

# Influence of temperature and upwelling intensity on Indian oil sardine (*Sardinella longiceps*) landings

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## Introduction

Environmental variability is known to be a key driver of population variability of small forage fish such as sardines, anchovy and herring (Bakun 1996, Alheit and Hagen 1997, Cury et al. 2000, Checkley et al. 2017). In particular, ocean temperature and upwelling dynamics, along with density-dependent feedback, have been identified as important in affecting recruitment success and biomass of European and Pacific sardines (*Sardina pilchardus* and *Sardinops sagax*) (Jacobson and MacCall 1995, Rykaczewski and Checkley 2008, Alheit et al. 2012, Lindegren and Checkley 2012, Lindegren et al. 2013). Like other sardines, the Indian oil sardines show strong interannual fluctuations and larger decadal booms and busts. The Indian oil sardine offers an instructive case study to investigate the effects of environmental variability, particularly temperature and upwelling dynamics, as they occupy an ocean system that is warmer than that occupied by other sardines and have a strong seasonal cycle driven by the Indian summer monsoon.

The Indian oil sardine (*Sardinella longiceps* Valenciennes, 1847) is one of the most commercially important fish resources along the southwest coast of India (Figure 1) and historically has comprised approximately 25% of the catch biomass (Vivekanandan et al. 2003). Landings of the Indian oil sardine are highly seasonal and peak during and after the summer monsoon period (June through September), in conjunction with the onset and early relaxation of coastal upwelling. At the same time, the landings of this small pelagic finfish are highly variable from year to year. Small pelagics are well known to exhibit high variability in biomass due to the effects of environmental conditions on survival and recruitment (Bakun 1996, Alheit and Hagen 1997, Cury et al. 2000, Checkley et al. 2017). In this fishery, however, environmental conditions also affect exposure of sardines to the fishery. Until recently, the Indian oil sardine

fishery was artisanal and based on small human or low powered boats with no refrigeration. The fishery was confined to nearshore waters, and thus migration of sardines in and out of the coastal zone greatly affected exposure to the fishery.

Researchers have examined a variety of environmental variables for their correlation with landings of the Indian oil sardine in order to understand the factors that drive landings variability. Precipitation during the southwest monsoon (Antony Raja 1969, 1974, Murty and Edelman 1971, Jayaprakash 2002) and the day of the monsoon arrival (Jayaprakash 2002) is thought to act as either a direct or indirect cue for spawning. Many studies have looked for correlations between precipitation, however the reported effects are positive in some studies and negative in others (Madhupratap et al. 1994). Researchers have also looked for and found correlations with various metrics of upwelling intensity, such as sea level at Cochin (Murty and Edelman 1971, Longhurst and Wooster 1990, Madhupratap et al. 1994, Jayaprakash 2002, Thara 2011), salinity and bottom sea temperature (Krishnakumar et al. 2008), and with direct measures of productivity, such as nearshore zooplankton and phytoplankton abundance (Hornell 1910, Nair 1952, Nair and Subrahmanyam 1955, Madhupratap et al. 1994, George et al. 2012, Piontkovski et al. 2015, Menon et al. 2019). Researchers have also found correlations with near-shore sea surface temperature (SST). SST can affect both somatic growth rates and juvenile survival but also can cause fish to move off-shore and away from the shore-based fishery (Annigeri 1969, Prabhu and Dhulkhed 1970, Pillai 1991). The El Niño/Southern Oscillation (ENSO) has a cascading effect on all the aforementioned environmental parameters (SST, prediction, upwelling) which in turn impact oil sardines, and correlations have been found between ONI peaks and landings peaks with a 9- to 12-month lag (Rohit et al. 2018).

In this paper, we study the utility of environmental covariates from remote sensing to explain year-to-year variability in landings using the time series of quarterly Indian oil sardine landings from the southwest coast of India. This time series is derived from a stratified sampling design that surveys landing sites along the coast and was first implemented in the 1950s (Srinath et al. 2005), however Yearly catch-at-length data are not available prior to 2001 and neither are stock size estimates nor fisheries independent data. Thus traditional length- or age-structured models (e.g. Virtual Population Analysis) are not possible. Instead we use time-series models with covariates to model landings. Modeling and forecasting landings data using time-series models has a long tradition in fisheries and has been applied to many species (Mendelsohn 1981, Cohen and Stone 1987, Nobel and Sathianandan 1991, Stergiou and Christou 1996, Lloret et al. 2000, Georgakarakos et al. 2006, Prista et al. 2011, Lawer 2016), including oil sardines (Srinath 1998, Venugopalan and Srinath 1998). These models

can be used to understand the variables associated with catch fluctuations and can be used to provide forecasts that assist fisheries planning. Unlike prior work on landings models with covariates, we use non-linear time-series models to allow a flexible effect of covariates and past catch on current landings. We also specifically focus on environmental covariates measured via remote sensing. Remote sensing data provide long time series of environmental data over a wide spatial extent at a daily and monthly resolution. A better understanding of how and whether remote sensing data explains variation in seasonal catch will support future efforts to use remote sensing data to improve catch forecasts.

## **Study Area**

Our analysis focuses on the Kerala coast (Figure 1) region of India, where the majority of the Indian oil sardines are landed and where oil sardines comprise ca. 40% of the marine fish catch (Srinath 1998, Vivekanandan et al. 2003). This area is in the Southeast Arabian Sea (SEAS), one of world's major upwelling zones, with seasonal peaks in primary productivity driven by upwelling caused by winds during the Indian summer monsoon (Madhupratap et al. 2001, Habeebrehman et al. 2008) between June and September. Within the SEAS, the coastal zone off Kerala between 9°N to 13°N has especially intense upwelling due to the combined effects of wind stress and remote forcing (Smitha et al. 2008, Smitha 2010). The result is a strong temperature differential between the near-shore and off-shore and high primary productivity and surface chlorophyll in this region during summer and early fall (Madhupratap et al. 2001, Habeebrehman et al. 2008, Jayaram et al. 2010, Raghavan et al. 2010, Smitha 2010, Chauhan et al. 2011). The primary productivity peaks subside after September while mesozooplankton abundances increase and remain high in the post-monsoon period (Madhupratap et al. 2001).

## **Oil sardine life cycle and fishery**

The Indian oil sardine fishery is restricted to the narrow strip of the western India continental shelf, within 20 km from the shore (Figure 1). The yearly cycle (Figure 2) of the fishery begins at the start of spawning during June to July, corresponding with the onset of the southwest monsoon (Chidambaram 1950, Antony Raja 1969) when the mature fish migrate from offshore to coastal spawning areas. The spawning begins during the southwest monsoon period when temperature, salinity and suitable food availability are conducive for larval survival (Chidambaram 1950, Murty and Edelman 1971, Jayaprakash and Pillai 2000, Krishnakumar et

al. 2008, Nair et al. 2016). Although peak spawning occurs in June to July, spawning continues into September (Hornell 1910, Hornell and Nayudu 1923, Antony Raja 1969, Prabhu and Dhulkhed 1970) and early- and late-spawning cohorts are evident in the length distributions of the 0-year fish. Spawning occurs in shallow waters outside of the traditional range of the fishery (Antony Raja 1964), and after spawning the adults migrate closer to the coast and the spent fish become exposed to the fishery.

After eggs are spawned, they develop rapidly into larvae (Nair 1959). The phytoplankton bloom that provide sardine larvae food is dependent upon nutrient influx from coastal upwelling and runoff from rivers during the summer monsoon and early fall. The blooms start in the south near the southern tip of India in June, increase in intensity and spread northward up the coast (Smitha 2010). Variation in the bloom initiation time and intensity leads to changes in the food supply and to corresponding changes in the growth and survival of larvae and in the later recruitment of 0-year sardines into the fishery (George et al. 2012). Oil sardines grow rapidly during their first few months, and 0-year fish (40mm to 100mm) from early spawning appear in the catch in August and September in most years (Antony Raja 1970, Nair et al. 2016). As the phytoplankton bloom spreads northward, the oil sardines follow, and the oil sardine fishery builds from south to north during the post-monsoon period. Oil sardines remain inshore feeding throughout the winter months, until March to May when the inshore waters warm considerably and sardines move off-shore to deeper waters (Chidambaram 1950). Catches of sardines are correspondingly low during this time for all size classes. The age at first maturity occurs at approximately 150 mm size (Nair et al. 2016), which is reached within one year. When the summer monsoon returns, the oil sardine cycle begins anew.

Catches along the Kerala coast are high throughout the year except during quarter 2, March-May (Figure 3). The age-distribution caught by the fishery varies through the year. The fishery is closed during June to mid-July during the monsoon and peak spawning, and when it resumes in mid-July, it is first dominated by 1-2.5 year old mature fish (Bensam 1964, Antony Raja 1969, Nair et al. 2016). In August or September a spike of 0-year (40mm) juveniles from the June spawning typically appears in the catch (Antony Raja 1969, Nair et al. 2016) and another spike of 0-year fish is sometimes seen in February from the late fall spawning (Prabhu and Dhulkhed 1967, 1970). From October through July, the catch is dominated by fish from 120mm-180mm (Antony Raja 1970, Prabhu and Dhulkhed 1970, Nair et al. 2016) which is a mix of 0-year, 1-year and 2-year fish (Nair et al. 2016, Rohit et al. 2018).

