

Waste and Material Footprints in prospective LCA: a macro study of 1593 activities from 2020-2050

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Abstract (max 400 words)

Purpose (max 75 words)

Advancing a circular economy requires system-wide quantification of waste and material flows. Yet Waste and Material Footprints (WMFs) remain under-reported in LCA and sparsely integrated into pLCA; moreover, waste treatment is poorly represented in prospective databases, limiting decision relevance. We quantify WMFs across ecoinvent and examine their evolution under contrasting pathways to (i) reveal sectoral and supply-chain hotspots, (ii) position WMFs alongside LCIA endpoints, and (iii) assess how scenario-aligned backgrounds modify footprint magnitudes and circularity.

Methods (max 125 words)

We built prospective LCI databases with *premise* (ecoinvent v3.9.1) aligned to two REMIND pathways—SSP1-PkBudg500 and SSP5-PkBudg500—for 2020–2050. Using T-reX, we tracked over 70 waste and material categories and computed WMFs for 1,593 market activities. In parallel, we calculated ReCiPe 2016 endpoints. We grouped activities into sectors to identify hotspots and explore temporal/scenario contrasts. We also calculated a waste circularity ratio, the share of each activity's waste footprint routed to material recovery and biological cycling (recycling+composting), excluding energy-recovery incineration and unmanaged disposal (e.g. open burning).

Results and discussion (max 125 words)

Metals/alloys, chemicals, and ores–minerals–fuels dominate total and hazardous waste footprints and anchor the upper tails, with outliers in rare-earths, precious metals, and nuclear-fuel routes. From 2020 to 2050, total waste and landfilling increase at the median in both pathways, while recycling and composting rise more modestly; the circularity ratio declines slightly and the hazardousness share falls modestly. Scenario differences are nuanced: SSP1 tends to show higher 2050 medians for total waste and landfilling yet a larger reduction in hazardousness; SSP5 exhibits slightly less deterioration in circularity. Across indicators, between-activity heterogeneity dwarfs pathway effects. WMFs correlate with ReCiPe human-health and ecosystem damage at activity and sector levels.

33 **Conclusions and recommendations (max 150 words)**

34 WMFs complement LCIA endpoints in prospective LCA by making material throughput and waste generation
35 explicit, revealing hotspots that impact profiles may obscure. Temporal signals in current IAM-linked pLCIs are
36 incremental: recovery routes expand, but not fast enough to offset rising disposal and gently declining circularity
37 for the median activity. A key constraint is the limited and uneven representation of waste systems in many
38 LCI/pLCI datasets—coarse treatment typologies, sparse regionalisation, inconsistent hazardous labels, static
39 assumptions for collection/sorting yields and quality losses, and minimal representation of secondary-market
40 uptake. We recommend embedding scenario-dependent circularity modules (prevention, capture, sorting,
41 recovery yields/qualities, substitution, leakage abatement) into future pLCIs, and reporting WMFs alongside the
42 circularity ratio to track whether recovery keeps pace with total waste growth.

43

44 **Keywords**

45 Circular economy, Waste footprints, Material footprints, Prospective life cycle assessment, Scenario-based life
46 cycle modelling, Integrated assessment models, Critical raw materials, Integrated assessment models

List of abbreviations

| | | |
|----|---------------|------------------------------------------------------|
| | AgriForeAnim | Agriculture, forestry, live animals & their products |
| 47 | CE | Circular Economy |
| 48 | Chemical | Chemical products |
| 49 | CPC | Cooperative Patent Classification |
| 50 | CRM | Critical Raw Material |
| 51 | CRT | Cathode Ray Tube |
| 52 | EF | Ecological Footprint |
| 53 | EoL | End-of-Life |
| 54 | GlasNonMetal | Glass and other non-metallic products |
| 55 | GLO | Global (ecoinvent location designation) |
| 56 | IAM | Integrated Assessment Model |
| 57 | IMAGE | Integrated Model to Assess the Global Environment |
| 58 | LCA | Life Cycle Assessment |
| 59 | LCI | Life Cycle Inventory |
| 60 | LCIA | Life Cycle Impact Assessment |
| 61 | LLDPE | Linear low-density polyethylene |
| 62 | MachElecTrans | Machinery, metal/electronic, transport equipment |
| 63 | MetalAlloy | Basic metals & alloys, incl. semi-finished products |
| 64 | MF | Material Footprint |
| 65 | MFA | Material Flow Analysis |
| 66 | OreMinFuel | Ores, minerals & fuels |
| 67 | PlastRub | Plastics & rubber products |
| 68 | pLCA | Prospective Life Cycle Assessment |
| 69 | ProcBio | Processed biobased products |
| 70 | PVC | Polyvinyl chloride |
| 71 | RCP | Representative Concentration Pathway |
| 72 | ReCiPe | A standard LCIA method set |

| | | |
|----|--------|------------------------------------------------------------------------------------------|
| 73 | REE | Rare Earth Element |
| 74 | REO | Rare Earth Oxide |
| 75 | REMIND | REgional Model of Investment and Development |
| 76 | re-X | A broad set of circular economy strategies (“reduce”, “reuse”, “repair”, “recycle” etc.) |
| 77 | RoW | Rest of World (ecoinvent location designation) |
| 78 | SDG | Sustainable Development Goal |
| 79 | SSP | Shared Socioeconomic Pathway |
| 80 | T-reX | The Tool for analysing re-X in LCA |
| 81 | UNFC | United Nations Framework Classification for Resources |
| 82 | WF | Waste Footprint |
| 83 | WMF | Waste and Material Footprint |

84 1 Introduction (1200 words)

85 Environmental context: why circularity and waste matter

86 Human activities continue to exceed key planetary boundaries, intensifying climate change, biodiversity loss,
 87 and resource depletion. In response, the transition to a circular economy has become a central pillar of
 88 sustainability policy (Ellen MacArthur Foundation, 2015; European Commission, 2020; Pardo & Schweitzer,
 89 2018). Circular strategies seek to decouple well-being from primary material extraction by reducing material
 90 demand and preventing waste across value chains through ‘re-X’ measures—refuse, rethink, repair,
 91 remanufacturing, and recycling (Kirchherr et al., 2017; Reike et al., 2018). Recent geopolitical tensions further
 92 underscore the vulnerability of globalised supply chains and the need for material efficiency and system
 93 resilience (Carrara et al., 2023; Hartley et al., 2024).

94 Waste and material footprints in LCA

95 Footprints provide compact indicators of environmental pressure. The Ecological and Carbon Footprints initiated
 96 this “footprint family” (Čuček et al., 2015; Wackernagel, 1994), which has since expanded without fully
 97 converging on a coherent framework (Giampietro & Saltelli, 2014; B. G. Ridoutt & Pfister, 2013; Vanham et al.,
 98 2019). The Material Footprint (MF)—the total supply-chain material use attributable to products, sectors, or
 99 economies—correlates strongly with human-health and biodiversity damage and is recognised by the United
 100 Nations for SDG monitoring (Lenzen et al., 2021; Wiedmann et al., 2013). By contrast, the Waste Footprint
 101 (WF)—the mass or volume of waste generated along value chains, including hazardous fractions—remains less
 102 developed and is often overlooked, despite evidence linking waste burdens to environmental damage and social
 103 inequity (Akese & Little, 2018; Laurenti et al., 2023; Steinmann et al., 2017). Considering WF alongside MF
 104 highlights where material use translates into waste generation, where hazardous waste arises, and where
 105 interventions may yield the greatest returns for circularity.

106 Life Cycle Assessment (LCA) is the prevailing method to quantify environmental impacts across product and
 107 service life cycles (Guinée et al., 2010). In standard practice, life cycle impact assessment (LCIA) methods (e.g.,
 108 ReCiPe, CML) convert inventory flows—elementary exchanges between technosphere and biosphere—into
 109 impact scores (Guinée et al., 2002; Huijbregts et al., 2016). Several LCIA frameworks incorporate aspects of
 110 waste and material use (e.g., Swiss Eco-Factors, EDIP, EN15804, Crustal Scarcity Indicator) (Arvidsson et al.,
 111 2020; CEN (European Committee for Standardization), 2019; Hauschild & Potting, 2004; Swiss Federal Office
 112 for the Environment (FOEN), 2021), yet few provide transparent, mass-consistent accounting of MF and WF.
 113 Some also rely on abstract units (e.g., Umweltbelastungspunkte in the Swiss Eco-Factors) that can complicate
 114 interpretation. Moreover, because waste is commonly modelled as a service (treatment), the magnitude and
 115 distribution of waste generation along supply chains can remain obscured, making upstream waste effectively
 116 “invisible” (Beylot et al., 2018; Guinée & Heijungs, 2021).

117 In practice, waste is often defined as material with negative economic value, but its significance extends far
 118 beyond treatment emissions (Bisinella et al., 2024; Guinée et al., 2004; Laurenti et al., 2023). Empirical studies

119 confirm associations between waste burdens, environmental damage, and disproportionate impacts on vulnerable
 120 communities (Akese & Little, 2018; Pellow, 2023; B. Ridoutt et al., 2010). Reporting WF and MF alongside
 121 conventional LCIA indicators can therefore make material throughput and waste generation explicit, reveal
 122 hidden hotspots, and improve prioritisation of circular economy strategies.

123 Future-oriented LCA and prospective background databases

124 Emerging technologies required for deep decarbonisation will scale over coming decades, often after substantial
 125 learning and capital investment (International Energy Agency (IEA), 2021, 2022). Prospective LCA (pLCA)—
 126 also called ex-ante or anticipatory LCA—assesses likely environmental implications early enough to inform
 127 design and policy (Cucurachi et al., 2018; Van Der Giesen et al., 2020). Robust pLCAs require background data
 128 that reflect plausible future economic, technological, and policy conditions. Prospective life cycle inventory
 129 (pLCI) databases therefore combine current LCI data (e.g., ecoinvent) with scenario information from integrated
 130 assessment models (IAMs) and other sources (Sacchi et al., 2022; Steubing et al., 2023).

131 IAM scenario frameworks typically pair a shared socio-economic pathway (SSP)—a narrative of societal
 132 development from sustainability-oriented (SSP1) to fossil-intensive (SSP5)—with a representative concentration
 133 pathway (RCP) that specifies a climate outcome via radiative forcing, corresponding to temperature goals such
 134 as 1.5–2 °C (Aboumabhoub et al., 2020; Meinshausen et al., 2020; Stehfest et al., 2014; Van Vuuren et al.,
 135 2017). Implemented in IAMs, SSP×RCP pairings generate region- and sector-specific trajectories for technology
 136 deployment and emissions (Sacchi et al., 2022). These scenarios are bounded by resource availability,
 137 infrastructure lock-in, and policy constraints such as carbon pricing, which shape feasible transitions (Pauliuk et
 138 al., 2017).

