

## **parent\_report.Rmd**

### **Introduction**

#### **About This Report**

This report is for the New England Fishery Management Council (NEFMC). The purpose of this report is to synthesize ecosystem information to allow the NEFMC to better meet fishery management objectives. The major messages of the report are synthesized on pages 1-3, with highlights of 2024 ecosystem events on page 4.

The information in this report is organized into two main sections; **performance measured against ecosystem-level management objectives** (Table 1), and potential **risks to meeting fishery management objectives** (Table 2: climate change and other ocean uses). A final section highlights **notable 2024 ecosystem observations**.

#### **Report structure**

A glossary of terms<sup>1</sup>, detailed technical methods documentation<sup>2</sup>, indicator data<sup>3</sup>, and detailed indicator descriptions<sup>4</sup> are available online. We recommend new readers first review the details of standard figure formatting (Fig. 54a), categorization of fish and invertebrate species into feeding guilds (Table 7), and definitions of ecological production units (EPUs, including the Gulf of Maine (GOM) and Georges Bank (GB); Fig. 54b) provided at the end of the document.

The two main sections contain subsections for each management objective or potential risk. Within each subsection, we first review observed trends for indicators representing each objective or risk, including the status of the most recent data year relative to a threshold (if available) or relative to the long-term average. Second, we identify potential drivers of observed trends, and synthesize results of indicators related to those drivers to outline potential implications for management. For example, if there are multiple drivers related to an indicator trend, do indicators associated with the drivers have similar trends, and can any drivers be affected by management action(s)? We emphasize that these implications are intended to represent testable hypotheses at present, rather than “answers,” because the science behind these indicators and syntheses continues to develop.

Table 1: Ecosystem-scale fishery management objectives in New England

Objective categories	Indicators reported
<b>Objectives: Provisioning and Cultural Services</b>	
Seafood Production	Landings; commercial total and by feeding guild; recreational harvest
Commercial Profits	Revenue decomposed to price and volume
Recreational Opportunities	Angler trips; recreational fleet diversity
Stability	Diversity indices (fishery and ecosystem)
Social & Cultural	Community fishing engagement and social vulnerability status
Protected Species	Bycatch; population (adult and juvenile) numbers; mortalities
<b>Potential Drivers: Supporting and Regulating Services</b>	
Management	Stock status; catch compared with catch limits
Biomass	Biomass or abundance by feeding guild from surveys
Environment	Climate and ecosystem risk indicators listed in Table 2

<sup>1</sup><https://noaa-edab.github.io/tech-doc/glossary.html>

<sup>2</sup><https://noaa-edab.github.io/tech-doc/>

<sup>3</sup><https://noaa-edab.github.io/ecodata/>

<sup>4</sup><https://noaa-edab.github.io/catalog/index.html>



# State of the Ecosystem 2025: New England

March 24, 2025

Table 2: Risks to meeting fishery management objectives in New England

Risk categories	Observation indicators reported	Potential driver indicators reported
<b>Climate and Ecosystem Risks</b>		
Risks to Managing Spatially	Managed species (fish and cetacean) distribution shifts	Benthic and pelagic forage distribution; ocean temperature, changes in currents and cold pool
Risks to Managing Seasonally	Managed species spawning and migration timing changes	Habitat timing: Length of ocean summer, cold pool seasonal persistence
Risks to Setting Catch Limits	Managed species body condition and recruitment changes	Benthic and pelagic forage quality & abundance: ocean temperature & acidification
<b>Other Ocean Uses Risks</b>		
Offshore Wind Risks	Fishery revenue and landings from wind lease areas by species and port	Wind development speed; Protected species presence and hotspots

## Performance Relative to Fishery Management Objectives

In this section, we examine indicators related to broad, ecosystem-level fishery management objectives. We also provide hypotheses on the implications of these trends—why we are seeing them, what's driving them, and potential or observed regime shifts or changes in ecosystem structure. Identifying multiple drivers, regime shifts, and potential changes to ecosystem structure, as well as identifying the most vulnerable resources, can help managers determine whether anything needs to be done differently to meet objectives and how to prioritize upcoming issues/risks.

### Seafood Production

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##### Indicators: Landings; commercial and recreational

This year, we present updated indicators for total [commercial landings](#), U.S. seafood landings (includes seafood, bait, and industrial landings), and Council-managed U.S. seafood landings through 2023. There are long-term declines in all New England landings time series except for total commercial landings on GB (Fig. 1). There exist long-term declines in commercial seafood landings and NEFMC managed seafood landings for both the GOM and GB, but over the last decade there is no trend in managed seafood landings in the GOM.

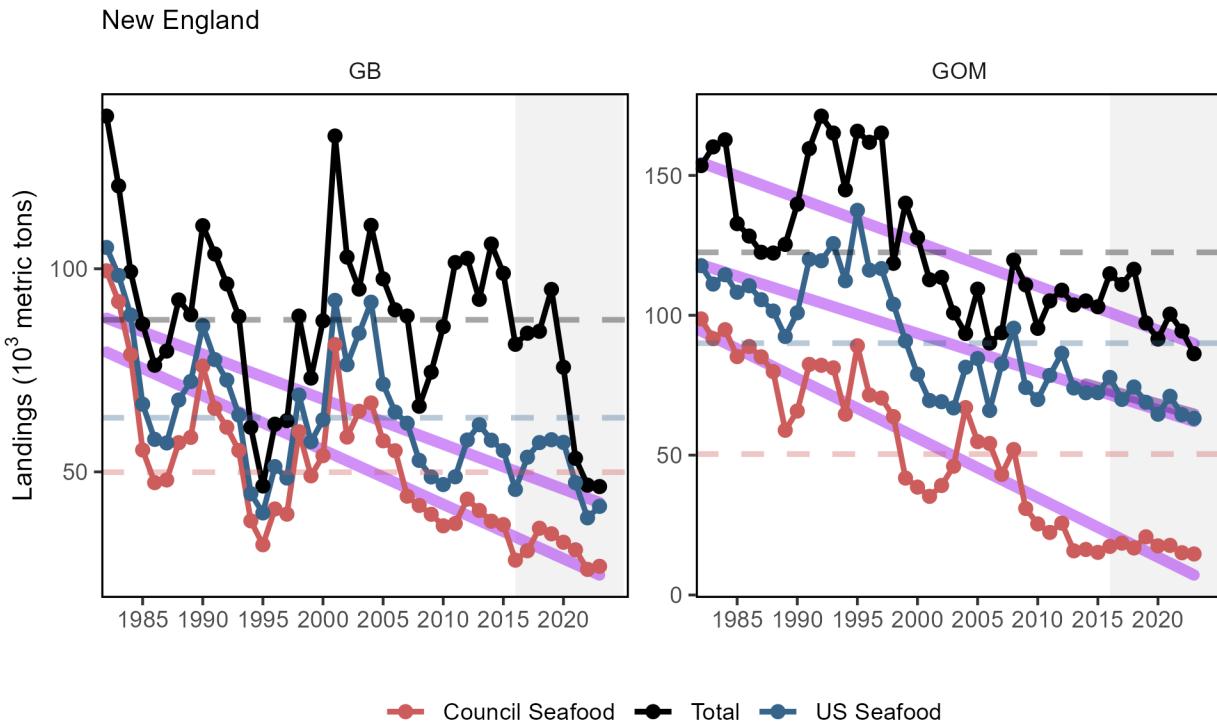


Figure 1: Total commercial landings (black), total U.S. seafood landings (blue), and New England managed U.S. seafood landings (red) for Georges Bank (GB) and the Gulf of Maine (GOM).

Commercial landings by guild include all species and all uses, and are reported as total for the guild and the NEFMC managed species within the [guild](#). As reported in previous years, downward trends persist for a number of guilds in both regions. Current high total landings for benthivores (GOM) are attributable to American lobster, and a significant long term increase in benthos landings (GB) is attributable to clams and scallops (Fig. 2). Current landings of planktivores are still below the long term mean.

[Aquaculture production](#) is not yet included in total seafood landings.

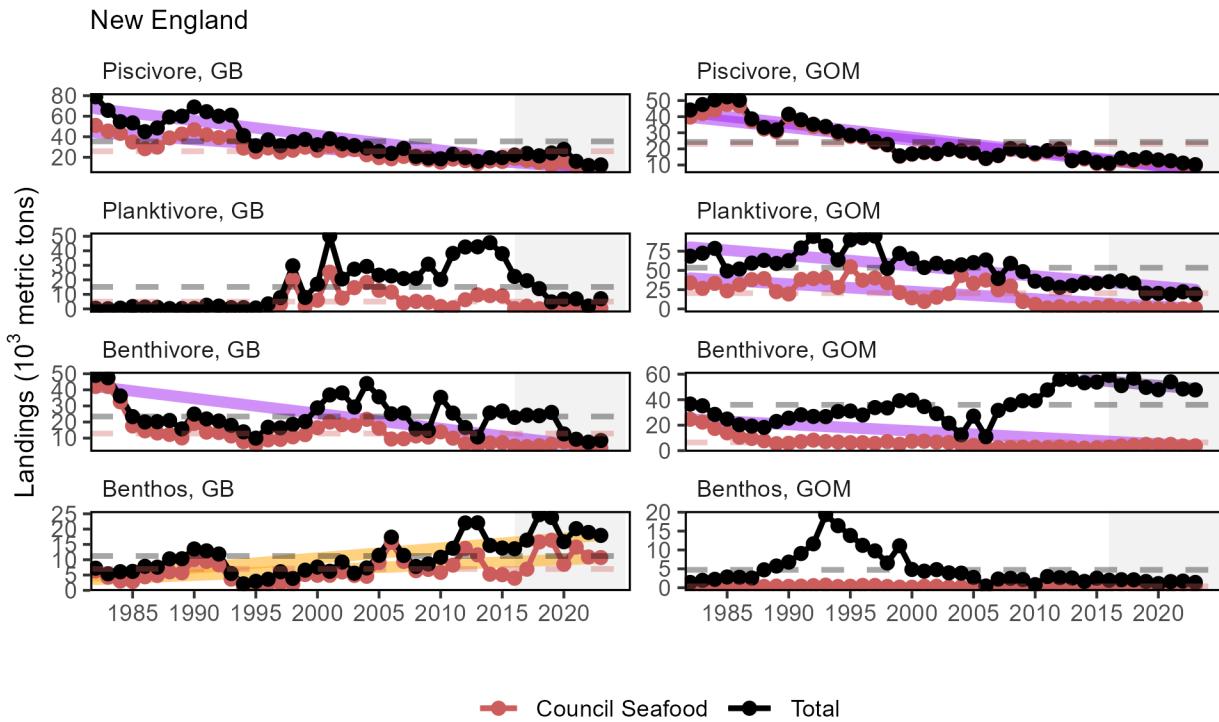


Figure 2: Total commercial landings (black) and NEFMC managed U.S. seafood landings (red) by feeding guild for the Gulf of Maine (GOM, right) and Georges Bank (GB, left).

**Total Community Climate Change Risk** is a measure of to what degree a region's landings (or revenue) is dependent on sensitivity and exposure factors that relate to species' risk to temperature or ocean acidification changes as the result of future climate change. For New England, the total climate vulnerability of landings (Fig. 3) was moderate in 2022 with no long-term trend suggesting a moderate reliance on climate-sensitive species. This proportion has not significantly changed since 2000.

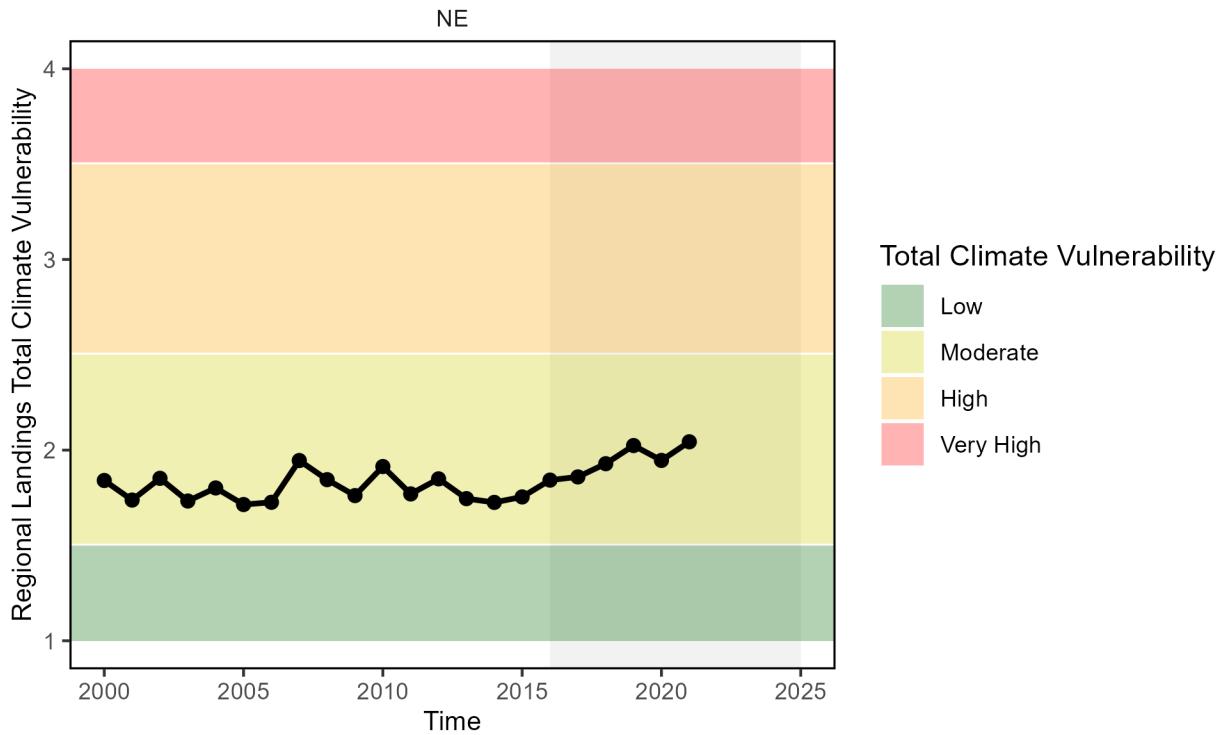


Figure 3: Total climate vulnerability on New England landings from 2000 to 2022. Horizontal colored bars show different climate risk levels.

Overall, [recreational harvest](#) (retained fish presumed to be eaten) has declined in New England (Fig. 4). However, recent harvest has remained above the historical low level in 2020. Recreational [shark landings](#) of pelagic and prohibited sharks have declined since 2018 (Fig ??), which is likely influenced by regulatory changes implemented in 2018 intended to rebuild shortfin mako stocks and comply with binding recommendations by the International Commission for the Conservation of Atlantic Tunas (ICCAT).

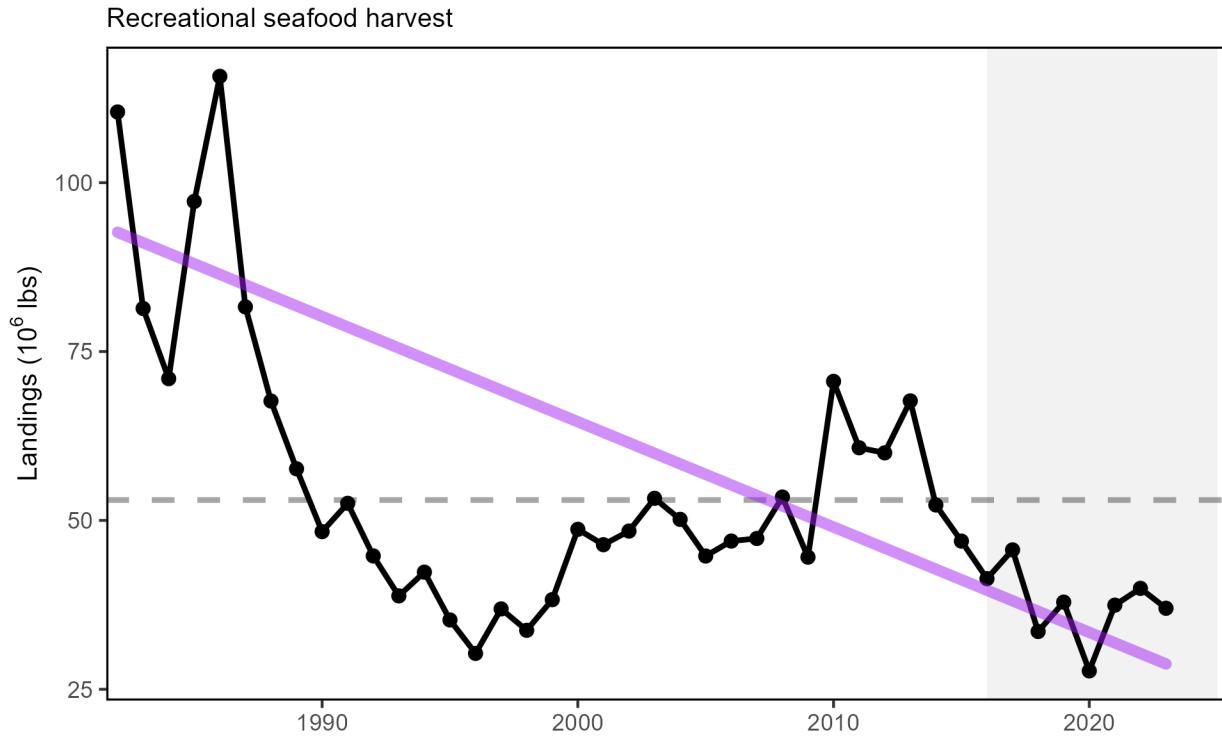


Figure 4: Total recreational seafood harvest (millions of pounds) in the New England region.

## Implications

Declining commercial (total and seafood) landings and recreational harvest can be driven by many interacting factors, including combinations of ecosystem and stock production, management actions, market conditions, and environmental change. While we cannot evaluate all possible drivers at present, here we evaluate the extent to which stock status, management, and system biomass trends may play a role.

**Stock Status** Single species [management objectives](#) (1. maintaining biomass above minimum thresholds and 2. maintaining fishing mortality below overfishing limits) are not being met for some NEFMC managed species. Thirteen stocks are currently estimated to be below  $B_{MSY}$  (Fig. 5), while status relative to  $B_{MSY}$  could not be assessed for 13 additional stocks (Table 3). Therefore, stock status and associated management constraints are likely contributing to decreased landings. To better address the role of management in future reports, we could examine how the total allowable catch (TAC) and the percentage of the TAC taken for each species has changed through time.

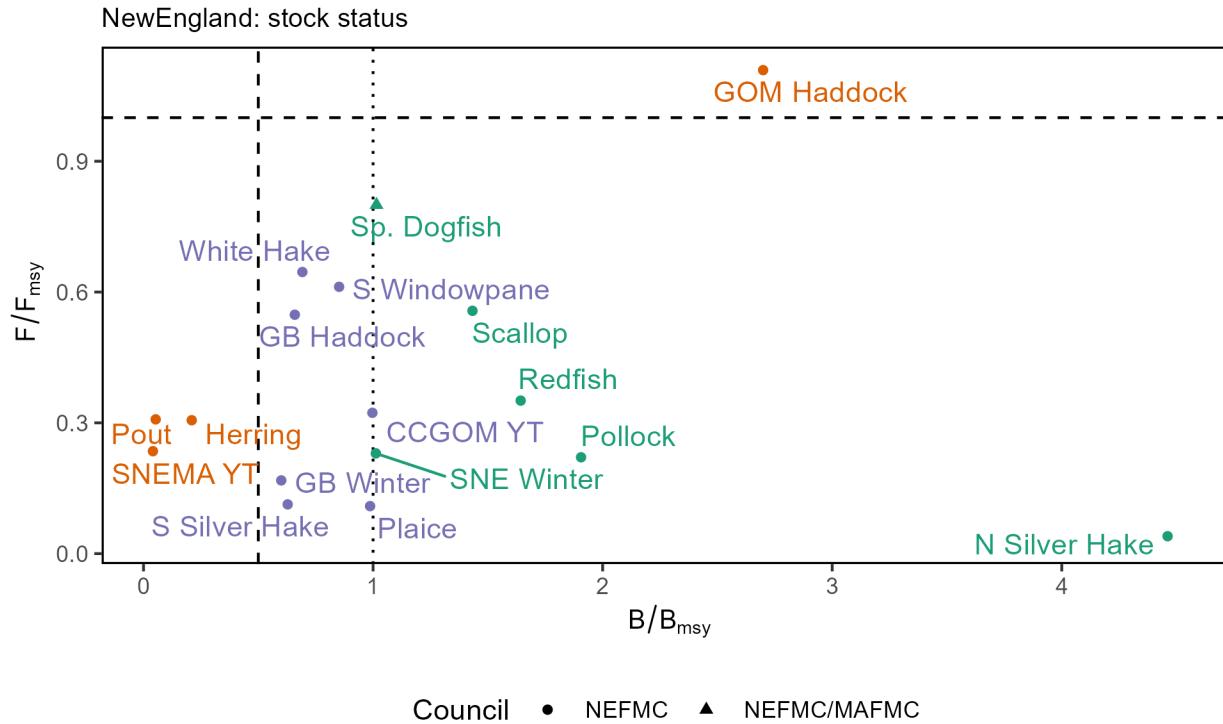


Figure 5: Summary of single species status for NEFMC and jointly federally managed stocks (goosefish and spiny dogfish). The dotted vertical line at one is the target biomass reference point of B. The dashed lines are the management thresholds of B (vertical) or F (horizontal). Colors denote stocks with  $B/B_{MSY} < 0.5$  or  $F/F_{MSY}$  (orange), stocks  $0.5 < B/B_{MSY} < 1$  (blue), and stocks  $B/B_{MSY} > 1$  (green). CCGOM = Cape Cod Gulf of Maine, GOM = Gulf of Maine, GB = Georges Bank, SNEMA = Southern New England Mid Atlantic

Table 3: Unknown or partially known stock status for NEFMC and jointly managed species.

Stock	F/F <sub>msy</sub>	B/B <sub>msy</sub>
Atlantic cod - Georges Bank <sup>1</sup>	-	-
Atlantic cod - Gulf of Maine <sup>1</sup>	-	-
Atlantic halibut - Northwestern Atlantic Coast	-	-
Barndoor skate - Georges Bank / Southern New England	-	1.070
Clearnose skate - Southern New England / Mid-Atlantic	-	0.802
Little skate - Georges Bank / Southern New England	-	0.580
Offshore hake - Northwestern Atlantic Coast	-	-
Red deepsea crab - Northwestern Atlantic	-	-
Red hake - Gulf of Maine / Northern Georges Bank	-	-
Red hake - Southern Georges Bank / Mid-Atlantic	-	-
Rosette skate - Southern New England / Mid-Atlantic	-	1.075
Smooth skate - Gulf of Maine	-	0.696
Thorny skate - Gulf of Maine	-	0.035
Windowpane - Gulf of Maine / Georges Bank	-	-
Winter flounder - Gulf of Maine	-	-
Winter skate - Georges Bank / Southern New England	-	1.120
Witch flounder - Northwestern Atlantic Coast	-	-
Yellowtail flounder - Georges Bank	0.09	-
Goosefish - Gulf of Maine / Northern Georges Bank	-	-
Goosefish - Southern Georges Bank / Mid-Atlantic	-	-

Table 3: Unknown or partially known stock status for NEFMC and jointly managed species.

Stock	F/Fmsy	B/Bmsy
<sup>†</sup> The most recent cod assessment made stock status recommendations for the four new stocks (Eastern Gulf of Maine, Western Gulf of Maine, Georges Bank, and Southern New England) but were not available yet for this report.		

**System Biomass** Aggregate biomass trends derived from scientific resource surveys have been stable to increasing in both regions (Fig. ?? & Fig. ??). The benthivores group spiked during the last decade, due to a large haddock recruitment, but appears to be returning to average levels. Planktivore biomass on GB continues to rise with the highest fall biomass observed since 1968. There are mixed trends in piscivores on GB, and increasing trends for planktivores across both regions and seasons and benthos on GB in both seasons. The New Hampshire/Maine state survey time series is too short to estimate trends, while the Massachusetts state survey shows the increasing trend in planktivores in the fall but a decrease in piscivores in the spring and benthos in both seasons (Fig. ??). While managed species comprise varying proportions of aggregate biomass, trends in landings are not mirroring shifts in the overall trophic structure of survey-sampled fish and invertebrates. Therefore, major shifts in feeding guilds or ecosystem trophic structure are unlikely to be driving the decline in landings.

**Effect on Seafood Production** With the poor or unknown stock status of many managed species, the decline in commercial seafood landings in the Gulf of Maine most likely reflects lower catch quotas implemented to rebuild overfished stocks, as well as market dynamics.

The decline in recreational seafood harvest stems from multiple drivers. Some of the decline, such as for recreational shark landings, continues to be driven by tightening regulations. However, changes in demographics and preferences for recreational activities likely play a role in non-HMS (Highly Migratory Species) declines in recreational harvest, with current harvests well below the time series average.

Other environmental changes require monitoring as they may become important drivers of future landings:

- Climate is trending into uncharted territory. Globally, 2024 was the warmest year on record<sup>5</sup> (see 2024 Highlights section).
- Stocks are shifting their distribution, moving towards the northeast and into deeper waters throughout the Northeast US Large Marine Ecosystem (Fig. 24, Climate Risks section).
- Ecosystem composition and production changes have been observed (see Stability section).
- Some fishing communities are affected by social vulnerabilities (see Social Vulnerability section).

