Trees in many US cities may indirectly increase atmospheric carbon

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3 Abstract

- 4 Urban trees are a critical part of the "green infrastructure" intended to make our growing
- 5 cities more sustainable in an era of climate change. The potential for urban trees to modify
- 6 microclimates and thereby reduce building energy use and the associated carbon emissions is
- 7 a commonly cited ecosystem service used to justify million tree planting campaigns across the
- 8 country. However, what we know of this ecosystem service comes primarily from unvalidated
- 9 simulation studies.
- 10 Using the first dataset of actual heating and cooling energy use combined with tree cover
- 11 data, we show that that contrary to the predictions of the most commonly used simulations,
- 12 trees in a cool climate city increase carbon emissions from residential building energy use.
- 13 This is driven primarily by near east (< 20m from building) tree cover. Further analysis
- 14 of urban areas in the US shows that this is likely the case in cool climates throughout the
- 15 country, encompassing approximately 39% of the US population and 62% of its area (56%,

- 16 excluding Alaska). Our results add geographic nuance to quantification of the effect of urban
- 17 trees on the carbon budget and could have major implications for tree planting programs in
- 18 cool climates.

19 Introduction

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Two global trends of the 21st century, climate change and increasing urbanization, have 20 deepened our need to make cities more sustainable, and urban trees are championed as 21 a means to that end. Several large cities in the U.S. have recently committed to large 22 tree planting programs (see Million Trees New York City and Million Trees Los Angeles). 23 Spending hundreds of millions of dollars, these cities hope that the environmental benefits, 24 25 particularly the reduction in building energy use and the associated carbon (C) emissions from power plants, will outweigh the cost (Young, 2011). 26 A single urban tree has a much stronger impact on the carbon cycle than a non-urban 27 28 counterpart because an urban tree induces or reduces more C emitting human behaviors than a rural one does. Both trees sequester carbon from the atmosphere, but the urban 29 tree requires more management (planting, watering, pruning, removal, chipping) and, by 30 modifying the microclimate, it can alter building energy use and the associated C emissions 31 (ACE) from power plants. 32 33 Trees primarily alter micro climates by 1) shading, 2) reducing wind speed, and 3) cooling via transpiration. With the exception of transpirative cooling, which is mostly active 34 in summer, these effects can both increase or decrease ACE. Shading to the west of build-35 ings greatly reduces summer cooling loads, but shading to the south of buildings, even by 36 deciduous trees, may increase winter heating loads (Heisler, 1986). Reduced wind speeds 37 have complex effects. They: 1) decrease convective heat loss, which is beneficial for winter 38 heating but detrimental for summer cooling, 2) decrease air infiltration which decreases both 39 heating and cooling energy use, and 3) decrease natural ventilation, increasing the need for 40 mechanical cooling (Huang et al., 1990). The strength of the effect of a tree on ACE attenu-41 42 ates with distance to a building. Trees far from a house have little affect on ACE via shading and wind reduction, but they likely affect ACE via evapotranspiration and the associated 43 44 reduction in temperature Ziter et al. (2019).

Whether the net effect of trees is to increase or decrease ACE depends on the balance of

beneficial and detrimental effects on heating and cooling energy use. This is largely mediated by the location of tree cover, the prevailing climate (e.g. number of heating- and coolingdegree days), building characteristics (orientation, insulation, size), occupant behavior and the carbon content of a kWh, which varies depending on the fuel mix in the electrical grid.