# Materials and Methods

## Sardine landing data

Quarterly fish landing data have been collected by the Central Marine Fisheries Research Institute (CMFRI) in Kochi, India, since the early 1950s using a stratified multi-stage sample design that takes into account landing centers, number of fishing days, and boat net combinations in fishing operations (Srinath et al. 2005). The quarterly landings for oil sardine landed from all gears in Kerala were obtained from CMFRI reports (1956-1984) and online databases (1985-2015) (CMFRI 1969, 1995, 2016, Pillai 1982, Jacob et al. 1987). The quarterly landing data were log-transformed to stabilize the seasonal variance.

## Remote sensing data

We analysed monthly composites of the following environmental data derived from satellite data: sea surface temperature (SST), chlorophyll-a (CHL), upwelling (UPW), the Oceanic Niño Index (ONI) and precipitation. The monthly means of the covariate time series are shown in Figure 4.

For sea surface temperature, we used Advanced Very-High Resolution Radiometer (AVHRR) data, which provides accurate nearshore SST values. Although the ICOADS product provides SST values for earlier years, ICOADS does not provide accurate nearshore temperatures. For 1981 to 2003, we used the Pathfinder Version 5.2 product on a 0.0417 degree grid. These data were developed by the Group for High Resolution Sea Surface Temperature (GHRSSST) and served by the US National Oceanographic Data Center. For 2004 to 2016, we used the NOAA CoastWatch SST products derived from NOAA's Polar Operational Environmental Satellites (POES).

For chlorophyll-a, we used the chlorophyll-a products developed by the Ocean Biology Processing Group in the Ocean Ecology Laboratory at the NASA Goddard Space Flight Center. For 1997 to 2002, we used the chlorophyll-a 2014.0 Reprocessing (R2014.0) product from the Sea-viewing Wide Field-of-view Sensor (SeaWiFS) on the Orbview-2 satellite. These data are on a 0.1 degree grid. For 2003 to 2017, we used the MODIS-Aqua product on a 0.05 degree grid. These CHL data are taken from measurements by the Moderate Resolution Imaging Spectroradiometer (MODIS) on NASA's Aqua Spacecraft. The SST and CHL data were averaged across thirteen 1 degree by 1 degree boxes which roughly parallel the bathymetry (Figure

1). The SST and CHL satellite data were retrieved from the NOAA ERDDAP server (Simons 2017).

For an index of coastal upwelling, we used the sea-surface temperature differential between near shore and 3 degrees offshore as described by Naidu et al. (1999) and Smitha et al. (2008). The index was computed as the average SST in box 4 off Kochi (Figure 1) minus the average SST in box 13. For SST, we used the remote sensing sea-surface temperature data sets described above. This SST-based upwelling index has been validated as a more reliable metric of upwelling off the coast of Kerala compared to wind-based upwelling indices (Smitha et al. 2008). The SST-based upwelling index and chlorophyll-a blooms are strongly correlated (Figure 5).

Precipitation data were obtained from two different sources. The first was an estimate of the monthly precipitation (in mm) over Kerala from land-based rain gauges (Kothawale and Rajeevan 2017); these data are available from the Indian Institute of Tropical Meteorology and the data are available from the start of our landing data (1956). The second was a remote sensing precipitation product from the NOAA Global Precipitation Climatology Project (Adler et al. 2016). This provides estimate of precipitation over the ocean using a global 2.5 degree grid. We used the 2.5 by 2.5 degree box defined by latitude 8.75 to 11.25 and longitude 73.25 to 75.75 for the precipitation off the coast of Kerala. These data are available from 1979 forward (NCEI 2017). The land and nearshore ocean precipitation data are highly correlated (Appendix E), supporting the use of the land time series as a proxy for the precipitation over the ocean off the Kerala coast.

The Oceanic Niño Index (ONI) is a measure of the SST anomaly in the east-central Pacific and is a standard index of the El Niño/Southern Oscillation (ENSO) cycle. The ONI index is 3-month running mean of ERSST.v5 SST anomalies in the Niño 3.4 region, based on centered 30-year base periods updated every 5 years. The ONI was downloaded from the NOAA National Weather Service Climate Prediction Center.

## Hypotheses

Our statistical analyses were structured around specific hypotheses (Table 1) concerning which remote sensing covariates in which months should correlate with landings in specific quarters. These hypotheses were based on biological information concerning how environmental conditions affect sardine survival and recruitment and affect exposure of Indian oil sardines to the

coastal fishery as reviewed in the introduction. The quarter 3 (Jul-September) catch overlaps the summer monsoon and the main spawning months. This is also the quarter where small 0-year fish from early spawning (June) often appear in the catch, sometimes in large numbers. Variables that affect or are correlated with movement of sardines inshore should be correlated with quarter 3 landings. In addition, pre-spawning (March-May) environmental conditions should be correlated with the spawning strength as adult oil sardines experience an acceleration of growth during this period along with egg development. The post-monsoon catch (October-May) is a mix of 0-year fish (less than 12 months old) and mature fish (greater than 12 months old). Variables that are correlated with spawning strength and larval and juvenile survival should correlate with the post-monsoon catch both in the current year and in future years, one to two years after.

Our hypotheses (Table 1) focus mainly on two drivers: upwelling and ocean temperature. We also test hypotheses concerning precipitation as this has historically been an environmental covariate considered to influence the timing of oil sardine landings. More recently, researchers have highlighted the influence of large-scale ocean processes, specifically the El Niño/Southern Oscillation, on sardine fluctuations; therefore we test the Ocean Niño Index (ONI) also. Chlorophyll density is directly correlated with sardine food availability and chlorophyll fronts are known to influence sardine shoaling. However our chlorophyll time series is short (1997-2015) and the statistical power for testing correlation with landings is low. Tests of chlorophyll are shown in the appendices but are not the focus of our analyses.

## Statistical models

We modeled the catches during the late-monsoon season (quarter 3, July-September) separately from the post-monsoon season (October-March). Thus there is no seasonality in our catch time series, as we analyzed a yearly time series of quarter 3 catches separately from a yearly time series of post-monsoon catches. We divided the catch in this way for biological and statistical reasons. Catch in quarter 3 (July-September) captures a mix of spawning age fish as it overlaps with the tail end of the spawning season, is affected by a fishery closure from July to mid-August during the summer monsoon, and is periodically inflated by the appearance of small 0-year fish from early summer spawning. In addition, the covariates that affect the timing of spawning, movement of post-spawning mature fish inshore, and early egg and larval survival may be different than those that affect later growth, survival and shoaling that exposes fish to the inshore fishery. Analyzing catch and covariate time series without seasonality also had

an important statistical benefit—we removed the problem of seasonality in the catch and all the covariates. The oil sardine life-cycle is seasonal and driven by the strong seasonality in this monsoon influenced system. A simple statistical model with quarters will explain much of the quarterly catch data since most of the yearly variability is due to seasonality and any environmental covariate with a similar seasonality will also show high correlation with the landings. Our goal was to explain year-to-year variability thus eliminating the confounding effect of seasonality in the data was important.

We tested ARIMA models on both quarter 3 and post-monsoon catch time series and found little support for auto-regressive errors (ARIMA models with a MA component) based on diagnostic tests of the residuals and model selection. The best supported ARIMA models were simple AR models ( $x_t = bx_{t-1} + \varepsilon_t$ ). This lack of strong autocorrelation in residuals has been found in other studies that tested ARIMA models for forecasting small pelagic catch (Stergiou and Christou 1996). We thus used AR-only models, however we tested both linear and non-linear models using generalized additive models (GAM) of the form  $x_t = s(x_{t-1}) + \varepsilon_t$ . The landings models were fit using conditional sum of squares (conditioning on the first 2 landings values in the time series). We investigated correlations between environmental variables and sardine catch using generalized additive models (GAMs, Wood 2017) to allow one to model the effect of a covariate as a flexible non-linear function. It was known that the effects of the environmental covariates were likely to be non-linear, albeit in an unknown way. Our approach is analogous to that taken by Jacobson and MacCall (1995) in a study of the effects of SST on Pacific sardine recruitment.

The first step in our analysis was to determine the catch model: the model for current catch as a function of the past catch. One feature of GAMs is that they allow the smoothing parameter of the response curve to be estimated. However we fixed the smoothing parameter at an intermediate value so that reasonably smooth responses were achieved and to limit the flexibility of the models being fit. Multi-modal or overly flexible response curves would not be realistic for our application. We used GAMs with smooth terms represented by penalized regression splines (Wood 2011, using the mgcv package in R) and fixed the smoothing term at an intermediate value ( $sp=0.6$ ).

