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Title: Principles of cross congruence do not apply in naturally disturbed dune slack habitats: implications for conservation monitoring.

3 Authors: Aoife Delaney^{*1}, Jane Stout¹

4 1: School of Natural Sciences, Trinity College Dublin

⁵ *Corresponding author: Botany Building, School of Natural Sciences, Trinity College Dublin,
⁶ Dublin 2, Ireland; e-mail: amdelane@tcd.ie; ph. no. +353 (0)1 896 3068

7 Declarations of interest: None

8 **Abstract**

9 Cross congruence, where diversity or composition of multiple species follow similar patterns,
10 underlies the use of indicator species in conservation practice. However, there are
11 circumstances in which cross congruence has been shown to break down, for example after
12 disturbance events. If cross congruence does not occur in habitats which experience natural
13 disturbance, then conservation measures based on indicator species may yield misleading
14 results. We assessed the degree of cross congruence among three biological taxa in dune
15 slacks, which are temporary ponds found in coastal sand dunes. We also investigated the
16 efficacy of a national monitoring assessment based on a single taxonomic group, plants, to
17 predict the diversity and composition of two other taxonomic groups: snails and water
18 beetles. We found no evidence of cross congruence among these three groups and the
19 plant-based monitoring system did not predict the composition or diversity of snails or water
20 beetles. The potential for dune slacks to support species of conservation interest was
21 demonstrated here as eight snail species and two water beetle species listed as Near
22 threatened or Vulnerable on the Irish Red Lists were found within the 24 sites surveyed, and
23 the Irish populations of two of these are of international significance (*Vertigo angustior*,
24 *Leiostyla anglica*). Some of these species of conservation interest were found in dune
25 slacks which were of poor conservation status according to the monitoring methodology
26 applied. Our results show that indicator species drawn from a single taxonomic group are
27 not adequate to monitor the general habitat condition of dune slacks. Dune slacks are
28 among many habitats of conservation interest which experience natural disturbance, and this
29 research has implications for conservation practice in other habitats such as seasonal
30 wetlands and fire-dependent habitats.

31 **Highlights**

- 32 • Conservation assessment protocols based on indicator species might be undermined
33 in habitats which experience natural disturbances because cross-taxon congruence
34 breaks down after disturbance events.
- 35 • We found that diversity of plants, snails and water beetles in dune slacks, a habitat
36 characterised by regular disturbance, were not congruent. A habitat assessment
37 based on plant indicator species failed to reliably distinguish between dune slack
38 sites on the basis of the richness or composition of their snail and water beetle
39 assemblages.
- 40 • Indicator species should be taken from a range of different taxonomic groups to
41 enhance conservation assessments in habitats which experience natural
42 disturbances.

43 **Keywords**

44 Cross-congruence; indicator species; dune slacks; invertebrates; snails; water beetles;
45 plants; conservation; assessment

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49 J. Denyer for their help with identifying beetle, snail and plant specimens. This work was
50 supported by the National Parks and Wildlife Service of Ireland and by a Trinity College
51 Dublin Studentship.

52 **1. Introduction**

53 Global biodiversity decline has been associated with reductions in ecosystem productivity
54 (Liang et al., 2016) and consequent ecosystem service provision (Cardinale et al., 2012),
55 across many different habitat types. Identifying, preserving and restoring habitats of
56 conservation interest have been recognised as means to slow biodiversity loss. These
57 activities require knowledge of the status of organisms using the habitat, but resource
58 constraints rarely allow exhaustive studies of all the living things within systems of interest
59 (Heino, 2010). Using a small number of surrogate or indicator species to infer the
60 conservation status or biological diversity of a habitat or area has been suggested as an
61 achievable alternative (Lambeck, 1997; Landres et al., 1988; Hillard et al., 2017; Cheyne et
62 al., 2016).

Using species as indicators or surrogates in this way assumes cross-congruence, where the diversity or composition of many taxa are linked, so that changes to the diversity or composition of the broader biological community can be determined by surveying the indicator species alone (Westgate et al., 2017; Lindenmayer et al., 2014; Hunter et al., 2016). Cross congruence has been shown to occur in nature and is very influential in conservation planning and policy (Gioria et al., 2010; Howard et al., 1998b; Lund & Rahbek, 2002; Wolters et al., 2006). However, there is a growing body of published data which indicates that relationships between diversity of different taxa are variable (Westgate et al., 2014). For example, taxa which are very strongly associated with a particular habitat type are likely to be more congruent than generalist species (Prendergast, 1997), but in highly heterogeneous habitats, cross congruence may be inflated due to species turnover between habitats, regardless of their conservation value (Ekroos et al., 2013). Therefore species are most effective as indicators of habitat conditions or diversity if they are used within a specific, homogeneous habitat. In addition, habitat scale can affect the strength of cross congruence (Hess et al., 2006), but this can be mitigated by using taxa with small body-sizes in habitats characterised by small patches (Wolters et al., 2006). Disturbance is an important factor (Heino, 2010), as was demonstrated in a survey of Canadian wetlands where weak cross-congruence was associated with historical human disturbance, and individual species varied in their response to stress gradients (Rooney and Bayley, 2012).

Despite the observed variability in the strength and occurrence of cross congruence, biological indicators and surrogate species are embedded in conservation policies all over the world (US EPA 2002; "Fauna Wetland Indicator Species List" 2018; Ramsar Convention Secretariat 2010), including international conservation policies such as the European Union (EU) Habitats Directive (Council Directive 92/43/EEC). The ultimate aim of the Habitats Directive is that all of the habitats and species of conservation interest achieve favourable conservation status (Evans and Arvela, 2011). Conservation status is assessed under four parameters: range, area, structure and functions and future prospects. The use of positive and negative indicator species is central to the habitat structure and functions assessment. The characteristic species listed in guidance documents are predominantly plants, and this appears to have had a strong influence on the choice of indicator species among signatory states, although the authors encourage the inclusion on non-plant species in the structure and functions assessment (European topic centre on biological diversity). Clearly since this principle is being applied for monitoring of high-level conservation objectives, its efficacy should be assessed, particularly in habitats which experience frequent natural disturbance.

Thus in this study, we specifically addressed two key questions:

98 1) is cross congruence maintained in habitats of conservation interest where natural
99 disturbance occurs?

100 2) in a habitat which experiences natural disturbances, do indicator species provide reliable
101 information regarding diversity and habitat use by the wider community of organisms?

102 We sought to answer these questions by comparing the composition and diversity of taxa
103 with contrasting biological and habitat requirements in dune slacks, a habitat which
104 experiences natural disturbance. We applied the EU Habitats Directive habitat monitoring
105 guidelines for Irish dune slacks (Delaney et al., 2013) to assess dune slacks in Ireland and
106 tested the ability of this assessment to discriminate between dune slacks on the basis of two
107 invertebrate groups: water beetles and snails.

108 **1.2. Study System and target taxa**

109 Dune slacks are damp depressions in sand dune systems ranging in size from a few square
110 metres to several hectares. They arise from areas of low-lying bare sand such as former
111 beaches or erosion features within larger sand dune systems. Unlike the dry dunes that
112 surround them, the water table in dune slacks is close to the surface and it typically exceeds
113 the slack floor when groundwater rises in late winter (Davy et al., 2006). Although dune
114 slacks are dynamic and experience succession over their lifetime (Bossuyt et al., 2003;
115 Smith, 2006), extreme habitat heterogeneity is rare within an individual dune slack at a
116 single point in time unless there is a history of different management regimes or disturbance
117 patterns within the slack (Adema et al., 2002). The clear delimitation between dune slacks
118 and dry dune habitats and their relative lack of human disturbance correspond with the
119 conditions in which strong cross congruence might be expected. Conversely, annual
120 flooding and desiccation events act as natural disturbances and the presence of a dry phase
121 and an aquatic phase means that the habitat experiences temporal heterogeneity. This
122 natural disturbance could result in a breakdown of cross congruence despite the relatively
123 low human impacts on most dune slacks.

124 The conservation value of dune slack habitats has been recognised under the EU Habitats
125 Directive and they are designated as 2190 Humid dune slacks (Council Directive
126 92/43/EEC). They are recognised as being important refuges for some animal species and
127 the inclusion of animal indicator species in the structure and functions assessment is
128 encouraged (Evans and Arvela, 2011). Within the Atlantic biogeographical region of the EU,
129 the typical species listed for Ireland (Delaney et al., 2013), the UK (Common standards
130 monitoring guidance), France and Portugal do not contain any animal species (European
131 topic centre on biological diversity), while Germany, the Netherlands and Belgium have

132 included some bird species or the natterjack toad in their dune slack habitat assessments
133 (European topic centre on biological diversity) but typical species for Denmark are not
134 available.

