

# Trees in cool climate cities may increase atmospheric carbon by altering building energy use

Tedward Erker<sup>\*</sup>, Philip A. Townsend

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## 3 **Abstract**

4 Urban trees are a critical part of the "green infrastructure" intended to make our growing  
5 cities more sustainable in an era of climate change. The potential for urban trees to modify  
6 microclimates and thereby reduce building energy use and the associated carbon emissions is  
7 a commonly cited ecosystem service used to justify million tree planting campaigns across the  
8 country. However, what we know of this ecosystem service comes primarily from unvalidated  
9 simulation studies.

10 Using the first dataset of actual heating and cooling energy use combined with tree cover  
11 data, we show that that contrary to the predictions of the most commonly used simulations,  
12 trees in a cool climate city increase carbon emissions from residential building energy use.  
13 This is driven primarily by near east ( $< 20\text{m}$  from building) tree cover. Further analysis  
14 of urban areas in the US shows that this is likely the case in cool climates throughout the  
15 country, encompassing approximately 39% of the US population and 62% of its area (56%,

16 excluding Alaska). Our results add geographic nuance to quantification of the effect of urban  
17 trees on the carbon budget and could have major implications for tree planting programs in  
18 cool climates.

## 19 Introduction

20 Two global trends of the 21st century, climate change and increasing urbanization, have  
21 deepened our need to make cities more sustainable, and urban trees are championed as  
22 a means to that end. Several large cities in the U.S. have recently committed to large  
23 tree planting programs (see Million Trees New York City and Million Trees Los Angeles).  
24 Spending hundreds of millions of dollars, these cities hope that the environmental benefits,  
25 particularly the reduction in building energy use and the associated carbon (**C**) emissions  
26 from power plants, will outweigh the cost (Young, 2011).

27 A single urban tree has a much stronger impact on the carbon cycle than a non-urban  
28 counterpart because an urban tree induces or reduces more C emitting human behaviors  
29 than a rural one does. Both trees sequester carbon from the atmosphere, but the urban  
30 tree requires more management (planting, watering, pruning, removal, chipping) and, by  
31 modifying the microclimate, it can alter building energy use and the associated C emissions  
32 (**ACE**) from power plants.

33 Trees primarily alter micro climates by 1) shading, 2) reducing wind speed, and 3) cool-  
34 ing via transpiration. With the exception of transpirative cooling, which is mostly active  
35 in summer, these effects can both increase or decrease ACE. Shading to the west of build-  
36 ings greatly reduces summer cooling loads, but shading to the south of buildings, even by  
37 deciduous trees, may increase winter heating loads (Heisler, 1986). Reduced wind speeds  
38 have complex effects. They: 1) decrease convective heat loss, which is beneficial for winter  
39 heating but detrimental for summer cooling, 2) decrease air infiltration which decreases both  
40 heating and cooling energy use, and 3) decrease natural ventilation, increasing the need for  
41 mechanical cooling (Huang et al., 1990). The strength of the effect of a tree on ACE attenu-  
42 ates with distance to a building. Trees far from a house have little affect on ACE via shading  
43 and wind reduction, but they likely affect ACE via evapotranspiration and the associated  
44 reduction in temperature Ziter et al. (2019).

45 Whether the net effect of trees is to increase or decrease ACE depends on the balance of

46 beneficial and detrimental effects on heating and cooling energy use. This is largely mediated  
47 by the location of tree cover, the prevailing climate (e.g. number of heating- and cooling-  
48 degree days), building characteristics (orientation, insulation, size), occupant behavior and  
49 the carbon content of a kWh, which varies depending on the fuel mix in the electrical grid.

## 50 Previous research

51 Decades worth of research primarily by two research groups, the US Forest Service's (USFS)  
52 Southwest Research Station Urban Ecosystems and Processes group and the Lawrence Berke-  
53 ley National Lab Heat Island Group (LBNL), have reported that, on average, trees reduce  
54 C emissions. In 2002, Akbari published a paper summarizing their group's findings: "Shade  
55 trees reduce building energy use and CO<sub>2</sub> emissions from power plants". In 1999, McPherson  
56 and Simpson wrote a technical report that was the basis of the iTree software, which has  
57 been used by thousands of communities around the U.S. to estimate ACE avoided. Their  
58 methodology was recently applied to estimate the effects of trees on ACE for the entire  
59 conterminous US (Nowak et al., 2017). Despite the number of publications on the topic, the  
60 length of time we have been researching the matter, and the many large cities with massive  
61 tree planting initiatives, our uncertainty about the effects of trees on building energy use is  
62 actually quite high (Pataki et al., 2006; McPherson and Simpson, 1999). The effect of trees  
63 on nearby building energy use is difficult and expensive to measure directly and complex to  
64 model.

65 Direct measures of the effect of trees on building energy use are rare, focused on cooling  
66 energy use, and limited in their ability to be extrapolated. To our knowledge, there are the  
67 only 5 studies that test the effect of trees on measured building energy use data (Akbari  
68 et al., 1997; Donovan and Butry, 2009; DeWalle et al., 1983; Parker, 1983; McPherson et al.,  
69 1989). Only two of these studies were of actual houses (not mobile homes nor models) and  
70 both are from Sacramento, CA and did not measure heating energy use (Akbari et al., 1997;

71 Donovan and Butry, 2009). Only one of the studies was from a cool, heating dominated  
72 climate (typical of much of the US) and it studied a single mobile home in a forest (DeWalle  
73 et al., 1983).

74 Given the challenges inherent in collecting direct measurements, simulation studies are  
75 useful attempts to extend our understanding of how trees affect building energy use and  
76 ACE. But these simulations necessarily contain simplifications and generalizations which are  
77 sometimes unrealistic or untestable due to lack of data.