<sup>5</sup>[https://noaa-edab.github.io/catalog/observation\\_synthesis.html](https://noaa-edab.github.io/catalog/observation_synthesis.html)

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### Commercial Profits

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#### Indicators: revenue (a proxy for profits)

Total [commercial revenues](#) from all species is below the long-term mean for both the GB and GOM regions in 2023 (Fig. 6). In addition, revenue from NEFMC managed species shows a long-term decline in the GOM. GB continues to exhibit a cyclical nature with regards to revenue, largely driven by rotational management of Atlantic sea

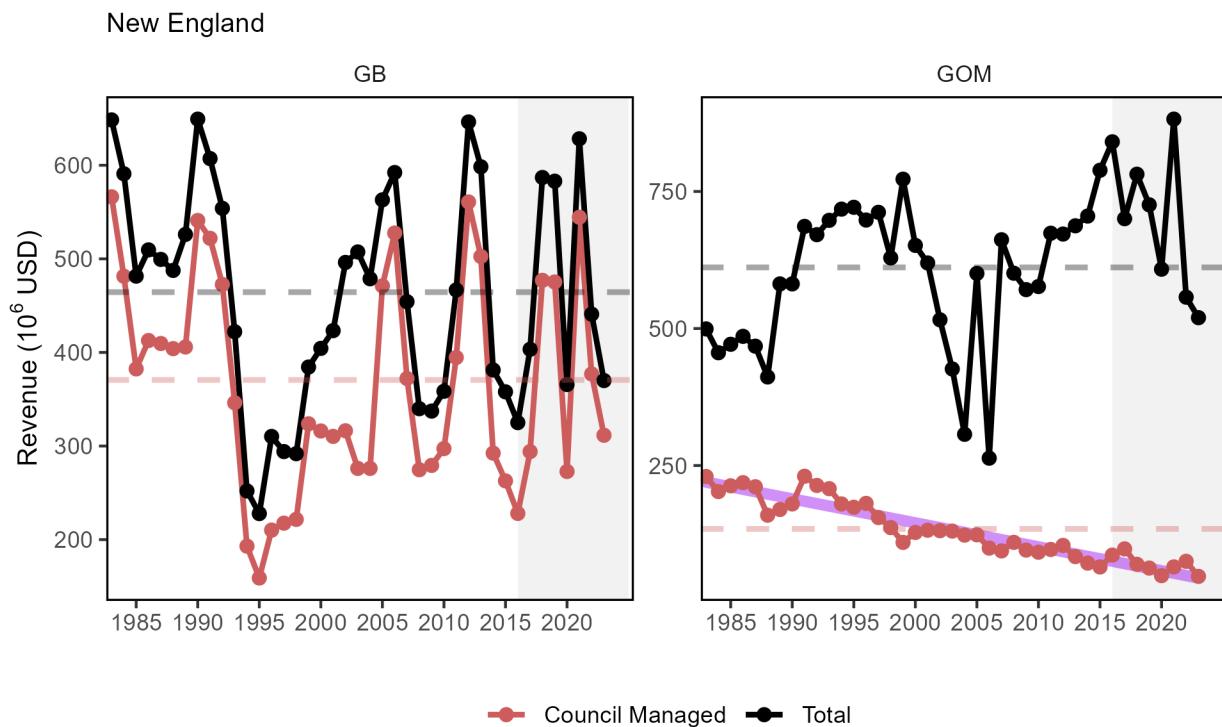


Figure 6: Revenue through 2023 for the New England region: total (black) and from NEFMC managed species (red).

Revenue earned by harvesting resources is a function of both the quantity landed of each species and the prices paid for landings. Beyond monitoring yearly changes in revenue, it is even more valuable to determine what drives these changes: harvest levels, the mix of species landed, price changes, or a combination of these. The [Bennet Indicator](#) decomposes revenue change into two parts, one driven by changing quantities (volumes), and a second driven by changing prices. All changes are in relation to a base year (1982).

In the GB region, revenues have been consistently lower than the 1982 baseline throughout the time series. The changes in total revenue in GB was primarily driven by volumes prior to 2010, and then by prices (Fig. 7). In the GOM, revenues have been above the 1982 baseline in all but four years, largely due to changing prices in most years. Breaking down the revenue by guild (Fig. 8), or GB, both the volume and price trend have been largely driven by benthos (scallops, quahogs and surfclams). In the GOM region, increased prices for benthivores (lobster) drove the year-over-year increases in overall prices. Benthivores also had a large influence on the overall volume indicator in the GOM.

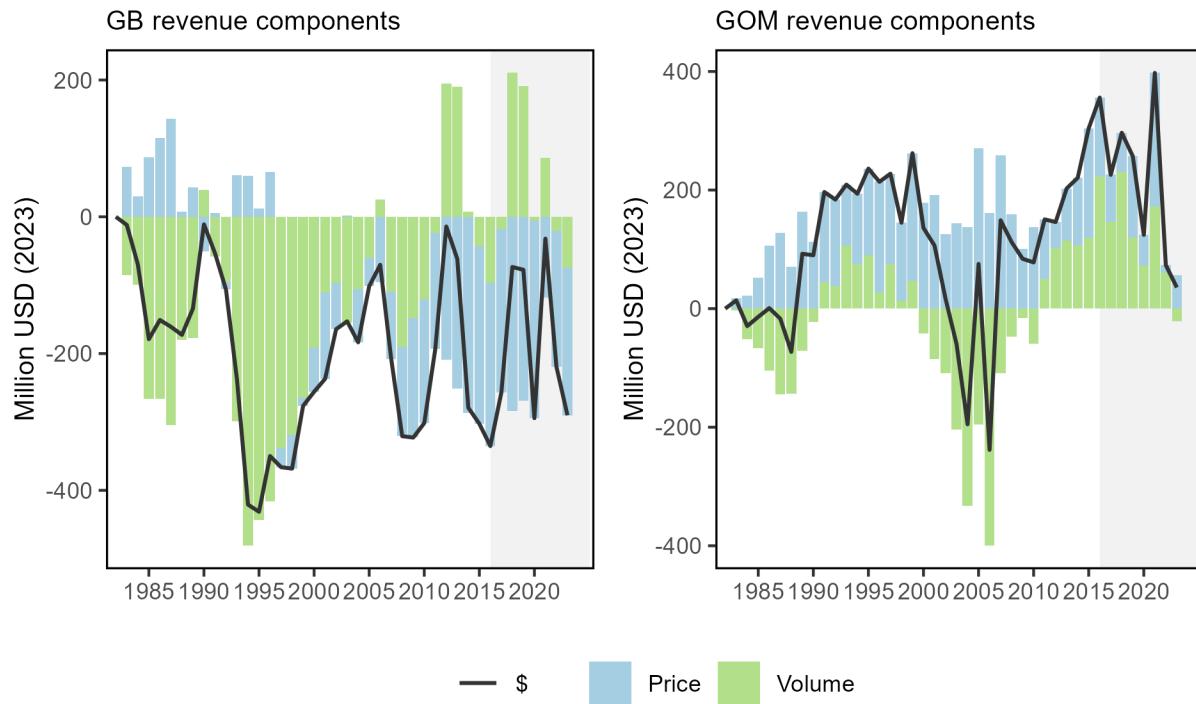


Figure 7: Revenue change from the 1982 baseline in 2023 dollars (black), price, and volume for commercial landings from Georges Bank (GB: left) and the Gulf of Maine (GOM: right)

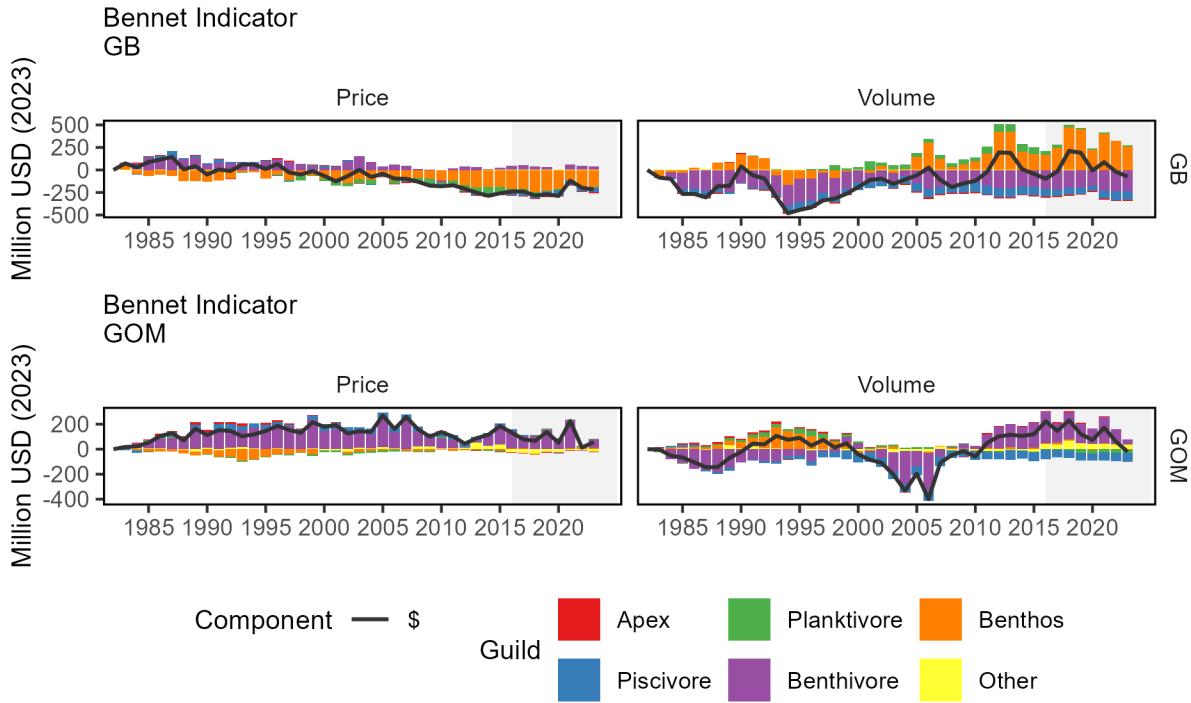


Figure 8: Revenue change from the long-term mean in 2023 dollars (black), price, and volume for commercial landings from Georges Bank (GB: top panels) and the Gulf of Maine (GOM: bottom panels)

For New England, [total climate vulnerability](#) of revenue was moderate in 2022 with no long-term trend (Fig. 9). This suggests that while New England commercial fishing is moderately reliant on climate-sensitive species, this proportion has not significantly changed since 2000.

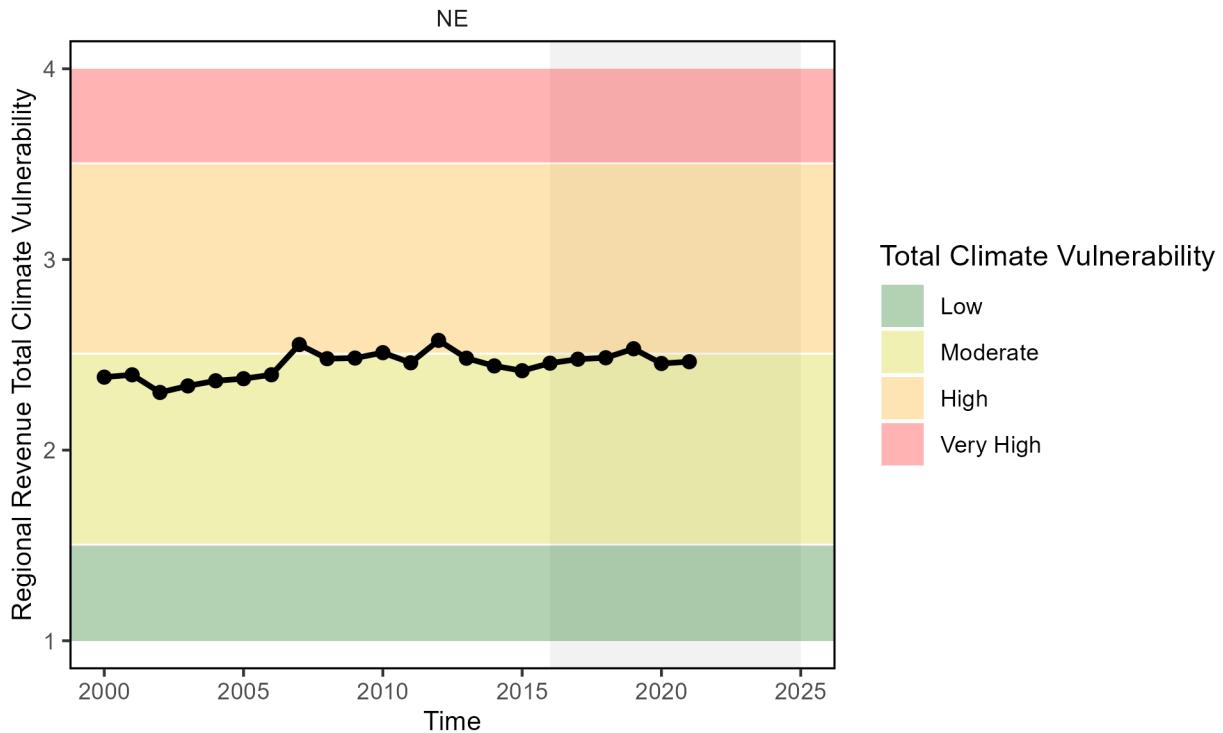


Figure 9: Total climate vulnerability on New England revenue from 2000 to 2022. Horizontal colored bars show different climate risk levels.

### Implications

The continued dependence on lobster in the GOM and sea scallops on GB is affected by multiple drivers including resource availability and market conditions. As both species are sensitive to ocean warming and acidification, it is important to monitor these and other climate drivers.

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### Recreational Opportunities

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### Indicators: Angler trips, fleet diversity

Recreational effort (angler trips) increased during 1982–2010, but has since declined to the long-term average (Fig. 10). Recreational fleets are defined as private vessels, shore-based fishing, or party-charter vessels. Recreational fleet diversity, or the relative importance of each fleet type, has remained relatively stable over the latter half of the time series (Fig. 11).

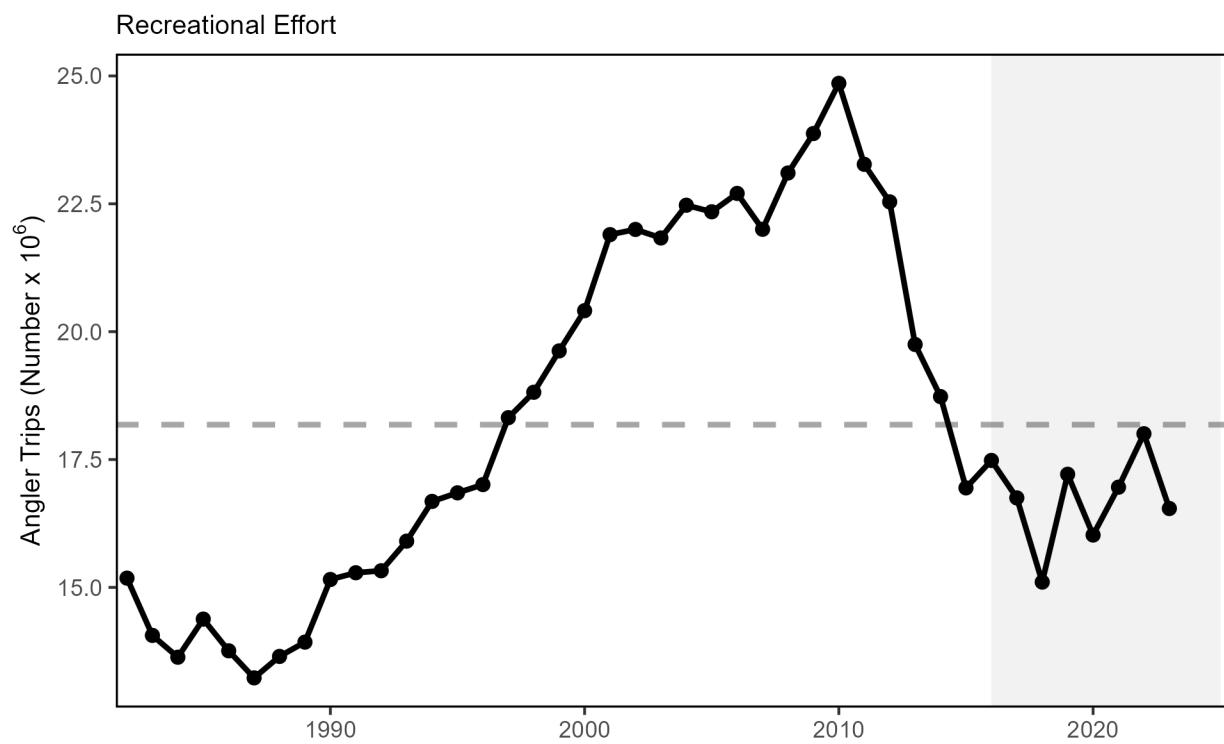


Figure 10: Recreational effort in New England.

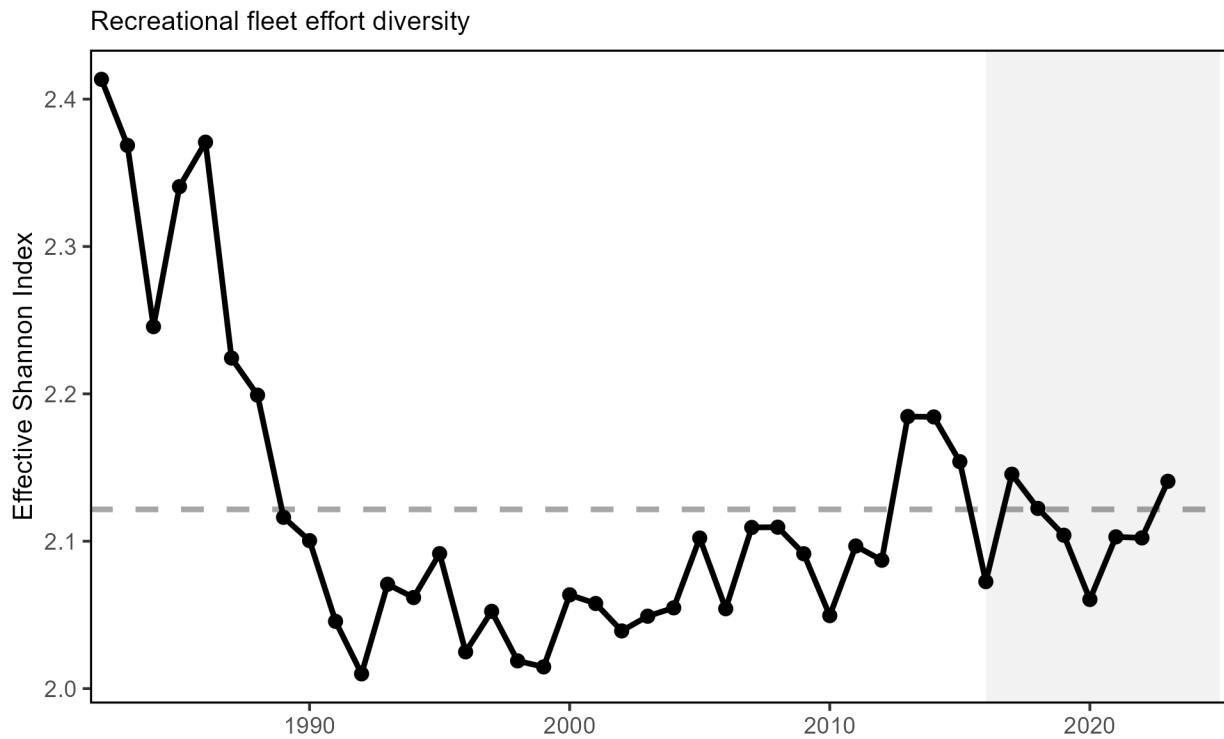


Figure 11: Recreational fleet effort diversity in New England.

## Implications

The absence of a long term trend in recreational angler trips and fleet effort diversity suggests relative stability in the overall number of recreational opportunities in the region.

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### Stability

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### **Indicators: fishery fleet and catch diversity, ecological component diversity, total primary production**

While there are many potential metrics of stability, we use diversity indices to evaluate overall stability in fisheries and ecosystems. In general, diversity that remains constant over time suggests a similar capacity to respond to change over time. A significant change in diversity over time does not necessarily indicate a problem or an improvement, but does indicate a need for further investigation. We examine diversity in commercial fleet and species catch, and recreational species catch (with fleet effort diversity discussed above), zooplankton, and adult fishes.

**Fishery Stability** Diversity estimates have been developed for species landed by commercial vessels with New England permits and fleets landing managed species. Although the effective number of species being landed in the commercial fleet rebounded slightly from the historical low of 2021, the diversity in catch is still well below the series average (Fig. 12). Commercial fishery fleet count is also below the time series average.

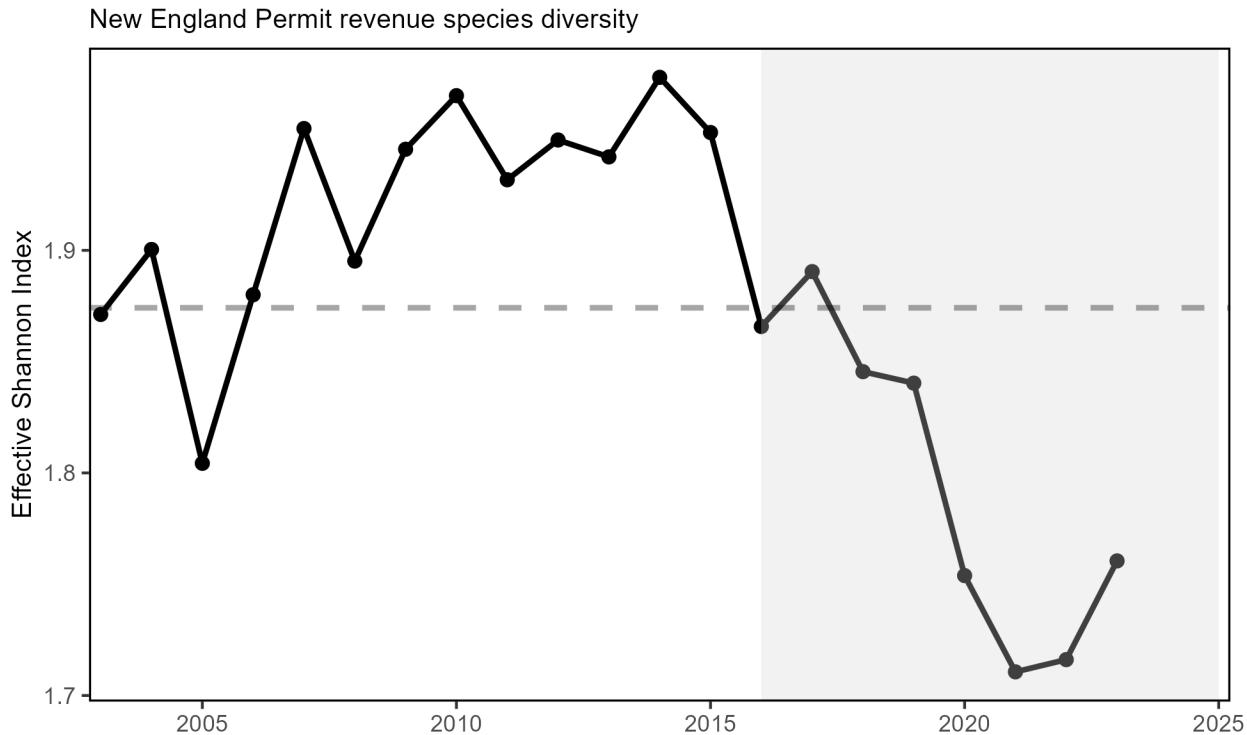


Figure 12: Species revenue diversity in New England.

As noted above, [recreational fleet effort diversity](#) is stable. However, recreational species catch diversity has been above the time series average since 2008 with a long-term positive trend (Fig. 13).

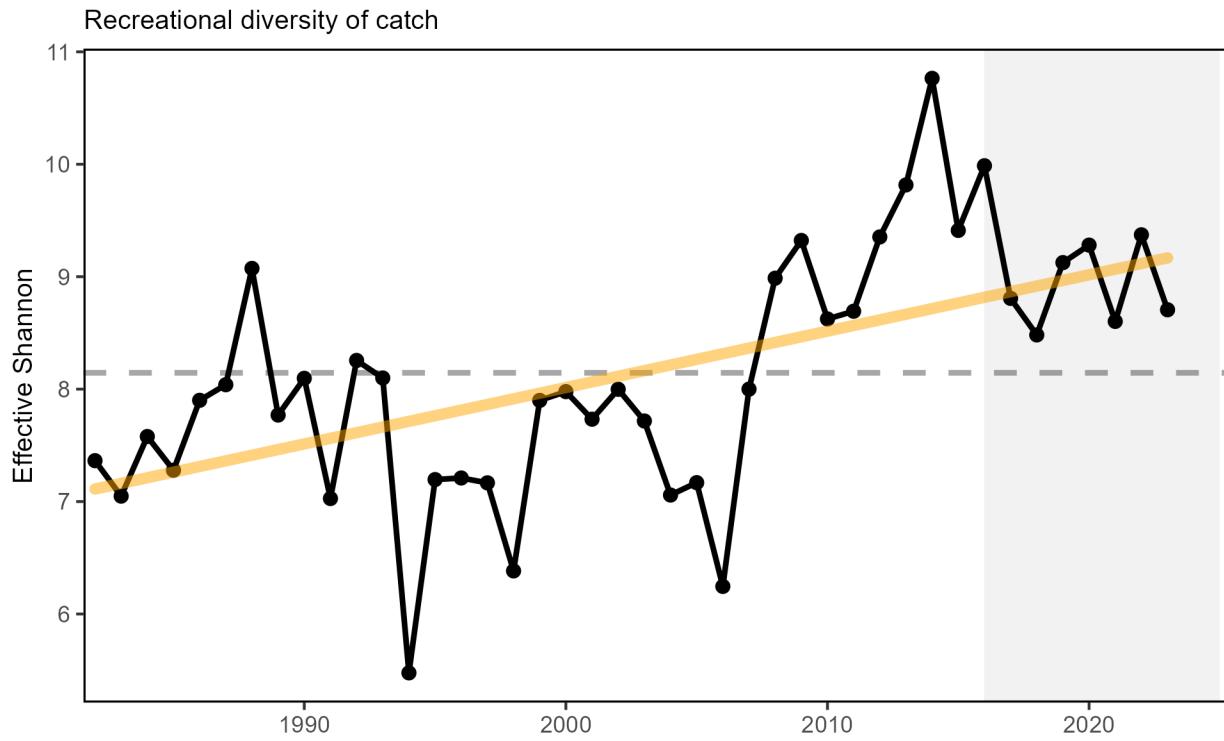


Figure 13: Diversity of recreational catch in New England.

**Ecological Stability** Ecological diversity indices show mixed trends. Total annual [primary production](#) (TPP) is a measure of the total amount of carbon (i.e. energy) produced by phytoplankton per year. 2023 saw record high TPP in the GOM due to a highly unusual phytoplankton bloom, but it is currently unknown how much of that primary production was incorporated into the upper trophic levels. Preliminary 2024 values were near the long-term average (Fig. 14).

