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Pitch Pine Stands Recover, Adapt and Persist In The Absence of Fire Under The Influence of Elevation and Topographic Factors

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Key words

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**ABSTRACT**

Globally rare pitch pine (*Pinus rigida* Miller) is thought to depend on intermittent fire, which encourages reproduction and niche preservation. At Mt. Desert island in Acadia National Park (ME, USA) a stand-replacing conflagration enveloped a portion of the island in 1947, however, there has been no recurrence. Other populations on the island are unaffected by fire disturbance for over one hundred years. Fire history is shown to influence plant form and functioning, yet these are not well quantified for pitch pine in relation to factors such as elevation and topography which impact traits of this species. We studied the impact of fire history, elevation and their interaction on traits of pitch pine individuals across elevation (9.5 to 404 m) and topographic (flat, ledge and cliff) gradients. Growth, stand density, photosynthetic water use efficiency, soil and foliar nutrients, soil water retention dimensions, elevation, slope and aspect were measured. Elevation was found to be a greater driver of persistence than fire history or an interaction between elevation and fire history. Crucially, we found fire return intervals are so infrequent as to reduce recovery capacity traits; instead, persistence capacity traits increase in anticipation of low intensity disturbances. Trees at upper elevations were smaller, narrower and less canopied than those at lower elevations and had greater water use efficiency, indicating a preference for stress resistance over growth driven by greater atmospheric pressure. At modest elevation trees exhibited greater soil moisture retention than those in high elevation, fire-exposed situations. Twenty-four statistically significant results manifested as criteria for management of pitch pine survival where persistence capacity is greatest at flat and cliff locations in both burned and unburned populations.

**INTRODUCTION**

On Mt. Desert Island (MDI) in Maine USA, globally threatened pitch pine (*Pinus rigida* Miller) dwell at the edge of their northeastern range. Some fire ecologists in Northeast USA suggest wildfire is required every six to twenty-five years to perpetuate and rejuvenate pitch pine (Jordan *et al* 2003) such as these. Yet, on MDI, pitch pine survive despite fire suppression (Miller *et al* 2017) which has persisted in the aftermath of the infamous 1947 conflagration (Fig. 1). It is not entirely clear what factors contribute to recovery and persistence in fire absence (Patterson Saunders and Horton 1983; Parshall *et al* 2003). In coastal Atlantic states further south, there is less concern since natural fire (Foereid *et al* 2015) and anthropogenic controlled burns (Carlo *et al* 2016) comprise a system where there is reduced fuel, removal of evergreen competition and opening of canopies (Neill *et al* 2007).

At MDI, trees in low-lying undisturbed (non-1947 fire) areas (Fig. 2) appear engaged in seedling recruitment superior to trees in the upper elevation fire zone. Indeed, reaction to fire absence manifests as physiological and morphological adaptations over many years (Little 1953). On the island, cone serotiny (Givnish 1981), thick bark and epicormic re-sprouting (Renninger *et al* 2013) are all but disappeared in formerly fire-prone pitch pine ecosystems (Jordan *et al* 2003). Suppression makes it likely former fire adaptive mechanisms like these (no longer required for stand sustenance) are wasted investments which likely encourage under-competition with other evergreens (Buma *et al* 2013). Critically, it is not clear if, in reaction to the former pyrogenic forest effect, recovery accelerates the disappearance of adaptive traits (Heuss 2018) or if this is a result of extrinsic responses to other, low intensity disturbances. Previous studies (e.g., Ibáñez *et al* 2019) have framed resilience as an artifact of recovery (Charpentier 2020) specifically in the absence of further fire disturbance and adaptation to climate change (Swanston *et al* 2018). We elect to distinguish between recovery and persistence capacity through a new qualitative model which explains pitch pine responses to fire history and non-fire environmental constraints through the lenses of elevation gradients and topography (Dunne *et al* 2004) (Fig. 3). It is informed by earlier resilience theories (Jordan Patterson and Windisch 2003; Howard and Stelacio 2011; and Ibáñez *et al* 2019). Here, we posit pitch pine populations exhibit non-mutually exclusive responses: (1) **recovery capacity** (*RECC*) as a measure of plant performance (Ibáñez *et al* 2019) long after fire disturbance (Patterson Saunders and Horton 1983) and (2) **persistence capacity** (*PERC*) that reflects reaction to the shock of disturbance absence (Brand and Jax 2007).

