

Land-use change and biodiversity: Challenges for assembling evidence on the greatest threat to nature

Charles W. Davison¹  | Carsten Rahbek^{1,2,3,4,5}  | Naia Morueta-Holme¹ 

¹Center for Macroecology, Evolution and Climate, GLOBE Institute, University of Copenhagen, Copenhagen, Denmark

²Center for Global Mountain Biodiversity, GLOBE Institute, University of Copenhagen, Copenhagen, Denmark

³Institute of Ecology, Peking University, Beijing, China

⁴Department of Life Sciences, Imperial College London, Ascot, UK

⁵Danish Institute for Advanced Study, University of Southern Denmark, Odense, Denmark

Correspondence

Charles W. Davison, Center for Macroecology, Evolution and Climate, GLOBE Institute, University of Copenhagen, Universitetsparken 15, 2100 Copenhagen, Denmark.
Email: charles.w.davison@gmail.com

Funding information

H2020 Marie Skłodowska-Curie Actions, Grant/Award Number: 746334; Danmarks Grundforskningsfond, Grant/Award Number: DNRF96

Abstract

Land-use change is considered the greatest threat to nature, having caused worldwide declines in the abundance, diversity, and health of species and ecosystems. Despite increasing research on this global change driver, there are still challenges to forming an effective synthesis. The estimated impact of land-use change on biodiversity can depend on location, research methods, and taxonomic focus, with recent global meta-analyses reaching disparate conclusions. Here, we critically appraise this research body and our ability to reach a reliable consensus. We employ named entity recognition to analyze more than 4000 abstracts, alongside full reading of 100 randomly selected papers. We highlight the broad range of study designs and methodologies used; the most common being local space-for-time comparisons that classify land use in situ. Species metrics including abundance, distribution, and diversity were measured more frequently than complex responses such as demography, vital rates, and behavior. We identified taxonomic biases, with vertebrates well represented while detritivores were largely missing. Omitting this group may hinder our understanding of how land-use change affects ecosystem feedback. Research was heavily biased toward temperate forested biomes in North America and Europe, with warmer regions being acutely underrepresented despite offering potential insights into the future effects of land-use change under novel climates. Various land-use histories were covered, although more research in understudied regions including Africa and the Middle East is required to capture regional differences in the form of current and historical land-use practices. Failure to address these challenges will impede our global understanding of land-use change impacts on biodiversity, limit the reliability of future projections and have repercussions for the conservation of threatened species. Beyond identifying literature biases, we highlight the research priorities and data gaps that need urgent attention and offer perspectives on how to move forward.

KEYWORDS

biodiversity, climate change, geographic bias, land-use change, land-use history, named entity extraction, research priorities, taxonomic bias

1 | INTRODUCTION

The IPBES 2019 global report ranked land-use change as the greatest driver of declines in nature and biodiversity. Combined with the direct exploitation of nature through hunting, fishing, logging, and harvesting, these threats are thought to account for more than 50% of the human impacts on terrestrial and freshwater ecosystems (IPBES, 2019). During the past 300 years, the terrestrial biosphere has transitioned from mostly wild to mostly anthropogenic (Ellis et al., 2010); humanity's influence now extends over three quarters of the terrestrial world (Venter et al., 2016). Agricultural expansion for cropping, plantations, and animal rearing is attributed as the leading cause of global land-use change (IPBES, 2018). The expansion and intensification of human land uses has increased our share of the planet's resources at the expense of biodiversity and the ecosystem services it provides (Díaz et al., 2019; Foley et al., 2005). It is estimated that the biodiversity of terrestrial communities has declined on average by more than one-fifth (Hill et al., 2018). The global biomass of vegetation has halved (Erb et al., 2018) and that of wild mammals has dropped by more than 75%, now only accounting for 4% of mammalian biomass, the remainder being livestock and humans (Bar-On et al., 2018).

Land-use change can have direct impacts on species through the destruction of habitat and modification of the environment (Andren, 1994; Bender et al., 1998; Fahrig, 1997). However, there can also be widespread enigmatic ecological impacts that are small and cumulative, spatially removed, or difficult to detect (Raiter et al., 2014). Land-use change also interacts with other global change drivers. For example, it is known that road building facilitates the spread of invasive species (Forys et al., 2002; Schmidt, 1989) and increases accessibility for direct exploitation (Coffin, 2007; Peres, 2000). Interactions between land-use change and climate change are apparent but insufficiently researched (Santos et al., 2021; Sirami et al., 2017). The projected rise of both these threats may jeopardize the persistence of threatened species and communities. Despite land-use change presenting arguably the greatest threat to biodiversity (IPBES, 2019; Sala et al., 2000), research into this topic has lagged behind investigations into climate change impacts (Titeux et al., 2016). In the climate change and species distribution literature, the role of land use has been largely ignored (Taheri et al., 2021).

Several meta-analyses and syntheses have investigated the global impacts of land-use change on biodiversity. These studies generally report that land-use change substantially reduces local species richness (Beckmann et al., 2019; Gerstner et al., 2014; Murphy & Romanuk, 2014; Newbold et al., 2015), alters species composition (Dornelas et al., 2014; Newbold et al., 2016), and diminishes abundance (Collen et al., 2009; Gibson et al., 2011; Newbold et al., 2015). Species functional groups also show diverse responses to land-use change (Newbold et al., 2020). A meta-analysis on the effects of agricultural and silvicultural intensification found an overall decline in species richness across many production systems and species groups (Beckmann et al., 2019). Similarly, urbanization was found to correlate with sharp declines in bird and plant density (Aronson et al., 2014).

A frequent observation of these syntheses is that the response of biodiversity depends on numerous factors including biome, taxonomic group, type of disturbance, and the biodiversity metric used (e.g., Beckmann et al., 2019; Gibson et al., 2011). Furthermore, many acknowledge limitations in the analysis and extrapolation of conclusions due to geographic and taxonomic biases inherent in the literature (e.g., De Palma et al., 2016; Murphy & Romanuk, 2014). These issues make global synthesis and projection difficult. Moving this research field forward will require recognizing, reducing, and accounting for these challenges.

Here, we review the land-use change and biodiversity literature to describe the range of study designs, land-use measurements, and species responses that are often used. We investigate the taxonomic, geographic, climatic, and historical distribution of the research by using named entity recognition applied to a large body of papers. Our aim is to critically evaluate the evidence base that is being used to assess the global impact of land-use change on species and to suggest ways for moving this key research forward.

2 | METHODS

We identified relevant studies by searching Scopus for articles that investigated the impacts of land use or land cover change on any aspect of species biology. We limited our search to terrestrial studies published prior to 2020, excluding reviews, meta-analyses, and studies from the distant past (search terms in Figure 1d). The abstracts of all identified papers were used for automatically extracting species and location mentions, which were used in all subsequent analyses. This method may favor journals or papers that include this information in the abstract, although we assume these are still representative of wider trends. Additionally, the relative uniformity of abstracts versus full papers allowed us to assess a large chunk of the literature while minimizing uncertainties and irregularities. A subset of papers was also randomly selected for full reading and manual annotation (methods diagram Figure S1).

2.1 | Extracting taxonomic and location data

We applied automated text mining methods to the abstracts of all papers identified by the Scopus search to determine the taxonomic focus and geographic distribution of research investigating land-use change impacts on species. Our analysis used and built on the methods of Millard et al. (2019) who applied named entity recognition to study the animal pollination literature. To identify the mentions of species binomial names across all abstracts, we used the R package "Taxize," which incorporates dictionary string matching (Taxonfinder) and machine learning (Neti Neti) algorithms (Chamberlain & Szoecs, 2013; Chamberlain et al., 2020). The extracted taxonomic entities were matched with records in the 2019 Catalogue of Life database (Roskov et al., 2019), resolving to currently accepted names when necessary. Where extracted names were not directly matched with entries in the

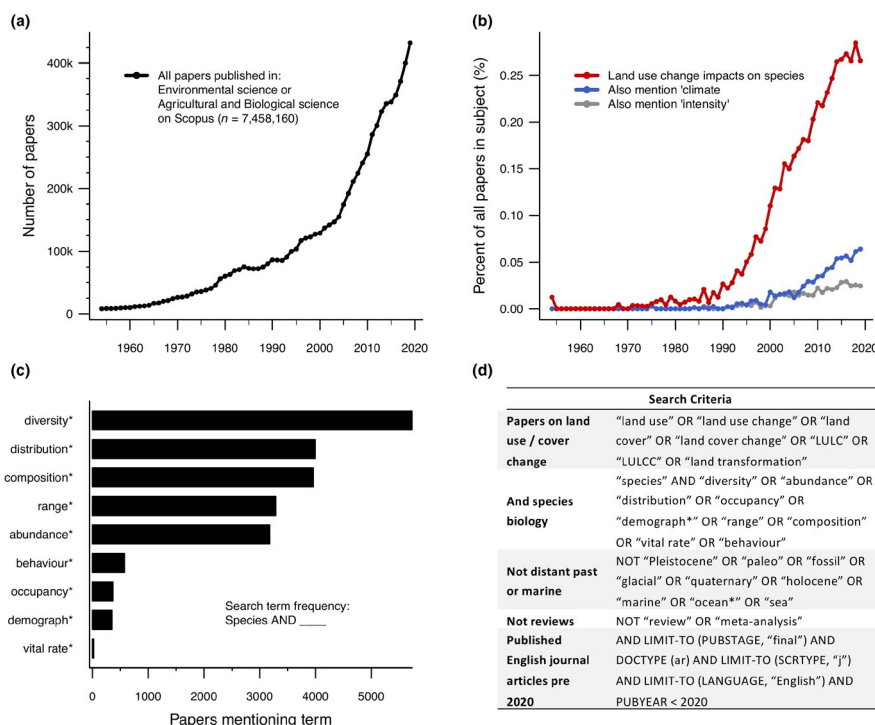


FIGURE 1 Search results. (a) The background rate of papers published in the subject areas "Environmental" or "Agricultural and Biological science" on Scopus searched 16/07/2020 (more than 93% of papers in our search have one of these subject areas). (b) Publications on the impact of land-use change on species are increasing faster than the background rate, that is, as a percentage of (a), and mentions of "climate" and "intensity" are slowly increasing. (c) The frequency of papers mentioning different aspects of species biology from the search criteria. (d) Full Scopus search terms

Catalogue of Life, we attempted a series of adjusted matches following Millard et al. (2019). Firstly, all punctuation was removed from the extracted name, and if a match was still not found, the abbreviation "spp" was also removed. For abbreviated extracted names (e.g., *Sus barbatus*), we found all matches with similarly abbreviated Catalogue of Life names, but only retained matches where the same genus was also mentioned in the abstract. For example, if an abstract mentioned *Acer rubrum* and *Acer saccharum* then we kept all Catalogue of Life entries for *A. saccharum* (including subspecies and varieties).

The process of matching extracted names with Catalogue of Life entries sometimes resulted in an extracted name being resolved to several accepted species, for example when the species was an ambiguous synonym or occasionally misapplied name. In these cases (9.9%), we kept only the directly matched accepted name from the same abstract (8.5%) or all candidate names if there were no direct matches (1.5%). Furthermore, all main analyses were conducted at genus level and above. We calculated the number of papers mentioning different taxonomic groups and used genus-level mentions to assess taxonomic representation. Additionally, species counts were aggregated into taxonomic groups and compared with the number of registered species per group in the Catalogue of Life 2019 (Figure S2; Roskov et al., 2019).

We extracted and geolocated all place name mentions (toponyms) from abstracts mentioning a species binomial using the Edinburgh Geoparser (Alex et al., 2015; Grover et al., 2010). Firstly, we removed from all abstracts any text following the copyright

symbol to eliminate locations associated with copyright information rather than the study itself (Millard et al., 2019). Next, we created individual text files of each abstract, which were batch processed with the Edinburgh geoparser using shell scripts. We kept only unique locations for each abstract and cleaned the results by removing any locations that were misidentified genera, species, or subspecies names, and by removing clearly spurious place names, for example, "Monte-Carlo," "Chao," and "Taxa" (full list in Table S2). To illustrate the geographic distribution of papers, we excluded mentions of continents and split all other geolocated places into two levels, country name mentions, and specific location mentions.

