# SPATIOTEMPORAL MODELLING & AUTOMATED IN-SITU SENSORS TO MONITOR HARMFUL ALGAL BLOOMS (HABS)

Case Study Lake Victoria

by

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Project report to the department of Geomatic Engineering and Geospatial Information Systems degree of Bachelor of Science in Geospatial Information Sciences (GIS), 2021



## **DECLARATION**

I declare that this project is my own work and has not been submitted by anybody else in any other university for the award of any degree to the best of my knowledge.

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## **Abstract**

This is meant to cover the whole project content in brief. I therefore request to update this section at project completion and closure stage.

-Thank you

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# Acronyms and abbreviations

1. DO	Dissolved Oxygen
2. DN	Digital Numbers
3. GDP	Gross Domestic Product
4. GEGIS	Geomatic Engineering and Geospatial Information Systems
5. HAB	Harmful Algal Bloom
6. FAI	Floating Algae Index
7. FLH	Fluorescence Line Height
8. IoT	Internet of Things
9. KMFRI	Kenya Marine and Fisheries Research Institute
10.KTN	Kenya Television News
11.LSAT	Lake Surface Air Temperature
12.LSWT	Lake Surface Water Temperature
13.MCI	Maximum Chlorophyll Index
14.MERIS	MEdium Resolution Imaging Spectroradiometer
15.ML	Machine Learning
16.MODIS	MODerate Resolution Imaging Spectrometer
17.OLI	Operational Land Imager
18.OC	Ocean Color
19.Quick-SCAT	Quick Scatterometer
20.SeaWiFS	Sea-viewing Wide-Field- of-View Sensor
21.SST	Sea Surface Temperature
22.TIRS	Thermal InfraRed Sensor
23.USGS	United States Geological Survey

#### 1 Introduction

#### 1.1 Background

Richardson K. (1997) defined Algal bloom as "the rapid growth of one or more species which leads to an increase in biomass of the species". This is normally associated with high concentrations of phytoplankton (algae). If the rapid growth is related to a toxic or harmful species, then the phenomenon can be termed as Harmful Algal Bloom (HAB). A species can be harmful due to the release of toxic substances e.g., *Cyanotoxins spp.* which are the most frequent in the Lake Victoria region (Okello et al., 2011). Often termed as "Red Tides", HABs have attracted a significant worldwide attention in research over the last two decades (W. Song et al., 2015; Manuel et al., 2020).

Development, stability, and density of the phenomenon are related to some environmental factors such as wind velocity and direction, Lake Surface Water Temperature (LSWT), Lake Surface Air Temperature (LSAT), Sea Surface Temperature (SST), currents, and adequate nutrient concentration, enough sunlight, warm temperatures (Tang et al, 2006) to be transported over a water body e.g., Kisumu Bay by local circulations and winds especially in large aggregates called *colonies* (Okello & Kurmayer, 2011).

A lot of scholars in the geosciences have put forward a bunch of approaches to detect and monitor HABs in both inland and ocean waters including generating indices from spectral band ratio algorithms e.g., Empirical visible-NIR band ratios (Matthews et al, 2012; Allan et al, 2015), blue-green band ratios (O'Reilly et al., 1998), red-edge (RE) region (685–720 nm) band ratios (Mittenzwey et al., 1992), thermal band based assessment (Tang et al., 2006) can be exploited to detect and monitor occurrence of a variety of HABs.

Currently, remote sensing of HAB detection and monitoring methods are primarily designed for sensors like MODIS, MERIS and SeaWiFS, (Kurekin et al., 2014) which despite allowing us to continuously monitor the behavior of the phenomenon (due to high temporal resolution (about 1 day revisit time), limits us from a detailed examination of HABs (due to coarse spatial resolutions (250~1130 meters) and only large scale HABs can be monitored by using them (Blondeau, 2014).

Landsat 8 OLI being able to discriminate finer details, therefore possesses the great value and potential to provide for the retrieval of Chl-a at 3om spatial resolution (Allan et al., 2015; Watanabe et al., 2015; Concha and Schott 2016; Manuel et al., 2020).

The advent and uptake of Internet of Things (IoT) further provides for quick development of geo-intelligent automated in-situ sensors that collects near real-time water quality data e.g., LSWT, LSAT which are thermal proxies and indicators of HAB presence thereby enabling a step-change in data availability for HAB monitoring in Lake Victoria.

Therefore, an approach that utilizes high spatiotemporal sensor's capability to detect and monitor HABs coupled by in-situ sensors technically sounds essential. Landsat 8 high spatiotemporal satellite images (30 & 100 meters for spatial resolution and 16 days for temporal) have advanced the space-based capabilities in detection and monitoring of HABs comparatively more accurately at relatively smaller inland water bodies e.g., Nyanza Gulf of Lake Victoria.

On that regard, this study intends to demonstrate the potential of using spectral bands 2, 3, 4, 5 and TIR band 10 of L8 OLI in detecting and monitoring of HABs using empirical statistical methods.

#### 1.2 Motivation and problem statement

Harmful Algal Blooms continue to be a major source of concern, not only because of their huge environmental and socioeconomic consequences, but also because of a recent dramatic increase in the number of cases documented around the world (Hill et al., 2020). HABs have the potential to cause serious environmental and human health issues, as well as other complications in deterioration of economic status thereby impacting a region's GDP. Environmental impacts include depletion of dissolved oxygen (DO) in the aquatic habitat causing mass fish stock (Tang et al, 2006) (*Fig 1,2,3 below*). Impacts posed to human inhabitants include toxic reactions to affected seafood and even in extreme cases, fatalities. Related economic impacts include adverse effects on coastal based industries e.g., fishing (Smith et al., 2019), tourism (*Fig 4*).

Lately, coastal regions of Lake Victoria especially around Nyanza Gulf (Kisumu Bay) have shown deterioration in its water quality as seen in severe signs of eutrophication with blooms (Simiyu et al., 2018). HABs can be caused by a variety of circumstances, but they are most commonly associated with favorable environmental conditions, such as increased nutrient levels-eutrophication (Santoleri et al., 2003), which are generally driven by urbanization, deforestation and poor agricultural practices (Gohin F. et al., 2006; Hecky et al., 2010).

#### 1.3 Justification

One of the bays of Lake Victoria that is mostly affected by nutrient enrichment is Nyanza Gulf (Gikuma-Njuru, P. 2013) which is discharged from the densely populated catchment with mostly subsistence agriculture (Calamari, D. 1995; Hecky, R.E. 2010). This has led to regular occurrence of bloom-forming cyanobacteria which has been associated with mass fish kills (*Fig 1, 2 & 3 below*) due to depletion of dissolved oxygen (DO) in the aquatic habitats and to some extremes, temporary shutdown of drinking water supply, i.e., from January to March 2004 (Sitoki et al., 2012).

For this purpose, a number of Landsat 8 images with some acquired in a bloom event and some during a non-bloom condition will be quantitatively analyzed. By comparing the statistically derived numerical values of the spectral indices in blooming and nonblooming condition, indices and thermal information will be extracted.



Fig 1: Standard Media Kenya reports through KTN about HABs causing Mass Fish stocks in Dunga Beach Kisumu region in February 2021, (Image: Standard Media KE)



Fig 2: Mass fish stock in Lake Victoria Shores impacted by the HABs (Image: Standard Media KE)



Fig 3: High valued Fish lost to HAB in Lake Victoria Shores (Image Source: KMFRI)



Fig 4: Tourism and therefore Income halted to HABs

#### 1.4 Research identification and Objectives

The main objective of this research is to detect, monitor and report the occurrence of Harmful Algal Blooms (HABs) and Cyanobacteria in Lake Victoria. This is achievable through the following specific objectives:

#### 1.4.1 Research objectives

- To monitor the concentration of chlorophyl-a(Chl-a) concentrations from L8 OLI data as HAB proxies in L. Victoria from 2015 to 2021.
- To monitor Lake Surface Air Temperature (LSAT) and Lake Surface Water Temperature (LSWT) from L8 TIRS images as another HAB indicator in L. Victoria.
- To develop automated Internet of Things (IoT) *in situ* sensors, applicable in near real-time to monitor and report geo-tagged Water quality data (e.g., LSWT) from 2021 onwards.

#### 1.4.2 Research questions

The following questions are formulated with respect to aforementioned objectives:

- 1) Can space based observation systems be used to detect and monitor HABs in Lake Victoria Inland water body?
- 2) Does HAB occurrence have a direct impact on the LSAT and LSWT at their point of influence?
- 3) Can IoT be utilized to monitor HAB occurrences inland water Lakes?

#### 1.5 Study outline

This research study is intended to be divided into 6 chapters whereby the first chapter introduces the study by detailing into the background, motivation and problem statement, justifying the problem, objectives and research questions; Chapter 2 contains the reviewed literature which relate to this thesis. Further, Chapter 3 will show the data and methods used in the study with Chapter 4 highlighting the results for the findings from the methods. Chapter 5 will discuss on the findings and finally, Chapter 6 will conclude and recommend for future research that might not be addressed at this level of geoscientific expertise.

