Supporting information for:

A high-resolution spatial model to predict exposure to pharmaceuticals in European surface

waters - ePiE

Authors

Rik Oldenkamp*1,2, Selwyn Hoeks1, Mirza Čengić1, Valerio Barbarossa1, Emily E. Burns2,

Alistair B.A. Boxall², Ad M.J. Ragas^{1,3}

¹Department of Environmental Science, Radboud University Nijmegen, 6500GL, Nijmegen, The Netherlands

²Environment Department, University of York, Heslington, York YO10 5DD, United Kingdom

³Faculty of Management, Science & Technology, Open Universiteit, Valkenburgerweg 177, 6419 AT Heerlen, The

Netherlands

* Corresponding author, at r.oldenkamp@science.ru.nl

Number of pages: 47

Number of figures: 3

Number of tables: 5

S1. Curation of UWWTD-Waterbase

The UWWTD-Waterbase database¹ contains spatial information on the location and characteristics of 27,695 European agglomerations, and 30,043 urban WWTPs and waste water collection systems. Agglomerations with a generated load below 2,000 p.e. (population equivalents) are not included in the database, since the directive does not require reporting on them.

The agglomeration characteristics reported in the database relevant for the ePiE model are their location (longitude and latitude coordinates), their generated load (p.e.), and the fraction connected to WWTP (-). Similarly, the UWWTD database contains for each WWTP the location (longitude and latitude coordinates), the load entering (p.e.), as well as its design capacity (p.e.). It must be noted that some WWTPs receive a combination of urban and industrial wastewater, of which the database only reports the combined load. Consequently, API loads towards such WWTPs might deviate from the estimations. In addition, each WWTP has been assigned an identification number related to the agglomeration connected to it, and the level of treatment is indicated.

Although the UWWTD-Waterbase database is extensive, it does contain some erroneous, missing, or ambiguous entries. Therefore, a curation of the database was performed. First, both agglomerations and WWTPs with missing coordinates or mixed up latitude and longitude were identified. This was the case for 1.2% of the agglomerations in the database. A few of these (16 records) had mixed up latitude and longitude, which was corrected. For another 13 agglomerations no coordinates were reported because their generated load was below 2,000 p.e. (population equivalents). These were excluded from further calculations. The remaining 292 agglomerations without correct coordinates (1.1% of the initial set) were also excluded

from further calculations. As a result, coordinates were available for 27,381 agglomerations in the database. From these, 799 were indicated in the database as 'inactive', generally due to merging of agglomerations into one record that were separately included in a previous version of the database, or because agglomerations dropped below the 2,000 p.e. size threshold for reporting. These agglomerations were also excluded from further calculations, resulting in a total of 26,582 agglomerations used in further calculations.

Of the wastewater treatment plants (WWTPs) in the database, 9.4% had missing or erroneous coordinates. Of these records, a few (11 records) had mixed up latitude and longitude, which were corrected. The exclusion of all WWTPs indicated in the database as being inactive reduced the database to 27,432 WWTPs and increased the percentage with correct coordinates to 94.2%. Of the remaining 5.8% of WWTP-records without correct coordinates (1,596 records), 784 were assigned coordinates of the agglomeration to which they were linked via the agglomeration identification number. Although this likely has led to a loss of accuracy in the estimation of the emission point for these WWTPs (location of WWTP now equals location of agglomeration linked to it), it was considered preferable over their exclusion. This resulted in a database with 26,620 WWTP-records with coordinates. The remaining 818 WWTP-records were removed from the database and excluded from further calculations. These were mainly Southern Italian, wastewater collecting systems without treatment connected to them. This resulted in 26,614 WWTP-records with coordinates in the database.

In principle, each WWTP included in the database is assigned an agglomeration identification number, ensuring that it is linked to an agglomeration in the database. In practice, however, only 12,765 of the 26,614 WWTP-records in the database (after curation for missing and

erroneous coordinates) had been assigned such an agglomeration identification number, leaving 13,849 without. In order to make a WWTP-based approach possible, the agglomerations connected to these WWTPs had to be deduced based on similarities in names, codes, coordinates, or loads generated. By doing this, we managed to increase the number of WWTPs with an agglomeration ID attached to it to 25,697. The remaining WWTPs were mainly located in Southern Italy. The 1,735 WWTP records that were excluded from the total database (818 due to missing/erroneous coordinates; 917 due to lack of agglomeration ID), represent ~3% of the total wastewater load entering all WWTPs in Europe. These WWTPs were included in the spatial distribution of the total consumption over agglomerations and WWTPs but were excluded from further concentration calculations.

After linking each WWTP to an agglomeration, 3,961 agglomerations were not yet connected to any WWTP-record. For 2,762 of these agglomerations, this is due to their 0% connectivity to wastewater collection and WWTPs. Instead, their wastewater might be (partly) addressed in independent appropriate systems (IAS), and might be (partly) discharged untreated. While the database does distinguish between these two options, we conservatively assumed direct discharge without treatment for all wastewater not directed towards a WWTP. From the remaining 1,199 of the agglomerations not connected to any WWTP, 562 could be linked to a WWTP serving multiple agglomerations (to which another agglomeration had already been linked). The final 637 agglomerations could not be linked to any WWTP. These were located in Southern Italy (428) and Croatia (209). For the Croatian records, this was due to the lack of information on their WWTP-connectivity (i.e., all Croatian agglomerations reported zero connectivity). Without knowing their WWTP-connectivity, it was not possible to determine whether they were actually connected to a WWTP. Furthermore, it should be noted that the

database does not contain any WWTPs in the vicinity of the Italian city of Napoli. Finally, a set of 26,607 unique combinations of agglomeration and WWTP was constructed.

S2. Model construction

River networks in ePiE were constructed for individual drainage basins as delineated by the global database HydroSHEDS.² Per basin, its borders were used to crop the river network, available as spatial line object at 30 arcseconds from the HydroSHEDS database, which were then translated into a binary 30 arcseconds raster with information on the presence or absence a river element. Next, the raster cells containing river elements were used to create a spatial point file, with each point representing a network node. All nodes were automatically classified as either:

- 1. junction (node where two streams meet, has two upstream nodes);
- 2. mouth (node where river flows into the sea);
- 3. start (nodes representing a river source);
- 4. regular node.

Via overlay of the river network with the lakes and reservoirs in the global database HydroLAKES³, available as spatial polygons, nodes located within a lake or reservoir were identified. These nodes are skipped during model computations. Additionally, intersections between the river network and lakes and reservoirs were determined, and were added as additional nodes to the network, classified as either:

- 1. outlet (the one intersection per lake/reservoir closest to the river mouth);
- 2. inlet (all other intersections)

Properties associated to the lakes and reservoirs in HydroLAKES, i.e., depth and hydraulic retention time, were added to their corresponding outlet node.

After the curation steps described in SI1, the WWTPs and agglomerations from the UWWTD-Waterbase database¹ were snapped to the river network and incorporated as emission sources. Any WWTPs and agglomerations located within lakes or reservoirs were allocated to these as direct emission source. Direct emissions into the sea were excluded from the model. The up- and downstream node for all WWTP and agglomeration specific nodes was determined using its relative placement in the river network.

Using the flow direction raster available from HydroSHEDS, the hydrological interconnectivity of the nodes in the network was determined, including distances between them. Finally, gridded information on air temperature, wind speed, slope, and streamflow was extracted to all nodes in the network.

S3. Loss processes

Extrapolation of degradation rates

The ePiE model accounts for three degradation processes, i.e. biodegradation, photolysis and hydrolysis. It allows for the correction of individual degradation rate constants for differences between temperature under test conditions (T_{test} ; K) and under field conditions (T_{field} ; K), using a correction factor f_{temp} (-), based on Arrhenius' temperature dependence of reaction rates (Equation S3.1). When georeferenced water temperatures are not available, ePiE assumes a default value for T_{field} of 285 K.8

$$f_{temp} = 2^{\frac{\left(T_{field} - T_{test}\right)}{10}}$$
 Equation S3.1

Additionally, degradation rate constants are corrected for sorption to suspended solids and dissolved organic carbon, assuming degradation processes only apply to the dissolved fraction of the API (f_{diss} ; -):

$$f_{diss} = \frac{1}{1 + K_{p,susp} \cdot C_{susp} + K_{p,DOC} \cdot C_{DOC}}$$
 Equation S3.2

In Equation S3.2, sorption to suspended solids is characterized with the chemical's suspended solids-water partition coefficient $K_{p,susp}$ (L/kg) and the local concentration of suspended solids C_{susp} (kg/L). Similarly, sorption to dissolved organic carbon is characterized with the chemical's dissolved organic carbon-water partition coefficient $K_{p,DOC}$ (L/kg) and the local concentration of dissolved organic carbon C_{DOC} (kg/L). ePiE uses default values of $0.015*10^{-3}$ kg/L⁹ and $0.005*10^{-3}$ kg/L¹⁰ for C_{susp} and C_{DOC} , respectively, when geographical information on these parameters is not available.

Moreover, because values for $K_{p,susp}$ and $K_{p,DOC}$ are generally not available for many APIs, ePiE estimates them under the assumption that sorption is mainly driven by the organic carbon content. Values for $K_{p,DOC}$ are directly estimated from their octanol-water partition coefficient ($K_{OW,n}$; -) and acid dissociation constant (pK_a ; -), according to Burkhard. Values for $K_{p,susp}$ are based on a default value of 0.1 for the organic carbon content of suspended solids or ($f_{OC,susp}$)¹². For nonionizing chemicals, the relationship derived by Sabljić et al. 13 between $K_{OW,n}$ and the organic carbon-water partitioning coefficient ($K_{OC,n}$; -) is then used to estimate $K_{p,susp}$:

$$K_{p,susp} = K_{OC,n} \cdot f_{OC,susp} = 1.26 \cdot K_{OW,n}^{0.81} \cdot f_{OC,susp}$$
 Equation S3.3

For ionizing chemicals, estimation of $K_{p,susp}$ is more complex because their neutral and ionized forms show different sorption behaviour (Equation S3.4). Therefore, K_{OC} values are

first estimated for the individual ionization forms using quantitative structure-activity relationships (QSAR) as derived by Franco and Trapp.¹⁴ Subsequently, these are combined based on the mass fractions of the individual forms (f_n , f_{cat} , f_{an} ; -), determined via Henderson-Hasselbalch equations:

$$K_{p,susp} = (f_n \cdot K_{OC,n} + f_{cat} \cdot K_{OC,cat} + f_{an} \cdot K_{OC,an}) \cdot f_{OC,susp}$$
 Equation S3.4

Because experimental photolysis rates are determined at water surface and at constant light, ePiE corrects them for reduced light intensity at local field conditions with correction factor f_{light} (-) (Equation S3.5).¹⁵ Light intensity depends on the time fraction of light per day (f_{day} ; -), for which we used a default value of 0.5, and reduces with local water depth (h_w ; m) and turbidity. The latter is expressed as the ratio between the light's average path length through the water column and the water depth (D_{λ} ; -), which we assigned a proposed default value for non-turbid waters of 1.2.¹⁵ Finally, penetration of light through the water column differs for different wavelengths, as reflected by the wavelength-specific beam attenuation coefficient α_{λ} (cm⁻¹). The relationship between wavelength and α_{λ} as provided in Table 13.6 in Schwarzenbach et al.¹⁵ was used to assign a value to α_{λ} for individual APIs based on their respective maximum absorption wavelengths λ_{max} (nm).

$$f_{light} = f_{day} \cdot \frac{1 - 10^{-D} \lambda^{\cdot \alpha} \lambda^{\cdot 100 \cdot h_W}}{\ln(10) \cdot D_{\lambda^{\cdot \alpha} \lambda^{\cdot 100 \cdot h_W}}}$$
 Equation S3.5

Intermedia transport rates

The ePiE model accounts for two intermedia transport processes, i.e. sedimentation and volatilization. Concentrations in sediments or air cannot be explicitly calculated with single-media models like ePiE, since they do not include separate compartments for these media.

Therefore, transport between media was directly accounted for via estimation of mass transport velocities.¹⁶

Sedimentation rates are derived by combining adsorption, desorption, sedimentation and resuspension velocities of suspended particulate into one equation describing the local rate of sedimentation k_{sed} (s⁻¹) (Equation S3.6). For estimation of k_{sed} , the sediment phase was treated as a homogeneous phase consisting of a water sub-phase and a solid sub-phase. Equilibrium was assumed between the pore water and solid sub-phases of the sediment phase, and the top layer of the sediment h_{sd} (m), with a default value of 0.03 m,¹² was considered to be well-mixed.

$$k_{sed} = \left(\frac{v_{ads} + v_{sed}}{h_w}\right) - \frac{\left(\frac{v_{ads} + v_{sed}}{h_w}\right) * \left(\frac{v_{res} + v_{des}}{h_{sd}}\right)}{\left(\frac{v_{res} + v_{des} + v_{sed,acc}}{h_{sd}} + k_{bio,sed}\right)}$$
Equation S3.6

in which k_{sed} is the local sedimentation rate (s⁻¹), and individual local velocities for adsorption, desorption, sedimentation, and resuspension (all expressed as m/s) are represented by v_{ads} (Equation S3.8), v_{des} (Equation S3.9), v_{sed} (Equation S3.10), and v_{res} , respectively (Equation S3.11). Additionally, k_{sed} depends on the net sediment accumulation rate in water $v_{sed,acc}$ (m/s), for which ePiE applies a default value $8.6*10^{-11}$ m/s, the same value that is used in the USEtox 2.0 model under default conditions. Moreover, k_{sed} is inversely related to the biodegradation of the chemical that might take place in the sediment layer ($k_{bio,sed}$; s⁻¹), since reduction of chemical in the sediment layer due to degradation will increase the mass flux towards the sediment layer to restore the equilibrium between water and sediment. In the absence of an experimental value for $k_{bio,sed}$, it is estimated from the biodegradation rate constant in surface water ($k_{bio,w}$; s⁻¹), with an extrapolation factor of 0.1, 17 a correction factor

for temperature (f_{temp} ; Equation S3.1), and a correction factor for sorption ($f_{sed,diss}$; -) as proposed by Honti et al.¹⁸:

$$f_{sed,diss} = \frac{1}{1 + Kp_{tot,sed} \cdot \rho_{s,sed} \cdot \frac{(1 - \theta_{sed})}{\theta_{sed}}}$$
 Equation S3.7

in which $Kp_{tot,sed}$ is the partitioning coefficient between water and sediment solids (L_{water}/kg_{solids}) ; $\rho_{s,sed}$ is the local mineral density of the sediment solids (kg_{solids}/L_{solids}) ; and θ_{sed} is the local porosity of the sediment $(L_{water}/L_{sediment})$. For $\rho_{s,sed}$ and θ_{sed} , ePiE applies default values of 2.33 $kg_{solids}/L_{solids}^{12, 18}$ and 0.8 $L_{water}/L_{sediment}^{19}$, respectively.

$$v_{ads} = \frac{v_{m,w,w|sd} \cdot v_{m,sd,w|sd}}{v_{m,w,w|sd} + v_{m,sd,w|sd}} \cdot f_{diss}$$
 Equation S3.8

$$v_{des} = \frac{v_{m,w,w|sd} \cdot v_{m,sd,w|sd}}{v_{m,w,w|sd} + v_{m,sd,w|sd}} \cdot \frac{1}{K_{sd|w}}$$
Equation S3.9

Adsorption velocities v_{ads} and desorption velocities v_{des} , are both estimated via partial mass transfer coefficients at the water/sediment interface, with $v_{m,w,w|sd}$ (m/s) representing the coefficient at the water side and $v_{m,sd,w|sd}$ (m/s) the coefficient at the sediment side. As proposed by Mackay 20 , $v_{m,w,w|sd}$ and $v_{m,sd,w|sd}$ were assigned default values of 2.778*10⁻⁶ m/s and 2.778*10⁻⁸ m/s. The affinity of the chemical for sorption to the sediment is reflected in Equation S3.9 by the dimensionless sediment/water partition coefficient $K_{sd|w}$ (-).

$$v_{sed} = f_{V_{solid,sd}} \cdot v_{sed,gross} \cdot \rho_{sd} \cdot K_{p,susp} \cdot f_{diss}$$
 Equation S3.10
$$v_{res} = v_{sed,gross} - v_{sed,acc}$$
 Equation S3.11

Local sedimentation velocities v_{sed} and resuspension velocities v_{res} both depend on the gross local sedimentation rate from water $v_{sed,aross}$ (m/s), which is derived as follows:

If
$$v_{sed,susp} \cdot \frac{c_{susp}}{\rho_{sd,bulk}} > v_{sed,acc}$$
 then:

$$v_{sed,gross} = v_{sed,susp} \cdot \frac{c_{susp}}{\rho_{sd,bulk}}$$
 else: Equation S3.12

 $v_{sed,gross} = v_{sed,acc}$

in which $\rho_{sd,bulk}$ is the local bulk density of the sediment (kg/L; see Equation S3.13), and $v_{sed,susp}$ is the settling velocity of suspended particles (m/s), for which we applied a default value of $2.89*10^{-5}$ m/s.²¹ Moreover, $v_{sed,acc}$ is the aforementioned net sediment accumulation rate in water (see Equation S3.6). Equations S3.10 through S3.12 imply that, if the sedimentation of particles from the water column is greater than the resuspension (i.e., there is a net sedimentation), the top layer is continuously refreshed. The older sediment layer, and with it the chemicals that are associated with the sediment, then gets buried under the freshly deposited material.

$$\rho_{sd,bulk} = 1 - f_{V_{solid} sd} + f_{V_{solid} sd} \cdot \rho_{sd}$$
 Equation S3.13

In which the volume fraction solids in sediment is represented by $f_{V_{solid,sd}}$ (-), and the mineral density of the sediment by ρ_{sd} (kg/L). These were assigned default values of 0.2¹⁹ and 2.1633 kg/L¹², respectively.

