

TECHNICAL ADVANCE

Coral bleaching response index: a new tool to standardize and compare susceptibility to thermal bleaching

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Abstract

As coral bleaching events become more frequent and intense, our ability to predict and mitigate future events depends upon our capacity to interpret patterns within previous episodes. Responses to thermal stress vary among coral species; however the diversity of coral assemblages, environmental conditions, assessment protocols, and severity criteria applied in the global effort to document bleaching patterns creates challenges for the development of a systemic metric of taxon-specific response. Here, we describe and validate a novel framework to standardize bleaching response records and estimate their measurement uncertainties. Taxon-specific bleaching and mortality records (2036) of 374 coral taxa (during 1982–2006) at 316 sites were standardized to average percent tissue area affected and a taxon-specific bleaching response index (taxon-BRI) was calculated by averaging taxon-specific response over all sites where a taxon was present. Differential bleaching among corals was widely variable (mean taxon-BRI = $25.06 \pm 18.44\%$, \pm SE). Coral response may differ because holobionts are biologically different (intrinsic factors), they were exposed to different environmental conditions (extrinsic factors), or inconsistencies in reporting (measurement uncertainty). We found that both extrinsic and intrinsic factors have comparable influence within a given site and event (60% and 40% of bleaching response variance of all records explained, respectively). However, when responses of individual taxa are averaged across sites to obtain taxon-BRI, differential response was primarily driven by intrinsic differences among taxa (65% of taxon-BRI variance explained), not conditions across sites (6% explained), nor measurement uncertainty (29% explained). Thus, taxon-BRI is a robust metric of intrinsic susceptibility of coral taxa. Taxon-BRI provides a broadly applicable framework for standardization and error estimation for disparate historical records and collection of novel data, allowing for unprecedented accuracy in parameterization of mechanistic and predictive models and conservation plans.

Keywords: biological response to climate change, coral bleaching, response index, *Symbiodinium*, symbiosis, thermal stress

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Introduction

Reef-building corals depend on mutualistic symbioses with photosynthetic dinoflagellates representing the genus *Symbiodinium* to support their metabolic requirements (Muscatine, 1990). The disruption of these associations (coral bleaching) result in increases in mortality and reductions in resistance to disease, predation, and bioerosion, and reduced capacity for damage repair, competition, growth, and reproduction (Jokiel, 2004; Jones, 2008). As global temperatures increase, corals are experiencing symbiosis-disrupting thermal stress (Hughes *et al.*, 2003; Pandolfi *et al.*, 2003) at increasing

frequencies and intensities (Hoegh-Guldberg *et al.*, 2007; Wilkinson, 2008) and are being eliminated at unsustainable rates (19% of coral reefs have been lost and about 35% are severely threatened; Wilkinson, 2008).

Within a bleaching event, coral susceptibility to stress is highly uneven. Coral colonies, inhabiting the same reef and apparently exposed to identical conditions, will bleach and die at different rates (Marshall & Baird, 2000; Loya *et al.*, 2001; Obura, 2001; Done *et al.*, 2003; McClanahan, 2004; Jones, 2008; van Woesik *et al.*, 2011; Guest *et al.*, 2012). Differential bleaching susceptibility among taxa has been attributed to many factors intrinsic to the holobiont (e.g., thermotolerance of *Symbiodinium*, coral physiological, morphological, and optical characteristics, and the interaction between coral,

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Symbiodinium, and their microbiota; Coles & Jokiel, 1977; Bhagooli & Hidaka, 2003; Baird *et al.*, 2009; Leggat *et al.*, 2011; van Woesik *et al.*, 2011; Cuning & Baker, 2013; Krediet *et al.*, 2013; Marcelino *et al.*, 2013) or extrinsic (*sensu* West & Salm, 2003) from the environment (e.g., site-specific environmental conditions, thermal stress, and frequency of thermal anomalies; McClanahan & Maina, 2003; Guest *et al.*, 2012; Pratchett *et al.*, 2013). Knowing which of many potential factors determine differential bleaching susceptibility among taxa will be essential to understanding bleaching mechanisms, and to preserve, manage, or reconstruct coral assemblages.

Although it is clear that bleaching is not uniform, ranking of responses and mechanisms driving them remain partially obscured. There are more than 40 years of coral bleaching data available; however the majority of available records do not include taxon-level bleaching and mortality response information and marshaling the historical data into a systematic summary is complicated due to data nonuniformity. The data are inherently inconsistent due to the diversity of situations encountered, observation periods relative to episode onset, sampling protocols (e.g., assessing the average affected colony, percent of affected colonies, or proportion of affected coral cover), severity criteria (e.g., data binned into categories ranging from increments of 10% change in color to broad definitions of pale, bleached, or dead), and taxonomic uncertainty. Additionally, coral taxa may appear to be differently affected because responses are intrinsically different among coral holobionts, the stresses that corals face are extrinsically different among events and locations, or response is inconsistently measured across reports. Several indices of taxon-specific differential bleaching and mortality (Gleason, 1993; Marshall & Baird, 2000; Done *et al.*, 2003; McClanahan, 2004; McClanahan *et al.*, 2004; Guest *et al.*, 2012; Pratchett *et al.*, 2013) or coral assemblage susceptibility (Manzello *et al.*, 2007; McClanahan *et al.*, 2007a) have been introduced that focus on specific events, taxon sets, or geographic areas; but a consensus of global data remains elusive.

Here, we propose a new framework to compare taxon-specific bleaching and mortality records across diverse reports to create a taxon-specific bleaching response index (taxon-BRI) and estimate uncertainty in their measurement. This new analysis significantly extends a previous index from 96 (Marcelino *et al.*, 2013) to 374 coral taxa (58 genera, 316 species) using 2036 records from 316 sites across seven biogeographic regions covering the years 1983–2006 (including the pantropical mass bleaching of 1997–1998, and regional events in 2002, GBR, and 2005, Caribbean) (Figs 1 and 2). We created a matrix B_{jk} of standardized taxon-

specific bleaching and mortality records j per site k , allowing us to address the following questions: (i) what is the range of differential bleaching responses among taxa; (ii) does a unified estimate of taxon-specific bleaching response (taxon-BRI or average B_{jk} for same j over different k) accurately reflect differential bleaching as reported by independent sources; (iii) does bleaching response of a coral assemblage (average B_{jk} of the assemblage with different j at site k) reflect the intensity and duration of thermal anomalies; (iv) how much of the variance in bleaching response among coral taxa is due to error of measurement within and inconsistencies across reports; (v) how much of the variance in bleaching response among coral taxa is due to factors extrinsic or intrinsic to the holobiont; and (vi) how much of the bleaching response of a given coral taxon at a site is the result of extrinsic and intrinsic factors?

Materials and methods

Data selection

We searched the historical record of mass bleaching events documented in peer-reviewed literature, scientific reports, unpublished data (collected by A. Baird), and electronic databases of events occurring from 1983 to 2006. We focused exclusively on taxon-specific (genus or species) records that included explicit severity criteria for both bleaching and bleaching-associated mortality. We selected 42 sources containing 2036 records, which applied a diversity of data collection protocols and bleaching severity criteria (Tables S1 and S2). Estimates of both bleaching and mortality are necessary for an accurate assessment, as neither is fully reflective of the damage caused by a bleaching event (McClanahan, 2004; Obura, 2009; Suggett & Smith, 2011). Similarly, time of observation relative to onset of bleaching is highly variable across reports (ranging from 1–3 to 6–12 months; Table S2) and using both bleaching (greater in early observations) and mortality (greater in later observations) will more accurately reflect the extent of physiological damage caused by an event. See Text S1.1 and Tables S1 and S2 for descriptions of reports used and considerations on bleaching-induced and associated mortality.

Construction of a matrix, B_{jk} , of standardized taxon-specific bleaching and mortality data per site

A bleaching response matrix, B_{jk} , was built using 2036 taxon-specific bleaching and mortality records of 374 coral taxa j surveyed at 316 sites k . Reports used different severity criteria, with some using a linear scale proportional to the percent cover affected (e.g., BLAGRRA dataset) while others used weighing coefficients for different levels of response (e.g., McClanahan *et al.*, 2007a; Baird-Palm Island; A. Baird, unpublished data) such that the contribution of the least affected colonies is augmented relative to the most affected (Table S1).

