Draft – Effect of Atmospheric Heatwaves on Reflectance and Pigment Composition of Intertidal *Nanozostera noltei* – Draft

Simon Oiry

Bede Ffinian Rowe Davies

Philippe Rosa

Augustin Debly

Pierre Gernez

Laurent Barillé

2024-10-17

Abstract

To be written

# 1. Introduction

Intertidal seagrasses play a crucial role in the ecosystem by providing habitats and feeding grounds for various marine species, supporting rich marine biodiversity, and contributing significantly to primary production and carbon sequestration (Sousa et al., 2019; Unsworth et al., 2022). These seagrasses are essential in maintaining the health of coastal ecosystems by stabilizing sediments, filtering water, and serving as indicators of environmental changes due to their sensitivity to water quality variations (Zoffoli et al., 2021). The interactions between seagrass meadows and their associated herbivores further enhance the delivery of ecosystem services, including coastal protection and fisheries support (Gardner and Finlayson, 2018; Jankowska et al., 2019; Zoffoli et al., 2023). Understanding and preserving these ecosystems are vital for maintaining the biodiversity and productivity of coastal regions (Ramesh and Mohanraju, 2020; Scott et al., 2018).

Despite their crucial role in marine ecosystems, intertidal seagrasses face numerous threats that compromise their health and functionality. Coastal development and human activities are primary threats. These activities not only reduce the available habitat for seagrasses but also increase water turbidity, which limits light penetration and hampers photosynthesis (Waycott et al., 2009). Seagrasses are also threatened by nutrient enrichment from agricultural and urban runoff, which can lead to eutrophication. This condition promotes the overgrowth of algal blooms that compete with seagrasses for light and nutrients, further stressing these important plants (Thomsen et al., 2023) (Oiry et al. 2024). Pollution from industrial and agricultural fields sources introduces harmful chemicals and heavy metals into coastal waters, posing toxic risks to seagrass health. These pollutants can affect the physiological processes of seagrasses, reducing their growth and survival rates (Sevgi and Leblebici, 2022) Additionally, invasive species can out compete native seagrasses for resources, altering community structure and function (Simpson et al., 2016).

Heatwaves, exacerbated by climate change, pose a growing threat to seagrasses. Marine Heatwaves (MHW), defined by Hobday et al. (2016) as prolonged discrete anomalously warm water events, and Atmospheric Heatwaves (AHW), defined by Perkins and Alexander (2013) as periods of at least three consecutive days with temperatures exceeding the 90th percentile, cause severe physiological stress on seagrasses (Deguette et al., 2022; Sawall et al., 2021). At the interface between the land and oceans, intertidal seagrasses are exposed to both MHW and AHW. Heatwaves have profound impacts on seagrasses, with their effects varying based on species and geographic location. For instance, the seagrass *Zostera marina* exhibits high susceptibility to elevated sea surface temperatures during winter and spring, leading to advanced flowering, high mortality rates, and reduced biomass (Sawall et al., 2021). Similarly, *Cymodocea nodosa* shows increased photosynthetic activity during heatwaves but suffers negative effects on photosynthetic performance and leaf biomass during recovery (Deguette et al., 2022). Additionally, different populations of *Zostera marina* along the European thermal gradient exhibit varied photophysiological responses during the recovery phase of heatwaves, indicating differential adaptation capabilities among populations (Winters et al., 2011). These events intensify other stressors, such as overgrazing and seed burial, compromising sexual recruitment (Guerrero-Meseguer et al., 2020).