Same as for sediments, concentrations in air are not explicitly calculated by ePiE, since it does not include a separate air compartment. Therefore, the intermedia volatilization rate between water and air k_{vol} (s⁻¹) was directly accounted for in Equation S3.14 via estimation of the volatilization velocity (Equation S3.15):

$$k_{vol} = rac{v_{vol}}{h_w}$$
 Equation S3.14

$$v_{vol} = \frac{v_{m,a,a|w} \cdot v_{m,w,a|w}}{v_{m,a,a|w} \cdot K_{a|w} + v_{m,w,a|w}} \cdot K_{a|w} \cdot f_{diss}$$
 Equation S3.15

The volatilization velocity v_{vol} (m/s) was estimated via local partial mass transfer coefficients at the air/water interface, with $v_{m,a,a|w}$ (m/s) representing the coefficient at the air side (Equation S3.16) and $v_{m,a,a|w}$ (m/s) the coefficient at the water side (Equation S3.17). Moreover, $K_{a|w}$ represents the dimensionless chemical-specific air/water partition coefficient, estimated via Equation S3.18.

$$v_{m,a,a|w} = 0.01 \cdot (0.3 + 0.2 \cdot v_{wind}) \cdot \frac{18}{MW}^{(0.67 \cdot 0.5)}$$
 Equation S3.16

$$v_{m,w,a|w} = 0.01 \cdot (0.0004 + 0.00004 \cdot v_{wind}^{2}) \cdot \frac{32}{MW}^{(0.5 \cdot 0.5)}$$
 Equation S3.17

$$K_{a|w} = \frac{P_{v} \cdot MW}{S_{w} \cdot R \cdot T_{air}}$$
 Equation S3.18

In these equations, environmental parameters v_{wind} and T_{air} represent the local wind speed (m/s) and air temperature (K), respectively. Physicochemical properties required are molecular weight MW (g/mol), vapour pressure P_v (Pa), and water solubility S_w (mg/L). R is the universal gas constant (8.314 Pa/m³/mol/K).

S4. Chemical model parameterization

Table S4.1. Physicochemical parameterisation of APIs.

API	CAS RN	MW (g/mol)	Ref	Pv (Pa)	Ref	S (mg/L)	Ref	Class ^a	pK _a (-)	Ref	K _{OW,n} (-)	Ref
Acetaminophen	103-90-2	151.16	22	2.59*10-4	23	1.40*10 ⁴	22	acid	9.46	22	1.86	23
Amitriptyline	28981-97-7	277.40	22	4.83*10 ⁻⁵	23	9.71	22	base	9.76	22	8.89*104	23
Atenolol	29122-68-7	266.34	22	1.03*10 ⁻⁷	23	1.33*10 ⁴	22	base	9.60	22	9.42*10 ⁻¹	23
Bezafibrate	41859–67–0	361.82	22	8.15*10-9	23	1.22	23	acid	3.83	22	1.78*104	23
Carbamazepine	298-46-4	236.27	22	1.17*10 ⁻⁵	23	1.77*10 ¹	22	neutral	NA	22	1.77*10 ²	23
Cimetidine	51481-61-9	252.34	22	1.85*10 ⁻⁷	23	9.38*10 ³	22	base	6.80	22	3.75	23
Citalopram	59729-33-8	324.39	22	1.51*10 ⁻⁵	23	3.11*10 ¹	23	base	9.78	22	5.52*10 ³	23
Codeine	76-57-3	299.36	22	2.55*10 ⁻⁸	23	9.00*10 ³	23	base	9.19	22	1.89*10 ¹	23
Desvenlafaxine	93413-62-8	263.38	22	9.13*10 ⁻⁶	23	3.67*10 ³	23	base	8.87	22	5.25*10 ²	23
Diazepam	439–14–5	284.74	22	1.36*10 ⁻⁵	23	5.00*10 ¹	22	base	3.40	22	5.01*10 ²	23
Diclofenac	15307-86-5	296.15	22	8.19*10 ⁻⁶	23	2.37	22	acid	4.15	22	3.24*10 ⁴	23
Diltiazem	42399-41-7	414.52	22	3.97*10 ⁻⁹	23	4.65*10 ²	22	base	8.18	22	6.23*10 ²	23
Erythromycin	114-07-8	733.93	22	2.83*10 ⁻²³	23	2.00*10 ³	22	base	8.38	22	3.01*10 ²	23
Gabapentin	60142-96-3	171.24	22	3.92*10 ⁻⁸	23	4.49*10 ³	22	acid	4.63	22	4.27*10 ⁻²	23
Hydrocodone	125-29-1	299.36	22	1.61*10 ⁻⁵	23	1.79*10 ³	23	base	8.61	22	1.46*10 ²	23
Ibuprofen	15687-27-1	206.28	22	2.48*10-2	23	2.10*10 ¹	22	acid	4.85	22	9.33*10 ³	23
Indomethacin	53-86-1	357.79	22	6.82*10 ⁻⁸	23	9.37*10 ⁻¹	22	acid	3.8	22	1.86*104	23
Lidocaine	137–58–6	234.34	22	9.01*10-4	23	7.97*10 ⁻¹	22	base	7.75	22	4.56*10 ¹	23
Loratadine	79794–75–5	382.88	22	1.21*10-6	23	4.10*10 ³	22	neutral	NA	22	4.54*10 ⁵	23
Metformin	657–24–9	129.16	22	1.01*10-2	23	1.00*10 ⁶	23	base	12.4	22	2.29*10 ⁻³	23
Naproxen	22204-53-1	230.26	22	1.70*10-4	23	1.59*10 ¹	22	acid	4.15	22	1.51*10 ³	23
Norethindrone	68–22–4	298.42	22	3.15*10 ⁻⁷	23	7.04	22	neutral	NA	22	9.74*10 ²	23
Oseltamivir	196618-13-0	312.40	22	1.75*10 ⁻⁶	23	6.86*10 ⁻¹	22	acid	7.7	22	8.91	23
Oxazepam	604-75-1	286.71	22	8.04*10 ⁻¹⁰	23	1.79*10 ²	22	neutral	NA	22	2.10*10 ²	23
Propranolol	525-66-6	259.34	22	1.26*10 ⁻⁵	23	6.17*10 ¹	22	base	9.67	22	3.96*10 ²	23

Ranitidine	66357–35–5	314.40	22	1.86*10-4	23	2.47*10 ¹	22	base	8.08	22	1.97	23
Sertraline	79617–96–2	306.23	22	1.56*10-4	23	3.50	22	base	9.85	22	1.93*10 ⁵	23
Sitagliptin	486460–32–6	407.31	22	3.88*10-6	23	3.40*10-2	22	base	8.78	22	2.45*10 ¹	23
Sulfamethoxazole	723–46–6	253.28	22	1.73*10-5	23	6.10*10 ²	22	acid	6.16	22	3.05	23
Temazepam	846-50-4	300.74	22	2.27*10 ⁻⁸	23	1.64*10 ²	22	neutral	NA	22	1.41*10 ²	23
Tramadol	27203-92-5	263.38	22	6.09*10 ⁻⁵	23	1.15*10 ³	23	base	9.23	22	1.03*10 ³	23
Triamterene	396-01-0	253.26	22	1.45*10 ⁻⁹	23	4.82*10 ¹	22	neutral	NA	22	6.29	23
Trimethoprim	738–70–5	290.32	22	1.00*10-6	23	4.00*10 ²	22	base	7.12	22	5.36	23
Venlafaxine	93413-69-5	277.40	22	3.28*10 ⁻⁵	23	5.72*10 ⁵	22	base	8.91	22	1.91*10 ³	23
Verapamil	52-53-9	454.60	22	5.57*10 ⁻⁷	23	4.47	22	base	9.68	22	6.26*10 ⁴	23

MW: molecular weight; Pv: vapour pressure; S: water solubility; pK_a : acid dissociation constant; $K_{OW,n}$: octanol-water partition coefficient of the neutral form. ^a Classification as neutral/acid/base was based on ionization states at pH 7 of strongest acid and strongest base pKa values.

Table S4.2. Environmental fate parameterisation of APIs.

API	CAS RN	f _{pc} (-)	Ref	f _{met} (-)	Ref	K _{p,ps} (L/kg)	n	Ref	K _{p,as} (L/kg)	n	Ref	K _{p,sed} (L/kg)	n	Ref
Acetaminophen	103-90-2	0.52	24	NA	-	5.11*10 ¹	6	25-28	3.15*10 ²	5	25, 27-30	3.27	7	31-34
Amitriptyline	28981-97-7	0.03	35	NA	-	1.19*10 ⁴	3	28, 36, 37	3.46*10 ³	15	28, 36-39	NA	0	-
Atenolol	29122-68-7	0.86	24	NA	-	2.00*10 ²	3	27, 28, 36	4.13*10 ²	9	27, 28, 36, 38, 40-43	7.82	11	31, 33, 34, 44-46
Bezafibrate	41859–67–0	0.63	24	NA	-	1.79*10 ¹	2	25	8.70*10 ¹	1	47	3.18*10 ¹	3	48
Carbamazepine	298–46–4	0.15	24	NA	-	1.16*10 ²	13	25-28, 37, 49-51	5.89*10 ²	29	25-30, 37, 39-41, 47, 49, 51-54	1.24*10 ¹	14	31-34, 55-58
Cimetidine	51481–61–9	0.48	59	NA	-	NA	0	-	3.60*10 ²	6	39	1.43*10 ²	6	33, 58
Citalopram	59729–33–8	0.97	24	NA	-	1.28*10 ⁴	2	36, 37	2.56*10 ³	7	36-38, 41	1.37*10 ⁴	2	60
Codeine	76-57-3	0.40	61	NA	-	NA	0	-	1.40*10 ¹	2	30, 40	8.10	2	62
Desvenlafaxine	93413-62-8	0.65	63	0.55	64	1.40*10 ²	1	37	1.01*10 ²	4	37	NA	0	-
Diazepam	439–14–5	0.11	24	NA	-	1.68*10 ²	2	28, 51	1.42*10 ²	11	28, 39, 40, 51	7.43	4	45, 55, 62
Diclofenac	15307-86-5	0.06	24	NA	-	2.42*10 ²	8	25, 27, 50, 51, 65, 66	2.16*10 ²	13	27, 29, 36, 38, 39, 51- 53, 65-69	2.94*10 ¹	9	48, 58
Diltiazem	42399–41–7	0.03	35	NA	-	2.19*10 ²	1	26	1.96*10 ²	6	26, 30, 36, 41	4.33*10 ²	3	33, 70
Erythromycin	114-07-8	0.98	24	NA	-	1.74*10 ²	4	25, 27, 71	9.49*10 ¹	5	25, 27, 38, 71	5.32*10 ²	3	72-74
Gabapentin	60142-96-3	1.00	35	NA	-	NA	0	-	NA	0	-	NA	0	-
Hydrocodone	125-29-1	0.12	35	0.11	75	NA	0	-	1.19*10 ²	5	39	NA	0	-
Ibuprofen	15687-27-1	0.20	24	NA	-	1.08*10 ¹	3	25, 27	2.00*10 ²	9	36, 39, 51-53, 68, 69	1.50*10 ¹	15	34, 48, 57, 58
Indomethacin	53-86-1	0.20	76	NA	-	NA	0	-	1.26*10 ²	2	38, 52	2.81	3	34
Lidocaine	137–58–6	0.08	24	NA	-	NA	0	-	3.30*10 ¹	1	43	NA	0	-
Loratadine	79794–75–5	0.20	77	NA	-	2.34*10 ³	1	27	3.32*10 ³	1	27	NA	0	-
Metformin	657–24–9	1.00	63	NA	-	NA	0	-	1.62*10 ¹	2	29, 30	NA	0	-
Naproxen	22204–53–1	0.70	63	NA	-	1.26*10 ¹	1	66	6.03*10 ¹	7	38, 39, 52, 53, 66, 68, 69	1.73	7	31, 45, 57, 58
Norethindrone	68-22-4	0.05	35	NA	-	5.15*10 ¹	2	50	NA	0	-	1.28*10 ²	1	78
Oseltamivir	196618-13-0	0.25	24	NA	-	NA	0	-	9.96	2	79	NA	0	-
Oxazepam	604-75-1	1.00	24	0.03	63	7.90*10 ²	1	36	9.91*10 ²	4	36, 40, 41	9.23	3	55, 62

Propranolol	525–66–6	0.26	24	NA	-	3.92*10 ³	4	26, 27	6.11*10 ²	14	26, 27, 38, 40-43, 49, 80	1.78*10 ²	10	33, 34, 44, 56, 70, 81
Ranitidine	66357–35–5	0.38	24	NA	-	NA	0	-	4.25*10 ²	2	30, 82	NA	0	-
Sertraline	79617–96–2	0.14	63	NA	-	3.00*104	2	36, 37	1.97*104	7	36-38, 41	1.71*10 ²	2	60
Sitagliptin	486460-32-6	0.79	35	NA	-	NA	0	-	NA	0	-	NA	0	-
Sulfamethoxazole	723–46–6	0.35	24	NA	-	5.82*10 ¹	7	25, 27, 28, 36, 50, 71	1.74*10 ²	21	25-28, 30, 36, 38, 39, 47, 53, 71, 82, 83	5.06*10 ¹	10	31, 62, 72, 74, 84, 85
Temazepam	846-50-4	0.05	35	0.07	86	NA	0	-	NA	0	-	1.66*10 ¹	2	62
Tramadol	27203-92-5	0.27	61	NA	-	1.10*10 ²	1	36	1.19*10 ²	2	36, 40	5.05	2	62
Triamterene	396-01-0	0.21	35	NA	-	NA	0	-	NA	0	-	NA	0	-
Trimethoprim	738–70–5	0.50	63	NA	-	2.42*10 ²	6	25, 27, 28, 36, 71	2.09*10 ²	22	25-28, 30, 36, 38, 39, 47, 71, 83, 87	1.21*10 ³	8	45, 58, 70, 74, 88
Venlafaxine	93413-69-5	0.05	35	NA	-	1.50*10 ³	1	37	2.42*10 ²	7	36, 37, 41, 43	NA	0	-
Verapamil	52-53-9	0.04	35	NA	-	1.72*10 ³	2	28, 36	3.15*10 ³	5	28, 36, 41	1.70*10 ²	1	58

