

DENVar: A global Dynamic Estimate of NDVI Variance

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Abstract

Since the 1980s, remote-sensing estimates of habitat productivity such as the Normalized Difference Vegetation Index (NDVI) have helped develop and support a variety of hypotheses and inform policy and industry-related decisions. The free availability of such data has allowed widespread access and use of well-established metrics of habitat productivity. However, historically, users tended to focus solely on (predictable) trends in productivity. In recent years, people have been recognizing the importance of environmental unpredictability (stochasticity), particularly as climate change and human-induced rapid environmental change transform the ecosystems species have evolved in and adapted to. We address this gap by presenting a new global Dynamic Estimate of NDVI Variance, *DENVar*. We estimate spatiotemporal trends in mean NDVI and the variance around the mean using Hierarchical Generalized Additive Models, which provide a flexible yet transparent model structure, unlike machine learning approaches such as neural networks. We show that *DENVar* can be used as a reliable proxy of environmental stochasticity, strongly correlates with **XXX**, and can be used to test many hypotheses related to environmental stochasticity, such as forage stochasticity, regime shifts, We conclude by offering some considerations around quantifying and interpreting environmental stochasticity, and we provide a publicly and freely available shiny app for estimating mean NDVI and the variance around it, *DENVar*, for given coordinates and optional dates. The app requires no knowledge of statistics or coding. Additionally, we provide a temporally static raster of mean NDVI and *DENVar* for the years 1981-2025 that can be used for GIS applications and does not require any knowledge of R.

22 Introduction

23 Developments in remote sensing since the 1970s have greatly expanded our ability to quantify
24 and monitor landscapes using satellite imagery

25 • NDVI is an estimate of greenness, NPP, forage availability
26 • NDVI has many uses and is often convenient because it's free, well-established, and
27 many people focus on changes in the mean
28 ability to predict and respond to changes in mean conditions (forage availability, phe-
29 nology, etc.) depends on the scale of changes relative to the perception scale (Levin,
30 1992; Frankham & Brook, 2004; Riotte-Lambert & Matthiopoulos, 2020; Steixner-Kumar &
31 Gläscher, 2020) and memory (Abrahms *et al.*, 2019). Consequently, environmental stochas-
32 ticity is also an important driver, besides changes in the mean (Mezzini *et al.*, 2025)

33 • variance is hard to calculate, as it depends on a correct estimate of the mean
34 • paper on estimating the asymptote of environmental variance globally, but variance
35 for a closed system must converge towards a finite value (Bachmaier & Backes, 2011;
36 Fleming *et al.*, 2014)
37 • we present DENVar
38 • HGAMs are highly flexible models that allow us to estimate trends in the data without
39 imposing preconceived expectations
40 • more transparent than black-box machine learning methods like neural networks, etc.
41 • <https://www.sciencedirect.com/science/article/abs/pii/S0034425718305625>
42 • <https://silvis.forest.wisc.edu/data/dhis/>
43 • we provide global rasters of the mean and variance in NDVI, averaged across 1981-2025
44 • we provide a shiny app for people to calculate mean and variance without any knowl-
45 edge of R

Methods

Choice of color schemes

We represent NDVI using a modified version of Fabio Crameri’s divergent *bukavu* palette (Crameri, 2018; Crameri, Shephard & Heron, 2020), which has high-contrast for deuteranope and protanope vision. For NDVI values between 0 and 1 (the right half of the palette), the colors also have sufficient contrast for the colors to be distinguishable by both tritanope and achromatic vision. Appendix A contains an approximate representation of the color palette for each vision type. We obtained all color palettes from the `khroma` package (v. 1.14.0, Frerebeau, 2024) for R (v. 4.4.1, R Core Team, 2024).

Input data

We obtained NDVI data using the image composites from the AVHRR and VIIRS sensors (Vermote & NOAA CDR Program, 2018, 2022). The code for downloading the data directly from the NOAA server is available at <https://github.com/QuantitativeEcologyLab/ndvi-stochasticity/blob/main/analysis/002-download-ndvi-rasters.R>. The number of non-NA land raster cells was substantially larger than the maximum data frame size in R ($2.28 \times 10^{10} > 2^{31} - 1 \approx 0.21 \times 10^{10}$), so we reduced the dataset size by calculating 15-day averages of NDVI (see Fig. A2) and aggregating the averaged rasters by a factor of 2×2 with `terra::aggregate(fact = 2)`. Although this resulted in a reduction in sample size, the temporal averaging also reduced signal-related noise.