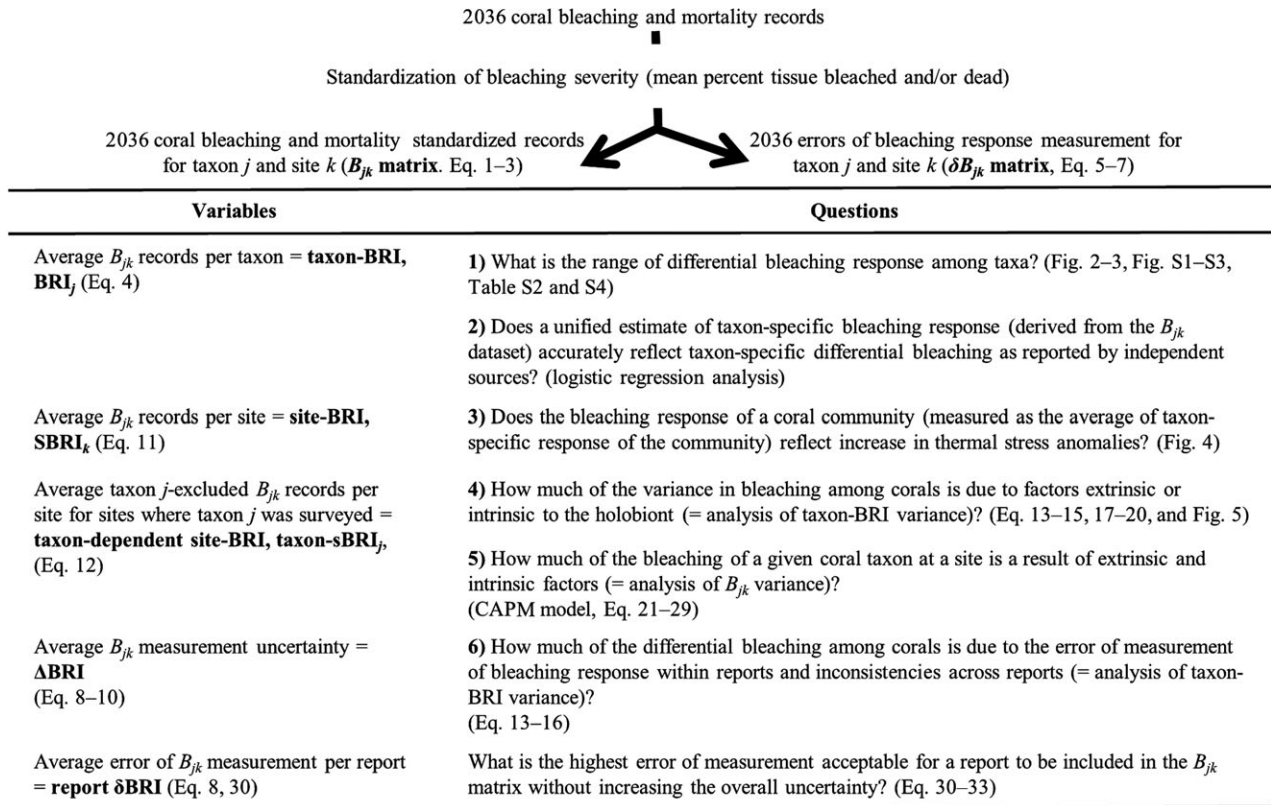


Fig. 1 Standardization process for bleaching and mortality records with a map of the equations and models used in the calculation of variables used in this study. Rationale: differential coral bleaching is due to an unknown combination of intrinsic (coral-*Symbiodinium* dependent) factors, extrinsic (environmental and thermal stress factors, approximated as $SBRI_k$) and uncertainty in measuring bleaching response.

Therefore, all records were standardized to be an estimate of the proportion of affected coral tissue (i.e., mean percent tissue bleached and/or dead). There were three types of reporting methodologies encountered, each requiring a unique conversion algorithm for standardization (see Table S1 for details on individual reports).

Methodology A. For reports that specified numbers of colonies in predetermined bleaching severity categories: let c be the vector of bleaching category limits such that $c_i, i = 1, \dots, M + 1$. For example, the dataset in Floros *et al.* (2004) defined bleaching categories as 0, 1–30%, 31–70%, 71–100% of colonies (Table S1). In our standardization, a dead colony is equivalent to 100% bleached; therefore, if we use these categories as an example, $c = (0, 0.3, 0.7, 1, 1)$. Furthermore, let b be the vector of the mean of bleaching categories: $b_i = (c_i + c_{i+1})/2, i = 1, \dots, M$ and $N_i, i = 1, \dots, M$ the number of taxon-specific coral colonies within each bleaching category (i.e., with bleaching score between c_i and c_{i+1}). Bleaching response B for a given taxon, (fraction of bleached tissue), is then estimated as

$$B = \frac{\sum_{i=1}^M N_i b_i}{\sum_{i=1}^M N_i} = \sum_{i=1}^M w_i b_i \quad (1)$$

where $w_i = N_i / \sum_{i=1}^M N_i$ is the probability (or percent of colonies) in bleaching category i .

Methodology B. An example of this is ReefBase (Table S1), where bleaching response for a given taxon is provided as a single score, which we define as s . This score ranged from 0 to 3: $s = \{0, 1, 2, 3\}$ and is determined by the portion of affected colonies. In this case (see Text S1.2 for derivation), bleaching response can be expressed as

$$B = wb + w_d b_d, \quad (2)$$

where $w = 1/2(w_{s+1} + w_s)$, if $s \geq 1$ or 0 if $s = 0$ is the portion of affected colonies with $\mathbf{w} = \{0, 0.1, 0.3, 1\}$ corresponding to scores $\mathbf{s} = \{0, 1, 2, 3\}$, $w_d = 1/2(w_{d+1} + w_d)$, if $s_d \geq 1$ or 0 if $s_d = 0$ is the portion of dead colonies with $\mathbf{w}_d = \{0, 0.1, 0.3, 1\}$ corresponding to the death scores $\mathbf{s}_d = \{0, 1, 2, 3\}$, and b and b_d are the expectations of the portion of bleached and dead tissue of these affected colonies. For the ReefBase data, $b = b_d = 0.55$.

Methodology C. These reports only provide the final weighted average of scores n , which we denote here as $B_n = \sum_{i=0}^n s_i w_i$, with no information on how the bleaching is