Previous research

Decades worth of research primarily by two research groups, the US Forest Service's (USFS) 51 Southwest Research Station Urban Ecosystems and Processes group and the Lawrence Berke-52 ley National Lab Heat Island Group (LBNL), have reported that, on average, trees reduce 53 C emissions. In 2002, Akbari published a paper summarizing their group's findings: "Shade 54 trees reduce building energy use and CO₂ emissions from power plants". In 1999, McPherson 55 and Simpson wrote a technical report that was the basis of the iTree software, which has 56 been used by thousands of communities around the U.S. to estimate ACE avoided. Their 57 methodology was recently applied to estimate the effects of trees on ACE for the entire 58 conterminous US (Nowak et al., 2017). Despite the number of publications on the topic, the 59 60 length of time we have been researching the matter, and the many large cities with massive tree planting initiatives, our uncertainty about the effects of trees on building energy use is 61 actually quite high (Pataki et al., 2006; McPherson and Simpson, 1999). The effect of trees 62 on nearby building energy use is difficult and expensive to measure directly and complex to 63 model. 64 Direct measures of the effect of trees on building energy use are rare, focused on cooling 65 energy use, and limited in their ability to be extrapolated. To our knowledge, there are the 66 only 5 studies that test the effect of trees on measured building energy use data (Akbari 67 et al., 1997; Donovan and Butry, 2009; DeWalle et al., 1983; Parker, 1983; McPherson et al., 68 1989). Only two of these studies were of actual houses (not mobile homes nor models) and 69 both are from Sacramento, CA and did not measure heating energy use (Akbari et al., 1997;

- 71 Donovan and Butry, 2009). Only one of the studies was from a cool, heating dominated
- 72 climate (typical of much of the US) and it studied a single mobile home in a forest (DeWalle
- 73 et al., 1983).
- Given the challenges inherent in collecting direct measurements, simulation studies are
- 75 useful attempts to extend our understanding of how trees affect building energy use and
- 76 ACE. But these simulations necessarily contain simplifications and generalizations which are
- 77 sometimes unrealistic or untestable due to lack of data.
- 78 The work from LBNL assumes: millions more trees are planted in an urban area (ex-
- 79 tremely ambitious); trees are planted to the west and south of buildings (ideal placement for
- 80 reducing cooling loads); and winter tree canopy transmissivity is 0.9 (0.7 is more realistic,
- 81 Heisler, 1986). In later work, microclimate wind effects are ignored (Akbari and Konopacki,
- 82 2005), and in earlier work, they use a three parameter equation fit to four data points to
- 83 estimate how wind speed is reduced by canopy cover (Heisler, 1990; Huang et al., 1990). Fi-
- 84 nally, the LBNL work uses potential evapotranspiration to predict cooling, and their model
- 85 uses parameters derived from crops. Given these assumptions, the authors note that their
- 86 work provides an upper boundary for the indirect effect of trees (Akbari and Konopacki,
- 87 2005; Huang et al., 1987).
- 88 USFS studies assume: lookup tables for the effect of tree shade on building energy use
- 89 are reliable (even though they may deviate from more detailed simulations by up to 10%,
- 90 Simpson, 2002); wind reduction only affects heating use in the winter, even though we know
- 91 cooling use is also affected, and they also use an overfit summertime leaf-on equation from
- 92 Heisler (1990). Evergreen trees are modeled as if they are windbreaks for rural farmhouses
- 93 in winter, even in suburban neighborhoods where other buildings and trees already block
- 94 significant winds; and estimated evapotranspirative cooling is optimistically high, higher
- 95 even than the self declared upper limit of Huang et al. (1987) (McPherson and Simpson,
- 96 1999).
- 97 The consequence of these assumptions is that simulations may overestimate the energy

