

# Rethinking the Road Effect Zone: The Probabilistic Effect of Roads on Ecosystems

*Manuscript elements:* Figure 1, figure 2, figure 3. Figure 1, 2, and 3 are to print in color.

*Keywords:* Anthropocene, Anthropogenic impacts, GPS tracking, Home range, Movement ecology, Road ecology, Space use.

*Manuscript type:* Note.

Prepared using the suggested L<sup>A</sup>T<sub>E</sub>X template for *Am. Nat.*

## Abstract

2 Roads are important for human socio-economic growth, but they carry substantial ecological  
3 impacts that can extend far beyond their physical footprint. These impacts have given rise to  
4 the so called ‘Road Effect Zone’. Although the foundational concept of the Road Effect Zone  
5 has proven useful in measuring patterns of biodiversity loss near roads, it focuses on changes in  
6 species presence/abundance, rather than the potential effects on ecological interactions. Here, we  
7 introduce a more mechanistic approach to the road effect zone that brings together the probability  
8 of a wildlife-vehicle collision with the probability of an ecological interaction occurring across  
9 an animal’s home range. We demonstrate the utility of a probabilistic representation of the road  
10 effect for giant anteaters living near a highway in Brazil. We then conclude with a brief discussion  
11 of how this probabilistic road effect can be employed in practice to inform species conservation.

## Introduction

13 The ca. 64,000,000 km of roads distributed across the globe are important for human socio-  
14 economic growth (Ibisch et al., 2016). Yet, while the area that roads occupy might be small, the  
15 ecological impacts they carry are substantial (Ascensão and Desbiez, 2022; Coffin, 2007; Fahrig  
16 and Rytwinski, 2009), and can extend far beyond their physical footprint (Forman and Alexan-  
17 der, 1998; Forman et al., 2003). From an ecological perspective, roads and roadside ecotones are  
18 considered high disturbance systems with non-natural chemical, physical, hydrological, and au-  
19 ditory properties (Brady and Richardson, 2017; Forman and Alexander, 1998; Reijnen et al., 1996).  
20 Roads have been shown to alter population densities (Andrasi et al., 2021; Fahrig and Rytwinski,  
21 2009; Reijnen et al., 1996), community composition (Truscott et al., 2005), evolutionary trajectories  
22 (Brady and Richardson, 2017; Brown and Brown, 2013), and are a serious source of non-natural  
23 mortality for many species (Ascensão and Desbiez, 2022; Carter et al., 2020; Desbiez et al., 2020;  
24 Silva et al., 2020). Fully understanding the ecological footprint of roads is thus of the utmost  
25 importance if we are to design well-informed ecological mitigation strategies.

26 Roads can cause a broad range of ecological impacts, but their effects are usually considered  
27 strongest with increasing proximity to the road surface. The strength or impact of these road  
28 effects decay at varying rates with increasing distance from the road. This relationship between  
29 the strength of effect and proximity to the road forms the basis for the concept of the 'Road Effect  
30 Zone' (Forman, 2000; Forman and Alexander, 1998), hereafter referred to as the REZ (Fig. 1).  
31 Ecologists and conservation practitioners regularly quantify the REZ for different species (e.g.,  
32 Andrasi et al., 2021; Eigenbrod et al., 2009; Semlitsch et al., 2007), and these distances are often  
33 used to make conservation recommendations (e.g., Ford et al., 2020; Forman and Deblinger, 2000;  
34 Peaden et al., 2015).

35 Although the REZ has proven useful in measuring patterns of biodiversity loss, it's corre-  
36 lative basis combines several mechanisms that shape where and why roads affect nature. The  
37 REZ concept is also a static measure that does account for the dynamic effects animal movement