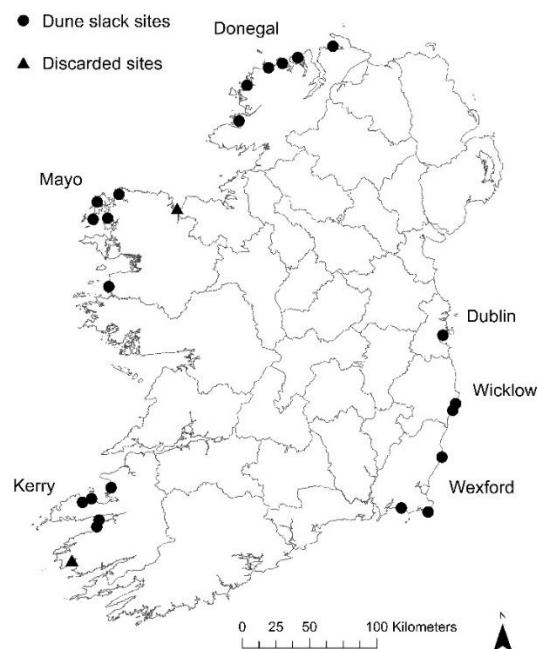
135 The three taxa chosen for this study were plants (vascular plants and bryophytes), snails
136 and water beetles. Both of the invertebrate taxonomic groups represent a subset of species
137 within large phyla which share specific characteristics; snails (class Gastropoda, phylum
138 Mollusca) have a shell into which their body can retract (Cameron, 2008) and water beetles
139 (Order Coleoptera, Class Insecta, Phylum Arthropoda) live underwater for part of their
140 lifecycle and so are likely to be caught in sweep-net samples (Foster et al., 2014). Vascular
141 plants and bryophytes are currently used to assess habitat condition in dune slacks in
142 accordance with the EU Habitats Directive in all of the signatory states. Water beetles and
143 snails were chosen because they have small body size, because they are well described
144 and can be identified to species, and because they are diverse groups which are sensitive to
145 environmental changes (Bilton et al., 2006). Both have been demonstrated to exhibit cross-
146 taxon congruence in ponds (Bilton et al., 2006; Gioria et al., 2010) and they are commonly
147 found in temporary waterbodies in Britain and Ireland (Nicolet et al., 2004; Reynolds, 1985).
148 However, because of differences in dispersal and lifecycles, water beetles and snails are
149 likely to differ in their responses to flooding and desiccation. Although records suggest that
150 dune slacks represent an important refuge for aquatic species due to the loss of inland
151 wetland habitats (Cameron, 2008; Foster et al., 2014; Foster & Friday, 2011; Davidson,
152 2014), comprehensive surveys of aquatic invertebrates in dune slacks have not been carried
153 out on a large scale in Europe or elsewhere. There is therefore no list of the aquatic
154 invertebrates associated with good habitat condition in dune slacks.

155 **2 Methods**

156 **2.1 Site selection**

157 We used pre-existing habitat maps (Delaney et al., 2013; Ryle et al., 2004) to select twenty-
158 four dune slacks across the Republic of Ireland (Figure 1). They were evenly distributed
159 among the regions containing the greatest concentration of dune slacks in Ireland: counties
160 Donegal, Mayo, Kerry, and the east coast from Dublin to Wexford. Selected slacks
161 conformed to a similar range in area from small (0.2–0.29 ha) to large (6–8 ha) within each
162 region. This research focussed on calcareous dune slacks with predominantly herbaceous
163 vegetation and a shallow water table which were located inside EU Natura 2000 Special
164 Areas of Conservation (SACs). Slacks with permanent water bodies, acidic conditions or
165 very sparse vegetation were not included. We did not include any slacks where human
166 interference in the substrate such as sand or sod removal was known to have occurred.

167 These exclusion criteria ensured that the most common dune slack types in Ireland were
168 available for selection while variability in the history, anthropogenic disturbance and
169 fundamental abiotic conditions affecting the slacks was limited. We selected only one slack
170 from any sand dune system to ensure that all sites were independent.



171
172 Figure 1 Dune slack sites included in this research. Reasons for discarding sites are given in section
173 3. NOTE: 1 COLUMN

174 **2.2 Vegetation survey**

175 We recorded vegetation in 2 m x 2 m quadrats from June to September 2015. Sampling
176 intensity was deemed adequate when the number of new vascular plant species recorded in
177 each new quadrat was equal to one or zero, resulting in 3 – 9 quadrats per site. We
178 recorded plant species cover in percent digitally in the field using *Turboveg for Windows*
179 2.120 (Hennekens and Schaminée, 2001) and nomenclature followed the Ireland2008v2
180 species list. If several vegetation communities, such as reed bed or *Salix repens* dunes,
181 occurred in a single slack, at least one quadrat was placed in each microhabitat, but
182 otherwise we placed the quadrats at random.

183 **2.3 Invertebrate surveys**

184 We collected snail samples once during the dry phase from July to September 2014 and
185 twice during the flooded phase in March and April 2015. During the dry phase, we removed
186 all leaf litter, vegetation and loose soil within a 25 cm x 25 cm quadrat adjacent to each

187 vegetation quadrat. We sampled snails at nine points evenly distributed in the dune slack
188 during the flooded phase by using a 0.5 mm mesh net to collect all loose material within a
189 box quadrat (23 cm x 27 cm) following O'Connor et al. (2004). All material collected was
190 dried and gastropods were isolated by passing the material through a 0.5 mm sieve (Long et
191 al., 2012). Individuals which had more than three complete whorls and a developed mouth
192 were treated as adults and identified to species following Cameron (2008). Juveniles could
193 not be identified reliably for all species, and so were not considered during the analysis.
194 Nomenclature followed Anderson (2005).

195 Water beetles were not surveyed during the dry phase as they do not remain in the habitat
196 when water is not present (Reynolds, 2003). In accordance with O'Connor et al. (2004) a
197 high sided quadrat measuring 40 cm x 35 cm was placed adjacent to each of the aquatic
198 snail quadrats in March and April 2015. Loose material was collected with a 0.5 mm net and
199 placed in a tray where the beetles were removed and stored in alcohol for identification.
200 Identification was carried out using Foster and Friday (2011), Foster et al. (2014) and Friday
201 (1988).

202 When comparing the diversity and composition patterns of water beetles with those of plants
203 and snails, we considered only those sites which flooded in the winter of 2014/2015 as water
204 beetles were only surveyed in flooded sites.

205 **2.4 Quantifying diversity**

206 Species-area curves approached the asymptote for plants at most sites, but water beetles
207 and snails were sparsely distributed at some sites and this resulted under-sampling of
208 waterbeetles at 56% and under-sampling of snails at 40% of sites. We used the iNEXT
209 package for R (R Development Core Team, 2013; Chao et al., 2014) to standardise samples
210 and calculate Hill numbers corresponding to species richness (H_0) and diversity (inverse
211 Simpson's concentration: H_2) (Chao et al., 2014). Inverse Simpson's concentration is a
212 useful complement to species richness because it is highly sensitive to dominance (Jost et
213 al., 2011). Standardisation was achieved by first estimating the coverage of the sample (the
214 likelihood that any individual taken from the total species pool at a site will belong to one of
215 the species included in the sample) at each site based on relative abundance (Chao and
216 Jost, 2012). The species richness or diversity was then rarefied or extrapolated to
217 correspond to 50% coverage to facilitate comparisons between sites. Fifty percent is the
218 minimum coverage required for an accurate comparison between groups, and did not
219 require extrapolating beyond the threshold for accuracy (extrapolation to twice the number of
220 samples actually recorded) (Chao and Lee, 1992). Estimated species richness represents a
221 minimum value, whereas the metric corresponding to Simpson's diversity is relatively

222 unbiased (Chao et al., 2014). At sites where four or fewer individuals were present, the
223 diversity index calculations should be interpreted with caution as they may not be an
224 accurate reflection of diversity. We consulted the Irish Red Lists to determine the IUCN
225 threat status category in Ireland for each species of plant, snail and water beetle we found
226 (Byrne et al., 2009; Foster et al., 2009; Lockhart et al., 2012; Wyse Jackson et al., 2016).