2.2 | Deriving additional data and validating

Based on the geolocated mentions identified in the previous step, we derived several attributes to assess the representativeness of the literature in a global context. Firstly, using a map of global biomes (Olson et al., 2001), we extracted the biome type at each of the specific location mentions and compared this with the global coverage of each biome. Secondly, we calculated the climate envelope covered by the studies and compared this with the global terrestrial climate using WorldClim version 2 Bioclimatic variables: Annual mean temperature and annual precipitation across 1970–2000 (10 arc minute resolution, Fick & Hijmans, 2017). Lastly, we assessed the land-use history of study locations using data derived from the KK10 past

human land-use model to determine the representation of historical land-use patterns (Ellis et al., 2013; Kaplan et al., 2011).

We manually assessed our search results and accuracy of automated text methods (full information and results in the Supplementary Material). We manually identified and resolved species in a random 1% of papers, and manually geoparsed 100 random abstracts mentioning a species to assess the accuracy of the automated geoparser and the relevance of papers.

2.3 | In-depth review of selected papers

To evaluate the methodological approaches used for studying land-use change impacts on species, we read and annotated a random selection of 100 papers. These were selected from papers that mention the term “land use” in their title, as these constituted a particularly relevant subset. These 100 papers were downloaded and processed using Zotero reference manager (Roy Rosenzweig Center for History & New Media, 2020). We extracted detailed information on the studies by manually adding keyword tags to each entry, which were then aggregated and summarized in R. Details included the geographic extent of studies, the experimental design used (following De Palma et al., 2018), the measure of land-use change, and overall findings. All data processing, analysis, and visualization were conducted in R version 4.0.3 (R Core Team, 2020).

3 | RESULTS AND DISCUSSION

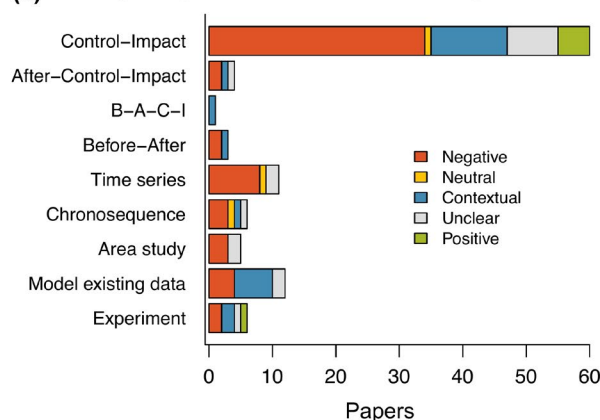
3.1 | Approaches for studying land-use change impacts on biodiversity

Our Scopus search yielded 12,192 papers investigating the impacts of land-use change on species. This research represents a rapidly growing field, with the rate of publication surpassing the background rate for the broader subject area (Figure 1a/b).

3.1.1 | Common study designs

A range of study designs are used in this field, with papers sometimes combining approaches; however, space-for-time substitutions (control-impact) are by far the most common, accounting for 61 out of the 100 papers in our detailed analysis (Figure 2a). Space-for-time studies are often the most straightforward to implement, resulting in an abundance of static data that when combined can facilitate broad-scale hypothesis testing. The PREDICTS project, for example, merges spatial comparisons of land-use impacts on species presence and abundance (Hudson et al., 2014). Analyses of this data set suggest reductions in local richness and abundance in response to land-use change, shifts in species composition, and unequal sensitivity of functional groups (Newbold et al., 2015, 2016, 2020). However, spatial comparisons make the often problematic assumption that sites are equal in all aspects besides

(a) Study design and effect of land use change on biodiversity



(b) Study extent and land use/cover data source

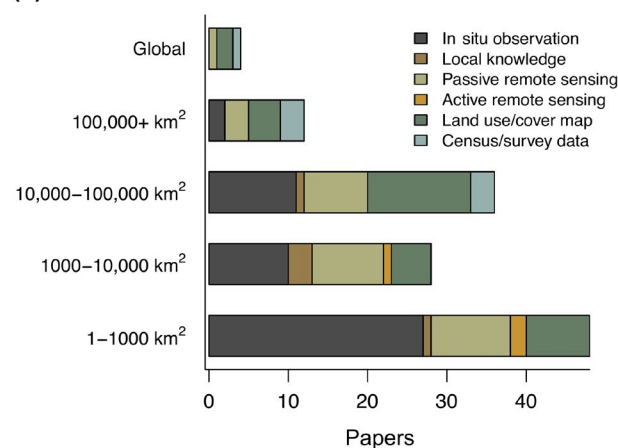


FIGURE 2 Features of manually assessed papers. (a) The frequency of different study designs and their general findings regarding the impact of land-use change on species attributes (e.g., richness, abundance, range). (b) The focal geographic extent and land-use/cover data source of papers

current land use, thereby overlooking important contextual information such as site history and the fact that land-use change is not randomly distributed (Damgaard, 2019). Static comparisons miss temporal interactions with climate change and fail to capture biotic lag effects that may take decades to unfold (e.g., Wearn et al., 2012).

Time-series were the second most common design, although only accounting for 11% of papers. Studies using this design can provide valuable information on rates of change if they run for long enough to encompass time lags are revisited at an ecologically relevant frequency and incorporate sites fully representative of the whole landscape (Collen et al., 2009; De Palma et al., 2018). Even so, it is difficult to ascribe changes to individual drivers as there can be many other influences including anthropogenic effects and natural population variability (Shoemaker et al., 2020). Global syntheses of time-series data have used various approaches to aggregate population trends, either by assessing the proportional changes in richness from the first to last year (Vellend et al., 2013), the slopes of individual time-series (Dornelas et al., 2014) or fitting

models to abundance trends (Collen et al., 2009). Syntheses are affected by the length and sampling frequency of studies (Cardinale et al., 2018; Gonzalez et al., 2016) as well as changes in research focus over time (Collen et al., 2009). For example, a shift from studying large, stable populations to studying small, threatened populations of conservation concern could artificially conflate temporal trends. When collating time-series studies, it is, therefore, especially important to establish clear historical baselines and to include studies investigating the same process (Cardinale et al., 2018; Gonzalez et al., 2016).

Detailed before-after-control-impact (BACI) assessments can capture temporal patterns and compare changes to a dynamic baseline. For example, França et al. (2016) showed that compared with BACI, a space-for-time substitution may underestimate the impacts of human disturbance on biodiversity by half. Yet BACI designs require a far greater investment of time and resources—resulting in limited replication and therefore limited utility in syntheses (De Palma et al., 2018). In our subset of 100 papers, only one was full BACI. Study designs that involved sampling prior to disturbance were generally the least common in our subset, probably because they require prior knowledge of any land-use change events, or the alteration of habitat for the sake of experiment. Our understanding of ecological processes could be enhanced by pursuing a joint framework that merges detailed causal models derived from experimental and BACI surveys with broad-scale models of static spatial variation.

All study approaches entail some trade-offs, either logistical or theoretical, and make key assumptions that can influence conclusions and impede efforts to synthesize results (reviewed in De Palma et al., 2018). Furthermore, all syntheses suffer from the fact that they combine studies that were designed for a variety of different purposes. The overall finding of manually read studies was for a negative impact of land-use change on species biology; however, a low sample size for all except control-impact studies precluded us from identifying differences in the distribution of findings across study designs. The impact of land-use change on species depends on context, for example, intensifying the use of an abandoned meadow leads to increased species diversity, but intensification of a heavily grazed meadow causes declines (Beckmann et al., 2019). This emphasizes the importance of using appropriate baselines and considering reference natural ecosystems when the interest lies in the conservation of undisturbed conditions (Trimble & van Aarde, 2012). Meta-analyses using space-for-time studies generally find negative impacts of land-use change on local species richness (e.g., Newbold et al., 2015); however, compilations of time-series have suggested changing community composition is more prevalent (e.g., Dornelas et al., 2014). Considering the disparate conclusions reached by these meta-analyses, the influence of study design is a question that deserves further attention.

3.1.2 | Measuring land-use change

Land-use change is challenging to quantify and often involves sorting complex habitats into coarse anthropocentric cover categories such

as forest, agriculture, and urban—a simplification that can overlook ecologically significant within class variation. A protected, floristically diverse grassland, for example, may support more and different species than a newly created grass field used for recreation. From an organism's perspective land-use classifications may be arbitrary, as they respond to myriad changes in difficult to measure environmental conditions such as light availability, resources, microclimate, and competition (Lauber et al., 2008; Stevens et al., 2011; Vallecillo et al., 2009).

There are many methods used to measure land-use change (De Palma et al., 2018; Verburg et al., 2011). We found that most studies used direct local observations (Figure 2b), which are typically accurate but can be limited in spatial scale. Remote sensing has been a boon for detecting broad-scale patterns of land-use change; we found this to be a commonly used method at all spatial scales, both directly and as a component of existing land-use maps (Figure 2b). However, the history of an area is crucial for biodiversity, yet satellite images of a 200-year-old forest and one that was deforested 40 years ago may appear broadly similar. Despite covering a shorter period than optical imagery, new active remote sensing technologies, such as LiDAR, let us make increasingly detailed 3D habitat measurements over large expanses (Simonson et al., 2014), and thus better identify structural differences, for example, between old and young forests. These active sensors were the least common method of describing land use in our analysis, which may reflect the relative novelty and expense of this data source.

Data on land-use intensity, such as fertilizer application or livestock per hectare, attempt to reconcile observed changes in biodiversity to within land-use variation, and capture more detailed facets of land-use change (Kehoe et al., 2015). The availability of intensity data is currently scarce, especially at the global level (Kuemmerle et al., 2013). Human appropriation of net primary productivity is an example metric that tries to capture intensity of land use with a single, continuous, remotely sensed variable (Haberl et al., 2004). A recent meta-analysis of the impact of increasing land-use intensity on biodiversity lamented the lack of studies reporting detailed land-use intensity measures such as input of fertilizers or pesticides (Beckmann et al., 2019). Only 22 of the 100 papers we read considered some measurable aspect of land-use intensity in their analyses; this is echoed in the fact that across all papers from the Scopus search, only 9% mention the word "intensity" in their abstract (Figure 1b). Moving forward, researchers should focus on the ecological processes that link land use and biodiversity change and seek out biologically relevant land-use metrics tailored to the taxa and habitat in question. To tackle inconsistencies in the classification of land-use types, we echo the call for a globally harmonized land-use classification system to aid synthesis (Gerstner et al., 2014; Verburg et al., 2011).

3.1.3 | Measuring biodiversity

The choice of species response metric also influences our interpretation of land-use change impacts, with the perceived effects completely reversing in some scenarios. To illustrate this, consider an

agricultural area beside a national park, with both habitats containing equal density of a particular species. Considering occupancy or abundance, one might conclude that agriculture is suitable habitat. However, if vital demographic rates such as births, deaths, and migration are examined, one may find that the agricultural area constitutes a sink habitat, with an elevated death rate that is concealed by immigration (e.g. Lamb et al., 2017). Information on animal species behavior may even elucidate that individuals use disturbed areas but are unable to shelter or reproduce within them (Love et al., 2018; Luskin et al., 2017). In these situations, the underlying dynamics may have been missed, and the long-term persistence of the population jeopardized.

Of the search terms used to identify the species level effects of land-use change, "diversity" was the most common, with terms related to species composition, distribution, and abundance also occurring frequently across the 12,192 papers (Figure 1c). Terms associated with more in-depth investigations such as demography, behavior, and vital rates were much scarce. Our findings are consistent with a meta-analysis of disturbed and undisturbed sites in tropical forests, which found that richness and abundance were the most often reported measures, whereas demographics were the least (Gibson et al., 2011). The authors also showed that the magnitude of the negative impact of disturbance varied according to the ecological metric used. For example, richness responded more strongly to land-use change than abundance, likely due to increasing abundance of some generalist species (Gibson et al., 2011). Aggregate community metrics, such as richness, ignore species identity and can lead to erroneous conclusions when shifts in community composition are missed (Mendenhall et al., 2012). Indeed, some authors suggest that changes in beta-diversity may be the main impact of land-use change (Dornelas et al., 2014; Vellend et al., 2013).