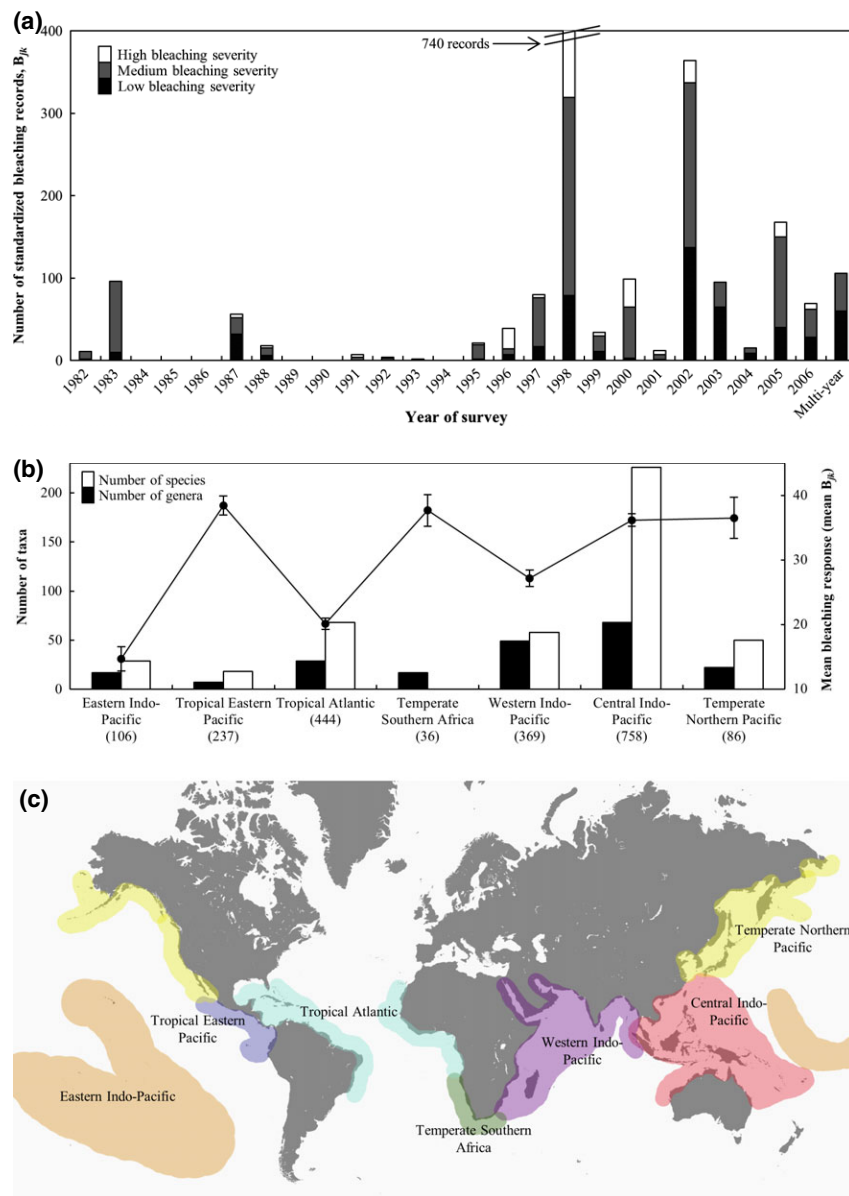


Fig. 2 Years, locations, and responses of the 2036 standardized bleaching and mortality records (B_{jk}) used in this study. Number of records per year categorized as high (≥ 75 th percentile; open bars), medium (grey bars), and low (≤ 25 th percentile; black bars) bleaching severity (a). Number of taxa (genera solid and species open bars) and mean (and standard error; line with error bars) response (b) per biogeographic realm after Spalding *et al.* (2007) (c). Parenthetical values are the total number of records.

distributed over different categories. B_a in this case has to be converted into B (see Text S1.2 for derivation):

$$B \approx \left(\frac{B_a}{n} \right)^2 \frac{b_n}{\alpha}, \quad (3)$$

where constant $\alpha \approx 1$ is found as a fitting parameter in equation $s_i \approx n \sqrt{\alpha \frac{b_i}{b_n}}$ (average $r^2 = 0.98$ s vs. b data fit).

Using these three methods, we estimated B for each taxon at each site, thus forming matrix B_{jk} , where j is the taxon, and k is the site (Table S2).

Construction of the bleaching response index per taxa (taxon-BRI or BRI_j)

After construction of matrix B_{jk} , the bleaching response index per-taxon (taxon-BRI) for species j is found as the average of the bleaching values over all bleaching sites:

$$BRI_j = \frac{1}{K_j} \sum_{k=1}^{K_j} B_{jk} \quad (4)$$

where, as above, the sites are indexed with subscript k , and K_j is the total number of sites with bleaching response for taxon j .

Taxa were reported at the genus- or species-level in the original reports, which is mirrored here. BRI_j values for genera are an average of all B_{jk} records for the genus and its daughter species at each site; however, if small numbers of species-level records are available within a genus, no genus-level BRI_j is calculated (resulting in values for 58 of 90 genera).

Two-approach validation of matrix B_{jk}

Concordance with independent assessments. We compared taxon-BRI (continuous) to the relative bleaching susceptibility of 181 taxa that match this study from the independent assessment of Carpenter *et al.* (2008) (categorized as ‘moderately or highly susceptible to bleaching’ or ‘moderately or highly resistant to bleaching’) through logistic regression.

Correlation between site-BRI and thermal anomalies. Bleaching responses of coral assemblages (measured as individual site-BRI) were compared to intensity of thermal anomaly per site \times bleaching-month (measured as degree heating weeks, DHW) through regression analysis for a subset of data from the wider Caribbean bioregion ($n = 263$ B_{jk} unique-taxon/site \times bleaching-month) records, for 39 taxa, at 35 sites, in years 1998–2006). DHW is a metric of the magnitude and duration of accumulated thermal stress (product of $^{\circ}\text{C}$ above the highest monthly mean sea-surface temperature for a location and its duration in weeks); values above four trigger bleaching and above eight cause mass bleaching and mortality (<http://coralreefwatch.noaa.gov>). Bleaching-month was selected as the 6-month max DHW prior to bleaching observation in order to capture the peak of stress and accommodate differences in observation period, and were compiled from satellite measurements of the U.S. National Oceanic and Atmospheric Administration (NOAA).

Estimation of error of measurements of individual bleaching response records, δB_{jk}

Bleaching response values B_{jk} are known only approximately. For example, if a report indicates a taxon has N colonies within the category of 30–70% bleached cover, the exact value is unknown, resulting in an error of measurement for each element, δB_{jk} , calculated through error propagation analysis.

Methodology A, the greatest possible error for a given report and coral taxon is $\delta b_i = (c_{i+1} - c_i)/2$, $i = 1, \dots, M$ because a colony in category i may have bleaching values with the mean $b_i \pm \delta b_i$. The error of measurement of a bleaching value for site k ($k = 1, \dots, K$ with K being the total number of sites in the database) and taxon j ($j = 1, \dots, J_k$ with J_k being the total number of taxa in site k) is then found using error propagation applied to equation:

$$\delta B_{jk} = \sqrt{\sum_{i=1}^{M_k} (w_{ijk} \delta b_{ijk})^2}. \quad (5)$$

$$\text{For methodology B, } \delta B = \sqrt{b^2(\delta w^2 + \delta w_d^2) + \delta b^2(w^2 + w_d^2)}, \quad (6)$$

where $\delta w = 1/2(w_{s+1} - w_s)$, if $s \geq 1$ or 0 if $s = 0$, $\delta w_d = 1/2(w_{d_{s+1}} - w_{d_s})$, if $s_d \geq 1$ or 0 if $s_d = 0$, and $\delta b_d = \delta b = \frac{0.45}{\sqrt{3}}$. The derivation is presented in Text S1.3.

For *methodology C* (see Text S1.3 for derivation)

$$\delta B \approx 2B_a \frac{b_n}{\alpha n^2}. \quad (7)$$

Uncertainty of taxon-BRI, ΔBRI_j

Since BRI_j for a given taxon j is found as the mean of all B_{jk} over all sites k for which data on taxon j is available, uncertainty of BRI_j can be quantified by two error metrics; the error of measurements of BRI_j , which is estimated using error propagation applied to BRI_j ,

$$\delta BRI_j = \frac{1}{K_j} \sqrt{\sum_{k=1}^{K_j} \delta B_{jk}^2}. \quad (8)$$

and standard error of the mean B_{jk} ,

$$SE [BRI_j] = \frac{1}{\sqrt{K_j}} \text{stdev}_k[B_{jk}], \quad (9)$$

where stdev_k stands for standard deviation of B_{jk} over site index k .

Although related, these two metrics have distinct uses. The error of measurements quantifies accuracy with which BRI_j is calculated as average of B_{jk} over all sites k where taxon j is surveyed. Standard error of the mean indicates how well this average is likely to approximate the true average in the limit of $K_j \rightarrow \infty$. Both types of uncertainty decrease with K_j as $\propto 1/\sqrt{K_j}$. When the number of observations K_j is large, standard error is larger than error of measurements and inherently takes into account variability of observations due to measurement error. However, in certain cases, and especially when the number of observations is small, estimated standard error might be smaller than error of measurements. Therefore, confidence interval (CI) or total uncertainty of BRI_j (ΔB_{jk} , i.e., how well the mean of a finite number of B_{jk} , each having its own error of measurements δB_{jk} , approximates the true mean for bleaching responses that would be observed in the limit of $K_j \rightarrow \infty$ and no measurement error), is calculated as the greater of two error metrics:

$$\Delta BRI_j = \max\{SE [BRI_j]|_{K_j \geq K_{\min}}, \delta BRI_j\}, \quad (10)$$

where standard error is calculated only if the taxon is found in at least K_{\min} sites; if $K_j < K_{\min}$, standard error is not calculated for small K_j due to low accuracy of estimation.