227 **2.5 Data analysis**

228 We performed data analyses in R (R Development Core Team 2013) with the aid of the data
229 analysis package Vegan (Oksanen et al., 2016). We used Kendall rank order correlation to
230 test whether a relationship could be detected between species richness and diversity of
231 plants, snails and water beetles in dune slack sites. A non-parametric method was
232 appropriate because sample sizes were uneven and the data distribution was non-normal
233 (Dytham, 2003). We visualised the composition of snail and plant species records at each
234 site with non-metric multidimensional scaling (NMS) which is a reliable method for zero-rich
235 data (Kent, 2011). Beetles were sparsely distributed and many species occurred only once
236 within the dataset. To compensate, we aggregated the beetle records to genus level before
237 performing NMS. This is reasonable since water beetle species within genera play
238 ecologically similar roles according to data describing their functional traits (Tachet et al.,
239 2010). We calculated the mean number of each species/genus recorded per quadrat at
240 each site and log-transformed the result ($\log_{10}(x + 1)$) to reduce the influence of extreme
241 values (McCune & Grace, 2002). We removed species/genera of plants or invertebrates
242 which only occurred once in the dataset prior to running the NMS, and any site which was
243 left with no species or a single species was also removed. Removing singleton species or
244 sites with a single species is recommended in ordination as they do not contribute useful
245 information for generating a matrix of similarity (McCune & Grace, 2002). We set the
246 number of axes to be generated in each case to 2 and the number analysis runs with
247 different random start configurations to 20. We used Procrustean rotation to assess how
248 similar the NMS plots for plants, snails and water beetles were and assessed the
249 significance of any similarities with the permutational test PROTEST.

250 **2.6 Cross-applicability of a conservation assessment based on plant communities**

251 We applied the 2190 Humid Dune Slacks Annex I habitat assessment used in Ireland
252 (Delaney et al. 2013) to data recorded in the field for each slack. The assessment was
253 designed to be applied over the entire dune slack habitat within a sand dune system, and we
254 modified it for application in a single dune slack. We retained all of the structure and
255 functions criteria with the exception of the criterion specifying the acceptable quantity of bare
256 ground (Table 1), as lack of bare ground within an individual dune slack is not an indication

257 of habitat degradation. We assessed the presence of positive indicator species listed on the
258 habitat assessment protocol (Delaney et al. 2013), which is dominated by plants of the
259 species-rich pioneer phase. We also assessed the presence of negative indicator species
260 which are typical of agricultural activities, abandonment and nutrient enrichment. We
261 considered the cover of bryophytes, scrub, non-native species and forb species and
262 environmental factors such as disturbance due to human activities and coastal defences
263 which have altered the sediment processes of the dune system. Ground water chemistry,
264 flood duration and dune slack morphology are not taken into account as part of the structure
265 and functions assessment.

266 We used the non-parametric Mann-Whitney U statistic to compare the diversity (estimated
267 richness: H_0 ; and diversity: H_2) of snails and beetles at sites which failed with those that
268 passed the habitat assessment. PERMANOVA analyses were carried out to compare the
269 distributions of plants, snails and water beetle species in sites which passed and failed the
270 2190 Humid Dune Slacks habitat assessment (distance measure: Bray Curtis, 1000
271 permutations).

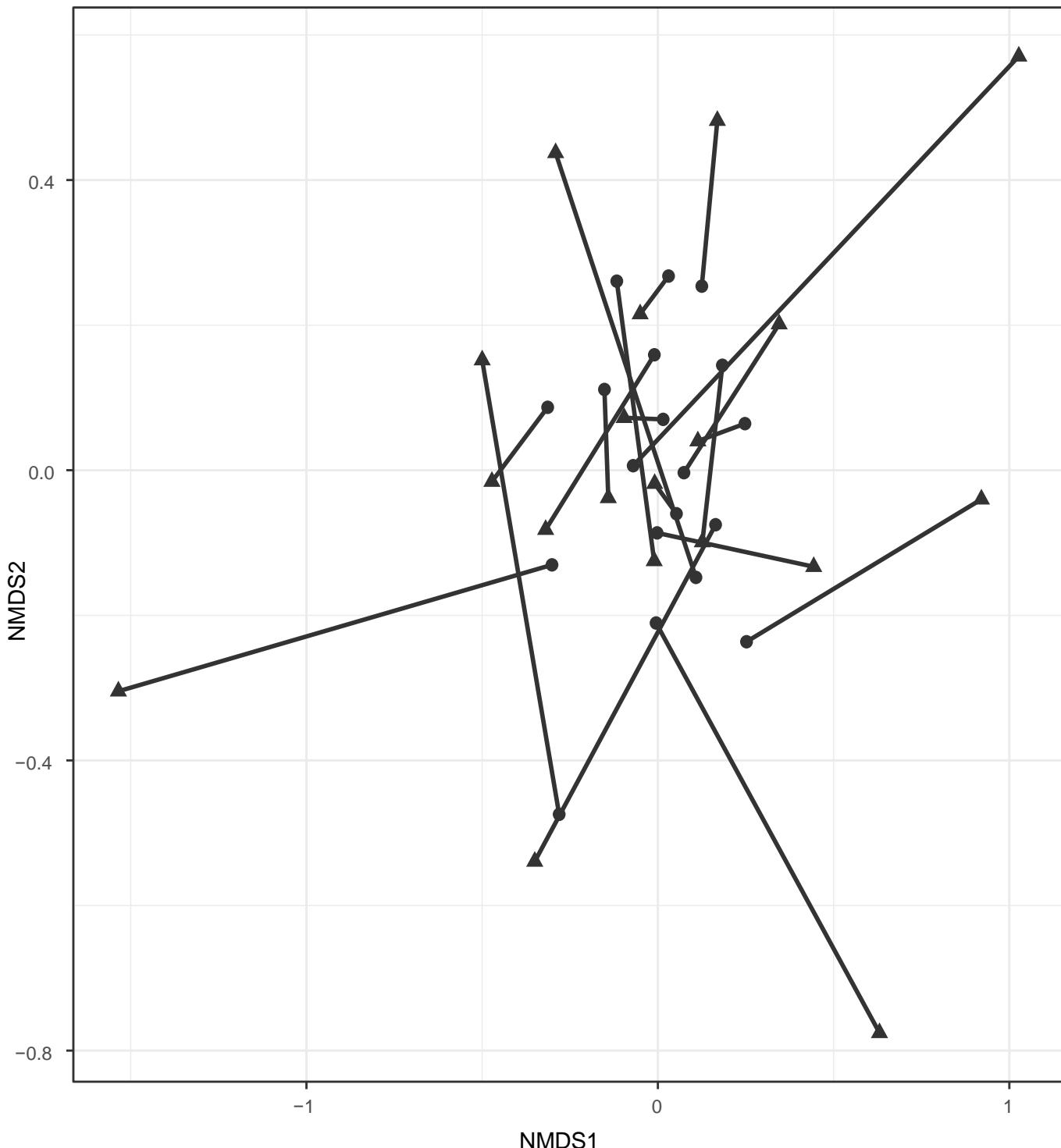
272 **3 Results**

273 The data recorded from quadrats at each site have been submitted to the Irish National
274 Biodiversity Data Centre. We excluded one site from the project because it was inaccessible
275 for much of the year and another because, on inspection, it did not contain dune slack
276 habitat. Six of the sites failed to flood during the monitoring year, so while snails and plants
277 were collected at 22 sites, water beetles sampling was carried out at 16 sites. Plants were
278 the most species rich group with 195 species, followed by snails (33 species) and water
279 beetles (23 species). Abundance of beetles and snails collected at each site was very
280 variable. No beetles were found at three sites and fewer than four individuals were collected
281 at seven of the remaining sites. More than four snails were sampled at 18 sites, and no
282 snails were recovered at one site.

283 There was no significant correlation between estimated richness (H_0) and diversity (H_2) of
284 plant, snail and water beetle assemblages within a site (Table 2).

285 The best fit NMS plots for plants had a stress value of 0.15, for snails 0.14 and for beetles
286 0.09, and stress values of less than 0.2 indicate a good two-dimensional representation of
287 multivariate data (Clarke, 1993). The procrustean rotation plots for plants and snails show
288 that there is considerable distance between the locations of sites in the ordination plot
289 when it was at the optimal or closest matching rotation (Figure 2) and no evidence for cross-
290 congruence was detected (sum of squares = 0.85, Procrustes correlation = 0.39, $p = 0.13$).