API	k _{bio,wwtp} (s ⁻¹)	n	Ref	k _{bio,sw} (s ⁻¹)	n	Ref	k _{photo} (s ⁻¹)	n	Ref	λ _{max} (nm)	Ref	k _{hyd,sw} (s ⁻¹)	n	Ref
Acetaminophen	5.01*10-4	4	30, 89-91	8.36*10 ⁻⁶	3	34, 92, 93	4.55*10 ⁻⁶	5	34, 92-94	250	94	0.00	3	92-94
Amitriptyline	NA	0	-	0.00	1	93	0.00	1	93	270	95	2.35*10 ⁻⁷	2	93, 95
Atenolol	5.37*10 ⁻⁵	5	40, 42, 43, 96, 97	5.42*10 ⁻⁷	2	34, 93	3.33*10 ⁻⁶	14	34, 93, 98-106	275	95, 100, 101, 107	7.25*10 ⁻⁷	7	93, 95, 101-103, 106
Bezafibrate	5.95*10 ⁻⁵	2	47, 90	NA	0	-	NA	0	-	NA	-	NA	0	-
Carbamazepine	1.75*10 ⁻⁶	11	30, 40, 52, 54, 68, 82, 89, 108- 110	2.31*10 ⁻⁸	5	34, 92, 93, 111, 112	9.48*10 ⁻⁶	20	34, 92, 93, 99, 111, 113-120	285	95, 114, 115, 118	1.67*10 ⁻⁹	5	92, 93, 95, 111, 114
Cimetidine	NA	0	-	0.00	2	121	1.74*10-4	2	105, 122	218	122	0.00	1	122
Citalopram	5.56*10 ⁻⁵	1	110	NA	0	-	3.88*10 ⁻⁷	2	60	285	95	8.59*10 ⁻⁸	2	60, 95
Codeine	2.28*10 ⁻⁵	2	30, 40	2.38*10 ⁻⁷	1	123	6.83*10 ⁻⁵	2	105, 123	300	124	0.00	1	123
Desvenlafaxine	NA	0	-	1.57*10 ⁻⁸	3	112, 125, 126	8.96*10 ⁻⁶	1	126	275	126	4.06*10 ⁻⁷	1	126
Diazepam	1.85*10-5	1	68	3.85*10 ⁻¹¹	1	127	4.67*10 ⁻⁶	3	128-130	315	128, 130	4.06*10 ⁻⁷	3	127, 128, 130
Diclofenac	1.41*10 ⁻⁵	9	47, 52, 59, 68, 89, 90, 109, 110	8.28*10 ⁻⁸	6	70, 93, 131-134	4.84*10-4	7	70, 93, 113, 132, 133, 135-137	275	133, 137-139	1.16*10-8	4	70, 133, 134, 140
Diltiazem	8.15*10 ⁻⁶	3	30	1.08*10-6	2	121	2.06*10 ⁻⁵	1	98	280	98	1.43*10 ⁻⁶	1	141

Erythromycin	4.63*10 ⁻⁵	4	54, 68, 90, 110, 142	8.02*10 ⁻⁸	1	143	4.51*10 ⁻⁶	2	105, 144	214	144	NA	0	-
Gabapentin	1.15*10-6	1	91	8.02*10-8	1	145	2.00*10-5	1	106	276	146	1.33*10 ⁻⁵	1	106
Hydrocodone	NA	0	-	4.34*10 ⁻⁷	2	121	NA	0	-	NA	-	NA	0	-
Ibuprofen	1.97*10 ⁻⁴	8	47, 52, 59, 68, 90, 109, 110, 147	8.27*10 ⁻⁷	4	34, 93, 134, 147	3.85*10 ⁻⁶	6	34, 93, 119, 120, 148, 149	265	119, 150	2.31*10 ⁻⁷	2	134, 148
Indomethacin	4.73*10 ⁻⁵	2	52, 90	2.69*10 ⁻⁷	3	34, 93, 94	2.46*10 ⁻⁵	3	34, 93, 94	265	94	0.00	1	94
Lidocaine	1.62*10 ⁻⁷	1	43	7.22*10 ⁻⁸	1	126	4.36*10 ⁻⁶	1	126	260	126	4.06*10 ⁻⁷	1	126
Loratadine	NA	0	-	NA	0	-	NA	0	-	NA	-	NA	0	-
Metformin	1.05*10-4	2	30, 151	8.02*10 ⁻⁸	1	152	3.33*10 ⁻⁷	1	106	232	153	1.50*10 ⁻⁵	1	106
Naproxen	1.63*10 ⁻⁵	9	47, 52, 59, 68, 90, 108	7.72*10 ⁻⁷	2	93, 134	1.15*10-4	3	93, 148, 149	260	149	5.02*10 ⁻⁸	3	134, 148
Norethindrone	NA	0	-	2.55*10 ⁻⁷	1	131	NA	0	-	NA	-	0.00	1	131
Oseltamivir	0.00	1	79	5.02*10-8	3	79, 154	8.08*10 ⁻⁶	2	155, 156	225	156	6.19*10 ⁻⁸	3	95, 155, 156
Oxazepam	0.00	1	40	NA	0	-	7.82*10 ⁻⁶	2	128, 130	315	95, 128, 130	0.00	3	95, 128, 130
Propranolol	1.83*10 ⁻⁵	4	40, 42, 43, 80	1.24*10 ⁻⁵	3	34, 93, 111	3.91*10 ⁻⁵	14	34, 70, 93, 99- 101, 103, 105, 111, 113, 149, 157	288	100, 101	0.00	3	93, 111, 121
Ranitidine	2.55*10 ⁻⁵	3	30, 82, 97	6.80*10-8	2	121	4.55*10 ⁻⁵	1	122	315	122, 158	0.00	1	122
Sertraline	NA	0	-	0.00	1	92	1.23*10 ⁻⁶	2	92, 129	205	159	0.00	1	92
Sitagliptin	NA	0	-	NA	0	-	NA	0	-	NA	-	NA	0	-
Sulfamethoxazole	2.63*10 ⁻⁵	10	30, 68, 82, 89, 90, 110, 142, 160, 161	8.90*10 ⁻⁶	6	85, 92, 93, 111, 132, 162	1.11*10-4	15	92, 99, 108, 111, 113, 115, 144, 163-166	274	115, 144, 165	4.95*10 ⁻⁸	5	85, 92, 95, 111, 131
Temazepam	NA	0	-	NA	0	-	5.39*10 ⁻⁶	1	130	315	130	0.00	1	130
Tramadol	0.00	1	40	NA	0	-	4.33*10 ⁻⁴	2	126, 167	271	126, 167	1.47*10 ⁻⁷	3	95, 126, 167
Triamterene	NA	0	-	NA	0	-	NA	0	-	NA	-	NA	0	-
Trimethoprim	2.53*10 ⁻⁶	7	54, 68, 87, 110, 142, 161	0.00	2	92, 93	7.85*10 ⁻⁶	5	92, 105, 108, 166, 168	283	95	2.70*10 ⁻⁸	4	92, 93, 95, 169
Venlafaxine	8.09*10 ⁻⁸	2	43, 147	1.94*10 ⁻⁸	3	125, 126, 147	3.41*10 ⁻⁶	1	126	275	95, 126	2.10*10 ⁻⁷	2	95, 126
Verapamil	NA	0	-	NA	0	-	NA	0	-	NA	-	NA	0	-

 f_{pc} : fraction of administered parent compound excreted/egested unchanged or as reversible conjugates via urine and feces; f_{met} : fraction of prodrug metabolized to the API of interest and excreted/egested via urine and feces; $K_{p,ps}$: sorption coefficient to primary sewage; $K_{p,as}$: sorption coefficient to activated sludge; $K_{p,sed}$: sorption coefficient to sediment; $k_{bio,wwtp}$: (pseudo-)first order biodegradation rate in WWTP; $k_{bio,sw}$: (pseudo-)first order biodegradation rate in surface water; k_{photo} : first order photolysis rate in surface water; k_{max} : maximum absorption wavelength; $k_{hyd,sw}$: first order hydrolysis rate in surface water.

S5. Consumption data

Table S5.1. Consumption data per API for each country and year as included in the model evaluation (kg/yr).

	Austria				Belgium				France			
API	2011	2013	2014	Ref	2011	2013	2014	Ref	2011	2013	2014	Ref
Acetaminophen	1.58*10 ⁵ a	-	-	-	2.08*10 ⁵ a	-	-	-	3.45*10 ^{6 b}	-	-	170
Amitriptyline	7.00*10 ² a	-	-	-	9.20*10 ² a	-	-	-	5.43*10 ³ a	-	-	-
Atenolol	1.39*10 ^{3 a}	1.41*10 ^{3 a}	1.42*10 ^{3 a}	-	1.83*10 ³ a	1.85*10 ^{3 a}	1.86*10 ³ a	-	1.91*10 ^{4 b}	1.93*10 ^{4 b}	1.94*10 ^{4 b}	170
Bezafibrate	1.28*10 ^{3 a}	-	-	-	1.68*10 ³ a	-	-	-	2.18*10 ^{4 b}	-	-	170
Carbamazepine	4.40*10 ^{3 a}	4.39*10 ³ a	4.39*10 ³ a	-	5.78*10 ^{3 a}	5.78*10 ^{3 a}	5.77*10 ^{3 a}	-	3.50*10 ^{4 b}	3.53*10 ^{4 b}	3.55*10 ^{4 b}	170
Cimetidine	-	-	-	-	-	-	-	-	-	-	-	-
Citalopram	5.52*10 ² a	5.62*10 ² a	5.69*10 ² a	-	7.25*10 ^{2 a}	7.40*10 ^{2 a}	7.47*10 ^{2 a}	-	3.64*10 ^{3 b}	3.67*10 ^{3 b}	3.69*10 ^{3 b}	170
Codeine	-	-	-	-	-	-	-	-	-	-	-	-
Desvenlafaxine	-	-	-	-	-	-	-	-	-	-	-	-
Diazepam	4.91*10 ^{1 a}	-	-	-	6.45*10 ¹ a	-	-	-	5.49*10 ^{2 b}	-	-	170
Diclofenac	4.67*10 ^{3 a}	-	-	-	6.14*10 ^{3 a}	-	-	-	1.03*10 ^{4 b}	-	-	170
Diltiazem	-	1.64*10 ^{3 a}	1.56*10 ^{3 a}	-	-	2.16*10 ^{3 a}	2.05*10 ^{3 a}	-	-	1.27*10 ⁴ a	1.21*10 ⁴ a	-
Erythromycin	4.31*10 ² a	4.26*10 ² a	4.21*10 ² a	-	5.66*10 ² a	5.61*10 ² a	5.53*10 ^{2 a}	-	3.34*10 ³ a	3.30*10 ³ a	3.26*10 ³ a	-
Gabapentin	4.39*10 ^{3 a}	4.47*10 ^{3 a}	4.53*10 ³ a	-	5.76*10 ^{3 a}	5.89*10 ^{3 a}	5.96*10 ^{3 a}	-	3.40*10 ⁴ a	3.47*10 ⁴ a	3.51*10 ⁴ a	-
Hydrocodone	-	-	-	-	-	-	-	-	-	-	-	-
Ibuprofen	3.32*10 ^{4 a}	-	-	-	4.36*10 ⁴ a	-	-	-	2.50*10 ^{5 b}	-	-	170
Indomethacin	1.05*10 ² a	-	-	-	1.38*10 ² a	-	-	-	1.04*10 ^{3 b}	-	-	170
Lidocaine	4.54*10 ^{3 a}	-	-	-	5.96*10 ^{3 a}	-	-	-	3.52*10 ⁴ a	-	-	-
Loratadine	-	-	-	-	-	-	-	-	-	-	-	-
Metformin	1.05*10 ⁵ a	-	-	-	1.38*10 ⁵ a	-	-	-	7.48*10 ^{5 b}	-	-	170
Naproxen	3.36*10 ^{3 a}	-	-	-	4.41*10 ^{3 a}	-	-	-	3.89*10 ^{4 b}	-	-	170
Norethindrone	-	-	-	-	-	-	-	-	-	-	-	-

	Germany				Luxembourg	;			Netherland	s		
Verapamil	1.74*10 ^{3 a}	-	-	-	2.28*10 ^{3 a}	-	-	-	1.35*10 ^{4 a}	-	-	-
Venlafaxine	1.32*10 ^{3 a}	1.35*10 ^{3 a}	1.37*10 ^{3 a}	-	1.73*10 ^{3 a}	1.79*10 ³ a	1.80*10 ³ a	-	1.02*10 ^{4 b}	1.05*10 ^{4 b}	1.06*10 ^{4 b}	170
Trimethoprim	4.26*10 ^{2 a}	4.24*10 ^{2 a}	4.22*10 ² a	-	5.60*10 ² a	5.59*10 ² a	5.55*10 ² a	-	3.49*10 ^{3 b}	3.52*10 ^{3 b}	3.54*10 ^{3 b}	170
Triamterene	-	-	-	-	-	-	-	-	-	-	-	-
Tramadol	2.43*10 ^{3 a}	-	-	-	3.19*10 ^{3 a}	-	-	-	2.70*10 ^{4 b}	-	-	170
Temazepam	-	-	-	-	-	-	-	-	-	-	-	-
Sulfamethoxazole	2.28*10 ^{3 a}	2.18*10 ^{3 a}	2.17*10 ^{3 a}	-	3.00*10 ^{3 a}	2.88*10 ^{3 a}	2.85*10 ^{3 a}	-	1.75*10 ^{4 b}	1.76*10 ^{4 b}	1.77*10 ^{4 b}	170
Sitagliptin	3.66*10 ^{2 a}	4.19*10 ² a	4.10*10 ² a	-	4.81*10 ^{2 a}	5.52*10 ^{2 a}	5.39*10 ² a	-	2.84*10 ^{3 a}	3.25*10 ^{3 a}	3.18*10 ^{3 a}	-
Sertraline	-	-	-	-	-	-	-	-	-	-	-	-
Ranitidine	2.04*10 ^{3 a}	-	-	-	2.68*10 ^{3 a}	-	-	-	1.22*10 ^{4 b}	-	-	170
Propranolol	8.60*10 ^{2 a}	8.69*10 ² a	8.71*10 ^{2 a}	-	1.13*10 ^{3 a}	1.14*10 ^{3 a}	1.14*10 ^{3 a}	-	1.30*10 ^{4 b}	1.32*10 ^{4 b}	1.32*10 ^{4 b}	170
Oxazepam	-	-	-	-	-	-	-	-	-	-	-	-
Oseltamivir	-	-	-	-	-	-	-	-	-	-	-	-

	Germany				Luxembourg	5			Netherland	s		
API	2011	2013	2014	Ref	2011	2013	2014	Ref	2011	2013	2014	Ref
Acetaminophen	5.52*10 ^{5 c}	-	-	170	9.68*10 ^{3 a}	-	-	-	5.52*10 ⁵	-	-	171
Amitriptyline	7.34*10 ^{3 c}	-	-	170	4.28*10 ^{1 a}	-	-	-	1.26*10 ³	-	-	171
Atenolol	7.32*10 ^{3 c}	7.35*10 ^{3 c}	7.37*10 ^{3 c}	170	8.52*10 ^{1 a}	9.11*10 ¹ a	9.33*10 ¹ a	-	2.50*10 ³	2.20*10 ³	2.05*10 ³	171
Bezafibrate	1.51*10 ^{4 c}	-	-	170	7.83*10 ^{1 a}	-	-	-	3.96*10 ²	-	-	171
Carbamazepine	6.29*10 ^{4 c}	6.31*10 ^{4 c}	6.33*10 ^{4 c}	170	2.69*10 ^{2 a}	2.79*10 ² a	2.83*10 ² a	-	7.56*10 ³	7.23*10 ³	7.01*10 ³	171
Cimetidine	-	-	-	-	-	-	-	-	-	-	-	-
Citalopram	4.89*10 ^{3 c}	4.91*10 ^{3 c}	4.92*10 ^{3 c}	170	3.37*10 ^{1 a}	3.57*10 ¹ a	3.67*10 ¹ a	-	8.20*10 ²	8.64*10 ²	8.90*10 ²	171
Codeine	-	-	-	-	-	-	-	-	-	-	-	-
Desvenlafaxine	-	-	-	-	-	-	-	-	-	-	-	-
Diazepam	4.89*10 ^{2 c}	-	-	170	3.00 a	-	-	-	9.44*10 ¹	-	-	171
Diclofenac	8.96*10 ^{4 c}	-	-	170	2.86*10 ^{2 a}	-	-	-	5.13*10 ³	-	-	171