Construction of site-specific bleaching response index (site-BRI or $SBRI_k$)

Reports differentially determined site-specific (coral assemblage-specific) bleaching response by (i) calculating average bleaching response index of taxa at a site in a similar manner to this study (McClanahan *et al.*, 2005), (ii) multiplying taxon-specific bleaching response scores for each taxon by their relative abundance (e.g. McClanahan *et al.*, 2007b, 2015) or (iii) by

choosing the most abundant taxa and determining site-bleaching susceptibility as the average product of susceptibility scores and relative abundances of each taxon (Manzello *et al.*, 2007).

Using matrix B_{jk} , coral assemblage-level response was determined by calculating average taxon-BRI of all taxa j at site k (site-BRI),

$$SBRI_k = \frac{1}{J_k} \sum_{j=1}^{J_k} B_{jk} \quad (11)$$

Construction of taxon-dependent site-BRI (taxon- $sBRI_j$)

Taxon-dependent site-BRI was calculated as the average bleaching response across communities where taxon j was surveyed,

$$\text{taxon} - sBRI_j = \frac{1}{K_j} \sum_{k=1}^{K_j} sB_{jk}, \text{ where } sB_{jk} = \frac{1}{J_k - 1} \sum_{i=1, i \neq j}^{J_k} B_{ik}. \quad (12)$$

Inclusion of the target taxon in sB_{jk} artificially amplifies the correlation between BRI_j and taxon- $sBRI_j$, particularly at sites with small J_k ; therefore, we excluded the target taxon from sB_{jk} . While $SBRI_k$ allows comparisons of bleaching response across coral assemblages at different sites, taxon- $sBRI_j$ allows comparisons of bleaching response across specific coral assemblages that include the target taxon.

Considering intrinsic and extrinsic factors (*sensu* West & Salm, 2003)

Intrinsic factors are defined as biological characteristics of the holobiont that affect bleaching resistance (e.g., *Symbiodinium* thermotolerance; coral morphology, physiology such as early heat shock or oxidative stress response, or skeletal and tissue light-scattering properties that affect internal light amplification; Baird *et al.*, 2009) and were evaluated as a whole (without isolating specific factors). This effect was quantified by analysis of variance of BRI_j (see Materials and methods).

Extrinsic factors are defined as environmental characteristics that affect bleaching resistance of the entire coral assemblage at a site (e.g., currents or turbidity, thermal stress, thermal history; West & Salm, 2003; Guest *et al.*, 2012) and were evaluated as a whole (without isolating individual factors). We reasoned that taxon- $sBRI_j$ (Eqn 12) is a reasonable approximation of the average effect of extrinsic factors at sites containing taxon j (but see section below, Portion of BRI_j variance due to extrinsic factors, for detailed considerations). Text S1.4 examines possible interactions between intrinsic and extrinsic factors.

Effect of uncertainty, extrinsic, and intrinsic factors on differential bleaching among coral (Analysis of BRI_j variance)

We consider differential bleaching among taxa to result from a combination of factors intrinsic (taxon-specific) or extrinsic (environmental) to the holobiont and uncertainty of bleaching response measurement (error of measurement within, and inconsistencies across, reports). The effect of each can be

assessed through analysis of variance of BRI_j for each taxon in the dataset $\{BRI_j\}$:

$$\text{Var} = \text{Var}_j[BRI_j] = (\text{stdev}_j[B_j])^2 = \text{Var}_{\text{int}} + \text{Var}_{\text{ext}} + \text{Var}_{\delta}, \quad (13)$$

where Var_{int} , Var_{ext} , and Var_{δ} are variances due to intrinsic factors, extrinsic factors, and uncertainty, respectively. The portion of the BRI_j variance explained by each of these factors is expressed as:

$$P_{\text{int}} = \frac{\text{Var}_{\text{int}}}{\text{Var}}, P_{\text{ext}} = \frac{\text{Var}_{\text{ext}}}{\text{Var}}, \text{ and } P_{\delta} = \frac{\text{Var}_{\delta}}{\text{Var}}. \quad (14)$$

When the portion of the BRI_j variance related to uncertainty is calculated, the portion of uncertainty-corrected BRI_j variance (portion of BRI_j variance that would be observed in the absence of error of measurements), explained by intrinsic and extrinsic factors can be estimated as:

$$P'_{\text{int}} = \frac{\text{Var}_{\text{int}}}{\text{Var} - \text{Var}_{\delta}}, \text{ and } P'_{\text{ext}} = \frac{\text{Var}_{\text{ext}}}{\text{Var} - \text{Var}_{\delta}}. \quad (15)$$

Each component of BRI_j variance was thus evaluated as discussed below.

Portion of BRI_j variance due to measurement uncertainty. The uncertainty component was determined by the total uncertainty ΔBRI_j , and Var_{δ} can be estimated as

$$\text{Var}_{\delta} = \frac{1}{T} \sum_{j=1}^T \Delta BRI_j^2, \quad (16)$$

where T is the total number of taxa in the dataset with ΔBRI_j given by equation (10).

Portion of BRI_j variance due to extrinsic factors. The extrinsic factors component was determined by performing a correlation analysis between BRI_j and taxon- $sBRI_j$. R^2 -statistic of the correlation between BRI_j and taxon- $sBRI_j$ is the portion of the variance of BRI_j due to extrinsic factors:

$$P_{\text{ext}} \approx R^2[BRI_j, sBRI_j]. \quad (17)$$

Note that this is only valid if sites contain taxa with a wide range of bleaching responses. If, for example, a substantial portion of the sites where taxon j is found has predominantly bleaching-susceptible taxa, this would increase taxon- $sBRI_j$ regardless of the effect of extrinsic factors. In this case, R^2 provides an overestimation of true P_{ext} . This bias may be lowered by restricting the analysis to sites with a large number of taxa (J_k) to include a wide range of bleaching responses. We addressed this by performing correlation analysis as a function of J_k :

$$R^2(n) = R^2[BRI_j, sBRI_j]_{J_k \geq n} \quad (18)$$

As the number of taxa n increases, the number of sites ($N(n)$) that satisfy the criterion $J_k \geq n$ decreases, which may lower the accuracy of correlation analysis. This was assessed by significance of the regression of BRI_j vs. taxon- $sBRI_j$ for $J_k \geq n$. We found that although significance decreased with n , the regression remained significant ($P < 0.01$) for $n < 45$ taxa ($N = 4$ sites, see Results). Thus, the asymptotic behavior of function $R^2(n)$ as n increases can be used as the estimate of P_{ext} :

$$P_{\text{ext}} = \lim_{n \rightarrow \infty} R^2(n) \approx \lim_{n \rightarrow \max_k J_k} R^2(n). \quad (19)$$

As discussed above, for any finite n , $R^2(n)$ is an overestimation of the true P_{ext} and thus can be used as an estimate of the upper bound of P_{ext} .

Portion of BRI_j variance due to intrinsic factors. With the uncertainty and extrinsic portions of the BRI variance estimated, variance due to intrinsic biological differences among taxa can be found using the additive variance property:

$$P_{\text{int}} = 1 - \frac{\text{Var}_{\text{ext}} - \text{Var}_{\delta}}{\text{Var}}. \quad (20)$$

Effects of intrinsic and extrinsic factors on bleaching response of a given taxon at a site (B_{jk})

The effects of intrinsic and extrinsic factors on bleaching response of taxon j at site k , were assessed by performing analysis of variance on elements of matrix B_{jk} . This leverages principles of the capital asset pricing model (CAPM), used in stock market analysis. In CAPM, ‘extrinsic factors’ are equivalent to global market movements and ‘intrinsic factors’ are equivalent to a stock and how it reacts to market changes; parallel to bleaching susceptibility where intrinsic properties define taxon-specific susceptibility to bleaching given the extrinsic environment. Applying CAPM, we express

$$B_{jk} = u_{jk} + \beta_j S_k + \alpha_j, \quad (21)$$

where u_{jk} is variability in estimation of B_{jk} due to error of measurements with $\text{Var}[u_{jk}] = \text{Var}[\delta B_{jk}]$, S_k is the measure of extrinsic factors that influenced all taxa at that site, and coefficients β_j and α_j are taxon-specific, intrinsic parameters. α_j quantifies baseline variations in the absence of anomalous stress for both coral (e.g., variation in tissue thickness or colony size/shape; Loya *et al.*, 2001; Stambler & Dubinsky, 2005) and *Symbiodinium* (e.g., seasonal variation density in the absence of anomalous stress (Fitt *et al.*, 2000; Nir *et al.*, 2014). β_j quantifies bleaching response of taxon j compared to other corals in the assemblage at site k ; all corals would bleach identically if taxa are not intrinsically different and $\beta_j = \text{const}$.