# 2. Literature review 2.1 A Synthesis on the Occurrence and Negative Impacts of HABs

During the last few decades (especially the last two decades), eutrophication of lakes as a result of HABs has been a common environmental problem observed in both freshwater (Jiang et al. 2015; Luo et al. 2016) and salty water bodies (Blondeau-Patissier et al., 2014), raising a global concern due to their ability to rapidly produce various hepatotoxic and neurotoxic substances (W. Song et al.,2015). This has so been the case even at a local water bodies like Lake Victoria in East Africa which showed a potential significance of algal blooms reported as early as 1980's (Ochumba, 1984).

Proliferation of algae causes eutrophication of freshwater lakes under certain environmental conditions and float on the water surface to cause abnormal watercolor amongst a variety of aquatic-related threats (Huisman et al. 2005; Liu and Yang 2012; Qin et al. 2016). These phenomena have been observed to be severely detrimental to aquatic environment in global inland water bodies which generally follows to the growing discharge of domestic or industrial wastewater as well as agriculture and fertilizer runoff (Glibert et al., 2005; Tang et al, 2006).

The ecological phenomenon has apparently been related to a variety of socioeconomic havocs like widespread occurrence of mass fish stocks and fish booms (Ochumba, 1985, 1987; Caballero et al., 2020). Being actively photosynthetic, the blooms, not only deplete the dissolved oxygen (DO) in waters that result in mass death to water lives but also release toxins, e.g., microcystin, that cause health risks to wildlife, livestock, pets, and humans who get in contact with the HABs (Hallegraeff, 1993), affecting the safety of drinking water supply (Guo 2007; Qin et al. 2010) leading to the disruption of the water food chain as it affects both trophic levels in the chain (Diaz and Rosenberg 2008).

Furthermore, surface foams or scums formed by CyanoHABs and their odorous compounds foul up water quality and surrounding recreational environment (Anderson et al., 2002).

# 2.2 Space-based Optical Remote Sensing in the Monitoring of Chlorophyll-a as an indicator of HAB event.

Numerous geoscience scholars have observed that space-based observations are of great significance and importance in quantifying in detailed the spatiotemporal concentrations of HABs in inland water bodies.

Surface-level concentrations of Chl-a, a proxy for HAB, has been quantitatively observed in aquatic ecosystems via the exploitation of optical remote sensing for many years (Clarke et al., 1970; Wezernak et al., 1976; Smith and Baker 1982; Gordon et al., 1983; O'Reilly et al., 1998; Tang et al., 2006; Gitelson et al., 2007; Allan et al., 2015; Watanabe et al., 2015; Concha and Schott 2016). This technique has routinely supported the generation of global Chl-a concentration maps for the oceans for more than two decades (W. Song et al., 2015; Manuel et al., 2020).

A wide variety of heritage algorithms have been developed in that regard. For example, the blue-green band-ratio models have been used to quantitatively asses Chl-a (Gordon et al., 1980; Bukata et al., 1995;). However, in optically more complex inland and coastal waters, the color of water is further modified by the dominance of organic and inorganic particles, as well as colored dissolved organic matter (CDOM) (Mittenzwey et al., 1992; Han et al., 1994; Harding et al., 1994) that do not generally co-vary with Chl-a.

Consequently, to improve the estimation of Chl-a in these eutrophic and turbid environments, more sensitive methodologies have been proposed and implemented. For instance, sensors that image within the Red Edge region (690–715 nm) of the EM spectrum (Vos et al., 1986; Mittenzwey et al., 1992), when combined with red bands have manifested at a greater confidence to correlate well with Chl-a in eutrophic and/or turbid waters (Munday and Zubkoff 1981; Gower et al., 1984; Khorram et al., 1987; Gitelson 1992; Rundquist et al., 1996; Gitelson et al., 2007).

The Operational Land Imager sensor aboard Landsat-8 which was commissioned into orbit in 2013, February to capture changes at relatively high (30 m) spatial resolution which is relatively high (Irons et al., 2012) has offered significant space observatory improvements in both data quality and quantity (i.e., both spectral and spatial

coverage) over other previous heritage optical sensors (Markham et al., 2014; Pahlevan et al., 2014; Markham et al., 2015) and is relatively capable of monitoring bimonthly HAB dynamics (Pahlevan et al., 2014; Allan et al., 2015).

Several methods have been developed to retrieve Chl-a from the four OLI visible bands (Markham et al., 2015; Watanabe et al., 2015; Concha and Schott 2016; Manuel et al., 2020).

#### 2.2.1 Retrieval of Chl-a from Landsat 8 OLI Optical Spectral Bands

O'Reilly et al. (1998) observed that for satellite missions that do not inherently support measurements in the Red Edge portion of the EMR Spectrum, Chl-a algorithms tend to alternatively rely on blue-green band ratio algorithms coupled by some complex ML approaches. This is true even for Landsat-8 (Le et al., 2013; Markham et al., 2015; Freitas and Dierssen 2019).

Most research on Chl-a retrieval has heavily relied on missions like the MERIS, inherently designed to image in the RE portions (Gitelson 1992; Gower et al., 2005; Gitelson et al., 2007); These algorithms have however been noted not to score well to missions that are limited to such measurements (RE) like MODIS (Esaias et al., 1998), VIIRS (Wang et al., 2014), Geostationary Ocean Color Image (GOCI) (Ryu et al., 2012) and Landsat 8 (Snyder et al., 2017). Thus, the adoption of Empirical standard algorithms of the Ocean Color (OC) spectral band family for Chl-a estimation (e.g., OC3), (Neil et al., 2019) which has been proven to improve Chl-a retrieval from OLI products in highly eutrophic and even clear aquatic regions (Cao et al., 2020) like Kisumu Bay.

The Ocean Color 3 regional algorithms specific to OLI spectral imageries have been successfully implemented in lakes and reservoirs (Allan et al., 2015; Watanabe et al., 2015; Snyder et al., 2017) for chl-a retrieval.

# 2.3 Capabilities of Space-based observations in the Monitoring of Lake Surface Air Temperature (LSAT) as an indicator of HAB event.

With the advent of thermal imaging remote sensing, it has been observed that Sea Surface Temperature (SST) observations derived from space borne sensors like the Advanced very high-resolution radiometer (AVHRR) in conjunction with optically derived Chl-a data have great potential to detect and monitor HAB at an improved accuracy (Tang et al., 2003a; Thomas et al., 2012).

Thermal infrared bands are potentially capable of quantifying the amount of infrared radiant heat emitted from aquatic surfaces that have environmental and economic import (Anderson, J. et al., 1984; Haakstad, M. et al., 1994; River S et al., 2004).

It has been noted that Satellite Sea Surface Temperature (SST) data are often used in combination with chlorophyll-a to relate to bloom events (Villareal et al., 2012) at a higher accuracy. This has also been observed in inland water bodies where Lake Surface Air and Water Temperatures (LSAT and LSWT) are taken in considerations (Thomas et al., 2012; Shi and Wang, 2007). Peñaflor et al. 2007 also examined the correlation of seasonal phytoplankton bloom in the Luzon Strait off the Philippines using MODIS Chl and SST, in addition to Quick-SCAT wind data. Similar findings were also documented by Binding et al., 2012, 2013 by using MERIS MCI in particular by focusing in the FLH and FAI, as well as other indices, used in conjunction with satellite estimates of Chl-a and SST to improve our image analysis in a very efficient way.

Blooms of blue-green algae are reportedly often associated with increase in water temperatures (Hutchinson, 1967). During a bloom observed in Lake Victoria, the average Lake Surface Water temperature of the open waters rose from (23.9-24.8°C) to (26.9-29.5°C) in May of 1986 within two weeks of the bloom period (Kilham, 1991; Gasse, Talling & Kilham, 1994).

### 2.4.1 A Review of the Potential of Smart IoT Solutions in Locationbased Water quality monitoring.

The advent of smart solutions and of Internet of Things (IoT) has lately shown an escalating curve in their great capability to monitor water quality particularly with advancement in communication technology (Jan et al., 2021). This has greatly improved the capabilities to remotely gather and disseminate the in-situ water parameters e.g., Lake Surface Water Temperature (LSWT) and Lake Surface Air Temperature (LSAT) (Cloete et al., 2016), which are correlated with an algal bloom event in inland water bodies.

In this regard, legislations have been passed through the relevant government and non-governmental agencies for example the Kenya Marine and Fisheries Research Institute (KMFRI), African Great Lakes to set thresholding standards in water quality parameters that relate to HAB events. These parameters include Lake Victoria Surface Air Temperature which is set to vary with the proximity from the shore, (25.5 +/- 5°C), Suspended solids (30 mg/L), Pathogens and bacteria (Nil/100ml), Fluoride (1.5mg/L), Total dissolved solids (1200mg/L), Ammonia (0.5 mg/L), Nitrates (10mg/L) (Budyko, M. I. 1974; Yin & Nicholson, n.d.), among many other water quality parameters.