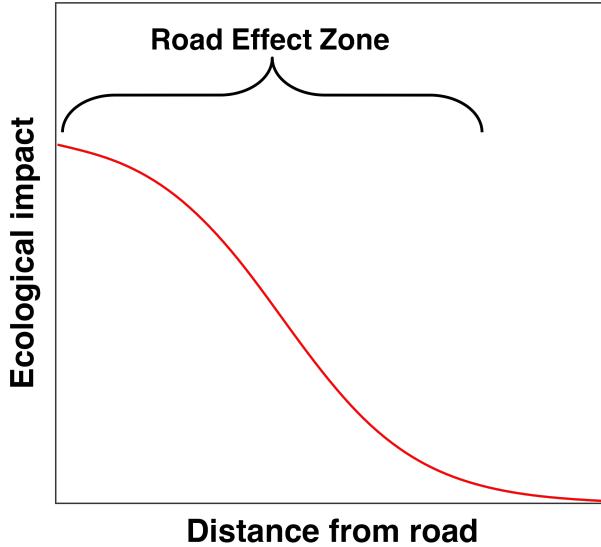


Figure 1: Theoretical depiction of the road effect zone as originally defined by Forman and Alexander (1998).

and species interactions. This is a noteworthy limitation. Consider for instance, migratory ungulates that move over vast distance between seasonal home ranges (Kauffman et al., 2020, 2021). The effects of road mortality on these highly vagile animals might decrease nutrient transfer (Subalusky et al., 2017), prey densities (Walton et al., 2017), and/or grazing pressure (Augustine and McNaughton, 1998) hundreds or even thousands of kilometers away from the road, yet these effects would not be captured within the current REZ framework. Similarly, for territorial species, the death of a roadside animal can have a cascading effect on the species' socio-spatial arrangement over vastly larger distances than the conventional REZ might suggest as individuals disperse and restructure themselves around newly unoccupied territories (Mumme et al., 2000). The reduction in density of frugivorous and/or seed-caching animals around roads can also have cascading effects on seed dispersal over larger areas than described by the REZ (Tucker et al., 2021). Here, we introduce a more mechanistic probabilistic road effect (PRE) that describes the broader ecological impacts that the mortality of an animal might have on the landscape.

51

## The Road Effect as a Joint Probability

52 As noted above, our focus is on the ecological impact of a wildlife mortality caused by a collision  
 53 with a vehicle. Our framework begins from the concept of animal space use and movement  
 54 ecology. An individual's home range describes the space it uses to undergo '*its normal activities*  
 55 *of food gathering, mating, and caring for young*' (Burt, 1943). Ecologists have long recognised the  
 56 utility of the home-range concept in describing patterns of space use (e.g., Kie et al., 2010), and  
 57 routinely estimate home ranges from animal tracking data (see Noonan et al., 2019). Statistically,  
 58 home-range estimation results in a probability distribution function ( $\text{PDF}_{\text{HR}}$ ) satisfying

$$\iint_{-\infty}^{\infty} \text{PDF}_{\text{HR}} \, dx \, dy = 1. \quad (1)$$

59 The  $\text{PDF}_{\text{HR}}$  provides information on the locations where an animal is likely to be found.  
 60 Importantly for the context of understanding the effects of roads on ecological processes,  $\text{PDF}_{\text{HR}}$   
 61 also represents the sample space over which an individual's ecological interactions (e.g., foraging,  
 62 mating, defecating, engaging in territorial defence, etc.) are expected to occur. This PDF can thus  
 63 be considered to be proportional to an animal's impact on or from the ecosystem, with core  
 64 (i.e., high probability) areas of the PDF being more heavily impacted by or from an animal than  
 65 tail (i.e., low probability) areas. Under the assumption that the probability of an animal being  
 66 roadkilled  $P\{\text{Roadkilled}\}$  is proportional to the amount of time it spends on the road, we can  
 67 quantify  $P\{\text{Roadkilled}\}$  by integrating  $\text{PDF}_{\text{HR}}$  over the area the falls on road surfaces

$$P\{\text{Roadkilled}\} = \iint_r^{r^i} \text{PDF}_{\text{HR}} \, dx \, dy, \quad (2)$$

68 where  $r$  and  $r^i$  represent the road margins. Similarly, the probability of an animal engaging in  
 69 normal ecological interactions,  $P\{\text{Ecological Interactions}\}$ , is proportional to the amount of time  
 70 it spends in locations other than on the road

$$P\{\text{Ecological Interactions}\} = \iint_{-\infty}^{\infty} \text{PDF}_{\text{HR}} \, dx \, dy - P\{\text{Roadkilled}\}. \quad (3)$$