139 premise, REMIND, and sectoral transformations

140 The premise workflow connects IAM projections to ecoinvent, producing pLCIs that regionalise markets and
 141 update process and supply-chain parameters for selected sectors (Sacchi et al., 2022; Sacchi et al., 2023). The
 142 most widely used IAMs are the REgional Model of Investment and Development (REMIND) (Aboumabhoub et
 143 al., 2020) and the Integrated Model to Assess the Global Environment (IMAGE) (Stehfest et al., 2014). Neither
 144 IAM scenarios nor LCI databases currently provide full, high-resolution coverage across all sectors and regions.
 145 IAMs are detailed for electricity but sparser for agriculture, chemicals, and material cycles; standard LCIs
 146 prioritise current technologies, leaving emerging options under-represented (Pauliuk et al., 2017; Sacchi et al.,
 147 2023). The current default transformation domains include electricity generation and markets (with storage),
 148 cement (clinker ratio, kiln efficiency, optional CCS), iron and steel (process efficiency and CCS), fuels (refining,
 149 synthetic and biofuels, hydrogen), road freight (powertrain shares and fleet relinking), batteries (mass/energy-
 150 density scaling and market composition), heat supply (CO_2 factors), air-pollutant factors, and biomass markets
 151 distinguishing purpose-grown from residual feedstocks (Sacchi et al., 2023). Additional research has produced
 152 additional scenarios that can be integrated into pLCA databases with premise for sectors such as cobalt (Van Der
 153 Meide et al., 2022), hydrogen (Wei et al., 2024) and cement (Müller et al., 2024).

154 While the aforementioned sectoral transformations can result in indirect changes to future waste flows

155 (McDowall et al., 2025), waste management is not yet a dedicated transformation domain and other waste-sector
156 inventories remain largely as they appear in the base database (Bisinella et al., 2024).

157 **Aim and contribution of this study**

158 Prospective analyses in LCA rely on the completeness and consistency of pLCIs, though, currently, they
159 insufficiently represent waste-sector dynamics, creating a ‘waste gap’ that limits interpretation of future
160 scenarios. Addressing this gap requires first clarifying how waste and material flows are represented in existing
161 LCA and pLCA databases at both macro and activity levels.

162 This study applies a purpose built python tool T-reX (McDowall et al., 2025)—integrated within Brightway and
163 compatible with premise-based pLCIs—to explore and quantify waste generation (including hazardous waste)
164 and material consumption (especially CRMs) across activities and sectors. Our objectives are to: (i) compute
165 waste and material footprints at multiple levels of aggregation, (ii) identify hotspots along supply chains under
166 present and prospective background conditions, and (iii) illustrate how results support circular-economy
167 strategies and supply-chain risk management.

168 Rather than developing a new LCIA method or prospective database, we demonstrate how targeted footprint
169 accounting complements existing indicators. By reporting total waste, hazardous waste, and material
170 consumption, and highlighting sectoral hotspots, our analysis shows how footprint accounting makes hidden
171 burdens visible, clarifies interpretive limits, and delivers actionable insights for circular-economy policies and
172 resource-risk management. Importantly, this work also provides a step toward embedding explicit waste-sector
173 dynamics in future pLCA databases, where dedicated transformation modules could capture prevention,
174 recycling, and secondary-material pathways alongside energy and transport transitions.

175 **2 Methodology (1900 words)**

176 **2.1 Selection and creation of pLCA databases**

177 Using the LCI database *ecoinvent* (version 3.9.1) (Wernet et al., 2016) as a basis, we constructed pLCI databases
 178 using *premise* (Sacchi et al., 2022) over ten-year intervals from 2020 to 2050. *premise* links IAM outputs to
 179 background LCI data by regionalising markets and updating technology efficiencies, fuel mixes, and emissions
 180 profiles; in our case, REMIND outputs drove these updates. REMIND is a global energy–economy–climate
 181 model that produces internally consistent projections of energy demand, technology portfolios, and greenhouse-
 182 gas emissions under alternative socio-economic narratives (Aboumaboub et al., 2020). We selected two
 183 contrasting REMIND pathways: SSP1-PkBudg500 and SSP5-PkBudg500. SSP1 (“sustainability”) represents
 184 low challenges to mitigation, rapid diffusion of clean technologies, and lower energy and material intensities.
 185 SSP5 (“fossil-fuelled development”) represents high economic growth coupled with high energy demand and a
 186 strong reliance on fossil fuels, thereby raising mitigation challenges (Bauer et al., 2017; Kriegler et al., 2017;
 187 Van Vuuren et al., 2017).

188 Within the SSP–RCP framework, the “PkBudg500” constraint imposes a stringent cumulative CO₂ budget
 189 consistent with 1.5 °C-class mitigation (often associated with RCP1.9 in the literature), which forces both worlds
 190 to meet a comparable climate target (Van Vuuren et al., 2011). We deliberately use the same PkBudg500
 191 constraint for SSP1 and SSP5 to enhance interpretability of pLCI comparisons. Using the same carbon budget
 192 (PkBudg500) for SSP1 and SSP5 holds climate ambition constant, so differences in the resulting pLCIs reflect
 193 socio-economic and technological structure rather than target stringency. This improves attribution: contrasts in
 194 waste and material footprints stem from patterns of demand, fuel mixes, and infrastructure, not from divergent
 195 radiative-forcing goals. An approximately 500 Gt CO₂ century-scale budget is a 1.5 °C-class constraint (often
 196 associated with RCP1.9), ensuring major energy transitions with material implications (electrification, CCS,
 197 hydrogen, storage) appear in both pathways, though to different extents. Thus, SSP1-PkBudg500 and SSP5-
 198 PkBudg500 share a common climate constraint but diverge structurally, providing a controlled basis for
 199 comparing footprints in prospective LCA (Intergovernmental Panel On Climate Change (IPCC), 2023).

200 **2.2 Waste and material footprinting with T-reX**

201 T-reX is a python program that operates directly on the technosphere to generate inventory-based waste and
 202 material footprints that can be computed like LCIA indicators while preserving exchange-level traceability
 203 (McDowell et al., 2025). After prospective databases are created (Section 2.1), the background is deconstructed
 204 to a flat, exchange-level list (via Brightway/wurst), which makes every technosphere flow addressable by name,
 205 unit, location, and metadata (Mutel, 2017b, 2017a). Pattern-based rules are then applied in two passes. First,
 206 waste detection targets exchanges whose names/units and treatment-chain context denote wastes, including
 207 routings to recycling, composting, anaerobic digestion, incineration, hazardous treatment, and landfill; “hidden”
 208 wastes that would otherwise be consumed inside treatment chains are surfaced at the point of generation, and
 209 hazardousness is taken only from explicit flags in the source inventories to avoid over-tagging from process
 210 names. Second, material demand is inferred from purchases of “market for ...” activities corresponding to single

211 materials or grouped families (e.g., rare earths, critical raw materials), so that footprints reflect supply-chain
 212 demand (including primary and secondary supply, co-production and substitution) rather than extraction events.

213 For each footprint category, the matched technosphere exchanges are mirrored one-to-one into an auxiliary
 214 “pseudo-biosphere” with unit-consistent characterisation factors. This preserves Brightway’s calculation
 215 mechanics while yielding inventory totals (mass or volume) rather than impact-characterised scores; mirrored
 216 flows retain pointers to their source exchanges, enabling decomposition by sector, activity, or individual flow
 217 with full auditability. The same mirroring logic is applied to all database variants (current and premise-aligned),
 218 so temporal and scenario differences arise solely from underlying inventories. In the default configuration, T-reX
 219 provides ten waste categories (duplicated across mass and volume units) and a configurable panel of material-
 220 demand categories (with defaults aligned to the EU CRM list (European Commission, 2023)); both sets are
 221 easily extended by user rules. Together, these design choices allow footprint computation at activity, sector, or
 222 whole-database levels under current or prospective backgrounds while remaining faithful to the system model
 223 and allocation choices embedded in ecoinvent.

224 Operationally, T-reX’s workflow comprises: (i) optional configuration of waste/material categories; (ii) optional
 225 generation of premise-aligned prospective databases; (iii) database expansion to an exchange list; (iv)
 226 identification and categorisation of target exchanges; (v) construction of a T-reX “pseudo-biosphere” database;
 227 (vi) creation of pseudo-LCIA methods; (vii) exchange editing to mirror technosphere flows; and (viii)
 228 verification. The result is a Brightway project containing both the original biosphere and the T-reX pseudo-
 229 biosphere alongside one or more manipulated technosphere databases, ready for footprint calculation using
 230 standard LCA methodology.

231 2.3 Selection of activities in the LCA/pLCA databases

232 We restricted the analysis to a transparent, comparable set of background “market” activities from each LCI
 233 database (baseline ecoinvent and its prospective variants), then harmonised, classified, and merged them.

234 Filters were applied to isolate the activities of interest. By default, we selected only activities whose names begin
 235 with “market for …” and whose activity type equals “market activity”, thereby focusing on market supply nodes
 236 rather than transformation or site-specific producer datasets. To avoid duplication we further restricted locations
 237 to ecoinvent’s global aggregates: GLO (global) and RoW (rest-of-world). We excluded activities that are waste
 238 or service oriented (name or classification containing “recovery”, “treatment”, “disposal”, “waste”, “services”,
 239 “scrap”, “site preparation”, “construction”, “maintenance”) to avoid conflating technosphere waste management
 240 with product supply. Finally, we limited reference units to mass and volume commodities (kilogram, cubic
 241 meter) so that material and waste footprints could be interpreted consistently across the activity set. After
 242 filtering, a total of 1593 activities remained in the selection.

243 This approach intentionally prioritises (i) market-level representativeness over plant-level specificity; (ii)
 244 globally comparable inventories over regional differentiation; and (iii) physically interpretable commodities over
 245 service or energy-only flows. Limitations include potential omission of region-specific markets, energy carriers

246 with non-mass units (e.g., kWh), and any product supplied exclusively via non-market datasets.

247 2.4 Categorisation of activities

248 To enable robust benchmarking across sectors, and within sectors and sub-sectors, we grouped activities using
 249 the Cooperative Patent Classification (CPC) codes stored in the ecoinvent metadata. CPC is the international
 250 standard for product taxonomy that organises goods and services by their material/functional characteristics
 251 (European Patent Office (EPO), 2025). In LCA databases it provides a stable, key for harmonising
 252 heterogeneous activity names (and thus facilitates comparisons that are otherwise noise-prone at the activity
 253 level). We follow prior large-N LCA work that aggregates products to analyse cross-category patterns (e.g.,
 254 (Laurenti et al., 2023)), and rely on the CPC fields available in ecoinvent v3.x (Wernet et al., 2016). Table 1 lists
 255 the number of activities for each category.