78 The work from LBNL assumes: millions more trees are planted in an urban area (ex-  
79 tremely ambitious); trees are planted to the west and south of buildings (ideal placement for  
80 reducing cooling loads); and winter tree canopy transmissivity is 0.9 (0.7 is more realistic,  
81 Heisler, 1986). In later work, microclimate wind effects are ignored (Akbari and Konopacki,  
82 2005), and in earlier work, they use a three parameter equation fit to four data points to  
83 estimate how wind speed is reduced by canopy cover (Heisler, 1990; Huang et al., 1990). Fi-  
84 nally, the LBNL work uses potential evapotranspiration to predict cooling, and their model  
85 uses parameters derived from crops. Given these assumptions, the authors note that their  
86 work provides an upper boundary for the indirect effect of trees (Akbari and Konopacki,  
87 2005; Huang et al., 1987).

88 USFS studies assume: lookup tables for the effect of tree shade on building energy use  
89 are reliable (even though they may deviate from more detailed simulations by up to 10%,  
90 Simpson, 2002); wind reduction only affects heating use in the winter, even though we know  
91 cooling use is also affected, and they also use an overfit summertime leaf-on equation from  
92 Heisler (1990). Evergreen trees are modeled as if they are windbreaks for rural farmhouses  
93 in winter, even in suburban neighborhoods where other buildings and trees already block  
94 significant winds; and estimated evapotranspirative cooling is optimistically high, higher  
95 even than the self declared upper limit of Huang et al. (1987) (McPherson and Simpson,  
96 1999).

97 The consequence of these assumptions is that simulations may overestimate the energy

98 reducing power of trees. What little validation we have has confirmed the general effects  
99 of trees on energy use that we expect in hot climates, but also highlight the imprecision  
100 of simulations as well as occasional discrepancies from empirical observations. Simulations  
101 of Akbari et al. (1997) were off by 2-fold, though trees were about twice as beneficial as  
102 predicted for the two houses studied. Donovan and Butry (2009) found trees to the north  
103 actually increasing electricity use, unlike the predictions of McPherson and Simpson (1999).

104 Despite providing estimates for the effects of trees on building energy use and ACE for  
105 anywhere in the country (Akbari and Konopacki, 2005) and the entire country (Nowak et al.,  
106 2017), we still have no empirical validation of the effect of urban trees in a cool climate. More  
107 than 3 out of every 4 people in the U.S. live in places with more heating degree days than  
108 cooling degree days, and Americans use much more energy for heating than for cooling (U.S.  
109 Department of Energy, 2009). To properly assess simulations of the role of urban trees  
110 in the C budget, comprehensive analyses are needed to test the relationship between tree  
111 location and energy usage (both heating and cooling). Our work in Madison, WI was the  
112 first to begin address this need. In 2016, we downloaded average annual energy use data  
113 for approximately 32 thousand single family residential homes and built a regression model  
114 between the amount of tree cover near each house and the C produced from electricity and  
115 natural gas use, controlling for other factors such as building characteristics.

116 In an extensive review of the effect of the urban forest on CO<sub>2</sub> emissions, Weissert et al.  
117 (2014) did not consider that trees could increase ACE. In a paper critical of many ecosystem  
118 services provided by trees, Pataki et al. (2011) nevertheless state that trees reduce energy  
119 use and ACE



Figure 1: Simulated shadows of trees on a house at the latitude of Madison, WI. In the summer, trees to the west of buildings provide the most effective shade since solar angles are lower and cooling demand highest in the afternoon. In winter, even deciduous trees can significantly reduce solar gain.

# Results

## Effect of trees on building associated C emissions

Trees increased C emissions associated with residential building energy use (**ACE**) in Madison, WI. This effect was the result of a trade-off between their electricity (cooling) saving and gas (heating) penalty. We estimated that 100m<sup>2</sup> of tree cover within 20m of a house increased ACE from gas use by 0.77% (95% CI: 0.68%, 0.85%), and decreased ACE from electricity use by 0.21% (95% CI: 0.34%, 0.080%). Our model for net ACE estimated that 100m<sup>2</sup> of tree cover increased ACE by 0.17% (95% CI: .09%, .27%).

The magnitude and direction of the effect depended on tree location relative to the building. Figure 2 shows the percent change in the ACE from 100m<sup>2</sup> of tree cover. Trees reduced ACE from electricity for all near regions except the east. Trees increased ACE from gas for all regions, especially in the near south and east. For net ACE, tree cover in the near east was the most important, having the only estimate with a 95% CI that excluded 0.

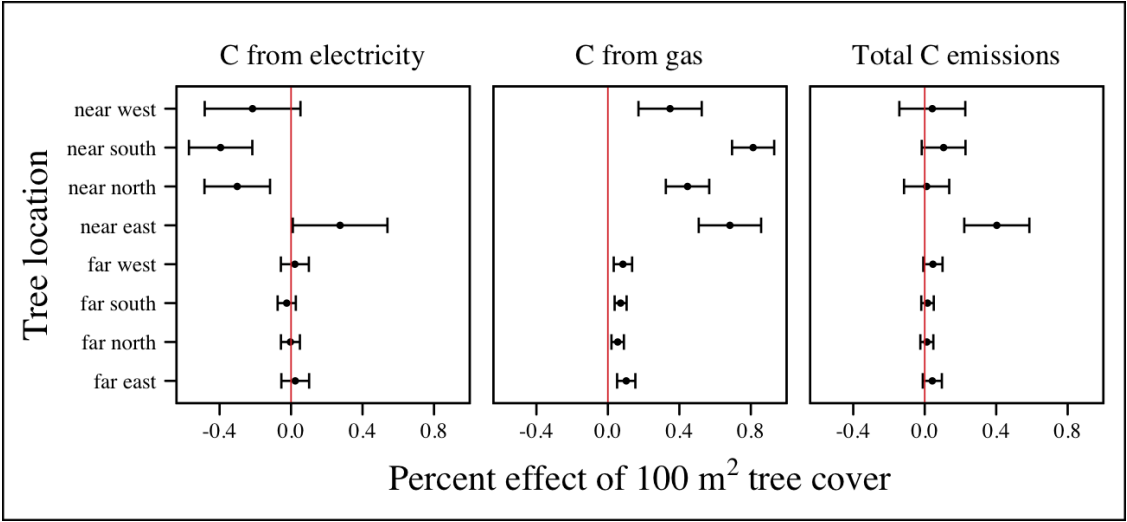


Figure 2: The percent effect of 100m<sup>2</sup> tree cover in different locations on C emissions from residential building energy use. n = 25095, bars indicate standard errors.