We estimate S_k as taxon- sB_{jk} . As above (see Eqn 12), when the number of taxa per site is large (as is the case with the number of stocks in a market), S_k can be estimated by averaging all taxa for the site, $S_k = (1/J_k) \sum_{i=1}^{J_k} B_{ik}$. Since our data include sites with few taxa, the taxon in question is excluded from the average in order to not confound the relationship between B_{jk} and S_k . Therefore, we estimate S_k as taxon- $sB_{jk} = (1/J_k - 1) \sum_{i=1, i \neq j}^{J_k} B_{ik}$ instead of $(1/J_k) \sum_{i=1}^{J_k} B_{ik}$.

Coefficients β_j and α_j are found by regression of B_{jk} on S_k . Accuracy and reliability of this regression depend on the number taxa per site and number of sites a taxon is present. Accuracy of sB_{jk} and S_k estimation increases with number of taxa per site (J_k). Accuracy of β_j and α_j estimation increases with number of sites reporting a taxon (K_j), which also increases significance of the regression.

Effect of extrinsic and intrinsic factors can be evaluated by two approaches:

Per-taxon diversity analysis: the contribution of extrinsic and intrinsic factors is first determined for each taxon, and then the cumulative contribution for all taxa is determined as the average of all taxa (with each taxon weighted equally). The portion of bleaching response variance explained by both intrinsic and extrinsic factors is:

$$\begin{aligned} P_{\text{ext,int}} &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - S_k \beta_j]}{\text{Var}_k[B_{jk}]} \right] \\ &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - sB_{jk} \beta_j]}{\text{Var}_k[B_{jk}]} \right]. \end{aligned} \quad (22)$$

$P_{\text{ext,int}}$ also quantifies the model (Eqn 21) fit to B_{jk} data. In order to estimate the effect of extrinsic factors alone, we modify the model (Eqn 21) by omitting taxon-specific effects:

$$B_{jk} = u_{jk} + S_k. \quad (23)$$

Consequently, the portion of response variance explained by extrinsic factors only is

$$\begin{aligned} P_{\text{ext}} &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - S_k]}{\text{Var}_k[B_{jk}]} \right] \\ &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - sB_{jk}]}{\text{Var}_k[B_{jk}]} \right]. \end{aligned} \quad (24)$$

The difference between portions of variance explained by model (Eqn 21) and explained by model (Eqn 23) estimates the influence of intrinsic factors:

$$P_{\text{int}} = P_{\text{ext,int}} - P_{\text{ext}}. \quad (25)$$

Finally, we can find uncertainty-corrected portions of the B_{jk} variance explained by the extrinsic and intrinsic factors (i.e., ‘true’ effects that would be observed in the absence of B_{jk} measurement uncertainty):

$$\begin{aligned} P'_{\text{ext,int}} &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - S_k \beta_j]}{\text{Var}_k[B_{jk}] - \text{average}_k[\delta B_{jk}^2]} \right], \\ P'_{\text{ext}} &= \text{average}_j \left[1 - \frac{\text{Var}_k[B_{jk} - S_k]}{\text{Var}_k[B_{jk}] - \text{average}_k[\delta B_{jk}^2]} \right], \\ P'_{\text{int}} &= P'_{\text{ext,int}} - P'_{\text{ext}}. \end{aligned} \quad (26)$$

Per-taxon abundance analysis: the contribution of each factor is calculated for each taxon j in each site (i.e., pair j,k) with the cumulative contribution given by the average over all taxa and sites (with taxa not weighted equally, text S1.5, Eqns 27–29). These two metrics are equivalent if all taxa showed similar abundances, but the per-taxon abundance analysis is predominantly influenced by more abundant taxa.

Evaluation of inclusion of a given report into the meta-analysis based on its uncertainty

Taxon-specific δB_{jk} can be used to evaluate the contribution of each report to overall uncertainty. Reports with large δB_{jk}

could potentially be excluded (after considering number and diversity of taxa, region, and year representation). Each additional report may reduce (in Eqn 8, for δB_{jk} a new report increases the denominator by 1) or increase total variability if the added term δB_{jk}^2 dominates (Text S1.6, Eqn 30). A threshold of report exclusion based on increase of total uncertainty was determined (Text S1.6, Eqns 31–33), but not applied here.

Results and Discussion

Here, we report a new framework for analysis of the sources of differential bleaching among corals which involves (i) standardizing taxon-specific bleaching and mortality records from disparate surveys, (ii) estimating the uncertainty of bleaching response measurement per-taxon due to error of measurement within- and inconsistencies across-surveys and (iii) calculating the effect of the main sources on bleaching variability: varied exposure to distinct environmental factors ('extrinsic'), biological differences among taxa ('intrinsic') and uncertainty of bleaching response measurement ('uncertainty').

What is the range of differential bleaching response among taxa?

Matrix, B_{jk} , of standardized (mean % affected tissue) coral taxon-specific bleaching and mortality responses for each taxon j ($n = 374$) and site k ($n = 316$) was con-

structed from 2036 records from 1983 to 2006 (Fig. 1, Tables S1 and S2). Standardization of records was specific to original data collection methods [See Table S3 as example of standardization of records reported in Obura, 2001 using Eqn (1)]. Records included data on mass bleaching events at pantropical (1997–1998), regional (2002 Great Barrier Reef and 2005 Caribbean), and sub-regional scales in seven biogeographic realms and identify the 1997–1998 event as the most severe in the dataset (Fig. 2). Taxon-specific records are either species- (63%) or genus-level (37%). Of the 90 genera surveyed, only six genera had >75 records: *Acropora* ($n = 320$), *Montipora* ($n = 94$), *Orbicella* ($n = 76$), *Pavona* ($n = 110$), *Pocillopora* ($n = 184$), and *Porites* ($n = 236$); these genera are represented by 53, 23, 3, 10, 8, and 27 species respectively (Figs S1 and S2, Table S2). Sites contain 6.5 ± 13.7 (mean \pm SD) taxa and range from 1 (60 sites) to 199 (1 combined site) taxa (Fig. S3, Table S2). Average B_{jk} over all sites k where taxon j was surveyed yields the taxon-Bleaching Response Index (taxon-BRI or BRI_j , Eqn 4, Table S4). BRI_j is highly variable across the 374 taxa assessed (25.1 ± 18.4 , average and standard error), where some genera are highly resistant (e.g., genera *Madracis*, *Montastraea*, *Symphyllia*; $BRI_j \leq 10\%$) and others highly susceptible (e.g., genera *Millepora*, *Seriatopora*, *Stylophora*; $BRI_j \geq 40\%$), but most genera are highly variable (mean coefficient of variation = 0.89 for 37 genera with >10 records, Fig. 3). Large intra-genus variability is

