- 1 Upwelling periodically disturbs the ecological assem-
- ² bly of microbial communities in the Laurentian Great
- 3 Lakes
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Abstract

The Laurentian Great Lakes hold 21% of the world's surface freshwater and sup-24 ply drinking water to nearly 40 million people. Here we provide the first evidence that 25 wind-driven upwelling fundamentally restructures microbial communities in Lake Ontario, 26 with its effects extended and redistributed by an internal Kelvin wave propagating along 27 the shoreline for >2 weeks. While thermal stratification is known to organize microbial 28 communities by depth and season, we show that this vertical structure arises from con-29 trasting mechanisms: homogenizing selection in surface waters and dispersal limitation with drift in the hypolimnion. Kelvin wave-driven upwelling disrupts this scaffold, dis-31 placing rare taxa into the surface and creating novel coastal communities enriched in 32 methane oxidation and sulfur metabolism genes—functional traits absent elsewhere in 33 the lake. We observed a Kelvin wave lasting over two weeks and propagating eastward 34 at ~ 60 km day ¹. Given the $\sim 10-12$ day recurrence of wind events, at any time, at least 35 one segment of Lake Ontario's coastline is typically experiencing upwelling, producing 36 pulses frequent and sustained enough to remodel microbial communities on ecologically 37 relevant timescales. Recurrent upwellings, sustained and redistributed by Kelvin waves, 38 act as a biological disturbance that overrides stratification, mobilizes rare functional po-39 tential, and assembles novel coastal microbial communities. As climate change lengthens 40 stratified periods and reshapes large-lake circulation, understanding how physical forc-41 ing governs microbial assembly is essential for forecasting the biogeochemical future of 42 Earth's great lakes, especially in shoreline zones where ecological shifts directly affect 43 human communities.

Data Availability Statement 45

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All raw and processed data for this project are publicly available. The main GitHub 46 repository for this project is available at https://github.com/MarschmiLab/Pendleton_ 47 2025 Ontario Publication Repo, which includes all processed data (Amplicon Sequence Variants (ASVs), CTD casts, water quality data, all other metadata, etc.), the code for all figures, tables, and summary statistics, including the generation of ASVs using the DADA2 workflow, and a reproducible renv environment. All the software versions and citations used for data processing and statistics are described in Table S2. The raw compressed 16S rRNA gene sequencing fastq files are available in the NCBI Sequence Read Archive under the BioProject accession numbers PRJNA1212049. All flow cytometry data are available in the FlowRepository database with the ID number FR-FCM-Z8SJ.

56 Introduction

Microbial communities fuel aquatic ecosystems by transforming energy, cycling nutrients, purifying water, and anchoring aquatic food webs [1–4]. In lakes, microbes recycle organic matter and support higher trophic levels via the microbial loop [5]. Yet in Earth's largest lakes, we still lack a predictive understanding of how microbial communities are structured or how they respond to the powerful physical forces of stratification and circulation that define these systems and are shifting under climate change [6–9].

Classic ecological theory holds that community assembly reflects a balance between deterministic forces like environmental selection and stochastic processes such as drift and dispersal limitation [3, 10]. In hydrodynamically active systems, like Earth's largest lakes and the global oceans, these ecological processes unfold within a dynamic physical environment shaped by ecosystem-scale drivers like stratification, currents, and wind-driven upwelling [6, 11, 12]. Large lakes, upon which hundreds of millions depend, thus offer a tractable but underused model for understanding how physics rewires microbial biogeography and biogeochemical function, with insights that scale to the oceans, where mesoscale circulation shapes the global cycling of carbon, nitrogen, and sulfur [13, 14].

Community assembly mechanisms, including selection, dispersal, diversification, and 72 drift, interact dynamically in response to environmental heterogeneity [3, 4, 15]. Selection 73 may homogenize communities under stable conditions or drive divergence across gradients, 74 while stochasticity introduces probabilistic variation in community structure [10]. These processes shift across spatial and temporal scales, and may respond strongly to physical mixing or isolation. Although microbial communities in smaller lakes and along watershed gradients have been widely studied [16–19], studies of large lakes have more often focused 78 on surface waters or treated them as spatially homogeneous systems [20, 21]. As a result, the influence of vertical structure and hydrodynamic circulation on microbial community 80 assembly in large lakes remains underexplored [22–25]. 81