Diltiazem	-	4.91*10 ^{3 c}	4.92*10 ^{3 c}	-	-	1.04*10 ² a	1.01*10 ² a	-	-	3.25*10 ³	3.08*10 ³	171
Erythromycin	9.23*10 ^{3 c}	9.26*10 ^{3 c}	9.29*10 ^{3 c}	170	2.63*10 ¹ a	2.71*10 ¹ a	2.72*10 ¹ a	-	4.38*10 ²	3.87*10 ²	3.41*10 ²	171
Gabapentin	6.85*10 ^{4 c}	6.88*10 ^{4 c}	6.90*10 ^{4 c}	170	2.68*10 ² a	2.84*10 ² a	2.93*10 ² a	-	6.55*10 ³	6.83*10 ³	7.05*10 ³	171
Hydrocodone	-	-	-	-	-	-	-	-	-	-	-	-
Ibuprofen	6.08*10 ^{5 c}	-	-	170	2.03*10 ³ a	-	-	-	1.97*10 ⁴	-	-	171
Indomethacin	9.78*10 ^{2 c}	-	-	170	6.44 a	-	-	-	1.44*10 ²	-	-	171
Lidocaine	7.34*10 ^{3 c}	-	-	170	2.77*10 ^{2 a}	-	-	-	2.33*104	-	-	171
Loratadine	-	-	-	-	-	-	-	-	-	-	-	-
Metformin	9.78*10 ^{5 c}	-	-	170	6.43*10 ³ a	-	-	-	2.80*10 ⁵	-	-	171
Naproxen	1.44*10 ^{4 c}	-	-	170	2.05*10 ² a	-	-	-	1.16*104	-	-	171
Norethindrone	-	-	-	-	-	-	-	-	-	-	-	-
Oseltamivir	-	-	-	-	-	-	-	-	-	-	-	-
Oxazepam	-	-	-	-	-	-	-	-	-	-	-	-
Propranolol	2.45*10 ^{3 c}	2.45*10 ^{3 c}	2.46*10 ^{3 c}	170	5.26*10 ^{1 a}	5.52*10 ¹ a	5.63*10 ¹ a	-	1.38*10 ³	1.39*10 ³	1.37*10 ³	171
Ranitidine	2.45*10 ^{4 c}	-	-	170	1.25*10 ² a	-	-	-	5.87*10 ³	-	-	171
Sertraline	-	-	-	-	-	-	-	-	-	-	-	-
Sitagliptin	4.89*10 ^{3 c}	4.91*10 ^{3 c}	4.92*10 ^{3 c}	170	2.24*10 ^{1 a}	2.66*10 ^{1 a}	2.65*10 ^{1 a}	-	4.41*10 ²	6.40*10 ²	5.95*10 ²	171
Sulfamethoxazole	3.42*10 ^{4 c}	3.43*10 ^{4 c}	3.44*10 ^{4 c}	170	1.39*10 ² a	1.39*10 ² a	1.40*10 ² a	-	4.42*10 ³	3.49*10 ³	3.26*10 ³	171
Temazepam	-	-	-	-	-	-	-	-	-	-	-	-
Tramadol	3.28*10 ^{4 c}	-	-	170	1.48*10 ² a	-	-	-	3.40*10 ³	-	-	171
Triamterene	-	-	-	-	-	-	-	-	-	-	-	-
Trimethoprim	7.31*10 ^{3 c}	7.34*10 ^{3 c}	7.36*10 ^{3 c}	170	2.60*10 ^{1 a}	2.70*10 ¹ a	2.73*10 ¹ a	-	4.37*10 ²	3.95*10 ²	3.58*10 ²	171
Venlafaxine	1.17*10 ^{4 c}	1.17*10 ^{4 c}	1.17*10 ^{4 c}	170	8.04*10 ^{1 a}	8.61*10 ¹ a	8.84*10 ¹ a	-	3.27*10 ³	3.45*10 ³	3.49*10 ³	171
Verapamil	2.45*10 ^{4 c}	-	-	170	1.06*10 ^{2 a}	-	-	-	3.12*10 ³	-	-	171

	Switzerland				United Kingd	om	
API	2011	2013	2014	Ref	2016	Ref	

Acataminanhan	1.02*10 ^{5 c}			170	2.52*10 ^{6 d}	172
Acetaminophen		-	<u>-</u>	170		172
Amitriptyline	6.58*10 ² a	-	-		1.28*10 ^{4 d}	
Atenolol	1.02*10 ^{3 c}	1.05*10 ^{3 c}	1.06*10 ^{3 c}	170	1.63*10 ^{4 d}	172
Bezafibrate	5.11*10 ^{2 c}	-	-	170	-	-
Carbamazepine	2.55*10 ^{3 c}	2.62*10 ^{3 c}	2.64*10 ^{3 c}	170	3.65*10 ^{4 d}	172
Cimetidine	-	-	-	-	1.11*10 ^{3 d}	172
Citalopram	7.66*10 ^{2 c}	7.85*10 ^{2 c}	7.93*10 ^{2 c}	170	1.04*10 ^{4 d}	172
Codeine	-	-	-	-	5.62*10 ^{4 d}	172
Desvenlafaxine	-	-	-	-	0.00	172
Diazepam	2.55*10 ^{1 c}	-	-	170	6.38*10 ^{2 d}	172
Diclofenac	5.11*10 ^{3 c}	-	-	170	-	-
Diltiazem	-	1.56*10 ^{3 a}	1.49*10 ³ a	-	2.06*10 ^{4 d}	172
Erythromycin	1.02*10 ^{2 c}	1.05*10 ^{2 c}	1.06*10 ^{2 c}	170	1.88*10 ^{4 d}	172
Gabapentin	2.55*10 ^{3 c}	2.62*10 ^{3 c}	2.64*10 ^{3 c}	170	1.60*10 ^{5 d}	172
Hydrocodone	-	-	-	-	9.42*10 ^{3 d}	172
Ibuprofen	2.55*10 ^{4 c}	-	-	170	-	-
Indomethacin	1.02*10 ^{2 c}	-	-	170	-	-
Lidocaine	1.02*10 ^{3 c}	-	-	170	8.15*10 ^{3 d}	172
Loratadine	-	-	-	-	8.58*10 ^{2 d}	172
Metformin	7.66*10 ^{4 c}	-	-	170	1.15*10 ^{6 d}	172
Naproxen	1.02*10 ^{3 c}	-	-	170	-	-
Norethindrone	-	-	-	-	1.66*10 ^{2 d}	172
Oseltamivir	-	-	-	-	6.66*10 ^{1 d}	172
Oxazepam	-	-	-	-	8.04 ^d	172
Propranolol	7.66*10 ^{2 c}	7.85*10 ^{2 c}	7.93*10 ^{2 c}	170	1.24*10 ^{4 d}	172
Ranitidine	1.02*10 ^{3 c}	-	-	170	4.63*10 ^{4 d}	172
Sertraline	-	-	-	-	2.88*10 ^{4 d}	172

Sitagliptin	3.44*10 ² a	3.99*10 ^{2 a}	3.92*10 ² a	-	1.91*10 ^{4 d}	172
Sulfamethoxazole	1.02*10 ^{3 c}	1.05*10 ^{3 c}	1.06*10 ^{3 c}	170	2.17*10 ^{3 d}	172
Temazepam	-	-	-	-	4.81*10 ^{2 d}	172
Tramadol	1.02*10 ^{3 c}	-	-	170	4.55*10 ^{4 d}	172
Triamterene	-	-	-	-	1.65*10 ^{1 d}	172
Trimethoprim	2.55*10 ^{2 c}	2.61*10 ^{2 c}	2.64*10 ^{2 c}	170	7.34*10 ^{3 d}	172
Venlafaxine	1.02*10 ^{3 c}	1.05*10 ^{3 c}	1.06*10 ^{3 c}	170	1.56*10 ^{4 d}	172
Verapamil	1.02*10 ^{3 c}	-	-	170	5.80*10 ^{3 d}	172

^a: average per capita consumption from other countries in basin, for which consumption data were available, extrapolated to country and year of interest based on demographics from Eurostat ¹⁷³; ^b: per capita consumption in 2004 ¹⁷⁰ extrapolated to year of interest based on demographics from Eurostat ¹⁷³; ^c: per capita consumption in 2009 ¹⁷⁰ extrapolated to year of interest based on demographics from Eurostat ¹⁷³; ^d: per capita consumption in England ¹⁷² extrapolated to whole of United Kingdom based on demographics from Office for National Statistics ¹⁷⁴.

Table S5.2. Consumption data per API for each country and year as included in the model evaluation (g/capita).

	Austria					Belgium				France			
API	2011	2013	2014	Ref	2011	2013	2014	Ref	2011	2013	2014	Ref	
Acetaminophen	1.89*10 ¹ a	-	-	-	1.89*10 ¹ a	-	-	-	5.30*10 ¹ b	-	-	170	
Amitriptyline	8.36*10 ^{-2 a}	-	-	-	8.36*10 ^{-2 a}	-	-	-	8.36*10 ^{-2 a}	-	-	-	
Atenolol	1.66*10 ^{-1 a}	1.67*10 ^{-1 a}	1.67*10 ^{-1 a}	-	1.66*10 ^{-1 a}	1.66*10 ^{-1 a}	1.66*10 ^{-1 a}	-	2.94*10 ^{-1 b}	2.94*10 ^{-1 b}	2.94*10 ^{-1 b}	170	
Bezafibrate	1.53*10 ^{-1 a}	-	-	-	1.53*10 ^{-1 a}	-	-	-	3.35*10 ^{-1 b}	-	-	170	
Carbamazepine	5.25*10 ^{-1 a}	5.19*10 ^{-1 a}	5.16*10 ^{-1 a}	-	5.25*10 ^{-1 a}	5.19*10 ^{-1 a}	5.16*10 ^{-1 a}	-	5.38*10 ^{-1 b}	5.38*10 ^{-1 b}	5.38*10 ^{-1 b}	170	
Cimetidine	-	-	-	-	-	-	-	-	-	-	-	-	
Citalopram	6.59*10 ^{-2 a}	6.65*10 ^{-2 a}	6.68*10 ^{-2 a}	-	6.59*10 ^{-2 a}	6.65*10 ^{-2 a}	6.68*10 ^{-2 a}	-	5.60*10 ^{-2 b}	5.60*10 ^{-2 b}	5.60*10 ^{-2 b}	170	
Codeine	-	-	-	-	-	-	-	-	-	-	-	-	
Desvenlafaxine	-	-	-	-	-	-	-	-	-	-	-	-	
Diazepam	5.86*10 ^{-3 a}	-	-	-	5.86*10 ^{-3 a}	-	-	-	8.44*10 ^{-3 b}	-	-	170	
Diclofenac	5.58*10 ^{-1 a}	-	-	-	5.58*10 ^{-1 a}	-	-	-	1.59*10 ^{-1 b}	-	-	170	
Diltiazem	-	1.94*10 ^{-1 a}	1.83*10 ^{-1 a}	-	-	1.94*10 ^{-1 a}	1.83*10 ^{-1 a}	-	-	1.94*10 ^{-1 a}	1.83*10 ^{-1 a}	-	
Erythromycin	5.15*10 ^{-2 a}	5.04*10 ^{-2 a}	4.94*10 ^{-2 a}	-	5.15*10 ^{-2 a}	5.04*10 ^{-2 a}	4.94*10 ^{-2 a}	-	5.15*10 ^{-2 a}	5.04*10 ^{-2 a}	4.94*10 ^{-2 a}	-	
Gabapentin	5.24*10 ^{-1 a}	5.29*10 ^{-1 a}	5.33*10 ^{-1 a}	-	5.24*10 ^{-1 a}	5.29*10 ^{-1 a}	5.33*10 ^{-1 a}	-	5.24*10 ^{-1 a}	5.29*10 ^{-1 a}	5.33*10 ^{-1 a}	-	
Hydrocodone	-	-	-	-	-	-	-	-	-	-	-	-	
Ibuprofen	3.96*10 ⁰ a	-	-	-	3.96*10 ⁰ a	-	-	-	3.85*10 ^{0 b}	-	-	170	
Indomethacin	1.26*10 ^{-2 a}	-	-	-	1.26*10 ^{-2 a}	-	-	-	1.61*10 ^{-2 b}	-	-	170	
Lidocaine	5.42*10 ^{-1 a}	-	-	-	5.42*10 ^{-1 a}	-	-	-	5.42*10 ^{-1 a}	-	-	-	
Loratadine	-	-	-	-	-	-	-	-	-	-	-	-	
Metformin	1.26*10 ^{1 a}	-	-	-	1.26*10 ¹ a	-	-	-	1.15*10 ¹ b	-	-	170	
Naproxen	4.01*10 ^{-1 a}	-	-	-	4.01*10 ^{-1 a}	-	-	-	5.99*10 ^{-1 b}	-	-	170	
Norethindrone	-	-	-	-	-	-	-	-	-	-	-	-	
Oseltamivir	-	-	-	-	-	-	-	-	-	-	-	-	

Oxazepam	-	-	-	-	-	-	-	-	-	-	-	-
Propranolol	1.03*10 ⁻¹ a	1.03*10 ⁻¹ a	1.02*10 ^{-1 a}	-	1.03*10 ^{-1 a}	1.03*10 ⁻¹ a	1.02*10 ⁻¹ a	-	2.00*10 ^{-1 b}	2.00*10 ^{-1 b}	2.00*10 ^{-1 b}	170
Ranitidine	2.44*10 ^{-1 a}	-	-	-	2.44*10 ^{-1 a}	-	-	-	1.87*10 ^{-1 b}	-	-	170
Sertraline	-	-	-	-	-	-	-	-	-	-	-	-
Sitagliptin	4.37*10 ^{-2 a}	4.96*10 ^{-2 a}	4.82*10 ^{-2 a}	-	4.37*10 ^{-2 a}	4.96*10 ^{-2 a}	4.82*10 ^{-2 a}	-	4.37*10 ^{-2 a}	4.96*10 ^{-2 a}	4.82*10-2 a	-
Sulfamethoxazole	2.72*10 ^{-1 a}	2.58*10 ^{-1 a}	2.55*10 ^{-1 a}	-	2.72*10 ^{-1 a}	2.58*10 ^{-1 a}	2.55*10 ^{-1 a}	-	2.69*10 ^{-1 b}	2.69*10 ^{-1 b}	2.69*10 ^{-1 b}	170
Temazepam	-	-	-	-	-	-	-	-	-	-	-	-
Tramadol	2.90*10 ⁻¹ a	-	-	-	2.90*10 ^{-1 a}	-	-	-	4.16*10 ^{-1 b}	-	-	170
Triamterene	-	-	-	-	-	-	-	-	-	-	-	-
Trimethoprim	5.09*10 ^{-2 a}	5.02*10 ^{-2 a}	4.97*10 ^{-2 a}	-	5.09*10 ^{-2 a}	5.02*10 ^{-2 a}	4.97*10 ^{-2 a}	-	5.37*10 ^{-2 b}	5.37*10 ^{-2 b}	5.37*10 ^{-2 b}	170
Venlafaxine	1.57*10 ^{-1 a}	1.60*10 ^{-1 a}	1.61*10 ^{-1 a}	-	1.57*10 ^{-1 a}	1.60*10 ^{-1 a}	1.61*10 ^{-1 a}	-	1.57*10 ^{-1 b}	1.60*10 ^{-1 b}	1.61*10 ^{-1 b}	170
Verapamil	2.07*10 ^{-1 a}	-	-	-	2.07*10 ^{-1 a}	-	-	-	2.07*10 ^{-1 a}	-	-	-