Each capacity metric reflects physical responses such as the tradeoff between growth and stress avoidance and colony retreat or expansion (Swanston *et al* 2018), as a function of stand density (Churchill *et al* 2012) and as an ‘ecologically stable strategy’(Day and Greenwood 2011). *RECC* is the better understood of the two pathways given the extensive literature covering resilience (Ibáñez *et al* 2019) and tree health (Fuller and Quine 2016). *PERC,* on the other hand, explains a trend where shade-intolerant pitch pine out-compete other evergreens such as red spruce (*Picea rubens*), hemlock (*Tsuga canadensis*) and balsam fir (*Abies balsamea*) expanding into new, mostly uninhabited territory. There is a paucity of biogeochemical, elevation and topographical data at MDI to inform the functionalities within this model. To remedy this, we consider soil and plant nutrition including C and N, stable isotope analysis to derive intrinsic water use efficiency (iWUEδ13C), soil moisture retention, growth and stand density set against fire history, elevation and topography.

Fire history is revealed by subsurface charcoal remnants found in previous fire zones. Charcoal pyrogenic carbon (PyC) festoons selected areas of the north side of Cadillac Mountain (Patterson Edwards and Maguire 1987) and downslope pockets along the Park Loop Road at Gorham cliffs. These deposits suggest the possibility of recalcitrant soil C and mineral capacity long after a fire event (Pingree and DeLuca 2017). Some evidence exists to establish the effects of an interaction between fire history artifacts like charcoal and elevation on tree physiological performance where, as elevation increases, so does water use efficiency (Wang *et al* 2017). Given the difference in fire histories between populations, this provides a nearly ideal testbed to better clarify aspects of chemical geography (Verma and Jayakumar 2012) and address a number of key questions with regard to leaf traits, tree growth and stand dynamics.

Our foremost aim is to examine a proposed model through an analysis of four key pitch pine populations according to the influence of fire, elevation and their combination. Given previous studies of fire involvement and post-fire photosynthetic response (Chen Wang and Jia 2017) we hypothesize pitch pine growth and population expansion is greater at low elevation, non-fire sites tied to higher soil C, foliar C (Butak 2014), more negative iWUEδ13C (Licht and Smith 2020) and greater soil water retention (positive indicators of *PERC*). We further theorize trees at low elevations display a greater propensity for growth in height, canopy and DBH through more competitive resource conservation (Stambaugh *et al* 2015). Concomitantly, we conjecture ubiquitous contours (Howard and Stelacio 2011) at upper ledge elevations which limit clustering (stand density) elsewhere, will be less stressful for colonization (Lafon *et al* 2014) and expansion on lower elevation flat and cliff surfaces which we predict will increase pioneering efforts at gentler slopes and more southerly aspect.

**METHODS**

**Study Extraction Sites**

We investigated fifteen pitch pine specimens at each of four sites at MDI, factorially crossed in a fire history (Miller *et al* 2014) by elevation design: (1) Wonderland trail between 9 and 25 m elevation (low elevation, outside the footprint of the 1947 fire), (2) Gorham cliffs between 24 and 36 m (low elevation, within the footprint), (3) St. Sauveur trail between 134 and 198 m (high elevation, outside the footprint) and (4) South Cadillac trail between 188 and 417 m (high elevation within the footprint).