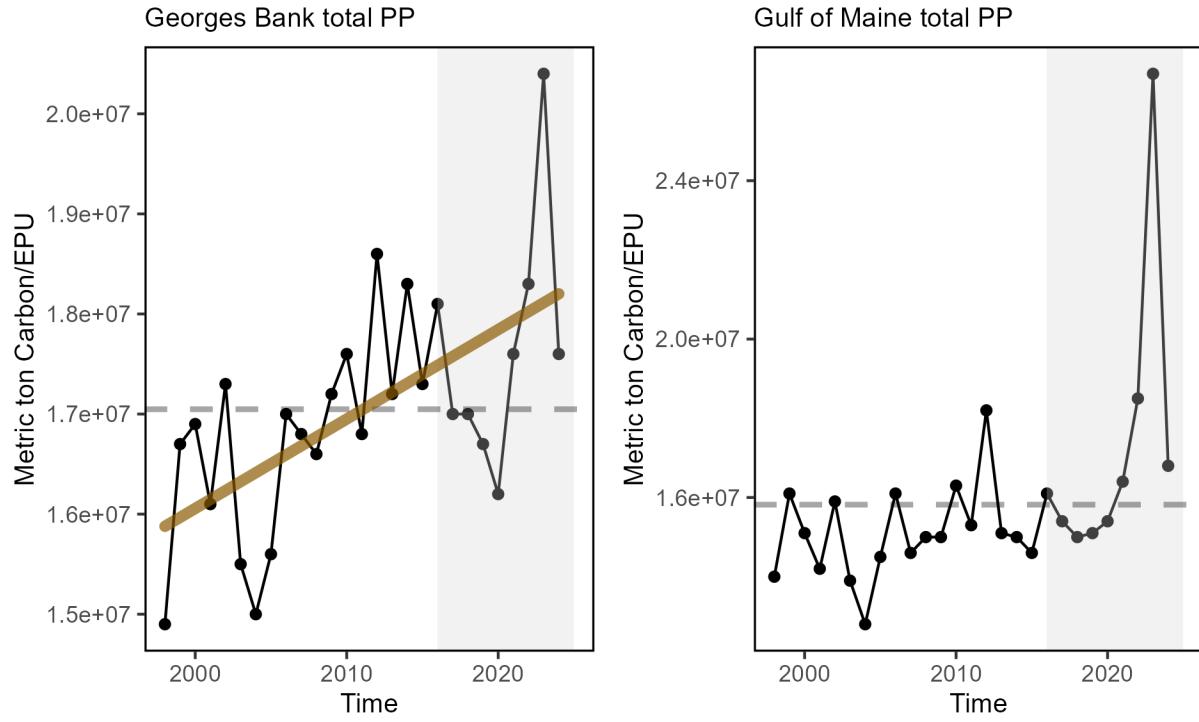


Figure 14: Total areal annual primary production by ecological production unit. The dashed line represents the long-term (1998-2023) annual mean.

[Zooplankton diversity](#) is increasing on GB, while no trend is evident in the GOM (Fig. 15). However, it is worth noting that the 2021 index for the GOM is the highest observed. [Adult fish diversity](#) shows an increasing trend in the GOM and no trend on GB (Fig. 16).

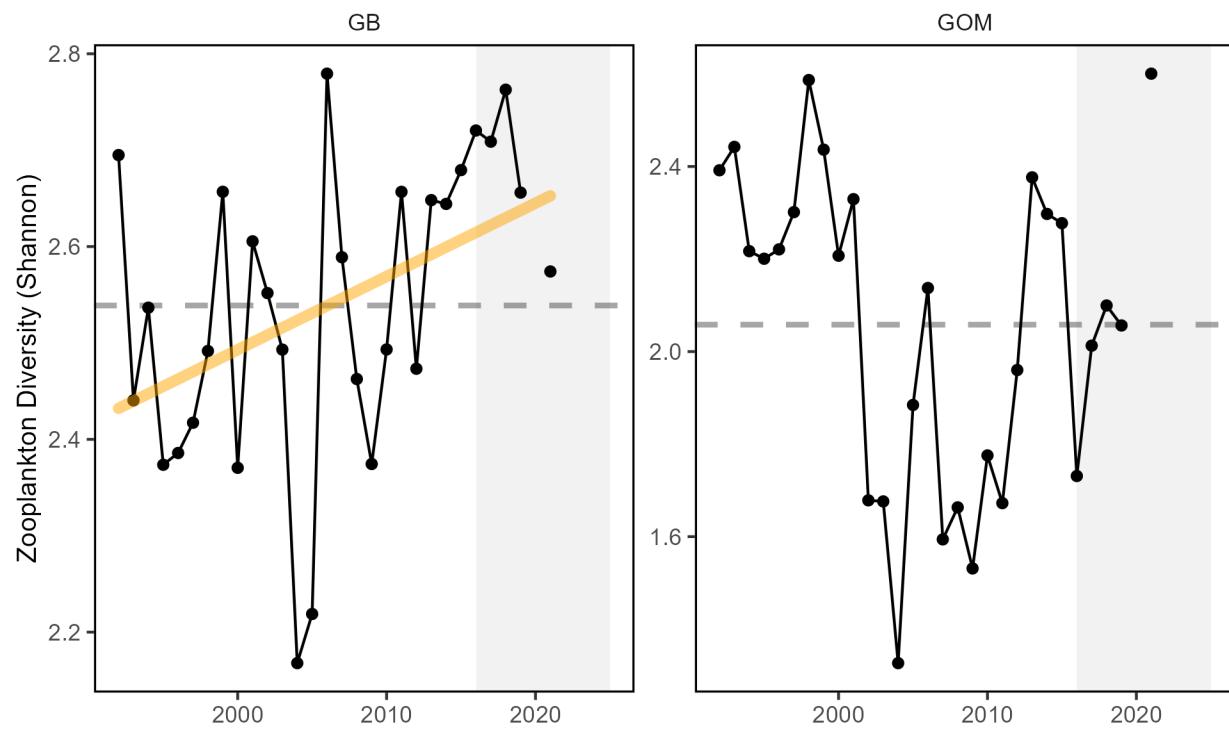


Figure 15: Zooplankton diversity on Georges Bank and in the Gulf of Maine, based on Shannon diversity index. 2020 surveys were incomplete due to COVID-19.

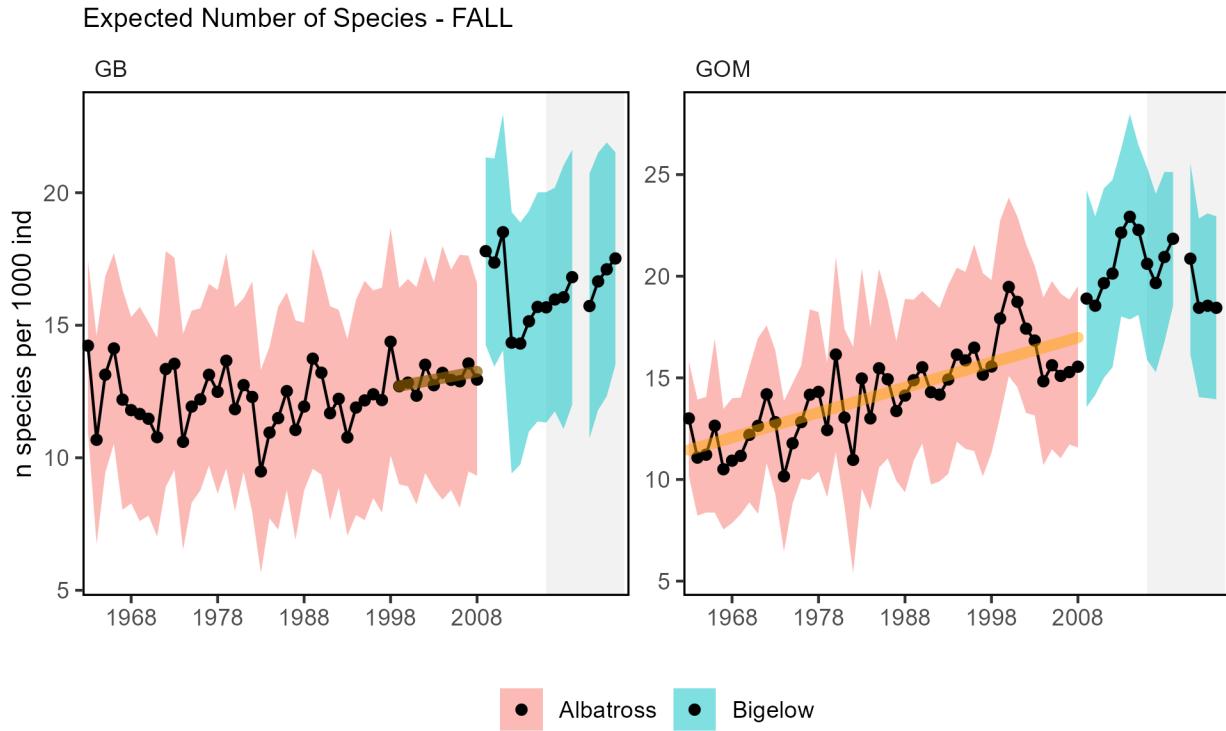


Figure 16: Adult fish diversity for Georges Bank and in the Gulf of Maine, based on expected number of species in a standard number of individuals. Results from survey vessels Albatross and Bigelow are reported separately due to catchability differences.

[Functional traits](#), such as length at maturity, asymptotic body size, or fecundity, can synthesize change across complex, diverse communities. Monitoring changes in functional trait distributions can provide a means of assessing ecosystem-scale resilience. There is evidence of long term change in trait distributions in the GOM with an increase in [pace of life](#). Both fall and spring finfish communities are showing declines in fecundity in GB and GOM (Fig. 17).

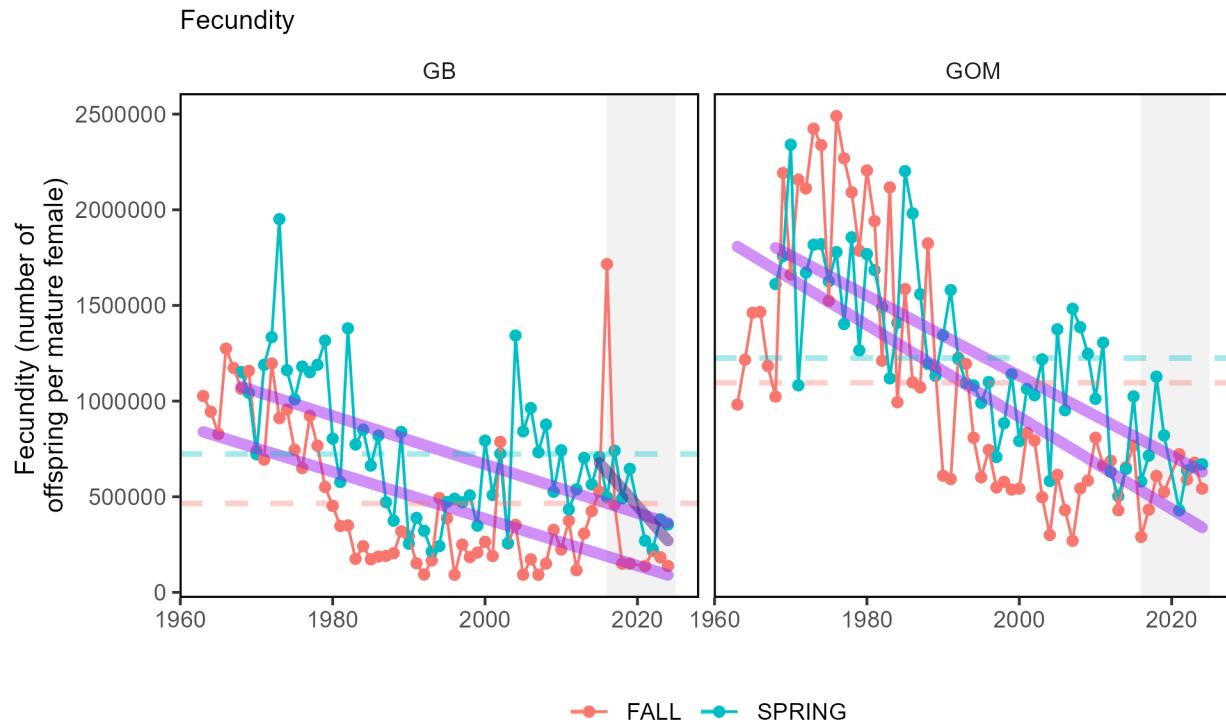


Figure 17: Fish community functional traits in New England based on Fall (red) and Spring (blue) survey data.

### Implications

Fleet diversity indices can be used to evaluate stability objectives as well as risks to fishery resilience and to maintain equity in access to fishery resources. The relatively low diversity estimates for the commercial fishery are likely driven by the continued reliance on a few species, such as sea scallops and lobster. This trend could diminish the capacity to respond to future fishing opportunities. Meanwhile, the increase in recreational species catch diversity is due to recent increases in Atlantic States Fisheries Management Council (ASFMC) and MAFMC managed species within the region, offsetting decreased limits on more traditional regional species.

Ecological diversity indices can provide insight into ecosystem structure. Changes in ecological diversity over time may indicate altered ecosystem structure with implications for fishery productivity and management. Increasing zooplankton diversity in GB is attributed to an overall increase in zooplankton abundance and the declining dominance of the calanoid copepod *Centropages typicus*. Stable adult fish diversity on GB suggests the same overall number and evenness over time, but does not rule out species substitutions (e.g., warm-water species replacing cold-water ones). Increasing adult diversity in the GOM suggests an increase in warm-water species and should be closely monitored.

As a whole, the examined diversity indicators suggest changes in commercial and recreational fisheries, likely driven by changes in the mix of species landed. However, there seems to be overall stability in ecosystem components. Increasing diversity in the recreational catch, GB zooplankton, and GOM adult fish accompanied by lows in commercial fleet diversity metrics, suggests warning signs of a potential regime shift or ecosystem restructuring and warrants continued monitoring to determine if managed species are affected.

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### **Community Social and Climate Vulnerability**

Providing for sustained participation of fishing communities, and avoiding adverse economic impacts to fishing communities are objectives of fishery management. We report the top communities most engaged in commercial and recreational fisheries and the degree to which these communities may be vulnerable to change based on their socioeconomic conditions using data for the most recent available year (2022).

Coastal fishing communities worldwide have or are likely to experience social, economic, and cultural impacts from climate change, both negative (e.g., loss of infrastructure, fish stock decline) and positive (e.g., increased abundance of valuable species). Changes in marine fisheries as a consequence of climate change will require adaptation by coastal fishing communities and fisheries managers alike. The Community Climate Change Risk Indicators were developed to help examine trends in climate change vulnerability in U.S. coastal fishing communities in the Northeast Region using indicators developed to understand fishing community level risk to climate change as based on species dependency.

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### **Indicators: Fishing Engagement and Community Social Vulnerability**

The [engagement indices](#) demonstrate the importance of commercial and recreational fishing to a given community relative to other coastal communities in a region. Social vulnerability indicators measure social factors that shape a community's ability to adapt to change.

For this report, we focus on top communities with the highest engagement scores, the top communities with the highest population relative engagement scores, and on three socio-demographic indicators within the CSVI toolset (poverty, personal disruption, population composition).

In 2022, New Bedford, MA stands out as having a particularly high engagement in commercial fishing, while Frenchboro, ME is much more engaged in commercial fishing relative to its population size (Fig. ??). Of particular concern among top communities are New Bedford and Boston, MA as they both have medium or higher scores for all three socio-demographic indicators, while Port Clyde-Tenants Harbor, ME ranked medium for two of the three indicator (Table 4).

Table 4: Socio-demographic indicator rankings (ranging from low = low vulnerability to high = high vulnerability) for Mid-Atlantic communities most engaged in commercial fishing, 2022. Blank spaces indicate no data available.

Community	Personal Disruption	Population Composition	Poverty
Cape May, NJ	low	med	low
Reedville, VA	low	low	low
Montauk, NY	low	low	low
Point Pleasant Beach, NJ	med	low	low
Hampton Bays/Shinnecock, NY	low	high	low
Barnegat Light, NJ	low	low	low
Bronx/City Island, NY	high	high	high
Newport News, VA	med	low	med
Hampton, VA	med	low	med
Wanchese, NC	low	low	low
Atlantic City, NJ	high	high	high

Table 4: Socio-demographic indicator rankings (ranging from low = low vulnerability to high = high vulnerability) for Mid-Atlantic communities most engaged in commercial fishing, 2022. Blank spaces indicate no data available.

Community	Personal Disruption	Population Composition	Poverty
Ocean City, MD	med	low	low
Swan Quarter, NC	low	low	low
Wachapreague, VA	low	low	low
Quinby, VA	med	low	low
Bowers, DE	low	low	low
Little Creek, DE	high	low	high
Oak Beach, NY	low	low	

Narragansett/Point Judith, RI; Newington, NH; and Gloucester, MA ranked as top communities for both commercial and recreational indices (Fig. ??), suggesting that they may be impacted simultaneously (to a greater degree than others) by commercial and recreational regulatory changes. Of the top-ranked recreational communities, only Provincetown, MA and Falmouth, MA had medium or higher ranks for more than one socio-demographic indicator (Table. 5) examined here (poverty, personal disruption, population composition). This suggests that future changes to recreational fishing conditions may disproportionately impact Provincetown and Falmouth.

Table 5: Socio-demographic indicator rankings (ranging from low = low vulnerability to high = high vulnerability) for Mid-Atlantic communities most engaged in recreational fishing, 2022. Blank spaces indicate no data available.

Community	Personal Disruption	Population Composition	Poverty
Cape May, NJ	low	med	low
Montauk, NY	low	low	low
Point Pleasant Beach, NJ	med	low	low
Barnegat Light, NJ	low	low	low
Ocean City, MD	med	low	low
Virginia Beach, VA	low	low	low
Morehead City, NC	med	low	med
Hatteras township, NC	low	low	low
Wachapreague, VA	low	low	low
Avon, NC	high	low	
Atlantic Highlands, NJ	low	low	low
Babylon, NY	low	low	low
Nags Head, NC	low	low	low
Point Lookout, NY	low	low	low
Nanticoke, MD	med high	low	low
Orient, NY	low	low	
Bivalve, MD	med	high	
Rodanthe, NC	low	low	
Topsail Beach, NC	low	low	low
Solomons Island/Solomons/Lusby, MD	low	low	low
Stevensville, MD	med	low	low

### Indicators: Community Climate Vulnerability in the New England

The [Community Climate Change Risk Indicators](#) are calculated by multiplying the percent contribution of species to the total value landed in a community by their respective Total Vulnerability scores (based on NOAA's Climate Vulnerability Assessment) for different sensitivity and exposure factors and then summing the resulting values by year. As a community (or region) shifts towards climate vulnerable species, its risk score increases. While there is not a long-term trend in total climate vulnerability across New England communities as a whole, the proportion of communities with moderate vulnerability is decreasing and shifting more towards high or very high vulnerability scores (Fig. 18). This suggests that some communities are shifting towards being more dependent on climate-vulnerable species, particularly shellfish.

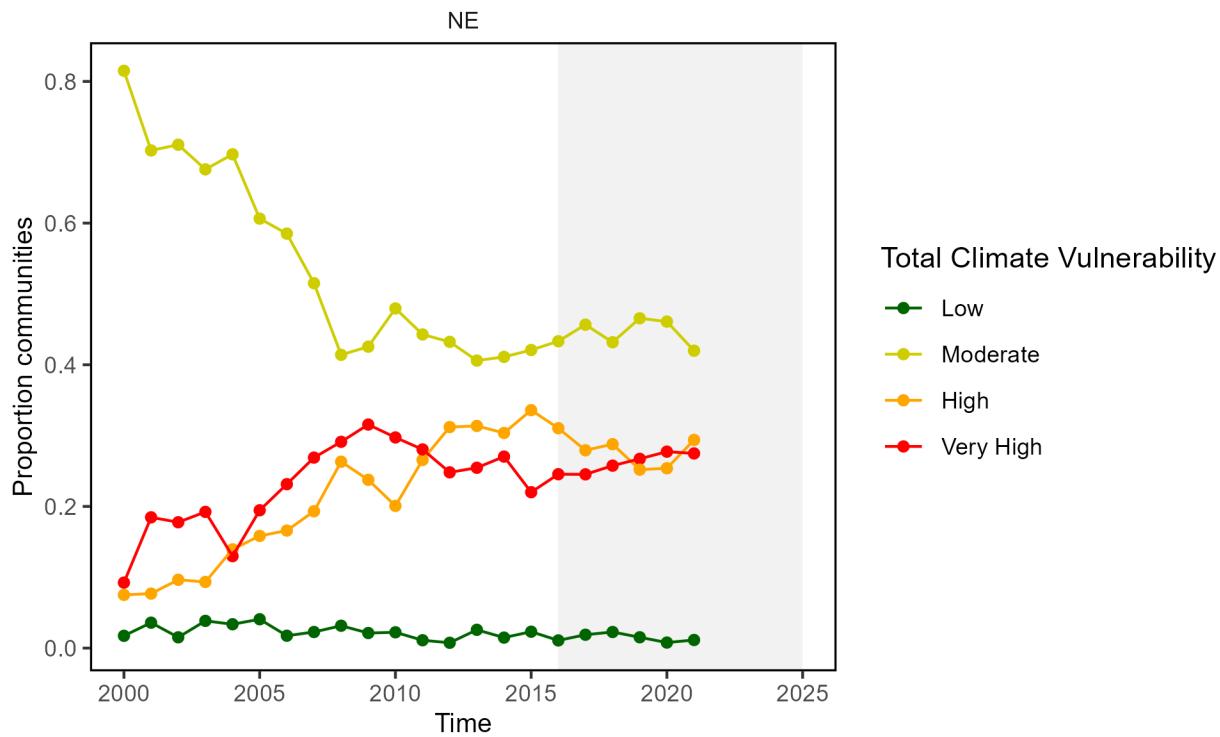


Figure 18: Proportion of New England communities at each revenue climate vulnerability level over time.

## Implications

These indicators provide a snapshot of the presence of socio-demographic concerns in the most highly engaged commercial and recreational fishing communities in New England. These communities may be especially vulnerable to changes in fishing patterns due to regulations and/or ecosystem changes. Several of these top fishing communities, both commercial and recreational fishing communities, demonstrated medium to high socio-demographic vulnerability, indicating that they may be at a disadvantage responding to change.

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### Protected Species

Fishery management objectives for protected species generally focus on reducing threats and on habitat conservation/restoration. Protected species include marine mammals protected under the Marine Mammal Protection Act, endangered and threatened species protected under the Endangered Species Act, and migratory birds protected under the Migratory Bird Treaty Act. In the Northeast U.S., endangered/threatened species include Atlantic salmon, Atlantic and shortnose sturgeon, all sea turtle species, giant manta ray, oceanic whitetip shark, and five baleen whales. Protected species objectives include managing bycatch to remain below potential biological removal (PBR) thresholds, recovering endangered populations, and monitoring unusual mortality events (UMEs). Here we report on performance relative to these objectives with available indicator data, as well as indicating the potential for future interactions driven by observed and predicted ecosystem changes in the Northeast U.S.

### Indicators: bycatch, population (adult and juvenile) numbers, mortalities

Average indices for both [harbor porpoise](#) (Fig. 19) and [gray seal](#) bycatch (Fig. 20) are below current PBR thresholds, meeting management objectives, although uncertainty in the gray seal bycatch estimate has increased recently, and gray seal bycatch is among the highest for marine mammals in the U.S.

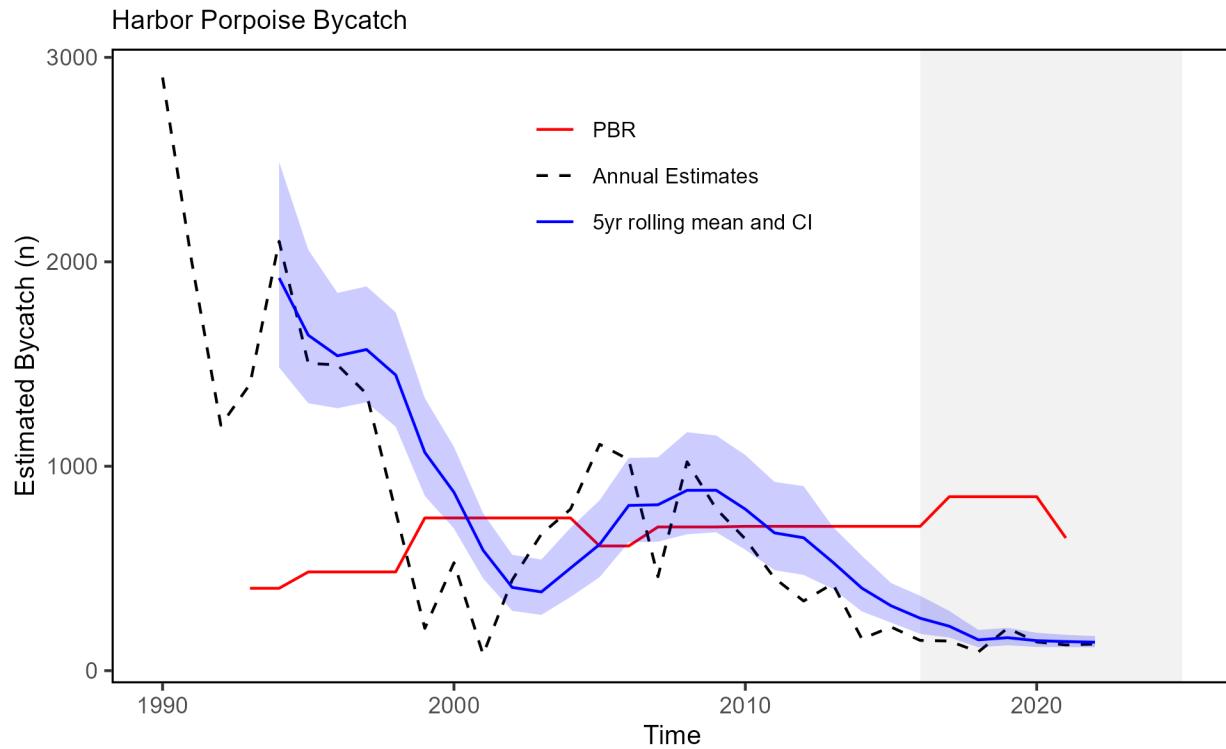


Figure 19: Harbor porpoise average bycatch estimate for Mid-Atlantic and New England gillnet fisheries (blue) and the potential biological removal (red).

The annual estimate for gray seal bycatch, most of which occurs in New England, has declined since 2019, in part driven by declining gillnet landings. In addition, estimates since 2019 have greater uncertainty stemming from low observer coverage in some times and areas since 2019. The rolling mean confidence interval remains just below the PBR threshold.