Germany					Luxembourg				Netherlands			
API	2011	2013	2014	Ref	2011	2013	2014	Ref	2011	2013	2014	Ref
Acetaminophen	6.89*10 ^{0 c}	-	-	170	1.89*10 ¹ a	-	-	-	2.73*10°	-	-	171
Amitriptyline	9.15*10 ^{-2 c}	-	-	170	8.36*10 ^{-2 a}	-	-	-	7.58*10-2	-	-	171
Atenolol	9.12*10 ^{-2 c}	9.12*10 ^{-2 c}	9.12*10 ^{-2 c}	170	1.66*10 ^{-1 a}	1.70*10 ^{-1 a}	1.70*10 ^{-1 a}	-	1.50*10 ⁻¹	1.31*10 ⁻¹	1.22*10-1	171
Bezafibrate	1.88*10 ^{-1 c}	-	-	170	1.53*10 ^{-1 a}	-	-	-	2.37*10-2	-	-	171
Carbamazepine	7.84*10 ^{-1 c}	7.84*10 ^{-1 c}	7.84*10 ^{-1 c}	170	5.25*10 ^{-1 a}	5.19*10 ^{-1 a}	5.16*10 ^{-1 a}	-	4.54*10 ⁻¹	4.31*10 ⁻¹	4.16*10 ⁻¹	171
Cimetidine	-	-	-	-	-	-	-	-	-	-	-	-
Citalopram	6.10*10 ^{-2 c}	6.10*10 ^{-2 c}	6.10*10 ^{-2 c}	170	6.59*10 ^{-2 a}	6.65*10 ^{-2 a}	6.68*10 ^{-2 a}	-	4.92*10-2	5.15*10-2	5.29*10-2	171
Codeine	-	-	-	-	-	-	-	-	-	-	-	-
Desvenlafaxine	-	-	-	-	-	-	-	-	-	-	-	-
Diazepam	6.10*10 ^{-3 c}	-	-	170	5.86*10 ^{-3 a}	-	-	-	5.67*10 ⁻³	-	-	171
Diclofenac	1.12*10 ^{0 c}	-	-	170	5.58*10 ^{-1 a}	-	-	-	3.08*10 ⁻¹	-	-	171
Diltiazem	-	6.10*10 ^{-2 c}	6.10*10 ^{-2 c}	-	-	1.94*10 ^{-1 a}	1.83*10 ^{-1 a}	-	-	1.94*10 ⁻¹	1.83*10-1	171

Erythromycin	1.15*10 ^{-1 c}	1.15*10 ^{-1 c}	1.15*10 ^{-1 c}	170	5.15*10 ^{-2 a}	5.04*10 ^{-2 a}	4.94*10 ^{-2 a}	-	2.63*10-2	2.31*10-2	2.03*10-2	171
Gabapentin	8.54*10 ^{-1 c}	8.54*10 ^{-1 c}	8.54*10 ^{-1 c}	170	5.24*10 ^{-1 a}	5.29*10 ^{-1 a}	5.33*10 ^{-1 a}	-	3.93*10 ⁻¹	4.07*10 ⁻¹	4.19*10 ⁻¹	171
Hydrocodone	-	-	-	-	-	-	-	-	-	-	-	-
Ibuprofen	7.58*10 ^{0 c}	-	-	170	3.96*10 ⁰ a	-	-	-	1.18*100	-	-	171
Indomethacin	1.22*10 ^{-2 c}	-	-	170	1.26*10 ^{-2 a}	-	-	-	8.66*10-3	-	-	17:
Lidocaine	9.15*10 ^{-2 c}	-	-	170	5.42*10 ^{-1 a}	-	-	-	1.40*100	-	-	17
Loratadine	-	-	-	-	-	-	-	-	-	-	-	-
Metformin	1.22*10 ^{1 c}	-	-	170	1.26*10 ¹ a	-	-	-	1.68*10 ¹	-	-	17
Naproxen	1.80*10 ^{-1 c}	-	-	170	4.01*10 ^{-1 a}	-	-	-	6.95*10 ⁻¹	-	-	17:
Norethindrone	-	-	-	-	-	-	-	-	-	-	-	-
Oseltamivir	-	-	-	-	-	-	-	-	-	-	-	-
Oxazepam	-	-	-	-	-	-	-	-	-	-	-	-
Propranolol	3.05*10 ^{-2 c}	3.05*10 ^{-2 c}	3.05*10 ^{-2 c}	170	1.03*10 ⁻¹ a	1.03*10 ⁻¹ a	1.02*10 ⁻¹ a	-	8.26*10-2	8.29*10-2	8.12*10-2	17
Ranitidine	3.05*10 ^{-1 c}	-	-	170	2.44*10 ^{-1 a}	-	-	-	3.52*10 ⁻¹	-	-	17
Sertraline	-	-	-	-	-	-	-	-	-	-	-	-
Sitagliptin	6.10*10 ^{-2 c}	6.10*10 ^{-2 c}	6.10*10 ^{-2 c}	170	4.37*10 ^{-2 a}	4.96*10 ^{-2 a}	4.82*10 ⁻² a	-	2.65*10-2	3.81*10-2	3.54*10 ⁻²	17
Sulfamethoxazole	4.26*10 ^{-1 c}	4.26*10 ^{-1 c}	4.26*10 ^{-1 c}	170	2.72*10 ^{-1 a}	2.58*10 ^{-1 a}	2.55*10 ^{-1 a}	-	2.65*10 ⁻¹	2.08*10 ⁻¹	1.94*10 ⁻¹	17
Temazepam	-	-	-	-	-	-	-	-	-	-	-	-
Tramadol	4.08*10 ^{-1 c}	-	-	170	2.90*10 ^{-1 a}	-	-	-	2.04*10 ⁻¹	-	-	17
Triamterene	-	-	-	-	-	-	-	-	-	-	-	-
Trimethoprim	9.12*10 ^{-2 c}	9.12*10 ^{-2 c}	9.12*10 ^{-2 c}	170	5.09*10 ^{-2 a}	5.02*10 ^{-2 a}	4.97*10 ^{-2 a}	-	2.62*10-2	2.35*10-2	2.13*10-2	17
Venlafaxine	1.45*10 ^{-1 c}	1.45*10 ^{-1 c}	1.45*10 ^{-1 c}	170	1.57*10 ^{-1 a}	1.60*10 ^{-1 a}	1.61*10 ^{-1 a}	-	1.96*10-1	2.06*10 ⁻¹	2.07*10 ⁻¹	17
Verapamil	3.05*10 ^{-1 c}	-	-	170	2.07*10 ^{-1 a}	-	-	-	1.87*10-1	-	-	17

	Switzerland		United Kingdom			
API	2011	2013	2014	Ref	2016	Ref
Acetaminophen	1.30*10 ^{1 c}	-	-	170	3.86*10 ^{1 d}	172

Amitriptyline	8.36*10 ^{-2 a}	-	-	170	1.96*10 ^{-1 d}	172
Atenolol	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	170	2.50*10 ^{-1 d}	172
Bezafibrate	6.49*10 ^{-2 c}	-	-	170	-	-
Carbamazepine	3.25*10 ^{-1 c}	3.25*10 ^{-1 c}	3.25*10 ^{-1 c}	170	5.58*10 ^{-1 d}	172
Cimetidine	-	-	-	-	1.70*10 ^{-2 d}	172
Citalopram	9.74*10 ^{-2 c}	9.76*10 ^{-2 c}	9.74*10 ^{-2 c}	170	1.59*10 ^{-1 d}	172
Codeine	-	-	-	-	8.59*10 ^{-1 d}	172
Desvenlafaxine	-	-	-	-	0.00*10 ^{0 d}	172
Diazepam	3.25*10 ^{-3 c}	-	-	170	9.76*10 ^{-3 d}	172
Diclofenac	6.49*10 ^{-1 c}	-	-	170	-	-
Diltiazem	-	1.94*10 ^{-1 a}	1.83*10 ^{-1 a}	-	3.15*10 ^{-1 d}	172
Erythromycin	1.30*10 ^{-2 c}	1.30*10 ^{-2 c}	1.30*10 ^{-2 c}	170	2.87*10 ^{-1 d}	172
Gabapentin	3.25*10 ^{-1 c}	3.25*10 ^{-1 c}	3.25*10 ^{-1 c}	170	2.45*10 ^{0 d}	172
Hydrocodone	-	-	-	-	1.44*10 ^{-1 d}	172
Ibuprofen	3.25*10 ^{0 c}	-	-	170	-	-
Indomethacin	1.30*10 ^{-2 c}	-	-	170	-	-
Lidocaine	1.30*10 ^{-1 c}	-	-	170	1.25*10 ^{-1 d}	172
Loratadine	-	-	-	-	1.31*10 ^{-2 d}	172
Metformin	9.74*10 ^{0 c}	-	-	170	1.76*10 ^{1 d}	172
Naproxen	1.30*10 ^{-1 c}	-	-	170	-	-
Norethindrone	-	-	-	-	2.54*10 ^{-3 d}	172
Oseltamivir	-	-	-	-	1.02*10 ^{-3 d}	172
Oxazepam	-	-	-	-	1.23*10 ^{-4 d}	172
Propranolol	9.74*10 ^{-2 c}	9.76*10 ^{-2 c}	9.74*10 ^{-2 c}	170	1.89*10 ^{-1 d}	172
Ranitidine	1.30*10 ^{-1 c}	-	-	170	7.07*10 ^{-1 d}	172
Sertraline	-	-	-	-	4.40*10 ^{-1 d}	172
Sitagliptin	4.37*10 ^{-2 a}	4.97*10 ^{-2 a}	4.82*10 ^{-2 a}	-	2.92*10 ^{-1 d}	172

Sulfamethoxazole	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	170	3.32*10 ^{-2 d}	172
Temazepam	-	-	-	-	7.36*10 ^{-3 d}	172
Tramadol	1.30*10 ^{-1 c}	-	-	170	6.95*10 ^{-1 d}	172
Triamterene	-	-	-	-	2.52*10 ^{-4 d}	172
Trimethoprim	3.25*10 ^{-2 c}	3.25*10 ^{-2 c}	3.25*10 ^{-2 c}	170	1.12*10 ^{-1 d}	172
Venlafaxine	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	1.30*10 ^{-1 c}	170	2.38*10 ^{-1 d}	172
Verapamil	1.30*10 ^{-1 c}	-	-	170	8.87*10 ^{-2 d}	172

^a: average per capita consumption from other countries in basin, for which consumption data were available, extrapolated to country and year of interest based on demographics from Eurostat ¹⁷³; ^b: per capita consumption in 2004 ¹⁷⁰ extrapolated to year of interest based on demographics from Eurostat ¹⁷³; ^c: per capita consumption in 2009 ¹⁷⁰ extrapolated to year of interest based on demographics from Eurostat ¹⁷³; ^d: per capita consumption in England ¹⁷² extrapolated to whole of United Kingdom based on demographics from Office for National Statistics ¹⁷⁴.

S6. Additional model results

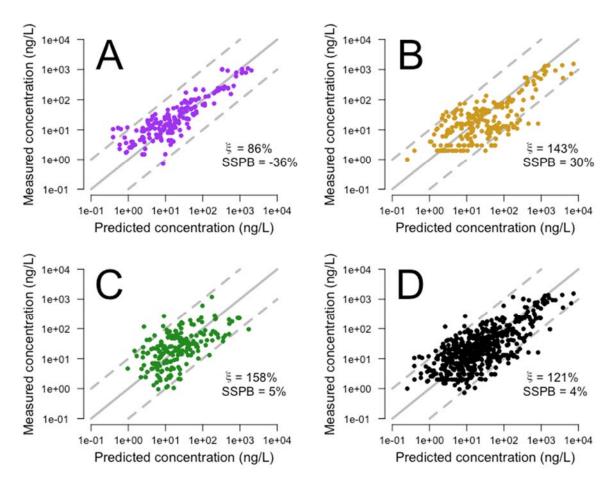


Figure S6.1 Predicted concentrations (i.e., > 0) versus detects (i.e., <40% of the measurements above below LOD) and non-detects if predicted concentration > LOD. Measured data originate from Burns et al.⁴¹ (purple; A), Ruff et al.⁴² (golden; B), Munz et al.⁴³ (green; C), and for all studies combined (black; D). Concentrations predicted under annual mean flow conditions (A) or lowest monthly mean flow conditions (B and C). Solid line represents 1:1 relationship; dashed lines represent 1:10 and 10:1 relationships. \(\xi_{\circ} \) median symmetric accuracy; SSPB: symmetric signed percentage bias.

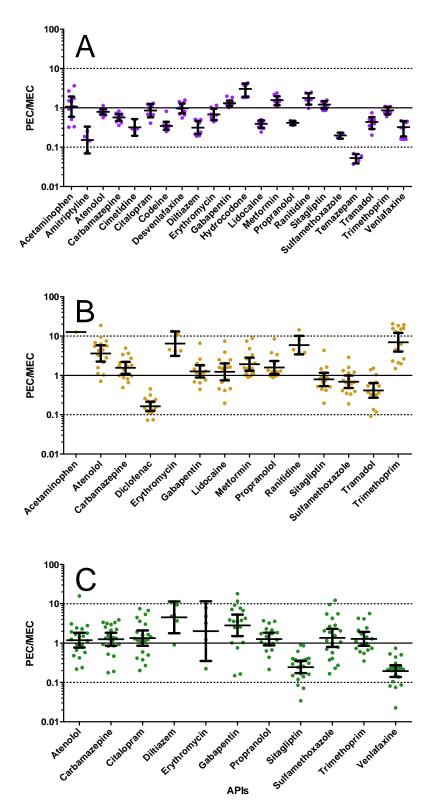


Figure S6.2 Ratios of predicted over measured concentrations (PEC/MEC), reported by Burns et al., ¹⁷⁵ (A), Ruff et al., ¹⁷⁶ (B) and Munz et al. ¹⁷⁷ (C). Coloured dots are individual combinations of API and location, measured above LOD; black bars represent 95th percentile and median over all locations per API measured. Concentrations predicted under annual mean flow conditions (A) or lowest monthly mean flow conditions (B and C).

Table S6.1. Predicted and measured mean annual flow (in m³ s⁻¹) at two gauging stations in the rivers Ouse and Foss, respectively.

	River Ouse	River Foss
Measured (2016)	51.83	0.990
Predicted (based on 2015 data)	52.00	1.145

S7. Interactive html-maps

Interactive html-maps with concentrations and risks per API resulting from the model application exercise can be found as online supplementary file, accessible via the file "Supporting Information S7.html".