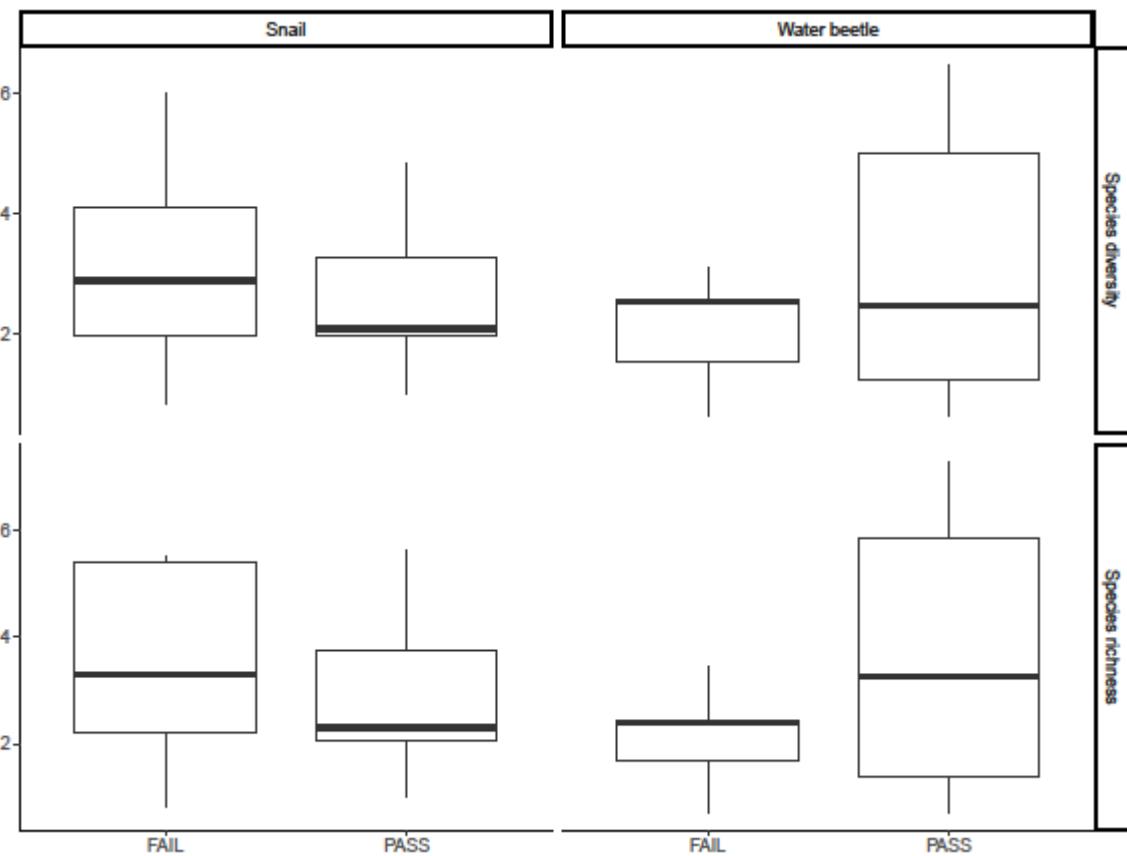
291 Nor was there evidence of cross congruence between water beetle assemblages and those
292 of plants (sum of squares = 0.74, Procrustes correlation = 0.51, p = 0.23) or snails (sum of
293 squares = 0.80, Procrustes correlation = 0.45, p = 0.39).



294
295 Figure 2 Ordination plot of plant assemblages at the sites surveyed (triangles). Triangles which are
296 located close to each other in ordination space represent sites with similar plant composition. The
297 ordination plot for snail species (dots) has been rescaled, rotated and superimposed on the plant
298 assemblage ordination in the configuration showing the greatest possible similarity between plots as

299 described by their plant and snail assemblages. Each site is represented as a dot (snail assemblage)
300 and a triangle (plant assemblage) joined by a line. The presence of long lines in relation to the NMS
301 axis lengths shows that the configuration of plant and snail assemblages within the sites surveyed
302 were not similar.

303 Twelve sites passed and 10 sites failed the habitat assessment based on vascular plants.
304 No significant difference was detected in the estimated species richness (H_0) of snails ($U =$
305 61, $p < 0.67$) or water beetles ($U = 17$, $p = 0.71$) or in the diversity (H_2) of snails ($U = 62$, $p =$
306 0.62) or water beetles ($U = 17$, $p > 0.71$) recorded at sites which passed and failed the
307 assessment (Figure 3). The species compositions of snails ($F = 0.67$, $p = 0.79$) and beetles
308 ($F = 1.55$, $p = 0.17$) did not differ significantly at sites which passed and failed the habitat
309 assessment. NOTE: 1 COLUMN



310
311 Figure 3 Diversity (Inverse Simpson's concentration, H_2) and richness (H_0) of snail and water beetle
312 species in sites which passed and failed the conservation assessment. The bold horizontal line
313 indicates the median value and boxes represent the interquartile range. The total range of values
314 recorded is shown as a vertical line above and below the box. NOTE: 2 COLUMNS

315 Snail species of conservation concern were found at eleven sites which passed and seven
316 sites which failed the habitat assessment. Of these species, six were listed as Vulnerable
317 and two as Near Threatened on the Irish Red List for snails (Byrne et al., 2009). Among

318 water beetles, *Enochrus halophilus* is listed as Vulnerable and was found in a site which
319 passed the assessment while *Dryops similis* is listed as Near Threatened and was found in
320 a site which failed the habitat assessment (Foster et al., 2009). The remaining species were
321 listed as being of Least Concern. All of the plant species found were of Least Concern
322 according to the Irish Red List for vascular plants and bryophytes (Lockhart et al., 2012;
323 Wyse Jackson et al., 2016).

324 **4 Discussion**

325 The dune slacks we surveyed generally contained diverse plant communities of wetlands
326 and sandy places which are typical of the habitat in Ireland and Britain (Fossitt, 2000;
327 Rodwell et al., 2000). There are no published records of large scale surveys of aquatic
328 invertebrate species in temperate dune slacks, and the number of coleopteran and
329 gastropod species recorded in other temporary water bodies can be highly variable. That
330 being said, snail abundance and diversity were high in relation to some other temporary
331 waterbodies (Nicolet et al., 2004; Porst and Irvine, 2009), although this may be partly due to
332 having surveyed both during the wet and dry phases rather than focussing on one or the
333 other. In comparison to surveys of other temporary freshwaters in Ireland and Britain, the
334 dune slacks we surveyed were relatively poor in water beetle species (Nicolet, 2004;
335 Collinson et al., 1995; Porst and Irvine, 2009). Plants and snails were more abundant than
336 water beetles in the dune slacks we surveyed and the results of analyses carried out on
337 them should carry greater weight.

338 Although we minimised factors associated with reduced cross-congruence by using
339 invertebrates with a small body size, conducting the research in a specific habitat type and
340 selecting sites within a conservation network to minimise the likelihood of human impacts
341 (Prendergast, 1997; Wolters et al., 2006; Rooney and Bayley, 2012), we did not find
342 evidence of cross-congruence in estimated diversity or in composition among plants, snails
343 and water beetles in Irish dune slacks. The taxa may be subject to different community
344 assembly processes due to their differing biology, or they may respond differently to the
345 abiotic forces affecting the dune slacks (Saetersdael & Gjerde, 2011). A likely explanation is
346 that the three taxonomic groups respond differently to the natural disturbance of periodic
347 flooding due to differences in their biological requirements. A fluctuating water table is
348 required to maintain the typical plant community of dune slacks (Davy et al., 2006, Currelli et
349 al., 2013, Grootjans et al., 1991), whereas the end of the flooded phase is a major
350 disturbance for water beetles, and dune slacks must be recolonised each flood season. A
351 thorough investigation of the role of community assembly forces including habitat filtering is
352 beyond the scope of this paper.

353 Because plant, snail and water beetle species are not cross-congruent in dune slacks,
354 managing the habitat to protect or restore a specific plant community will not necessarily
355 protect the snails and water beetles found there. For example, three of the six sites which
356 were not observed to flood in 2014/2015 passed the conservation assessment. According to
357 data from Met Éireann, the Irish meteorological service, rainfall in 2014 exceeded the annual
358 average of the preceding 30 years in all of the regions where we selected sites, so lower
359 than average water tables were not expected during the survey period. Four of the sites
360 which failed to flood were on the east coast of Ireland where water resources are thought to
361 experience greater pressure due to urban and tourist developments (Ryle et al., 2009,
362 Delaney et al., 2013), and so reduced frequency or duration of flooding could be linked to
363 human activities. Drying out is a major threat to European dune slacks (Grootjans and
364 Stuyfzand, 1998) and it is of concern that this monitoring system failed to identify sites that
365 no longer flood on an annual basis and therefore do not reliably provide habitat for aquatic
366 species. This could lead to missed opportunities for conservation interventions and
367 subsequent loss of species from some sites. In addition, annual flooding has been shown to
368 be important for the long term maintenance of the calcareous wetland vegetation community
369 of dune slacks elsewhere in Europe. By the time major vegetation changes are noted, the
370 dune slack system may have been damaged beyond the point of restoration to its normal
371 functioning (Grootjans et al., 1991). These results suggest that the use of a set of indicator
372 species composed entirely of plants is inadequate to protect the long-term functioning of
373 dune slack habitats.