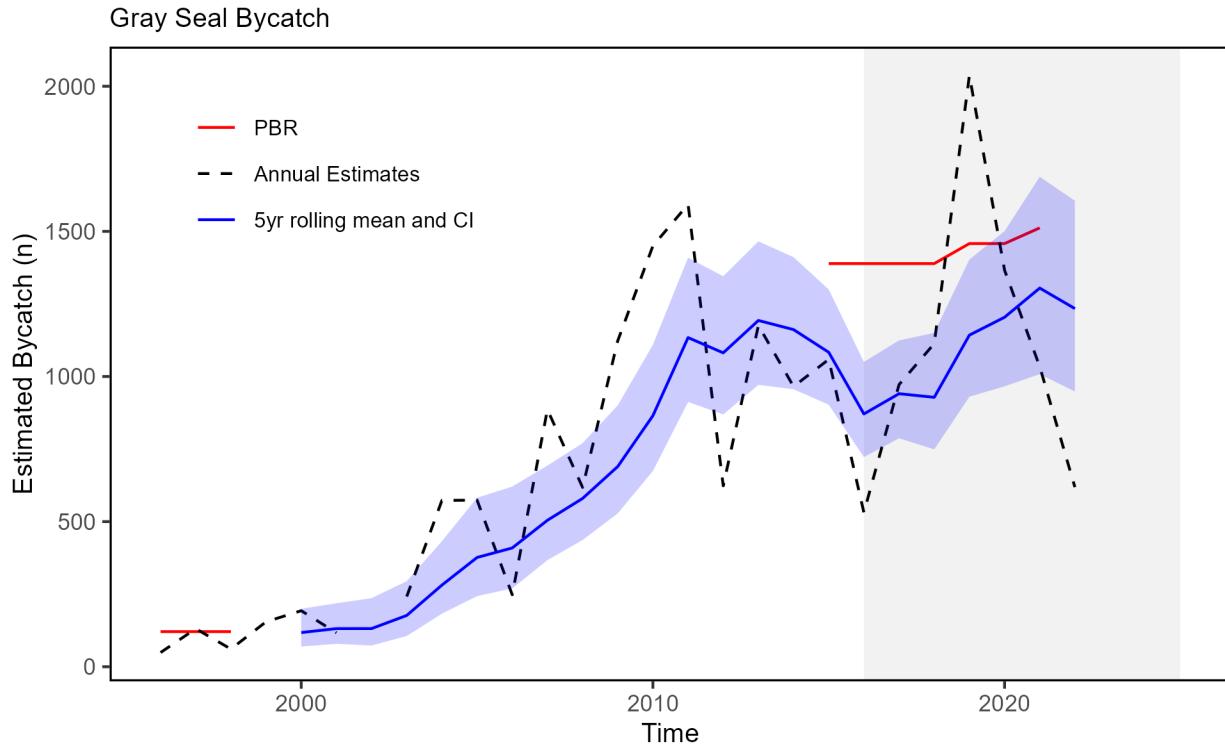


Figure 20: Gray Seal average bycatch estimate for gillnet fisheries (blue) and the potential biological removal (red).

The [North Atlantic right whale population](#) was on a recovery trajectory until 2010, but has since declined (Fig. 21). The sharp decline observed from 2015-2020 appears to have slowed, although the right whale population continues to experience annual mortalities above recovery thresholds. Reduced survival rates of adult females lead to diverging abundance trends between sexes. It is estimated that there are fewer than 70 adult females remaining in the population.

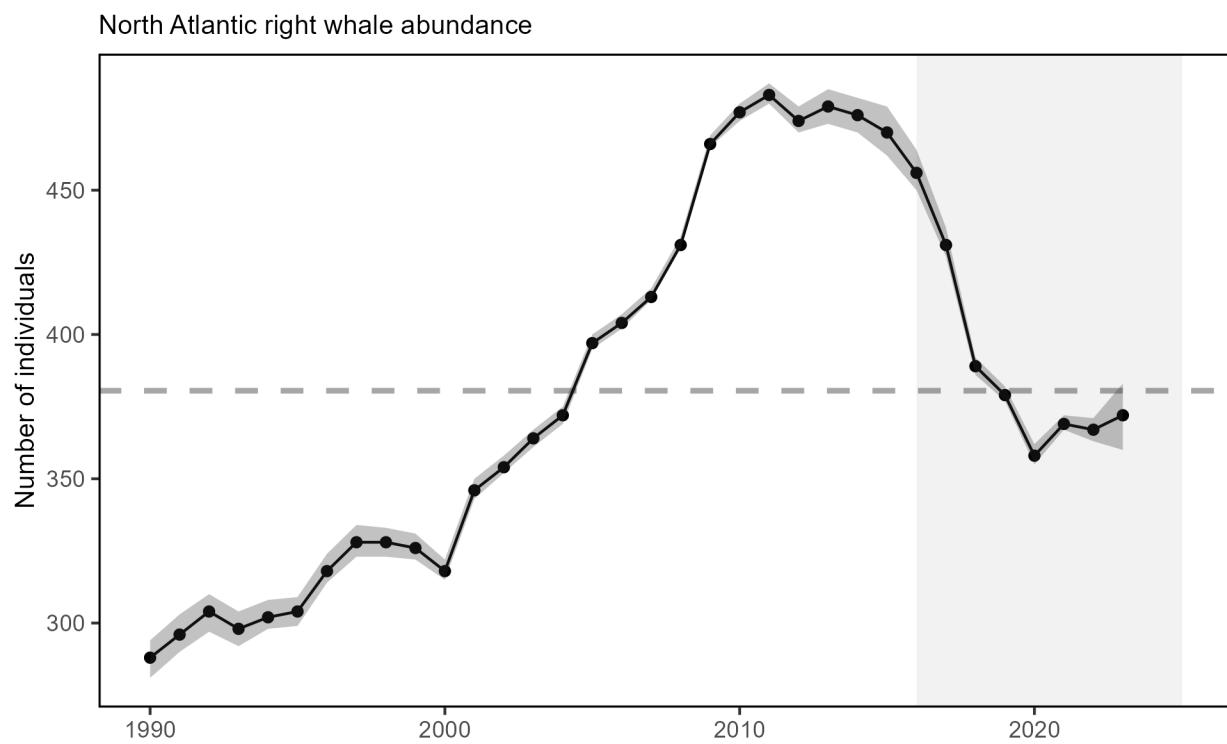


Figure 21: Estimated North Atlantic right whale abundance on the Northeast Shelf.

North Atlantic right whale [calf counts](#) have generally declined after 2009 to the point of having zero new calves observed in 2018 (Fig. 22). However, since 2020, calf births have been closer to the long-term average.

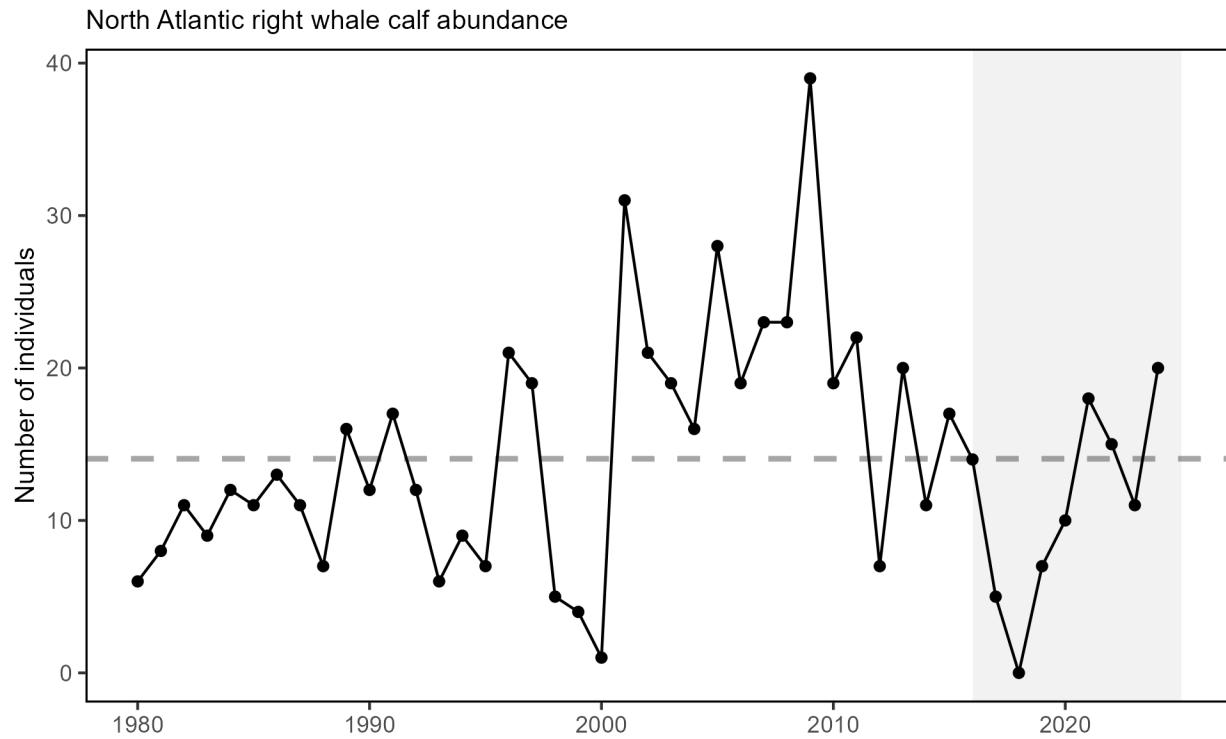


Figure 22: Number of North Atlantic right whale calf births, 1990 - 2022.

This year, the Unusual Mortality Event (UME) for North Atlantic right whales continued. From 2017 through 2 January 2025, the total UME right whale mortalities includes 41 dead stranded whales, 19 in the US and 22 in Canada. When alive but seriously injured whales (39) and sublethal injuries or ill whales (71) are taken into account, 151 individual whales are included in the UME. Recent research suggests that many mortalities go unobserved and the true number of mortalities are about three times the count of the observed mortalities. The primary cause of death is “human interaction” from entanglements or vessel strikes.

A UME continued from previous years for humpback whales (2016-present) and Atlantic minke whales (2018-present); suspected causes include human interactions. A UME for Northeast pinnipeds that began in 2018 for infectious disease is pending closure as of February 2025.

### **Implications**

Bycatch management measures have been implemented to maintain bycatch below PBR thresholds. The downward trend in harbor porpoise bycatch could also be due to a decrease in harbor porpoise abundance in U.S. waters, reducing their overlap with fisheries, and a decrease in gillnet effort. The increasing trend in gray seal bycatch may be related to an increase in the gray seal population ([U.S. pup counts](#)), supported by the dramatic rise over the last three decades in observed numbers of gray seal pups born at U.S. breeding sites plus an increase in adult seals at the breeding sites, some of which are supplemented by Canadian adults.

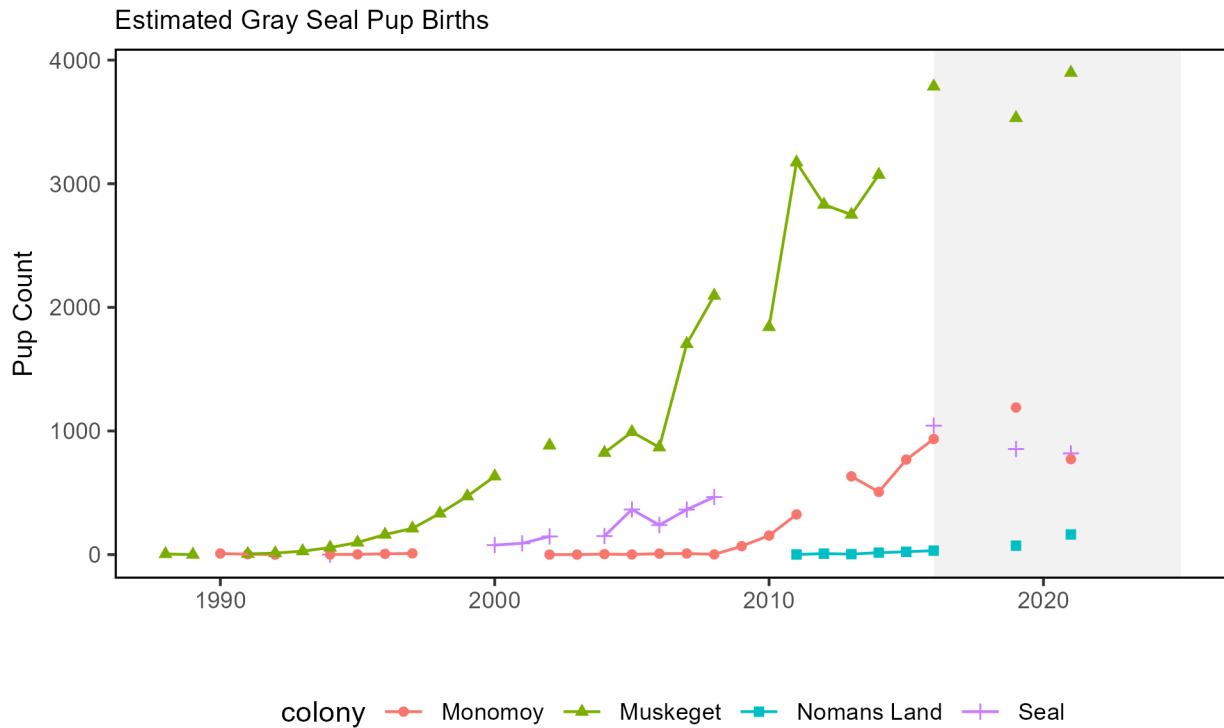


Figure 23: Estimated number of gray seal pups born at four United States pupping colonies at various times from 1988 to 2021. Recreated from Wood et al. 2022 (Figure 5).

Strong evidence exists to suggest that interactions between right whales and both the fixed gear fisheries in the U.S. and Canada and vessel strikes in the U.S. are contributing substantially to the decline of the species. Further, right whale distribution has changed since 2010. [Recent research](#) suggests that recent climate driven changes in ocean circulation have resulted in right whale distribution changes driven by increased warm water influx through the Northeast Channel, which has reduced the primary right whale prey (the copepod *Calanus finmarchicus*) in the central and eastern portions of the Gulf of Maine. Additional potential stressors include offshore wind development, which overlaps with important habitat areas used year-round by right whales, including mother and calf migration corridors and foraging habitat. This area is also a primary right whale winter foraging habitat. Additional information can be found in the [offshore wind risks section](#).

The UMEs are under investigation and are likely the result of multiple drivers. For all large whale UMEs, human interaction appears to have contributed to increased mortalities, although investigations are not complete.

A climate vulnerability assessment is published for Atlantic and Gulf marine mammal populations.

## Risks to Meeting Fishery Management Objectives

### Climate and Ecosystem Change

#### Risks to managing spatially

Shifting species distributions (changes in spatial extent or center of gravity) alter both species interactions and fishery interactions. In particular, shifting species distributions can affect expected management outcomes from spatial allocations and bycatch measures based on historical fish and protected species distributions. Species availability to surveys can also change as distributions shift within survey footprints, complicating the interpretation of survey trends.

Coastwide indicators are reviewed in this section to evaluate spatial change throughout the Northeast US shelf. Indicators are identical between the Mid Atlantic and New England reports.

**Indicators: Fish and protected species distribution shifts** As noted in the [Seafood Production Implications](#) section above, the center of [distribution](#) for a suite of 48 commercially or ecologically important fish species combined along the entire Northeast Shelf continues to show movement towards the northeast and generally into deeper water (Fig. 24). Distribution shifts have been noted for several [highly migratory species](#), including sharks, billfish and tunas between 2002 and 2019.

[Habitat model-based species richness](#) suggests shifts of both cooler and warmer water species to the northeast. Similar patterns have been found for [marine mammals](#), with multiple species shifting northeast between 2010 and 2017 in most seasons (Fig. 25).

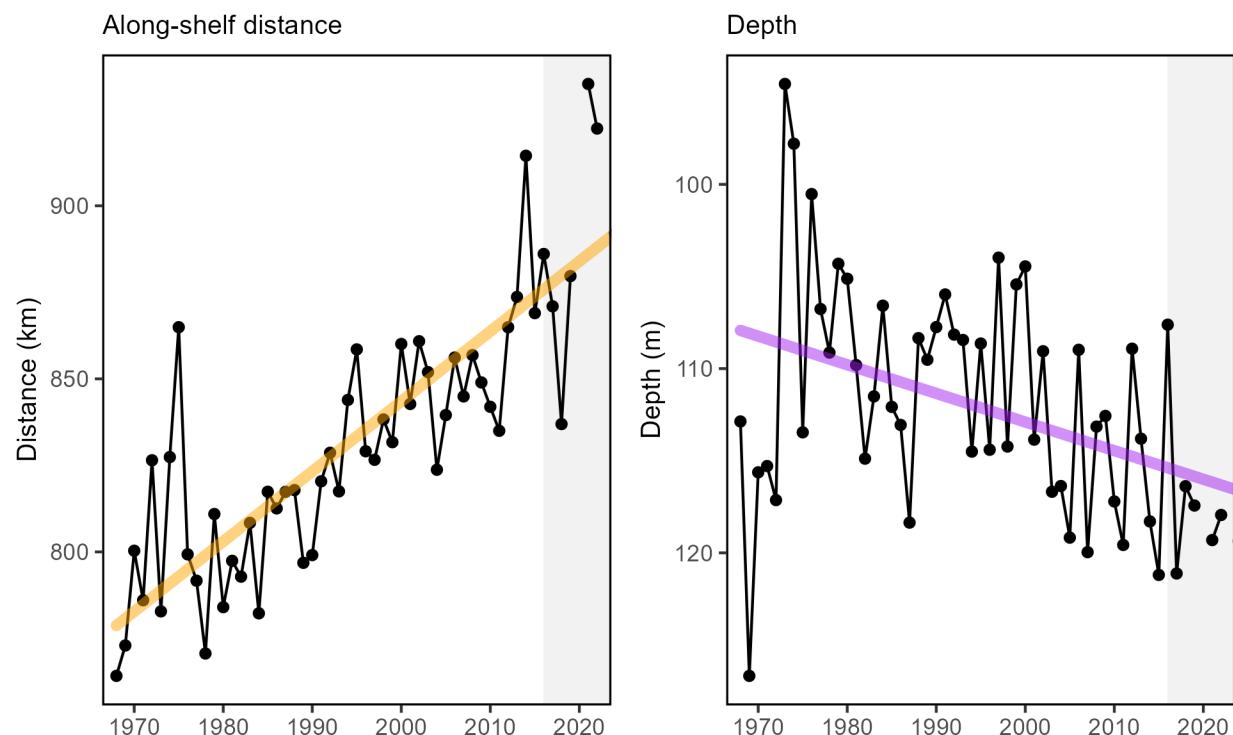


Figure 24: Aggregate species distribution metrics for species in the Northeast Large Marine Ecosystem: along shelf distance with increasing trend (orange), and depth with decreasing trend indicating deeper water (purple).

## Whale and Dolphin Distribution Shifts

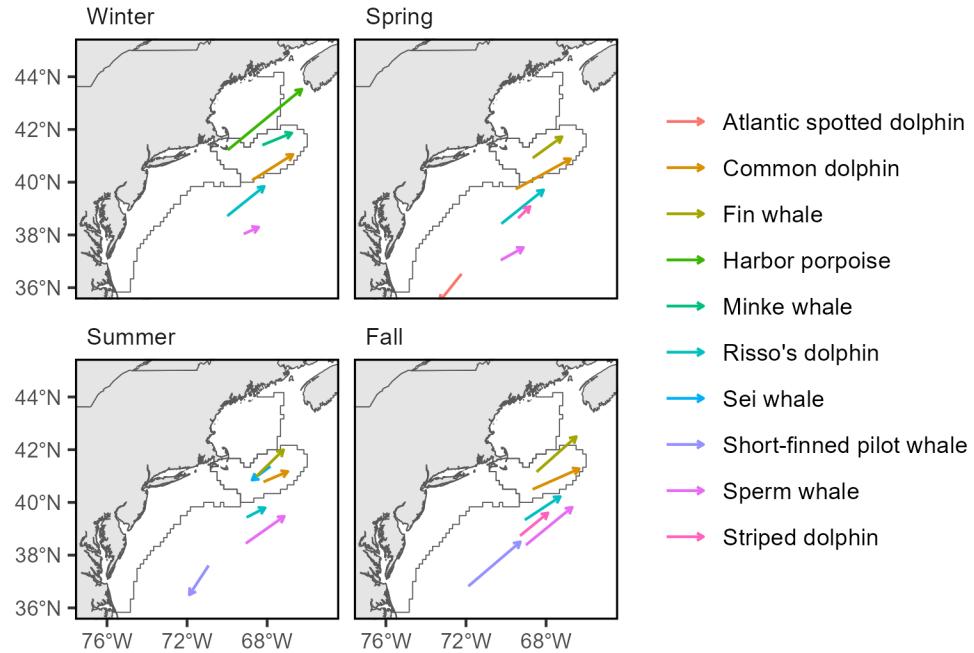


Figure 25: Direction and magnitude of core habitat shifts, represented by the length of the line of the seasonal weighted centroid for species with more than 70 km difference between 2010 and 2017 (tip of arrow).

**Drivers:** Mobile populations shift distributions to maintain suitable temperature and prey fields, possibly expanding ranges if new suitable habitat exists. Changes in managed species distribution is related, in part, to the [distribution of forage biomass](#). Since 1982, the fall center of gravity of forage fish (20 species combined) has moved to the north and east (Fig. 27). Spring forage fish center of gravity has been more variable over time. [Small copepods](#), widespread prey of many larval and juvenile fish, show a similar shift in center of gravity as forage fish, to the north and east in the fall, as well as northward in spring.

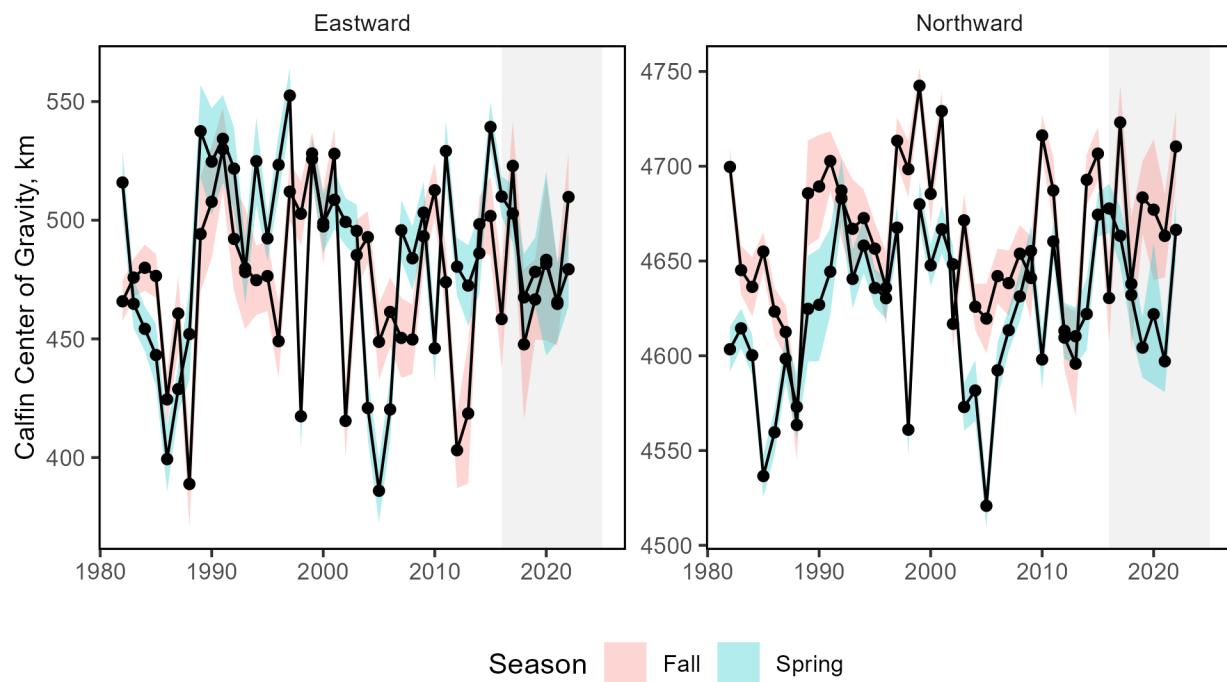


Figure 26: Eastward (left) and northward (right) shifts in the center of gravity for *\*Calanus finmarchicus\** on the Northeast U.S. Shelf.

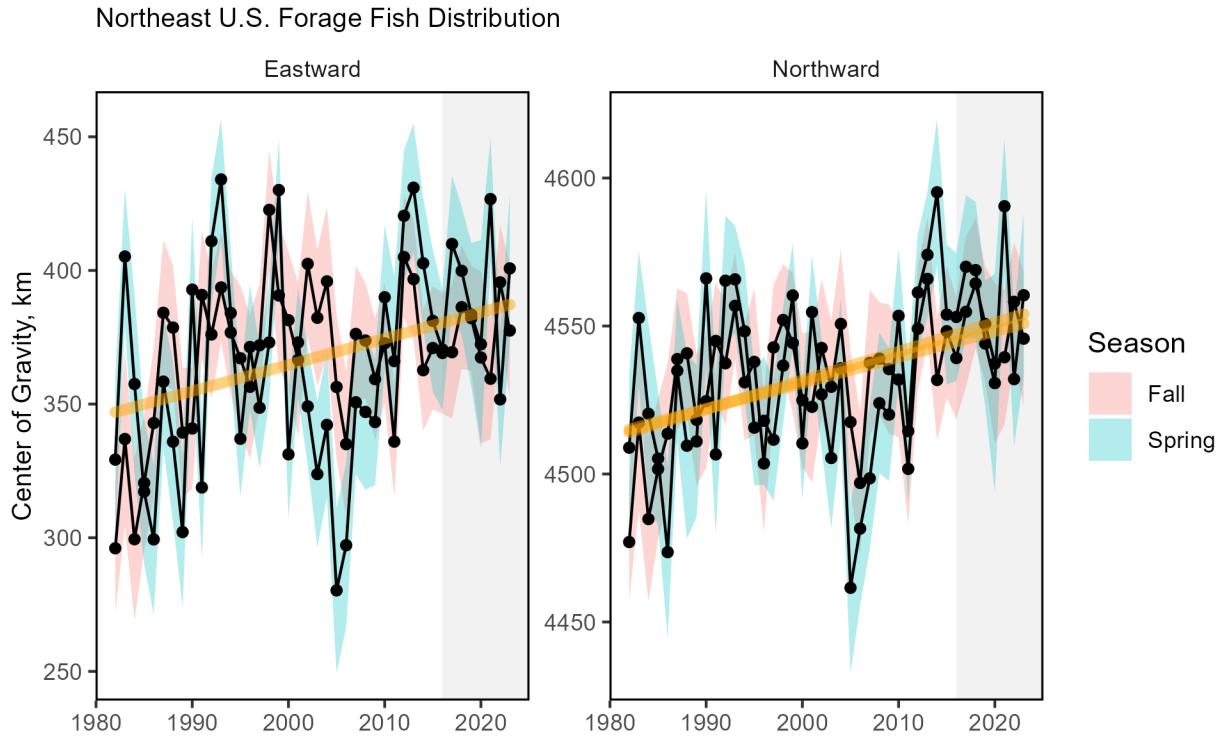


Figure 27: Eastward (left) and northward (right) shifts in the center of gravity for 20 forage fish species on the Northeast U.S. Shelf, with increasing trend (orange) for fall eastward and northward center of gravity.

