- Lower Athabasca Surface Water and Sediment Quality
- 2 Criteria for Protection of Indigenous Use
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16

Executive Summary

Surface water and sediment quality criteria were defined to protect Indigenous water use by 23 Athabasca Chipewyan First Nation (ACFN), Fort McKay First Nation (FMFN) and Mikisew 24 Cree First Nation (MCFN) members in the Lower Athabasca Region (LAR) using two ap-25 proaches: current condition and risk-based. Current condition values were developed by collat-26 27 ing and analyzing surface water and sediment quality monitoring data from multi-stakeholder, government and community-based programs and identifying representative values for three 28 seasons (high flow, open water and under ice). Health risk criteria were defined by identifying 29 valued components that reflect use of surface water by Indigenous community members; con-30 sumption of traditional foods, medicine and surface water, trapping furbearing mammals that 31 consume aquatic biota, the health of wildlife (birds and mammals) from ingesting surface water 32 and diet items, and aquatic ecosystem health. Available surface water and sediment quality 33 34 guidelines were reviewed to identify level of protection for the traditional valued components. When unavailable, health risk criteria were derived using methods prescribed by regulatory 35 agencies, using community specific ingestion rates of traditional foods (fish, and medicinal 36 plants) estimated from a traditional food survey of 230 community members. 37 The study found that goals reflecting current condition of surface water in the LAR indi-38 cated relatively good water and sediment quality, with some exceptions. Current conditions 39 were generally lower than the calculated risk-based criteria, with some exceptions especially 40 for metals and metalloids. For risk-based protection goals, surface water quality guidelines for the protection of human health were available but not from governments in Alberta or Canada. 42 Adopting human health water quality criteria from the United States Environmental Protection 43 Agency provided a good starting point for protection for of community members consuming 44 fish and drinking water from surface water bodies. However, the traditional food consumption 45 rates were higher than those used to derive US EPA criteria and therefore the adoption of this 46 approach in the WQCIUs required modification to account for the higher consumption rates of 47 ACFN, FMFN, and MCFN members. The collection of statistically representative community 48 survey results enabled the risk assessor to analyze and calculate community members' ingestion 49 rates of traditional foods and medicines for the three participating Indigenous communities. 50 The WQCIUs (for surface water and sediment) can be used by Indigenous communities, 51 government and regulatory agencies and industry stakeholders to assess potential changes in surface water and sediment conditions and risks to human and ecological receptors from releases 53 of contaminants from oil sands to the Athabasca River and downstream within Lake Athabasca 54 and the Athabasca Delta. The WQCIUs were developed for constituents characterized in 55

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- 56 oil sands mine water (OSMW), as well as for several additional common constituents and
- 57 measures. As a result, the health risk criteria can be used to assess risks from the placement
- 58 of tailings and OSMW in aquatic closure (reclamation) features such as constructed wetlands
- 59 and End Pit Lakes (EPLs).
- This report is structured as follows: Chapter 1 includes a summary of the study findings,
- 61 and applies health risk criteria to the calculated current conditions in the Lower Athabasca
- 62 River, the Athabasca River Delta and Lake Athabasca; Chapter 2 details the development of
- 63 the current conditions, Chapters 3 and 4 detail the development of the health risk criteria; and
- 64 Chapter 5 provides some detail about the community consumption surveys conducted with
- 65 and by ACFN, MCFN and FMFN.
- 66 Keywords: Indigenous, protection, goals, Indigenous land use, traditional food, community
- 67 survey, ingestion rate, monitoring, non-degradation, risk, health, human, wildlife, aquatic
- 68 biota, ecosystem, oil sands, tailings, OSPW, wetlands, end pit lakes, Athabasca River,
- 69 Athabasca River Delta, Lake Athabasca.

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Chapter 1

33 Summary and Application of

Findings

This document outlines an approach for the development of health risk criteria and establish-235 ment of current conditions against which chemical parameters in surface water and sediment 236 can be assessed to identify potential health risks as well as changes in conditions over time and 237 space. These Water and Sediment Quality Criteria for the Protection of Indigenous Use (WQ-238 CIUs) were developed for the protection of water use by Athabasca Chipewyan First Nation 239 (ACFN), Fort McKay First Nation (FMFN) and Mikisew Cree First Nation (MCFN) in the 240 Lower Athabasca River region (LAR) of Alberta. This chapter describes key results from this 241 study and provides a comparison of the current condition of the Athabasca River, Athabasca 242 River Delta and Lake Athabasca to the health risk criteria. 243 The WQCIUs were developed to address gaps in existing government water, sediment 244 and tissue guidelines and water quality management frameworks. ACFN, FMFN and MCFN 245 246 expectations for establishment of current condiditions were that they would be season or flowspecific, and that they would be established for the entire Lower Athabasca Region (river, 247 delta, lake). ACFN, FMFN and MCFN expectations for establishment of health risk criteria 248 were that they would include all constituents of concern in the region, that they would account 249 for bioaccumulation and biomagnification effects, that they would include humans, wildlife and 250 plants as receptors, and that they would account for Indigenous community water uses. 251 The WQCIUs were developed to specifically consider the rights of Indigenous Peoples¹ and 252 to support the evaluation of environmental conditions relative to tiers, triggers, limits, thresh-253

¹Indigenous peoples possess the same rights as all people, and specific rights as Indigenous people, such as Aboriginal and Treaty Rights enshrined in the Constitution Act, 1982, and through UNDRIP.

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olds or other "limits of change" that ensure ecosystem components are sustainable, ecosystems are healthy, and effects to human health and well-being are avoided, minimized, or reduced as defined under the Oil Sands Monitoring Program (OSM)² Program.

More broadly, the health risk criteria and current conditions provide government and industry stakeholders with a framework and criteria for assessing performance of treatment technologies, produced effluents, and remediation and reclamation activities that reflect the values and interests of participating Indigenous communities. This includes risk tolerances and protection requirements for establishing and maintaining safe and usable environments to support exercising Aboriginal Rights, as defined by ACFN, FMFN and MCFN.

The WQCIUs should not necessarily be adopted as guidelines or objectives, which are prescribed under provincial policy and may be applied as legislative requirements³. Rather the WQCIUs reflect performance criteria which should be used to assess the health and safety of aquatic ecosystems to support Indigenous water uses.

1.1 Ecosystem Approach to Water Management

Health risk criteria and current conditions were developed for protection of ecosystem function which includes ecological and human receptors and their interactions with abiotic components of the environment (Keen et al., 2012) as described in Figure 1.1).

Environmental management decisions which consider the complex interactions within ecosystems more closely resemble the world views of Indigenous communities and traditional strategies for assessing and managing natural resources and minimizing health risks (Liboiron, 2021).

³Guidelines are science-based recommendations that form a cornerstone of water quality and aquatic ecosystem management. They are not legal instruments, however, guidelines and the site-specific objectives derived from them can be used in developing legally binding effluent limits under the Environmental Protection and Enhancement Act (EPEA). They can also be used in management frameworks as part of Regional Plans developed under the Land-use Framework (GoA, 2008) and the Alberta Land Stewardship Act, as well as other management tools. They are an integral component of the GOA Integrated Resource Management system that operates in accordance with the principle of cumulative effects management. The guidelines in this document support the Water Quality Based Effluent Limits Procedures Manual (AEP, 1995), the Alberta Tier 1 Soil and Groundwater Remediation Guidelines (Alberta Environment and Parks (AEP, 2016a), and the Alberta Tier 2 Soil and Groundwater Remediation Guidelines (AEP, 2016b). The recreation and aesthetic guidelines also support those in use by Alberta Health under the Public Health Act.

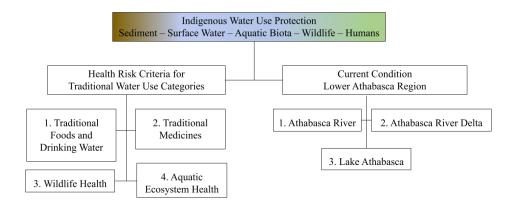


Figure 1.1: Ecosystem health approach to developing health risk criteria and current conditions for the protection of Indigenous water use and interactions with surface water and sediment.

275 1.2 Water Use by Indigenous Communities

276 Four water use categories, as presented in Table 1.1 were defined based on descriptions of water use described by community members from ACFN, FMFN and MCFN. The four categories 277 278 were used to develop a conceptual model linking community members to the environment through exposure pathways, as well as identifying protection goals for surface water and sedi-279 ment (see Section 3.4.1 of this report for more details of this process). In the development of 280 Indigenous water use categories, water use by gender or age were not considered and further 281 study may be necessary to understand exposure pathways by gender or age across the com-282 munity. However, gender and age were considered in understanding community consumption 283 patterns, barriers to consuming traditional foods and medicines and in the development of 284 health risk criteria which considered consumption of traditional foods. Water is a core com-285 ponent of all aspects of life for ACFN, FMFN, and MCFN members. Each of the water use 286 categories identified below should be understood as inextricably linked to ACFN, FMFN, and 287 MCFN's cultural and spiritual value of water. 288

Table 1.1: Indigenous community water uses and health protection goals used to define water use criteria.

| Indigenous water use | Protection Goal |
|--------------------------------------|---|
| Traditional foods and drinking water | Safe foods consumption |
| | Safe natural surface water consumption |
| Traditional medicines | Safe medicine consumption |
| Aquatic ecosystem health | Aquatic community consumption unchanged |
| | Robust populations |
| | Natural behaviours and patterns |
| Wildlife health | Healthy wildlife |
| | Robust populations |
| | Natural behaviours and patterns |
| | Good quality pelts |

Exposure pathways, indicators and endpoints linked to water protection goals were then used to evaluate the level of protection offered by applying provincial and federal surface water quality guidelines. The results indicate that exposure pathways (ingestion of traditional foods, medicine, and surface water) and endpoints (e.g., carcinogenicity) for the protection of human health are not considered under environmental quality guidelines for the protection of surface water in Alberta or Canada (GoA, 2018; CCME, 2021). Protection goals linked to wildlife species are either less sensitive or not considered as frequently as aquatic biota, which was identified as the key protection endpoint. No reference to the protection of surface water for the spiritual and cultural needs of Indigenous communities were identified, as this was beyond the scope of this study. However, these are important components for inclusion in future work aimed at protecting all community water uses holistically.

Sediment is an integral component of aquatic ecosystems providing a substrate for fish and invertebrates to reproduce and live in and plants to grow but also a source of nutrients and energy supporting ecosystem production that supports the energy needs of food webs. Sediments act as sources and sinks for environmental contaminants, which can directly affect the health and diversity of benthos (plants and animals living at the bottom of a water body) interacting with the sediment and contribute to the biomagnification of persistent contaminants in aquatic and terrestrial food webs.

A review of sediment quality guidelines adopted in Alberta indicates a low level of protection

both for benthic organisms and overlaying surface water due to limitations in available sediment toxicity test data and derivation methods.

310 1.3 Water and Sediment Quality Criteria for Indigenous

311 Use Protection

- Review of provincial water quality management tools under policy and regulations revealed that the following are not currently considered by Alberta when assessing the condition of surface water to support management decisions.
- Surface water is not assessed as a drinking water source (GoA, 2018)
- Assessing the partitioning of contaminants to sediments and subsequent deposition and downstream transport is not required (AEP, 1995)
- Persistence and biomagnification of contaminants within aquatic and semi-aquatic food webs is not assessed (AEP, 1995; GoA, 2018)
- Risk to human health from ingestion of surface water and aquatic biota do not need to be assessed beyond application of Alberta surface water guidelines for aquatic life and recreation use (GoA, 2018)
- Current guidance on releases allow for impacts to acute and chronic mixing zone areas within natural receiving water (AEP, 1995)
- Water, sediment and tissue quality guidelines have not been published for each contaminant identified as having intrinsically toxic properties and characterized in oil sands mine water (i.e. naphthenic acids, low and high molecular weight PAHs).
- The identified limitations in the provincial system for assessing and managing environmental and human risks from contaminants in surface water and sediment were addressed by developing health risk criteria for those media which allows for an assessment of potential impacts to Indigenous water use pathways; traditional foods and drinking water, traditional medicines, aquatic ecosystem health, and wildlife health.
- Figure 1.2 (below) summarizes findings from a review of federal, provincial and international water quality guidelines for the protection of freshwater life/ aquatic biota (US EPA, AEP, CCME), wildlife (AEP, CCME, Sample et al. (1996)) and humans (US EPA, Health Canada, WHO). The pie chart indicates the percentage of published water quality guidelines that were developed to protect the most sensitive receptor group from the contaminants of interest evaluated in this study. The results indicate that humans are the most sensitive re-

ceptor group from exposure to 52% of the contaminants for which published water quality guidelines are available. Aquatic biota are the next most sensitive receptor group (45%) and finally wildlife species are generally less sensitive than human and aquatic receptors (3% of available guidelines noted wildlife species as the most sensitive receptors). It is important to note that there was a lack of wildlife watering guidelines available for several parameters and additional health risk criteria were not derived, only available guidelines for livestock were adopted.

This is an important finding which supports the inclusion of guidelines derived for the protection of human health (Health Canada, US EPA, WHO), specifically for carcinogenic substances, which are not an assessment endpoint considered in protection of aquatic life or wildlife/ livestock water quality guidelines (AEP, CCME).

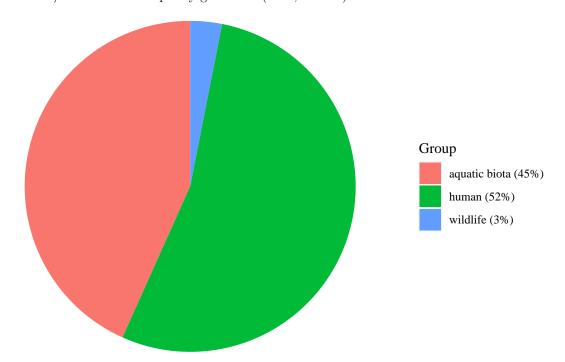


Figure 1.2: Number (percentage) of published human and environmental quality guidelines that are driven by human, aquatic biota or wildlife species as the most sensitive receptor group (n = 308)

Modifications of the published guidelines were also used to achieve a higher degree of protection for consumers of traditional foods from the communities of ACFN, FMFN, and MCFN, as previously reported consumption rates representing the general population (22 g/d; (US EPA, 2015a) and Northern Alberta Indigenous communities (27.8 g/d; (Chan et al., 2016)) were lower than those reported through the community surveys for fish (388 g/d), and rat root (6.8 g/d).

A generic health risk criteria for surface water quality that identifies the most sensitive

water use by contaminant is proposed as a conservative approach similar to that adopted for assessing soil and groundwater contamination (GoA, 2018). The generic health risk criteria should be applied unless a specific water use category is being assessed to answer community or research study questions and each water use category is not being assessed individually. A single health risk criteria for sediment quality (mg/kg) is proposed for the protection of sediment associated biota and biomagnification within aquatic food webs.

Importantly, the health risk criteria proposed here address gaps in published surface water and sediment quality guidelines that could underestimate potential effects in the ambient environment from mixtures of PAHs from varying modes of action. The WQCIUs were developed to assess surface water and sediment quality monitoring data by various groupings. The first, carcinogenicity (as BaP equivalents) is used to assess potential risks to human receptors from exposure to carcinogenic PAH congeners in the traditional foods and drinking water and medicines water use categories. The second are non-carcinogenic effect-based groups for high and low molecular PAH congeners and are used assess risks to ecological receptors in the aquatic ecosystem and wildlife health water use categories.

Together, the Indigenous criteria for water (generic) and sediment presented in Table 1.2 and Table 1.3, will allow ACFN, FMFN and MCFN to assess the ability for surface water bodies to meet their needs by ensuring water, animals, and plants are safe to consume and that populations are healthy and available to support Indigenous use.

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories.

| | | Generic (All water uses protected) | | | |
|---------------------------------|--------------------|------------------------------------|-------------------|--------------------|---|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source |
| .alphaEndosulfan | | ug/L | 0.056 | aquatic biota | US EPA Aquatic Life Criteria |
| .betaEndosulfan | | $\mathrm{ug/L}$ | 0.056 | aquatic biota | US EPA Aquatic Life Criteria |
| 1,1,1-Trichloroethane* | | ug/L | 200 | human | US EPA DWR |
| 1,1,2,2- Tetrachloroethane* | | $\mathrm{ug/L}$ | 2 | human | HH DW+Org (US EPA) |
| 1,1,2-Trichloroethane | | ug/L | 3 | human | US EPA DWR |
| 1,1-Dichloroethylene | | ug/L | 7 | human | US EPA DWR |
| 1,2,3,4- Tetrachlorobenzene | | m ug/L | 0.03 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| 1,2,3-Trichlorobenzene | | $\mathrm{ug/L}$ | 8 | aquatic biota | AEP Water PAL CCME Water PAL |
| 1,2,4-Trichlorobenzene | | ug/L | 0.071 | human | HH DW+Org (US EPA) |
| 1,2-Dibromo-3- chloropropane | | ug/L | 0.2 | human | US EPA DWR |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | | Generic (All water use | es protected) |
|--|--------------------|-----------------|-------------------|------------------------|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source |
| 1,2-Dibromoethane | | ug/L | 0.4 | human | WHO DW |
| 1,2-Dichlorobenzene | | ug/L | 0.7 | aquatic biota | AEP Water PAL |
| 1,2-Dichloroethane* | | m ug/L | 5 | human wildlife | Health Canada DW AEP Water Ag CCME Water Ag (limited) US EPA DWR |
| 1,2-Dichloroethene | | ug/L | 50 | human | WHO DW |
| 1,2-Dichloropropane* | | ug/L | 5 | human | US EPA DWR |
| 1,2- Diphenylhydrazine [*] | | $\mathrm{ug/L}$ | 0.3 | human | HH DW+Org (US EPA) |
| 1,3-Dichlorobenzene | | ug/L | 7 | human | HH DW+Org (US EPA) |
| 1,3-Dichloropropene* | | ug/L | 2.7 | human | HH DW+Org (US EPA) |
| 1,4-Dichlorobenzene | | ug/L | 26 | aquatic biota | AEP Water PAL |
| 1,4-Dioxane | | $\mathrm{ug/L}$ | 50 | human | WHO DW |
| 2,3,4,6- Tetrachlorophenol | | ug/L | 1 | human | USEPA WQC AO |
| 2,3-Dichlorophenol | | $\mathrm{ug/L}$ | 0.04 | human | USEPA WQC AO |
| 2,4,5-Trichlorophenol | | ug/L | 1 | human | USEPA WQC AO |
| 2,4,6-Trichlorophenol* | | ug/L | 2 | human | USEPA WQC AO |
| 2,4-D | | ug/L | 4 | aquatic biota | CCME Water PAL AEP Water PAL |
| 2,4-DB | | ug/L | 25 | aquatic biota | AEP Water PAL |
| 2,4-Dichlorophenol | | ug/L | 0.3 | human | USEPA WQC AO |
| 2,4-Dimethylphenol | | ug/L | 100 | human | HH DW+Org (US EPA) |
| 2,4-Dinitrophenol | | $\mathrm{ug/L}$ | 10 | human | HH DW+Org (US EPA) |
| 2,4-Dinitrotoluene* | | $\mathrm{ug/L}$ | 0.49 | human | HH DW+Org (US EPA) |
| 2,5-Dichlorophenol | | ug/L | 0.5 | human | USEPA WQC AO |
| 2,6-Dichlorophenol | | ug/L | 0.2 | human | USEPA WQC AO |
| 2-Chloronaphthalene | | ug/L | 800 | human | HH DW+Org (US EPA) |
| 2-Chlorophenol | | ug/L | 0.1 | human | USEPA WQC AO |
| 2-Methyl-4,6- Dinitrophenol | | $\mathrm{ug/L}$ | 2 | human | HH DW+Org (US EPA) |
| 2-Methyl-4- Chlorophenol | | $\mathrm{ug/L}$ | 1800 | human | USEPA WQC AO |
| 3,3'-Dichlorobenzidine | | $\mathrm{ug/L}$ | 0.49 | human | HH DW+Org (US EPA) |
| 3,4-Dichlorophenol | | ug/L | 0.3 | human | USEPA WQC AO |
| 3-Chlorophenol | | ug/L | 0.1 | human | USEPA WQC AO |
| 3-Iodo-2-propynyl butyl carbamate | | ug/L | 1.9 | aquatic biota | CCME Water PAL AEP Water PAL |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|--|--------------------|-----------------|------------------------------------|--------------------|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| 3-Methyl-4- Chlorophenol | | ug/L | 500 | human | HH DW+Org (US EPA) | |
| 3-Methyl-6- Chlorophenol | | ug/L | 20 | human | USEPA WQC AO | |
| 4-Chlorophenol | | ug/L | 0.1 | human | USEPA WQC AO | |
| Acenaphthene [§] | | $\mathrm{ug/L}$ | 4.79 | human | HH DW+Org (derived) | |
| Acridine | | ug/L | 4.4 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Acrolein | | ug/L | 2.87 | human | HH DW+Org (derived) | |
| Acrylamide | | ug/L | 0.07 | human | HH DW+Org (derived) | |
| Acrylonitrile* | | ug/L | 0.53 | human | HH DW+Org (derived) | |
| Alachlor | | $\mathrm{ug/L}$ | 2 | human | US EPA DWR | |
| Alcohol ethoxylates | | $\mathrm{ug/L}$ | 70 | aquatic biota | FEQG Water PAL | |
| Aldicarb | | ug/L | 1 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Aldrin^* | | ug/L | 0.0000077 | human | USEPA WQC HH Or HH DW+Org (US EPA) | |
| Aldrin and dieldrin | | ug/L | 0.03 | human | WHO DW | |
| Alkalinity, total | | mg/L | 20 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | |
| alpha-Endosulfan | | ug/L | 1.82 | human | HH DW+Org (derived) | |
| alpha- ${\rm Hexachlorocyclohexane}^*$ | | ug/L | 0.0002 | human | HH DW+Org (derived) | |
| Aluminum | Total | ug/L | 18 | wildlife | US DOE Wildlife | |
| Aluminum | Dissolved | $\mathrm{ug/L}$ | 50 | aquatic biota | AEP Water PAL | |
| Ammonia | | mg/L | 0.67 | human | HH DW+Org (derived) | |
| Ammonia, unionized | | $\mathrm{mg/L}$ | 0.016 | aquatic biota | AEP Water PAL | |
| Aniline | | ug/L | 2.2 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Anthracene | | ug/L | 0.012 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Antimony | Total | ug/L | 4.59 | human | HH DW+Org (derived) | |
| Arsenic* | Total | ug/L | 0.03 | human | HH DW+Org (derived) | |
| Arsenic ^{*††} | Dissolved | ug/L | 150 | aquatic biota | US EPA Aquatic Life Criteria | |
| Asbestos | | ug/L | 7 | human | US EPA DWR HH DW+Org (US EPA) | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | | Generic (All water use | All water uses protected) | | |
|--|--------------------|-----------------|-------------------|------------------------|--|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | | |
| Atrazine | | $\mathrm{ug/L}$ | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL | | |
| Atrazine and its chloro-s-triazine metabolites | | $\mathrm{ug/L}$ | 100 | human | WHO DW | | |
| Azinphos-methyl | | $\mathrm{ug/L}$ | 0.01 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | | |
| Barium | Total | $\mathrm{ug/L}$ | 1000 | human | HH DW+Org (US EPA) Health Canada DW | | |
| Benzene* | | ug/L | 2.11 | human | HH DW+Org (derived) | | |
| Benzidine* | | ug/L | 0.001 | human | HH DW+Org (derived) | | |
| $Benzo(a) anthracene^{*\dagger}$ | | $\mathrm{ug/L}$ | 0.001 | human | HH DW+Org (derived) | | |
| $Benzo(a)pyrene^{*\dagger}$ | | $\mathrm{ug/L}$ | 0.0001 | human | HH DW+Org (derived) | | |
| $Benzo(b) fluoranthene^{*\dagger}$ | | $\mathrm{ug/L}$ | 0.001 | human | HH DW+Org (derived) | | |
| $Benzo(k) fluoranthene^{*\dagger}$ | | ug/L | 0.01 | human | HH DW+Org (derived) | | |
| Beryllium | Total | $\mathrm{ug/L}$ | 3.27 | human | HH DW+Org (derived) | | |
| beta-Endosulfan | | $\mathrm{ug/L}$ | 2.87 | human | HH DW+Org (derived) | | |
| $\begin{array}{l} \text{beta-} \\ \text{Hexachlorocyclohexane}^* \end{array}$ | | ug/L | 0.01 | human | HH DW+Org (derived) | | |
| Bis(2-Chloro-1- methylethyl) Ether | | ug/L | 127.99 | human | HH DW+Org (derived) | | |
| Bis(2-Chloroethyl) Ether* | | ug/L | 0.25 | human | HH DW+Org (derived) | | |
| Bis(2-Ethylhexyl) Phthalate | | ug/L | 0.21 | human | HH DW+Org (derived) | | |
| $\begin{array}{l} \operatorname{Bis}(\operatorname{Chloromethyl}) \\ \operatorname{Ether}^* \end{array}$ | | ug/L | 0.001 | human | HH DW+Org (derived) | | |
| Bisphenol A-d6 | | ug/L | 3.5 | aquatic biota | FEQG Water PAL | | |
| Boron | Total | ug/L | 1333.33 | human | HH DW+Org (derived) | | |
| Bromacil | | ug/L | 5 | aquatic biota | AEP Water PAL CCME Water PAL | | |
| Bromate | | $\mathrm{ug/L}$ | 10 | human | Health Canada DW US EPA DWR WHO DW | | |
| Bro-modichloromethane | | $\mathrm{ug/L}$ | 6.33 | human | HH DW+Org (derived) | | |
| Bromoform | | $\mathrm{ug/L}$ | 7 | human | HH DW+Org (US EPA) | | |
| | | | | | | | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | | Generic (All water use | es protected) |
|--|--------------------|-----------------|-------------------|------------------------|---|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source |
| Bromoxynil | | ug/L | 5 | aquatic biota human | AEP Water PAL CCME Water PAL Health Canada DW |
| Butylbenzyl Phthalate* | | ug/L | 0.06 | human | HH DW+Org (derived) |
| Cadmium [‡] | Total | ug/L | 0.002 | human | HH DW+Org (derived) |
| Cadmium [‡] †† | Dissolved | ug/L | 0.824 | aquatic biota | US EPA Aquatic Lif Criteria |
| Calcium | | $\mathrm{mg/L}$ | 1000 | wildlife | CCME Water Ag (limited) AEP Water Ag |
| Captan | | ug/L | 1.3 | aquatic biota | CCME Water PAL AEP Water PAL |
| Carbamazepine | | ug/L | 10 | aquatic biota | CCME Water PAL AEP Water PAL |
| Carbaryl | | ug/L | 0.2 | aquatic biota | AEP Water PAL CCME Water PAL |
| Carbofuran | | ug/L | 1.8 | aquatic biota | CCME Water PAL AEP Water PAL |
| Carbon tetrachloride | | ug/L | 1.9 | human | HH DW+Org (derived) |
| Chloramines | | $\mathrm{ug/L}$ | 0.5 | aquatic biota | CCME Water PAL |
| Chlorate | | $\mathrm{ug/L}$ | 700 | human | WHO DW |
| Chlordane | | ug/L | 0.001 | human | HH DW+Org (derived) |
| Chloride | | mg/L | 120 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chlorinated paraffins, long-chain, C18-C20 | | ug/L | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL |
| Chlorinated paraffins, medium-chain, C14-C17 | | $\mathrm{ug/L}$ | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL |
| Chlorinated paraffins, short-chain, C10-C13 | | ug/L | 2.4 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Chlorine | | ug/L | 0.5 | aquatic biota | AEP Water PAL |
| Chlorine dioxide | | ug/L | 800 | human | US EPA DWR |
| Chlorite | | ug/L | 700 | human | WHO DW |
| Chlorobenzene | | ug/L | 1.3 | aquatic biota | AEP Water PAL |
| Chlorodibro- momethane | | ug/L | 8 | human | HH DW+Org (US EPA) |
| Chloroform | | ug/L | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL |
| Chlorophenol | | ug/L | 7 | aquatic biota | AEP Water PAL CCME Water PAL |
| Chlorophenoxy Herbicide (2,4,5-TP) [Silvex] | | $\mathrm{ug/L}$ | 20.55 | human | HH DW+Org (derived) |
| Chlorothalonil | | ug/L | 0.18 | aquatic biota | CCME Water PAL AEP Water PAL |
| | | | | | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | | Generic (All water uses protected) | | |
|--------------------------------|--------------------|-----------------|-------------------|------------------------------------|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Chlorotoluron | | ug/L | 30 | human | WHO DW | |
| Chlorpyrifos | | ug/L | 0.002 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Chromium | Total | ug/L | 50 | human | WHO DW Health Canada DW | |
| Chromium (III) [‡] | Total | ug/L | 8.9 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Chromium (III) [‡] †† | Dissolved | ug/L | 100.92 | aquatic biota | US EPA Aquatic Life Criteria | |
| Chromium (VI) | Total | ug/L | 1 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Chromium (VI) | Dissolved | ug/L | 5 | aquatic biota | FEQG Water PAL | |
| Chrysene*† | | ug/L | 0.07 | human | HH DW+Org (derived) | |
| cis-1,2- Dichloroethylene | | ug/L | 70 | human | US EPA DWR | |
| Cobalt [‡] | Total | ug/L | 1.10 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Copper* [‡] | Total | ug/L | 2.76 | aquatic biota | CCME Water PAL | |
| Copper | Dissolved | ug/L | 0.53 | aquatic biota | FEQG Water PAL | |
| Cyanazine | | ug/L | 0.6 | human | WHO DW | |
| Cyanide | | ug/L | 3.62 | human | HH DW+Org (derived) | |
| Cyanobacterial toxins | | ug/L | 1.5 | human | Health Canada DW | |
| Dalapon | | ug/L | 200 | human | US EPA DWR | |
| DDT and metabolites* | | ug/L | 0.000004 | wildlife | US DOE Wildlife | |
| Deltamethrin | | ug/L | 0.0004 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Demeton | | ug/L | 0.1 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | |
| Di(2-ethylhexyl) adipate | | $\mathrm{ug/L}$ | 400 | human | US EPA DWR | |
| Di(2-ethylhexyl) phthalate | | ug/L | 6 | human | US EPA DWR | |
| Di-n-Butyl Phthalate | | $\mathrm{ug/L}$ | 0.15 | wildlife | US DOE Wildlife | |
| Diazinon | | $\mathrm{ug/L}$ | 0.17 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | |
| $Dibenzo(a,h)anthracene^*$ | | ug/L | 0.0001 | human | HH DW+Org (derived) | |
| Dibromoacetonitrile | | $\mathrm{ug/L}$ | 70 | human | WHO DW | |
| Dibro- mochloromethane | | ug/L | 5.21 | human | HH DW+Org (derived) | |
| Dicamba | | ug/L | 10 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Dichloroacetate | | $\mathrm{ug/L}$ | 50 | human | WHO DW | |
| ${\bf Dichloroacetonitrile}^*$ | | $\mathrm{ug/L}$ | 20 | human | WHO DW | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|------------------------------------|--------------------|-----------------|------------------------------------|--------------------|---|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Dichlorobro- momethane | | $\mathrm{ug/L}$ | 9.5 | human | HH DW+Org (US EPA) | |
| ${\bf Dichloromethane}^*$ | | ug/L | 5 | human | US EPA DWR | |
| Dichlorophenol | | $\mathrm{ug/L}$ | 0.2 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Dichlorprop | | ug/L | 100 | human | WHO DW | |
| Diclofop-methyl | | ug/L | 6.1 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Didecyl dimethyl ammonium chloride | | ug/L | 1.5 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Dieldrin* | | ug/L | 0.00001 | human | HH DW+Org (derived) HH DW+Org (US EPA) | |
| Diethanolamine | | ug/L | 450 | aquatic biota | AEP Water PAL | |
| Diethyl Phthalate | | $\mathrm{ug/L}$ | 35.61 | human | HH DW+Org (derived) | |
| Diethylene glycol | | ug/L | 150000 | aquatic biota | AEP Water PAL | |
| Diisopropanolamine | | $\mathrm{ug/L}$ | 1600 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Dimethoate | | $\mathrm{ug/L}$ | 3 | wildlife | CCME Water Ag (limited) AEP Water Ag | |
| Dimethyl Phthalate | | ug/L | 102.91 | human | HH DW+Org (derived) | |
| Dinitrophenols | | $\mathrm{ug/L}$ | 10 | human | HH DW+Org (US EPA) | |
| Dinoseb | | ug/L | 0.05 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Dioxin (2,3,7,8-TCDD) | | ug/L | 0.00000002 | 134 wildlife | US DOE Wildlife | |
| Diquat | | ug/L | 20 | human | US EPA DWR | |
| Diuron | | ug/L | 150 | human | Health Canada DW | |
| Edetic acid | | ug/L | 600 | human | WHO DW | |
| Endosulfan | | ug/L | 0.003 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Endosulfan Sulfate | | $\mathrm{ug/L}$ | 2.63 | human | HH DW+Org (derived) | |
| Endothall | | ug/L | 100 | human | US EPA DWR | |
| Endrin | | ug/L | 0.001 | wildlife | US DOE Wildlife | |
| Endrin Aldehyde | | $\mathrm{ug/L}$ | 0.11 | human | HH DW+Org (derived) | |
| Epichlorohydrin | | ug/L | 0.4 | human | WHO DW | |
| Ethanol | | | 123377 | wildlife | US DOE Wildlife | |
| Ethinyl estradiol | | ng/L | 0.5 | aquatic biota | AEP Water PAL | |
| Ethyl acetate | | | 136465 | wildlife | US DOE Wildlife | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|--|--------------------|-----------------|------------------------------------|------------------------|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Ethylbenzene | | $\mathrm{ug/L}$ | 2.4 | wildlife | AEP Water Ag CCME Water Ag (limited) | |
| Ethylene dibromide | | ug/L | 0.05 | human | US EPA DWR | |
| Ethylene glycol | | $\mathrm{ug/L}$ | 192000 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Fenoprop | | $\mathrm{ug/L}$ | 9 | human | WHO DW | |
| Fluoranthene [§] | | ug/L | 0.04 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Fluorene [§] | | ug/L | 3 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Fluoride | | $\mathrm{mg/L}$ | 0.12 | aquatic biota | CCME Water PAL | |
| Formaldehyde | | | 73910 | wildlife | US DOE Wildlife | |
| gamma- Hexachlorocyclohexane [Lindane] | | ug/L | 0.01 | aquatic biota | AEP Water PAL | |
| Glyphosate | | ug/L | 280 | human wildlife | AEP Water Ag Health Canada DW CCME Water Ag (limited) | |
| Haloacetic acids | | ug/L | 60 | human | US EPA DWR | |
| neptaBDE | | ng/L | 14 | aquatic biota | FEQG Water PAL | |
| Heptachlor* | | ug/L | 0.00004 | human | HH DW+Org (derived) | |
| Heptachlor epoxide [*] | | ug/L | 0.0001 | human | HH DW+Org (derived) | |
| hexaBDE | | ng/L | 120 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Hexabromocyclodode- cane | | ug/L | 0.56 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Hexachlorobenzene* | | ug/L | 0.0001 | human | HH DW+Org (derived) | |
| Hexachlorobutadiene* | | ug/L | 0.001 | human | HH DW+Org (derived) | |
| $\operatorname{Hexachlorocyclohexane}^*$ | | $\mathrm{ug/L}$ | 0.01 | aquatic biota human | HH DW+Org (derived) CCME Water PAL | |
| Hexachlorocyclopenta- diene | | ug/L | 0.4 | human | HH DW+Org (derived) | |
| Hexachloroethane* | | ug/L | 0.02 | human | HH DW+Org (derived) | |
| Hydrazine | | ug/L | 2.6 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Hydrogen Sulfide | | ug/L | 2 | aquatic biota | US EPA Aquatic Lis Criteria | |
| Hydroxyatrazine | | $\mathrm{ug/L}$ | 200 | human | WHO DW | |
| Imidacloprid | | ug/L | 0.23 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Indeno(1,2,3- $^{\circ}$ cd)pyrene *† | | ug/L | 0.001 | human | HH DW+Org (derived) | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|--|--------------------|-----------------|------------------------------------|------------------------|---|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Inorganic nitrogen (nitrate and nitrite) | Dissolved | $\mathrm{mg/L}$ | 100 | wildlife | CCME Water Ag (limited) AEP Water Ag | |
| Iron | Total | $\mathrm{ug/L}$ | 300 | aquatic biota human | CCME Water PAL USEPA WQC AO | |
| Iron | Dissolved | ug/L | 300 | aquatic biota | AEP Water PAL | |
| Isophorone* | | $\mathrm{ug/L}$ | 268.41 | human | HH DW+Org (derived) | |
| Isoproturon | | ug/L | 9 | human | WHO DW | |
| $\operatorname{Lead}^{\ddagger}$ | Total | ug/L | 4.01 | aquatic biota | AEP Water PAL CCME Water PAL | |
| $\operatorname{Lead}^{\ddagger \ \dagger \dagger}$ | Dissolved | $\mathrm{ug/L}$ | 3.07 | aquatic biota | US EPA Aquatic Lif Criteria | |
| Linuron | | ug/L | 7 | aquatic biota | CCME Water PAL AEP Water PAL | |
| m-Dichlorobenzene | | ug/L | 150 | aquatic biota | CCME Water PAL | |
| Malathion | | $\mathrm{ug/L}$ | 0.1 | aquatic biota | AEP Water PAL US EPA Aquatic Lif Criteria | |
| Manganese | Total | ug/L | 50 | human | HH DW+Org (US EPA) | |
| MCPA | | $\mathrm{ug/L}$ | 2.6 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Mecoprop | | $\mathrm{ug/L}$ | 10 | human | WHO DW | |
| Mercury | Total | $\mathrm{ug/L}$ | 0.0016 | wildlife | US DOE Wildlife | |
| Mercury ^{††} | Dissolved | $\mathrm{ug/L}$ | 0.77 | aquatic biota | US EPA Aquatic Lif Criteria | |
| Mercury (methyl) | Total | ug/L | 0.001 | aquatic biota | AEP Water PAL | |
| Mercury (methyl) | Dissolved | ug/L | 0.004 | aquatic biota | CCME Water PAL | |
| Methanol | | ug/L | 1500 | aquatic biota | AEP Water PAL | |
| Methoprene | | $\mathrm{ug/L}$ | 0.09 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Methoxychlor | | $\mathrm{ug/L}$ | 0.001 | human | HH DW+Org (derived) | |
| Methyl Bromide | | ug/L | 100 | human | HH DW+Org (US EPA) | |
| Methyl tert-butyl ether | | ug/L | 10 | aquatic biota | AEP Water PAL | |
| Methylene chloride* | | ug/L | 32.62 | human | HH DW+Org (derived) | |
| Metolachlor | | ug/L | 7.8 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Metribuzin | | ug/L | 1 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Microcystin-LR | | $\mathrm{ug/L}$ | 1 | human | WHO DW | |
| Mirex | | ug/L | 0.001 | aquatic biota | US EPA Aquatic Lif Criteria AEP Water PAL | |
| Molinate | | $\mathrm{ug/L}$ | 6 | human | WHO DW | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | - | Generic (All water use | es protected) |
|--|--------------------|-----------------|-------------------|--|---|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source |
| Molybdenum | Total | ug/L | 33.33 | human | HH DW+Org (derived) |
| Monochloramine | | ug/L | 3000 | human | WHO DW |
| Monochloroacetate | | ug/L | 20 | human | WHO DW |
| Monochlorobenzene | | ug/L | 1.3 | aquatic biota | CCME Water PAL AEP Water PAL |
| Monoethanolamine | | ug/L | 75 | aquatic biota | AEP Water PAL |
| N-Nitrosodi-n- Propylamine* | | ug/L | 0.05 | human | HH DW+Org (US EPA) HH DW+Org (derived) |
| N- Nitrosodimethylamine* | | ug/L | 0.007 | human | HH DW+Org (US EPA) |
| N- Nitrosodiphenylamine* | | ug/L | 33 | human | HH DW+Org (US EPA) |
| Naphthalene [§] | | ug/L | 1 | aquatic biota | AEP Water PAL |
| Naphthenic acids (Lower Athabasca River) | Total | ug/L | < 0.05 | Adopted current condition (Oil Sands Monitoring Program Reporting Limit) | |
| Naphthenic acids (Athabasca River Delta) | Total | ug/L | 230 | Adopted current condition (50th percentile, high flow) | |
| Naphthenic acids (Lake Athabasca) | Total | $\mathrm{ug/L}$ | 140 | Adopted current condition (50th percentile, open water) | |
| Nickel [‡] | Total | ug/L | 7.35 | human | HH DW+Org (derived) |
| Nickel [‡] †† | Dissolved | ug/L | 60.68 | aquatic biota | US EPA Aquatic Lit Criteria |
| Nitrate | Dissolved | $\mathrm{mg/L}$ | 3 | aquatic biota | CCME Water PAL AEP Water PAL |
| Nitrilotriacetic acid | | ug/L | 200 | human | WHO DW |
| Nitrite | Dissolved | $\mathrm{mg/L}$ | 0.06 | aquatic biota | CCME Water PAL |
| Nitrobenzene | | ug/L | 9.72 | human | HH DW+Org (derived) |
| Nitrosamines | | ug/L | 0.008 | human | HH DW+Org (US EPA) |
| Nitrosodibutylamine | | ug/L | 0.05 | human | HH DW+Org (derived) |
| Nitrosodiethylamine | | ug/L | 0.002 | human | HH DW+Org (derived) |
| Nitrosopyrrolidine | | ug/L | 0.16 | human | HH DW+Org (US EPA) HH DW+Org (derived) |
| Nonylphenol | | ug/L | 6.6 | aquatic biota | US EPA Aquatic Lis Criteria |
| | | | | | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|---|--------------------|-----------------|------------------------------------|---------------------------------|---|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Nonylphenol and its ethoxylates | | ug/L | 1 | aquatic biota | CCME Water PAL | |
| o-Dichlorobenzene | | ug/L | 0.7 | aquatic biota | AEP Water PAL CCME Water PAL | |
| octaBDE | | ng/L | 14 | aquatic biota | FEQG Water PAL | |
| Oxamyl (Vydate) | | ug/L | 200 | human | US EPA DWR | |
| p,p - Dichlorodiphenyldichlo (DDD)* | roethane | $\mathrm{ug/L}$ | 0.001 | human | HH DW+Org (US EPA) | |
| p,p - Dichlorodiphenyldichlo (DDE)* | rc | $\mathrm{ug/L}$ | 0.00018 | human | USEPA WQC HH Or | |
| p-Dichlorobenzene | | ug/L | 5 | human | Health Canada DW | |
| Paraquat | | ug/L | 10 | human | Health Canada DW | |
| Parathion | | $\mathrm{ug/L}$ | 0.013 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | |
| Pendimethalin | | ug/L | 20 | human | WHO DW | |
| pentaBDE | | ng/L | 0.2 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| pentaBDE (BDE-100) | | ng/L | 0.2 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| pentaBDE (BDE-99) | | ng/L | 4 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| Pentachlorobenzene | | ug/L | 0.01 | human | HH DW+Org (derived) | |
| Pentachloronitroben- zene | | | 4 | wildlife | US DOE Wildlife | |
| Pentachlorophenol | | ug/L | 0.1 | human | HH DW+Org (derived) | |
| Perchlorate | | ug/L | 70 | human | WHO DW | |
| Perfluorooctanesul- fonate | | ug/L | 0.6 | human | Health Canada DW | |
| Perfluorooctanoic acid | | ug/L | 0.2 | human | Health Canada DW | |
| Permethrin | | ug/L | 0.004 | aquatic biota | AEP Water PAL CCME Water PAL | |
| рН | | pH units | 7-9 | aquatic biota human human | US EPA Aquatic Life Criteria HH DW+Org (US EPA) AEP Water PAL CCME Water PAL Health Canada DW | |
| Phenanthrene [§] | | ug/L | 0.4 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Phenol | | $\mathrm{ug/L}$ | 2 | wildlife | CCME Water Ag (limited) AEP Water Ag | |
| Phorate | | ug/L | 2 | human | Health Canada DW | |
| Picloram | | ug/L | 29 | aquatic biota | CCME Water PAL AEP Water PAL | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|--------------------------------------|--------------------|--------------------|------------------------------------|--------------------|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Polychlorinated Biphenyls (PCBs)* | | ug/L | 0.00064 | human | USEPA WQC HH Org | |
| Propylene glycol | | ug/L | 500000 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Pyrene [§] | | ug/L | 0.025 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Quinoline | | ug/L | 3.4 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Selenium | Total | ug/L | 0.24 | wildlife | US DOE Wildlife | |
| Silver | Total | ug/L | 0.25 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Simazine | | $\mathrm{ug/L}$ | 2 | human | WHO DW | |
| Sodium dichloroisocyanurate | | ug/L | 40000 | human | WHO DW | |
| Solids Dissolved and Salinity | | ug/L | 250000 | human | HH DW+Org (US EPA) | |
| Strontium | Total | $\mathrm{ug/L}$ | 4000 | human | HH DW+Org (derived) | |
| Styrene | | ug/L | 20 | human | WHO DW | |
| Sulfate | | $\mathrm{mg/L}$ | 250 | human | WHO DW | |
| Sulfide | | $\mathrm{mg/L}$ | 0.0019 | aquatic biota | AEP Water PAL | |
| Sulfolane | | ug/L | 50 | aquatic biota | AEP Water PAL | |
| Tebuthiuron | | ug/L | 1.6 | aquatic biota | CCME Water PAL | |
| Terbufos | | ug/L | 1 | human | Health Canada DW | |
| Terbuthylazine | | ug/L | 7 | human | WHO DW | |
| tetraBDE | | ng/L | 24 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Tetrabromobisphenol A | | ug/L | 3.1 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Tetrachloroethane | | ug/L | 13.3 | aquatic biota | CCME Water PAL | |
| ${\it Tetrachloroethylene}^*$ | | $\mathrm{ug/L}$ | 4.48 | human | HH DW+Org (derived) | |
| Tetrachlorophenol | | ug/L | 1 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Thallium | Total | ug/L | 0.02 | human | HH DW+Org (derived) | |
| Toluene | | ug/L | 0.5 | aquatic biota | AEP Water PAL | |
| Total dissolved solids | | $\mathrm{mg/L}$ | 3000 | wildlife | AEP Water Ag CCME Water Ag (limited) | |
| Toxaphene | | ug/L | 0.0002 | aquatic biota | US EPA Aquatic Life Criteria | |
| Toxicity (acute) ^{††*} | | Toxic Units (a) | 0.3 | aquatic biota | AEP Water PAL | |
| Toxicity (chronic) ^{††**} | | Toxic Units (c) | 1 | aquatic biota | AEP Water PAL | |
| trans-1,2- Dichloroethylene | | $\mathrm{ug/L}$ | 100 | human | US EPA DWR | |
| | | | | | | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. (continued)

| | | | Generic (All water uses protected) | | | |
|--------------------------------|--------------------|-----------------|------------------------------------|--------------------|--|--|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source | |
| Triallate | | $\mathrm{ug/L}$ | 0.24 | aquatic biota | CCME Water PAL AEP Water PAL | |
| triBDE | | ng/L | 46 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| Tribromomethane | | $\mathrm{ug/L}$ | 100 | wildlife | CCME Water Ag (limited) | |
| Tributyltin | | ug/L | 0.008 | aquatic biota | CCME Water PAL | |
| Trichlorfon | | ug/L | 0.009 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Trichloroacetate | | ug/L | 200 | human | WHO DW | |
| ${\it Trichloroethylene}^*$ | | $\mathrm{ug/L}$ | 1.38 | human | HH DW+Org (derived) | |
| Trichlorophenol | | $\mathrm{ug/L}$ | 18 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Triclosan | | ug/L | 0.47 | aquatic biota | FEQG Water PAL | |
| Tricyclohexyltin | | ug/L | 250 | wildlife | CCME Water Ag (limited) AEP Water Ag | |
| Triethylene glycol | | ug/L | 350000 | aquatic biota | AEP Water PAL | |
| Trifluralin | | $\mathrm{ug/L}$ | 0.2 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Trihalomethanes | | ug/L | 80 | human | US EPA DWR | |
| Triphenyltin | | $\mathrm{ug/L}$ | 0.022 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Uranium | Total | $\mathrm{ug/L}$ | 15 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Vanadium | Total | ug/L | 100 | wildlife | AEP Water Ag CCME Water Ag (limited) | |
| Vinyl chloride* | | $\mathrm{ug/L}$ | 0.18 | human | HH DW+Org (derived) | |
| Xylene | | ug/L | 28 | wildlife | US DOE Wildlife | |
| Xylenes (total) | | ug/L | 10000 | human | US EPA DWR | |
| Zinc^{\ddagger} | Total | $\mathrm{ug/L}$ | 12.72 | human | HH DW+Org (derived) | |
| Zinc^{\ddagger} | Dissolved | $\mathrm{ug/L}$ | 31.35 | aquatic biota | CCME Water PAL | |
| Low Moelcular Weight PAHs \P | | | | | | |

Table 1.2: Generic health risk criteria for the protection of all Indigenous water use categories. *(continued)*

| | | | | Generic (All water use | es protected) |
|-----------|--------------------|-------|-------------------|------------------------|---------------|
| Parameter | Sample Fraction | Units | Most Stringent | Sensitive Receptor | Source |

 $\begin{array}{l} {\rm High~Molecular} \\ {\rm Weight~PAHs}^{**} \end{array}$

Note:

HH DW + Org and Org were adjusted to reflect carcinogenity of 1 in 1000,000 (1 x 10^{-5}) ILCR levels (Alberta Health (2019))

HH DW+Org: Human Health (HH) criteria from consuming surface water (SW) and aquatic organisms (O)

AO; Aesthetic Objectives, DW; Drinking Water; PAL; Protection of Aquatic Life, Ag; Agriculture Aquatic biota; invertebrates, plants and fish

Wildlife; bird and mammalian species

* Known human carcinogen via oral exposure route (Health Canada (2021))

- † The following known human carcinogens and must be converted to Provisional Benzo[a]pyrene RPF and summed as per Health Canada (2021) then compared to the Benzo(a)pyrene and equivalents health risk criteria: Anthanthrene, Benzo[c]chrysene, Benzo[g]chrysene, Benzo[c]phenanthrene, Cyclopenta[c,d]pyrene, Dibenzo[a,e]fluoranthene Dibenzo[a,e]pyrene, Dibenzo[a,h]pyrene, Dibenzo[a,i]pyrene, Dibenzo[a,l]pyrene, 9,10- Dimethylanthracene, 7,12- Dimethylbenzo[a]anthracene, 1,2- Dimethylbenzo[a]pyrene, 1,6- Dimethylbenzo[a]pyrene, 3,6- Dimethylbenzo[a]pyrene, 4,5- Dimethylbenzo[a]pyrene, 5,6- Dimethylchrysene, 5,7- Dimethylchrysene, 5,11- Dimethylchrysene, 1,4- Dimethylphenanthrene, 4,10- Dimethylphenanthrene, 5- Ethylchrysene, Fluoranthene, 7- Methylbenzo[a]anthracene, Methylbenzo[a]anthracene, 11- Methylbenzo[a]anthracene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, 12- Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, 12- Methylbenzo[a]pyrene, M
- [‡] Calculated using modifying factors presented in Table 3.1.
- § Sum identified LMW PAH congeners (Anthracene, Acenaphthene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) and compare to Naphthalene health risk criteria (adopted as surrogate) (CCME (2010))
- ¶ Sum identified LMW PAH congeners (Anthracene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) (CCME (2010))
- ** Sum of identified HMW PAH congeners (Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene) (CCME (2010))
- †† Comparison of water quality data must be presented for both Dissolved and total fractions
- ^{‡‡} Toxic Unit-Acute (TUa) is the reciprocal of the effluent concentration (i.e., TUa = 100/LC50) that causes 50 percent of the organisms to die by the end of an acute toxicity test (US EPA (2000c))
- §§ Toxic Unit-Chronic (TUc) is the reciprocal of the effluent concentration (e.g., TUc = 100/NOEC) that causes no observable effect (NOEC) on the test organisms by the end of a chronic toxicity test (US EPA (2000c)).

