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**Marcus W. Beck<sup>1</sup>, John C. Lehrter<sup>1</sup>**

<sup>1</sup>*USEPA National Health and Environmental Effects Research Laboratory*

*Gulf Ecology Division, 1 Sabine Island Drive, Gulf Breeze, FL 32561*

*Phone: 850-934-2480, Fax: 850-934-2401*

*Emails: [beck.marcus@epa.gov](mailto:beck.marcus@epa.gov), [lehrter.john@epa.gov](mailto:lehrter.john@epa.gov)*

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## Abstract

Bio-geo-chemical models are useful tools in environmental sciences that can guide management and policy-making. Significant time and resources are spent developing these models in system-specific contexts. The optimization of model parameters to maximize precision, including transferability of these models to different systems, are fundamental concerns in the development and application of these tools. This study describes quantitative limitations of coupled hydrodynamic-ecological modelling by contrasting numeric and ecological certainty with a systematic framework for characterizing parameter uncertainty. We evaluate a simple bio-geo-chemical model that is the 1-dimensional unit of a larger spatio-temporal model of hypoxia on the Louisiana Shelf of Gulf of Mexico as an example. Results from analysis of the 1D model are used to infer larger trends in dissolved oxygen dynamics over time, having implications for understanding factors that contribute to environmental conditions that are detrimental to aquatic resources. In particular, we focus on issues of parameter identifiability using global sensitivity analyses to provide quantitative descriptions of numerical constraints on model precision. We argue that quantitative and ecological certainty in model calibration are often at odds and the practitioner must explicitly choose model components to optimize given tradeoffs between the two. We further conclude that numerically optimal parameter sets are often small subsets of the complete parameter set for a model because of redundancies in the unique effects of parameter perturbations on model output. As a result, we demonstrate that use of a model for inference into ecological mechanisms of observed or predicted changes in environmental condition can be potentially misguided in the absence of quantitative descriptions of identifiability. Although these concerns have been expressed in the literature, they are rarely explicitly addressed or included in model evaluations. In addition to immediate implications for regional models, we provide a framework for describing the effects of parameter uncertainty and identifiability that can be applied to similar models to better inform environmental management.

## 1 Introduction

1. Simulation/biogeochemical/process-based models overview, contrast with statistical models
2. What models seek to provide - generality, precision, realism [Levins \(1966\)](#), there is a tradeoff so models are 1) developed in partial independence and dependence on the world

and theory, 2) function autonomously from both, or 3) represent both at the same time, from Morrison and Morgan (1999), cited in Ganju et al. (2016). This is similar to the bias-variance tradeoff for statistical models, e.g., overparameterization of a model makes it very biased as it fits the data (the world) exactly, tradeoff between sensitivity and error with changes in model complexity (more complexity is less error but increasing sensitivity) described in Snowling and Kramer (2001)

3. How is model performance/uncertainty evaluated regarding what they should provide - structural, observational, parameter Beck (1987)?

4. Parameter uncertainty as low-hanging fruit - can do post-hoc and from inner to outer level of complexity, parameter uncertainty is the most common, e.g., marine ecological model Mateus and Franz (2015), but stopped short, global sensitivity analysis of eutrophication model Estrada and Diaz (2010)

5. Challenges related to uncertainty - similar to degrees of freedom, identifiability definition from Brun et al. (2001) and need to evaluate identifiability Fasham et al. (2006), Omlin et al. (2001) did a similar analysis with freshwater biogeochem model

6. objectives of study, definition of 'optimum'

(a) What parameters have the greatest influence on dissolved oxygen as identified by local and global sensitivity analyses?

(b) What is the 'parameter space' (from sensitivity analysis) and 'identifiability' of the model parameters?

(c) Can we get a better estimate of identifiability other than 10-15 according to Brun et al. (2001), Soetaert and Petzoldt (2010)?

(d) Can we reconcile differences in numeric uncertainty and ecological uncertainty, e.g., given a finite parameter space, what is an appropriate way to choose parameters given question of interest?