reducing power of trees. What little validation we have has confirmed the general effects of trees on energy use that we expect in hot climates, but also highlight the imprecision 99 of simulations as well as occasional discrepancies from empirical observations. Simulations 100 of Akbari et al. (1997) were off by 2-fold, though trees were about twice as beneficial as 101 predicted for the two houses studied. Donovan and Butry (2009) found trees to the north 102 actually increasing electricity use, unlike the predictions of McPherson and Simpson (1999). 103 Despite providing estimates for the effects of trees on building energy use and ACE for 104 anywhere in the country (Akbari and Konopacki, 2005) and the entire country (Nowak et al., 105 2017), we still have no empirical validation of the effect of urban trees in a cool climate. More 106 107 than 3 out of every 4 people in the U.S. live in places with more heating degree days than cooling degree days, and Americans use much more energy for heating than for cooling (U.S. 108 Department of Energy, 2009). To properly assess simulations of the role of urban trees 109 110 in the C budget, comprehensive analyses are needed to test the relationship between tree location and energy usage (both heating and cooling). Our work in Madison, WI was the 111 first to begin address this need. In 2016, we downloaded average annual energy use data 112 for approximately 32 thousand single family residential homes and built a regression model 113 between the amount of tree cover near each house and the C produced from electricity and 114 natural gas use, controlling for other factors such as building characteristics. 115

116 Results

117 Effect of trees on building associated C emissions

Trees increased C emissions associated with residential building energy use (ACE) in Madison, WI. This effect was the result of a trade-off between their electricity (cooling) saving and gas (heating) penalty. We estimated that 100m² of tree cover within 20m of a house increased ACE from gas use by 0.77% (95% CI: 0.68%, 0.85%), and decreased ACE from electricity use by 0.21% (95% CI: 0.34%, 0.080%). Our model for net ACE estimated that



Figure 1: Simulated shadows of trees on a house at the latitude of Madison, WI. In the summer, trees to the west of buildings provide the most effective shade since solar angles are lower and cooling demand highest in the afternoon. In winter, even deciduous trees can significantly reduce solar gain.

123 100m^2 of tree cover increased ACE by 0.17% (95% CI: .09%, .27%).

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The magnitude and direction of the effect depended on tree location relative to the building. Figure 2 shows the percent change in the ACE from 100m² of tree cover. Trees reduced ACE from electricity for all near regions except the east. Trees increased ACE from gas for all regions, especially in the near south and east. For net ACE, tree cover in the near east was the most important, having the only estimate with a 95% CI that excluded 0.

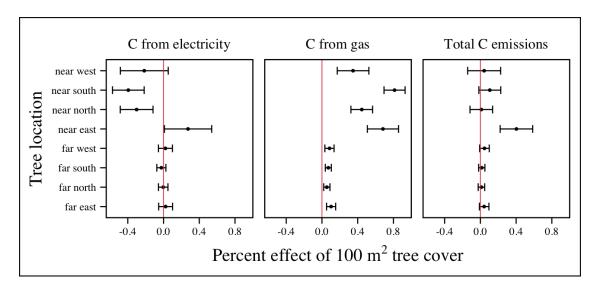


Figure 2: The percent effect of 100m^2 tree cover in different locations on C emissions from residential building energy use. n=25095, bars indicate standard errors.

129 Effect of existing tree cover on a typical house

The median house in our sample was responsible for 1084 and 954 kg C annual emissions 130 131 due to electricity use and gas use, respectively. Multiplying the median tree cover in each region (see table 1) by its coefficient we estimated the effects of typical tree cover on a typical 132 house in Madison: electricity C emissions were reduced by 33.8 kg C / yr (95% CI: 14.7, 133 134 52.7), but gas C emissions were increased by 102.3 kg C / year (95% CI: 92.9, 111.8). Our combined model estimated the net effect of existing tree cover is to increase C emissions by 135 about 62 kg C/year (95% CI: 38.7, 85.3) for a typical house. This is 2.5% of the median 136 house's annual ACE. 137

Table 1: Summary statistics for amount of tree cover (m²) in each region around houses in Madison, WI.

Region	min	mean	median	max
near west	0	193	179	742
near south	0	372	363	1443
near north	0	357	345	1197
near east	0	193	179	764
far west	0	974	960	2640
far south	0	1676	1653	4376
far north	0	1673	1661	4602
far east	0	967	955	2677

While tree cover in far regions had smaller per unit area effects than in near regions, there was more tree cover in farther regions, so when median tree cover was multiplied by the smaller coefficients some of the farther regions had larger typical effects than near ones (figure 3). Typical tree cover in the far east and far west regions had a greater estimated effect than cover in the near north and near west.