Our catch models took the following forms

- random walk:  $\ln(C_{i,t}) = \alpha + \ln(C_{j,t-1}) + \varepsilon_t$
- AR-1:  $\ln(C_{i,t}) = \alpha + \phi_1 \ln(C_{j,t-1}) + \varepsilon_t$
- AR-2:  $\ln(C_{i,t}) = \alpha + \phi_1 \ln(C_{j,t-1}) + \phi_2 \ln(C_{k,t-2}) + \varepsilon_t$

- non-linear:  $\ln(C_{i,t}) = \alpha + s_1(\ln(C_{j,t-1})) + s_2(\ln(C_{k,t-2})) + \varepsilon_t$

where  $\ln(C_{i,t})$  is the log catch in the current year  $t$  in season  $i$ . We modeled two different catches: 3rd quarter catch  $S_t$  (July-September), which is during the late part of the summer monsoon, and post-monsoon catch  $N_t$  (October-June). The catches were logged to stabilize and normalize the variance.  $s()$  is a non-linear function estimated by the GAM algorithm. The model is primarily statistical, meaning it should not be thought of as being a population growth model. We tested models with prior year post-monsoon catch ( $N_{t-1}$ ) and 3rd quarter catch ( $S_{t-1}$ ) as the explanatory catch variable.  $S_t$  was not used as a predictor for  $N_t$ ;  $S_t$  is the quarter immediately prior to  $N_t$  and would not be available for a forecast model since time is required to process landings data. The catch models were fit to 1982 to 2015 catch data, corresponding to the years where the SST, upwelling and precipitation data were available. F-tests and AIC on nested sets of models (Wood et al. 2016) were used to evaluate the support for the catch models and later for the covariate models. After selection of the best model with the 1982-2015 data, the fitting was repeated with the 1956-1981 and 1956-2015 catch data to confirm the form of the catch models.

Once the catch models were determined, the covariates were studied individually and then jointly. As with the catch models, F-tests and AIC on nested sets of GAM models were used to evaluate the support for models with covariates. The smoothing term was fixed at an intermediate value (sp=0.6) instead of treated as an estimated variable. Our models for catch with covariates typically took the form  $\ln(C_{i,t}) = M + s_1(V_{1,t}) + s_2(V_{2,t-1}) + \varepsilon_t$  or  $\ln(C_{i,t}) = M + \beta_1 V_{1,t} + \beta_2 V_{2,t-1} + \varepsilon_t$  where  $M$  was the best catch model from step 1. Thus models with covariates modeled both as a linear and non-linear effect were compared. The covariates tested are those discussed in the section on covariates that have been hypothesized to drive the size of the sardine biomass exposed to the fishery. We tested both models with one and two covariates, and did not use correlated covariates in the same model.

## Results

### Catches in prior seasons as explanatory variables

The monsoon catch models were compared against a “naive” model which was the “last year’s catch” model (Table 2). The “naive” model has no estimated parameters and is a standard null model for time series modeling. Models with  $\ln(N_{t-1})$  (post-monsoon catch in prior year),

whether linear or non-linear, as explanatory covariate were strongly supported over the naive model and over models with  $\ln(S_{t-1})$  (monsoon catch in prior year) as the explanatory variable. Addition of the catch two years prior,  $\ln(N_{t-2})$  or  $\ln(S_{t-2})$ , did not reduce AIC and for  $\ln(N_{t-2})$  led to either no decrease in the residual error (MASE) or an increased the residual error for the model with linearity (Table 2, Linearity test). Addition of  $\ln(S_{t-2})$  did decrease the residual errors, but the was not warranted given the increased number of estimated parameters based on AIC. We also tested the support for non-linearity in the effect of the prior year catch on the monsoon catch. This was done by comparing models with  $\ln(N_{t-1})$  or  $\ln(S_{t-1})$  included as a linear term or as a non-linear function  $s()$  (Table 2, Linearity test). The residual error decreased using a non-linear response at the cost increased degrees of freedom. The result was only weak (non-significant) support for allowing a non-linear response based on AIC and the F-test. The full set of models tested, including tests using catch during the spawning months in previous seasons as a covariate are shown in Tables A1 and A2. The results were the same if we used the full landings data set from 1956 to 2015 (Table A3). Overall, the landings in prior seasons was only weakly explanatory for the monsoon catch, and the maximum adjusted  $R^2$  for these models was less than 30% (Table 2).

The results on model structure were similar for models of the post-monsoon landings ( $N_t$ ) during the post-summer monsoon months (Table 3), but the models explained much more of the variance (adjusted  $R^2 = 57.0$ ). The most supported model for  $N_t$  (Table 3) used a non-linear response to landings during the post-monsoon months of the previous season  $\ln(N_{t-1})$  with a non-linear response to quarter 3 landings two years prior  $\ln(S_{t-2})$ . There was low support for including landings earlier than two seasons prior or for using the quarter 3 landings during in the immediately prior season (Tables A4, A5, and A6). We did not test models for the October-June catch using the quarter 3 (July-September) catch in the current fishing season, so immediately prior. These data would not be available in a forecasting setting as the data require time to process.

## **Environmental covariates as explanatory variables**

There was no support for using precipitation during the summer monsoon (June-July) or pre-monsoon period (April-May) as an explanatory variable for the quarter 3 (July-September) or post-monsoon (October-March) catch (hypotheses S1 and S2; Tables B1 and B2). This was the case whether precipitation in the current or previous season was used, if precipitation was included as non-linear or non-linear effect, or if the smoothing term (degree of non-linearity

allowed) was estimated and thus not constrained, and if either precipitation during monsoon (June-July) or pre-monsoon (April-May) were used as the covariate. Quarter 3 overlaps with the spawning period and precipitation is often thought to trigger spawning, however we were unable to find any consistent association of catch during these spawning and early-post spawning months with precipitation. We also found no correlation between post-monsoon SST in the prior year (hypothesis L1) or the ONI index (hypothesis A2) for either July-September or post-monsoon catch (Tables B1 and B2).

However, we found significant correlation between the summer monsoon upwelling indices in the current season: average nearshore SST along the Kerala coast during June-September and the average SST nearshore versus offshore differential (UPW) off Kochi in June-September (Table 4, Table B3 and Table B4). These two upwelling indices are correlated but not identical. The model with average June-September nearshore SST was more supported than the model using the SST differential off Kochi. For July-September catch, this model with a non-linear response had an adjusted  $R^2$  of 41.0 versus an adjusted  $R^2$  of 24.4 for the model with no covariates (Table B3), and for October-March catch, the adjusted  $R^2$  was 61.8 versus 56.6 (Table B4). Note, that this covariate is June-September in the current season overlaps with the July-September catch. Thus this model cannot be used to forecast July-September catch but does help us understand what factors may be influencing catch during the monsoon.

The sea-surface temperature before spawning (March-May) has been speculated to be correlated with successful egg development and spawning behavior (hypothesis S4 and S5) and extreme heat events pre-spawning have been associated with low recruitment. This suggests that March-May in the current and prior years should be associated with low catch. The sea-surface temperature during larval and early juvenile development (October-December) affects survival and growth in multiple ways and thus could correlate with biomass in future years (hypothesis L1). There was no support for either of these variates as explanatory variables for the July-September catch and only weak support (based on AIC) for March-May SST in the current season for explaining variability in post-monsoon catch. The fall average SST in the prior season did not explain variability in either July-September or October-March catch. See Tables B3 and B4.

The average sea surface temperature over multiple prior years has been found to be correlated with sardine recruitment in Pacific sardines (Jacobson and MacCall 1995, Lindegren et al. 2013, Checkley et al. 2017) and southern African sardines (Boyer et al. 2001). We tested as a model covariate the average SST for 2.5 years prior to the July-September catch, so January-June in the current calendar year and the two prior calendar years for a 30-month average. This

covariate can be used for forecasting since it does not overlap with either July-September or October-March catch. This variate with a non-linear response was best covariate for both the July-September and the post-monsoon catch. For post-monsoon catch, the model with SST had an adjusted  $R^2$  of 67.5 versus 56.6 without. For the July-September catch, the adjusted  $R^2$  was 41.0 with SST and 24.4 without. The response curve was step-like with a negative effect at low temperatures and then an positive flat effect at higher temperatures (Figure 6). This is similar to the step-response found in studies of the correlation between average SST and recruitment in Pacific sardines (Jacobson and MacCall 1995).

Chlorophyll-a density is speculated to be an important predictor of larval sardine survival and growth. In addition, sardines shoal in response to coastal chlorophyll blooms, which brings them in contact with the coastal fisheries. Thus chlorophyll-a density is assumed to be an important driver of future or current sardine catches. We had chlorophyll-a remote sensing data only from 1998 onward. Our simplest covariate model required 5 degrees of freedom, thus we were limited in the analyses we could conduct. In addition, the years, 1998-2014, have relatively low variability in catch sizes; the logged catch sizes during this period range from 10-11 during quarter 3 and 11-12 during the other three quarters. Second degree polynomial models were fit (Appendix C) to the average log chlorophyll-a density in the current and prior season from quarter 3 (July-September), 4 (October-December), and 1 (January-March). Chlorophyll-a density was not a significant predictor for the spawning catch for any of the tested combinations of current or prior season and quarter. The only significant effect was seen for post-summer monsoon catches using chlorophyll-a density in October-December of the current and prior season (Table C1). This matches results which found that the upwelling index in October-December of the prior season was a predictor for the post-summer monsoon catch. The SST-based upwelling index and chlorophyll-a density are both indices of low-trophic level productivity.