**Allometry**Individuals were further characterized according to tree height, canopy spread and stem diameter of the bole at breast height (DBH). Tree height was estimated using nested, 2 m calibrated, aluminum rods (Garelick, St. Paul, MN, USA. DBH was measured at 1.06 m using a ProSkit electronic digital caliper (Amelia, VA, USA). Canopy spread was measured using the span between the same calibrated aluminum rods fixed with two landscape flags as a ground truth reference.

**Clustering**Mean distances between sampled trees (*N*=167) up to five of their nearest, reproductively mature conspecific neighbors (within 5 m) in the same clump (Churchill et all 2012) were calculated as mean neighbor distance—a surrogate, but inverse, measure for stand density (Mosseler Rajora and Major 2004).  
**Topographic features**

We used a Kodak Trimble Juno 3B unit to obtain horizontal resolution of data plotted using between five and seven satellite telecommunication vehicles to maintain a maximum PDOP (Position Dilution of Precision). These data were differentially corrected and have estimated accuracies in the horizontal and vertical direction of 2 meters, while SA (selective availability) is set to zero. Multiple satellite-configured GPS data (USGS 2m LIDAR 2010) determined coordinates for individual trees (Lubinski Hop and Gawler 2003) as well as slope and aspect attributes using ArcGIS (version 10). Mapping of this type of data is used in the past to compare physiography and recalcitrant chemical biogeography, particularly in fire prone contexts (Szpakowski and Jensen 2019).

**Isotopic analysis**

We obtained C isotopic data (δ13C) and N isotopic data (δ15N) of fully expanded leaves (needle cluster) of 15 individual at each site. Sample fascicles were separated and dried for two days at 60 ◦C ground in a SPEX ball mill (Metuchen, NJ, USA), weighed to +/- 2 mg for leaf tissue and +/- 5 mg for soil using a Cole-Palmer (Vernon Hills, IL, USA) micro analytic balance and rolled in Costech (Valencia, CA, USA) 5 x 9 mm tin capsules. A Thermo Delta (Waltham, MA, USA) V+ IR-MS continuous flow isotope ratio mass spectrometer with a universal triple collector was used. Combustion gasses were separated on a gas chromatograph column, passed through a diluter and reference gas box, and introduced into the spectrometer. C abundance or depletion according to δ13C was used to indicate water use efficiency (iWUEδ13C) (Farquhar et al. 1989).

**Foliar tissue analysis**

Leaf tissue was obtained from excision of basal fascicle bundles at 1.06 m. 50 mL samples of needles were separated, cut and dried for two days at 60 ◦C. Then they were ground in a SPEX ball mill (Metuchen, NJ, USA), sieved to <10 mm, and <2 mL were fed to a Leco CN-2000 Carbon-Nitrogen Analyzer (Leco Corp., St. Joseph, MI) coupled with the spectrometer to determine C and N concentrations. 35 mL aliquots were submitted for standard plant tissue nutrient analysis using a TJA Model 975 AtomComp ICP-AES (Thermo Jarrell-Ash Corp., Franklin, MA). The method comprised submersion in a 5 mL trace-metal-grade HNO3 treatment, then refluxed on hot block at 80 ◦C for two hours and diluted to 25 mL with 0.4 micron PTFE syringe filters to access extractable macro and micro inorganics.

**Soil analysis**

Soils (porous and acidic hornblende granite or Ellsworth schist) were uniformly shallow, homogeneous, low fertility (varying between .7-2.5 cm) and overlain with rapidly drying needle duff (Day *et al* 2005). All sites were excavated similarly by hand trowel and soil probe (Accuproducts, Saline, MI, USA); aliquots were extracted from Oa-Ab horizon soil pockets of organic and mineral deposits within 50 cm of the tree base. 250 mL soil samples were sieved (#10) and measured in 2019. Drying was performed in an oven at 100°C for two days. Analysis was performed using a modified Mehlich method using inductively coupled plasma spectroscopy, pH measurement via proton activity of a 1:1 slurry and effective soil cation exchange capacity (CEC) via formic acid extraction. These methods are described more fully elsewhere (Lichtand Smith 2018). Soil C and N were calculated using elemental analysis in a similar fashion to method above.