374 There is no reference list of invertebrates associated with ideal conditions in dune slacks, but
375 species classified as Vulnerable on the Irish Red Lists were found at six of the ten sites
376 which failed the habitat assessment. Although these are not necessarily dune slack
377 specialists, two were aquatic species and four were wetland species (Foster et al., 2009;
378 Byrne et al., 2009), so they are likely to be particularly associated with the humid conditions
379 in dune slacks rather than the sand dune habitat as a whole. Because eight of the species
380 of conservation interest occurred at sites which failed the conservation assessment, they
381 may be subjected to conservation measures which are not designed to protect them (e.g.
382 NPWS, 2014).

383 The presence of eight snail species and two water beetle species listed as Near Threatened
384 or Vulnerable on Irish Red Lists indicates that dune slacks provide important habitat for
385 invertebrates whose populations are in decline. The Irish populations of two of the species
386 recorded here are of international significance (*Vertigo angustior*, *Leiostyla anglica*). Many
387 of the wetland and aquatic species on the Irish Red List were formerly found throughout
388 Ireland but are now most commonly found in coastal locations (Byrne et al., 2009, Foster et

389 al., 2009), and the importance of coastal wetlands as refuges for declining species is not
390 merely an Irish phenomenon. Globally, wetland losses are estimated at between 54% and
391 57 % since 1700, most losses have occurred since 1900 and loss has occurred at a greater
392 rate inland than in coastal regions (Davidson, 2014). Coastal wetlands such as dune slacks,
393 Mediterranean temporary ponds and marshes therefore represent important refuges for
394 freshwater wetland species affected by habitat loss. Measures such as the creation of areas
395 of conservation in line with the EU Habitats Directive have the potential to prevent further
396 biodiversity loss on an international scale (Hochkirch et al. 2013) but guidance on the choice
397 of monitoring tools including indicator species should take the dual aquatic and terrestrial
398 aspects of wetlands into account to avoid the implementation of inappropriate conservation
399 measures.

400 The increasingly precarious position of much life on earth has been highlighted in the
401 publication of the World Wildlife Fund's Living Planet Report (World Wildlife Fund, 2016), a
402 report that echoes themes in academic literature (Cardinale et al., 2012; Díaz et al., 2006).
403 Wetlands and their dependent taxa are among many that have suffered losses. Large-scale
404 interventions to conserve habitats and species such as the Habitats Directive are to be
405 welcomed but it is important to test them and ensure that the resources invested in
406 conservation yield the best possible result. In contrast to the results reported here, the
407 species recommended for protection under Annexes II and IV of the Habitats Directive have
408 proved to be potentially reliable umbrella species in conservation planning (Lund, 2002).
409 This is likely to relate to the fact that Articles II and IV include a wide range of taxa, so the
410 conservation sites chosen on the basis of these species would contain a range of
411 environmental conditions suitable for a variety of taxonomic groups. Our results show that
412 choosing indicator species from more than a single taxonomic group would improve the
413 sensitivity of monitoring schemes in dune slacks. Further research is required to determine
414 whether this is also the case in other habitats which experience disturbance as part of their
415 natural functioning.

416

417 **References**

- 418 Anderson, R., (2005) An annotated list of the non-marine molluscs of Britain and Ireland.
419 Journal of Conchology. 38 (6): 607–637.
- 420 Adema, E.B., Grootjans, A.P., Petersen, J., Grijpstra, J., (2002) Alternative stable states in a
421 wet calcareous dune slack in The Netherlands. J. Veg. Sci. 13, 107–114.
422 <https://doi.org/10.1111/j.1654-1103.2002.tb02027.x>
- 423 Bilton, D.T., McAbendroth, L., Bedford, A., Ramsay, P.M., (2006) How wide to cast the net?
424 Cross-taxon congruence of species richness, community similarity and indicator taxa in
425 ponds. Freshw. Biol. 51, 578–590. <https://doi.org/10.1111/j.1365-2427.2006.01505.x>
- 426 Bossuyt, B., Honnay, O., Hermy, M., (2003) An island biogeographical view of the
427 successional pathway in wet dune slacks. J. Veg. Sci. 14, 781–788.
- 428 Byrne, A., Moorkens, E.A., Anderson, R., Killeen, I.J. & Regan, E.C., (2009) Ireland Red List
429 No. 2 – Non-Marine Molluscs. National Parks and Wildlife Service, Department of the
430 Environment, Heritage and Local Government, Dublin, Ireland.
- 431 Cameron, R., (2008) Land Snails in the British Isles. Field Studies Council, Ivor, United
432 Kingdom.
- 433 Cardinale, B.J., Duffy, J.E., Gonzalez, A., Hooper, D.U., Perrings, C., Venail, P., Narwani,
434 A., Mace, G.M., Tilman, D., Wardle, D.A., Kinzig, A.P., Daily, G.C., Loreau, M., Grace, J.B.,
435 Larigauderie, A., Srivastava, D.S., Naeem, S., (2012) Biodiversity loss and its impact on
436 humanity. Nature 486, 59–67. doi: 10.1038/nature11148
- 437 Chao, A., Gotelli, N. J., Hsieh, T. C., Sander, E. L., Ma, K. H., Colwell, R. K. and Ellison, A.
438 M. (2014) Rarefaction and extrapolation with Hill numbers: A framework for sampling and
439 estimation in species diversity studies, Ecological Monographs, 84(1), pp. 45–67. doi:
440 10.1890/13-0133.1.
- 441 Chao, A., & Jost, L. (2012) Coverage-based rarefaction and extrapolation: Standardizing
442 samples by completeness rather than size. Ecology, 93(12), 2533–2547.
443 <http://doi.org/10.1890/11-1952.1>
- 444 Chao, A., Lee, S. M. (1992) Estimating the number of classes via sample coverage. J. Am.
445 Stat. Assoc. 87, 210–217. doi:10.1080/01621459.1992.10475194

- 446 Cheyne, S. M., Sastramidjaja, W. J., Muhalir, Rayadin, Y., & Macdonald, D. W. (2016)
447 Mammalian communities as indicators of disturbance across Indonesian Borneo. Global
448 Ecology and Conservation, 7, 157–173. <https://doi.org/10.1016/j.gecco.2016.06.002>
- 449 Clarke, K. R. (1993) Non-parametric multivariate analysis of changes in community
450 structure. Austral J Ecol 18:117-143.
- 451 Collinson, N. H., Biggs, J., Corfield, A., Hodson, M. J., Walker, D., Whitfield, M. and
452 Williams, P. J., (1995) Temporary and permanent ponds :an assessment of the effects of
453 drying out on the conservation value of aquatic macroinvertebrate communities, Biological
454 Conservation, 74, pp. 125–133.
- 455 Joint Nature Conservation Committee (2004) Common standards monitoring guidance for
456 sand dune habitats http://jncc.defra.gov.uk/pdf/CSM_coastal_sand_dune.pdf
- 457 Curreli, A., Wallace, H., Freeman, C., Hollingham, M., Stratford, C., Johnson, H. and Jones,
458 L. (2013) Eco-hydrological requirements of dune slack vegetation and the implications of
459 climate change. Science of the total environment. Elsevier B.V., 443, pp. 910–9. doi:
460 10.1016/j.scitotenv.2012.11.035.
- 461 Davidson, N.C., (2014) How much wetland has the world lost? Long-term and recent trends
462 in global wetland area. Mar. Freshw. Res. 65, 934. doi: 10.1071/MF14173
- 463 Davy, A. J., Grootjans, A. P., Hiscock, K. and Petersen, J. (2006) Development of eco-
464 hydrological guidelines for dune habitats – Phase 1, English Nature Research Reports,
465 (696).
- 466 Delaney, A., Devaney, F.M, Martin, J.M. and Barron, S.J. (2013) Monitoring survey of Annex
467 I sand dune habitats in Ireland. Irish Wildlife Manuals, No. 75. National Parks and Wildlife
468 Service, Department of Arts, Heritage and the Gaeltacht, Dublin, Ireland.
- 469 Diaz, S., Fargione, J., Chapin, F.S., Tilman, D., (2006) Biodiversity loss threatens human
470 well-being. PLoS Biol. 4, e277. doi:10.1371/journal.pbio.0040277
- 471 Dytham, C., (2003) Choosing and using statistics a biologists guide 2nd edition, Blackwell
472 publishing, Oxford 179:180
- 473 Ekroos, J., Kuussaari, M., Tiainen, J., Heliölä, J., Seimola, T., Helenius, J., (2013)
474 Correlations in species richness between taxa depend on habitat, scale and landscape
475 context. Ecol. Indic. 34, 528–535. doi:10.1016/j.ecolind.2013.06.015