## Discussion

Sardines in all the world's ecosystems exhibit large fluctuations in abundance (Baumgartner et al. 1992). These small forage fish are strongly influenced by natural variability in the ocean environment. Upwelling, influenced by both large-scale forces such as regimes shifts and El Niño/Southern Oscillation patterns (Alheit and Hagen 1997, Schwartzlose et al. 1999) and by seasonal wind and current patterns, brings nutrient rich and oxygen rich waters to the surface. This drives the seasonal variability in phytoplankton resources and in turn sardine prey (Bakun

et al. 2008). Local variability in temperature, salinity, and oxygen levels have both direct and indirect on sardine reproduction, recruitment and survival (Checkley et al. 2017). Sardines are also influenced by competition and predation by other species and well-known for their sensitivity to over-fishing which has been linked to many fishery collapses (Kripa et al. 2018).

Many studies on Pacific sardines have looked at the correlation between ocean temperature (SST) and recruitment. Temperature can have direct effect on larval survival and growth and an indirect effect on food availability. Studies in the California Current System, have found that SST explains year-to-year variability in Pacific sardine recruitment (Jacobson and MacCall 1995, Checkley et al. 2009, 2017, Lindegren and Checkley 2012) and that the average nearshore temperature over multiple seasons is the explanatory variable. Similar to these studies, we found that the average nearshore SST over multiple seasons was the covariate that explained the most variability in catch both in the monsoon and post-monsoon months. McClatchie et al. (2010) found no SST relationship with SST and Pacific sardine recruitment, however their analysis used a linear relationship while the other studies, and ours, that found a relationship (Jacobson and MacCall 1995, Checkley et al. 2017) allowed a non-linear relationship. Both Jacobson and MacCall (1995) and Checkley et al (2017) found a step-like response function for temperature: below a threshold value the effect of temperature was linear and above the threshold, the effect was flat and at lower temperatures the effect was negative and became positive as temperature increased. Our analysis found a similar pattern with a negative effect when the 2.5-year average temperature was below 28.35°C and positive above and with the positive effect leveling off above 28.5°C (Figure 6).

There were four outlier years when catch were much lower than expected based on prior catches: 1986, 1991, 1994 and 2013. The 2.5-year average SST predicted the collapses in 1986 and 1991 (Figure 7); the size of the residual with the covariate was much smaller than without the covariate. The largest collapse was in 1994 and the most recent, in our dataset, was 2014. The 2.5-year average SST did not predict the 1994 nor 2013 collapse. There was no change in the size of the residual with and without the covariate. In fact, none of the covariates we tested changed the size of the model residuals for 1994 nor 2013. The causes of these unusual declines appear either unrelated to the environmental factors we studied. This suggests either that other factors, biological or anthropogenic, drove these declines or that a particular combination of environmental factors led to the declines. It should also be noted that our upwelling indices captured on one aspect of upwelling: the nearshore intensity. Other aspects of upwelling also affect oil sardines, such as the spatial extent both along the coast and off the coast and the timing of the start of upwelling.

Seasonal productivity in the SE Arabian Sea upwelling is driven by the summer monsoon, which causes strong coastal upwelling that moves from the south to the north over the summer. This drives a strong seasonal pattern of zooplankton abundance (Figure 5). Despite the strong connection between sardine recruitment, growth and survival with upwelling, we found no correlation with upwelling in the prior season with landings. We did find a correlation between upwelling in the current season with landings in the current season. The biological reasons behind a positive relationship with upwelling are clear. Upwelling drives productivity and higher food resources in the current season leads to higher recruitment and higher numbers of 0-year fish in the landings or may bring sardines into the nearshore to feed where they are exposed to the fishery. However, the explanatory power of the upwelling indices was mainly due to the negative effect of extremely high upwelling (Figure 6). Extremely high upwelling transports larval sardines offshore and can create regions of low oxygen which sardines avoid.

## Conclusions

Remote sensing satellites can be used to detect changes in ocean physical, biological and chemical properties, such as surface temperature, winds, surface height, surface waves, rainfall and surface salinity, as well as the ecosystem and water quality changes. Unlike in-situ measurements, environmental measures from remote-sensing can be acquired rapidly and over large regions. However, which environmental covariates will improve forecasts is not obvious from oil-sardine life-history alone. We tested using many of the covariates known or suspected to have an effect on sardine spawning, growth and survival: precipitation, upwelling indices, ocean temperature and chlorophyll-a in various critical months of the sardine life-cycle. We found that the multi-year average nearshore ocean temperature explained the most variability in the landings. This covariate is not as directly tied to stages of the oil-sardine life-cycle as the other covariates we tested, though it does integrate over multiple influences (upwelling strength and temperature) over multiple years.

The temperature of the Western Indian Ocean, of which the Southeast Arabian Sea is a part, has been increasing over the last century at a rate higher than any other tropical ocean (Roxy et al. 2014) and the warming has been most extreme during the summer monsoon months. This ocean climate change is affecting oil sardine distributions with significant landings now occurring north of Goa (Vivekanandan et al. 2009). Continued warming is expected to affect the productivity of the region via multiple pathways, including both the direct effects of temperature change on the physiology and behavior of organisms and a multitude of indirect

effects (Moustahfid et al. 2018). These indirect effects includes changes to salinity, oxygen concentrations, currents, wind patterns, ocean stratification and upwelling spatial patterns, phenology, and intensity. Incorporating environmental covariates into landings forecasts has the potential to improve fisheries management for small pelagics such as oil sardines in the face of a changing ocean environment (Tommasi et al. 2016, Haltuch et al. 2019). However, monitoring forecast performance and covariate performance in models will be crucial as a changing ocean environment may also change the association between landings and average sea surface temperature.

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## Figure Legends

Figure 1. Close up of Kerala State with the latitude/longitude boxes used for the satellite data. Kerala State is marked in grey and the oil sardine catch from this region is being modeled.

Figure 3. Quarterly catch data 1956-2014 from Kerala. The catches have a strong seasonal pattern with the highest catches in quarter 4 Note that quarter 3 is July-Sept and that the fishery is closed July 1 to Aug 15, thus the fishery is only open 1.5 months in quarter 3. The mean catch (metric tonnes) in quarters 1 to 4 are 38, 19.2, 30.9, and 59.9 metric tonnes respectively.

Figure 2. The sardine life-cycle in the Southeast Arabian Sea and how it interacts with the fishery.

Figure 4. Remote sensing covariates used in the analysis. All data are monthly averages over Box 4 in Figure 1 on the Kerala coast off of Kochi. Panel A) Upwelling Index. The upwelling index is the difference between the near-shore sea surface temperature (SST) and the off-shore SST defined as 3 degrees longitude offshore. Panel B) Surface chlorophyll-a (CHL). The CHL data are only available from 1997 onward. Panel C) Sea surface temperature constructed from Advanced Very High Resolution Radiometer (AVHRR). Panel D) Average daily rainfall (mm/day) off the Kerala coast.

Figure 5. Key oil sardine life-history events overlaid on the monthly sea surface temperature in the nearshore and offshore and the nearshore chlorophyll density.

Figure 6. Effects of covariates estimated from the GAM models. Panel A) Effect of the 2.5 year average nearshore SST on catch during the catch during July-September (late spawning and early post-spawning) months. Panel B) Effect of upwelling (inshore/off-shore SST differential) during June-September in the current season on July-September catch. The index is the difference between offshore and inshore SST, thus a negative value indicates warmer coastal surface water than off-shore. Panel C) Effect of the 2.5 year average nearshore SST on catch during the catch during October-March (post-monsoon, age-0, -1, -2 year fish). Panel D) Effect of upwelling (inshore/off-shore SST differential) during June-September in the current season on October-March catch. Strong upwelling (positive upwelling index) in the larval and juvenile high growth period (Oct-Dec) is associated with higher early survival and larger cohorts of age-0 fish in the catch.

Figure 7. Fitted versus observed catch with models with and without environmental covariates. Panel A) Fitted versus observed log catch in the spawning months with only non-spawning catch in the previous season as the covariate:  $S_t = s(N_{t-1}) + \varepsilon_t$ . Panel B) Fitted

versus observed log catch in July-September with the 2.5-year average nearshore SST added as a covariate to the model in panel A. This model was:  $S_t = s(N_{t-1}) + s(V_t) + \varepsilon_t$ . Panel C) Fitted versus observed log catch in the post-monsoon months with only post-monsoon catch in the previous season and July-September catch two seasons prior as the covariates:  $N_t = s(N_{t-1}) + s(S_{t-2}) + \varepsilon_t$ . Panel D) Fitted versus observed log catch in the post-monsoon months with 2.5-year average nearshore SST ( $V$ ) added as covariates. This model was  $N_t = s(N_{t-1}) + s(S_{t-2}) + s(V_t) + \varepsilon_t$ .

