



Species-specific effects of near-future CO₂ on the respiratory performance of two tropical prey fish and their predator



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ABSTRACT

Ocean surface CO₂ levels are increasing in line with rising atmospheric CO₂ and could exceed 900 μatm by year 2100, with extremes above 2000 μatm in some coastal habitats. The imminent increase in ocean pCO₂ is predicted to have negative consequences for marine fishes, including reduced aerobic performance, but variability among species could be expected. Understanding interspecific responses to ocean acidification is important for predicting the consequences of ocean acidification on communities and ecosystems. In the present study, the effects of exposure to near-future seawater CO₂ (860 μatm) on resting ($\dot{M}O_{2\text{rest}}$) and maximum ($\dot{M}O_{2\text{max}}$) oxygen consumption rates were determined for three tropical coral reef fish species interlinked through predator-prey relationships: juvenile *Pomacentrus moluccensis* and *Pomacentrus amboinensis*, and one of their predators: adult *Pseudochromis fuscus*. Contrary to predictions, one of the prey species, *P. amboinensis*, displayed a 28–39% increase in $\dot{M}O_{2\text{max}}$ after both an acute and four-day exposure to near-future CO₂ seawater, while maintaining $\dot{M}O_{2\text{rest}}$. By contrast, the same treatment had no significant effects on $\dot{M}O_{2\text{rest}}$ or $\dot{M}O_{2\text{max}}$ of the other two species. However, acute exposure of *P. amboinensis* to 1400 and 2400 μatm CO₂ resulted in $\dot{M}O_{2\text{max}}$ returning to control values. Overall, the findings suggest that: (1) the metabolic costs of living in a near-future CO₂ seawater environment were insignificant for the species examined at rest; (2) the $\dot{M}O_{2\text{max}}$ response of tropical reef species to near-future CO₂ seawater can be dependent on the severity of external hypercapnia; and (3) near-future ocean pCO₂ may not be detrimental to aerobic scope of all fish species and it may even augment aerobic scope of some species. The present results also highlight that close phylogenetic relatedness and living in the same environment, does not necessarily imply similar physiological responses to near-future CO₂.

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1. Introduction

The average concentration of carbon dioxide (CO₂) in the atmosphere has increased from approximately 280 ppm in pre-industrial times (Barnola et al., 1987) to >400 ppm in 2013 (Dlugokencky and Tans, 2013) and is projected to exceed 900 ppm by the year 2100 if the current emission trajectory is maintained (Meinshausen et al., 2011). Because atmospheric and ocean surface pCO₂ are in equilibrium, CO₂ in the ocean is also increasing at approximately the same rate as in the atmosphere (Doney, 2010). Moreover, due to the hydrolysis of CO₂ in seawater, ocean surface pH is 0.1 unit lower today than preindustrial values and is predicted to be a further 0.3–0.4 units lower by 2100, which translates to a 100–150% increase in [H⁺] (Solomon et al., 2007). Some coastal regions could experience changes to [H⁺] that are at least 2–3 times the global average due to CO₂ enhancement from eutrophication

(Melzner et al., 2012) and amplification of natural CO₂ and pH variation (Shaw et al., 2013). Such changes in ocean chemistry are predicted to affect physiological functions of many marine organisms, with potentially far-reaching effects on marine diversity and ecosystem processes (Fabry et al., 2007; Pörtner, 2008; Gattuso and Hansson, 2011).

Ocean acidification has been hypothesized to have negative consequences for the performance of marine fishes, primarily through an effect of the capacity for oxygen supply and delivery (Pörtner et al., 2004). Aerobic scope, which represents the oxygen available for any activities beyond that required for basic maintenance (Fry, 1947, 1971; Fry and Hart, 1948), is expected to decline with increasing pCO₂ (Pörtner and Farrell, 2008). Reduced aerobic scope could affect individual fitness, since less energy can be devoted to digestion, growth and reproduction (Munday et al., 2009b; 2012). Reduced aerobic scope could also affect the outcome of key ecological interactions and ultimately the structure of ecological communities (Pörtner, 2008; Nilsson et al., 2009).

In accordance with these predictions, the aerobic capacity of two cardinalfish species from the Great Barrier Reef (GBR) was significantly reduced by exposure to CO₂-acidified water (Munday et al., 2009a). However, it may be expected that not all fish species of coral reef

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ecosystems will be similarly affected by CO₂-induced ocean acidification. Fishes are the most diverse group of vertebrates and the ontogenetic and lifestyle traits of some species could provide pre-adaptation to high ambient CO₂ (Ishimatsu et al., 2008; Melzner et al., 2009b). Indeed, the aerobic scope of some fish species appears to be unaffected by hypercapnia (McKenzie et al., 2003; Ishimatsu et al., 2008; Baker and Brauner, 2012). Interspecific differences in the response of aerobic scope to near-future seawater CO₂ could have important ecological ramifications, especially for species that interact through competitive or predator–prey relationships (Munday et al., 2012). Many studies implicate species interactions to be an important proximate cause of extinction due to climate change, particularly due to decreases in food availability (Cahill et al., 2012).

The objective of the present study was to assess species-specific effects of near-future seawater CO₂ on aerobic performance among tropical reef fish species involved in predator and prey relationships. The model predator species investigated was the brown dottyback (*Pseudochromis fuscus* Müller & Troschel 1849), a common mesopredator on the GBR. Two closely related damselfishes, the lemon damselfish (*Pomacentrus moluccensis* Bleeker 1853) and the Ambon damselfish (*Pomacentrus amboinensis* Bleeker 1868) were chosen as the model prey species. *P. fuscus* is known to be a major predator of recently settled juveniles of these two damselfishes on the GBR (Holmes and McCormick, 2010). Respirometry was utilized to measure the effects of exposure to elevated CO₂ on resting ($\dot{M}O_{2\text{rest}}$) and maximum ($\dot{M}O_{2\text{max}}$) oxygen consumption rates. In fish, physiological alterations in response to elevated ambient CO₂ can occur within minutes (i.e., ventilatory responses; Gilmour and Perry, 2006), hours (i.e., blood acid-base regulation; Brauner and Baker, 2009; Heisler, 1993; Esbaugh et al., 2012) to days (i.e., neurological disruptions; Nilsson et al., 2012). In terms of neurological impairments, four days of near-future CO₂ exposure has been shown to disrupt a number of sensory systems and alter the behavior of reef fishes, including the three species examined in the present study (Cripps et al., 2011; Ferrari et al., 2011a; Ferrari et al., 2011b; Nilsson et al., 2012). Longer exposure to elevated CO₂ does not induce further behavioral effects (Munday et al., 2010). Therefore, in our first experiment, we measured oxygen consumption after exposing the three species to ambient or elevated CO₂ for four days, to enable direct comparisons with previous studies on coral reef fish. In a second experiment, we measured oxygen consumption of the damselfishes, following acute exposure to near-future seawater CO₂ to determine if exposure to elevated CO₂ induces an immediate effect on aerobic performance. In both experiments, the elevated CO₂ treatment (860 µatm) was selected to approximate the level predicted for the atmosphere and ocean surface in 2100 under the IPCC A2 emissions scenario (Meehl et al., 2007). Finally, a third experiment was conducted to understand the effects of more extreme fluctuations in seawater pCO₂ and consequently pH that may occur in some coastal habitats, including shallow coral reef flats (Shaw et al., 2013). Four groups of *P. amboinensis* were exposed to one of four pH levels spanning from the present-day control pH of 8.1 down to pH 7.5 at 0.2 unit increments. The desired pH level was obtained by adding increasing volumes of 100 mM hydrochloric acid to the seawater. The addition of a strong acid to a closed system, such as a closed respirometer, has similar consequences on pCO₂ and pH as equilibrating water of an open system with CO₂ gas (Gattuso and Lavigne, 2009). Corresponding pCO₂ levels were approximately 450 µatm at pH 8.1, 860 µatm at pH 7.9, 1400 µatm at pH 7.7 and 2400 µatm at pH 7.5.