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Table 1.3: Risk based sediment quality criteria for the protection of Indigenous use.

| Parameter | Alberta ISQG (mg/kg) | SQC (mg/kg) | Source |
|--|------------------------|--|--|
| letals | | | |
| Arsenic* | 5.9 | 4.1 | Quebec (DSEE)-REL |
| Cadmium | _ | 0.33 | Quebec (DSEE)-REL |
| Chromium (total) | 37.3 | 25 | Quebec (DSEE)-REL |
| Copper | 35.7 | 8.6 | SST Benchmark Approach (Derived) |
| Lead | 35 | 25 | Quebec (DSEE)-REL |
| Manganese | _ | 460 | Ontario (OMOE) LEL |
| Mercury | 0.17 | 0.094 | Quebec (DSEE)-REL |
| Molybdenum | _ | 718 | SST Benchmark Approach (Derived) |
| Nickel | _ | 16 | Ontario (OMOEE) - LEL |
| Selenium | 2 | 2 | Alberta ISQG |
| Silver | _ | 0.57 | Washington WSDOE |
| Thallium | _ | 0.86 | Health Canada (2020) |
| Uranium | _ | 0.594 | SST Benchmark Approach (Derived) |
| Vanadium | _ | 125 | SST Benchmark Approach (Derived) |
| Zinc | 123 | 7.4 | SST Benchmark Approach (Derived) |
| olycyclic Aromatic Hydroc Low MW PAHs | arbons — | 0.552 | US EPA (OSWER)-ER-L |
| High MW PAHs | _ | 0.655 | US EPA (Region IV - FDEP)-TEL |
| Total PAHs | _ | 1.684 | US EPA (Region IV - FDEP)-TEL |
| Acenaphthene | 0.00671 | 0.0037 | Quebec (DSEE)-REL |
| Acenaphthylene | 0.00587 | 0.0033 | Quebec (DSEE)-REL |
| Anthracene | 0.0469 | 0.0087 | US DOE-EqP secondary |
| Benz[a]anthracene* | 0.0317 | 0.0079 | Derived EqP fish tissue, carcinogenicity |
| Benzo[a]pyrene* | 0.0319 | 6e-04 | Derived EqP fish tissue, carcinogenicity |
| Chrysene* | | | |
| | 0.0571 | 0.079 | Derived EqP fish tissue, carcinogenicity |
| Dibenz[a,h]anthracene* | 0.0571 | 0.079 0.00062 | Derived EqP fish tissue, carcinogenicity Derived EqP fish tissue, carcinogenicity |
| • | | | |
| Dibenz[a,h]anthracene* Fluoranthene Fluorene | _ | 0.00062 | Derived EqP fish tissue, carcinogenicity |
| Dibenz[a,h]anthracene* Fluoranthene | — 0.111 | $0.00062 \\ 0.047$ | Derived EqP fish tissue, carcinogenicity Quebec (DSEE)-REL |
| Dibenz[a,h]anthracene* Fluoranthene Fluorene | — 0.111 | 0.00062 0.047 0.01 | Derived EqP fish tissue, carcinogenicity Quebec (DSEE)-REL Quebec (DSEE)-OEL |
| Dibenz[a,h]anthracene* Fluoranthene Fluorene 2-Methylnaphthalene | — 0.111 | 0.00062 0.047 0.01 0.016 | Derived EqP fish tissue, carcinogenicity Quebec (DSEE)-REL Quebec (DSEE)-OEL Quebec (DSEE)-REL |
| Dibenz[a,h]anthracene* Fluoranthene Fluorene 2-Methylnaphthalene Naphthalene | — 0.111 | 0.00062 0.047 0.01 0.016 0.017 | Derived EqP fish tissue, carcinogenicity Quebec (DSEE)-REL Quebec (DSEE)-OEL Quebec (DSEE)-REL Quebec (DSEE)-REL |

Table 1.3: Risk based sediment quality criteria for the protection of Indigenous use. (continued)

| Parameter | Alberta ISQG (mg/kg) | SQC (mg/kg) | Source |
|-----------|----------------------|-------------|----------------------------------|
| Phenols | _ | 0.23 | Derived EqP fish tissue tainting |

Note:

Sum identified LMW PAH congeners (Anthracene, Acenaphthene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) (CCME (2010))

Sum of identified HMW PAH congeners (Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene) (CCME (2010))

^{*} Denotes carcinogenic substance

The following sections provide illustrations of how the health risk criteria and current con-376 ditions may be applied by users to assess potential health risks and changes in environmental conditions. Other applications, not discussed here, may include assessing risks to the envi-378 ronment and Indigenous land users from contaminants in treated tailings deposits used to create closure and reclamation landscapes, assessments of oil sands project applications (and 380 amendments), and oilsands mine water effluent releases to the ambient environment. 381

1.4 **Current Conditions**

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Existing, accessible water and sediment quality data collected through various monitoring 383 and research programs in the lower Athabasca River, the Athabasca River Delta and Lake 384 Athabasca were used to determine the current condition in monitored water and sediment 385 quality parameters (see Chapter 2 of this report). Specifically, normal (i.e., median) and 386 unusually low or high (i.e., 5th and 95th percentiles) values for these parameters were calculated 387 for the high flow, open water and under ice seasons (water) and annually (sediment) in the 388 River, Delta and Lake. The data used to define these current conditions were obtained between 389 390 2011 and 2020, except for sediment quality in the Delta where data obtained between 2000 and 2016. 391

Current State: Comparison of Current Conditions to Health 1.4.1 Risk Criteria

The following section provides an overview of the state of the Lower Athabasca River, 395 Athabasca Lake and Athabasca River Delta by comparing the current conditions to the health risk criteria established in Chapters 3 to 4 of this study. 396

Specific reference has been made to whether a chemical parameter exceeding the proposed health risk criteria is a known human carcinogen or not. This is an important component of the health risk criteria which addresses provincial gaps in the assessment of surface water and sediment quality (that do not currently include humans as a receptor and therefore have excluded an assessment of potential carcinogenicity) and directly addresses concerns around elevated cancer rates which ACFN, FMFN, and MCFN members have identified (McLachlan, 2014), and which led to the 2009 and 2014 investigations by researchers (Eggertson, 2009; Colquhoun et al., 2010) and Alberta Health (ACB, 2009; Chen, 2009; Services, 2014).

The comparison presented below is an illustration of how the health risk criteria are intended to be applied to surface water and sediment quality data and provides a preliminary assessment

- 407 of the current condition of water and sediment quality in the LAR, ARD, and Lake Athabasca.
- 408 Exceedances of the current condition 50th percentile values means that ambient conditions are
- 409 exceeding the health risk criteria about half of the time in a given season. It is important to
- 410 note that exceedances for other constituents may also be occurring, but less frequently, and
- 411 the comprehensive current condition tables presented in Table 2 can be used together with
- 412 the health risk criteria to, for example, determine whether exceedances are occurring but less
- 413 frequently (compare to 95th percentile), or even more frequently than half of the time (compare
- 414 to 5th percentile).
- The results presented below are an indication of potential risk drivers but have not been as-
- 416 sessed to understand health risks, sources of contaminants (i.e., oilsands development, natural),
- 417 or changes over time.
- The information therefore has limitations which must be addressed through follow up stud-
- 419 ies to understand potential health risks to community members, fish, and wildlife and to under-
- 420 stand how oil sands development and other sources have contributed (or not) to contaminants
- 421 in the LAR, ARD, and Lake Athabasca.

422 1.4.1.1 Athabasca River – Water Quality

- 423 The concentrations of most constituents of concern related to oil sands mining and natural oil
- 424 sands deposits are lower than the generic health risk criteria identified for each parameter (see
- 425 Table 1.4), with some exceptions discussed below.
- 426 Most of the current condition median values for PAHs with applicable health risk criteria
- 427 were not measured above detection limit in the river, and none of these exceeded the calculated
- 428 health risk criteria.
- 429 The majority of health risk criteria exceedances were related to metal concentrations with a
- 430 higher frequency of exceedances noted for total fractions compared to dissolved, and during high
- 431 flow time periods compared to periods of open water and under ice (see Table 1.4). Dissolved
- 432 copper was an exception, with consistent exceedances of the health risk criteria in all seasons.
- 433 In addition, it should be noted that all dissolved arsenic and cadmium concentrations exceed
- 434 the health risk criteria for the corresponding total fraction, which results from the guideline
- 435 development process discussed in Section 3.3.3 of this report. Importantly, for both arsenic
- 436 and cadmium, median dissolved fraction concentrations represent approximately one third to
- 437 one half of the median total fraction concentration. Similarly, median dissolved fractions of
- 438 copper exceed generic health risk criteria under all flow conditions and represent a significant
- 439 fraction of the median total fraction.

The median total arsenic, cadmium, iron and mercury concentrations exceed the generic 440 health risk criteria in all seasons. The consistency of these exceedances indicates a year-round 441 source(s) of these elements to the river, although all three have highest median concentrations 442 in the high flow season. 443 Median concentrations of other metals in river water exceed the generic health risk criteria 444 only during high flow conditions (i.e., total cobalt, copper, manganese, mercury, thallium, 445 zinc), while total aluminum exceeds the generic health risk criteria during both the high flow 446 and open water seasons. 447 These exceedances are likely related to the increased loads of trace elements that are bound 448 to suspended sediments and particles that are carried in Athabasca River water during spring 449 runoff and snow melt. Such particles can be contributed by erosion and sedimentation from 450 catchments, including both undisturbed areas and areas impacted by human development. 451 452 However, since dissolved arsenic and cadmium concentrations also consistently exceed the total fraction health risk criteria, it is unlikely that association with suspended particles are 453 the only, or even dominant, control over concentrations of these two elements in the river. 454 Since current conditions indicate elevated concentrations (i.e., exceedances of health risk 455 criteria) of some trace elements and historically members of ACFN, FMFN and MCFN con-456 sume untreated drinking water from the Lower Athabasca Region, additional studies are rec-457 ommended to more comprehensively assess how the identified exceedances could affect human, 458 aquatic biota and wildlife species health. Also, management of oil sands releases of these con-459 taminants may be required to mitigate potential risks from the elevated condition currently 460

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identified in the Athabasca River.

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Table 1.4: Comparison of health risk criteria to current conditions (Athabasca River).

| | | Generic health risk criteria (All water uses protected) | | | ected) Current Condition | | | |
|---|-----------------|---|--|---------------------------------|--------------------------|-----------------|----------------|--|
| Parameter | Unit | Helath Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50tl | |
| Conventional Variables | | | | | | | | |
| Alkalinity, total as CaCO3 | $\mathrm{mg/L}$ | 20.00 | AEP Water PAL US EPA Aquatic Life Criteria | aquatic biota | 89.00 | 101.00 | 163.00 | |
| Dissolved Metals | | | | | | | | |
| Aluminum, Filtered | ug/L | 50.00 | AEP Water PAL | aquatic biota | 32.35 | 16.00 | 13.20 | |
| Arsenic, Filtered * | ug/L | 150.00 | US EPA Aquatic Life Criteria | aquatic biota | 0.55 | 0.49 | 0.46 | |
| Cadmium, Filtered * | ug/L | 0.82 | US EPA Aquatic Life Criteria | aquatic biota | 0.011 | 0.010 | 0.015 | |
| Copper, Filtered | ug/L | 0.53 | FEQG Water PAL | aquatic biota | 1.28 | 0.66 | 0.58 | |
| Iron, Filtered | ug/L | 300.00 | AEP Water PAL | aquatic biota | 190.50 | 157.00 | 255.00 | |
| Lead, Filtered | ug/L | 3.07 | US EPA Aquatic Life Criteria | aquatic biota | 0.089 | 0.039 | 0.032 | |
| Nickel, Filtered | ug/L | 60.68 | US EPA Aquatic Life Criteria | aquatic biota | 1.38 | 0.91 | 0.94 | |
| Zinc, Filtered | ug/L | 31.35 | CCME Water PAL | aquatic biota | 0.60 | 0.40 | 1.30 | |
| Field | | | US EPA Aquatic Life Criteria | | | | | |
| рН | pH units | 7-9 | HH DW+Org (US EPA) AEP Water PAL CCME Water PAL Health Canada DW | aquatic biota human human | 7.97 | 8.20 | 7.52 | |
| General Organics | | | | | | | | |
| Toluene | ug/L | 0.50 | AEP Water PAL | aquatic biota | • | 0.031 | • | |
| Nutrients and BOD Ammonia and ammonium, Unfiltered as N | $\mathrm{mg/L}$ | 0.67 | HH DW+Org (derived) | human | 0.011 | 0.0080 | 0.048 | |
| PAHs | | | | | | | | |
| Chrysene | $_{ m ng/L}$ | 70.00 | HH DW+Org (derived) | human | 2.51 | • | • | |
| Fluoranthene | ng/L | 40.00 | AEP Water PAL CCME Water PAL | aquatic biota | 2.14 | • | • | |
| Naphthalene | ng/L | 1000.00 | AEP Water PAL | aquatic biota | 23.78 | 43.05 | 26.65 | |
| Phenanthrene | ng/L | 400.00 | CCME Water PAL AEP Water PAL | aquatic biota | 10.64 | • | • | |
| Pyrene | $\mathrm{ng/L}$ | 25.00 | CCME Water PAL AEP Water PAL | aquatic biota | 3.34 | • | • | |
| Total Metals | | | | | | | | |
| Aluminum, Unfiltered | ug/L | 18.00 | US DOE Wildlife | wildlife | 2530.00 | 316.00 | 54.00 | |
| Antimony, Unfiltered | ug/L | 4.59 | HH DW+Org (derived) | human | 0.11 | 0.060 | 0.056 | |
| Arsenic, Unfiltered | ug/L | 0.030 | HH DW+Org (derived) | human | 1.98 | 0.71 | 0.56 | |
| Barium, Unfiltered | ug/L | 1000.00 | HH DW+Org (US EPA) Health Canada DW | human | 73.80 | 53.70 | 85.20 | |

Table 1.4: Comparison of health risk criteria to current conditions (Athabasca River). (continued)

| | | Generic health risk criteria (All water uses protected) | | | Current Condition | | |
|----------------------------------|------|---|---------------------------------|------------------------|-------------------|-----------------|----------------|
| Parameter | Unit | Helath Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50tl |
| Beryllium, Unfiltered | ug/L | 3.27 | HH DW+Org (derived) | human | 0.14 | 0.020 | 0.0070 |
| Boron, Unfiltered | ug/L | 1333.33 | HH DW+Org (derived) | human | 25.30 | 23.60 | 36.40 |
| Cadmium, Unfiltered | ug/L | 0.0020 | HH DW+Org (derived) | human | 0.050 | 0.017 | 0.016 |
| Chromium, Unfiltered | ug/L | 50.00 | WHO DW Health Canada DW | human | 3.56 | 0.45 | 0.18 |
| Cobalt, Unfiltered | ug/L | 1.10 | FEQG Water PAL AEP Water PAL | aquatic biota | 1.65 | 0.27 | 0.09 |
| Copper, Unfiltered | ug/L | 2.76 | CCME Water PAL | aquatic biota | 4.40 | 0.91 | 0.66 |
| Iron, Unfiltered | ug/L | 300.00 | CCME Water PAL USEPA WQC AO | aquatic biota human | 4290.00 | 709.00 | 430.50 |
| Lead, Unfiltered | ug/L | 4.01 | AEP Water PAL CCME Water PAL | aquatic biota | 2.15 | 0.27 | 0.09 |
| Manganese, Unfiltered | ug/L | 50.00 | HH DW+Org (US EPA) | human | 114.00 | 38.50 | 15.85 |
| Mercury, Unfiltered | ng/L | 1.58 | US DOE Wildlife | wildlife | 10.00 | 1.90 | 0.68 |
| Methylmercury(1+), Unfiltered | ng/L | 1.00 | AEP Water PAL | aquatic biota | 0.18 | 0.060 | 0.037 |
| Molybdenum, Unfiltered | ug/L | 33.33 | HH DW+Org (derived) | human | 0.75 | 0.73 | 0.90 |
| Nickel, Unfiltered | ug/L | 7.35 | HH DW+Org (derived) | human | 5.23 | 1.32 | 1.03 |
| Selenium, Unfiltered | ug/L | 0.24 | US DOE Wildlife | wildlife | 0.22 | 0.14 | 0.21 |
| Silver, Unfiltered | ug/L | 0.25 | AEP Water PAL CCME Water PAL | aquatic biota | 0.023 | 0.0040 | 0.0020 |
| Strontium, Unfiltered | ug/L | 4000.00 | HH DW+Org (derived) | human | 214.00 | 223.00 | 352.00 |
| Thallium, Unfiltered | ug/L | 0.020 | HH DW+Org (derived) | human | 0.053 | 0.010 | 0.0050 |
| Uranium, Unfiltered | ug/L | 15.00 | CCME Water PAL AEP Water PAL | aquatic biota | 0.45 | 0.37 | 0.57 |
| Vanadium, Unfiltered | ug/L | 100.00 | AEP Water Ag CCME Water Ag | wildlife | 6.92 | 1.07 | 0.36 |
| Zinc, Unfiltered | ug/L | 12.72 | HH DW+Org (derived) | human | 13.10 | 2.00 | 1.85 |

Note:

Refer to Tables 1.2 and 1.3 for health risk criteria calculation methods

Bolded values indicate exceedances of the corresponding water quality criteria for Indigenous use

Where under-ice conditions were calculated for individual sites (not merged), the maximum value across those sites is displayed

^{*} Dissolved current condition concentrations exceed health risk criteria for total fraction. See discussion in Section 3.3.3

Table 1.5: Comparison of health risk criteria for carcinogenic (BaP and equivalents) and non-carcinogenic (Naphthalene and equivalents) polycyclic aromatic hydrocarbon (PAH) congeners to current conditions (Athabasca River)

| | | Current Condition | | | |
|-------------------------------|-----------------|---------------------------------|-------------------|--------------------|-------------------|
| Parameter | Unit | Generic health risk criteria | High Flow 50th | Open Water 50th | Under Ice 50th |
| BaP (and equivalents) | ug/L | 0.0001 | 0.00011 | 0.00000 | 0.00000 |
| Naphthalene (and equivalents) | $\mathrm{ug/L}$ | 1.0000 | 0.02078 | 0.02078 | 0.02078 |

^{*} Known human carcinogens must be converted to provisional Benzo[a]pyrene RPF and summed (Health Canada (2021))

462 1.4.1.2 Athabasca River – Sediment

The median current condition sediment concentrations in the River exceeded the generic health risk criteria for sediment (also referred to as the SQC) for manganese, uranium and zinc and the carcinogenic substances benzo(a)pyrene, dibenz[a,h]anthracene, and arsenic (see Table 1.6 below).

Table 1.6: Comparison of Indigenous use Sediment Quality Criteria to current conditions (Athabasca River).

| Parameter | Unit | Health Risk Criteria | Annual 50th |
|-----------------------|-----------------------------------|-------------------------|-------------|
| General Organics | | | |
| Naphthenic acids | ug/g | 3.30 | 136.50 |
| PAHs | | | |
| 2-Methylnaphthalene | ng/g | 16.00 | 10.98 |
| Acenaphthene | ng/g | 3.70 | 0.70 |
| Anthracene | ng/g | 8.70 | 0.61 |
| Benz[a]anthracene | ng/g | 7.85 | 2.82 |
| Benzo[a]pyrene | ng/g | 0.62 | 4.05 |
| Chrysene | ng/g | 26.00 | 12.60 |
| Dibenz[a,h]anthracene | ng/g | 0.62 | 1.69 |
| Fluoranthene | ng/g | 47.00 | 3.43 |
| Fluorene | ng/g | 10.00 | 1.24 |
| Naphthalene | ng/g | 17.00 | 4.00 |
| Phenanthrene | ng/g | 25.00 | 11.10 |
| Pyrene | ng/g | 29.00 | 6.85 |
| Total Metals | | | |
| Arsenic | $\mathbf{u}\mathbf{g}/\mathbf{g}$ | 4.10 | 4.21 |
| Cadmium | ug/g | 0.33 | 0.14 |
| Chromium | ug/g | 25.00 | 10.90 |
| Copper | ug/g | 8.60 | 6.75 |
| Lead | ug/g | 11.00 | 5.34 |

[†] Sum identified LMW PAH congeners (Anthracene, Acenaphthene, Acenaphthylene, Fluorene, Naphthalene, Phenanthrene, Pyrene) and compare to Naphthalene health risk criteria (adopted as surrogate) (CCME (2010))

Table 1.6: Comparison of Indigenous use Sediment Quality Criteria to current conditions (Athabasca River). (continued)

| Parameter | Unit | Health Risk Criteria | Annual 50th |
|------------|------|-------------------------|-------------|
| Manganese | ug/g | 28.00 | 289.00 |
| Molybdenum | ug/g | 718.00 | 0.44 |
| Nickel | ug/g | 16.00 | 13.30 |
| Silver | ug/g | 0.57 | 0.05 |
| Thallium | ug/g | 0.86 | 0.10 |
| Uranium | ug/g | 0.59 | 0.67 |
| Vanadium | ug/g | 125.00 | 17.10 |
| Zinc | ug/g | 7.40 | 39.90 |

Note:

Refer to Tables 1.2 and 1.3 for health risk criteria calculation methods Bolded rows indicate exceedances of the corresponding water quality criteria for Indigenous use

Comparison of the sum of median annual concentrations of low and high molecular weight and total PAH groupings to the respective SQC proposed for each group indicates that exceedances are unlikely using this "average" measure of sediment quality in the Athabasca River (see Table 1.7). The high MW group includes the known carcinogenic PAHs.

Table 1.7: Comparison of median concentrations (ng/g) of PAH groups (high and low molecular weight; total PAHs) measured in the Athabasca River to proposed sediment health risk criteria.

| | High MW PAH | Low MW PAH | Total PAH |
|----------------|-------------|------------|-----------|
| River | 33 | 39 | 72 |
| SQC - sediment | 655 | 552 | 1,684 |

Note:

High MW PAHs and carcinogens Sum of 50%ile for Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene)

Low MW PAHs Sum of 50%ile for Acenaphthene, Acenaphthylene, Anthracene, Fluoranthene, Fluorene, 2-methylnapthalene, Naphthalene, Phenanthrene, Pyrene

The sediment health risk criteria (also referred to as SQCs) were developed to consider the protection of sediment associated biota from direct exposure and exposure through consuming diet items from the bioaccumulation of these contaminants within aquatic food webs. Comparison of these SQC with the current condition in the Athabasca River Table 1.6 indicate that there may be risks to sediment associated biota from exposure to PAHs and certain metals as well as risks of exposure through ingestion of aquatic biota, however, additional studies are required to better understand the risk potential and what management actions could be required.

Athabasca River Delta – Water 1.4.1.3

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Concentrations of chemical parameters appear to be elevated in the Athabasca River Delta 480 surface water compared to the river and Lake Athabasca. Like the river, median trace ele-481 482 ment concentrations measured in total fractions in the delta exceeded health risk criteria more frequently compared to dissolved fractions (see Table 1.8). However, as noted for the river, dis-483 solved arsenic and cadmium concentrations exceed the health risk criteria for the corresponding 484 485 total fraction (see dicussion in Section 3.3.3 of this report). Seasonal conditions did not appear to vary to the same extent as in the river, because exceedances were more frequently identified 486 in all seasons and for upper, median and lower values in each range (e.g., arsenic (carcinogenic 487 substance), cadmium and total iron, as well as chlorine). 488 489 Median concentrations of total mercury, cobalt, copper and thallium exceeded generic health risk criteria in the delta during high flow only, while median total aluminum and man-490 ganese exceeded during both high flow and open water. Notably, and in contrast to conditions 491 in the river, for many of these total metal parameters, the lower bound of their concentration 492 range also exceeded the generic health risk criteria. These patterns were not present for most of 493 the corresponding dissolved metals in delta water, indicating particle-associated fractions play 494 495 a significant role in these consistent exceedances. However, median concentrations of dissolved copper in all seasons exceeded the generic health risk criteria, indicating that relevant copper, 496 and arsenic and cadmium concentrations in water in the delta are not predominantly driven 497 498 by particle-associated fractions. The median concentration of the ion fluoride and the composite measure total dissolved 499 solids also exceeded the generic health risk criteria during the under ice season in the Delta. 500 501 This pattern generally indicates a lack of dilution power in these Delta channels during the winter, and the fluoride exceedance mirrors the elevated concentration in the River under ice. 502 503 The substantive number of chemical parameters exceeding the generic water quality health risk criteria indicates that there may be risks to community members, fish and wildlife consum-504 ing, interacting with, and ingesting aquatic biota within the ARD, however, a risk assessment 505 to verify potential health risk was beyond the scope of this study. 506 Future studies to address monitoring gaps (see Chapter 2), assess potential risks to human 507 and environmental health, and understand the contribution of oilsands development to the 508 current state of the Athbasca River Delta are recommended.

Table 1.8: Comparison of health riskcriteria to current conditions (Athabasca River Delta).

| | | Generic Health Risk Criteria | | | Current Condition | | |
|--|-----------------|------------------------------|---|---------------------------------|-------------------|-----------------|----------------|
| Parameter | Unit | Health Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50tl |
| Conventional Variables | | | | | | | |
| Alkalinity, total as CaCO3 | $\mathrm{mg/L}$ | 20.00 | AEP Water PAL US EPA Aquatic Life Criteria | aquatic biota | 89.00 | 110.00 | 140.00 |
| Total dissolved solids, Filtered | mg/L | 250.00 | HH DW+Org (US EPA) | human | 140.00 | 180.00 | 250.00 |
| Dissolved Metals | | | | | | | |
| Aluminum, Filtered | ug/L | 50.00 | AEP Water PAL | aquatic biota | 16.20 | 7.96 | 4.23 |
| Arsenic, Filtered * | ug/L | 150.00 | US EPA Aquatic Life Criteria | aquatic biota | 0.55 | 0.50 | 0.42 |
| Cadmium, Filtered * | ug/L | 0.82 | US EPA Aquatic Life Criteria | aquatic biota | 0.009 | 0.009 | 0.014 |
| Copper, Filtered | ug/L | 0.53 | FEQG Water PAL | aquatic biota | 1.56 | 0.97 | 0.75 |
| Iron, Filtered | ug/L | 300.00 | AEP Water PAL | aquatic biota | 121.50 | 95.00 | 178.00 |
| Lead. Filtered | ug/L | 3.07 | US EPA Aquatic Life Criteria | aquatic biota | 0.084 | 0.038 | 0.052 |
| Mercury, Filtered | ng/L | 770.00 | US EPA Aquatic Life Criteria | aquatic biota | 0.001 | • | 0.50 |
| Methylmercury(1+), Filtered | ng/L | 4.00 | CCME Water PAL | aquatic biota | 0.061 | 0.039 | 0.028 |
| Nickel, Filtered | ug/L | 60.68 | US EPA Aquatic Life Criteria | aquatic biota | 1.43 | 0.75 | 0.76 |
| Zinc, Filtered | ug/L | 31.35 | CCME Water PAL | aquatic biota | 0.62 | 0.53 | 1.58 |
| Field | | | | | | | |
| рН | pH units | 7-9 | US EPA Aquatic Life Criteria HH DW+Org (US EPA) AEP Water PAL CCME Water PAL Health Canada DW | aquatic biota human human | 7.89 | 8.00 | 7.44 |
| Major Ions | | | | | | | |
| Chloride, Unfiltered | $\mathrm{mg/L}$ | 120.00 | CCME Water PAL AEP Water PAL | aquatic biota | 6.00 | 12.00 | 25.00 |
| Fluoride, Unfiltered | mg/L | 0.12 | CCME Water PAL | aquatic biota | 0.10 | 0.10 | 0.12 |
| Sulfate, Unfiltered as SO4 | $\mathrm{mg/L}$ | 250.00 | WHO DW | human | 23.00 | 28.00 | 36.00 |
| Nutrients and BOD | | | | | | | |
| Ammonia and ammonium, Unfiltered as N | $\mathrm{mg/L}$ | 0.67 | HH DW+Org (derived) | human | • | 0.022 | 0.052 |
| Total Metals | | | | | | | |
| Mercury, Unfiltered | ng/L | 1.58 | US DOE Wildlife | wildlife | 8.90 | 2.99 | 0.82 |
| Methylmercury $(1+)$, Unfiltered | ng/L | 1.00 | AEP Water PAL | aquatic biota | 0.16 | 0.072 | 0.039 |
| Total Recoverable Metals | | | | | | | |
| Aluminum, Unfiltered | ug/L | 18.00 | US DOE Wildlife | wildlife | 2770.00 | 792.00 | 97.50 |
| Antimony, Unfiltered | ug/L | 4.59 | HH DW+Org (derived) | human | 0.10 | 0.065 | 0.052 |

Table 1.8: Comparison of health riskcriteria to current conditions (Athabasca River Delta). (continued)

| | | (| Generic Health Risk Criteria | | Current Condition | | |
|------------------------|------|----------------------|--|------------------------|-------------------|-----------------|---------------|
| Parameter | Unit | Health Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50t |
| Arsenic, Unfiltered | ug/L | 0.030 | HH DW+Org (derived) | human | 1.75 | 0.86 | 0.57 |
| Barium, Unfiltered | ug/L | 1000.00 | HH DW+Org (US EPA) Health Canada DW | human | 86.15 | 56.90 | 64.05 |
| Beryllium, Unfiltered | ug/L | 3.27 | HH DW+Org (derived) | human | 0.14 | 0.043 | 0.0080 |
| Boron, Unfiltered | ug/L | 1333.33 | HH DW+Org (derived) | human | 24.80 | 24.70 | 32.85 |
| Cadmium, Unfiltered | ug/L | 0.0020 | HH DW+Org (derived) | human | 0.058 | 0.020 | 0.020 |
| Chlorine, Unfiltered | mg/L | 0.00050 | AEP Water PAL | aquatic biota | 4.13 | 8.40 | 20.80 |
| Chromium, Unfiltered | ug/L | 50.00 | WHO DW Health Canada DW | human | 3.22 | 0.92 | 0.22 |
| Cobalt, Unfiltered | ug/L | 1.10 | FEQG Water PAL AEP Water PAL | aquatic biota | 1.36 | 0.41 | 0.12 |
| Copper, Unfiltered | ug/L | 2.76 | CCME Water PAL | aquatic biota | 3.65 | 1.42 | 0.91 |
| Iron, Unfiltered | ug/L | 300.00 | CCME Water PAL USEPA WQC AO | aquatic biota human | 4240.00 | 1050.00 | 565.50 |
| Lead, Unfiltered | ug/L | 4.01 | AEP Water PAL CCME Water PAL | aquatic biota | 2.13 | 0.47 | 0.16 |
| Manganese, Unfiltered | ug/L | 50.00 | HH DW+Org (US EPA) | human | 104.40 | 54.70 | 30.75 |
| Molybdenum, Unfiltered | ug/L | 33.33 | HH DW+Org (derived) | human | 0.52 | 0.60 | 0.65 |
| Nickel, Unfiltered | ug/L | 7.35 | HH DW+Org (derived) | human | 4.33 | 1.55 | 1.02 |
| Selenium, Unfiltered | ug/L | 0.24 | US DOE Wildlife | wildlife | 0.26 | 0.22 | 0.30 |
| Silver, Unfiltered | ug/L | 0.25 | AEP Water PAL CCME Water PAL | aquatic biota | 0.023 | 0.0060 | 0.0030 |
| Strontium, Unfiltered | ug/L | 4000.00 | HH DW+Org (derived) | human | 174.50 | 206.00 | 275.00 |
| Thallium, Unfiltered | ug/L | 0.020 | HH DW+Org (derived) | human | 0.048 | 0.016 | 0.0060 |
| Uranium, Unfiltered | ug/L | 15.00 | CCME Water PAL AEP Water PAL | aquatic biota | 0.49 | 0.41 | 0.44 |
| Vanadium, Unfiltered | ug/L | 100.00 | AEP Water Ag CCME Water Ag | wildlife | 6.73 | 2.04 | 0.43 |
| Zinc, Unfiltered | ug/L | 12.72 | HH DW+Org (derived) | human | 10.36 | 3.10 | 2.58 |

Note:

Refer to Tables 1.2 and 1.3 for health risk criteria calculation methods

Bolded values indicate exceedances of the corresponding water quality criteria for Indigenous use

Where under-ice conditions were calculated for individual sites (not merged), the maximum value across those sites is displayed

*Dissolved current condition concentrations exceed health risk criteria for total fraction. See discussion in Section 3.3.3

510 1.4.1.4 Athabasca River Delta – Sediment

In terms of sediment quality, the concentrations of trace elements, as well as PAHs in the 511 Athabasca River Delta sediment were relatively high compared to the lower Athabasca River. 512 This coincided with a higher median proportion of finer particles, specifically silt and clay, 513 in the delta sediments compared to the river sediments (see Table 1.9). This makes sense, 514 because these finer sediments are more likely to drop out of the water column in the relatively 515 lower-energy environment of delta channels compared to the river. Finer sediments are also 516 more likely to have these associated constituents compared to sand, which made up a larger 517 proportion of river sediment. 518

Table 1.9: Comparison of median small sediment particle size distributions measured in the Athabasca River and Athabasca River Delta.

| | % Clay* | % Silt [†] | % Sand [‡] |
|-------|---------|---------------------|---------------------|
| River | 7 | 19 | 72 |
| Delta | 16 | 48 | 34 |

 $^{^* &}lt; 2 \text{ um}$

Median sediment concentrations of the carcinogenic substances benzo(a)pyrene and arsenic exceeded the calculated health risk criteria for Indigenous use. Several other non-carcinogenic parameters also exceeded the generic health risk criteria under median conditions, specifically copper, manganese, nickel and zinc.

Table 1.10: Comparison of Indigenous use Sediment Quality Criteria to current conditions (Athabasca River Delta).

| Parameter | Unit | Health Risk Criteria | Annual 50th |
|----------------|--------------------------|-------------------------|-------------|
| PAHs | | | |
| Benzo[a]pyrene | m ng/g | 0.62 | 5.88 |
| Chrysene | ng/g | 26.00 | 17.75 |
| Fluoranthene | ng/g | 47.00 | 3.87 |
| Fluorene | ng/g | 10.00 | 2.30 |
| Naphthalene | ng/g | 17.00 | 7.75 |
| Phenanthrene | ng/g | 25.00 | 15.95 |
| Pyrene | ng/g | 29.00 | 10.45 |
| Total Metals | | | |
| Arsenic | ug/g | 4.10 | 4.95 |
| Chromium | ug/g | 25.00 | 14.95 |
| Copper | ${f ug/g}$ | 8.60 | 13.10 |
| Lead | ug/g | 11.00 | 7.90 |
| Manganese | ug/g | 28.00 | 392.00 |

 $^{^{\}dagger}$ > or = 2 um to < 63 um

 $^{^{\}ddagger} > \text{or} = 63 \text{ um to} < 2000 \text{ um}$

Table 1.10: Comparison of Indigenous use Sediment Quality Criteria to current conditions (Athabasca River Delta). (continued)

| Parameter | Unit | Health Risk Criteria | Annual 50th |
|---------------------------|--------------------------|-------------------------|-------------|
| Mercury | ug/g | 0.09 | 0.04 |
| Nickel | ug/g | 16.00 | 18.75 |
| Selenium | ug/g | 0.63 | 0.41 |
| $\operatorname{Thallium}$ | ug/g | 0.86 | 0.16 |
| Vanadium | ug/g | 125.00 | 21.70 |
| ${f Zinc}$ | ug/g | 7.40 | 59.35 |

Note:

Refer to Tables 1.2 and 1.3 for health risk criteria calculation methods Bolded rows indicate exceedances of the corresponding water quality criteria for Indigenous use

In addition, the PAH data available for the delta included far fewer parameters compared to PAH data from the river. Comparison of the sum of median annual concentrations of low and high molecular weight and total PAH groupings to the respective SQC proposed for each group indicates that exceedances are unlikely using this "average" measure of sediment quality in the Athabasca River Delta (see Table 1.11).

Table 1.11: Comparison of median concentrations (ng/g) of PAH groups (high and low molecular weight; total PAHs) measured in the Athabasca River Delta to proposed sediment health risk criteria.

| | High MW PAH | Low MW PAH | Total PAH | | |
|--|-------------|------------|-----------|--|--|
| River | 30 | 40 | 70 | | |
| SQC - sediment | 655 | 552 | 1,684 | | |
| Note: High MW PAHs and carcinogens Sum of 50 Low MW PAHs Sum of 50 | | | | | |

Given that several carcinogenic and noncarcinogenic parameters exceeded the most stringent (generic) health risk criteria for sediment using upper and lower ranges of the data, it is recommended that future studies on health risks and establishing contributions from oil sands development include an assessment and additional monitoring for chemical parameters in sediments (as recommended under the ARD water discussion).

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1.4.1.5 Lake Athabasca - Water

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The available water quality data for Lake Athabasca were more limited in terms of the number 534 of parameters and the number of observations in under ice and high flow seasons. There were 535 no sediment quality data available for Lake Athabasca. 536

Exceedances of health risk criteria in the lake were observed for total metal fractions under 537 open water conditions (see Table 1.12). Aluminum, arsenic (carcinogenic substance), and iron exceeded under median conditions and may present the most likely risk potential although 539 upper ranges of other total copper, manganese, nickel and zinc as well as total dissolved 540 solids exceeded health risk criteria (refer to Chapter 3 for complete current condition tables). 541

Dissolved metals data were not available for the lake. 542

543 It is important to recognize the community of Ft. Chipewyan has access to treated 544 Athabasca Lake water as a drinking water source and the concentrations of the above noted parameters may be decreased through the municipal water treatment process. It is unclear 545 546 to what degree ACFN, FMFN and MCFN members consume untreated water from Lake Athabasca and if there could be risks to community members, fish and wildlife from water 547 quality conditions reported here. It is recommended that a focused study to better understand 548 549 the results presented here be completed in the future.

Table 1.12: Comparison of health risk criteria to current conditions (Lake Athabasca).

| | | | Generic health risk criteria | Current Condition | | | | |
|---|-----------------|----------------------|---|---------------------------------|----------------|-----------------|----------------|--|
| Parameter | Unit | Health Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50th | |
| Conventional Variables Total dissolved solids, Filtered | ${ m mg/L}$ | 250.00 | HH DW+Org (US EPA) | human | ٠ | 57.00 | • | |
| Field | | | | | | | | |
| рН | pH units | 7-9 | US EPA Aquatic Life Criteria HH DW+Org (US EPA) AEP Water PAL CCME Water PAL Health Canada DW | aquatic biota human human | 8.22 | 8.13 | • | |
| Major Ions | | | | | | | | |
| Chloride, Unfiltered | mg/L | 120.00 | CCME Water PAL AEP Water PAL | aquatic biota | • | 3.70 | • | |
| Sulfate, Unfiltered as SO4 | $\mathrm{mg/L}$ | 250.00 | WHO DW | human | • | 6.00 | • | |
| Total Metals | | | | | | | | |
| Aluminum, Unfiltered | ug/L | 18.00 | US DOE Wildlife | wildlife | • | 591.00 | • | |
| Arsenic, Unfiltered | ug/L | 0.030 | HH DW+Org (derived) | human | • | 0.70 | • | |
| Barium, Unfiltered | $\mathrm{ug/L}$ | 1000.00 | HH DW+Org (US EPA) Health Canada DW | human | • | 29.90 | • | |
| Beryllium, Unfiltered | $_{ m ug/L}$ | 3.27 | HH DW+Org (derived) | human | • | 0.032 | • | |
| Chromium, Unfiltered | ug/L | 50.00 | WHO DW Health Canada DW | human | • | 0.90 | • | |
| Copper, Unfiltered | $_{ m ug/L}$ | 2.76 | CCME Water PAL | aquatic biota | • | 1.45 | • | |
| Iron, Unfiltered | ug/L | 300.00 | CCME Water PAL USEPA WQC AO | aquatic biota human | • | 953.00 | • | |
| Lead, Unfiltered | $\mathrm{ug/L}$ | 4.01 | AEP Water PAL CCME Water PAL | aquatic biota | • | 0.55 | • | |
| Manganese, Unfiltered | ug/L | 50.00 | HH DW+Org (US EPA) | human | • | 21.10 | • | |
| Molybdenum, Unfiltered | ug/L | 33.33 | HH DW+Org (derived) | human | • | 0.30 | • | |
| Nickel, Unfiltered | ug/L | 7.35 | HH DW+Org (derived) | human | • | 1.50 | • | |
| Vanadium, Unfiltered | $\mathrm{ug/L}$ | 100.00 | AEP Water Ag CCME Water Ag | wildlife | • | 1.90 | • | |

Table 1.12: Comparison of health risk criteria to current conditions (Lake Athabasca). (continued)

| | | Generic health risk criteria | | | | Current Condition | |
|-------------------------|------|------------------------------|---------------------|----------|----------------|-------------------|----------------|
| Parameter | Unit | Health Risk Criteria | Source | Receptor | High Flow 50th | Open Water 50th | Under Ice 50th |
| Zinc, Unfiltered | ug/L | 12.72 | HH DW+Org (derived) | human | • | 4.05 | • |
| Note: | - | | - , , | | | | |
| Refer to Tables 1.2 and | | | methods | | | | |

Bolded values indicate exceedances of the corresponding water quality criteria for Indigenous use

Where under-ice conditions were calculated for individual sites (not merged), the maximum value across those sites is displayed

550 1.4.2 Athabasca River Delta current condition - Comparison to 551 LARP Surface Water Quality Management Framework (trig552 gers)

There is another comparison that can be made with the Athabasca River Delta sites, which is with the current conditions calculated for the Lower Athabasca Regional Plan (LARP) Surface Water Quality Management Framework. Mean and peak (95th percentile) water quality triggers under LARP were calculated using data from the same sites used in this study. However, in the case of the development of LARP triggers, monitoring data from before 2009 were used whereas in this study, data from after 2011 were used to calculate current conditions (see Chapter 2).

A comparison between these values is provided in Table 1.13 below. Comparison of the current conditions to the LARP triggers indicates that the LARP annual mean values are often lower in value – generally meaning more conservative – than the high flow median current condition values calculated here, but are often higher in value – generally meaning less conservative – for the open water and under ice seasons.

LARP trigger values for dissolved beryllium, total boron, dissolved and total cadmium, and dissolved thallium are very high in comparison to this study's current conditions. Specifically, neither the median or 95th percentile values calculated in this study exceed the LARP trigger for these parameters (see bolded values in Table 1.13). In addition, the LARP trigger for ammonia is high compared to the current condition for high flow and open water, and LARP triggers for total phosphosurs and total dissolved phosphorus are high compared to current conditions for open water and under ice. These differences may reflect a change in Delta water quality since the LARP values were released using data obtained before 2009, since the data used to calculate the current condition were obtained after 2011. Alternatively, these differences may be related to the different statistical methods used in the LARP and this study's current condition calculation. Whatever the cause, these LARP triggers should be re-examined to ensure that they are statistically robust and that they are currently relevant to the lower Athabasca River.

The consequences of the lack of seasonal specificity in the calculated LARP triggers is particularly clear when comparing them to the seasonal current conditions, and it is recommended that LARP triggers are re-calculated using the seasonal approach. This would ensure that relevant and reasonable triggers are applied for the majority of the year (i.e., during open water and under ice) when concentrations are generally lower than the LARP triggers.