References

- 1. EEA, UWWTD-WaterBase. Available from: http://www.eea.europa.eu/data-and-maps/data/waterbase-uwwtd-urban-waste-water-treatment-directive-4. Latest update at 19 February 2015. In European Environmental Agency: 2015.
- 2. Lehner, B.; Verdin, K.; Jarvis, A. New global hydrography derived from spaceborne elevation data. *Eos, Transactions American Geophysical Union* **2008**, *89* (10), 93-94.
- 3. Messager, M. L.; Lehner, B.; Grill, G.; Nedeva, I.; Schmitt, O. Estimating the volume and age of water stored in global lakes using a geo-statistical approach. **2016**, *7*, 13603.
- 4. Jones, P. D.; Harris, I. C. CRU TS3.10: Climatic Research Unit (CRU) Time-Series (TS) Version 3.10 of High Resolution Gridded Data of Month-by-month Variation in Climate (Jan. 1901 Dec. 2009). Available at:
- http://catalogue.ceda.ac.uk/uuid/ac3e6be017970639a9278e64d3fd5508. Accessed at 02/05/2016. In 30/03/2013 ed.; University of East Anglia Climatic Research Unit. NCAS British Atmospheric Data Centre: 2013.
- 5. Kalnay, E.; Kanamitsu, M.; Kistler, R.; Collins, W.; Deaven, D.; Gandin, L.; Iredell, M.; Saha, S.; White, G.; Woollen, J.; Zhu, Y.; Leetmaa, A.; Reynolds, R.; Chelliah, M.; Ebisuzaki, W.; Higgins, W.; Janowiak, J.; Mo, K. C.; Ropelewski, C.; Wang, J.; Jenne, R.; Joseph, D. The NCEP/NCAR 40-Year Reanalysis Project. *Bulletin of the American Meteorological Society* **1996**, *77* (3), 437-471.
- 6. Verdin, K. L.; Godt, J. W.; Funk, C. C.; Pedreros, D.; Worstell, B.; Verdin, J. *Development of a global slope dataset for estimation of landslide occurrence resulting from earthquakes*; 2007-1188; Reston, VA, 2007.
- 7. Barbarossa, V.; Huijbregts, M. A. J.; Beusen, A. H. W.; Beck, H. E.; King, H.; Schipper, A. M. FLO1K, global maps of mean, maximum and minimum annual streamflow at 1 km resolution from 1960 through 2015. *Scientific Data* **2018**, *5*, 180052.
- 8. ECHA Guidance on information requirements and Chemical Safety Assessment. Chapter R.16: Environmental exposure assessment; European Chemicals Agency ECHA: Helsinki, Finland, 2016.
- 9. Humbert, S.; Marshall, J. D.; Shaked, S.; Spadaro, J. V.; Nishioka, Y.; Preiss, P.; McKone, T. E.; Horvath, A.; Jolliet, O. Intake Fraction for Particulate Matter: Recommendations for Life Cycle Impact Assessment. *Environmental Science & Technology* **2011**, *45* (11), 4808-4816.
- 10. Asselman, N. E. M. Suspended sediment in the river Rhine: the impact of climate change on erosion, transport, and deposition. Koninklijk Nederlands Aardrijkskundig Genootschap/Faculteit Ruimtelijke Wetenschappen Universiteit Utrecht: 1997; Vol. 234.
- 11. Burkhard, L. P. Estimating Dissolved Organic Carbon Partition Coefficients for Nonionic Organic Chemicals. *Environmental Science & Technology* **2000**, *34* (22), 4663-4668.
- 12. Fantke, P.; Bijster, M.; Guignard, C.; Hauschild, M.; Huijbregts, M.; Jolliet, O.; Kounina, A.; Magaud, V.; Margni, M.; McKone, T. E.; Posthuma, L.; Rosenbaum, R. K.; van de Meent, D.; van Zelm, R. *USEtox* (R) 2.0 Documentation (Version 1), http://usetox.org. USEtox (R) is a registered trademark of the USEtox (R) Team in the European Union and the United States. All rights reserved. (C) USEtox (R) Team. 2016.
- 13. Sabljić, A.; Güsten, H.; Verhaar, H.; Hermens, J. QSAR modelling of soil sorption. Improvements and systematics of log KOC vs. log KOW correlations. *Chemosphere* **1995**, *31*, (11–12), 4489-4514.

- 14. Franco, A.; Trapp, S. Estimation of the soil—water partition coefficient normalized to organic carbon for ionizable organic chemicals. *Environmental Toxicology and Chemistry* **2008**, *27* (10), 1995-2004.
- 15. Schwarzenbach, R. P.; Gschwend, P. M.; Imboden, D. M. Photochemical transformation reactions. In *Environmental organic chemistry*, Wiley-Interscience: New York, New York, USA, 1993; pp 436-484.
- 16. Margni, M.; Pennington, D. W.; Bennett, D. H.; Jolliet, O. Cyclic Exchanges and Level of Coupling between Environmental Media: Intermedia Feedback in Multimedia Fate Models. *Environmental Science & Technology* **2004**, *38* (20), 5450-5457.
- 17. Rosenbaum, R. K.; Bachmann, T. M.; Gold, L. S.; Huijbregts, M. A. J.; Jolliet, O.; Juraske, R.; Koehler, A.; Larsen, H. F.; MacLeod, M.; Margni, M.; McKone, T. E.; Payet, J.; Schuhmacher, M.; van de Meent, D.; Hauschild, M. Z. USEtox—the UNEP-SETAC toxicity model: recommended characterisation factors for human toxicity and freshwater ecotoxicity in life cycle impact assessment. *The International Journal of Life Cycle Assessment* 2008, 13 (7), 532.
- 18. Honti, M.; Hahn, S.; Hennecke, D.; Junker, T.; Shrestha, P.; Fenner, K. Bridging across OECD 308 and 309 Data in Search of a Robust Biotransformation Indicator. *Environmental Science & Technology* **2016**, *50* (13), 6865-6872.
- 19. Paterson, S.; Mackay, D. Interpreting chemical partitioning in soil-plant-air systems with a fugacity model. In *Plant Contamination*, Trapp, S.; McFarlane, J. C., Eds. Lewis Publishers: Boca Raton, Florida, USA, 1995; pp 191-214.
- 20. Mackay, D. *Multimedia Environmental Models: The Fugacity Approach*. Second Edition ed.; CRC Press: Chelsea, Michigan, USA, 2001; p 272.
- 21. den Hollander, H. A.; van Eijkeren, J. C. H.; van de Meent, D. *SimpleBox 3.0: Multimedia mass balance model for evaluating the fate of chemicals in the environment;* National Institute for Public Health and the Environment (RIVM): Bilthoven, the Netherlands, 2004.
- 22. Wishart, D. S.; Knox, C.; Guo, A. C.; Shrivastava, S.; Hassanali, M.; Stothard, P.; Chang, Z.; Woolsey, J. DrugBank: a comprehensive resource for in silico drug discovery and exploration. *Nucleic Acids Research* **2006**, *34* (suppl_1), D668-D672.
- 23. US EPA, Estimation Programs Interface Suite™ for Microsoft® Windows, v 4.1. . In United States Environmental Protection Agency, Washington, DC, USA., 2017.
- 24. Lienert, J.; Gudel, K.; Escher, B. I. Screening method for ecotoxicological hazard assessment of 42 pharmaceuticals considering human metabolism and excretory routes. *Environ. Sci. Technol.* **2007**, *41* (12), 4471-4478.
- 25. Yan, Q.; Gao, X.; Chen, Y.-P.; Peng, X.-Y.; Zhang, Y.-X.; Gan, X.-M.; Zi, C.-F.; Guo, J.-S. Occurrence, fate and ecotoxicological assessment of pharmaceutically active compounds in wastewater and sludge from wastewater treatment plants in Chongqing, the Three Gorges Reservoir Area. *Science of The Total Environment* **2014**, *470*–*471*, 618-630.
- 26. Okuda, T.; Yamashita, N.; Tanaka, H.; Matsukawa, H.; Tanabe, K. Development of extraction method of pharmaceuticals and their occurrences found in Japanese wastewater treatment plants. *Environment International* **2009**, *35* (5), 815-820.
- 27. Radjenovic, J.; Petrovic, M.; Barceló, D. Fate and distribution of pharmaceuticals in wastewater and sewage sludge of the conventional activated sludge (CAS) and advanced membrane bioreactor (MBR) treatment. *Water Research* **2009**, *43* (3), 831-841.

- 28. Stevens-Garmon, J.; Drewes, J. E.; Khan, S. J.; McDonald, J. A.; Dickenson, E. R. V. Sorption of emerging trace organic compounds onto wastewater sludge solids. *Water Research* **2011**, *45* (11), 3417-3426.
- 29. Berthod, L. M. C. Mechanistic approach to predicting the sorption characteristics of pharmaceuticals. University of Portsmouth, 2015.
- 30. Blair, B.; Nikolaus, A.; Hedman, C.; Klaper, R.; Grundl, T. Evaluating the degradation, sorption, and negative mass balances of pharmaceuticals and personal care products during wastewater treatment. *Chemosphere* **2015**, *134*, 395-401.
- 31. Martinez-Hernandez, V.; Meffe, R.; Herrera, S.; Arranz, E.; de Bustamante, I. Sorption/desorption of non-hydrophobic and ionisable pharmaceutical and personal care products from reclaimed water onto/from a natural sediment. *Science of the Total Environment* **2014**, *472*, 273-281.
- 32. Hari, A. C.; Paruchuri, R. A.; Sabatini, D. A.; Kibbey, T. C. G. Effects of pH and cationic and nonionic surfactants on the adsorption of pharmaceuticals to a natural aquifer material. *Environmental Science & Technology* **2005**, *39* (8), 2592-2598.
- 33. Williams, M.; Saison, C. L. A.; Williams, D. B.; Kookana, R. S. Can aquatic distribution of human pharmaceuticals be related to pharmacological data? *Chemosphere* **2006**, *65* (11), 2253-2259.
- 34. Yamamoto, H.; Nakamura, Y.; Moriguchi, S.; Honda, Y.; Tamura, I.; Hirata, Y.; Hayashi, A.; Sekizawa, J. Persistence and partitioning of eight selected pharmaceuticals in the aquatic environment: Laboratory photolysis, biodegradation, and sorption experiments. *Water Research* **2009**, *43* (2), 351-362.
- 35. Dong, Z.; Senn, D. B.; Moran, R. E.; Shine, J. P. Prioritizing environmental risk of prescription pharmaceuticals. *Regulatory Toxicology and Pharmacology* **2013**, *65* (1), 60-67.
- 36. Hörsing, M.; Ledin, A.; Grabic, R.; Fick, J.; Tysklind, M.; Jansen, J. I. C.; Andersen, H. R. Determination of sorption of seventy-five pharmaceuticals in sewage sludge. *Water Research* **2011**, *45*, (15) 4470-4482.
- 37. Lajeunesse, A.; Smyth, S. A.; Barclay, K.; Sauvé, S.; Gagnon, C. Distribution of antidepressant residues in wastewater and biosolids following different treatment processes by municipal wastewater treatment plants in Canada. *Water Research* **2012**, *4*, (17), 5600-5612.
- 38. Barron, L.; Havel, J.; Purcell, M.; Szpak, M.; Kelleher, B.; Paull, B. Predicting sorption of pharmaceuticals and personal care products onto soil and digested sludge using artificial neural networks. *Analyst* **2009**, *134* (4), 663-670.
- 39. Hyland, K. C.; Dickenson, E. R. V.; Drewes, J. E.; Higgins, C. P. Sorption of ionized and neutral emerging trace organic compounds onto activated sludge from different wastewater treatment configurations. *Water Research* **2012**, *46* (6), 1958-1968.
- 40. Wick, A.; Fink, G.; Joss, A.; Siegrist, H.; Ternes, T. A. Fate of beta blockers and psychoactive drugs in conventional wastewater treatment. *Water Research* **2009**, *43* (4), 1060-1074.
- 41. Subedi, B.; Kannan, K. Occurrence and fate of select psychoactive pharmaceuticals and antihypertensives in two wastewater treatment plants in New York State, USA. *Science of The Total Environment* **2015**, *514*, 273-280.
- 42. Maurer, M.; Escher, B. I.; Richle, P.; Schaffner, C.; Alder, A. C. Elimination of β -blockers in sewage treatment plants. *Water Research* **2007**, *41* (7), 1614-1622.

- 43. Gulde, R.; Helbling, D. E.; Scheidegger, A.; Fenner, K. pH-Dependent Biotransformation of Ionizable Organic Micropollutants in Activated Sludge. *Environmental Science & Technology* **2014**, *48* (23), 13760-13768.
- 44. Ramil, M.; El Aref, T.; Fink, G.; Scheurer, M.; Ternes, T. A. Fate of Beta Blockers in Aquatic-Sediment Systems: Sorption and Biotransformation. *Environmental Science & Technology* **2010**, *44* (3), 962-970.
- 45. Schaffer, M.; Boxberger, N.; Börnick, H.; Licha, T.; Worch, E. Sorption influenced transport of ionizable pharmaceuticals onto a natural sandy aquifer sediment at different pH. *Chemosphere* **2012**, *87* (5), 513-520.
- 46. Schaffer, M.; Börnick, H.; Nödler, K.; Licha, T.; Worch, E. Role of cation exchange processes on the sorption influenced transport of cationic β -blockers in aquifer sediments. *Water Res.* **2012**, *46* (17), 5472-5482.
- 47. Abegglen, C.; Joss, A.; McArdell, C. S.; Fink, G.; Schlusener, M. P.; Ternes, T. A.; Siegrist, H. The fate of selected micropollutants in a single-house MBR. *Water Res* **2009**, *43* (7), 2036-2046.
- 48. Agunbiade, F. O.; Moodley, B. Occurrence and distribution pattern of acidic pharmaceuticals in surface water, wastewater, and sediment of the Msunduzi River, Kwazulu-Natal, South Africa. *Environmental Toxicology and Chemistry* **2016**, *35* (1), 36-46.
- 49. Martín, J.; Camacho-Muñoz, D.; Santos, J. L.; Aparicio, I.; Alonso, E. Occurrence of pharmaceutical compounds in wastewater and sludge from wastewater treatment plants: Removal and ecotoxicological impact of wastewater discharges and sludge disposal. *Journal of Hazardous Materials* **2012**, *239*–*240*, 40-47.
- 50. Morissette, M. F.; Duy, S. V.; Arp, H. P. H.; Sauve, S. Sorption and desorption of diverse contaminants of varying polarity in wastewater sludge with and without alum. *Environ. Sci.-Process Impacts* **2015**, *17* (3), 674-682.
- 51. Ternes, T. A.; Herrmann, N.; Bonerz, M.; Knacker, T.; Siegrist, H.; Joss, A. A rapid method to measure the solid–water distribution coefficient (Kd) for pharmaceuticals and musk fragrances in sewage sludge. *Water Research* **2004**, *38* (19), 4075-4084.
- 52. Urase, T.; Kikuta, T. Separate estimation of adsorption and degradation of pharmaceutical substances and estrogens in the activated sludge process. *Water Research* **2005**, *39* (7), 1289-1300.
- 53. Carballa, M.; Omil, F.; Lema, J. M. Calculation Methods to Perform Mass Balances of Micropollutants in Sewage Treatment Plants. Application to Pharmaceutical and Personal Care Products (PPCPs). *Environmental Science & Technology* **2007**, *41* (3), 884-890.
- 54. Xue, W.; Wu, C.; Xiao, K.; Huang, X.; Zhou, H.; Tsuno, H.; Tanaka, H. Elimination and fate of selected micro-organic pollutants in a full-scale anaerobic/anoxic/aerobic process combined with membrane bioreactor for municipal wastewater reclamation. *Water Research* **2010**, *44* (20), 5999-6010.
- 55. Loffler, D.; Rombke, J.; Meller, M.; Ternes, T. A. Environmental fate of pharmaceuticals in water/sediment systems. *Environmental Science & Technology* **2005**, *39* (14), 5209-5218.
- 56. Zhou, J.; Broodbank, N. Sediment-water interactions of pharmaceutical residues in the river environment. *Water Research* **2014**, *48* 61-70.
- 57. Scheytt, T.; Mersmann, P.; Lindstädt, R.; Heberer, T. Determination of sorption coefficients of pharmaceutically active substances carbamazepine, diclofenac, and ibuprofen, in sandy sediments. *Chemosphere* **2005**, *60* (2), 245-253.