2. Materials and methods

2.1. Experimental fish

The experiments were conducted at Lizard Island Research Station (LIRS; 14°40'S, 145°28'E) between December and January (austral summer). Juvenile *P. moluccensis* (mean ± sd, 40.9 ± 5.6 mg) and

P. amboinensis (53.9 ± 12.4 mg) were caught at night using light traps moored 2 m below the surface and approximately 100 m off the reef (Meekan et al., 2001). In this location, the fish are trapped immediately before their arrival to the reef at the end of their planktonic larval stage (Meekan et al., 1993). Every morning, juveniles were collected from the traps and transferred to the laboratory where they were exposed to ambient or elevated CO₂ for four days (see Section 2.2). Adult *P. fuscus* (4.51 ± 0.78 g) were collected from shallow reefs (<6 m) in the Lizard Island lagoon using a hand-net after lightly anesthetizing them with a mixture of clove oil, ethanol and seawater (Munday and Wilson, 1997). Captured fish were transported to the research station where they were maintained for two days prior to exposure to CO₂ treatments. Fish were maintained at ambient ocean temperatures, which ranged from 28.3 to 30.4 °C (Table 1; 29.4 ± 0.1 °C) during the experimental period. Damselfishes were fed freshly hatched *Artemia* nauplii three times daily, and *P. fuscus* were fed twice daily to satiation with INVE Aquaculture Nutrition pellets. Feeding was discontinued 18–24 h prior to resting oxygen consumption measurements (see section 2.3.1). Animal care and experimental protocols complied with regulations at James Cook University and Lizard Island Research Station, and were approved by the James Cook University Ethic Committee (Approval # A1722). Fish were collected under permit G10/33239.1 from the GBR Marine Park Authority.

2.2. CO₂ exposure

Fish were exposed to either aerated control water (pCO₂ = 451 µatm) or 860 µatm CO₂ water (termed near-future seawater CO₂ in the present study) for four days (Table 1). Near-future seawater CO₂ concentrations were maintained by CO₂-dosing to a set pH_{NBS} (National Bureau of Standards) following standard techniques for ocean acidification research (Gattuso et al., 2010). Seawater was pumped from the ocean into two 60 l header tanks, one equilibrated with air (ambient control) and the other with CO₂ to achieve the pH expected to correspond to the ocean CO₂ concentration projected for 2100 (Meehl et al., 2007). The pH level was based upon preliminary observations of total alkalinity, salinity and temperature of seawater at Lizard Island. A pH-controller (Aqua Medic GmbH, Bissendorf, Germany) was attached to the CO₂-treated header tank to maintain pH at the desired level. A solenoid injected a slow stream of CO₂ into a submersible pump at the bottom of the header tank whenever the seawater pH rose above the set point. The pump ensured rapid dissolution of CO₂ into the seawater and also served as a vigorous stirrer. The pump in the control seawater header tank was injected with a slow stream of air. Seawater from each header tank was supplied at a rate of ca 500 ml min⁻¹ to four replicate 35 l aquaria for each species. pCO₂ in the aquaria was checked twice daily with a CO₂-permeable membrane connected to an infrared CO₂ probe (Vaisala GMP343, Vaisala, Helsinki, Finland) in a closed loop (Hari et al., 2008). Water samples were collected at the start, middle and end of the experiment in order to precisely determine pCO₂. Total alkalinity (A_T) of seawater was estimated by Gran titration (Gran, 1950; 1952) using certified reference material from Dr. A. G. Dickson (Scripps Institution of Oceanography), and average seawater pCO₂ was calculated with CO2SYS (<http://cdiac.ornl.gov/oceans/co2rpt.html>) from measured A_T and pH and using the constants of Mehrbach et al. (1973) refit by Dickson and Millero (1987).

Table 1

Mean (±SE) seawater parameters in the experimental system. pCO₂ was estimated with the program CO2SYS from measured pH, salinity and total alkalinity (A_T) of water samples.

Treatment	pH _{NBS}	Temperature (°C)	Salinity (ppt)	A_T (µmol kg ⁻¹ SW)	pCO ₂ (µatm)
Control	8.11–8.17	29.4 ± 0.1	34.5	2272 ± 13	451 ± 15
Near-future CO ₂	7.90–7.92	29.4 ± 0.1	34.5	2267 ± 2	860 ± 14

2.3. Experimental set-up and protocol

$\dot{M}O_{2\text{rest}}$ and $\dot{M}O_{2\text{max}}$ were used as proxies for resting and maximum metabolic rates and were measured by respirometry as previously utilized for assessing the effects of climate change variables on other fish (see Ishimatsu et al., 2008 for review), especially tropical reef fish species (Nilsson and Östlund-Nilsson, 2004; Nilsson et al., 2007a; Nilsson et al., 2007b; Munday et al., 2009a; Gardiner et al., 2010; Nilsson et al., 2010). Respirometry chambers were immersed in temperature-controlled (29 °C) aquaria continuously supplied with either air- or near-future CO₂-equilibrated seawater.

2.3.1. Resting oxygen consumption

Cylindrical 26.7-ml static respirometers were used for *P. moluccensis* and *P. amboinensis* juveniles. After four days of exposure to current day or near-future CO₂ seawater, one fish was transferred to each respirometry chamber. The chamber was left open and the fish left undisturbed to habituate to the chamber for 1–2 h whereupon the chamber was cautiously closed without disturbing the fish. Previous experiments have shown that habituation periods longer than 2 h in the chamber do not further reduce $\dot{M}O_2$ (Nilsson and Östlund-Nilsson, 2004; Nilsson et al., 2010). All fishes included in this study settled down rapidly and remained virtually motionless during the measuring period. Once the chamber had been sealed, water oxygen concentration was recorded continuously with an oxygen probe (CellOx 325, WTW, Germany; calibrated daily) connected to an oxygen meter (OXI 340i, WTW, Germany). The oxygen probe was fitted with a magnetic propeller (BOD stirring accessory, WTW, Germany) set in motion with a magnetic stir plate situated outside the aquarium along the glass wall. The propeller ensured gentle water mixing inside the respirometer and water renewal along the O₂ probe membrane during the habituation and recording periods. The oxygen meters were connected to a data acquisition system (PowerLab 4/20, ADInstruments, Colorado Springs, USA). $\dot{M}O_{2\text{rest}}$ was calculated from the steady rate of oxygen consumption observed between 100 and 90% of air saturation. The decrease of water oxygen concentration was recorded until it reached approximately 10% of air saturation in order to calculate the critical oxygen concentration (O_{2crit}), which is the lowest O₂ concentration where the fish is still able to maintain $\dot{M}O_{2\text{rest}}$. O_{2crit} was reached 2.5–3 h after the chamber was sealed. For each species, two parallel setups allowed for the simultaneous recording of one fish in near-future CO₂ water and another fish in control conditions. Similar to a number of other prior experiments conducted to determine the O_{2crit} of a fish using this classic protocol, the measurement of O_{2crit} required the respirometer to remain closed until almost all O₂ was depleted from the system. Consequently, CO₂ concomitantly increased in the respirometer due to the respiration of the fish and the fish simultaneously experienced hypoxia and increasing hypercapnia during the O_{2crit} measurement. However, most of the excreted CO₂ would convert rapidly to bicarbonate. Assuming a respiratory quotient of 1.0, calculations of the rise in dissolved inorganic carbon (DIC) in the respirometers caused by the respiration of the fish estimate that the maximum build-up of CO₂ for each 10% fall in O₂ was 9.7 and 12.2% for control and near-future CO₂ fish respectively.

For *P. fuscus*, $\dot{M}O_{2\text{rest}}$ was measured in 1615-ml intermittent-flow respirometers. Fish were first habituated to the chambers for 90 min. Preliminary experiments determined that 90 min was ample time for this species to ensure O₂ consumption rates had reached the lowest possible values under the experimental conditions. Beyond this time, O₂ consumption rates did not significantly vary. Submersible pumps supplied a water flow (150 l h⁻¹) from the aquaria through the chambers and after the habituation period, water flow to each chamber was stopped for 15 min every 30 min over a period of 90 min. The time the water flow was interrupted was short enough to ensure O₂ did not fall below 80% of air saturation. Water oxygen concentration (mg l⁻¹) was continuously recorded at a frequency of 1 Hz using oxygen-sensitive REDFLASH dye on contactless spots (2 mm) adhered to the