143 Comparing C emissions from energy use due to trees to C stored and 144 sequestered.

145 For comparison, consider a green ash tree with a crown area of 100m². This tree would store approximately 1360 kg C in above ground biomass and it could sequester around 34 kg C / year. That same tree in the near east region of a typical house in Madison was estimated to increase C emissions by 9.8 kg C/yr (95% CI: 6.7, 12.9). In the near west the estimated effect was 1.0 kg C/yr (95% CI: -2.1, 4.1).

Discussion

151 Interpreting Tree Effects

152 In the cool climate city of Madison, with 7283 HDD and 597 CDD and a electricity emission 153 factor of $0.206 \ \mathrm{kg} \ \mathrm{C} \ / \ \mathrm{kwh}$, the effect of trees on ACE was clear: trees increased ACE from

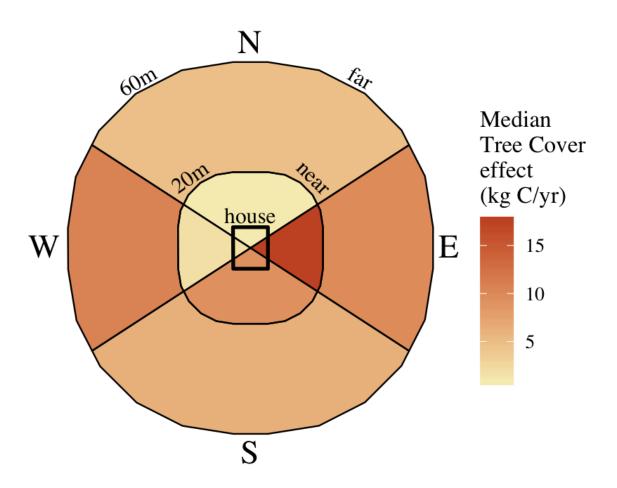


Figure 3: Effect of typical tree cover on a typical building's C emissions.

gas use more than they decreased ACE from electricity use, resulting in a net increase in ACE. This result suggests that shading was the most important process and that whatever gas savings trees may have provided in winter by reducing wind speeds was swamped by the penalty in reduced solar radiation.

By separating tree cover into different locations, it appeared that for the most regions, the beneficial effects of trees on electricity ACE mostly canceled out the detrimental effects of trees on gas ACE, with the exception of the near east. This suggests that trees to the east may have been responsible for most of the net increase in ACE. Eastern trees did not provide electricity savings since houses require less cooling in the morning hours, but still caused an increased gas use in winter. This agrees with Donovan and Butry (2009) who also found trees to the east had no effect on electricity use.

As expected, trees to the near south had a strong effect on electricity savings, but they also had a stronger gas penalty. Trees in the near west and near north had the weakest gas penalty, which may have been due to the savings they provided by reducing wind speed. Somewhat surprising was the weakness of the estimated electricity savings of trees in the near west, which all simulations have predicted has the strongest effect.

170 Comparing to past work

Our findings agreed with some though not all of the past simulation studies, and the modeling 171 of wind is the main cause of discrepancies. Theyer Jr and Maeda (1985) modeled the shading effects of south trees on building energy use and reported that trees increased emissions in 173 cities with more heating degree days than cooling degree days. McPherson et al. (1988) 174 investigated the shading and wind effects on building energy use in 4 cities, one of which 175 was Madison, WI. Converting their results into C, trees in Madison caused a small increase 176 in emissions, though their method for modeling wind was later criticized and abandoned 177 (Simpson and McPherson, 1998). Akbari and Konopacki (2005) developed a method to 178 179 predict the effect of a tree planting program and increasing roof albedo for any city in the U.S. Figure 4 illustrates an application of their method to every census tract in the conterminous US for pre-1980s houses using updated energy emission factors. About 40% of the US population live in areas where the Akbari and Konopacki (2005) model predicts that trees increase C emissions. While their methods were limited as mentioned above, and they modeled theoretical, not existing, tree cover, their work suggests that many large cities especially in New England, the Northwest, the Mountains and the Upper Midwest would need to carefully consider the C implications of large tree planting programs.