### 133 Effect of existing tree cover on a typical house

134 The median house in our sample was responsible for 1084 and 954 kg C annual emissions  
 135 due to electricity use and gas use, respectively. Multiplying the median tree cover in each  
 136 region (see table 1) by its coefficient we estimated the effects of typical tree cover on a typical  
 137 house in Madison: electricity C emissions were reduced by 33.8 kg C / yr (95% CI: 14.7,  
 138 52.7), but gas C emissions were increased by 102.3 kg C / year (95% CI: 92.9, 111.8). Our  
 139 combined model estimated the net effect of existing tree cover is to increase C emissions by  
 140 about 62 kg C/year (95% CI: 38.7, 85.3) for a typical house. This is 2.5% of the median  
 141 house's annual ACE.

Table 1: Summary statistics for amount of tree cover ( $\text{m}^2$ ) in each region around houses in Madison, WI.

Region	min	mean	median	max
near west	0	193	179	742
near south	0	372	363	1443
near north	0	357	345	1197
near east	0	193	179	764
far west	0	974	960	2640
far south	0	1676	1653	4376
far north	0	1673	1661	4602
far east	0	967	955	2677

142 While tree cover in far regions had smaller per unit area effects than in near regions,  
 143 there was more tree cover in farther regions, so when median tree cover was multiplied by  
 144 the smaller coefficients some of the farther regions had larger typical effects than near ones  
 145 (figure 3). Typical tree cover in the far east and far west regions had a greater estimated  
 146 effect than cover in the near north and near west.

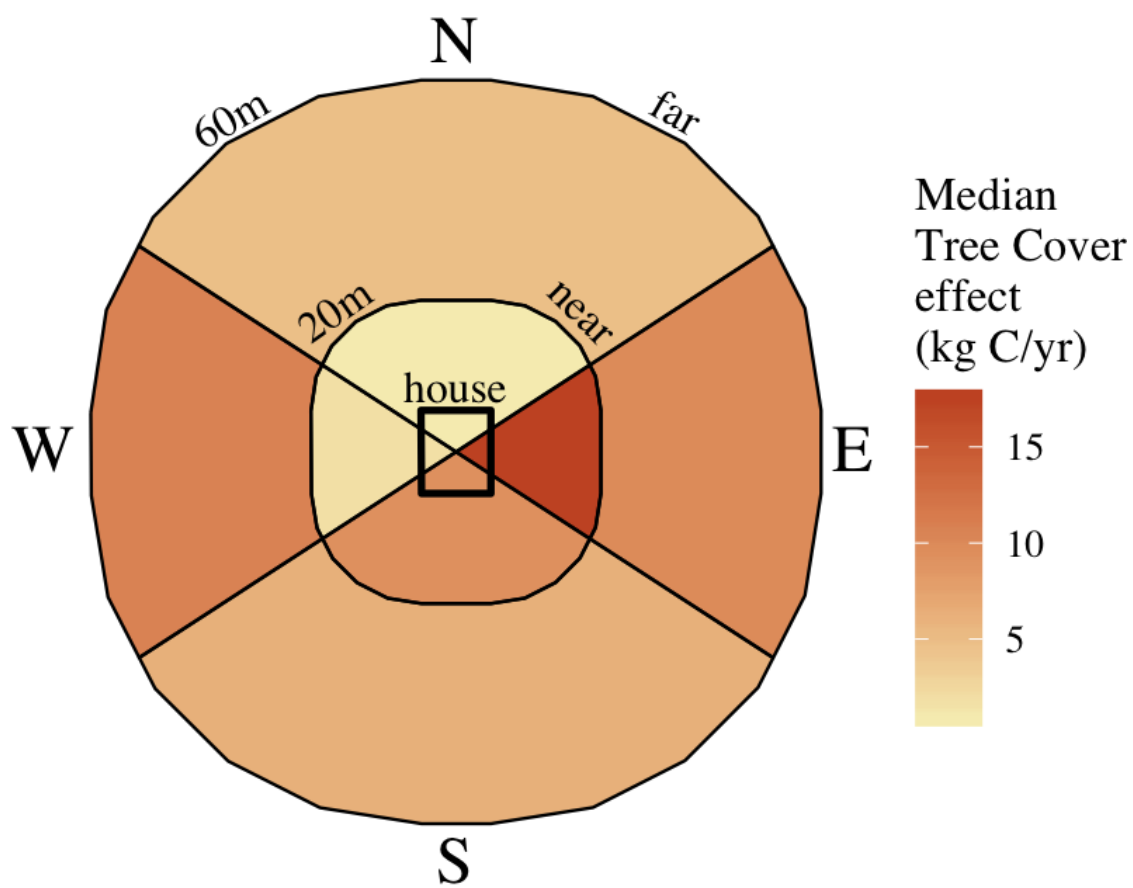


Figure 3: Effect of typical tree cover on a typical building's C emissions.

147 **Comparing C emissions from energy use due to trees to C stored and**  
148 **sequestered.**

149 For comparison, consider a green ash tree with a crown area of 100m<sup>2</sup>. This tree would store  
150 approximately 1360 kg C in above ground biomass and it could sequester around 34 kg C /  
151 year. That same tree in the near east region of a typical house in Madison was estimated  
152 to increase C emissions by 9.8 kg C/yr (95% CI: 6.7, 12.9). In the near west the estimated  
153 effect was 1.0 kg C/yr (95% CI: -2.1, 4.1). The the transfer of carbon from atmosphere  
154 to the biosphere is an order of magnitude larger than the transfer from the lithosphere to  
155 atmosphere.