**Soil Water Retention (SWR)**

70 mL soil samples were extracted at fifteen tree locations at four sites, from <7.5 cm (Oa-Ab) horizon above bedrock. In a laboratory 50 g H2O were added to each aliquot to assess net water retention as a subset of soil moisture evaporation (*ψ*g) to determine net evaporative loss or adsorption to surfaces. Soil water retention analysis was conducted according to the Fields method described previously (Lichtand Smith 2018). Retention effects of gravitational and evaporation forces was made on a wet basis where Wm=g H2O **●** g-1 moist soil, as described previously (Jingfang and Wenwei 2018).

**Statistical Analysis**

All data were analyzed using a similar linear model structure with elevation (high or low) and presence of the 1947 fire (yes or no) as categorical fixed factors. The interaction between elevation and presence of the 1947 fire was also included in each model. In total, twenty four models were fit with the following dependent variables: tree height (m), canopy spread (m), DBH (cm), foliar: C (%), N (%), C/N (unitless), δ13C (‰), δ15N (‰), Ca (g g-1), P (g g-1), K (g g-1), Mg (g g-1), Al (g g-1), Zn (g g-1), soil: C (g g-1), N (g g-1), C/N (unitless), Ca (g g-1), P (g g-1), K (g g-1), Mg (g g-1), Al (g g-1), Zn (g g-1) and water retention (%). Tree height, canopy spread, DBH, foliar P, foliar K, foliar Zn, soil P, soil Al, soil Zn, and soil CN were log transformed to meet model assumptions of normality and heterogeneity of variances. Soil water retention was arcsin square root transformed to meet model assumptions. Analysis of circular aspect data was achieved through Watson’s Two-Sample Test of Homogeneity (1962) to yield a *t*=test statistic. This test was applied in order to tackle problems originating from the analysis of circular aspect data, performed on pairs of four sites to determine differences between aspect (ratio). Except for Watson, all models were fit using the ‘lm’ function in R. Significance tests for each fixed factor was performed using the ‘anova’ function in R. Post-hoc Tukey’s tests were done to examine significant interactions between elevation and the presence of the 1947 fire using the ‘emmeans’ package (Lenth 2018) in R. All analyses, except for Watson, were performed with R version 3.5.1 (R Core Team 2019).

**RESULTS**

Elevation, as opposed to fire or elevation x fire interaction, was the dominant factor in outcomes which favored *PERC* over *RECC*.We reasoned *PERC* would occur more frequently on level or modest ledge slopes at low elevation. That assumption was confirmed in stark differences in elevation (*P* < 0.001, Tab. 1), fire (*P* < 0.001, Tab. 1) and an interaction between fire and elevation (*P* < 0.001, Tab. 1) for both elevation and slope measurements. An ANOVA was not performed to analyze aspect function due to problems originating from the analysis of circular data. Instead, a Watson Two-Sample Test of Homogeneity (Fig. 4) was employed. We found significant differences in aspect between Cadillac and St. Sauveur (*P* < 0.001, Fig. 4, Tab. 2), Cadillac and Wonderland (*P* < 0.05, Fig. 4, Tab. 2), Gorham Cliffs and St. Sauveur (*P* < 0.01, Fig. 4, Tab. 2), Gorham Cliffs and Wonderland (*P* < 0.05, Fig. 4 Tab. 2) and St. Sauveur and Wonderland (*P* < 0.01, Fig. 4, Tab. 2). Wonderland and St. Sauveur, were oriented fairly evenly between South and Northwest.Mean slope (Fig. 5 and Tab. 1) averages ranged from 4.5-13° for mostly level Wonderland, and slightly steeper ledge populations at St. Sauveur and South Cadillac. Trees at Gorham cliffs sat on much steeper pitches at an average of 33.5° and were more tightly connected.