One powerful but largely overlooked driver of microbial dynamics in large lakes is the wind-upwelling-Kelvin wave cascade. Wind initiates coastal upwelling through Ek-

man transport, tilting the thermocline. When the wind relaxes, this displaced density interface releases its stored energy as baroclinic internal waves. In offshore waters, these 85 manifest as near-inertial oscillations [26]; along the coast, they generate internal Kelvin 86 waves—coastally trapped gravity waves that roll anticlockwise (in the Northern Hemi-87 sphere) within a narrow 3–5 km shoreline band [27–29]. Propagating with the wave is 88 a baroclinic coastal jet: a shore-parallel current, strongest near the surface, that redistributes upwelling and downwelling water along the lake's perimeter, a disruption of the 90 thermocline that can last for weeks. This process repeats in Lake Ontario every ~10–12 91 days [30] and every 5-10 days in Lake Erie [29]. While physical limnologists have long 92 studied these features for their role in transporting nutrients, pollutants, and thermal discharges [11, 27], their ecological consequences remain largely unexamined [31]. 94

Here, we leverage Lake Ontario, a 19,000 km² freshwater basin with well-characterized 95 seasonal circulation, to test how physical and ecological forcings shape microbial com-96 munity structure and function. The lake stratifies annually and experiences wind-driven 97 upwelling events that transport cold, nutrient-rich water to surface zones [6, 32–34], creating transient and chemically distinct habitats. Using an interdisciplinary framework that gq integrates 16S rRNA gene sequencing, flow cytometry, environmental profiling across 100 47 stations and two seasons, and physical limnology, we reveal that Kelvin wave-driven 101 upwelling fundamentally reshapes microbial biogeography, mobilizes rare functional po-102 tential, and challenges existing models of ecological assembly. We show that microbial 103 communities stratify by depth due to homogenizing selection rather than dispersal limita-104 tion. Strikingly, when upwelling disrupts this scaffold, it assembles novel shoreline com-105 munities enriched in rare taxa and distinctive functional traits. Together, these findings 106 reveal how physical forcing structures microbial ecosystems and challenge assumptions of 107 spatial homogeneity in large lakes. 108

$_{109}$ Methods

110 Sample collection

Samples were collected from multiple stations throughout Lake Ontario aboard the R/V111 Lake Guardian in May and September of 2023 as part of the US Environmental Protection 112 Agency's (EPA) Lower Food Web Cooperative Science and Monitoring Initiative (CSMI) 113 survey (Fig. 1A; Table S1). The stations were arranged in 5 north-south transects 114 across the lake, which were sampled in an east-west fashion over May $15^{\rm th}$ - $19^{\rm th}$ and September 23rd-27th. Water samples were collected using a rosette while depth profiles 116 were taken with a CTD meter equipped with fluorometer, dissolved oxygen, spherical 117 photosynthetic active radiation sensor, and turbidity sensors. "Surface" samples were 118 from 5 m below the surface. "Mid" samples were collected at stations >30 m deep, either at the fluorescence maximum or thermocline in May, or at the station's average depth 120 in September. "Bottom" samples were taken 2 m above the sediment. Samples were 121 pre-filtered consecutively through sterilized 200 μ m and 20 μ m Nitex mesh (Wildco) into 122 sterile 10L bottles (Nalgene). Microbial samples were filtered onto a polyethersulfone $0.22~\mu\mathrm{m}$ filter (Millipore Sigma). Filters were flash frozen in liquid N_2 and stored at -80 C. 124

125 Environmental data

Water chemistry and chlorophyll a data were generated by the U.S. EPA's Great Lakes National Program Office according to their standard protocols [35]. See the Supporting Information for details on environmental data, including temperature monitors and meteorological records.

130 Quantifying cells with flow cytometry

Flow cytometry samples were prepared in the field by adding 1 μ L (in May) or 5 μ L (in September) of 25% glutaraldehyde to 1 mL of sample water with a 10 minute incubation. Fixed samples were stained with SYBR I green dye (1X concentration) and cell abundance was measured in DNA-positive cells with a BD Accuri C6 Flow Cytometer in triplicate (Fig. S1). Additional details are provided in the supporting information.

DNA extraction & Illumina sequencing

All DNA extractions were carried out using the Qiagen DNeasy PowerWater kit, mostly following the manufactuer recommendations (more details in Supporting Information). Half a filter was used for May samples and a quarter filter was used for September samples, translating to a minimum of 540mL extracted, with a median of 1320mL extracted. An extraction negative was produced for each kit using a blank filter.