In-situ Water quality monitoring is defined as the collection of information at set locations and at regular intervals in order to provide data which may be used to assess current conditions and even establish trends etc. (Yang et al., 2011; Encinas et al., 2017). This can be used to ascertain the abnormalities in the pre-set standards or provide early warning identification of hazards (Raju et al., 2017)

The proposed power efficient, simple solution *in-situ* monitoring solution is set to provide near real time analysis and dissemination of LSAT and LSWT collected. The system is also designed to connect to a remote user and provide them with an alert based on the pre-configured thresholds. E.g., Lake Water Authorities who are remotely located in the offices when there is a significant deviation of LSAT and LSWT from the pre-defined set of standard values which is an associated with a bloom.

Integration of various methodologies such as the Risk Quotient Approach to show color coded hazards and each hazard level for each parameter assessed at each location e.g., Surface temperature at a given location mapped using Geographical Information Systems (Wan M. et al, 2019).

Location-based approach in assessing lake and water resource pollution has been implemented by Waspmote in Georgia, USA (Waspmote). Sample points were established using field survey, lab analyses and geospatial techniques to monitor the various water quality parameters. The results were tables, graphs and maps showing the concentrations of the parameters.

Boddula et al. (2017) proposed a wireless sensor system, CyanoSense, which provided a low footprint, low power and low-cost solution for the monitoring algal bloom remotely in Lake Oconee, Georgia, USA.

A Smart GIS-based in-situ model for assessing aquifer vulnerability was implemented in Kakamigahara Heights, Gifu Prefecture, central Japan (Babiker et al. 2010)

A GIS-based emergency response system for sudden water pollution was developed in China. (Ma, Xu, and Wang 2014; Zhang 2014). It uses GIS technology and a hydraulic water quality model to represent the levels and extents of pollution. Maps showing the spatial distribution of the sample points are generated.

#### 2.4.2 Importance of IoT concept in this Scope

Beyond providing ground truth data especially for validation of the remotely sensed data (LSAT in this regard), the internet of things

- I. Helps the geoscientific world to work and generally live smarter at large.
- II. On the same note, the automated IoT system provides the GIS and Remote sensing researcher with complete control over the entire monitoring with little to no human-t-man or human-to-machine interaction and to some extent, even prediction.
- III. In addition to automating the whole process, IoT essentially comes in by providing a near real-time monitoring of how the entire system operates. By sending the near real time GPS position of the entire system and housekeeping data, this saves us the anguish arising from unknown state of affairs.
- IV. Consumption of IoT inherently reduces the manual processes involved, thereby cutting down labour costs to be incurred. The amount, in form of money, time and labour that would otherwise be set aside by the authority e.g., the Kenya Marine and Fisheries Institute KMFRI for physical field collection of that Ground Truthing and validation data is saved on.
  - V. All said and done, IoT stands out as a few of the chief principal technologies of everyday GIS and Remote sensing research life.

#### 3. Materials and methods

#### 3.1 Study area

Lake Victoria, with an extensive surface area of about 68,800 KM² and an average depth of 40m at a maximum depth of 79m ranks the second largest fresh water lake in the world after Lake Superior and the Largest in Africa. Lying between 3°S to 0° 30`N latitude and 31°40`E to 34°50`E is distributed among these three East African countries viz Tanzania 51%, Uganda 43% and Kenya the remaining 6% (africangreatlakes.org).

That in place, the lake is privileged to serve as economical home of about 40 million residents (Calamari et al.,1995) in those riparian reserves. These millions of individuals solely bank on the lake for all aspects of their daily economic livelihood ranging from, fishing, agriculture, and industrial applications just to barely highlight but a few. On that regard, it's ecological monitoring should be of great geoscientific interest.

Being located in Equatorial regions of the globe, the lake has an alternating climatic condition varying from tropical rain forest with rainfall over the lake for a better portion of the year to a semi dry climate with sporadically discontinuous droughts over some locations.

This provides ambient temperatures varying between 12-26°C which therefore provides an optimum host condition for the growth and development of the *Cyanobacteria spp.* in this scope (Okello et al., 2011).

Figure 5 shows the location and extent of the extract of the study area particularly relevant to this study.

#### LAKE VICTORIA

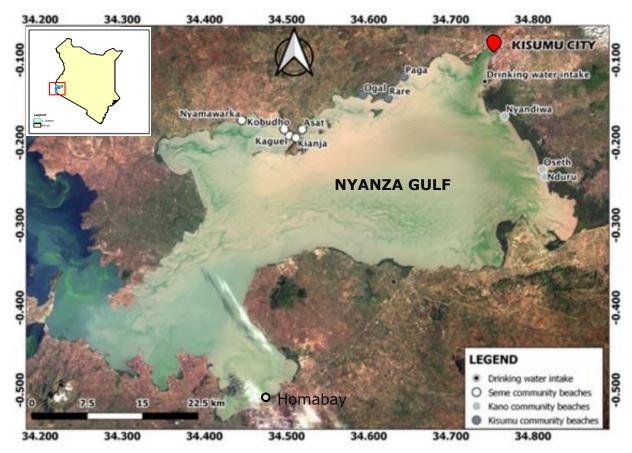


Fig. 5 Map of Winam Gulf with study sites.

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#### 3.2 Data

Five (5) datasets are exploited in this study as displayed in table 1 below.

- 1. From Landsat 8 Operational Land Imager Sensor from where spectral bands 2 to 5 centered on Blue  $_{(480\ nm)}$ , Green $_{(560nm)}$ , Red  $_{(655\ nm)}$  and NIR  $_{(865\ nm)}$  will be used in the extraction of Chl-a.
- 2. Thermal data extracted from the L8 TIR sensor from which  $TIR1_{(10,895nm)}$  band10 will be for the extraction of LSAT.
- 3. In-situ field data form the KMFRI department to aid validate the remote sensing processes.
- 4. ESRI GIS Shapefiles for delineation of study areas
- 5. In-situ data from the locally deployable autonomous IoT sensors.

Table 1: Data Sources and their roles

Data Type	Source	Role/Use
Landsat 8 OLI (30m, 16 days)	Google Earth Engine (2015-2021)	Spatiotemporal HAB Monitoring
Landsat 8 TIR (100m, 16 days)	Google Earth Engine (2015-2021)	Lake Surface Air Temperature Monitoring (LSAT)
Meteorological Data	Kenya Marine & Fisheries Research Institute-KMFRI (2015-2021)	Water Quality assessment
Shapefiles	Geodatabase of Global Administrative areas- GADM	Delineate the Study area
In-Situ Data	In-situ Sensors 2021 Onwards	Continued In-Situ Algal Monitoring

Table 2: Tools and Materials used in the study

Tool/Material	Role	Availability
Google Earth Engine (GEE)	Download and Processing	Freely Available
QGIS, R	Further Analysis & Maps	Free
Microcontroller & Sensors	In-Situ data Monitoring	Local Purchase

#### 3.3 Methodology

#### 3.3.1 Chl-a Estimation from Landsat 8 OLI

#### 3.3.1.1 Relevance of Landsat-8 in Chl-a estimation

The L8 satellite supplies multispectral images comprising of 11 spectral bands, 9 in the Optical imaging region and 2 in the Thermal region (USGS). Chl-a is known to have distinctly prominent absorption behavioral patterns between certain wavelengths of the visible portion of the EM spectrum within which this space vehicle image. Notably, around the blue (450 - 475 nm) and red (650-675nm) regions, Chl-a exhibits high absorption tendency. Further, at the green and NIR regions of the EM spectrum, Chl-a exhibits high reflectance values that could reach 500 and 700 nm, respectively (Beck et al., 2016; Tuuli et al., 2020). This information has been widely exploited by researchers to develop Chl-a quantification algorithm, (Richard et al., 2018) and this study is not an exception. The presence of narrower bandwidths and at a finer spatial resolution of 30m improves L8's pigment discrimination ability even in water bodies despite the fact that it was purposely designed for terrestrial applications (Pahlevan et al., 2014). Further, the second Thermal InfraRed (TIR) sensor on board L8 provides for the measuring of the Lake Surface Air Temperature which is another significant proxy for HABs that is investigated in this study.

#### 3.3.1.2 Satellite Data Preprocessing

Landsat-8 OLI level-1 collection-1 data product consists of quantized and calibrated scaled DN values. Retrieval of Chl-a from Landsat-8 sensor over the study region involved the following three steps,

- (i) Obtaining absolute scaled DN values for all the required bands (B2, B3, B4 and B5).
- (ii) The images were then subjected for atmospheric correction to minimize the atmospheric attenuation effects in the quite humid Lake Victoria regions.
- (iii) Then finally Chl-a were retrieved by utilizing Ocean Chlorophyll (OC) algorithms (OC-2 shown in equation 1 below).