Information on demographics and vital rates help explain how some species persist or thrive in disturbed environments (e.g., Srinivasan et al., 2015). Linking these measurements to functional traits may help us determine what characterizes winners and losers. Examining species behavior can also provide insights into subtle land-use change effects (Chapman et al., 2019; Davison et al., 2019). Still, investigating these detailed variables is difficult, resource demanding, and subject to unique local variation. Basic presence data is easier to collect and is available globally for many taxa (e.g., BIEN - Enquist et al., 2016; GBIF, 2021). Nonetheless, species presence data overlooks important variation in abundance, and the sampling procedures can be biased or inaccurate (Anderson, 2012; Araújo et al., 2019). Abundance data facilitates quantitative synthesis across comparable studies (e.g., Newbold et al., 2020), yet in-depth research can also be integrated to provide important insights. Gaynor et al. (2018) for example, conducted a meta-analysis showing that many mammals are becoming increasingly nocturnal in response to human disturbance. Going forward, the trade-off between data volume and the level of insight must be carefully balanced to maximize our understanding of the distribution, and underlying mechanisms, of land-use change impacts.

3.2 | Biases in the literature

3.2.1 | Taxonomic biases

For understanding the global effects of land-use change on biodiversity, it is essential to sample a diverse range of species. Broad taxonomic coverage helps us capture complex indirect impacts and trophic feedback (Barnes et al., 2017). Excessive focus on charismatic species can be problematic from a conservation standpoint as their effectiveness as umbrella species is debated (Simberloff, 1998; Williams et al., 2000). Our application of taxonomic entity recognition discovered 9864 species binomial names from 4108 abstracts, 34% of the total papers. We found 77.4% of species were matched to an accepted name in the Catalogue of Life, and 13.5% were matched to an abbreviated accepted name; 7.8% were matched to a synonym, and 1.1% were matched to an abbreviated synonym. Species mentioned belonged to 2721 different genera, with most genera (60%) being mentioned only once (mean = 2.90, SD = 6.63, Figure 3b). Overall, we found strong taxonomic biases in the land-use change literature (Figure 3; Figure S2), in line with findings from the fields of biodiversity monitoring and conservation research (Di Marco et al., 2017; Troudet et al., 2017).

Plants were mentioned in 45% of all papers with a species in their abstract. Several genera of large trees including *Pinus*, *Quercus*, and *Eucalyptus* dominated (Figure 3b). The bias toward forested systems, may partly explain the prevalence of plants in this research area. However, we must interpret these findings with caution due to the role of forestry as a driver of land-use change and the practice of using plant species to describe study habitats (e.g., *Pinus sylvestris* plantations). In a random subset of 100 abstracts mentioning plant species, we found that 30% did not look at the response of plants to land-use change, indicating that the bias toward plants is not as extreme as estimated. Nonetheless, plants have been found to be the most studied group in other assessments of the literature dealing with invasion ecology (Pyšek et al., 2008), climate change (Felton et al., 2009), and tropical forest fragmentation (Deikumah et al., 2014). Fungi were the least studied group and were only mentioned in 1.7% of papers. Our choice of search strings may have some influence on these patterns if researchers of particular groups traditionally use different terminology instead of "land use," for example, "management" in fungal biodiversity studies along forestry intensity gradients.

Vertebrates were mentioned in 43% of papers, almost equaling plants, and were, therefore, overrepresented considering species richness of the group (Figure 3; Figure S2). Birds and mammals dominated among the vertebrates, which aligns with the observations from other meta- and literature analyses (e.g., Collen et al., 2009). Amphibians are one of the animal lineages with the highest share of threatened species (González-del-Pliego et al., 2019), yet they require far more attention in this field. Reptiles featured in the fewest papers of any vertebrate, despite having more described species than mammals and amphibians. The scarcity of papers investigating fish may have exacerbated effects given that terrestrial and freshwater biodiversity do not align spatially (Darwall et al., 2011). Rerunning the search

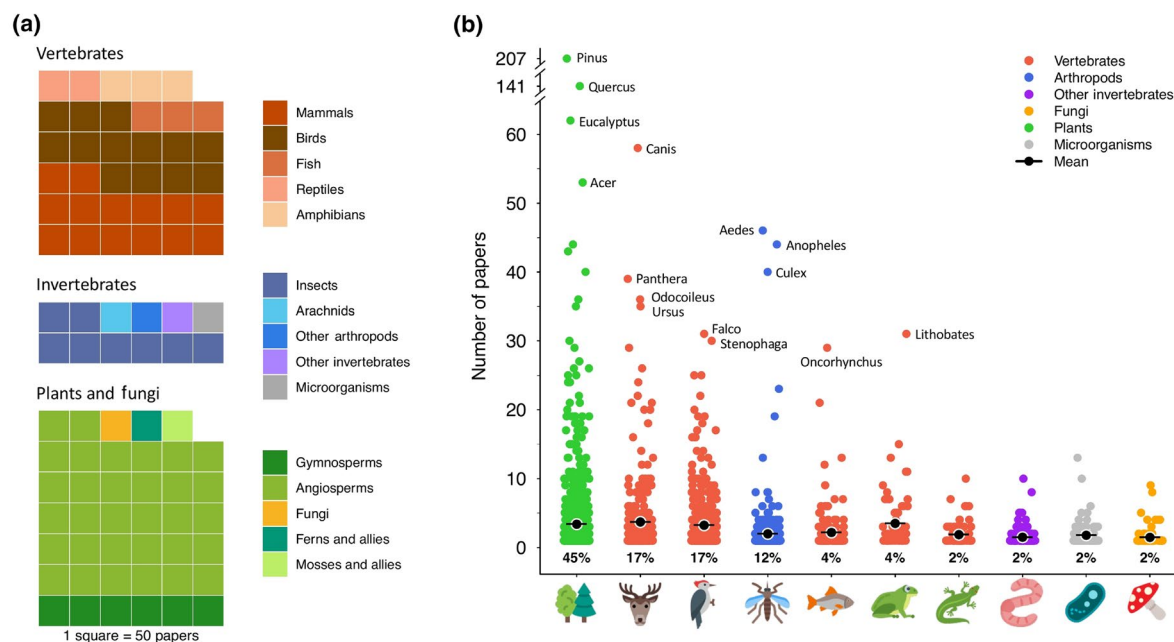


FIGURE 3 Taxonomic analysis. (a) Distribution of taxonomic mentions across papers in the literature sample. Some papers mention more than one group and are counted more than once; taxa present in less than 50 papers still occupy one square. (b) The number of papers mentioning different genera, ordered by percentage of papers mentioning each taxonomic group: plants, mammals, birds, arthropods, fish, amphibians, reptiles, other invertebrates, microorganisms, and fungi. Black lines indicate the average number of papers per genus in each group, some well represented genera are annotated

including terms “biomass,” “catch,” and “harvest” did not change the representation of fish in the results (original: 4.56%; new: 4.54%).

A few mosquito genera were the focus of many arthropod studies, likely due to their role as vectors of human disease and the relationship between land-use change and their prevalence (Conn et al., 2002). The stark difference between research on vertebrates and invertebrates matches the pattern seen in global biodiversity monitoring (Butchart et al., 2010; Troudet et al., 2017) and on evaluated groups in the IUCN red list (Cardoso et al., 2011). This disparity is concerning given the conservation of vertebrates does not inherently protect invertebrates (Lawton et al., 1998; Prendergast et al., 1993). A lack of past research on invertebrates has also reduced our ability to make reliable risk assessments (Karam-Gemael et al., 2020). Considering the dearth of studies on Fungi and invertebrates, the representation of detritivores in the literature appears deficient. Missing this key functional group may have implications for synthesizing how land-use change affects core ecosystem processes and feedback.

Taxonomic biases are a caveat to the conclusions of many land-use change meta-analyses. Some syntheses resort to investigating a handful of well-studied groups, trading generality for data availability (e.g., Aronson et al., 2014; De Palma et al., 2016). The PREDICTS project presents arguably the most taxonomically complete database but is still biased toward vertebrates and certain invertebrate groups (e.g., butterflies). Although equal representation of all taxonomic groups in research on land-use change is unrealistic, encompassing more groups may improve the selection of conservation areas (Kier et al., 2009). Rare species, despite constituting a large fraction of diversity (Enquist et al., 2019) are regularly omitted from analyses due

to low detectability; their omission means we are building our understanding of biodiversity change processes on common, wide-ranging species (Jetz & Rahbek, 2002). Integrating understudied taxa will benefit our understanding of the global process and further insights into how species and functional groups respond differently (Gibson et al., 2011; Newbold et al., 2020). Above all, a holistic ecosystem level approach will be possible, allowing consideration of indirect effects (Chillo et al., 2018), ecosystem functioning (Dislich et al., 2017), and trophic feedback (Barnes et al., 2017).

3.2.2 | Geographic biases

A well-distributed literature body across space helps us infer global patterns of land-use change, identify regional mechanisms, and highlight threatened areas. Of the 4108 abstracts mentioning species, 3213 also mentioned a geographic location. Of the 147 countries mentioned, the United States appeared most often, with roughly the same number of mentions (480) as the next five countries combined (Figure S4). The highest prevalence of studies clearly occurs in Europe and North America (Figure 4; Figure S3). The trend toward wealthy western countries matches that found in other literature bodies, particularly conservation research (Velasco et al., 2015). For instance, Martin et al. (2012) found that study prevalence in ecology research was strongly associated with Gross National Income—90% of study locations were within the wealthiest 30% of countries. Derived meta-analyses are, thus, rarely truly global. For instance, 68% of the data used in a global analysis of plant responses to land-use change

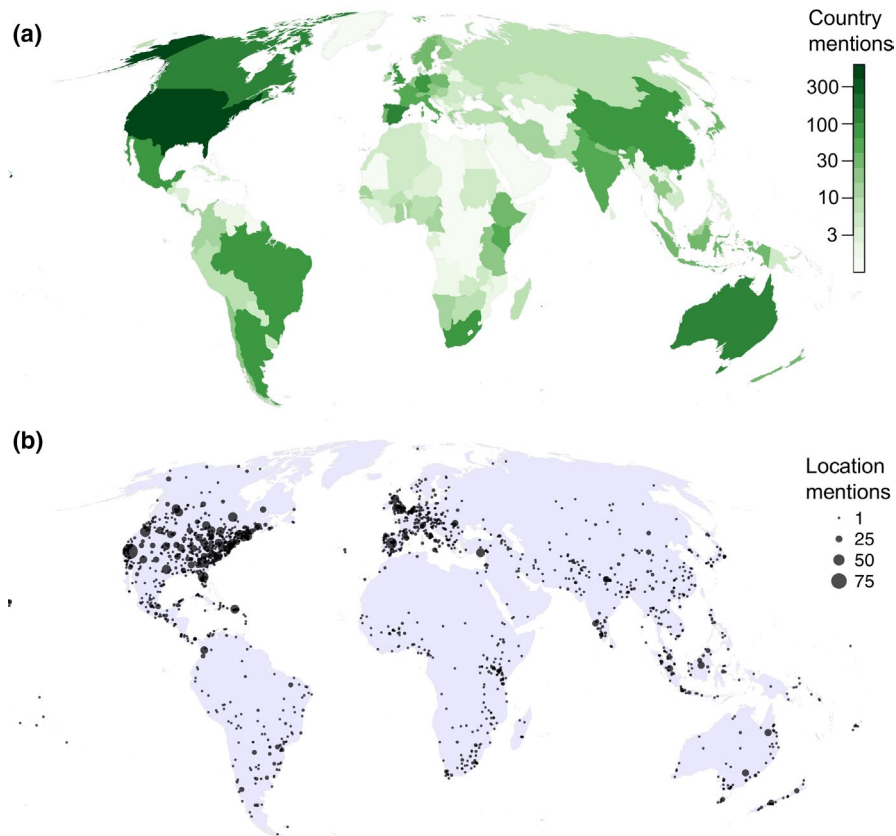


FIGURE 4 Geographic locations of papers. (a) Countries are colored by the number of papers mentioning them. (b) Black circles show the distribution and number of papers mentioning specific locations. A list of top countries mentioned, and a map of study density are in Figures S3 and S4