256 Each activity was assigned a category and sub-category from CPC ranges, with explicit overrides for edge cases.
 257 Where CPC ranges overlap, later rules supersede earlier ones (e.g., plastics/rubber overrides chemicals). The
 258 resulting alignment used in the study is:

- 259 • **AgriForeAnim**
 - 260 • Agricultural & forestry products: CPC 00000–01999, 03000–03999, 39000–39999
 - 261 • Live animals, fish & their products: CPC 02000–02999, 04000–04999
- 262 • **ProcBio**
 - 263 • Food & beverages, animal feed: CPC 21000–23999, 42000–42999
 - 264 • Textile: CPC 26000–28199
 - 265 • Wood, straw & cork: CPC 31000–31999 (plus CPC 38100)
 - 266 • Pulp & paper: CPC 32000–32999 (plus CPC 38450→Textile)
- 267 • **OreMinFuel**
 - 268 • Ores, minerals & fuels: CPC 11000–17999, 33000–33999, 60000–69999
- 269 • **Chemical**
 - 270 • Chemical products: CPC 18000–18999, 34000–34699, 34800–35499
- 271 • **PlastRub**
 - 272 • Plastics & rubber products: CPC 34700–34799, 35500–36999
- 273 • **GlasNonMetal**
 - 274 • Glass & other non-metallic products: CPC 37000–37999
- 275 • **MetalAlloy**
 - 276 • Basic metals & alloys (incl. semi-finished): CPC 40000–41999
- 277 • **MachElecTrans**
 - 278 • Metal/electronic equipment & parts: CPC 43000–48999, 49941–49999 (plus CPC 38150→Furniture)
 - 279 • Transport vehicles: CPC 49000–49940

280 2.5 Extraction of activity price data

281 Market price data was obtained from the ecoinvent database using brightway2 and a python script written by the
 282 authors (included in the supplementary information, section S3). The first step was to partially import the
 283 ‘ecospold2’ files in the uncompressed database, each of which represents a single activity. Our python script
 284 then scanned the metadata for each activity to extract the price attribute, saving it in a csv file along with the
 285 name and unique identifying code. This data is available in the supplementary information, section S2.

286 2.6 Calculations with LCIA and Waste and Resource Footprint methods

287 For every activity–year–scenario combination, we computed a panel of T-reX footprints together with
 288 benchmark LCIA endpoints. The T-reX panel comprised ten waste footprints (total, hazardous, and route-
 289 specific recovery/disposal categories, each in mass and volume units) and sixty material-demand footprints
 290 (single materials and aggregated classes). Each method was instantiated by creating the corresponding pseudo-
 291 biosphere flows and characterisation tables and then running Brightway calculations on the filtered “market”
 292 activity set (Section 2.3). Scores are returned in the physical units of the mirrored exchanges. Negative material
 293 scores (arising where co-product supply offsets purchases) were retained to reflect the database’s
 294 allocation/substitution logic rather than truncated, and were handled explicitly in interpretation. Quality-
 295 assurance checks covered unit consistency, exclusivity of hazardous tagging to explicitly flagged exchanges, and
 296 routing partitions (recovery vs disposal) summing to total waste within numerical tolerances.

297 In parallel, we calculated ReCiPe 2016 (H) endpoints (human health (DALY), ecosystems (species·year), and
 298 resource scarcity (USD2013)) for the same activity set and all background years/scenarios (Huijbregts et al.,
 299 2016). ReCiPe 2016 was chosen because it is widely adopted, methodologically harmonised, provides a compact
 300 set of interpretable endpoints, and offers a well-documented mapping from midpoints to endpoints at a global
 301 scale. Using endpoints, rather than a large basket of midpoints, supports concise comparison with inventory-level
 302 signals (waste/material footprints) when examining whether waste- or material-intensive sectors are also
 303 damage-intensive.

304 All calculations were executed in Brightway on ecoinvent v3.9.1 and on premise-generated prospective
 305 backgrounds aligned to REMIND scenarios, ensuring that regionalised markets, technology efficiencies, fuel
 306 mixes, and emissions profiles propagate identically into both sets of indicators. This design enables like-for-like
 307 comparisons across activities, sectors, years, and scenarios, and allows interrogation of divergences between
 308 LCIA damage and inventory-based footprints—for example, those driven by allocation effects in ecoinvent or by
 309 the current under-specification of future waste-system transformations in prospective databases.

310 2.7 Calculation of waste circularity ratio

311 Waste circularity (C_w) was calculated as the proportion of total waste that is routed to recovery-oriented
 312 treatment rather than final disposal. For each activity, total waste generation (W_{total}) was compared against the
 313 summed quantities of waste that are recycled, composted, or anaerobically digested. The indicator was defined
 314 as:

$$315 \quad C_w = \frac{W_{recycled} + W_{composted} + W_{digested}}{W_{total}} * 100$$

316 This formulation captures the share of waste that remains circulating within the technosphere, providing a simple
 317 mass-balance measure of material recovery. $W_{recycled}$ includes both mechanical and chemical recycling processes;
 318 $W_{composted}$ represents organic fractions entering aerobic composting; and $W_{digested}$ covers biogenic waste treated
 319 through anaerobic digestion. All three components were identified from *T-reX* waste exchange data based on
 320 process names and CPC classifications. Activities with $C_w=0$ correspond to fully linear waste pathways, whereas
 321 higher percentages indicate greater reintegration of materials into productive use and thus higher degrees of
 322 circularity.

323 2.8 Calculation of waste hazardousness ratio

324 The waste hazardousness ratio (H_w) was calculated to indicate the share of total waste that is classified as
 325 hazardous within each activity. It expresses the proportion of all outgoing waste flows identified as hazardous
 326 (e.g., toxic, corrosive, flammable, or otherwise regulated) relative to the total waste generated. The indicator was
 327 defined as:

$$328 \quad H_w = \frac{W_{hazardous}}{W_{total}} * 100$$

329 where $W_{hazardous}$ represents the mass of all waste exchanges labelled as hazardous in the database metadata, and
 330 W_{total} is the sum of all waste outputs from the activity, irrespective of classification. This metric provides a
 331 normalised measure of waste toxicity potential at the inventory level, allowing comparisons across sectors
 332 independent of total waste magnitude. A higher H_w value denotes a larger fraction of hazardous waste within an
 333 activity's total waste profile, while lower values indicate predominantly non-hazardous material streams.

334

335 **3 Results (3600 words)**

336 **3.1 Total waste footprints across sectors**

337 Table 2 and Figure 1 together depict the distribution of total waste footprints across the main industrial
 338 categories. Both the descriptive statistics and the boxplot highlight the extreme skewness of waste generation
 339 within the technosphere: while most activities produce relatively modest quantities of waste, a small subset of
 340 heavy-industrial processes contributes disproportionately large amounts. Metals and alloys dominate, exhibiting
 341 median values two to three orders of magnitude higher than most other sectors and an extended upper tail driven
 342 by mining, smelting, and refining processes. The chemical and machinery–electronics–transport categories also
 343 display broad interquartile ranges and numerous outliers, underscoring their structural complexity and diversity
 344 of production scales. In contrast, agriculture, forestry, and animal products and non-metallic minerals cluster
 345 tightly around low median values, indicating generally limited waste generation per functional unit. The log-
 346 scaled spread observed in Figure 1 emphasises that even within individual categories, waste intensity can vary by
 347 up to six orders of magnitude, reflecting differences in process technology, regional supply-chain composition,
 348 and allocation effects. Overall, these patterns confirm that waste formation is highly concentrated in material-
 349 and energy-intensive industries, reinforcing the need for targeted circularity interventions in metallurgical and
 350 chemical value chains rather than diffuse, economy-wide measures.

351 The activity-level maxima reported in Table 3 identify the processes that anchor these upper tails and clarify why
 352 sectoral aggregates skew so strongly. In chemicals, the top entries are lutetium oxide, thulium oxide, and heavy
 353 water, each with extraordinary waste intensities—on the order of 10^8 kg waste per kg product (6.04×10^8 ; $1.61 \times$
 354 10^8 ; 1.57×10^8 , respectively)—and high prices (€165–620 in 2005 euros per kg). These values are consistent
 355 with ultra-selective separations from dilute feeds (e.g., multi-stage solvent extraction for rare earths; isotope
 356 separation for D₂O), where low yields, extensive reagent use, and large raffinate streams dominate the footprint.
 357 In metals and alloys, gold–silver ingots (5.90×10^8 kg/kg), unrefined silver (5.37×10^8 kg/kg), and platinum
 358 (2.42×10^8 kg/kg; €20,600/kg) likewise exhibit extreme intensities aligned with very low ore grades and residue-
 359 rich pyrometallurgical–hydrometallurgical chains; these few activities materially shape the category’s long upper
 360 tail. Machinery–electronics–transport is led by integrated circuits (logic and memory types: $1.75\text{--}1.76 \times 10^7$
 361 kg/kg) and active electronic components (1.53×10^7 kg/kg; high unit prices), a pattern compatible with clean-
 362 room manufacturing that relies on ultra-pure inputs, high consumable use, and yield losses across many steps.

363 Other categories show the same mechanism—outlier processes dominate within otherwise modest distributions.
 364 In ores–minerals–fuels, enriched uranium products ($\sim 1.05\text{--}1.09 \times 10^7$ kg/kg; €586/kg) top the list, reflecting
 365 enrichment tails and extensive upstream processing. In processed bio-based products, silk items—reeled raw silk
 366 hank (2.79×10^6 kg/kg; €18.88/kg) and silk yarn (7.75×10^5 kg/kg; €31.01/kg)—and large-fish canning ($1.02 \times$
 367 10^6 kg/kg; €0.65/kg) point to high volumes of aqueous effluents and organic residues per kilogram of high-value
 368 output. For plastics and rubber, high-volume commodities such as PVC (emulsion and bulk polymerisation) and
 369 LLDPE occupy the top three ($\sim 4.1\text{--}4.4 \times 10^5$ kg/kg) despite low prices (€1.29/kg), indicating that large absolute
 370 waste burdens can arise even where unit values are low. Non-metallic minerals are led by legacy and specialised

371 glass products—CRT panel glass (8.03×10^4 kg/kg), solar collector glass tubes with silver mirrors (4.79×10^4
 372 kg/kg), and glass fibre (2.91×10^4 kg/kg), where coating, forming, and cullet management contribute
 373 disproportionately relative to unit mass. Agriculture, forestry, and animal products show a similar outlier
 374 structure: cocoons (2.66×10^5 kg/kg; €8.26/kg), swine for slaughter (1.34×10^5 kg/kg; €5.48/kg), and greasy
 375 sheep fleece (5.88×10^4 kg/kg; €2.82/kg) concentrate aqueous and organic by-product streams in a handful of
 376 items, while most agricultural commodities remain near the low category median.

377 Two cross-cutting implications follow from Table 3. First, the sectoral tails are shaped by processes
 378 characterised by either extreme selectivity (rare-earth oxides, heavy water, semiconductor devices) or very low
 379 natural concentrations (precious metals, platinum-group metals, nuclear fuels), where large material throughputs
 380 and auxiliary inputs are intrinsic to achieving specification, hence high waste per kilogram of final product.
 381 Second, price and waste intensity are only loosely coupled: some of the highest waste intensities coincide with
 382 very high prices (platinum, integrated circuits), but others occur in low-price, high-volume goods (PVC,
 383 LLDPE), implying that prioritisation should consider both mass-based contributions and economic leverage.
 384 These observations reinforce the case for targeted interventions: improving yields and reagent recovery in
 385 separation-intensive chains (rare earths, precious metals, nuclear fuels), and scaling process-control and scrap-
 386 looping strategies in polymers, glass, and electronics, where small fractional improvements at very large scales
 387 can meaningfully suppress the long-tail contribution to the technosphere's aggregate waste footprint.

388 3.2 Waste circularity across sectors

389 Figure 2a illustrates the distribution of waste circularity (C_w) across the eight aggregated industrial categories.
 390 Overall, circularity remains low, with medians below 5% in every category (agriculture/forestry/animal products
 391 2.48%, processed bio-based products 1.11%, chemicals 0.862%, glass/non-metallics 0.843%, ores/minerals/fuels
 392 0.613%, plastics/rubber 0.421%, metals/alloys 0.492%, and machinery-electronics-transport 0.286%). This
 393 confirms that (as modelled by ecoinvent 3.9.1) only a small share of waste is presently routed to recovery via
 394 recycling, composting, or anaerobic digestion.