The V4 region of the 16S rRNA gene was amplified using the Earth Microbiome 142 Project protocol [36, 37]. Polymerase chain reaction (PCR) was performed, with blanks, in triplicate 25 μ L reactions using KAPA HiFi 2X MasterMix (Roche), 0.2 μ M 515F 144 (5'-GTGYCAGCMGCCGCGGTAA) and 806R (5'-GGACTACNVGGGTWTCTAAT) 145 primers with Illumina Nextera adapter overhangs, and 5 ng of sample DNA. A Zymo-146 BIOMICS Microbial Community DNA Standard (Zymo) was included to assess the sequencing error rate. Samples and indexing blanks were sequenced using a 2 x 250 148 bp paired-end strategy on an Illumina MiSeq at the Cornell Biotechnology Resources 149 Center, generating 15.72 million reads and a median sequencing depth of 105,500 reads 150 (excluding blanks). 151

152 Microbial bioinformatic processing

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Raw Illumina sequences were processed into amplicon sequence variants (ASVs) using 153 the standard workflow in dada2 package in R [38]. The phyloseq package was used to 154 organize the ASV count table, metadata, and later taxonomic table and phylogeny [39]. 155 Taxonomy was first assigned using the TaxAss freshwater database [40] and then with 156 the Silva 16S rRNA database [41]. Mitochondrial, chloroplast, and ASVs with a higher 157 relative abundance in blanks were removed. In the mock community, three spurious ASVs 158 were detected, representing a sequencing error rate of 0.03%. The resulting minimum 159 number of reads in a sample was 7,931 reads, with a maximum of 73,652 reads, and a 160 total of 7,280 unique ASVs. 161

A phylogenetic tree was constructed using MAFFT for alignment and FastTree

under a generalized time-reversible model [42, 43]. The tree was rooted at the mostrecent-common-ancestor (MRCA) of the Archaea using the ape package [44]. Several polyphyletic, anomalous tips had exceptionally long branch lengths and no taxonomic assignment and were removed by filtering for tips with node heights greater than 2 as calculated using the phytools::nodeheight function [45].

168 Ecological Statistics

Between-sample (beta) diversity was calculated using functions from the vegan or 169 GUniFrac packages (modified to take into account absolute abundances) [46, 47]. Unless 170 otherwise noted, community dissimilarity refers to the Generalized weighted Unifrac dis-171 tance ($\alpha = 0.5$) calculated using absolute abundances. The relative abundance of each 172 ASV in a given sample was normalized by the cell concentration of that sample, providing 173 the absolute abundance of each ASV. Ordinations were constructed using principal coordi-174 nates analysis implemented in phyloseq [48]. Replicates were highly similar to each other 175 (Fig. S2) and were merged by summing reads, resulting in a minimum of 13,986 reads 176 and a median of 82,765 reads. Depth-month groups were defined with UPGMA cluster-177 ing based on community dissimilarity, cutting the tree at 3 groups (Fig. S3A). Variance 178 partitioning and post-hoc tests were performed with Vegan functions varpart, dbrda, 179 and anova.cca [46]. Permutational multivariate analysis of variance (PERMANOVA) 180 and multivariate homogeneity of groups dispersions tests were run with adonis2 and 181 betadisper from the vegan package [46]. 182

Within-sample (alpha) diversity was estimated using the iNEXT package [49, 50].
All rarefaction curves were asymptotic, confirming sufficient sequencing depth for representative sample richness (Fig. S4). Differential abundance of microbial ASVs between depth-month groups was calculated at the Class level using the pairwise setting in the ANCOM-BC2 package [51]. Significant differences had an FDR-corrected p-value less than 0.05 and passed the sensitivity test across multiple pseudocount values; other parameters were kept at default settings. Microbial community assembly processes were quantified with iCAMP [52]. Full details and explanantion of iCAMP are available in

the supplemental. We used a maximum bin size of 24 ASVs, a maximum bin distance of 0.4, and a confidence threshold of 0.975 compared to a randomized null distribution with 1000 iterations [52]. Microbial functional traits were inferred using FAPROTAX [53], using the absolute abundance of each ASV.

Statistical testing, including Two-Sample Wilcoxon tests following Kruskal-Wallis tests; correlation tests; and linear models were done in R, using base functions and the ggpubr package [54]. If multiple comparisons occurred within a given plot panel, all p-values were corrected with a Bonferroni Correction.

199 Spatial Analysis

Mapping and spatial interpolation was performed either in R using the packages sf, terra, and tmap, or in QGIS using GRASS [55–59]. Data files for Lake Ontario's bathymetry were downloaded from the National Oceanic and Atmospheric Administration [60].

204 Results

Upwelling events are frequent drivers of spatial heterogeneity and persist on microbially relevant timescales

We detected upwelling events of cold water along the northern shore in May (stations 207 29 & 43) and the southern shore in September (stations 35 and 38) (Fig. 1B, Fig. S5). In 208 May, the warm southern shore likely reflected a combination of downwelling and relatively 209 warmer water flowing in from the Niagara River (Fig. S6A-C). In September, strong easterly winds in the days leading up to sampling likely drove the pronounced southern 211 upwelling observed at station 38, as reflected in both temperature (Fig. 1B, Fig. S6D) and 212 nutrient data (Fig. S7). While prevailing winds in Lake Ontario are typically westerly, 213 driving upwelling near Toronto, these easterlies reversed the usual pattern, generating 214 upwelling near Rochester instead. We estimate this upwelling formed near the Niagara 215 outlet on September 23rd and propagated eastward as a Kelvin wave along the southern

shore, concluding at the eastern terminus by October 9th (>2 weeks later), at which point winds shifted back from the west-its prevailing direction-causing upwellings along the northern shore (Fig. S6D-E).