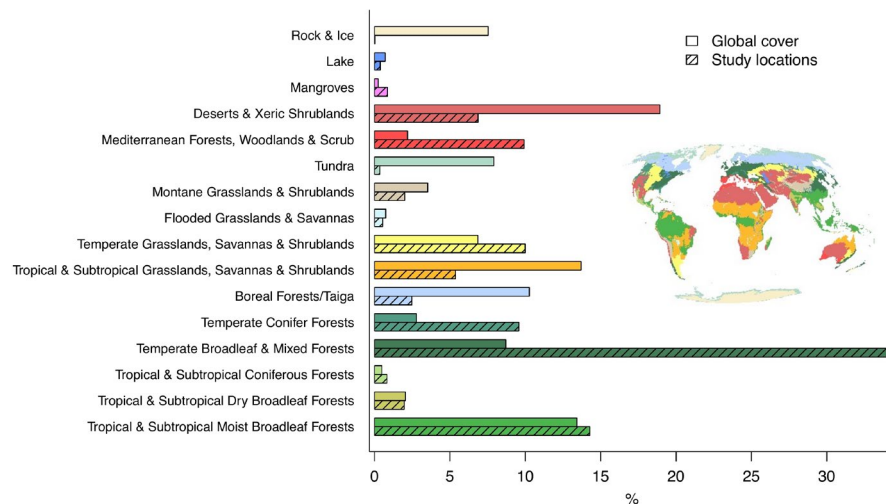


FIGURE 5 Representation of Earth's biomes: Horizontal bars compare the percent of all location mentions within each biome to the global cover of that biome. Map adapted from: (Olson et al., 2001). Locations are counted once for each study that mentions them

stemmed from Europe and North America and included no data from 73% of the world's countries (Gerstner et al., 2014).

In Russia, the Middle East, and Africa we found very few studies despite some hotspots in South Africa and Tanzania. The blank spots on the map—where there are few or no identified studies—reveal

gaps in our knowledge of region-specific land-use change drivers and how they may be impacting ecological communities. In the conservation literature at least, research in Africa and Southeast Asia appears to be stagnant or even decreasing over time, exacerbating their underrepresentation (Di Marco et al., 2017). Access to existing

research may also be hampering our understanding of some countries like Russia, where large amounts of biodiversity data exist but remain fragmented and digitally inaccessible to researchers (Ivanova & Shashkov, 2017).

The uneven geographic distribution of studies has consequences for the representation of biomes (*sensu* Olson et al., 2001, Figure 5). Based on the coordinates of subcountry locations, we detected a surplus of studies in all temperate biomes, accounting for over half of all locations. This abundance, and disparity with global area, was most apparent for temperate forest biomes. The most underrepresented biomes regarding land area are deserts and xeric shrublands, tundra, and boreal forests. Insufficient sampling of arid regions could hamper conservation efforts as Beckmann et al. (2019) found that arid areas, alongside tropical forests, had the highest loss of biodiversity following increased land-use intensity.

Community responses to land-use change can vary geographically due to the specific nature of threats and differences in species sensitivity (De Palma et al., 2016; Murphy & Romanuk, 2014). Indeed, turnover of assemblages across land-use classes is more pronounced in the tropics (Newbold et al., 2016) and the sensitivity of biota varies even between tropical regions (Gibson et al., 2011). Considering this, the data vacuum over the Afrotropics, which includes the second largest contiguous natural forest, is particularly concerning. Dependence on research from a limited geographic area, which is biased regarding the distribution of biodiversity and of land-use threats, hampers our global understanding of how land-use change impacts species. Future research should arguably be focused in areas with greater biological complexity, extinction risk, and/or current and future land-use pressure. Tropical regions emerge as a key candidate in all these aspects; they harbor the most species-rich assemblages on earth and are subject to increasing human pressure and rates of extinctions (Bradshaw et al., 2009). Improved sampling across biomes will be key to producing reliable syntheses of how land-use change affects biodiversity globally.

3.3 | Further dimensions of land-use change

3.3.1 | The role of climate

Land use and climate change interact with the implications for the distribution and persistence of species (Guo et al., 2018; Oliver & Morecroft, 2014; Santos et al., 2021). Climate change may shift the productive range of crop species poleward and uphill, influencing regional land-use trajectories. Land-use change, on the other hand, can affect regional and global climate directly through changes to the surface energy budget and indirectly through the carbon cycle (Pielke et al., 2002). Climate and land-use change may exacerbate and modify the negative effects of one another on biodiversity (Oliver & Morecroft, 2014), with the quantity and arrangement of suitable habitats in a landscape influencing the strength of this interaction (Pyke, 2004). Given this interaction the scarcity of studies in our analysis that even mention the word “climate” in their abstract highlights a concerning research gap (Figure 1b; see also Titeux et al., 2017).

In our analysis, we found that the climatic distribution of land-use study locations did not match with the pattern of global terrestrial climate (Figure 6). The global climate is dominated by large arid areas with high mean temperature ($>20^{\circ}\text{C}$); areas of low mean temperature ($\sim 0^{\circ}\text{C}$) and moderate precipitation; and areas where warmth and high rainfall coincide. In contrast, land-use change studies were concentrated in areas that had a mild mean annual temperature (~ 10 to 15°C) and total annual precipitation of approximately 1000 mm, that is, temperate conditions. The identified study locations underrepresent hot and cold arid regions and warm tropical areas.

The narrow climatic range of this research body may have several consequences as the richness of taxonomic groups peak in different climate zones (Prendergast et al., 1993) and species vary in their sensitivity to climate change (Angert et al., 2011). Mantyka-Pringle et al. (2012) found that the negative impacts of habitat loss and fragmentation on biodiversity were greatest in areas with higher maximum temperatures

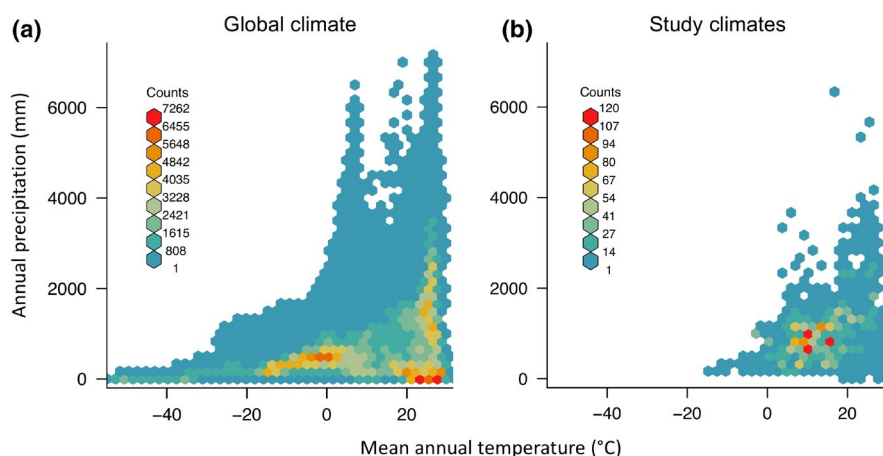


FIGURE 6 Representation of Earth's climate in the land-use change literature. The left-hand side shows the number of 10 arc minute cells that fall within each precipitation and temperature combination, while the right shows the number of study locations mentioned in each zone

Regions without significant human use in AD2000:

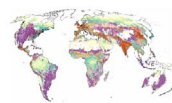
Recovery from peak use

- 0–1%
- 1–5%
- 5–10 %
- 10–20%
- 20–50%
- >50%
- NSU

Regions with significant human use in AD2000:

Years of significant use

- >8000
- 5000–8000
- 3000–5000
- 2000–3000
- 1000–2000
- 500–1000
- 250–500
- 100–250
- <100



Supplementary Figure S5

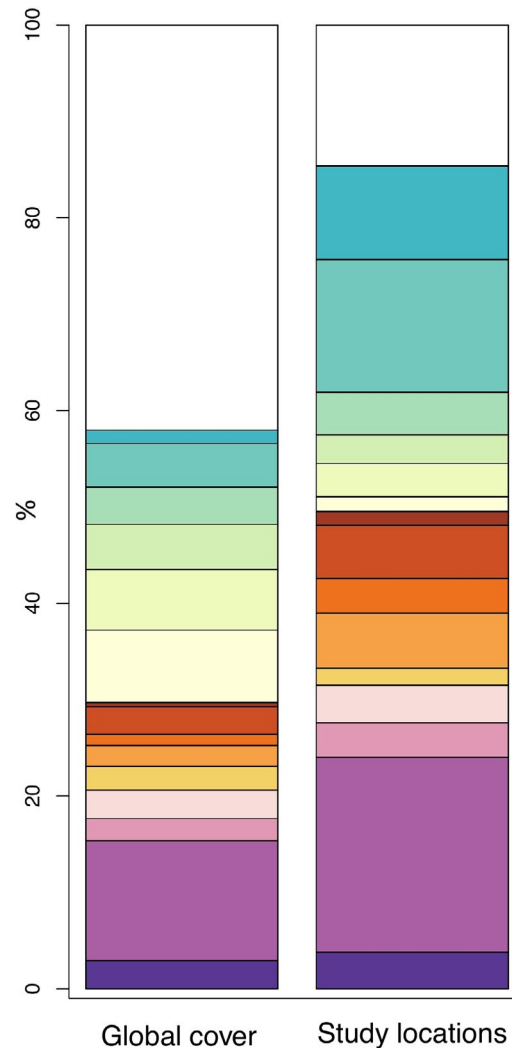


FIGURE 7 Representation of land-use histories as the percent of location mentions within each category versus global cover. Land-use histories identified from Ellis et al. (2013) based on the KK10 historical land-use model (Kaplan et al., 2011). Land-use histories are split depending on whether cells have significant (>20% of area) human use in the year AD 2000, in which case the number of years of sustained use above this threshold is given. For areas with less than 20% human use in AD 2000 the percent recovery from peak land use is given. Some areas have never been significantly used (NSU). Full map is shown in Figure S5

and lower precipitation. This finding may help explain why the impact of increasing land-use intensity on biodiversity may be greater in arid and tropical regions (Beckmann et al., 2019). Guo et al. (2018) also showed that forest loss and temperature positively interact to drive greater upslope movement of species in warmer regions. Surveying a broader range of climate conditions will allow us to capture confounding effects of land use and climate and may even help us predict the future dynamics of land-use change impacts under novel climates.

3.3.2 | Land-use history

Extensive and persistent influences of humans date back thousands of years in some regions (Ellis et al., 2013; Miede et al., 2014). The legacy effects of this past land use are detectable in current patterns of biodiversity from local to global scales (Debinski et al., 2011;

Polaina et al., 2019). Due to extinction debts and immigration credits, the time-lagged loss and gain of species following a disturbance, it is possible that not all species responses to today's disturbances are yet detectable (Jackson & Sax, 2010; Tilman et al., 1994). These transient states may influence study conclusions and can result in the underestimation of long-term threat levels (Hanski & Ovaskainen, 2002). Global maps of land-use history can help us identify prime regions for investigating the role of legacy effects and recovery times (Figure S5).

