LRH: Granville & Banks-Leite

RRH: Limited dispersal in mangroves

Mangrove propagules are limited in their capacity to disperse across long distances

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Abstract

Mangroves are subject to rapid and large-scale habitat changes which threaten their unique genetic diversity and provision of critically important ecosystem services. Habitat fragmentation reduces connectivity which can impair dispersal and lead to genetic isolation. However, it is unclear whether fragmentation could impact mangrove genetic isolation, as mangrove propagules can disperse long distances. Here, we conducted a meta-analysis of studies reporting a correlation between geographic distance and genetic distance in mangrove plants. From the 22 studies that met the inclusion criteria, we found a significant isolation-by-distance effect; geographic distance was significantly associated with Nei's genetic distance and F_{ST}. Our results show how mangrove propagules may be limited in their capacity to disperse across long distances, which highlights the importance of maintaining close proximity between habitat patches and reducing habitat fragmentation.

1 Introduction

2 Mangroves are of exceptionally high ecological and societal value as they provide crucial ecosystem services

3 and unique genetic diversity which must be maintained. However, mangroves are threatened by anthropogenic

4 disturbances and sea level rise which are leading to substantial losses of habitats. Habitat loss is well known to

5 impact dispersal across habitats, which is vital for maintaining genetic diversity and for range shifts in response

6 to environmental changes (Van der Stocken et al. 2019a). Therefore, understanding the relationship between

7 spatial and genetic structuring of populations is fundamental in developing effective conservation strategies that

8 maintain dispersal across habitats (Durrant et al. 2014, Taylor et al. 2021, Wright et al. 2015).

9 Buoyant mangrove propagules have the capacity for long-distance dispersal by water because they remain viable

10 for extended periods of time, therefore they can drift in ocean currents (Binks et al. 2019). This could be

11 expected to lead to high connectivity between habitat patches, which could mean that increasing distance

12 between patches would have little effect on genetic isolation. However, field studies have shown that mangrove

13 propagules tend not to disperse far from their release point, leading to patterns of isolation-by-distance (Binks et

14 al. 2019, Clarke 1993, Yan et al. 2016). Therefore, we quantified the relationship between geographic distance

15 and genetic distance in mangrove plant communities to better understand how far mangrove propagules disperse.

16 While it is generally accepted that habitat loss has negative effects on biodiversity, the effects of habitat

17 fragmentation *per se* (independent of habitat loss) are more variable and context-dependent (Fahrig 2003). On

18 the one hand, Wilcove et al's 1986 definition of habitat fragmentation as a process whereby *"a large expanse of*

19 *habitat is transformed into a number of smaller patches of smaller total area, isolated from each other by a*

20 *matrix of habitats unlike the original"* (Wilcove 1986) implies that fragmentation leads to increased isolation of

21 patches (Fahrig, 2003). On the other hand, fragmentation *per se* could result in many small patches that are less

22 isolated from each other, assuming that dispersal across the matrix is possible (Fahrig 2003). In mangroves,

23 however, the matrix between habitat patches is composed of water, which can be easily transposed by

24 propagules and human-modified land use, which cannot be colonised by mangroves.

1   Mangrove forests are intertidal wetlands, found along coastlines in tropical, sub-tropical and warm-temperate

2   climates (Bryan-Brown et al. 2020). Globally, there are two main hotspots of mangrove biodiversity; the Indo-

3   West Pacific contains approximately 54 mangrove species, and the Atlantic-East Pacific contains approximately

4   17 mangrove species (Van der Stocken et al. 2019a). Local- and regional- scale threats to mangroves include

5   aquaculture, agriculture, urban development and pollution. Broader scale threats include long-term processes

6   such as sea-level rise and climate change. Therefore, anthropogenic activity impacts mangroves both directly and

7   indirectly (Friess et al. 2019). Between 2000 and 2012, mangrove forests were lost at an average rate of 0.18 %

8   per year (Richards & Friess 2016). Mangrove forests are of exceptionally high ecological value because the

9   plants that inhabit them are among the few woody plant species that tolerate the salinity of open sea and are

10   adapted to living in intertidal zones (Mantiquilla et al. 2021). Therefore, it is important to maintain the unique

11   genetic diversity found in mangroves. Furthermore, mangroves have an estimated economic value of \$194,000

12   per hectare per year (Costanza et al. 2014) as they provide crucial ecosystem services, including carbon storage,

13   fisheries, and erosion protection. The provision of these services depends on the size and arrangement of forest

14   patterns (Bryan-Brown et al. 2020).