Bleaching and browning events of seagrass beds have been observed following episodes of intense heat along the Brittany coast of France (Pers. obs.) then affecting leaf color, which are expected to alter leaf reflectance. Remote sensing is increasingly being utilized to monitor marine ecosystems, including seagrass meadows. By using spectral indices, such as the Normalized Difference Vegetation Index (NDVI) and the Soil-Adjusted Vegetation Index (SAVI), or by analyzing specific spectral patterns, remote sensing can effectively quantify vegetation health over time (Akbar et al., 2020; Cârlan et al., 2020; Huete, 2012; Kloos et al., 2021). Through the Water Framework Directive and the Marine Strategy Framework Directive, Europe is promoting remote sensing techniques for habitat mapping, as these methods enable the monitoring of extensive areas over time (Papathanasopoulou et al., 2019). This study will experimentally test the hypothesis that warm events modify the pigment composition and reflectance of seagrass, linking these changes with satellite remote sensing.

# 2. Material & Methods

## 2.1 Heatwaves detection and characterisation

### 2.1.1 Air temperature

Since January 1, 2024, Meteo France weather data has been freely and openly accessible. Hourly air temperature data (°C) for the French Atlantic and Channel coasts was retrieved using a [custom script](https://github.com/SigOiry/HeatWave_Seagrasses/blob/main/Scripts/MeteoFrance_Extraction.qmd) as no API was available at the time of this study. Weather stations located within 10 kilometers of the coastline were considered, but only those with at least 30 years of data were included to ensure reliable climatological reconstruction. Of the 156 weather stations within 10 kilometers of the coast, only 36 had sufficient data for climatology reconstruction. The hourly data was then aggregated into daily mean temperatures for each station.

Heatwave detection was performed using the HeatwaveR package in R (Schlegel and Smit, 2018). This package utilizes the methodology proposed by Hobday et al. (2016) to detect heatwave events. The climatology for the year was computed using the temperature time series. An event was considered a heatwave each time the temperature exceeded the 90th percentile of the climatology for five consecutive days. The severity of each event has been assessed using the methodology proposed by Hobday et al. (2018).

### 2.1.2 Water temperature

(CMEMS, 2024)

## 2.2 Experiment

### 2.2.1 Sampling and acclimation of seagrasses

Seagrass was sampled from a *Nanozostera noltei* (dwarf eelgrass, syn. *Zostera noltei*) meadow on Noirmoutier Island, France (46°57’32.0”N 2°10’37.0”W) at low tide in June 2024. A home-made inox sampling box was used to sample seagrass from an area of 30 cm by 15 cm and 5 cm deep, maintaining the sediment structure and avoiding damage to the rhizomes and the leafs of the seagrass ([Figure 1](#fig-design) A). This sampling box allowed to limitate sampling variability between replicates. The seagrass, along with sediment, meiofauna, and macrofauna, was placed in plastic trays. To avoid hydric stress during transportation, seawater was added to each tray. Simultaneously, seawater was sampled from a nearby site and transported to the lab, where it was filtered using a 0.22 µm nitrocellulose filter to remove all suspended particulate matter. This seawater was used in the acclimation tank and the intertidal chambers. The seagrasses were acclimated at high tide for one weeks with a water temperature of 17°C, matching the temperature at the time of sampling, and with light of 150 µmol.s-1.m-2 of PAR photons (Akbar et al., 2020). A wave generator was used in the tank to circulate and reoxygenate the water.

### 2.2.2 Experimental design

|  |
| --- |
| Figure 1: Illustrations of the various steps of the experiment. A: Field sampling of seagrass using a homemade sampling box; B: Intertidal chambers used during the experiment; C: Seagrass sample inside a chamber during the experiment at high tide; D: Photo of the treatment sample at the start of the experiment; E: Photo of the treatment sample at the end of the experiment. |