Synthesizing land-use change research requires covering the full diversity of land-use histories and understanding time-lagged biotic responses. We found that studies in areas with sustained human use (>20% of area used) until the present all had a higher share of studies than their global coverage suggests (Figure 7). The disparity in magnitude is particularly evident for areas with greater than 2000 years of sustained use, and areas with 100–250 years of sustained use. Coverage of these two extremes

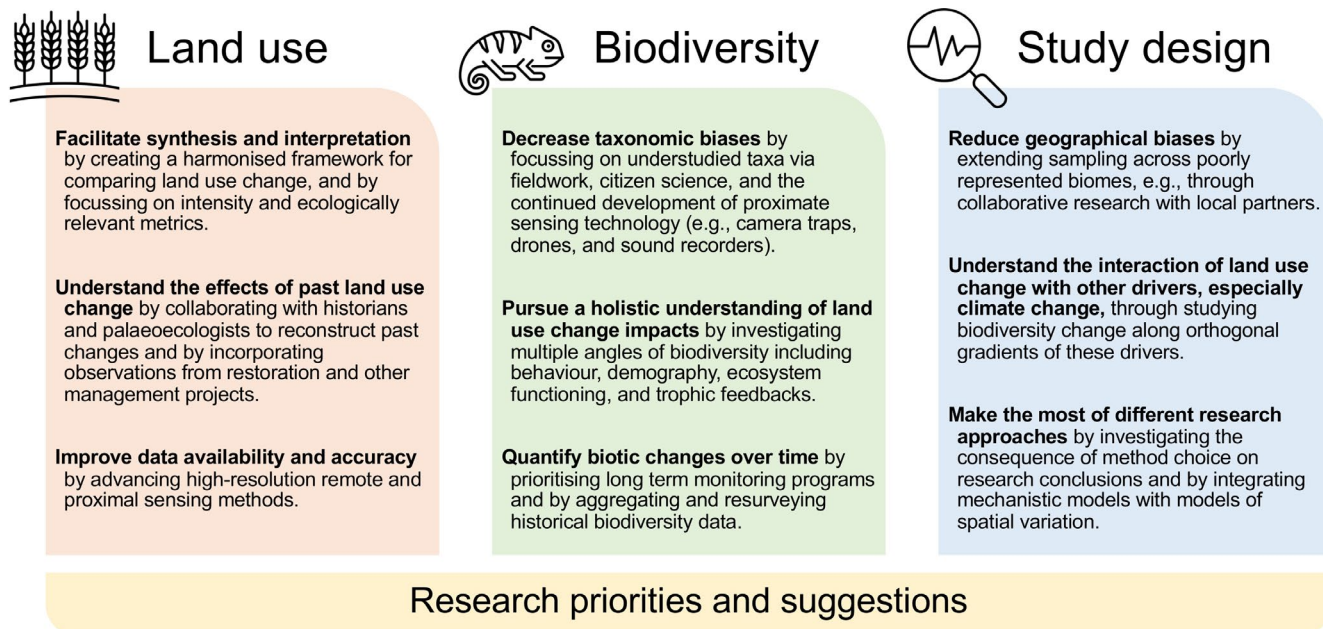


FIGURE 8 Nine recommended research priorities for developing a more comprehensive and balanced understanding of the impacts of land-use change on biodiversity

mainly results from the geographical bias toward Europe and North America, which respectively, have long and short histories of intensive use.

In some regions, human impacts have declined from previous levels and what appear to be pristine ecosystems may actually be recovering. Forests regrowing after agricultural abandonment, for example, can display the legacy effects lasting millennia (Bürgi et al., 2017; Dupouey et al., 2002). We found that land with less than 10% recovery from peak use was underrepresented in the literature, whereas areas with greater than 20% recovery were overrepresented. Greater focus on the response of species in recently recovering areas may provide valuable insights for conservation and ecosystem regeneration. Considering the remoteness of unused land (NSU, Figure S5) it featured a surprising number of studies. Martin et al. (2012) showed that ecologists typically place sites in pristine areas, even when they constitute small pockets in disturbed landscapes. However, the assumptions behind this tendency disregard the influence of humans in practically all terrestrial ecosystems (Faurby & Svenning, 2015; Martin et al., 2012). Moving forward, researchers must select appropriate baselines, encompass fluctuating population dynamics, and investigate the full range of land-use histories and recovering areas. Studies across orthogonal gradients of land use, history, and climate may ultimately help us isolate the effects of land use and its interaction with other drivers.

4 | CONCLUSION

Our comprehensive review of >4000 abstracts and 100 papers in the field of land-use change impacts on biodiversity shows that a

multitude of study designs and measurement characteristics are used, each with their strengths and weaknesses. Combining methods that complement each other and studying a range of species responses will aid synthesis and allow us to detect more subtle indirect effects of land-use change and understand ecological mechanisms. Making comparisons with appropriate baselines, both spatially and temporally, will clarify the direction and magnitude of biodiversity impacts. The taxonomic bias in this field restricts our understanding of biotic feedback and trophic cascades, and the dearth of information on some of the most sensitive and at-risk groups will hamper conservation efforts.

The geographic bias of studies limits the applicability of findings and the reliability of future projections. Investigating the full diversity of regional land-use mechanisms will be key to determining a global outlook. A bias toward forest ecosystems may be at the expense of other diverse habitats including tropical and subtropical savannas and grasslands. A key challenge moving forward will be exploring the interaction of land-use change with other global drivers as climate change is set to become an increasingly important factor into the future (IPBES, 2019). Forthcoming studies should focus on how climate interacts with land use, particularly in understudied tropical and arid climates.

The importance of land-use history and legacy effects is clear; the use of different baselines can lead to miscalculating the direction and magnitude of disturbance impacts. Although the literature reflected the breadth of land-use histories relatively well, recently recovering areas require more attention and may provide valuable insights for conservation. With North American and European studies dominating the research, we may be neglecting regional differences in current and historical land-use pressures. Failure to cultivate a balanced body of literature, accounting for all these inherent challenges, will

negatively impact our global understanding of land-use change impacts on species. These biases jeopardize the reliability of spatial and temporal projections of global change impacts on biodiversity and have repercussions for the conservation of imperiled species.

5 | PERSPECTIVES

To improve our understanding of how land-use change affects biodiversity, we need to investigate both sides of the paradigm in greater detail through improvements in land use and biodiversity data as well as study designs (Figure 8). Assessments of land-use change should incorporate within class variation in use intensity (Dullinger et al., 2021) and prioritize biologically relevant metrics over coarse, human-centered classes (Santos et al., 2021). Regarding biodiversity, more studies should include multiple trophic levels and focus on behavior and demographics to reveal potential mechanisms behind land-use impacts and uncover temporal trends within communities. Land use and biodiversity surveys across large extents (e.g., nationally) are important but need to be aligned spatially and temporally. Moving forward, we must also put more attention on the legacies of historical human disturbance when monitoring biodiversity and pursue further research into the interaction of land-use change with other global drivers (Titeux et al., 2017).

Geographic and taxonomic biases underpin many of the challenges facing this field. However, neither can be solved with the data that we currently have. Filling these gaps will require concerted fieldwork on neglected taxonomic groups in understudied regions, with an emphasis on collaborative research efforts. The advancement of space-borne remote sensing will also be pivotal. Global sub-meter resolution maps of vegetation and land use will be invaluable for tackling geographic biases when coupled with fieldwork in understudied systems. Proximal sensing methods such as camera traps, drones, sound recorders, and climate loggers will also help increase the effectiveness of data collection. We cannot make effective conservation and policy decisions on what we do not fully understand. In this paper, we have highlighted the research priorities and data gaps that need urgent attention.

ACKNOWLEDGMENTS

We thank Bennett Van Hoff for valuable assistance with data collection. We acknowledge the Danish National Research Foundation for funding for the Center for Macroecology, Evolution and Climate, grant no. DNR96. NMH was also supported by the European Union's Horizon 2020 research and innovation program under the Marie Skłodowska-Curie grant agreement no. 746334.

CONFLICT OF INTEREST

The authors declare no conflict of interest.

DATA AVAILABILITY STATEMENT

The data that support the findings of this study are available from the corresponding author on reasonable request.

ORCID

Charles W. Davison  <https://orcid.org/0000-0001-9709-0951>

Carsten Rahbek  <https://orcid.org/0000-0003-4585-0300>

Naia Morueta-Holme  <https://orcid.org/0000-0002-0776-4092>

REFERENCES

- Alex, B., Byrne, K., Grover, C., & Tobin, R. (2015). Adapting the Edinburgh geoparser for historical georeferencing. *International Journal of Humanities and Arts Computing*, 9(1), 15–35. <https://doi.org/10.3366/ijhac.2015.0136>
- Anderson, R. P. (2012). Harnessing the world's biodiversity data: Promise and peril in ecological niche modeling of species distributions. *Annals of the New York Academy of Sciences*, 1260(1), 66–80. <https://doi.org/10.1111/j.1749-6632.2011.06440.x>
- Andren, H. (1994). Effects of habitat fragmentation on birds and mammals in landscapes with different proportions of suitable habitat. *Oikos*, 71(3), 355–366. <https://doi.org/10.2307/3545823>
- Angert, A. L., Crozier, L. G., Rissler, L. J., Gilman, S. E., Tewksbury, J. J., & Chunco, A. J. (2011). Do species' traits predict recent shifts at expanding range edges? *Ecology Letters*, 14(7), 677–689. <https://doi.org/10.1111/j.1461-0248.2011.01620.x>
- Araújo, M. B., Anderson, R. P., Barbosa, A. M., Beale, C. M., Dormann, C. F., Early, R., Garcia, R. A., Guisan, A., Maiorano, L., Naimi, B., O'Hara, R. B., Zimmermann, N. E., & Rahbek, C. (2019). Standards for distribution models in biodiversity assessments. *Science Advances*, 5(1), eaat4858. <https://doi.org/10.1126/sciadv.aat4858>
- Aronson, M. F. J., La Sorte, F. A., Nilon, C. H., Katti, M., Goddard, M. A., Lepczyk, C. A., Warren, P. S., Williams, N. S. G., Cilliers, S., Clackson, B., Dobbs, C., Dolan, R., Hedblom, M., Klotz, S., Kooijmans, J. L., Kühn, I., MacGregor-Fors, I., McDonnell, M., Mörtberg, U., ... Winter, M. (2014). A global analysis of the impacts of urbanization on bird and plant diversity reveals key anthropogenic drivers. *Proceedings of the Royal Society B: Biological Sciences*, 281(1780), 20133330. <https://doi.org/10.1098/rspb.2013.3330>
- Barnes, A. D., Allen, K., Kreft, H., Corre, M. D., Jochum, M., Veldkamp, E., Clough, Y., Daniel, R., Darras, K., Denmead, L. H., Farikhah Haneda, N., Hertel, D., Knohl, A., Kotowska, M. M., Kurniawan, S., Meijide, A., Rembold, K., Edho Prabowo, W., Schneider, D., ... Brose, U. (2017). Direct and cascading impacts of tropical land-use change on multi-trophic biodiversity. *Nature Ecology & Evolution*, 1(10), 1511–1519. <https://doi.org/10.1038/s41559-017-0275-7>
- Bar-On, Y. M., Phillips, R., & Milo, R. (2018). The biomass distribution on Earth. *Proceedings of the National Academy of Sciences of the United States of America*, 115(25), 6506–6511. <https://doi.org/10.1073/pnas.1711842115>
- Beckmann, M., Gerstner, K., Akin-Fajiye, M., Ceaușu, S., Kambach, S., Kinlock, N. L., Phillips, H. R. P., Verhagen, W., Gurevitch, J., Klotz, S., Newbold, T., Verburg, P. H., Winter, M., & Seppelt, R. (2019). Conventional land-use intensification reduces species richness and increases production: A global meta-analysis. *Global Change Biology*, 25(6), 1941–1956. <https://doi.org/10.1111/gcb.14606>
- Bender, D. J., Contreras, T. A., & Fahrig, L. (1998). Habitat loss and population decline: A meta-analysis of the patch size effect. *Ecology*, 79(2), 517. <https://doi.org/10.2307/176950>
- Bradshaw, C. J. A., Sodhi, N. S., & Brook, B. W. (2009). Tropical turmoil: A biodiversity tragedy in progress. *Frontiers in Ecology and the Environment*, 7(2), 79–87. <https://doi.org/10.1890/070193>
- Bürgi, M., Östlund, L., & Mladenoff, D. J. (2017). Legacy effects of human land use: Ecosystems as time-lagged systems. *Ecosystems*, 20(1), 94–103. <https://doi.org/10.1007/s10021-016-0051-6>
- Butchart, S. H. M., Walpole, M., Collen, B., Van Strien, A., Scharlemann, J. P. W., Almond, R. E. A., Baillie, J. E. M., Bomhard, B., Brown, C., Bruno, J., Carpenter, K. E., Carr, G. M., Chanson, J., Chenery, A.