Table 1.13: Surface water quality triggers from the LARP Surface Water Quality Management Framework and seasonal current condition values calculated as part of this study for sites in the Athabasca River Delta. LARP values that appear to be an overestimate compared to the current condition values calculated in this study are bolded. Note that LARP central tendency measures are annual means, whereas this study used seasonal medians.

| | | LARP W | ater Quality Triggers | High | flow | Open | water | Under ice | |
|----------------------------|-------|---------|---------------------------|--------|--------|--------|--------|-----------|--------|
| Parameter Name | Units | Mean | Peak (95th percentile) | Median | 95%ile | Median | 95%ile | Median | 95%ile |
| Nutrients | | | | | | | | | |
| Total ammonia | mg/L | 0.05 | 0.12 | < | < | 0.022 | 0.08 | 0.052 | 0.096 |
| Nitrate | mg/L | 0.09 | 0.26 | 0.046 | 0.11 | - | - | 0.17 | 0.27 |
| Total nitrogen | mg/L | 0.60 | 1.04 | - | - | - | - | - | |
| Total dissolved phosphorus | mg/L | 0.02 | 0.03 | 0.014 | 0.027 | 0.008 | 0.018 | 0.013 | 0.019 |
| Total phosphorus | mg/L | 0.07 | 0.26 | 0.11 | 0.228 | 0.041 | 0.192 | 0.024 | 0.046 |
| ons | | | | | | | | | |
| Calcium | mg/L | 34.70 | 48.90 | 27.5 | 33.8 | 32.5 | 37.8 | 42 | 49.2 |
| Chloride | mg/L | 20.20 | 45.00 | 6 | 124 | 12 | 21.4 | 25 | 40 |
| Magnesium | mg/L | 9.50 | 13.70 | 7.9 | 9.7 | 9.4 | 11.8 | 12-13 | 14-15 |
| Potassium | mg/L | 1.40 | 2.10 | 1.3 | 2.6 | 1.2 | 1.5 | 1.8 | 2.3 |
| Sodium | mg/L | 21.50 | 43.70 | 9.4 | 15.8 | 16 | 20 | 29 | 40.2 |
| Sulfate | mg/L | 26.70 | 41.40 | 23 | 28.8 | 28 | 39 | 36 | 47.1 |
| Metals and Metalloids | | | | | | | | | |
| Aluminum - dissolved | ug/L | 16.00 | 49.00 | 16.2 | 104.85 | 7.96 | 39.06 | 4.23 | 18.39 |
| Aluminum - total | ug/L | 1533.00 | 6454.00 | 2770 | 13475 | 792 | 5480 | 97.5 | 1202.2 |
| Antimony - dissolved | ug/L | 0.11 | 0.20 | 0.087 | 0.129 | < | < | < | < |
| Antimony - total | ug/L | 0.15 | 0.39 | 0.1 | 0.152 | 0.065 | 0.285 | 0.051 | 0.125 |
| Arsenic - dissolved | ug/L | 0.50 | 0.70 | 0.546 | 0.787 | 0.504 | 0.799 | 0.424 | 0.596 |
| Arsenic - total | ug/L | 1.10 | 2.50 | 1.75 | 2.908 | 0.862 | 1.954 | 0.574 | 0.825 |
| Barium - dissolved | ug/L | 52.60 | 73.70 | 42.95 | 49.55 | 45.6 | 53.3 | 59.75 | 70.34 |

Table 1.13: Surface water quality triggers from the LARP Surface Water Quality Management Framework and seasonal current condition values calculated as part of this study for sites in the Athabasca River Delta. LARP values that appear to be an overestimate compared to the current condition values calculated in this study are bolded. Note that LARP central tendency measures are annual means, whereas this study used seasonal medians. (continued)

| | | LARP W | ater Quality Triggers | High | flow | Open | water | Und | Under ice | |
|---------------------------|-----------------|---------|---------------------------|--------|--------|---------|--------|----------------------|-------------|--|
| Parameter Name | Units | Mean | Peak (95th percentile) | Median | 95%ile | Median | 95%ile | Median | 95%ile | |
| Barium - total | ug/L | 79.30 | 147.60 | 86.15 | 239.25 | 56.9 | 141.06 | 64.05 | 77.965 | |
| Beryllium - dissolved | m ug/L | 0.08 | 0.27 | 0.006 | 0.022 | 0.001 | 0.043 | 0.003 | 0.046 | |
| Bismuth - total | ug/L | 0.02 | 0.06 | 0.017 | 0.06 | 0.009 | 0.023 | 0.002 | 0.021 | |
| Boron - dissolved | ug/L | 26.00 | 40.00 | 22.2 | 30.925 | 22.6 | 29.2 | 31.75 | 37.77 | |
| Boron - total | m ug/L | 48.00 | $\boldsymbol{69.00}$ | 24.8 | 41.775 | 24.7 | 40.54 | 32.85 | 39.78 | |
| ${\bf Cadmium-dissolved}$ | m ug/L | 0.10 | 0.52 | 0.009 | 0.022 | 0.009 | 0.109 | 0.014 | 0.033 | |
| Cadmium – total | m ug/L | 0.30 | 1.20 | 0.058 | 0.274 | 0.02 | 0.126 | 0.02 | 0.093 | |
| Chromium - dissolved | ug/L | 0.41 | 0.65 | 0.235 | 0.756 | 0.148 | 0.543 | 0.24 | 0.476 | |
| Chromium - total | ug/L | 3.00 | 8.00 | 3.215 | 11.71 | 0.919 | 6.314 | 0.216 | 0.685 | |
| Cobalt - dissolved | ug/L | 0.07 | 0.11 | 0.067 | 0.127 | 0.067 | 0.217 | 0.058 - 0.078 | 0.137 - 0.1 | |
| Cobalt - total | ug/L | 0.80 | 2.20 | 1.355 | 4.942 | 0.414 | 1.874 | 0.124 | 0.426 | |
| Copper - dissolved | $\mathrm{ug/L}$ | 1.60 | 3.60 | 1.555 | 2.46 | 0.97 | 2.184 | 0.75 | 1.353 | |
| Copper - total | ug/L | 3.10 | 7.20 | 3.645 | 10.127 | 1.42 | 4.812 | 0.905 | 1.897 | |
| Iron - dissolved | ug/L | 185.00 | 372.00 | 121.5 | 426.5 | 95 | 293.6 | 178 | 367.4 | |
| Iron - total | ug/L | 1899.00 | 5821.00 | 4240 | 13625 | 1050 | 4414 | 565.5 | 1294.5 | |
| Lead – dissolved | $\mathrm{ug/L}$ | 0.56 | 0.56 | 0.084 | 0.259 | 0.038 | 0.228 | 0.052 | 0.756 | |
| Lead - total | ug/L | 3.30 | 7.00 | 2.125 | 10.55 | 0.466 | 2.806 | 1.16 | 2.564 | |
| Lithium - dissolved | ug/L | 6.00 | 9.00 | 5.21 | 7.4 | 6.09 | 7.204 | 8.59 | 10.785 | |
| Lithium - total | ug/L | 9.00 | 12.00 | 7.455 | 16.95 | 6.83 | 8.132 | 8.92 | 11.085 | |
| Manganese - dissolved | ug/L | 12.00 | 36.00 | 1.725 | 6.015 | 1.4 | 8.228 | 18.8 | 35.095 | |
| Manganese - total | ug/L | 65.00 | 141.00 | 104.4 | 320.5 | 54.7 | 113.8 | 30.75 | 51.665 | |
| Mercury - total | ug/L | 0.01 | 0.02 | 0.0089 | 0.0238 | 0.00299 | 0.0137 | 0.00082 | 0.00425 | |
| Molybdenum - dissolved | ug/L | 0.70 | 1.20 | 0.494 | 0.7 | 0.629 | 0.984 | 0.638 | 0.752 | |
| | | | | | | | | | | |

Table 1.13: Surface water quality triggers from the LARP Surface Water Quality Management Framework and seasonal current condition values calculated as part of this study for sites in the Athabasca River Delta. LARP values that appear to be an overestimate compared to the current condition values calculated in this study are bolded. Note that LARP central tendency measures are annual means, whereas this study used seasonal medians. (continued)

| | | LARP Water Quality Triggers | | High | High flow | | Open water | | Under ice | |
|-----------------------|--------|-----------------------------|---------------------------|--------|-----------|--------|------------|---------------|-------------|--|
| Parameter Name | Units | Mean | Peak (95th percentile) | Median | 95%ile | Median | 95%ile | Median | 95%ile | |
| Molybdenum - total | ug/L | 0.90 | 1.60 | 0.516 | 0.73 | 0.602 | 0.985 | 0.649 | 0.769 | |
| Nickel - dissolved | ug/L | 1.60 | 4.70 | 1.425 | 3.475 | 0.749 | 1.334 | 0.764 | 1.473 | |
| Nickel - total | ug/L | 3.40 | 8.20 | 4.325 | 13.172 | 1.55 | 4.968 | 1.015 | 2.245 | |
| Selenium - dissolved | ug/L | 0.23 | 0.41 | 0.114 | 0.259 | 0.239 | 0.3 | 0.247 | 0.454 | |
| Selenium - total | ug/L | 0.33 | 0.58 | 0.26 | 0.467 | 0.22 | 0.3 | 0.3 | 0.5 | |
| Silver - total | ug/L | 0.02 | 0.07 | 0.022 | 0.329 | 0.006 | 0.027 | 0.002 - 0.003 | 0.011-0.01 | |
| Strontium - dissolved | ug/L | 215.00 | 361.00 | 162.5 | 213 | 206 | 253 | 266 | 339.4 | |
| Strontium - total | ug/L | 225.00 | 361.00 | 174.5 | 227.5 | 206 | 256.6 | 275 | 343.4 | |
| Thallium - dissolved | m ug/L | 0.02 | 0.11 | 0.006 | 0.008 | 0.005 | 0.014 | 0.005 | 0.019 | |
| Thallium - total | ug/L | 0.05 | 0.18 | 0.048 | 0.211 | 0.016 | 0.107 | 0.006 | 0.045 | |
| Thorium - dissolved | ug/L | 0.03 | 0.09 | 0.026 | 0.131 | 0.014 | 0.058 | 0.007 | 0.05 | |
| Thorium - total | ug/L | 0.35 | 1.44 | 0.415 | 2.51 | 0.135 | 0.882 | 0.024 | 0.204 | |
| Titanium - dissolved | ug/L | 2.00 | 7.00 | 1.905 | 9.209 | 1.03 | 4.722 | 1.175 | 2.328 | |
| Titanium - total | ug/L | 30.00 | 104.00 | 33.9 | 127 | 11.6 | 69.98 | 2.53 | 22.63 | |
| Uranium - dissolved | ug/L | 0.31 | 0.38 | 0.344 | 0.385 | 0.353 | 0.434 | 0.39 - 0.42 | 0.48-0.49 | |
| Uranium - total | ug/L | 0.40 | 0.70 | 0.487 | 1.274 | 0.414 | 0.646 | 0.4 - 0.44 | 0.53 - 0.52 | |
| Vanadium - dissolved | ug/L | 0.45 | 0.70 | 0.435 | 0.673 | 0.306 | 0.649 | 0.171 | 0.329 | |
| Vanadium - total | ug/L | 4.00 | 16.00 | 6.73 | 21.225 | 2.04 | 12.248 | 0.43 | 2.043 | |
| Zinc - dissolved | ug/L | 4.50 | 12.40 | 0.615 | 1.73 | 0.531 | 1.109 | 1.03 - 1.58 | 3.51 - 7.75 | |
| Zinc - total | ug/L | 12.30 | 25.60 | 10.355 | 32.95 | 3.1 | 15.626 | 1.65 - 2.58 | 6.98-13.2 | |

Note:

- data insufficient

< too highly censored

1.5 Conclusions and Next Steps

Along with the current conditions, the health risk criteria for water and sediment quality address limitations in the provincial water quality assessment and management system. Addressing these limitations is critical to protect Indigenous community members who rely on the aquatic ecosystem to live and exercise their rights as Indigenous Peoples.

The comparison of current conditions established in this report to the health risk criteria for surface water and sediment indicate that there are conditions in each of the Athabasca River, Athabasca River Delta and Lake Athabasca which warrant further investigation. This may be accomplished through studies assessing health risks from consuming traditional foods and untreated surface water, and by ongoing efforts to better understand the contribution of oil sands development to the current condition.

While surface water quality criteria to protect consumers of fish were identified, there are uncertainties associated with the methods employed (United States Environmental Protection Agency (US EPA), 2021; Sample et al., 1996) and there is an outstanding need to develop fish tissue specific criteria to ensure community members and wildlife consuming fish are sufficiently protected. Development of fish tissue residues for persistent and bioaccumulative substances would allow for an assessment of monitoring data currently available through various Community Based Monitoring (CBM) programs. Due to limited scope, this component was not integrated into the risk based criteria and future studies in this area are recommended.

The research presented here can be used by Indigenous communities, governments and regulatory agencies, and industry stakeholders to aid in answering community questions around how current and future oil sands development may affect the health of the environment and of Indigenous community members, as well as their ways of life, and cumulatively impact and further deteriorate conditions in the Athabasca River, Athabasca River Delta and Lake Athabasca. However, answering these questions requires implementation of this research and application of the WQCIUs in industry, community, and government led studies and assessments.

Specifically, the proposed health risk criteria and current conditions can be used assess potential changes in surface water and sediment conditions and risks to human and ecological receptors posed by releases of contaminants from oil sands developments to the Athabasca River and downstream within the Athabasca Delta and Lake Athabasca. The health risk criteria can also be used to guide decision making regarding the placement of tailings and OSMW in aquatic closure (reclamation) features such as constructed wetlands and end pit lakes (EPLs).

$_{6}$ Chapter 2

617 Current Conditions

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620 2.1 Introduction

The following describes the development of current conditions for application as surface water 621 and sediment quality criteria or limits of change. This reflects Indigenous communities' con-622 cerns that the condition of the Athabasca River, Athabasca River Delta and Lake Athabasca 623 should not be degraded any further from current condition, recognizing that the communities 624 have established that the current condition is already deteriorated from conditions prior to 625 1967. The objective of this study is to use existing, accessible water and sediment quality data 626 collected through various monitoring and research programs in the lower Athabasca River, the 627 628 Athabasca River Delta and Lake Athabasca to determine the range and variability in water and sediment quality parameters. This exercise will determine what normal (i.e., median) and 629 unusually low or high (i.e., 5th and 95th percentiles) values for these parameters are in recent 630 years at these locations. These values will be based on conditions during the period of record 631 for the data used in this study. It is important to note that in the view of ACFN, FMFN and 632 MCFN, the current conditions developed here are meant to serve as a baseline and not an ideal 633 state 634

635 2.2 Request from communities for current conditions

- 636 Athabasca Chipewyan First Nation (ACFN), Mikisew Cree First Nation (MCFN) and Fort
- 637 McKay First Nation (FMFN), three First Nations with territories located along the lower

Athabasca River (LAR), at Lake Athabasca and in the Peace-Athabasca Delta are concerned about water quality in these surface water systems. Since the onset of oil sands mining along the LAR along with other stressors on water quality related to upstream effluent release and landscape change, water quality in the LAR and its downstream environment has changed (Glozier et al., 2009; Hebben, 2009; Tondu, 2017; Glozier et al., 2018). In some cases, these changes have been in step with the nature and magnitude of these stressors, while in others the causes have not been identified.

In the face of ongoing development and land disturbance in the Lower Athbasca Region, including oil sands extraction operations, there is a desire to understand the quality of water and sediment in the lower Athabasca River, the Athabasca River Delta and Lake Athabasca in its current state. The variability in constituent concentrations and other measures of water and sediment quality across years and locations can be characterized and described using relatively simple statistics, which is one way to establish "antidegradation" quality criteria. This type of approach involves establishing what normal water and sediment quality at these locations is so that future monitoring results can be compared against these normal conditions, in order to detect when measured environmental quality is different from normal.

As part of the WQCIU project, ACFN, MCFN and FMFN have requested that this benchmark approach be taken in order to create a mechanism to ensure that water and sediment quality in the lower Athabasca River, its delta and Lake Athabasca do not deteriorate from current conditions. However, these communities have established that water and sediment quality in these locations has already deteriorated compared to conditions before human development in the region expanded significantly after 1967. Establishment of what is normal in these surface water systems using monitoring data that were collected after anthropogenic impacts have occurred means that this normal scenario does not represent natural or unimpacted conditions.

2.3 Long-term monitoring programs

The province of Alberta operates a long-term river network (LTRN) monitoring program which maintains four water quality monitoring sites on the lower Athabasca River and its delta, along with three upstream in the Athabasca Basin and many more throughout the province. Currently, this program involves approximately once-a-month sampling at the monitored sites, including the "Old Fort" station located in the Athabasca River Delta downstream of all oil sands development (historically, actually two stations - AB07DD0010 and AB07DD0105). The

- available water quality data record from this site runs from 1987 to present, although historically the program often missed certain months, especially during winter. Data from the Old Fort sites were used to establish current condition water quality triggers for the Surface Water Quality Management Framework of the Lower Athabasca Regional Plan (LARP)(Alberta Environment and Sustainable Resource Development (AESRD), 2012).
- Similarly, there is one long term monitoring station maintained by Environment and Climate Change Canada on the lower Athabasca River, also located downstream of all current oil
 sands development. This site is known as Athabasca River at 27 Baseline (AL07DD0001, or
 site M9) and has an available record of water quality data from 1989 to present day, collected
 monthly. Data from this station were included in the most recent federal reporting on water
 quality in the major rivers around Wood Buffalo National Park, specifically the Peace, Slave
 and Athabasca Rivers (using data up to 2006, (Glozier et al., 2009).
- Finally, since 2011, the Mikisew Cree First Nation (MCFN) and Athabasca Chipewyan First
 Nation (ACFN) have conducted a water quality monitoring program in the lower Athabasca
 River Delta and Lake Athabasca, as well as in the larger Peace-Athabasca Delta(PAD).

685 2.4 Regional monitoring programs targeting Oil Sands

686 2.4.1 Alberta Oil Sands Environmental Research Program (AOSERP)

The Alberta Oil Sands Environmental Research Program (AOSERP) was run by Alberta Environment and Parks between 1975 and 1985. The Program goal was to establish baseline conditions and assess terrestrial, aquatic, air and human impacts of oil sands developments, and numerous AOSERP reports 4 are available online. Unfortunately, the availability of AEOSERP data, especially in an electronic format, is limited. Many of the data sets are available only in published reports.

693 2.4.2 Regional Aquatics Monitoring Program (RAMP)

The Regional Aquatics Monitoring Program (RAMP) was initiated in 1997 as a multistakeholder organization, with funding provided by oil sands industry members. On its website, the RAMP lists Fort McKay First Nation and Fort McKay Métis Local No. 63 as members of its Steering Committee5, and in its organizational chart Fort McMurray First Nation is included as a member6, however it isn't clear when these memberships were in effect. In addition, the Steering Committee membership list includes municipal, provincial and federal government agencies

- 701 The objectives of the RAMP program 7 were as follows:
- Monitor aquatic environments in the Athabasca oil sands region to detect and assess cumulative effects and regional trends;
- Collect baseline data to characterize natural variability in the aquatic environment in the Athabasca oil sands region;
- Collect and compare data against which predictions contained in Environmental Impact
 Assessments (EIAs) can be assessed;
- Collect data that satisfy the monitoring required by regulatory approvals of oil sands and other developments;
- Collect data that satisfy the monitoring requirements of company-specific community
 agreements;
- Recognize and incorporate traditional environmental knowledge into monitoring and assessment activities;
- Communicate monitoring and assessment activities, results and recommendations to communities in the Regional Municipality of Wood Buffalo, regulatory agencies and other interested parties;
- Continuously review and adjust the program to incorporate monitoring results, technological advances, community concerns, and new or changed project approval conditions; and
- Conduct a periodic peer review of the program's results against its objectives, and recommend adjustments necessary for the program's continued success.
- 722 The RAMP was focused on monitoring both potential oil sands development stressors, such as water and sediment quality and hydrology, and potential oil sands development effects, such 723as in benthic invertebrate communities and fish populations. The RAMP program classified 724 sampling sites as baseline or test, depending on their location relative to oil sands development, 725 but also made extensive use of the idea of a regional baseline against which ongoing monitoring 726 results were compared. The RAMP regional study area8 included the lower Athabasca River 727 and the Athabasca River Delta, as well as Lake Athabasca (Figure 2.1). The water quality 728 729 regional baseline for the Athabasca River mainstem and Delta sites was based on data collected 730 in the fall from the Athabasca River upstream Fort McMurray, downstream of Fort McMurray and its wastewater treatment plant outfall but upstream of oil sands activity, as well as from 731 several tributaries of the lower Athabasca River (Hatfield Consultants, 2009). Unlike water 732 quality, sediment quality data were not compared to a regional baseline, but were compared 733 to data previously collected from the same stations. 734

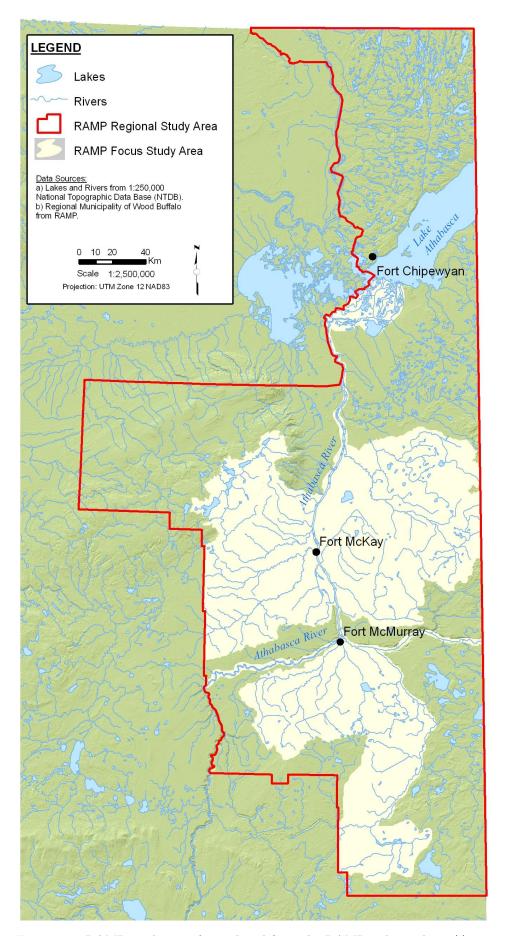


Figure 2.1: RAMP study area (reproduced from the RAMP website: http://www.rampalberta.org/ramp/design+and+monlitoring/approach/study+areas.aspx)

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Water and sediment quality monitoring was conducted at a maximum of 26 sites in the lower Athabasca River Mainstem, although sediment quality monitoring occurred only during certain time periods. In the Athabasca River Delta, sediment quality monitoring and limited water quality monitoring occurred in the Fletcher Channel, Goose Island Channel, Big Point Channel and the Embarras River. The RAMP did not include water or sediment quality monitoring of Lake Athabasca. A schematic diagram¹ produced by the RAMP of the relative water inflows from tributaries in the LAR is shown in Figure 2.2 below:

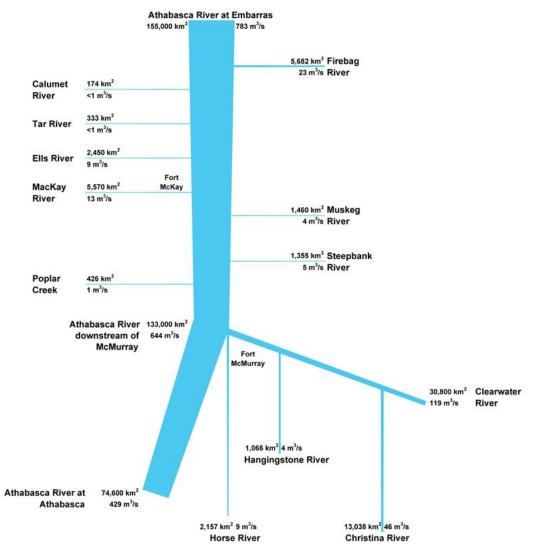


Figure 2.2: Relative water inflows from tributaries in the LAR (figure taken from the RAMP website: http://www.rampalberta.org/river/hydrology/river+hydrology.aspx).

The final standalone report from the RAMP was for the 2012 sampling year and was released in 2013. In 2010 and 2011, two scientific peer reviews of the RAMP program were conducted and identified several areas of concern in terms of the program's ability to detect change

¹http://www.ramp-alberta.org/river/hydrology/river+hydrology.aspx

over time and space (e.g., lack of statistical confidence or power), and especially its ability to identify change as impacts of oil sands development activity (e.g., poorly or undefined baseline conditions) (Dowdeswell et al., 2010). The RAMP issued a response to the AITF peer review (Burn et al., 2011), outlining changes to its monitoring, reporting and communication practices and providing additional explanation and information (Regional Aquatics Monitoring Program (RAMP), 2011). RAMP data was also made publicly available on the program website.

751 2.4.3 Joint Oil Sands Monitoring Program/Oil Sands Monitoring 752 Program (JOSM/OSM)

The Joint Oil Sands Monitoring Program (JOSM) was a cooperative effort between the governments of Canada and Alberta to monitor the environment in the lower Athabasca River/mineable oil sands region. The JOSM program was developed in response to criticisms of the RAMP program discussed above. The JOSM program officially operated between 2012 and 2015, working with many of the same consulting companies that had operated the RAMP program, and publishing collaborative annual reports. After 2015, the JOSM program transitioned to the Oil Sands Monitoring (OSM) Program, which retained some but not all of the RAMP water quality sampling sites.

The design of the JOSM program included several core elements, including an integrated monitoring program that would aim to measure "accumulated state," or changes in the aquatic environment that are outside of both local and regional baseline. Measuring accumulated state requires the establishment of a baseline state, however the JOSM design document acknowledged that establishing baseline water quality condition in the mineable oil sands region (OSR) would be challenging due to the low number of long-term water quality monitoring stations in the OSR, the general lack of water or sediment quality data from the time before oil sands development, and the changing nature of oil sands development stressors (mines and other facilities being built and expanding over time) (Wrona et al., 2011). In order to better estimate baseline conditions, the JOSM water quality program design suggested using modeling exercises, data mining existing reports for historic data, and using sediment cores from surface waters to provide information about historical conditions. The water quality design document also indicated that the JOSM program should include establishment of additional baseline or unimpacted reference sites to the extent possible, as well as include efforts to monitor impacted areas before and after development occurs in the future.

Measuring accumulated state also requires monitoring of landscape change over space and time, including changes in point and non-point source loadings of substances to surface waters

- 778 (Wrona et al., 2011). The separate types of oil sands development compliance and performance
- 779 (i.e., follow-up) monitoring were mentioned in the JOSM water quality program design. It was
- 780 noted that this monitoring data must be integrated into a standardised and accessible electronic
- 781 reporting system that is shared with the larger regional monitoring program. Performance
- 782 monitoring in particular was included as a requirement to verify or validate predictions made
- 783 in Environmental Impact Assessments (Wrona et al., 2011).
- The core results proposed for the JOSM water quality monitoring program were:
- Assessment of accumulated environmental condition or state;
- Improved understanding of the relationships between system drivers and environmental
- 787 response; and,
- Cumulative effects assessment. (Wrona et al., 2011)
- According to the JOSM design document, in the absence of these core results, "cumulative
- 790 change cannot be detected, predicted, managed or mitigated." (p. 9).
- 791 Ten monitoring locations were selected for the mainstem Athabasca River, from the inflow-
- 792 ing "boundary condition" M0 site at the town of Athabasca downstream to M9 the downstream
- 793 boundary condition, closest to the Athabasca River Delta at Lake Athabasca and downstream
- 794 of all oil sands development (see Figure 2.3 below). These sites incorporated several existing
- 795 provincial and federal long-term monitoring program locations.

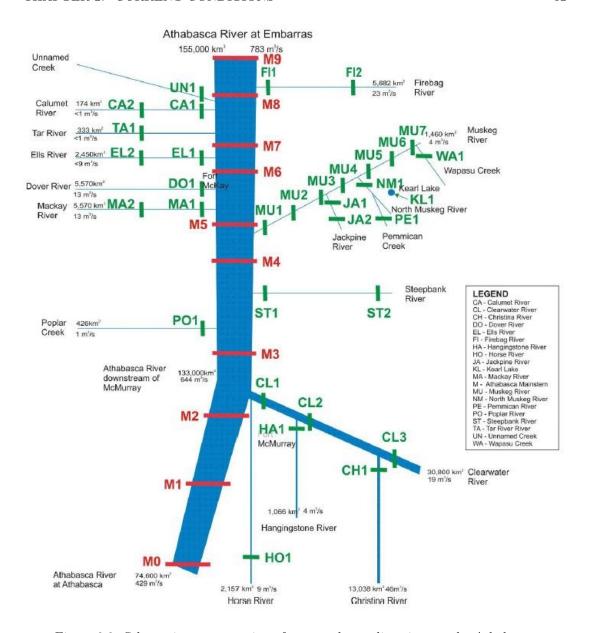


Figure 2.3: Schematic representation of proposed sampling sites on the Athabasca River mainstem and major tributaries (reproduced from Wrona et al. (2011), Figure 6).

The JOSM water quality program was designed to be integrated and coordinated with a hydrometric and sediment monitoring program, since it was recognized that sediment dynamics in the Athabasca River can be a significant driver of contaminant dynamics in the River and of contaminant loadings to downstream environments (Wrona et al., 2011). Groundwater quality monitoring was also meant to be coordinated with surface water quality monitoring as part of the program design, especially focused around oil sands mine tailings impoundments. Naphthenic acids, as a complex mixture of compounds that are a significant source of toxicity in oil sands process water, were targeted for further characterization, including by a fingerprinting

804 research program conducted by Environment Canada (Wrona et al., 2011).

The JOSM program and its successor program, OSM, have been operating up to present 805 day. In 2018, a series of summary reports were published for the JOSM aquatics program using 806 data collected up to 2015. At that time, only one statistically significant longitudinal (upstream 807 to downstream) trend in water quality was noted - a gradual increase in dissolved selenium 808 between M3 and M6, after which concentrations stabilized downstream (Cooke et al., 2018). 809 Those authors also noted a decreasing trend or stabilization of several nitrogen and phosphorus 810 measures between the years 2000 and 2014 at the long-term monitoring site M9. These trends 811 were linked by the authors to several changes in anthropogenic inputs, both upstream of Fort 812 McMurray as well as at the Fort McMurray wastewater treatment plant when the treatment 813 process was improved significantly in 2010 (Cooke et al., 2018). Increasing trends between 2000 814 and 2014 in certain metal concentrations, including dissolved arsenic, aluminum and iron, as 815 816 well as total selenium were also noted, as were decreasing and increasing trends for certain ions. After a water quality monitoring network rationalization exercise conducted in 2016, sampling 817 at some of the mainstem Athabasca River monitoring sites was discontinued. 818

2.4.4 Other Monitoring in the LAR, the PAD and Lake Athabasca

Several other large multi-year monitoring and research programs have been completed over the years, with support from provincial and federal government agencies and to varying extents the involvement of Indigenous communities. These include the Northern River Basins Study (1991-1996), the Peace-Athabasca Delta Technical Studies (1993-1996), and the Northern Rivers Ecosystem Initiative (1998-2004). Similar to the AEOSERP program data, the availability of monitoring and sampling data generated by these programs is limited, with many of the data sets available only in published reports.

The province of Alberta has historically collected water quality data from Lake Athabasca, especially in the late 1980's and 1990's. This data is available from the province's surface water quality website under the "Lake Water Quality" program name, which includes data from lakes located across Alberta.

In addition to these long-term studies and monitoring programs, there have been many focused field programs and studies conducted by Indigenous communities, academic institutions, private industry and governments that encompassed water and sediment quality in the lower Athabasca River region. The vast majority of these studies' data are not readily available in a digital format, and were not included in this study. However, digitizing these historical data sets for inclusion in an enhanced water and sediment quality characterization effort would be

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837 a worthwhile future project.

838 2.5 Methods

839 2.5.1 Data used in this Study

840 2.5.1.1 RAMP data

- The RAMP water quality data is available for download from a dedicated website that is 841 maintained by Alberta Environment and Parks. Both water and sediment quality data are 842 843 available from the RAMP program for sites in the lower Athabasca River and the Athabasca River Delta channels. For all data used in this study, including RAMP data, it was assumed 844 that data review and quality control was completed by the responsible program. Sediment 845 quality samples were collected once per year in the fall. Water quality samples were collected 846 847 from the Athabasca River and Delta in the fall, with one site sampled four times per year (ATR-DD). Water quality samples were also collected multiple times per year at two sites, 848 upstream of Fort McMurray and at "Old Fort," but this actually reflects provincial long-term 849 monitoring (Hatfield Consultants, 2009). Sediment quality was generally no longer sampled in 850 the Athabasca River after 2004, and water quality was no longer sampled at most sites in the 851 Athabasca River Delta channels after 2004. 852 Water samples were generally collected as near-surface grab samples, with the sample bottle 853 uncapped and recapped at depth where possible (Hatfield Consultants, 2009). Field measures 854 of water quality were obtained using a multiparameter sonde, a Winkler titration kit, a pH 855 meter and a turbidity meter. Sediment samples were collected mainly with grab samplers or 856 dredges (e.g., Ekman or Ponar grab), from depositional environments within river channels. 857 At certain times, for example at some Athabasca Delta sites in 2005, a sediment corer was 858 859 used to collect sediment samples for analysis (Hatfield Consultants, 2009). 860 The number of water quality parameters measured by RAMP also varied over time, but generally included basic chemical and physical properties, major ions, nutrients, metals, naph-861 thenic acids and some polycyclic aromatic compounds (PACs). While the parameters analysed 862
- addition of "ultra-trace" analysis of total mercury in water in 2002 (effectively lowers the detection limit, can detect lower concentrations)

important changes to the analysed water quality parameters, including:

did not change substantially over the course of the program up until 2012, there were a few

discontinuation of PAC analysis in water in 2005 due to non-detectable or very low

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- some concentrations in nearly all water samples
- discontinuation of chlorophyll analyses in water from streams and rivers in 2006 due to
- frequent non-detectable concentrations and a lack of correlation with nutrient parameters
- (chlorophyll continued to be measured in periphyton or algae from the bottom of streams
- and rivers)
- a switch in the laboratory conducting metals analysis in 2002 (Hatfield Consultants,
- 874 2009)
- In 2006, the RAMP sediment quality monitoring program was modified to better align
- 876 with sampling of benthic invertebrates, and a one-time extensive sediment quality program
- 877 was conducted in the Athabasca River Delta (Hatfield 2009). The parameters analysed in
- 878 the RAMP sediment quality program generally included physical properties, carbon content,
- 879 metals, various organic compounds, and 'parent' and alkylated polycyclic aromatic compounds
- 880 (PACs). The analysed parameters changed over time as follows:
- addition of particle size distribution, total inorganic carbon, and total carbon in 1998
- addition of total volatile hydrocarbons (TVH) and total extractable hydrocarbons (TEH)
- 883 in 2000

- switch to the Canadian Council of Ministers of the Environment (CCME) four-fraction
- hydrocarbon assay in 2005.
- Analytical methods, and specifically VMV method codes, for RAMP water and sediment
- 887 quality samples were taken from Table 1 and Table 2 of the Addenda to the RAMP Technical
- 888 Design and Rationale Document (Hatfield Consultants, 2011), and verified through discussions
- 889 with Hatfield Consultants personnel (M. Davies, pers. comm. October and September 2020)
- 890 and staff of AXYS Analytical Services Ltd. (G. Brooks, pers. comm. December 2020).

2.5.1.2 LTRN and LWQ provincial data

- 892 The province of Alberta maintains two water quality sampling stations in the lower Athabasca
- 893 River mainstem, as part of the provincial Long-Term River Network (LTRN) water quality
- 894 monitoring program. The furthest upstream site is just upstream of Fort McMurray and the
- 895 confluences of the Horse and Clearwater Rivers (AB07CC0030, also known in the JOSM/OSM
- 896 program as site M2). Further downstream is the next site, which is upstream of the confluence
- 897 with the Firebag River (AB07DA0980, also known in the JOSM/OSM program as site M8).
- 898 Downstream in the Athabasca River Delta, two more LTRN sites together make up the station
- 899 known as "Old Fort" (AB07DD0010, AB07DD0105). The annual water quality record for Old

Fort from before 2016 is actually the combined monthly sampling at site AB07DD0010 during 900 the open water season, and at AB07DD0105 during the ice-covered season (Kruk & Ballard, 901 2020). The two stations are separated by about 20 km and the confluence of the Richardson 902 River. In 2016, year-round monthly sampling began at site AB07DD0010 ("Athabasca River 903 at Old Fort - Right Bank") but site AB07DD0105 ("Athabasca River downstream of Devil's 904 Elbow at Winter Road Crossing") remains a seasonal sampling site with data collected for the 905 ice-covered season only. 906 Monthly sampling has been conducted either seasonally or year-round at the lower 907 Athabasca River LTRN sites as early as 1987 upstream of Fort McMurray, since 1989 at 908 Old Fort, and since 2008 at the site upstream of the Firebag River. LTRN water quality 909 sampling has involved the analysis of hundreds of parameters, including basic chemical and 910 physical properties, major ions, nutrients, metals, naphthenic acids, parent, alkylated and 911 912 nitrogen-containing polycyclic aromatic compounds (PACs), pesticides, bacteriological measures, general organics, organohalides, phthalates, and phenolics. Not all of these parameters 913 have been measured for the entire duration of the program, however. LTRN water samples 914 in the lower Athabasca River were generally collected as near-surface grab samples or as 915 vertically integrated samples (sample bottle on a sampling iron lowered through the water 916 column) (GoA, 2019b). 917 LTRN water quality data are available for download via a dedicated website that is main-918 tained by Alberta Environment and Parks10,11. However, for the purposes of this study, data 919 were obtained directly via an email request to the Alberta Environment and Parks surface wa-920 921 ter data request email 12, which provided a more comprehensive dataset with more measured parameters compared to what is available online. 922 The province of Alberta also maintains a website with water quality data obtained from 923 lakes in the province, including from Lake Athabasca13, although provincial lake water quality 924 (LWQ) data availability is not as consistent over time as the LTRN program. Water quality 925 data from ten sites on Lake Athabasca were obtained by direct email request from Alberta 926 Environment and Parks, and the majority of the data were collected in the late 1980's and 927 early 1990's. There were dozens of water quality parameters measured, including basic chem-928 929 ical and physical properties, major ions, nutrients, chlorophyll a, metals, parent polycyclic aromatic compounds (PACs), bacteriological measures, general organics, organohalides, ph-930 thalates, phenolics and radium radiation. Vertical profile data for basic field measures were 931

collected at some of the Lake Athabasca sites.

2.5.1.3 ECCC long-term monitoring data

Environment and Climate Change Canada (ECCC) maintains a water quality monitoring site 934 on the lower Athabasca River as part of its National Long-Term Water Quality Monitoring 935 Program. The site (AL07DD0001) is located North of the confluence with the Firebag River 936 in the south-western corner of Wood Buffalo National Park, and is referred to as Athabasca 937 River at 27 Baseline. The monitoring site has been maintained since 1989, but the official data 938 set available from the ECCC website includes data from the year 2000 to present. Water is 939 sampled at the site monthly, except in November and December, for basic chemical and physical 940 properties, major ions, nutrients, metals, parent and alkylated polycyclic aromatic compounds 941 (PACs), and pesticides. This site was incorporated into the JOSM/OSM program as M9 (see 942 943 below), and is considered to reflect improvement or "recovery" conditions from impacts of oil sands development and WWTP-related impacts to water quality and other aquatic ecosystems 944 (Glozier et al., 2018). 945

946 **2.5.1.4** JOSM/OSM data

The Joint Oil Sands Monitoring (JOSM) and Oil Sands Monitoring (OSM) Programs, now just OSM, involved sampling for water quality in the lower Athabasca River mainstream and its tributaries. There are over a dozen sites on the River that are referred to as OSM sites, however in actuality, several of these overlapped with AEP LTRN sites (M0, M1, M2, M8) and ECCC long-term monitoring sites (M9). There were therefore five water quality sites that were established specifically for the JOSM-OSM program (M3 through M7), and in some cases these sites are in the vicinity of former RAMP sites.

Water quality data generated by the JOSM-OSM program were obtained from the federal Oil Sands Monitoring website 14. Data were downloaded from the "mainstem" lower Athabasca River water quality dataset, which was collected starting in 2011 and with data available up to 2018.

The JOSM mainstem water quality program began with a comprehensive investigation of sampling methods and data variability in the River, from 2011 to 2014 (Glozier et al., 2018).

Different field sampling methods and data treatments were investigated using a 10-panel cross-

channel approach at each sampling site (Figure 2.4).

| West Shore | | Panel | | | | | | | | | East Shore |
|--------------------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|----------------------|
| | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | |
| A) Ten Panel Isokinetic Composite | \otimes | Physically Pooled |
| B) Ten Panel Sampling Iron Composite | \otimes | Physically Pooled |
| C) Ten Panel Sampling Iron Grab | \otimes | Statistically Pooled |
| D) 3 Panel Sampling Iron Grab | | | \otimes | | | \otimes | | \otimes | | \otimes | Statistically Pooled |
| E) Thalweg Sampling Iron Grab | | | | | | \otimes | | | | | Individual Grab |

Figure 2.4: Schematic of multi-panel sampling approaches, categories and data treatment for statistical analyses (reproduced from Glozier et al. (2018), Figure 18).

The results of the methods investigation indicated that cross-channel variability in water quality was significant at OSM sites M3 through M7 in the mainstem. For this reason, the JOSM researchers recommended that vertically integrated water samples (taken from the top of the River water column down to the River bed) at the deepest point of the River in each cross-section site (the thalweg) become the standard JOSM water quality sampling method for the lower Athabasca River. Importantly, the JOSM researchers determined that water quality samples taken from just below the River water surface, usually from shore or even from the middle of the River, are not comparable to samples collected according to the JOSM standard (Glozier et al., 2018). This difference is most likely associated with the larger amount of suspended sediment and other particles that are carried in the River due to the different hydrodynamic forces through the water column at the thalweg, compared with at the water surface and especially along the shoreline, where water flow energy is lower (N. Glozier, personal communication, January 22 2021; C. Cooke, personal communication, January 28 2021).

A water quality network rationalization workshop was attended by JOSM researchers and others in 2016, and as a result sampling at sites M4, M5 and M6 were suspended after March 2017 (Cooke et al., 2018; Glozier et al., 2018). Water quality at these three sites was determined to be essentially the same, apart from an increase in dissolved selenium concentrations with distance downstream (Glozier et al., 2018). Sites M4-M6 were originally intended to monitor flow and water quality including constituent loads up and downstream of major tributary rivers, and the recommendation to suspend monitoring at these sites noted that conditions at M7 capture all inputs from major tributary rivers (Glozier et al., 2018). Sampling at sites M1 was also suspended as part of the program rationalization (sampled from shore by Alberta Environment and Parks, AB07CC0100). The program rationalization confirmed that site M0 and the "Grand Rapids" site upstream of the McMurray oil sands geological formation and

Fort McMurray are necessary to characterize conditions upstream of the oil sands region. Both 986 of these sites are sampled by Alberta Environment and Parks (site codes M0 = AB07BE0010, 987 Grand Rapids = AB07CC0130). The rationalization also identified a step-change in water 988 quality parameters between sites M2 and M3 (Glozier et al., 2018). Both M2 and M3 are 989 located within the McMurray formation and upstream of oil sands development, but site M2 990 is upstream of the wastewater treatment plant (WWTP) effluent release location while M3 991 is downstream of that location and therefore influenced by this effluent release. Site M2 is 992 sampled from the shore by Alberta Environment and Parks (AB07CC0030), while sampling at 993 M3 is conducted using the OSM depth-integrated at the thalweg and shoreline panel method. 994 Sampling at M7 in the OSM program continues and water quality at that site is characterized 995 as capturing cumulative effects of all oil sands development as well as inputs from major LAR 996 tributaries (Glozier et al., 2018). There is also water quality data for the lower Athabasca River 997 998 mainstem available as part of the OSM benthic invertebrate monitoring program, however that data was not used in this study. This is because the sampling methods used were best suited for 999 characterization of the local habitat conditions, specifically erosional habitats where benthic 1000 1001 invertebrates could be effectively sampled, rather than for characterization of the River as a whole. 1002

1003 2.5.1.5 MCFN and ACFN CBM data

MCFN and ACFN began water quality collection in 2011 as part of community-based mon-1004 itoring (CBM) programs. These programs have several sites located throughout the Peace-1005 1006 Athabasca Delta, as well as the Athabasca River and Lake Athabasca. Sampling is ongoing and generally occurs throughout the open water season. Water quality data from these pro-1007 grams were obtained from the program manager (B. Maclean and C. Bampfylde, pers. comm.), 1008 and are also available online (MCFN15 and ACFN16). Generally speaking, these programs 1009 have involved the approximately weekly collection of "field" water quality data using a multi-1010 sensor sonde during the open water season, as well as more detailed near-surface grab water 1011 samples for laboratory analyses approximately four times a year, although this approach has 1012 varied over the years. Finalized data for this monitoring program were obtained directly from 1013 the program managers, for sampling between 2014 and 2019. Field-measured water quality 1014 data for both the ACFN and MCFN CBM programs are reported as water-column average 1015 1016 values.

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2.5.1.6 Enhanced Monitoring Program data

The Enhanced Monitoring Program is a focused study of water and sediment quality in the 1018 lower Athabasca that was initiated as part of the work of the Oil Sands Process Water (OSPW) 1019 Science Team and has been funded by the Oil Sands Monitoring (OSM) program. The En-1020 hanced Monitoring program collected water and sediment quality samples during 2018 and 2019 in a localized area near a proposed mine water release site, in addition to sites further 1022 up- and downstream in the Athabasca River. Because bed sediment quality data for the lower Athabasca River in recent years is not otherwise readily available, data from this program 1024 was used in part to characterize sediment quality in the mainstem Athabasca River. Water 1025 quality data for this program are currently available through a publicly accessible website sup-1026 1027 ported by the OSM program, however, sediment quality data were provided by the study's 1028 lead researcher (K. Hicks, pers. comm).

2.5.1.7 Compiled Sites – Water

Table 2.1 below lists all of the monitoring site locations by water quality monitoring program, for all data compiled in this study. The sites from which data were used to calculate current conditions are indicated in bold text in the table, and all data compiled from all programs are presented in Appendix A.1.

Table 2.1: Names and locations of monitoring sites that were included in the water quality data compilation. Bolded rows indicate locations used in the calculation of current conditions. The selection rationale for these locations is explained in the data selection methods sections below.

| Section | Site Name | Program | Latitude | Longitude |
|--------------------------|-----------------|-----------------|-----------|------------|
| Athabasca River | AB07CC0030 | LTRN | 56.720280 | -111.40556 |
| Athabasca River | AB07DA0980 | LTRN | 57.723610 | -111.37917 |
| Athabasca River | AL07DD0002 | JOSM | 56.720611 | -111.40283 |
| Athabasca River | AL07DD0004 (M4) | \mathbf{JOSM} | 57.127639 | -111.60003 |
| Athabasca River | AL07DD0005 (M5) | JOSM | 57.157583 | -111.62394 |
| Athabasca River | AL07DD0007 (M7) | \mathbf{JOSM} | 57.313950 | -111.66737 |
| Athabasca River | AL07DD0008 (M3) | JOSM | 56.839910 | -111.41164 |
| Athabasca River | AL07DD0009 (M6) | \mathbf{JOSM} | 57.215300 | -111.60727 |
| Athabasca River | Snowbirds | ACFN/MCFN | 58.355402 | -111.54556 |
| Athabasca River Delta | AB07DD0010 | LTRN | 58.382780 | -111.51778 |
| Athabasca River Delta | AB07DD0105 | LTRN | 58.447220 | -111.18583 |
| Athabasca River Delta | Athabasca River | ACFN/MCFN | 58.657433 | -110.77628 |

Table 2.1: Names and locations of monitoring sites that were included in the water quality data compilation. Bolded rows indicate locations used in the calculation of current conditions. The selection rationale for these locations is explained in the data selection methods sections below. *(continued)*

| Section | Site Name | Program | Latitude | Longitude |
|-------------------------------|----------------------------------|-----------------------|--------------|------------|
| Athabasca River Delta | Athabasca River at Cutoff | ACFN/MCFN | 58.397113 | -111.52733 |
| Athabasca River Delta | Athabasca at Embarras Portage | ACFN/MCFN | 58.397113 | -111.52733 |
| Athabasca River Delta | Embarras Lowpoint | ACFN/MCFN | 58.472286 | -111.48958 |
| Athabasca River Delta | Embarras River | ACFN/MCFN | 58.685627 | -111.05304 |
| Athabasca River Delta | Fisherman's Channel | ACFN/MCFN | 58.661893 | -110.77168 |
| Athabasca River Delta | Goose Island Channel | ACFN/MCFN | 58.669596 | -110.87028 |
| Lake Athabasca | Dock Site | ACFN/MCFN | 58.690843 | -111.15889 |
| Lake Athabasca | Lake Athabasca | ACFN/MCFN | 58.711461 | -111.08976 |
| Lake Athabasca | Water Intake | ACFN/MCFN | 58.710816 | -111.14499 |
| Note: Bolded rows indicates t | hat the site contributed to | the current condition | calculation. | |

2.6 Compiled Sites – Sediments

Table 2.2 below lists all of the monitoring site locations by sediment quality monitoring program, for all data compiled in this study. The sites from which data were used to calculate current conditions are indicated in bold text in the table, and all data compiled from all programs are presented in Appendix A.1.