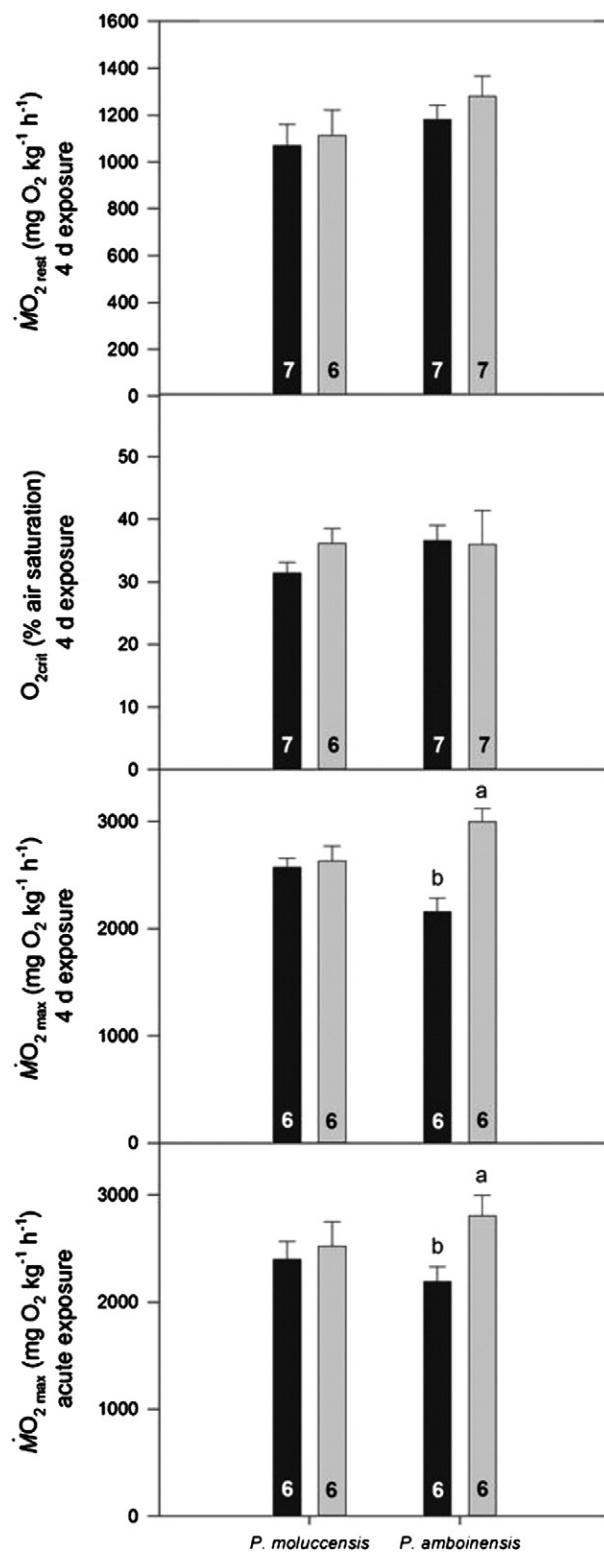


Fig. 1. Respiratory performance (mean \pm SE) of *P. moluccensis* and *P. amboinensis*. A. Resting oxygen consumption rate ($\dot{M}O_{2\text{rest}}$; $\text{mg O}_2 \text{ kg}^{-1} \text{ h}^{-1}$), B. critical oxygen concentration ($O_{2\text{crit}}$; % of air saturation), C. maximum oxygen consumption rate ($\dot{M}O_{2\text{max}}$; $\text{mg O}_2 \text{ kg}^{-1} \text{ h}^{-1}$) after a 4 d exposure to control seawater (451 μatm ; ■) or near-future CO₂ seawater (860 μatm ; □) and D. $\dot{M}O_{2\text{max}}$ after an acute exposure to near-future CO₂ seawater. Letters that differ indicate statistically significant differences (see text for *P* values). N numbers are indicated at the bottom of each bar.

inside of each chamber and connected via fiber-optic cable to a Firestim Optical Oxygen Meter (Pyro Science e. K., Aachen, Germany). Data were analyzed using LabChart 6.1.3 (ADInstruments, Colorado Springs, USA).

2.3.2. Maximum rate of oxygen consumption

$\dot{M}O_{2\max}$ was measured in custom made cylindrical swimming chambers previously utilized by a large number of studies for small tropical coral reef fish species (see Nilsson et al., 2007b for diagram and detailed description of the set-up; Gardiner et al., 2010; Nilsson et al., 2009; Munday et al., 2009a; Nilsson et al., 2007a; Nilsson et al., 2010; Nilsson and Östlund-Nilsson, 2004). The volume of the chambers ranged between 215 and 267 ml for *P. moluccensis* and *P. amboinensis* juveniles and was 1594 ml for *P. fuscus*. The decrease in $[O_2]$ was measured with an O_2 electrode (WTW CellOx 325, as above). It has been suggested that the aerobically fuelled muscle mass in some fish is not large enough to force them to reach the maximum rate of oxygen uptake during maximal swimming performance (Goolish, 1991). Therefore, fish were fed *ad libitum* prior to estimating $\dot{M}O_{2\max}$, since it is likely that the combined oxygen needs of digestion and maximal swimming would be high enough to engage the full capacity of the respiratory system (Bennett and Hicks, 2001; Gardiner et al., 2010). Up to three *P. amboinensis* (to increase the total fish weight so that reliable recordings could be made), one *P. moluccensis* or one *P. fuscus* (same fish as for $\dot{M}O_{2\text{rest}}$ measurements) were placed in the swimming chamber. The water speed was regulated with a magnetic stirrer located beneath the chamber. As soon as the water was set in motion, the fishes started swimming against the current, apparently guided by landmarks provided by items such as the oxygen electrode and the edges of the surrounding aquarium. The speed was consistently increased to a point where it was assumed that the fishes swam at their maximum speed. The water speed corresponded to the point at which the fishes could just barely maintain a steady position in the chamber. At a slightly higher speed, the fishes were no longer able to maintain position for more than a few seconds and stopped swimming. Nilsson et al. (2007b) showed that water speeds of approximately 50 and 125 cm s⁻¹ could be achieved near the inner and outer wall of the chamber, respectively. The speeds are more than sufficient for each of the species examined to reach their maximal swimming speed. Pre-settlement larvae of both *Pomacentrus* species as well as *P. fuscus*, which are more efficient swimmers than the post-settlement larvae and adults studied here (Nilsson et al., 2007b) exhibit an average maximum sustained swimming speed ranging from less than 30 cm s⁻¹ to a maximum of 36 cm s⁻¹ (Fisher et al., 2005). The decrease in oxygen concentration was recorded at a frequency of 1 Hz in the chamber at the maximum swimming speed for up to 6 min, during which time oxygen concentration remained above 90% of air saturation.

First, $\dot{M}O_{2\max}$ was measured in the three species that had been maintained in control or near-future CO_2 seawater for four days. Then, in order to assess an acute effect of near-future CO_2 , $\dot{M}O_{2\max}$ was measured in *P. moluccensis* and *P. amboinensis* maintained in control conditions for four days and acutely exposed to near-future CO_2 while in the respirometer. Finally, to further understand the effects of increased CO_2 and low pH in more extreme habitats, $\dot{M}O_{2\max}$ of four groups of *P. amboinensis* was measured after acute exposure of the fish to pH levels of 8.1 (control, present day level, corresponding to approximately 450 μatm CO_2), 7.9 (corresponding to approximately 860 μatm CO_2), 7.7 (corresponding to approximately 1400 μatm CO_2) and 7.5 (corresponding to approximately 2400 μatm CO_2). The desired pH level was obtained by diluting increasing volumes of 100 mM hydrochloric acid in the seawater utilized in the experimental set-up. The water was prepared immediately prior to the experiment, and within 2 min, the fish was placed in the respirometry chamber, which was sealed, thereby preventing the water from equilibrating with the atmosphere. $\dot{M}O_{2\max}$ was then measured as described above. Average seawater pCO_2 was calculated with CO2SYS (<http://cdiac.ornl.gov/oceans/co2rpt.html>) from measured pH and assuming the same A_T as in the control conditions. The addition of a strong acid to a closed system like a closed respirometer has rather similar consequences on water chemistry as equilibrating water of an open system with CO_2 gas, thereby allowing reasonable comparison between both techniques (Gattuso and Lavigne, 2009). However, without the addition of CO_3^{2-} or HCO_3^- , the technique leads to a slightly

lower A_T . For example, Gattuso et al. (2010) report that the A_T of seawater with a salinity of 35 ppt will decrease by 6% when pH is decreased from 8.1 to 7.8 via the addition of HCl in a closed system. This would reduce pCO_2 estimates in our system by 6%. Therefore, our estimates of pCO_2 from hydrochloric acid addition may be marginally higher than what was achieved in the respirometer prior to the start of the experiment and addition of respiratory CO_2 from the fish.