The predictor data for the models included: WWF biome (Olson & Dinerstein, 2002, see Fig. A3), ecoregion (the individual polygons from Olson & Dinerstein, 2002, see Fig. A4), integer day of year (1 to 366), integer year (1981 to 2025), and elevation above sea level (Fig. A5). We excluded all polygons that did not have any NDVI data (0.06% of land, 29% of the polygons, maximum area: $452 \text{ km}^2 \approx (21 \text{ km})^2$). We distinguished between biomes in the northern and southern hemisphere to allow for different seasonal trends between hemispheres.

The 15-day averages allowed us to have non-repeating values of day of year across years, since 15 is not a factor of 365 or 366. We downloaded the global digital elevation model using the `get_elev_raster()` function from the `elevatr` package for R (v. 0.99.0, Hollister *et al.*, 2023) with a resolution of 0.076 degrees.

Modeling

We estimated spatiotemporal trends in mean NDVI using Hierarchical Generalized Additive Models (HGAMs) via the `mgcv` package for R (v. 1.9-3, Wood, 2017). To reduce modeling fitting times with negligible losses to model accuracy, we used the `bam()` function with fast REstricted Marginal Likelihood (`method = fREML`) and covariate discretization (`discrete = TRUE`). See Wood, Goude & Shaw (2015) and Wood *et al.* (2017) for more information. Ideally, NDVI should be modeled using beta location-scale models (after the linear transformation $Y^* = \frac{Y+1}{2}$) to account for the fact that: (1) NDVI is bounded between -1 and 1, and (2) the variance in NDVI is dependent on the mean (and vice-versa), since ecosystems with very low NDVI (e.g., rock, ice, or concrete) or very high NDVI (e.g., dense forest) tend to have lower variance in NDVI. However, preliminary tests showed that fitting times for beta models were prohibitive, especially for beta location-scale models. In contrast, Gaussian models fit substantially faster and provided very similar spatialtemporal estimates of mean NDVI (see the `analysis/000-sardinia-test.R` script in the GitHub repository). The global model for the mean NDVI had the structure below:

```
m_mean <- bam(
  ndvi_15_day_mean ~
    biome + # to avoid intercept shrinkage
    s(poly_id, bs = 'mrf', xt = list(nb = nbs)) +
    s(doy, biome, bs = 'fs', xt = list(bs = 'cc'), k = 10) +
    s(year, biome, bs = 'fs', xt = list(bs = 'cr'), k = 10) +
    ti(doy, year, biome, bs = c('cc', 'cr', 're'), k = c(5, 5)) +
    s(elevation_m, bs = 'cr', k = 5),
  family = gaussian(),
  data = d,
  method = 'fREML',
```

```

knots = list(doy = c(0.5, 366.5)),
drop.unused.levels = TRUE,
discrete = TRUE,
samfrac = 0.001, # find initial guesses with a subset of the data
nthreads = future::availableCores(logical = FALSE) - 2,
control = gam.control(trace = TRUE))

```

Biome-specific fixed-effect intercepts allowed to account for the differences in mean NDVI through biomes without coefficient shrinkage (i.e., reversion towards the global mean). A Markov Random Field of the ecoregions (`bs = mrf`, see page 240? of Wood, 2017) estimated smoothed, ecoregion-specific deviations from the biome-level spatial means. Two factor-smooth interaction terms (`bs = 'fs'`) accounted for biome-specific seasonal and yearly trends. Seasonal terms were made cyclical by using cyclical cubic splines (`xt = list(bs = 'cc')`), which are continuous up to and including second derivative at the edge knots (0.5 and 366.5, which correspond to 00:00 of January 1st and 24:00 of December 31st, respectively). A tensor product interaction smooth for each biome estimated the change in seasonal trends over the years. Finally, a smooth of elevation above sea level accounted for the effect of altitude. See Pedersen *et al.* (2019) for more information on hierarchical modeling and the use and interpretation of factor smooth interaction terms.

will we subtract the average residual from each pixel?