In contrast, [macrobenthos](#) center of gravity has shifted westward (Fig. 28). Macrobenthos are small bottom-dwelling invertebrates including polychaete worms, small crustaceans, bivalves (non-commercial), gastropods, nemerteans, tunicates, cnidarians, brittle stars, sea cucumbers, and sand dollars, and are prey for many managed species. [Large copepods](#) have a similar pattern to macrobenthos, trending westward in fall.

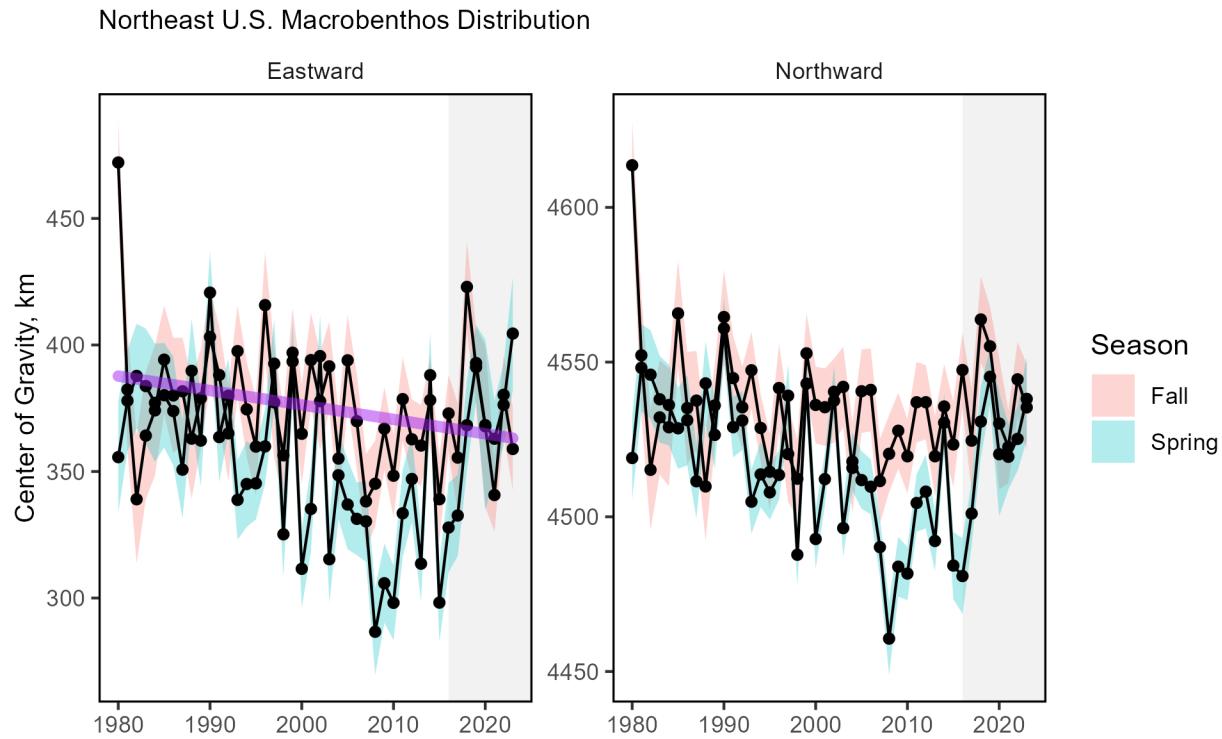


Figure 28: Eastward (left) and northward (right) shifts in the center of gravity for macrobenthos species on the Northeast U.S. Shelf

Ocean temperatures influence the distribution, seasonal timing, and productivity of managed species (see sections below). The Northeast US shelf, including the Mid-Atlantic, has experienced a continued warming trend for both the [long term annual sea surface](#) (Fig. 29) and [seasonal surface](#) and [bottom temperature](#). However, 2024 surface and bottom temperatures were near normal to cooler than normal conditions in all seasons in the MAB (see also the [2024 Highlights section](#)).

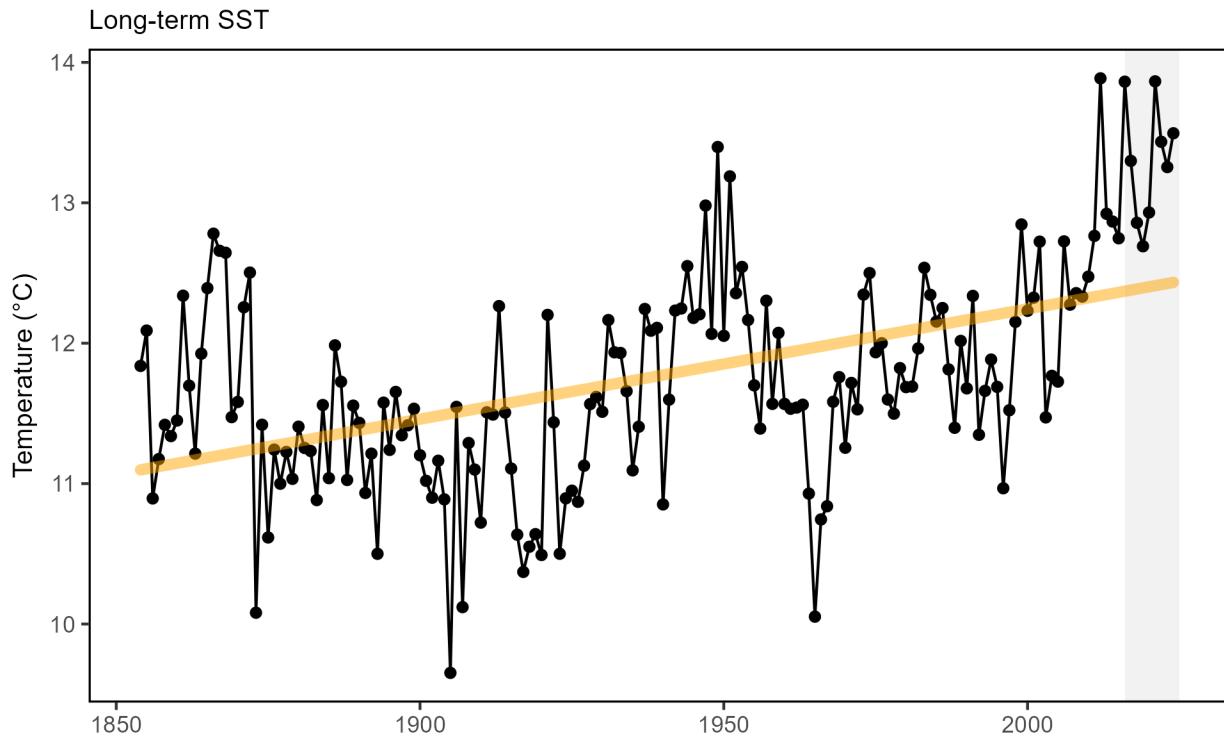


Figure 29: Northeast US annual sea surface temperature (SST, black), with increasing trend (orange).

Species suitable habitat can expand or contract when changes in temperature and major oceanographic conditions alter distinct water mass habitats. The variability of the Gulf Stream is a major driver of the predominant oceanographic conditions of the Northeast U.S. continental shelf. As the [Gulf Stream](#) has become less stable and shifted northward in the last decade (Fig. 30), warmer ocean temperatures have been observed on the northeast shelf and a higher proportion of [Warm Slope Water](#) has been present in the Northeast Channel. Since 2008, the Gulf Stream has moved closer to the Grand Banks, reducing the supply of cold, fresh, and oxygen-rich Labrador Current waters to the Northwest Atlantic Shelf. In 2024, however, the [eastern portion of the Gulf Stream was further south](#), which could affect the composition of the source water entering the Gulf of Maine through the Northeast Channel (see [2024 Highlights](#)).

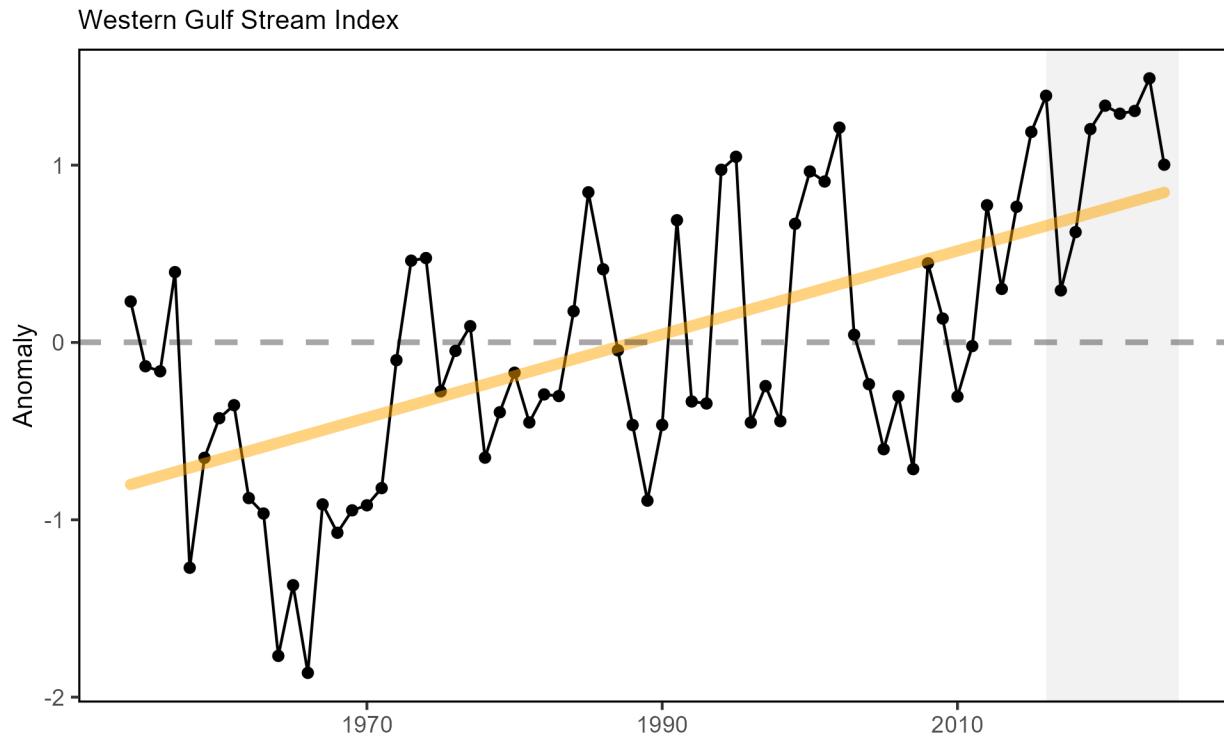


Figure 30: Index representing changes in the location of the western (between 64 and 55 degrees W) Gulf Stream north wall (black). Positive values represent a more northerly Gulf Stream position, with increasing trend (orange).

Changes in ocean temperature and circulation alter habitat features such as the Middle Atlantic Bight [Cold Pool](#), a band of relatively cold near-bottom water from spring to fall over the northern MAB. The cold pool represents essential fish spawning and nursery habitat, and affects fish distribution and behavior. The cold pool has been getting warmer and its areal extent has been shrinking over time (Fig. ??). In 2024, however, the cold pool index and extent were near the long-term average, likely due to the influx of Labrador Slope and Scotian Shelf waters into the system.

**Future Considerations** Distribution shifts caused by changes in thermal habitat and ocean circulation are likely to continue as long as long-term trends persist. Episodic and short-term events (see [2024 Highlights](#)) may increase variability in the trends, however species distributions are unlikely to reverse to historical ranges in the short term. Increased mechanistic understanding of distribution drivers is needed to better understand future distribution shifts: species with high mobility or short lifespans react differently from immobile or long lived species.

Long-term oceanographic projections forecast a temporary pause in warming over the next decade due to internal variability in circulation and a southward shift of the Gulf Stream. Near-term forecasts are being evaluated to determine how well they are able to predict episodic and anomalous events that are outside of the long-term patterns.

Adapting management to changing stock distributions and dynamic ocean processes will require continued monitoring of populations in space and evaluating management measures against a range of possible future spatial distributions. Processes like the [East Coast Climate Scenario Planning](#), and subsequent formation of the [East Coast Climate Coordination Group](#), can help coordinate management.

#### Risks to managing seasonally

The effectiveness of seasonal management actions (fishing seasons or area opening/closing) depends on a proper alignment with the seasonal life cycle events (phenology) of fish stocks (e.g. migration timing and spawning). Changes in the timing of these biological cycles can reduce the effectiveness of management measures if not accounted for. The

timing of seasonal patterns can also change the interactions between fisheries and non-target species thus influencing the amount of bycatch and the availability of species to surveys.

**Indicators: Timing shifts** Spawning timing is shifting earlier for multiple stocks, including haddock and yellowtail flounder. Spawning of both haddock stocks occurred earlier in the year, as indicated by more resting (post-spawning) stage fish in the 2010s as compared to earlier in the time series (Fig. 31). The northern (Cape Cod/GOM) yellowtail flounder stock shows earlier active spawning in recent years with a decline in pre-spawning resting females. The recent increase in resting females in the southern (SNE) stock also indicates a shift to earlier spawning (i.e. more post-spawn fish). Yellowtail flounder spawning is related to bottom temperature, week of year, and decade sampled for each of the three stocks.

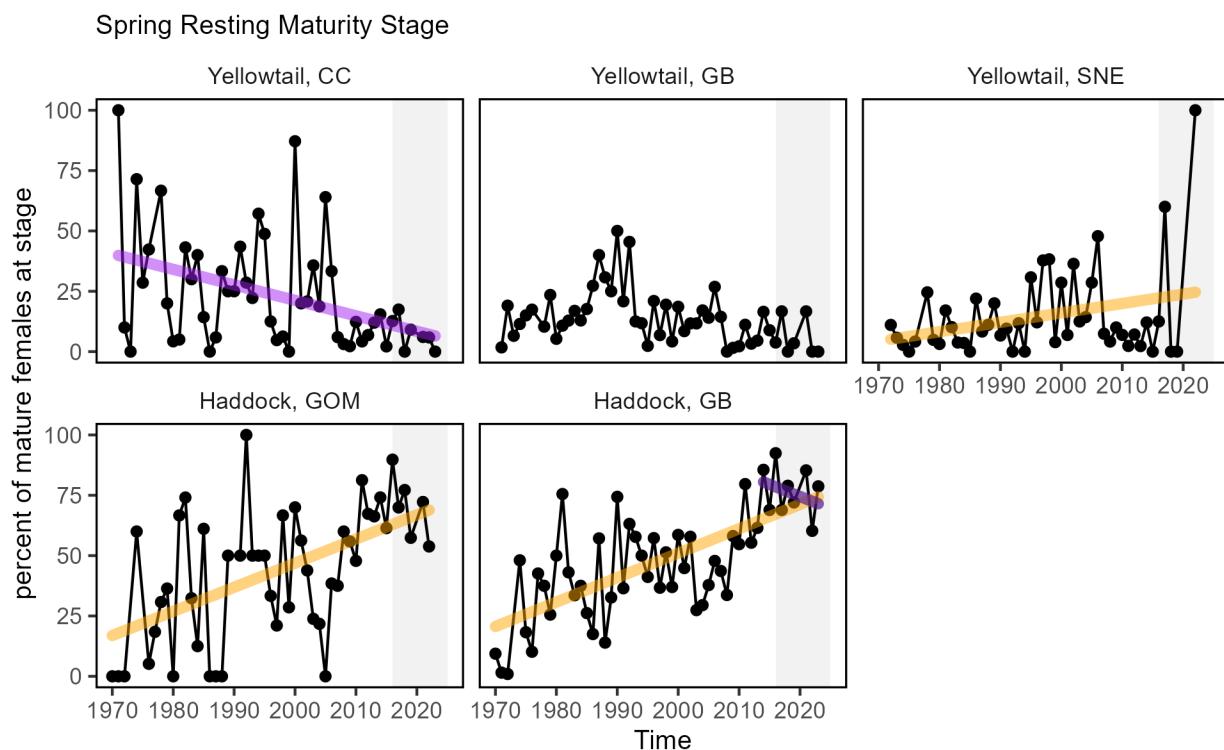


Figure 31: Percent resting stage (non-spawning) mature female fish (black) from spring NEFSC bottom trawl survey with significant increases (orange) and decreases (purple) from two haddock and three yellowtail flounder stocks: CC = Cape Cod Gulf of Maine, GOM = Gulf of Maine, GB = Georges Bank, SNE = Southern New England.

Migration timing of some tuna and large whale migrations has changed. An analysis of recreational fishing data between 2019 and 2022 identified multiple shifts in important HMS species. For example, Bigeye tuna were caught 50 days earlier; small and large bluefin tuna were caught 38 and 80 days earlier respectively in Massachusetts; and blue marlin in New York were caught 27 days earlier. In Cape Cod Bay, peak spring habitat use by right and humpback whales has shifted 18-19 days later over time.

Understanding whether seasonal patterns are changing for stocks requires regular observations throughout the year. For example, baseline work on [cetacean presence in Southern New England](#) shows different seasonal use patterns for whale and dolphin species. Despite the importance of understanding seasonal patterns, we have few indicators that directly assess timing shifts of species. We plan on incorporating more indicators of timing shifts and phenology in future reports.

**Drivers:** The drivers of timing shifts in managed stocks are generally coupled to shifts in environmental or biological conditions, since these can result in changes in habitat quality or food availability within the year. Changes in the

timing of fall phytoplankton blooms and seasonal shifts in zooplankton communities are indicators of changes in seasonal food availability to stocks.

Along with the overall warming trends in New England, ocean summer conditions have been lasting longer (Fig. 32) due to the later transition from warm summer conditions to cooler fall temperatures. These transition dates relate how daily temperatures compare to the seasonal norm. Changes in the timing of seasonal environmental cycles can alter biological processes (migrations, spawning, etc.) that are triggered by seasonal events.

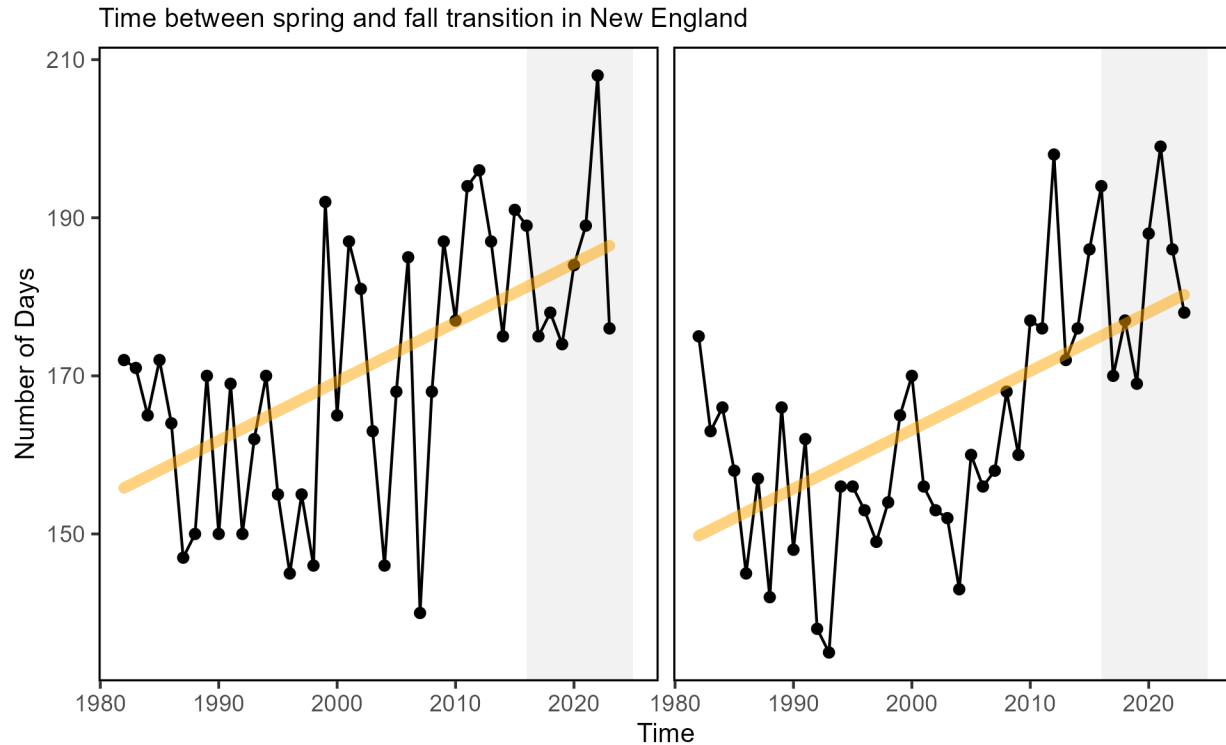


Figure 32: Ocean summer length: the annual total number of days between the spring thermal transition date and the fall thermal transition date.

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The Middle Atlantic Bight [Cold Pool](#) is a summer to early fall feature that creates seasonally suitable habitat for many species, including some managed by the NEFMC. Since the mid-2000s, the Cold Pool has persisted for a shorter portion of the year (Fig. ??). In 2024, however, all Cold Pool indices were near the long-term average and likely related to the influx of northern waters into the system (see 2024 highlights section). A change in the timing of the autumn breakdown of the Cold Pool may impact the recruitment of species that rely on it for seasonal cues and habitat. Southern New England-Mid Atlantic yellowtail flounder recruitment and settlement are related to the strength of the MAB Cold Pool (a factor of extent and persistence). The correlation of pre-recruit settlers to the Cold Pool is thought to represent a bottleneck in yellowtail flounder life history, whereby a local and temporary increase in bottom temperature can negatively impact the survival of settlers. Including the effect of Cold Pool variations on yellowtail recruitment reduced retrospective patterns and improved predictive skill in a stock assessment model. This is especially important given the long-term decline in the duration of the Cold Pool.

The seasonal timing of [phytoplankton](#) blooms shows a tendency towards an increased fall bloom over time in the GOM and GB, with chlorophyll significantly increasing October and November (GB) and December (GOM) (Fig. ??). January concentrations are trending higher since the late 1990s, but they are still below the mean spring and fall bloom values.

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**Future Considerations** For species reliant on environmental processes to dictate the timing of their behavior (e.g. phytoplankton bloom timing, thermal transition, or the duration of the cold pool), it is possible that some changes are episodic and have interannual variability, while other timing effects can change on scales of years to decades. Other species may rely on the general seasonal succession of environmental drivers (e.g. the timing of the fall turnover) to cue biological processes, and these long-term trends are unlikely to reverse in coming years. Such timing shifts in migration or spawning may continue. Management actions that rely on effective alignment of fisheries availability and biological processes should continue to evaluate whether prior assumptions on seasonal timings still hold, and new indicators should be developed to monitor timing shifts for stocks.

### **Risks to setting catch limits**

The efficacy of short-term stock projections and rebuilding plans rely on accurate understanding of processes affecting stock growth, reproduction, and natural mortality. These biological processes are often driven by underlying environmental change. When observed environmental change occurs, there is a risk that established stock-level biological reference points may no longer reflect the current population and short-term projections become more uncertain.

## **07\_risk\_setting\_catch\_limits\_newengland.Rmd**

**Indicators: Fish productivity and condition shifts** Indicators of [fish productivity](#) are derived from observations (surveys) or models (stock assessments). With the exception of two years (2006 and 2013), fish productivity has been below the long-term average in the Gulf of Maine since the early 2000s, as described by the small-fish-per-large-fish anomaly indicator (derived from NEFSC bottom trawl survey)(Fig. 33). This decline in fish productivity is also shown by a similar analysis based on stock assessment model outputs (recruitment per spawning stock biomass anomaly). Other signs of changing productivity in New England are the declines in [common tern chicks](#) per nest (Fig. 34) and continued low returns of hatchery [Atlantic salmon](#)(Fig. 35) despite short-term increases in adult salmon numbers.

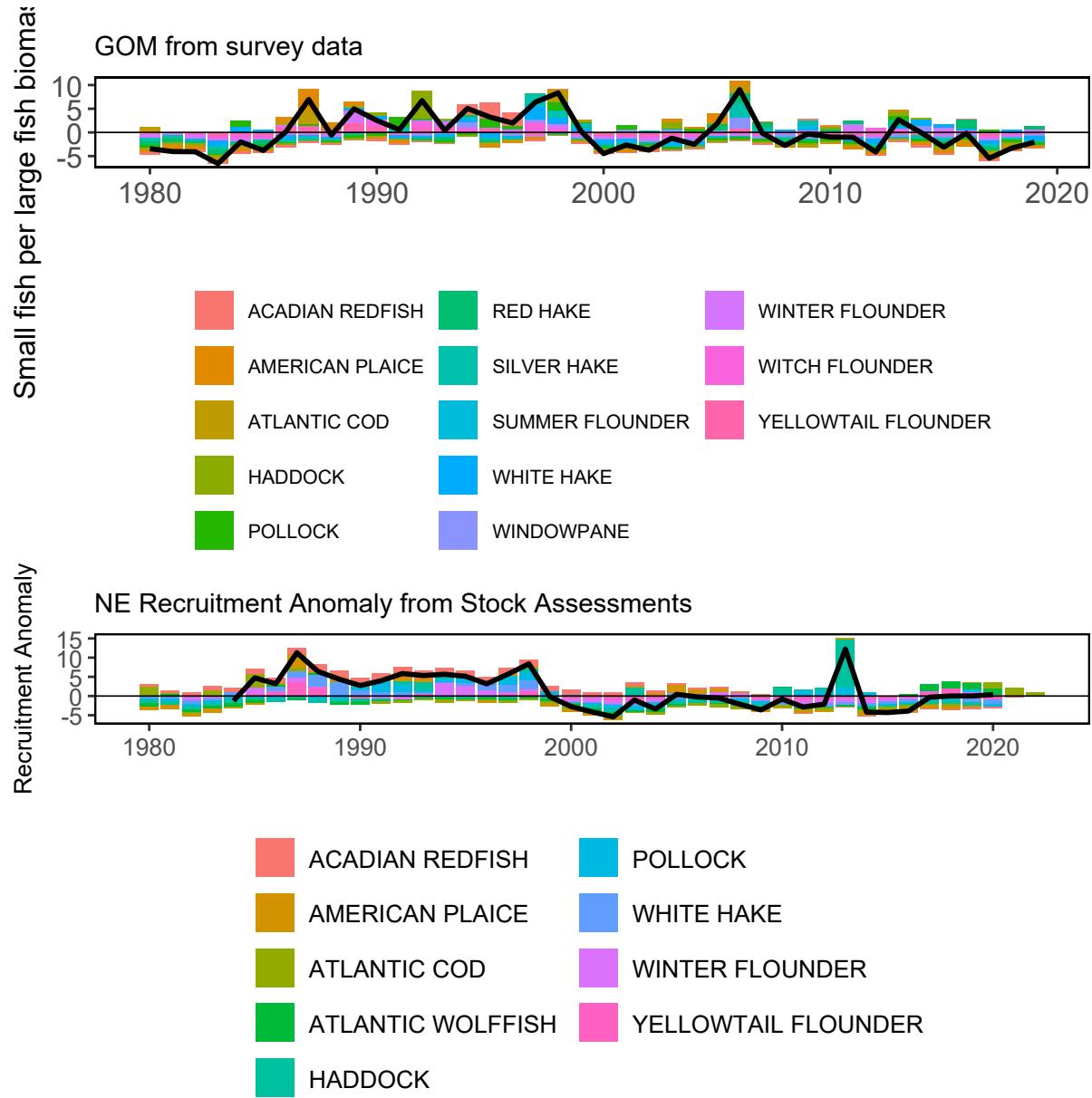


Figure 33: Fish productivity measures. Top: Small-fish-per-large-fish survey biomass anomaly in the Gulf of Maine. Bottom: assessment recruitment per spawning stock biomass anomaly for stocks managed by the New England Fishery Management Council region. The summed anomaly across species is shown by the black line, drawn across all years with the same number of stocks analyzed.