Because the measurements of $\dot{M}O_{2\max}$ in the small damselfish (*P. amboinensis*) required the pooling of individuals (see above), $\dot{M}O_{2\text{rest}}$ and $\dot{M}O_{2\max}$ had to be measured on different sets of fish. The procedure precluded the calculation of individual aerobic scope in damselfish. As for the measurement of $\dot{M}O_{2\text{rest}}$, two parallel setups allowed for the simultaneous measurement of $\dot{M}O_{2\max}$ of fish exposed to near-future or higher CO_2 , and those exposed to control conditions.

2.4. Data analyses and statistics

Oxygen consumption ($\dot{M}O_2$ in $\text{mgO}_2 \text{ kg}^{-1} \text{ h}^{-1}$) was calculated for each fish or pool of fish using the following formula:

$$\dot{M}O_2 = \Delta [O_2] \times \Delta t^{-1} \times VOL_{\text{resp}} \times M^{-1}$$

where $\Delta [O_2]$ is the decrease in water oxygen concentration ($\Delta \text{mgO}_2 \text{ l}^{-1}$), Δt the recording time (h), VOL_{resp} is the volume of the respirometer minus the volume of the fish (l), and M the mass of the fish (kg). Fish $\dot{M}O_2$ was corrected for background (microbial) respiration measured after each recording. For *P. moluccensis* and *P. amboinensis*, $\dot{M}O_{2\text{rest}}$ was measured in a closed respirometer from $\Delta [O_2]$ data above 90% of air saturation. $O_{2\text{crit}}$ was calculated for each damselfish as the concentration of O_2 at the intersection of the regression lines from the data recorded >80% of air saturation (well above $O_{2\text{crit}}$) and the data recorded <20% air saturation (well below $O_{2\text{crit}}$ as evident from the break in the curve at $O_{2\text{crit}}$). For *P. fuscus*, three slopes ($\Delta [O_2] \times \Delta t^{-1}$) were averaged to calculate $\dot{M}O_{2\text{rest}}$. $\dot{M}O_{2\max}$ was calculated for all fish from within the first minute of recording. Unpaired *t*-tests were utilized to assess the effect of seawater pCO_2 on $\dot{M}O_{2\text{rest}}$ and $\dot{M}O_{2\max}$ for each species, and the effect of seawater CO_2 concentration on $O_{2\text{crit}}$ in *P. moluccensis* and *P. amboinensis*. In the experiment examining the effect of extreme pH on $\dot{M}O_{2\max}$ in *P. amboinensis*, no significant differences were found between the three control groups ($P = 0.236$). Therefore, the control data were pooled. Then, a one-way ANOVA was used to test for the effect of pH on $\dot{M}O_{2\max}$. In all instances, $P < 0.05$ was considered significant.

3. Results

$\dot{M}O_{2\text{rest}}$ of the two damselfish prey species was unaltered after exposure to near-future CO_2 for four days (Fig. 1A). Similarly, four days of exposure to near-future CO_2 had no significant effect on $O_{2\text{crit}}$ for either of the two damselfish prey species (Fig. 1B).

$\dot{M}O_{2\max}$ of *P. moluccensis* was unaffected by four days exposure as well as acute exposure to near-future CO_2 (Fig. 1C and D). In contrast, $\dot{M}O_{2\max}$ of *P. amboinensis* was significantly higher after four days (+39%; $t_{[10]} = 4.665$; $P = 0.009$) as well as acute (+28%; $t_{[10]} = 2.597$; $P = 0.029$) exposure to near-future CO_2 when compared to controls (Fig. 1C and D). Furthermore, the two CO_2 treatments resulted in a similar increase in $\dot{M}O_{2\max}$. $\dot{M}O_{2\max}$ of *P. amboinensis* measured after four days exposure to near-future CO_2 was not significantly different than $\dot{M}O_{2\max}$ measured after acute exposure to near-future CO_2 ($P = 0.417$). Similarly, $\dot{M}O_{2\max}$ of the respective control groups did not differ significantly ($P = 0.858$).

Exposure to near-future CO_2 for four days had no effect on either $\dot{M}O_{2\text{rest}}$ or $\dot{M}O_{2\max}$ of the predator, *P. fuscus* (Fig. 2). Consequently, there was no significant effect of near-future CO_2 on net aerobic scope or factorial aerobic scope in this species.

When acutely exposed to pH 7.9 (approximately 860 μatm pCO_2), *P. amboinensis* exhibited a $\dot{M}O_{2\max}$ that was significantly higher than

that of the control fish, as well that of fish acutely exposed to pH 7.7 (approximately 1400 μatm $p\text{CO}_2$) or 7.5 (approximately 2400 μatm $p\text{CO}_2$; Fig. 3; $P < 0.0001$). The increase in $\dot{M}\text{O}_{2\text{max}}$ (+32%) was quantitatively similar to the increased $\dot{M}\text{O}_{2\text{max}}$ displayed by *P. amboinensis* after the four-day and acute exposures to near-future CO_2 . However, at pH 7.7 and 7.5, $\dot{M}\text{O}_{2\text{max}}$ of *P. amboinensis* was no longer elevated and was not significantly different from the control (Fig. 3).

4. Discussion

CO_2 -driven ocean acidification has been predicted to have detrimental effects on marine organisms by reducing the scope for aerobic performance (Pörtner and Farrell, 2008; Pörtner and Knust, 2007; Seibel and Walsh, 2001). However, contrary to expectations, none of the three tropical reef fish examined here, *P. moluccensis*, *P. amboinensis* or *P. fuscus*, exhibited an elevated $\dot{M}\text{O}_{2\text{rest}}$ or reduced $\dot{M}\text{O}_{2\text{max}}$ when exposed to the average $p\text{CO}_2$ projected to occur in the ocean surface by the year 2100. Rather, $\dot{M}\text{O}_{2\text{rest}}$ of the two juvenile damselfish prey species (*P. moluccensis* and *P. amboinensis*) and the adult predator (*P. fuscus*) was maintained after a four-day exposure to near-future seawater CO_2 . Early life stages of fish and other marine organisms are believed to be more sensitive to pH changes because of their high metabolic demand (Brown and Sadler, 1989; Pörtner et al., 2005). The unchanged $\dot{M}\text{O}_{2\text{rest}}$ of juvenile *P. moluccensis* and *P. amboinensis* after four day exposure to near-future CO_2 thus suggests that the hypothesized metabolic costs of living in a high CO_2 environment, namely altered acid-base balance, ionoregulation and cardiorespiratory function (Pörtner et al., 2004), were insignificant for these species at rest. Nevertheless, the unchanged $\dot{M}\text{O}_{2\text{rest}}$ of the species examined after four days exposure to near-future CO_2 does not preclude that physiological changes occurred, including compensatory ones. For example, gill ionoregulatory machinery is rapidly altered (within 8 h to 2 d) in response to hypercapnia exposure in the eelpout (*Zoarces viviparous*), without measurable effect on resting metabolic rate over the same time period (Deigweiler et al., 2008). Similarly, long-term (4–12 months) exposure to hypercapnia leads to upregulated gill Na^+/K^+ -ATPase activity and protein expression in the Atlantic cod (*Gadus morhua*), but resting and low activity metabolic rates remain unaltered (Melzner et al., 2009a). Furthermore, the gilthead bream

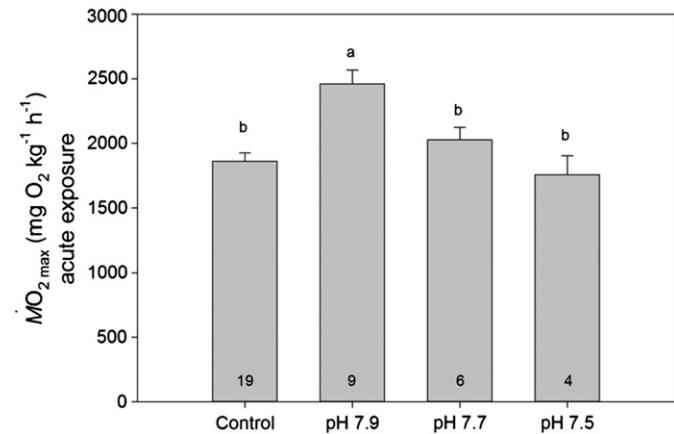


Fig. 3. Maximum oxygen consumption rate ($\dot{M}\text{O}_{2\text{max}}$; $\text{mg O}_2 \text{ kg}^{-1} \text{ h}^{-1}$; mean \pm se) of *P. amboinensis* at different levels of seawater pH. N numbers are indicated at the bottom of each bar.