*Allometry and Stand Density*

Notably, an interaction between fire and elevation was significant (*P* < 0.01) for tree height (Fig. 6A and Tab. 3). Trees which experienced the 1947 fire, stationed at lower altitude, as well as those which were not, enjoyed a wider canopy (*P* < 0.01, Fig. 6B and Tab. 3) than those in the mountains. Elevation (*P* < 0.01, Fig. 6C and Tab. 3) and an interaction between fire and elevation (*P* < 0.05, Fig. 6C and Tab. 3) were also significant for DBH. Trees which experienced the 1947 fire at high elevation had smaller DBH (cm) values than trees at lower elevation sites (Tukey’s HSD: Fig. 6C and Tab. 3). We hypothesized shoreline populations would comprise denser colonies of conspecific neighbors within 5m clusters, a surrogate value for stand density; this was borne out in our analysis (*P* < 0.01, Fig. 6D and Tab. 3) differences between populations was evident with the greatest clustering occurring at Gorham cliffs.

*iWUEδ13C*

An accurate measure of intrinsic water use efficiency is calculated using a C stable isotope as an index of C abundance or depletion. Trees at higher elevations experienced more negative iWUEδ13C (*P* < 0.01, Fig. 7A and Tab. 6), reflecting greater C depletion there but this response was not fire-dependent since it occurred at St. Sauveur as well. There were no significant differences between tree populations per δ15N (*P* > 0.05, Fig. 7B and Tab. 6)

*Foliar organics*

Foliar C was more readily available at upper elevations, however the results were not statistically significant (*P* > 0.05, Fig. 8A and Tab. 4); nor were they significant with regard to C/N (*P* > 0.05, Fig. 8C and Tab. 4). However, fire accounted for a significant influence on N availability (*P* < 0.05, Fig. 8B and Tab. 4)

*Foliar macronutrient*

Foliar mineral availability was impacted negatively as elevation increased as witnessed by a 48% decline in Ca+2 (*P* < 0.001, Fig. 9A and Tab. 5). Foliar P was significantly higher (*P* < 0.01, Fig. 9B and Tab. 5) at fire-involved sites as was true for foliar K+ according to fire (Tukey’s HSD: *P* < 0.05, Fig. 9C and Tab. 5) and elevation x fire (Tukey’s HSD: *P* < 0.05, Fig. 9C and Tab. 5). Neither Al+ nor Mg+ availability stood out but Zn foliar (*P* < 0.01, Fig. 9F and Tab. 5) concentrations were 9% lower in the subalpine zone than on the shoreline.

*Soil organics*

Soil C was more readily available at upper elevations, however it was 37% (significantly) lower (*P* < 0.05, Fig. 10A and Tab. 7) at coastal (non-fire) elevations compared. Contrary to our hypothesis soil C recalcitrance at one of two high elevation sites which experienced fire in 1947 was significantly depleted (*P* < 0.05, Fig. 10A and Tab. 7). Soil N was nearly twice as abundant at Wonderland as on the slopes of South Cadillac trail but differences in soil N between sites were not statistically significant (*P* > 0.05, Fig. 10B and Tab. 7). C/N was 15% lower at high elevation sites (*P* < 0.05, Fig. 10C and Tab. 7) but we found no significant disparity in C/N when either fire history or fire history x elevation interactions were examined.

*Soil macronutrient*

Soil Ca+2 (*P* < 0.001, Fig. 11A and Tab. 8) was 33% lower at altitude than at Wonderland. P, Mg and Zn were not significantly different (Tab. 8). However, fire accounted for a 48% reduction in Kat fire-experienced sites (*P* < 0.01, Fig. 11C and Tab. 8) and Al+ (*P* < 0.01, Fig. 11E and Tab. 8) was highest where a fire x elevation factor was considered.