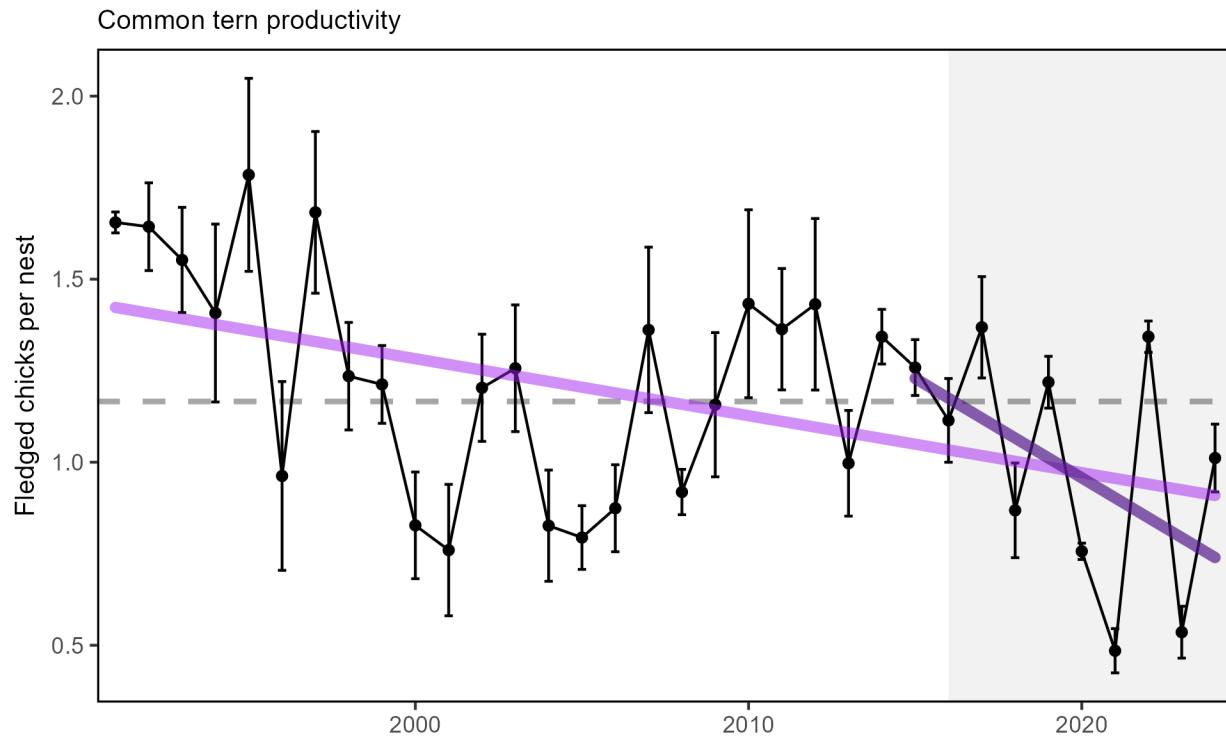


Figure 34: Productivity of Common terns in the Gulf of Maine.

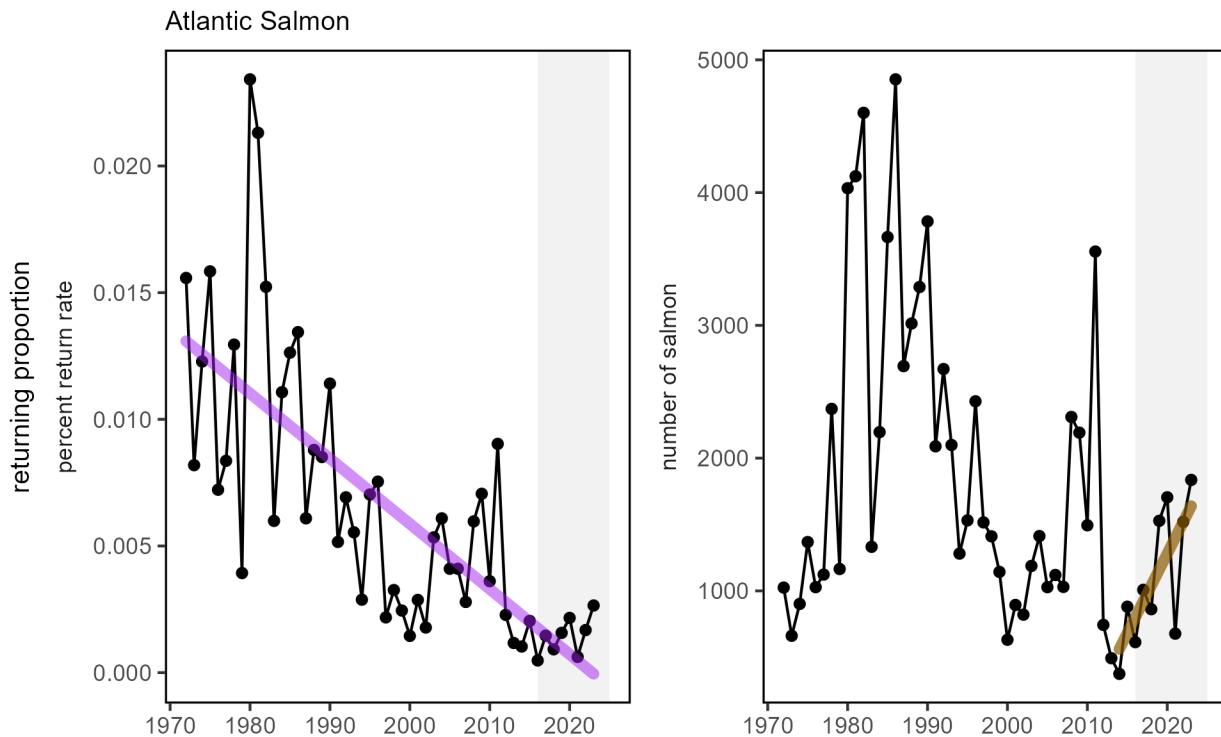


Figure 35: Return rate proportions and abundance of Atlantic salmon.

The health of individual fish (i.e., fish condition) can contribute to population productivity through improved growth, reproduction, and survival. [Fish condition](#) in the Gulf of Maine and Georges Bank regions were generally good prior to 2000, poor from 2001-2010 (concurrent with declines in fish productivity, Fig. 33), and a mix of good and poor since 2011. In 2024, fish condition was poor for most species on both Georges Bank and in the Gulf of Maine (Fig. 36). Preliminary analyses show that changes in temperature, zooplankton, fishing pressure, and population size influence the condition of different fish species.

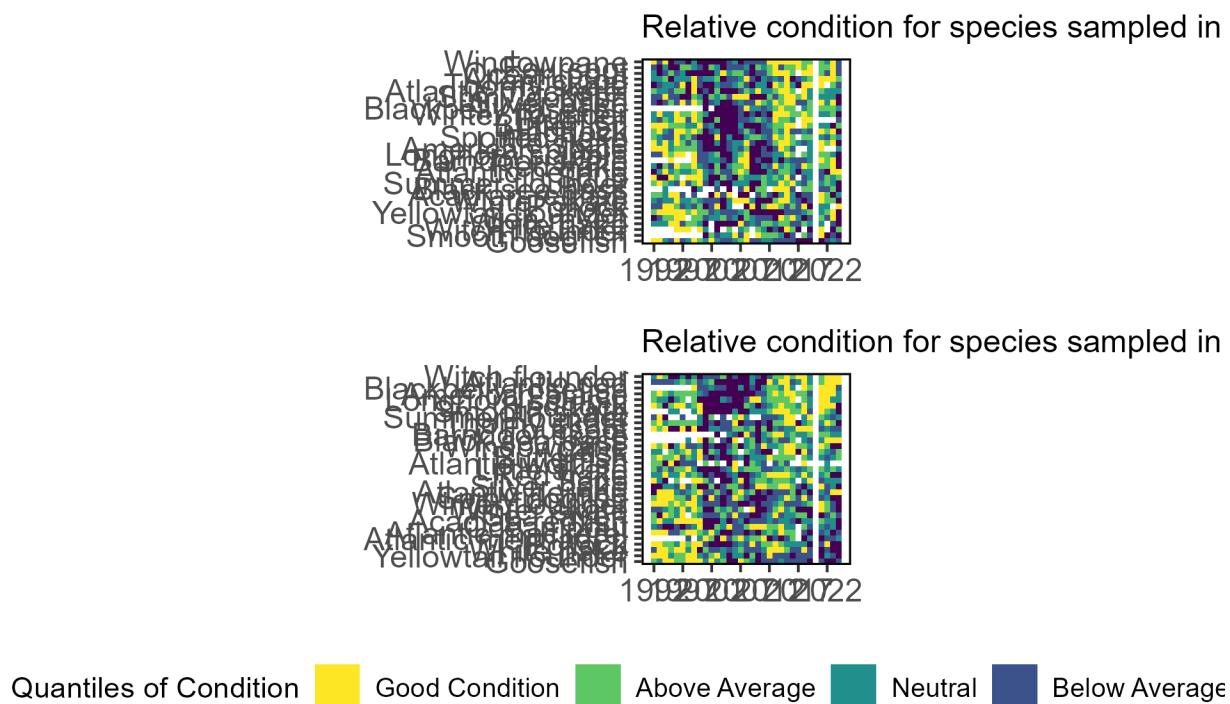


Figure 36: Return rate proportions and abundance of Atlantic salmon.

**Drivers** Fish productivity and condition are affected by increasing metabolic demands from increasing temperature, combined with changes in the availability and quality of prey. Long-term environmental trends and episodic extreme temperatures, ocean acidification, and low oxygen events represent multiple stressors that can affect growth rates, reproductive success and recruitment, and cause mortality.

**Biological Drivers: Forage quality and abundance** The amount of forage fish available in the ecosystem combined with the energy content of the forage species determines the amount of energy potentially available to predators in the ecosystem. Changes in the forage fish base can drive managed and protected species production and condition.

The [energy content](#) of juvenile and adult forage fish as prey is related to forage fish growth and reproductive cycles, as well as environmental conditions. The energy content of Atlantic herring was estimated to be highest of any forage species in the 1980s and 1990s, based on very small numbers of fish. Most observations from the NEFSC trawl surveys are below the previous estimates (Fig. 37). However, a recent study that included samples from additional sources indicated herring energy density peaked in summer, with some values closer to the historic estimates. Silver hake, longfin squid (*Loligo* in figure), and shortfin squid (*Illex* in figure) remain lower than previous estimates.

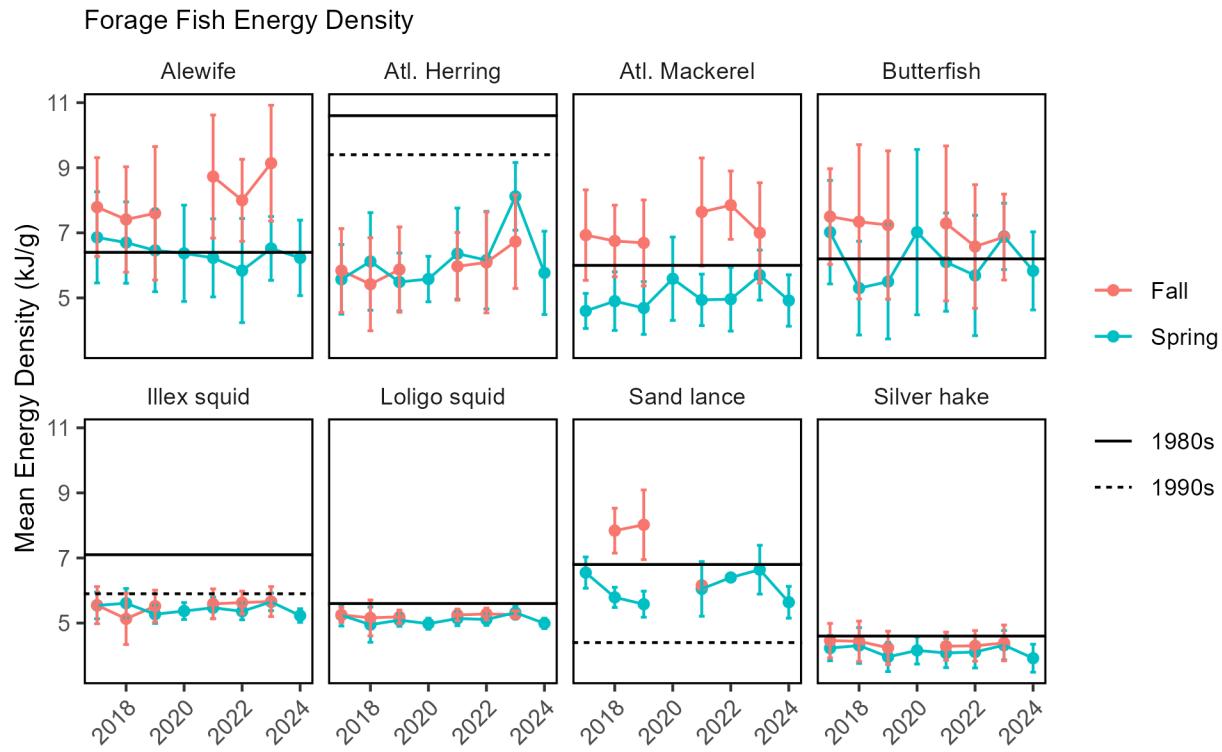


Figure 37: Forage fish energy density mean and standard deviation by season and year, compared with 1980s (solid line; Steimle and Terranova 1985) and 1990s (dashed line; Lawson et al. 1998) values.

Changes in the overall abundance of forage fish can influence managed species productivity as it relates to changes in food availability. New England [fall forage biomass](#) is stable with long-term increases in the spring GOM (Fig. 38). Forage biomass was highest during fall in the 1980s.

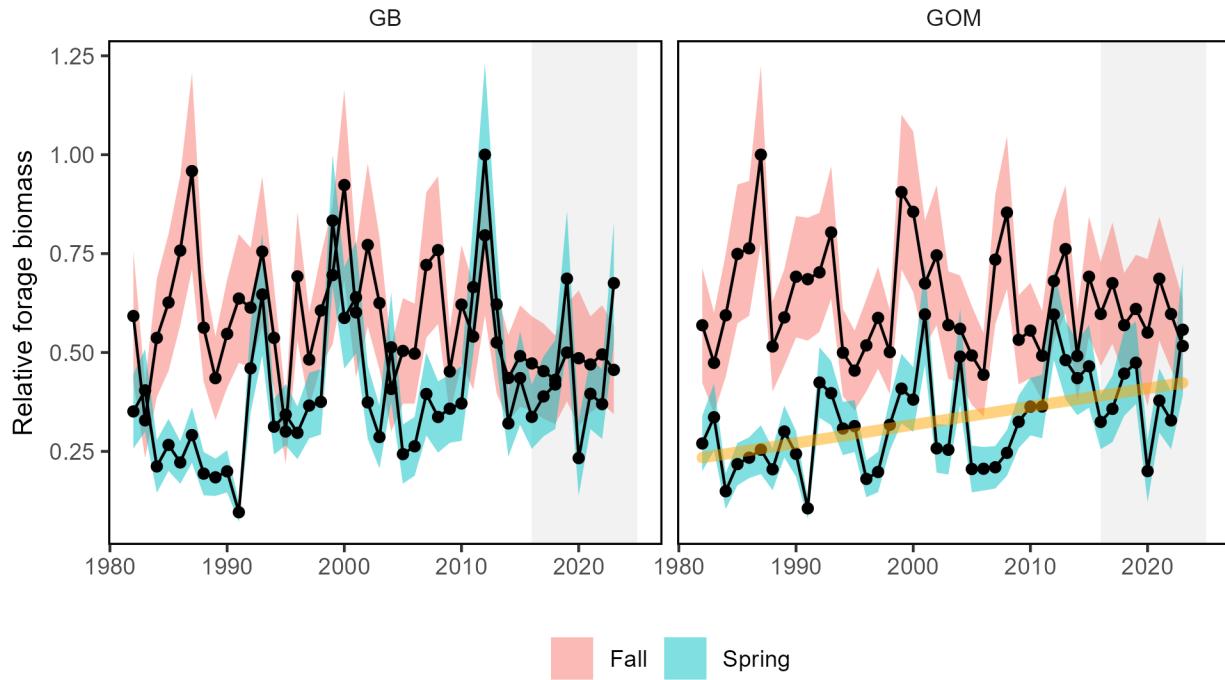


Figure 38: Forage fish index in GB (left) and GOM (right) for spring (blue) and fall (red) surveys. Index values are relative to the maximum observation within a region across surveys.

[Benthic invertebrates](#) are extremely important forage for some managed species (e.g. flatfish, juvenile cod and haddock) Macrofauna indices show long term declines in spring. In contrast, megafauna indices show long-term increases during the fall in both GB and GOM (Fig. 39).

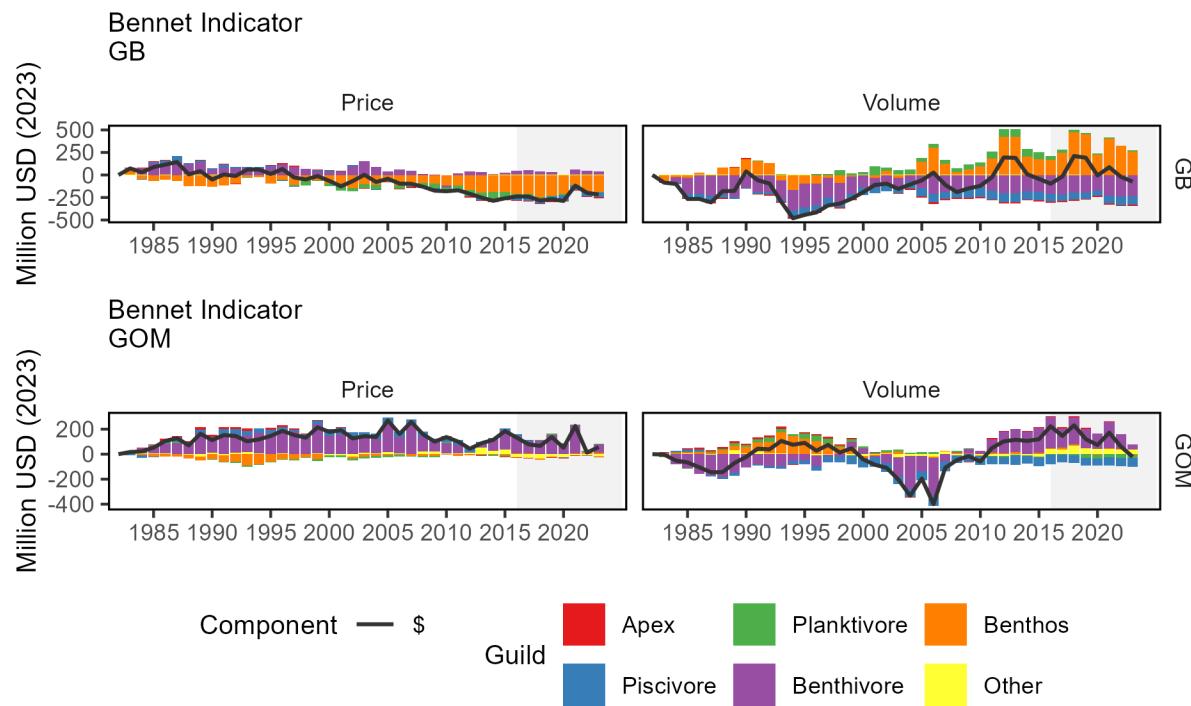


Figure 39: Changes in benthos abundance in New England for megabenthos (top) and macrobenthos (bottom).

**Biological Drivers: Lower trophic levels** Phytoplankton are the foundation of the marine food web and are the primary food source for zooplankton and filter feeders such as shellfish. Multiple environmental and oceanographic drivers affect the abundance, composition, spatial distribution, and productivity of phytoplankton. While changes in phytoplankton productivity could affect fish productivity (including forage), there is no clear long-term trend in New England total primary production (Fig. 14).

Zooplankton communities in New England have been changing in composition. There has been an increase in GOM small bodied copepods, a recent decline in large bodied copepods in GB, and long-term increases in euphausiids in GB and GOM (Fig. 40). A changing mix of zooplankton prey can impact forage fish energy content and abundance, as well as the prey field of filter feeding whales.

Since 2010, the abundance of the lipid-rich older stages of *Calanus finmarchicus* in the GOM has declined. Observations from a fixed time series station in Wilkinson Basin indicate that *Calanus* seasonal abundance in late summer-winter between 2020-2022 has declined to 20-40% of its population level in 2005-2008 but has rebounded to 60-70% in 2024 due to the presence of colder water (Fig. 41). However, spring abundances are still the same as 15-20 years ago. The seasonal differences in abundance change reflect differences in influence of primary seasonal drivers:

1. *Calanus* reproductive output is tied to phytoplankton availability in late winter/early spring.
2. Gulf of Maine source waters drive *Calanus* supply (high *Calanus* in Scotian Shelf/Labrador shelf water (LSW) and less in warm slope water (WSW))
3. Predation is likely higher with warmer temperatures

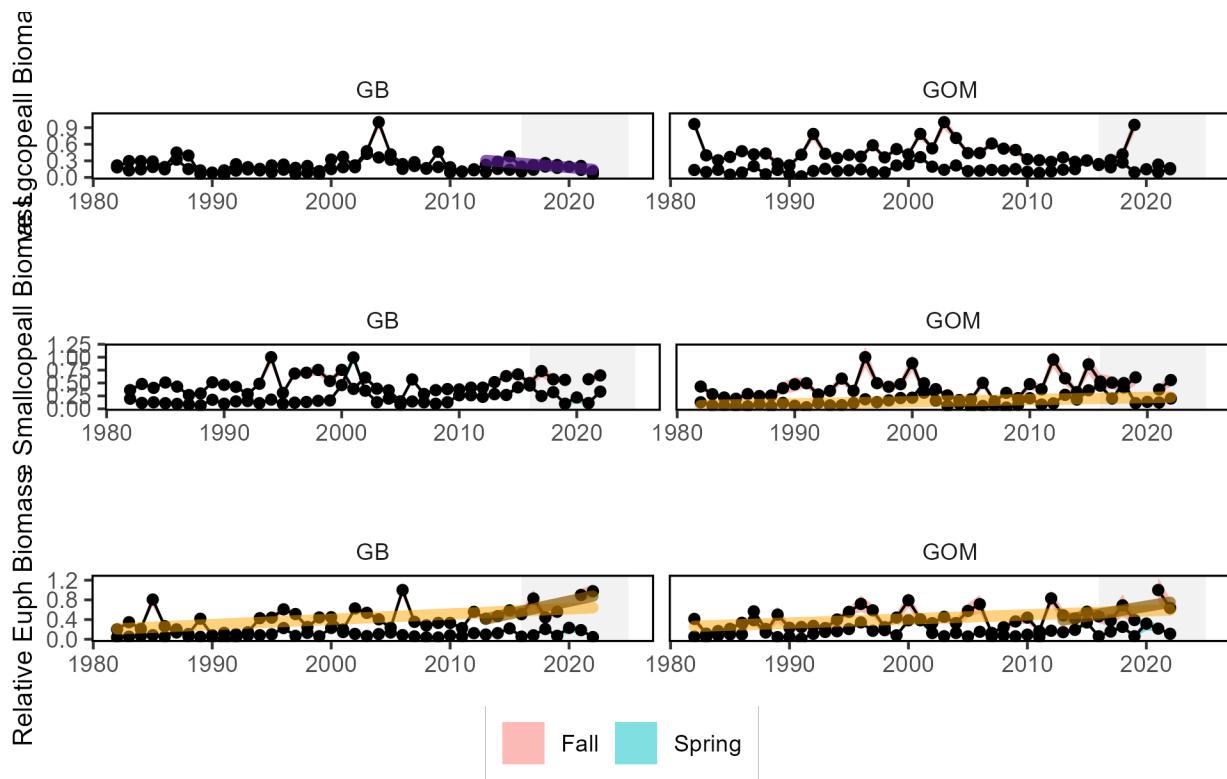


Figure 40: Georges Bank (GB) and Gulf of Maine (GOM) abundance anomalies three dominant zooplankton (*Calanus finmarchicus*, *Calanus typicus*, and *Pseudocalanus spp.*).