395 The wide spreads in a few categories reflect identifiable outliers. In agriculture/forestry/animal products, several
 396 biogenic commodities exceed 10%—notably vanilla (14.58%), green coffee (14.01%), and processing tomatoes
 397 (13.76%). Processed bio-based products show the highest maxima overall—cottonseed oil (16.16%) and
 398 cottonseed meal (15.99%), which is consistent with well-established by-product recovery chains in the industry.
 399 More modest but still notable recoveries occur in glass/non-metallics (borosilicate glass tubes 6.26%),
 400 ores/minerals/fuels (tungsten concentrate 5.01%; steatite 6.01%), and a handful of machinery/electronics items
 401 (electron gun for CRT displays 4.26%). Chemicals are mostly near zero but include a few recovery-rich lines
 402 (e.g., helium, crude stockpiling 9.65%). By contrast, metals/alloys and plastics/rubber rarely exceed 2–3%, with
 403 isolated cases such as molybdenum trioxide (2.25%) and phenolic resin (2.17%) marking the upper tails.

404 Taken together, these statistics reinforce a predominantly linear metabolism: even where outliers exist, most
 405 activities in metals, chemicals, and high-volume manufacturing sit near zero circularity. Improving
 406 representation of future waste-management transformations in prospective LCA databases (and targeting the

407 specific hotspots identified above) will be essential if circularity gains are to be credibly reflected in scenario
 408 analyses.

409 3.3 Waste hazardousness across sectors

410 Figure 2b shows the share of each activity's total waste that is classified as hazardous (H_w). Across the
 411 technosphere, hazardous fractions are generally small. Most categories cluster close to zero with medians around
 412 0–2% (plastics/rubber 0.198%, chemicals 0.081%, machinery-electronics-transport 0.105%, metals/alloys
 413 0.033%, ores/minerals/fuels 0.226%, glass/non-metallics 0.085%, processed bio-based products 0.086%, and
 414 agriculture/forestry/animal products 0.109%). Distributions are nevertheless fat-tailed. Plastics and rubber has
 415 the highest central tendency (mean 3.85%) and the broadest spread, with a long upper tail reaching into the tens
 416 of percent; the top activities include styrene-acrylonitrile (42.6%), ABS (40.9%), and PVDC granulate (26.5%),
 417 consistent with solvent- and additive-rich streams. Chemicals retain a low median but show persistent double-
 418 digit outliers, e.g., tebuconazole (11.5%), semiconductor-grade gallium (11.3%), and carbon tetrachloride
 419 (10.0%). Ores/minerals/fuels also exhibit high outliers despite a low median, led by pipeline olefins such as
 420 ethylene (24.6%) and propylene (23.3%).

421 Machinery-electronics-transport features sporadic peaks (aluminium collector foil for Li-ion cells 6.09%;
 422 carbon-fibre reinforced plastic 5.74%; LCD polariser stacks 1.39%), while metals/alloys remains tightly centred
 423 but includes forming/drawing steps with elevated shares (aluminium sheet rolling 5.20%; steel pipe drawing
 424 5.04%; copper wire drawing 2.63%). Glass/non-metallics is low-centred yet contains bituminous adhesive
 425 compounds among its highest values (3.61% hot; 3.61% cold) alongside ceramic tiles (0.84%). Agriculture and
 426 processed bio-products cluster near zero but still present isolated cases—marine fish (0.97%), tropical hardwood
 427 sawlogs (0.93%), reeled raw silk (1.26%), and certain fish products (1.1%)—that should not be overlooked.

428 Taken together, these results indicate that hazardousness is weakly coupled to total waste magnitude: material-
 429 intensive sectors dominate in tonnes, but hazardous fractions are concentrated in specific sub-processes within
 430 plastics/rubber, chemicals, selected ore/fuel supply chains, and niche manufacturing steps. Prioritisation should
 431 therefore consider both dimensions—volume and H_w —to avoid overlooking small but risk-relevant streams.

432 3.4 Material demand footprints across sectors

433 While sixty material-demand footprints were computed (full results in Supplementary Information S2), we focus
 434 here on two policy-salient indicators with contrasting patterns of concentration and pervasiveness: natural gas
 435 and rare-earth elements (REEs).

436 3.4.1 Natural gas demand

437 Figure 4 indicates that natural-gas demand is pervasive across the technosphere yet strongly right-skewed, with a
 438 handful of activity types anchoring the upper tail. Medians reveal the broadly distributed baseline—machinery-
 439 electronics-transport highest (4.95×10^3 kg gas per kg product), then metals and alloys (1.14×10^3), chemicals
 440 (6.43×10^2), plastics and rubber (3.77×10^2), processed bio-based products (7.85×10^1), ores-minerals-fuels
 441 (5.56×10^1), glass/non-metallics (3.30×10^1), and agriculture/forestry/animal products (3.18×10^1) but means

442 are pulled upward by extreme outliers. In metals and alloys the tail is dominated by precious-metal refining, with
 443 gas intensities of 1.15×10^7 , 7.65×10^6 , and 3.85×10^6 kg/kg for unrefined gold, gold, and platinum,
 444 respectively; these alone explain the large mean–median separation in that category. Electronics exhibits
 445 similarly elevated hotspots—integrated circuits (logic and memory) and active components at 3.00×10^5 , $2.32 \times$
 446 10^5 , and 2.41×10^5 kg/kg—consistent with multi-step, yield-sensitive thermal processing. In ores–minerals–
 447 fuels, enriched-uranium products cluster around 4.01×10^5 kg/kg, reflecting enrichment and fuel-element
 448 fabrication. Chemicals show a modest median but wide spread due to gas’s dual role as heat and feedstock, with
 449 lutetium oxide, scandium oxide, and heavy water at 9.20×10^5 , 5.04×10^5 , and 4.39×10^5 kg/kg. Categories with
 450 lower central tendencies still present specialised high-gas outliers, such as glass tubes with silver mirrors ($1.67 \times$
 451 10^3 kg/kg), sanitary ceramics and basic refractories ($1.0\text{--}1.03 \times 10^3$), PVF films and dispersions ($5.6\text{--}5.0 \times 10^3$),
 452 and silk products (1.25×10^4 and 3.35×10^3). The agricultural category, while more clustered and having the
 453 lowest median value, still presents high demand outliers such as cocoons, cashew, and tilapia which reach $1.47 \times$
 454 10^3 , 1.02×10^3 , and 7.32×10^2 kg/kg. Occasional small negative minima are numerically negligible and reflect
 455 allocation/crediting artefacts rather than genuine net production. Overall, natural-gas use is diffuse at baseline
 456 but aggregate burdens are dominated by a narrow set of thermal-intensive hotspots in metallurgical, electronic,
 457 nuclear-fuel, and selected specialty lines—implying that targeted efficiency upgrades and fuel switching in these
 458 tails will deliver the largest system-wide reductions.

459 **3.4.2 Rare earth element demand**

460 Figure 5 shows that rare-earth element (REE) demand is highly concentrated and strongly right-skewed. Medians
 461 reveal the underlying pattern: machinery–electronics–transport sits highest (3.24×10^{-1} kg REE per kg product),
 462 followed by metals and alloys (7.14×10^{-2}), chemicals (2.79×10^{-2}), and plastics/rubber (2.01×10^{-2}); all other
 463 categories cluster near the floor (processed bio-products 3.65×10^{-3} ; agriculture/forestry/animal products $3.12 \times$
 464 10^{-3} ; ores–minerals–fuels 2.22×10^{-3} ; glass/non-metallics 1.16×10^{-3}). Means, however, are dominated by a
 465 small number of REE-specific markets, most starkly in chemicals, where the mean rises to 1.78×10^4 kg/kg
 466 despite a near-zero median. This tail is anchored by scandium oxide (1.16×10^7 kg/kg), lutetium oxide ($2.52 \times$
 467 10^5), and thulium oxide (6.74×10^4). Metals and alloys show similar tail behaviour (mean 1.36×10^2 ; max $8.80 \times$
 468 10^3), driven by alloying and catalyst lines such as ferroniobium 66% (8.80×10^3), platinum (1.30×10^3), and
 469 metal catalysts for catalytic converters (9.52×10^2). Machinery–electronics–transport combines the highest
 470 central tendency with long upper tails (mean 3.99×10^1 ; max 2.64×10^3), reflecting magnet and battery supply
 471 chains (e.g., LaNi₅ positive electrodes (2.64×10^3), NiMH prismatic batteries (8.85×10^2), and permanent
 472 magnets for electric motors (7.42×10^2)). Ores–minerals–fuels exhibit sporadic but high-intensity demands
 473 (mean 7.47×10^1 ; max 5.68×10^3), led by pyrochlore concentrate (5.68×10^3) and enriched-uranium fuel
 474 elements (3.87×10^2). Plastics/rubber, glass/non-metallics, and processed bio-products have low medians and
 475 modest spreads but include identifiable outliers such as tetrafluoroethylene film on glass (7.29×10^1), LCD glass
 476 (1.30), and reeled raw silk hank (1.01). Agriculture/forestry/animal products remain close to the floor overall,
 477 though seed and cocoon markets register small but non-negligible purchases (cocoons 1.21×10^{-1} ; fodder beet
 478 and sugar beet seed 9.5×10^{-2}). Occasional negative minima (down to about -3×10^{-4} kg/kg) are numerically

479 negligible and reflect allocation or substitution credits rather than genuine negative demand. In sum, the REE
 480 footprint is tail-dominated: database-wide purchases are governed by a narrow set of specialised activities in
 481 REE processing, alloying, magnets, and advanced components, implying that targeted interventions in these
 482 chains will be far more effective than diffuse, economy-wide measures.

483 3.5 ReCiPe LCIA results across sectors

484 Values from ReCiPe's endpoint LCIA methods are not directly comparable in magnitude to inventory footprints;
 485 we therefore discuss rankings and trends only.

486 3.5.1 ***Damage to resource availability***

487 For damage to resource availability (Figure 6), medians indicate the broad centre of pressure sits in machinery–
 488 electronics–transport (2.52×10^3), followed by plastics/rubber (7.20×10^2) and metals/alloys (7.27×10^2), with
 489 chemicals (5.39×10^2) close behind; ores–minerals–fuels (1.88×10^2), processed bio-products (5.25×10^1),
 490 glass/non-metallics (4.03×10^1), and agriculture/forestry/animal products (3.35×10^1) form a lower tier. Means,
 491 however, reveal extreme right tails, most pronounced in metals/alloys (mean 2.30×10^5 ; max 6.49×10^6) and
 492 chemicals (mean 2.83×10^4 ; max 4.65×10^6), driven by a narrow set of activities. In metals/alloys the tail is
 493 anchored by platinum and gold (6.49×10^6 , 4.17×10^6 , 3.99×10^6), while in chemicals it is rare-earth oxides
 494 (samarium–europium–gadolinium, praseodymium–neodymium, and REO concentrates at 4.65×10^6 , 1.53×10^6 ,
 495 1.48×10^6). Machinery–electronics–transport combines a high median with notable outliers tied to magnet and
 496 battery chains (LaNi electrodes 1.14×10^6 ; permanent magnets 4.02×10^5 ; NiMH batteries 3.83×10^5). Ores–
 497 minerals–fuels show sporadic but sizeable points (enriched uranium products 2.49×10^5). Plastics/rubber's
 498 relatively high median is shaped by fluoropolymer lines (tetrafluoroethylene film/monomer and
 499 polyvinylfluoride dispersion at 4.84×10^3 – 4.54×10^3). In glass/non-metallics, LCD glass and hard materials
 500 (silicon carbide, battery-grade synthetic graphite) sit atop the distribution (1.06×10^3 ; 600). Processed bio-
 501 products and agriculture feature much lower medians but still contain expensive, high-scarcity items (reeled raw
 502 silk 8.46×10^3 ; cocoons 996).