- 476 Evans, D., Arvela, M., (2011) Assessment and reporting under Article 12 of the Birds
477 Directive Explanatory Notes & Guidelines for the period 2008-2012 Final Version December
478 2011.
- 479 European topic centre on biological diversity, <https://bd.eionet.europa.eu/>, accessed 23
480 October 2016.
- 481 Fauna Wetland Indicator Species List, WetlandInfo, Department of Environment and
482 Heritage Protection, Queensland,
483 <<https://wetlandinfo.ehp.qld.gov.au/wetlands/ecology/components/fauna/fauna-indicator-species-list.html>>, accessed 12 February 2018.
- 485 Fossitt, J.A., (2000) A Guide to Habitats in Ireland. The Heritage Council An Chomhairle
486 Oidhreachta, Kilkenny.
- 487 Foster, G. N., Nelson, B. H. & O Connor, Á., (2009) Ireland Red List No. 1 – Water beetles.
488 National Parks and Wildlife Service, Department of Environment, Heritage and Local
489 Government, Dublin, Ireland.
- 490 Foster, G. N., Bilton, D. T., Friday, L., (2014) Keys to adults of the water beetles of Britain
491 and Ireland. Part 2, (Coleoptera: Polyphaga: Hydrophiloidea - both aquatic and terrestrial
492 species). Telford: Published for the Royal Entomological Society by the Field Studies
493 Council.
- 494 Foster, G. N., Friday, L., (2011) Keys to adults of the water beetles of Britain and Ireland.
495 Part 2. Telford : Published for the Royal Entomological Society by the Field Studies Council.
- 496 Friday, L., (1998) A key to the adults of British water beetles, Field Studies Council, Taunton,
497 Somerset.
- 498 Gioria, M., Schaffers, A., Bacaro, G., Feehan, J., (2010) The conservation value of farmland
499 ponds: Predicting water beetle assemblages using vascular plants as a surrogate group.
500 Biol. Conserv. 143, 1125–1133. doi:10.1016/j.biocon.2010.02.007
- 501 Grootjans, A. P., Hartog, P. S., Fresco, L. F. M. and Esselink, H., (1991) ‘Succession and
502 fluctuation in a wet dune slack in relation to hydrological changes’, Journal of Vegetation
503 Science, 2, pp. 545–554.
- 504 Grootjans, A.P., Stuyfzand, P.J., (1998) European dune slacks : strong interactions of
505 biology, pedogenesis and hydrology. TREE 13, 96–100.

- 506 Heino, J., (2010) Are indicator groups and cross-taxon congruence useful for predicting
507 biodiversity in aquatic ecosystems? *Ecol. Indic.* 10, 112–117.
508 doi:10.1016/j.ecolind.2009.04.013
- 509 Hennekens, S.M., Schaminée, J.H.J., (2001) TURBOVEG, a comprehensive data base
510 management system for vegetation data. *J. Veg. Sci.* 12, 589–591. doi: 10.2307/3237010
- 511 Hess, G.R., Bartel, R.A., Leidner, A.K., Rosenfeld, K.M., Rubino, M.J., Snider, S.B., Ricketts,
512 T.H., (2006) Effectiveness of biodiversity indicators varies with extent, grain, and region.
513 *Biol. Conserv.* 132, 448–457. doi:10.1016/j.biocon.2006.04.037
- 514 Hillard, M., Nielsen C.K., Groninger, J.W., (2017) Swamp rabbits as indicators of wildlife
515 habitat quality in bottomland hardwood forest ecosystems. *Ecological Indicators* 79, 47 - 53
- 516 Hochkirch, A., Schmitt, T., Beninde, J., Hiery, M., Kinitz, T., Kirschen, J., Matenaar, D.,
517 Rohde, K., Stoefen, A., Wagner, N., Zink, A., Lötters, S., Veith, M., Proelss, A., (2013)
518 Europe needs a new vision for a Natura 2020 network. *Conserv. Lett.* 6, 462–467.
519 doi:10.1111/conl.12006
- 520 Howard, P.C., Viskanic, P., Davenport, T.R.B., Kigenyi, F.W., Baltzer, M., Dickinson, C.J.,
521 Lwanga, J.S., Matthews, R.A., Balmford, A., (1998) Complementarity and the use of
522 indicator groups for reserve selection in Uganda. *Nature* 394, 472–475. doi: 10.1038/28843;
- 523 Hunter, M., Westgate, M., Barton, P., Calhoun, A., Pierson, J., Tulloch, A., Beger, M.,
524 Branquinho, C., Caro, T., Gross, J., Heino, J., Lane, P., Longo, C., Martin, K., McDowell,
525 W.H., Mellin, C., Salo, H., Lindenmayer, D., (2016) Two roles for ecological surrogacy:
526 Indicator surrogates and management surrogates. *Ecol. Indic.* 63, 121–125.
527 doi:10.1016/j.ecolind.2015.11.049
- 528 Jost, L., Chao., A., Chazdon L.R., (2011) Compositional similarity and beta diversity. In
529 Magurran & McGill (eds) *Biological diversity: frontiers in measurement and assessment*,
530 Oxford University Press pp. 66 - 84
- 531 Kent, M., (2011) *Vegetation description and data analysis: A practical approach*, 2nd Edition,
532 Wiley-Blackwell, Oxford
- 533 Lambeck, R.J., (1997) A Multi-Species Umbrella for Nature Conservation. *Conserv. Biol.* 11,
534 849–856.
- 535 Landres, P.B., Verner, J., Thomas, J.W., (1988) Ecological uses of vertebrate indicator
536 species : A Critique. *Conserv. Biol.* 2, 316–329.

- 537 Liang, J., Crowther, T.W., Picard, N., Wiser, S., Zhou, M., Alberti, G., Schulze, E.-D.,
538 McGuire, A.D., Bozzato, F., Pretzsch, H., de-Miguel, S., Paquette, A., Héault, B., Scherer-
539 Lorenzen, M., Barrett, C.B., Glick, H.B., Hengeveld, G.M., Nabuurs, G.-J., Pfautsch, S.,
540 Viana, H., Vibrans, A.C., Ammer, C., Schall, P., Verbyla, D., Tchebakova, N., Fischer, M.,
541 Watson, J. V., Chen, H.Y.H., Lei, X., Schelhaas, M.-J., Lu, H., Gianelle, D., Parfenova, E.I.,
542 Salas, C., Lee, E., Lee, B., Kim, H.S., Bruehlheide, H., Coomes, D.A., Piotto, D., Sunderland,
543 T., Schmid, B., Gourlet-Fleury, S., Sonké, B., Tavani, R., Zhu, J., Brandl, S., Vayreda, J.,
544 Kitahara, F., Searle, E.B., Neldner, V.J., Ngugi, M.R., Baraloto, C., Frizzera, L., Bałazy, R.,
545 Oleksyn, J., Zawiła-Niedźwiecki, T., Bouriaud, O., Bussotti, F., Finér, L., Jaroszewicz, B.,
546 Jucker, T., Valladares, F., Jagodzinski, A.M., Peri, P.L., Gonmadje, C., Marthy, W., O'Brien,
547 T., Martin, E.H., Marshall, A.R., Rovero, F., Bitariho, R., Niklaus, P.A., Alvarez-Loayza, P.,
548 Chamuya, N., Valencia, R., Mortier, F., Wortel, V., Engone-Obiang, N.L., Ferreira, L. V.,
549 Odeke, D.E., Vasquez, R.M., Lewis, S.L., Reich, P.B., (2016) Positive biodiversity-
550 productivity relationship predominant in global forests. *Science* 354 (3609).
- 551 Lindenmayer, D.B., Lane, P.W., Westgate, M.J., Crane, M., Michael, D., Okada, S., Barton,
552 P.S., (2014) An Empirical Assessment of the Focal Species Hypothesis. *Conserv. Biol.* 28,
553 1594–1603. doi:10.1111/cobi.12330
- 554 Lockhart, N., Hodgetts, N. & Holyoak, D., (2012) Ireland Red List No.8: Bryophytes. National
555 Parks and Wildlife Service, Department of Arts, Heritage and the Gaeltacht, Dublin, Ireland.
- 556 Long M.P., Moorkens E.A., Kelly, D.L., (2012) The short-term impacts of cessation of
557 grazing on plants and land snails in grasslands in the west of Ireland. *Dry Grasslands of*
558 *Europe: Grazing and Ecosystem Services. Proceedings of 9th European Dry Grassland*
559 *Meeting (EDGM)*. Eds: Vrahnakis M., A.P. Kyriazopoulos, D. Chouvardas and G. Fotiadis,
560 Prespa, Greece, pp 48 – 54.
- 561 Lund, M.P., (2002) Performance of the species listed in the European Community “Habitats”
562 Directive as indicators of species richness in Denmark. *Environ. Sci. Policy* 5, 105–112. doi:
563 10.1016/S1462-9011(02)00031-X
- 564 Lund, M.P., Rahbek, C., (2002) Cross-taxon congruence in complementarity and
565 conservation of temperate biodiversity. *Anim. Conserv.* 5, 163–171. doi:
566 10.1017/S1367943002002226
- 567 McCune, B. and Grace, J.B., (2002) Analysis of Ecological Communities. MjM Software,
568 Gleneden Beach, Oregon, USA (www.pcord.com) 304 pages. ISBN: 0-9721290-0-6