Table 1. Hypotheses for covariates affecting landings.  $S_t$  is quarter 3 (July-September) catch in the current season,  $S_{t-1}$  is quarter 3 catch in the previous season.  $N_t$  is the post-monsoon October-March catch in the current season and  $N_{t-1}$  is the October-March catch in the prior season. Because the fishing season is July-June,  $N_t$  spans two calendar years. DD = hypotheses related to effects of past abundance (landings) on current abundance. S = hypotheses related to spawning. L = hypotheses related to larval and juvenile growth and survival. A = hypotheses affecting all ages.

Hypothesis	Resp.	Covariates
DD1. $S_t$ is dominated by age 2+ fish, thus abundance of the 1-yr and 2-yr ages in the prior season (Oct-Mar catch) should be correlated with the abundance of mature fish this year.	$S_t$	$N_{t-1}$
DD2. Abundance of 1-yr and 2-yr fish should be correlated with strength of the cohorts from the previous two seasons.	$S_t$ and $N_t$	$S_{t-1}$ and $S_{t-2}$
DD3. Because age 2 fish appear in the post-monsoon catch, we expect the post-monsoon catch (dominated by age 1 and 2) in the previous season to be correlated with the post-monsoon catch in the current season.	$N_t$	$N_{t-1}$
S1. The onset of monsoon precipitation triggers movement of adults from offshore to spawning areas due to changes in salinity, turbulence or noise. Spent adults migrate inshore and are exposed to the fishery.	$S_t$	Jun-Jul precipitation in year $t$
S2. The level of precipitation in pre-monsoon months predicts spawning strength.	$S_t$	Apr-May precipitation in year $t$
S3. Precipitation initiates and supports spawning. Spawning affects post-monsoon catch in current and future seasons.	$N_t$	Apr-May and Jun-Jul precipitation in year $t$ and $t - 1$
S4. Extremely high upwelling brings poorly oxygenated water and very low temperatures to the surface causing mature fish to avoid nearshore areas and leads to lower exposure to the fishery.	$S_t$	Jun-Sep upwelling index in year $t$
S5. Extreme heat events in the pre-spawning months cause mature fish to move offshore away from productive feeding areas leading to poor spawning condition. Poor recruitment leads to few 0-age in post-monsoon catch and few 1-age fish in next season catch.	$S_t$ and $N_t$	Nearshore Mar-May SST in year $t$ and $t - 1$

Table 1. Continued.

Hypothesis	Resp.	Covariates
L1. Larval growth and survival is highest in an intermediate temperature window. The prior year post-monsoon larval survival and growth is associated with higher current year biomass.	$N_t$ and $S_t$	Nearshore SST during Oct-Dec in year t-1
L2. Upwelling is associated with higher productivity and higher density of zooplankton, which leads to better larval and juvenile growth and survival. The strength of summer upwelling should be associated with higher biomass in future years and the appearance of 0-age fish in post-monsoon catch. However, extremely strong upwelling brings poorly oxygenated water to the surface causing larval mortality and offshore advection and causes mature fish to move offshore.	$N_t$ and $S_t$	Jun-Sep upwelling index in year $t - 1$ and $t$
L3. Chlorophyll blooms are signatures of high productivity from nutrient influx either due to upwelling or coastal inputs. The monsoon bloom intensity should be associated with 0-year fish abundance in year $t$ and future sardine biomass.	$N_t$ and $S_t$	Chl-a density Jun-Sep in year $t - 1$ and $t$ (for $N_t$ )
A1. The multi-year average SST is associated with a variety of factors which affect spawning and early survival and has been found to correlate with sardine recruitment (Checkley et al. 2017) and thus future biomass available to the fishery.	$N_t$ and $S_t$	3-year average SST
A2. The changes brought about by the El Niño/Southern Oscillation (ENSO) cycle have a variety of effects on environmental parameters (precipitation, SST, thermal fronts, wind) which impacts spawning and early survival with a 1-year lag.	$N_t$ and $S_t$	ONI in year t-1

Table 2. Model selection tests of time-dependency and linearity for the monsoon (Jul-Sep) catch model using F-tests and AIC of nested models fit to log of landings data.  $S_t$  is the catch during the monsoon (Jul-Sep) of season  $t$ .  $N_{t-1}$  is the post-monsoon (Oct-Mar) catch in the prior sardine season.  $N_{t-2}$  is the same for two seasons prior.  $s()$  is a non-linear function of the response variable. MASE is mean absolute squared error of the residuals. The tests are nested and the numbers before the models indicates the nest level. In the list with numbers 1, 2, 3a, 3b, two nested tests were done; one on 1, 2, 3a and one on 1, 2, 3b. A simple base model may be indicated as M0 or M1, and the nested model as M0 (or M1) + an extra component. See tables in Appendix A for the full set of time-dependency and linearity tests. This table only shows the nested model sets that were most supported. The top model based on smallest AIC is marked with an arrow.

Model	Residual df	MASE	Adj. R2	F	p value	AIC
Naive Model 1984-2015 data						
$\ln(S_t) = \ln(S_{t-1}) + \varepsilon_t$	32		1			122.85
AR-1 Model						
$\ln(S_t) = \alpha + \beta \ln(S_{t-1}) + \varepsilon_t$	30	0.814				114.14
Time dependency test						
1. $\ln(S_t) = \alpha + \ln(N_{t-1}) + \varepsilon_t$	31	0.856	14.2			111.78
2. M0 $\ln(S_t) = \alpha + \beta \ln(N_{t-1}) + \varepsilon_t$	30	0.794	22.2	4.06	0.053	109.59
3a. $\ln(S_t) = M0 + \beta_2 \ln(N_{t-2})$	29	0.797	19.6	0.01	0.919	111.57
3b. $\ln(S_t) = M0 + \beta_2 \ln(S_{t-2})$	29	0.778	20.8	0.45	0.508	111.09
Linearity and time-dependency test						
1. $\ln(S_t) = \alpha + \beta \ln(N_{t-1}) + \varepsilon_t$	30	0.794	22.2			109.59
$\Rightarrow$ 2. M1 $\ln(S_t) = \alpha + s(\ln(N_{t-1})) + \varepsilon_t$	28.6	0.761	24.4	1.26	0.287	109.52
3a. $\ln(S_t) = M1 + s(\ln(N_{t-2}))$	26.4	0.761	21.2	0.28	0.785	112.42
3b. $\ln(S_t) = M1 + s(\ln(S_{t-2}))$	25.9	0.724	26.3	1.09	0.367	110.63

Table 3. Model selection tests for the post-monsoon (Oct-Mar) catch model ( $N_t$ ) using F-tests and AIC.  $S_t$  is the catch during the monsoon (Jul-Sep).  $N_t$  is the catch during the post-monsoon period (Oct-Mar) of season  $t$ ; note the fishing season is defined as Jul-Jun not calendar year.  $S_{t-1}$  and  $N_{t-1}$  are the catch during the prior sardine season during and after the monsoon respectively.  $S_{t-2}$  and  $N_{t-2}$  are the same for two seasons prior. See Table 2 for more details.

Model	Residual df	MASE	Adj. R2	F	p value	AIC
Naive Model 1984-2014 data						
$\ln(N_t) = \ln(N_{t-1}) + \varepsilon_t$	31		1			90.87
Time dependency test						
1. $\ln(N_t) = \alpha + \beta \ln(S_{t-1}) + \varepsilon_t$	30	1.018	26.2			93.17
2. $\ln(N_t) = \alpha + \beta \ln(N_{t-1}) + \varepsilon_t$	28.6	0.978	37			88.28
Linearity and time-dependency test						
1. $\ln(N_t) = \alpha + \beta \ln(N_{t-1}) + \varepsilon_t$	29	0.978	37			88.28
2. M1 $\ln(N_t) = \alpha + s(\ln(N_{t-1})) + \varepsilon_t$	27.6	0.874	45	4.87	0.026	84.75
3a. $\ln(N_t) = M1 + s_2(\ln(N_{t-2}))$	25.4	0.805	46	1.09	0.357	86.11
$\Rightarrow$ 3b. $\ln(N_t) = M1 + s_2(\ln(S_{t-2}))$	24.8	0.743	57	11.05	0.007	79.53

Table 4. Top covariates for the monsoon (Jul-Sep) and post-monsoon (Oct-Mar) catch ( $S_t$  and  $N_t$ ) models. The models are nested; the number indicates the level of nestedness. Models at levels 2 and higher are shown with the component that is added to the base level model (M0 or M1) at top. The full set of covariate models tested are given in Appendix B. The fitted versus observed catches from the covariate models are shown in Figure 7.