A second HGAM estimated the variance in NDVI around the mean estimated by the model above. The model had an identical structure to the model for the mean, with the exception that (1) the response variable was the squared residual from the first model, such that the model estimated the mean squared residual (i.e., the variance) for a given point in time and space, and (2) a smooth term of the estimated mean NDVI. The model is available below:

```

#' *CHANGE TO SOS BASIS*
m_var <- bam(
  e_2 ~
    biome +
    s(poly_id, bs = 'mrf', xt = list(nb = nbs)) +
    s(doy, biome, bs = 'fs', xt = list(bs = 'cc'), k = 10) +
    s(year, biome, bs = 'fs', xt = list(bs = 'cr'), k = 10) +
    ti(doy, year, biome, bs = c('cc', 'cr', 're'), k = c(5, 5)) +

```

```

s(elevation_m, bs = 'cr', k = 5) +
s(mu_hat, bs = 'cr', k = 5),
family = gaussian(),
data = d,
method = 'fREML',
knots = list(doy = c(0.5, 366.5)),
drop.unused.levels = TRUE,
discrete = TRUE,
samfrac = 0.001,
nthreads = future::availableCores(logical = FALSE) - 2,
control = gam.control(trace = TRUE))

```

Results

General results

- two Orthographic projections of DENVar: Northern and southern hemispheres on spring solstice

Correlations with other metrics

- temperature
- seasonal temperature range
- precip
- seasonal temperature range
- spp diversity
- gross primary productivity
- max animal weight
- <https://silvis.forest.wisc.edu/globalwui/>
- <https://silvis.forest.wisc.edu/data/dhis/>: NDVI16 (Normalized Difference Vegetation Index), EVI16 (Enhanced Vegetation Index), FPAR8 (Fraction absorbed Photosynthetically Active Radiation), LAI8 (Leaf Area Index), GPP8 (Gross Primary Productivity)