Table 2.2: Names and locations of monitoring site that were included in the sediment quality data compilation. Bolded rows indicate locations used in the calculation of current conditions. The selection rationale for these locations is explained in the data selection methods sections below.

| Section | Site Name | Program | Latitude | Longitude |
|-----------------|-------------------|-----------------|-----------|------------|
| Athabasca River | AB07DA0062 | OSPW | 56.850200 | -111.42064 |
| Athabasca River | AB07DA0800 | \mathbf{OSPW} | 57.330470 | -111.67964 |
| Athabasca River | ${ m AB07DA3008}$ | \mathbf{OSPW} | 57.122941 | -111.60156 |
| Athabasca River | AB07DA3009 | \mathbf{OSPW} | 57.070580 | -111.53305 |
| Athabasca River | ${ m AB07DA3015}$ | \mathbf{OSPW} | 57.047184 | -111.50941 |
| Athabasca River | AB07DA3016 | \mathbf{OSPW} | 57.047853 | -111.51138 |
| Athabasca River | ${ m AB07DA3017}$ | \mathbf{OSPW} | 57.039101 | -111.50832 |
| Athabasca River | AB07DA3018 | OSPW | 57.037512 | -111.50970 |
| Athabasca River | AB07DA3020 | \mathbf{OSPW} | 57.034986 | -111.50558 |

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Table 2.2: Names and locations of monitoring site that were included in the sediment quality data compilation. Bolded rows indicate locations used in the calculation of current conditions. The selection rationale for these locations is explained in the data selection methods sections below. *(continued)*

| Section | Site Name | Program | Latitude | Longitude |
|-----------------------|-------------|-----------------|-----------|------------|
| Athabasca River | AB07DA3021 | OSPW | 57.033723 | -111.50386 |
| Athabasca River | AB07DA3022 | \mathbf{OSPW} | 57.029219 | -111.50218 |
| Athabasca River | AB07DA3023 | \mathbf{OSPW} | 57.009880 | -111.47409 |
| Athabasca River | AB07DA3024 | \mathbf{OSPW} | 56.939911 | -111.44329 |
| Athabasca River | ATR-DC-CC | RAMP | 56.826557 | -111.40931 |
| Athabasca River | ATR-DC-E | RAMP | 56.826562 | -111.40767 |
| Athabasca River | ATR-DC-M | RAMP | 56.826538 | -111.40839 |
| Athabasca River | ATR-DC-W | RAMP | 56.826540 | -111.40796 |
| Athabasca River | ATR-DD-CC | RAMP | 57.453661 | -111.60622 |
| Athabasca River | ATR-DD-E | RAMP | 57.452778 | -111.60232 |
| Athabasca River | ATR-DD-W | RAMP | 57.455284 | -111.60981 |
| Athabasca River | ATR-ER | \mathbf{RAMP} | 58.353316 | -111.54185 |
| Athabasca River | ATR-FC-CC-D | RAMP | 57.407729 | -111.64489 |
| Athabasca River | ATR-FC-E | RAMP | 57.407625 | -111.64035 |
| Athabasca River | ATR-FC-E-D | RAMP | 57.409593 | -111.64048 |
| Athabasca River | ATR-FC-M | RAMP | 57.407759 | -111.64527 |
| Athabasca River | ATR-FC-W | RAMP | 57.407621 | -111.64987 |
| Athabasca River | ATR-FC-W-D | RAMP | 57.410182 | -111.64984 |
| Athabasca River | ATR-FR-CC | RAMP | 57.740747 | -111.36842 |
| Athabasca River | ATR-FR-E | RAMP | 57.744557 | -111.36186 |
| Athabasca River | ATR-FR-W | RAMP | 57.746842 | -111.36907 |
| Athabasca River | ATR-MR-E | RAMP | 57.131901 | -111.60292 |
| Athabasca River | ATR-MR-E-D | RAMP | 57.133029 | -111.60510 |
| Athabasca River | ATR-MR-M | RAMP | 57.131120 | -111.60509 |
| Athabasca River | ATR-MR-W | RAMP | 57.130189 | -111.60786 |
| Athabasca River | ATR-MR-W-D | RAMP | 57.132301 | -111.60898 |
| Athabasca River | ATR-SR-E | RAMP | 57.019199 | -111.47867 |
| Athabasca River | ATR-SR-M | RAMP | 57.017546 | -111.48007 |
| Athabasca River | ATR-SR-W | RAMP | 57.015363 | -111.48112 |
| Athabasca River | ATR-UFM | RAMP | 56.718330 | -111.40307 |
| Athabasca River Delta | ARD-1 | \mathbf{RAMP} | 58.590791 | -110.79524 |
| Athabasca River Delta | ARD-2 | RAMP | 58.439591 | -111.29812 |
| Athabasca River Delta | ATR-OF | \mathbf{RAMP} | 58.408734 | -111.50990 |
| Athabasca River Delta | BEC | RAMP | 58.452500 | -111.06111 |
| Athabasca River Delta | BPC-1 | \mathbf{RAMP} | 58.590791 | -110.79524 |

Table 2.2: Names and locations of monitoring site that were included in the sediment quality data compilation. Bolded rows indicate locations used in the calculation of current conditions. The selection rationale for these locations is explained in the data selection methods sections below. (continued)

| Section | Site Name | Program | Latitude | Longitude |
|------------------------------|-----------|-----------------|-----------|------------|
| Athabasca River Delta | BPC-2 | RAMP | 58.462714 | -110.85983 |
| Athabasca River Delta | EMR-1 | \mathbf{RAMP} | 58.358268 | -111.55015 |
| Athabasca River Delta | EMR-2 | RAMP | 58.567500 | -111.09222 |
| Athabasca River Delta | FLB-1 | \mathbf{RAMP} | 58.447996 | -110.91532 |
| Athabasca River Delta | FLC-1 | RAMP | 58.564539 | -111.06220 |
| Athabasca River Delta Note: | GIC-1 | RAMP | 58.588101 | -110.83525 |

Bolded rows indicates that the site contributed to the current condition calculation.

2.7 Calculation of Current Conditions

1040 2.7.1 Data standardization

One of the most significant challenges in assembling water and sediment quality data from multiple sources is to standardize the data descriptions to ensure that the same or similar measurement and analytical methods are used for the compiled parameter-specific data sets Sprague et al. (2017). This allows for a comparison of "apples to apples" in terms of each specific parameter across all programs.

The United States Environmental Protection Agency (US EPA) has created a data standard framework for discrete non-continuous water quality dataset reporting, known as WQX, or Water Quality Exchange². This framework was adopted by the DataStream initiative in Canada, an open access platform for sharing surface water quality and sediment quality data developed and maintained by the non-profit Gordon Foundation³. As part of its program, DataStream produced an upload template ⁴ as well as nutrient data standardization guidance ⁵. This template was used in this study to compile water and sediment quality data from all of the source data sets. The nutrient guidance document was also followed, specifically the separation of filtration status and extraction/sample preparation status, in order to avoid ambiguity and ensure comparability. According to that guidance, the terms "filtered," "unfiltered" and "non-filterable" were assigned to account for the more conventional sample fraction

²https://www.epa.gov/waterdata/water-quality-data

³https://gordonfoundation.ca/initiatives/datastream/

 $^{^4} https://datastream.cdn.prismic.io/datastream/8af9357f-b1aa-40dd-ba5c-59fa990c01f2_DataStream+Upload+Template+2.5_Jan2021.xlsb$

 $^{^5} https://datastream.cdn.prismic.io/datastream\%2F9d12bb3f-e456-4de0-9613-f8f7e50f221a_datastream+nutrient+data+best+practices+guide_march2019.pdf$

descriptions "dissolved," "total" and "particulate." At the same time the term "total" was assigned to encompass multiple forms including organic/inorganic, ionic/biological, etc. For example, the parameter "Total nitrogen, mixed forms" refers to multiple forms of nitrogen (i.e., organic nitrogen, ammonia, nitrate, nitrite) and is accompanied by an additional sample fraction qualifier, namely filtered, unfiltered or non-filterable. These combinations would therefore correspond to the more conventional terms total dissolved nitrogen, total nitrogen and total particulate nitrogen, respectively. Care was taken to ensure that reported method speciation aligned or were converted to equivalence (e.g., all forms of nitrogen reported 'as N,' and not separately as N, NO3, NH4, etc., when combining and comparing across data sets).

A similar approach was taken for trace elements and metals, where the filtration status was reported separately, as the sample fraction, while the characteristic name indicated the type of extraction methods used. Generally, little to no extraction was conducted for dissolved metals, acidification over time was used for extractable metals, acidification and heat were applied for total metals, and acidification, heat and increased pressure for total recoverable metals.

Detailed method descriptions were consulted to determine the preparation and analytical methods used for each parameter, and clarifications were made with the data holder. For almost all programs, valid method variable, or VMV codes, were provided for each observation. VMV codes are specific to several aspects of laboratory analysis, including sample preparation and analysis methods, and detection limits. VMV dictionary files were provided by both Alberta Environment and Parks and Environment and Climate Change Canada researchers (N. Glozier, pers comm.), to account for differences between VMV schemes in use by the two agencies. For certain data from the RAMP program, as well as for ACFN and MCFN CBM water data, VMV codes were not provided in the original data sets. Instead, other standardized methods contexts, including US EPA and American Public Health Association (APHA) method numbers, are provided wherever possible. Additional method information was obtained from the data holders and responsible laboratories where possible. Where it wasn't possible to determine aspects of the methods used, especially for sample fraction (filtration status), the label "unknown" was added to the parameter name instead. No outliers were removed from datasets, and only finalized data that had undergone program-specific quality control measures were used in this study (please refer to each program for details of these measures).

A purpose-built PostgreSQL database was created to house all of the compiled data sets, with native support for International System of Units (SI) units. This means that the original source data along with the respective unit and method speciation were imported as a complete

| 1091 | observation, and were converted to a standard unit for analysis and display as required. Each |
|------|---|
| 1092 | parameter in the database was differentiated for analysis and reporting as a unique combination |
| 1093 | of basic parameter name, method speciation and sample fraction. The integrity of data in |
| 1094 | the database was controlled through automated data subset checks including unit conversion |
| 1095 | checks, before-and-after aggregate counts and value sum tests. This data flow is illustrated in |
| 1096 | Figure 2.5 below. |

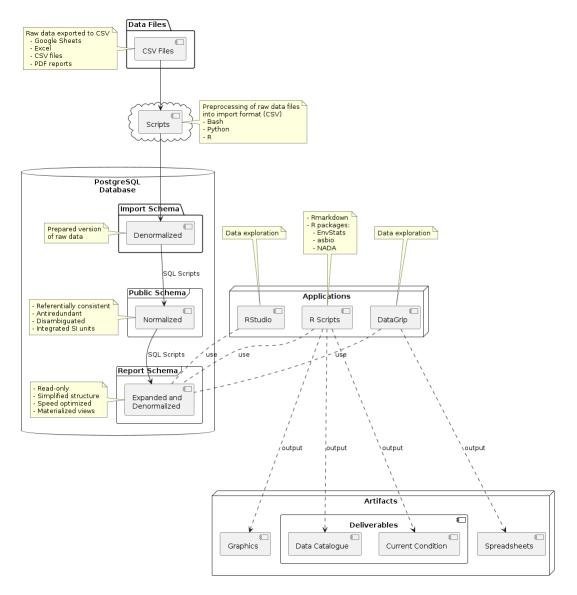


Figure 2.5: High-level data flow used to generate the current conditions.

While only a subset of the compiled water and sediment quality data were used to calculate current conditions (see selection criteria below), all of the compiled data are presented in Appendix A.1 using summary tables and figures.

2.7.2 Treatment of censored data

Water quality datasets often include what is referred to as "censored" data points or nondetects. Censored data are data that are reported as above or below some threshold value, without an actual specific value (Helsel et al., 2020). This usually occurs in water quality data that are reported as below or above a method detection limit. In general, detection limits, sometimes referred to as quantitation limits, refer to the lowest or highest constituent

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concentration that can be accurately measured. This can apply to measures collected using 1106 equipment or sensors in the field, or to laboratory analyses. If a sample is reported as having 1107 a concentration of a certain water quality constituent below a detection limit, then the actual 1108 concentration is somewhere between zero and the detection limit. However, the exact value 1109 is unknown. Dealing with censored data correctly is a very important step in water quality 1110 data analysis, especially when the goal is to characterize the range in values for a parameter 1111 from a dataset that includes censored data points. This is because the value of those censored 1112 data points is unknown, however data analysts will often assign a value to them in order to 1113 facilitate statistical analysis. This results in an estimated value that is usually an overestimate 1114 1115 or underestimate of the real value and, especially where the detection limit is much higher or lower than the real values, the resulting findings and conclusions can be unacceptably 1116 inaccurate. 1117

In this study, censored data are not removed from datasets and they are not substituted with another value before conducting statistical analyses. Instead, censored data points were replaced with the detection limit value or with the highest detection limit value in that compiled dataset (i.e., recensoring), depending on the input requirements of the statistical test conducted (after (Helsel, 2011)). Non-parametric rank-based analysis was used for censored data sets, which does not rely on estimating the actual value of censored data points. Non-parametric statistical analyses are often most appropriate because water and sediment quality data in general and censored data specifically often don't meet the requirements of parametric analysis.

1126 2.7.3 Seasons (high flow, open water, under ice)

In this study, water quality data for the Athabasca River and its Delta as well as Lake 1127 Athabasca are considered in the context of the hydrological seasons outlined in Glozier et 1128 al. (2009). There is significant variation in water quality in the Athabasca River with varia-1129 1130 tion in flow, especially during high flows in spring, in response to storm events during summer and fall, and in the winter under ice. Table 2.3 below outlines the months that are included 1131 in these seasons, along with the season names used by (2009) and in this study. Consultations 1132 1133 with the program manager of the ACFN and MCFN CBM program confirmed that these seasons also reflect seasonal changes in Lake Athabasca, although the specific conditions may not 1134 be the same. 1135

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Table 2.3: Season names

| Months | Season name in Glozier et al | Season name in this study |
|----------------|---------------------------------|------------------------------|
| May-July | Spring/Summer | High Flow |
| August-October | Fall | Open Water |
| November-April | Winter | Under Ice |

2.7.4 Monitoring Location Categories

Water and sediment quality data from the lower Athabasca River, its Delta and Lake Athabasca were assigned to overarching locations, based on these spatial designations. The focal length of the Athabasca River reaches from just upstream (south and west) of the city of Fort McMurray downstream (north) to the separation of the Embarras River from the Athabasca River. This separation also defined the beginning of the Athabasca River Delta, and the focus in this study was the Athabasca River Delta channels. Data from lakes and other rivers and tributaries in the Delta were not included in this study, despite the fact that those aquatic ecosystems have important connections to the channels and the River basin as a whole. Finally, data from Lake Athabasca defined the most downstream (northerly) location category used in this study.

1146 2.7.5 Statistical Methods

In order to characterize water and sediment quality compiled for each study area, the data were first tested for differences across laboratory analysis methods and sampling sites, where more than one method per parameters and multiple sampling sites were included in the data set. Before analysis, censored data points were re-censored to the highest detection limit in the dataset. Then a non-parametric Brunner-Dette-Munk (BDM) test was performed for each water and sediment quality parameter (Helsel et al., 2020). The BDM tests for differences in cumulative distributions between parameter - specific data sets, and does not require that the tested data sets follow a normal distribution or that the compared datasets have equivalent variability (i.e., are 'homoscedastic'). In this case, a two-factor BDM test was conducted to test for differences in distributions between values of the two factors "analysis method" and "sampling site" (Aho 2015; Helsel et al. 2020). The BDM test compares distribution functions, and specifically the frequency of high vs. low values, between data subsets for each identified factor (Helsel et al. 2020). In this study a significant difference was determined where p values <0.05. If a significant difference in data distribution was found according to the analysis method factor, the smaller or less consistent over time data set(s) was removed from

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the analysis, so that only a single method remained. In practice, this situation only occurred in the LTRN water quality data for the Athabasca River Delta current condition calculations.

Data for total dissolved solids (VMV 10451, n=6), manganese (VMV 102089, n=103, and iron (VMV 102090, n=103) were removed in favour of alternative method data with relatively more post-2011 observations. If a difference was found according to sampling site, then the data were separated into site-specific sets for further analysis and reporting. Where no differences were found, data were pooled across methods and/or sites for further analysis.

After data groupings were determined, parameter and season-specific quantiles were calculated and reported, specifically the 5th, 50th, and 95th percentile. These percentiles represent the parameter value at which 95%, 50% and 5% of the parameter data points have a greater value. Therefore, the 5th percentile value indicates a very low parameter value, the 50th percentile the middle or median parameter value, and the 95th percentile a very high parameter value. In other words, these percentiles indicate the lowest, middle and highest parameter values, or a range of 'normal' parameter values, for a given location. The 5th and 95th percentiles are used to define the end values instead of the minimum and maximum values because the latter can include very extreme values registered under exceptional circumstances, and may also include values that reflect errors such as sample contamination or equipment malfunction. Such extreme values will unavoidably be reported in the future, however, they should make up no more than the upper and lower 5% of a data set. Both the lower and upper bounds of parameter value ranges are important because impacts on aquatic ecosystems can occur both where concentrations of constituents are too high or too low (e.g., alkalinity, dissolved oxygen). In addition, the upper and lower bounds of certain parameter values are important in determining the extent to which they modify the toxicity of other constituents (e.g., pH, temperature, dissolved organic carbon). The use of percentiles in water and sediment quality data summaries is common in environmental impact assessments, and the 95th percentile is used to define water quality triggers in the Surface Water Quality Management Framework of the Lower Athabasca Regional Plan (Alberta Environment and Sustainable Resource Development (AESRD), 2012).

For non-censored data sets, a straightforward quantile method was used to determine these percentile values using a "weibull" plotting position approach ("quantile' function in R with type=6, formula (i)/(n+1), where i = rank of observation and n = sample size)(Helsel et al. 2020). For censored data, a robust regression on order statistics (robust ROS) method was used to estimate the 5th, 50th and 95th percentiles, except where the data set size (n) was greater than 50 and the level of data censoring was between 50% and 80%. In the latter case,

a maximum likelihood estimate (MLE) method for censored data was used (after guidance in Bolks, DeWire, and Harcum (2014)). For datasets that were more than 80% censored, no estimation of quantiles was performed. Both the robust ROS and censored MLE methods involve interpolation approaches to estimate quantile values, including below the uncensored detection limit value. In other words, these methods estimate the frequency distribution below (or above, as applicable) the detected data values, usually including the 5th percentile value and, in some cases, the 50th percentile value.

In cases where the censored MLE method was used to estimate quantile values, grouped or non-grouped (as required) parameter data were tested to determine the best-fit distribution from the following possibilities; normal (Gaussian), lognormal, and gamma. This was done by calculating and maximizing a probability plot correlation coefficient (PPCC) for each distribution type after Helsel (2011). If the normal distribution was identified as the best fit, the dataset 5th percentile was examined to determine whether it was non-negative. If it was negative, then the normal distribution was discarded in favour of the next best fit distribution.

2.7.6 Lower Athabasca River Data Selection

This study uses the water quality data collected by the JOSM/OSM programs in the lower Athabasca River using the vertically-integrated-at-the-thalweg field sampling method to char-acterize current water quality in the River. While there was also extensive LTRN and RAMP program data available for water quality in the lower Athabasca River, the sampling method employed by those programs (generally nearshore via wading and often just below the water surface) meant that it was not suitable to be combined with the JOSM/OSM program data (C. Cooke and N. Glozier, pers. comms.). The JOSM/OSM data were favoured in this case because the sampling method used - vertically integrated sampling at the thalweg - was shown to best reflect and encompass the variability in lateral and vertical constituent concentrations, and therefore, to also best approximate and align with constituent loads in the River (Glozier et al., 2018).

The drawback of using the JOSM/OSM water quality data to characterize conditions in the lower Athabasca River is that the data are limited in terms of the period of record, which begins in 2012 and continues up to the most recently available data from 2019. In comparison, the period of record for the two LTRN sites in the lower Athabasca River begins much earlier, in 1987, and continues up to the most recently available data from 2019. The longer period of record for LTRN is a valuable record of conditions over that time period, and would be more amenable to an evaluation of trends over time (N. Glozier, pers. comm.). Therefore, the

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water quality conditions characterized using the JOSM/OSM data reflect recent and current conditions, and not historical conditions such as pre-development or during the increasing levels of anthropogenic and industrial development that occurred prior to 2012.

The analytical methods used in the JOSM/OSM program include two different methods for analysis of total metals or trace elements. These are a 34-element suite that is "in-bottle digest" as well as a 45-element suite referred to as "modified EPA 200.8 ICP-MS." Data from the two different methods are not combinable (N. Glozier, pers. comm.), and therefore data derived using the "in-bottle digest" 34-element suite methods were removed from this analysis. Sediment data for the lower Athabasca mainstem consisted of RAMP and OSM-funded Enhanced Monitoring Program data. The RAMP sediment data were collected from the Athabasca mainstem in the fall over the years 1997 through 2005, with additional limited sampling between 2007 and 2013. The Enhanced Monitoring Program sediment data were collected in the fall of 2018 and 2019 as grab samples from sites along a roughly 60 km river length, centred around a potential future discharge location adjacent to the Syncrude Mildred Lake mine site. In order to align with the time span considered for the Athabasca River water quality analysis, post-2011 data were included in the sediment quality analysis. Where data were obtained using methods that were not appropriate for grouping, the methods with the shortest period of record and/or the smallest sample size were removed from the analysis. For the most part, this meant that the Enhanced Monitoring program data was favoured, due to the much higher number of samples collected in recent years.

1249 2.7.7 Athabasca River Delta Data Selection

The longest water quality data set in the Athabasca River Delta channels is for the provincial 1250 LTRN sites AB07DD0010 and AB07DD0105, also known as Athabasca River at Old Fort and 1251 downstream of Devil's Elbow at Winter Road Crossing, respectively. These sites combined 1252 are the composite "Old Fort" provincial water quality site that serves as the focal point for 1253 the Lower Athabasca Regional Plan (LARP) Surface Water Quality Management Framework. 1254 Several of the methods used by the LTRN and by the MCFN and ACFN CBM programs 1255 1256 to measure the same parameter were not compatible for grouping, and many of the multiple methods used over time within the LTRN program were also not combinable. Given the longer 1257 period of record, more frequent sampling, and larger number of parameters measured, the 1258 1259 LTRN data was used for this analysis. The LTRN data set was truncated to include only post-2011 data in the analysis, since several analytical methods for multiple parameters were 1260 changed between the years 2008 and 2010 and were not combinable. 1261

Sediment quality data were available from the RAMP program for the Athabasca River

Delta. Those data were collected in the fall between 2000 and 2016, and the analytical methods

used were consistent over time.

1265 2.7.8 Lake Athabasca Data Selection

The longest water quality dataset in Lake Athabasca is for sites from the ACFN and MCFN 1266 CBM programs. Data from the two sites, near the Fort Chipewyan water intake and at the 1267 Dock site, have been collected about four times a year since 2011. The available provincial 1268 water quality data for Lake Athabasca didn't generally consist of long-term data sets, but did 1269 include data from eight locations on the lake. In addition, while the CBM data is relatively 1270 recent, the provincial LWQ data is strictly more historical, collected between the late 1980's 1271 and early 1990's. For both the ACFN and MCFN CBM programs, the sampling and analytical 1272 1273 methods used were the same, and in particular the field-measured parameter data are average values from water column profile data taken at 1m intervals. Given that it is a long-term and 1274 recent dataset, the ACFN MCFN CBM data were used to calculate current conditions in Lake 1275 1276 Athabasca. There were no sediment quality data obtained for Lake Athabasca from the monitoring 1277

$_{79}$ 2.8 Results

programs surveyed in this study.

1280 2.8.1 Lower Athabasca River Current Conditions

The current condition (5th, 50th, and 95th percentile values) for each water and sediment quality parameter and each season are presented for the lower Athabasca River in Table 2.4 (water) and Table 2.5 (sediment). Note that additional information, including sample size, analytical method codes, and quantile estimation method for each suite of current conditions are provided in Appendix A.2.

Table 2.4: Current Conditions, Athabasca River water.

| | | | | High Flow | | (| Open Wat | er | Under Ice | | | |
|--|-----------------|------------|--------|-----------|--------|--------|----------|--------|-----------|--------|------|--|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95t | |
| entional Variables | | | | | | | | | | | | |
| Alkalinity, Phenolphthalein (total hydroxide+1/2 carbonate) as CaCO3 | mg/L | all sites | - | - | - | 1.00 | 6.40 | 7.06 | - | - | | |
| Alkalinity, total as CaCO3 | mg/L | all sites | 61.05 | 89.00 | 99.09 | 81.54 | 101.00 | 122.00 | + | + | | |
| | mg/L | AL07DD0004 | + | + | + | + | + | + | - | - | | |
| | mg/L | AL07DD0005 | + | + | + | + | + | + | - | - | | |
| | mg/L | AL07DD0007 | + | + | + | + | + | + | 133.00 | 147.00 | 165. | |
| | mg/L | AL07DD0008 | + | + | + | + | + | + | 89.00 | 163.00 | 199. | |
| | mg/L | AL07DD0009 | + | + | + | + | + | + | - | - | | |
| Fixed suspended solids, Non-Filterable (Particle) | mg/L | all sites | 30.50 | 166.00 | 661.80 | 3.95 | 20.40 | 125.70 | < | < | | |
| Organic carbon, Filtered | mg/L | all sites | 3.53 | 12.20 | 16.36 | 4.24 | 7.90 | 17.50 | 5.49 | 7.43 | 10 | |
| Organic carbon, Non-Filterable (Particle) | $\mathrm{mg/L}$ | all sites | 1.23 | 4.01 | 13.17 | 0.39 | 0.98 | 5.07 | 0.09 | 0.23 | 0 | |
| Specific conductivity | uS/cm | all sites | 160.90 | 216.00 | 263.10 | 213.20 | 266.00 | 322.20 | 318.85 | 409.50 | 484 | |
| Total suspended solids, Non-Filterable (Particle) | $\mathrm{mg/L}$ | all sites | 37.04 | 183.00 | 719.90 | 9.64 | 24.00 | 141.50 | < | < | | |
| True colour, Filtered | TCU | all sites | - | - | - | - | - | - | - | - | | |
| True colour, Supernate | rel units | all sites | 5.00 | 60.00 | 98.25 | 6.00 | 25.00 | 88.00 | 5.00 | 15.00 | 35 | |
| Turbidity | NTU | all sites | 18.49 | 69.00 | 219.00 | 5.28 | 12.20 | 95.20 | 1.84 | 3.65 | 6 | |
| pH, lab | pH units | all sites | 7.79 | 8.09 | 8.32 | 7.94 | 8.22 | 8.38 | 7.65 | 7.84 | 8 | |
| olved Metals | | | | | | | | | | | | |
| Aluminum, Filtered | ug/L | all sites | 7.68 | 32.35 | 117.90 | 5.06 | 16.00 | 56.68 | 3.83 | 13.20 | 28 | |
| Antimony, Filtered | ug/L | all sites | 0.04 | 0.07 | 0.12 | 0.03 | 0.05 | 0.11 | + | + | | |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | ${\bf High\ Flow}$ | | C | pen Wate | er | Under Ice | | |
|---------------------|----------------------------|------------|-------|--------------------|-------|-------|------------------|-------|-----------|-------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95t |
| | $\mathrm{ug/L}$ | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | -ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.04 | 0.06 | 0. |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.02 | 0.05 | 0. |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Arsenic, Filtered | ug/L | all sites | 0.37 | 0.55 | 0.81 | 0.36 | 0.49 | 0.73 | 0.32 | 0.46 | 0. |
| Barium, Filtered | ug/L | all sites | 24.52 | 43.75 | 55.41 | 27.22 | 49.10 | 63.38 | + | + | |
| | $_{ m ug/L}$ | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | $\overline{\mathrm{ug/L}}$ | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | $\overline{\mathrm{ug/L}}$ | AL07DD0007 | + | + | + | + | + | + | 62.30 | 71.90 | 79 |
| | $\mathrm{ug/L}$ | AL07DD0008 | + | + | + | + | + | + | 24.90 | 86.65 | 109 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Beryllium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0 |
| Bismuth, Filtered | ug/L | all sites | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | < | < | |
| Boron, Filtered | ug/L | all sites | 12.84 | 21.60 | 30.28 | 15.18 | 23.30 | 31.22 | 30.39 | 36.35 | 41 |
| Cadmium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0 |
| Cerium, Filtered | ug/L | all sites | 0.04 | 0.18 | 0.60 | 0.02 | 0.07 | 0.27 | 0.02 | 0.06 | 0 |
| Cesium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0 |
| Chromium, Filtered | $\mathrm{ug/L}$ | all sites | 0.05 | 0.10 | 0.25 | 0.03 | 0.06 | 0.14 | 0.06 | 0.08 | 0 |
| Cobalt, Filtered | $\mathrm{ug/L}$ | all sites | 0.04 | 0.07 | 0.17 | 0.04 | 0.08 | 0.12 | + | + | |
| | $_{ m ug/L}$ | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | $_{ m ug/L}$ | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.04 | 0.06 | 0 |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | Open Water | | | Under Ice | | |
|----------------------|----------------------------|------------|-------|-----------|--------|------------|--------|--------|-----------|------------------|-------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95t |
| | -ug/L | AL07DD0008 | + | + | + | + | + | + | 0.04 | 0.05 | 0.0 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Copper, Filtered | ug/L | all sites | 0.62 | 1.28 | 2.41 | 0.42 | 0.66 | 1.56 | + | + | - |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.28 | 0.58 | 0.9 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.31 | 0.56 | 1.2 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Gallium, Filtered | ug/L | all sites | 0.01 | 0.02 | 0.04 | 0.00 | 0.01 | 0.06 | 0.00 | 0.01 | 0.0 |
| Germanium, Filtered | ug/L | all sites | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | + | + | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.01 | 0.01 | 0.0 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.01 | 0.01 | 0.0 |
| | $\overline{\mathrm{ug/L}}$ | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Indium, Filtered | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Iron, Filtered | ug/L | all sites | 22.64 | 190.50 | 572.75 | 37.76 | 157.00 | 445.60 | 72.11 | 255.00 | 563.5 |
| Lanthanum, Filtered | ug/L | all sites | 0.02 | 0.10 | 0.28 | 0.01 | 0.04 | 0.15 | 0.01 | 0.03 | 0.0 |
| Lead, Filtered | ug/L | all sites | 0.02 | 0.09 | 0.30 | 0.01 | 0.04 | 0.13 | 0.02 | 0.03 | 0.0 |
| Lithium, Filtered | ug/L | all sites | 3.98 | 5.39 | 7.37 | 4.80 | 6.03 | 8.58 | 7.96 | 9.98 | 11.5 |
| Manganese, Filtered | ug/L | all sites | 0.58 | 2.71 | 5.57 | 0.71 | 2.06 | 5.84 | 2.20 | 7.91 | 12.0 |
| Molybdenum, Filtered | ug/L | all sites | + | + | + | 0.33 | 0.69 | 0.91 | + | + | |
| | ug/L | AL07DD0004 | 0.40 | 0.59 | 2.88 | + | + | + | - | - | |
| | ug/L | AL07DD0005 | 0.50 | 0.63 | 0.73 | + | + | + | - | - | |
| | | | | | | | | | | | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | (| Open Wate | er | Under Ice | | |
|---------------------|--------------|------------|-------|-----------|--------|--------|------------------|--------|-----------|--------|-------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95t |
| | -ug/L | AL07DD0008 | 0.26 | 0.53 | 0.81 | + | + | + | 0.23 | 0.89 | 1.1 |
| | $_{ m ug/L}$ | AL07DD0009 | - | - | - | + | + | + | - | - | |
| Nickel, Filtered | ug/L | all sites | 0.74 | 1.38 | 2.52 | 0.68 | 0.91 | 1.74 | 0.49 | 0.94 | 1.4 |
| Niobium, Filtered | ug/L | all sites | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.0 |
| Palladium, Filtered | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Platinum, Filtered | ug/L | all sites | < | < | < | < | < | < | < | < | < |
| Rubidium, Filtered | ug/L | all sites | 0.56 | 0.89 | 1.16 | 0.68 | 0.84 | 0.98 | 1.07 | 1.44 | 1.9 |
| Scandium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.14 | 0.00 | 0.01 | 0.06 | 0.00 | 0.01 | 0.0 |
| Selenium, Filtered | ug/L | all sites | 0.07 | 0.15 | 0.22 | 0.08 | 0.12 | 0.17 | + | + | - |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.11 | 0.16 | 0.2 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.05 | 0.20 | 0.3 |
| | $_{ m ug/L}$ | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Silver, Filtered | ug/L | all sites | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 |
| Strontium, Filtered | ug/L | all sites | 81.89 | 170.00 | 241.05 | 123.20 | 226.00 | 303.60 | + | + | - |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 278.00 | 322.00 | 388.0 |
| | $_{ m ug/L}$ | AL07DD0008 | + | + | + | + | + | + | 134.00 | 364.00 | 489.0 |
| | $_{ m ug/L}$ | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Tellurium, Filtered | ug/L | all sites | 0.01 | 0.01 | 0.01 | < | < | < | + | + | - |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | $_{ m ug/L}$ | AL07DD0007 | + | + | + | + | + | + | 0.01 | 0.01 | 0.0 |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | <u> </u> | | Open Water | | | Under Ice | | |
|-----------------------|----------------------------|------------|--------|------------------|----------|--------|------------------|--------|-------|-----------|-----|--|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95 | |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.00 | 0.00 | 0.0 | |
| | $_{ m ug/L}$ | AL07DD0009 | + | + | + | + | + | + | - | - | | |
| Thallium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0. | |
| Tin, Filtered | $\mathrm{ug/L}$ | all sites | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 | 0.05 | 0.00 | 0.01 | 0. | |
| Titanium, Filtered | ug/L | all sites | 0.10 | 1.00 | 4.54 | 0.10 | 0.50 | 1.50 | 0.10 | 0.50 | 1. | |
| Tungsten, Filtered | ug/L | all sites | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0. | |
| Uranium, Filtered | ug/L | all sites | 0.13 | 0.34 | 0.48 | 0.14 | 0.36 | 0.48 | + | + | | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.40 | 0.45 | 0 | |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.10 | 0.57 | 0 | |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | | |
| Vanadium, Filtered | ug/L | all sites | 0.21 | 0.39 | 0.74 | 0.15 | 0.31 | 0.64 | 0.13 | 0.20 | 0 | |
| Yttrium, Filtered | ug/L | all sites | 0.05 | 0.18 | 0.42 | 0.04 | 0.08 | 0.26 | 0.05 | 0.07 | 0 | |
| Zinc, Filtered | ug/L | all sites | 0.27 | 0.60 | 2.15 | 0.16 | 0.40 | 1.20 | + | + | | |
| | $\overline{\mathrm{ug/L}}$ | AL07DD0004 | + | + | + | + | + | + | - | - | | |
| | $\overline{\mathrm{ug/L}}$ | AL07DD0005 | + | + | + | + | + | + | - | - | | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.60 | 1.30 | 3 | |
| | $_{ m ug/L}$ | AL07DD0008 | + | + | + | + | + | + | 0.60 | 1.30 | 3 | |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | | |
| Zirconium, Filtered | $\mathrm{ug/L}$ | all sites | 0.08 | 0.20 | 0.50 | 0.05 | 0.10 | 0.30 | 0.07 | 0.10 | 0 | |
| | | | | | | | | | | | | |
| Dissolved oxygen (DO) | $\mathrm{mg/L}$ | all sites | 8.15 | 8.72 | 10.75 | 8.07 | 9.86 | 13.01 | 11.54 | 12.39 | 13. | |
| Specific conductivity | uS/cm | all sites | 153.70 | 222.00 | 269.35 | 225.20 | 268.00 | 319.40 | + | + | | |
| | uS/cm | AL07DD0004 | + | + | + | + | + | + | - | - | | |

 $_{\infty}^{\infty}$

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | Open Water | | | Under Ice | | |
|-------------------------|-----------------|------------|-------|-----------|--------|------------|------------------|-------|-----------|--------|-------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95t |
| | uS/cm | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | uS/cm | AL07DD0007 | + | + | + | + | + | + | 373.00 | 417.00 | 484.0 |
| | uS/cm | AL07DD0008 | + | + | + | + | + | + | 266.00 | 432.00 | 521.0 |
| | uS/cm | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Temperature, water | $\deg C$ | all sites | 10.46 | 18.79 | 22.14 | 2.44 | 12.68 | 22.62 | + | + | |
| | $_{\rm degC}$ | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | $_{\rm degC}$ | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | $_{\rm degC}$ | AL07DD0007 | + | + | + | + | + | + | -0.32 | -0.13 | -0. |
| | $_{\rm degC}$ | AL07DD0008 | + | + | + | + | + | + | -0.80 | -0.25 | -0. |
| | $_{\rm degC}$ | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Turbidity | NTU | all sites | 20.25 | 64.65 | 321.95 | 2.43 | 12.15 | 71.75 | 0.00 | 1.50 | 101. |
| рН | pH units | all sites | 7.74 | 7.97 | 8.29 | 7.83 | 8.20 | 8.41 | 7.06 | 7.51 | 8. |
| ral Organics Benzene | $\mathrm{ug/L}$ | all sites | < | < | < | _ | _ | _ | < | < | |
| C10-C16 Hydrocarbons | ug/L | all sites | 23.15 | 52.59 | 133.06 | < | < | < | < | < | |
| C16-C34 Hydrocarbons | ug/L | all sites | < | < | < | < | < | < | < | < | |
| C34-C50 Hydrocarbons | ug/L | all sites | < | < | < | < | < | < | < | < | |
| C6-C10 Hydrocarbons | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Cyanide | $\mathrm{mg/L}$ | all sites | < | < | < | < | < | < | < | < | |
| Ethylbenzene | ug/L | all sites | < | < | < | - | - | - | < | < | |
| Hydrocarbons, petroleum | $\mathrm{mg/L}$ | all sites | 0.02 | 0.08 | 0.40 | < | < | < | < | < | |
| Naphthenic acids | $\mathrm{mg/L}$ | all sites | < | < | < | < | < | < | < | < | |
| Toluene | ug/L | all sites | + | + | + | 0.01 | 0.03 | 0.14 | < | < | |
| | ug/L | AL07DD0004 | - | - | - | + | + | + | + | + | |
| | ug/L | AL07DD0005 | - | - | - | + | + | + | + | + | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | C | pen Wate | er | Under Ice | | |
|--------------------|-----------------|------------|-------|-----------|-------|-------|------------------|-------|-----------|-------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95t |
| | -ug/L | AL07DD0007 | - | - | - | + | + | + | + | + | - |
| | ug/L | AL07DD0008 | < | < | < | + | + | + | + | + | |
| | ug/L | AL07DD0009 | - | - | - | + | + | + | + | + | |
| m,p-Xylene | ug/L | all sites | < | < | < | - | - | - | < | < | |
| o-Xylene | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Iajor Ions | | | | | | | | | | | |
| Calcium, Filtered | $_{ m mg/L}$ | all sites | + | + | + | 23.47 | 32.15 | 38.89 | 24.26 | 43.20 | 57. |
| | $_{ m mg/L}$ | AL07DD0004 | - | - | - | + | + | + | + | + | |
| | $\mathrm{mg/L}$ | AL07DD0005 | - | - | - | + | + | + | + | + | |
| | $_{ m mg/L}$ | AL07DD0007 | - | - | - | + | + | + | + | + | |
| | $_{ m mg/L}$ | AL07DD0008 | 15.80 | 23.15 | 33.20 | + | + | + | + | + | |
| | $_{ m mg/L}$ | AL07DD0009 | - | - | - | + | + | + | + | + | |
| Calcium, Unknown | $\mathrm{mg/L}$ | all sites | 22.40 | 27.10 | 29.80 | 19.80 | 32.00 | 36.00 | 26.10 | 38.40 | 48 |
| Chloride, Filtered | $\mathrm{mg/L}$ | all sites | 1.15 | 4.52 | 12.93 | 1.52 | 8.13 | 18.04 | + | + | |
| | $_{ m mg/L}$ | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | $_{ m mg/L}$ | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | $_{ m mg/L}$ | AL07DD0007 | + | + | + | + | + | + | 14.70 | 17.90 | 24. |
| | $_{ m mg/L}$ | AL07DD0008 | + | + | + | + | + | + | 5.38 | 13.16 | 36. |
| | $_{ m mg/L}$ | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Fluoride, Filtered | $\mathrm{mg/L}$ | all sites | + | + | + | 0.06 | 0.09 | 0.11 | + | + | |
| | $_{ m mg/L}$ | AL07DD0004 | 0.07 | 0.09 | 0.09 | + | + | + | - | - | |
| | $_{ m mg/L}$ | AL07DD0005 | 0.06 | 0.09 | 0.09 | + | + | + | - | - | |
| | $_{ m mg/L}$ | AL07DD0007 | 0.08 | 0.09 | 0.10 | + | + | + | 0.10 | 0.11 | 0 |
| | $_{ m mg/L}$ | AL07DD0008 | 0.07 | 0.08 | 0.09 | + | + | + | 0.09 | 0.11 | 0. |
| | | | | | | | | | | | |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | ${\bf High\ Flow}$ | | C | pen Wate | er | Under Ice | | |
|--|--------------------|------------|-------|--------------------|-------|----------------|------------------|-------|-----------|------------------|------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | $5\mathrm{th}$ | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95t |
| | $\mathrm{mg/L}$ | AL07DD0009 | - | - | - | + | + | + | - | - | |
| Magnesium, Filtered | mg/L | all sites | + | + | + | 6.73 | 8.55 | 11.40 | + | + | |
| | $_{ m mg/L}$ | AL07DD0004 | 4.76 | 7.13 | 8.55 | + | + | + | - | - | |
| | $_{\mathrm{mg/L}}$ | AL07DD0005 | 5.59 | 6.97 | 7.84 | + | + | + | - | - | |
| | $_{ m mg/L}$ | AL07DD0007 | 6.73 | 8.32 | 9.40 | + | + | + | 10.10 | 12.30 | 14.0 |
| | $_{\mathrm{mg/L}}$ | AL07DD0008 | 4.29 | 6.48 | 9.35 | + | + | + | 7.08 | 13.35 | 17.1 |
| | mg/L | AL07DD0009 | - | - | - | + | + | + | - | - | |
| Potassium, Filtered | mg/L | all sites | 0.79 | 1.03 | 1.75 | 0.95 | 1.11 | 1.41 | 1.27 | 2.03 | 2.5 |
| Silica, Filtered as SiO2 | $\mathrm{mg/L}$ | all sites | 3.06 | 5.89 | 9.02 | 1.92 | 4.51 | 7.91 | 5.63 | 8.85 | 12. |
| Silica, Unknown as SiO2 | mg/L | all sites | 4.63 | 5.39 | 6.62 | 3.71 | 5.74 | 8.40 | 7.88 | 9.17 | 11.5 |
| Sodium, Filtered | $\mathrm{mg/L}$ | all sites | 6.12 | 8.63 | 13.06 | 6.99 | 12.20 | 18.22 | 21.49 | 27.80 | 32. |
| Sulfate, Filtered as SO4 | mg/L | all sites | + | + | + | 9.67 | 24.00 | 37.26 | + | + | |
| | mg/L | AL07DD0004 | 9.91 | 16.60 | 24.10 | + | + | + | - | - | |
| | mg/L | AL07DD0005 | 10.60 | 17.00 | 20.70 | + | + | + | - | - | |
| | mg/L | AL07DD0007 | 15.60 | 21.75 | 29.00 | + | + | + | 31.50 | 38.70 | 52. |
| | mg/L | AL07DD0008 | 6.61 | 13.20 | 30.40 | + | + | + | 11.60 | 44.05 | 65.5 |
| | mg/L | AL07DD0009 | - | - | - | + | + | + | - | - | |
| Ammonia and ammonium, Unfiltered as N | mg/L | all sites | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 | 0.02 | 0.02 | 0.05 | 0.0 |
| Inorganic nitrogen (nitrate and nitrite), Filtered | $\mathrm{mg/L}$ | all sites | 0.01 | 0.03 | 0.07 | 0.00 | 0.01 | 0.03 | + | + | |
| | mg/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | $_{\mathrm{mg/L}}$ | AL07DD0007 | + | + | + | + | + | + | 0.21 | 0.26 | 0. |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | C | pen Wate | r | J | Jnder Ice | |
|--|-----------------|------------|------|------------------|------|------|----------|------|------|------------------|-----|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 951 |
| | $\mathrm{mg/L}$ | AL07DD0008 | + | + | + | + | + | + | 0.18 | 0.22 | 0. |
| | mg/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Organic Nitrogen, Non-Filterable (Particle) as N | $\mathrm{mg/L}$ | all sites | 0.11 | 0.31 | 1.00 | 0.03 | 0.11 | 0.31 | + | + | |
| | mg/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0007 | + | + | + | + | + | + | 0.01 | 0.02 | 0 |
| | mg/L | AL07DD0008 | + | + | + | + | + | + | 0.01 | 0.02 | C |
| | mg/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Total Nitrogen, mixed forms, Filtered as N | mg/L | all sites | 0.12 | 0.30 | 0.61 | 0.11 | 0.22 | 0.62 | 0.39 | 0.53 | (|
| Total Nitrogen, mixed forms, Non-Filterable (Particle) as N | mg/L | all sites | - | - | - | 0.07 | 0.10 | 0.47 | - | - | |
| Total Nitrogen, mixed forms, Unknown as N | $\mathrm{mg/L}$ | all sites | 0.29 | 0.45 | 0.59 | 0.22 | 0.34 | 0.52 | + | + | |
| | mg/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0007 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0008 | + | + | + | + | + | + | - | - | |
| | mg/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Total Phosphorus, mixed forms, Filtered as P | $\mathrm{mg/L}$ | all sites | 0.01 | 0.02 | 0.03 | 0.00 | 0.01 | 0.03 | 0.01 | 0.02 | (|
| Total Phosphorus, mixed forms, Unfiltered as P | mg/L | all sites | 0.05 | 0.19 | 0.58 | 0.02 | 0.05 | 0.19 | 0.02 | 0.04 | (|
| nohalides | | | | | | | | | | | |
| 2-Chloronaphthalene | ng/L | AL07DD0004 | < | < | < | - | - | - | - | - | |
| | ng/L | AL07DD0005 | - | - | - | - | - | - | - | - | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | O | pen Wate | r | Ţ | Under Ice | |
|-------------------------------|-----------------|------------|------|-----------|-------|------|----------|------|------|------------------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95t |
| | $\mathrm{ng/L}$ | AL07DD0007 | - | - | - | - | - | - | - | - | |
| | $_{\rm ng/L}$ | AL07DD0008 | - | - | - | - | - | - | - | - | |
| | ng/L | AL07DD0009 | - | - | - | - | - | - | - | - | |
| Hs | | | | | | | | | | | |
| 1,2,3,4-Tetrahydronaphthalene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| 1,6,7-Trimethylnaphthalene | ng/L | all sites | 0.46 | 1.64 | 4.15 | 0.35 | 1.00 | 3.11 | 0.11 | 0.43 | 2. |
| 1-Methylnaphthalene | ng/L | all sites | 1.17 | 4.70 | 18.66 | < | < | < | < | < | |
| 2-Isopropylnaphthalene | ng/L | all sites | < | < | < | < | < | < | - | - | |
| 2-Methylnaphthalene | $\mathrm{ng/L}$ | all sites | 2.48 | 9.19 | 35.30 | < | < | < | < | < | |
| 3-Methylcholanthrene | $\mathrm{ng/L}$ | all sites | 1.24 | 4.26 | 13.78 | 0.13 | 0.52 | 2.49 | < | < | |
| 7,10-Dimethylbenzo[a]pyrene | ng/L | all sites | < | < | < | < | < | < | - | - | |
| 7-Methylbenzo[a]pyrene | ${ m ng/L}$ | all sites | < | < | < | < | < | < | - | - | |
| 9-Ethylfluorene | ng/L | all sites | < | < | < | < | < | < | - | - | |
| 9-Methylfluorene | ${ m ng/L}$ | all sites | 0.10 | 0.56 | 3.92 | < | < | < | < | < | |
| Acenaphthene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Acenaphthylene | $\mathrm{ng/L}$ | AL07DD0004 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0005 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0007 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0008 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0009 | - | - | - | - | - | - | - | - | |
| Anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Benz[a]anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | C | pen Wate | r | J | Jnder Ice | |
|------------------------------|------|------------|-------|------------------|-------|----------------|----------|-------|----------------|------------------|------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | $50 \mathrm{th}$ | 95t] |
| Benzo(b)fluoranthene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Benzo[a]pyrene | ng/L | all sites | < | < | < | < | < | < | < | < | 4 |
| Benzo[e]pyrene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Benzo[ghi]perylene | ng/L | AL07DD0004 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0005 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0007 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0008 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0009 | - | - | - | - | - | - | - | - | |
| Benzo[k]fluoranthene | ng/L | AL07DD0004 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0005 | < | < | < | < | < | < | - | - | |
| | ng/L | AL07DD0007 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0008 | < | < | < | < | < | < | < | < | |
| | ng/L | AL07DD0009 | - | - | - | - | - | - | - | - | |
| Biphenyl | ng/L | all sites | - | - | - | - | - | - | - | - | |
| C1-Dibenzothiophenes | ng/L | all sites | - | - | - | - | - | - | - | - | |
| C1-Fluoranthenes/pyrenes | ng/L | all sites | 23.36 | 30.50 | 45.02 | - | - | - | - | - | |
| C2-1,6-Dimethylnaphthalene | ng/L | all sites | 4.48 | 6.21 | 27.16 | 0.50 | 1.89 | 8.97 | 1.05 | 2.23 | 5. |
| C2-1,9-Dimethylfluorene | ng/L | all sites | 0.07 | 0.42 | 3.40 | < | < | < | - | - | |
| C2-3-Ethylfluoranthene | ng/L | all sites | < | < | < | < | < | < | - | - | |
| C2-Benzopyrenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C2-Chrysenes | ng/L | all sites | 4.13 | 7.42 | 14.61 | < | < | < | < | < | |
| C2-Dibenzothiophenes | ng/L | all sites | 6.26 | 21.00 | 50.82 | - | - | - | - | - | |
| C2-Dimethyldibenzothiophenes | ng/L | all sites | 3.95 | 16.56 | 60.42 | 0.32 | 1.70 | 26.69 | 0.39 | 0.75 | 2. |
| C2-Fluoranthenes/pyrenes | ng/L | all sites | 5.39 | 6.87 | 9.07 | < | < | < | < | < | |
| C2-Fluorenes | ng/L | all sites | 14.00 | 21.90 | 50.10 | - | - | - | - | - | |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | O | pen Wate | er | J | Jnder Ice | |
|------------------------------------|------|------------|-------|------------------|-------|------|----------|-------|----------------|-----------|-----|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95t |
| C2-Naphthalenes | ng/L | all sites | - | - | - | - | - | - | - | - | |
| C2-Phenanthrenes | ng/L | all sites | 7.91 | 26.20 | 85.24 | 0.09 | 1.44 | 29.99 | - | - | |
| C3-2,4,7-Trimethyldibenzothiophene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C3-4-Propyldibenzothiophene | ng/L | all sites | 0.07 | 0.45 | 3.73 | < | < | < | < | < | |
| C3-Chrysenes | ng/L | all sites | 9.57 | 10.60 | 11.90 | - | - | - | - | - | |
| C3-Dibenzothiophenes | ng/L | all sites | 16.40 | 18.50 | 27.50 | - | - | - | - | - | |
| C3-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C3-Fluorenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C3-N-Propylfluorene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C3-Naphthalenes | ng/L | all sites | 5.53 | 15.23 | 50.65 | < | < | < | < | < | |
| C3-Phenanthrenes | ng/L | all sites | 5.99 | 15.65 | 49.18 | - | - | - | - | - | |
| C4-Chrysenes | ng/L | all sites | 11.58 | 12.65 | 13.84 | - | - | - | - | - | |
| C4-Dibenzothiophenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C4-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C4-Fluorenes | ng/L | all sites | < | < | < | < | < | < | < | < | |
| C4-Naphthalenes | ng/L | all sites | 11.51 | 22.00 | 39.20 | - | - | - | - | - | |
| C4-Phenanthrenes | ng/L | all sites | + | + | + | < | < | < | < | < | |
| | ng/L | AL07DD0004 | - | - | - | + | + | + | + | + | |
| | ng/L | AL07DD0005 | 4.66 | 8.95 | 14.55 | + | + | + | + | + | |
| | ng/L | AL07DD0007 | - | - | - | + | + | + | + | + | |
| | ng/L | AL07DD0008 | - | - | - | + | + | + | + | + | |
| | ng/L | AL07DD0009 | - | | - | + | + | + | + | + | |
| Chrysene | ng/L | all sites | 0.36 | 2.51 | 23.46 | - | - | - | - | - | |
| Dibenz[a,h]anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Dibenzothiophene | ng/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flov | V | (| Open Wat | er | | Under Ice | 9 |
|-------------------------------|---|---------------------|---|--|--|--|--|--|--|---|--|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | $95 \mathrm{th}$ |
| Fluoranthene | ng/L | all sites | 0.67 | 2.14 | 7.11 | < | < | < | < | < | < |
| Fluorene | ng/L | all sites | - | - | - | - | - | - | - | - | - |
| Indene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Indeno[1,2,3-cd]fluoranthene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Indeno[1,2,3-cd]pyrene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Methylbenzopyrene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Methylchrysene | ng/L | all sites | 37.07 | 59.20 | 91.20 | < | < | < | - | - | - |
| Methyldibenzothiophene | ng/L | all sites | 1.52 | 3.55 | 17.76 | 0.24 | 0.93 | 4.47 | 0.30 | 0.82 | 2.60 |
| Methylfluoranthene | ng/L | all sites | 4.24 | 7.70 | 30.77 | 0.18 | 1.17 | 7.91 | < | < | < |
| Methylfluorene | ng/L | all sites | 14.61 | 30.30 | 57.48 | - | - | - | - | - | - |
| Methylnaphthalene | ng/L | all sites | 19.11 | 48.03 | 148.13 | - | - | _ | - | - | - |
| Methylphenanthrene | ng/L | all sites | 6.21 | 30.20 | 110.19 | < | < | < | - | - | - |
| Naphthalene | ng/L | all sites | 3.16 | 23.78 | 251.85 | 11.84 | 43.05 | 123.20 | 4.51 | 26.65 | 200.50 |
| Perylene | ng/L | all sites | 1.59 | 9.09 | 71.88 | < | < | < | < | < | < |
| Phenanthrene | ng/L | all sites | 2.95 | 10.64 | 34.80 | < | < | < | - | - | - |
| Pyrene | ng/L | all sites | 0.67 | 3.34 | 24.60 | < | < | < | < | < | < |
| Retene | ng/L | all sites | 1.86 | 10.25 | 67.50 | < | < | < | < | < | < |
| lics | | | | | | | | | | | |
| Phenol | ug/L | all sites | < | < | < | < | < | < | < | < | < |
| PANHs | | | | | | | | | | | |
| Acridine | | | < | < | < | < | < | < | < | < | < |
| Carbazole | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Metals Aluminum Unfiltered | 110 /L | all sites | 142.40 | 2530 00 | 8576 00 | 110.89 | 316.00 | 3154 00 | 15 18 | 54 00 | 127.85 |
| Antimony, Unfiltered | ug/L | all sites | 0.05 | 0.11 | 0.20 | 0.02 | 0.06 | 0.15 | 0.01 | 0.06 | 0.09 |
| | Fluoranthene Fluorene Indene Indeno[1,2,3-cd]fluoranthene Indeno[1,2,3-cd]pyrene Methylbenzopyrene Methylchrysene Methyldibenzothiophene Methylfluoranthene Methylfluoranthene Methylphenanthrene Methylphenanthrene Perylene Phenanthrene Phenanthrene Retene Iics Phenol PANHs Acridine Carbazole Metals Aluminum, Unfiltered | Fluoranthene ng/L | Fluoranthene ng/L all sites Fluorene ng/L all sites Indene ng/L all sites Indeno[1,2,3-cd]fluoranthene ng/L all sites Indeno[1,2,3-cd]pyrene ng/L all sites Indeno[1,2,3-cd]pyrene ng/L all sites Methylbenzopyrene ng/L all sites Methylchrysene ng/L all sites Methyldibenzothiophene ng/L all sites Methylfluoranthene ng/L all sites Methylfluoranthene ng/L all sites Methylfluorene ng/L all sites Methylphenanthrene ng/L all sites Methylphenanthrene ng/L all sites Methylphenanthrene ng/L all sites Perylene ng/L all sites Phenanthrene ng/L all sites Phenanthrene ng/L all sites Phenanthrene ng/L all sites Phenol ug/L all sites Methylphenanthrene ng/L all sites Methylphenanthrene ng/L all sites Phenol ug/L all sites Methylphenanthrene ng/L all sites Methylphenanthrene ng/L all sites Phenol ug/L all sites Methylphenanthrene ng/L all sites Metals Aluminum, Unfiltered ug/L all sites | Fluoranthene ng/L all sites 0.67 Fluorene ng/L all sites - 1 Indene ng/L all sites - 2 Indene ng/L all sites < 2 Indeno[1,2,3-cd]fluoranthene ng/L all sites < 3 Indeno[1,2,3-cd]pyrene ng/L all sites 3 Indeno[1,2,3-cd]fluoranthene ng/L all sites 1 Indeno[1,2, | Fluoranthene ng/L all sites 0.67 2.14 Fluorene ng/L all sites - - Indene ng/L all sites - Indene ng/L all sites < < Indene ng/L all sites 37.07 59.20 Indene ng/L all sites 31.52 3.55 Indene ng/L all sites 31.61 30.30 Indene ng/L all sites 31.61 Indene ng/L all sites | Fluoranthene ng/L all sites 0.67 2.14 7.11 Fluorene ng/L all sites | Fluoranthene ng/L all sites 0.67 2.14 7.11 < Fluorene ng/L all sites | Fluoranthene ng/L all sites 0.67 2.14 7.11 < < < < < < < < < | Fluoranthene ng/L all sites 0.67 2.14 7.11 < < < < < < < < < | Phoranthene ng/L all sites 0.67 2.14 7.11 < < < < < < < < < | Fluoranthene ng/L all sites 0.67 2.14 7.11 < |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flo | W | (| Open Wat | er | | Under Ice | 9 |
|-----------------------|-----------------|------------|--------|------------------|----------|--------|-----------------|---------|--------|------------------|-------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | $50\mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95tł |
| Arsenic, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.64 | 1.98 | 5.43 | 0.50 | 0.71 | 2.63 | 0.38 | 0.56 | 0.7 |
| Barium, Unfiltered | ug/L | all sites | 48.02 | 73.80 | 174.00 | 34.70 | 53.70 | 104.24 | + | + | + |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 63.30 | 69.50 | 79.3 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 26.00 | 85.20 | 107.0 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Beryllium, Unfiltered | ug/L | all sites | 0.03 | 0.14 | 0.46 | 0.01 | 0.02 | 0.17 | 0.00 | 0.01 | 0.03 |
| Bismuth, Unfiltered | ug/L | all sites | 0.01 | 0.03 | 0.14 | 0.00 | 0.00 | 0.04 | 0.00 | 0.00 | 0.0 |
| Boron, Unfiltered | ug/L | all sites | 13.96 | 25.30 | 34.60 | 16.26 | 23.60 | 31.56 | 31.14 | 36.40 | 43.0 |
| Cadmium, Unfiltered | ug/L | all sites | 0.02 | 0.05 | 0.17 | 0.01 | 0.02 | 0.07 | 0.01 | 0.02 | 0.0 |
| Cerium, Unfiltered | ug/L | all sites | 0.99 | 5.59 | 17.62 | 0.29 | 0.64 | 6.50 | 0.07 | 0.18 | 0.5 |
| Cesium, Unfiltered | ug/L | all sites | 0.07 | 0.49 | 1.67 | 0.02 | 0.06 | 0.58 | 0.01 | 0.01 | 0.0 |
| Chromium, Unfiltered | ug/L | all sites | 0.26 | 3.56 | 11.80 | 0.20 | 0.45 | 4.41 | 0.04 | 0.18 | 0.3 |
| Cobalt, Unfiltered | ug/L | all sites | 0.39 | 1.65 | 5.23 | 0.17 | 0.27 | 1.94 | 0.08 | 0.09 | 0.1 |
| Copper, Unfiltered | $\mathrm{ug/L}$ | all sites | 1.14 | 4.40 | 12.36 | 0.53 | 0.91 | 5.69 | + | + | - |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.29 | 0.66 | 0.9 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.17 | 0.59 | 2.0 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Gallium, Unfiltered | ug/L | all sites | 0.07 | 0.78 | 2.72 | 0.05 | 0.10 | 0.91 | 0.01 | 0.03 | 0.0 |
| Germanium, Unfiltered | ug/L | all sites | 0.02 | 0.07 | 0.22 | 0.01 | 0.02 | 0.06 | 0.01 | 0.01 | 0.0 |
| Indium, Unfiltered | ug/L | all sites | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.01 | < | < | < |
| Iron, Unfiltered | ug/L | all sites | 631.40 | 4290.00 | 12800.00 | 308.00 | 709.00 | 5302.00 | 132.90 | 430.50 | 863.5 |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | | pen Wat | er | 1 | Under Ice | |
|-------------------------------|------|------------|-------|-----------|--------|-------|------------------|--------|------|------------------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95t |
| Lanthanum, Unfiltered | ug/L | all sites | 0.45 | 2.58 | 8.40 | 0.13 | 0.31 | 3.05 | 0.04 | 0.09 | 0.2 |
| Lead, Unfiltered | ug/L | all sites | 0.45 | 2.15 | 6.85 | 0.11 | 0.27 | 2.48 | 0.03 | 0.09 | 0.3 |
| Lithium, Unfiltered | ug/L | all sites | 5.47 | 7.88 | 13.52 | 5.75 | 6.91 | 9.95 | 8.32 | 9.97 | 11. |
| Manganese, Unfiltered | ug/L | all sites | 48.26 | 114.00 | 289.00 | 16.30 | 38.50 | 135.00 | 5.38 | 15.85 | 26.7 |
| Mercury, Unfiltered | ng/L | all sites | 2.85 | 10.00 | 28.90 | 0.98 | 1.90 | 12.63 | 0.47 | 0.68 | 0.9 |
| Methylmercury(1+), Unfiltered | ng/L | all sites | 0.07 | 0.18 | 0.33 | 0.02 | 0.06 | 0.22 | 0.03 | 0.04 | 0.0 |
| Molybdenum, Unfiltered | ug/L | all sites | 0.39 | 0.75 | 1.24 | 0.36 | 0.73 | 1.01 | + | + | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.69 | 0.77 | 3. |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.23 | 0.90 | 1. |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Nickel, Unfiltered | ug/L | all sites | 1.45 | 5.23 | 16.32 | 0.90 | 1.32 | 6.39 | + | + | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.75 | 1.03 | 1. |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.45 | 0.96 | 2. |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Niobium, Unfiltered | ug/L | all sites | 0.00 | 0.10 | 0.23 | 0.00 | 0.01 | 0.11 | 0.00 | 0.00 | 0. |
| Palladium, Unfiltered | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Platinum, Unfiltered | ug/L | all sites | 0.00 | 0.00 | 0.00 | < | < | < | < | < | |
| Rubidium, Unfiltered | ug/L | all sites | 1.49 | 5.93 | 18.42 | 1.06 | 1.40 | 6.71 | 1.18 | 1.57 | 1. |
| Scandium, Unfiltered | ug/L | all sites | 0.02 | 0.44 | 2.52 | 0.00 | 0.05 | 0.66 | 0.00 | 0.02 | 0. |
| Selenium, Unfiltered | ug/L | all sites | 0.14 | 0.22 | 0.59 | 0.10 | 0.14 | 0.29 | + | + | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |

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Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | | High Flow | | (| Open Wat | er | | Under Ice | ; |
|-----------------------|-----------------|------------|--------|------------------|--------|--------|------------------|--------|--------|-----------|-----|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95t |
| | -ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.13 | 0.18 | 0. |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.04 | 0.20 | 0. |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Silver, Unfiltered | ug/L | all sites | 0.00 | 0.02 | 0.07 | 0.00 | 0.00 | 0.04 | 0.00 | 0.00 | 0 |
| Strontium, Unfiltered | ug/L | all sites | + | + | + | 123.00 | 223.00 | 293.00 | + | + | |
| | ug/L | AL07DD0004 | 111.00 | 177.00 | 222.00 | + | + | + | - | - | |
| | ug/L | AL07DD0005 | 136.00 | 182.00 | 205.00 | + | + | + | - | - | |
| | ug/L | AL07DD0007 | 162.00 | 214.00 | 246.00 | + | + | + | 275.00 | 316.00 | 384 |
| | ug/L | AL07DD0008 | 81.60 | 137.00 | 248.00 | + | + | + | 134.00 | 352.00 | 481 |
| | ug/L | AL07DD0009 | - | - | - | + | + | + | - | - | |
| Tellurium, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.00 | 0.01 | 0.06 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 | 0 |
| Thallium, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.01 | 0.05 | 0.18 | 0.01 | 0.01 | 0.05 | 0.00 | 0.01 | 0 |
| Tin, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.03 | 0.09 | 0.39 | 0.00 | 0.02 | 0.14 | 0.00 | 0.01 | 0 |
| Titanium, Unfiltered | ug/L | all sites | 3.02 | 36.00 | 98.38 | 1.80 | 5.30 | 50.18 | 0.40 | 1.10 | 2 |
| Tungsten, Unfiltered | ug/L | all sites | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0 |
| Uranium, Unfiltered | ug/L | all sites | 0.27 | 0.45 | 1.03 | 0.18 | 0.37 | 0.57 | + | + | |
| | ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 0.38 | 0.45 | 0 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.10 | 0.57 | 0 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | |
| Vanadium, Unfiltered | ug/L | all sites | 0.88 | 6.92 | 23.36 | 0.57 | 1.07 | 8.98 | 0.22 | 0.36 | 0 |
| Yttrium, Unfiltered | ug/L | all sites | 0.48 | 2.07 | 6.49 | 0.15 | 0.31 | 2.49 | 0.09 | 0.11 | 0 |
| Zinc, Unfiltered | ug/L | all sites | 2.52 | 13.10 | 41.38 | 0.98 | 2.00 | 14.64 | + | + | |

Table 2.4: Current Conditions, Athabasca River water. (continued)

| | | | High Flow | | | Open Water | | | Under Ice | | |
|-----------------------|-------|------------|-----------|------|------|-----------------|------|------|----------------|------------------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5 \mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | $50 \mathrm{th}$ | 95th |
| | -ug/L | AL07DD0004 | + | + | + | + | + | + | - | - | - |
| | ug/L | AL07DD0005 | + | + | + | + | + | + | - | - | - |
| | ug/L | AL07DD0007 | + | + | + | + | + | + | 1.00 | 1.60 | 2.00 |
| | ug/L | AL07DD0008 | + | + | + | + | + | + | 0.70 | 1.85 | 6.90 |
| | ug/L | AL07DD0009 | + | + | + | + | + | + | - | - | - |
| Zirconium, Unfiltered | ug/L | all sites | 0.36 | 1.80 | 4.40 | 0.20 | 0.30 | 2.82 | 0.10 | 0.20 | 0.30 |

Note:

- data insufficient
- < too highly censored;
- + grouped differently (merged sites vs individual site);

 ${\bf Table~2.5:~Current~Conditions,~Athabasca~River~sediment.}$

| Pa | urameter | Unit | Site | 5th | 50th | 95th |
|----------------------------|---|-----------------------|---|--------------------------------------|--------------------------------------|--------------------------------------|
| Conventiona | l Variables id Neutralization Potential as %CaCO3 | % | all sites | | | |
| | | | all sites | 0.99 | 7.00 | 15.48 |
| | rain size, clay (<2 um) | | | | | |
| | rain size, sand (>=63 um to 2000 um) | % | all sites | 30.50 | 72.00 | 98.80 |
| | rain size, silt (>=2 to 63 um) | % | all sites | 1.48 | 19.40 | 48.44 |
| | organic carbon | % | all sites | - | - | - |
| | ss on Ignition @ 375 C | % | all sites | 0.64 | 1.50 | 3.23 |
| Mo | oisture content | - % | AB07DA0062 | - | - | |
| | | % | AB07DA0800 | - | - | |
| | | % | AB07DA3008 | - | - | |
| | | % | AB07DA3009 | - | - | |
| | | % | AB07DA3015 | - | - | |
| | | % | AB07DA3016 | - | - | |
| | | % | AB07DA3017 | - | - | |
| | | % | AB07DA3018 | - | - | |
| | | % | AB07DA3020 | - | - | |
| | | % | AB07DA3021 | - | - | |
| | | % | AB07DA3022 | - | - | |
| | | % | AB07DA3023 | - | - | |
| | | % | AB07DA3024 | - | - | |
| | | | ATR-ER | - | - | |
| Or | ganic Matter | % | all sites | 0.68 | 1.40 | 2.7 |
| Or | ganic carbon | % | all sites | - | - | |
| To | otal carbon | % | all sites | - | - | |
| Extractable Me | Metals ethylmercury(1+), Extractable | ng/g | all sites | 0.02 | 0.31 | 1.19 |
| General Org | | | | | | |
| BT | ΓEX, Total | ug/g | all sites | - | - | |
| Be | enzene | ug/g | all sites | - | - | |
| C1 | 0-C16 Hydrocarbons | ug/g | all sites | - | - | |
| C1 | 10H16O2 | % | all sites | 0.00 | 0.01 | 0.0 |
| C1 | 10H18O2 | % | all sites | 0.01 | 0.04 | 0.1 |
| C1 | 10H20O2 | % | all sites | 0.07 | 0.39 | 1.6 |
| C1 | 11H14O2 | % | all sites | 0.01 | 0.03 | 0.0 |
| | 11H16O2 | % | all sites | 0.00 | 0.00 | 0.0 |
| C1 | | % | all sites | 0.00 | 0.01 | 0.0 |
| | 11H18O2 | | | 0.01 | 0.00 | 0.1 |
| C1 | 11H18O2 11H20O2 | % | all sites | 0.01 | 0.06 | |
| C1 | | % % | all sites | 0.01 | 0.06 | |
| C1 C1 | 1H20O2 | | | | | 0.78 |
| C1 C1 C1 | 11H20O2 11H22O2 | % | all sites | 0.21 | 0.45 | 0.78 |
| C1 C1 C1 C1 | 11H20O2 11H22O2 12H16O2 | % | all sites | 0.21 | 0.45 0.01 | 0.78 |
| C1 C1 C1 C1 C1 | 11H20O2 11H22O2 12H16O2 12H18O2 | % % % | all sites all sites | 0.21 0.00 0.00 | 0.45 0.01 0.00 | 0.78 0.00 0.00 0.28 |
| C1 C1 C1 C1 C1 C1 C1 C1 C1 | 11H20O2 11H22O2 12H16O2 12H18O2 12H20O2 | % % % % | all sites all sites all sites all sites | 0.21 0.00 0.00 0.01 | 0.45 0.01 0.00 0.06 | 0.78 0.00 0.00 0.28 0.65 |
| C1 C1 C1 C1 C1 C1 C1 C1 C1 | 11H20O2 11H22O2 12H16O2 12H18O2 12H20O2 | % % % % % | all sites all sites all sites all sites all sites | 0.21 0.00 0.00 0.01 0.11 | 0.45 0.01 0.00 0.06 0.31 | 0.73 0.00 0.00 0.23 0.63 |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| Parameter | Unit | Site | 5th | $50\mathrm{th}$ | 95th |
|----------------------------------|--------|------------|--------------|-----------------|------------------------|
| C13H20O2 | % | all sites | 0.01 | 0.03 | 0.14 |
| C13H22O2 | % | all sites | 0.00 | 0.03 | 0.20 |
| C13H24O2 | % | all sites | 0.04 | 0.10 | 0.20 |
| C13H26O2 | % | all sites | 0.38 | 0.77 | 0.94 |
| C14H16O2 | % | all sites | < | < | < |
| C14H18O2 | % | all sites | 0.00 | 0.01 | 0.08 |
| C14H20O2 | % | all sites | 0.00 | 0.03 | 0.09 |
| C14H22O2 | % | all sites | 0.05 | 0.10 | 1.61 |
| C14H24O2 | % | all sites | 0.06 | 0.14 | 2.64 |
| C14H26O2 | % | all sites | 0.42 | 0.79 | 1.31 |
| C14H28O2 | | AB07DA0062 | - | - | - |
| | | AB07DA0800 | - | - | _ |
| | % | AB07DA3008 | _ | - | |
| | % | AB07DA3009 | - | - | - |
| | % | AB07DA3015 | - | - | - |
| | % | AB07DA3016 | - | - | - |
| | | AB07DA3017 | - | - | - |
| | | AB07DA3018 | - | - | _ |
| | | AB07DA3020 | - | - | _ |
| | | AB07DA3021 | - | - | - |
| | | AB07DA3022 | - | - | - |
| | | AB07DA3023 | - | - | _ |
| | | AB07DA3024 | - | - | - |
| C15H14O2 | % | all sites | 0.00 | 0.01 | 0.02 |
| C15H16O2 | % | all sites | 0.00 | 0.01 | 0.03 |
| C15H18O2 | % | all sites | 0.00 | 0.00 | 0.03 |
| C15H20O2 | % | all sites | 0.00 | 0.04 | 0.17 |
| C15H22O2 | % | all sites | 0.02 | 0.10 | 1.44 |
| C15H24O2 | % | all sites | 0.03 | 0.15 | 2.12 |
| C15H26O2 | % | all sites | 0.07 | 0.18 | 1.90 |
| C15H28O2 | % | all sites | 0.83 | 2.01 | 3.51 |
| C15H30O2 | % | all sites | 2.61 | 4.24 | 6.84 |
| C16-C34 Hydrocarbons | ug/g | all sites | - | - | - |
| C16H14O2 | % | all sites | 0.00 | 0.01 | 0.04 |
| C16H16O2 | % | all sites | < | < | < |
| C16H18O2 | % | all sites | 0.00 | 0.01 | 0.05 |
| C16H20O2 | % | all sites | 0.00 | 0.03 | 0.14 |
| C16H22O2 | % | all sites | 0.01 | 0.06 | 0.22 |
| | % | all sites | 0.33 | 2.17 | 3.93 |
| C16H24O2 | | | 0.47 | 0.70 | 4.55 |
| C16H24O2 C16H26O2 | % | all sites | 0.47 | 2.79 | 4.00 |
| | % | all sites | 0.47 | 3.03 | |
| C16H26O2 | | | | | 4.71 |
| C16H26O2 C16H28O2 | % | all sites | 0.76 | 3.03 | 4.71 20.71 25.45 |
| C16H26O2 C16H28O2 C16H30O2 | % % | all sites | 0.76 6.65 | 3.03 13.70 | 4.71 20.71 |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| Parameter | Unit | Site | $5\mathrm{th}$ | 50th | 95th |
|-----------|------|-----------|----------------|------|-------|
| C17H22O2 | % | all sites | 0.00 | 0.04 | 0.22 |
| C17H24O2 | % | all sites | 0.01 | 0.07 | 0.26 |
| C17H26O2 | % | all sites | 0.04 | 0.12 | 0.46 |
| C17H28O2 | % | all sites | 0.08 | 0.27 | 0.69 |
| C17H30O2 | % | all sites | 0.13 | 0.30 | 0.68 |
| C17H32O2 | % | all sites | 1.66 | 2.94 | 7.08 |
| C17H34O2 | % | all sites | 1.42 | 2.92 | 8.32 |
| C18H20O2 | % | all sites | 0.00 | 0.01 | 0.10 |
| C18H22O2 | % | all sites | 0.01 | 0.04 | 0.14 |
| C18H24O2 | % | all sites | 0.03 | 0.09 | 0.17 |
| C18H26O2 | % | all sites | 0.08 | 0.14 | 0.64 |
| C18H28O2 | % | all sites | 0.32 | 1.77 | 5.47 |
| C18H30O2 | % | all sites | 0.62 | 1.93 | 3.47 |
| C18H32O2 | % | all sites | 1.47 | 2.78 | 6.48 |
| C18H34O2 | % | all sites | 4.56 | 7.01 | 25.26 |
| C18H36O2 | % | all sites | 0.12 | 0.61 | 24.95 |
| C19H20O2 | % | all sites | 0.00 | 0.00 | 0.09 |
| C19H22O2 | % | all sites | 0.03 | 0.14 | 0.48 |
| C19H24O2 | % | all sites | 0.01 | 0.05 | 0.10 |
| C19H26O2 | % | all sites | 0.02 | 0.08 | 0.33 |
| C19H28O2 | % | all sites | 0.03 | 0.15 | 0.38 |
| C19H30O2 | % | all sites | 0.05 | 0.16 | 0.35 |
| C19H32O2 | % | all sites | 0.03 | 0.15 | 0.61 |
| C19H34O2 | % | all sites | 0.07 | 0.32 | 1.09 |
| C19H36O2 | % | all sites | 0.22 | 0.46 | 1.16 |
| C19H38O2 | % | all sites | 0.20 | 0.32 | 0.56 |
| C20H22O2 | % | all sites | 0.00 | 0.01 | 0.12 |
| C20H24O2 | % | all sites | 0.01 | 0.03 | 0.11 |
| C20H26O2 | % | all sites | 0.02 | 0.12 | 0.29 |
| C20H28O2 | % | all sites | 0.45 | 1.06 | 4.85 |
| C20H30O2 | % | all sites | 0.95 | 7.21 | 13.09 |
| C20H32O2 | % | all sites | 0.39 | 1.19 | 2.14 |
| C20H34O2 | % | all sites | 0.13 | 0.32 | 0.69 |
| C20H36O2 | % | all sites | 0.22 | 0.41 | 1.42 |
| C20H38O2 | % | all sites | 0.11 | 0.29 | 0.52 |
| C20H40O2 | % | all sites | 0.30 | 0.85 | 1.25 |
| C21H24O2 | % | all sites | 0.01 | 0.05 | 0.10 |
| C21H26O2 | % | all sites | 0.00 | 0.01 | 0.05 |
| C21H28O2 | % | all sites | 0.00 | 0.02 | 0.10 |
| C21H30O2 | % | all sites | 0.01 | 0.06 | 0.12 |
| C21H32O2 | % | all sites | 0.02 | 0.07 | 0.24 |
| C21H34O2 | % | all sites | 0.03 | 0.11 | 0.40 |
| C21H36O2 | % | all sites | 0.02 | 0.20 | 0.82 |
| C21H38O2 | % | all sites | 0.04 | 0.29 | 1.37 |
| C21H40O2 | % | all sites | 0.01 | 0.10 | 0.48 |
| | | | | | |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| Parameter | Unit | Site | $5\mathrm{th}$ | $50 \mathrm{th}$ | $95 \mathrm{th}$ |
|----------------------|------|-----------|----------------|------------------|------------------|
| C21H42O2 | % | all sites | 0.21 | 0.39 | 0.96 |
| C22H32O2 | % | all sites | 0.12 | 0.80 | 2.45 |
| C22H34O2 | % | all sites | 0.08 | 0.24 | 0.81 |
| C22H36O2 | % | all sites | 0.04 | 0.12 | 0.50 |
| C22H38O2 | % | all sites | 0.03 | 0.10 | 0.30 |
| C22H40O2 | % | all sites | 0.06 | 0.28 | 1.39 |
| C22H42O2 | % | all sites | 0.12 | 0.34 | 1.11 |
| C22H44O2 | % | all sites | 0.01 | 0.60 | 1.86 |
| C23H32O2 | % | all sites | 0.00 | 0.02 | 0.07 |
| C23H34O2 | % | all sites | 0.00 | 0.03 | 0.10 |
| C23H36O2 | % | all sites | 0.00 | 0.04 | 0.12 |
| C23H38O2 | % | all sites | 0.01 | 0.06 | 0.30 |
| C23H40O2 | % | all sites | 0.02 | 0.15 | 0.85 |
| C23H42O2 | % | all sites | 0.04 | 0.27 | 1.38 |
| C23H44O2 | % | all sites | 0.05 | 0.19 | 0.85 |
| C23H46O2 | % | all sites | 0.12 | 0.41 | 0.92 |
| C24H36O2 | % | all sites | 0.00 | 0.02 | 0.10 |
| C24H38O2 | % | all sites | 0.01 | 0.03 | 0.08 |
| C24H40O2 | % | all sites | 0.01 | 0.04 | 0.12 |
| C24H42O2 | % | all sites | 0.04 | 0.20 | 1.23 |
| C24H44O2 | % | all sites | 0.06 | 0.24 | 1.34 |
| C24H46O2 | % | all sites | 0.03 | 0.23 | 0.38 |
| C24H48O2 | % | all sites | 0.01 | 0.75 | 2.04 |
| C25H38O2 | % | all sites | 0.00 | 0.00 | 0.05 |
| C25H40O2 | % | all sites | 0.01 | 0.04 | 0.08 |
| C25H42O2 | % | all sites | 0.01 | 0.03 | 0.12 |
| C25H44O2 | % | all sites | 0.01 | 0.08 | 0.28 |
| C25H46O2 | % | all sites | 0.04 | 0.15 | 0.49 |
| C25H48O2 | % | all sites | 0.04 | 0.09 | 0.38 |
| C25H50O2 | % | all sites | 0.01 | 0.39 | 0.80 |
| C34-C50 Hydrocarbons | ug/g | all sites | - | - | - |
| C5H10O2 | % | all sites | 0.00 | 0.03 | 0.12 |
| C6H12O2 | % | all sites | 0.00 | 0.02 | 0.14 |
| C7H12O2 | % | all sites | 0.00 | 0.01 | 0.03 |
| C7H14O2 | % | all sites | 0.01 | 0.04 | 0.19 |
| C8H14O2 | % | all sites | 0.01 | 0.02 | 0.07 |
| C8H16O2 | % | all sites | 0.04 | 0.18 | 0.69 |
| C9H14O2 | % | all sites | 0.00 | 0.01 | 0.06 |
| C9H16O2 | % | all sites | 0.00 | 0.03 | 0.07 |
| C9H18O2 | % | all sites | 0.13 | 0.47 | 1.38 |
| Ethylbenzene | ug/g | all sites | - | - | - |
| Hydrocarbons | ug/g | all sites | - | - | - |
| Naphthenic acids | ug/g | all sites | 52.91 | 136.50 | 458.90 |
| Toluene | ug/g | all sites | - | - | - |
| Total xylenes | ug/g | all sites | - | - | - |
| | | | | | |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| | Parameter | Unit | Site | 5th | 50th | 95tl |
|---------|---|------|-----------|--------|---------|----------|
| | m,p-Xylene | ug/g | all sites | - | - | |
| | o-Xylene | ug/g | all sites | - | - | |
| Nutrier | its and BOD | | | | | |
| | Ammonium, Available as N | ng/g | all sites | 819.46 | 6550.00 | 25800.00 |
| | Kjeldahl nitrogen, Total | % | all sites | 0.01 | 0.04 | 0.1 |
| PAHs | 1,2,6-Trimethylphenanthrene | ng/g | all sites | 1.05 | 3.15 | 8.6 |
| | 1,2-Dimethylnaphthalene | ng/g | all sites | 0.22 | 1.53 | 2.9 |
| | 1,4,6,7-Tetramethylnaphthalene | -, - | all sites | 1.65 | 4.55 | 8.0 |
| | 1,6,7-Trimethylnaphthalene | ng/g | all sites | 1.41 | 6.21 | 10.2 |
| | 1,7-Dimethylfluorene | ng/g | | | | |
| | | ng/g | all sites | 0.53 | 1.62 | 4.6 |
| | 1,7-Dimethylphenanthrene | ng/g | all sites | 2.05 | 6.92 | 22.4 |
| | 1,8-Dimethylphenanthrene | ng/g | all sites | 0.51 | 1.75 | 4.9 |
| | 1-Methylchrysene | ng/g | all sites | 1.55 | 4.68 | 29.0 |
| | 1-Methylnaphthalene | ng/g | all sites | 1.40 | 6.79 | 16.6 |
| | 1-Methylphenanthrene | ng/g | all sites | 1.70 | 6.16 | 21.4 |
| | 2,3,6-Trimethylnaphthalene | ng/g | all sites | 1.71 | 7.29 | 14.2 |
| | 2,4-Dimethyldibenzothiophene | ng/g | all sites | 1.59 | 4.05 | 26.1 |
| | 2,6-Dimethylnaphthalene | ng/g | all sites | 1.56 | 6.96 | 18. |
| | 2,6-Dimethylphenanthrene | ng/g | all sites | 1.08 | 3.13 | 17.5 |
| | 2-Methylanthracene | ng/g | all sites | 0.47 | 1.19 | 19.0 |
| | 2-Methyldibenzothiophenes/3- Methyldibenzothiophenes | ng/g | all sites | 1.12 | 3.58 | 45.0 |
| | 2-Methylfluorene | ng/g | all sites | 0.46 | 1.09 | 3.0 |
| | 2-Methylnaphthalene | ng/g | all sites | 2.15 | 10.98 | 32.0 |
| | 2-Methylphenanthrene | ng/g | all sites | 2.50 | 9.30 | 48.6 |
| | 3,6-Dimethylphenanthrene | ng/g | all sites | 1.34 | 3.92 | 12.3 |
| | 3-Methylfluoranthene/Benzo[a]fluorene | ng/g | all sites | 3.29 | 8.38 | 31.8 |
| | 3-Methylphenanthrene | ng/g | all sites | 2.07 | 6.86 | 29.4 |
| | 4,6-Dimethyldibenzothiophene | ng/g | all sites | _ | _ | |
| | 5,9-Dimethylchrysene | ng/g | all sites | 4.84 | 11.90 | 56.5 |
| | 5-Methylchrysene/6-Methylchrysene | ng/g | all sites | 1.00 | 2.84 | 11.9 |
| | 7-Methylbenzo[a]pyrene | ng/g | all sites | 1.03 | 2.54 | 12.0 |
| | 9-Methylphenanthrene/4- | ng/g | all sites | 2.57 | 7.95 | 22.9 |
| | Methylphenanthrene | | | | | |
| | Acenaphthene | ng/g | all sites | 0.23 | 0.69 | 1.5 |
| | Acenaphthylene | ng/g | all sites | - | - | |
| | Anthracene | ng/g | all sites | 0.07 | 0.61 | 4.5 |
| | Benz[a]anthracene | ng/g | all sites | 0.16 | 2.82 | 44.5 |
| | Benzo(b)fluoranthene | ng/g | all sites | 2.38 | 7.83 | 22.3 |
| | Benzo(j+k)fluoranthene | ng/g | all sites | 1.10 | 2.73 | 13.8 |
| | Benzo[a]pyrene | ng/g | all sites | 0.30 | 4.05 | 51.7 |
| | Benzo[b,j,k] fluoranthene | ng/g | all sites | | | |
| | Benzo[e]pyrene | ng/g | all sites | 2.87 | 8.22 | 46.9 |
| | Benzo[ghi]perylene | ng/g | all sites | 0.72 | 7.17 | 35.8 |
| | Biphenyl | ng/g | all sites | 0.45 | 3.51 | 6.3 |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| | Parameter | Unit | Site | 5th | 50th | 95t |
|--------------|------------------------------------|------|-----------|--------|----------|--------|
| - | C1-Acenaphthenes | ng/g | all sites | 0.08 | 0.21 | 0.3 |
| - | C1-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 11.20 | 35.15 | 262.0 |
| - | C1-Benzofluoranthenes/benzopyrenes | ng/g | all sites | 2.68 | 36.90 | 239.0 |
| - | C1-Biphenyls | ng/g | all sites | 0.35 | 5.20 | 9.7 |
| - | C1-Dibenzothiophenes | ng/g | all sites | 0.35 | 10.70 | 109.8 |
| - | C1-Fluoranthenes/pyrenes | ng/g | all sites | 5.23 | 27.90 | 121.0 |
| = | C1-Fluorenes | ng/g | all sites | 0.55 | 4.31 | 14.1 |
| - | C1-Naphthalenes | ng/g | all sites | 0.71 | 15.30 | 46. |
| = | C1-Phenanthrenes/anthracenes | ng/g | all sites | 1.18 | 20.10 | 133.9 |
| - | C2-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 4.07 | 39.70 | 209. |
| - | C2-Benzofluoranthenes/benzopyrenes | ng/g | all sites | 1.46 | 19.40 | 129.0 |
| _ | C2-Biphenyls | ng/g | all sites | 1.06 | 4.44 | 7.9 |
| - | C2-Dibenzothiophenes | ng/g | all sites | 2.30 | 54.40 | 321.5 |
| _ | C2-Fluoranthenes/pyrenes | ng/g | all sites | 10.37 | 48.20 | 159.0 |
| - | C2-Fluorenes | ng/g | all sites | 0.51 | 19.40 | 48. |
| _ | C2-Naphthalenes | ng/g | all sites | 2.23 | 27.50 | 68. |
| _ | C2-Phenanthrenes/anthracenes | ng/g | all sites | 1.59 | 38.40 | 147. |
| _ | C3-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 5.91 | 16.30 | 49. |
| _ | C3-Dibenzothiophenes | ng/g | all sites | 4.40 | 103.00 | 364. |
| - | C3-Fluoranthenes/pyrenes | ng/g | all sites | 9.05 | 38.20 | 96. |
| _ | C3-Fluorenes | ng/g | all sites | 1.73 | 38.30 | 96. |
| - | C3-Naphthalenes | ng/g | all sites | 1.55 | 26.20 | 53. |
| - | C3-Phenanthrenes/anthracenes | ng/g | all sites | 2.67 | 50.00 | 127. |
| - | C4-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 2.43 | 8.35 | 17. |
| - | C4-Dibenzothiophenes | ng/g | all sites | 6.23 | 82.00 | 274. |
| _ | C4-Fluoranthenes/pyrenes | ng/g | all sites | 7.32 | 22.05 | 47. |
| - | C4-Naphthalenes | ng/g | all sites | 1.24 | 28.80 | 50. |
| _ | C4-Phenanthrenes/anthracenes | ng/g | all sites | 16.61 | 215.00 | 895. |
| _ | Chrysene | ng/g | all sites | 1.03 | 12.60 | 73. |
| _ | Dibenz[a,h]anthracene | ng/g | all sites | 0.33 | 1.69 | 5. |
| _ | Dibenzothiophene | ng/g | all sites | 0.14 | 1.76 | 23. |
| _ | Fluoranthene | ng/g | all sites | 0.19 | 3.43 | 10. |
| _ | Fluorene | ng/g | all sites | 0.06 | 1.24 | 3. |
| _ | Indeno[1,2,3-cd]pyrene | ng/g | all sites | 0.37 | 3.82 | 13. |
| _ | Naphthalene | ng/g | all sites | 0.51 | 4.00 | 14. |
| - | Perylene | ng/g | all sites | 22.10 | 68.75 | 129. |
| | Phenanthrene | ng/g | all sites | 0.55 | 11.10 | 35. |
| - | Pyrene | ng/g | all sites | 0.62 | 6.85 | 36.9 |
| - | Retene | ng/g | all sites | 2.82 | 42.20 | 89.: |
| henolics | | | | | | |
| | Phenols, Extractable | ng/g | all sites | < | < | |
| otal Me | | , | 11 | 640.00 | F0.40.00 | 0000 |
| - | Autimore | ug/g | all sites | 848.00 | 5340.00 | 9890.0 |
| | Antimony | ug/g | all sites | 0.09 | 0.20 | 6.6 |
| - | Arsenic | ug/g | all sites | 1.96 | 4.21 | |
| | | | | | | |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| ug/g ug/g ug/g ug/g | AB07DA0062 AB07DA0800 AB07DA3008 | - | - | - |
|---------------------|---|---|---|--|
| ug/g ug/g | | - | - | _ |
| ug/g | AB07DA3008 | | | |
| | | - | - | - |
| / | AB07DA3009 | - | - | - |
| ug/g | AB07DA3015 | - | - | - |
| ug/g | AB07DA3016 | - | - | - |
| ug/g | AB07DA3017 | - | - | - |
| ug/g | AB07DA3018 | - | - | - |
| ug/g | AB07DA3020 | - | - | - |
| ug/g | AB07DA3021 | - | - | - |
| ug/g | AB07DA3022 | - | - | - |
| ug/g | AB07DA3023 | - | - | - |
| ug/g | AB07DA3024 | - | - | - |
| ug/g | ATR-ER | - | - | - |
| ug/g | all sites | 0.19 | 0.35 | 0.56 |
| ug/g | all sites | < | < | < |
| ug/g | all sites | 1.28 | 5.25 | 8.42 |
| | all sites | 0.06 | 0.13 | 0.23 |
| | AB07DA0062 | - | - | _ |
| | AB07DA0800 | - | _ | _ |
| ug/g | AB07DA3008 | - | - | _ |
| | AB07DA3009 | - | _ | |
| ug/g | AB07DA3015 | - | _ | _ |
| | AB07DA3016 | _ | _ | _ |
| | AB07DA3017 | - | _ | _ |
| | AB07DA3018 | - | _ | _ |
| | AB07DA3020 | _ | _ | _ |
| | AB07DA3021 | _ | _ | _ |
| | AB07DA3022 | _ | _ | _ |
| | AB07DA3023 | - | _ | _ |
| ug/g | AB07DA3024 | _ | _ | _ |
| ug/g | all sites | 2.29 | 10.90 | 17.35 |
| | all sites | 2.00 | 6.03 | 8.80 |
| ug/g | all sites | 1.02 | 6.75 | 15.65 |
| ug/g | all sites | 4000.00 | 13000.00 | 20300.00 |
| ug/g | all sites | 1.47 | 5.34 | 9.41 |
| ug/g | all sites | 4.25 | 8.12 | 12.36 |
| ug/g | AB07DA0062 | - | _ | - |
| ug/g | AB07DA0800 | - | - | _ |
| ug/g | AB07DA3008 | - | _ | _ |
| ug/g | AB07DA3009 | - | _ | _ |
| | AB07DA3015 | _ | _ | _ |
| | AB07DA3016 | _ | _ | |
| ug/g | | | _ | |
| ug/g ug/g | AB07DA3017 | | | |
| | ug/g ug/g ug/g ug/g ug/g ug/g ug/g ug/g | ug/g AB07DA3020 ug/g AB07DA3021 ug/g AB07DA3022 ug/g AB07DA3023 ug/g AB07DA3024 ug/g AB07DA3024 ug/g ATR-ER ug/g all sites ug/g all sites ug/g all sites ug/g AB07DA0062 ug/g AB07DA3008 ug/g AB07DA3008 ug/g AB07DA3009 ug/g AB07DA3015 ug/g AB07DA3015 ug/g AB07DA3016 ug/g AB07DA3018 ug/g AB07DA3020 ug/g AB07DA3021 ug/g AB07DA3022 ug/g AB07DA3023 ug/g All sites ug/g AB07DA062 ug/g AB07DA0062 ug/g AB07DA0080 <td>ug/g AB07DA3020 - ug/g AB07DA3021 - ug/g AB07DA3022 - ug/g AB07DA3023 - ug/g AB07DA3024 - ug/g ATR-ER - ug/g all sites 0.19 ug/g all sites 0.06 ug/g all sites 0.06 ug/g AB07DA0062 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3015 - ug/g AB07DA3015 - ug/g AB07DA3016 - ug/g AB07DA3018 - ug/g AB07DA3020 - ug/g AB07DA3021 - ug/g AB07DA3022 - ug/g AB07DA3024 - ug/g all sites 2.29</td> <td>ug/g AB07DA3020 - - ug/g AB07DA3021 - - ug/g AB07DA3022 - - ug/g AB07DA3023 - - ug/g AB07DA3024 - - ug/g ATR-ER - - ug/g all sites 0.19 0.35 ug/g all sites 0.19 0.35 ug/g all sites 0.19 0.35 ug/g all sites 0.09 0.35 ug/g all sites 0.06 0.13 ug/g AB07DA0062 - - ug/g AB07DA3008 - - ug/g AB07DA3009 - - ug/g AB07DA3015 - - ug/g AB07DA3017 - - ug/g AB07DA3021 - - ug/g AB07DA3022 - - ug/g AB07DA3023 - -</td> | ug/g AB07DA3020 - ug/g AB07DA3021 - ug/g AB07DA3022 - ug/g AB07DA3023 - ug/g AB07DA3024 - ug/g ATR-ER - ug/g all sites 0.19 ug/g all sites 0.06 ug/g all sites 0.06 ug/g AB07DA0062 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3008 - ug/g AB07DA3015 - ug/g AB07DA3015 - ug/g AB07DA3016 - ug/g AB07DA3018 - ug/g AB07DA3020 - ug/g AB07DA3021 - ug/g AB07DA3022 - ug/g AB07DA3024 - ug/g all sites 2.29 | ug/g AB07DA3020 - - ug/g AB07DA3021 - - ug/g AB07DA3022 - - ug/g AB07DA3023 - - ug/g AB07DA3024 - - ug/g ATR-ER - - ug/g all sites 0.19 0.35 ug/g all sites 0.19 0.35 ug/g all sites 0.19 0.35 ug/g all sites 0.09 0.35 ug/g all sites 0.06 0.13 ug/g AB07DA0062 - - ug/g AB07DA3008 - - ug/g AB07DA3009 - - ug/g AB07DA3015 - - ug/g AB07DA3017 - - ug/g AB07DA3021 - - ug/g AB07DA3022 - - ug/g AB07DA3023 - - |

Table 2.5: Current Conditions, Athabasca River sediment. (continued)

| Parameter | Unit | Site | $5 \mathrm{th}$ | $50 \mathrm{th}$ | 95th |
|------------|------|------------|-----------------|------------------|---------|
| | ug/g | AB07DA3020 | - | - | - |
| | ug/g | AB07DA3021 | - | - | - |
| | ug/g | AB07DA3022 | - | - | - |
| | ug/g | AB07DA3023 | - | - | - |
| | ug/g | AB07DA3024 | - | - | - |
| | ug/g | ATR-ER | - | - | - |
| Manganese | ug/g | all sites | 78.35 | 289.00 | 555.50 |
| Mercury | ug/g | all sites | < | < | < |
| Molybdenum | ug/g | all sites | 0.15 | 0.44 | 0.82 |
| Nickel | ug/g | all sites | 3.37 | 13.30 | 21.15 |
| Phosphorus | ug/g | AB07DA0062 | - | - | - |
| | ug/g | AB07DA0800 | - | - | - |
| | ug/g | AB07DA3008 | - | - | - |
| | ug/g | AB07DA3009 | - | - | - |
| | ug/g | AB07DA3015 | - | - | - |
| | ug/g | AB07DA3016 | - | - | - |
| | ug/g | AB07DA3017 | _ | _ | - |
| | ug/g | AB07DA3018 | - | - | - |
| | ug/g | AB07DA3020 | - | - | - |
| | ug/g | AB07DA3021 | - | - | - |
| | ug/g | AB07DA3022 | - | - | - |
| | ug/g | AB07DA3023 | - | - | - |
| | ug/g | AB07DA3024 | - | - | - |
| Potassium | ug/g | all sites | 222.10 | 767.50 | 1261.50 |
| Silver | ug/g | all sites | 0.03 | 0.05 | 0.09 |
| Sodium | ug/g | all sites | < | < | < |
| Strontium | ug/g | all sites | 7.95 | 46.70 | 75.55 |
| Thallium | ug/g | all sites | 0.04 | 0.10 | 0.16 |
| Thorium | ug/g | all sites | 0.89 | 3.33 | 5.25 |
| Tin | ug/g | all sites | 0.11 | 0.25 | 0.41 |
| Titanium | ug/g | all sites | 34.41 | 63.90 | 96.81 |
| Tungsten | ug/g | all sites | < | < | < |
| Uranium | ug/g | all sites | 0.12 | 0.67 | 1.00 |
| Vanadium | ug/g | all sites | 4.21 | 17.10 | 27.40 |
| Zinc | ug/g | all sites | 9.45 | 39.90 | 65.40 |
| Zirconium | ug/g | all sites | 1.32 | 3.95 | 5.95 |

Note:

1286 2.8.2 Athabasca River Delta Current Conditions

1287 The current condition (5th, 50th, and 95th percentile values) for each water and sediment

1288 quality parameter and each season are presented for the Athabasca River Delta in Table 2.6

⁻ data insufficient

< too highly censored;

- 1289 (water) and Table 2.7 (sediment). Note that additional information, including sample size,
- analytical method codes, and quantile estimation method for each suite of current conditions
- 1291 are provided in Appendix A.2.