(*Sparus auratus*) exhibits a shift from aerobic to anaerobic metabolic pathways during hypercapnic exposure (Michaelidis et al., 2007). Whether similar compensatory physiological changes also occur in coral reef fish during exposure to near-future climate change CO_2 levels remains to be investigated.

Like observed for $\dot{M}\text{O}_{2\text{rest}}$, $\dot{M}\text{O}_{2\text{max}}$ of *P. moluccensis* and *P. fuscus* was also unaffected by exposure to near-future CO_2 . In contrast, $\dot{M}\text{O}_{2\text{max}}$ of *P. amboinensis* was higher under near-future CO_2 conditions than under control conditions. The lack of a detrimental effect of increased CO_2 on $\dot{M}\text{O}_{2\text{rest}}$ and $\dot{M}\text{O}_{2\text{max}}$, as reported here for *P. moluccensis* and *P. fuscus*, is not unprecedented for fish (McKenzie et al., 2003; Ishimatsu et al., 2008; Melzner et al., 2009a). However, to the best of our knowledge, an augmented aerobic capacity of a juvenile marine teleost in response to elevated CO_2 has not been documented previously. The validity of the present results for *P. amboinensis* are supported by the findings that $\dot{M}\text{O}_{2\text{max}}$ was increased with elevated CO_2 in separate groups of fish that were exposed to near-future CO_2 using three different experimental techniques. The specific treatments consisted of a four-day exposure to near-future CO_2 maintained by CO_2 -dosing to a set pH_{NBS} , an acute exposure to near-future CO_2 maintained by CO_2 -dosing to a set pH_{NBS} and an acute exposure to near-future CO_2 obtained by the addition of strong acid into a closed system. Across the three methodologies, all 21 individuals exposed to near-future CO_2 and the resulting acidosis exhibited an increased $\dot{M}\text{O}_{2\text{max}}$. Moreover, the magnitude of the increase in $\dot{M}\text{O}_{2\text{max}}$ was consistent among the different experimental protocols (+28–39%). Finally, $\dot{M}\text{O}_{2\text{max}}$, $\dot{M}\text{O}_{2\text{rest}}$ and O_2crit of *P. amboinensis* in control conditions were comparable to data from a previous study using similar size fish (Nilsson et al., 2007b), indicating that the elevated $\dot{M}\text{O}_{2\text{max}}$ under near-future CO_2 did not arise from a comparison to control fish that were underperforming.

The physiological mechanism(s) underlying the increased $\dot{M}\text{O}_{2\text{max}}$ of *P. amboinensis* under near-future seawater CO_2 conditions remain to be elucidated. However, the consistent increase of $\dot{M}\text{O}_{2\text{max}}$ exhibited by *P. amboinensis* across the three experimental protocols, two of which constituted an acute exposure to near-future CO_2 , suggests that the phenomenon did not arise from physiological acclimation to the elevated CO_2 , but rather from the consequences of an acute exposure to increased CO_2 on the physiology of the fish. One explanation for the increased $\dot{M}\text{O}_{2\text{max}}$ of *P. amboinensis* when exposed to near-future CO_2 is that the maximum swimming speed of the fish was greater. Indeed, it is well established that $\dot{M}\text{O}_{2\text{max}}$ of fishes is positively correlated with swimming speed (Smith, 1965; Fry, 1971; Torres and Childress, 1983; Bushnell et al., 1984; Lee et al., 2003). A greater swimming speed could arise from a number of possibilities.

A greater maximal swimming speed could stem from a change in the motivation of the fish to swim fast. Recent studies have revealed that

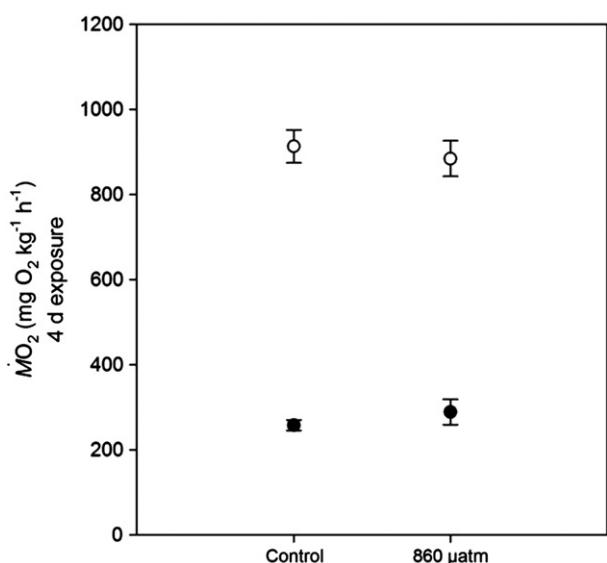


Fig. 2. Respiratory performance (mean \pm se) of *P. fuscus*. Resting oxygen consumption rate ($\dot{M}\text{O}_{2\text{rest}}$; $\text{mg O}_2 \text{ kg}^{-1} \text{ h}^{-1}$) (●) and maximum oxygen consumption rate ($\dot{M}\text{O}_{2\text{max}}$; $\text{mg O}_2 \text{ kg}^{-1} \text{ h}^{-1}$) (○) after a 4 d exposure to control seawater (451 μatm ; n = 8) or near-future CO_2 seawater (860 μatm ; n = 7).

tropical reef fish exhibit an array of behavioral and sensory disruptions when exposed near-future CO₂ for several days. These range from reversal or loss of olfactory and auditory preferences, loss of behavioral lateralization to increased boldness and activity levels (Dixson et al., 2010; Munday et al., 2010; Simpson et al., 2011; Domenici et al., 2012; Nilsson et al., 2012). The behavioral changes are believed to arise from alterations of the normal flow of Cl⁻ and HCO₃⁻ through GABA-A receptors caused by disruptions of transmembrane Cl⁻ and HCO₃⁻ ion gradients (Nilsson et al., 2012). Nevertheless, we find it unlikely that the increased $\dot{M}O_{2\max}$ of *P. amboinensis* arises from a motivational drive to swim faster due to the effects of CO₂ exposure on neurotransmitter function because the previously documented behavioral alterations only occurred after several days' exposure to increased CO₂ (Dixson et al., 2010; Munday et al., 2010; Simpson et al., 2011; Domenici et al., 2012; Nilsson et al., 2012). Short-term fluctuations in CO₂ did not induce behavioral effects (Munday et al., 2010). In comparison, in the present study, *P. amboinensis* displayed an elevated $\dot{M}O_{2\max}$ immediately (within minutes) of exposure to near-future CO₂. Moreover, a recent study investigating acid-base balance of the gulf toadfish (*Opsanus beta*) during exposure to near-future seawater CO₂, reported that full compensation of the respiratory acidosis and elevation of plasma HCO₃⁻ did not occur until after two hours of exposure (Esbaugh et al., 2012). Again, the time course of physiological change is at odds with alacritous increase of $\dot{M}O_{2\max}$ displayed by *P. amboinensis* in the present study.

Alternatively, a greater swimming speed could arise from an increased oxygen delivery to the swimming muscles and oxygen uptake at the gills. Recent *in vitro* and *in vivo* studies have revealed that moderate acidosis can serve to increase the delivery of oxygen to the red muscle of teleost fish (Rummer and Brauner, 2011; Rummer et al., 2013). Briefly, in fish exposed to a stressor or mild acidosis, catecholamine release stimulates the activation of Na⁺/H⁺ exchange across the erythrocyte membrane, thereby increasing red blood cell intracellular pH relative to the plasma and thus facilitating hemoglobin-O₂ binding at the gills (Boutilier et al., 1986; Nikinmaa, 1986). However, if plasma accessible carbonic anhydrase, which catalyzes the reversible conversion of HCO₃⁻ and H⁺ to CO₂, is present in muscle capillaries, it serves to short-circuit the Na⁺/H⁺ exchange, reduce red blood cell pH and hemoglobin-O₂ affinity and enhance O₂ unloading. For rainbow trout (*Oncorhynchus mykiss*) exposed to less than 1% CO₂, the mechanism increased red muscle pO₂ by 65% (Rummer et al., 2013). In the present study, the possibility exists that the combined exposure of *P. amboinensis* to near-future CO₂ and maximal exercise led to catecholamine release. If carbonic anhydrase is present in the red muscle of *P. amboinensis*, an increased oxygen delivery to the swimming muscles would have also likely ensued. Moreover, external CO₂ can rapidly induce cardiorespiratory responses via gill chemoreceptors (Reid et al., 2005). In particular, environmental CO₂ elicits increased ventilation in most water-breathing fishes. An increased oxygen delivery to the muscles combined with an increased ventilation rate could translate to the observed greater oxygen uptake.