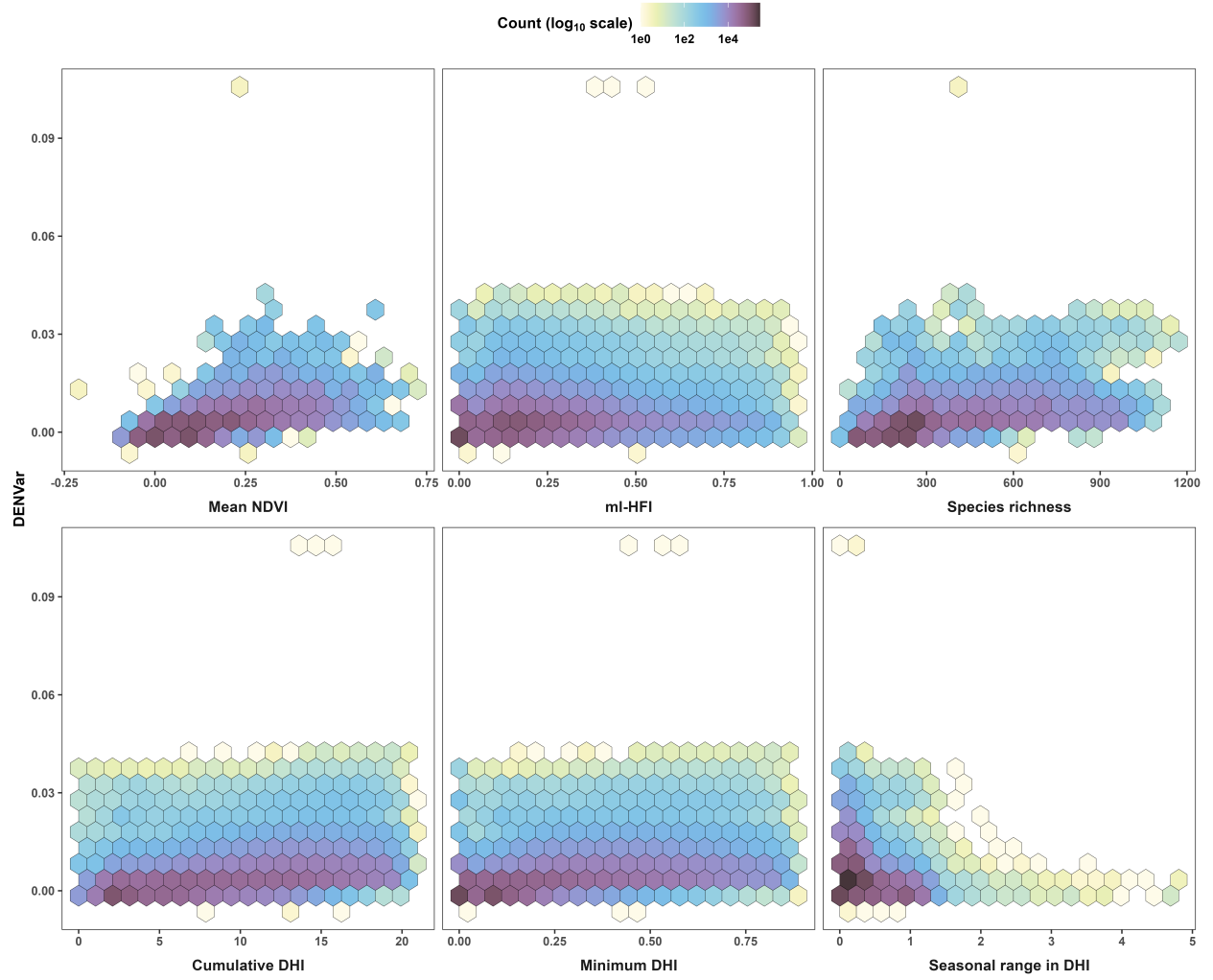


Figure 1: Hex plots of the relationship between the global Dynamic Estimate of NDVI Variance (DENVar) and common ecological metrics. The fill of the hexagonal cells represents the log-transformed number of points within the cell. (A) Mean NDVI (as estimated in this manuscript); (B) the machine-learning Human Footprint Index of Keys, Barnes & Carter (2021); (C-E) the 2015 cumulative, minimum, and seasonal range in the Dynamic Habitat Index of XXX, (F).

How to obtain the data

Static rasters of mean and variance (1981-2025)

- for GIS users (tif)

Shiny app for obtaining values

- get values for specific coordinates: 1981-2025 average
- get values for specific coordinates and dates (upload as a CSV, download as a CSV)
- screenshot of the shiny app

Discussion

We have presented a new spatiotemporally dynamic metric of environmental stochasticity, which we have called the Dynamic Estimate of NDVI Variance, DENVar. In this section, we discuss the interpretation of DENVar in greater depth, and we argue the advantages and limitations of the statistical methods used. We conclude by suggesting some use cases of DENVar within animal movement ecology, landscape ecology, and phenology, but we invite readers to consider how DENVar may be of use in their field and adjacent disciplines.

Why use complicated statistics instead of machine learning?

Admittedly, the title of this section is misleading for two reasons. Firstly, some may argue that Generalized Additive Models are a form of machine-learning (here abbreviated as ML), but we intend to contrast GAMs with more “black-box” types of machine learning, such as neural networks and random forests. The second reason for which the title is misleading is that it implies that machine learning is not complicated, or at least that it is less complicated than the models we used. However, we argue that while traditional statistical models may seem more intimidating to some, their complexity is what allows analysts to produce interpretable results and measures of uncertainty (Wood, 2017). And while the theoretical

basis for GAMs can be discouraging for those who do not have a strong background in mathematics or statistics, outputs from GAMs can be presented in an accessible manner even to those without a strong quantitative background through the use of well-designed figures, especially in a Bayesian context (when the aim is to estimate a parameter). In contrast, while ML methods will often give understandable results, it may not be possible to discern *why* such results were “chosen” by the model. Instead, GAMs produce flexible yet interpretable estimated effects that can be represented with relatively simple figures, even if the mathematical formula(e) for the model may be complicated.

Ideally, a model should estimate or explore the relationships among variables such that results can be interpreted in a transparent and unambiguous manner, such that readers can scrutinize the underlying assumptions, methods, and decisions. We recognize that the assessing GAMs requires a certain level of expertise, but we hope the readers can appreciate the option to assess the GAMs more easily than a series of consecutive, weighted regressions (as in the case of neural networks). In contrast, GAMs learn from data with no initial starting assumptions other than user-imposed limitations of wiggleness, which are clearly outlined in each basis size of the models’ terms. The structure and wiggleness of GAMs can thus be grossly designed based on the system’s complexity and structure while allowing the data to inform the model trends, wiggleness, and coefficient size.