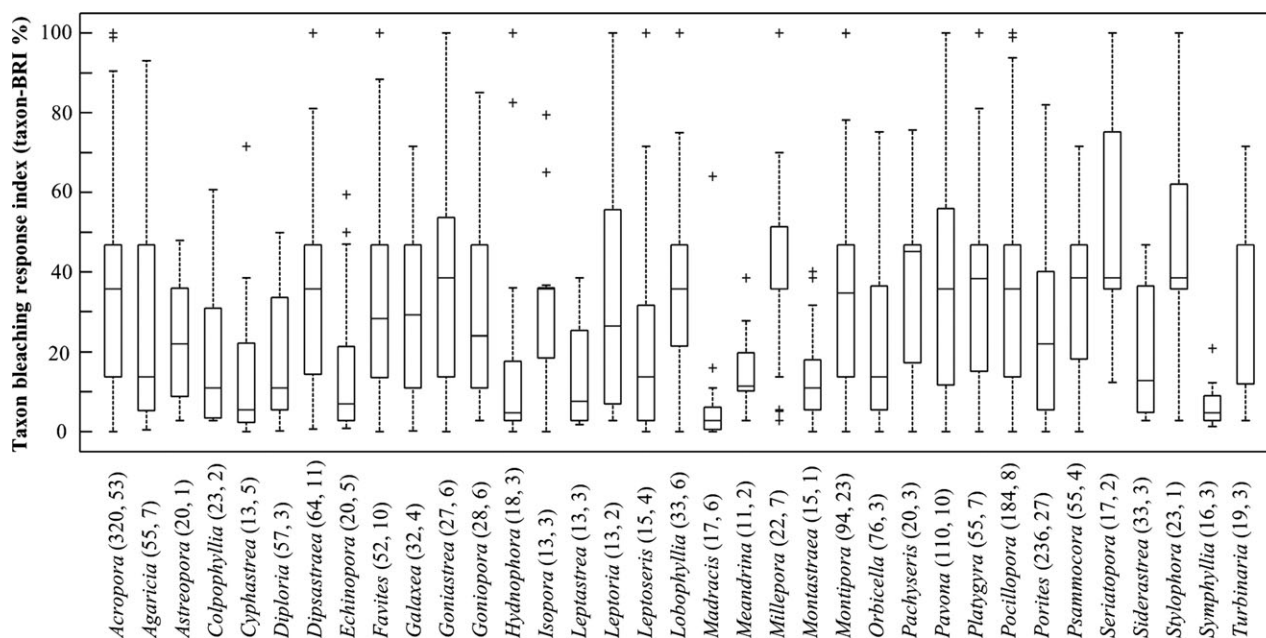


Fig. 3 Taxon-BRI values representing genera with >10 records and showing the median (centerline), 25–75th percentiles (box), range (whiskers), and outliers beyond 1.5 X the interquartile range (+). Only 35 genera are shown as the variance within *Astrea* and *Cycloseris* is minute; 247 species are represented from 1819 records. Parenthetical values are the number of records and species per genus.

observed (within the 37 genera that have >10 records; Fig. 3) where some genera have a uniform bleaching response (e.g., *Astrea*, *Cycloseris*, *Madracis*) while others have large inter-species variability (e.g., *Leptoria*, *Pavona*, *Seriatopora*).

Our results demonstrate substantial variation among individual responses. Such variation is typical (Marshall & Baird, 2000; Baird & Marshall, 2002; Oxenford *et al.*, 2008) and is the material upon which natural selection acts. Our results suggest that an adaptive response via natural selection is to be expected in response to changing climate. Furthermore, as individual sites experience serial bleaching episodes, species may acclimate or adapt (McClanahan & Maina, 2003; Guest *et al.*, 2012; Pratchett *et al.*, 2013; Grottoli *et al.*, 2014; Logan *et al.*, 2014). Repeatedly bleached sites could be targeted for evidence of chronological change, as additional datasets are added to the index.

Does a unified estimate of taxon-BRI accurately reflect taxon-specific differential bleaching as reported by independent sources?

Response indices based on meta-analysis of bleaching records could be invalid because of significant disagreement among source reports due to differential severity of events, assemblage structures, environmental conditions, observation periods, data collection protocols, and severity criteria. If BRI_j is an accurate reflection of the true taxon-specific bleaching response, then independent assessments should report similar patterns. Comparing BRI_j to the conclusions of Carpenter *et al.* (2008) for the 181 species found in both datasets, there is a significant positive correlation (logistic regression, $r^2 = 0.03$, $P < 0.02$) indicating congruence between BRI_j and an independent assessment. Comparing BRI_j to the data of Loya *et al.* (2001) and the follow-up by van Woesik *et al.* (2011), we see many examples of similarly categorized species: High-susceptible *Seriatopora hystrix* (taxon-BRI = 61.46), *Stylophora pistillata* (56.42), *Pocillopora damicornis* (42.29), and *Porites nigrescens* (41.33); medium-susceptible *Dipsastraea favius* (32.05), *Favites pentagona* (27.07), *Favites halicora* (25.66), and *Galaxea fascicularis* (25.43); and low-susceptible *Montipora digitata* (18.52), *Leptastrea transversa* (16.27), *Leptastrea purpurea* (15.73), and *Coelastrea aspera* (14.27).

Does the bleaching response of a coral assemblage (site-BRI) reflect the intensity and duration of thermal anomalies?

The bleaching response of all taxa surveyed in a site was averaged to calculate the coral assemblage bleaching response (site-BRI or $SBRI_k$, Eqn 11, average

33.76 ± 1.33 over 316 sites), which has been used to compare across sites and identify particular extrinsic factors affecting bleaching response (e.g., water flow, thermal history, acute thermal stress, and temperature variability; Manzello *et al.*, 2007; McClanahan *et al.*, 2007b, 2005). If $SBRI_k$ truly reflects assemblage bleaching response at a site, it should account for thermal-stress intensity and thermal-stress history of that assemblage. Comparing $SBRI_k$ of 35 Caribbean sites (limited to sites with >2 taxa per site; 39 unique taxa, mean of 7.27 ± 0.7 taxa per site) with the maximum DHW recorded over 6 months prior to the survey, there is a significant positive correlation (linear regression, $r^2 = 0.21$, $P = 0.006$; Fig. 4) indicating that assemblage response (as measured by $SBRI_k$) is reflective of thermal stress, which is similar to the correlation reported for thermal anomalies and assemblage bleaching responses of 360 Caribbean sites in 2005 (Eakin *et al.*, 2010).

How much of the variance in bleaching response among coral taxa is due to the error of measurement within reports and inconsistencies across reports?

Bleaching response values B_{jk} are estimates of the true bleaching response due to biological and experimental

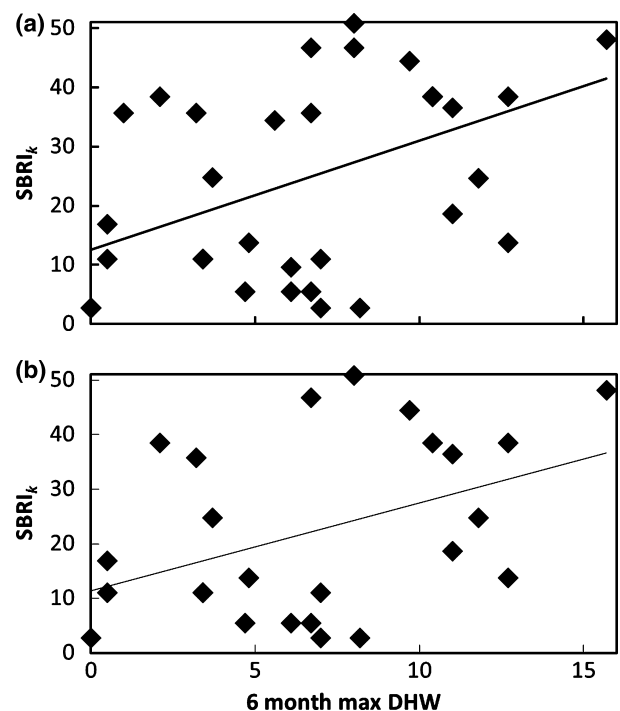


Fig. 4 Regressions of site-specific BRI ($SBRI_k$, Eqn 11) on 6 month-max degree heating weeks (DHW) for Caribbean sites during 1998–2006. Regressions shown using sites with at least three (35 sites; $r^2 = 0.21$, $P = 0.006$) (a) or five (25 sites; $r^2 = 0.17$, $P = 0.038$) (b) taxa per site, which contain a total of 39 or 38 unique taxa and 240 or 209 unique records, respectively.