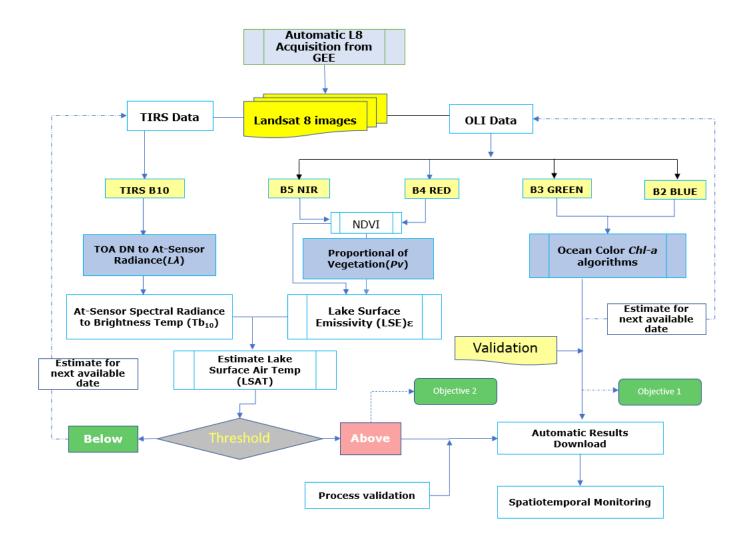


Fig 6: Overall Methodology Workflow for Objectives 1 and 2

#### 3.3.1.3 Estimating Chl-a Concentration

Based on band positioning in the OLI sensor and previous good performance in retrieval of chl-a in inland waters (Augusto-Silva et al.,) Ocean Color 2 (OC2) algorithm was settled upon for this retrieval purpose.

OC-2 is a modified fourth order polynomial algorithm which was originally developed for the SeaWiFS data (O'Reilly et al., 1998) but can be tuned to perform relatively well with Landsat 8 (Watanabe et al., 2017; Richard et al., 2018)

Mathematically, OC-2 is an empirical two-band ratio algorithm (NASA 2011) for the quantitative estimation of chl-a and is given by:

$$Chl - a = 10^{(0.2511 - 2.0853R + 1.5035R^2 - 3.1747R^3 + 0.3383R^4)}$$
 ..... (eq. 1)

Where 
$$R = log_{10} \frac{Rrs(490)}{Rrs(555)}$$
 (eq. 2)

#### Where:

Ch-a is the estimated quantified value of Chlorophyll-a

R is the quantifying coefficient derived from the spectral bands

Rrs is Reflectance in the specified wavelength region

For OC-2, ratio of Rrs at Blue band at wavelength 490nm and Rrs Green band at 555nm are used to estimate the chl-a as reported in the following dates.

Year	Date and Month	Reporting body
2015	12 <sup>th</sup> January, 22 <sup>nd</sup> February	Nasa Earth data, KMFRI
2016	23 <sup>rd</sup> February	KMFRI
2017	04 <sup>th</sup> September	Africa great Lakes
2018	27 <sup>th</sup> January	KMFRI, Nasa Earth Data
2019	18 <sup>th</sup> August	KMFRI
2020	29 <sup>th</sup> August,	KMFRI
2021	No Data	None Reported

Table 3: HABs reported in Lake Victoria, (KMFRI, RCMRD, NASA Earth Data)

The results from the chl-a estimations from this algorithm are presented in section 4 of this write-up.

#### 3.3.1.4 Accuracy Assessment

To validate the obtained chl-a values, chl-a validation dates and data were retrieved from KMFRI and some from NASA Giovanni Chlorophyll-a portal, managed by Nasa Earth Data. This portal has been for so long a rich archive for global chl-a generated from MODIS, Sentinel-3 OCLI, SeaWiFS just to mention but a few. Chl-a for the study area for years 2015 through 2021 were therefore acquired from the portal through Google Earth Engine. The spatially enabled data was then used to assess the accuracy of the estimated Chl-a from Landsat 8 OLI.

The statistical metrics used for validation was coefficient of determination (R<sup>2</sup>) as given by the equations 3 below.

$$R^{2} = \frac{\sum (Modeled_{Chl-a,i} - Mean_{Chl-a,i})^{2}}{\sum (Actual_{ActChl-a,i} - Mean_{ActChl-a,i})^{2}}$$
 (eq. 3)

where  $Modeled_{Chl-a,i}$  is the estimated Chl-a at point i,  $Mean_{chl-a,i}$  is the mean of the modeled chl-a from Landsat 8 at point i,  $Actual_{ActChl-a,i}$  represents the reference chl-a from Sentinel-3 OLCI at points i, while  $Mean_{ActLChl-a}$  is the average of the sampled actual chl-a from Sentinel-3 OLCI sensor.

#### 3.3.2 Lake Surface Air Temperature Estimation from Landsat 8 OLI

This part of the methodology workflow mainly intends to estimate the LSAT as another proxy of HABs (Owen M et al., 2017)

#### 3.3.2.1 Relevance of Landsat-8 in retrieval of LSAT

Thomas et al. (2012), after extensive research in the aquatic environments, observed that certain thermal conditions favor the growth, stability and spread of a wide variety of algal species. Further, after their domination and colonization, these microscopic and photosynthetically active organisms increase the LSAT of the region.

L8 TIR supplies two thermal bands (10 and 11) for the retrieval of LSAT. Wang et al. (2015) proposed the Mono-Window Algorithm for the retrieval of Land surface Temperature using the Landsat 8 band 10. Moreover, this algorithm scores equally acceptable when applied in the retrieval of LSAT in inland water bodies, hence the adoption of the same in this study. It should however be noted that data from the Landsat 8 TIRS Band 11 have large uncertainty and Wang et al., (2015) suggested the use of TIRS band 10 data as a single spectral band for LSAT estimation; for this reason, present work provides the use of L8 band 10 alone.

The information acquired is processed in a multi-software platform, in order to calculate the lake surface air temperature. For this purpose, R programming environment and ArcGIS 10.8 were used.

For the successful retrieval of LSAT, the study includes the major following steps as indicated in the workflow presented in (Figure 7). That is: -

- Conversion of L8 DN pixel values to at-sensor spectral radiance (Lλ)
- $\bullet$  Transformation of at-sensor spectral radiance to at-sensor brightness temperature (Tb<sub>10</sub>)
- Estimation of Lake Surface Emissivity
- Estimation of Lake Surface Air Temperature by adopting MW algorithm (Wang et at., 2015).

#### 3.3.2.2 Retrieval of Lake Surface Air Temperature

#### 3.3.2.2.1 TOA spectral radiance (L $\lambda$ )

The first step is to convert the raw DN (Digital Number) values of band10 to obtain the TOA (top of atmospheric) spectral radiance ( $L\lambda$ ) by multiplying the multiplicative radiometric rescaling factor (ML) of TIR bands with its corresponding TIR band and adding additive rescaling factor ( $A_L$ ) using equation (4) below

$$L\lambda = ML * QCal + A_L - (4)$$

#### Where:

L $\lambda$  is the spectral radiance in  $\frac{\text{watts}}{m^2 srad^{-1} um^{-1}}$ 

**ML** is the Band<sub>10</sub> multiplicative rescaling factor obtained from metadata (e.g.,0.0003342);

**QCal** is the DN value for the quantized and calibrated standard product pixel of band 10.

 $\mathbf{A}_{\mathbf{L}}$  is the band-specific additive rescaling factor obtained from the metadata (0.1);

#### 3.3.2.2.2 Brightness temperature (Tb<sub>10</sub>) in °C

As Latif 2014 puts it, brightness temperature is the Electromagnetic Radiance travelling upward from the top of the Earth's atmosphere. Therefore, The Brightness Temperature is not a temperature on the ground rather is the temperature at the satellite (M Z Dahiru et al., 2020). The spectral radiance value  $\mathbf{L}\lambda$  obtained in equation (4) above, was then converted to brightness temperature BT by adopting equation (5) below

$$Tb_{10} = \frac{K2}{Ln[(\frac{K1}{L\lambda})+1]} - 273.15 ----- (5)$$

#### Where:

 $\mathsf{Tb}_{10}$  is the brightness temperature;

K1 and K2 are thermal constants, obtained from the metadata file of the L\* OLI;  $L\lambda$  is top of atmospheric radiance.

#### 3.3.2.2.3. Normalized Difference Vegetation Index-NDVI

This mathematical algorithm which ranges between -1.0 to +1.0 is essential to identify different land surface cover types of the study which is further necessary to calculate proportional vegetation (Pv) and Lake Surface Emissivity ( $\epsilon$ ) (M Z Dahiru et al., 2020). NDVI is calculated on per-pixel basis as the normalized difference between the red band (0.64 - 0.67 $\mu$ m) and near infrared band (0.85-0.88 $\mu$ m) of the images using the formula in equation (6) below.

$$NDVI = \frac{NIR - RED}{NIR + RED} - (6)$$

#### 3.3.2.2.4 Proportion of Vegetation (Pv)

The NDVI for the study area derived from equation (6) above was then used to estimate the thermal emitting target under each lake surface cover type denoted as proportional vegetation which is a direct function of NDVI.

The vegetation and bare soil proportions are acquired from the NDVI of pure pixels.

Pv was calculated using the equation (7) below.

$$PV = \left[ \frac{NDVI - NDVImin}{NDVImax - NDVImin} \right]^{2} - (7)$$

#### **3.3.2.2.5.** Lake surface emissivity ( $\varepsilon$ )

This is the radiative properties of lake features/targets which characterizes the ability of a body to emit thermal radiation energy across the lake surface into the air atmosphere (Rhinane et al. 2012). The knowledge of lake surface emissivity is further exploited to estimate lake surface air temperature as shown in equation (8) below

$$E = 0.004 * Pv + 0.986 ----- (8)$$

#### 3.3.2.2.6. Lake Surface Air Temperature (LSAT)

We finally calculate the retrieval of LSAT using brightness temperature (BT) obtained from band 10 and Lake Surface Emissivity derived above.