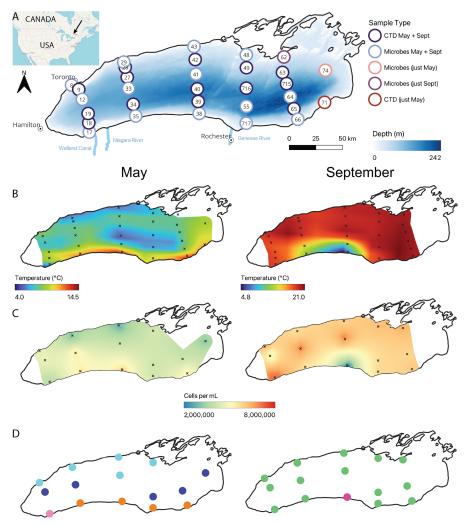


Figure 1. Upwelling reshapes surface microbial abundance and composition. (A) Microbial samples were collected at fifteen stations (light blue) in both months, though station 74 was only sampled in September, and station 62 only in May. All stations with microbial data had associated chemical data collected. Additional stations with only chemical data (dark blue) were collected. Station 71 was only collected in May, and CTD data was unreliable for station 65 in September. (B) Surface (5 m) temperatures in Lake Ontario interpolated spatially across the lake in May and September using multilevel B-splines. (C) Surface cell counts interpolated across May and September using multilevel B-splines. (D) Surface community clusters (shown by color) formed via UPGMA clustering on the community dissimilarity (Fig. S8). Trees were cut at the same height for both months (h = 0.32), resulting in four groups in May and two groups in September.

Upwelling reshapes surface microbial abundance and composition and enriches rare taxa

Spatial patterns of cell abundance varied across the lake surface in both May and
September and correlated to surface temperatures (Fig. 1C). In May, cold, upwelling
areas along the northern shore had lower cell counts while warmer offshore areas and
the southern shore downwelling and Niagara-river impacted regions exhibited higher cell
counts. In September, the upwelling zone at station 38 had the lowest cells abundances,
whereas station 17, influenced by Welland Canal water, had the highest cell abundances.

Surface communities differentiated spatially in response to patterns in surface temperature cause by upwelling (Fig. 1D). Using UPGMA clustering based on community

240 dissimilarity and cutting the dendrogram at the same height (Fig. S8), we defined 4 241 clusters in May and two clusters in September, consistent with higher dispersion in May 242 (Fig. S3C, PERMDISP, p = 0.002). In May, these clusters corresponded to the northern 243 upwelling, an offshore pelagic group, the southern downwelling, and an isolated station 244 near the Welland canal. In contrast, surface communities across September were comparatively homogeneous, besides in the upwelling zone (Fig. 1D, Fig. S3C). The high 246 dissimilarity was driven by an influx of hypolimnion-associated taxa (Fig. S9A), likely 247 due to mass effects, and many rare taxa not observed elsewhere (Fig. S9B, S10A-D), 248 which collectively increased the potential for unique microbial interactions (Fig. S9C). Rare taxa were also abundant in May near the Welland Canal (Fig. S10B) but not along 250 the northern shore upwelling (Fig. S10B). 251

Environmental conditions shaped by stratification, not geographic distance, determine community composition

To address what ecological processes structured the spatial patterns shown in Fig. 1, we explored the relative importance of distance- and depth-decay in Lake Ontario's microbial communities. Community similarity (1 - weighted UniFrac dissimilarity) was not correlated with geographic distance but exhibited a strong relationship with depth in both months, particularly when the thermocline intensified in September (Fig. 2A-

B). Variance partitioning revealed that depth, month, and other environmental factors together explained 73% of the variation in microbial community composition (weighted UniFrac; Fig. 2C), a signal echoed in the PCA of environmental gradients (Fig. S3B). Geographic location, even when combined with environmental variables, accounted for only 3% of the variance.