*Soil Water Retention, CEC and pH*

Soil Water Retention (SWR) was significantly different between sites (*P* <0.01, Fig. 12A and Tab. 9) and markedly higher at Gorham cliffs which experienced fire as compared to other sites. Soil pH was not a factor (Tab. 9) but CEC was statistically significant (*P* < 0.05, Fig. 12C and Tab. 9) where elevation was concerned.

**DISCUSSION**

Four populations, representative of eighteen on MDI, were more responsive to non-fire environmental disturbance than fire disturbance, at least over the last one hundred years according to twenty-four statistically significant response variable data. Our findings underscored differences between *RECC* and *PERC* pathways and provided an explanation to resolve an enigma of persistence in fire absence and the dominant influence of elevation. The most visible effects of elevation-fire interactions were notable where biometric (height, DBH) and soil water retention factors were considered. Investigators noted a connection between the greatest pitch pine verticality (*µ*=6.2 m) and DBH (*µ*=40.1 cm) at Gorham cliffs reflective of *PERC* tendencies and consistent with results of the Watson test. Fire-experienced site aspects at Gorham cliffs and South Cadillac were oriented between East and South, while non-fire location topography and aspect characteristics favored Southern and Western orientations.

Elevation played a role in several meaningful ways. We confirmed an hypothesis that certain factors, such as iWUEδ13C, depleted foliar nutrients and growth limitations were linked to higher elevation gradients (as described by Wang *et al* 2017; Chen, Wang and Jia 2017). Further, we found distance between neighbors at higher elevation was greater (37.7%) compared to lower elevation sites, thus generating sparser tree numbers aggregated within *RECC* compared to *PERC.* At the highest elevation, South Cadillac trail, we expected to find the steepest slopes but they were far less inclined than those at Gorham cliffs. We anticipated low elevation (<50 m) populations would feature a greater number of conspecific neighbors as a function of no fire history, flat or cliff orientation, evidenced by a gentle slope (<10°), and tendency towards a southerly aspect (*µ*=180°). In fact, this theory was found correct, as Wonderland, which was by far the flattest terrain, trees achieved the widest distribution with the greatest distance between trees over the widest contiguous area as a function of *RESC.* There was sufficient variability in some of the data, like stand density, to produce valuable information about differences between flat, cliff and ledge communities, even if at a small spatial scale.

Fire was less meaningful than elevation even at high elevation ledge communities as a determinant of future prosperity. We found no compelling empirical evidence that it is a necessary ingredient in perpetuation at MDI. Fire return intervals have lengthened at MDI to the point where they are quite likely to be too intermittent to perpetuate previous fire resistance traits. A shift back to fire, accompanied by a re-introduction of serotinous characteristics, is not impossible in the future; however, current climate projections advise against this occurrence. Despite the absence of fire, there is no absence of fire remnants. The current study builds on previous work by the authors (Licht and Smith 2020) to take an obvious post-fire effect—anthropogenic fire charcoal (PyC)—and its association with greater (more negative) iWUEδ13C, negligible consumption of Ca, K and Mgand enhanced soil moisture holding capacity changes according to elevation and topography. In this study we hypothesized water use efficiency, stored plant C and water retention would change significantly according to those factors which they did. When we dove more deeply into that interaction, we found greater stand density at flat and cliff sites associated with more positive iWUEδ13C and photosynthetic growth, consistent with reported outcomes for non-glaciated populations in flat, sand plain New Jersey Pine Barrens (Mikita-Barbato *et al* 2015; Schafer and Bohrer 2016). We also found congruence between low elevation outcomes in the present study and those reported (*µ*=33m above sea level) in the New Jersey pine barrens (Carlo *et al* 2016). However, we could not draw any other useful comparisons between the studies due to a lack of data concerning clustering (stand density), slope or aspect factors. However, there is evidence from previous investigations that post-fire PyC remnants which endure in the soil layer (DeBano 1981), increase alkali cations (Kolden *et al* 2017), and solubilized minerals (Caldwell and Richards 1989) likely linked to thermal exfoliation (Shakesby and Doerr 2006). From the standpoint of reproduction, the absence of fire, at locations such as *PERC*-oriented Wonderland trail (Butak 2014), does not appear to be slowing down expansion in that flat-sloped region nor in other cliff locations which we and others studied (Howard 2010).