LSAT has a linear relationship with at sensor brightness temperature and Lake Surface Emissivity and can therefore be retrieved using the equation (9) below:

$$LSAT = \frac{Tb10}{\left\{1 + \left[(\lambda Tb10)\right] \middle| \rho * Ln(\varepsilon)\right\}} - \cdots (9)$$

Were,

LSAT is the LSAT in Celsius (°C),

 $\mathbf{Tb}10$  is at-sensor BT (°C),

 $\lambda$  is the average wavelength of band 10,

E is the emissivity calculated from equation (v) above

This methodological workflow algorithm finally derives Lake Surface Air Temperature for the region under study and the results were as presented in section 4 of this write-up.

#### 3.3.2.2.7 Accuracy Assessment

To validate the estimated LSAT values, validation data was retrieved from NASA's MODIS which collects global Sea Surface Temperature on a near-daily basis as one of its products. Lake Surface Air Temperature for the study area for years 2015 through 2021 were then used to assess the accuracy of the derived LSAT.

The statistical metrics used for validation was the coefficient of determination (R<sup>2</sup>) as given by the equations below

$$R^{2} = \frac{\sum (Modeled_{LSAT,i} - Mean_{LSAT,i})^{2}}{\sum (Actual_{LSAT,i} - Mean_{ActLSAT,i})^{2}}$$
 (eq. 10)

where  $Modeled_{LSAT,i}$  is the estimated LSAT at point i,  $Mean_{LSAT,i}$  is the mean of the modeled LSAT from Landsat 8 at point i,  $Actual_{LSAT,i}$  represents the reference LSAT at points I, while  $Mean_{ActLSAT,i}$  is the average of the sampled actual LSAT from MODIS.

#### 3.3.3 Automated In-situ Internet of Things System

The automated *in-situ* system consists of a variety of hardware and software that work in conjunction to collect and disseminate LSAT remotely.

The internet of things, or IoT, is a network of interconnected computing devices, mechanical and digital machinery or people having unique identifiers (UIDs) and the ability to transfer data without requiring human-to-human or human-to-computer interaction. (2019, Alexander S.)

Narrowing down into the specifics of this research project, the word IoT has been used to refer to a man-made system that has been assigned a unique and private Internet Protocol (IP) address, loaded with sensors and is able to transfer data e.g.,

- Lake Surface Air Temperature
- Lake Surface Salinity
- GPS Location of the system at the time of the collection of such data
- The whole system in-house status including working condition
- Unexpected system failure
- System power level.

All these data are packaged into one socket and transmitted over a wireless network between the system and a remote monitoring computer.

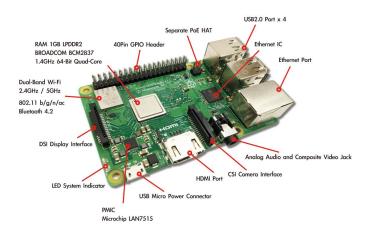
#### 3.3.3.1 Devices and Sensors Used

At least the following components and sensors have been used in the achievability of collection of the In-Situ data:

#### 1. Microcontroller (MCU):

This can be loosely defined as an integrated circuit that contains a brain or rather the microprocessor that has a memory and integrated circuits and that has the potential to control the functionalities of an embedded system or electronic device (such as our Water Quality Monitoring system). There do exist a vast variety of Commercial Off The Shelf-(COTS) MCUs e.g.,

- ATmega (Arduinos)
- ESP32
- BCM
- Single Board Computers (SBC) e.g., Raspberry Pi
- ARM Cortex e.g., STM32



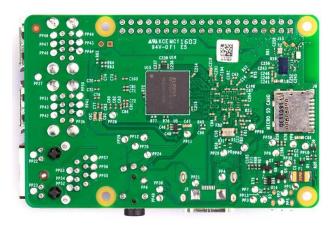


Figure 7 Raspberry Pi model 3B+ used in this project Front and Back views

#### 3.3.3.1.1 Reason for choice of Raspberry Pi in this project

#### 1. Powerfulness:

Having the capacity to perform relatively similar to a full-blown computer, it ranks the highest at the time of submission of this thesis by being able to simultaneously run multiple tasks at ago and further operate as a server gives it more potential score and rating.

#### 2. Little to no electronics prerequisites needed

Little prior embedded programming languages knowledge and its components is needed here. Unlike for Arduino and other ARM Cortex MCUs, you definitively need a very good and deep electronic background, and need to know about embedded programming languages e.g., Object oriented C++.

#### 3.3.3.2. Sensors Used

At least the following sensors have been integrated and used as part of the in-situ data collection system.

#### 3.3.3.2.1 Lake Surface Air Temperature (DHT11 Sensor)

The DHT11 is a widely accepted relatively cheap, a digital and easy-to-use air temperature sensor.

#### 3.3.3.2.1.1 Features of DHT11 Air Temperature Sensor

- $_{\odot}$  Measures air temperature within the ranges 0-50°C with an accepted accuracy of  $\pm 2$  °C.
- $_{\odot}$  The sensor need be interfaced with a filtering capacitor and a pull-up resistor of 10KΩ connected in series with the VCC output pin of the sensor
- Input voltage is ~3.3 V to 5.0V

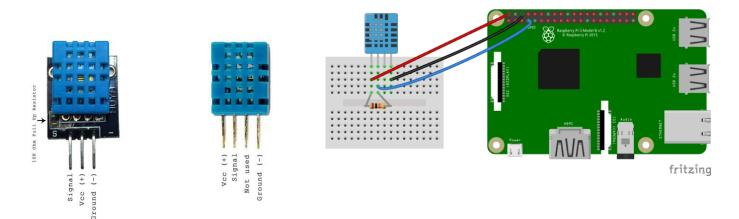


Figure 8. DHT 11 Sensor module and how it hooks up to the Raspberry-pi Microcontroller

#### 3.3.3.2.1.2 Mode of operation of DHT11 Air Temperature Sensor

The DHT11 basically measures the Air Temperature under the concept that, with increase in the air temperatures, the resistance of the Negative Temperature Coefficient NTC thermistor installed in it decreases, hence recoding higher temperature values (Electronics Lovers, 2020)

#### 3.3.3.2.2 Global Positioning System (GPS) Sensor.

The Neo-6M- Ublox GPS sensor Module was settled on to solely play the role of collection of the (x, y) coordinate location information where the whole system will be at any given time of collection and dissemination of LSAT data.

#### 3.3.3.2.2.1 Basic mode of Operation

A basic GPS signal receiver capable of collecting the GPS pseudo frequency signals broadcast at 1575 MHz from GPS space segments and computationally determine the location of any system that it's embedded to. This will tell us precisely where in Lake Victoria there exists abnormality in temperature rises, calling for an emergency mitigation action.



Figure 9 The Neo-6M GPS Sensor Module

#### 3.3.3.2 Working and Operation of the Automated IoT System.

An IoT ecosystem integrates an array of smart devices that use embedded systems technology, such as: -

- 1. In-built processors for example Intel, BCM (For our case).
- 2. An array of sensors (as previously discussed in section 3.3.3.2 above) that collect and send the data from their point of attachments.
- 3. Communication hardware, that disseminates the collected data.

Upon establish a secure connection with the IoT gateway device, the sensors will be able to share the sensor data they collect where data is either sent to the cloud server as commonly seen in IoT systems that run on Raspberry Pi (using POST, GET mechanisms) for analysis or the sensor data can just be locally are analysed. Our IoT system and functionality are fully automated with little to no human intervention as illustrated in the *Fig. 9* below.

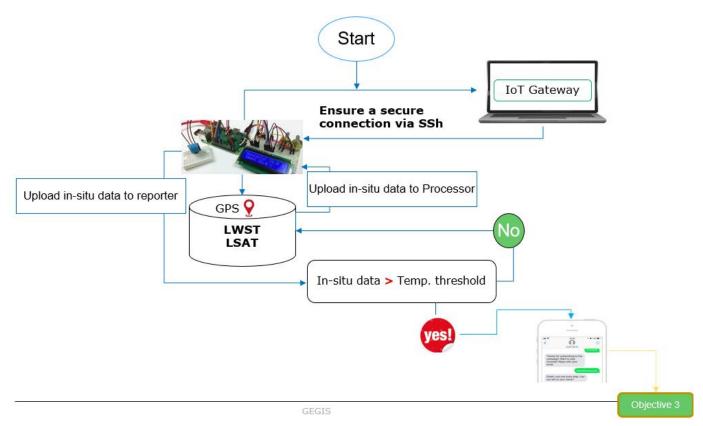


Figure 10: Overall Methodology Workflow for the IoT system.

#### 4. Results

#### 4.1.1 Chl-a concentration Maps

OC-2 algorithm was proposed and used to estimate the HAB concentrations as reported. The following displayed results were obtained.