503 In relation to our inventory footprints, resource scarcity concentrates even more sharply in precious-metal, REE
 504 and magnet/battery chains, overlapping with REE demand tails and parts of the waste tails, but diverging from
 505 diffuse natural-gas dependence.

506 3.5.2 ***Damage to human health***

507 For damage to human health (Figure 7), medians place machinery–electronics–transport highest (1.04×10^{-1}),
 508 followed by metals/alloys (5.52×10^{-2}), with plastics/rubber (8.12×10^{-3}) and chemicals (9.33×10^{-3}) forming a
 509 middle tier; ores–minerals–fuels (1.68×10^{-3}), processed bio-products (2.73×10^{-3}), glass/non-metallics ($1.31 \times$
 510 10^3) and agriculture/forestry/animal products (2.39×10^3) cluster lower. Means, however, expose extreme right
 511 tails, most striking in metals/alloys (mean 3.43×10^1 ; max 1.09×10^3), reflecting precious metal chains that
 512 dominate category totals (platinum 1.09×10^3 ; metal catalyst for catalytic converters 6.55×10^2 ; gold $5.09 \times$
 513 10^2). Machinery–electronics–transport combines a high centre with large outliers tied to semiconductor and
 514 component manufacture (integrated circuits and active components at 8.23, 5.69, and 3.31). Ores–minerals–fuels

515 show a modest median yet sizeable extremes from nuclear-fuel steps (enriched uranium fuel elements 20.4).
 516 Chemicals display a near-zero median but contain REE oxide hotspots (lutetium 17.9; scandium 11.7; thulium
 517 4.77). Plastics/rubber, glass/non-metallics, and processed bio-products remain low-centred but include
 518 identifiable high lines (e.g., tetrafluoroethylene film 0.161; LCD glass 0.024; reeled raw silk 0.635).
 519 Agriculture's top entries are ruminant liveweight markets (weaned calves/heifers and cattle 0.46–0.71), but most
 520 activities sit near the lower tail.

521 Relative to the inventory footprints, human-health damage overlaps strongly with the waste and material
 522 hotspots for metals/alloys and parts of machinery–electronics–transport, where precious metals, REEs, and
 523 complex processing co-locate. Divergences are also evident: categories with diffuse natural-gas dependence do
 524 not necessarily translate into high DALY medians, and some high-waste lines (bulk polymers, glass) contribute
 525 less to endpoint damage than precious-metal and REE chains. This suggests pairing tail-targeted material
 526 strategies (in precious metal, REE and semiconductor routes) with cross-cutting energy and emissions controls,
 527 recognising that waste tonnage and health damage are governed by related but distinct mechanisms.

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529 **3.5.3 *Damage to ecosystems***

530 For damage to ecosystems (Figure 8), on median values, machinery–electronics–transport sits highest (1.54×10^{-4}), followed by metals/alloys (6.15×10^{-5}). A lower tier clusters around $1\text{--}2 \times 10^{-5}$ —agriculture/forestry/animal
 531 products (1.79×10^{-5}), chemicals (1.82×10^{-5}), plastics/rubber (1.62×10^{-5}) and processed bio-products ($1.50 \times$
 532 10^{-5})—while glass/non-metallics and ores–minerals–fuels lie near the floor (3×10^{-6}). Means reveal a strongly
 533 right-skewed distribution dominated by metals/alloys (mean 3.74×10^{-2} ; max 1.20), with notable but much
 534 smaller tails in ores–minerals–fuels (mean 9.36×10^{-4}) and machinery–electronics–transport (mean 5.23×10^{-4}).

536 The upper tails are anchored by a narrow set of activities. In metals/alloys, platinum (1.20), metal catalysts for
 537 catalytic converters (0.729), and gold (0.474) dominate category totals—consistent with precious/PGM supply
 538 chains driving ecosystem damage. In machinery–electronics–transport, integrated circuits and active components
 539 sit at the top (9.8×10^{-3} , 7.2×10^{-3} , 5.3×10^{-3}), reflecting semiconductor fabrication's energy- and chemical-
 540 intensive steps. Chemicals show REE oxides as clear hotspots (lutetium 0.041, thulium 0.011, scandium 0.010),
 541 while ores–minerals–fuels register enrichment and fuel-element steps (0.011–0.0113). Categories with low
 542 centres still feature identifiable outliers: fluoropolymer lines in plastics/rubber (4.24×10^{-4} and 4.02×10^{-4}), LCD
 543 glass and hard materials in glass/non-metallics (3.82×10^{-5} to 3.11×10^{-5}), and silk products in processed bio-
 544 products (1.93×10^{-3}). Agriculture's tail is led by ruminant liveweight and fleece markets (6.99×10^{-4} to $4.56 \times$
 545 10^{-4}), though the median remains low.

546 Relative to the waste and material footprints, ecosystem damage concentrates where precious metals, REEs, and
 547 advanced components co-locate, overlapping with REE demand and parts of the waste tails, yet it diverges from
 548 diffuse natural-gas dependence (which raises energy use broadly but does not uniformly translate into high
 549 species-year damage). This implies pairing tail-targeted measures in REE, precious metal, and semiconductor
 550 routes with cross-cutting controls on emissions and process chemicals to address ecosystem risks most

551 effectively.

3.6 Temporal and scenario trends in waste footprints

553 Figure 3 synthesises temporal trends in nine inventory-based waste and material indicators, reported as
 554 distributions across the 1593 selected market activities for 2020–2050 under SSP1-PkBudg500 and SSP5-
 555 PkBudg500. We describe medians and dispersion (min–max, standard deviation) to emphasise central tendencies
 556 while acknowledging persistently fat-tailed behaviour.

557 Captured CO₂ routed to storage (a) grows from a near-zero median in 2020 to substantial magnitudes by 2050 in
 558 both pathways, reflecting widespread deployment of CCS under the shared carbon budget. The median rises to
 559 472 kg/kg (SSP1) and 388 kg/kg (SSP5) by 2050, with very wide spreads that expand over time (maxima reach
 560 5.5×10^7 kg/kg and 3.4×10^7 kg/kg, respectively), indicating a small set of CCS-intensive activities emerging as
 561 outliers.

562 Total waste generation (b) increases steadily in both scenarios. Median values move from 3.55×10^3 in 2020 to
 563 4.89×10^3 kg/kg (SSP1) and 4.51×10^3 kg/kg (SSP5) by 2050, i.e., +38% and +27%. Distributions remain broad
 564 and heavy-tailed throughout (stable maxima around 3.0×10^8 kg/kg), suggesting that sectoral heterogeneity
 565 persists even as backgrounds evolve.

566 The hazardousness share (c) declines modestly over time, with a stronger reduction in SSP1. Medians fall from
 567 0.094% to 0.076% in SSP1 and to 0.089% in SSP5. Despite lower medians, variability remains large (standard
 568 deviations around 2 across years) and maxima stay high (declining from 41% to 34–36%), consistent with a long
 569 tail of hazardous-waste-intensive processes that standard scenario updates do not remove.

570 The waste circularity ratio (d) slips slightly in both pathways. SSP1 medians move from 0.880 to 0.819 (−6.9%),
 571 while SSP5 shifts from 0.879% to 0.844%. This soft deterioration occurs alongside rising total waste and only
 572 modest gains in specific recovery routes, implying that disposal grows faster than recovery for the median
 573 activity. Spread narrows only marginally (standard deviation drops from 1.95 to 1.48–1.59), indicating limited
 574 convergence.

575 Route-specific categories confirm these patterns. Landfilling (e) increases at the median by 41% in SSP1 (301 to
 576 424 kg/kg) and 28% in SSP5 (301 to 384), with maxima extending from 7.0×10^6 to $9.6\text{--}9.7 \times 10^6$ kg/kg.
 577 Recycling (f) rises more gently (16% in SSP1 (40.1 to 46.7 kg/kg) and 13% in SSP5 (40.1 to 45.2 kg/kg)),
 578 insufficient to offset the faster growth in total waste. Composting (g) exhibits the sharpest relative gain from a
 579 very low base: medians increase from 6.93×10^{-3} kg/kg to 1.14×10^{-2} kg/kg in SSP1 and to 9.01×10^{-3} kg/kg in
 580 SSP5; however, the absolute levels remain negligible for most activities, and dispersion is dominated by a small
 581 number of large organic streams (max 4.53×10^4 kg/kg throughout). Waste incineration (h) grows modestly
 582 (~15% in both SSPs), with medians rising from 11.4 kg/kg to 13.1 kg/kg and stable, wide ranges (max 3.5×10^5
 583 kg/kg). Open burning (I) edges upward by 10% in both scenarios (0.888 kg/kg to 0.982 kg/kg in SSP1; 0.888
 584 kg/kg to 0.975 kg/kg in SSP5). Its tail remains high: maxima hover around 2.9×10^4 , pointing to persistent
 585 uncontrolled disposal hotspots.

586 Across indicators, SSP1 tends to show slightly higher 2050 medians for total waste and landfilling than SSP5,
587 despite the sustainability narrative, while achieving a larger reduction in hazardousness share. SSP5 often shows
588 slightly less deterioration in circularity. The shared CO₂ budget drives CCS growth in both cases, with the SSP1
589 median exceeding SSP5 by 2050, yet variability is so large that pathway differences are dwarfed by between-
590 activity heterogeneity. Overall, the box-plot distributions indicate incremental change rather than step-changes:
591 recovery routes expand, but not fast enough to prevent a gradual decline in circularity and rising disposal for the
592 median activity; fat tails remain for nearly all categories, underscoring the importance of targeted interventions
593 in the most waste-intensive and poorly managed nodes of the technosphere.

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4 Discussion (1200 words)

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4.1 What this study adds

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This study provides a macro-level prospective quantification of waste and material footprints (WMFs) across 1593 market activities, offering system-wide benchmarks from 2020 to 2050. The analysis reveals strong sectoral contrasts. Mining, metals, and basic chemicals dominate both total and hazardous waste outputs, while services and light manufacturing generate lower footprints per unit output. These sectoral patterns align with ReCiPe endpoint results, indicating that WMFs serve as credible proxies for environmental damage, especially in human health and ecosystem quality categories. This reinforces conclusions from Laurenti et al. (2023), who found consistent correlations between waste intensity and endpoint damage scores.