In this regard, the species-specific responses to near-future CO₂ in terms of $\dot{M}O_{2\max}$ may indicate different response times for catecholamine release, regulatory capacities or tolerance to changes in blood H⁺ and/or differences among the species with regards to the presence of plasma accessible carbonic anhydrase in the red muscle. The lack of increase of $\dot{M}O_{2\max}$ of *P. amboinensis* when the fish were exposed to pH 7.7 (corresponding to approximately 1400 μ atm CO₂) or pH 7.5 (corresponding to approximately 2400 μ atm CO₂) could arise from negative physiological effects of acidosis at the more extreme lower pH levels (Brauner and Baker, 2009) nullifying the enhanced physiological capacity to swim faster. This possibility could explain why many previous studies that exposed fish to much higher CO₂ levels (3000–60000 μ atm) than those employed in the present study did not report any positive effects on $\dot{M}O_{2\max}$ or aerobic scope (McKenzie et al., 2003; Melzner et al., 2009a). The juvenile damselfish and adult *P. fuscus* studied were much too small to enable blood sampling to test the above hypotheses.

An alternate explanation for the increased $\dot{M}O_{2\max}$ of *P. amboinensis* under near-future CO₂ conditions is that maximal swimming speed was unchanged at $\dot{M}O_{2\max}$, but an additional demand for oxygen to maintain homeostasis arose that was not apparent at $\dot{M}O_{2\text{rest}}$ or during maximal swimming under control conditions. For example, the 'osmo-respiratory compromise' almost doubles with exercise (Randall et al., 1972; Nilsson, 1986). In this scenario, *P. amboinensis* would have incurred a greater cost to swim at its maximum speed under near-future CO₂ conditions. Clearly, future studies incorporating the measurement of swimming speed at $\dot{M}O_{2\max}$ are needed to differentiate the possible explanations for the increased $\dot{M}O_{2\max}$ of *P. amboinensis* under near-future CO₂. The use of a swimming flume was not feasible in the present study because such a system would have required a much larger volume of water than appropriate to accurately measure oxygen consumption of the extremely small juvenile fish (Steffensen, 1989). Rather, a cylindrical swimming chamber with a small volume, but with the capacity to swim the fish at their maximal swimming speed (Nilsson and Östlund-Nilsson, 2004; Fisher et al., 2005; Nilsson et al., 2007a; Munday et al., 2009a; Nilsson et al., 2009; Gardiner et al., 2010; Nilsson et al., 2010) was utilized to obtain reliable measurements of oxygen consumption.

Regardless of the possible mechanistic determinant(s) of the elevated $\dot{M}O_{2\max}$ under near-future CO₂, the differing response to near-future CO₂ between *P. amboinensis* and its congener *P. moluccensis* could foreseeably have consequences for ecological interactions and the relative abundance of species within coral reef fish communities. With a higher aerobic metabolic capacity (and potentially maximum swimming speed), *P. amboinensis* would have the potential for increased individual performance in any energetically demanding behavior, such as swimming against a current, repaying O₂ debt after repeated anaerobic burst-swimming escapes from a threat, foraging or digesting (i.e., specific dynamic action). Concurrently, *P. amboinensis* should still be able to enter and remain in the hypoxic waters found deep inside coral colonies at night in order to escape predation (Nilsson et al., 2007a). The unchanged O_{2crit} of *P. amboinensis* after exposure to near-future CO₂ suggests that no trade-off exists for this species between its higher aerobic capacity and its hypoxia tolerance. In addition, in face of warming ocean surface temperatures, the enhanced $\dot{M}O_{2\max}$ of juvenile *P. amboinensis* under near-future CO₂ may enable it to maintain its thermal tolerance window, and perhaps geographical distribution, as thermal tolerance is thought to be guided by aerobic scope in many species (Pörtner and Knust, 2007). On the contrary, if *P. amboinensis* incurs a greater metabolic cost while swimming at its maximum speed under near-future CO₂ conditions, energy expenditure for any other energetically demanding process would be reduced. Consequently, individual performance would be decreased in any energetically demanding behavior, leading to potentially negative consequences for the species.

Interestingly, the results of a recent experiment that examined mortality rate of a number of juvenile damselfish species when facing *P. fuscus* in a mesocosm after a four-day exposure to CO₂-acidified water found that *P. amboinensis* showed a similar mortality rate to *P. moluccensis*, as well as to other damselfish species (Ferrari et al., 2011b). The findings suggest no benefit or disadvantage of the increased $\dot{M}O_{2\max}$ displayed by *P. amboinensis* under near-future CO₂ conditions. It may be that larger scale and longer term studies that encompass a variety of other variables such as temperature, current, life stages and the presence of other predators are required to reveal the implications of species-specific $\dot{M}O_{2\max}$ responses to near-future CO₂ on fitness and mortality rate. Another recent study reported that in CO₂ acidified water, *P. fuscus* had a slower response to prey detection than in control water, but higher activity levels (Cripps et al., 2011). The higher activity level had been suggested to compensate for slower prey detection by increasing the chance of prey encounter (Cripps et al., 2011). The present findings indicate that a higher activity level of *P. fuscus* in CO₂-acidified water is not linked to a higher aerobic performance (aerobic scope). The higher activity levels of *P. fuscus* might be one of the consequences of increased neural excitation in CO₂-acidified water (Nilsson et al., 2012).

5. Concluding remarks

In summary, the present study reports an increased aerobic capacity of a juvenile marine teleost, *P. amboinensis*, in response to near-future pCO₂, but no effect on its congener *P. moluccensis* or on their predator *P. fuscus*. While the mechanistic basis for the species-specific responses and potential for differences in the swimming performance of the fish under near-future pCO₂ remains to be investigated, the results emphasize that being of the same genus, sharing similar ecology and life history, or living in the same environment, does not necessarily imply similar physiological responses to near-future CO₂. The results highlight that understanding interspecific variability is an important component of predicting the consequences of ocean acidification on marine communities and ecosystems. Additional studies assessing the effects of a number of other environmental factors in conjunction with near-future CO₂ exposure are required to more fully understand the interesting finding of the increased $\dot{M}O_{2\text{max}}$ of *P. amboinensis* in response to near-future seawater CO₂. Most importantly, future experiments should assess how maximum swimming ability is affected and if species-specific responses to elevated CO₂ persist with elevated temperature. Elevated pCO₂ in the future will not occur independently of temperature. Likewise, investigations into the effects of near-future CO₂ and temperature on different life stages of tropical coral reef species, as well as the capacity of both prey and predators to adapt to ocean acidification over the long-term, are needed to understand key ecological interactions and ultimately how the structure of ecological communities will be affected by climate change variables.

