signal (Flynn et al. 2011; Khalil et al. 2018; Tucker et al. 2018). For instance, the hunting guild differs among families and hence shows strong phylogenetic signal (Cardoso et al. 2011).

TD and PD of ground beetles responded positively to salinity and negatively to bare ground. For ground beetles, contrary to spiders, salinity has a higher and positive impact on PD than on TD. This would mean that salinity resistance is not restricted to a single or few phylogenetic branches but evolved convergently across distant branches of the phylogeny (consistent with suggestions of Baulechner et al. 2020 for other traits). Again, the absence of a direct effect of salinity on FD but only through PD can be explained by the spectrum of traits used here that may not affect salinity tolerance. The increase of TD with salinity appears inconsistent with existing literature from hypersaline habitats (Desender and Maelfait 1999; Pétillon et al. 2007). The high-cliff heathlands we studied are less saline and salinity might not exclude species, but only permit the additional occurrence of haloresistant species, therefore rising TD. The negative effect of bare ground on PD but not on TD suggests that bare ground selected for one, albeit species-rich phylogenetic lineage. Bare ground is anthropogenic here, and TD of ground beetles not declining under anthropogenic perturbation was previously reported (Verschoor and Krebs 1995), and possibly reflects quick recolonization post-perturbation by eurytopic and highly-dispersive species (Varet et al. 2013) although the species richness of specialist species is often reduced by disturbances for this taxon (see the reviews Niemla and Kotz 2009; Magura et al. 2010).

Ants were the only taxon in which FD was controlled not by PD but by environmental constraints and TD. Presence of a negative, or absence of a positive relationship between PD and FD were expected in particular in ants due to existing competition between many ant species (Cerda et al. 2013). Little relationships of FD to the environment was also expected given capacity of niche construction, albeit strong relationship may emerge if niche construction as such is possible only in some environments, consistent with earlier observations on responses of ant assemblages FD to environmental filters (Campbell et al. 2010; Flynn et al. 2011; Arnan et al. 2015; Fichaux et al. 2019). A positive effect of soil depth on FD is consistent with the fact that most temperate species of ants nest underground (Torossian 1997) and hence profit from deep soils. Moreover, the increase of FD with TD, but not with PD suggests that traits accounted for in FD are not phylogenetically conserved nor convergent in ants. Instead, adding a species seems to, on average, add trait values to the community independent of the selection pressure of environmental constraints. This result appears to be more consistent with a neutral model, with any new species potentially representing a new trait state (Purves and Turnbull 2010; Novack-Gottshall 2016; Stevens and Grimshaw 2020) or superposition of patterns resulting from phylogenetic conservatism and of mutual exclusion of similar related species.

Conclusions

Overall, among the models formulated in the Introduction, results are consistent with phylogenetic conservatism in traits in spiders and ground beetles: PD relates positively to FD (Webb et al. 2002), even when accounting for TD. Our results are not consistent with models of strong character displacement among close relatives that should lead to negative relationships between PD and FD (Prinzing et al. 2008), or reduced niche packing under environmental constraints (Safi et al. 2011). These models invoke strong competitive interactions, stronger than found among grounddwelling predators (see Wise 2006 for spiders and Fichaux et al. 2019 for ants). Overall, characterizing multiple aspects of biological diversity and of the environment, and relating them through structural equation modeling, revealed major differences among ground-dwelling predators in what drives their functional diversity. In particular, the positive relationships between PD and FD found among spiders and ground beetles disappears among ants, possibly resulting from somewhat higher competition among ants. Future studies might explicitly test the relationship between competition and the loss of positive FD/PD relationships using either null



model approaches (e.g. Magura and Lövei 2019) or using our approach of structural equation modeling but applied across a larger number of taxa.

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