## 156 **Discussion**

### 157 **Interpreting Tree Effects**

158 In the cool climate city of Madison, with 7283 HDD and 597 CDD and a electricity emission  
159 factor of 0.206 kg C / kwh, the effect of trees on ACE was clear: trees increased ACE from  
160 gas use more than they decreased ACE from electricity use, resulting in a net increase in  
161 ACE.

162 In simulation studies, if shade were the only affect on ACE (winter wind speed reduction  
163 was not included) trees in cool climate cities would cause an increase in ACE. Since we found  
164 an increase in ACE with increased tree cover

165 This result suggests that shading was the most important process and that whatever gas  
166 savings trees may have provided in winter by reducing wind speeds was swamped by the  
167 penalty in reduced solar radiation.

168 By separating tree cover into different locations, it appeared that for the most regions,  
169 the beneficial effects of trees on electricity ACE *mostly* canceled out the detrimental effects  
170 of trees on gas ACE, with the exception of the near east. This suggests that trees to the

171 east may have been responsible for most of the net increase in ACE. Eastern trees did not  
172 provide electricity savings since houses require less cooling in the morning hours, but still  
173 caused an increased gas use in winter. This agrees with Donovan and Butry (2009) who also  
174 found trees to the east had no effect on electricity use.

175 As expected, trees to the near south had a strong effect on electricity savings, but they  
176 also had a stronger gas penalty. Trees in the near west and near north had the weakest gas  
177 penalty, which may have been due to the savings they provided by reducing wind speed.  
178 Somewhat surprising was the weakness of the estimated electricity savings of trees in the  
179 near west, which all simulations have predicted has the strongest effect.

180 Trees to the north and gas use..... doesn't make much sense, and could be wrong.  
181 But consider that over buildings is included, so there is some shading. it's north of the  
182 building's centroid. transpiration possible effect? there are leaves on trees for part of the  
183 heating season and there are some evergreen trees that can transpire for much of year.

184 More here a la reviewer 3

## 185 **Comparing to past work**

186 Our findings agreed with some though not all of the past simulation studies, and the modeling  
187 of wind is the main cause of discrepancies. Thayer Jr and Maeda (1985) modeled the shading  
188 effects of south trees on building energy use and reported that trees increased emissions in  
189 cities with more heating degree days than cooling degree days. McPherson et al. (1988)  
190 investigated the shading and wind effects on building energy use in 4 cities, one of which  
191 was Madison, WI. Converting their results into C, trees in Madison caused a small increase  
192 in emissions, though their method for modeling wind was later criticized and abandoned  
193 (Simpson and McPherson, 1998). Akbari and Konopacki (2005) developed a method to  
194 predict the effect of a tree planting program and increasing roof albedo for any city in  
195 the U.S. Figure 4 illustrates an application of their method to every census tract in the  
196 conterminous US for pre-1980s houses using updated energy emission factors. Clearly climate

197 largely drives the relationship between ACE and trees at large scales, but there is significant  
 198 regional variation due to differences in electricity C emission factors. For example, despite  
 199 its cool climate, trees in Chicago reduce ACE because the electricity reduction benefit is  
 200 larger with more C per kwh.  
 201 while  
 202 across the rust  
 203 About 40% of the US population live in areas where the Akbari and Konopacki (2005)  
 204 model predicts that trees increase C emissions. While their methods were limited as men-  
 205 tioned above, and they modeled theoretical, not existing, tree cover, their work suggests that  
 206 many large cities especially in New England, the Northwest, the Mountains and the Upper  
 207 Midwest would need to carefully consider the C implications of large tree planting programs.

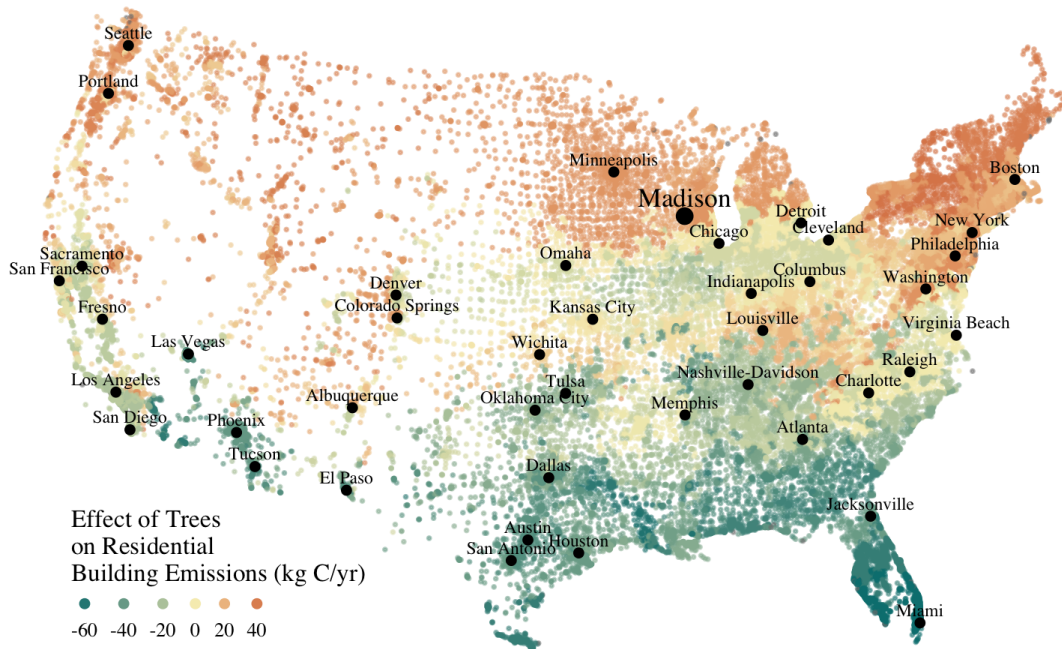


Figure 4: Each census tract in the conterminous US shaded by magnitude of building C emissions effect of trees planted to west and south of a pre-1980's home and increasing roof albedo. Differences in regional emission factors (C/kWh) cause deviations from climate trend. New England has especially high ACE for the climate because their electricity is cleaner (low C/kWh). About 40% of Americans live in places where trees increase ACE. Model based on Akbari and Konopacki (2005).