TOE

Table 2.6: Current Conditions, Athabasca River Delta water.

| | | | | High Flow | | (| Open Wate | r | | Under Ice | |
|--|----------------------|------------|---------|-----------|---------|---------|-----------|---------|---------|-----------|--------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95tl |
| acteria | | | | | | | | | | | |
| Escherichia coli | No/100 mL | all sites | 1.37 | 5.48 | 30.00 | < | < | < | < | < | • |
| Fecal Coliform | $No/100~\mathrm{mL}$ | all sites | 1.24 | 6.50 | 39.80 | 0.09 | 1.53 | 29.00 | < | < | |
| Total Coliform | No/100~mL | all sites | - | - | - | - | - | - | - | - | |
| onventional Variables Alkalinity, Phenolphthalein (total hydroxide+1/2 carbonate) as CaCO3 | ${ m mg/L}$ | all sites | < | < | < | < | < | < | < | < | |
| Alkalinity, total as CaCO3 | $\mathrm{mg/L}$ | all sites | 68.80 | 89.00 | 100.00 | 90.40 | 110.00 | 128.00 | 100.00 | 140.00 | 160.0 |
| Deuterium/Hydrogen ratio | o/oo VSMOW | all sites | -152.40 | -144.25 | -135.60 | -142.20 | -139.30 | -133.80 | -144.57 | -139.95 | -136.0 |
| Dissolved oxygen (DO) | mg/L | all sites | - | - | - | - | - | - | - | - | |
| Organic carbon, Filtered | mg/L | all sites | 4.60 | 12.00 | 19.60 | 5.42 | 7.90 | 16.80 | 4.48 | 7.50 | 13. |
| Organic carbon, Unfiltered | mg/L | all sites | - | - | - | - | - | - | - | - | |
| Organic carbon, Unknown | mg/L | all sites | 4.30 | 12.50 | 19.00 | 4.47 | 9.10 | 20.50 | 5.03 | 8.20 | 14. |
| Oxidation reduction potential (ORP) | mV | all sites | 162.30 | 288.50 | 547.90 | 107.00 | 208.50 | 421.25 | + | + | |
| | mV | AB07DD0010 | + | + | + | + | + | + | 105.20 | 193.00 | 426. |
| | mV | AB07DD0105 | + | + | + | + | + | + | 104.30 | 227.50 | 553. |
| Oxygen-18 | o/oo VSMOW | all sites | -19.02 | -18.18 | -16.98 | -17.76 | -17.30 | -16.70 | -18.21 | -17.32 | -16. |
| Specific conductivity | uS/cm | all sites | 172.00 | 220.00 | 286.00 | 232.00 | 290.00 | 362.00 | 289.00 | 420.00 | 493. |
| Temperature, air | $\deg C$ | all sites | 6.00 | 17.00 | 34.00 | -4.00 | 8.00 | 22.00 | -26.50 | -7.00 | 6. |
| Total dissolved solids, Filtered | $\mathrm{mg/L}$ | all sites | 101.00 | 140.00 | 180.00 | 141.00 | 180.00 | 267.00 | 178.00 | 250.00 | 302. |
| Total suspended solids, Non-Filterable (Particle) | $\mathrm{mg/L}$ | all sites | 34.00 | 160.00 | 612.00 | 10.40 | 32.00 | 206.00 | 1.30 | 4.00 | 17. |
| True colour, Filtered | rel units | all sites | 15.60 | 66.00 | 126.00 | 16.20 | 32.00 | 97.80 | 17.80 | 28.00 | 57. |
| Turbidity | NTU | all sites | 4.12 | 65.00 | 246.00 | 4.20 | 13.00 | 77.80 | 2.88 | 3.70 | 14. |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flow | | C | pen Water | r | | Under Ice | |
|--|-----------------|------------|-------|-----------|--------|-------|-----------|--------|--------|-----------|-------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95th |
| pH, lab | pH units | all sites | 7.63 | 8.02 | 8.17 | 7.60 | 8.04 | 8.20 | + | + | + |
| | pH units | AB07DD0010 | + | + | + | + | + | + | 7.78 | 7.96 | 8.00 |
| | pH units | AB07DD0105 | + | + | + | + | + | + | 7.64 | 7.88 | 8.02 |
| Dissolved Metals Aluminum, Filtered | $\mathrm{ug/L}$ | all sites | 3.55 | 16.20 | 104.85 | 1.84 | 7.96 | 39.06 | 1.92 | 4.23 | 18.39 |
| Antimony, Filtered | $\mathrm{ug/L}$ | all sites | 0.06 | 0.09 | 0.13 | < | < | < | + | + | + |
| | $_{ m ug/L}$ | AB07DD0010 | + | + | + | + | + | + | < | < | < |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | < | < | < |
| Arsenic, Filtered | ug/L | all sites | 0.35 | 0.55 | 0.79 | 0.33 | 0.50 | 0.80 | 0.30 | 0.42 | 0.6 |
| Barium, Filtered | $\mathrm{ug/L}$ | all sites | 34.70 | 42.95 | 49.55 | 40.78 | 45.60 | 53.30 | 44.51 | 59.75 | 70.3 |
| Beryllium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.04 | 0.00 | 0.00 | 0.0 |
| Bismuth, Filtered | $\mathrm{ug/L}$ | all sites | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.02 | < | < | |
| Boron, Filtered | $\mathrm{ug/L}$ | all sites | 15.62 | 22.20 | 30.93 | 17.86 | 22.60 | 29.20 | 24.36 | 31.75 | 37.7 |
| Cadmium, Filtered | $\mathrm{ug/L}$ | all sites | 0.01 | 0.01 | 0.02 | 0.00 | 0.01 | 0.11 | 0.01 | 0.01 | 0.0 |
| Calcium, Filtered | $\mathrm{mg/L}$ | all sites | 17.65 | 25.75 | 31.07 | 25.12 | 31.40 | 36.80 | 29.55 | 40.20 | 48.6 |
| Chlorine, Filtered | $\mathrm{mg/L}$ | all sites | 1.56 | 4.09 | 7.83 | 4.03 | 8.22 | 16.48 | 10.29 | 20.80 | 37.0 |
| Chromium, Filtered | ug/L | all sites | 0.08 | 0.23 | 0.76 | 0.05 | 0.15 | 0.54 | 0.10 | 0.24 | 0.4 |
| Cobalt, Filtered | $\mathrm{ug/L}$ | all sites | 0.04 | 0.07 | 0.13 | 0.04 | 0.07 | 0.22 | + | + | - |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 0.04 | 0.08 | 0.1 |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | 0.02 | 0.06 | 0.1 |
| Copper, Filtered | ug/L | all sites | 0.83 | 1.55 | 2.46 | 0.65 | 0.97 | 2.18 | 0.50 | 0.75 | 1.3 |
| Iron, Filtered | ug/L | all sites | 29.55 | 121.50 | 426.50 | 23.60 | 95.00 | 293.60 | 116.65 | 178.00 | 367.4 |
| Lead, Filtered | $\mathrm{ug/L}$ | all sites | 0.02 | 0.08 | 0.26 | 0.01 | 0.04 | 0.23 | 0.01 | 0.05 | 0.7 |
| Lithium, Filtered | ug/L | all sites | 3.75 | 5.21 | 7.40 | 4.73 | 6.09 | 7.20 | 6.78 | 8.59 | 10.7 |
| Manganese, Filtered | ug/L | all sites | 0.55 | 1.73 | 6.01 | 0.31 | 1.40 | 8.23 | 4.68 | 18.80 | 35.0 |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flow | | (| Open Water | r | | Under Ice | |
|-----------------------------|-----------------|------------|-------|------------------|--------|--------|------------|--------|--------|-----------|--------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95th | 5th | 50th | 95tl |
| Mercury, Filtered | $\mathrm{ng/L}$ | all sites | - | - | - | - | - | - | 0.33 | 0.50 | 1.29 |
| Methylmercury(1+), Filtered | ng/L | all sites | 0.02 | 0.06 | 0.11 | 0.02 | 0.04 | 0.12 | 0.02 | 0.03 | 0.0 |
| Molybdenum, Filtered | ug/L | all sites | 0.15 | 0.49 | 0.70 | 0.38 | 0.63 | 0.98 | 0.52 | 0.64 | 0.75 |
| Nickel, Filtered | ug/L | all sites | 0.36 | 1.43 | 3.48 | 0.29 | 0.75 | 1.33 | 0.07 | 0.76 | 1.4 |
| Selenium, Filtered | ug/L | all sites | 0.05 | 0.11 | 0.26 | 0.18 | 0.24 | 0.30 | 0.14 | 0.25 | 0.4 |
| Silver, Filtered | ug/L | all sites | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.0 |
| Strontium, Filtered | ug/L | all sites | 99.12 | 162.50 | 213.00 | 128.20 | 206.00 | 253.00 | 195.80 | 266.00 | 339.40 |
| Thallium, Filtered | ug/L | all sites | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.0 |
| Thorium, Filtered | ug/L | all sites | 0.00 | 0.03 | 0.13 | 0.00 | 0.01 | 0.06 | 0.00 | 0.01 | 0.0 |
| Tin, Filtered | ug/L | all sites | < | < | < | < | < | < | < | < | |
| Titanium, Filtered | ug/L | all sites | 0.64 | 1.91 | 9.21 | 0.44 | 1.03 | 4.72 | 0.81 | 1.18 | 2.3 |
| Uranium, Filtered | ug/L | all sites | 0.25 | 0.34 | 0.39 | 0.26 | 0.35 | 0.43 | + | + | - |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 0.27 | 0.42 | 0.4 |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | 0.31 | 0.39 | 0.4 |
| Vanadium, Filtered | ug/L | all sites | 0.26 | 0.43 | 0.67 | 0.19 | 0.31 | 0.65 | 0.07 | 0.17 | 0.3 |
| Zinc, Filtered | ug/L | all sites | 0.23 | 0.61 | 1.73 | 0.22 | 0.53 | 1.11 | + | + | - |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 0.75 | 1.02 | 3.5 |
| | $_{ m ug/L}$ | AB07DD0105 | + | + | + | + | + | + | 0.59 | 1.58 | 7.7 |
| actable Metals | | | | | | | | | | | |
| Aluminum, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Antimony, Unfiltered | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Arsenic, Unfiltered | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Barium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Beryllium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bismuth, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | J | Jnder Ice | |
|------------------------|-----------------|-----------|------|------------------|------|-----------------|-----------|------|-----------------|-----------|-----|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | $5 \mathrm{th}$ | 50th | 95th | $5 \mathrm{th}$ | 50th | 95t |
| Boron, Unfiltered | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Cadmium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Calcium, Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Chromium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Cobalt, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Copper, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Iron, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Lead, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Lithium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Manganese, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Molybdenum, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Nickel, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Selenium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Silver, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Strontium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Thallium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Thorium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Tin, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Titanium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Uranium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Vanadium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Zinc, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| i | | | | | | | | | | | |
| Colour (visual) | 1 | all sites | 0.20 | 1.00 | 2.00 | 0.20 | 1.00 | 1.80 | 0.00 | 1.00 | 1. |
| Depth, snow cover | m | all sites | - | - | - | - | - | - | 0.03 | 0.16 | 0. |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flow | | (| Open Water | r | | ${\bf Under\ Ice}$ | |
|--|-----------------|------------|--------|-----------|--------|--------|------------|--------|--------|--------------------|-------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95tl |
| Dissolved oxygen (DO) | $\mathrm{mg/L}$ | all sites | 7.64 | 9.05 | 11.28 | 7.88 | 10.40 | 13.16 | + | + | - |
| | mg/L | AB07DD0010 | + | + | + | + | + | + | 9.87 | 11.32 | 13.4 |
| | mg/L | AB07DD0105 | + | + | + | + | + | + | 8.79 | 10.78 | 12.9 |
| Floating solids or foam | 1 | all sites | 0.00 | 1.00 | 3.00 | 0.00 | 1.00 | 2.00 | 0.00 | 0.00 | 0.0 |
| Ice cover | % | all sites | - | - | - | - | - | - | 88.25 | 100.00 | 100.0 |
| Ice thickness | m | AB07DD0010 | + | + | + | + | + | + | 0.10 | 0.50 | 0.7 |
| | m | AB07DD0105 | + | + | + | + | + | + | 0.26 | 0.70 | 1.3 |
| Odor | 1 | all sites | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 |
| Snow cover | % | all sites | - | - | - | - | - | - | 80.00 | 100.00 | 100.0 |
| Specific conductivity | uS/cm | all sites | 150.06 | 228.60 | 287.38 | 217.25 | 286.20 | 362.00 | + | + | |
| | uS/cm | AB07DD0010 | + | + | + | + | + | + | 137.18 | 425.40 | 510.4 |
| | uS/cm | AB07DD0105 | + | + | + | + | + | + | 271.09 | 401.20 | 486.5 |
| Temperature, water | $\deg C$ | all sites | 7.40 | 17.27 | 21.82 | 1.59 | 10.95 | 21.91 | -0.21 | 0.01 | 0.1 |
| Turbidity, visual | 1 | all sites | 1.00 | 2.00 | 3.00 | 0.00 | 1.00 | 2.00 | 0.00 | 1.00 | 1.1 |
| рН | pH units | all sites | 7.51 | 7.88 | 8.20 | 7.47 | 8.00 | 9.05 | + | + | |
| | pH units | AB07DD0010 | + | + | + | + | + | + | 6.97 | 7.43 | 8.2 |
| | pH units | AB07DD0105 | + | + | + | + | + | + | 6.33 | 7.25 | 7.6 |
| neral Organics 12-Chlorodehydroabietic acid | $\mathrm{ug/L}$ | all sites | - | - | - | _ | _ | - | - | _ | |
| 14-Chlorodehydroabietic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4-Dinitrotoluene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,6-Dinitrotoluene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2-Chloroethyl vinyl ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 3,4,5-Trichlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 3,4,5-Trichloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | Ţ | Under Ice | |
|-------------------------|-----------------|-----------|-----|------------------|------|-----------------|-----------|------|-----|-----------|------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | $5 \mathrm{th}$ | 50th | 95th | 5th | 50th | 95tl |
| 3,4,6-Trichlorocatechol | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 3,4,6-Trichloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 3,4-Dichlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 3,4-Dichloroguaiacol | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 3,5-Dichlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 3,6-Dichlorocatechol | mg/L | all sites | - | - | - | - | - | - | - | - | |
| 4,5,6-Trichloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4,5,6-Trichlorosyringol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4,5-Dichlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4,5-Dichloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4,5-Dichloroveratrole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4,6-Dichloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4-Chlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4-Chloroguaiacol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Abietic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Arachidic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| BTEX, Total | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | < | < | |
| Benzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Benzidine | ug/L | all sites | - | - | - | - | - | - | - | - | |
| C10-C16 Hydrocarbons | ug/L | all sites | - | - | - | - | - | - | < | < | |
| C16-C34 Hydrocarbons | ug/L | all sites | < | < | < | < | < | < | < | < | |
| C34-C50 Hydrocarbons | ug/L | all sites | - | - | - | - | - | - | < | < | |
| C6-C10 Hydrocarbons | ug/L | all sites | - | - | - | - | - | - | < | < | |
| Cumene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Cyanide, Unknown | mg/L | all sites | < | < | | | | | | | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | Ţ | Under Ice | |
|-------------------------------|-----------------|-----------|------|-----------|------|----------------|-----------|------|----------------|-----------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95tl |
| Dehydroabietic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Ethylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Isophorone | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Isopimaric acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Levopimaric acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Linoleic acid | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Methyl tert-butyl ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Myristic acid | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| N-Nitrosodi-n-propylamine | ug/L | all sites | - | - | - | - | - | - | - | - | |
| N-Nitrosodiphenylamine | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Naphthenic acids | mg/L | all sites | 0.07 | 0.23 | 0.41 | 0.07 | 0.14 | 0.27 | 0.05 | 0.19 | 0.52 |
| Neoabietic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Nitrobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Oilsands extractable organics | mg/L | all sites | 0.28 | 0.66 | 6.95 | 0.15 | 0.40 | 2.93 | 0.14 | 0.50 | 1.66 |
| Oleic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Palmitic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Palustric acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Pimaric acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| S-Ethyl dipropylthiocarbamate | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Sandaracopimaric acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Stearic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Styrene | ug/L | all sites | - | - | - | - | - | - | < | < | < |
| Tetrachlorocatechol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Tetrachloroguaiacol | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Tetrachloroveratrole | ug/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flow | | O | pen Water | | 1 | Under Ice | |
|---|-----------------|------------|-------|------------------|-------|-------|------------------|-------|-------|------------------|-----|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95t |
| Toluene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Vinyl chloride | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Xylene | ug/L | all sites | - | - | - | - | - | - | < | < | |
| m,p-Xylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| n-Butylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| n-Propylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| o-Xylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Cymene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| sec-Butylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| tert-Butylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| or Ions | | | | | | | | | | | |
| Calcium, Filtered | $\mathrm{mg/L}$ | all sites | 20.40 | 27.00 | 33.80 | 26.00 | 33.00 | 37.80 | 32.00 | 42.00 | 49. |
| Chlorate, Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Chloride, Unfiltered | $\mathrm{mg/L}$ | all sites | 3.70 | 6.00 | 12.40 | 6.04 | 12.00 | 21.40 | 13.90 | 25.00 | 40. |
| Fluoride, Unfiltered | $\mathrm{mg/L}$ | all sites | 0.08 | 0.10 | 0.12 | 0.09 | 0.10 | 0.13 | 0.10 | 0.12 | 0. |
| Magnesium, Filtered | $\mathrm{mg/L}$ | all sites | 4.84 | 7.90 | 9.74 | 8.32 | 9.40 | 11.80 | + | + | |
| | $\mathrm{mg/L}$ | AB07DD0010 | + | + | + | + | + | + | 9.42 | 13.00 | 15. |
| | $\mathrm{mg/L}$ | AB07DD0105 | + | + | + | + | + | + | 9.65 | 12.00 | 14. |
| Potassium, Filtered | $\mathrm{mg/L}$ | all sites | 0.74 | 1.30 | 2.60 | 0.96 | 1.20 | 1.48 | 1.29 | 1.80 | 2. |
| Sodium, Filtered | $\mathrm{mg/L}$ | all sites | 8.20 | 9.40 | 15.80 | 10.20 | 16.00 | 20.00 | 20.70 | 29.00 | 40. |
| Sulfate, Unfiltered as SO4 | $\mathrm{mg/L}$ | all sites | 14.00 | 23.00 | 28.80 | 19.40 | 28.00 | 39.00 | 27.80 | 36.00 | 47. |
| Sulfide, Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| rients and BOD Ammonia and ammonium, Unfiltered as N | mg/L | all sites | < | < | < | 0.01 | 0.02 | 0.08 | 0.02 | 0.05 | 0. |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | I | High Flow | | O | pen Water | | Ţ | Under Ice | |
|---|-----------------|-----------|------|-----------|-------|-----------------|-----------|-------|-----------------|------------------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | $5 \mathrm{th}$ | 50th | 95th | $5 \mathrm{th}$ | $50 \mathrm{th}$ | 95t |
| Biochemical oxygen demand, standard conditions, Filtered | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Carbonaceous biochemical oxygen demand, non-standard conditions | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Chlorophyll a | ug/L | all sites | 1.32 | 6.21 | 11.22 | 4.02 | 6.40 | 13.02 | 0.26 | 0.40 | 4.5 |
| Inorganic nitrogen (nitrate and nitrite), Unfiltered as N | mg/L | all sites | 0.02 | 0.05 | 0.11 | - | - | - | 0.03 | 0.17 | 0. |
| Kjeldahl nitrogen, Unfiltered as N | mg/L | all sites | 0.33 | 0.70 | 1.70 | 0.18 | 0.45 | 0.86 | 0.26 | 0.41 | 0. |
| Nitrate, Unfiltered as N | mg/L | all sites | 0.02 | 0.05 | 0.11 | - | - | - | 0.03 | 0.17 | 0. |
| Nitrite, Unfiltered as N | mg/L | all sites | - | - | - | - | - | - | < | < | |
| Orthophosphate, Filtered as P | mg/L | all sites | 0.00 | 0.00 | 0.01 | < | < | < | 0.00 | 0.00 | 0. |
| Silica, reactive, Unknown | mg/L | all sites | 3.20 | 5.80 | 6.40 | - | - | - | - | - | |
| Total Phosphorus, mixed forms, Filtered as P | $\mathrm{mg/L}$ | all sites | 0.01 | 0.01 | 0.03 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0 |
| Total Phosphorus, mixed forms, Unfiltered as P | $\mathrm{mg/L}$ | all sites | 0.04 | 0.11 | 0.23 | 0.01 | 0.04 | 0.19 | 0.02 | 0.02 | 0 |
| nohalides | | | | | | | | | | | |
| 1,1,1,2-Tetrachloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,1,1-Trichloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,1,2,2-Tetrachloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,1,2-Trichloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,1-Dichloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,1-Dichloroethylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,2,3-Trichlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,2,3-Trichloropropane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,2,4-Trichlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,2,4-Trimethylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | 1 | High Flow | | O | pen Water | | Ţ | Under Ice | |
|-----------------------------------|-----------------|-----------|-----|-----------|------|----------------|-----------|------|----------------|-----------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95th |
| 1,2-Dibromo-3-chloropropane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,2-Dichloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| 1,2-Dichloropropane | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| 1,2-Diphenylhydrazine | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,3,5-Trimethylbenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,3-DICHLOROPROPANE | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1,3-Dichlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 1-Propene, 1,1-dichloro- | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 12,14-Dichlorodehydroabietic acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,2-Dichloropropane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4,6-Trichloroanisole | mg/L | all sites | - | - | - | - | - | - | - | - | |
| 2,6-Dichlorosyringaldehyde | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 2-Chloronaphthalene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| 2-Chlorosyringaldehyde | mg/L | all sites | - | - | - | - | - | - | - | - | |
| 4-Bromophenyl phenyl ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 5,6-Dichlorovanillin | $_{ m mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 5-Chlorovanillin | mg/L | all sites | - | - | - | - | - | - | - | - | |
| 6-Chlorovanillin | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 9,10-Dichlorostearic Acid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Adsorbable Organic Halide | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bis(2-chloroethoxy)methane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bis(2-chloroethyl) ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bis(2-chloroisopropyl) ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bromobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| CFC-11 | ug/L | all sites | _ | _ | _ | _ | _ | _ | _ | _ | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | Ţ | Under Ice | |
|-----------------------------|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|-----------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95t |
| Carbon tetrachloride | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chlorodibromomethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chloroform | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chloromethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Dibromomethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Dichlorobromomethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Ethylene dibromide | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Hexachlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Hexachlorobutadiene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Hexachlorocyclopentadiene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Hexachloroethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Methyl bromide | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Methylene chloride | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Tetrachloroethylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Tribromomethane | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Trichloroethylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| cis-1,2-Dichloroethylene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| cis-1,3-Dichloropropene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| o-Chlorotoluene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| o-Dichlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Chlorophenyl phenyl ether | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Chlorotoluene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Dichlorobenzene | ug/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | I | High Flow | | O | pen Water | | Ţ | Under Ice | |
|--|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|-----------|----|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95 |
| trans-1,2-Dichloroethene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| trans-1,3-Dichloropropene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| S | | | | | | | | | | | |
| 1-Methylnaphthalene | ng/L | all sites | - | - | - | - | - | - | < | < | |
| 2-Methylnaphthalene | ng/L | all sites | - | - | - | - | - | - | < | < | |
| 3-Methylcholanthrene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| $7,\!12\text{-}Dimethylbenz[a] anthracene$ | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Acenaphthene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Acenaphthylene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Benz[a]anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | |
| Benzo(b)fluoranthene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Benzo[a]pyrene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Benzo[b,j,k]fluoranthene | ug/L | all sites | - | - | - | - | - | - | < | < | |
| Benzo[c]phenanthrene | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Benzo[e]pyrene | ng/L | all sites | - | - | - | - | - | - | < | < | |
| Benzo[ghi]perylene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Benzo[k]fluoranthene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| C1-Dibenzothiophenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C1-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C2-Chrysenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C2-Dibenzothiophenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C2-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C2-Fluorenes | ng/L | all sites | < | < | < | - | - | - | < | < | |
| C2-Naphthalenes | ng/L | all sites | < | < | < | - | - | - | < | < | |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | Under Ice | | |
|------------------------------|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95th |
| C2-Phenanthrenes/anthracenes | ug/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Chrysenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Dibenzothiophenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Fluorenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Naphthalenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C3-Phenanthrenes/anthracenes | ug/L | all sites | < | < | < | - | - | - | < | < | < |
| C4-Chrysenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C4-Dibenzothiophenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C4-Fluoranthenes/pyrenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C4-Fluorenes | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| C4-Naphthalenes | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| C4-Phenanthrenes/anthracenes | ug/L | all sites | < | < | < | - | - | - | < | < | < |
| Chrysene | ng/L | all sites | - | - | - | - | - | - | - | - | - |
| Dibenz[a,h]anthracene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Dibenzo[a,h]pyrene | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Dibenzo[a,i]pyrene | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Dibenzo[a,l]pyrene | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Fluoranthene | ng/L | all sites | - | - | - | - | - | - | - | - | - |
| Fluorene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Indeno[1,2,3-cd]pyrene | ng/L | all sites | < | < | < | < | < | < | < | < | < |
| Methylchrysene | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| Methylfluorene | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| Methylphenanthrene | ng/L | all sites | < | < | < | - | - | - | < | < | < |
| Naphthalene | ng/L | all sites | - | - | - | - | - | - | - | - | - |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | O | pen Water | | Under Ice | | |
|--|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95t |
| Perylene | ng/L | all sites | - | - | - | - | - | - | < | < | |
| Phenanthrene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Pyrene | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Retene | ng/L | all sites | - | - | - | - | - | - | < | < | |
| sticide | | | | | | | | | | | |
| .alphaEndosulfan | ug/L | all sites | < | < | < | < | < | < | - | - | |
| .lambdaCyhalothrin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4-D | ug/L | all sites | < | < | < | < | < | < | - | - | |
| 2,4-DB | ug/L | all sites | < | < | < | < | < | < | - | - | |
| 2-Chloro-4-isopropylamino-6-amino-s-triazine | ug/L | all sites | < | < | < | < | < | < | - | - | |
| 2-Choro-6-ethylamino-4-amino-s-triazine | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Aldicarb | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Aldicarb sulfone | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Aldicarb sulfoxide | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Aldrin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Aminocarb | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Aminopyralid | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Atrazine | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Atrazine de-ethylated | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Azinphos-methyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Azoxystrobin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Benomyl | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Bentazon | ug/L | all sites | < | < | < | < | < | < | _ | - | |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | I | High Flow | | O | pen Water | | Under Ice | | |
|-----------------------------------|------|-----------|-----|-----------|------|-----------------|-----------|------|----------------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5 \mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95tl |
| Benzene Hexachloride, Alpha (BHC) | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Bromacil | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Bromoxynil | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Carbaryl | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Carbofuran | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Carboxin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Chlorothalonil | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Chlorpyrifos | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Clodinafop acid metabolite | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Clodinafop-propargyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Clopyralid | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Clothianidin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Cyanazine | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Deltamethrin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Diazinon | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Dicamba | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Dichlorprop | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Diclofop methyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Dieldrin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Difenoconazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Dimethoate | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Disulfoton | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Diuron | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Ethalfluralin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Ethion | ug/L | all sites | < | < | < | < | < | < | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | |] | High Flow | | | pen Water | | Under Ice | | |
|-----------------------|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95tl |
| Ethofumesate | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Fenoxaprop-p-ethyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Fenoxaprop-p-methyl | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Fluazifop-P-butyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Fluroxypyr | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Hexaconazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Imazamethabenz-methyl | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Imazamox | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Imazethapyr | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Imidacloprid | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Iprodione | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Lindane | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Linuron | ug/L | all sites | < | < | < | < | < | < | - | - | |
| MCPA | ug/L | all sites | < | < | < | < | < | < | - | - | |
| MCPB | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Malathion | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Mecoprop | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Metalaxyl-M | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Metconazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Methomyl | ug/L | all sites | < | < | < | - | - | - | - | - | |
| Methoxychlor | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Metolachlor | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Metribuzin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Monuron | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Napropamide | ug/L | all sites | < | < | < | < | < | < | _ | - | - |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | 1 | High Flow | | O | pen Water | | Under Ice | | |
|-------------------------------------|------|-----------|-----|-----------|------|----------------|-----------|------|----------------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5\mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95tl |
| OH-Carbofuran | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Oxycarboxin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Parathion | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Permethrin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Phorate | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Picloram | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Picoxystrobin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Propiconazole | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Prothioconazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Pyraclostrobin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Pyridaben | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Quinclorac | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Quizalofop | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Simazine | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Tebuconazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Terbufos | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Thiamethoxam | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Triallate | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Triclopyr | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Trifloxystrobin | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Trifluralin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| Triticonazole | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Vinclozolin | ug/L | all sites | < | < | < | < | < | < | - | - | |
| oolics 2,3,4,6-Tetrachlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |

Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | 1 | High Flow | | О | pen Water | | Under Ice | | |
|---|-----------------|-----------|------|-----------|------|-----------------|-----------|------|----------------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | $5 \mathrm{th}$ | 50th | 95th | $5\mathrm{th}$ | 50th | 95th |
| 2,4,5-Trichlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| 2,4,6-Trichlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4-Dichlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4-Dichlorophenol/2,5- Dichlorophenol | mg/L | all sites | - | - | - | - | - | - | - | - | |
| 2,4-Dimethylphenol | $\mathrm{ug/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 2,4-Dinitrophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 2,6-Dichlorophenol | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | |
| 4,6-Dinitro-o-cresol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| 4-Chloro-2-methylphenol | ug/L | all sites | < | < | < | < | < | < | - | - | |
| 4-Chlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Pentachlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Phenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Phenolics | $\mathrm{mg/L}$ | all sites | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.0 |
| o-Chlorophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| o-Nitrophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Chloro-m-cresol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| p-Nitrophenol | ug/L | all sites | - | - | - | - | - | - | - | - | |
| chalates Butyl benzyl phthalate | ug/L | all sites | _ | _ | _ | _ | _ | _ | _ | _ | |
| Di(2-ethoxylhexyl) phthalate | ug/L | all sites | _ | | | _ | _ | _ | | | |
| Di-n-octyl phthalate | ug/L | all sites | | | | | | | | | |
| Dibutyl phthalate | | all sites | - | | | | | | - | - | |
| Diethyl phthalate | ug/L | all sites | - | - | - | - | - | - | - | - | |
| | ug/L | | | - | | | | - | | - | |
| Dimethyl phthalate | ug/L | all sites | _ | - | - | - | - | - | - | - | |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flov | 7 | (| Open Wate | er | | Under Ice | |
|---|-----------------|-----------|---------|------------------|----------|--------|-----------|---------|--------|-----------|---------|
| Parameter | Unit | Site | 5th | $50 \mathrm{th}$ | 95th | 5th | 50th | 95th | 5th | 50th | 95th |
| Target PANHs | | | | | | | | | | | |
| Acridine | ug/L | all sites | - | _ | - | - | _ | - | - | _ | - |
| Total Metals | | | | | | | | | | | |
| Chromium(VI), Unknown | $\mathrm{mg/L}$ | all sites | < | < | < | - | - | - | - | - | - |
| Mercury, Unfiltered | ng/L | all sites | 3.42 | 8.90 | 23.80 | 0.80 | 2.99 | 13.70 | 0.46 | 0.82 | 4.25 |
| Methylmercury(1+), Unfiltered | $\mathrm{ng/L}$ | all sites | 0.03 | 0.16 | 0.25 | 0.04 | 0.07 | 0.19 | 0.03 | 0.04 | 0.10 |
| Cotal Recoverable Metals Aluminum, Unfiltered | m ug/L | all sites | 396.75 | 2770.00 | 13475.00 | 142.40 | 792.00 | 5480.00 | 26.60 | 97.50 | 1202.25 |
| Antimony, Unfiltered | ug/L | all sites | 0.07 | 0.10 | 0.15 | 0.03 | 0.07 | 0.28 | 0.04 | 0.05 | 0.12 |
| Arsenic, Unfiltered | ug/L | all sites | 0.72 | 1.75 | 2.91 | 0.50 | 0.86 | 1.95 | 0.42 | 0.57 | 0.83 |
| Barium, Unfiltered | $\mathrm{ug/L}$ | all sites | 55.85 | 86.15 | 239.25 | 46.06 | 56.90 | 141.06 | 49.84 | 64.05 | 77.97 |
| Beryllium, Unfiltered | ug/L | all sites | 0.03 | 0.14 | 0.47 | 0.01 | 0.04 | 0.23 | 0.00 | 0.01 | 0.11 |
| Bismuth, Unfiltered | ug/L | all sites | 0.01 | 0.02 | 0.06 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.02 |
| Boron, Unfiltered | ug/L | all sites | 17.00 | 24.80 | 41.77 | 20.70 | 24.70 | 40.54 | 24.30 | 32.85 | 39.78 |
| Cadmium, Unfiltered | ug/L | all sites | 0.02 | 0.06 | 0.27 | 0.01 | 0.02 | 0.13 | 0.01 | 0.02 | 0.09 |
| Calcium, Unfiltered | $\mathrm{mg/L}$ | all sites | 19.57 | 27.85 | 35.48 | 25.82 | 32.40 | 38.18 | 29.82 | 40.50 | 50.23 |
| Chlorine, Unfiltered | $\mathrm{mg/L}$ | all sites | 1.58 | 4.12 | 7.88 | 4.06 | 8.40 | 16.74 | 10.89 | 20.80 | 38.17 |
| Chromium, Unfiltered | ug/L | all sites | 0.69 | 3.21 | 11.71 | 0.15 | 0.92 | 6.31 | 0.05 | 0.22 | 0.68 |
| Cobalt, Unfiltered | ug/L | all sites | 0.39 | 1.35 | 4.94 | 0.17 | 0.41 | 1.87 | 0.06 | 0.12 | 0.43 |
| Copper, Unfiltered | ug/L | all sites | 1.63 | 3.65 | 10.13 | 0.94 | 1.42 | 4.81 | 0.54 | 0.91 | 1.90 |
| Iron, Unfiltered | ug/L | all sites | 1292.50 | 4240.00 | 13625.00 | 454.20 | 1050.00 | 4414.00 | 412.75 | 565.50 | 1294.50 |
| Lead, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.54 | 2.12 | 10.55 | 0.17 | 0.47 | 2.81 | 0.07 | 0.16 | 2.56 |
| Lithium, Unfiltered | $\mathrm{ug/L}$ | all sites | 5.16 | 7.46 | 16.95 | 5.83 | 6.83 | 8.13 | 7.04 | 8.92 | 11.09 |
| Manganese, Unfiltered | ug/L | all sites | 44.25 | 104.40 | 320.50 | 19.80 | 54.70 | 113.80 | 16.82 | 30.75 | 51.66 |
| Molybdenum, Unfiltered | ug/L | all sites | 0.15 | 0.52 | 0.73 | 0.38 | 0.60 | 0.98 | 0.54 | 0.65 | 0.77 |

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Table 2.6: Current Conditions, Athabasca River Delta water. (continued)

| | | | | High Flow | | (| Open Water | r | Under Ice | | | |
|-----------------------|-----------------------|------------|--------|-----------|--------|--------|------------|--------|-----------|--------|--------|--|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95th | |
| Nickel, Unfiltered | $\mathrm{ug/L}$ | all sites | 1.50 | 4.33 | 13.17 | 0.60 | 1.55 | 4.97 | 0.10 | 1.01 | 2.25 | |
| Selenium, Unfiltered | ug/L | all sites | 0.15 | 0.26 | 0.47 | 0.15 | 0.22 | 0.30 | 0.19 | 0.30 | 0.50 | |
| Silver, Unfiltered | ug/L | all sites | 0.01 | 0.02 | 0.33 | 0.00 | 0.01 | 0.03 | + | + | + | |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 0.00 | 0.00 | 0.01 | |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | 0.00 | 0.00 | 0.02 | |
| Strontium, Unfiltered | ug/L | all sites | 111.00 | 174.50 | 227.50 | 129.40 | 206.00 | 256.60 | 197.10 | 275.00 | 343.40 | |
| Thallium, Unfiltered | ug/L | all sites | 0.02 | 0.05 | 0.21 | 0.01 | 0.02 | 0.11 | 0.00 | 0.01 | 0.05 | |
| Thorium, Unfiltered | ug/L | all sites | 0.09 | 0.42 | 2.51 | 0.03 | 0.14 | 0.88 | 0.01 | 0.02 | 0.20 | |
| Tin, Unfiltered | $\mathrm{ug/L}$ | all sites | 0.02 | 0.05 | 0.11 | < | < | < | 0.01 | 0.04 | 0.10 | |
| Titanium, Unfiltered | $\mathrm{ug/L}$ | all sites | 6.74 | 33.90 | 127.00 | 2.78 | 11.60 | 69.98 | 1.73 | 2.53 | 22.63 | |
| Uranium, Unfiltered | ug/L | all sites | 0.36 | 0.49 | 1.27 | 0.32 | 0.41 | 0.65 | + | + | + | |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 0.28 | 0.44 | 0.52 | |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | 0.31 | 0.40 | 0.52 | |
| Vanadium, Unfiltered | ug/L | all sites | 1.58 | 6.73 | 21.23 | 0.64 | 2.04 | 12.25 | 0.25 | 0.43 | 2.04 | |
| Zinc, Unfiltered | ug/L | all sites | 3.27 | 10.36 | 32.95 | 1.40 | 3.10 | 15.63 | + | + | + | |
| | ug/L | AB07DD0010 | + | + | + | + | + | + | 1.02 | 1.65 | 6.98 | |
| | ug/L | AB07DD0105 | + | + | + | + | + | + | 1.05 | 2.58 | 13.22 | |

Note:

- data insufficient
- < too highly censored;
- + grouped differently (merged sites vs individual site);

Table 2.7: Current Conditions, Athabasca River Delta sediment.

| Parameter | Unit | Site | 5th | 50th | 95t |
|--|------|-----------|--------|--------|--------|
| Conventional Variables Acid Neutralization Potential as %CaCO3 | % | all sites | 1.61 | 5.51 | 8.3 |
| Grain size, clay (<2 um) | % | all sites | 3.07 | 16.10 | 33.2 |
| Grain size, sand (>=63 um to 2000 um) | % | all sites | 3.39 | 34.50 | 92.0 |
| Grain size, silt (>=2 to 63 um) | % | all sites | 4.57 | 48.20 | 72.3 |
| Inorganic carbon | % | all sites | 0.24 | 0.74 | 1.0 |
| Moisture content | % | all sites | 22.25 | 34.20 | 56.3 |
| Organic carbon | % | all sites | 0.53 | 1.44 | 2.5 |
| Total carbon | % | all sites | 0.77 | 2.10 | 3.5 |
| General Organics | | | | | |
| AEP Total recoverable hydrocarbons | ug/g | all sites | 600.00 | 700.00 | 1400.0 |
| BTEX, Total | ug/g | all sites | - | - | |
| Benzene | ug/g | all sites | < | < | |
| C10-C16 Hydrocarbons | ug/g | all sites | 15.48 | 26.65 | 48.6 |
| C11-C30 AEP Total extractable hydrocarbons | ug/g | all sites | 54.00 | 200.00 | 500.0 |
| C16-C34 Hydrocarbons | ug/g | all sites | 33.42 | 216.00 | 394.5 |
| C34-C50 Hydrocarbons | ug/g | all sites | 33.45 | 172.00 | 424.5 |
| C5-C10 AEP Total volatile hydrocarbons | ug/g | all sites | 0.79 | 2.35 | 8.5 |
| Ethylbenzene | ug/g | all sites | < | < | |
| Hydrocarbons | ug/g | all sites | 85.25 | 405.50 | 715. |
| Styrene | ug/g | all sites | - | - | |
| Toluene | ug/g | all sites | < | < | |
| Total xylenes | ug/g | all sites | - | - | |
| m,p-Xylene | ug/g | all sites | < | < | |
| o-Xylene | ug/g | all sites | < | < | |
| AHs 1,2,6-Trimethylphenanthrene | ng/g | all sites | - | _ | |
| 1,2-Dimethylnaphthalene | ng/g | all sites | - | - | |
| 1,4,6,7-Tetramethylnaphthalene | ng/g | all sites | _ | _ | |
| 1,6,7-Trimethylnaphthalene | ng/g | all sites | | _ | |
| 1,7-Dimethylfluorene | ng/g | all sites | _ | _ | |
| 1,7-Dimethylphenanthrene | ng/g | all sites | | _ | |
| 1,8-Dimethylphenanthrene | ng/g | all sites | _ | _ | |
| 1-Methylchrysene | ng/g | all sites | | _ | |
| 1-Methylnaphthalene | ng/g | all sites | | _ | |
| 1-Methylphenanthrene | ng/g | all sites | - | _ | |
| 2,3,6-Trimethylnaphthalene | ng/g | all sites | _ | | |
| 2,4-Dimethyldibenzothiophene | ng/g | all sites | - | _ | |
| 2,6-Dimethylnaphthalene | ng/g | all sites | _ | | |
| 2,6-Dimethylphenanthrene | ng/g | all sites | - | - | |
| 2-Methylanthracene | ng/g | all sites | _ | _ | |
| 2-Methyldibenzothiophenes/3-Methyldibenzothiophenes | ng/g | all sites | - | - | |
| 2-Methylfluorene | ng/g | all sites | - | _ | |
| 2-Methylnaphthalene | ng/g | all sites | | | |

Table 2.7: Current Conditions, Athabasca River Delta sediment. (continued)

| Parameter | Unit | Site | 5th | 50th | 95t |
|---|-------|-----------|-------|-------|-------|
| 2-Methylphenanthrene | ng/g | all sites | - | | |
| 3,6-Dimethylphenanthrene | ng/g | all sites | - | - | |
| 3-Methylfluoranthene/Benzo[a]fluorene | ng/g | all sites | - | - | |
| 3-Methylphenanthrene | ng/g | all sites | - | - | |
| 5,9-Dimethylchrysene | ng/g | all sites | - | - | |
| 5-Methylchrysene/6-Methylchrysene | ng/g | all sites | - | - | |
| 7-Methylbenzo[a]pyrene | ng/g | all sites | _ | - | |
| 9-Methylphenanthrene/4- Methylphenanthrene | ng/g | all sites | - | - | |
| Acenaphthene | ng/g | all sites | < | < | |
| Acenaphthylene | ng/g | all sites | < | < | |
| Anthracene | ng/g | all sites | < | < | |
| Benz[a]anthracene | ng/g | all sites | < | < | |
| Benzo(b)fluoranthene | ng/g | all sites | _ | - | |
| Benzo(j+k)fluoranthene | ng/g | all sites | - | - | |
| Benzo[a]pyrene | ng/g | all sites | 3.39 | 5.88 | 10.2 |
| Benzo[b,j,k]fluoranthene | ng/g | all sites | 3.30 | 15.65 | 27.7 |
| Benzo[e]pyrene | ng/g | all sites | _ | _ | |
| Benzo[ghi]perylene | ng/g | all sites | 3.44 | 10.45 | 18.4 |
| Biphenyl | ng/g | all sites | 1.69 | 5.87 | 10.6 |
| C1-Acenaphthenes | ng/g | all sites | < | < | |
| C1-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 7.73 | 67.95 | 256.7 |
| C1-Benzofluoranthenes/benzopyrenes | ng/g | all sites | 17.39 | 47.45 | 87.6 |
| C1-Biphenyls | ng/g | all sites | 3.30 | 6.80 | 14.4 |
| C1-Dibenzothiophenes | ng/g | all sites | 3.46 | 11.35 | 22.9 |
| C1-Fluoranthenes/pyrenes | ng/g | all sites | 17.90 | 46.25 | 135.5 |
| C1-Fluorenes | ng/g | all sites | 3.26 | 8.54 | 25.5 |
| C1-Naphthalenes | ng/g | all sites | 5.87 | 26.25 | 48.4 |
| C1-Phenanthrenes/anthracenes | ng/g | all sites | 7.01 | 37.80 | 77.2 |
| C2-Benzo[a]anthracenes/chrysenes | ng/g | all sites | < | < | |
| C2-Benzofluoranthenes/benzopyrenes | ng/g | all sites | 9.50 | 21.15 | 39.2 |
| C2-Biphenyls | ng/g | all sites | 2.97 | 8.62 | 25.8 |
| C2-Dibenzothiophenes | ng/g | all sites | 15.80 | 49.45 | 108.8 |
| C2-Fluoranthenes/pyrenes | ng/g | all sites | 31.49 | 80.80 | 243.7 |
| C2-Fluorenes | ng/g | all sites | 8.81 | 26.50 | 55.4 |
| C2-Naphthalenes | ng/g | all sites | 11.60 | 43.00 | 78.9 |
| C2-Phenanthrenes/anthracenes | ng/g | all sites | 5.43 | 52.25 | 96.1 |
| C3-Benzo[a]anthracenes/chrysenes | ng/g | all sites | - | - | 00.1 |
| C3-Dibenzothiophenes | ng/g | all sites | 27.12 | 92.50 | 253.5 |
| C3-Fluoranthenes/pyrenes | ng/g | all sites | 28.47 | 78.20 | 198.9 |
| C3-Fluorenes | ng/g | all sites | 12.00 | 37.75 | 104.2 |
| C3-Naphthalenes | ng/g | all sites | 10.54 | 37.35 | 61.7 |
| C3-Phenanthrenes/anthracenes | ng/g | all sites | 19.91 | 59.00 | 144.7 |
| C4-Benzo[a]anthracenes/chrysenes | ng/g | all sites | 10.01 | - | 177.1 |
| CT Denzo[a]anomacenes/ em ysenes | 118/8 | an altes | - | - | |

Table 2.7: Current Conditions, Athabasca River Delta sediment. (continued)

| Parameter | | | | | | |
|------------------|----------------|------|-----------|---------|----------|---------|
| 1 0101110001 | | Unit | Site | 5th | 50th | 95t |
| C4-Fluoranthene | es/pyrenes | ng/g | all sites | - | - | |
| C4-Naphthalene | s | ng/g | all sites | 10.15 | 27.80 | 55.8 |
| C4-Phenanthren | es/anthracenes | ng/g | all sites | 24.50 | 248.00 | 543.7 |
| Chrysene | | ng/g | all sites | 3.43 | 17.75 | 30.3 |
| Dibenz[a,h]anth | racene | ng/g | all sites | < | < | |
| Dibenzothiopher | ne | ng/g | all sites | < | < | |
| Fluoranthene | | ng/g | all sites | 1.14 | 3.87 | 7.1 |
| Fluorene | | ng/g | all sites | 0.38 | 2.30 | 4.5 |
| Indeno[1,2,3-cd] | pyrene | ng/g | all sites | 2.25 | 6.22 | 11.5 |
| Naphthalene | | ng/g | all sites | 2.17 | 7.75 | 20.2 |
| Perylene | | ng/g | all sites | _ | _ | |
| Phenanthrene | | ng/g | all sites | 3.72 | 15.95 | 27.2 |
| Pyrene | | ng/g | all sites | 3.22 | 10.45 | 18.5 |
| Retene | | ng/g | all sites | 12.88 | 52.10 | 132.7 |
| otal Metals | | 5, 5 | | | | |
| Aluminum | | ug/g | all sites | 3314.00 | 7800.00 | 14340.0 |
| Antimony | | ug/g | all sites | 0.13 | 0.22 | 0.3 |
| Arsenic | | ug/g | all sites | 2.97 | 4.95 | 8.1 |
| Barium | | ug/g | all sites | 66.33 | 149.50 | 213.5 |
| Beryllium | | ug/g | all sites | < | < | |
| Bismuth | | ug/g | all sites | < | < | |
| Boron | | ug/g | all sites | 4.00 | 10.00 | 23.4 |
| Cadmium | | ug/g | all sites | < | < | |
| Calcium | | ug/g | all sites | 9030.00 | 21100.00 | 27880.0 |
| Chromium | | ug/g | all sites | 7.65 | 14.95 | 32.8 |
| Cobalt | | ug/g | all sites | 5.03 | 7.70 | 11.2 |
| Copper | | ug/g | all sites | 4.54 | 13.10 | 22.2 |
| Iron | | ug/g | all sites | 8956.00 | 17500.00 | 26380.0 |
| Lead | | ug/g | all sites | 3.85 | 7.91 | 12.1 |
| Lithium | | ug/g | all sites | 2.19 | 10.70 | 20. |
| Magnesium | | ug/g | all sites | 3518.00 | 7340.00 | 9310.0 |
| Manganese | | ug/g | all sites | 172.80 | 392.00 | 632.6 |
| Mercury | | ug/g | all sites | 0.02 | 0.04 | 0.0 |
| Molybdenum | | ug/g | all sites | < | < | |
| Nickel | | ug/g | all sites | 10.19 | 18.75 | 29.4 |
| Phosphorus | | ug/g | all sites | 185.50 | 610.50 | 767.5 |
| Potassium | | ug/g | all sites | 525.50 | 1200.00 | 2100.0 |
| Selenium | | ug/g | all sites | 0.19 | 0.41 | 1.0 |
| Silver | | ug/g | all sites | - | - | |
| Sodium | | ug/g | all sites | 72.89 | 140.00 | 277.5 |
| Dodiani | | ug/g | all sites | 26.70 | 60.50 | 80.5 |
| Strontium | | | | | | |
| - | | ug/g | all sites | 0.09 | 0.16 | 0.2 |
| Strontium | | ug/g | all sites | 0.09 | 0.16 | 0.2 |
| Strontium | | | | | | |

Table 2.7: Current Conditions, Athabasca River Delta sediment. (continued)

| Parameter | Unit | Site | $5\mathrm{th}$ | 50th | 95th |
|-----------|------|-----------|----------------|-------|-------|
| Vanadium | ug/g | all sites | 12.82 | 21.70 | 36.10 |
| Zinc | ug/g | all sites | 29.82 | 59.35 | 83.53 |
| Zirconium | ug/g | all sites | - | - | - |

Note:

- data insufficient
- < too highly censored;

1292 2.8.3 Lake Athabasca Current Conditions

The current condition (5th, 50th, and 95th percentile values) for each water quality parameter and each season are presented for Lake Athabasca in Table 2.8 (water). Note that additional information, including sample size, analytical method codes, and quantile estimation method for each suite of current conditions are provided in Appendix A.2.