- 58. Williams, M.; Ong, P. L.; Williams, D. B.; Kookana, R. S. Estimating the sorption of pharmaceuticals based on their pharmacological distribution. *Environmental Toxicology and Chemistry* **2009**, *28* (12), 2572-2579.
- 59. Khan, S. J.; Ongerth, J. E. Modelling of pharmaceutical residues in Australian sewage by quantities of use and fugacity calculations. *Chemosphere* **2004**, *54* (3), 355-367.
- 60. Kwon, J.-W.; Armbrust, K. L. Aqueous Solubility, n-Octanol–Water Partition Coefficient, and Sorption of Five Selective Serotonin Reuptake Inhibitors to Sediments and Soils. *Bulletin of Environmental Contamination and Toxicology* **2008**, *81* (2), 128-135.
- 61. Lienert, J.; Burki, T.; Escher, B. I. Reducing micropollutants with source control: substance flow analysis of 212 pharmaceuticals in faeces and urine. *Water Science and Technology* **2007**, *56* (5), 87-96.
- 62. Stein, K.; Ramil, M.; Fink, G.; Sander, M.; Ternes, T. A. Analysis and sorption of psychoactive drugs onto sediment. *Environmental Science & Technology* **2008**, *42* (17), 6415-6423.
- 63. Besse, J.-P.; Kausch-Barreto, C.; Garric, J. Exposure Assessment of Pharmaceuticals and Their Metabolites in the Aquatic Environment: Application to the French Situation and Preliminary Prioritization. *Human and Ecological Risk Assessment: An International Journal* **2008**, *14* (4), 665-695.
- Drugs.com, Venlafaxine Information from Drugs.com; accessed on December 10, 2017; available from: https://www.drugs.com/pro/venlafaxine.html. In 2017.
- 65. Stasinakis, A. S.; Thomaidis, N. S.; Arvaniti, O. S.; Asimakopoulos, A. G.; Samaras, V. G.; Ajibola, A.; Mamais, D.; Lekkas, T. D. Contribution of primary and secondary treatment on the removal of benzothiazoles, benzotriazoles, endocrine disruptors, pharmaceuticals and perfluorinated compounds in a sewage treatment plant. *Science of The Total Environment* **2013**, *463*, 1067-1075.
- 66. Reif, R.; Santos, A.; Judd, S. J.; Lema, J. M.; Omil, F. Occurrence and fate of pharmaceutical and personal care products in a sewage treatment works. *Journal of Environmental Monitoring* **2011**, *13* (1), 137-144.
- 67. Jelic, A.; Fatone, F.; Di Fabio, S.; Petrovic, M.; Cecchi, F.; Barcelo, D. Tracing pharmaceuticals in a municipal plant for integrated wastewater and organic solid waste treatment. *Science of The Total Environment* **2012**, *433* (0), 352-361.
- 68. Fernandez-Fontaina, E.; Pinho, I.; Carballa, M.; Omil, F.; Lema, J. M. Biodegradation kinetic constants and sorption coefficients of micropollutants in membrane bioreactors. *Biodegradation* **2013**, *24* (2), 165-177.
- 69. Petrie, B.; McAdam, E. J.; Lester, J. N.; Cartmell, E. Obtaining process mass balances of pharmaceuticals and triclosan to determine their fate during wastewater treatment. *Science of the Total Environment* **2014**, *497*, 553-560.
- 70. Hanamoto, S.; Nakada, N.; Yamashita, N.; Tanaka, H. Modeling the Photochemical Attenuation of Down-the-Drain Chemicals during River Transport by Stochastic Methods and Field Measurements of Pharmaceuticals and Personal Care Products. *Environmental Science & Technology* **2013**, *47* (23), 13571-13577.
- 71. Senta, I.; Terzic, S.; Ahel, M. Occurrence and fate of dissolved and particulate antimicrobials in municipal wastewater treatment. *Water Research* **2013**, *47* (2), 705-714.
- 72. Kim, S.-C.; Carlson, K. Temporal and Spatial Trends in the Occurrence of Human and Veterinary Antibiotics in Aqueous and River Sediment Matrices. *Environmental Science & Technology* **2007**, *41* (1), 50-57.

- 73. Xu, W. H. H.; Zhang, G.; Wai, O. W. H.; Zou, S. C. C.; Li, X. D. D. Transport and adsorption of antibiotics by marine sediments in a dynamic environment. *J. Soils Sediments* **2009**, *9* (4), 364-373.
- 74. Xu, J.; Zhang, Y.; Zhou, C. B.; Guo, C. S.; Wang, D. M.; Du, P.; Luo, Y.; Wan, J.; Meng, W. Distribution, sources and composition of antibiotics in sediment, overlying water and pore water from Taihu Lake, China. *Science of the Total Environment* **2014**, *497*, 267-273.
- 75. Oyler, J. M.; Cone, E. J.; Joseph, J. R. E.; Huestis, M. A. Identification of Hydrocodone in Human Urine Following Controlled Codeine Administration. *Journal of Analytical Toxicology* **2000**, *24* (7), 530-535.
- 76. Hucker, H. B.; Zacchei, A. G.; Cox, S. V.; Brodie, D. A.; Cantwell, N. H. R. Studies on the absorption, distribution and excretion of indomethacin in various species. *Journal of Pharmacology and Experimental Therapeutics* **1966**, *153* (2), 237-249.
- 77. Ebadi, M. *Desk reference of clinical pharmacology*. CRC Press: 2007.
- 78. Lopez de Alda, M. J.; Gil, A.; Paz, E.; Barcelo, D. Occurrence and analysis of estrogens and progestogens in river sediments by liquid chromatography-electrospray-mass spectrometry. *Analyst* **2002**, *127* (10), 1299-1304.
- 79. Straub, J. O. An environmental risk assessment for oseltamivir (Tamiflu®) for sewage works and surface waters under seasonal-influenza- and pandemic-use conditions. *Ecotoxicology and Environmental Safety* **2009**, *72* (6), 1625-1634.
- 80. Popple, T.; Williams, J. B.; May, E.; Mills, G. A.; Oliver, R. Evaluation of a sequencing batch reactor sewage treatment rig for investigating the fate of radioactively labelled pharmaceuticals: Case study of propranolol. *Water Research* **2016**, *88*, 83-92.
- 81. Kibbey, T. C.; Paruchuri, R.; Sabatini, D. A.; Chen, L., Adsorption of beta blockers to environmental surfaces. *Environmental science & technology* **2007**, *41* (15), 5349-5356.
- 82. Vasiliadou, I. A.; Molina, R.; Martínez, F.; Melero, J. A. Biological removal of pharmaceutical and personal care products by a mixed microbial culture: Sorption, desorption and biodegradation. *Biochemical Engineering Journal* **2013**, *81*, 108-119.
- 83. Gobel, A.; Thomsen, A.; McArdell, C. S.; Joss, A.; Giger, W. Occurrence and sorption behavior of sulfonamides, macrolides, and trimethoprim in activated sludge treatment. *Environ Sci Technol* **2005**, *39* (11), 3981-9.
- 84. Hou, J.; Pan, B.; Niu, X.; Chen, J.; Xing, B. Sulfamethoxazole sorption by sediment fractions in comparison to pyrene and bisphenol A. *Environmental Pollution* **2010**, *158* (9), 2826-2832.
- 85. Xu, B.; Mao, D.; Luo, Y.; Xu, L. Sulfamethoxazole biodegradation and biotransformation in the water—sediment system of a natural river. *Bioresource Technology* **2011**, *102* (14), 7069-7076.
- 86. Chiba, K.; Horii, H.; Chiba, T.; Kato, Y.; Hirano, T.; Ishizaki, T. Development and preliminary application of high-performance liquid chromatographic assay of urinary metabolites of diazepam in humans. *Journal of Chromatography B: Biomedical Sciences and Applications* **1995**, *668* (1), 77-84.
- 87. Halling-Sørensen, B.; Holten Lützhøft, H. C.; Andersen, H. R.; Ingerslev, F. Environmental risk assessment of antibiotics: comparison of mecillinam, trimethoprim and ciprofloxacin. *J Antimicrob Chemother* **2000**, *46 Suppl A*, 53-58.
- 88. Lara-Martín, P. A.; González-Mazo, E.; Petrovic, M.; Barceló, D.; Brownawell, B. J. Occurrence, distribution and partitioning of nonionic surfactants and pharmaceuticals in the urbanized Long Island Sound Estuary (NY). *Mar. Pollut. Bull.* **2014**, *85* (2), 710-719.

- 89. Majewsky, M.; Gallé, T.; Yargeau, V.; Fischer, K. Active heterotrophic biomass and sludge retention time (SRT) as determining factors for biodegradation kinetics of pharmaceuticals in activated sludge. *Bioresource Technology* **2011**, *102* (16), 7415-7421.
- 90. Joss, A.; Zabczynski, S.; Göbel, A.; Hoffmann, B.; Löffler, D.; McArdell, C. S.; Ternes, T. A.; Thomsen, A.; Siegrist, H. Biological degradation of pharmaceuticals in municipal wastewater treatment: Proposing a classification scheme. *Water Research* **2006**, *40* (8), 1686-1696.
- 91. Yu, J. T.; Bouwer, E. J.; Coelhan, M. Occurrence and biodegradability studies of selected pharmaceuticals and personal care products in sewage effluent. *Agricultural Water Management* **2006**, *86* (1–2), 72-80.
- 92. Lam, M. W.; Young, C. J.; Brain, R. A.; Johnson, D. J.; Hanson, M. A.; Wilson, C. J.; Richards, S. M.; Solomon, K. R.; Mabury, S. A. Aquatic persistence of eight pharmaceuticals in a microcosm study. *Environ Toxicol Chem* **2004**, *23* (6), 1431-40.
- 93. Baena-Nogueras, R. M.; González-Mazo, E.; Lara-Martín, P. A. Degradation kinetics of pharmaceuticals and personal care products in surface waters: photolysis vs biodegradation. *Science of The Total Environment* **2017**, *590*–*591*, 643-654.
- 94. Kawabata, K.; Sugihara, K.; Sanoh, S.; Kitamura, S.; Ohta, S. Photodegradation of pharmaceuticals in the aquatic environment by sunlight and UV-A, -B and -C irradiation. *The Journal of Toxicological Sciences* **2013**, *38* (2), 215-223.
- 95. Blum, K. M.; Norström, S. H.; Golovko, O.; Grabic, R.; Järhult, J. D.; Koba, O.; Lindström, H. S. Removal of 30 active pharmaceutical ingredients in surface water under long-term artificial UV irradiation. *Chemosphere* **2017**, *176*,175-182.
- 96. Ribeiro, A. R.; Afonso, C. M.; Castro, P. M. L.; Tiritan, M. E. Enantioselective HPLC analysis and biodegradation of atenolol, metoprolol and fluoxetine. *Environmental Chemistry Letters* **2013**, *11* (1), 83-90.
- 97. Kern, S.; Baumgartner, R.; Helbling, D. E.; Hollender, J.; Singer, H.; Loos, M. J.; Schwarzenbach, R. P.; Fenner, K. A tiered procedure for assessing the formation of biotransformation products of pharmaceuticals and biocides during activated sludge treatment. *Journal of Environmental Monitoring* **2010**, *12* (11), 2100-2111.
- 98. Andrisano, V.; Gotti, R.; Leoni, A.; Cavrini, V. Photodegradation studies on Atenolol by liquid chromatography. *Journal of Pharmaceutical and Biomedical Analysis* **1999**, *21* (4), 851-857.
- 99. Jasper, J. T.; Sedlak, D. L. Phototransformation of Wastewater-Derived Trace Organic Contaminants in Open-Water Unit Process Treatment Wetlands. *Environmental Science & Technology* **2013**, *47* (19), 10781-10790.
- 100. Liu, Q.-T.; Williams, H. E. Kinetics and Degradation Products for Direct Photolysis of β-Blockers in Water. *Environmental Science & Technology* **2007**, *41* (3), 803-810.
- 101. Piram, A.; Salvador, A.; Verne, C.; Herbreteau, B.; Faure, R. Photolysis of β -blockers in environmental waters. *Chemosphere* **2008**, *73* (8), 1265-1271.
- 102. Wang, L.; Xu, H.; Cooper, W. J.; Song, W. Photochemical fate of beta-blockers in NOM enriched waters. *Science of The Total Environment* **2012**, *426*, 289-295.
- 103. Liu, Q.-T.; Cumming, R. I.; Sharpe, A. D. Photo-induced environmental depletion processes of beta-blockers in river waters. *Photochem Photobiol Sci* **2009**, *8* (6), 768-777.
- 104. Zeng, C.; Ji, Y.; Zhou, L.; Zhang, Y.; Yang, X. The role of dissolved organic matters in the aquatic photodegradation of atenolol. *Journal of Hazardous Materials* **2012**, *239*–*240*, 340-347.

- 105. Wang, X. H.; Lin, A. Y. C. Is the phototransformation of pharmaceuticals a natural purification process that decreases ecological and human health risks? *Environ. Pollut.* **2014**, *186*. 203-215.
- 106. Neamţu, M.; Grandjean, D.; Sienkiewicz, A.; Le Faucheur, S.; Slaveykova, V.; Colmenares, J. J. V.; Pulgarín, C.; de Alencastro, L. F. Degradation of eight relevant micropollutants in different water matrices by neutral photo-Fenton process under UV254 and simulated solar light irradiation A comparative study. *Applied Catalysis B: Environmental* **2014**, *158–159*, 30-37.
- 107. Chen, Y.; Li, H.; Wang, Z.; Li, H.; Tao, T.; Zuo, Y. Photodegradation of selected β -blockers in aqueous fulvic acid solutions: Kinetics, mechanism, and product analysis. *Water Research* **2012**, *46* (9), 2965-2972.
- 108. Chen, X.; Vollertsen, J.; Nielsen, J. L.; Gieraltowska Dall, A.; Bester, K. Degradation of PPCPs in activated sludge from different WWTPs in Denmark. *Ecotoxicology* **2015**, *24* (10), 2073-2080.
- 109. Kruglova, A.; Ahlgren, P.; Korhonen, N.; Rantanen, P.; Mikola, A.; Vahala, R. Biodegradation of ibuprofen, diclofenac and carbamazepine in nitrifying activated sludge under 12 °C temperature conditions. *Science of The Total Environment* **2014**, *499*, 394-401.
- 110. Suarez, S.; Lema, J. M.; Omil, F. Removal of Pharmaceutical and Personal Care Products (PPCPs) under nitrifying and denitrifying conditions. *Water Research* **2010**, *44* (10), 3214-3224.
- 111. Kunkel, U.; Radke, M. Fate of pharmaceuticals in rivers: Deriving a benchmark dataset at favorable attenuation conditions. *Water Research* **2012**, *46* (17), 5551-5565.
- 112. Aymerich, I.; Acuña, V.; Barceló, D.; García, M. J.; Petrovic, M.; Poch, M.; Rodriguez-Mozaz, S.; Rodríguez-Roda, I.; Sabater, S.; von Schiller, D.; Corominas, L. Attenuation of pharmaceuticals and their transformation products in a wastewater treatment plant and its receiving river ecosystem. *Water Research* **2016**, *100*, 126-136.
- 113. Andreozzi, R.; Marotta, R.; Paxeus, N. Pharmaceuticals in STP effluents and their solar photodegradation in aquatic environment. *Chemosphere* **2003**, *50*, (10), 1319-1330.
- 114. Andreozzi, R.; Marotta, R.; Pinto, G.; Pollio, A., Carbamazepine in water: persistence in the environment, ozonation treatment and preliminary assessment on algal toxicity. *Water Research* **2002**, *36* (11), 2869-2877.
- 115. Lam, M.; Mabury, S. Photodegradation of the pharmaceuticals atorvastatin, carbamazepine, levofloxacin, and sulfamethoxazole in natural waters. *Aquatic Sciences* **2005**, *67* (2), 177-188.
- 116. Calisto, V.; Domingues, M. R. M.; Erny, G. L.; Esteves, V. I. Direct photodegradation of carbamazepine followed by micellar electrokinetic chromatography and mass spectrometry. *Water Research* **2011**, *45* (3), 1095-1104.
- 117. Calza, P.; Medana, C.; Padovano, E.; Giancotti, V.; Minero, C. Fate of selected pharmaceuticals in river waters. *Environmental Science and Pollution Research* **2013**, *20* (4), 2262-2270.
- 118. Doll, T. E.; Frimmel, F. H. Fate of pharmaceuticals-photodegradation by simulated solar UV-light. *Chemosphere* **2003**, *52* (10), 1757-1769.
- 119. Peuravuori, J.; Pihlaja, K. Phototransformations of selected pharmaceuticals under low-energy UVA–vis and powerful UVB–UVA irradiations in aqueous solutions—the role of natural dissolved organic chromophoric material. *Analytical and Bioanalytical Chemistry* **2009**, *394* (6), 1621-1636.