Our empirical findings disagree with those simulation studies that model the relationship between tree cover and wind speed following Heisler (1990) and McPherson and Simpson (1999). When the beneficial effects of wind are excluded for models of several cool climate cities: Toronto (Akbari and Taha, 1992), Chicago (Jo and McPherson, 2001), Minneapolis, Sacramento, and Washington (Huang et al., 1990), trees either have no effect or increase energy use and ACE, which agrees with our general findings. The iTree model which uses the methods of McPherson and Simpson (1999) predicts that the shading effects of a large deciduous tree in the Northern Tier, North Central, Mountains, Pacific Northwest, and California Coast regions increases ACE of a 1950-1980 vintage house by 0.136 to 9.52 kg, depending on the region. This is comparable to our results. However, the wind effect in the iTree model of that same tree on the same house decreases heating ACE by 1.23 to 66.14 kg depending on the region and existing canopy: an order of magnitude greater savings for gas ACE from wind reduction than the penalty from shading. Given that our model coefficients show that trees increases ACE, it suggests that shading is a more important process than wind speed reduction. In other words, our results agree with the shading but not wind reduction effects proposed by others, and therefore may suggest that shading is being more accurately modeled than wind in existing simulations. McPherson and Simpson (1999) note that the uncertainty in their methods was high, and, given our contradictory findings, it is clear that more data and improved models are needed to better parameterize the complex and uncertain relationship between tree cover, wind, and building energy use.

## Considering the larger C cycle

The effect on ACE of a tree with a 100 m<sup>2</sup> canopy area is an order of magnitude smaller than that tree's C sequestration. However, it is important to make the distinction between different pools of C. Discounting increased ACE as irrelevant because C sequestration more than compensates, fails to recognize that ACE is an input of fossilized C while sequestration is a temporary transfer of C from the atmosphere to biosphere. In the short term, sequestration

may assist in climate change mitigation, but unless forested land is permanently expanded or wood products are forever prevented from decay, in the long run (hundreds of years) sequestration by trees can never offset fossil C emissions. Indeed this same conclusion was made for fossilized C emissions due to tree management (Nowak et al., 2002). The avoided ACE from trees had been estimated to more than offset these management emissions in a life-cycle analysis of the Million Trees Los Angeles program (McPherson and Kendall, 2014). However, our results suggest that for cool climate communities, shade trees actually increase ACE and, especially when combined with the C emissions from management, are atmospheric C sources in the long term.

**Trees relative to other factors that affect ACE and the ACE effect of trees relative to other ecosystem services/disservices.**

Considering all of the factors that determine building energy use and ACE, trees play a very minor role, which we estimated to be about 2.5% of the ACE of a median house. As buildings become better built and insulated the effect of trees on ACE will decrease. Far greater ACE savings are possible with improved construction and savvy occupant behavior. However, the effect of trees on energy use and ACE is one of the most often cited ecosystem services of trees (Roy et al., 2012), and evidence that ACE is increased by trees highlights the large uncertainty in software used by thousands of communities to justify urban forest costs.

Still, effects on ACE are just one of the ecosystem effects that trees have in cities. Trees may also improve air quality, reduce stormwater runoff, reduce noise, and provide wildlife habitat. The aesthetic value of trees is often far greater than the value of the ecosystem services or disservices provided (McPherson et al., 2005). Even after publishing that trees reduced ACE on average, Akbari (2002) noted that this benefit alone may not justify the cost of tree planting. Our opposing results have a similar caveat: even after finding the detrimental impacts of trees on ACE in cool climates, management decisions need to consider

these results as just one of the many benefits and costs of trees. Our results suggest that trees planted on all but the near east side of a house are net neutral in terms of ACE, so that the other benefits of tree planting, such as aesthetics, could be accomplished in cool climates through careful selection of planting locations.

## **Future work**

Using actual energy use data from over 25,000 houses, we provide a much needed complement to simulation models of tree effects on ACE in cool climates. However, there is need for continuing work to address remaining shortcomings. The observational nature of our data is strengthened by the size of the dataset, but ultimately causal inference depends on our physical knowledge of how trees alter building energy use. More experimental studies are needed especially in cool climate cities to better understand that relationship. Not all coefficients in our model agree with our existing physical understanding of how trees affect building energy use. For example, it is surprising that trees to the near west have such a weak effect on electricity use. Our data on tree cover was also limited by a lack of information about tree height, which means we could not address how adjusting the size of trees planted in an urban area affects ACE. Incorporating lidar could provide more accurate estimates of tree shading and wind reduction. Furthermore, the scale of the effects that our study could detect is much smaller than the city-wide effects many simulation studies address. Ultimately, this work is a sample of one year from one city with the accompanying limitations. The warm December during the sampling period may mean the effect of trees is even more detrimental than we report, but more years are needed to say. The location of Madison near the boundary that Akbari and Konopacki (2005) identified between trees being a sink and a source is useful, but more cities are needed to empirically determine this boundary.