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In addition to total waste flows, our study differentiates waste circularity and hazardousness across sectors. Sectors with established recycling networks (e.g. ferrous metals) exhibit high circularity, while others (e.g. mixed municipal and construction waste) remain predominantly linear. Our results confirm that waste quality attributes such as circularity percentage and hazardousness meaningfully influence associated impact profiles. For example, high hazardous waste does not always translate to high damage scores, reflecting controlled treatment processes that limit environmental release—an insight consistent with Laurenti et al.'s findings.

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We also quantify sectoral material demand footprints, which typically mirror waste outputs but provide additional insights—particularly for high-tech products that use critical materials in small volumes. These results capture potential raw material supply risks not evident from waste alone and offer a second, complementary lens on throughput.

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Crucially, we assess WMF trajectories under IAM-aligned prospective scenarios. While absolute waste volumes tend to increase with economic activity, waste intensity shows only modest improvement over 2020–2050 in both SSP1 and SSP5-based pathways. Small scenario differences do emerge, however, many sectors remain static in their waste intensities, reflecting limited representation of circularity transformations in the premise-modified pLCIs. This underlines a methodological blind spot: unlike the energy system, the waste system remains largely “frozen” in today’s conditions. Without explicit modelling of future waste system trajectories, pLCA may underestimate the dynamics of waste-related environmental burdens.

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4.2 Strengths of the approach

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This study advances methodological practice by integrating WMF accounting directly into LCA using a “pseudo-LCIA” approach via T-reX. By tracking waste and material flows from technosphere inventories in physical units, our method maintains full compatibility with standard LCA workflows while improving interpretability. Unlike damage-based footprints that rely on complex weighting or cause-effect chains, our results offer direct observables such as kilograms of waste produced or kilograms of natural gas consumed. This makes them more communicable, especially for circular economy audiences focused on mass flows and reuse potential.

631 Moreover, we align WMF tracking with scenario-aligned background databases using the premise framework.
632 This ensures that footprint results reflect upstream system decarbonisation and technological shifts from IAM
633 outputs. While IAM-pLCA integration has typically focused on carbon and energy flows, our study broadens its
634 scope by tracing non-emission material and waste flows across time and policy futures.

635 Interpretively, the T-reX model repositions waste not as a background service flow (as in ecoinvent's disposal
636 treatment convention) but as an observable technosphere output. This realigns LCA with a more physical and
637 intuitive accounting structure—closer to mass-balance logic and consistent with recent critiques of the “waste-
638 as-service” model (Guinée & Heijungs, 2021).

639 Overall, the approach improves transparency, strengthens system boundaries for resource accounting, and links
640 pLCA to circular economy metrics in a credible and actionable way.

641 **4.3 Limitations and caveats**

642 Several limitations temper the interpretation of our results. First, the footprint calculations depend on how waste
643 is classified and allocated in the underlying LCI. We inherit ecoinvent's economic allocation conventions, which
644 can suppress mass-based waste signals in high-value co-product systems and exaggerate them in low-value
645 processes. For example, by-products with minimal market value may appear heavily waste-intensive, even if
646 physically minor. While such allocation rules are standard, they limit the physical interpretability of absolute
647 WMF values. Future work should explore physical or consequential allocation to better align with material mass
648 flows.

649 Second, the classification of what counts as “waste” is context- and time-dependent. Industrial residues reused in
650 another process (e.g. blast furnace slag) are excluded from waste tallies, while functionally similar materials
651 discarded in landfills are counted. As material markets evolve, these boundaries may shift, requiring adaptive
652 definitions.

653 Third, our method does not yet differentiate between the environmental risk of different wastes. One tonne of
654 inert mining waste and one tonne of toxic sludge carry very different impacts but contribute equally to the total
655 WMF. Users must therefore interpret footprint magnitudes in the context of waste composition and fate.
656 Incorporating fate-specific indicators or hazard-weighted metrics would increase the decision relevance of the
657 method.

658 A further limitation is the limited coverage of sectoral transformation in the scenario-linked pLCIs. Premise
659 currently modifies key sectors (energy, cement, steel, transport) but does not adjust many manufacturing or
660 waste management processes. As a result, WMFs remain static across many sectors, despite plausible
661 expectations of circularity gains. Bisinella et al. (2024) and van der Giesen et al. (2020) have flagged this same
662 gap, arguing that pLCA models must evolve to capture circular economy dynamics explicitly, especially in waste
663 recovery, substitution, and material efficiency.

664 Uncertainty also remains intrinsic to ex-ante LCA. While we use established SSP-based scenarios, actual

665 developments could diverge significantly. Moreover, spatial differentiation is limited: our results are global
666 averages, and local waste impacts or recycling potentials are not captured. A kilogram of waste in a region with
667 landfill scarcity or weak regulation may cause much more harm than the same mass in a highly regulated
668 context.

669 Together, these caveats point to key directions for future refinement: dynamic waste classification, hazard
670 weighting, greater scenario coverage, and spatial resolution.

671 **4.4 Outlook and use**

672 Our findings highlight three areas for immediate uptake and future research.

673 First, WMFs offer a vital complement to standard impact categories in LCA, particularly for assessing alignment
674 with circular economy strategies. As governments and corporations adopt CE targets (e.g. halving residual
675 waste, increasing secondary material shares), they need metrics that connect supply chain performance with
676 waste and material throughput. Our method provides such metrics, grounded in LCI and compatible with
677 standard software. Policymakers can use WMFs to identify high-priority sectors, benchmark progress, and track
678 improvements in circularity over time.

679 Second, scenario-modified WMFs enable policy foresight. Analysts can assess how different decarbonisation or
680 material transition pathways might affect not just emissions but also resource and waste outcomes. This aligns
681 LCA with broader sustainability goals, including resource security and pollution prevention, and supports whole-
682 economy transition planning.

683 Third, our results point to the need for expanded scenario coverage in pLCA. This includes explicit modelling of
684 waste collection, sorting, reuse, and recycling systems—modules that remain underdeveloped in current IAM-
685 pLCA pipelines. Developing such modules, or linking with existing circular economy models (e.g. MFA or
686 stock-flow models), would allow future LCI datasets to better reflect CE policy ambitions.

687 Overall, this study demonstrates that WMFs can be rigorously and transparently integrated into prospective
688 LCA, revealing meaningful patterns across sectors and scenarios. While limitations remain—especially around
689 allocation, waste characterisation, and scenario scope—the methodological advances offer a concrete foundation
690 for embedding circularity into forward-looking environmental assessments.

691

692

693 **5 Conclusions and recommendations (300 words)**

694 This study introduces a systematic framework for quantifying waste and material footprints (WMFs) in
695 prospective life cycle assessment, integrating 1593 market activities and multiple scenario-aligned inventories to
696 track flows from 2020 to 2050. The results demonstrate that WMFs vary widely across sectors, with mining,
697 metals, and chemicals contributing disproportionately to total and hazardous waste generation. These patterns
698 align closely with conventional LCIA damage profiles, reinforcing the interpretive and policy value of WMFs as
699 complementary indicators in sustainability assessments.

700 By distinguishing waste quality attributes (such as circularity potential and hazardousness) and linking them to
701 scenario-based changes in supply chains, the approach reveals important blind spots in current pLCA practices.
702 While energy-related transformations are well captured in IAM-linked databases, circularity measures and waste
703 system evolution remain under-represented. As a result, WMFs show only modest improvements over time,
704 underscoring the need for targeted integration of waste-sector dynamics into future pLCI development.

705 Methodologically, the combination of transparent, physically grounded footprint indicators with scenario-aligned
706 inventories represents a step forward for both LCA and circular economy modelling. The T-reX framework
707 allows practitioners to extract and interpret resource throughput and waste burdens in a manner consistent with
708 established LCA workflows but with far greater clarity on material cycling and discard.

709 Looking ahead, further development is needed to incorporate fate-differentiated waste flows, refine allocation
710 choices for mass-based inference, and extend scenario coverage beyond the currently modelled sectors.
711 Nonetheless, the WMF framework presented here offers a robust platform for linking life cycle thinking to
712 material efficiency, waste prevention, and critical raw material strategies—supporting informed decisions in the
713 context of global sustainability transitions. As circular economy targets become increasingly central to policy,
714 tools that trace waste and material flows through prospective supply chains will be essential.

715

716 **Supplementary Material**

717 The supplementary material supplied in the appendices of this manuscript contain the following sections:

718 S1. Additional figures referenced in the text

719 S2. Complete tabulated data

720 S3. Python scripts used for the production of results

721 **Data availability**

722 All publicly available data related to this manuscript is available in online repositories hosted by Zenodo
723 (<https://doi.org/10.5281/zenodo.16995460>) and Github (https://github.com/Stew-McD/T-reX_LCA-MacroStudy)
724

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729 **CRediT authorship contribution statement**

730 **Stewart Charles McDowall:** Conceptualisation, Methodology, Investigation, Data curation, Formal analysis,
731 Validation, Visualisation, Writing: original draft, Writing: review & editing, Visualisation.

732 **Carlos Felipe Blanco:** Conceptualisation, Methodology, Validation, Writing: review & editing, Funding
733 acquisition, Supervision.

734 **Stefano Cucurachi:** Conceptualisation, Methodology, Validation, Writing: review & editing, Funding
735 acquisition, Supervision.

736 **CRediT authorship visualisation**



738 **Declarations**

739 **Competing interests**

740 The authors declare that they have no known competing financial interests or personal relationships that could
 741 have appeared to influence the work reported in this paper.

742 **Open access**

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751 **Use of artificial intelligence**

752 The authors declare that no generative artificial intelligence tools were used in the generation of the research data
 753 or results reported in this paper. Generative AI was used solely to assist in the editing and refinement of the
 754 manuscript text, with all content reviewed and approved by the authors.