Model	Residual df	MASE	Adj. R2	F	p value	AIC
Jul-Sep catch models with covariates						
$V_t$ = Jun-Sep SST current season						
$W_t$ = Jun-Sep UPW current season						
$Z_t$ = 2.5-year average SST						
1. M0 $\ln(S_t) = \alpha + s(\ln(N_{t-1})) + \varepsilon_t$	28.6	0.761	24			109.52
2a. $\ln(S_t) = M0 + s(V_t)$	25.9	0.683	41	3.84	0.025	103.43
2b. $\ln(S_t) = M0 + \beta W_t$	27.6	0.706	33	4.96	0.034	106.32
$\Rightarrow$ 2c. $\ln(S_t) = M1 + s(Z_t)$	23.7	0.641	47	5.43	0.01	101.65
Oct-Mar catch models with covariates						
$V_t$ = Mar-May SST current season						
$W_t$ = Jun-Sep upwelling current season						
$Z_t$ = 2.5-year average SST						
1. M1 $\ln(N_t) = \alpha + s(\ln(N_{t-1})) + s(\ln(S_{t-2})) + \varepsilon_t$	24.8	0.45	57			79.53
2a. $\ln(N_t) = M1 + s(V_t)$	22	0.413	63	2.53	0.089	76.01
2b. $\ln(N_t) = M1 + \beta W_t$	23.8	0.46	62	4.91	0.037	76
$\Rightarrow$ 2c. $\ln(N_t) = M1 + s(Z_t)$	22.7	0.36	67	4.98	0.016	71.88