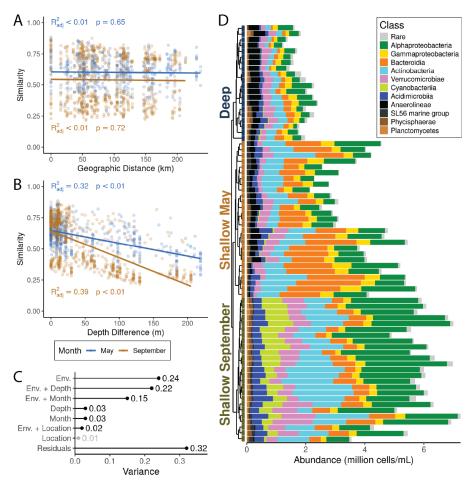


Figure 2. Thermal stratification—not geographic distance—drives microbial community structure in Lake Ontario. Microbial community (A) distance- and (B) depth-decay relationships by month using similarity defined as 1- community dissimilarity. (C) Variance partitioning of the weighted UniFrac dissimilarity. Environmental variables (Env.) corresponded to scaled physical and chemical data from each sample. Variables which were significant (all p < 0.001) when tested using an ANOVA-like permutation test for Constrained Correspondence Analysis post-dbRDA are in black whereas those that were not (i.e., location) are in light gray. (D) Community composition of all samples in absolute abundance, colored at the Class-level. Samples are arranged via hierarchical clustering as shown in Fig. S3A.

As a result of stratification, Lake Ontario bacterial communities were clustered into three distinct groups (Fig. S3A-D). These groups closely hew to sample depth (typically >30 m for deep samples, depending on thermocline depth) and month of collection, and are thus referred to as "depth-month groups." This pattern was consistent across

both hierarchical clustering (Fig. S3A) and principal coordinates analysis (Fig. S3C).

Deep samples from both months grouped together, while shallow samples from May and

September formed distinct clusters (p < 0.001, $R^2 = 0.55$, PERMANOVA). Therefore, we

categorized bacterial communities into three groups for analysis: Shallow May, Shallow

September, and Deep.

Microbial cell abundances across these groups were strongly correlated with temperature (Fig. S11A; Spearman's R =0.92, p < 2.2e-16). Cell abundances were highest in Shallow September, followed by Shallow May, and lowest in Deep samples (Fig. S11B). Unexpectedly, deep September samples had fewer cells than deep May samples (p = 0.031, Two-Sample Wilcoxon).

There were a core set of cosmopolitan taxa shared across all three groups, including
the Alphaproteobacteria LD12, Actinobacteria acI lineages, Gammaproteobacteria LD28
and PnecB, and Verrucomicrobiae LD19 (Fig. 2D, Fig. S12B). Differentially abundant
taxa included an enrichment of Chloroflexi in Deep samples, Bacteroidia in shallow May,
and Cyanobacteria in shallow September (Fig. S12C), including potentially harmful taxa
(Fig. S13). A more thorough discussion of abundant microbial groups can be found in
the Supporting Information.

Season and depth modulate the strength of ecological selection across the lake

We deepened our analysis with community assembly processes across stratification using iCAMP to quantify selection, dispersal, and drift (Fig. 3). Drift dominated Deep samples (52%), with some influence from dispersal limitation (13%) (Fig. 3A). Dispersal limitation was more pronounced when comparing Deep to Shallow samples (12%), increasing to 23% in September during stronger stratification. In Shallow samples, homogenizing selection was greater, especially in September (67%) (Fig. 2A). Across all depth-month groups, homogenizing selection explained 36%–55% of turnover, indicating consistent selection for a core microbial community throughout the lake (Fig. 2B).

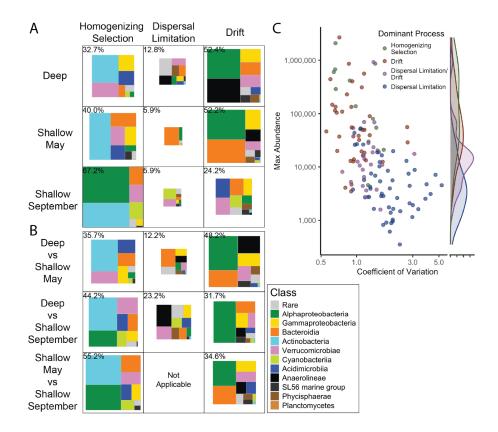


Figure 3. Contrasting forces of selection and drift structure microbial communities across depth. (A-B) Relative contribution of different taxonomic groups to microbial community assembly processes, as determined by iCAMP. The area of each box represents the relative contribution of that process (columns) to turnovers between samples (A) within a given depth-month group, or (B) between two depth-month groups. The percentage contribution of a given process to community turnover is in the top left corner, and boxes are scaled relative to the most influential process (i.e. homogenizing selection in shallow September at 62.7%), and the area represented by each Class is scaled relative to that Class's contribution to that process. (C) The relative importance of community assembly processes within each phylogenetic bin, calculated by iCAMP. Absolute ASV counts were summed per bin per sample while the coefficient of variation was calculated as the standard deviation of abundances across samples by the mean abundance of that bin across all samples. The density plot on the right shows the distribution of each process across the maximum phylogenetic bin abundances. Three bins dominated by heterogeneous selection or a combination of drift and homogenizing selection were excluded.