569 Nicolet, P., Biggs, J., Fox, G., Hodson, M.J., Reynolds, C., Whitfield, M., Williams, P., (2004)
570 The wetland plant and macroinvertebrate assemblages of temporary ponds in England and
571 Wales. *Biol. Conserv.* 120, 265–282. doi:10.1016/j.biocon.2004.03.010

572 NPWS, (2014) Ballyness Bay SAC (site code 1090) Conservation objectives supporting
573 document -coastal habitats.
574 [https://www.npws.ie/sites/default/files/publications/pdf/Ballyness%20Bay%20SAC%20\(001090\)%20Conservation%20objectives%20supporting%20document%20-%20marine%20habitats%20\[Version%201\].pdf](https://www.npws.ie/sites/default/files/publications/pdf/Ballyness%20Bay%20SAC%20(001090)%20Conservation%20objectives%20supporting%20document%20-%20marine%20habitats%20[Version%201].pdf), accessed 23 October 2016

577 O'Connor, Á., Bradish, S., Reed, T., Moran, J., Regan, E., Visser, M., Gormally, M. and
578 Skeffington, M. S., (2004) 'A comparison of the efficacy of pond-net and box sampling
579 methods in turloughs – Irish ephemeral aquatic systems', *Hydrobiologia*, 524(1), pp. 133–
580 144. doi: 10.1023/B:HYDR.0000036128.83998.44.

581 Oksanen, J., Blanchet, G.F., Friendly, M., Kindt R., Legendre, P., McGlinn, D., Minchin P. R.,
582 O'Hara, R. B., Simpson G.L., Solymos P., Stevens, M.H., Szoecs E. and Wagner H., (2016).
583 vegan: Community Ecology Package. R package version 2.4-0. <https://CRAN.R-project.org/package=vegan>, Accessed: 22 October 2017

585 Porst, G., Irvine, K., (2009). Distinctiveness of macroinvertebrate communities in turloughs
586 (temporary ponds) and their response to environmental variables. *Aquat. Conserv. Mar.*
587 *Freshw. Ecosyst.* 19, 456–465. doi: 10.1002/aqc

588 Prendergast, J., (1997). Species richness covariance in higher taxa : empirical tests of the
589 biodiversity indicator concept. *Ecography (Cop.)*. 210–217.

590 R Core Team (2013). R: A language and environment for statistical computing. R
591 Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.

592 Ramsar Convention Secretariat 2010,
593 <https://www.ramsar.org/sites/default/files/documents/pdf/lib/hbk4-18.pdf>, accessed 12
594 February 2018

595 Reynolds, J.D., (2003) Fauna of turloughs and other wetlands. In: Otte, R. (ed.) *Irish*
596 *Wetlands: Distribution, ecology, uses and economic value*. University College Dublin Press.

597 Rodwell, J.S., Pigott, D., Joint Nature Conservation Committee (Great Britain), (2000).
598 British plant communities. Volume 5, Maritime communities and vegetation of open habitats.
599 Cambridge University Press.

- 600 Rooney, R.C., Bayley, S.E., (2012). Community congruence of plants, invertebrates and
601 birds in natural and constructed shallow open-water wetlands: Do we need to monitor
602 multiple assemblages? *Ecol. Indic.* 20, 42–50. doi:10.1016/j.ecolind.2011.11.029
- 603 Ryle, T., Murray, A., Connolly, K., Swann, M., (2009). Coastal Monitoring Project 2004-2006.
604 Unpublished report of the Irish National Parks and Wildlife Service.
605 https://www.npws.ie/sites/default/files/publications/pdf/Ryle_et_al_2009_Coastal_Monitoring_Project.pdf
- 607 Sætersdal, M., Gjerde, I., 2011. Prioritising conservation areas using species surrogate
608 measures: Consistent with ecological theory? *J. Appl. Ecol.* 48, 1236–1240.
609 doi:10.1111/j.1365-2664.2011.02027.x
- 610 Smith, P.H., 2006. Changes in the floristic composition of sand-dune slacks over a twenty-
611 year period. *Watsonia* 26, 41–49.
- 612 Tachet, H., Richoux, P., Bournaud, M. & Usseglio-Polatera, P., (2000) Invertébrés d'eau
613 douce: systématique, biologie, écologie. CNRS Editions, Paris
- 614 U.S. EPA, 2002. Biological Assessments and Criteria: Crucial Components of Water Quality
615 Programs. United States Environmental Protection Agency, Office of Water, Washington,
616 DC.
- 617 Westgate, M.J., Barton, P.S., Lane, P.W., Lindenmayer, D.B., (2014) Global meta-analysis
618 reveals low consistency of biodiversity congruence relationships. *Nat. Commun.* 5, 3899.
619 doi: 10.1038/ncomms4899
- 620 Westgate, M.J., Tulloch, A.I.T., Barton, P.S., Pierson, J.C., Lindenmayer, D.B., (2017)
621 Optimal taxonomic groups for biodiversity assessment: A meta-analytic approach.
622 *Ecography* (Cop.). 40, 539–548.
- 623 Wolters, V., Bengtsson, J., Zaitsev, A.S., Zaitsev, S., (2006) Relationship among the species
624 richness of different taxa. *Ecology* 87, 1886–1895.
- 625 WWF (2016) Living Planet Report 2016. Risk and resilience in a new era. WWF
626 International, Gland, Switzerland
- 627 Wyse Jackson, M., FitzPatrick, U., Cole, E., Matthew, J., McFerran, D., Skeffington, M.S.,
628 Wright, M., (2016) Vascular Plants. Irel. Red List No. 10 141

629 Table 1 Habitat assessment criteria modified from Delaney et al. (2013) for application at a single
 630 dune slack. NOTE: 2 COLUMNS

Assessment Criterion	Pass requirement
Positive indicator species	Four species present in over 40% of quadrats and two further species present in more than 20% of quadrats. At least three positive indicator species must be present in every quadrat.
Negative indicator species	No species present in more than 60% of quadrats or with a cover within a quadrat of greater than 25%. Combined cover of negative indicators 5% or less.
Cover of scrub	Scrub present in no more than 40% of quadrats and combined cover of 5% or less
Cover of <i>Salix repens</i>	A mean of less than 40% cover in quadrats
Frequency of bryophytes	Present in more than 20% of quadrats
Non-native species	No non-native species present in more than 20% of quadrats
Forb: grass cover ratio	Forb cover over 30%, grass cover below 70%
Disturbance	Impacts such as trampling, vehicle damage etc. affecting no more than 20% of habitat
Coastal defences	No anthropogenic impacts on the substrate/mobility of the system resulting from coastal defence works put in place or modified subsequent to the site's designation as an SAC, or from sediment extraction in the same period.