## 2 Methods

### 2.1 Model description

GGEM, FishTank, parameters and parameter categories, how they compare to other models

### 2.2 Local sensitivity analysis

Methods from [Reichert and Vanrolleghem \(2001\)](#), [Soetaert and Petzoldt \(2010\)](#)

Sensitivity of the model to parameter changes is evaluated using the difference in model results before and after perturbing each parameter value. Each parameter is perturbed by the same value as a percentage of the whole. The default value (as  $1 + 1e-8$  proportion, default value for ‘tiny’ argument in ‘sensFun’). A sensitivity value  $S$  is estimated for each time step  $i$  given a set value for parameter  $j$  as:

$$S_{ij} = \frac{\partial y_i}{\partial \Theta_j} \cdot \frac{w_{\Theta_j}}{w_{y_i}} \quad (1)$$

where the estimate depends on the change in the predicted value for response variable  $y$  divided by the change in the parameter  $\Theta_j$  multiplied by the quotient of scaling factors  $w$  for each. The scaling factors,  $w_{\Theta_j}$  for the parameter  $\Theta_j$  and  $w_{y_i}$  for response variable  $y$ , are set as the default value of the unperturbed parameter and the predicted value of  $y_i$  after perturbation ([Soetaert and Petzoldt 2010](#)). The scaling ensures the estimates are unitless such that the relative magnitudes provide a comparison for model sensitivity to parameter changes that may vary in scale. The FME package summarizes sensitivity as  $L1$  and  $L2$  across the time series:

$$L1 = \sum |S_{ij}|/n \quad (2)$$

$$L2 = \sqrt{\sum (S_{ij}^2)}/n \quad (3)$$

The mean, minimum, and maximum  $S_{ij}$  values for each parameter are also reported. In general, positive mean sensitivity estimates generally indicate that a parameter has a positive effect on the model results for a given increase in the parameter. However, the effect can change

over time so a plot of the 'sensitivity function' should be viewed which shows the difference from the values of the response variable before and after changing a parameter value. Note that the perturbation factor in 'sensFun' depends on the default value of each parameter such perturbations are not consistent for parameter values less than or greater than the perturbation factor. This can produce results that are inconsistent for different levels of perturbation. A custom function was keeps the perturbation factors constant regardless of the magnitude of the default values. Sensitivity for each parameter using the custom function are estimated using the above equations from the FME package. All parameters that are considered 'sensitive' should be further evaluated by plotting the predictions before and after perturbation and across the total range of the parameter.

Each parameter was perturbed by 50% using 'sensfun' to identify model sensitivity. Parameters that produced different model results from the default values are shown below, by category.

The parameters that have the greatest effect on the model by category (optics, organics, phytoplankton, zooplankton) are as follows.

Plotting the raw values from the sensitivity analysis provides a visual assessment of changes.

## 2.3 Global sensitivity and identifiability

Identifiability describes the ability to estimate a parameter in relation to variation among the remaining parameters. A parameter is identifiable if all parameters within the set can be uniquely estimated based on the observed data. Parameters that are unidentifiable typically produce similar model outputs for a given relative perturbation, i.e., the effect of altering one parameter can be undone by altering one or more other parameters. Model calibration will not converge for parameters sets that are unidentifiable. Identifiability is estimated from the minimum eigenvector of the cross-product of a model's sensitivity matrix (Brun et al. 2001, Omlin et al. 2001):

$$\gamma = \frac{1}{\sqrt{\min \left( \text{EV}[\hat{S}^T \hat{S}] \right)}} \quad (4)$$

where  $\gamma$  ranges from one to infinity for perfectly identifiable (orthogonal) or unidentifiable

(perfectly collinear) results for a set of parameters in the sensitivity matrix  $S$ . Identifiability can be estimated for any combination of parameters for a model. Values less than 10-15 indicate parameters are generally identifiable. The FME package provides the 'collin' function to estimate identifiability of output from the 'sensFun'. For reasons described above, a custom function was created to determine identifiability from model residuals, where the residuals were based on a relative perturbation of the model parameters. A comparison of results of identifiability using the custom functions and those from the FME package were generally in agreement. Standard output from the identifiability function is shown below, where identifiability is displayed as a matrix with increasing combinations of subsets of the parameters.

The identifiability functions evaluate the ability to identify all subsets from pairwise to all parameters in the subset. As an example, the identifiability of pairwise combinations for each of the categories are shown below.

Pairwise identifiability of the top two most sensitive parameters in each category is shown, first as a table showing all unique combinations from two to all parameters and second as a figure showing pairwise combinations.

## **2.4 Identifying subsets and testing calibration**

Model calibration must consider the competing objectives of parameter inclusion and identifiability. Estimating parameters for novel datasets increases model complexity and the ability to identify parameters decreases with the inclusion of parameters to estimate. The above analyses indicated that estimating all of the parameters was impossible because of redundancies in the the model output. Calibration to novel datasets should balance the competing objectives of identifiability while including parameters for which model output is most sensitive. The plots below show different scenarios of parameter inclusion for model calibration. Reading each plot from left to right can be interpreted as including additional parameters, where each parameter is ranked by relative sensitivity. The inset in each plot shows the identifiability of including parameters, up to a maximum where additional inclusion exceeds the identifiability threshold of fifteen. The scenarios for including parameters all begin with the parameters that have the greatest effect on the model output. The inclusion of additional parameters depends on the scenario. The first scenario selected parameters by decreasing sensitivity within each category (i.e., four separate models calibrated for optics, organics, phytoplankton, or zooplankton), the second

scenario selected parameters by sensitivity regardless of category, and the third scenario selects parameters by sensitivity with equal representation between categories. Boxes in each plot represent a unique parameter and the numbers in each box represent the relative rank of a parameter's sensitivity to the model output within each category.