References

- Aboumaboub, T., Auer, C., Bauer, N., Baumstark, L., Bertram, C., Bi, S., Dietrich, J., Dirnaichner, A., Giannousakis, A., Haller, M., Hilaire, J., Klein, D., Koch, J., Körner, A., Kriegler, E., Leimbach, M., Levesque, A., Lorenz, A., Luderer, G., ... Ueckerdt, F. (2020, March 27). *REMIND - REgional Model of INvestments and Development—Version 2.1.0*. <https://www.pik-potsdam.de/research/transformation-pathways/models/remind>
- Akese, G. A., & Little, P. C. (2018). Electronic waste and the environmental justice challenge in Agbogbloshie. *Environmental Justice*, 77–83. <https://doi.org/10.1089/env.2017.0039>
- Arvidsson, R., Söderman, M. L., Sandén, B. A., Nordelöf, A., & others. (2020). A crustal scarcity indicator for long-term global elemental resource assessment in LCA. *The International Journal of Life Cycle Assessment*. <https://doi.org/10.1007/s11367-020-01781-1>
- Bauer, N., Calvin, K., Emmerling, J., Fricko, O., Fujimori, S., Hilaire, J., Eom, J., Krey, V., Kriegler, E., Mouratiadou, I., Sytze De Boer, H., Van Den Berg, M., Carrara, S., Daioglou, V., Drouet, L., Edmonds, J. E., Gernaat, D., Havlik, P., Johnson, N., ... Van Vuuren, D. P. (2017). Shared Socio-Economic Pathways of the Energy Sector – Quantifying the Narratives. *Global Environmental Change*, 42, 316–330. <https://doi.org/10.1016/j.gloenvcha.2016.07.006>
- Beylot, A., Muller, S., Descat, M., Ménard, Y., & others. (2018). Life cycle assessment of the French municipal solid waste incineration sector. *Waste Management*. <https://doi.org/10.1016/j.wasman.2018.08.037>
- Bisinella, V., Schmidt, S., Varling, A., Laner, D., & others. (2024). Waste LCA and the future. *Waste Management*, 53–75. <https://doi.org/10.1016/j.wasman.2023.11.021>
- Carrara, S., Bobba, S., Blagoeva, D., Alves Dias, P., Cavalli, A., Georgitzikis, K., Grohol, M., Itul, A., Kuzov, T., Latunussa, C., Lyons, L., Malano, G., Maury, T., Prior Arce, A., Somers, J., Telsnig, T., Veeh, C., Wittmer, D., Black, C., ... Christou, M. (2023). *Supply chain analysis and material demand forecast in strategic technologies and sectors in the EU – A foresight study*. Publications Office of the European Union. <https://doi.org/10.2760/334074>
- CEN (European Committee for Standardization). (2019). *EN 15804: Sustainability of construction works—Environmental product declarations—Core rules for the product category of construction products*. https://standards.cencenelec.eu/dyn/www/f?p=205:7:0:::FSP_ORG_ID:481830
- Čuček, L., Klemeš, J. J., & Kravanja, Z. (2015). Overview of environmental footprints. In J. J. Klemeš (Ed.), *Assessing and Measuring Environmental Impact and Sustainability* (pp. 131–193). Butterworth-Heinemann. <https://doi.org/10.1016/B978-0-12-799968-5.00005-1>
- Cucurachi, S., van der Giesen, C., & Guinée, J. (2018). Ex-ante LCA of emerging technologies. *Procedia CIRP*, 463–468. <https://doi.org/10.1016/j.procir.2017.11.005>
- Ellen MacArthur Foundation. (2015). *Towards a Circular Economy: Business rationale for an accelerated transition*. <https://www.ellenmacarthurfoundation.org/towards-a-circular-economy-business-rationale-for-an-accelerated-transition>
- European Commission. (2020). *A New Circular Economy Action Plan For a Cleaner and More Competitive Europe*. European Commission. <https://doi.org/10.2779/05068>
- European Commission. (2023). *European Critical Raw Materials Act*. European Commission. https://single-market-economy.ec.europa.eu/publications/european-critical/raw-materials-act_en
- European Patent Office (EPO). (2025). *Cooperative Patent Classification System*. <https://www.cooperativepatentclassification.org/home>
- Giampietro, M., & Saltelli, A. (2014). Footprints to nowhere. *Ecological Indicators*, 610–621. <https://doi.org/10.1016/j.ecolind.2014.01.030>

- Guinée, J. B., Gorrée, M., Heijungs, R., & others. (2002). *Handbook on Life Cycle Assessment. Operational Guide to the ISO Standards*. <https://www.universiteitleiden.nl/en/research/research-projects/science/cml-new-dutch-lca-guide>
- Guinée, J. B., & Heijungs, R. (2021). Waste is not a service. *The International Journal of Life Cycle Assessment*, 1538–1540. <https://doi.org/10.1007/s11367-021-01955-5>
- Guinée, J. B., Heijungs, R., & Huppes, G. (2004). Economic allocation: Examples and derived decision tree. *International Journal of Life Cycle Assessment*. <https://doi.org/10.1007/BF02978533>
- Guinée, J. B., Heijungs, R., Huppes, G., Zamagni, A., & others. (2010). Life cycle assessment: Past, present, and future. *Environmental Science & Technology*, 90–96. <https://doi.org/10.1021/es101316v>
- Hartley, K., Baldassarre, B., & Kirchherr, J. (2024). Circular economy as crisis response: A primer. *Journal of Cleaner Production*, 434, 140140. <https://doi.org/10.1016/j.jclepro.2023.140140>
- Hauschild, M. Z., & Potting, J. (2004). *Spatial differentiation in life cycle impact assessment: The EDIP-2003 methodology. Guidelines from the Danish EPA* (pp. 1–195). Danish Environmental Protection Agency. <https://api.semanticscholar.org/CorpusID:113556375>
- Huijbregts, M. A. J., Steinmann, Z. J. N., Elshout, P. M. F., Stam, G., & others. (2016). ReCiPe2016: A harmonised life cycle impact assessment method at midpoint and endpoint level. *The International Journal of Life Cycle Assessment*. <https://doi.org/10.1007/s11367-016-1246-y>
- Intergovernmental Panel On Climate Change (IPCC) (Ed.). (2023). Mitigation Pathways Compatible with Long-term Goals. In *Climate Change 2022—Mitigation of Climate Change* (1st edn, pp. 295–408). Cambridge University Press. <https://doi.org/10.1017/9781009157926.005>
- International Energy Agency (IEA). (2021). *Net Zero by 2050*. IEA. <https://doi.org/10.1787/c8328405-en>
- International Energy Agency (IEA). (2022). *Renewables 2022*. International Energy Agency (IEA). <https://www.iea.org/reports/renewables-2022>
- Kirchherr, J., Reike, D., & Hekkert, M. (2017). Conceptualizing the circular economy: An analysis of 114 definitions. *Resources, Conservation and Recycling*. <https://doi.org/10.1016/j.resconrec.2017.09.005>
- Kriegler, E., Bauer, N., Popp, A., Humpenöder, F., Leimbach, M., Strefler, J., Baumstark, L., Bodirsky, B. L., Hilaire, J., Klein, D., Mouratiadou, I., Weindl, I., Bertram, C., Dietrich, J.-P., Luderer, G., Pehl, M., Pietzcker, R., Piontek, F., Lotze-Campen, H., ... Edenhofer, O. (2017). Fossil-fueled development (SSP5): An energy and resource intensive scenario for the 21st century. *Global Environmental Change*, 42, 297–315. <https://doi.org/10.1016/j.gloenvcha.2016.05.015>
- Laurenti, R., Demirer Demir, D., & Finnveden, G. (2023). Analyzing the relationship between product waste footprints and environmental damage—A life cycle analysis of 1,400+ products. *Science of The Total Environment*. <https://doi.org/10.1016/j.scitotenv.2022.160405>
- Lenzen, M., Geschke, A., West, J., Fry, J., & others. (2021). Implementing the material footprint to measure progress towards Sustainable Development Goals 8 and 12. *Nature Sustainability*, 157–166. <https://doi.org/10.1038/s41893-021-00811-6>
- McDowall, S. C., Lanphear, E., Cucurachi, S., & Blanco, C. F. (2025). T-reX: Quantifying waste and material footprints in current and future Life Cycle Assessment (LCA) databases. *Resources, Conservation and Recycling*, 222, 108464. <https://doi.org/10.1016/j.resconrec.2025.108464>
- Meinshausen, M., Nicholls, Z. R. J., Lewis, J., Gidden, M. J., Vogel, E., Freund, M., Beyerle, U., Gessner, C., Nauels, A., Bauer, N., Canadell, J. G., Daniel, J. S., John, A., Krumbel, P. B., Luderer, G., Meinshausen, N., Montzka, S. A., Rayner, P. J., Reimann, S., ... Wang, R. H. J. (2020). The shared socio-economic pathway (SSP) greenhouse gas concentrations and their extensions to 2500. *Geoscientific Model Development*. <https://doi.org/10.5194/gmd-13-3571-2020>
- Müller, A., Harpprecht, C., Sacchi, R., Maes, B., Van Sluisveld, M., Daioglou, V., Šavija, B., & Steubing, B.

- (2024). Decarbonizing the cement industry: Findings from coupling prospective life cycle assessment of clinker with integrated assessment model scenarios. *Journal of Cleaner Production*, 450, 141884. <https://doi.org/10.1016/j.jclepro.2024.141884>
- Mutel, C. (2017a). Brightway: An open source framework for life cycle assessment. *Journal of Open Source Software*. <https://doi.org/10.21105/joss.00236>
- Mutel, C. (2017b). *Wurst documentation*. <https://buildmedia.readthedocs.org/media/pdf/wurst/stable/wurst.pdf>
- Pardo, R., & Schweitzer, J. P. (2018). *A Long-term Strategy for a European Circular Economy – Setting the Course for Success* [Policy Paper]. Think2030.
- Pauliuk, S., Arvesen, A., Stadler, K., & Hertwich, E. G. (2017). Industrial ecology in integrated assessment models. *Nature Climate Change*, 7(1), 13–20. <https://doi.org/10.1038/nclimate3148>
- Pellow, D. N. (2023). Environmental justice. In *Handbook on Inequality and the Environment* (pp. 71–85). Edward Elgar Publishing. <https://doi.org/10.4337/9781800881136.00014>
- Reike, D., Vermeulen, W. J. V., & Witjes, S. (2018). The circular economy: New or Refurbished as CE 3.0? — Exploring Controversies in the Conceptualization of the Circular Economy through a Focus on History and Resource Value Retention Options. *Resources, Conservation and Recycling*. <https://doi.org/10.1016/j.resconrec.2017.08.027>
- Ridoutt, B. G., & Pfister, S. (2013). Towards an integrated family of footprint indicators. *Journal of Industrial Ecology*, 337–339. <https://doi.org/10.1111/jiec.12026>
- Ridoutt, B., Juliano, P., Sanguansri, P., & Sellahewa, J. (2010). The water footprint of food waste: Case study of fresh mango in Australia. *Journal of Cleaner Production*, 1714–1721. <https://doi.org/10.1016/j.jclepro.2010.07.011>
- Sacchi, R., Terlouw, T., Siala, K., Dirnaichner, A., & others. (2022). PRospective EnvironMental Impact asSEment (premise): A streamlined approach to producing databases for prospective life cycle assessment using integrated assessment models. *Renewable and Sustainable Energy Reviews*. <https://doi.org/10.1016/j.rser.2022.112311>
- Sacchi, R., Terlouw, T., Siala, K., Dirnaichner, A., & others. (2023). *Premise | Documentation*. <https://premise.readthedocs.io/>
- Stehfest, E., van Vuuren, D., Bouwman, L., Kram, T., & others. (2014). *Integrated assessment of global environmental change with IMAGE 3.0: Model description and policy applications*. <https://www.pbl.nl/en/publications/integrated-assessment-of-global-environmental-change-with-image-30-model-description-and-policy-applications>
- Steinmann, Z. J. N., Schipper, A. M., Hauck, M., Giljum, S., & others. (2017). Resource footprints are good proxies of environmental damage. *Environmental Science & Technology*. <https://doi.org/10.1021/acs.est.7b00698>
- Steubing, B., Mendoza Beltran, A., & Sacchi, R. (2023). Conditions for the broad application of prospective life cycle inventory databases. *The International Journal of Life Cycle Assessment*, 28(9), 1092–1103. <https://doi.org/10.1007/s11367-023-02192-8>
- Swiss Federal Office for the Environment (FOEN). (2021). *Swiss Eco-Factors 2021 according to the Ecological Scarcity Method: Methodological fundamentals and their application in Switzerland*. <https://www.bafu.admin.ch/bafu/en/home/topics/economy-consumption/economy-and-consumption-publications/publications-economy-and-consumption/eco-factors-switzerland.html>
- Van Der Giesen, C., Cucurachi, S., Guinée, J., Kramer, G. J., & Tukker, A. (2020). A critical view on the current application of LCA for new technologies and recommendations for improved practice. *Journal of Cleaner Production*, 259, 120904. <https://doi.org/10.1016/j.jclepro.2020.120904>
- Van Der Meide, M., Harpprecht, C., Northey, S., Yang, Y., & Steubing, B. (2022). Effects of the energy