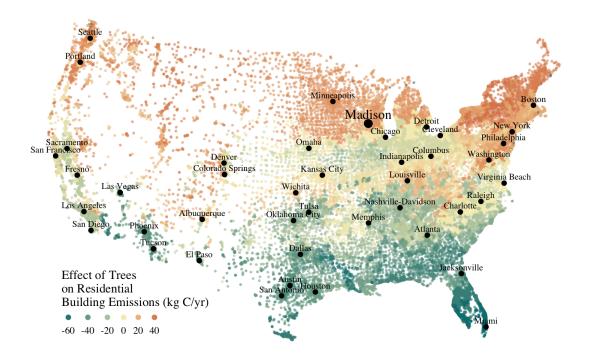


Figure 4: Each census tract in the conterminous US shaded by magnitude of building C emissions effect of trees planted to west and south of a pre-1980's home and increasing roof albedo. Differences in regional emission factors (C/kWh) cause deviations from climate trend. New England has especially high ACE for the climate because their electricity is cleaner (low C/kWh). About 40% of Americans live in places where trees increase ACE. Model based on Akbari and Konopacki (2005).

Our empirical findings disagree with those simulation studies that model the relationship between tree cover and wind speed following Heisler (1990) and McPherson and Simpson (1999). When the beneficial effects of wind are excluded for models of several cool climate

cities: Toronto (Akbari and Taha, 1992), Chicago (Jo and McPherson, 2001), Minneapolis, 190 Sacramento, and Washington (Huang et al., 1990), trees either have no effect or increase 191 energy use and ACE, which agrees with our general findings. The iTree model of McPherson 192 and Simpson (1999) predicts that the shading effects of a large deciduous tree in the Norther 193 Tier, North Central, Mountains, Pacific Northwest, and California Coast regions increases 194 ACE of a 1950-1980 vintage house by 0.136 to 9.52 kg, depending on the region. This is 195 comparable to our results. However, the wind effect in the iTree model of that same tree 196 on the same house decreases heating ACE by 1.23 to 66.14 kg depending on the region and 197 existing canopy: an order of magnitude greater savings for gas ACE from wind reduction than 198 the penalty from shading. However, our model coefficients derived from measured gas use 199 suggest shading is a more important process than wind shielding. McPherson and Simpson 200 (1999) note that the uncertainty in their methods was high, and, given our contradictory 201 202 findings, it is clear that more data and improved models are needed to better parameterize the complex and uncertain relationship between tree cover, wind, and building energy use. 203

204 Considering the larger C cycle

The effect on ACE of a tree with a 100 m² canopy area is an order of magnitude smaller 205 than that tree's C sequestration. However, it is important to make the distinction between 206 different pools of C. Discounting increased ACE as irrelevant because C sequestration more 207 than compensates, fails to recognize that ACE is an input of fossilized C while sequestra-208 tion is a temporary transfer of C from the atmosphere to biosphere. Unless forested land is 209 210 permanently expanded or wood products are forever prevented from decay, in the long run (hundreds of years) sequestration by trees can never offset fossil C emissions. Indeed this 211 same conclusion was made for fossilized C emissions due to tree management (Nowak et al., 212 2002). The avoided ACE from trees had been estimated to more than offset these manage-213 ment emissions in a life-cycle analysis of the Million Trees Los Angeles program (McPherson 214 215 and Kendall, 2014). However, our results suggest that for cool climate communities, trees actually increase ACE and, especially when combined with the C emissions from management,are atmospheric C sources.