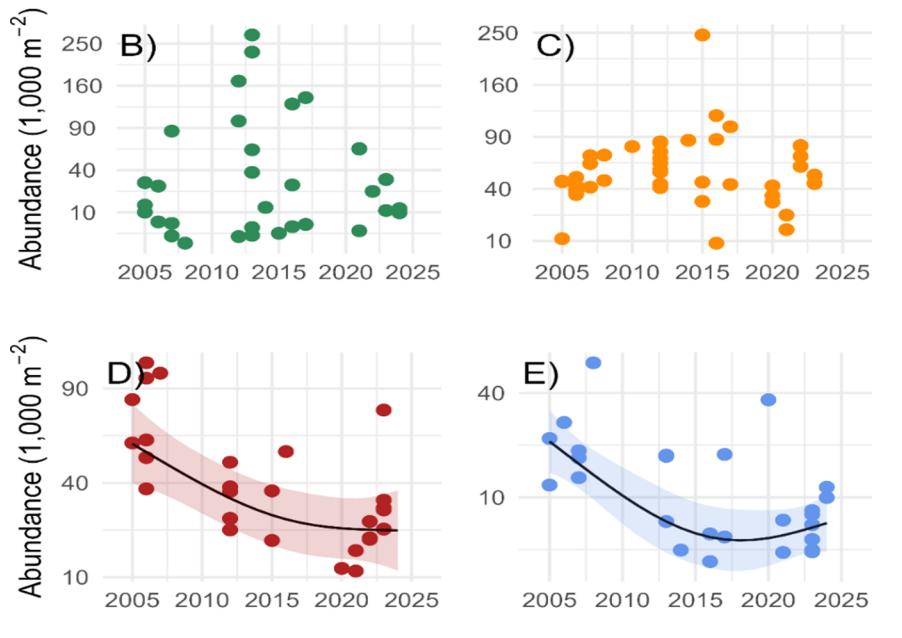


Figure 41: Abundance ( $\text{no m}^{-2}$ ) of *\*C. finmarchicus\** C3-C6 estimated from  $200\mu$  vertical ring net tows. Individual data with fitted lines. Data from 2005-2010 (circles), 2011-2021 (triangles), and 2022-2024 (squares). WBTS station seasonal abundance time series for B) spring, C) summer, D) fall, E) winter. Vertical lines denote season boundaries. If the seasonal abundance time series is significant, GAM predictions are calculated with day of year set to 1, 100, 200, and 300 for winter, spring, summer, and fall, respectively.

**Environmental Drivers** Fish production can also be directly related to the prevailing environmental conditions by altering metabolic (growth) and reproductive processes. Many species possess thermal tolerances and can experience stressful or lethal conditions if temperatures exceed certain levels. Extreme temperature at both the [surface](#) (Fig. 29) and [bottom](#) can exceed [thermal tolerance](#) limits for some fish. For example, 2012 had among the warmest surface and bottom temperatures (GB) in New England. A large proportion of the Georges Bank and Mid-Atlantic regions had bottom temperatures above the  $15^\circ\text{C}$  thermal tolerance for most groundfish, with some days in the Mid-Atlantic exceeding the  $24^\circ\text{C}$  potential mortality limit (Fig. 42).

In 2024, only one [surface marine heatwave](#) occurred throughout the entire U.S. Northeast Shelf due to the cooler ocean conditions observed in the region. This surface marine heatwave occurred in the Gulf of Maine starting on May 29th, peaking on June 7th, and lasting 12 days. This marine heatwave was not within the top 10 on record in terms of intensity.

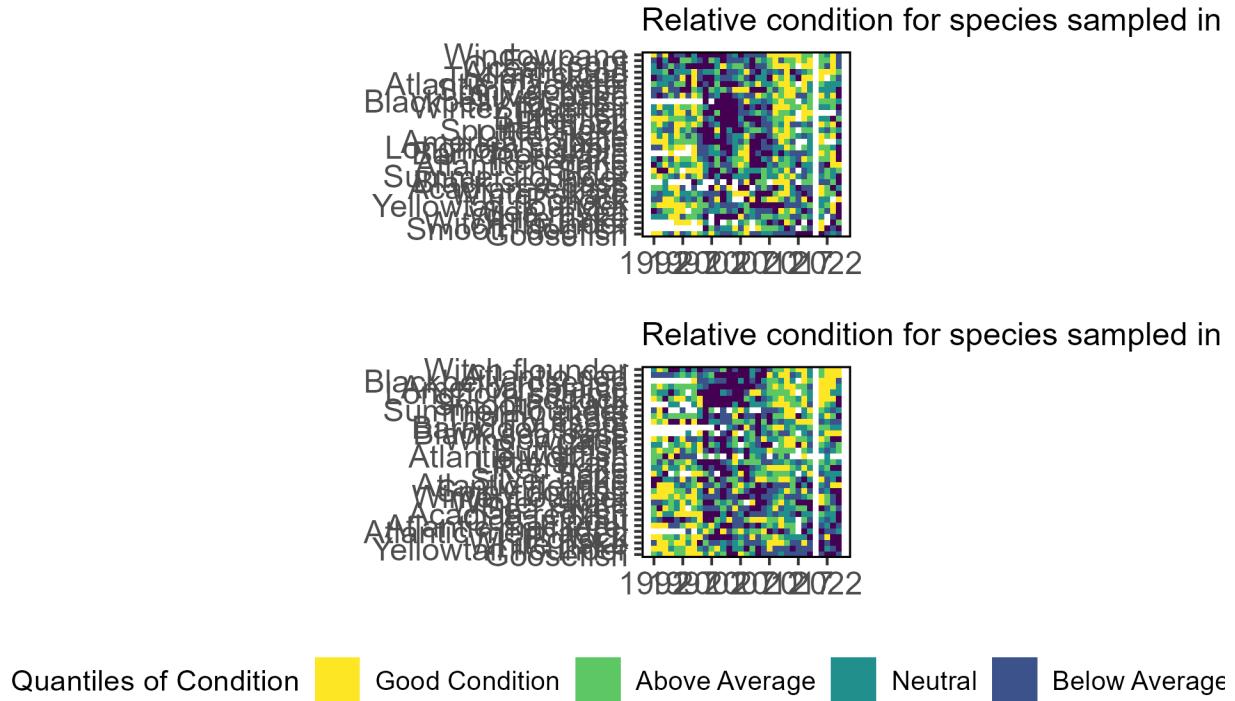


Figure 42: The number of days in 2024 where bottom temperature exceeds 15°C (left) and 24°C (right) based on the GLORYS 1/12 degree grid.

**Ocean acidification** risks vary among species and include reduced survival, growth, reproduction, and productivity, where high OA risk indicates potential negative effects to species. OA risk can also be heightened during colder conditions due to increased  $CO_2$  absorption by the water or by transport of high  $CO_2$  water masses (see [highlights section](#)). Higher OA risk conditions were observed for Atlantic sea scallop and longfin squid in Long Island Sound and the nearshore and mid shelf regions of the New Jersey shelf during summers of 2016, 2018, 2019, 2023, and 2024 (Fig. 43). The OA indicator observed on the Mid-Atlantic coastal shelf during summer 2024 was the most extreme recorded when compared to all of the years sampled (since 2007).

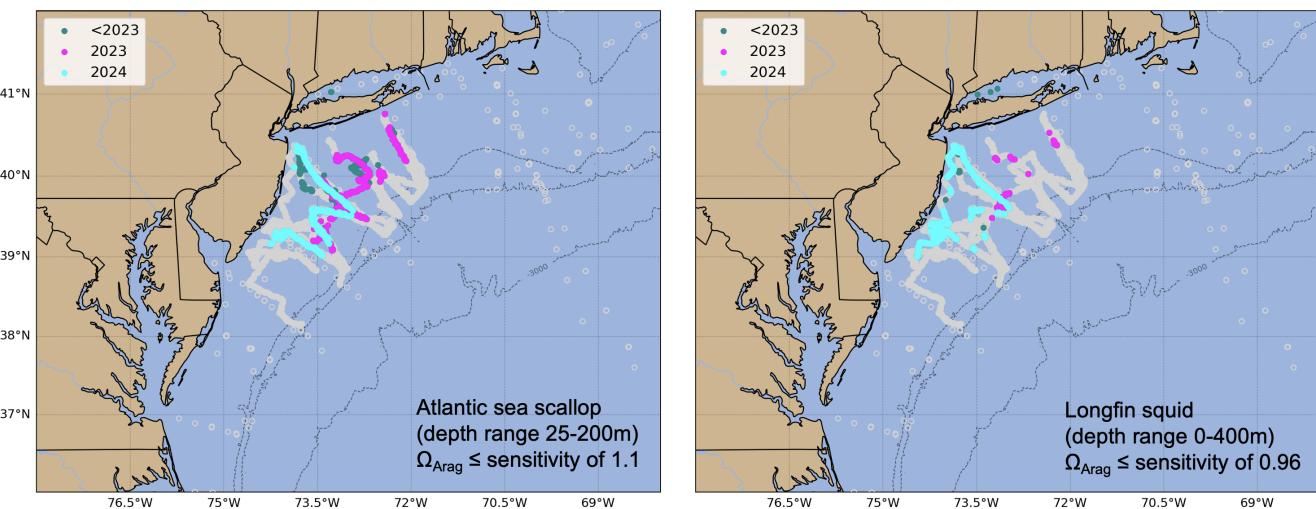


Figure 43: Locations where bottom aragonite saturation state ( $\Omega_{Arag}$ ; summer only: June-August) were at or below the laboratory-derived sensitivity level for Atlantic sea scallop (left panel) and longfin squid (right panel) for the time periods 2007-2022 (dark cyan), 2023 only (magenta) and 2024 only (cyan). Gray circles indicate locations where bottom  $\Omega_{Arag}$  values were above the species specific sensitivity values..

Biological and oceanographic processes can affect the amount of oxygen present in the water column. During low oxygen (hypoxic) events, species' growth is negatively affected and very low oxygen can result in mortality. The duration and extent of hypoxic events is being monitored, but long-term shelf-wide observations are not yet available. However, [hypoxic events](#) were detected off the coast of New Jersey in 2023 and were potentially responsible for fish, lobster, and crab [mortalities](#). No hypoxic events were observed on the NE shelf in 2024.

**Drivers: Predation** The abundance and distribution of predators can affect both the productivity and mortality rates on managed stocks. Predators can consume managed species or compete for the same resources resulting in increased natural mortality or declining productivity, respectively. The northeast shift in some [highly migratory species](#) (Fig. 25) indicates a change in the overlap between predators and prey. Since we also observe distribution shifts in both managed and forage species, the effect of changing predator distributions alone is difficult to quantify.

[Gray seals](#) are fish predators with increasing populations in New England, however they are broad generalist feeders that do not generally target commercially-sized managed species. [Stock status](#) is mixed for Atlantic Highly Migratory Species (HMS) stocks (including sharks, swordfish, billfish, and tunas) occurring throughout the Northeast U.S. shelf. While there are several HMS species considered to be overfished or that have unknown stock status, the population status for some managed Atlantic sharks and tunas is at or above the biomass target, suggesting the potential for robust (or rebuilt) predator populations among these managed species. Stable predator populations suggest stable predation pressure on managed species, but increasing predator populations may reflect increasing predation pressure.

**Future Considerations** The processes that control fish productivity and mortality are dynamic, complex, and the result of the interactions between multiple system drivers. There is a real risk that short-term predictions in assessments and rebuilding plans that assume unchanging underlying conditions will not be as effective, given the observed ecological and environmental process changes documented throughout the report. Assumptions for species' growth, reproduction, and natural mortality should continue to be evaluated for individual species. With observations of system-wide productivity shifts of multiple managed stocks, more research is needed to determine whether regime shifts or ecosystem reorganization are occurring, and how this should be incorporated into management.

## **07\_risk\_setting\_catch\_limits\_newengland.Rmd**

**Future Considerations** The processes that control fish productivity and mortality are dynamic, complex, and are the result of the interactions between multiple system drivers. There is a real risk that short-term predictions in assessments and rebuilding plans that assume unchanging underlying conditions will be highly uncertain and not be as effective, given the observed change documented in the prior sections in both ecological and environmental processes. Assumptions for species' growth, reproduction, and natural mortality should continue to be evaluated for individual species. With observations of system-wide productivity shifts of multiple managed stocks, more research is needed to determine whether regime shifts or ecosystem reorganization are occurring, and how this should be incorporated into management.

### **Other Ocean Uses: Offshore Wind**

#### **Indicators: development timeline, revenue in lease areas, coastal community vulnerability**

All reported potential offshore wind projected development timelines and data are subject to change and have been based on BOEM Environmental Impact Statements. Offshore wind development schedule and areas are subject to change based on the Executive Order [Temporary Withdrawal of All Areas on the Outer Continental Shelf from Offshore Wind Leasing and Review of the Federal Government's Leasing and Permitting Practices for Wind Projects](#).

As of January 2025, 30 offshore [wind development](#) projects are proposed for construction over the next decade in the Northeast (timelines and project data for 2025 are based on the Maryland Offshore Wind Final Environmental Impact Statement, Appendix D). Offshore wind areas are anticipated to cover more than 2.3 million acres by 2030 in the Greater Atlantic region (Fig. 44). An additional 800,000 lease acres are proposed for development beyond 2030 and 17 million acres are identified by BOEM as designated planning areas (Fig. 45).

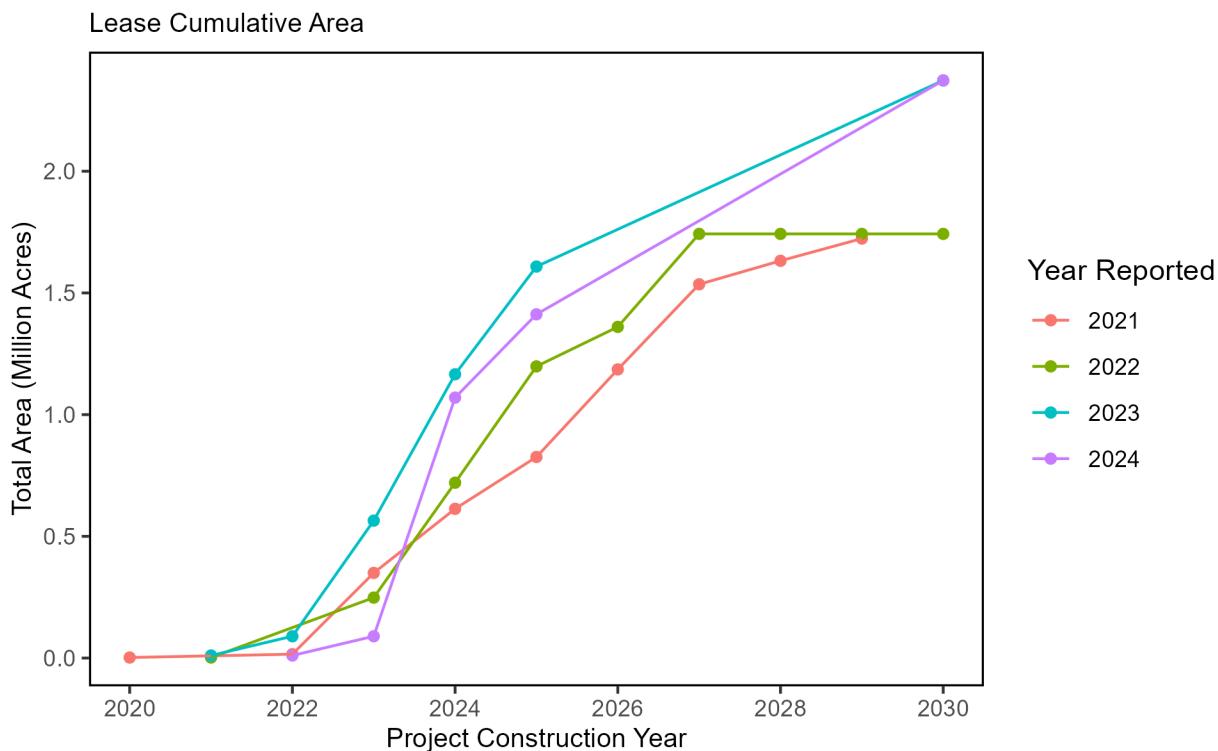


Figure 44: Total area proposed for wind development on the northeast shelf through 2030.

Just over 3,200 foundations and more than 12,000 miles of inter-array and offshore export cables are proposed to date (Fig. 45). Based on current timelines, the areas affected would be spread out such that it is unlikely that any

one region would experience full development at one time. Construction of three projects in Southern New England (Vineyard Wind, South Fork Wind Farm, and Revolution Wind) and two more in the Mid-Atlantic/New York Bight (Coastal Virginia Offshore Wind and Empire Wind 1) during 2024 affected fisheries managed by the Mid-Atlantic Fishery Management Council. It is likely that construction will begin on other projects in Southern New England and possibly the New York Bight during 2025 that will further affect regional fisheries.

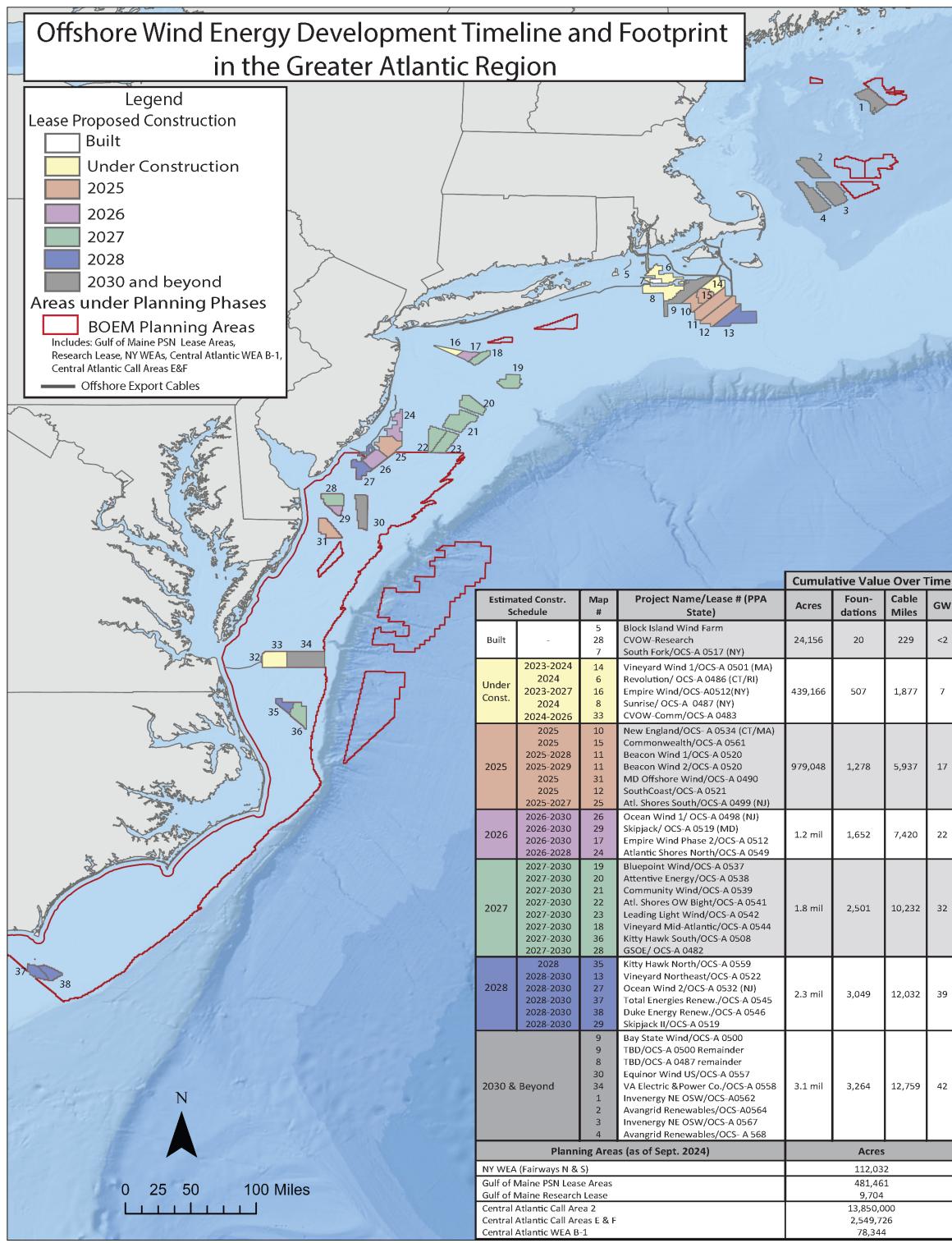


Figure 45: All Northeast Project areas by year construction ends (each project has a 2 year construction period).

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NewEngland: Fishery Revenue in Lease Areas

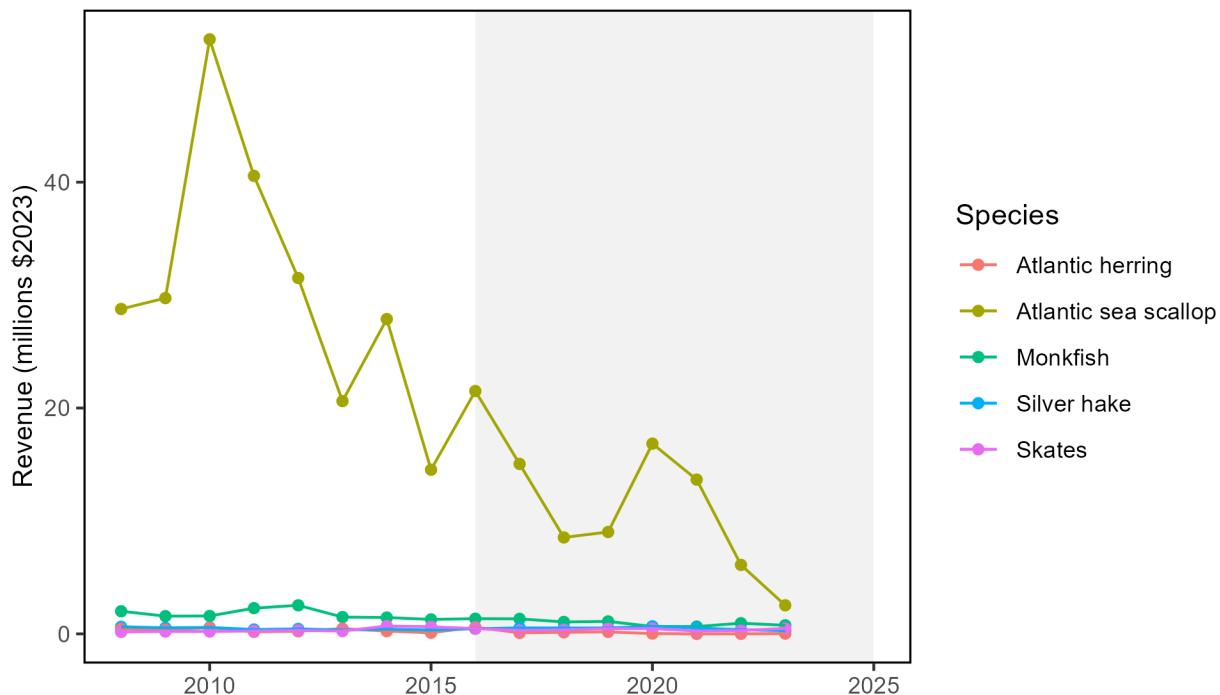


Figure 46: Fishery revenues from NEFMC managed species in the Wind energy lease areas.

Table 6: New England managed species Landings and Revenue from Wind Energy Areas. Skates includes barndoor, winter, clearnose, smooth, little, and general skates reported in logbooks. \*Less than a maximum of 50,000 lb was reported landed annually in wind energy lease areas for these species..

NEFMC, MAFMC, and ASMFC Managed Species	Maximum Percent Total Annual Regional Species Landings	Maximum Percent Total Annual Regional Species Revenue
Blueline Tilefish*	13	16
Atlantic Surfclam	16	15
Ocean Quahog	12	11
Black Sea Bass	9	10
Scup	8	9
Atlantic Mackerel	8	8
Chub Mackerel	15	8
Longfin Squid	8	8
Monkfish	9	8
Butterfish	8	7
Golden Tilefish	6	6
Summer Flounder	5	5
Bluefish*	4	4
Spiny Dogfish	2	3
Illex Squid	2	2

Social vulnerabilities of communities are priority concerns with offshore wind development and fisheries impacts in the Northeast, and the impacts of offshore wind development are expected to differentially [impact specific coastal](#)

**communities.** Additionally, impacts of offshore wind development may unevenly affect individual operators, with some permit holders deriving a much higher proportion of revenue from wind areas than the port-based mean.

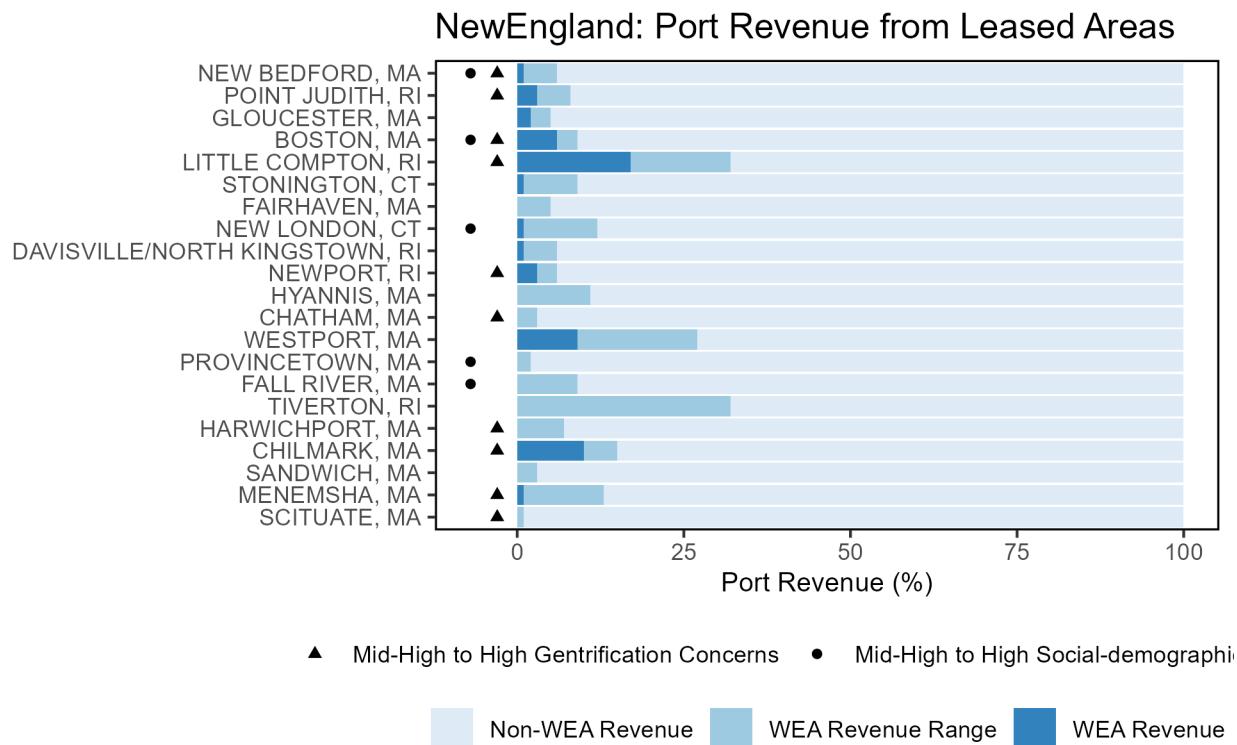


Figure 47: Percent of port fisheries revenue from Wind Energy Areas (WEA) in descending order from most to least port fisheries revenue from WEA.