71 The ecological impact of wildlife vehicle collision is equivalent to the outcome of an animal  
 72 not being able to engage in normal behaviour across its home range. This impact can thus be  
 73 quantified via the joint probability of (1) an animal encountering a vehicle on a road and it  
 74 being killed on the road,  $P\{\text{Roadkilled}\}$ , and (2) the probability of an animal engaging in ecolog-  
 75 ically relevant behaviour,  $P\{\text{Ecological Interactions}\}$ , or  $P\{\text{Roadkilled, Ecological Interactions}\}$ .

76 Assuming independence, the PRE is given by the joint probability of these two events

$$P\{\text{Road Effect}\} = P\{\text{Roadkilled, Ecological Interactions}\} = P\{\text{Roadkilled}\}P\{\text{Ecological Interactions}\} \quad (4)$$

77 If an animal occupies a home range that is far away from a road  $P\{\text{Roadkilled}\}$  will be  $\approx 0$ ,  
 78 resulting no road effect. Similarly, if an animal spends all of its time on roads,  $P\{\text{Roadkilled}\}$  may  
 79 be high, but  $P\{\text{Ecological Interactions}\}$  in areas away from the road will be  $\approx 0$ , also resulting  
 80 in no road effect. For animals where  $P\{\text{Roadkilled}\} \neq 0$  and  $P\{\text{Ecological Interactions}\} \neq 0$ ,  
 81 however, the road effect will be some non-zero probability with a strength that is a function of  
 82 how much time an animal spends on roads relative to the rest of its home range. Thus, the  
 83 strongest PRE must occur for animals that strongly interact with both roads and areas away from  
 84 roads. Similarly, organisms with large home ranges have more diffuse ecological interactions,  
 85 whereas animals with smaller home ranges have denser interactions over an equivalent time  
 86 period. While the general concept of a PRE is useful, we can also integrate over areas of interest  
 87 to calculate the local PRE. For instance, the PRE within some distance threshold of a road can be  
 88 quantified as

$$\iint_z^{z^i} P\{\text{Roadkilled, Ecological Interactions}\} \, dx \, dy, \quad (5)$$

89 where  $z$  and  $z^i$  represent distance thresholds from the road edge. For example, setting  $z$  to 0m

and  $z^i$  to 1m would provide the probability of a road effect within 1m of the road edge. Defining the road effect zone in this way allows it to flexibly take individual-specific forms based on each animal's movement ecology (Fig. 2). Notably, although our focus here is on the ecological impacts of roadkilled animals, the concepts presented herein can be readily extended to other spatially explicit ecological processes, such as the area a root or mycelial network diffuses through (Bielčík et al., 2019).