Table 2.8: Current Conditions, Lake Athabasca water.

| | Parameter | Unit Site | |] | High Flow | | Open Water | | | Under Ice | | |
|-------|---|-----------------|-----------|---------|-----------|--------|------------|--------|--------|-----------|------|------|
| | | | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95tl |
| Conve | entional Variables | | | | | | | | | | | |
| | Alkalinity, total | $\mathrm{mg/L}$ | all sites | - | - | - | 30.20 | 35.20 | 99.30 | - | - | |
| | Hardness as CaCO3 | $\mathrm{mg/L}$ | all sites | - | - | - | 31.20 | 38.54 | 104.00 | - | - | |
| | Organic carbon, Filtered | $\mathrm{mg/L}$ | all sites | - | - | - | 3.30 | 4.35 | 13.50 | - | - | |
| | Organic carbon, Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | 3.50 | 4.15 | 13.10 | - | - | |
| | Specific conductivity | uS/cm | all sites | - | - | - | 79.70 | 92.35 | 234.00 | - | - | |
| _ | Total dissolved solids, Filtered | mg/L | all sites | - | - | - | 22.00 | 57.00 | 268.00 | - | - | |
| _ | Total suspended solids, Non-Filterable (Particle) | mg/L | all sites | - | - | - | 1.11 | 20.00 | 212.85 | - | - | |
| | Turbidity, Unfiltered | NTU | all sites | - | - | - | 6.08 | 25.95 | 158.00 | - | - | |
| _ | pH, lab | pH units | all sites | - | - | - | 7.58 | 7.72 | 8.11 | - | - | |
| ield | | | | | | | | | | | | |
| _ | Conductivity | uS/cm | all sites | 73.19 | 170.52 | 248.91 | 45.57 | 136.13 | 226.60 | - | - | |
| _ | Depth, Secchi disk depth | cm | all sites | 1.50 | 10.12 | 55.50 | 10.03 | 21.59 | 81.10 | - | - | |
| | Dissolved oxygen (DO) | mg/L | all sites | 6.24 | 9.04 | 12.67 | 7.96 | 9.80 | 13.92 | - | - | |
| | Dissolved oxygen saturation | % | all sites | 62.93 | 94.62 | 113.90 | 84.33 | 95.27 | 117.30 | - | - | |
| _ | Oxidation reduction potential (ORP) | mV | all sites | -286.94 | 135.50 | 319.68 | -447.32 | 108.72 | 286.20 | - | - | |
| _ | Salinity | ppt | all sites | 0.04 | 0.09 | 0.17 | 0.03 | 0.10 | 0.14 | - | - | |
| | Temperature, water | $\deg C$ | all sites | 7.79 | 17.55 | 22.28 | 1.17 | 14.00 | 21.50 | - | - | |
| _ | Turbidity | NTU | all sites | 9.70 | 48.80 | 198.70 | 7.54 | 24.70 | 80.70 | - | - | |
| | рН | pH units | all sites | 7.75 | 8.22 | 9.39 | 7.67 | 8.13 | 8.55 | - | - | |
| lene | ral Organics Silica gel treated n-hexane extractable material | m mg/L | all sites | - | - | - | < | < | < | - | - | |
| Iajo | r Ions Calcium, Unfiltered | mg/L | all sites | _ | - | - | _ | _ | _ | _ | _ | |

Table 2.8: Current Conditions, Lake Athabasca water. (continued)

| | | | | High Flow | | | Open Water | | | Under Ice | | |
|---|-----------------|-----------|-----|-----------|------|--------|------------|---------|-----|-----------|------|--|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95th | |
| Chloride, Unfiltered | mg/L | all sites | - | - | - | 3.30 | 3.70 | 4.70 | - | - | - | |
| Fluoride, Unfiltered | mg/L | all sites | - | - | - | < | < | < | - | - | | |
| Magnesium, Unfiltered | mg/L | all sites | - | - | - | - | - | - | - | - | | |
| Potassium, Unfiltered | mg/L | all sites | - | - | - | - | - | - | - | - | | |
| Sodium, Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | - | - | - | - | - | | |
| Sulfate, Unfiltered as SO4 | $\mathrm{mg/L}$ | all sites | - | - | - | 3.00 | 6.00 | 20.00 | - | - | | |
| Ammonia and ammonium, Unfiltered as N | $\mathrm{mg/L}$ | all sites | - | - | - | < | < | < | - | - | | |
| Inorganic nitrogen (nitrate and nitrite), Unfiltered as N | mg/L | all sites | - | - | - | 0.02 | 0.10 | 0.22 | - | - | | |
| Nitrate, Unfiltered as N | $\mathrm{mg/L}$ | all sites | - | - | - | 0.01 | 0.10 | 0.22 | - | - | | |
| Nitrite, Unfiltered as N | mg/L | all sites | - | - | - | 0.00 | 0.00 | 0.04 | - | - | | |
| Orthophosphate, Unfiltered as P | $\mathrm{mg/L}$ | all sites | - | - | - | 0.00 | 0.00 | 0.00 | - | - | | |
| Total Nitrogen, mixed forms, Filtered as N | mg/L | all sites | - | - | - | 0.17 | 0.20 | 0.47 | - | - | | |
| Total Nitrogen, mixed forms, Unfiltered as N | mg/L | all sites | - | - | - | 0.20 | 0.25 | 0.65 | - | - | | |
| Total Phosphorus, mixed forms, Filtered as P | $\mathrm{mg/L}$ | all sites | - | - | - | 0.00 | 0.00 | 0.01 | - | - | | |
| Total Phosphorus, mixed forms, Unfiltered as P | mg/L | all sites | - | - | - | 0.01 | 0.04 | 0.27 | - | - | | |
| al Metals Aluminum, Unfiltered | ug/L | all sites | - | - | - | 137.00 | 591.00 | 3100.00 | - | _ | | |
| Antimony, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | | |
| Arsenic, Unfiltered | ug/L | all sites | - | - | - | 0.30 | 0.70 | 2.40 | - | - | | |
| Barium, Unfiltered | ug/L | all sites | - | - | - | 19.10 | 29.90 | 92.60 | - | - | | |

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Table 2.8: Current Conditions, Lake Athabasca water. (continued)

| | | | High Flow | | | | Under Ice | | | | |
|-------------------------------|-----------------|-----------|-----------|------|------|--------|------------------|---------|-----|------------------|-----|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | $50 \mathrm{th}$ | 95th | 5th | $50 \mathrm{th}$ | 95t |
| Beryllium, Unfiltered | ug/L | all sites | - | - | - | 0.01 | 0.03 | 0.14 | - | - | |
| Bismuth, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Boron, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Cadmium, Unfiltered | ug/L | all sites | - | - | _ | < | < | < | - | - | |
| Cesium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Chromium, Filtered | ug/L | all sites | - | - | - | < | < | < | - | - | |
| Chromium, Unfiltered | ug/L | all sites | - | - | - | 0.30 | 0.90 | 4.90 | - | - | |
| Chromium(VI), Unfiltered | $\mathrm{mg/L}$ | all sites | - | - | - | < | < | < | - | - | |
| Cobalt, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Copper, Unfiltered | ug/L | all sites | - | - | - | 0.90 | 1.45 | 7.20 | - | - | |
| Iron, Unfiltered | ug/L | all sites | - | - | - | 236.00 | 953.00 | 6700.00 | - | - | |
| Lead, Unfiltered | ug/L | all sites | - | - | - | 0.10 | 0.55 | 3.60 | - | - | |
| Lithium, Unfiltered | ug/L | all sites | - | - | - | 3.00 | 3.85 | 8.00 | - | - | |
| Manganese, Unfiltered | ug/L | all sites | - | - | - | 6.70 | 21.10 | 162.00 | - | - | |
| Mercury, Unfiltered | ng/L | all sites | - | - | - | - | - | - | - | - | |
| Methylmercury(1+), Unfiltered | $\mathrm{ng/L}$ | all sites | - | - | - | - | - | - | - | - | |
| Molybdenum, Unfiltered | ug/L | all sites | - | - | - | 0.10 | 0.30 | 0.70 | - | - | |
| Nickel, Unfiltered | ug/L | all sites | - | - | - | 0.60 | 1.50 | 8.70 | - | - | |
| Rubidium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Selenium, Unfiltered | ug/L | all sites | - | - | - | < | < | < | - | - | |
| Silver, Unfiltered | ug/L | all sites | - | - | - | < | < | < | - | - | |
| Strontium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Thallium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Tin, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | |
| Titanium, Unfiltered | ug/L | all sites | - | - | _ | - | - | - | - | - | |

Table 2.8: Current Conditions, Lake Athabasca water. (continued)

| | | | High Flow | | | Open Water | | | Under Ice | | |
|----------------------|------|-----------|-----------|------|------|------------|------|-------|-----------|------|------|
| Parameter | Unit | Site | 5th | 50th | 95th | 5th | 50th | 95th | 5th | 50th | 95th |
| Uranium, Unfiltered | ug/L | all sites | - | - | - | - | - | - | - | - | - |
| Vanadium, Unfiltered | ug/L | all sites | - | - | - | 0.50 | 1.90 | 9.20 | - | - | - |
| Zinc, Unfiltered | ug/L | all sites | - | - | - | 1.02 | 4.05 | 20.70 | - | - | |

Note:

- data insufficient
- < too highly censored;
- + grouped differently (merged sites vs individual site);

2.9 Discussion

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2.9.1 Water and Sediment Quality

In the lower Athabasca River, the Athabasca River Delta and Lake Athabasca, median concen-1299 trations of nitrogen species, including ammonia and nitrate, are generally below guidelines for 1300 the protection of aquatic life. Median total phosphorus measures are mostly below the level at 1301 which eutrophication becomes a concern, however, high flow median and other peak values (i.e., 1302 95th percentile) are above that level, up to 0.59 mg/L in the lower Athabasca River. However, 1303 similarly high peak concentrations of total phosphorus in the Athabasca River Delta do not 1304 1305 correspond to high concentrations of chlorophyll a, which is an indicator of algal biomass in the water column. Instead, median and peak chlorophyll a measures in the Athabasca River Delta 1306 during the high flow and open water seasons indicate mesotrophic conditions. No measures of 1307 1308 benthic or epiphytic chlorophyll were available for any of the locations in this study. 1309 Field and laboratory measures of pH indicate that the River, Delta and Lake water is neutral to moderately basic, with moderate to high hardness levels, moderate conductivity measures 1310 including significant contributions from sodium, calcium and sulfate ions. An exception to 1311 this is in the Delta and Lake during the under ice season, where some 5th percentile values 1312 were slightly acidic. Dissolved oxygen concentrations are above the required concentration to 1313 support aquatic life, although it can be relatively low during the high flow season in Lake 1314 Athabasca, presumably in early winter after the ice cover has been in place for many months. 1315 In general, Lake Athabasca water is slightly less alkaline with lower concentrations of chloride 1316 and sulfate compared to River and Delta water. 1317 1318 Certain median metals and trace element concentrations in water are above provincial guidelines for the protection of aquatic life. This includes total cobalt, total and dissolved cop-1319 per, total lead, total manganese, total selenium, total thallium and total zinc in the Athabasca 1320 River and Delta, especially in the high flow seasons but also in others. Total mercury ex-1321 ceeds these guidelines in the River, but insufficient data are available for the Delta. In Lake 1322 Athabasca, where total metals and trace elements data were available for the open water season 1323 only, fewer guideline exceedances were noted. Those exceedances included total copper and 1324 1325 lead (peak values only). For many trace elements and metals, data for Lake Athabasca were insufficient to calculate summary statistics. 1326 The pattern of trace element exceedances in water in the Athabasca River and Delta occur-1327 ring especially in the high flow season, indicates that these constituents are likely associated 1328 with suspended particles that are transported in the water column predominantly during high 1329

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flows. The majority of total trace elements measured in the Athabasca River follow this pattern, including total lead, total mercury, total nickel, total selenium, total uranium, and total vanadium. Measures of total suspended solids in these locations are highest in the high flow season, lower in the open water season, and lowest in the under ice season, coinciding with these exceedances and supporting the importance of the association of particles and certain trace elements. In addition, in the Athabasca River, there are examples of non-particle as-1335 sociated, or dissolved, trace element concentrations that peak during the high flow season, 1336 including dissolved aluminum, dissolved chromium, dissolved copper, dissolved lead, and dissolved nickel. Not all trace element concentrations peak during the high flow season, however, 1338 1339 for example, in the Athabasca River, dissolved barium, dissolved boron, dissolved lithium, dissolved manganese, dissolved strontium, dissolved uranium, total boron and total strontium 1340 concentrations peak in the under ice season. Other trace elements, both dissolved and total, 1342 do not exhibit distinct peaks in any season. In some cases in the Athabasca River, the seasonal pattern of trace element concentrations is site-specific, indicating the importance of local conditions. The seasonal patterns of trace element and other constituent concentrations can help to understand the sources and delivery pathways of these constituents to the Athabasca River, Athabasca River Delta and Lake Athabasca when paired with information about water 1346 and sediment delivery to these systems. For example, the proportion of water inflows made up by groundwater, snow melt, overland runoff generated during storms and from upstream flow 1348 generally changes predictably through the seasons. 1349

Pesticides and organohalides were generally not measured in water above the relevant detection limits in the Athabasca River and the Delta. This was also true for the vast majority of measured PAHs and general organic measures in the River, with the exception of certain hydrocarbon measures, toluene, and certain mainly alkylated PAHs (the latter mainly during high flows). In the Delta, PAHs and general organic constituents were not measured above the relevant detection limits, with the exception of naphthenic acids and the related measure, oil sands extractable organics, which were consistently detected. Pesticides were not measured in Lake Athabasca water, and organohalide data were minimal.

Certain trace elements and metals were detected at elevated levels in sediment in the River and Delta, however most median concentrations did not exceed the provincial guidelines for the protection of aquatic life, with the exception of nickel in the Delta. For those PAHs with provincial sediment quality guidelines for the protection of aquatic life, no exceedances in the current conditions were noted. It is important to keep in mind however, that most of the measured metals, trace elements and PAHs do not have applicable sediment quality guidelines. For example, in the Athabasca River Delta, 20 non-alkylated PAHs, 27 alkylated PAHs, 27 alkylated PAHs, 27 alkylated PAHs alkylated PAH groups and dibenzothiphene were measured in sediments, however Alberta sediment quality guidelines for the protection of aquatic life apply to only 11 non-alkylated PAHs (GoA, 2018).

2.9.2 The Effect of Location

It should be kept in mind that in many cases, different detection limits were in effect for water quality measures from the Athabasca River, the Delta and the Lake. The lack of detection in one system does not necessarily mean that it is a lower concentration than in the other system, where it may have been detected. In addition, no statistical tests were conducted to test for differences between these locations, but it should also be remembered that not all available data for each location were used to create current conditions due to incompatible sampling and analytical methods.

Notwithstanding the above, some trace elements appear to have higher median concentrations in water in the Athabasca River compared to the Athabasca River Delta (e.g., dissolved aluminum, dissolved iron), while for others the reverse is true (e.g., dissolved chromium, dissolved copper, dissolved thallium, dissolved titanium). For other trace elements, there is no consistent difference apparent between these locations. Other than these general observations, little in the way of differences between the Athabasca River, Delta and Lake water quality were noted. There are insufficient data currently available for Lake Athabasca to establish high flow and under ice current conditions for most measured parameters. For the open water season, median concentrations for most trace elements in Lake Athabasca were similar to those in the River and Delta, with some exceptions such as somewhat higher chromium, copper and zinc compared to the River and lower aluminum, molybdenum and zinc compared to the Delta.

In terms of sediment quality, the River and Delta locations are distinguished by particle size, with a relatively greater proportion of silt and clay in the Delta and a greater proportion of sand in the River. Most measured trace element concentrations in the Delta are also higher than in the River sediment, including aluminum, boron, chromium, cobalt, copper, iron, lead, lithium, manganese, nickel, strontium, thallium, vanadium and zinc, while the reverse was true for titanium. Many PAHs were also present in higher concentrations in the Delta sediment compared to the River, especially for alkylated PAHs that were consistently measured in both locations. The smaller sediment particle size in the Delta compared to the River are likely related to this increased concentrations of trace elements and PAHs in the Delta, since PAHs are preferentially associated with smaller sediment particles (CCME, 1999), although other

influences may also be present.

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2.9.3 The Effect of Season

Generally, major ions concentrations and related measures such as alkalinity and specific con-1399 ductivity are highest in the River and Delta in the under ice season. This is a common 1400 phenomenon, given the lower water flows and lower dilution potential. There may also be an 1401 increased proportion of high-solute groundwater inflows during the winter, when surface water 1402 inputs are lowest. 1403 Ammonia and nitrogen are also highest in the under ice season, with most total nutrient 1404 measures highest in the high flow season. The latter is quite common where total nitrogen 1405 and phosphorus are associated with particles in the water, which are generally at their highest 1406 concentration during high flow. 1407 1408 Surprisingly, in both the River and Delta, field measured dissolved oxygen concentrations are highest during the ice covered season. This is counter-intuitive, given that ice covers 1409 generally reduce the potential for oxygen to be entrained in the water column and that algae 1410 1411 are not usually as photosynthetically active during winter months. However, colder water can accommodate more dissolved oxygen and the ice covered season as defined in this report may 1412 very well include ice free periods, both of which can contribute to higher dissolved oxygen 1413 1414 concentrations. Dissolved oxygen data for the under ice season were not available for Lake Athabasca. 1415 Dissolved and total metals and trace element concentrations are variable across seasons. 1416 Notably, in the Athabasca River, concentrations values for these parameters are most often 1417 significantly different across sampling sites during the high flow season and especially the under 1418 ice season. In the Delta, site-specific percentile values were calculated for the under ice season. 1419 This suggests that local differences or influences are most consequential during the under ice 1420 1421 season, at least in terms of metals and trace elements concentrations. Otherwise, most total measures (more associated with particles) are at their highest concentrations during high flow, 1422

2.10 Application

The current conditions calculated in this study serve as a "baseline" range for water and sediment quality in the Athabasca River, the Athabasca River Delta and Lake Athabasca.

Sediment data were not collected seasonally and are not included in this discussion.

while dissolved measures were more variable across seasons.

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They characterize water and sediment quality for the specific sampling sites or the reaches 1428 across which the sampling sites span, using data collected by the selected monitoring programs 1429 between 2011 and 2020, as available. This study has not identified change in or impacts 1430 to water or sediment quality in these locations, nor has it inferred sources of the measured 1431 constituents. The intended application of these current conditions is to serve as "no change" 1432 criteria in the absence of risk-based guideline values formulated in other sections of this report. 1433 The current conditions can serve as a benchmark against which past or future conditions can 1434 be compared, with relevance to impact prediction and assessment projects, water and sediment 1435 quality monitoring, or risk assessment, for example. 1436

1437 2.11 Limitations

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2.11.1 Potential to Rehabilitate Long-term Datasets

As has already been discussed, this study was limited by the incompatibility of sampling and analytical methods used to collect water and sediment quality data by different programs and even within programs at different times over the period of record. The setting of current conditions according to the methods used in this study would benefit from additional data points, many of which could be included in such an analysis if the differences introduced by variations in methods could be reconciled.

In addition to this additional potential improvement, further monitoring in Lake Athabasca would greatly contribute to establishing additional current conditions for water and sediment quality in that location, especially during the high flow and under ice seasons.

$_{ imes}$ Chapter 3

1449 Health Risk Criteria for the

Protection of Surface Water to

Support Indigenous Use

- 1452 MANDY L. OLSGARD MSC, P. BIOL. AND CHANEL YEUNG MSC, BIT
- 1453 Integrated Toxicology Solutions

1454 3.1 Introduction

- 1455 Community members from ACFN, MCFN, and FMFN have observed changes in the health
- and condition of surface water, aquatic biota, wildlife (birds and mammals) and community
- 1457 members since development of the oil sands began in the 1960s (Personal communications;
- 1458 Pinto, A. et., al., 2019; Droitsch, D. and Simieritsch, T., 2010)
- 1459 Health concerns expressed by community members include changes in the behavior and
- 1460 health of fish (i.e., soft/mushy muscle, increased parasites and tumors, increased and malfor-
- 1461 mations of gills and body parts), fewer and small and unhealthy furbearers, absence of inver-
- 1462 tebrate species used by fish and birds as food sources, decreased potency of medicinal plants
- 1463 and increased prevalence of human health morbidities such as cancer and skin disorders.
- 1464 ACFN, FMFN, and MCFN community members are concerned that the changes in health
- 1465 condition of humans, wildlife and aquatic biota are linked to the release of contaminants by oil
- 1466 sands mining operations (Personal communications; McLachlan (2014); Droitsch & Simieritsch
- 1467 (2010)).
- The health concerns described above have been observed and recorded by Indigenous com-

munity members during their time on the land while participating in activities, such as; trap-1469 ping fur bearing semi-aquatic mammals (i.e., beaver, mink, otter, muskrat), drinking from 1470 lakes, rivers and muskeg, fishing and hunting for food (i.e., walleye, pickerel, whitefish, moose, 1471 ducks) and harvesting medicines to treat various conditions (i.e., rat root). Through this 1472 connection with the land, members of ACFN, FMFN, and MCFN are guided by their knowl-1473 edge that the health of the "land" is directly related to their ability to sustain their way of life and their overall sense of wellbeing (Personal communications; Baker & Westman (2018); 1475 Cunningham & Stanley (2003)). 1476 In Alberta, risks to aquatic environments from exposure to chemical substances are assessed 1477 by comparing ambient monitoring data to environmental quality guidelines derived for the 1478 protection of aquatic life (GoA (2018); CCME (2021)). Surface water quality guidelines are 1479 also available to assess potential risks to livestock (GoA, 2018) and human health from the 1480 1481 consumption of drinking water (Health Canada, 2021). However, the latter guidelines are

Indigenous community members to chemical substances during Indigenous land use activities. Previous research by Olsgard & Thompson (2020) identified several surface water quality guidelines (GoA, 2018) which do not consider bioaccumulation and persistence of chemical substances which could limit the protection of higher trophic level species. Specifically beaver, northern pintail ducks, lesser scaup, muskrat, river otter and bald eagles could be at risk from biomagnification of methyl mercury, selenium, and thallium in aquatic food webs.

rarely applied to surface water in Alberta (GoA, 2018) resulting in a disconnect between the

provincial process for assessing risks posed by the quality of surface waters and the exposure of

Due to limitations in the comprehensiveness of the existing surface water quality guidelines in Alberta and Canada, a need to develop water quality criteria that protect the ways in which Indigenous people interact with and rely on surface water was identified.

The following describes the development of health risk criteria to assess potential risks to Indigenous community members and the environment on which they rely for exercising Aboriginal Rights. The health risk criteria can also be applied as limits of change which reflect Aboriginal Rights and health risk concerns related to the condition of the Athabasca River, 1496 Athabasca River Delta, and Lake Athabasca.

3.2Objective

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To address gaps in surface water quality guidelines which may limit the protection of Indigenous community members, aquatic receptors and wildlife by identifying and/or deriving health risk 1500

criteria which explicitly consider Indigenous use of water for constituents of concern that may
be naturally occurring, related to releases from non-oilsands industrial sectors, and present in
oil sands mine water (OSMW) which may seep or be actively released to surface water bodies
historically and currently used by ACFN, FMFN, and MCFN members while exercising their
Aboriginal Rights.

3.3 Methods

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- The following stages, described in detail below, were used to identify and/ or modify existing surface water quality guidelines and derive health risk criteria that consider protection of the aquatic environment to support Indigenous land use.
- Develop a Indigenous water use conceptual model and identify protection goals,
- Identify constituents of potential concern (COPCs),
- Identify available surface water guidelines by protection endpoint,
- Adopt available guidelines as Indigenous water use protection criteria in those cases where protection goals are met, and
- Derive criteria, when Indigenous water use protection was not considered.

1516 3.3.1 Indigenous Water Use Conceptual Model

Indigenous water use protection goals for health risks were identified by developing a conceptual model based on Indigenous knowledge shared by community members and staff from ACFN, FMFN and MCFN. The conceptual model identifies indicators (i.e., culturally important ecosystem components), exposure pathways for human and ecological indicators, and the protection criteria and endpoints for each Indigenous water use protection goal.

1522 3.3.2 Identification of Chemical Substances

- 1523 Chapter 2 provides a detailed description of monitoring data collected in ambient surface water 1524 in the Lower Athabasca Region. Surface water quality guidelines are not available for each of 1525 these parameters, nor are they required. Rather, the approach herein is to identify indicators 1526 of change and effect related to oil sands development pressures and compare concentrations of 1527 those indicator parameters to guidelines appropriate for Indigenous water use.
- For the purposes of this study OSMW refers to any water produced and/ or accumulated by oil sands mining activities, including oil sands process water (OSPW), expressed water from

- tailings impoundments, collected surface water runoff, industrial wastewater, sewage water, etc.
- Classes and species of chemical substances, which have been characterized in air emissions, tailings and OSMW were identified as indicator parameters and used to focus the development of health risk criteria. The following information sources were consulted:
- Peer reviewed literature,
- Ambient monitoring data, and
- Industry regulatory reporting.
- Additionally, measured parameters, which may not be identified in oil sands specific data sets, identified in the monitoring networks described in Chapter 2 were also considered. These parameters provide an indication of other sources of contaminants (i.e., naturally occurring; agriculture and municipal sectors) in the Athabasca River watershed which may cumulatively contribute to potential risks to human and environmental health.

3.3.3 Inventory of Surface Water Quality Guidelines

- Available surface water quality guidelines were identified through a jurisdictional scan of the
- 1545 regulatory agencies described below. Previous work completed by Olsgard & Thompson (2020)
- 1546 was also considered during this exercise.

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- 1547 Identified guidelines (and supporting technical documents) were reviewed and an inventory
- 1548 of existing surface water quality guidelines used by regulatory agencies was developed.

1549 Environmental Quality Guidelines for Alberta Surface Waters

- 1550 These guidelines are for application to surface water quality (to protect aquatic life (PAL),
- 1551 agricultural, and recreational uses), sediment quality, and tissue residue (to protect wildlife
- consumers and fish from direct toxicity) (GoA, 2018). The surface water quality guidelines do
- 1553 not apply to drinking water and the user is directed to Health Canada guidelines. The majority
- 1554 of guidelines have been adopted or modified from CCME, US EPA and British Columbia
- 1555 Canadian Environmental Quality Guideline for Water (CEQGs; CCME (2021)).

1556 Canadian Environmental Quality Guidelines (CEQG)

- The CEQGs provide science-based goals for water quality through published fact sheets and scientific criteria documents which describe the development of guidelines for the majority of
- 1559 substances with available surface water quality guidelines (to protect aquatic life, agricultural,

and recreational uses), sediment quality, and tissue residue (to protect wildlife consumers and fish from direct toxicity. Guidelines are developed using CCME (2007) protocol which updates to the previous development in 1987, which closely aligned with development of the National Water Quality Standards by the US EPA and adopted widely throughout Canada.

4 Federal Environmental Quality Guidelines (FEQG)

The FEQGs were developed to support federal initiatives and provide thresholds below which direct adverse effects from the chemical on aquatic life exposed via water or sediment, or bioaccumulative effects in wildlife (birds and mammals) that consume aquatic life should be unlikely.

The federal government identifies that FEQGs are not effluent limits nor are they "never to be exceeded" values. Seventeen FEQCs and scientific criteria documents have been developed to meet requirements of the federal environment Minister under Section 54 of CEPA, which goes beyond factors which were considered in development of the CCME CEQGs (Canadian Environmental Protection Act (CEPA), 1999).

Guidelines for Canadian Drinking Water Quality (CDWQG)

The CDWQGs were established by Health Canada (2020a) in collaboration with the Fed-1574 eral Provincial-Territorial Committee on Drinking Water based on current, published scientific 1575 1576 research related to health effects (defined as Maximum Acceptable Concentrations (MACs), aesthetic effects (i.e., taste, odour, colour), and operational (i.e., treatment) considerations). 1577 The CDWQGs are developed for substances which could result in toxicological effects in ex-1578 posed humans, have the potential to be present in drinking water supplies and have available 1579 1580 methods of quantification (i.e., lab analysis). Scientific criteria documents have been published for each substance with a Maximum Acceptable Concentration (MAC). 1581

National Drinking Water Regulations (DWR)

The US EPA DWRs (US EPA, 2021a) are legal limits for more than 90 chemical and microbial contaminants in United States drinking water. The legal limit for each substance reflects both human health protection and concentrations that are achievable using the best available technology.

National Recommended Water Quality Criteria (WQCs)

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The US EPA provides three Criteria under the National Recommended Water Quality Program 1588 (WQCs); aquatic life, human health, and organoleptic (i.e., aesthetic) (US EPA, 2021b). 1589 The Aquatic Life Criteria published in the National Recommended WQCs vary from those 1590 prescribed in Canada and Alberta as the data from freshwater species toxicity tests reported 1591 as total recoverable fractions have been converted to a dissolved fraction using Conversion 1592 Factors (CFs) (US EPA, 1993, 1996). The US EPA determined that dissolved guidelines are 1593 more appropriate as they represent the fraction of metals which is bioavailable to aquatic biota 1594 (as adsorption at gill surfaces required dissolved forms of metals) compared to particulate forms 1595 of metals which cannot be taken up as easily within biological organisms (US EPA, 1993). 1596 1597 The US EPA (1993) referenced studies which report that the toxicity of particulate metals 1598 is less compared to dissolved metals. To derive dissolved metal criteria the US EPA calculated CFs from toxicity tests in which both the total recoverable and dissolved fractions of the 1599 metal of interest was measured. The US EPA (1993) also states that the CF derived dissolved 1600 guidelines should be applied to conditions where pH ranges from 6.5-9 and total organic carbon 1601 and total suspended solids are less than 5 mg/L. Table 3.1 indicates that the median values for 1602 1603 open water season in the Lower Athabasca River are within the prescribed range for pH (8.2) but well above for total suspended solids (24 mg/L) and total organic carbon (8.9 mg/L). 1604 Aquatic Life (AL WQCs) describe criteria which are the highest contaminant specific con-1605 1606 centrations that are not expected to pose a significant risk to most aquatic species. The AL WQCs are reported in total concentrations. Conversion factors are available for estimating 1607 total metals when dissolved metals were measured. 1608 Human Health Ambient Water Quality Criteria (HH AWQCs) developed under United 1609 States legislation (Section 304(a) of the Clean Water Act) represent substance specific concen-1610 1611 trations that are not expected to cause adverse effects to human health from the consumption of drinking water alone or in combination with consuming organisms (i.e., fish). The HH 1612 AWQCs consider both carcinogenic and non-carcinogenic effects from exposure of humans to 1613 chemical substances in untreated surface water and wild organisms. Notably, the HH WQCs 1614 are recommended for consideration by "authorized tribes", comparable to First Nations in 1615 Canada when adopting criteria into their water quality standards. Methodology for deriving 1616 the HH AWQCs is also available (US EPA, 2000b). 1617 Organoleptic Effect (OE WQCs), similar to Health Canada Aesthetic Objectives (Health 1618 1619 Canada, 2020a), protect water against tainting and fouling from offensive odours, colour, and

taste (World Health Organization (WHO), 2017).

Guidelines for Drinking Water Quality (GDWQs; WHO, 2017 4th Ed)

The GDWQs for chemical, microbial, radiological and acceptability (i.e., aesthetics) aspects are based on over 50 years of WHO guidance on identifying safe drinking water quality and recognized internationally as formative regulations and standards for water safety in support of public health. In addition to health-based guidelines, the WHO provides guidance on developing a conceptual framework for implementation, water safety plans, and monitoring (World Health Organization (WHO), 2017).

Toxicological Benchmarks for Wildlife (US Department of Energy, 1996)

1630 The Oak Ridge National Laboratory (ORNL) reported No Observable Adverse Effect Levels (NOAELs) for 9 representative mammalian wildlife species or 11 avian wildlife which were 1631 then used to derive species-based toxicological benchmarks that represent concentrations of 1632 chemicals in environmental media (water, sediment, soil, food, etc.) that are presumed to 1633 nonhazardous for the listed wildlife species. The piscivore benchmarks reported as surface 1634 water quality concentrations (mg/L) can be used to assess the potential risks to mammals (i.e., 1635 mink and otter) and birds (i.e., kingfisher, mallard, great blue heron, osprey) from ingesting 1636 chemicals in surface water and fish (Sample et al., 1996). 1637

The combined food and water benchmarks for wildlife species primarily consuming aquatic organisms (piscivores) as reported in Sample et. al., (1996) were calculated using the following equation:

1641 Equation (3.1)
$$C_w = \frac{NOAEL_w \times bw_w}{W + (F \times BAF)} \tag{3.1}$$

Where:

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 C_m = Concentration of the contaminant in the drinking water of an animal (mg/L)

 $NOAEL_w$ = No Observable adverse Effects Level in wildlife species (mg/kg bw/d)

 bw_w = body weight of wildlife species

W = Water ingestion rate (L/d)

F = Food ingestion rate (kg/d)

BAF = ratio of concentration of a contaminant in tissue (mg/kg) over water (mg/L)

Adopting Existing Guidelines as Water Quality Criteria for the 1642 Protection of Indigenous Use 1643

1644 To determine whether available guidelines consider Indigenous water use protection goals, the inventory of guidelines for COPCs was compared to the protection goals for each Indigenous 1645 1646 water use category described in the Indigenous water use conceptual model as described in Section 3.4.1. 1647

If a currently available surface water quality guideline considered protection of Indigenous water use goals (indicators, exposure pathways and endpoints), the regulatory guideline was adopted as the health risk criteria for Indigenous use protection for that substance.

If the review exercise indicated that there were no available guidelines for a COPC or 1651 1652 that currently available surface water quality guidelines did not consider Indigenous water use protection goals it was not adopted, and health risk criteria were developed using the methods 1653 discussed below. 1654

Deriving Water Quality Criteria for the Protection of Indige-3.3.5 1655 nous Use 1656

Health risk criteria for the protection of humans consuming surface water and traditional foods 1657 were derived using guidance from the US EPA (2000b) "Methodology for Deriving Ambient 1658 Water Quality Criteria for the Protection of Human Health". 1659

Health risk criteria for Indigenous use protection were derived through modifications of the US EPA (2000b) Equation (3.2) to account for consumption of locally caught fish and river/lake/muskeg water as drinking water and the ingestion of medicinal plants Equation (3.2).1663

The US EPA (2015c) values for body weight (80 kg) and drinking water intake (2.4 L) were 1664 considered representative of ACFN, FMFN, and MCFN adult community members. 1665

Chemical-specific inputs used to develop the HH AWQC were adopted when available/pub-1666 1667 lished (US EPA, 2015b). When not available, values were sourced from resources specified in US EPA (2000b). 1668

References doses for non-cancer effects (RfD, mg/kg-d) and Risk-specific doses for carcino-1669 gens (RsD, mg/kg-d) were adopted from the current US EPA Integrated Risk Information 1670 System (US EPA IRIS). 1671

Bioaccumulation factors (BAFs), bioconcentration factors (BCFs), food chain multipliers 1672 (FCM), and lipid fractions for organic substances were adopted from US EPA (2015b) and 1673

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- $1674 \quad \text{inorganic substances were adopted from several US EPA ecological risk assessment documents};$
- 1675 BAFs (Sample et al., 1996), BCFs and FCMs (US EPA, 1999).
- As per Alberta Health (2019) the dose associated with an incremental lifetime cancer risk
- 1677 (ILCR) of 1 in 100,000 (1 x 10-5) is considered to be "essentially negligible" and was adopted
- 1678 rather than the acceptable risk level for cancer (1 x 10-6) used by the US EPA (2000b; 2015a).

Equation (3.2): Consumption of traditional foods and drinking water to derive health risk criteria (modified from US EPA US EPA (2000b)).

$$HRC\ TF + DW(ug/L) = \frac{toxicity\ value(\frac{mg}{kg} - d)xRSC \times BW(kg)x1,000(\mu\frac{g}{mg})}{DI(\frac{L}{d}) + \sum_{i=2}^{4}(FCRi(kg/d) \times BAFi(L/kg))} \eqno(3.2)$$

Where:

HRCTF + DW = health risk criteria for traditional foods and drinking water consumption toxicity value = RfD x RSC (mg/kg-d) for noncarcinogenic effects or 10-5/CSF (kg-d/mg) for carcinogenic effects

RSC = relative source contribution (applicable to only noncarcinogenic) (0.2, unless otherwise stated)

BW = body weight (80 kg)

DI = drinking water intake (2.4 L/d) = summation of values for a quatic trophic levels (TLs), where the letter i stands for the TLs to be considered, starting with TL2 and proceeding to TL4

FCR = Fish Consumption Rate (0.388 kg/d)

BAFi = bioaccumulation factor for aquatic TLs 2, 3, and 4

Equation (3.3): Equation to derive water quality criteria for human health protection from consumption of medicinal plants (modified from US EPA (2000b)).

$$HRC \ medicinal \ plants(ug/L) = \frac{toxicity \ value(\frac{mg}{kg} - d)xRSC \times BW(kg)x1,000(\frac{\mu g}{mg})}{PCRxBCF_{eS-P}} \ \ (3.3)$$

Where:

 $HRC\ medicinal\ plants = \text{health risk criteria for protection of}$ health risks from exposure to contaminants in medicinal plants $toxicity\ value = \text{RfD}\ x\ \text{RSC}\ (\text{mg/kg-d})\ \text{for noncarcinogenic effects or}$ $10\text{-}5/\text{CSF}\ (\text{kg-d/mg})\ \text{for carcinogenic effects}$ RSC = relative source contribution (applicable to only noncarcinogenic effects), $(0.2,\ \text{unless otherwise stated})$ BW = body weight (80 kg) $PCR = \text{medicinal plant consumption rate } (0.007\ \text{kg/d})$

bioconcentration factor sediment to plant

1683 **3.4** Results

1684 3.4.1 Indigenous Water Use Conceptual Model

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Indigenous water uses and exposure pathways for community members (human receptors) were 1685 identified through personal communications with community members and staff from ACFN, 1686 FMFN and MCFN. 1687 The community identified Indigenous water uses, cultural practices and species of impor-1688 1689 tance were integrated into a conceptual model with western science measures (quality focused criteria and endpoints) to define Indigenous water uses and protection goals. Each use and 1690 protection goal are discussed below to provide context for why each Indigenous water use must 1691 1692 be considered in developing surface water quality criteria to achieve protection goals. A visual depiction of the detailed conceptual model is provided in Figure 3.1 and each of the Indigenous 1693 1694 water uses and protection goals described further below.

Traditional foods

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Community members (human receptors) are exposed to contaminants through ingestion of 1696 culturally important wildlife and fish species. Fish are directly exposed to and take up con-1697 1698 taminants from the surface water but can also accumulate toxic substances through ingestion of prey items (invertebrates and smaller fish). Therefore, consideration of the trophic level 1699 from which fish are consumed is important in developing surface water quality criteria that 1700 protect humans from consumption of fish. This is a well-recognized exposure pathway and 1701 human health risk regulated for certain substances in Canada (Health Canada, 2020b) and 1702 used to set maximum consumption levels/advisories by GoA (2019a) and the US EPA (2000a). 1703 An often-overlooked exposure pathway is the uptake of contaminants by wildlife from con-1704 1705 suming surface water. This pathway was identified by community members as a potential cause 1706 of decreased health being observed in herbivorous mammals and waterfowl species (moose, mallard, scaup) relied on for traditional diets (as discussed under the wildlife health water use) 1707 but is also an exposure pathway for community members ingesting wildlife tissues. 1708 Exposure of human receptors to contaminants through ingestion of wildlife species (as 1709 traditional foods) is considered in human health risk assessment methods (Alberta Health 1710 1711 (2019); Health Canada (2021); Health Canada (2019); Health Canada (2018)) but not mirrored in surface water quality guidelines applied in Alberta. To ensure protection of community members (human receptors) from exposure to contam-1713 1714 inants in wildlife and fish water quality, guidelines must consider biomagnification of contaminants in food webs and carcinogenicity, which is a human health endpoint not considered in 1715 the derivation of environmental quality guidelines, such as those developed by the US EPA US 1716 1717 EPA (2015c). Surface water quality guidelines against which monitoring data can be compared when 1718 1719 collected under risk-based surveillance programs must consider Indigenous community health

Natural waterbodies as drinking water sources

Regardless of Health Canada and Alberta Health guidance on sources of drinking water, members of ACFN, FMFN and MCFN have traditionally and continue to consume untreated drinking water from surface water bodies in the Lower Athabasca Region (i.e., lakes, rivers, muskeg). As such, ambient water quality guidelines such as the (US EPA, 2015c) which consider ingestion of raw surface water must be applied to understand impacts to Indigenous water use.

exposure pathways and endpoints to understand impacts to Indigenous water use and protec-

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tion goals.

Traditional medicines

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- Through traditional knowledge guided practices Indigenous communities rely on the medicinal properties of several aquatic plant species for treating health maladies (i.e., cardiovascular health, kidney infections, respiratory problems). Aquatic plants such as wild mint and rat root may absorb and translocate chemical substances from surface water and sediments resulting in potential exposure of community members relying on these species for preparations of medicinal teas, powders, and poultices (Clemens (2006)).
- 1735 Community members have also noted that the potency of medicinal plants is decreasing
 1736 as is availability. Both of these concerns are thought to be linked to chemical emissions from
 1737 industrial development and the changes to the land (personal communications).
- The accumulation of contaminants from surface water and sediment in medical plants and exposure of community members must be considered in developing surface water quality criteria however, no guidelines which considered bioaccumulation in plant species were identified through publications from US EPA (1999; 2000b). This pathway is rarely assessed in human health risk assessments and may require further investigation.

Aquatic ecosystem health

- Members of ACFN, FMFN and MCFN have shared that their health is experiential and re-
- 1745 lational from an Indigenous world view and directly related to their sense of personal health
- 1746 and wellbeing. As such, water cannot be managed as a single component broken off from
- 1747 the environment or communities. Water is the giver of life and must be protected using tra-
- 1748 ditional knowledge and now due to industrial development, western science methods. But
- 1749 western science water management was unnecessary prior to industrial development in the
- 1750 Lower Athabasca Region (personal communications).
- While several of the identified guidelines (GoA (2018); CCME (2021); US EPA (2021b))
- 1752 consider protection of aquatic life through four main receptor groups (fish, amphibians, inver-
- 1753 tebrates, plants/algae) it is really the integration of these components that establishes and
- maintains a functional and healthy ecosystem from an indigenous perspective (Greenwood &
- 1755 Leeuw (2007); Arsenault et al. (2018)).

Wildlife health

- Wildlife health, like water health described above, is a community health indicator upon which
- 1758 members of ACFN, FMFN and MCFN view their personal sense of wellbeing. The quality of

moose and duck meat, abundance, and presence of wildlife species for trapping and hunting 1759 and population dynamics between predators and prey have been noted by community members 1760 as changing and as being of poorer quality overall since industrial development began. 1761 Community members are concerned that wildlife species are being exposed to contaminants 1762 though their drinking water and diet (aquatic plants, invertebrates, algae) and that these 1763 contaminants are directly affecting wildlife health but also human health through ingestion of 1764 traditional foods (personal communications) (Baker & Westman, 2018). 1765 Eccles et al. (2020) validated the community observation that contaminant concentrations 1766 are changing (increasing) in water in the oil sands region, and this could be impacting wildlife 1767 health. 1768 Exposure of wildlife to contaminants is a well described exposure pathway in the oil sands 1769 region (Rodríguez-Estival & Smits, 2016) and the requirement to assess potential risks to 1770 1771 wildlife species from exposure to contaminants is well defined in ecological risk assessment guidance (CCME, 2020) and subsequent exposure in humans consuming wildlife as traditional 1772 foods (Health Canada (2021); Health Canada (2012); Health Canada (2010)). However, water 1773 quality guidelines are limited to the protection of livestock for agricultural purposes again 1774 disconnecting the regulatory practice of risk assessment from the realities of Indigenous water 1776 use. Environmental and human health impacts from persistent and bioaccumulative substances 1777 which can biomagnify in aquatic ecosystems is well described (Arnot & Gobas (2004); Ali et al. (2019)) and exposure pathways linked to the contamination of traditional foods is described 1779 1780 above. 1781

However, wildlife support Indigenous community traditional lifestyles beyond provision of traditional foods. Trapping semi-aquatic furbearing species such as muskrat, beaver and otter are recognized Aboriginal Rights (Collins & Murtha (2009); Passelac-Ross (2005)) and the sale of pelts has long been an economic staple in Athabasca Region First Nation Communities (Baker & Westman, 2018).

Semi-aquatic mammals' diets are sustained by aquatic biota (invertebrates, plants, fish) and members from ACFN, FMFN and MCFN have noted that the health, quality of pelts, and abundance of muskrats has been declining over time. Members have attributed the decline in condition and quality of pelts to poor water quality and the decreasing populations to lower water levels in the PAD (Personal communications).

While not a common factor considered in the development of water quality guidelines, the health of aquatic fur-bearing mammals is directly linked to aquatic ecosystems and water

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793 quality criteria are required to protect this water use.

| Primary Use | | | Se | condary Use | Protection | |
|--------------------------------|--------------------------------------|--------------------------------------|--|--|---|----------------------------------|
| Receptor | Water use | Exposure pathway (human receptor) | Environmental Indicator | Exposure pathway (ecological receptor) | Goal | Endpoints |
| | | | Fish | Direct contact/ uptake Ingestion aquatic biota | Safe food consumption | |
| | Traditional foods and drinking water | Direct exposure - Ingestion | Plants Wildlife | Direct contact/ uptake | Safe natural surface water consumption | Carcinogenic Non-carcinogenic |
| | | | Water | Ingestion aquatic biota Water ingestion | | Aesthetic |
| Indigenous | 2. Traditional medicines | Direct exposure - Ingestion | Plants | Direct contact and uptake | Safe medicine consumption Potency of medicinal plants | |
| community member (Human) | 3. Aquatic ecosystem health | Indirect health determinant | Invertebrates Fish Plants Algae | Direct contact/ uptake Direct contact/ uptake Direct contact/ uptake Direct contact/ uptake | Aquatic community composition unchanged, healthy, and robust biota populations | _ |
| | 4. Wildlife health | Indirect health determinant | Mammals Birds | Water ingestion Fish Ingestion | Healthy wildlife, robust populations, natural behaviours, good quality pelts | Non-carcinogenic Aesthetic |

Figure 3.1: Indigenous Water Use Conceptual Model

3.4.2 Inventory of Contaminants

The inventory of contaminants for which health risk criteria were developed include constituents of concern that may be naturally occurring, related to releases from non-oilsands industrial sectors, and present in oil sands mine water (OSMW).

There are several sources of OSMW associated with mining activities. Tailings waste streams are comprised of sand, silt, clay, processed water, and residual bitumen which is a complex mixture of a multitude of chemicals (Allen, 2008). Mine water that accumulates from muskeg dewatering and collection of surface water runoff from mine sites has a different chemical signature than surface water bodies such as lakes and contains elevated trace elements and polycyclic aromatic hydrocarbons, both dissolved and bound to suspended solids and organic matter, which elicit toxicological responses in exposed receptors (Alexander, A.C. and Chambers, P. 2016; Kelly, E. et., al., 2009). Naturally saline basal groundwater is also accumulated in OSMW inventories during depressurization (Sawatsky et al., 2004) and the toxicity associated with exposing surface water biota to saline groundwater has been documented for decades (Giles & Klaverkamp (1979); Rogers & Lake (1979)).

The contaminants associated with the various sources of OSMW have also been identified as contributing to acute and chronic toxicity in biological organisms (Li et al. (2017); Mahaffey & Dubé (2017); Hughes et al. (2017)).

In addition to mine water, contaminants released from point and area source emissions from oil sands mines contribute deposition of acids (from transformation of gaseous compounds), and PAHs and trace elements (from particulate matter) (Lynam et al. (2015); Brook et al. (2019))

- Through this review the following classes of substances were identified in oil sands mine water, tailings, and air emissions (deposited in the ambient environment). The concentrations and types of chemical substances varies by oil sands operation as extraction, processing and treatment technologies differ by mine. Variability in composition of OSMW was indiscernible using externally available information sources, therefore, all identified contaminated classes were included for identifying Indigenous water use protection goals.
- Inorganic ions (such as salts, ammonia and nutrients),
- Trace elements and heavy metals,
- Volatile organic hydrocarbons (VOCs) including Benzene (B), Toluene (T), Ethylbenzene
- 1825 (E) and Xylene (X),
- Polycyclic aromatic hydrocarbons (PAHs),
- Petroleum hydrocarbon fractions (PHC F1-F4),
- Sulfates, sulfites, and sulfides,
- Nitrate and nitrites, and
- Organic compounds (such as phenols and naphthenic acids).