For example, Little Compton, RI had a minimum of 17% and maximum of 32% overlap of wind energy revenue to the total port revenue between 2008-2023 (Fig. 47). BOEM reports that cumulative offshore wind development (if all proposed projects are developed) could have moderate impacts on low-income members of vulnerable communities who work in the commercial fishing and for-hire fishing industry due to disruptions to fish populations, restrictions on navigation, and increased vessel traffic as well as existing vulnerabilities of low-income workers to economic impacts.

Top fishing communities with high **socio-demographic concerns** such as New Bedford, MA and New London, CT should be considered in decision making to reduce the social and economic impacts and aid in the resilience and adaptive capacity of underserved communities. These two ports are also undergoing significant changes to support offshore wind development port infrastructure needs. Socio-demographic concerns also highlight communities where further resources are needed to reach underserved and underrepresented groups and create opportunities for, and directly involve, these groups in the decision-making process.

Some ports in the Mid-Atlantic land New England-managed species from wind areas as well. For the maximum percent value reported in each Mid-Atlantic port, the majority (at least 50% based on both value and pounds) of those landings were New England managed species within wind areas for nine communities (Fig. ??).

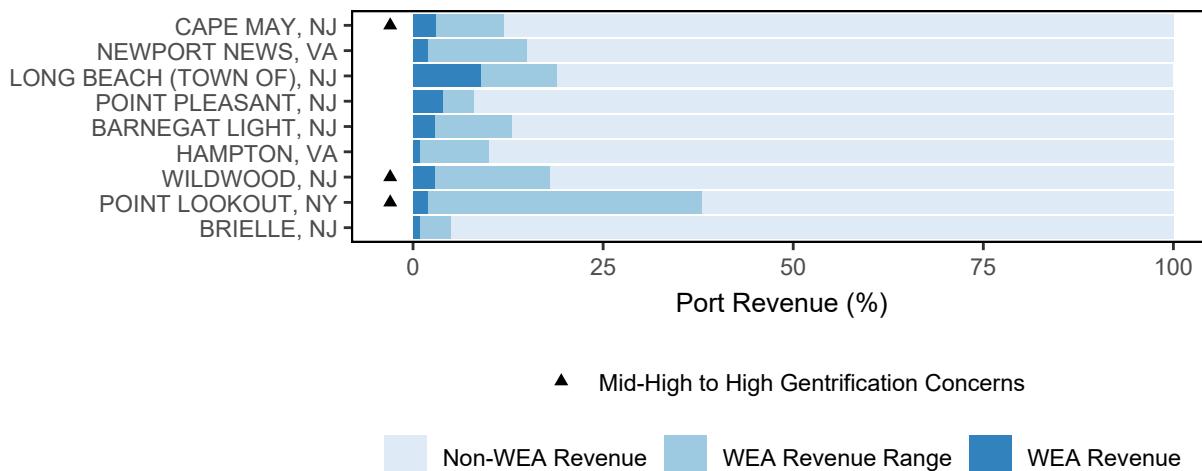


Figure 48: Percent of New England port revenue with majority MAFMC landings from Wind Energy Areas (WEA) in descending order from most to least port revenue from WEA.

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### **Implications**

Current plans for buildout of offshore wind in a patchwork of areas spreads the impacts differentially throughout the region (Fig. 45). Up to 12% of total average revenue for major New England commercial species in lease areas could be forgone or reduced and associated effort displaced if all sites are developed. Displaced fishing effort can alter historic fishing area, timing, and method patterns, which can in turn change habitat, species (managed and protected), and fleet interactions. Several factors, including fishery regulations, fishery availability, and user conflicts affect where, when, and how fishing effort may be displaced, along with impacts to and responses of affected fish species.

Planned development overlaps NARW mother and calf migration corridors and a significant foraging habitat that is used throughout the year (Fig. 49). Turbine presence and extraction of energy from the system could alter local oceanography and may affect right whale prey availability. For example, persistent foraging hotspots of right whales and seabirds overlap on Nantucket Shoals, where unique hydrography aggregates enhanced prey densities. Wind leases (OCS-A 0521 and OCS-A 0522) currently intersect these hotspots on the southwestern corner of Nantucket Shoals and a prominent tidal front associated with invertebrate prey swarms important to seabirds and possibly right whales. Proposed wind development areas also bring increased vessel strike risk from construction and operation vessels. In addition, there are a number of potential impacts to whales from pile driving and operational noise such as displacement, increased levels of communication masking, and elevated stress hormones.

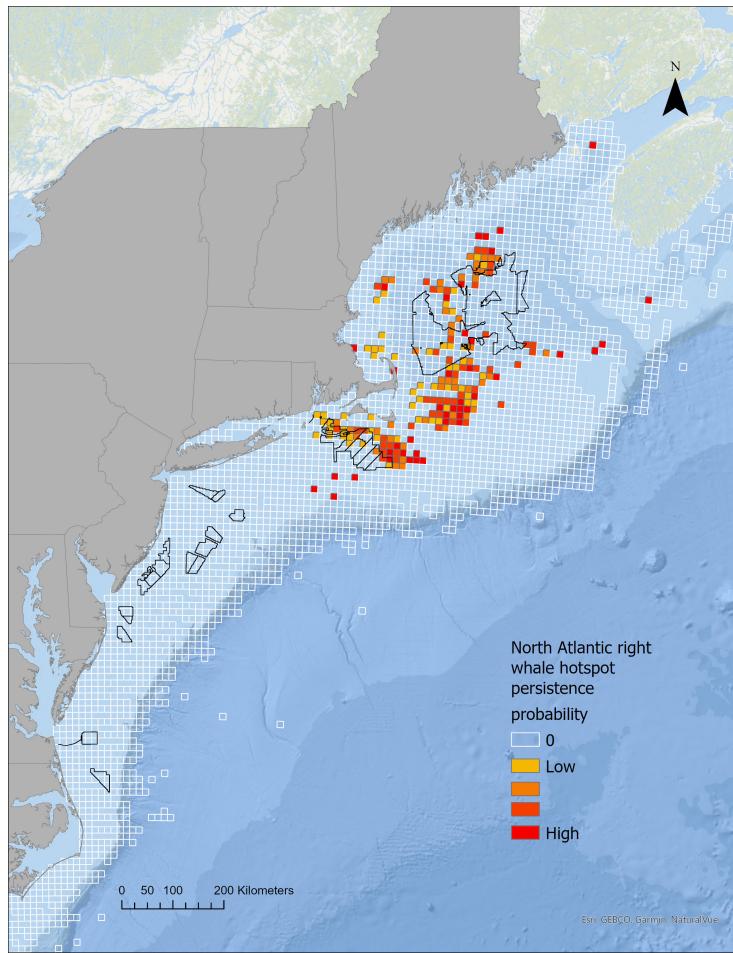


Figure 49: Northern Right Whale persistent hotspots (red shading) and Wind Energy Areas (black outlines).

Scientific data collection surveys for ocean and ecosystem conditions, fish, and protected species will be altered, potentially increasing uncertainty for stock assessments and associated management decision making.

The increase of offshore wind development can have both positive (e.g., employment opportunities) and negative (e.g., space-use conflicts) effects. Continued increase in coastal development and gentrification pressure has resulted in loss of fishing infrastructure space within ports. Understanding these existing pressures can allow for avoiding and mitigating negative impacts to our shore support industry and communities dependent on fishing. Some of the communities with the highest fisheries revenue overlap with offshore wind development areas that are also vulnerable to gentrification pressure are Point Judith and Newport, RI; and Boston and New Bedford, MA.

## 2024 Highlights

This section intends to provide a record of [noteworthy observations reported in 2024](#) across the Northeast U.S. region. The full ecosystem and fisheries impacts of many of these observations are still to be determined. They should, however, be noted and considered in future analyses and management decisions.

2024 global sea surface and air temperatures exceeded 2023 as the warmest year on record, but colder than average temperatures were observed in the Northeast U.S. Oceanographic and ecological conditions in the Northwest Atlantic were markedly different in 2024 compared to recent years.

**Northwest Atlantic Phenomena** Late 2023 and early 2024 observations indicate movement of cooler and fresher water into the Northwest Atlantic, although there are seasonal and local exceptions to this pattern. Anomalously cold (Fig. 50) and low salinity conditions were recorded throughout the Northeast Shelf and were widespread across the Slope Sea for much of the year. These cooler and fresher conditions are linked to the southward movement of the eastern portion of the [Gulf Stream](#) and possibly an increased influx of Labrador Slope and Scotian Shelf water into the system.

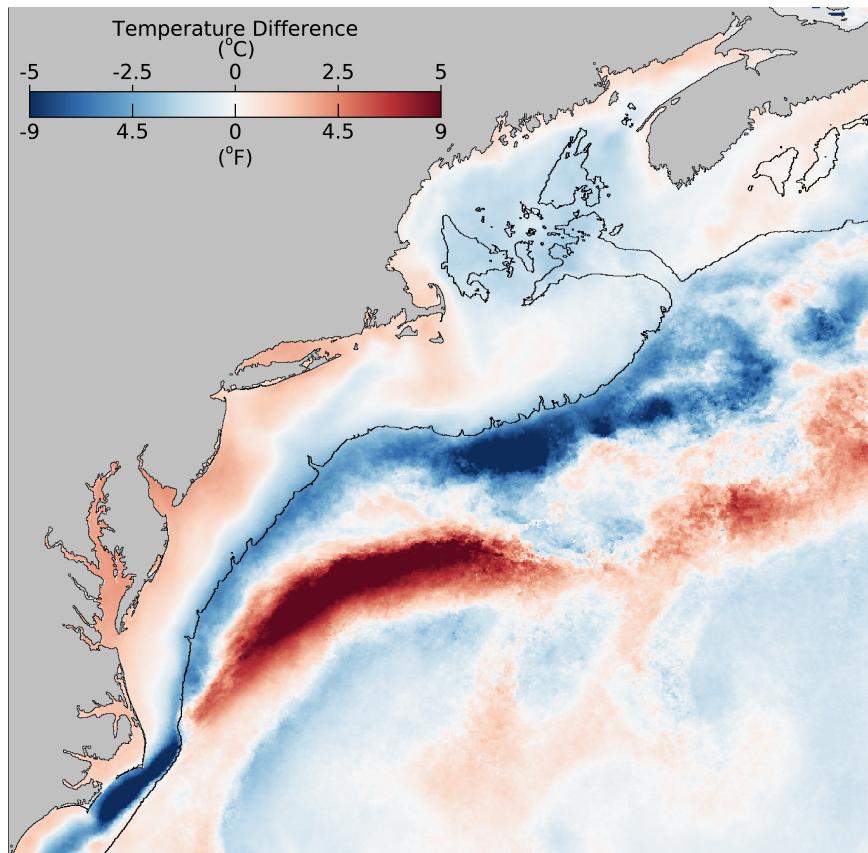


Figure 50: February 2024 sea surface temperature difference compared to the February 2000-2020 long-term mean from the NOAA Advanced Clear-Sky Processor for Ocean (ACSPO) Super-collated SST.

In 2023, Labrador Slope water accounted for more than 50% of the [source water](#) entering the Gulf of Maine through the Northeast Channel (Fig. 51); data are still being processed for 2024. Colder, fresher water detected deep in the Jordan Basin for the [first half of 2024](#) suggests an increased influx of Labrador Slope and Scotian Shelf water, which resulted in colder and fresher conditions throughout the Northwest Atlantic and contributed to the increased size and colder temperatures of the Mid-Atlantic [Cold Pool](#).

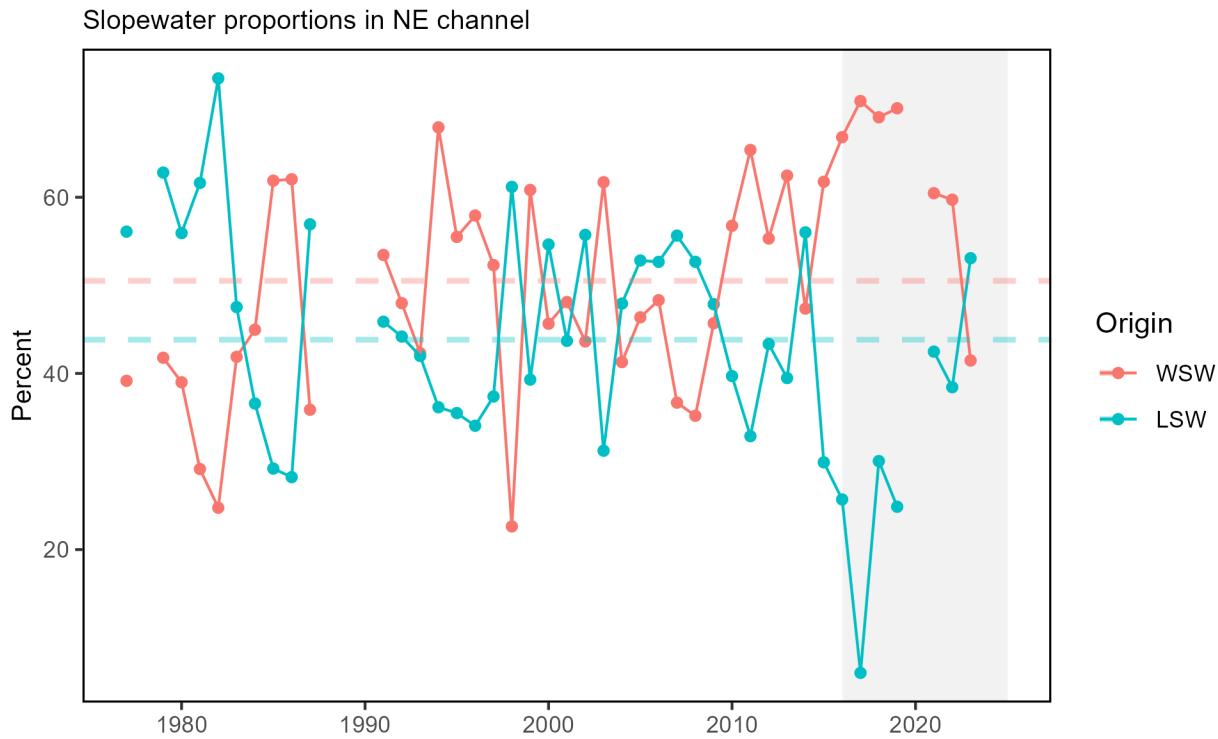


Figure 51: The proportion of Warm Slope Water (WSW) and Labrador Slope Water (LSW) enter the Gulf of Maine through the Northeast Channel from 1977 to 2023. The orange and teal dashed lines represent the long-term proportion averages for the WSW and LSW respectively.

**Northeast Shelf and Local Phenomena** The influx of the northern waters is likely linked to multiple observations across the Northeast Shelf including the uncommon presence of Arctic *Calanus* zooplankton species in the Gulf of Maine, delayed migration of many species, and redistribution of some species. Several members of the fishing community noted delayed migration of species into typical fishing grounds. In particular, they attributed the delayed migration of longfin squid, black sea bass, and haddock to the cooler water temperatures. Many also reported redistribution of some species. Specifically, pollock, bluefin tuna, Atlantic mackerel, longfin squid, bluefish, and bonito were observed in surprising or unusual locations. Some species, such as Atlantic mackerel, were reported outside of typical fishing grounds and in higher abundance compared to recent years. Anglers also reported good catches of red drum in Chesapeake Bay and record high (since 1995) numbers were observed at Poplar Island survey location.

In the summer, Chesapeake Bay recorded warm temperatures and low bottom water dissolved oxygen that resulted in less than suitable habitat for species such as striped bass and blue crabs. These poor conditions can affect their distribution, growth, and survival. Additionally, lower than average spring and summer salinity negatively impacted oyster hatchery operations and increased the area of available habitat for invasive blue catfish, potentially increasing predation on blue crabs and other important finfish species.

During the summer months there were multiple prolonged upwelling events that brought cold water to the surface off the New Jersey coast. There was also an atypical phytoplankton bloom south of Long Island in late June to early July 2024, possibly linked to an upwelling event (Fig. 52). The bloom was dominated by coccolithophores, which have an exoskeleton made up of calcium carbonate plates that can turn the water an opaque turquoise color. Large blooms of coccolithophores are unusual in this region, but they are not considered harmful and are grazed by zooplankton. Additionally, there were observations of multiple whale species aggregating near the Hudson Canyon between May and August.



Figure 52: An OLCI Sentinel 3A true color image with enhanced contrast captured on July 2, 2024. Coccolithophores shed their coccolith plates during the later stages of the bloom cycle, which results in the milky turquoise water color (Image credit: NOAA STAR, OCVView and Ocean Color Science Team).

Summer bottom [ocean acidification \(OA\)](#) risk in the Mid-Atlantic was the highest recorded since sampling began in 2007. High OA risk is measured as low aragonite saturation state( $\Omega$ ). Similarly, the winter/early spring [Gulf of Maine surface OA risk](#) was significantly above the climatological average and near the sensitivity levels for cod ( $\Omega < 1.19$ ) and lobster ( $\Omega < 1.09$ ) (Fig.53). These observations were likely driven by the greater volume of fresher, less-buffered Labrador Slope water entering the Gulf of Maine and Mid-Atlantic, as well as cooler conditions. The 2023 and 2024 high summer OA risk has increased the extent of potentially unfavorable habitat for Atlantic sea scallops ( $\Omega < 1.1$ ) and longfin squid ( $\Omega < 0.96$ ). Additionally, for the first time, high OA risk conditions were observed outside of summer (fall for both species and spring for Atlantic sea scallops).

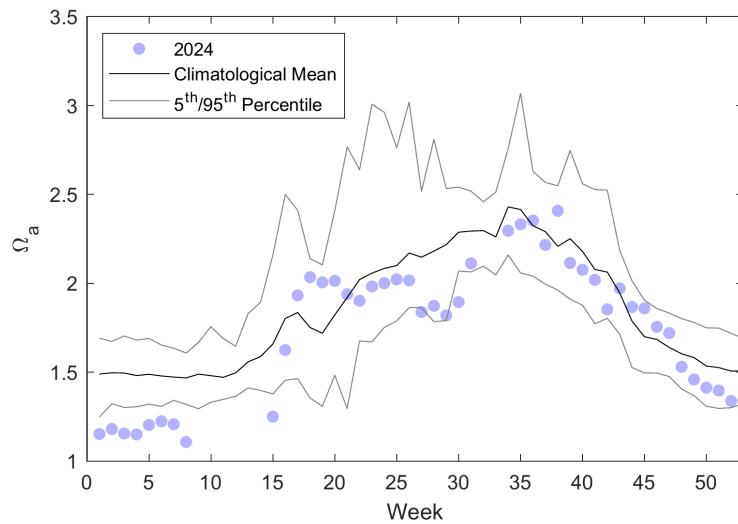


Figure 53: Weekly average surface aragonite saturation state measured at the long-term buoy location in the Gulf of Maine at 43.02 N and 70.54 W

In contrast to the documented die-off of scallops in the Mid-Atlantic Elephant Trunk region between the 2022 and 2023 surveys, in 2024 there was strong scallop recruitment in the southeastern portion of the Nantucket Lightship Area.

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## Document Orientation

The figure format is illustrated in Fig 54a. Trend lines are shown when slope is significantly different from 0 at the  $p < 0.05$  level. An orange line signifies an overall positive trend, and purple signifies a negative trend. To minimize bias introduced by small sample size, no trend is fit for  $< 30$  year time series. Dashed lines represent mean values of time series unless the indicator is an anomaly, in which case the dashed line is equal to 0. Shaded regions indicate the past ten years. If there are no new data for 2022, the shaded region will still cover this time period. The spatial scale of indicators is either coastwide, Mid-Atlantic states (New York, New Jersey, Delaware, Maryland, Virginia, North Carolina), or at the Mid-Atlantic Bight (MAB) Ecosystem Production Unit (EPU, Fig. 54b) level.

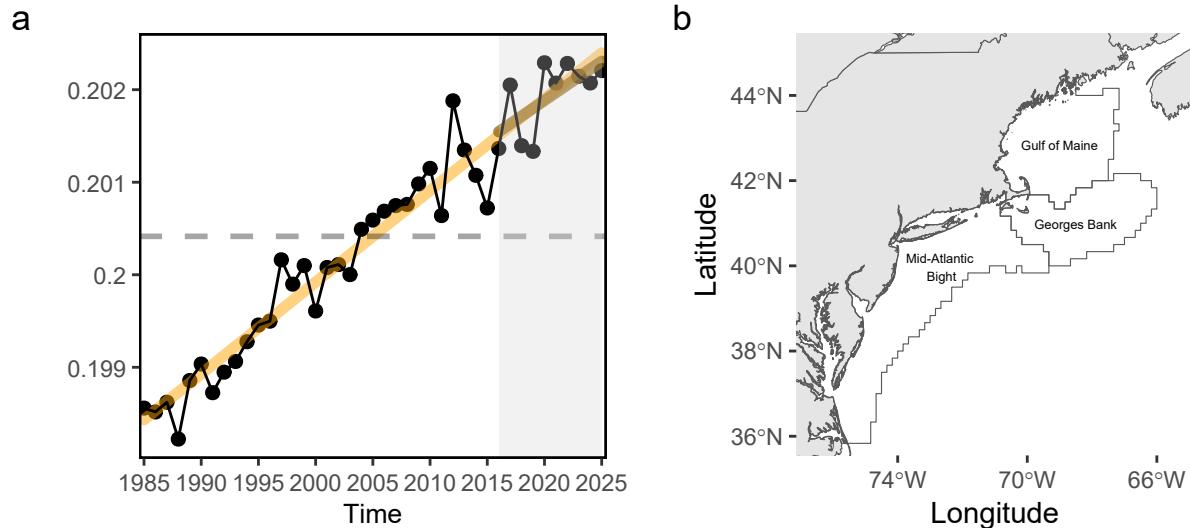


Figure 54: Document orientation. a. Key to figures. b. The Northeast Large Marine Ecosystem.

Fish and invertebrates are aggregated into similar feeding categories (Table 7) to evaluate ecosystem level trends in predators and prey.

Table 7: Feeding guilds and management bodies.

Guild	MAFMC	Joint	NEFMC	State or Other
Apex Predator				shark uncl, swordfish, yellowfin tuna, bluefin tuna
Piscivore	summer flounder, bluefish, northern shortfin squid, longfin squid	spiny dogfish, goosefish	winter skate, clearnose skate, thorny skate, offshore hake, silver hake, atlantic cod, red hake, atlantic halibut, acadian redfish	sea lamprey, sandbar shark, atlantic angel shark, atlantic torpedo, conger eel, spotted hake, cusk, fourspot flounder, windowpane, john dory, atlantic cutlassfish, blue runner, striped bass, weakfish, sea raven, northern stargazer, pollock, banded rudderfish, atlantic sharpnose shark, inshore lizardfish, atlantic brief squid, northern sennet, king mackerel, spanish mackerel
Planktivore	atlantic mackerel, chub mackerel, butterfish	atlantic herring	harvestfishes, smelts, round herring, alewife, blueback herring, american shad, menhaden, bay anchovy, striped anchovy, rainbow smelt, atlantic argentine, slender snipe eel, atlantic silverside, northern pipefish, atlantic moonfish, lookdown, blackbelly rosefish, lumpfish, northern sand lance, atlantic saury, mackerel scad, bigeye scad, round scad, rough scad, silver rag, weitzmans pearlsides, atlantic soft pout, sevenspine bay shrimp, pink glass shrimp, polar lebbeid, friendly blade shrimp, bristled longbeak, aesop shrimp, norwegian shrimp, northern shrimp, brown rock shrimp, atlantic thread herring, spanish sardine, atlantic bumper, harvestfish, striated argentine, silver anchovy	

Table 7: Feeding guilds and management bodies.

Guild	MAFMC	Joint	NEFMC	State or Other
Benthivore	black sea bass, scup, tilefish		barndoor skate, rosette skate, little skate, smooth skate, haddock, american plaice, yellowtail flounder, winter flounder, witch flounder, atlantic wolffish, ocean pout, crab,red deepsea	crab,unc, hagfish, porgy,red, sea bass,nk, atlantic hagfish, roughtail stingray, smooth dogfish, chain dogfish, bluntnose stingray, bullnose ray, southern stingray, longfin hake, fourbeard rockling, marlin-spike, gulf stream flounder, longspine snipefish, blackmouth bass, threespine stickleback, smallmouth flounder, hogchoker, bigeye, atlantic croaker, pigfish, northern kingfish, silver perch, spot, deepbody boarfish, sculpin uncl, moustache sculpin, longhorn sculpin, alligatorfish, grubby, atlantic seasnail, northern searobin, striped searobin, armored searobin, cunner, tautog, snakeblenny, daubed shanny, radiated shanny, red goatfish, striped cusk-eel, wolf eelpout, wrymouth, fawn cusk-eel, northern puffer, striped burrfish, planehead filefish, gray triggerfish, shortnose greeneye, beardfish, cownose ray, american lobster, cancer crab uncl, jonah crab, atlantic rock crab, blue crab, spider crab uncl, horseshoe crab, coarsehand lady crab, lady crab, northern stone crab, snow crab, spiny butterfly ray, smooth butterfly ray, snakefish, atlantic midshipman, bank cusk-eel, red cornetfish, squid cuttlefish and octopod uncl, spoonarm octopus, bank sea bass, rock sea bass, sand perch, cobia, crevalle jack, vermillion snapper, tomtate, jolthead porgy, saucereye porgy, whitebone porgy, knobbed porgy, sheepshead porgy, littlehead porgy, silver porgy, pinfish, red porgy, porgy and pinfish uncl, banded drum, southern kingfish, atlantic spadefish, leopard searobin, dusky flounder, triggerfish filefish uncl, blackcheek tonguefish, orange filefish, queen triggerfish, ocean triggerfish
Benthos	atlantic surfclam, ocean quahog		sea scallop	sea cucumber, sea urchins, snails(conchs), sea urchin and sand dollar uncl, channeled whelk, blue mussel