Interpreting DENVar: the importance of scale

- smoothness of mean impacts smoothness and size of variance: depends on scale of interest (e.g., animals’ ability to respond to, learn, and predict changes in the mean). Changes not attributed to trends in the mean are stochastic
- DENVar is a good metric as long as the spatiotemporal scale of interest is similar or larger to that of DENVar. May not work well for small-scale processes such as the movement dynamics of a small rodent, the phenology of a small orchard when spatiotemporal heterogeneity is finer than DENVar’s resolution

Environmental stochasticity and animal movement ecology

The field of animal movement ecology has long shown that animals alter their movement in response to resource abundance or availability (Burt, 1943; Charnov, 1976; Southwood, 1977; Broekman *et al.*, 2024). More recently, movement ecologists have started focusing more on resource stochasticity (e.g., Stephens & Charnov, 1982; Rizzuto *et al.*, 2021), but it is worth distinguishing between resource stochasticity and seasonality (Nilsen, Herfindal & Linnell, 2005). Seasonality is generally best interpreted as a cyclical change in the mean that many animals can learn to predict, as in the case of green-wave surfing (Middleton *et al.*, 2018; Geremia *et al.*, 2019). In contrast, environmental stochasticity refers to unpredictable changes, such as forest fires (ref?), sudden snow storms (Aikens *et al.*, 2025), and unusual phenology (ref?). Animals also respond to stochastic changes (Mezzini *et al.*, 2025), but their ability to adapt depends on behavioral plasticity (Rickbeil *et al.*, 2019; Steixner-Kumar & Gläser, 2020), and memory (Polansky, Kilian & Wittemyer, 2015; Abrahms *et al.*, 2019; Riotte-Lambert & Matthiopoulos, 2020; Falcón-Cortés *et al.*, 2021; Ranc, Cagnacci & Moorcroft, 2022).

Botero *et al.* (2015) tipping points

Polazzo *et al.* (2024)

extreme events, black swans (Logares & Nuñez, 2012; Anderson *et al.*, 2017)

- migration routes are key habitat (Ortega *et al.*, in prep), especially in stochastic environments
- can fail to adapt to changes in conditions Sawyer *et al.* (2019)
- <https://www.sciencedirect.com/science/article/pii/S0960982220308484>
- <https://onlinelibrary.wiley.com/doi/full/10.1111/gcb.15169>
- industrial development affects animals' ability to respond to both predictable and unpredictable change (<https://www.nature.com/articles/s41559-022-01887-9>)
- effects of sudden deep snow in red desert on proghorn (<https://www.sciencedirect.com/>

science/article/abs/pii/S0960982225002957)

- <https://www.nature.com/articles/s41467-023-37750-z>
- <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1002/ecy.4238>
- detecting regime shifts (in lakes: Bjorndahl *et al.*, 2022)

Environmental stochasticity and population dynamics

read Lande (1993)

Environmental stochasticity reduces habitats' energetic balance and carrying capacity (Chevin, Lande & Mace, 2010) as individuals struggle to rely on unpredictable resources and increase the instability of populations and communities (Lande, 1993). Additionally, stochastic environments reduce the ability of specialist species to adapt and evolve while selecting for more generalist species (Levins, 1974). Consequently, stochastic environments tend to have lower species diversity (**marcus?**).

Environmental stochasticity and fire ecology

Recent work by (**devries?**) has suggested that NDVI stochasticity is an important metric for evaluating forest recovery after large fires, while (**collinson?**) has suggested that variance in NDVI is an important predictor of the occurrence of forest fires due to ignition by lightning.

Environmental stochasticity and phenology

- Jonzén *et al.* (2006)
- effects can be compounded by cumulative effects (e.g., loss of species diversity, see Wolf, Zavaleta & Selman (2017))
- climate change affects plant phenology, but changes in situ may not match experimental estimates (Wolkovich *et al.*, 2012)
- reproduction of painted lady butterfly (Stefanescu, Ubach & Wiklund, 2021)
- Wessling *et al.* (2018)

Useful references

- Keith *et al.* (2008)
- Chevin *et al.* (2010)
- Rickbeil *et al.* (2019)
- Mueller *et al.* (2008)
- Pettoirelli *et al.* (2005)
- Keys, Barnes & Carter (2021)
- Nilsen *et al.* (2005)
- Merkle *et al.* (2016)
- Tian *et al.* (2015)
- Huang *et al.* (2021)
- Fan & Liu (2016)
- Wang *et al.* (2019)
- Pease (2024)
- Xu *et al.* (2021)
- Mezzini *et al.* (2025)

TO READ: * Site fidelity as a maladaptive behavior in the Anthropocene: <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1002/fee.2456> * Pettoirelli N., Vik J.O., Mysterud A., Gaillard J.-M., Tucker C.J. & Stenseth N.Chr. (2005). Using the satellite-derived NDVI to assess ecological responses to environmental change. *Trends in Ecology & Evolution* 20, 503–510. <https://doi.org/10.1016/j.tree.2005.05.011>

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