15   Even in areas with low rates of mangrove loss, there is a global trend towards ubiquitous fragmentation, which

16   can potentially pose a threat to mangrove biodiversity (Bryan-Brown et al. 2020). Therefore, there is a need to

17   quantify the impact of fragmentation on mangrove biodiversity. Measures of genetic distance such as Nei's

18   genetic distance (Nei 1972) and F_{ST} (Wright 1950) provide useful insights into the genetic structure and

19   connectedness of a habitat (Bohonak 1999). An association between geographic distance and genetic distance

20   can indicate isolation-by-distance (Wright 1943). If mangrove patches are more isolated from each other and

21   propagules cannot disperse between them, there will likely be a larger pairwise genetic distance between these

22   patches (Jaquiéry et al. 2011). Therefore, this meta-analysis explores the effects of geographic distance on

23   genetic distance in mangrove plants, with the aim of providing insight into the effect of patch isolation on

24   genetic isolation in mangroves.

Methods

Systematic literature search and inclusion criteria

In May 2021, an extensive search of the relevant literature was conducted, following the PRISMA (Preferred reporting items for systematic review and meta-analysis) statement which provides a standardised framework for reporting systematic reviews and meta-analyses (Moher et al. 2009). The following search terms were used in Web of Science and Scopus:

1. mangrove AND fragment* AND ('genetic diversity' OR 'genetic differentiation')
2. mangrove AND connect* AND ('genetic diversity' OR 'genetic differentiation')
3. mangrove AND isolat* AND ('genetic diversity' OR 'genetic differentiation')

An initial search, after removing duplicates, yielded 199 papers, which were screened for eligibility. This resulted in the exclusion of 148 non-relevant papers and 4 papers that could not be accessed. From the 47 papers that remained, we selected papers that reported the results of a Mantel test for the correlation between matrices of untransformed Euclidean geographic distance on Nei's genetic distance (11 papers, representing 13 case studies) or untransformed Wright's F_{ST} (8 papers, representing 9 case studies). Nei's genetic distance and F_{ST} were chosen as measures of genetic diversity because they were the most commonly reported so this helped maximise the sample size. The analysis of the effect of geographic distance on Nei's genetic diversity is henceforth referred to as 'Nei's meta-analysis' and the analysis of the effect of geographic distance on Wright's F_{ST} is henceforth referred to as 'F_{ST} meta-analysis'