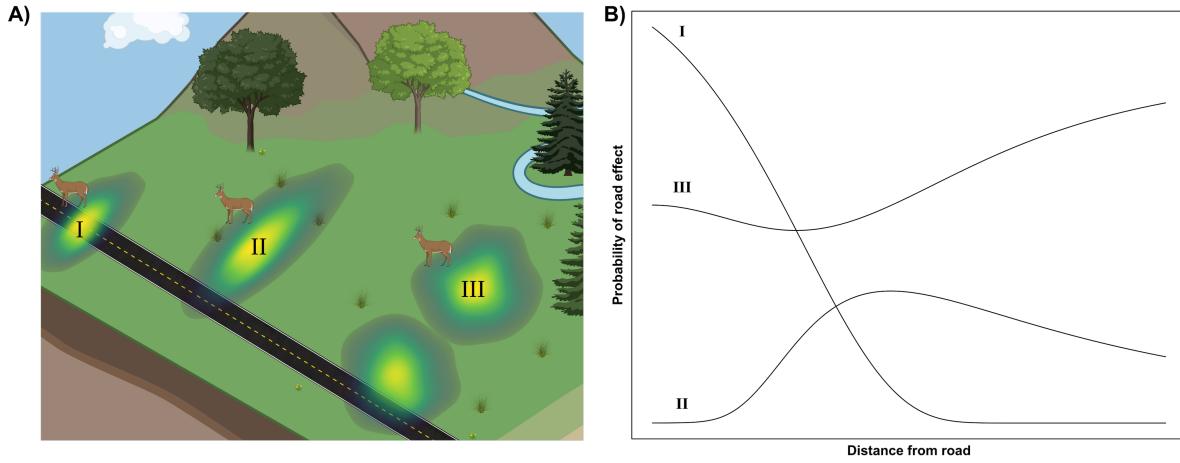


Figure 2: Theoretical depiction of the probabilistic road effect (PRE). Panel A shows three animal home ranges that vary in extent and overlap with a road. Animal I has a small home range with high road overlap, Animal II has a small home range with low road overlap, and Animal III has a large home range with intermediate overlap. Note how the three different patterns of space use shown in A result in different PREs in B. As more home range area overlaps with roads, the probability of a road mortality increases, thereby increasing the probability of a road effect occurring. Similarly, as an animal's home range increases, its ecological effects become more diffuse and the effects of a road mortality may be felt over a larger area. Panel A was created with BioRender.

## The Probabilistic Road Effect for Giant Anteaters

97 Here, we demonstrate the utility of a probabilistic representation of the road effect on a pair  
98 of giant anteaters (*Myrmecophaga tridactyla*) from the Brazilian Cerrado. Giant anteaters are the  
99 largest extant anteater, reaching over 2m and weighing up to 50kg (McNab, 1984) and are dis-  
100 tributed throughout Central and South America (Gardner, 2008). Giant anteater populations  
101 have suffered severe reductions and wildlife-vehicle-collisions are a major threat to their survival  
102 (Ascensão et al., 2021; Noonan et al., 2022). Wild giant anteaters were captured between 2017 and  
103 2018, in the vicinity of the three paved highways in the state of Mato Grosso do Sul, in the Cer-  
104 rado biome, and equipped with tracking collars that obtained GPS fixes at 20-min intervals (for  
105 full details see Noonan et al., 2022). A preliminary analysis on these data suggested that these  
106 individuals occupied fixed home ranges that regularly overlap paved highways. Following the  
107 workflow described above, we estimated the PRE for these two individuals in R (ver. 4.2.1, R Core  
108 Team, 2016). Home range areas were estimated using Autocorrelated Kernel Density Estimation  
109 (AKDE, Fleming et al., 2015) via the `ctmm` R package (ver. 1.1.0, Calabrese et al., 2016).

110 The two animals we estimated PREs for exhibited two different patterns in space use. One  
111 animal occupied a home range that was centered on the road (Fig. 3A), whereas the other oc-  
112 cupied the roadside and surrounding area, but spent little time on the road itself (Fig. 3B). As  
113 would be expected for the animal that lived right along the roadside, the ecological effects of a  
114 road mortality were greatest near the road (Fig. 3C). For the second animal, their home range was  
115 further from the road, resulting in a weaker overall road effect, but one that peaked in probability  
116 between 1-2km from the road (Fig. 3D). In other words, although  $P\{\text{Roadkilled}\}$  was greater for  
117 the first giant anteater, the road has the probability to impact the ecosystem more than 2km away  
118 from the road through the loss of the second giant anteater.