Trees relative to other factors that affect ACE and the ACE effect of trees relative to other ecosystem services/disservices.

Considering all of the factors that determine building energy use and ACE, trees play a 220 very minor role, which we estimated to be about 2.5% of the ACE of a median house. As 221 buildings become better built and insulated the effect of trees on ACE will decrease. Far 222 greater ACE savings are possible with improved construction and savvy occupant behavior. 223 224 However, the effect of trees on energy use and ACE is one of the most often cited ecosystem services of trees (Roy et al., 2012), and evidence that ACE is increased by trees highlights 225 the large uncertainty in software used by thousands of communities to justify urban forest 226 227 costs.

Still, effects on ACE are just one of the ecosystem effects that trees have in cities. Trees 228 229 may also improve air quality, reduce stormwater runoff, reduce noise, and provide wildlife habitat. The aesthetic value of trees is often far greater than the value of the ecosystem 230 services or disservices provided (McPherson et al., 2005). Even after publishing that trees 231 reduced ACE on average, Akbari (2002) noted that this benefit alone may not justify the 232 cost of tree planting. Our opposing results have a similar caveat: even after finding the 233 detrimental impacts of trees on ACE in cool climates, management decisions need to consider 234 these results as just one of the many benefits and costs of trees. Our results suggest that 235 trees planted on all but the near east side of a house are net neutral in terms of ACE, so 236 237 that the other benefits of tree planting, such as aesthetics, could be accomplished in cool 238 climates through careful selection of planting locations.

239 Future work

240 Using actual energy use data from over 25,000 houses, we provide a much needed complement to simulation models of tree effects on ACE in cool climates. However, there is need for 241 continuing work to address remaining shortcomings. The observational nature of our data is 242 243 strengthed by the size of the dataset, but ultimately causal inference depends on our physical 244 knowledge of how trees alter building energy use. More experimental studies are needed especially in cool climate cities to better understand that relationship. Not all coefficients in 245 our model agree with our existing physical understanding of how trees affect building energy 246 use. For example, it is surprising that trees to the near west have such a weak effect on 247 248 electricity use. Our data on tree cover was also limited by a lack of information about tree height, which means we could not address how adjusting the size of trees planted in an urban 249 250 area affects ACE. Incorporating lidar could provide more accurate estimates of tree shading 251 and wind reduction. Furthermore, the scale of the effects that our study could detect is much smaller than the city-wide effects many simulation studies address. Ultimately, this work is 252 a sample of one year from one city with the accompanying limitations. The warm December 253 during the sampling period may mean the effect of trees is even more detrimental than we 254 report, but more years are needed to say. The location of Madison near the boundary that 255 Akbari and Konopacki (2005) identified between trees being a sink and a source is useful, 256 but more cities are needed to empiracally determine this boundary. 257 Our work reveals a blind spot in urban forest ecosystem studies. In an extensive review of 258 259 the effect of the urban forest on CO₂ emissions, Weissert et al. (2014) did not consider that trees could increase ACE. In a paper critical of many ecosystem services provided by trees, 260 Pataki et al. (2011) nevertheless state that trees reduce energy use and ACE. While this 261 may be true in most of the US, and the potential ACE reduction is larger than the potential 262 ACE increase, it ignores geographical nuance (Akbari and Konopacki, 2005). In many ways 263 264 it is not surprising, given the climatic diversity across the country, that the effects of trees on ACE might also vary and that our prescriptions for how to plant trees to minimize ACE 265

could be different between Los Angeles and New York City. However our study is only the first study to use both gas and electric energy use observations, and the first study of its kind in a cool climate. Much more work with observed energy use is needed to identify the border between atmospheric C sink and source.