Statistical analysis

For the effect size, we extracted Pearson's r value from the reported Mantel test in each paper. Where R^2 was reported, it was converted to r by taking the square root (this was done in 2 papers for the Nei's meta-analysis and 8 papers for the F_{ST} meta-analysis). Each study was weighted according to the following formula developed by Reed & Frankham (2003) specifically for meta-analyses of genetic diversity: $\sqrt{(A-2) \times N}$, where $A =$ number of populations and $N =$ number of individuals for each paper.
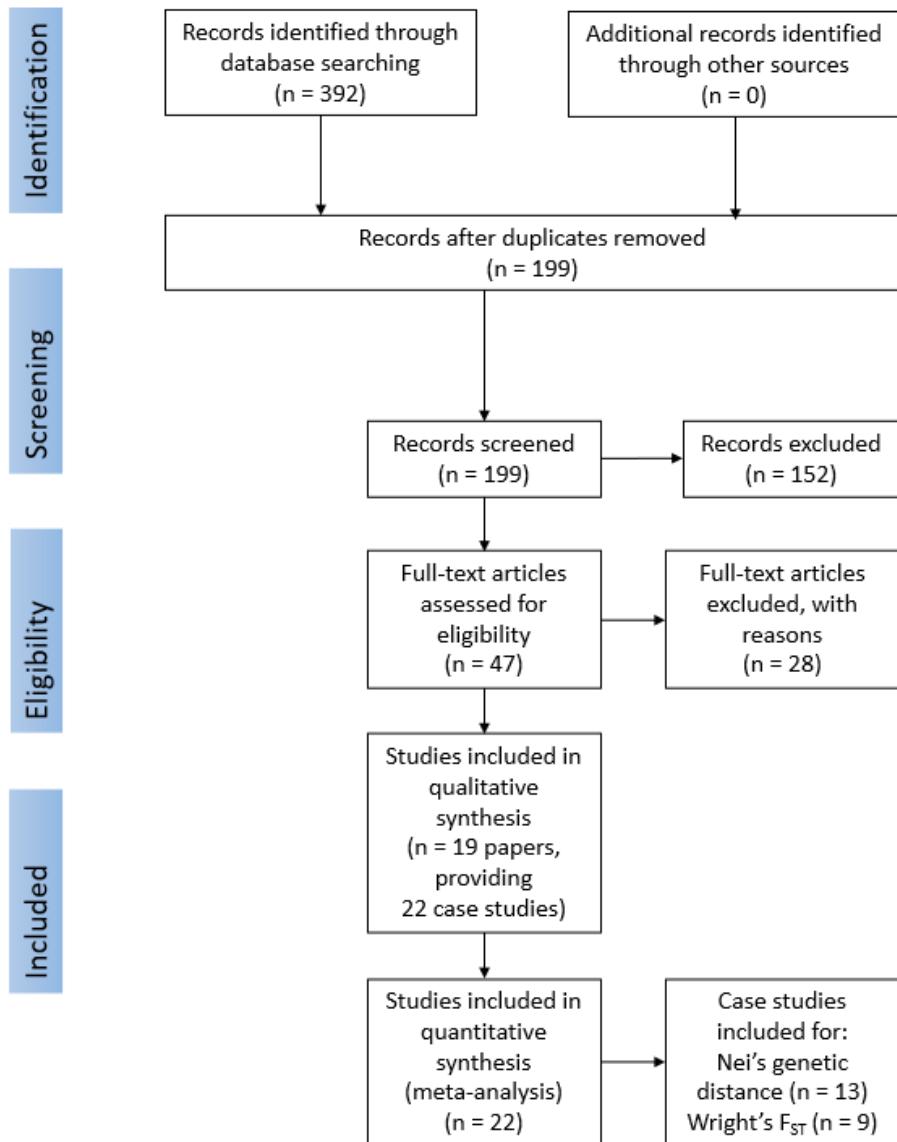
1   The meta-analysis was conducted using the metafor (Viechtbauer 2010) and robumeta (Fisher et al. 2017)

2   packages in R version 4.2.1 (R Core Team 2022). We used a random-effects model because studies were drawn

3   from different populations. For the analysis, Pearson's *r* was transformed to Fisher's *Z* to ensure normal

4   distribution. After performing meta-analytic calculations, Fisher's *Z* was converted back to Pearson's *r* for

5   reporting summary effect sizes (Quintana 2015).

6   The Q-statistic was calculated to assess heterogeneity among studies. The Q-statistic is the ratio of observed

7   variation to within-study variance. It evaluates the null hypothesis that all studies are examining the same effect

8   (Quintana 2015). Different studies used different mangrove plant species and different molecular markers to

9   assess genetic variation (Table S1). To assess the effect of this, we fitted separate meta-analytic models that

10 moderated for the effects of species and marker, respectively. To account for effect size dependency resulting

11 from the same study reporting multiple effect sizes (2 papers in the Nei's meta-analysis and 1 paper in the F_{ST}

12 meta-analysis), we used robust variance estimation as this is appropriate for meta-analyses with less than 40

13 studies and does not assume knowledge of within-study correlations (Quintana 2015). Egger's regression test

14 was used to assess publication bias by testing for funnel plot asymmetry. Publication bias is the phenomenon

15 whereby studies with larger effect sizes are more likely to be published, and therefore included in the meta-

16 analysis (Quintana 2015).

1 Results

2 Included studies

3   There were 13 case studies for Nei's genetic distance and 9 case studies for F_{ST} (Table S1, Figure 1). Egger's

4   regression test for funnel plot asymmetry showed no effect of publication bias (Nei's meta-analysis: z = 0.158, p