Two intertidal chambers from [ElectricBlue](https://electricblue.eu/intertidal-chamber) were used to simulate tidal cycles and control water temperature during high tide and air temperature during low tide ([Figure 1](#fig-design) B,C). One chamber served as the control, while the other was used for the experimental treatment. The control chamber was maintained at temperatures representative of the typical seasonal conditions: water temperatures between 18°C and 19°C and air temperatures between 18°C and 23°C, following circadian temperature variability ([Figure 2](#fig-Profile) left). For the experimental treatment, the air temperature was set to mimic an atmospheric heatwave that occurred over the seagrass meadow of Porh Saint-Guénël, Plouharnel, France (47°35’40.0”N, 3°07’30.0”W) from August 26, 2021, to September 6, 2021. On the first day of the experiment, air temperatures in the experimental chamber were set to range from 23°C at night to 35°C during the day, with a daily increase of 1°C. The water temperature in the experimental chamber was similarly adjusted to reflect the heatwave conditions, starting from the normal seasonal range (18°C) and gradually increasing to simulate the rising temperatures experienced during the heatwave (+0.5°C daily). This setup aimed to replicate the thermal stress experienced by the seagrass meadow during the actual heatwave event ([Figure 2](#fig-Profile) right).

|  |
| --- |
| Figure 2: Temperature profiles of both the control (left) and the treatment (right) followed during the heatwave experiment. The red line indicates air temperature, whereas the blue line indicates water temperature. Due to the tidal cycle followed during the experiment, the seagrasses only experience temperatures representade by solid lines. |

### 2.2.3 Measurement and sampling

#### 2.2.3.1 Hyperspectral measurement

Throughout the experiment, hyperspectral signatures of both the control and treatment seagrasses were taken using an ASD HandHeld 2 equipped with a fiber optic, allowing measurements to be taken directly inside the chamber without opening it. Automatic spectra acquisition has been done using the [RS3 softaware](https://www.malvernpanalytical.com/en/learn/knowledge-center/user-manuals/rs3-software-user-manual) developed by the intrument manufacturer. An average of five reflectance spectrum (), each with an integration time of 544 ms, was taken every minute. Every 10 minutes, the fiber optic was switched from one benthic chamber to the other, in order to measure reflectance in both treatment and control. Because light conditions were controlled inside of the chambers, reflectance calibration of the instrument was performed only each morning at the very first moment of low tide using a Spectralon white reference with 99% Lambertian reflectivity.

The second derivative of was calculated to retrieve absorption features and compare their variability over time. Two radiometric indices were also monitored throughout the experiment :

* The Normalized Difference Vegetation Index (NDVI, Rouse et al. (1974)), as a proxy of the concentration of chlorophyll-a ([Equation 1](#eq-ndvi))

where and are the reflectance at 840 nm and 668 nm respectively.

* The Green Leaf Index (GLI, Louhaichi et al. (2001)), as a measurement of the greenness of seagrass leafs ([Equation 2](#eq-gli))

where and are the reflectance in green at 550 nm and in the blue at 450 nm, respectively.

* The Mid-Infrared Water Absorption Index (MIWAI), proposed here and designed to measure water absorption at 970 nm (**REF**), estimates the difference between the reflectance at 970 nm and a linear interpolation of the reflectance values at 950 and 990 nm. This interpolation represents the expected reflectance value in the absence of water.

where , and are the reflectance in the infrared at 990, 970 and 950 nm, respectively.

#### 2.2.3.2 Multispectral imagery measurement

Parallel to hyperspectral measurements, multispectral images were taken at the beginning and end of each diurnal low tide (09:00 am and 03:00 pm). A Micasense RedEdge-MX Dual multispectral camera, originally designed to be mounted on a drone, was modified for use without a drone. A 3D-printed mount was designed to attach the camera to the intertidal chamber and ensure that each picture was captured under the same conditions. At each time step (09:00 am and 03:00 pm), a first picture of the Spectralon was taken to allow for image correction in reflectance, followed by a second picture of the target. [DISCOV](https://sigoiry.github.io/DISCOV-MicaSense/), a Neural Network classification model previously developed to map intertidal vegetation using drone imagery, has been applied to each image taken inside the intertidal chambers. To understand the behavior of the model on seagrasses affected by heatwaves, classification images from before and after the heatwave have been compared.