270 Conclusion

Using observed energy use data, we have shown that trees near residential houses in Madison, 271 WI increase energy use and associated C emissions and near east tree cover has the strongest 272 net effect. Extending past simulation studies, we show that this is likely the case for a large 273 area of the US. The magnitude and direction of the effect is dependent on tree location 274 relative to buildings, climate, building characteristics, occupant behavior, and the C content 275 of electricity. Disagreements between our results and past work is due to how wind effects 276 277 are modeled and much more work is needed to better understand this process. We add critical geographic nuance to research that could have major implications for tree planting 278 programs in cool climates. 279

280 Methods

281 Building Energy Use

In April 2016, we obtained the annual energy use summary table (April 2015 - April 2016) from Madison Gas and Electric's publicly available website for approximately 32 thousand single family residential houses in Madison, WI. This included average monthly gas and electricity use. This period exhibited a much warmer than average December (about 6° C) and had low snowfall. We removed from our sample outliers that used fewer than 120 therms (which is less than the 0.5% quantile) or fewer than 240 kWh (which is less than the 0.05% quantile) annually. We included only buildings that used natural gas for heating and had

289 central air conditioning. Our final sample size used to build models was 25095.

290 Carbon Emissions

291 We converted energy use to C emissions using emission factors published by the US EPA's Emissions & Generation Resource Integrated Database, eGRID (Emissions & Generation 292 293 Resource Integrated Database, 2016). 100% of the carbon in natural gas is oxidized to CO₂ when burned for heating. The carbon coefficient for natural gas is 1.446 kg C / therm 294 (United State Environmental Protection Agency, 2017). For electricity, Madison, WI is a 295 part of the Midwest Reliability Organization East (MROE) region of the North American 296 297 electric grid. The estimated carbon coefficient for power generated in this region is 0.2063698 298 kg C/kWh (Emissions & Generation Resource Integrated Database, 2016). We had originally used emission factor for MROE from 2012 (.1567988 kg C / kWh) and by switching to the 299 updated and higher 2016 emission factor (0.2063698 kg C/kWh), the overall detrimental 300 effects of trees on ACE was diminished from about 3.4% to 2.5%. 301

302 Building Characteristics

303 Energy use is strongly determined by building characteristics. For every address in the city, 304 the City of Madison releases the assessor's property information, which includes information on building age, size, materials, type of heating and cooling, as well as which schools serve 305 306 the address. We removed any houses that had bad or missing data. Many of the covariates, such as size and price, were strongly correlated. Given that our primary interest was how tree 307 cover affected building energy use, not how building characteristics affect building energy 308 use, we reduced the dimensionality of building characteristics using principal components 309 analysis. This reduced the number of building covariates from 20 (Lot area, length of water 310 frontage, year built, number of stories, number of bedrooms, number of bathrooms (full and 311 half), number of fireplaces, living area on each floor, finished attic area, finished basement 312 area, total basement area, crawl space area, year roof was replaced, number of stalls in each 313

314 garage, land value, improvement value) to 5 orthogonal vectors, accounting for 55% of the 315 variance.

For tree cover we used a 1m resolution landcover map derived from 2013 National Agriculture

316 Tree Canopy

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Inventory Program (NAIP) visible and near-infrared digital aerial imagery (Erker et. al, in 318 review). Using building footprints from the Dane county, for each house for which we had 319 energy use data, we divided the space around it into 8 regions defined by 2 buffers around 320 the house of distance 20 m and 60m and 4 rays from the building's centroid. Tree cover 321 322 closer than 20m was considered near, tree cover farther than 20m and closer than 60m was considered far. These buffers were subdivided into north, west, south, and east regions by 323 rays of angles 57, 123, 237, 303 degrees from north. These angles are within 1 degree of the 324 azimuth angle of sunrise and sunset at the two solstices. This defines the south region as 325 the region that is exposed to direct sunlight year-round, and the north region as the region 326 that is never exposed to direct sunlight (this relationship is approximate and complicated by 327 individual building geometry). Within each of the eight regions we summed the area covered 328 by trees, and then use the tree cover in each region as predictors in our models. 329 We tested buffers of different widths (every 3m from 3m to 60m), but found because 330 of the observational nature of our data that we needed to aggregate regions to remove 331 multicollinearity that caused unstable coefficient estimates. Using a distance of 18, 21, or 332 24 m instead of 20m to separate "near" from "far" cover only slightly changed coefficient 333 334 estimates.