5   = 0.875. F_{ST} meta-analysis: z = 1.31, p = 0.192. Figure S1).
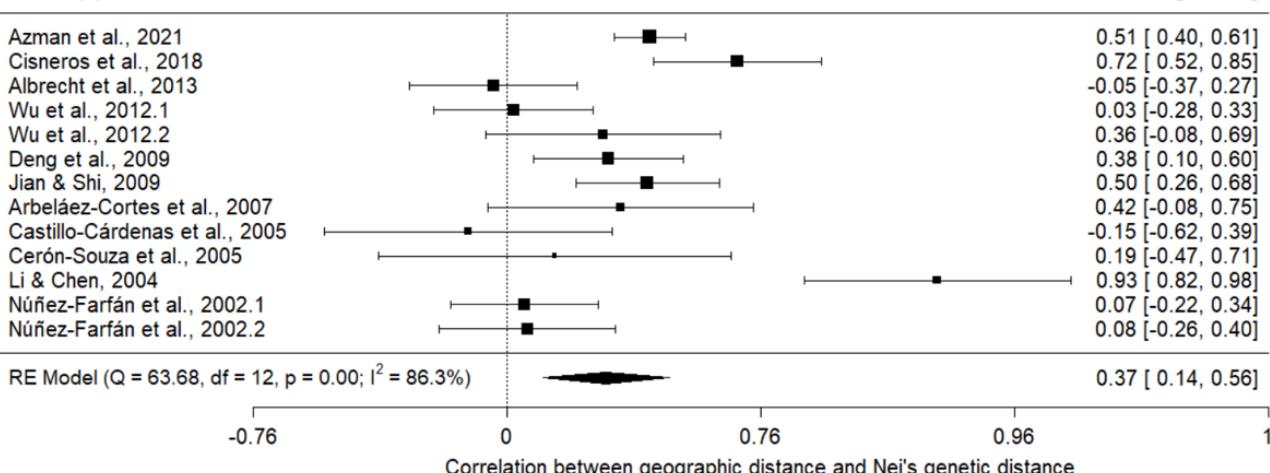
6   Figure 1. PRISMA flowchart (Moher et al. 2009) showing the sequence of selection of papers for meta-analysis

7   of the effects of geographic distance on genetic distance.

1 Effect sizes

2   We found a significant association between geographic distance and Nei's genetic distance (estimated model

3   coefficient = 0.37, 95 % CI = 0.14 - 0.56. $Z = 3.07$, p = 0.002. Figure 2A), which was not changed when

4   accounting for effect size dependency by robust variance estimation (estimated model coefficient = 0.39, 95 %

5   CI = 0.12 – 0.66). We also found a significant association between geographic distance and F_{ST} (estimated model

6   coefficient = 0.63, 95 % CI = 0.41 – 0.78. $Z = 4.77$, p < 0.0001. Figure 2b). This model coefficient was not

7   changed significantly when accounting for effect size dependency by robust variance estimation (estimated

8   model coefficient = 0.75, 95 % CI = 0.38 – 1.11).
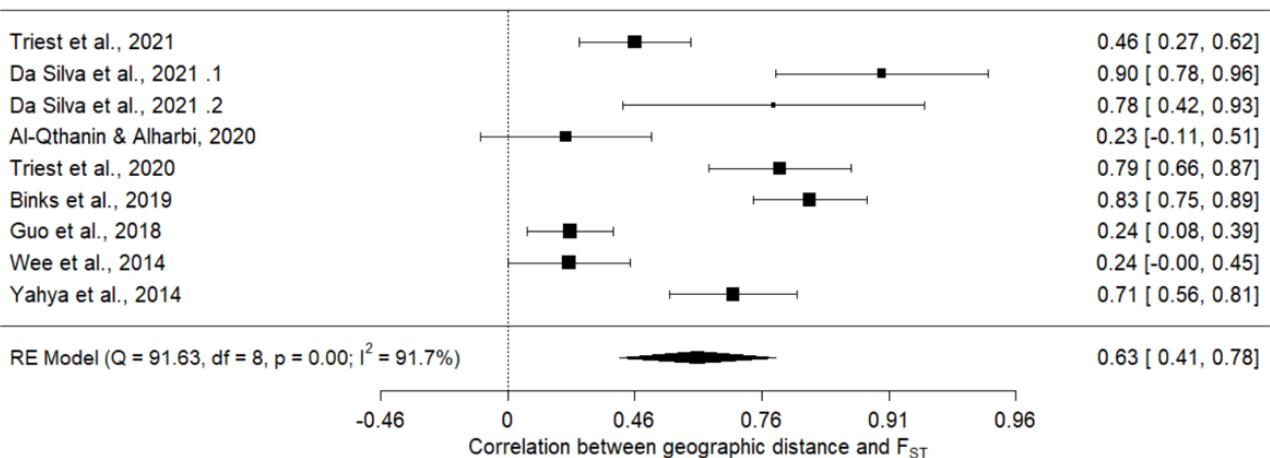
9   Figure 2. Effect of geographic distance on (A) Nei's genetic distance and (B) genetic distance measured by F_{ST}

10 for the original unadjusted models, with Fisher's Z transformed back to Pearson's r. The polygon at the bottom

8

1   of each plot shows the estimated model coefficient and its bounds represent the 95 % confidence intervals (CIs).