#### 2.2.3.3 Leaves sampling and HPLC measurement

At the beginning and the end of each diurn low tide (09:00 am and 03:00 pm) leaves samples have been took in both the test and the control. leaves sampled have been stored under -80°C waiting for analysis. Pigment composition and biomass were analyzed using high-performance liquid chromatography (HPLC). The HPLC system (Alliance HPLC 248 System, Waters) was equipped with a reverse-phase C-18 separating column (SunFire C-18 Column, 100Å, 3.5 µm, 2.1 mm x 50 mm, Waters), preceded by a precolumn (VanGuard 3.9 mm x 5 mm, Waters). The system also included a photodiode array detector (2998 PDA) and a fluorimeter (Ex: 425 nm, Em: 655 nm; RF-20A, SHIMADZU).

**Au secours Philippe !!!**

# 3. Bibliography

Akbar, M., Arisanto, P., Sukirno, B., Merdeka, P., Priadhi, M., Zallesa, S., 2020. Mangrove vegetation health index analysis by implementing NDVI (normalized difference vegetation index) classification method on sentinel-2 image data case study: Segara anakan, kabupaten cilacap, in: IOP Conference Series: Earth and Environmental Science. IOP Publishing, p. 012069.

Cârlan, I., Mihai, B.-A., Nistor, C., Große-Stoltenberg, A., 2020. Identifying urban vegetation stress factors based on open access remote sensing imagery and field observations. Ecological Informatics 55, 101032.

CMEMS, 2024. European north west shelf/iberia biscay irish seas – high resolution ODYSSEA sea surface temperature multi-sensor L3 observations reprocessed, e.u. Copernicus marine service information (CMEMS). Marine data store (MDS). (Accecced on 17-10-2024). <https://doi.org/10.48670/moi-00311>

Deguette, A., Barrote, I., Silva, J., 2022. Physiological and morphological effects of a marine heatwave on the seagrass cymodocea nodosa. Scientific Reports 12, 7950.

Gardner, R.C., Finlayson, C., 2018. Global wetland outlook: State of the world’s wetlands and their services to people, in: Ramsar Convention Secretariat. pp. 2020–5.

Guerrero-Meseguer, L., Marı́n, A., Sanz-Lázaro, C., 2020. Heat wave intensity can vary the cumulative effects of multiple environmental stressors on posidonia oceanica seedlings. Marine Environmental Research 159, 105001.

Hobday, A.J., Alexander, L.V., Perkins, S.E., Smale, D.A., Straub, S.C., Oliver, E.C., Benthuysen, J.A., Burrows, M.T., Donat, M.G., Feng, M., others, 2016. A hierarchical approach to defining marine heatwaves. Progress in oceanography 141, 227–238.

Hobday, A.J., Oliver, E.C., Gupta, A.S., Benthuysen, J.A., Burrows, M.T., Donat, M.G., Holbrook, N.J., Moore, P.J., Thomsen, M.S., Wernberg, T., others, 2018. Categorizing and naming marine heatwaves. Oceanography 31, 162–173.

Huete, A.R., 2012. Vegetation indices, remote sensing and forest monitoring. Geography Compass 6, 513–532.

Jankowska, E., Michel, L.N., Lepoint, G., Włodarska-Kowalczuk, M., 2019. Stabilizing effects of seagrass meadows on coastal water benthic food webs. Journal of Experimental Marine Biology and Ecology 510, 54–63.

Kloos, S., Yuan, Y., Castelli, M., Menzel, A., 2021. Agricultural drought detection with MODIS based vegetation health indices in southeast germany. Remote Sensing 13, 3907.

Louhaichi, M., Borman, M.M., Johnson, D.E., 2001. Spatially located platform and aerial photography for documentation of grazing impacts on wheat. Geocarto International 16, 65–70.