- 120. Matamoros, V.; Duhec, A.; Albaigés, J.; Bayona, J. M. Photodegradation of Carbamazepine, Ibuprofen, Ketoprofen and 17α -Ethinylestradiol in Fresh and Seawater. *Water, Air, and Soil Pollution* **2008**, *196* (1), 161.
- 121. Benotti, M. J.; Brownawell, B. J. Microbial degradation of pharmaceuticals in estuarine and coastal seawater. *Environmental Pollution* **2009**, *157* (3), 994-1002.
- 122. Latch, D. E.; Stender, B. L.; Packer, J. L.; Arnold, W. A.; McNeill, K. Photochemical fate of pharmaceuticals in the environment: cimetidine and ranitidine. *Environmental science & technology* **2003**, *37* (15), 3342-3350.
- 123. Lin, A. Y. C.; Lin, Y. C.; Lee, W. N. Prevalence and sunlight photolysis of controlled and chemotherapeutic drugs in aqueous environments. *Environmental Pollution* **2014**, *187*, 170-181.
- 124. Vasantharaju, S. G.; Prabu, S. L.; Jacob, A. Spectrofluorimetric Method for Determination of Citalopram in Bulk and Pharmaceutical Dosage Forms. *Indian Journal of Pharmaceutical Sciences* **2008**, *70* (5), 647-648.
- 125. Writer, J. H.; Antweiler, R. C.; Ferrer, I.; Ryan, J. N.; Thurman, E. M. In-stream attenuation of neuro-active pharmaceuticals and their metabolites. *Environmental science & technology* **2013**, *47* (17), 9781-9790.
- 126. Rúa-Gómez, P. C.; Püttmann, W. Degradation of lidocaine, tramadol, venlafaxine and the metabolites O-desmethyltramadol and O-desmethylvenlafaxine in surface waters. *Chemosphere* **2013**, *90* (6), 1952-1959.
- 127. Straub, J. O. Deterministic and probabilistic environmental risk assessment for diazepam. In *Pharmaceuticals in the Environment*, Kummerer, K., Ed. 2008.
- 128. Calisto, V.; Domingues, M. R. M.; Esteves, V. I. Photodegradation of psychiatric pharmaceuticals in aquatic environments Kinetics and photodegradation products. *Water Research* **2011**, *45* (18), 6097-6106.
- 129. Jakimska, A.; SliwkaKaszynska, M.; Nagorski, P.; KotWasik, A.; Namiesnik, J. Environmental fate of two psychiatric drugs, diazepam and sertraline: Phototransformation and investigation of their photoproducts in natural waters. *J Chromatogr Sep Tech* **2014**, *5* (253), 2.
- 130. West, C. E.; Rowland, S. J. Aqueous Phototransformation of Diazepam and Related Human Metabolites under Simulated Sunlight. *Environmental Science & Technology* **2012**, *46* (9), 4749-4756.
- 131. Cormier, G.; Barbeau, B.; Arp, H. P. H.; Sauve, S. The degradation behaviour of nine diverse contaminants in urban surface water and wastewater prior to water treatment. *Environmental Science: Processes & Impacts* **2015**, *1*, (12), 2051-2065.
- 132. Poirier-Larabie, S.; Segura, P. A.; Gagnon, C. Degradation of the pharmaceuticals diclofenac and sulfamethoxazole and their transformation products under controlled environmental conditions. *Science of The Total Environment* **2016**, *557*–*558*, 257-267.
- 133. Buser, H.-R.; Poiger, T.; Müller, M. D. Occurrence and Fate of the Pharmaceutical Drug Diclofenac in Surface Waters: Rapid Photodegradation in a Lake. *Environmental Science & Technology* **1998**, *32* (22), 3449-3456.
- 134. Koumaki, E.; Mamais, D.; Noutsopoulos, C. Environmental fate of non-steroidal anti-inflammatory drugs in river water/sediment systems. *Journal of Hazardous Materials* **2017**, *323*, *Part A*, 233-241.
- 135. Tixier, C.; Singer, H. P.; Oellers, S.; Müller, S. R. Occurrence and Fate of Carbamazepine, Clofibric Acid, Diclofenac, Ibuprofen, Ketoprofen, and Naproxen in Surface Waters. *Environmental Science & Technology* **2003**, *37* (6), 1061-1068.

- 136. Poiger, T.; Buser, H.-R.; Müller, M. D. Photodegradation of the pharmaceutical drug diclofenac in a lake: Pathway, field measurements, and mathematical modeling. *Environmental Toxicology and Chemistry* **2001**, *20* (2), 256-263.
- 137. Peuravuori, J. Aquatic photochemistry of diclofenac in the presence of natural dissolved organic chromophoric material and nitrate. *International Journal of Environmental Analytical Chemistry* **2012**, *92* (13), 1470-1492.
- 138. Eriksson, J.; Svanfelt, J.; Kronberg, L. A Photochemical Study of Diclofenac and Its Major Transformation Products. *Photochemistry and Photobiology* **2010**, *86* (3), 528-532.
- 139. Salgado, R.; Pereira, V. J.; Carvalho, G.; Soeiro, R.; Gaffney, V.; Almeida, C.; Cardoso, V. V.; Ferreira, E.; Benoliel, M. J.; Ternes, T. A.; Oehmen, A.; Reis, M. A. M.; Noronha, J. P. Photodegradation kinetics and transformation products of ketoprofen, diclofenac and atenolol in pure water and treated wastewater. *Journal of Hazardous Materials* **2013**, *244–245*, 516-527.
- 140. Packer, J. L.; Werner, J. J.; Latch, D. E.; McNeill, K.; Arnold, W. A. Photochemical fate of pharmaceuticals in the environment: Naproxen, diclofenac, clofibric acid, and ibuprofen. *Aquatic Sciences* **2003**, *65* (4), 342-351.
- 141. Suleiman, M. S.; Abdulhameed, M. E.; Najib, N. M.; Muti, H. Y. Effect of ultraviolet radiation on the stability of diltiazem. *International Journal of Pharmaceutics* **1989**, *50* (1), 71-73.
- 142. Li, B.; Zhang, T, Biodegradation and adsorption of antibiotics in the activated sludge process. *Environ. Sci. Technol.* **2010**, *44*,(9), 3468-73.
- 143. Alexy, R.; Kümpel, T.; Kümmerer, K. Assessment of degradation of 18 antibiotics in the Closed Bottle Test. *Chemosphere* **2004**, *57* (6), 505-512.
- 144. Batchu, S. R.; Panditi, V. R.; O'Shea, K. E.; Gardinali, P. R. Photodegradation of antibiotics under simulated solar radiation: Implications for their environmental fate. *Science of the Total Environment* **2014**, *470*, 299-310.
- 145. Herrmann, M.; Menz, J.; Olsson, O.; Kümmerer, K. Identification of phototransformation products of the antiepileptic drug gabapentin: Biodegradability and initial assessment of toxicity. *Water Research* **2015**, *85*, 11-21.
- 146. Abdellatef, H. E.; Khalil, H. M. Colorimetric determination of gabapentin in pharmaceutical formulation. *Journal of Pharmaceutical and Biomedical Analysis* **2003**, *31* (1), 209-214.
- 147. Boix, C.; Ibáñez, M.; Sancho, J. V.; Parsons, J. R.; Voogt, P. d.; Hernández, F. Biotransformation of pharmaceuticals in surface water and during waste water treatment: Identification and occurrence of transformation products. *Journal of Hazardous Materials* **2016**, *302*, 175-187.
- 148. Fono, L. J.; Kolodziej, E. P.; Sedlak, D. L. Attenuation of Wastewater-Derived Contaminants in an Effluent-Dominated River. *Environmental Science & Technology* **2006**, *40* (23), 7257-7262.
- 149. Lin, A. Y.-C.; Reinhard, M. Photodegradation of common environmental pharmaceuticals and estrogens in river water. *Environmental Toxicology and Chemistry* **2005**, *24* (6), 1303-1309.
- 150. Vione, D.; Maddigapu, P. R.; De Laurentiis, E.; Minella, M.; Pazzi, M.; Maurino, V.; Minero, C.; Kouras, S.; Richard, C. Modelling the photochemical fate of ibuprofen in surface waters. *Water Research* **2011**, *45* (20), 6725-6736.
- 151. Burgis, C. Predicting Biological Removal of Contaminants in Wastewater Treatment: QSBR Modeling. University of Virginia, Charlottesville, Virginia, USA, 2012.

- 152. Trautwein, C.; Kümmerer, K. Incomplete aerobic degradation of the antidiabetic drug Metformin and identification of the bacterial dead-end transformation product Guanylurea. *Chemosphere* **2011**, *85* (5), 765-773.
- 153. Arayne, M.; Sultana, N.; Zuberi, M. v.; Siddiqui, F. Spectrophotometric quantitation of metformin in bulk drug and pharmaceutical formulations using multivariate technique. *Indian journal of pharmaceutical sciences* **2009**, *71* (3), 331.
- 154. Accinelli, C.; Saccà, M. L.; Fick, J.; Mencarelli, M.; Lindberg, R.; Olsen, B. Dissipation and removal of oseltamivir (Tamiflu) in different aquatic environments. *Chemosphere* **2010**, *79* (8), 891-897.
- 155. Bartels, P.; von Tümpling Jr, W. The environmental fate of the antiviral drug oseltamivir carboxylate in different waters. *Science of The Total Environment* **2008**, *405* (1–3), 215-225.
- 156. Gonçalves, C.; Pérez, S.; Osorio, V.; Petrovic, M.; Alpendurada, M. F.; Barceló, D. Photofate of Oseltamivir (Tamiflu) and Oseltamivir Carboxylate under Natural and Simulated Solar Irradiation: Kinetics, Identification of the Transformation Products, and Environmental Occurrence. *Environmental Science & Technology* **2011**, *45* (10), 4307-4314.
- 157. Robinson, P. F.; Liu, Q.-T.; Riddle, A. M.; Murray-Smith, R. Modeling the impact of direct phototransformation on predicted environmental concentrations (PECs) of propranolol hydrochloride in UK and US rivers. *Chemosphere* **2007**, *66* (4), 757-766.
- 158. Addamo, M.; Augugliaro, V.; Di Paola, A.; Garcia-Lopez, E.; Loddo, V.; Marci, G.; Palmisano, L. Removal of drugs in aqueous systems by photoassisted degradation. *J. Appl. Electrochem.* **2005**, *35* (7-8), 765-774.
- 159. Walash, M.; Belal, F.; El-Enany, N.; El-Mansi, H. Spectrophotometric determination of the antidepressants sertraline and paroxetine HCl using 2, 4-dinitrofluorobenzene. *Int J Biomed Sci* **2010**, *6*, 252-259.
- 160. Plósz, B. G.; Leknes, H.; Thomas, K. V. Impacts of Competitive Inhibition, Parent Compound Formation and Partitioning Behavior on the Removal of Antibiotics in Municipal Wastewater Treatment. *Environmental Science & Technology* **2010**, *44* (2), 734-742.
- 161. Pérez, S.; Eichhorn, P.; Aga, D. S. Evaluating the biodegradability of sulfamethazine, sulfamethoxazole, sulfathiazole, and trimethoprim at different stages of sewage treatment. *Environmental Toxicology and Chemistry* **2005**, *24* (6), 1361-1367.
- 162. Lai, H.-T.; Hou, J.-H. Light and microbial effects on the transformation of four sulfonamides in eel pond water and sediment. *Aquaculture* **2008**, *283* (1–4), 50-55.
- 163. Bonvin, F.; Omlin, J.; Rutler, R.; Schweizer, W. B.; Alaimo, P. J.; Strathmann, T. J.; McNeill, K.; Kohn, T. Direct Photolysis of Human Metabolites of the Antibiotic Sulfamethoxazole: Evidence for Abiotic Back-Transformation. *Environmental Science & Technology* **2013**, *47* (13), 6746-6755.
- 164. Niu, J.; Zhang, L.; Li, Y.; Zhao, J.; Lv, S.; Xiao, K. Effects of environmental factors on sulfamethoxazole photodegradation under simulated sunlight irradiation: Kinetics and mechanism. *Journal of Environmental Sciences* **2013**, *25* (6), 1098-1106.
- 165. Boreen, A. L.; Arnold, W. A.; McNeill, K. Photochemical Fate of Sulfa Drugs in the Aquatic Environment: Sulfa Drugs Containing Five-Membered Heterocyclic Groups. *Environmental Science & Technology* **2004**, *38* (14), 3933-3940.
- 166. Ryan, C. C.; Tan, D. T.; Arnold, W. A. Direct and indirect photolysis of sulfamethoxazole and trimethoprim in wastewater treatment plant effluent. *Water Res* **2011**, *45* (3), 1280-6.

- 167. Antonopoulou, M.; Konstantinou, I. Photocatalytic degradation and mineralization of tramadol pharmaceutical in aqueous TiO2 suspensions: Evaluation of kinetics, mechanisms and ecotoxicity. *Applied Catalysis A: General* **2016**, *515*, 136-143.
- 168. Sirtori, C.; Agüera, A.; Gernjak, W.; Malato, S. Effect of water-matrix composition on Trimethoprim solar photodegradation kinetics and pathways. *Water Research* **2010**, *44* (9), 2735-2744.
- 169. Loftin, K. A.; Adams, C. D.; Meyer, M. T.; Surampalli, R. Effects of Ionic Strength, Temperature, and pH on Degradation of Selected Antibiotics All rights reserved. No part of this periodical may be reproduced or transmitted in any form or by any means, electronic or mechanical, including photocopying, recording, or any information storage and retrieval system, without permission in writing from the publisher. *J. Environ. Qual.* **2008**, *37* (2), 378-386.
- 170. Singer, H. P.; Wössner, A. E.; McArdell, C. S.; Fenner, K. Rapid Screening for Exposure to "Non-Target" Pharmaceuticals from Wastewater Effluents by Combining HRMS-Based Suspect Screening and Exposure Modeling. *Environmental Science & Technology* **2016**, *50* (13), 6698–6707
- 171. Zorginstituut Nederland, GIPdatabank; available at: https://www.gipdatabank.nl. In 2015.
- 172. National Health Service, Prescription Cost Analysis Data [online], available at: < https://www.nhsbsa.nhs.uk/prescription-data/dispensing-data/prescription-cost-analysis-pca-data>, last accessed: 11 Dec., 2017. In 2017.
- 173. Eurostat, Population on 1 January by age and sex [demo_pjan]. Available from: http://appsso.eurostat.ec.europa.eu/nui/show.do?dataset=demo_pjan&lang=en. Last update: 27-02-2018; accessed: 18-04-2018. In European Union: 2018.
- 174. Office for National Statistics, Population Estimates for UK, England and Wales, Scotland and Northern Ireland: Mid-2016. Available at:
- https://www.ons.gov.uk/releases/populationestimatesforukenglandandwalesscotlandandnorthernirelandmid2016. In Fareham, UK, 2017.
- 175. Burns, E. E.; Carter, L. J.; Kolpin, D. W.; Thomas-Oates, J.; Boxall, A. B. A. Temporal and spatial variation in pharmaceutical concentrations in an urban river system. *Water Research* **2018**, *137*, 72-85.
- 176. Ruff, M.; Mueller, M. S.; Loos, M.; Singer, H. P. Quantitative target and systematic non-target analysis of polar organic micro-pollutants along the river Rhine using high-resolution mass-spectrometry Identification of unknown sources and compounds. *Water Research* **2015**, *87*, 145-154.
- 177. Munz, N. A.; Burdon, F. J.; de Zwart, D.; Junghans, M.; Melo, L.; Reyes, M.; Schönenberger, U.; Singer, H. P.; Spycher, B.; Hollender, J.; Stamm, C. Pesticides drive risk of micropollutants in wastewater-impacted streams during low flow conditions. *Water Research* **2017**, *110*, 366-377.