In Figure 10 below, it is noted that on the event of HAB as reported, Chl-a values shoot significantly to up to 31.3 Mg/M³ in the affected regions as seen in red patched areas. This was unlike the case in the *blue* regions in the western parts on the study area as shown below. They reported very low chl-a score around 1.2 Mg/M³ especially for a non-turbid water body like Lake Victoria.

Figure 11 below as well indicates that HAB that occurred especially on the North-Eastern parts of the study area could be well explained by the high chl-a values as reported by the OC-2 algorithm.

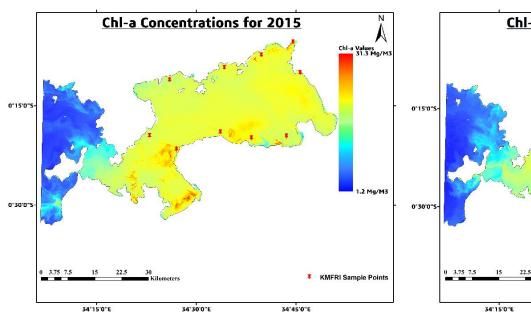


Fig 10. HAB concentration, 2015

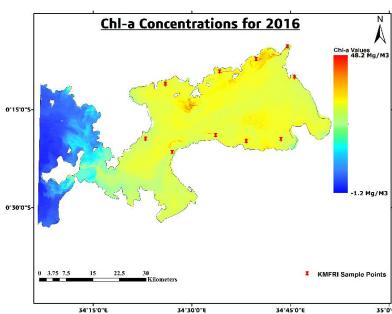


Fig 11. HAB concentration, 2016

In Figures 12 and 13 below, it is noted that on the event of HAB, Chl-a values are seen to rise significantly to up to  $39.6 \text{ Mg/M}^3$  and  $25.9 \text{ Mg/M}^3$  in the affected regions Notably, in 2018, the severity of the bloom was quite low as the highest chl-a values were  $25.9 \text{ Mg/M}^3$ .

The western regions of the study area still maintain comparatively low chl-a values indicating they were not significantly affected.

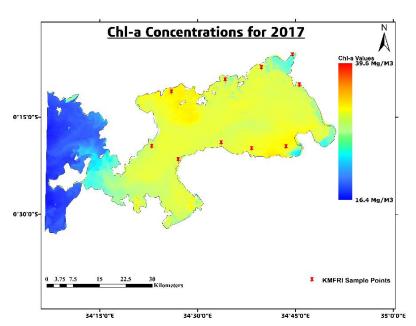


Fig 12. HAB concentration, 2017

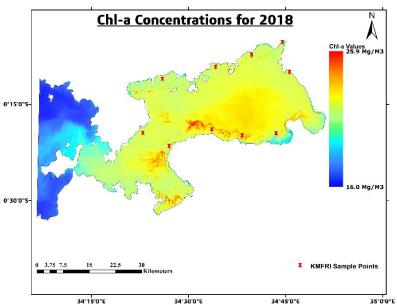


Fig 13. HAB concentration, 2018

Finally figures 14 and 15 for 2019 and 2020 below, show that the severity of the HAB went relatively high. This was realised with the increased eutrophication in the said region as Simiyu et al., 2018 observed. This research therefore agrees with the previously well-documented studies. Chl-a values are seen to rise significantly to up to 49.3 Mg/M³ and 57.1 Mg/M³ in 2019 and 2020. The cases were more severe.

The western regions of the study area still maintain comparatively low chl-a values indicating they were not significantly affected.

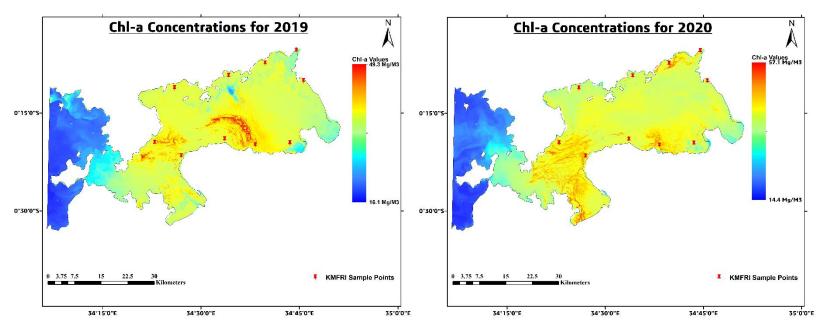


Fig 14. HAB concentration, 2019

Fig 15. HAB concentration, 2020

## 1.1.2 Accuracy assessment of the Chl-a Estimates.

As a general rule of remote sensing and quantitative estimations, the Chl-a estimates had to be validated. In order to validate the estimated results and analyze the relationship between Sentinel-3 OLCI chl-a data and those retrieved from Landsat8 OLI, scatter plots and linear fitting were done with the help of 1000 to 500 samples between Chl-a values and the corresponding Sentinel-3 values from the product.

Statistical analysis indicated (Figures 16 and 17) that the coefficients of determination ( $R^2$ ) were seen to be 0.836884 (2015) and 0.843381 for 2016.

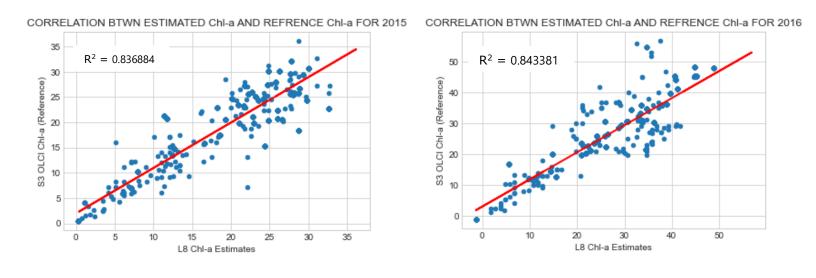


Fig 16 & 17. Scatter plots of correlation of L8 OLI estimates and S3-OLCI products for 2015 and 2016.

For estimated obtained in 2017 and 2018, statistical analysis showed (Figures 18 and 19) that the coefficients of determination ( $R^2$ ) were seen to be 0.891653 (2017) and 0.839463 for 2018.

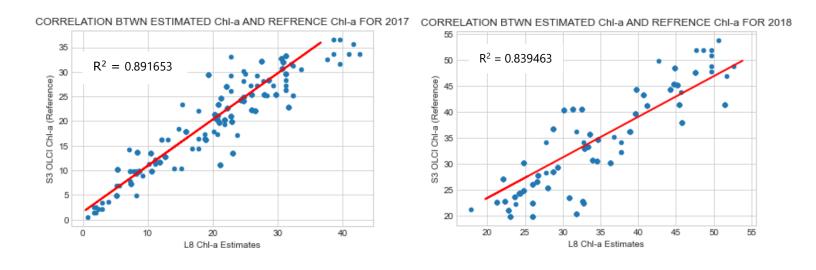


Fig 18 & 19. Scatter plots of correlation of L8 OLI estimates and S3-OLCI product for 2017 and 2018

For estimates obtained in 2019 and 2020, statistical analysis further showed (Figures 20 and 21) that strong positive coefficients of determination ( $R^2$ ) of 0.899546 (2019) and 0.882167 for 2020.

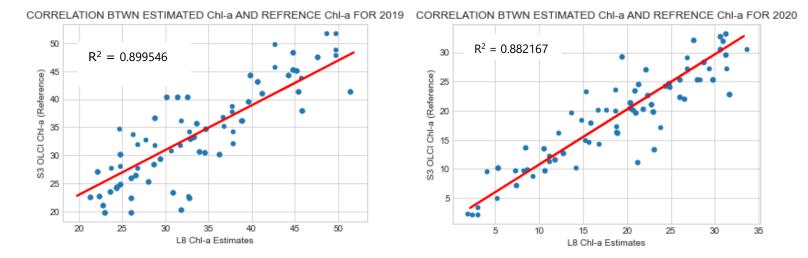


Fig 20 & 21. Scatter plots of correlation of L8 OLI estimates and S3-OLCI product for 2019 and 2020.

This essentially implies that there was a significant strong positive correlation between Chl-a estimated by OC-2 algorithm and ESA's data products provided by Copernicus.

That is to say, in regions with insufficient estimate parameters for Chl-a, it is truly reliable and accurate to employ the OC-2 algorithm to estimate using Landsat 8 OLI.

#### 4.2.1 LSAT For corresponding reported HABs Events

As indicated in the methodology, the Moving Window Algorithm was proposed and adopted to estimate the LSAT as another proxy for the HAB as reported. The following displayed results were obtained from the process.

In Figures 23 and 24 below, it can be observed that on the event of HAB as reported, there existed a relatively high thermal emissivity from the Lake surface especially in the affected regions.

This is manifested by the LSAT values rising significantly to up to 35.7  $^{\circ}$ C in 2015 and 36.6  $^{\circ}$ C for 2016 especially in the affected regions as shown in the figures below. The unaffected regions however maintained low LSAT values of 23.6  $^{\circ}$ C and 16.9  $^{\circ}$ C respectively.