The following plots show identifiability with the addition of parameters.

### **3 Results**

### **4 Discussion**

Questions specific to GOM - what initial conditions are important? How many phytoplankton groups do we need (e.g., related to structural uncertainty)?

How does the assimilation of additional parameters (e.g., other state variables) during calibration influence the conclusions?

How does uncertainty translate to what a model should provide (generality v precision)?  
The first step - find out what can be optimized but then do not overfit....

What about structural uncertainty - does sensitivity of a model to variation in a parameter imply parameter uncertainty and/or structural uncertainty?

A final point about optimization with identifiable parameter sets - optimization to fit the data still does not ensure a correct model. Failing in one way can be over-compensated by another feature, e.g., the parameter set that is optimized (see [Flynn \(2005\)](#), p. 1207, third paragraph)

[Omlin et al. \(2001\)](#) state that the sensitivity, identifiability, estimation process is iterative (p. 113), need to rinse and repeat for proper calibration.

How to improve identifiability - get more/better observed data, include obs from other state variables in RSS minimization (eqn q in [Omlin et al. \(2001\)](#))

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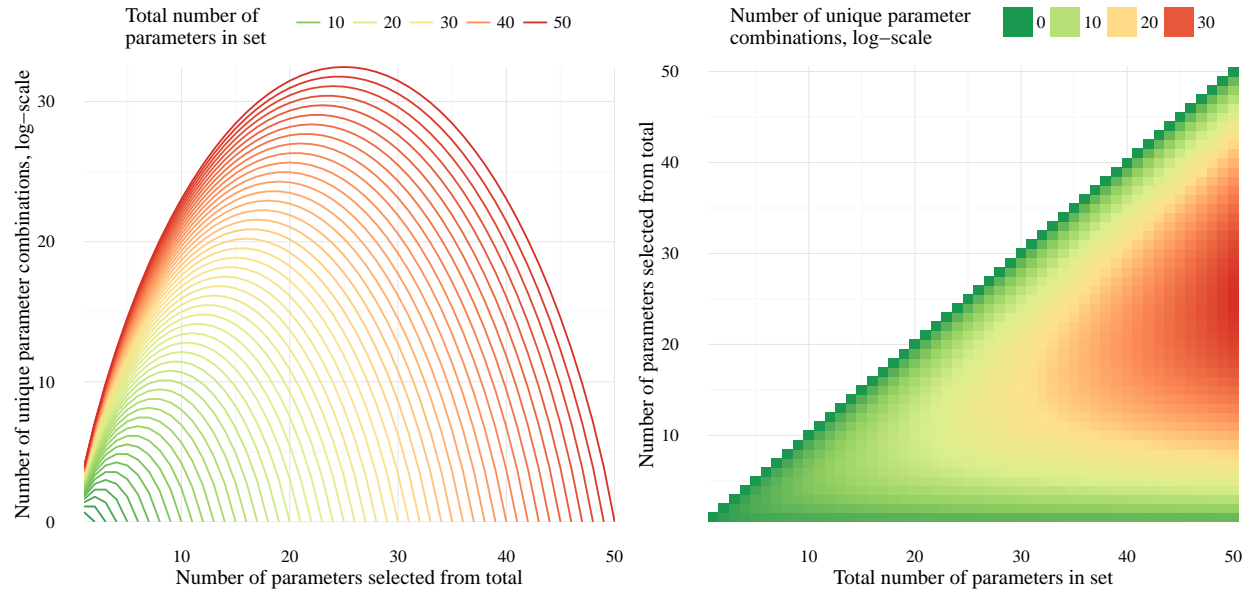


Fig. 1: Number of unique combinations that can occur by selecting parameters for optimization from different parameter sets. The number of combinations are shown for increasing numbers of selected parameters from the total in the set, where 100 parameter sets are shown each with one through 100 total parameters. For example, there are six unique parameter combinations that result from choosing two parameters from a set. Note that the number of unique combinations is shown as the natural-log. The same information is viewed differently in each plot.