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Distinct microbial taxa were responsible for the relative importance of each assembly process (Fig 3A-B). Homogenizing selection was primarily driven by the Actinobacteria, while drift was dominated by Alphaproteobacteria. Dispersal limitation between deep 319 and shallow groups was primarily influenced by Bacteroidia in May and Anaerolineae in September.

The contribution of taxonomic groups to community assembly was closely tied to 322 their abundance within samples, particularly for within-group comparisons (compare Fig. 2D and Fig. 3A). More abundant bins had lower abundance variation across samples 324

(Spearman's ranked correlation, $\rho = -0.61$, p < 0.0001) and were primarily shaped by drift and homogenizing selection (Fig. 3C). In contrast, low-abundance bins with high variance were dominated by dispersal limitation, while intermediate-abundance bins were influenced by both drift and dispersal limitation (Fig. 3C).

329 Upwellings are an opportunity for novel microbial metabolic potential

Our analyses revealed that stratification structures microbial communities by modu-330 lating the relative influence of selection and drift. But what happens to microbial biogeo-331 chemical cycling when upwelling disrupts this physical structure? We used FAPROTAX 332 to infer the abundance of genes associated with microbial metabolic functions (Fig. 4). 333 Upwelling zones exhibited a functional signature that blended traits from both the epil-334 imnion (e.g., photoautotrophy; Fig. 4A) and the hypolimnion (e.g., ammonia oxidation; 335 Fig. 4B). However, consistent with the presence of novel microbial taxa and interactions 336 in the September upwelling zone (Fig. S9), we also detected functions that were unique 337 to upwelling, including sulfate respiration (Fig. 4C) and methanotrophy (Fig. 4D). These 338 results suggest that upwelling does not just remix existing microbial functional potential— 339 it creates entirely distinct biogeochemical niches that may alter elemental cycling in the 340 Great Lakes.

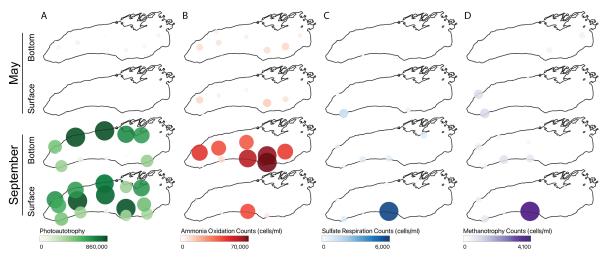


Figure 4. Upwelling creates distinct microbial functional potential. Functional traits were inferred using FAPROTAX [53] using the absolute abundance of each ASV as input, and results are shown for (A) photoautotrophy, (B) aerobic ammonia oxidation, (C) sulfate respiration and (D) methanotrophy. The size and color of each point is scaled by the inferred abundance of each functional trait, in units of predicted cells/ml. Note the range of values varies between functional traits.

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Discussion

Physical forcing and ecological processes interact to govern microbial community structure

Microbial community assembly in Lake Ontario is fundamentally shaped by physical forces that structure the water column. Stratification imposed vertical gradients, with depth and season modulating the balance between deterministic and stochastic processes.

Yet, when stratification was disrupted through upwelling, downwelling, or tributary mixing, community composition was rapidly restructured. These transient disturbances reveal how responsive microbial communities are to ephemeral but ecologically significant niches in large lakes.

Upwelling creates transient but consequential microbial niches

Lake Ontario's ocean-like circulation drives microbial turnover at spatial and temporal 359 scales rarely quantified in freshwater systems. In spring, thermal bars trap nutrients 360 and sculpt sharp near-shore gradients [61]. In summer, typically westerly wind events 361 force Ekman transport, tilt the thermocline, and inject cold, nutrient-rich hypolimnetic 362 water into the surface [27, 29, 32, 62, 63]. As winds relax, energy is released as internal 363 waves: near-inertial oscillations offshore and coastally trapped Kelvin waves [26, 28]. Propagating anticlockwise at ~60 km day ¹ [28], the Kelvin wave–jet complex drags the 365 upwelling signal laterally and, as we show, sustains an easterly microbial disturbance that persisted over two weeks (Fig. S6D-E). This is a timescale relevant for microbial growth, 367 succession, and metabolism. 368

Although physical limnologists have characterized these waves for decades, our data provide the first evidence that they restructure microbial communities, mobilizing rare taxa and their interaction networks and, in turn, reshaping functional potential. Given that upwelling-generating winds occur every ~10–12 days in Lake Ontario [30] and can persist for weeks, some segment of the shoreline is likely experiencing upwelling at any given time during the stratified season. Each event introduces a microbial disturbance

capable of restructuring community composition and ecosystem function. Upwellingfavorable winds have already increased by ~45% over the past three decades [34] and are projected to intensify further [6], underscoring the urgency of linking physical forcing to microbial ecological response.