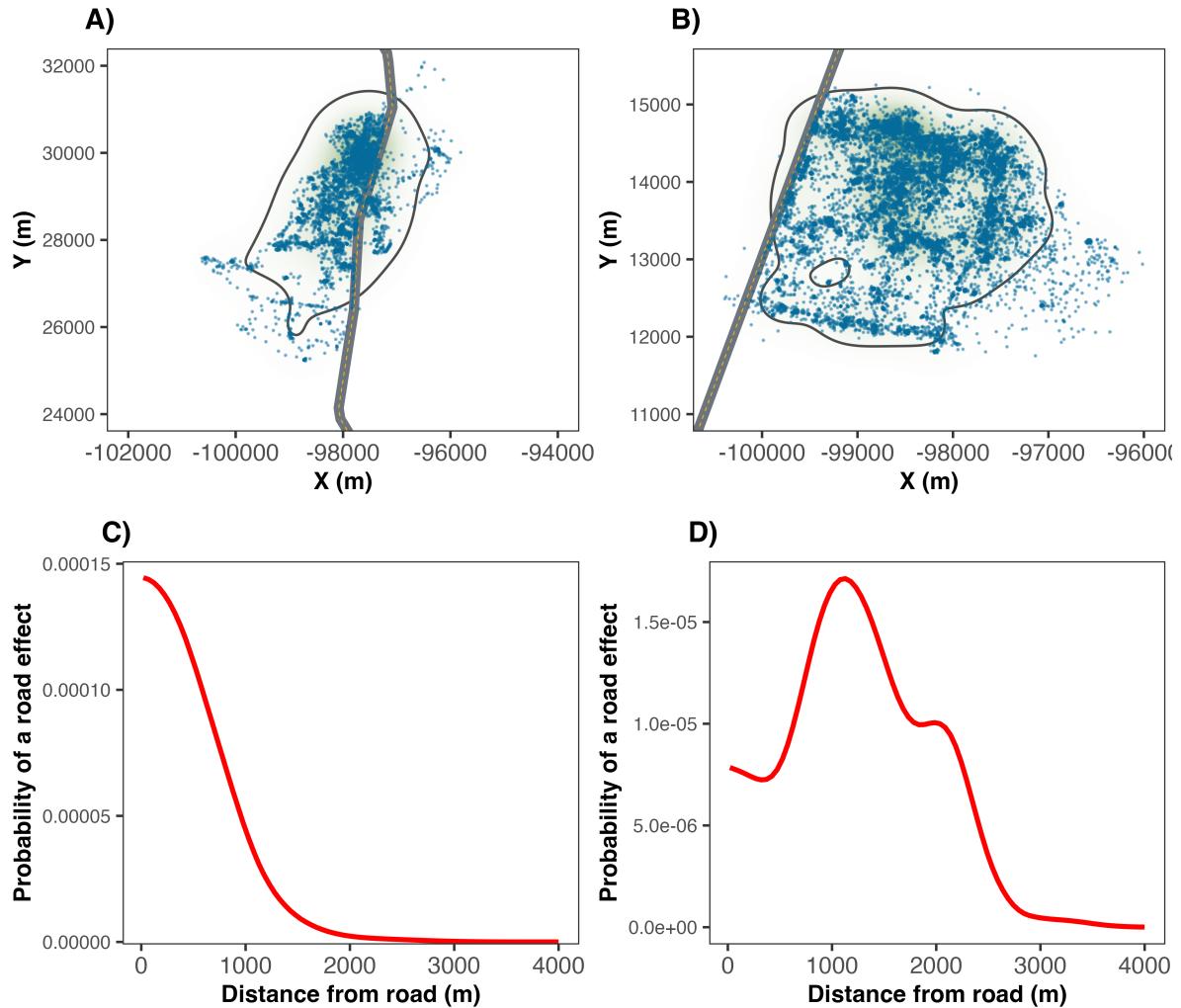


Figure 3: The relationship between space use and the road effect zone for two giant anteaters in the Brazilian Cerrado. Note how in panel A), the animal lives right along the roadside, so the ecological effects of a road mortality are greatest near the road, as shown in panel C). In Panel B), the animal's home range was further from the road and while the overall road effect was weaker, it peaked between 1-2km from the road.

119

## Discussion

120 The idea that ecological conditions will be more pristine the further one moves from away from  
 121 a road has been a focal point of road ecology research since Forman and Alexander (1998) first