- M., Csirke, J., Davidson, N. C., Dentener, F., Foster, M., Galli, A., ... Watson, R. (2010). Global biodiversity: Indicators of recent declines. *Science*, 328(5982), 1164–1168. <https://doi.org/10.1126/science.1187512>
- Cardinale, B. J., Gonzalez, A., Allington, G. R. H., & Loreau, M. (2018). Is local biodiversity declining or not? A summary of the debate over analysis of species richness time trends. *Biological Conservation*, 219, 175–183. <https://doi.org/10.1016/j.biocon.2017.12.021>
- Cardoso, P., Borges, P. A. V., Triantis, K. A., Ferrández, M. A., & Martín, J. L. (2011). Adapting the IUCN Red List criteria for invertebrates. *Biological Conservation*, 144(10), 2432–2440. <https://doi.org/10.1016/j.biocon.2011.06.020>
- Chamberlain, S., & Szoecs, E. (2013). taxize: Taxonomic search and retrieval in R. *F1000Research*, 2, 191. <https://doi.org/10.12688/f1000research.2-191.v2>
- Chamberlain, S., Szoecs, E., Foster, Z., Arendsee, Z., Boettiger, C., Ram, K., Bartomeus, I., Baumgartner, J., O'Donnell, J., Oksanen, J., Tzovaras, B. G., Marchand, P., Tran, V., Salmon, M., Li, G., & Grenié, M. (2020). taxize: Taxonomic information from around the web. R package version 0.9.98. <https://github.com/ropensci/taxize>
- Chapman, P. M., Loveridge, R., Rowcliffe, J. M., Carbone, C., Bernard, H., Davison, C. W., & Ewers, R. M. (2019). Minimal spillover of native small mammals from Bornean tropical forests into adjacent oil palm plantations. *Frontiers in Forests and Global Change*, 2(2). <https://doi.org/10.3389/ffgc.2019.00002>
- Chillo, V., Vázquez, D. P., Amoroso, M. M., & Bennett, E. M. (2018). Land-use intensity indirectly affects ecosystem services mainly through plant functional identity in a temperate forest. *Functional Ecology*, 32(5), 1390–1399. <https://doi.org/10.1111/1365-2435.13064>
- Coffin, A. W. (2007). From roadkill to road ecology: A review of the ecological effects of roads. *Journal of Transport Geography*, 15(5), 396–406. <https://doi.org/10.1016/j.jtrangeo.2006.11.006>
- Collen, B., Loh, J., Whitmee, S., McRae, L., Amin, R., & Baillie, J. E. M. (2009). Monitoring change in vertebrate abundance: The living planet index. *Conservation Biology*, 23(2), 317–327. <https://doi.org/10.1111/j.1523-1739.2008.01117.x>
- Conn, J. E., Wilkerson, R. C., Segura, M. N. O., De Souza, R. T. L., Schlichting, C. D., Wirtz, R. A., & Póvoa, M. M. (2002). Emergence of a new neotropical malaria vector facilitated by human migration and changes in land use. *American Journal of Tropical Medicine and Hygiene*, 66(1), 18–22. <https://doi.org/10.4269/ajtmh.2002.66.18>
- Damgaard, C. (2019). A critique of the space-for-time substitution practice in community ecology. *Trends in Ecology & Evolution*, 34(5), 416–421. <https://doi.org/10.1016/j.tree.2019.01.013>
- Darwall, W. R. T., Holland, R. A., Smith, K. G., Allen, D., Brooks, E. G. E., Katarya, V., Pollock, C. M., Shi, Y., Clausnitzer, V., Cumberlidge, N., Cuttelod, A., Dijkstra, K. D. B., Diop, M. D., García, N., Seddon, M. B., Skelton, P. H., Snoeks, J., Tweddle, D., & Vié, J. C. (2011). Implications of bias in conservation research and investment for freshwater species. *Conservation Letters*, 4(6), 474–482. <https://doi.org/10.1111/j.1755-263X.2011.00202.x>
- Davison, C. W., Chapman, P. M., Bernard, H., Ewers, R. M., Wearn, O. R., Bernard, H., & Ewers, R. M. (2019). Shifts in the demographics and behavior of bearded pigs (*Sus barbatus*) across a land-use gradient. *Biotropica*, 51(6), 1–11. <https://doi.org/10.1111/btp.12724>
- De Palma, A., Abrahamczyk, S., Aizen, M. A., Albrecht, M., Basset, Y., Bates, A., Blake, R. J., Boutin, C., Bugter, R., Connop, S., Cruz-López, L., Cunningham, S. A., Darvill, B., Diekötter, T., Dorn, S., Downing, N., Entling, M. H., Farwig, N., Felicioli, A., ... Purvis, A. (2016). Predicting bee community responses to land-use changes: Effects of geographic and taxonomic biases. *Scientific Reports*, 6(31153). <https://doi.org/10.1038/srep31153>
- De Palma, A., Sanchez Ortiz, K., Martin, P. A., Chadwick, A., Gilbert, G., Bates, A. E., Börger, L., Contu, S., Hill, S. L. L., & Purvis, A. (2018). Challenges with inferring how land-use affects terrestrial biodiversity: Study design, time, space and synthesis. *Advances in Ecological Research*, 58, 163–199. <https://doi.org/10.1016/b.s.aecr.2017.12.004>
- Debinski, D. M., Moranz, R. A., Delaney, J. T., Miller, J. R., Engle, D. M., Winkler, L. B., McGranahan, D. A., Barney, R. J., Trager, J. C., Stephenson, A. L., & Gillespie, M. K. (2011). A cross-taxonomic comparison of insect responses to grassland management and land-use legacies. *Ecosphere*, 2(12), 1–16. <https://doi.org/10.1890/ES11-00226.1>
- Deikumah, J. P., Mcalpine, C. A., & Maron, M. (2014). Biogeographical and taxonomic biases in tropical forest fragmentation research. *Conservation Biology*, 28(6), 1522–1531. <https://doi.org/10.1111/cobi.12348>
- Di Marco, M., Chapman, S., Althor, G., Kearney, S., Besancon, C., Butt, N., Maina, J. M., Possingham, H. P., von Biebertstein, K. R., Venter, O., & Watson, J. E. M. (2017). Changing trends and persisting biases in three decades of conservation science. *Global Ecology and Conservation*, 10, 32–42. <https://doi.org/10.1016/j.gecco.2017.01.008>
- Díaz, S., Settele, J., Brondizio, E. S., Ngo, H. T., Agard, J., Arneeth, A., Balvanera, P., Brauman, K. A., Butchart, S. H. M., Chan, K. M. A., Garibaldi, L. A., Ichii, K., Liu, J., Subramanian, S. M., Midgley, G. F., Miloslavich, P., Molnár, Z., Obura, D., Pfaff, A., ... Zayas, C. N. (2019). Pervasive human-driven decline of life on Earth points to the need for transformative change. *Science*, 366(6471), eaax3100. <https://doi.org/10.1126/science.aax3100>
- Dislich, C., Keyel, A. C., Salecker, J., Kisel, Y., Meyer, K. M., Auliya, M., Barnes, A. D., Corre, M. D., Darras, K., Faust, H., Hess, B., Klasen, S., Knohl, A., Kreft, H., Meijide, A., Nurdiansyah, F., Otten, F., Pe'er, G., Steinebach, S., ... Wiegand, K. (2017). A review of the ecosystem functions in oil palm plantations, using forests as a reference system. *Biological Reviews*, 92(3), 1539–1569. <https://doi.org/10.1111/brv.12295>
- Dornelas, M., Gotelli, N. J., McGill, B., Shimadzu, H., Moyes, F., Sievers, C., & Magurran, A. E. (2014). Assemblage time series reveal biodiversity change but not systematic loss. *Science*, 344(6181), 296–299. <https://doi.org/10.1126/science.1248484>
- Dullinger, I., Essl, F., Moser, D., Erb, K., Haberl, H., & Dullinger, S. (2021). Biodiversity models need to represent land-use intensity more comprehensively. *Global Ecology and Biogeography*, 30(5), 924–932. <https://doi.org/10.1111/geb.13289>
- Dupouey, J. L., Dambrine, E., Laffite, J. D., & Moares, C. (2002). Irreversible impact of past land use on forest soils and biodiversity. *Ecology*, 83(11), 2978. <https://doi.org/10.2307/3071833>
- Ellis, E. C., Goldewijk, K. K., Siebert, S., Lightman, D., & Ramankutty, N. (2010). Anthropogenic transformation of the biomes, 1700 to 2000. *Global Ecology and Biogeography*, 19(5), 589–606. <https://doi.org/10.1111/j.1466-8238.2010.00540.x>
- Ellis, E. C., Kaplan, J. O., Fuller, D. Q., Vavrus, S., Goldewijk, K. K., & Verburg, P. H. (2013). Used planet: A global history. *Proceedings of the National Academy of Sciences of the United States of America*, 110(20), 7978–7985. <https://doi.org/10.1073/pnas.1217241110>
- Enquist, B., Condit, R., Peet, R., Schildhauer, M., & Thiers, B. (2016). Cyberinfrastructure for an integrated botanical information network to investigate the ecological impacts of global climate change on plant biodiversity. *PeerJ*, 4, e2615v2. <https://doi.org/10.7287/peerj.preprints.2615>
- Enquist, B. J., Feng, X., Boyle, B., Maitner, B., Newman, E. A., Jørgensen, P. M., Roehrdanz, P. R., Thiers, B. M., Burger, J. R., Corlett, R. T., Couvreur, T. L. P., Dauby, G., Donoghue, J. C., Foden, W., Lovett, J. C., Marquet, P. A., Merow, C., Midgley, G., Morueta-Holme, N., ... McGill, B. J. (2019). The commonness of rarity: Global and future distribution of rarity across land plants. *Science Advances*, 5(11), 1–13. <https://doi.org/10.1126/sciadv.aaz0414>
- Erb, K. H., Kastner, T., Plutzer, C., Bais, A. L. S., Carvalhais, N., Fetzel, T., Gingrich, S., Haberl, H., Lauk, C., Niedertscheider, M., Pongratz, J., Thurner, M., & Luyssaert, S. (2018). Unexpectedly large impact