variability (e.g., the dynamics of bleaching response may be nonlinear, and variation in observation time relative to onset may yield atypical estimates) and error of measurement (e.g., categorical assessments are variously broad and imprecise). Standard error of the mean includes natural variability in bleaching response, inconsistencies across reports, and error of measurement within reports (Eqn 9) when taxa are frequently surveyed, but standard deviation (and therefore standard error) is not defined for taxa surveyed once (47.5% of taxa) and poorly defined for taxa surveyed twice (16.5% of taxa; Table S4). Error of measurement quantifies accuracy with which BRI_j is calculated as average of B_{jk} (Eqn 8), with measurement error of $\pm \delta B_{jk}$. Within-survey variability and error are compounded by the number of surveys used to calculate BRI_j from individual B_{jk} records, so error propagation analysis was used to calculate BRI_j error of measurement (Table S2 for individual B_{jk} and Table S4 for individual BRI_j). Total uncertainty of taxon-BRI (ΔBRI_j) was determined as the greater of two error metrics (standard error and error of measurement, Eqns 8–10, Table S4). The accuracy of determining ΔBRI_j is expected to increase with the number of times a taxon is surveyed at different sites. ΔBRI_j was determined for taxa surveyed in at least three sites ($K_{\min} = 3$) and for taxa surveyed at increasingly higher numbers of sites (K_{\min} was increased by $n + 1$ at each analysis; Eqn 10). The ΔBRI_j decreased from 11.30 ($K_{\min} = 3$), corresponding to 37% of the mean taxon-BRI, to a minimum of 10.44 ($K_{\min} = 8$), corresponding to 35% of the mean taxon-BRI. Furthermore, the portion of variance of BRI_j due to uncertainty P_δ decreased from 34% ($K_{\min} = 3$) to 29% ($K_{\min} = 8$ and higher). These results suggest that the bleaching response of a taxon (BRI_j) is, on average, at least 2.5 times higher than uncertainty of estimating that response and differential bleaching is due, in part, to uncertainty in estimating bleaching response (about 29%).

As more surveys of taxon-specific bleaching and mortality become available, accuracy of determining uncertainty ΔBRI_j is expected to increase. Taxon-specific δB_{jk} can also be used to evaluate how reports add to overall uncertainty of BRI_j against the information provided (e.g., number of new or rare taxa, new bleaching episode or site) so that reports with large δB_{jk} (report average uncertainty, or average- δB_{jk} ; Table S1) could be excluded if the information they provide adds little to the matrix (Eqns 30–33, Text S1.6). The threshold at which uncertainty of a report increases total uncertainty can be determined; reports with average- δB_{jk} more than twice the average of all reports for a taxon increase total

uncertainty and are candidates for exclusion (if minimizing the uncertainty is the goal, Eqns 30–33, Text S1.6).

How much of the variance in bleaching response among coral taxa is due to factors extrinsic or intrinsic to the holobiont?

We considered differential bleaching among corals (BRI_j variance) to result from taxon-specific intrinsic factors, extrinsic factors, and measurement uncertainty. The effect of each of these factors was assessed by analysis of BRI_j variance and estimating the portion of taxon-BRI variance caused by each (Eqns 13–20). Taxon-dependent site-BRI (taxon-sBRI_j, Eqn 12) was used as an approximation of the average effect of extrinsic factors experienced by the sites where taxon j is surveyed. The effect of extrinsic factors P_{ext} was found by correlation between BRI_j and taxon-sBRI_j (Eqn 17). Because we cannot eliminate effects of intrinsic factors on taxon-sBRI_j (e.g., a site may contain mostly bleaching-susceptible coral taxa), P_{ext} is the upper bound of the true variance due to extrinsic factors, and it is expected to increase in accuracy for sites with diverse bleaching responses (i.e., in the limit of a high number of taxa per site n , Eqns 17–19). We estimated the effect of extrinsic factors P_{ext} by correlation between BRI_j and taxon-sBRI_j (Eqn 17) for sites with two (most conservative estimate) and increasingly higher (n up to 44) numbers of taxa. Maximal R^2 (0.33) was observed for sites with two taxa, and it monotonically decreased to 0.057 as the number of taxa per site increased to 44 ($P < 0.01$, Fig. 5 – regression continues to decrease for higher taxa below 0.02, but is no longer significant). Therefore, the effect of extrinsic factors on BRI_j is negligibly small, $P_{\text{ext}} < 6\%$. Further analysis of distinct sites with well-characterized environmental factors, thermal history, and magnitude of coral bleaching responses could provide a more accurate estimate of the effect of extrinsic factors on differential bleaching among corals.

Effect of intrinsic factors (P_{int}), estimated as the remainder of BRI_j variance after uncertainty P_δ (29%) and extrinsic factors P_{ext} (<6%) are calculated (Eqns 13–20), is 65%. Therefore, BRI_j (or taxon-BRI) is a robust measure of innate biological differences among taxa with negligible site-bias and small measurement uncertainty. When averaging coral bleaching responses over several episodes and sites (with diverse environmental conditions and thermal-stress intensities) to determine BRI_j , the influence of extrinsic factors becomes minimal and intrinsic biological properties of the taxa themselves drive differences in bleaching response.

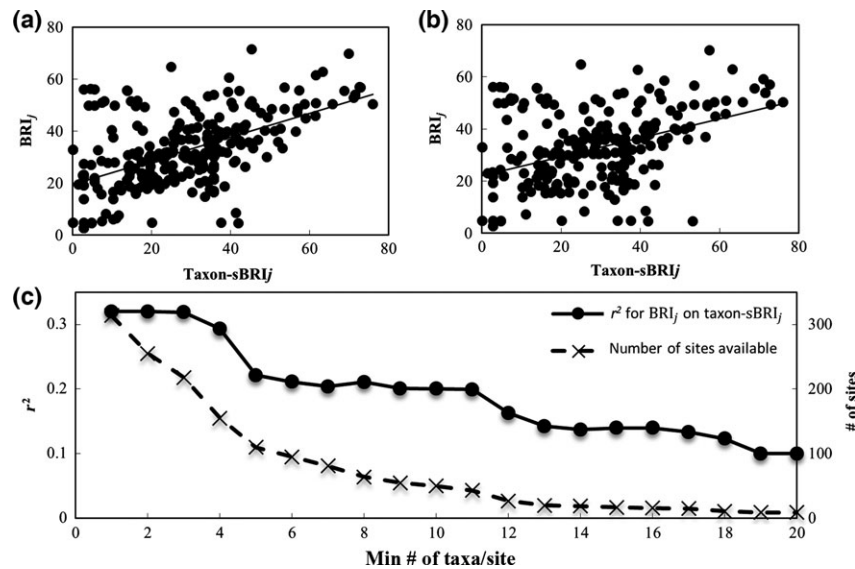


Fig. 5 Relationship between the bleaching response of taxa (taxon-BRI or BRI_j) and their communities (taxon-dependent site-BRI or taxon-s BRI_j , Eqn 12) and its dependence on sample sizes. Regressions of BRI_j on taxon-s BRI_j for all (314) sites ($r^2 = 0.33$, $P < 0.001$) (a) and for sites (43) with ≥ 10 taxa ($r^2 = 0.20$, $P < 0.001$) (b). As the minimum number of taxa per site increases, r^2 for BRI_j on taxon-s BRI_j regressions decrease (solid line), as well as the number of sites available (containing the minimum number of taxa) for analysis (dashed line).

How much of the bleaching response of a given coral taxon at a site is a result of extrinsic and intrinsic factors?

It is conceivable that the influence of extrinsic factors (e.g., thermal anomaly) at a site might be the dominant determinant of taxon-specific bleaching in a specific event (B_{jk}), and simultaneously the global average of bleaching response at multiple events and sites (assessed by BRI_j) might be primarily determined by intrinsic biological properties if the taxon is exposed to a diversity of events that average themselves out and thus minimize the effects of extrinsic conditions. In order to test this hypothesis, we assessed the effects of intrinsic and extrinsic factors on bleaching response of taxon j at site k by performing analysis of variance on elements on matrix B_{jk} using CAPM (where taxon-s B_{jk} is the independent- and B_{jk} the dependent-variable, Eqns 21–29). Accuracy and reliability of this model (Eqn 21) rely upon the number of taxa per site J_k and sites a taxon is present K_j . Considering only taxa and sites with large J_k ($J_k \geq n$) and K_j ($K_j \geq m$) decreases the number of elements available for analysis, therefore a subset of B_{jk} (897 elements) that maximized n ($n = 5$) and m ($m = 9$) while retaining sites (26 sites) was selected. This model (Eqn 21) explained all the uncertainty-corrected variance of B_{jk} ($P'_{ext, in}$), confirming the validity of the approach.