Papathanasopoulou, E., Simis, S., Alikas, K., Ansper, A., Anttila, J., Barillé, A., Barillé, L., Brando, V., Bresciani, M., Bučas, M., others, 2019. Satellite-assisted monitoring of water quality to support the implementation of the water framework directive. EOMORES white paper.

Perkins, S.E., Alexander, L.V., 2013. On the measurement of heat waves. Journal of climate 26, 4500–4517.

Ramesh, C., Mohanraju, R., 2020. Seagrass ecosystems of andaman and nicobar islands: Status and future perspective. Environmental & Earth Sciences Research Journal 7.

Rouse, J.W., Haas, R.H., Schell, J.A., Deering, D.W., others, 1974. Monitoring vegetation systems in the great plains with ERTS. NASA Spec. Publ 351, 309.

Sawall, Y., Ito, M., Pansch, C., 2021. Chronically elevated sea surface temperatures revealed high susceptibility of the eelgrass zostera marina to winter and spring warming. Limnology and Oceanography 66, 4112–4124.

Schlegel, R.W., Smit, A.J., 2018. heatwaveR: A central algorithm for the detection of heatwaves and cold-spells. Journal of Open Source Software 3, 821. <https://doi.org/10.21105/joss.00821>

Scott, A.L., York, P.H., Duncan, C., Macreadie, P.I., Connolly, R.M., Ellis, M.T., Jarvis, J.C., Jinks, K.I., Marsh, H., Rasheed, M.A., 2018. The role of herbivory in structuring tropical seagrass ecosystem service delivery. Frontiers in Plant Science 9, 127.

Sevgi, K., Leblebici, S., 2022. Bitkilerde ağır metal stresine verilen fizyolojik ve moleküler yanıtlar. Journal of Anatolian Environmental and Animal Sciences 7, 528–536.

Simpson, T.S., Wernberg, T., McDonald, J.I., 2016. Distribution and localised effects of the invasive ascidian didemnum perlucidum (monniot 1983) in an urban estuary. PLoS One 11, e0154201.

Sousa, A.I., Silva, J.F. da, Azevedo, A., Lillebø, A.I., 2019. Blue carbon stock in zostera noltei meadows at ria de aveiro coastal lagoon (portugal) over a decade. Scientific reports 9, 14387.

Thomsen, E., Herbeck, L.S., Viana, I.G., Jennerjahn, T.C., 2023. Meadow trophic status regulates the nitrogen filter function of tropical seagrasses in seasonally eutrophic coastal waters. Limnology and Oceanography 68, 1906–1919.

Unsworth, R.K., Cullen-Unsworth, L.C., Jones, B.L., Lilley, R.J., 2022. The planetary role of seagrass conservation. Science 377, 609–613.

Waycott, M., Duarte, C.M., Carruthers, T.J., Orth, R.J., Dennison, W.C., Olyarnik, S., Calladine, A., Fourqurean, J.W., Heck Jr, K.L., Hughes, A.R., others, 2009. Accelerating loss of seagrasses across the globe threatens coastal ecosystems. Proceedings of the national academy of sciences 106, 12377–12381.

Winters, G., Nelle, P., Fricke, B., Rauch, G., Reusch, T.B., 2011. Effects of a simulated heat wave on photophysiology and gene expression of high-and low-latitude populations of zostera marina. Marine Ecology Progress Series 435, 83–95.

Zoffoli, M.L., Gernez, P., Godet, L., Peters, S., Oiry, S., Barillé, L., 2021. Decadal increase in the ecological status of a north-atlantic intertidal seagrass meadow observed with multi-mission satellite time-series. Ecological Indicators 130, 108033.

Zoffoli, M.L., Gernez, P., Oiry, S., Godet, L., Dalloyau, S., Davies, B.F.R., Barillé, L., 2023. Remote sensing in seagrass ecology: Coupled dynamics between migratory herbivorous birds and intertidal meadows observed by satellite during four decades. Remote Sensing in Ecology and Conservation 9, 420–433.