Our work reveals a blind spot in urban forest ecosystem studies. In an extensive review of the effect of the urban forest on CO<sub>2</sub> emissions, Weissert et al. (2014) did not consider that trees could increase ACE. In a paper critical of many ecosystem services provided by trees,



286 Pataki et al. (2011) nevertheless state that trees reduce energy use and ACE. While this  
287 may be true in most of the US, and the potential ACE reduction is larger than the potential  
288 ACE increase, it ignores geographical nuance (Akbari and Konopacki, 2005). In many ways  
289 it is not surprising, given the climatic diversity across the country, that the effects of trees  
290 on ACE might also vary and that our prescriptions for how to plant trees to minimize ACE  
291 could be different between Los Angeles and New York City. However our study is only the  
292 first study to use both gas and electric energy use observations, and the first study of its  
293 kind in a cool climate. Much more work with observed energy use is needed to identify the  
294 border between atmospheric C sink and source. Planners and designers should

## 295 Conclusion

296 Using observed energy use data, we have shown that trees near residential houses in Madison,  
297 WI increase energy use and associated C emissions and near east tree cover has the strongest  
298 net effect. Extending past simulation studies, we show that this is likely the case for a large  
299 area of the US. The magnitude and direction of the effect is dependent on tree location  
300 relative to buildings, climate, building characteristics, occupant behavior, and the C content  
301 of electricity. Disagreements between our results and past work may be due to how wind  
302 effects are modeled and much more work is needed to better understand this process. We add  
303 critical geographic nuance to research that could have major implications for tree planting  
304 programs in cool climates.

305 Add that I don't necessarily invalidate other studies (a la reviewer 3)

## Methods

### Building Energy Use

In April 2016, we obtained the annual energy use summary table (April 2015 - April 2016) from Madison Gas and Electric's publicly available website for approximately 32 thousand single family residential houses in Madison, WI. This included average monthly gas and electricity use. This period exhibited a much warmer than average December (about 6° C) and had low snowfall. We removed from our sample outliers that used fewer than 120 therms (which is less than the 0.5% quantile) or fewer than 240 kWh (which is less than the 0.05% quantile) annually. We included only buildings that used natural gas for heating and had central air conditioning. Our final sample size used to build models was 25095.

### Carbon Emissions

We converted energy use to C emissions using emission factors published by the US EPA's Emissions & Generation Resource Integrated Database, eGRID (Emissions & Generation Resource Integrated Database, 2016). 100% of the carbon in natural gas is oxidized to CO<sub>2</sub> when burned for heating. The carbon coefficient for natural gas is 1.446 kg C / therm (United State Environmental Protection Agency, 2017). For electricity, Madison, WI is a part of the Midwest Reliability Organization East (MROE) region of the North American electric grid. The estimated carbon coefficient for power generated in this region is 0.2063698 kg C/kWh (Emissions & Generation Resource Integrated Database, 2016). We had originally used emission factor for MROE from 2012 (.1567988 kg C / kWh) and by switching to the updated and higher 2016 emission factor (0.2063698 kg C/kWh), the overall detrimental effects of trees on ACE was diminished from about 3.4% to 2.5%.

## 328 **Building Characteristics**

329 Energy use is strongly determined by building characteristics. For every address in the city,  
330 the City of Madison releases the assessor’s property information, which includes information  
331 on building age, size, materials, type of heating and cooling, as well as which schools serve  
332 the address. We removed any houses that had bad or missing data. Many of the covariates,  
333 such as size and price, were strongly correlated. Given that our primary interest was how tree  
334 cover affected building energy use, not how building characteristics affect building energy  
335 use, we reduced the dimensionality of building characteristics using principal components  
336 analysis. This reduced the number of building covariates from 20 (Lot area, length of water  
337 frontage, year built, number of stories, number of bedrooms, number of bathrooms (full and  
338 half), number of fireplaces, living area on each floor, finished attic area, finished basement  
339 area, total basement area, crawl space area, year roof was replaced, number of stalls in each  
340 garage, land value, improvement value) to 5 orthogonal vectors, accounting for 55% of the  
341 variance.

## 342 **Tree Canopy**

343 For tree cover we used a 1m resolution landcover map derived from 2013 National Agriculture  
344 Inventory Program (NAIP) visible and near-infrared digital aerial imagery with an accuracy  
345 of 85% (Erker et al., 2018). Using building footprints from the Dane county, for each house  
346 for which we had energy use data, we divided the space around it into 8 regions defined by 2  
347 buffers around the house of distance 20 m and 60m and 4 rays from the building’s centroid.  
348 Tree cover closer than 20m was considered near, tree cover farther than 20m and closer than  
349 60m was considered far. These buffers were subdivided into north, west, south, and east  
350 regions by rays of angles 57, 123, 237, 303 degrees from north. These angles are within  
351 1 degree of the azimuth angle of sunrise and sunset at the two solstices. This defines the  
352 south region as the region that is exposed to direct sunlight year-round, and the north region  
353 as the region that is never exposed to direct sunlight (this relationship is approximate and

354 complicated by individual building geometry). Within each of the eight regions we summed  
355 the area covered by trees, and then use the tree cover in each region as predictors in our  
356 models.

357 We tested buffers of different widths (every 3m from 3m to 60m), but found because  
358 of the observational nature of our data that we needed to aggregate regions to remove  
359 multicollinearity that caused unstable coefficient estimates. Using a distance of 18, 21, or  
360 24 m instead of 20m to separate "near" from "far" cover only slightly changed coefficient  
361 estimates. By fitting a model with all tree cover close to a house aggregated into one  
362 variable and then a model with the tree cover separated into 8 variables defined by distance  
363 and direction we tested the overall association of ACE with tree cover and then tested for  
364 specific associations by distance and direction.

## 365 **Building Cover**

366 Nearby buildings likely also affect the energy use of a building. To test this hypothesis  
367 we calculated the area of buildings in each of the eight regions around every building and  
368 included these as covariates in our modeling. We used building footprints from Dane County  
369 which consists of structures the size of a single car garage or larger. The horizontal accuracy  
370 is +/- 6.6 feet for well-defined points, at a ninety percent confidence level.