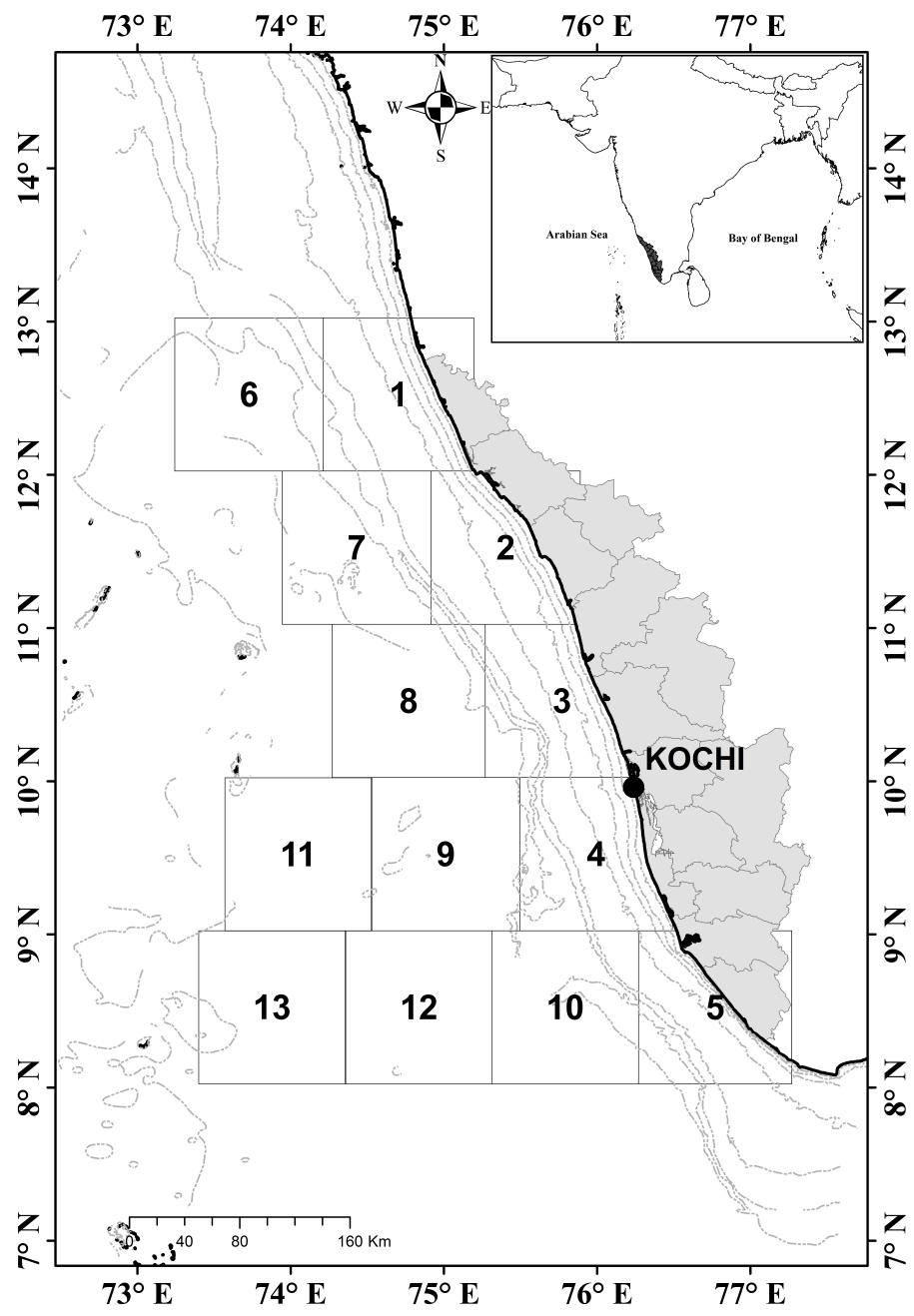


Figure 1

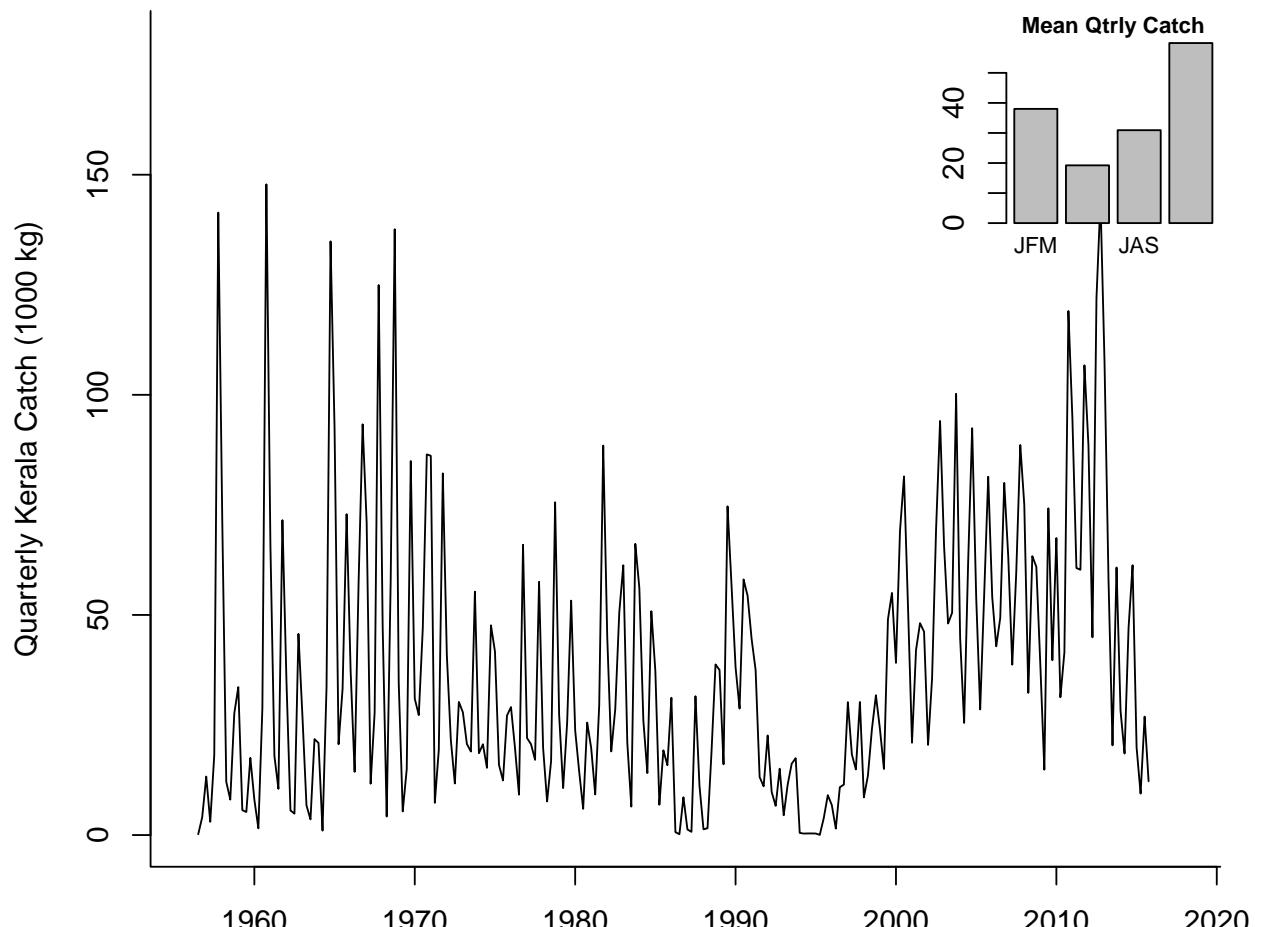


Figure 3

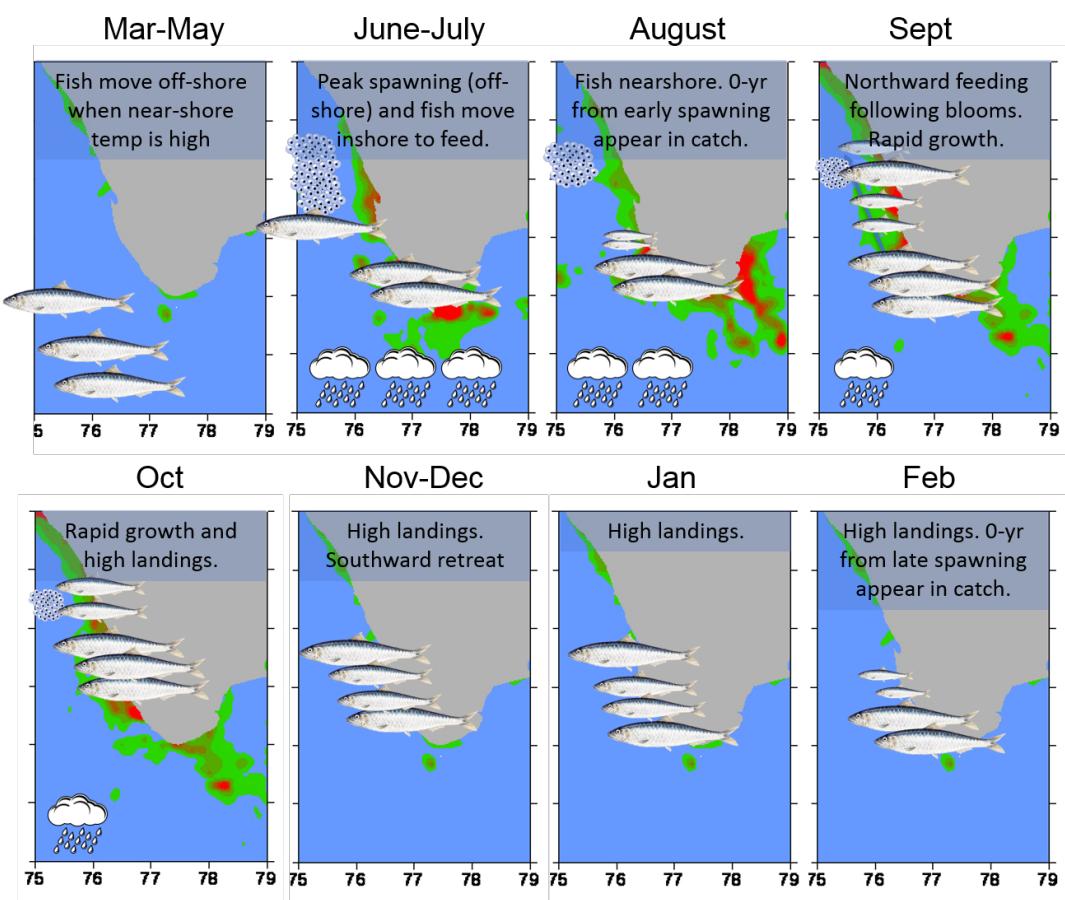


Figure 2

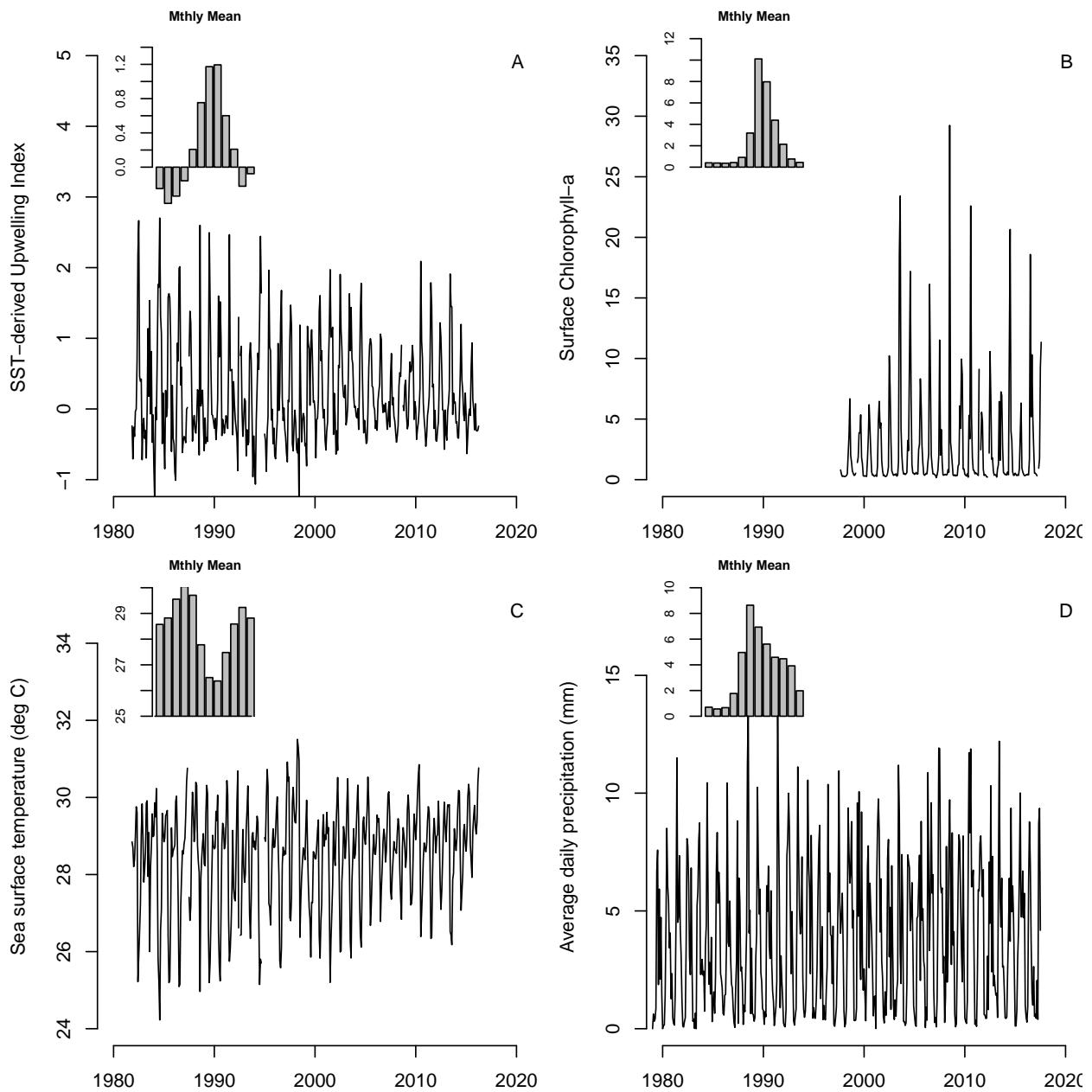