A 'per-taxon diversity' analysis (Eqns 22–26), where the portion of variance due to extrinsic and intrinsic

factors is identified for each taxon before finding the estimated cumulative contribution of these factors by unweighted averaging of those for each taxon, was able to explain 58 (P'_{ext}) and 42 (P'_{int}) % of the uncertainty-corrected variance of B_{jk} . A complimentary 'per-taxon abundance' analysis (Eqns 27–29, Text S1.6), where the effects of extrinsic and intrinsic factors are determined for all taxa at each site before averaging these values for all elements in matrix B_{jk} (such that the contribution of each taxon is not weighted equally), returned similar values ($P'_{ext} = 62\%$, $P'_{int} = 38\%$). If all taxa had the same representation of individuals in the ecosystem (i.e. $K_j = \text{const}$ (j)), the two analyses would be equivalent, but since the distribution is unequal variance of B_{jk} is predominantly affected by the bleaching response of the most common taxa.

These results indicate that both intrinsic and extrinsic factors play a significant role in influencing coral bleaching within a specific bleaching episode at a site (B_{jk}).

Limitations and general considerations

Taxon-BRI is an accurate measure of bleaching response if two conditions are satisfied: nonbleaching-related deaths and the probability of corals recovering before observation can be neglected. Given that most surveys are performed soon after bleaching, both conditions are reasonable and the probabilities of violating them are expected to be small (Obura, 2001; Baird & Marshall, 2002; McClanahan *et al.*, 2004).

Elucidating the range of bleaching responses among different taxa, communities, and geographic regions is essential for understanding physiological variability in bleaching within and among taxa (Obura, 2001; van Woesik *et al.*, 2011). To improve estimates of species susceptibility to bleaching, we need much better estimates of bleaching-induced mortality. For example, a survey conducted proximal to a thermal anomaly might overestimate bleaching and underestimate bleaching-induced mortality (with possible increase in B_{jk}) while a survey conducted distal to a thermal anomaly might be confounded by recovery (decrease in B_{jk}) or mortality that is bleaching-related but not bleaching-induced (increase in B_{jk}) (McClanahan *et al.*, 2004; Obura, 2005, 2009; Jones, 2008; Text S1.1). Because the distinction between these types of mortality was often unclear, we included all surveys, which should increase uncertainty of taxon-BRI. Very few studies measure bleaching at the species-level and even fewer follow the fate of individuals from bleaching to either mortality or recovery. This is not surprising as collecting these data requires multiple surveys over many months (e.g. Baird & Marshall, 2002). Consequently, future monitoring efforts should aim to collect data at the species-level and follow individual colonies through time.

Ambiguities in data collection and reporting may potentially underestimate the error of measurement. While some reporting allowed specific quantification of error, others required postulation of quantitative categories from qualitative descriptions (e.g., 'pale' corals were assumed to be <10% bleached). Additionally, when cumulative statistics are calculated (e.g., taxon-BRI), uncertainty factors not accounted for by δB_{jk} , should increase standard error of the mean, which is why total uncertainty (ΔB_{RI}) is estimated as the greater of the metrics. However, since accuracy of estimation of standard deviation, and therefore standard error, is decreased for rarely surveyed taxa, total uncertainty for these taxa was not accurately estimated (Fig. 2c, Table S4).

The effect of extrinsic factors on taxon-BRI (P_{ext}) can be estimated through taxon-dependent site-BRI, taxon-sBRI_{*j*} and the accuracy P_{ext} depends on whether taxon-sBRI_{*j*} is a good estimate of the severity of environmental stress at a site. This, in turn, is valid only if two conditions are met. First, the site contains high diversity of taxa with a wide range of bleaching responses (rarely the case in practice); to address this limitation, we considered sites containing a large number of taxa (n in Eqns 18 and 19) and evaluated the asymptomatic behavior of P_{ext} at large n . Our results show that P_{ext} is negligibly small indicating that BRI is primarily an intrinsic characteristic. Second, environmental exposure is uniform across the site. While most reports provide

bleaching response for specific locations, a few provided bleaching data aggregated from multiple independent locations and bleaching events. In these cases, it cannot be assumed that all locations were exposed to the same conditions, so we excluded from the initial analysis two reports with large aggregates of multisite/multiepisodic (Done *et al.*, 2003; McClanahan *et al.*, 2009). Five other reports provided data on extended sites covering more than 112 kilometers, (about one degree latitude x longitude, labeled with * in Table S1), but were not excluded from the initial analysis since accuracy of estimating P_{ext} and P_{int} depends on the number of reports. However, we tested the influence of extended-site reports on P_{ext} and P_{int} and found that exclusion of these five reports did not have a significant effect: P_{ext} did not change for per-taxon abundance calculations (62%) (changed from 58% to 56% for the per-taxon diversity) and P_{int} did not change for per-taxon abundance calculations (38%) (changed from 42% to 44% for per-taxon diversity). We therefore conclude that the analysis is robust to report selection.

We tested the effect of regional differences on taxon-BRI variance by excluding the Central Indo-Pacific biogeographic region (758 records, 23% of all surveys) from the analysis. The conclusions of the analysis appeared to be independent of the regional differences among corals: although having more records clearly increases the number of taxa for which the analysis can be performed and improves its accuracy, the main conclusions still hold regardless of the geographical location of the taxa (see Text S1.7 for quantitative analysis of regional differences).

Conclusion

The BRI presented here is the first pantropical assessment of bleaching and mortality, inclusive of measurement uncertainty, that attempts to build a unified comparison of taxon-specific response from the historical record. The index is both expandable to include new records, taxa, and sites, and customizable to target specific locations, events, or times. The quantification of bleaching response provides a tool for assessing traits associated with bleaching, bleaching mechanisms, and management, conservation, or mediation plans. Furthermore, the effect of the uncertainty of bleaching measurement on differential bleaching of corals identified in this study (~29%) suggests that standardization of protocols and reporting would help increase the precision of susceptibility estimates and allow a global repository of standardized bleaching surveys. These standardizations are particularly important as the third pantropical bleaching event, which is expected to affect 38% of all reefs and kill >12 000 km²

of coral (www.globalcoralbleaching.org), is currently unfolding. Estimates of bleaching susceptibility are fundamental to assessing the resilience potential of reef sites, which can then be used to inform management decisions (Maynard *et al.*, 2015). More specifically, accurate estimates of species susceptibility to disturbances, such as bleaching, will allow management to identify and afford protection to susceptible species. In addition, temporal trends of species susceptibility are required to reveal whether or not corals are adapting to climate change (Guest *et al.*, 2012; Pratchett *et al.*, 2013) and this information can assist in deciding whether drastic interventions, such as assisted migration or selective breeding programs are required to conserve susceptible species (van Oppen *et al.*, 2015).

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Supporting Information

Additional Supporting Information may be found in the online version of this article:

Figure S1. Number of standardized bleaching and mortality records (B_{jk}) per genus for genera with <50 records and >50 records (inset) categorized as high (≥ 75 th percentile; open bars), medium (grey bars), and low (≤ 25 th percentile; black bars).

Figure S2. Distribution of species for genera with ≥ 75 records; showing *Pocillopora* (a), *Montipora* (b), *Pavona* (c), *Orbicella* (d), *Acropora* (e), and *Porites* (f).

Figure S3. Frequency distribution of the number of taxa per site.

Table S1. Source information used in the compilation of the 2068 entries used to determine taxon-specific bleaching response index (BRI).

Table S2. List of 2036 standardized bleaching and mortality records B_{jk} for 374 taxa j surveyed at 316 sites k used to determine the taxon-BRI (taxon-specific bleaching response index).

Table S3. Standardization of (Obura, 2001) bleaching and mortality records according to Eqn 1.

Table S4. Taxon-specific BRI values for 374 taxa at 316 sites and associated uncertainties (Eqn 10).