335 Building Cover

Nearby buildings likely also affect the energy use of a building. To test this hypothesis we calculated the area of buildings in each of the eight regions around every building and included these as covariates in our modeling.

339 Modeling

We fit linear models where the response was log transformed annual ACE for gas use, for 340 electricity use, or for gas and electricity combined (net). Because a separate model was 341 built to explain net C emissions, coefficient estimates for the net model were not precisely 342 343 the sum of the coefficients from the electricity and gas models. ACE was log transformed to meet assumptions of normality and diagnostic plots were assessed to check other model 344 assumptions and potential sensitivity to influential observations. Variables in our model were: 345 5 principal components of building characteristics, building cover in each of the 8 regions, tree 346 cover in each of the 8 regions and a random effect for elementary school which might capture 347 neighborhood characteristics such as culture. We used AIC as a variable selection criterion 348 and in our final models only used the first 5 building characteristics principal components 349 350 and we dropped all the building cover covariates. Estimates for the coefficients of tree cover 351 were not sensitive to the inclusion or removal of these covariates, but model fit improved. Although some tree cover covariates increased AIC, we kept all tree cover covariates in the 352 353 model because we wanted estimates of their effects, however uncertain they might be. We fit models using the R package lme4 (Bates et al., 2015). 354

355 Interpreting coefficients

To improve interpretability of coefficients, we back transformed them to the original scale and expressed the multiplicative effects as a percentage (Gelman and Hill, 2007). We then multiplied this percent change by the median ACE (a better estimator of the central tendency because of the right skew in our data) to estimate the typical effect in absolute C terms. To get typical effects of tree cover, we multiplied median tree cover in each region by its coefficient estimate and back transformed to the original scale.

Estimating C storage and sequestration of a green ash with 100m² canopy

To estimate C storage and sequestration by a single green ash tree with a canopy cover of 100m², we used allometric equations to estimate that tree's diameter at breast height (DBH) and mass and then, assuming an annual DBH growth of 0.61 cm, predicted the change in mass to get C sequestration Nowak and Crane (2002); McPherson et al. (2016).

368 Extending Analyses from Published Literature

To compare our work to past simulation studies we converted results that were in Therms 369 370 or kWh to kg C. We did this for Thayer Jr and Maeda (1985), McPherson et al. (1988), and Huang et al. (1990) using updated emission factors corresponding to each study city's 371 eGrid subregion (Emissions & Generation Resource Integrated Database, 2016). To extend 372 Akbari and Konopacki (2005), we joined climate data (heating and cooling degree days) from 373 the nearest NOAA weather station to census tract centroids U.S. Census Tract Centroids 374 375 (2010); Arguez et al. (2012). Then for each census tract we predicted the effect of trees and increasing roof albedo on the energy use of a pre-1980's building with gas heating following 376 377 their table that bins houses according to heating degree-days and using emission factors corresponding to the eGrid subregion containing the census tract centroid. Separating out 378 the indirect effects of trees from the indirect effects of increasing roof albedo was not possible 379 380 because these were not modeled separately. However, the general trend would be similar, but with a decreased electricity savings and a decreased heating penalty. Akbari and Konopacki 381 (2005) found the effect of tree shade to be stronger than the indirect effects of increased roof 382 383 albedo and transpirative cooling. We also used the join of climate and census tract data to estimate approximately 77% of the U.S. population lives in places with more heating than 384 cooling degree-days. 385

386 Code

- 387 All of the code and data for these analyses are present on Github (https://github.com/
- 388 TedwardErker/energy). Code is provisional pending review.

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