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References

- Baker, D.W., Brauner, C.J., 2012. Metabolic changes associated with acid-base regulation during hypercarbia in the CO₂-tolerant chondrostean, white sturgeon (*Acipenser transmontanus*). *Comp. Biochem. Physiol. A Mol. Integr. Physiol.* 161, 61–68.
- Barnola, J.M., Raynaud, D., Korotkevich, Y.S., Lorius, C., 1987. Vostok ice core provides 160,000-year record of atmospheric CO₂. *Nature* 329, 408–414.
- Bennett, A.F., Hicks, J.W., 2001. Postprandial exercise: prioritization or additivity of the metabolic responses? *J. Exp. Biol.* 204, 2127–2132.
- Boutilier, R.G., Iwama, G.K., Randall, D.J., 1986. The promotion of catecholamine release in rainbow trout, *Salmo gairdneri*, by acute acidosis: interactions between red cell pH and haemoglobin oxygen-carrying capacity. *J. Exp. Biol.* 123, 145–157.
- Brauner, C.J., Baker, D.W., 2009. Patterns of acid-base regulation during exposure to hypercarbia in fishes. In: Glass, M.L., Wood, S.C. (Eds.), *Cardio-Respiratory Control in Vertebrates*. Springer, Berlin, Germany, pp. 43–63.
- Brown, D.J.A., Sadler, K., 1989. Fish survival in acid waters. In: Morris, R., Taylor, E.W., Brown, D.J.A., Brown, J.A. (Eds.), *Acid Toxicity and Aquatic Animals*. Cambridge University Press, Cambridge, UK, pp. 31–44.
- Bushnell, P.G., Steffensen, J.F., Johansen, K., 1984. Oxygen consumption and swimming performance in hypoxia-acclimated rainbow trout *Salmo gairdneri*. *J. Exp. Biol.* 113, 225–235.
- Cahill, A.E., Aiello-Lammens, M.E., Fisher-Reid, M.C., Hua, X., Karanewsky, C.J., Yeong Ryu, H., Sbeglia, G.C., Spagnolo, F., Waldron, J.B., Warsi, O., Wiens, J.J., 2012. How does climate change cause extinction? *Proc. R. Soc. B Biol. Sci.* 280, 20121890.
- Cripps, I.L., Munday, P.L., McCormick, M.I., 2011. Ocean acidification affects prey detection by a predatory reef fish. *PLoS One* 6, e22736.
- Deigweiler, K., Koschnick, N., Portner, H.O., Lucassen, M., 2008. Acclimation of ion regulatory capacities in gills of marine fish under environmental hypercapnia. *Am. J. Physiol. Regul. Integr. Comp. Physiol.* 295, R1660–R1670.
- Dickson, A.G., Millero, F.J., 1987. A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. *Deep-Sea Res. I* 34, 1733–1743.
- Dixson, D.L., Munday, P.L., Jones, G.P., 2010. Ocean acidification disrupts the innate ability of fish to detect predator olfactory cues. *Ecol. Lett.* 13, 68–75.
- Slagersonky, E., Tans, P., 2013. Trends in atmospheric carbon dioxide. <http://www.esrl.noaa.gov/gmd/ccgg/trends/global.html#global>.
- Domenici, P., Allan, B., McCormick, M.I., Munday, P.L., 2012. Elevated carbon dioxide affects behavioural lateralization in a coral reef fish. *Biol. Lett.* 8, 78–81.
- Doney, S.C., 2010. The growing human footprint on coastal and open-ocean biogeochemistry. *Science* 328, 1512–1516.
- Esbau, A., Heuer, R., Grosell, M., 2012. Impacts of ocean acidification on respiratory gas exchange and acid–base balance in a marine teleost, *Opsanus beta*. *J. Comp. Physiol. B Biochem. Syst. Environ. Physiol.* 1–14.
- Fabry, V.J., Seibel, B.A., Feely, R.A., Orr, J.C., 2007. Impacts of ocean acidification on marine fauna and ecosystem processes. *4th International Zooplankton Production Symposium*. Oxford Univ Press, Hiroshima, Japan, pp. 414–432.
- Ferrari, M.C.O., Dixson, D.L., Munday, P.L., McCormick, M.I., Meekan, M.G., Sih, A., Chivers, D.P., 2011a. Intragenetic variation in antipredator responses of coral reef fishes affected by ocean acidification: implications for climate change projections on marine communities. *Glob. Chang. Biol.* 17, 2980–2986.
- Ferrari, M.C.O., McCormick, M.I., Munday, P.L., Meekan, M.G., Dixson, D.L., Lonnstedt, O., Chivers, D.P., 2011b. Putting prey and predator into the CO₂ equation – qualitative and quantitative effects of ocean acidification on predator-prey interactions. *Ecol. Lett.* 14, 1143–1148.
- Fisher, R., Leis, J.M., Clark, D.L., Wilson, S.K., 2005. Critical swimming speeds of late-stage coral reef fish larvae: variation within species, among species and between locations. *Mar. Biol.* 147, 1201–1212.
- Fry, F.E.J., 1947. Effects of the Environment on Animal Activity. Publications of the Ontario Fisheries Research Laboratory, Biological Services, No 55. University of Toronto studies.
- Fry, F.E.J., 1971. The effect of environmental factors on the physiology of fish. In: Hoar, W.S., Randall, D.J. (Eds.), *Environmental Relations and Behavior*. Academic Press, London, pp. 1–98.
- Fry, F.E.J., Hart, J.S., 1948. The relation of temperature to oxygen consumption in the goldfish. *Biol. Bull.* 94, 66–77.
- Gardiner, N.M., Munday, P.L., Nilsson, G.E., 2010. Counter-gradient variation in respiratory performance of coral reef fishes at elevated temperatures. *PLoS One* 5 (e13299) (13291–13213).
- Gattuso, J.-P., Hansson, L., 2011. *Ocean Acidification*. Oxford University Press, London, UK 352.
- Gattuso, J.-P., Lavigne, H., 2009. Technical note: approaches and software tools to investigate the impact of ocean acidification. *Biogeosciences* 6, 2121–2133.
- Gattuso, J.-P., Kunshan, G., Lee, K., Rost, B., Schulz, K.G., 2010. Approaches and tools to manipulate the carbonate chemistry. In: Riebesell, U., Fabry, V.J., Hansson, L., Gattuso, J.-P. (Eds.), *Guide to Best Practices for Ocean Acidification Research and Data Reporting*. Publications Office of the European Union, Luxembourg, pp. 41–52.
- Gilmour, K.M., Perry, S.F., 2006. Branchial chemoreceptor regulation of cardiorespiratory function. In: Toshiaki, J.H., Barbara, S.Z. (Eds.), *Fish Physiology*, 25. Academic Press, pp. 97–151.
- Goolish, E.M., 1991. Aerobic and anaerobic scaling in fish. *Biol. Rev.* 66, 33–56.
- Gran, G., 1950. Determination of the equivalence point in potentiometric titrations. *Acta Chem. Scand.* 4, 559–577.
- Gran, G., 1952. Determination of the equivalence point in potentiometric titrations. Part II. *Analyst* 77, 661–671.
- Hari, P., Pumpanen, J., Huotari, J., Kolari, P., Grace, J., Vesala, T., Ojala, A., 2008. High-frequency measurements of photosynthesis of planktonic algae using rugged nondispersive infrared carbon dioxide probes. *Limnol. Oceanogr. Methods* 6, 347–354.
- Heisler, N., 1993. Acid-base regulation. In: Evans, D.H. (Ed.), *The Physiology of Fishes*. CRC Press Inc, Boca Raton, FL, pp. 343–377.
- Holmes, T.H., McCormick, M.I., 2010. Size-selectivity of predatory reef fish on juvenile prey. *Mar. Ecol. Prog. Ser.* 399, 273–283.
- Ishimatsu, A., Hayashi, M., Kikkawa, T., 2008. Fishes in high-CO₂, acidified oceans. *Mar. Ecol. Prog. Ser.* 373, 295–302.
- Lee, C.G., Farrell, A.P., Lotto, A., MacNutt, M.J., Hinch, S.G., Healey, M.C., 2003. The effect of temperature on swimming performance and oxygen consumption in adult sockeye (*Oncorhynchus nerka*) and coho (*O. kisutch*) salmon stocks. *J. Exp. Biol.* 206, 3239–3251.
- McKenzie, D.J., Piccolella, M., Dalla Valle, A.Z., Taylor, E.W., Bolis, C.L., Steffensen, J.F., 2003. Tolerance of chronic hypercapnia by the European eel *Anguilla anguilla*. *J. Exp. Biol.* 206, 1717–1726.
- Meehl, G.A., Stocker, T.F., Collins, W.D., Friedlingstein, P., Gaye, A.T., Gregory, J.M., Kitoh, A., Knutti, R., Murphy, J.M., Noda, A., Raper, S.C.B., Watterson, I.G., Weaver, A.J., Zhao, Z.-C., 2007. Global climate projections. In: Solomon, S., Qin, D., Manning, M., Chen, Z., Marquis, M., Averyt, K.B., Tignor, M., Miller, H.L. (Eds.), *Climate Change 2007: The Physical Science Basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, UK, pp. 747–846.
- Meekan, M.G., Milicich, M.J., Doherty, P.J., 1993. Larval production drives temporal patterns of larval supply and recruitment of a coral-reef damselfish. *Mar. Ecol. Prog. Ser.* 93, 217–225.
- Meekan, M.M., Wilson, S.W., Halford, A.H., Retzel, A.R., 2001. A comparison of catches of fishes and invertebrates by two light trap designs, in tropical NW Australia. *Mar. Biol.* 139, 373–381.