## 371 **Modeling**

372 We fit linear models where the response was log transformed annual ACE for gas use, for  
373 electricity use, or for gas and electricity combined (net). Because a separate model was  
374 built to explain net C emissions, coefficient estimates for the net model were not precisely  
375 the sum of the coefficients from the electricity and gas models. ACE was log transformed  
376 to meet assumptions of normality and diagnostic plots were assessed to check other model  
377 assumptions and potential sensitivity to influential observations. Our first models aggregated  
378 all tree cover near buildings into one variable, and subsequent models separated tree cover

379 based on direction and distance into eight variables. In addition to tree cover, variables in  
 380 our model were: 5 principal components of building characteristics, building cover in each of  
 381 the 8 regions, and a random effect for elementary school which might capture neighborhood  
 382 characteristics such as culture. We used AIC as a variable selection criterion and in our final  
 383 models only used the first 5 building characteristics principal components and we dropped all  
 384 the building cover covariates. Estimates for the coefficients of tree cover were not sensitive  
 385 to the inclusion or removal of these covariates, but model fit improved. Although some tree  
 386 cover covariates increased AIC, we kept all tree cover covariates in the model because we  
 387 wanted estimates of their effects, however uncertain they might be. We also fit models We  
 388 fit models using the R package lme4 (Bates et al., 2015).

### 389 **Interpreting coefficients**

390 To improve interpretability of coefficients, we back transformed them to the original scale  
 391 and expressed the multiplicative effects as a percentage (Gelman and Hill, 2007). We then  
 392 multiplied this percent change by the median ACE (a better estimator of the central tendency  
 393 because of the right skew in our data) to estimate the typical effect in absolute C terms.  
 394 To get typical effects of tree cover, we multiplied median tree cover in each region by its  
 395 coefficient estimate and back transformed to the original scale.

### 396 **Estimating C storage and sequestration of a green ash with 100m<sup>2</sup>** 397 **canopy**

398 To estimate C storage and sequestration by a single green ash tree with a canopy cover of  
 399 100m<sup>2</sup>, we used allometric equations to estimate that tree's diameter at breast height (DBH)  
 400 and mass and then, assuming an annual DBH growth of 0.61 cm, predicted the change in  
 401 mass to get C sequestration Nowak and Crane (2002); McPherson et al. (2016).

## Extending Analyses from Published Literature

To compare our work to past simulation studies we converted results that were in Therms or kWh to kg C. We did this for Thayer Jr and Maeda (1985), McPherson et al. (1988), and Huang et al. (1990) using updated emission factors corresponding to each study city's eGrid subregion (Emissions & Generation Resource Integrated Database, 2016). To extend Akbari and Konopacki (2005), we joined climate data (heating and cooling degree days) from the nearest NOAA weather station to census tract centroids U.S. Census Tract Centroids (2010); Arguez et al. (2012). It was from this join of climate and census data that we determined that 77% of the U.S. population lives in places with more heating than cooling degree days. Then for each census tract we predicted the effect of trees and increasing roof albedo on the energy use of a pre-1980's building with gas heating following their table that bins houses according to heating degree-days and using emission factors corresponding to the eGrid subregion containing the census tract centroid. Separating out the indirect effects of trees from the indirect effects of increasing roof albedo was not possible because these were not modeled separately. However, the general trend would be similar, but with a decreased electricity savings and a decreased heating penalty. Akbari and Konopacki (2005) found the effect of tree shade to be stronger than the indirect effects of increased roof albedo and transpirative cooling.

## Code

All of the code and data for these analyses are present on Github (<https://github.com/TedwardErker/energy>). Code is provisional pending review.

## References

Akbari, H. (2002). Shade trees reduce building energy use and CO<sub>2</sub> emissions from power plants. *Environmental Pollution*, 116(nil):S119–S126.

426 Akbari, H. and Konopacki, S. (2005). Calculating energy-saving potentials of heat-island  
427 reduction strategies. *Energy Policy*, 33(6):721–756.

428 Akbari, H., Kurn, D. M., Bretz, S. E., and Hanford, J. W. (1997). Peak power and cooling  
429 energy savings of shade trees. *Energy and buildings*, 25(2):139–148.

430 Akbari, H. and Taha, H. (1992). The impact of trees and white surfaces on residential  
431 heating and cooling energy use in four canadian cities. *Energy*, 17(2):141 – 149.

432 Arguez, A., Durre, I., Applequist, S., Vose, R. S., Squires, M. F., Yin, X., Heim, R. R., and  
433 Owen, T. W. (2012). Noaa’s 1981-2010 u.s. climate normals: An overview. *Bulletin of the*  
434 *American Meteorological Society*, 93(11):1687–1697.

435 Bates, D., Mächler, M., Bolker, B., and Walker, S. (2015). Fitting linear mixed-effects models  
436 using lme4. *Journal of Statistical Software*, 67(1):1–48.

437 DeWalle, D. R., Heisler, G. M., and Jacobs, R. E. (1983). Forest home sites influence heating  
438 and cooling energy. *Journal of Forestry*, 81(2):84–88.

439 Donovan, G. H. and Butry, D. T. (2009). The value of shade: Estimating the effect of urban  
440 trees on summertime electricity use. *Energy and Buildings*, 41(6):662–668.

441 Emissions & Generation Resource Integrated Database (2016). Accessed Jul. 24, 2018.

442 Erker, T., Townsend, P. A., Wang, L., Lorentz, L., and Stoltman, A. (2018). A statewide  
443 urban tree canopy mapping method. *Remote Sensing of Environment (in review)*.

444 Gelman, A. and Hill, J. (2007). *Data Analysis Using Regression and Multilevel/Hierarchical*  
445 *Models*. Analytical Methods for Social Research. Cambridge University Press.

446 Heisler, G. M. (1986). Effects of individual trees on the solar radiation climate of small  
447 buildings. *Urban Ecology*, 9(3-4):337–359.