Despite their frequency, the microbial consequences of upwelling remain under-379 characterised, especially in freshwater systems [64, 65]. During the September event 380 we tracked hypolimnetic lineages (e.g., Anaerolineae, Nitrospiria) surfacing at station 38 381 (Fig. S9A), which we estimate had coupled microbial and thermal signatures persist-382 ing for more than two weeks (Fig. S6D-E). Upwelling sites were uniquely enriched in 383 sulfate-respiration and methanotrophic genes (Fig. 4C–D), revealing that upwelling does 384 more than mix water: it creates novel biogeochemical niches capable of rewiring microbial ecosystem function. As wind regimes intensify under climate change, resolving these 386 physical-microbial linkages will be essential for predicting the function and resilience of 387 Earth's largest lakes. 388

Rare taxa contribute disproportionately to community novelty in upwelling zones

Upwelling and tributary-influenced zones hosted microbial communities distinct from surface or bottom waters, enriched in rare taxa and unique microbial interactions (Fig. S9). In September, low-abundance hypolimnetic microbes dominated at the southern upwelling site. In May, similar enrichment occurred near the Welland Canal but not in the upwelling zone, possibly due to colder temperatures at that station and consequent slower microbial growth rates [66].

These rare taxa are often overlooked, but may serve as a latent reservoir of ecological innovation—a microbial seed bank whose functional potential emerges under transient, physically driven conditions [67]. The distribution of these rare taxa is most likely to be structured by dispersal limitation, so their presence may be both spatially and environmentally constrained. While the long-term consequences remain unquantified, our findings point to a cryptic but consequential layer of microbial responsiveness. This

highlights upwelling as a driver of microbial dynamism with the potential to reshape ecosystem function in the world's largest lakes.

Tributary inputs did not generate persistent nearshore-offshore gradients, in contrast 405 to patterns in Lake Michigan [68]. Instead, strong offshore selection filtered rare taxa, 406 reinforcing a cosmopolitan microbial core structured by drift and selection (Fig. 3). How-407 ever, episodic enrichment near tributary inputs combined with downwelling and thermal 408 bar formation (Fig. 1A, Fig. 4) may temporarily boost microbial diversity and function 409 in nearshore zones. These findings suggest that rare taxa, though ephemeral, may play 410 outsized roles in microbial ecosystem response, revealing an under-appreciated layer of 411 adaptability in Great Lakes microbiomes. 412

Depth and season drive a shift from selection to drift and dispersal limitation

Environmental gradients from stratification and hydrodynamics, rather than geo-414 graphic distance, best explain microbial community structure. This is consistent with 415 metacommunity theory and findings that dispersal is rarely limiting in aquatic systems 416 [4]. Surface communities, especially in late summer, were shaped by homogenizing selec-417 tion, yielding consistent assemblages across sites. In contrast, hypolimnion communities 418 were more stable and governed by drift and dispersal limitation. In May, depth-decay was 419 weaker (Fig. 2B), and surface communities were more spatially variable (Fig. 1D) and 420 similar to those at depth, likely reflecting ongoing niche differentiation after spring mix-421 ing. By September, stratification and low turbulence reinforced surface selection. These 422 dynamics mirror trends in the Pacific Ocean, where surface microbes are selected and dis-423 persal limitation intensifies with depth [69]. Similar patterns occur in oligotrophic lakes, 424 where nutrient scarcity favors efficient, streamlined taxa [16, 18, 70], while productive 425 systems show more stochasticity [71–73]. 426

In the hypolimnion, stratification constrained dispersal (Figs. 2B & 3A–B), and limited mixing fostered site-specific communities that were consistent across space and time (Fig. S3C), despite environmental stability. These cold, oxygenated deep waters harbor slow-growing microbes, with low cell densities (Fig. S11B) [74], limited gene

flow, and dormancy-all weakening selection [75]—explaining the dominance of drift in hypolimnetic bins (Fig. 3C).

While cold, oligotrophic hypolimnia are often thought to be selection-dominated [see Supplemental Discussion: Selection in the Hypolimion; [76]; [24]; [77]; [78]], our results align with recent ocean studies showing that drift increases with depth due to sparse biomass and isolation [69]. These effects were evident in Lake Ontario's hypolimnion, where distance-decay was weak (Fig. S14E–F) but dispersal limitation was consistent. Rather than forming a continuous biogeographic gradient [20], deep microbial communities appeared as isolated parcels shaped by intermittent mixing, resulting in a spatial mosaic of stochastic processes.