Figure 4

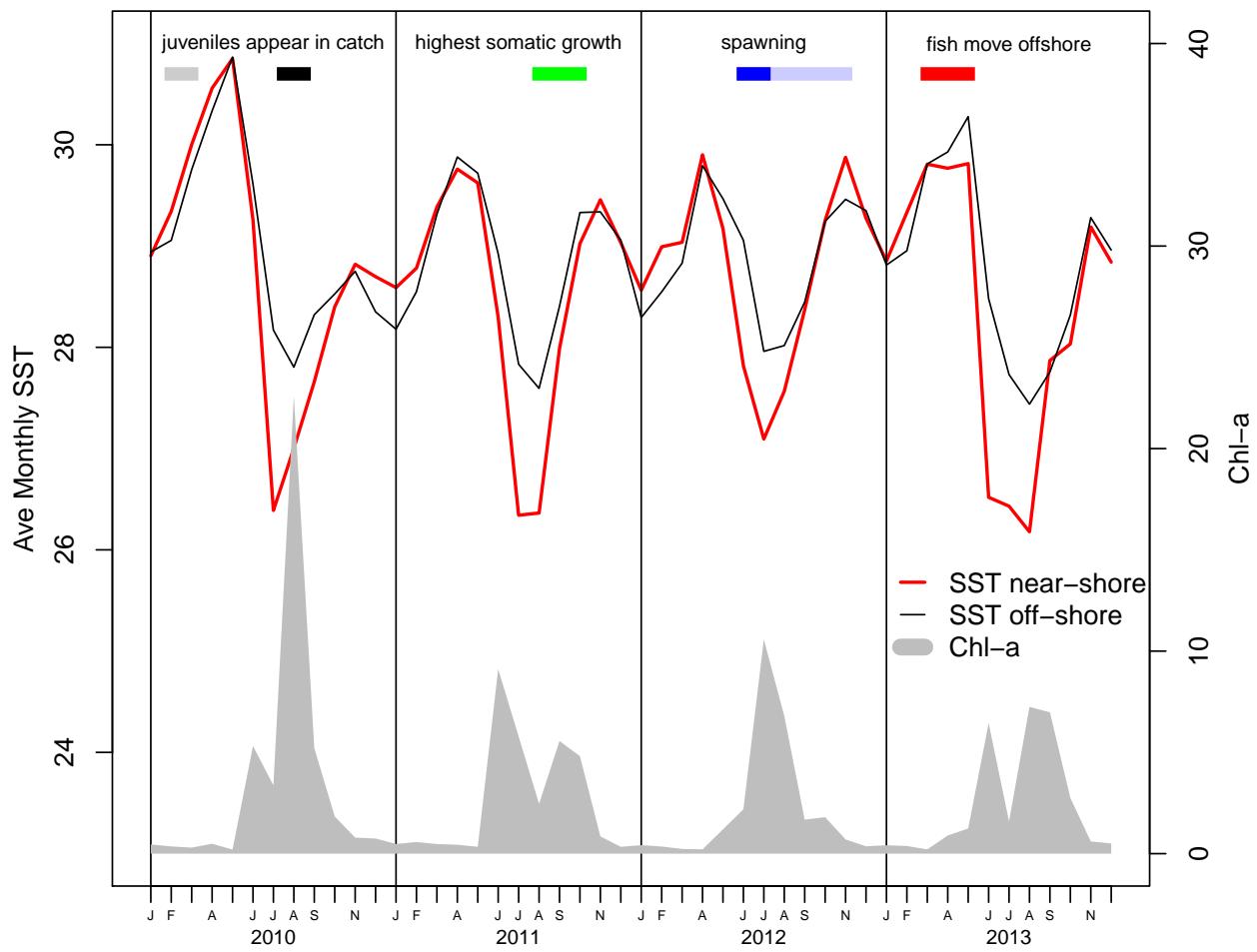


Figure 5

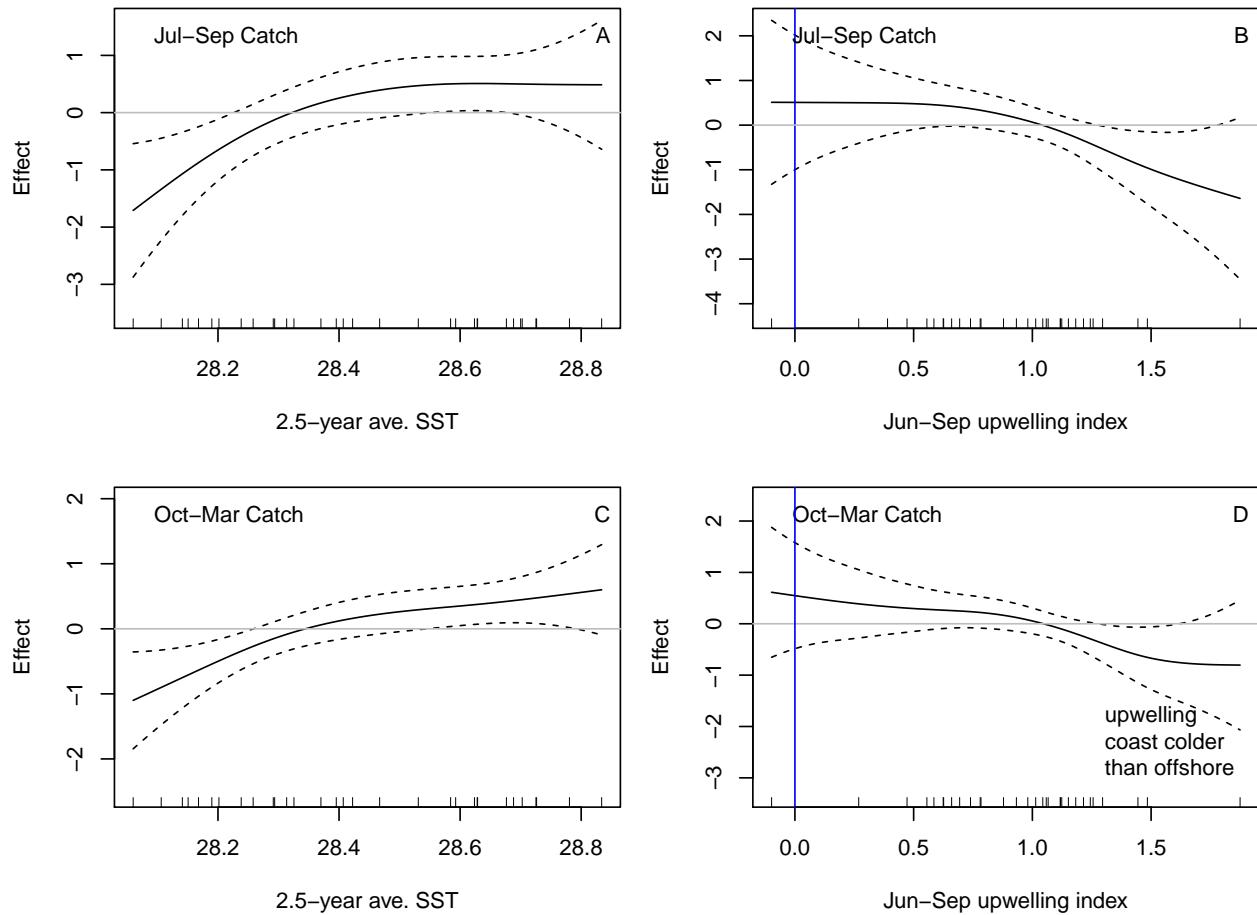


Figure 6

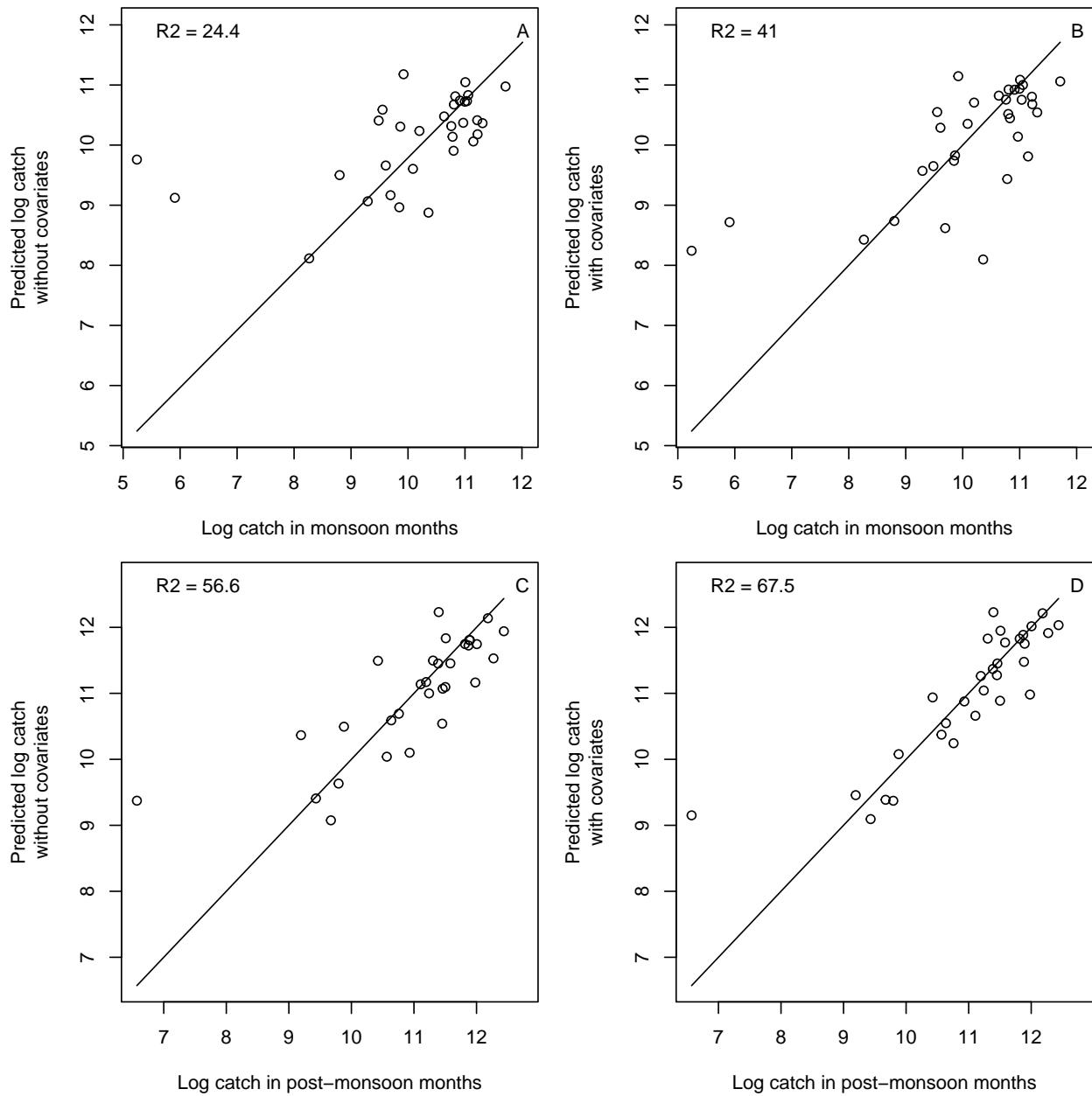


Figure 7