2   Each squared point corresponds to a different study, as labelled. The size of the square corresponds to the

3   contribution of the study to the estimated model coefficient.

4   **Heterogeneity and moderator analysis**

5   There was significant heterogeneity between the included studies (Nei's meta-analysis: Q = 63.7, df = 12, p <

6   0.0001. F_{ST} meta-analysis Q = 91.6, df = 8, p < 0.0001). Li & Chen (2004) contributed disproportionately to the

7   overall heterogeneity in Nei's meta-analysis. Therefore, for the purpose of comparison, a separate random-

8   effects model was fitted to the same data set but excluding Li & Chen (2004). This reduced the summary effect

9   size from 0.37 (95 % CI = 0.14 - 0.56) to 0.29 (95 % CI = 0.12 - 0.45). The overall heterogeneity was reduced,

10   but there was still significant heterogeneity (Q = 40.4, df = 11, p <0.0001). Since excluding this study did not

11   significantly remove the heterogeneity, all other analyses include this study.

12   To determine the source of heterogeneity, we conducted mixed-effects moderator analyses with taxon and

13   marker as separate moderator variables. The type of genetic marker used significantly moderated the correlation

14   between geographic distance and F_{ST} (QM = 12.0, df = 3, p = 0.0075), but not Nei's genetic distance (QM =

15   3.63, df = 4, p = 0.46). Whereas differences in the taxon investigated did not significantly moderate either of

16   these correlations (Nei's genetic distance: QM = 12.5, df = 7, p = 0.086. F_{ST}: QM = 7.50, df = 6, p = 0.28).

9

Discussion

Our global meta-analysis showed a significant correlation between geographic distance and genetic distance in mangrove plant communities. This isolation-by-distance effect could suggest that mangrove plants are limited in their capacity to disperse across habitat patches. This is consistent with the conclusions made by Binks et al. (2019) that habitat discontinuities lead to reduced gene flow between patches because mangrove propagules tend not to disperse far from their release point. Maintaining gene flow, which is critical for long-term population persistence (Salm et al. 2000, Wright et al. 2015), will depend on maintaining proximity among habitat patches, especially under conditions of habitat transformation which threaten mangrove biodiversity.

Isolation-by-distance indicates that spatial structure and genetic structure are highly correlated and suggests that dispersal limitation may be an important driver of mangrove community assembly. Dispersal is essential for enabling sessile organisms, such as plants, to move away from unfavourable conditions if they are unable to adapt to such conditions (Kinlan & Gaines 2003). Isolation-by-distance suggests that these important adaptive responses are constrained by natural dispersal mechanisms (Sexton et al. 2014). If mangrove propagules are limited in their dispersal capabilities, populations in habitat patches are more likely to become isolated from each other. This could result in a meta-population structure with smaller populations that are more vulnerable to demographic and environmental stochasticity (Lande 1993). While we recognise that dispersal depends on several biotic and abiotic factors that affect the release, transport and establishment of propagules (Van der Stocken et al. 2019b), the isolation-by-distance effect shown here highlights the importance of geographic distance in constraining gene flow. Therefore, when conserving and managing mangroves, the importance of maintaining close proximity between habitat fragments should be considered.

Our results indicate that the genetic structure of mangrove communities is dependent on spatial structure. For protected area networks to successfully maintain landscape connectivity, the size and arrangement of these networks should reflect the dispersal capabilities of the inhabiting species (Durrant et al. 2014, Shanks et al. 2003). Therefore, optimal design of protected area networks requires knowledge of effective dispersal distances. While the present study does not address exact distances, our results suggest that the realised dispersal

capabilities of mangrove propagules depend heavily on the geographic distances across which they are

dispersing. This emphasises the need for future studies to quantify effective dispersal distances in mangroves and

consider how mangrove dispersal could be affected by habitat change.

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Competing interest declaration

Competing interests: The authors declare none.

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