631

632 Table 2 Correlation tests for relationships between estimated species richness, diversity and
 633 composition of snails, plants and beetles. NOTE: 2 COLUMNS

Test	Subjects	Z-score	Tau	P value
Kendall rank correlation of estimated species richness (H_o)	Plants and snails	-0.33	-0.05	0.74
	Plants and beetles	0.62	0.13	0.54
	Snails and beetles	-0.24	-0.05	0.81
Kendall rank correlation of estimated diversity (H_2)	Plants and snails	-0.76	-0.11	0.45
	Plants and beetles	0.37	0.08	0.71
	Snails and beetles	-0.24	-0.05	0.81

634

● Dune slack sites

▲ Discarded sites

Donegal

Mayo

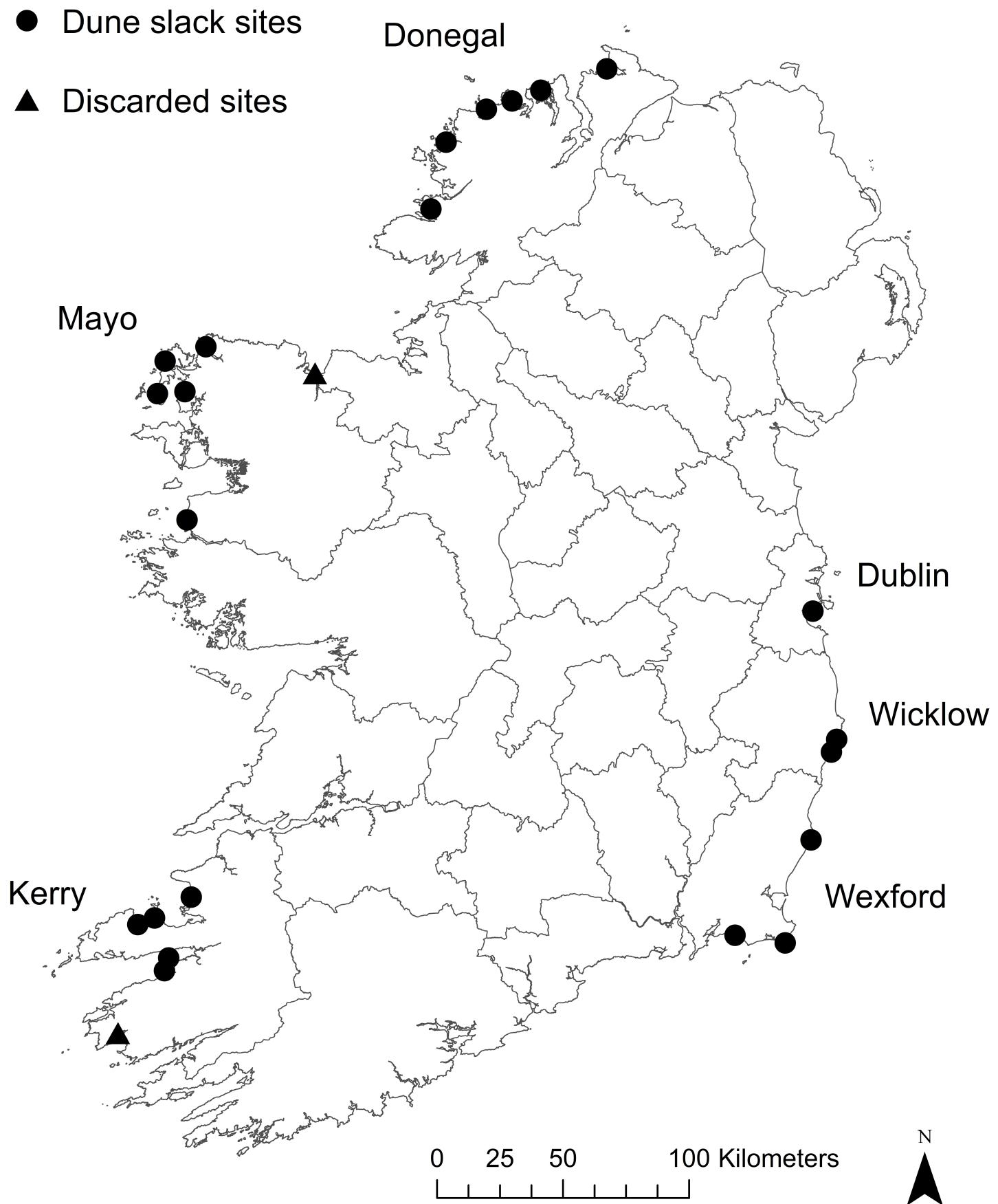
Dublin

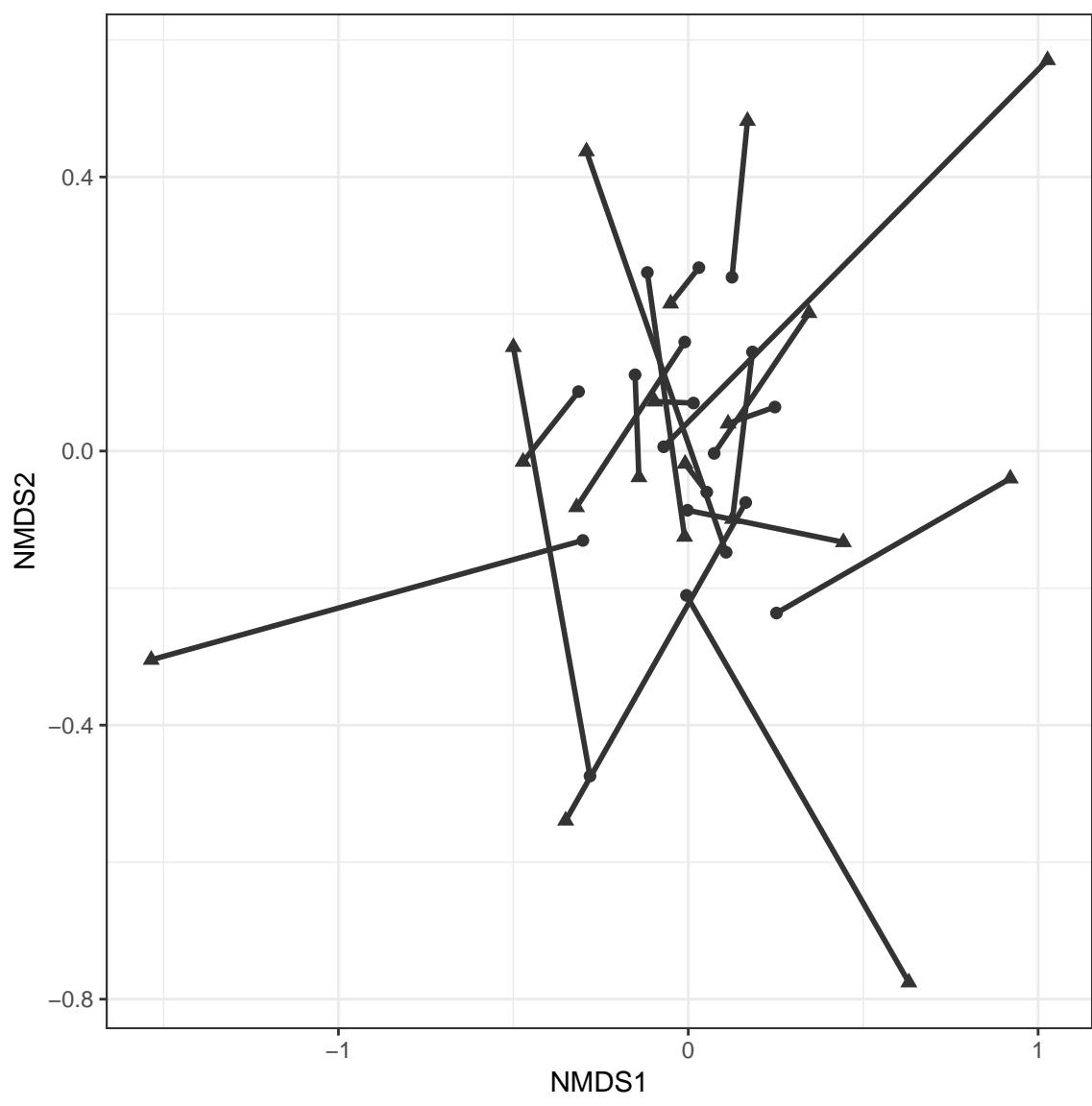
Wicklow

Kerry

Wexford

0 25 50 100 Kilometers





Snail

Water beetle

Species diversity

Species richness