Developing further insights into the ecological stoichiometry of pitch pine at MDI is a worthwhile goal given earlier discoveries by Fernandez (2008). Our nutrient analysis derived from burned and unburned trees was similar to a methodology in a New Jersey study (Renninger *et al* 2013) and produced some similar results. Since elevation is a key factor, we noted foliar Zn was substantial at lower elevations at MDI; this was consistent with findings by others in another New Jersey investigation (Kolker *et al* 2013). In both Maine and New Jersey, it is possible *PERC* may be enhanced by higher concentrations of foliar Ca2+, K+, Mg2+, P, Al+ and Zn (Mg2+ foliar availability was not significant in this study). We were particularly interested in P, a limiting factor (Verma and Jayakumar 2012); it is sometimes found to co-occur with pools of greater C (Preston and Schmidt 2006). Scientists elsewhere assert fire-induced sedimentary charcoal produces soil C enrichment (Patel *et al* 2016) and as measured in soil columns (Hart Horn and Grissino-Mayer 2008). One could argue that charcoal remnants likely play a role in *RECC* at burned-over Cadillac Brook (below the heights of South Cadillac trail) but it is not known to what extent these benefit the larger ecosystem. We conjecture soil C persistence since the 1947 perturbance at burned-over areas such as South Cadillac trail reflects a failure of pyrogenic carbon removal (Doerr *et al* 2018), though elsewhere lower C availability is attributable to greater consumption by fungi (Luo *et al* 2017). P concentration at upper elevations at MDI contrasted with more modest availability at other, low-lying pinelands at MDI and in New Jersey (Renninger *et al* 2013; Alkañiz *et al* 2018). One explanation for substantial soil P availability most likely derives from charcoal remnants; however, lacking mycorrhizal studies we were unable to confirm the extent of P liberated from the charcoal in the two 1947 fire precincts.

Climate is likely the final arbiter of decline rather than stand-replacing fire disturbance in the case of long-term pitch pine livelihood. Recent climate change models anticipate negative impact on future vegetative status at MDI (Fernandez *et al* 2015) but these do not specifically address the vicissitudes of pitch pine nor the extent to which plasticity (Day *et al* 2014) is shaping tendency towards *PERC* or *RECC*. What has been clear for at least two decades is the effect of global climate change on ecophysiology traits. Day Greenwood and White (2001) found an uptick in annual temperatures signaled increased leaf-air vapor pressure deficits which negatively impact pitch pine stomata response and limit gas exchange. In a related report, scientists found warming trends (Kunkel *et al* 2013) increase pitch pine difficulties in reproduction (Ledig Smouse and Hom 2015). This includes weather-related effects such as episodic drought, harsh winds and salt spray (Schmitt 2015; Fernandez *et al* 2015) and cold intolerance (Berang and Steiner 1985). Increases in annual winter temperatures (Lesk *et al* 2017) coupled with an absence of fire cause concern about a quite different disturbance—potential invasion within the next decade of an herbivore, Southern pine beetle (*Dendroctonus frontalis* Zimmer or ‘SPB’). Although deer browse and rodent damage (Ledig *et al* 2013) historically impeded tree survival in pine barrens, SPB has already paid a deadly visit to Long Island NY (Dodds *et al* 2018). Unless its progress is deterred by other insect predators like double checkered clerid (*Thanasimus dubius*) (Coulson and Klepzig 2011), it is possible pitch pines along with understory plants, butterflies and moth members of the Acadia ecosystem will suffer the same fate experienced in more southerly locations (Lesk *et al* 2017). According to several authors (Day *et al* 2005; Lee *et al* 2019) warming climate impacts the suitability of habitat; pitch pine tendencies to consolidate, regenerate or migrate may be in jeopardy. These effects are likely to eventually limit aspects of *PERC* such as niche expansion, if they have not already, through a combination of diminished open space capacity, loss of enriched substrates and elimination of ‘safe sites’.