- Mehrbach, C., Culvers, C.H., Hawley, J.E., Pytkowic, R.M., 1973. Measurement of apparent dissociation-constants of carbonic-acid in seawater at atmospheric pressure. Limnol. Oceanogr. 18, 897–907.
- Meinshausen, M., Smith, S., Calvin, K., Daniel, J., Kainuma, M., Lamarque, J.F., Matsumoto, K., Montzka, S., Raper, S., Riahi, K., Thomson, A., Velders, G., van Vuuren, D.P., 2011. The RCP greenhouse gas concentrations and their extensions from 1765 to 2300. Climatic Change 109, 213–241.
- Melzner, F., Gobel, S., Langenbuch, M., Gutowska, M.A., Portner, H.O., Lucassen, M., 2009a. Swimming performance in Atlantic cod (*Gadus morhua*) following long-term (4–12 months) acclimation to elevated seawater pCO₂. Aquat. Toxicol. 92, 30–37.
- Melzner, F., Gutowska, M.A., Langenbuch, M., Dupont, S., Lucassen, M., Thorndyke, M.C., Bleich, M., Poertner, H.O., 2009b. Physiological basis for high CO₂ tolerance in marine ectothermic animals: pre-adaptation through lifestyle and ontogeny? Biogeosciences 6, 2313–2331.
- Melzner, F., Thomsen, J., Koeve, W., Oschlies, A., Gutowska, M.A., Bange, H.W., Hansen, H.P., Körtzinger, A., 2012. Future ocean acidification will be amplified by hypoxia in coastal habitats. Mar. Biol. <http://dx.doi.org/10.1007/s00227-012-1954-1> (Epub ahead of print).
- Michaelidis, B., Spring, A., Pörtner, H., 2007. Effects of long-term acclimation to environmental hypercapnia on extracellular acid-base status and metabolic capacity in Mediterranean fish *Sparus aurata*. Mar. Biol. 150, 1417–1429.
- Munday, P.L., Wilson, S.K., 1997. Comparative efficacy of clove oil and other chemicals in anaesthetization of *Pomacentrus amboinensis*, a coral reef fish. J. Fish Biol. 51, 931–938.
- Munday, P.L., Crawley, N.E., Nilsson, G.E., 2009a. Interacting effects of elevated temperature and ocean acidification on the aerobic performance of coral reef fishes. Mar. Ecol. Prog. Ser. 388, 235–242.
- Munday, P.L., Donelson, J.M., Dixson, D.L., Endo, G.G.K., 2009b. Effects of ocean acidification on the early life history of a tropical marine fish. Proc. R. Soc. B Biol. Sci. 276, 3275–3283.
- Munday, P.L., Dixson, D.L., McCormick, M.I., Meekan, M., Ferrari, M.C.O., Chivers, D.P., 2010. Replenishment of fish populations is threatened by ocean acidification. Proc. Natl. Acad. Sci. U. S. A. 107, 12930–12934.
- Munday, P.L., McCormick, M.I., Nilsson, G.E., 2012. Impact of global warming and rising CO₂ levels on coral reef fishes: what hope for the future? J. Exp. Biol. 215, 3865–3873.
- Nikinmaa, M., 1986. Control of red cell pH in teleost fishes. Ann. Zool. Fenn. 23, 223–235.
- Nilsson, S., 1986. Control of gill blood flow. In: Nilsson, S., Holmgren, S. (Eds.), Fish Physiology: Recent Advances. Croom Helm, London, pp. 87–101.
- Nilsson, G.E., Östlund-Nilsson, S., 2004. Hypoxia in paradise: widespread hypoxia tolerance in coral reef fishes. Proc. R. Soc. B Biol. Sci. 271, S30–S33.
- Nilsson, G.E., Hobbs, J.-P.A., Östlund-Nilsson, S., 2007a. Tribute to P. L. Lutz: respiratory ecophysiology of coral-reef teleosts. J. Exp. Biol. 210, 1673–1686.
- Nilsson, G.E., Östlund-Nilsson, S., Penfold, R., Grutter, A.S., 2007b. From record performance to hypoxia tolerance: respiratory transition in damselfish larvae settling on a coral reef. Proc. R. Soc. B Biol. Sci. 274, 79–85.
- Nilsson, G.E., Crawley, N., Lunde, I.G., Munday, P.L., 2009. Elevated temperature reduces the respiratory scope of coral reef fishes. Glob. Chang. Biol. 15, 1405–1412.
- Nilsson, G.E., Östlund-Nilsson, S., Munday, P.L., 2010. Effects of elevated temperature on coral reef fishes: loss of hypoxia tolerance and inability to acclimate. Comp. Biochem. Physiol. A Mol. Integr. Physiol. 156, 389–393.
- Nilsson, G.E., Dixson, D.L., Domenici, P., McCormick, M.I., Sorensen, C., Watson, S.-A., Munday, P.L., 2012. Near-future carbon dioxide levels alter fish behaviour by interfering with neurotransmitter function. Nat. Clim. Chang. 2, 201–204.
- Pörtner, H.O., 2008. Ecosystem effects of ocean acidification in times of ocean warming: a physiologist's view. Mar. Ecol. Prog. Ser. 373, 203–217.
- Pörtner, H.O., Farrell, A.P., 2008. Physiology and climate change. Science 322, 690–692.
- Pörtner, H.O., Knust, R., 2007. Climate change affects marine fishes through the oxygen limitation of thermal tolerance. Science 315, 95–97.
- Pörtner, H.O., Langenbuch, M., Reipschläger, A., 2004. Biological impact of elevated ocean CO₂ concentrations: lessons from animal physiology and earth history. J. Oceanogr. 60, 705–718.
- Pörtner, H.O., Langenbuch, M., Michaelidis, B., 2005. Synergistic effects of temperature extremes, hypoxia, and increases in CO₂ on marine animals: from Earth history to global change. J. Geophys. Res. C 110, C09S10.
- Randall, D.J., Baumgarten, D., Malyusz, M., 1972. The relationship between gas and ion transfer across the gills of fishes. Comp. Biochem. Physiol. A Physiol. 41, 629–637.
- Reid, S.G., Sundin, L., Milsom, W.K., 2005. The cardiorespiratory system in tropical fishes: structure, function, and control. Fish Physiol. 21, 225–275.
- Rummer, J.L., Brauner, C.J., 2011. Plasma-accessible carbonic anhydrase at the tissue of a teleost fish may greatly enhance oxygen delivery: *in vitro* evidence in rainbow trout, *Oncorhynchus mykiss*. J. Exp. Biol. 214, 2319–2328.
- Rummer, J.L., McKenzie, D.J., Innocenti, A., Supuran, C.T., Brauner, C.J., 2013. Root effect hemoglobin may have evolved to enhance general tissue oxygen delivery. Science 340, 1327–1329.
- Seibel, B.A., Walsh, P.J., 2001. Carbon cycle - potential impacts of CO₂ injection on deep-sea biota. Science 294, 319–320.
- Shaw, E.C., McNeil, B.I., Tilbrook, B., Matear, R., Bates, M.L., 2013. Anthropogenic changes to seawater buffer capacity combined with natural reef metabolism induce extreme future coral reef CO₂ conditions. Glob. Chang. Biol. <http://dx.doi.org/10.1111/gcb.12154> (Epub ahead of print).
- Simpson, S.D., Munday, P.L., Wittenrich, M.L., Manassa, R., Dixson, D.L., Gagliano, M., Yan, H.Y., 2013. Ocean acidification erodes crucial auditory behaviour in a marine fish. Biol. Lett. 7, 917–920.
- Smith, H., 1965. Some experiments on the oxygen consumption of goldfish (*Carassius auratus* L.) in relation to swimming speed. Can. J. Zool.-Rev. Can. Zool. 43, 623–633.
- Solomon, S., Qin, D., Manning, M., Chen, Z., Marquis, M., Averyt, K.B., Tignor, M., Miller, H.L., 2007. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA 996.
- Steffensen, J.F., 1989. Some errors in respirometry of aquatic breathers: how to avoid and correct for them? Fish Physiol. Biochem. 6, 49–59.
- Torres, J.J., Childress, J.J., 1983. Relationship of oxygen consumption to swimming speed in *Euphausia pacifica* – 1. Effects of temperature and pressure. Mar. Biol. 74, 79–86.