441 Microbial abundance and taxonomy link niche breadth to assembly processes

Abundant taxa were governed by selection and drift (Fig. 3C), consistent with the idea that dominant microbes experience stronger ecological filtering [16, 22]. Rare taxa were more dispersal limited, reflecting restricted distributions, though it is unclear if this is a result of narrow niche breadth or a transient introduction from an external source [79]. Synthesizing multi-year datasets could help discern which rare taxa are transiently rare versus permanently rare.

Taxonomic identity also influenced assembly. Actinobacteria, a group characterized by diverse metabolic capabilities and small genomes [80], were primarily associated with homogenizing selection. In contrast, Alphaproteobacteria, dominated by the ubiquitous LD12 clade, were governed by drift. This contrast may reflect differences in ecological versatility. Widespread Actinobacteria may exploit more diverse niches via adaptive radiation [80]. In contrast, the distribution Alphaproteobacteria like LD12 were structured by both selection and drift (Fig. 3). This highlights the importance of drift in structuring microbial distributions, especially for highly abundant taxa.

These patterns underscore that community assembly cannot be explained by abundance alone. Phylogenetic identity, ecological function, and life history traits all shape

how taxa respond to environmental gradients and physical forcing. Yet, our interpre-458 tation is limited by the temporal resolution of our sampling, which captured only two 459 stratified timepoints. Key transitions such as thermocline formation, fall mixing, and 460 inverse stratification remain unresolved [7, 81]. Finer habitat delineations unexplored 461 here (such as shorelines, benthic surfaces, and suspended particles) are also likely to in-462 fluence assembly processes [68, 82, 83]. Finally, while iCAMP offers valuable insights, its 463 inferences reflect model assumptions and should be applied with appropriate caution (see 464 Supplemental Discussion: Biases in iCAMP). 465

466 Upwelling-driven microbial restructuring alters functional potential

Hydrodynamic restructuring of microbial communities carries clear biogeochemical 467 consequences. Upwelling, tributary mixing, and downwelling redistribute microbial taxa 468 and their metabolic potential, creating biogeochemical hot spots that are both spatially 469 and temporally dynamic [32, 84]. Our results suggest that upwelling introduces rare, 470 functionally distinct taxa into the surface waters. These include lineages associated with 471 sulfur cycling, methane oxidation, and ammonia oxidation and may contribute to shifts 472 in biogeochemical processing. These changes unfolded over weeks, as Kelvin wave-driven 473 upwelling swept laterally along the nearshore, suggesting that microbial metabolism re-474 sponds rapidly to these common physical forcings. 475

Functional inference based on metabarcoding, however, has inherent limitations. The
16S rRNA V4 region lacks the resolution to capture fine-scale genomic or functional
differences, and taxa with identical ASVs may differ in metabolic capacity [64, 85]. Additionally, variable gene copy numbers and high dormancy rates in freshwater microbes
can decouple gene presence from metabolic activity [86, 87]. As a result, the accuracy of
functional inference depends on the trait of interest, the environment, and the size and
curation of reference databases [53, 64, 88, 89]. For instance, nitrification rates in Lake
Erie have been shown to be uncoupled from nitrifier abundance [90, 91].

To stay conservative, we used FAPROTAX, a relatively stringent approach; only 24% of ASVs matched at least one annotated function, likely underestimating the true functional

potential of these communities. Future work using metatranscriptomics could directly capture how microbial metabolism responds to upwelling in real time, opening a window into functional shifts that remain invisible [86, 92, 93]. These insights could be powerfully validated by experimental incubations that quantify microbial biogeochemical rates in upwelled waters [94–96], transforming our understanding of how physical disturbance drives ecosystem-scale microbial processes.

492 Microbial biogeography emerges from physical-ecological coupling

Microbial communities in Lake Ontario are not static, but dynamically assembled by the interaction of physical forcing and ecological processes. As northern lakes warm, deepening and lengthening stratification will isolate hypolimnetic communities [97–99] while enhancing surface selection, reshaping microbial biogeochemical cycling in ways we are only beginning to quantify. At the same time, upwelling-favorable winds are becoming more frequent and are predicted to intensify, increasing the prevalence and impact of microbial disturbance zones alongshore.

Our data show that such disturbances are not rare, recurring 2-3 times a month, with their influence persisting for weeks. These pulses override stratification and assemble novel shoreline microbial communities, remodeling ecosystems on meaningful ecological timescales.

By pairing high-resolution microbial, ecological, and physical limnology data, we demonstrate that community structure and function in large lakes are governed not just by who arrives or survives, but by how water moves. These systems offer natural laboratories for testing metacommunity theory, bridging freshwater and marine paradigms, and forecasting how microbial life will shape and respond to a changing climate.

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