Pitch pine is considered an important guardian of underlying heath communities at MDI; it is foundational as a necessary ecosystem component in a stressed environment. Despite increasing climate pressures, tree retreat into ever more sparse conditions reinforces their facilitator status (Connell and Slatyer 1977) for this species explicitly maintains the livelihood of underlying flora through a sharing and distribution of ecoservices. Nevertheless, competitive advantages enjoyed currently may even give way to ‘mesophication’—negative feedback for shade intolerant trees like pitch pine (Nowacki and Abrams 2008), perhaps more widespread at the confluence of fire suppression, overabundance of deer and climate change.

The model we proposed is not built on a machine learning framework nor is it intended as a predictive model, *per se*, yet results attached to this model are useful in several ways. First, metrics provide a context for describing recovery or persistence in mathematical relationships along an adaptivity curve. Second, our method operationalizes recovery and persistence mechanisms fit to an ecological framework (Brand and Jax 2007). Finally, our model may be used to better understand how pitch pine in other ecosystems, outside of the Northeast U.S., discriminate between *RECC* and *PERC* at a given elevation and within certain topographic parameters. We anticipate the Watson test may be useful in these cases too with regard to multiple samples drawn from the same population where data located on a circle of geographic loci make conventional hypothetical tests are problematic.

While investigators did not determine the precise impact of *RECC* on seedling recruitment (Patterson *et al* 2014), for example, the present data provides a firmer purchase on current regeneration and expansion concepts—essential to an appreciation of influences on persistence in the absence of forest or prescribed fire.

**CONCLUSION**

The authors presented an explanatory model of pitch pine post-fire recovery and persistence capacities to analyze population status as a function of fire, elevation and topographical data. We found adaptivity effects (growth, expansion into greater stand density) account for greater growth and stand density in longer fire absent locales in anticipation of other disturbances. The authors determined fire return intervals are so infrequent as to reduce recovery features found elsewhere (e.g., cone serotiny). Flat and ledge pitch pine populations exhibited greater buoyancy than trees in more strenuous cliff situations. We also identified a selective preference for either growth or stress tolerance in order to conserve water use, soil and foliar C, macronutrients and soil moisture retention, depending on fire exposure, elevation or their interaction. Our findings across eighteen dimensions unravel an enigma about persistence in a post-fire milieu during a critical phase of the Anthropocene age (Crutzen and Stoermer 2000). At lower elevations, which represent the vast majority of pitch pine populations in the Northeast US, we predict newly pioneered locations reflect a continuation of *PERC* signaled by significant differences in density, slope and aspect. At a time when continued climate change may tip the scale away from survival, our findings encourage the use of a model by forest managers to better understand the imposition of fire absence on flat and ledge communities.

## Data Availability Statement

Data used in this article can be found at the following repository: <https://github.com/SmithEcophysLab/pitchpine/releases/tag/v1.0> (doi: [10.5281/zenodo.3600547](https://doi.org/10.5281/zenodo.3600547)).

## Author Contributions

JL and NS conceived the work, contributed substantially to the interpretation of the data and to drafting the manuscript, gave final approval of the version submitted, and agreed to be accountable for all aspects of the work. Questions related to the accuracy or integrity of any part of the work are appropriately investigated and resolved. JL carried out sample collection and field measurements, conducted soil water retention tests and prepared samples for EA-IRMS analysis. NS performed C/N foliar analysis, conducting statistical analyses and formulating figures and tables.

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