- of forest management and grazing on global vegetation biomass. *Nature*, 553(7686), 73–76. <https://doi.org/10.1038/nature25138>
- Fahrig, L. (1997). Relative effects of habitat loss and fragmentation on population extinction. *The Journal of Wildlife Management*, 61(3), 603. <https://doi.org/10.2307/3802168>
- Faurby, S., & Svenning, J. C. (2015). Historic and prehistoric human-driven extinctions have reshaped global mammal diversity patterns. *Diversity and Distributions*, 21(10), 1155–1166. <https://doi.org/10.1111/ddi.12369>
- Felton, A. A. M. A., Fischer, J., Lindenmayer, D. B., Montague-Drake, R., Lowe, A. R., Saunders, D., Felton, A. A. M. A., Steffen, W., Munro, N. T., Youngentob, K., Gillen, J., Gibbons, P., Bruzgul, J. E., Fazez, I., Bond, S. J., Elliott, C. P., Macdonald, B. C. T., Porfiri, L. L., Westgate, M., & Worthy, M. (2009). Climate change, conservation and management: An assessment of the peer-reviewed scientific journal literature. *Biodiversity and Conservation*, 18(8), 2243–2253. <https://doi.org/10.1007/s10531-009-9652-0>
- Fick, S. E., & Hijmans, R. J. (2017). WorldClim 2: New 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology*, 37(12), 4302–4315. <https://doi.org/10.1002/joc.5086>
- Foley, J. A., DeFries, R., Asner, G. P., Barford, C., Bonan, G., Carpenter, S. R., Stuart Chapin, F., Coe, M. T., Daily, G. C., Gibbs, H. K., Helkowski, J. H., Holloway, T., Howard, E. A., Kucharik, C. J., Monfreda, C., Patz, J. A., Colin Prentice, I., Ramankutty, N., & Snyder, P. K. (2005). Global consequences of land use. *Science*, 309(570), 570–574. <https://doi.org/10.1126/science.1111772>
- Forsy, E. A., Allen, C. R., & Wojcik, D. P. (2002). Influence of the proximity and amount of human development and roads on the occurrence of the red imported fire ant in the lower Florida Keys. *Biological Conservation*, 108(1), 27–33. [https://doi.org/10.1016/S0006-3207\(02\)00086-1](https://doi.org/10.1016/S0006-3207(02)00086-1)
- França, F., Louzada, J., Korasaki, V., Griffiths, H., Silveira, J. M., & Barlow, J. (2016). Do space-for-time assessments underestimate the impacts of logging on tropical biodiversity? An Amazonian case study using dung beetles. *Journal of Applied Ecology*, 53(4), 1098–1105. <https://doi.org/10.1111/1365-2664.12657>
- Gaynor, K. M., Hohnowski, C. E., Carter, N. H., & Brashares, J. S. (2018). The influence of human disturbance on wildlife nocturnality. *Science*, 360(6394), 1232–1235. <https://doi.org/10.1126/science.aar7121>
- GBIF: The Global Biodiversity Information Facility. (2021). *What is GBIF?* <https://www.gbif.org/what-is-gbif>
- Gerstner, K., Dormann, C. F., Stein, A., Manceur, A. M., & Seppelt, R. (2014). Effects of land use on plant diversity—A global meta-analysis. *Journal of Applied Ecology*, 51(6), 1690–1700. <https://doi.org/10.1111/1365-2664.12329>
- Gibson, L., Lee, T. M., Koh, L. P., Brook, B. W., Gardner, T. A., Barlow, J., Peres, C. A., Bradshaw, C. J. A., Laurance, W. F., Lovejoy, T. E., & Sodhi, N. S. (2011). Primary forests are irreplaceable for sustaining tropical biodiversity. *Nature*, 478(7369), 378–381. <https://doi.org/10.1038/nature10425>
- Gonzalez, A., Cardinale, B. J., Allington, G. R. H., Byrnes, J., Endsley, K. A., Brown, D. G., Hooper, D. U., Isbell, F., O'Connor, M. I., & Loreau, M. (2016). Estimating local biodiversity change: A critique of papers claiming no net loss of local diversity. *Ecology*, 97(8), 1949–1960. <https://doi.org/10.1890/15-1759.1>
- González-del-Piiego, P., Freckleton, R. P., Edwards, D. P., Koo, M. S., Scheffers, B. R., Pyron, R. A., & Jetz, W. (2019). Phylogenetic and trait-based prediction of extinction risk for data-deficient amphibians. *Current Biology*, 29(9), 1557–1563.e3. <https://doi.org/10.1016/j.cub.2019.04.005>
- Grover, C., Tobin, R., Byrne, K., Woollard, M., Reid, J., Dunn, S., & Ball, J. (2010). Use of the Edinburgh geoparser for georeferencing digitized historical collections. *Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences*, 368(1925), 3875–3889. <https://doi.org/10.1098/rsta.2010.0149>
- Guo, F., Lenoir, J., & Bonebrake, T. C. (2018). Land-use change interacts with climate to determine elevational species redistribution. *Nature Communications*, 9(1315). <https://doi.org/10.1038/s41467-018-03786-9>
- Haberl, H., Schulz, N. B., Plutzar, C., Erb, K. H., Krausmann, F., Loibl, W., Moser, D., Sauberer, N., Weisz, H., Zechmeister, H. G., & Zülka, P. (2004). Human appropriation of net primary production and species diversity in agricultural landscapes. *Agriculture, Ecosystems and Environment*, 102(2), 213–218. <https://doi.org/10.1016/j.agee.2003.07.004>
- Hanski, I., & Ovaskainen, O. (2002). Extinction debt at extinction threshold. *Conservation Biology*, 16(3), 666–673. <https://doi.org/10.1046/j.1523-1739.2002.00342.x>
- Hill, S., Gonzalez, R., Sanchez-Ortiz, K., Caton, E., Espinoza, F., Newbold, T., Tylianakis, J., Scharlemann, J., De Palma, A., & Purvis, A. (2018). Worldwide impacts of past and projected future land-use change on local species richness and the Biodiversity Intactness Index. *BioRxiv*, 311787. <https://doi.org/10.1101/311787>
- Hudson, L. N., Newbold, T., Contu, S., Hill, S. L. L., Lysenko, I., De Palma, A., Phillips, H. R. P., Senior, R. A., Bennett, D. J., Booth, H., Choimes, A., Correia, D. L. P., Day, J., Echeverría-Londoño, S., Garon, M., Harrison, M. L. K., Ingram, D. J., Jung, M., Kemp, V., ... Purvis, A. (2014). The PREDICTS database: A global database of how local terrestrial biodiversity responds to human impacts. *Ecology and Evolution*, 4(24), 4701–4735. <https://doi.org/10.1002/ece3.1303>
- IPBES. (2018). *Summary for policymakers of the assessment report on land degradation and restoration of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services*. In R. Scholes, L. Montanarella, A. Brainich, N. Barger, B. ten Brink, M. Cantele, B. Erasmus, J. Fisher, T. Gardner, D. J. G. Holland, F. Kohler, J. S. Kotiaho, G. Von Maltitz, G. Nangendo, R. Pandit, J. Parrotta, M. D. Potts, S. Prince, M. Sankaran, & L. Willemsen (Eds.). IPBES Secretariat.
- IPBES. (2019). *Summary for policymakers of the global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services*. In S. Díaz, J. Settele, E. Brondizio, H. T. Ngo, M. Guèze, J. Agard, A. Arneth, P. Balvanera, K. Brauman, S. Butchart, K. Chan, L. Garibaldi, K. Ichii, J. Liu, S. M. Subramanian, G. Midgley, P. Miloslavich, Z. Molnár, D. Obura, & C. Zayas (Eds.). IPBES Secretariat.
- Ivanova, N. V., & Shashkov, M. P. (2017). Biodiversity databases in Russia: Towards a national portal. *Arctic Science*, 3(3), 560–576. <https://doi.org/10.1139/as-2016-0050>
- Jackson, S. T., & Sax, D. F. (2010). Balancing biodiversity in a changing environment: Extinction debt, immigration credit and species turnover. *Trends in Ecology & Evolution*, 25(3), 153–160. <https://doi.org/10.1016/j.tree.2009.10.001>
- Jetz, W., & Rahbek, C. (2002). Geographic range size and determinants of avian species richness. *Science*, 297(5586), 1548–1551. <https://doi.org/10.1126/science.1072779>
- Kaplan, J. O., Krumhardt, K. M., Ellis, E. C., Ruddiman, W. F., Lemmen, C., & Goldewijk, K. K. (2011). Holocene carbon emissions as a result of anthropogenic land cover change. *Holocene*, 21(5), 775–791. <https://doi.org/10.1177/0959683610386983>
- Karam-Gemael, M., Decker, P., Stoev, P., Marques, M. I., & Chagas, A. (2020). Conservation of terrestrial invertebrates: A review of IUCN and regional red lists for myriapoda. *ZooKeys*, 930, 221–229. <https://doi.org/10.3897/zookeys.930.48943>
- Kehoe, L., Kuemmerle, T., Meyer, C., Levers, C., Václavík, T., & Kreft, H. (2015). Global patterns of agricultural land-use intensity and vertebrate diversity. *Diversity and Distributions*, 21(11), 1308–1318. <https://doi.org/10.1111/ddi.12359>
- Kier, G., Kreft, H., Tien, M. L., Jetz, W., Ibsch, P. L., Nowicki, C., Mutke, J., & Barthlott, W. (2009). A global assessment of endemism and species richness across island and mainland regions. *Proceedings of the National Academy of Sciences of the United States of America*, 106(23), 9322–9327. <https://doi.org/10.1073/pnas.0810306106>