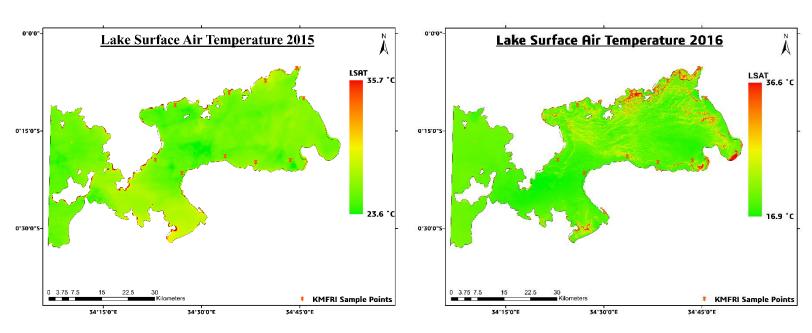


Fig 23. LSAT concentration, 2015

Fig 24. LSAT concentration, 2016

In Figures 25 and 26 below, it is noted that on the event of HAB, LSAT values are seen to rise significantly to up to  $35.1\,^{\circ}\text{C}$  and  $36.6\,^{\circ}\text{C}$  in the affected regions.

As seen in the Chl-a maps in 2017 and 2018, the severity of the bloom was almost evenly distributed in 2017, while in 2018 the distribution mainly dominated the south eastern part as supported by the LSAT maps below.

The western regions of the study area still maintain comparatively low LSAT values indicating they were not adversely affected by the phenomena.

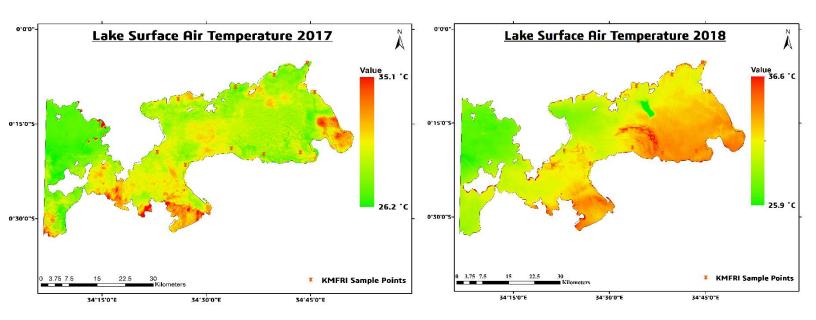


Fig 25. LSAT concentration, 2015

Fig 26. LSAT concentration, 2016

In Figures 27 and 28 below, it is as well noted that on the event of HAB, LSAT values are seen to rise significantly to up to 36.8 °C and 34.8 °C in the affected regions.

This research therefore agrees with the previously well-documented studies. LSAT values are as well seen to rise significantly in 2019 and 2020. The cases were more severe.

The western regions of the study area still continue to maintain comparatively low LSAT values indicating they were not adversely affected by the phenomena in all the reported dates.

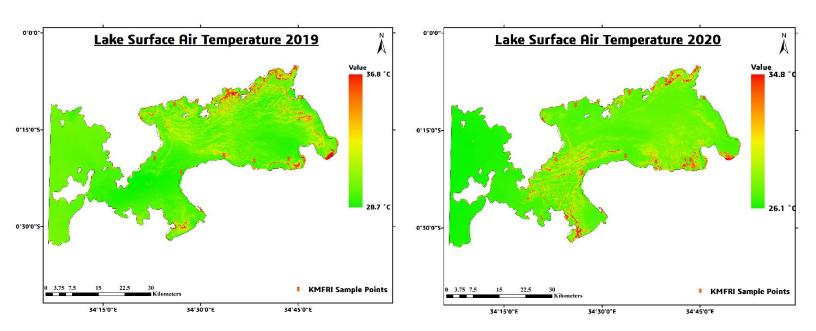


Fig 27. LSAT concentration, 2019

Fig 28. LSAT concentration, 2020

#### 1.1.2 Accuracy assessment of the LSAT Estimates.

In order to validate the estimated results and analyze the relationship between MODIS LSAT data and those retrieved from Landsat8 TIR, scatter plots and linear fitting were done with the help of randomly generated 1000 to 500 samples between LSAT values and the corresponding MODIS values from the product.

Statistical analysis showed (Figures 29 & 30) that the coefficients of determination  $(R^2)$  were seen to be 0.666916 (2015) and 0.843381 for 2016.

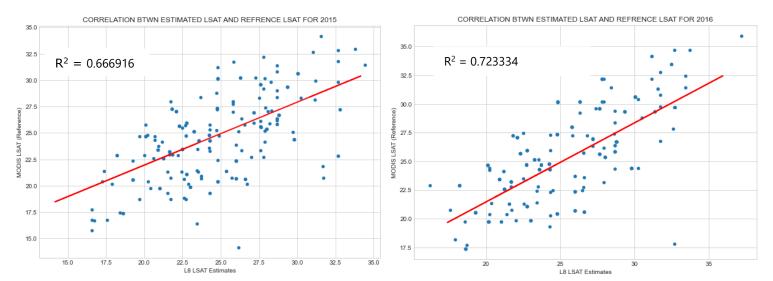


Fig 29 & 30. Scatter plots of correlation of L8 TIRs estimates and MODIS product for 2015 and 2016.

For estimates obtained in 2017 and 2018, statistical analysis showed (Figures 31 and 32) that the coefficients of determination ( $R^2$ ) were reported to be 0. 696916 (2017) and 0. 781505 for 2018.

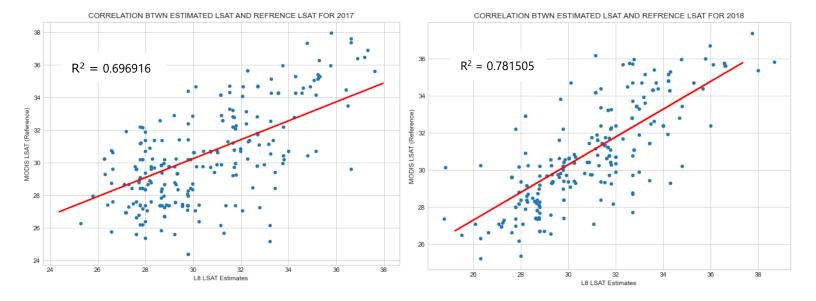


Fig 31 & 32. Scatter plots of correlation of L8 TIRs estimates and MODIS product for 2017 and 2018.

For estimates obtained in 2019 and 2020, statistical analysis further showed (Figures 33 and 34) strong positive coefficients of determination ( $R^2$ ) of 0.820665 (2019) and 0.754280 for 2020.

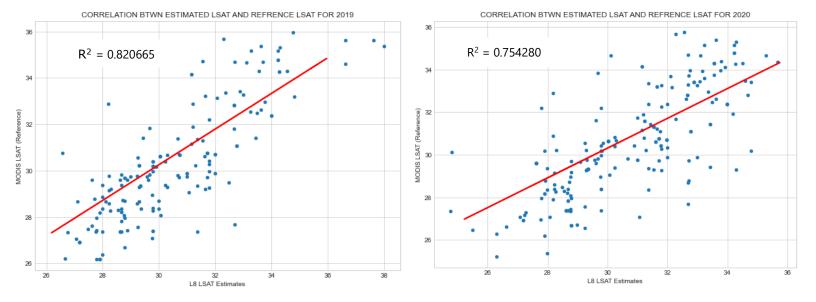


Fig 33 & 34. Scatter plots of correlation of L8 TIRs estimates and MODIS product for 2017 and 2018.

These correlations of determinations essentially implies that there did exist significant strong positive correlation between Lake Surface Air Temperatures estimated by Moving Window algorithm and NASA's MODIS data products provided by the Earth Data.

That is to say, in regions with insufficient estimate parameters for Lake Surface Air Temperature, it is truly reliable and accurate to employ the Moving Window algorithm to estimate using Landsat 8 TIRs.

# 4.3 Automated In-Situ IoT System

In order to come up with a near real-time ground-based monitoring system, an Automated In-situ IoT system was developed, tested and found to be fully operational.

This was primarily to collect geotagged Lake Surface Air Temperature and Lake Surface pH, in this case the salinity being the case.

The system was automatically able to collect the LSAT and send a geotagged notification via text SMS in case of LSAT condition anomalies. The Africastalking text message API was used to relay the near-real time messages from the IoT system to the remote users or authorities.

Figure 35, 36 and 37 indicates how the message upon arrival, appears in the GUI. The text comes along with a Google Maps link indicating the System location from which such data are coming from.

When the user clicks the link, a Google Map UI as indicated in Figures 38, 39 and 40 is realized, showing the Google Map being rendered on the users' screen to show the real-time map locating the system.