448 Heisler, G. M. (1990). Mean wind speed below building height in residential neighborhoods  
 449 with different tree densities. volume 96. Proceedings of the American Society of Heating,  
 450 Refrigeration and Air conditioning Engineers.

451 Huang, Y. J., Akbari, H., and Taha, H. (1990). The wind-shielding and shading effects  
 452 of trees on residential heating and cooling requirements. volume 96. Proceedings of the  
 453 American Society of Heating, Refrigeration and Air conditioning Engineers.

454 Huang, Y. J., Akbari, H., Taha, H., and Rosenfeld, A. H. (1987). The potential of vegetation  
 455 in reducing summer cooling loads in residential buildings. *Journal of Climate and Applied*  
 456 *Meteorology*, 26(9):1103–1116.

457 Jo, H.-K. and McPherson, E. (2001). Indirect carbon reduction by residential vegetation and  
 458 planting strategies in chicago, usa. *Journal of Environmental Management*, 61(2):165–177.

459 McPherson, E., Simpson, J. R., and Livingston, M. (1989). Effects of three landscape  
 460 treatments on residential energy and water use in tucson, arizona. *Energy and Buildings*,  
 461 13(2):127–138.

462 McPherson, E. G., Herrington, L. P., and Heisler, G. M. (1988). Impacts of vegetation on  
 463 residential heating and cooling. *Energy and Buildings*, 12(1):41–51.

464 McPherson, E. G. and Kendall, A. (2014). A life cycle carbon dioxide inventory of the  
 465 million trees los angeles program. *The International Journal of Life Cycle Assessment*,  
 466 19(9):1653–1665.

467 McPherson, E. G. and Simpson, J. R. (1999). Carbon dioxide reduction through urban  
 468 forestry. *Gen. Tech. Rep. PSW-171, USDA For. Serv., Pacific Southwest Research Station*,  
 469 *Albany, CA*.

470 McPherson, E. G., van Doorn, N. S., and Peper, P. J. (2016). Urban tree database and  
 471 allometric equations.



472 McPherson, G., Simpson, J. R., Peper, P. J., Maco, S. E., and Xiao, Q. (2005). Municipal  
 473 forest benefits and costs in five us cities. *Journal of Forestry*, 103(8):411–416.

474 Nowak, D. J., Appleton, N., Ellis, A., and Greenfield, E. (2017). Residential building energy  
 475 conservation and avoided power plant emissions by urban and community trees in the  
 476 united states. *Urban Forestry & Urban Greening*, 21:158–165.

477 Nowak, D. J. and Crane, D. E. (2002). Carbon storage and sequestration by urban trees in  
 478 the usa. *Environmental Pollution*, 116(3):381–389.

479 Nowak, D. J., Stevens, J. C., Sisinni, S. M., and Luley, C. J. (2002). Effects of urban tree  
 480 management and species selection on atmospheric carbon dioxide.

481 Parker, J. H. (1983). Landscaping to reduce the energy used in cooling buildings. *Journal*  
 482 *of Forestry*, 81(2):82–105.

483 Pataki, D. E., Alig, R. J., Fung, A. S., Golubiewski, N. E., Kennedy, C. A., McPherson,  
 484 E. G., Nowak, D. J., Pouyat, R. V., and Lankao, P. R. (2006). Urban ecosystems and the  
 485 north american carbon cycle. *Global Change Biology*, 12(11):2092–2102.

486 Pataki, D. E., Carreiro, M. M., Cherrier, J., Grulke, N. E., Jennings, V., Pincetl, S., Pouyat,  
 487 R. V., Whitlow, T. H., and Zipperer, W. C. (2011). Coupling biogeochemical cycles in  
 488 urban environments: Ecosystem services, green solutions, and misconceptions. *Frontiers*  
 489 *in Ecology and the Environment*, 9(1):27–36.

490 Roy, S., Byrne, J., and Pickering, C. (2012). A systematic quantitative review of urban tree  
 491 benefits, costs, and assessment methods across cities in different climatic zones. *Urban*  
 492 *Forestry & Urban Greening*, 11(4):351–363.

493 Simpson, J. and McPherson, E. (1998). Simulation of tree shade impacts on residential  
 494 energy use for space conditioning in sacramento. *Atmospheric Environment*, 32(1):69–74.

495 Simpson, J. R. (2002). Improved estimates of tree-shade effects on residential energy use.  
 496 *Energy and Buildings*, 34(10):1067–1076.  
 497 Thayer Jr, R. L. and Maeda, B. T. (1985). Measuring street tree impact on solar performance:  
 498 a five-climate computer modeling study. *Journal of arboriculture (USA)*.  
 499 United State Environmental Protection Agency (2017). Inventory of u.s. greenhouse gas  
 500 emissions and sinks: Annex 2 methodology and data for estimating CO<sub>2</sub> emissions from  
 501 fossil fuel combustion. (430-P-17-001).  
 502 U.S. Census Tract Centroids (2010). Accessed Jul. 24, 2018.  
 503 U.S. Department of Energy, E. I. A. (2009). Wisconsin household energy report.  
 504 Weissert, L., Salmond, J., and Schwendenmann, L. (2014). A review of the current progress in  
 505 quantifying the potential of urban forests to mitigate urban co2 emissions. *Urban Climate*,  
 506 8(nil):100–125.  
 507 Young, R. F. (2011). Planting the living city. *Journal of the American Planning Association*,  
 508 77(4):368–381.  
 509 Ziter, C. D., Pedersen, E. J., Kucharik, C. J., and Turner, M. G. (2019). Scale-dependent  
 510 interactions between tree canopy cover and impervious surfaces reduce daytime urban heat  
 511 during summer. *Proceedings of the National Academy of Sciences*, 116(15):7575–7580.

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 516 McPherson and Simpson (1999)  
 517 Heisler (1990)