- Kuemmerle, T., Erb, K., Meyfroidt, P., Müller, D., Verburg, P. H., Estel, S., Haberl, H., Hostert, P., Jepsen, M. R., Kastner, T., Levers, C., Lindner, M., Plutzer, C., Verkerk, P. J., van der Zanden, E. H., & Reenberg, A. (2013). Challenges and opportunities in mapping land use intensity globally. *Current Opinion in Environmental Sustainability*, 5(5), 484–493. <https://doi.org/10.1016/j.cosust.2013.06.002>
- Lamb, C. T., Mowat, G., McLellan, B. N., Nielsen, S. E., & Boutin, S. (2017). Forbidden fruit: Human settlement and abundant fruit create an ecological trap for an apex omnivore. *Journal of Animal Ecology*, 86(1), 55–65. <https://doi.org/10.1111/1365-2656.12589>
- Lauber, C. L., Strickland, M. S., Bradford, M. A., & Fierer, N. (2008). The influence of soil properties on the structure of bacterial and fungal communities across land-use types. *Soil Biology and Biochemistry*, 40(9), 2407–2415. <https://doi.org/10.1016/j.soilbio.2008.05.021>
- Lawton, J. H., Bignell, D. E., Bolton, B., Bloemers, G. F., Eggleton, P., Hammond, P. M., Hodda, M., Holt, R. D., Larsen, T. B., Mawdsley, N. A., Stork, N. E., Srivastava, D. S., & Watt, A. D. (1998). Biodiversity inventories, indicator taxa and effects of habitat modification in tropical forest. *Nature*, 391(6662), 72–76. <https://doi.org/10.1038/34166>
- Love, K., Kurz, D. J., Vaughan, I. P., Ke, A., Evans, L. J., & Goossens, B. (2018). Bearded pig (*Sus barbatus*) utilisation of a fragmented forest-oil palm landscape in Sabah, Malaysian Borneo. *Wildlife Research*, 44(8), 603–612. <https://doi.org/10.1071/WR16189>
- Luskin, M. S., Brashares, J. S., Ickes, K., Sun, I. F., Fletcher, C., Wright, S. J., & Potts, M. D. (2017). Cross-boundary subsidy cascades from oil palm degrade distant tropical forests. *Nature Communications*, 8(1), 2231. <https://doi.org/10.1038/s41467-017-01920-7>
- Mantyka-Pringle, C. S., Martin, T. G., & Rhodes, J. R. (2012). Interactions between climate and habitat loss effects on biodiversity: A systematic review and meta-analysis. *Global Change Biology*, 18(4), 1239–1252. <https://doi.org/10.1111/j.1365-2486.2011.02593.x>
- Martin, L. J., Blossey, B., & Ellis, E. (2012). Mapping where ecologists work: Biases in the global distribution of terrestrial ecological observations. *Frontiers in Ecology and the Environment*, 10(4), 195–201. <https://doi.org/10.1890/110154>
- Mendenhall, C. D., Daily, G. C., & Ehrlich, P. R. (2012). Improving estimates of biodiversity loss. *Biological Conservation*, 151(1), 32–34. <https://doi.org/10.1016/j.biocon.2012.01.069>
- Miehe, G., Miehe, S., Böhner, J., Kaiser, K., Hensen, I., Madsen, D., Liu, J. Q., & Opgenoorth, L. (2014). How old is the human footprint in the world's largest alpine ecosystem? A review of multiproxy records from the Tibetan Plateau from the ecologists' viewpoint. *Quaternary Science Reviews*, 86, 190–209. <https://doi.org/10.1016/j.quascirev.2013.12.004>
- Millard, J. W., Freeman, R., & Newbold, T. (2019). Text-analysis reveals taxonomic and geographic disparities in animal pollination literature. *Ecography*, 43(1), 44–59. <https://doi.org/10.1111/ecog.04532>
- Murphy, G. E. P., & Romanuk, T. N. (2014). A meta-analysis of declines in local species richness from human disturbances. *Ecology and Evolution*, 4(1), 91–103. <https://doi.org/10.1002/ece3.909>
- Newbold, T., Bentley, L. F., Hill, S. L. L., Edgar, M. J., Horton, M., Su, G., Şekercioğlu, Ç. H., Collen, B., & Purvis, A. (2020). Global effects of land use on biodiversity differ among functional groups. *Functional Ecology*, 34(3), 684–693. <https://doi.org/10.1111/1365-2435.13500>
- Newbold, T., Hudson, L. N., Hill, S. L. L., Contu, S., Gray, C. L., Scharlemann, J. P. W., Börger, L., Phillips, H. R. P., Sheil, D., Lysenko, I., & Purvis, A. (2016). Global patterns of terrestrial assemblage turnover within and among land uses. *Ecography*, 39(12), 1151–1163. <https://doi.org/10.1111/ecog.01932>
- Newbold, T., Hudson, L. N., Hill, S. L. L., Contu, S., Lysenko, I., Senior, R. A., Börger, L., Bennett, D. J., Choimes, A., Collen, B., Day, J., De Palma, A., Díaz, S., Echeverria-Londoño, S., Edgar, M. J., Feldman, A., Garon, M., Harrison, M. L. K., Alhussaini, T., ... Purvis, A. (2015). Global effects of land use on local terrestrial biodiversity. *Nature*, 520(7545), 45–50. <https://doi.org/10.1038/nature14324>
- Oliver, T. H., & Morecroft, M. D. (2014). Interactions between climate change and land use change on biodiversity: Attribution problems, risks, and opportunities. *Wiley Interdisciplinary Reviews: Climate Change*, 5(3), 317–335. <https://doi.org/10.1002/wcc.271>
- Olson, D. M., Dinerstein, E., Wikramanayake, E. D., Burgess, N. D., Powell, G. V. N., Underwood, E. C., Jennifer, A., D'amico, I. I., Strand, H. E., Morrison, J. C., Loucks, C. J., Allnutt, T. F., Ricketts, T. H., Kura, Y., Lamoreux, J. F., Wettengel, W. W., Hedao, P., & Kassem, K. R. (2001). Terrestrial ecoregions of the world: A new map of life on earth. *BioScience*, 51(11), 933–938. [https://doi.org/10.1641/0006-3568\(2001\)051](https://doi.org/10.1641/0006-3568(2001)051)
- Peres, C. A. (2000). Effects of subsistence hunting on vertebrate community structure in Amazonian forests. *Conservation Biology*, 14(1), 240–253. <https://doi.org/10.1046/j.1523-1739.2000.98485.x>
- Pielke, R. A., Marland, G., Betts, R. A., Chase, T. N., Eastman, J. L., Niles, J. O., Niyogi, D. D. S., & Running, S. W. (2002). The influence of land-use change and landscape dynamics on the climate system: Relevance to climate-change policy beyond the radiative effect of greenhouse gases. *Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences*, 360(1797), 1705–1719. <https://doi.org/10.1098/rsta.2002.1027>
- Polaina, E., González-Suárez, M., & Revilla, E. (2019). The legacy of past human land use in current patterns of mammal distribution. *Ecography*, 42(10), 1623–1635. <https://doi.org/10.1111/ecog.04406>
- Prendergast, J. R., Quinn, R. M., Lawton, J. H., Eversham, B. C., & Gibbons, D. W. (1993). Rare species, the coincidence of diversity hotspots and conservation strategies. *Nature*, 365(6444), 335–337. <https://doi.org/10.1038/365335a0>
- Pyke, C. R. (2004). Habitat loss confounds climate change impacts. *Frontiers in Ecology and the Environment*, 2(4), 178. <https://doi.org/10.2307/3868312>
- Pyšek, P., Richardson, D. M., Pergl, J., Jarošík, V., Sixtová, Z., & Weber, E. (2008). Geographical and taxonomic biases in invasion ecology. *Trends in Ecology & Evolution*, 23(5), 237–244. <https://doi.org/10.1016/j.tree.2008.02.002>
- R Core Team. (2020). *R: A language and environment for statistical computing*. R Foundation for Statistical Computing R Foundation for Statistical Computing. <https://doi.org/10.1007/978-3-540-74686-7>
- Raiter, K. G., Possingham, H. P., Prober, S. M., & Hobbs, R. J. (2014). Under the radar: Mitigating enigmatic ecological impacts. *Trends in Ecology & Evolution*, 29(11), 635–644. <https://doi.org/10.1016/j.tree.2014.09.003>
- Roskov, Y., Ower, G., Orrell, T., Nicolson, D., Bailly, N., Kirk, P. M., Bourgoin, T., DeWalt, R. E., Decock, W., van Nieukerken, E., Zarucchi, J., & Penev, L. (2019). *Species 2000 & ITIS catalogue of life, 2019 annual checklist*. Species 2000: Naturalis. www.catalogueoflife.org/annual-checklist/2019
- Roy Rosenzweig Center for History and New Media. (2020). *Zotero* (5.0.8.8). www.zotero.org
- Sala, O. E., Chapin, F. S., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-Sanwald, E., Huenneke, L. F., Jackson, R. B., Kinzig, A., Leemans, R., Lodge, D. M., Mooney, H. A., Oesterheld, M., Poff, N. L. R., Sykes, M. T., Walker, B. H., Walker, M., & Wall, D. H. (2000). Global biodiversity scenarios for the year 2100. *Science*, 287(5459), 1770–1774. <https://doi.org/10.1126/science.287.5459.1770>
- Santos, M. J., Smith, A. B., Dekker, S. C., Eppinga, M. B., Leitão, P. J., Moreno-Mateos, D., Morueta-Holme, N., & Ruggeri, M. (2021). The role of land use and land cover change in climate change vulnerability assessments of biodiversity: A systematic review. *Landscape Ecology*. <https://doi.org/10.1007/s10980-021-01276-w>
- Schmidt, W. (1989). Plant dispersal by motor cars. *Vegetatio*, 80(2), 147–152. <https://doi.org/10.1007/BF00048038>
- Shoemaker, L. G., Sullivan, L. L., Donohue, I., Cabral, J. S., Williams, R. J., Mayfield, M. M., Chase, J. M., Chu, C., Harpole, W. S., Huth,

- A., HilleRisLambers, J., James, A. R. M., Kraft, N. J. B., May, F., Muthukrishnan, R., Satterlee, S., Taubert, F., Wang, X., Wiegand, T., ... Abbott, K. C. (2020). Integrating the underlying structure of stochasticity into community ecology. *Ecology*, 101(2), e02922. <https://doi.org/10.1002/ecy.2922>
- Simberloff, D. (1998). Flagships, umbrellas, and keystones: Is single-species management passe in the landscape era? *Biological Conservation*, 83(3), 247–257. [https://doi.org/10.1016/S0006-3207\(97\)00081-5](https://doi.org/10.1016/S0006-3207(97)00081-5)
- Simonson, W. D., Allen, H. D., & Coomes, D. A. (2014). Applications of airborne lidar for the assessment of animal species diversity. *Methods in Ecology and Evolution*, 5(8), 719–729. <https://doi.org/10.1111/2041-210X.12219>
- Sirami, C., Caplat, P., Popy, S., Clamens, A., Arlettaz, R., Jiguet, F., Brotons, L., & Martin, J. L. (2017). Impacts of global change on species distributions: Obstacles and solutions to integrate climate and land use. *Global Ecology and Biogeography*, 26(4), 385–394. <https://doi.org/10.1111/geb.12555>
- Srinivasan, U., Hines, J. E., & Quader, S. (2015). Demographic superiority with increased logging in tropical understorey insectivorous birds. *Journal of Applied Ecology*, 52(5), 1374–1380. <https://doi.org/10.1111/1365-2664.12475>
- Stevens, A. F. J., Bayne, E. M., & Wellicome, T. I. (2011). Soil and climate are better than biotic land cover for predicting home-range habitat selection by endangered burrowing owls across the Canadian Prairies. *Biological Conservation*, 144(5), 1526–1536. <https://doi.org/10.1016/j.biocon.2010.10.032>
- Taheri, S., Naimi, B., Rahbek, C., & Araújo, M. B. (2021). Improvements in reports of species redistribution under climate change are required. *Science Advances*, 7(15), eabe1110. <https://doi.org/10.1126/sciadv.abe1110>
- Tilman, D., May, R. M., Lehman, C. L., & Nowak, M. A. (1994). Habitat destruction and the extinction debt. *Nature*, 371(6492), 65–66. <https://doi.org/10.1038/371065a0>
- Titeux, N., Henle, K., Mihoub, J. B., & Brotons, L. (2016). Climate change distracts us from other threats to biodiversity. *Frontiers in Ecology and the Environment*, 14(6), 291. <https://doi.org/10.1002/fee.1303>
- Titeux, N., Henle, K., Mihoub, J.-B., Regos, A., Geijzenendorffer, I. R., Cramer, W., Verbarg, P. H., & Brotons, L. (2017). Global scenarios for biodiversity need to better integrate climate and land use change. *Diversity and Distributions*, 23(11), 1231–1234. <https://doi.org/10.1111/ddi.12624>
- Trimble, M. J., & van Aarde, R. J. (2012). Geographical and taxonomic biases in research on biodiversity in human-modified landscapes. *Ecosphere*, 3(12), 1–16. <https://doi.org/10.1890/es12-00299.1>
- Troutet, J., Grandcolas, P., Blin, A., Vignes-Lebbe, R., & Legendre, F. (2017). Taxonomic bias in biodiversity data and societal preferences. *Scientific Reports*, 7(1), 1–14. <https://doi.org/10.1038/s41598-017-09084-6>
- Vallecillo, S., Brotons, L., & Thuiller, W. (2009). Dangers of predicting bird species distributions in response to land-cover changes. *Ecological Applications*, 19(2), 538–549. <https://doi.org/10.1890/08-0348.1>
- Velasco, D., García-Llorente, M., Alonso, B., Dolera, A., Palomo, I., Iñiesta-Arandia, I., & Martín-López, B. (2015). Biodiversity conservation research challenges in the 21st century: A review of publishing trends in 2000 and 2011. *Environmental Science and Policy*, 54, 90–96. <https://doi.org/10.1016/j.envsci.2015.06.008>
- Vellend, M., Baeten, L., Myers-Smith, I. H., Elmendorf, S. C., Beausejour, R., Brown, C. D., De Frenne, P., Verheyen, K., & Wipf, S. (2013). Global meta-analysis reveals no net change in local-scale plant biodiversity over time. *Proceedings of the National Academy of Sciences of the United States of America*, 110(48), 19456–19459. <https://doi.org/10.1073/pnas.1312779110>
- Venter, O., Sanderson, E. W., Magrath, A., Allan, J. R., Beher, J., Jones, K. R., Possingham, H. P., Laurance, W. F., Wood, P., Fekete, B. M., Levy, M. A., & Watson, J. E. M. (2016). Sixteen years of change in the global terrestrial human footprint and implications for biodiversity conservation. *Nature Communications*, 7(12558). <https://doi.org/10.1038/ncomms12558>
- Verbarg, P. H., Neumann, K., & Nol, L. (2011). Challenges in using land use and land cover data for global change studies. *Global Change Biology*, 17(2), 974–989. <https://doi.org/10.1111/j.1365-2486.2010.02307.x>
- Wearn, O. R., Reuman, D. C., & Ewers, R. M. (2012). Extinction debt and windows of conservation opportunity in the Brazilian Amazon. *Science*, 337(6091), 228–232. <https://doi.org/10.1126/science.1219013>
- Williams, P. H., Burgess, N. D., & Rahbek, C. (2000). Flagship species, ecological complementarity and conserving the diversity of mammals and birds in sub-Saharan Africa. *Animal Conservation*, 3(3), 249–260. <https://doi.org/10.1017/S1367943000000974>

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

How to cite this article: Davison, C. W., Rahbek, C., & Morueta-Holme, N. (2021). Land-use change and biodiversity: Challenges for assembling evidence on the greatest threat to nature. *Global Change Biology*, 00, 1–16. <https://doi.org/10.1111/gcb.15846>