<sup>122</sup> introduced the concept of the Road Effect Zone more than two decades ago (Andrasi et al., 2021;  
<sup>123</sup> Brady and Richardson, 2017; Eigenbrod et al., 2009; Forman and Alexander, 1998; Forman and  
<sup>124</sup> Deblinger, 2000; Peaden et al., 2015; Reijnen et al., 1996; Semlitsch et al., 2007). This conventional  
<sup>125</sup> REZ focuses on quantifying the area over which roads alter measurable ecological impacts, typi-  
<sup>126</sup> cally community composition or population densities (Andrasi et al., 2021; Fahrig and Rytwinski,  
<sup>127</sup> 2009; Reijnen et al., 1996; Truscott et al., 2005). While this concept has proven informative for un-  
<sup>128</sup> derstanding the ecological footprint of roads, the static nature of the REZ means that impacts  
<sup>129</sup> related to animal movement and species interactions are not captured. For instance, consider  
<sup>130</sup> a situation where there is a measurable reduction in the population density of a species near  
<sup>131</sup> a road. This effect may be due to increased road-induced mortality, but it may also be due to  
<sup>132</sup> road-avoidance and altered movement. Each of these mechanisms may reduce roadside popula-  
<sup>133</sup> tion densities, but each would have different ecological impacts and require different mitigation  
<sup>134</sup> strategies. Here we demonstrate how probability theory can provide a powerful tool for re-  
<sup>135</sup> thinking the road effect zone, and an individual-based framework for scaling up to population-,  
<sup>136</sup> or community-level effects.

<sup>137</sup> Of note is the fact that our PRE framework as defined here relies on the assumption that an  
<sup>138</sup> animal's probability of being road-killed is directly proportional to the time it spends on a road.  
<sup>139</sup> This is clearly an overly simplistic assumption as factors such as traffic volume and learning will  
<sup>140</sup> impact mortality rates (Ascensão and Desbiez, 2022; Ford and Fahrig, 2007; Mumme et al., 2000;  
<sup>141</sup> Noonan et al., 2022; van Langevelde and Jaarsma, 2005). Nonetheless, because our framework is  
<sup>142</sup> based on quantifying the joint probability of independent events, incorporating additional infor-  
<sup>143</sup> mation affecting  $P\{\text{Roadkilled}\}$  is straightforward. Indeed, this flexibility is one of the strengths  
<sup>144</sup> of this framework. For instance, if signage (e.g., altered speed limits, wildlife corridor notices,  
<sup>145</sup> etc.) is being considered as a mortality reduction strategy, the effect should be a measureable  
<sup>146</sup> reduction in  $P\{\text{Roadkilled}\}$ , and, consequently,  $P\{\text{Road Effect}\}$ . Similarly, if a practitioner is  
<sup>147</sup> tasked with making recommendations on the placement of a road-crossing structure, the con-  
<sup>148</sup>ventional REZ would suggest that the effects of reduced road mortality will be felt according to

149 some decay function as shown in figure 1. In reality, however, the capacity for animals to move  
150 freely across a roadway with  $P\{\text{Roadkilled}\} = 0$  (i.e., the benefit provided by the crossing struc-  
151 ture) can have ecosystem-wide benefits over vastly larger distances than the conventional REZ  
152 might suggest, depending on the vagility and behaviour of the animals the crossing-structure  
153 would benefit. The PRE thus provides ecologists and conservation practitioners with a dynamic  
154 metric for quantifying the ecosystem-wide impacts of roads, and/or the benefits of management  
155 strategies.

156 In this study, we have developed a framework for estimating road effect zones from ani-  
157 mal movement data. Using movement data from giant anteaters occupying roadside habitats  
158 in Brazil, we have demonstrated how this framework can be used in practice to understand the  
159 ecological impacts of a road, and help inform species management. Notably, this framework  
160 builds straightforwardly off of home-range estimation and requires no specialized data collec-  
161 tion protocols, allowing researchers to easily quantify the potential ecological impacts of roads  
162 for a variety of taxa.

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## Figure legends

Figure 1: Theoretical depiction of the road effect zone as originally defined by Forman and Alexander (1998).

Figure 2: The relationship between space use and the road effect zone for two giant anteaters in the Brazilian Cerrado. Note how in panel A), the animal lives right along the roadside, so the ecological effects of a road mortality are greatest near the road, as shown in panel C). In Panel B), the animal's home range was further from the road and while the overall road effect was weaker, it peaked between 1-2km from the road.

Figure 3: The relationship between space use and the road effect zone for two giant anteaters in the Brazilian Cerrado. Note how in panel A), the animal lives right along the roadside, so the ecological effects of a road mortality are greatest near the road, as shown in panel C). In Panel B), the animal's home range was further from the road and while the overall road effect was weaker, it peaked between 1-2km from the road.