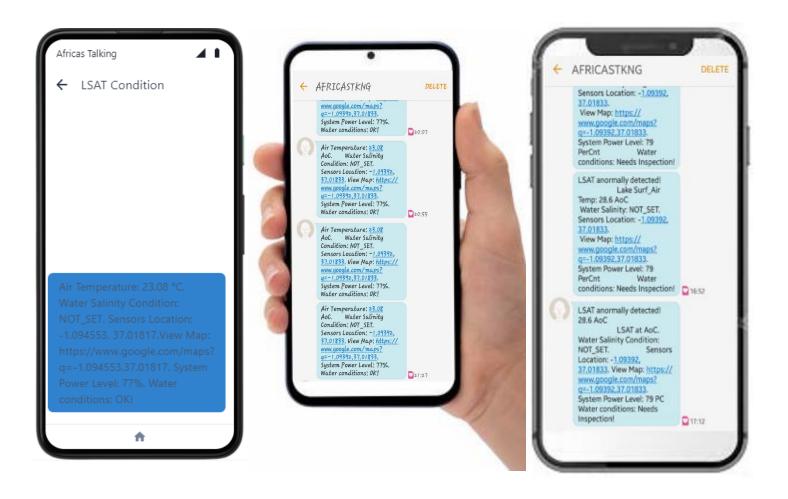


Figure 35, 36 and 37 The text message UI when received by client/user.

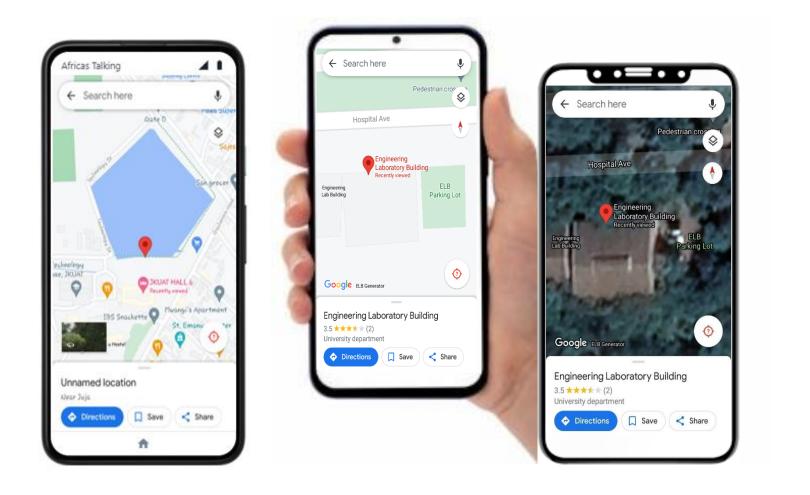


Figure 38, 39 and 40 Client/User viewing sensor located at JKUAT pond and Engineering Laboratory Building (ELB) via Google Maps API.

# 2. Discussion

As seen from the results presented in section four (4) above, Chl-a concentrations we reportedly high in the dates assessed and seen to be far above the minimum optimal value of non-turbid waters like Lake Victoria, which has been determined as 20.05m/gm³ (Ref). This situation took effect owing to the circumstantial elevated proliferation, predominance and therefore prevalence of the CyanoHABs species that produce the toxins [47,48]. The used Ocean Color-2 algorithm presented in this literature was able to model and represent this phenomenon.

It was noted that of all the reported incidences, it was noted that 2018 had the lowest magnitude of impact as manifested by the least concentrations of 25.9 m/gm³. 2020 bloom occurrence saw the highest concentrations of 57.1 m/gm³. Generally, the highest concentrations in most cases were located either at the North, South or Central part of the Lake Victoria basin, without necessarily manifesting any predictable trend. In the mentioned sections, the eutrophication caused by a high discharge of pollutions coming from the busy industrial sections of the riparian reserves played the greatest responsibility for the proliferation of harmful algae [2,12,44,45,75].

With regard to the Lake Surface Air Temperature (LSAT), the values were seen to corroborate with the Chl-a values since high temperatures spots were as well associated with relatively high Chl-a concentrations. This agrees with the hypothesis that algae occurrences are closely associated with high temperatures and upon their occurrence, they cause the LSAT to as well rise.

The estimated Chl-a and LSAT were correlated with the well accepted Sentinel 3 OLCI Chl-a (ref)and MODIS LSAT (ref) products. The Chl-a estimates were found to correlate positively strong with R<sup>2</sup> scores ranging from 0.83 to 0.899. Again, the estimated LSAT were correlated with the MODIS LSAT (ref) products. The Chl-a estimates were as well found to correlate positively strong with R<sup>2</sup> scores ranging from 0.67 to 0.82. These correlations of determinations essentially implies that there did exist significant strong positive correlation between the estimates and data products highlighted before.

That is to say, in regions with insufficient estimate parameters for Lake Surface Air Temperature and Chl-a, it is truly reliable and accurate to employ the Ocean Color - 2 Moving Window algorithm to estimate using Landsat 8 TIRs. These two can then be implemented on board Landsat 8 and enable automatic download of the results in the event of future HAB, which will therefore call for immediate and precautionary corrective measures.

The full-blown IoT system which is made up of various components like Lake Surface Air Temperature DHT11 sensor, Neo-Ublox GPS modules with a fixed antenna, Raspberry Pi Microprocessor, breadboard that serves as the circuit prototyping board and power supply that run the software. The system that was made up of the abovementioned hardware and software was able to measure and report in real time the Lake Surface Air Temperate, and giving the location in form of WGS84 latitude, longitude, and the time in Greenwich Meridian Time (GMT) and display in a remote hand-held phone as shown in the Figures 35 to 37.A google map showing the location of the system as at the time of LSAT collection was also implemented as seen in Figures 38, 39 and 40. The LSAT of the test environments in JKUAT ranged from 23.8 to 28.6 which were within optimum were reported. Therefore, this system can help in near-real time to monitor and track the LSAT of Lake Victoria and any other area of Study of related phenomena. The IoT system of hardware and software can therefore improve water quality monitoring and assessment enabling it to be in real time and give the spatial attribute of location which will couple the space borne monitoring system like Landsat space missions.

## 3. Conclusions and Outlook

In conclusion, the study was able to come up and present a methodological approach to detect and monitor the concentration of chlorophyl-a(Chl-a) from Landsat 8 Operational Land Imager data as HAB proxies in Lake Victoria from 2015 to 2021. This was implemented through the Ocean Colour 2 algorithm. Therefore, space-based observations play a significant role in the monitoring of Harmful Algal Blooms.

Further, Landsat 8 Thermal InfraRed (TIR) sensor abord the same space vehicle was able to provide for corresponding thermal images to estimate the rising levels of Lake Surface Air Temperatures in the advent of HAB occurrence.

As displayed in the result section, it is noted that the in occurrence of HAB, Chl-a values rose significantly to 31 to 457.1 Mg/M3 Similarly, the moving window algorithm LSAT estimates indicated that on a bloom event, the LSAT rose in the affected areas up to values 35.1 °C to 36.6 25 °C. Conversely, parts of the Study area that were never affected by the phenomena marked the lowest chl-a value (-1.2 mg/m³) and corresponding temperatures value (less than 27 °C)

It was further noted that a near-real time monitoring approach was adopted since the reporting of the occurrence of the phenomena ideally need to be as soon as they occur. An automatic cloud-based processing and download of Chl-a was then tested on a relatively small portion of the study area.

Finally, to ensure near-real time monitoring at field level, an automated in-situ system was developed to collect and monitor the Lake Surface Air Temperature (LSAT) as another proxy for HAB event. The system was tested and found operational was able to report remotely in case of any abnormal rise in the Lake Surface Air Temperature, calling for further response. This would help the maritime authorities to take effective response to the alert sent and take corrective measures. The system was tested in the local University Pond as indicated in Figures 35 to 40 and was able to report Surface Air Temperature values of ranges 23°C to 28.6°C.

With the unfortunate lack of synchronized measured ground truthing data from the authorities like KMFRI, the estimated Chl-a and LSAT products have been validated using in Chl-a products from Sentinel 3 -OLCI and LSAT with the MODIS (MOD11A1) data; Validation results from the scatter plots

confirmed that the acceptably high correlation coefficient values of 0.743 to 0.876 for LSAT and 0.845 to 0.902 for Chl-a, this confidently demonstrates the applicability of the proposed and adopted algorithm in Chl-a estimation and LSAT retrieval in the study area.

## **Recommendations**

After the implementation of a variety of proposals, the following are areas of improvement that need to be scaled for future betterment of this project:

- a) To increase the temporal resolutions of the monitoring and reporting, there need to do a Spatio-temporal image fusion of Landsat 8 and MODIS aqua to provide a near daily 30 M spatial resolution of Chla-a and LSAT estimates instead of waiting for the 16-day 30M Landsat as this current study implements.
- b) The Maritime and Environmental management authorities within the Lake Victoria riparian region should take more action in ensuring that the sewage from the Industrial regions is well treated before discharge to minimize the eutrophication of the Lake Victoria Basin.
- c) The presented System had relatively low mechanical integrity as there was no protective cover to shield it against environmental hazards like scorching sun, unfavorable rainy weather conditions among others. A protective cover to shelter all the components should therefore be fabricated.
- d) Furthermore, including another water quality sensor such as water pH (Salinity) which is another indicator of HAB should be incorporated in the study. The resent study proposed but was not within the might to implement the said.

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**Appendix**Add any extra materials that could not fit in the main text of your write up and is relevant to your study.