- transition on environmental impacts of cobalt supply: A prospective life cycle assessment study on future supply of cobalt. *Journal of Industrial Ecology*, 26(5), 1631–1645. <https://doi.org/10.1111/jiec.13258>
- Van Vuuren, D. P., Edmonds, J., Kainuma, M., Riahi, K., Thomson, A., Hibbard, K., Hurtt, G. C., Kram, T., Krey, V., Lamarque, J.-F., Masui, T., Meinshausen, M., Nakicenovic, N., Smith, S. J., & Rose, S. K. (2011). The representative concentration pathways: An overview. *Climatic Change*, 109(1–2), 5–31. <https://doi.org/10.1007/s10584-011-0148-z>
- Van Vuuren, D. P., Riahi, K., Calvin, K., Dellink, R., Emmerling, J., Fujimori, S., KC, S., Kriegler, E., & O'Neill, B. (2017). The Shared Socio-economic Pathways: Trajectories for human development and global environmental change. *Global Environmental Change*. <https://doi.org/10.1016/j.gloenvcha.2016.10.009>
- Vanhamb, D., Leip, A., Galli, A., Kastner, T., & others. (2019). Environmental footprint family to address local to planetary sustainability and deliver on the SDGs. *Science of The Total Environment*. <https://doi.org/10.1016/j.scitotenv.2019.133642>
- Wackernagel, M. (1994). *Ecological footprint and appropriated carrying capacity: A tool for planning toward sustainability* [PhD Thesis, University of British Columbia]. <https://doi.org/10.14288/1.0088048>
- Wei, S., Sacchi, R., Tukker, A., Suh, S., & Steubing, B. (2024). Future environmental impacts of global hydrogen production. *Energy & Environmental Science*, 17(6), 2157–2172. <https://doi.org/10.1039/D3EE03875K>
- Wernet, G., Bauer, C., Steubing, B., Reinhard, J., & others. (2016). The ecoinvent database version 3 (part I): Overview and methodology. *The International Journal of Life Cycle Assessment*. <https://doi.org/10.1007/s11367-016-1087-8>
- Wiedmann, T. O., Schandl, H., Lenzen, M., Moran, D., & others. (2013). The material footprint of nations. *Proceedings of the National Academy of Sciences*, 6271–6276. <https://doi.org/10.1073/pnas.1220362110>

Tables

Table 1 Categorisation and count of the selected market activities

| Category full name | Abbreviated name | Count |
|------------------------------------------------------|------------------|-------|
| Agriculture, forestry, live animals & their products | AgriForeAnim | 212 |
| Chemical products | Chemical | 669 |
| Glass and other non-metallic products | GlasNonMetal | 110 |
| Machinery, metal/electronic, transport equipment | MachElecTrans | 122 |
| Basic metals & alloys, incl. semi-finished products | MetalAlloy | 86 |
| Ores, minerals & fuels | OreMinFuel | 132 |
| Plastics & rubber products | PlastRub | 78 |
| Processed biobased products | ProcBio | 184 |

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Table 2 Waste footprint statistics for each category (total waste)

| Category | Mean | std | Min | Max |
|---------------|----------|----------|-----------|----------|
| AgriForeAnim | 5.32E+03 | 2.13E+04 | -1.04E+02 | 2.66E+05 |
| Chemical | 1.70E+06 | 2.50E+07 | 0.00E+00 | 6.04E+08 |
| GlasNonMetal | 5.13E+03 | 1.02E+04 | -2.89E+02 | 8.03E+04 |
| MachElecTrans | 1.27E+06 | 2.75E+06 | 9.62E+02 | 1.76E+07 |
| MetalAlloy | 2.08E+07 | 9.07E+07 | 2.43E+01 | 5.90E+08 |
| OreMinFuel | 9.35E+05 | 2.83E+06 | 0.00E+00 | 1.09E+07 |
| PlastRub | 5.14E+04 | 1.01E+05 | 6.28E+01 | 4.36E+05 |
| ProcBio | 9.35E+04 | 3.64E+05 | 0.00E+00 | 2.79E+06 |

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Table 3 Top three activities for each product category with the “Waste - Total” footprint method

| Category | Name | Waste – Total (kg/kg) | Price (EUR2005/kg) |
|---------------|-----------------------------------------------------------|-----------------------|--------------------|
| AgriForeAnim | market for cocoons | 2.66E+05 | 8.26 |
| AgriForeAnim | market for swine for slaughtering, live weight | 1.34E+05 | 5.48 |
| AgriForeAnim | market for sheep fleece in the grease | 5.88E+04 | 2.82 |
| Chemical | market for lutetium oxide | 6.04E+08 | 619.06 |
| Chemical | market for thulium oxide | 1.61E+08 | 165.4 |
| Chemical | market for heavy water | 1.57E+08 | 620 |
| GlasNonMetal | market for panel glass, for cathode ray tube display | 8.03E+04 | 0.8 |
| GlasNonMetal | market for solar collector glass tube, with silver mirror | 4.79E+04 | 3.78 |
| GlasNonMetal | market for glass fibre | 2.91E+04 | 0.8 |
| MachElecTrans | market for integrated circuit, logic type | 1.76E+07 | 1260.01 |
| MachElecTrans | market for integrated circuit, memory type | 1.75E+07 | 121.85 |
| MachElecTrans | market for electronic component, active, unspecified | 1.53E+07 | 745.98 |
| MetalAlloy | market for gold-silver, ingot | 5.90E+08 | 2337.81 |
| MetalAlloy | market for silver, unrefined | 5.37E+08 | 314.63 |
| MetalAlloy | market for platinum | 2.42E+08 | 20600 |
| OreMinFuel | market for enriched uranium, 4.2% | 1.09E+07 | 586 |
| OreMinFuel | market for uranium, enriched 4.2%, in fuel element... | 1.09E+07 | 586 |
| OreMinFuel | market for uranium, enriched 4%, in fuel element... | 1.05E+07 | 586 |
| PlastRub | market for polyvinylchloride, emulsion polymerised | 4.36E+05 | 1.29 |
| PlastRub | market for polyethylene, linear low density, granulate | 4.17E+05 | 1.29 |
| PlastRub | market for polyvinylchloride, bulk polymerised | 4.14E+05 | 1.29 |

| | | | |
|---------|-------------------------------------|----------|-------|
| ProcBio | market for reeled raw silk hank | 2.79E+06 | 18.88 |
| ProcBio | market for fish canning, large fish | 1.02E+06 | 0.65 |
| ProcBio | market for yarn, silk | 7.75E+05 | 31.01 |

Figure Captions

Figure 1. Distribution of total supply-chain waste per product (kg waste per kg product, \log_{10} scale) across major industrial categories for a total of 1593 activities in ecoinvent 3.9.1. Boxes show interquartile ranges with median lines; whiskers indicate $1.5 \times \text{IQR}$, and dots denote the individual activities.

Figure 2. Waste circularity and hazardousness ratios across industrial categories for a total of 1593 activities in ecoinvent 3.9.1. The subfigures are: (a) Waste circularity—the share of total waste routed to recovery by recycling, composting, or anaerobic digestion, and (b) Waste hazardousness—the fraction of total waste classified as hazardous.

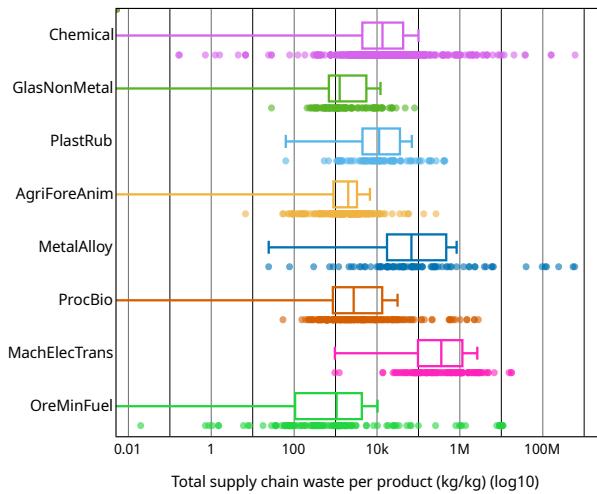
Boxes show interquartile ranges with medians; whiskers denote $1.5 \times \text{IQR}$; points are individual activities.

Figure 3. Scenario-based temporal trends in a selection of waste footprints, depicting the distribution of scores for 1593 activities in current and prospective LCA databases from 2020 to 2050 under the SSP1-PkBudg500 and SSP5-PkBudg500 scenario models. The subfigures are: (a) Waste – Carbon dioxide (CCS) (kg/kg), (b) Waste – Total (kg/kg), (c) Waste – Hazardousness (%), (d) Waste – Circularity ratio (%), (e) Waste – Landfilled (kg/kg), (f) Waste – Recycled (kg/kg), (g) Waste – Composted (kg/kg), (h) Waste – Incinerated (kg/kg) and (i) Waste – Openly burned (kg/kg).

Figures

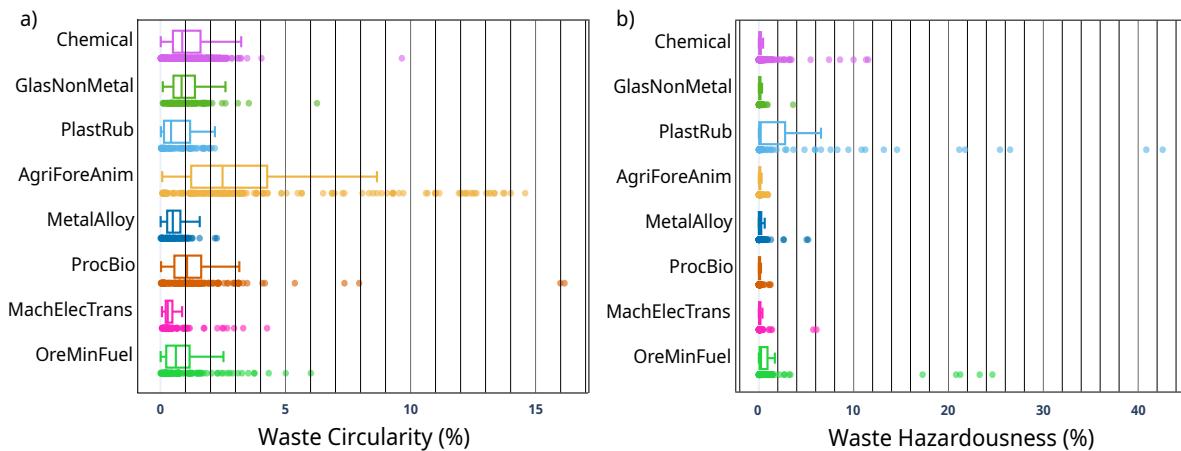
758

1.



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2.



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