1831 3.4.3 Available Surface Water Quality Guidelines

- 1832 As identified in the Indigenous water use conceptual model, water quality guidelines are re-
- 1833 quired for both human and ecological (aquatic, wildlife) receptors to meet community identified
- 1834 protection goals for four traditional water use categories; consumption of traditional foods and
- 1835 drinking water, consumption of traditional medicines, wildlife health, and aquatic ecosystem
- 1836 health (Figure 3.1).
- 1837 Chronic surface water quality guidelines for the protection of aquatic biota, wildlife and hu-
- 1838 man receptors were identified from multiple jurisdictions. Available guidelines, by jurisdiction,
- 1839 are briefly described below.
- 1840 Certain parameters (cadmium, copper, lead, nickel and zinc) require the guideline to be
- 1841 calculated using modifying factors for total hardness or alkalinity (as CaCO3 mg/L), pH,
- 1842 water temperature (C), chloride (mg/L) and/ or dissolved organic carbon (mg/L) from the
- 1843 area where guidelines are being applied. Modifying factors were adopted from 50th percentile
- 1844 values in open water season from multiple locations in the Athabasca River (see Chapter 2),
- 1845 summarized in Table 3.1 below.

Table 3.1: Modifying Factors calculated from median values measured during open water season at "Old Fort" from 2011-2019.

| Modifying Factor | Unit | Median |
|--------------------------|----------------------|--------|
| Alkalinity | as CaCO3 mg/L | 110.0 |
| Field pH | pH units | 8.0 |
| Water Temperature | $^{\circ}\mathrm{C}$ | 10.9 |
| Total suspended solids | $\mathrm{mg/L}$ | 24.0 |
| Chloride | m mg/L | 12.0 |
| Total hardness | as CaCO3 mg/L | 120.0 |
| Dissolved organic carbon | m mg/L | 7.9 |
| Total organic carbon | $\mathrm{mg/L}$ | 8.9 |

Generally, ambient water quality and drinking water quality guidelines for the protection of human health endpoints, including carcinogenicity, were prescribed by the US EPA, Health Canada and the WHO while those available from the GOA and CCME were limited to the protection of aquatic biota, livestock (agricultural uses) and wildlife consuming aquatic biota (for a single OSMW contaminant (mercury)).

A detailed comparison of available guidelines for each substance by jurisdiction and water use is provided in Appendix A.3.

Chronic surface water quality guidelines could not be identified for naphthenic acids, BTEX compounds, or petroleum hydrocarbons. For these substances, water use protection criteria are defined by the current conditions described in Chapter 2.

A comparison of available guidelines was used to identify the most sensitive use and/ or receptor group (i.e. aquatic biota, humans, livestock, wildlife) for surface water as shown in Table 3.2. Appendix A.3 should be consulted to determine which guidelines were available for each use.

Table 3.2 indicates that aquatic biota are the most sensitive receptor group for 45% of substances related to oil sands wastes and emissions. As commonly practiced in Alberta, adopting the protection of aquatic life (PAL) guidelines to assess risks from exposure to chemicals in OSMW would limit the protection of humans and wildlife (birds and mammals) which are the most sensitive receptors for exposure to 52% and 3% of the substances in oil sands with available guidelines. As shown in Table 3.2, approximately 52% of chemicals which have been detected in the ambient environment and characterized in OSMW present a higher risk potential to humans, which are not currently considered under provincial guidelines (GoA, 2018).

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies.

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------|--------------------|-----------------|-------|--------------------|---|
| .alphaEndosulfan | | | $\mathrm{ug/L}$ | 0.056 | aquatic biota | US EPA Aquatic Life Criteria |
| .betaEndosulfan | | | $\mathrm{ug/L}$ | 0.056 | aquatic biota | US EPA Aquatic Life Criteria |
| 1,1,1-Trichloroethane | | | $\mathrm{ug/L}$ | 200 | human | US EPA DWR |
| 1,1,2,2-Tetrachloroethane | | | $\mathrm{ug/L}$ | 2 | human | HH DW+Org (US EPA) |
| 1,1,2-Trichloroethane | | | $\mathrm{ug/L}$ | 3 | human | US EPA DWR |
| 1,1-Dichloroethylene | | | $\mathrm{ug/L}$ | 7 | human | US EPA DWR |
| 1,2,3,4-Tetrachlorobenzene | | | $\mathrm{ug/L}$ | 0.03 | human | HH DW+Org (US EPA) USEPA WQC HH Org |
| 1,2,3-Trichlorobenzene | | | $\mathrm{ug/L}$ | 8 | aquatic biota | AEP Water PAL CCME Water PAL |
| 1,2,4-Trichlorobenzene | | | ug/L | 0.071 | human | HH DW+Org (US EPA) |
| $1, 2\hbox{-}{\rm Dibromo-}3\hbox{-}{\rm chloropropane}$ | | | $\mathrm{ug/L}$ | 0.2 | human | US EPA DWR |
| 1,2-Dibromoethane | | | $\mathrm{ug/L}$ | 0.4 | human | WHO DW |
| 1,2-Dichlorobenzene | | | ug/L | 0.7 | aquatic biota | AEP Water PAL |
| 1,2-Dichloroethane | | | ${ m ug/L}$ | 5 | human wildlife | CCME Water Ag AEP Water Ag US EPA DWR Health Canada DW |
| 1,2-Dichloroethene | | | $\mathrm{ug/L}$ | 50 | human | WHO DW |
| 1,2-Dichloropropane | | | $\mathrm{ug/L}$ | 5 | human | US EPA DWR |
| 1,2-Diphenylhydrazine | | | $\mathrm{ug/L}$ | 0.3 | human | HH DW+Org (US EPA) |
| 1,3-Dichlorobenzene | | | ug/L | 7 | human | HH DW+Org (US EPA) |
| 1,3-Dichloropropene | | | ug/L | 2.7 | human | HH DW+Org (US EPA) |
| 1,4-Dichlorobenzene | | | $\mathrm{ug/L}$ | 26 | aquatic biota | AEP Water PAL |
| 1,4-Dioxane | | | $\mathrm{ug/L}$ | 50 | human | WHO DW |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|-----------------------------------|----------------------|--------------------|---------------------------|-------|--------------------|---------------------------------|
| 2,3,4,6-Tetrachlorophenol | | | $\mathrm{ug/L}$ | 1 | human | USEPA WQC AO |
| 2,3-Dichlorophenol | | | $\mathrm{ug/L}$ | 0.04 | human | USEPA WQC AO |
| 2,4,5-Trichlorophenol | | | ug/L | 1 | human | USEPA WQC AO |
| 2,4,6-Trichlorophenol | | | $\mathrm{ug/L}$ | 2 | human | USEPA WQC AO |
| 2,4-D | | | $\mathrm{ng/L}$ | 4 | aquatic biota | AEP Water PAL CCME Water PAL |
| 2,4-DB | | | ug/L | 25 | aquatic biota | AEP Water PAL |
| 2,4-Dichlorophenol | | | ug/L | 0.3 | human | USEPA WQC AO |
| 2,4-Dimethylphenol | | | $\mathrm{ug/L}$ | 100 | human | HH DW+Org (US EPA) |
| 2,4-Dinitrophenol | | | ug/L | 10 | human | HH DW+Org (US EPA) |
| 2,4-Dinitrotoluene | | | $\mathrm{ug/L}$ | 0.49 | human | HH DW+Org (US EPA) |
| 2,5-Dichlorophenol | | | ug/L | 0.5 | human | USEPA WQC AO |
| 2,6-Dichlorophenol | | | $\mathrm{mg/L}$ | 0.2 | human | USEPA WQC AO |
| 2-Chloronaphthalene | | | $\mathrm{ug/L}$ | 800 | human | HH DW+Org (US EPA) |
| 2-Chlorophenol | | | % satu- ra- tion | 0.1 | human | USEPA WQC AO |
| 2-Methyl-4,6-Dinitrophenol | | | $\mathrm{ug/L}$ | 2 | human | HH DW+Org (US EPA) |
| 2-Methyl-4-Chlorophenol | | | $\mathrm{ug/L}$ | 1800 | human | USEPA WQC AO |
| 3,3'-Dichlorobenzidine | | | ng/L | 0.49 | human | HH DW+Org (US EPA) |
| 3,4-Dichlorophenol | | | ug/L | 0.3 | human | USEPA WQC AO |
| 3-Chlorophenol | | | ug/L | 0.1 | human | USEPA WQC AO |
| 3-Iodo-2-propynyl butyl carbamate | | | ng/L | 1.9 | aquatic biota | AEP Water PAL CCME Water PAL |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|-----------------------------|----------------------|--------------------|-----------------|-----------|------------------------|---|
| 3-Methyl-4-Chlorophenol | | | $\mathrm{ug/L}$ | 500 | human | HH DW+Org (US EPA) |
| 3-Methyl-6-Chlorophenol | | | ug/L | 20 | human | USEPA WQC AO |
| 4-Chlorophenol | | | ug/L | 0.1 | human | USEPA WQC AO |
| Acenaphthene | | | ug/L | 5.8 | aquatic biota | CCME Water PAL AEP Water PAL |
| Acridine | | | $\mathrm{ug/L}$ | 4.4 | aquatic biota | AEP Water PAL CCME Water PAL |
| Acrolein | | | $\mathrm{ug/L}$ | 3 | aquatic biota human | US EPA Aquatic Life Criteria HH DW+Org (US EPA) AEP Water PAL |
| Acrylamide | | | ug/L | 0.5 | human | WHO DW US EPA DWR |
| Acrylonitrile | | | $\mathrm{ug/L}$ | 0.61 | human | HH DW+Org (US EPA) |
| Alachlor | | | ug/L | 2 | human | US EPA DWR |
| Alcohol ethoxylates | | | ug/L | 70 | aquatic biota | FEQG Water PAL |
| Aldicarb | | | $\mathrm{ug/L}$ | 1 | aquatic biota | CCME Water PAL AEP Water PAL |
| Aldrin | as N | | $\mathrm{mg/L}$ | 0.0000077 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Aldrin and dieldrin | | | $\mathrm{ug/L}$ | 0.03 | human | WHO DW |
| Alkalinity, total | | | ug/L | 20 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL |
| alpha-Endosulfan | | | ug/L | 20 | human | HH DW+Org (US EPA) |
| alpha-Hexachlorocyclohexane | | | $\mathrm{ug/L}$ | 0.0036 | human | HH DW+Org (US EPA) |
| Aluminum | | Total | ug/L | 100 | aquatic biota | CCME Water PAL |
| Aluminum | | Dissolved | $\mathrm{ug/L}$ | 50 | aquatic biota | AEP Water PAL |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------|--------------------|-----------------|--------|--------------------|---|
| Ammonia | | | $\mathrm{ug/L}$ | 0.794 | aquatic biota | AEP Water PAL |
| Ammonia, unionized | | | ug/L | 0.016 | aquatic biota | AEP Water PAL |
| Aniline | | | $\mathrm{ug/L}$ | 2.2 | aquatic biota | AEP Water PAL CCME Water PAL |
| Anthracene | | | $\mathrm{ug/L}$ | 0.012 | aquatic biota | AEP Water PAL CCME Water PAL |
| Antimony | | Total | $\mathrm{ug/L}$ | 5.6 | human | HH DW+Org (US EPA) |
| Arsenic | | Total | $\mathrm{ug/L}$ | 0.18 | human | HH DW+Org (US EPA) |
| Arsenic | | Dissolved | $\mathrm{ug/L}$ | 150 | aquatic biota | US EPA Aquatic Life Criteria |
| Asbestos | | | $\mathrm{ug/L}$ | 7 | human | US EPA DWR HH DW+Org (US EPA) |
| Atrazine | | | $\mathrm{ug/L}$ | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL |
| Atrazine and its chloro-s-triazine metabolites | | | $\mathrm{ug/L}$ | 100 | human | WHO DW |
| Azinphos-methyl | | | $\mathrm{ug/L}$ | 0.01 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria |
| Barium | | Total | $\mathrm{ug/L}$ | 1000 | human | HH DW+Org (US EPA) Health Canada DW |
| Benzene | | | $\mathrm{ug/L}$ | 5 | human | US EPA DWR Health Canada DW |
| Benzidine | | | ug/L | 0.0014 | human | HH DW+Org (US EPA) |
| Benzo(a)anthracene | | | ug/L | 0.012 | human | HH DW+Org (US EPA) |
| Benzo(a)pyrene | | | ug/L | 0.001 | human | HH DW+Org (US EPA) |
| Benzo(b)fluoranthene | | | ug/L | 0.012 | human | HH DW+Org (US EPA) |
| Benzo(k) fluoranthene | | | ug/L | 0.12 | human | HH DW+Org (US EPA) |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| | Total | ug/L ug/L ug/L | 4 20 0.08 | human human | US EPA DWR HH DW+Org (US EPA) |
|---|-------|-----------------|---|--|---|
| | | ug/L | | | HH DW+Org (US EPA) |
| | | -, | 0.08 | | |
| | | ug/L | | human | HH DW+Org (US EPA) |
| | | | 200 | human | HH DW+Org (US EPA) |
| | | ug/L | 0.3 | human | HH DW+Org (US EPA) |
| | | ug/L | 0.32 | human | HH DW+Org (US EPA) |
| | | ug/L | 0.002 | human | HH DW+Org (US EPA) |
| | | ug/L | 3.5 | aquatic biota | FEQG Water PAL |
| | | ug/L | | | |
| | Total | $\mathrm{ug/L}$ | 1500 | aquatic biota | AEP Water PAL CCME Water PAL |
| | | ug/L | 5 | aquatic biota | CCME Water PAL AEP Water PAL |
| | | ug/L | 10 | human | WHO DW Health Canada DW US EPA DWR |
| | | ug/L | 60 | human | WHO DW |
| | | ug/L | 7 | human | HH DW+Org (US EPA) |
| N | | $\mathrm{mg/L}$ | 5 | aquatic biota human | Health Canada DW AEP Water PAL CCME Water PAL |
| | | $\mathrm{ug/L}$ | 1 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| N | Total | mg/L 0. | 1843828121 | aquatic biota | CCME Water PAL AEP Water PAL |
| | | N | ug/L ug/L ug/L ug/L ug/L ug/L ug/L ug/L | ug/L 3.5 ug/L Total ug/L 1500 ug/L 5 ug/L 10 ug/L 60 ug/L 7 mg/L 5 ug/L 7 | $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|---------------------------|--------------------|-----------------|--------------|--------------------|---------------------------------|
| Cadmium | | Dissolved | $\mathrm{ug/L}$ | 0.8237781279 | aquatic biota | US EPA Aquatic Life Criteria |
| Calcium | | | ug/L | 1000 | wildlife | AEP Water Ag CCME Water Ag |
| Captan | | | ug/L | 1.3 | aquatic biota | AEP Water PAL CCME Water PAL |
| Carbamazepine | | | ug/L | 10 | aquatic biota | CCME Water PAL AEP Water PAL |
| Carbaryl | | | ug/L | 0.2 | aquatic biota | CCME Water PAL AEP Water PAL |
| Carbofuran | | | ug/L | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL |
| Carbon tetrachloride | | | ug/L | 2 | human | Health Canada DW |
| Chloramines | | | ug/L | 0.5 | aquatic biota | CCME Water PAL |
| Chlorate | | | ng/L | 700 | human | WHO DW |
| Chlordane | | | ug/L | 0.003 | human | HH DW+Org (US EPA) |
| Chloride | | | ug/L | 120 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chlorinated paraffins, long-chain, C18-C20 | | | ug/L | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL |
| Chlorinated paraffins, medium-chain, C14-C17 | | | ug/L | 2.4 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Chlorinated paraffins, short-chain, $C10\text{-}C13$ | as paraquat dichloride | | ug/L | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL |
| Chlorine | | | ug/L | 0.5 | aquatic biota | AEP Water PAL |
| Chlorine dioxide | | | ug/L | 800 | human | US EPA DWR |
| Chlorite | | | ng/L | 700 | human | WHO DW |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------|--------------------|-----------------|-------------|--------------------|---------------------------------|
| Chlorobenzene | | | $\mathrm{ng/L}$ | 1.3 | aquatic biota | AEP Water PAL |
| Chlorodibromomethane | | | ng/L | 8 | human | HH DW+Org (US EPA) |
| Chloroform | | | ug/L | 1.8 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chlorophenol | | | ug/L | 7 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chlorophenoxy Herbicide (2,4,5-TP) [Silvex] | | | ug/L | 50 | human | US EPA DWR |
| Chlorothalonil | | | ug/L | 0.18 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chlorotoluron | | | ug/L | 30 | human | WHO DW |
| Chlorpyrifos | | | $\mathrm{ug/L}$ | 0.002 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chromium | | Total | pH units | 50 | human | WHO DW Health Canada DW |
| Chromium (III) | | Total | ug/L | 8.9 | aquatic biota | CCME Water PAL AEP Water PAL |
| Chromium (III) | | Dissolved | $\mathrm{ug/L}$ | 100.9185723 | aquatic biota | US EPA Aquatic Life Criteria |
| Chromium (VI) | | Total | $\mathrm{ug/L}$ | 1 | aquatic biota | AEP Water PAL CCME Water PAL |
| Chromium (VI) | | Dissolved | $\mathrm{ug/L}$ | 5 | aquatic biota | FEQG Water PAL |
| Chrysene | | | $\mathrm{ug/L}$ | 1.2 | human | HH DW+Org (US EPA) |
| cis-1,2-Dichloroethylene | | | ug/L | 70 | human | US EPA DWR |
| Cobalt | | Total | $\mathrm{ug/L}$ | 1.099682588 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Copper | | Total | ug/L | 2.763433095 | aquatic biota | CCME Water PAL |
| | | | | | | |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Cyanazine ug/L 0.6 human WHO DW Cyanide ug/L 4 human HH DW+Org (US EPA) Cyanobacterial toxins ug/L 1.5 human Health Canada DW Dalapon ug/L 200 human US EPA DWR DDT and metabolites ug/L 0.0003 human US EPA WQC HH Org HH DW+Org (US EPA) Dehydroabietic acid ug/L ug/L CCME Water PAL AEP | Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------------|----------------------|--------------------|--------------------------|--------|--------------------|---|
| Cyanide ug/L 4 human HH DW+Org (US EPA) Cyanobacterial toxins ug/L 1.5 human Health Canada DW Dalapon ug/L 200 human US EPA DWR DDT and metabolites ug/L 0.0003 human US EPA WQC HH Org HH DW+Org (US EPA) Dehydroabietic acid ug/L Ug/L WH DW+Org (US EPA) Deltamethrin as SO4 mg/L 0.0004 aquatic biota CCME Water PAL AEP Water PAL AE | Copper | | Dissolved | $\mathrm{ug/L}$ | 0.53 | aquatic biota | FEQG Water PAL |
| Cyanobacterial toxins Dalapon Ug/L Dalapon Ug/L DDT and metabolites Ug/L Delydroabietic acid Deltamethrin as SO4 Deltamethrin Deltamethrin Deltamethrin Di(2-ethylhexyl) adipate Di(3-ethylhexyl) phthalate Di-n-Butyl Phthalate Di-n-Butyl Phthalate Dibenzo(a,h)anthracene Dibenz | Cyanazine | | | $\mathrm{ug/L}$ | 0.6 | human | WHO DW |
| Dalapon ug/L 200 human US EPA DWR DDT and metabolites ug/L 0.0003 human USEPA WQC HH Org HH DW+Org (US EPA) Dehydroabietic acid ug/L Deltamethrin as SO4 mg/L 0.0004 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL AEP Water PAL AEP Water PAL Oli(2-ethylhexyl) adipate ug/L 400 human US EPA DWR Di(2-ethylhexyl) phthalate ug/L 6 human US EPA DWR Di-n-Butyl Phthalate ug/L 19 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL Oliazinon ug/L 0.17 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL AEP Water PAL AEP Water PAL Oliazinon ug/L 0.17 aquatic biota US EPA Aquatic Life Criteria AEP Water PAL AEP Water PAL Oliazinon ug/L 0.001 human HH DW+Org (US EPA) Dibenzo(a,h)anthracene ug/L 70 human WHO DW Dibromocetonitrile ug/L 70 human WHO DW Dibromocetonitrile ug/L 100 human WHO DW AEP Water Ag WHO DW AEP Water Ag WHO DW AEP Water Ag CCME Water PAL AEP Wate | Cyanide | | | ug/L | 4 | human | HH DW+Org (US EPA) |
| DDT and metabolites ug/L ug/L Deltydroabietic acid ug/L Deltamethrin as SO4 mg/L Deltamethrin as SO4 mg/L Deltydroabietic acid mg/L Deltamethrin as SO4 mg/L Deltydroabietic acid Demeton mg/L Deltydroabietic acid aquatic biota Deltydroabieta Deltydroabi | Cyanobacterial toxins | | | ug/L | 1.5 | human | Health Canada DW |
| Dehydroabietic acid Deltamethrin as SO4 mg/L Demeton mg/L De | Dalapon | | | ug/L | 200 | human | US EPA DWR |
| Deltamethrin as SO4 mg/L 0.0004 aquatic biota CCME Water PAL AEP Water PAL Demeton mg/L 0.1 aquatic biota US EPA Aquatic Life Criteria AEP Water PAL US EPA Aquatic Life Criteria AEP Water PAL Di(2-ethylhexyl) adipate ug/L 400 human US EPA DWR Di(2-ethylhexyl) phthalate ug/L 6 human US EPA DWR Di-n-Butyl Phthalate ug/L 19 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL Diazinon ug/L 0.17 aquatic biota US EPA Aquatic Life Criteria AEP Water PAL Dibenzo(a,h)anthracene ng/L 0.001 human HH DW+Org (US EPA) Dibromoacetonitrile ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human CCME Water Ag WHO DW AEP Water Ag Water PAL AEP Water PAL A | DDT and metabolites | | | $\mathrm{ug/L}$ | 0.0003 | human | |
| Demeton mg/L 0.1 aquatic biota US EPA Aquatic Life Criteria AEP Water PAL Di(2-ethylhexyl) adipate ug/L 400 human US EPA DWR Di(2-ethylhexyl) phthalate ug/L 6 human US EPA DWR Di-n-Butyl Phthalate ug/L 19 aquatic biota CCME Water PAL AEP Water PAL Diazinon ug/L 0.17 aquatic biota CS EPA Aquatic Life Criteria AEP Water PAL Dibenzo(a,h)anthracene ng/L 0.001 human HH DW+Org (US EPA) Dibromoacetonitrile ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human wildlife WHO DW AEP Water Ag Dicamba ug/L 10 aquatic biota CCME Water Ag Di aquatic biota CCME Water PAL AEP Water PAL CCME Water Ag CCME Water Ag CCME Water Ag CCME Water Ag CCME Water PAL AEP Water PAL AEP Water PAL CCME Water PAL AEP Water PAL | Dehydroabietic acid | | | $\mathrm{ug/L}$ | | | |
| AEP Water PAL Di(2-ethylhexyl) adipate ug/L 400 human US EPA DWR Di(2-ethylhexyl) phthalate ug/L 6 human US EPA DWR Di-n-Butyl Phthalate ug/L 19 aquatic biota CCME Water PAL AEP Water PAL Diazinon ug/L 0.17 aquatic biota US EPA Aquatic Life Criteria AEP Water PAL Dibenzo(a,h)anthracene ng/L 0.001 human HH DW+Org (US EPA) Dibromoacetonitrile ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human CCME Water Ag Dicamba ug/L 10 aquatic biota CCME Water PAL AEP Water PAL CCME Water PAL CCME Water Ag WHO DW AEP Water Ag CCME Water Ag WHO DW AEP Water Ag CCME Water PAL AEP Water PAL AEP Water PAL | Deltamethrin | as SO4 | | $\mathrm{mg/L}$ | 0.0004 | aquatic biota | 0 0 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 |
| Di(2-ethylhexyl) phthalate Ug/L 19 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL Diazinon Ug/L 0.17 aquatic biota US EPA DWR US EPA DWR CCME Water PAL AEP Water PAL Dibenzo(a,h)anthracene ng/L 0.001 human HH DW+Org (US EPA) Dibromoacetonitrile Ug/L 0.001 human WHO DW Dibromochloromethane Ug/L 100 human CCME Water Ag WHO DW AEP Water Ag WHO DW AEP Water Ag COME Water Ag WHO DW AEP Water Ag COME Water PAL | Demeton | | | mg/L | 0.1 | aquatic biota | |
| Di-n-Butyl Phthalate ug/L 19 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL US EPA Aquatic Life Criteria AEP Water PAL US EPA Aquatic Life Criteria AEP Water PAL Dibenzo(a,h)anthracene ng/L 0.001 human HH DW+Org (US EPA) Dibromoacetonitrile ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human wildlife WHO DW AEP Water Ag WHO DW AEP Water Ag WHO DW AEP Water Ag CCME Water Ag WHO DW AEP Water Ag Dicamba Ug/L 10 aquatic biota CCME Water PAL AEP Water PAL | Di(2-ethylhexyl) adipate | | | $\mathrm{ug/L}$ | 400 | human | US EPA DWR |
| Diazinon ug/L 0.17 aquatic biota ug/L 0.18 EPA Aquatic Life Criteria AEP Water PAL Dibenzo(a,h)anthracene ug/L 0.001 human ug/L | Di(2-ethylhexyl) phthalate | | | $\mathrm{ug/L}$ | 6 | human | US EPA DWR |
| AEP Water PAL Dibenzo(a,h)anthracene ng/L ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human wildlife WHO DW AEP Water Ag WHO DW AEP Water Ag Dicamba ug/L 10 aquatic biota CCME Water PAL AEP Water PAL AEP Water PAL | Di-n-Butyl Phthalate | | | $\mathrm{ug/L}$ | 19 | aquatic biota | |
| Dibromoacetonitrile ug/L 70 human WHO DW Dibromochloromethane ug/L 100 human CCME Water Ag wildlife WHO DW AEP Water Ag Dicamba ug/L 10 aquatic biota CCME Water PAL AEP Water PAL | Diazinon | | | $\mathrm{ug/L}$ | 0.17 | aquatic biota | |
| Dibromochloromethane $ \begin{array}{ccccccccccccccccccccccccccccccccccc$ | Dibenzo(a,h)anthracene | | | $\mathrm{ng/L}$ | 0.001 | human | HH DW+Org (US EPA) |
| wildlife WHO DW AEP Water Ag Dicamba ug/L 10 aquatic biota CCME Water PAL AEP Water PAL | Dibromoacetonitrile | | | $\mathrm{ug/L}$ | 70 | human | WHO DW |
| AEP Water PAL | Dibromochloromethane | | | ug/L | 100 | | WHO DW |
| Dichloroacetate $\rm ug/L$ 50 human WHO DW | Dicamba | | | ug/L | 10 | aquatic biota | |
| | Dichloroacetate | | | $\mathrm{ug/L}$ | 50 | human | WHO DW |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|------------------------------------|----------------------|--------------------|-----------------|------------|--------------------|--|
| Dichloroacetonitrile | | | ug/L | 20 | human | WHO DW |
| Dichlorobromomethane | | | ug/L | 9.5 | human | HH DW+Org (US EPA) |
| Dichloromethane | | | $\mathrm{mg/L}$ | 5 | human | US EPA DWR |
| Dichlorophenol | | | ug/L | 0.2 | aquatic biota | AEP Water PAL CCME Water PAL |
| Dichlorprop | | | Toxic units | 100 | human | WHO DW |
| Diclofop-methyl | | | $\mathrm{ug/L}$ | 6.1 | aquatic biota | AEP Water PAL CCME Water PAL |
| Didecyl dimethyl ammonium chloride | | | $\mathrm{ug/L}$ | 1.5 | aquatic biota | AEP Water PAL CCME Water PAL |
| Dieldrin | | | ng/L | 0.00001 | human | HH DW+Org (US EPA) |
| Diethanolamine | | | ug/L | 450 | aquatic biota | AEP Water PAL |
| Diethyl Phthalate | | | ug/L | 600 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Diethylene glycol | | | ug/L | 150000 | aquatic biota | AEP Water PAL |
| Diisopropanolamine | | | $\mathrm{ug/L}$ | 1600 | aquatic biota | CCME Water PAL AEP Water PAL |
| Dimethoate | | | ug/L | 3 | wildlife | AEP Water Ag CCME Water Ag |
| Dimethyl Phthalate | | | ug/L | 2000 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Dinitrophenols | | | ug/L | 10 | human | HH DW+Org (US EPA) |
| Dinoseb | | | $\mathrm{ug/L}$ | 0.05 | aquatic biota | CCME Water PAL AEP Water PAL |
| Dioxin $(2,3,7,8\text{-TCDD})$ | | | ug/L | 0.00000005 | human | HH DW+Org (US EPA) |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--------------------|----------------------|--------------------|-----------------|--------|--------------------|--|
| Diquat | | | ug/L | 20 | human | US EPA DWR |
| Diuron | | | ug/L | 150 | human | Health Canada DW |
| Dummy | | | ug/L | 0 | medicinal | Derived traditional plant |
| Edetic acid | | | ug/L | 600 | human | WHO DW |
| Endosulfan | | | $\mathrm{ug/L}$ | 0.003 | aquatic biota | AEP Water PAL CCME Water PAL |
| Endosulfan Sulfate | | | ug/L | 20 | human | HH DW+Org (US EPA) |
| Endothall | | | ug/L | 100 | human | US EPA DWR |
| Endrin | | | $\mathrm{ug/L}$ | 0.0023 | aquatic biota | AEP Water PAL CCME Water PAL |
| Endrin Aldehyde | | | $\mathrm{ug/L}$ | 1 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Epichlorohydrin | | | ug/L | 0.4 | human | WHO DW |
| Ethinyl estradiol | | | ug/L | 0.5 | aquatic biota | AEP Water PAL |
| Ethylbenzene | | | $\mathrm{ug/L}$ | 2.4 | wildlife | AEP Water Ag CCME Water Ag |
| Ethylene dibromide | | | ug/L | 0.05 | human | US EPA DWR |
| Ethylene glycol | | | $\mathrm{ug/L}$ | 192000 | aquatic biota | CCME Water PAL AEP Water PAL |
| Fenoprop | | | ug/L | 9 | human | WHO DW |
| Fluoranthene | | | $\mathrm{ug/L}$ | 0.04 | aquatic biota | AEP Water PAL CCME Water PAL |
| Fluorene | | | $\mathrm{mg/L}$ | 3 | aquatic biota | AEP Water PAL CCME Water PAL |
| Fluoride | | | ug/L | 0.12 | aquatic biota | CCME Water PAL |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------|--------------------|-----------------|----------|--------------------|---|
| gamma-Hexachlorocyclohexane [Lindane] | | | $\mathrm{ug/L}$ | 0.01 | aquatic biota | AEP Water PAL |
| Gases (total Dissolved) | | | $\mathrm{ug/L}$ | | | |
| Glyphosate | | | ng/L | 280 | human wildlife | CCME Water Ag Health Canada DW AEP Water Ag |
| Haloacetic acids | | | ug/L | 60 | human | US EPA DWR |
| heptaBDE | | | $\mathrm{ug/L}$ | 14 | aquatic biota | FEQG Water PAL |
| Heptachlor | | | ng/L | 0.000059 | human | USEPA WQC HH Org |
| Heptachlor epoxide | | | $\mathrm{ug/L}$ | 0.00032 | human | HH DW+Org (US EPA) USEPA WQC HH Org |
| hexaBDE | | | $\mathrm{ug/L}$ | 120 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Hexabromocyclododecane | | | $\mathrm{ug/L}$ | 0.56 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Hexachlorobenzene | | | $\mathrm{ug/L}$ | 0.00079 | human | USEPA WQC HH Org |
| Hexachlorobutadiene | | | $\mathrm{ug/L}$ | 0.1 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Hexachlorocyclohexane | | | ug/L | 0.01 | aquatic biota | CCME Water PAL |
| Hexachlorocyclopentadiene | | | $\mathrm{ug/L}$ | 1 | human | USEPA WQC AO |
| Hexachloroethane | | | $\mathrm{ug/L}$ | 1 | human | USEPA WQC HH Org HH DW+Org (US EPA) |
| Hydrazine | | | $\mathrm{ug/L}$ | 2.6 | aquatic biota | FEQG Water PAL AEP Water PAL |
| Hydrogen Sulfide | | | $\mathrm{ug/L}$ | 2 | aquatic biota | US EPA Aquatic Life Criteria |
| Hydroxyatrazine | | | $\mathrm{ug/L}$ | 200 | human | WHO DW |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source | |
|--|----------------------|--------------------|-----------------|-------------|------------------------|---|--|
| Imidacloprid | as N | | $\mathrm{mg/L}$ | 0.23 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Indeno(1,2,3-cd)pyrene | | | ug/L | 0.012 | human | HH DW+Org (US EPA) | |
| Inorganic nitrogen (nitrate and nitrite) | | Dissolved | ug/L | 100 | wildlife | AEP Water Ag CCME Water Ag | |
| Iron | | Total | $\mathrm{ug/L}$ | 300 | aquatic biota human | USEPA WQC AO CCME Water PAL | |
| Iron | | Dissolved | ug/L | 300 | aquatic biota | AEP Water PAL | |
| Isophorone | | | ug/L | 340 | human | HH DW+Org (US EPA) | |
| Isoproturon | | | ug/L | 9 | human | WHO DW | |
| Lead | | Total | $\mathrm{ug/L}$ | 4.01275079 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Lead | | Dissolved | ug/L | 3.067487163 | aquatic biota | US EPA Aquatic Life Criteria | |
| Linuron | | | m ug/L 7 | 7 | aquatic biota | CCME Water PAL AEP Water PAL | |
| m-Dichlorobenzene | | | ug/L | 150 | aquatic biota | CCME Water PAL | |
| Malathion | | | $\mathrm{ug/L}$ | 0.1 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | |
| Manganese | | Total | ug/L | 50 | human | HH DW+Org (US EPA) | |
| MCPA | | | $\mathrm{ug/L}$ | 2.6 | aquatic biota | AEP Water PAL CCME Water PAL | |
| Mecoprop | | | ug/L | 10 | human | WHO DW | |
| Mercury | | Total | ug/L | 0.005 | aquatic biota | AEP Water PAL | |
| Mercury | | Dissolved | ug/L | 0.77 | aquatic biota | US EPA Aquatic Life Criteria | |
| Mercury (methyl) | | Total | ug/L | 0.001 | aquatic biota | AEP Water PAL | |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|---------------------------|----------------------|--------------------|--------------------------|-------|--------------------|---|
| Mercury (methyl) | | Dissolved | $\mathrm{ug/L}$ | 0.004 | aquatic biota | CCME Water PAL |
| Methanol | | | ug/L | 1500 | aquatic biota | AEP Water PAL |
| Methoprene | | | $\mathrm{ug/L}$ | 0.09 | aquatic biota | AEP Water PAL CCME Water PAL |
| Methoxychlor | | | $\mathrm{ug/L}$ | 0.02 | human | HH DW+Org (US EPA) USEPA WQC HH Org |
| Methyl Bromide | | | ug/L | 100 | human | HH DW+Org (US EPA) |
| Methyl tert-butyl ether | | | $\mathrm{ug/L}$ | 10 | aquatic biota | AEP Water PAL |
| Methylene chloride | | | $\mathrm{ug/L}$ | 98.1 | aquatic biota | CCME Water PAL AEP Water PAL |
| Metolachlor | | | $\mathrm{ug/L}$ | 7.8 | aquatic biota | CCME Water PAL AEP Water PAL |
| Metribuzin | | | $\mathrm{ug/L}$ | 1 | aquatic biota | AEP Water PAL CCME Water PAL |
| Microcystin-LR | | | $\mathrm{ug/L}$ | 1 | human | WHO DW |
| Mirex | | | $\mathrm{ug/L}$ | 0.001 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL |
| Molinate | | | $\mathrm{ug/L}$ | 6 | human | WHO DW |
| Molybdenum | | Total | $\mathrm{ug/L}$ | 73 | aquatic biota | AEP Water PAL CCME Water PAL |
| Monochloramine | | | $\mathrm{ug/L}$ | 3000 | human | WHO DW |
| Monochloroacetate | | | ug/L | 20 | human | WHO DW |
| Monochlorobenzene | | | $\mathrm{ug/L}$ | 1.3 | aquatic biota | AEP Water PAL CCME Water PAL |
| Monoethanolamine | | | ug/L | 75 | aquatic biota | AEP Water PAL |
| N-Nitrosodi-n-Propylamine | | | $\mathrm{ug/L}$ | 0.05 | human | HH DW+Org (US EPA) |
| | | | | | | |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--|----------------------|--------------------|-----------------|-------|--------------------|---------------------------------|
| N-Nitrosodimethylamine | | | $\mathrm{ug/L}$ | 0.007 | human | HH DW+Org (US EPA) |
| N-Nitrosodiphenylamine | | | $\mathrm{ug/L}$ | 33 | human | HH DW+Org (US EPA) |
| Naphthalene | as N | | mg/L | 1 | aquatic biota | AEP Water PAL |
| Nickel | | Total | $\mathrm{ug/L}$ | 60.86 | aquatic biota | AEP Water PAL |
| Nickel | as N | Dissolved | mg/L | 60.67 | aquatic biota | US EPA Aquatic Life Criteria |
| Nitrate | | Dissolved | $\mathrm{ug/L}$ | 3 | aquatic biota | CCME Water PAL AEP Water PAL |
| Nitrilotriacetic acid | | | ug/L | 200 | human | WHO DW |
| Nitrite | | Dissolved | $\mathrm{ug/L}$ | 0.06 | aquatic biota | CCME Water PAL |
| Nitrobenzene | | | ug/L | 10 | human | HH DW+Org (US EPA) |
| Nitrosamines | | | $\mathrm{ug/L}$ | 0.008 | human | HH DW+Org (US EPA) |
| Nitrosodibutylamine | | | ug/L | 0.063 | human | HH DW+Org (US EPA) |
| Nitrosodiethylamine | | | $\mathrm{ug/L}$ | 0.008 | human | HH DW+Org (US EPA) |
| Nitrosopyrrolidine | | | ug/L | 0.16 | human | HH DW+Org (US EPA) |
| Nonylphenol | | | $\mathrm{ng/L}$ | 6.6 | aquatic biota | US EPA Aquatic Life Criteria |
| Nonylphenol and its ethoxylates | | | ug/L | 1 | aquatic biota | CCME Water PAL |
| o-Dichlorobenzene | | | $\mathrm{ug/L}$ | 0.7 | aquatic biota | CCME Water PAL AEP Water PAL |
| octaBDE | | | ug/L | 14 | aquatic biota | FEQG Water PAL |
| Oxamyl (Vydate) | | | $\mathrm{ug/L}$ | 200 | human | US EPA DWR |
| p,p - Dichlorodiphenyldichloroethane (DDD) | | | $\mathrm{ug/L}$ | 0.001 | human | HH DW+Org (US EPA) |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source | |
|--|---------------------------|--------------------|------------------|---------|---------------------------------|---|--|
| p,p - Dichlorodiphenyldichloroethylene (DDE) | | | $\mathrm{ug/L}$ | 0.00018 | human | USEPA WQC HH Org | |
| p-Dichlorobenzene | as paraquat dichloride | | $\mathrm{ug/L}$ | 5 | human | Health Canada DW | |
| Paraquat | | | ug/L | 10 | human | Health Canada DW | |
| Parathion | | | ug/L 0.013 | | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | |
| Pendimethalin | | | ng/L | 20 | human | WHO DW | |
| pentaBDE | | | $\mathrm{ng/L}$ | 0.2 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| pentaBDE (BDE-100) | | | ${ m ng/L}$ | 0.2 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| pentaBDE (BDE-99) | | | $\mathrm{ug/L}$ | 4 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| Pentachlorobenzene | | | $\mathrm{ug/L}$ | 0.1 | human | USEPA WQC HH Org HH DW+Org (US EPA) | |
| Pentachlorophenol | | | ug/L | 0.3 | human | HH DW+Org (US EPA) | |
| Perchlorate | | | ug/L | 70 | human | WHO DW | |
| Perfluorooctanesulfonate | | | ug/L | 0.6 | human | Health Canada DW | |
| Perfluorooctanoic acid | | | ug/L | 0.2 | human | Health Canada DW | |
| Permethrin | | | $_{ m pH}$ units | 0.004 | aquatic biota | AEP Water PAL CCME Water PAL | |
| рН | | | m ug/L | 9-Jul | aquatic biota human human | HH DW+Org (US EPA) US EPA Aquatic Life Criteria CCME Water PAL AEP Water PAL Health Canada DW | |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|----------------------------------|----------------------|--------------------|-----------------|---------|--------------------|---------------------------------|
| Phenanthrene | | | $\mathrm{ug/L}$ | 0.4 | aquatic biota | AEP Water PAL CCME Water PAL |
| Phenol | | | $\mathrm{ug/L}$ | 2 | wildlife | AEP Water Ag CCME Water Ag |
| Phorate | | | ug/L | 2 | human | Health Canada DW |
| Picloram | | | $\mathrm{ug/L}$ | 29 | aquatic biota | AEP Water PAL CCME Water PAL |
| Polychlorinated Biphenyls (PCBs) | | | ug/L | 0.00064 | human | USEPA WQC HH Org |
| Propylene glycol | | | $\mathrm{ug/L}$ | 500000 | aquatic biota | AEP Water PAL CCME Water PAL |
| Pyrene | | | $\mathrm{ug/L}$ | 0.025 | aquatic biota | AEP Water PAL CCME Water PAL |
| Quinoline | | | $\mathrm{ug/L}$ | 3.4 | aquatic biota | AEP Water PAL CCME Water PAL |
| Selenium | | Total | ug/L | 1 | aquatic biota | CCME Water PAL |
| Silver | | Total | $\mathrm{ug/L}$ | 0.25 | aquatic biota | CCME Water PAL AEP Water PAL |
| Simazine | | | ug/L | 2 | human | WHO DW |
| Sodium dichloroisocyanurate | | | ug/L | 40000 | human | WHO DW |
| Solids Dissolved and Salinity | | | ug/L | 250000 | human | HH DW+Org (US EPA) |
| Strontium | | Total | ug/L | 7000 | human Health Ca | Health Canada DW |
| Styrene | as SO4 | | $\mathrm{mg/L}$ | 20 | human | WHO DW |
| Sulfate | | | $\mathrm{mg/L}$ | 250 | human | WHO DW |
| Sulfide | | | ug/L | 0.0019 | aquatic biota | AEP Water PAL |
| Sulfolane | | | ug/L | 50 | aquatic biota | AEP Water PAL |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source | |
|----------------------------|----------------------|--------------------|---|--------|--------------------|---------------------------------|--|
| Tebuthiuron | | | ug/L | 1.6 | aquatic biota | CCME Water PAL | |
| Terbufos | | | ug/L | 1 | human | Health Canada DW | |
| Terbuthylazine | | | ng/L | 7 | human | WHO DW | |
| tetraBDE | | | $\mathrm{ug/L}$ | 24 | aquatic biota | AEP Water PAL FEQG Water PAL | |
| Tetrabromobisphenol A | | | $\mathrm{ug/L}$ | 3.1 | aquatic biota | FEQG Water PAL AEP Water PAL | |
| Tetrachloroethane | | | ug/L | 13.3 | aquatic biota | CCME Water PAL | |
| Tetrachloroethylene | | | ug/L | 5 | human | US EPA DWR | |
| Tetrachlorophenol | | | $\mathrm{ug/L}$ | 1 | aquatic biota | CCME Water PAL AEP Water PAL | |
| Thallium | | Total | ug/L | 0.24 | human | HH DW+Org (US EPA) | |
| Toluene | | | $\mathrm{mg/L}$ | 0.5 | aquatic biota | AEP Water PAL | |
| Total Dissolved solids | | | $\mathrm{ug/L}$ | 3000 | wildlife | AEP Water Ag CCME Water Ag | |
| Toxaphene | | | ug/L | 0.0002 | aquatic biota | US EPA Aquatic Life Criteria | |
| Toxicity (acute) | | | $\begin{array}{c} \text{Toxic} \\ \text{Units} \\ \text{(TUa)} \end{array}$ | 0.3 | aquatic biota | AEP Water PAL | |
| Toxicity (chronic) | | | $\begin{array}{c} \text{Toxic} \\ \text{Units} \\ \text{(TUc)} \end{array}$ | 1.0 | aquatic biota | AEP Water PAL | |
| Trans-1,2-Dichloroethylene | | | ng/L | 100 | human | HH DW+Org (US EPA) | |
| Triallate | | | $\mathrm{ug/L}$ | 0.24 | aquatic biota | CCME Water PAL AEP Water PAL | |
| ${ m triBDE}$ | | | ug/L | 46 | aquatic biota | AEP Water PAL FEQG Water PAL | |

Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Value | Sensitive Receptor | Source |
|--------------------|----------------------|--------------------|-----------------|--------|--------------------|---------------------------------|
| Tribromomethane | | | $\mathrm{ug/L}$ | 100 | wildlife | CCME Water Ag |
| Tributyltin | | | ug/L | 0.008 | aquatic biota | CCME Water PAL |
| Trichlorfon | | | $\mathrm{ug/L}$ | 0.009 | aquatic biota | CCME Water PAL AEP Water PAL |
| Trichloroacetate | | | $\mathrm{ug/L}$ | 200 | human | WHO DW |
| Trichloroethylene | | | $\mathrm{ug/L}$ | 5 | human | Health Canada DW US EPA DWR |
| Trichlorophenol | | | $\mathrm{ug/L}$ | 18 | aquatic biota | AEP Water PAL CCME Water PAL |
| Triclosan | | | ug/L | 0.47 | aquatic biota | FEQG Water PAL |
| Tricyclohexyltin | | | $\mathrm{ug/L}$ | 250 | wildlife | AEP Water Ag CCME Water Ag |
| Triethylene glycol | | | ug/L | 350000 | aquatic biota | AEP Water PAL |
| Trifluralin | | | $\mathrm{ug/L}$ | 0.2 | aquatic biota | CCME Water PAL AEP Water PAL |
| Trihalomethanes | | | ug/L | 80 | human | US EPA DWR |
| Triphenyltin | | | $\mathrm{ug/L}$ | 0.022 | aquatic biota | AEP Water PAL CCME Water PAL |
| Uranium | | Total | $\mathrm{ug/L}$ | 15 | aquatic biota | AEP Water PAL CCME Water PAL |
| Vanadium | | Total | $\mathrm{ug/L}$ | 100 | wildlife | AEP Water Ag CCME Water Ag |
| Vinyl chloride | | | $\mathrm{ug/L}$ | 0.22 | human | HH DW+Org (US EPA) |
| Xylene | | | $\mathrm{ug/L}$ | 30 | aquatic biota | AEP Water PAL |
| Xylenes (total) | | | $\mathrm{ug/L}$ | 10000 | human | US EPA DWR |
| Zinc | | Total | $\mathrm{ug/L}$ | 30 | aquatic biota | AEP Water PAL |

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Table 3.2: Identification of most stringent surface water quality guidelines and sensitive receptor as published by provincial, federal and international regulatory agencies. *(continued)*

| Parameter | Method Speciation | Sample Fraction | - | | Sensitive Receptor | Source |
|-----------|----------------------|--------------------|------|-------|--------------------|----------------|
| Zinc | | Dissolved | ug/L | 31.34 | aquatic biota | CCME Water PAL |

1869 3.4.4 Water Quality Criteria for the Protection of Indigenous Use 1870 (adopted)

- Based on review of available guidelines described in Section 3.4.3 existing guidelines can offer a degree of protection for the goals, and endpoints identified for Indigenous water uses (Figure 3.1) and were adopted as health risk criteria when appropriate. As discussed above, the degree of health protection varies by agency and substance and available guidelines could only be adopted for two two Indigenous water use categories; wildlife health and aquatic ecosystem health (Figure 3.1), as described below.
- For wildlife health and aquatic ecosystem health water use categories, individual PAH congeners should be compared to indicated criteria, when available. However, criteria could not be established for all PAH congeners. In these cases, the sum of low and high molecular weight (MW) congeners should be compared to the criteria for naphthalene and BaP, respectively. The equations below can be used to estimate concentrations of low and high MW PAH mixtures which exert toxicity through the same mechanism of action (CCME, 2010).
- Low MW PAHs = (Anthracene, Acenaphthene, Acenaphthylene, Fluoranthene, Fluorene,
 Naphthalene, Phenanthrene, Pyrene)
- High MW PAHs = (Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene,

 Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene)

1887 Wildlife Health

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- Surface water concentrations for the protection of piscivorous wildlife species consuming surface water and fish were identified in Sample et al. (1996).
- Additionally, in Alberta, Tier 1 soil and groundwater remediation guidelines consider the protection of surface water for wildlife watering (via hydraulically connected groundwater) by modifying the livestock/ agriculture guidelines to account for contaminant migration from groundwater to surface water (AEP, 2019).
 - Aligning with Alberta guidance, livestock watering guidelines for agricultural water uses were also considered applicable to wildlife species to assess potential risks to wildlife health from ingestion of contaminants in water sources. Review of the protocol for deriving livestock watering guidelines for agricultural uses indicates that livestock watering guidelines were developed, where possible, for both agricultural bird (i.e. poultry) and large mammal (i.e. cattle) species (CCME, 2021). The agricultural species are similar to wildlife species of cultural importance to Indigenous communities (i.e., mallard, lesser scaup, moose) further supporting the

| application | of liv | vestock | watering | guide | lines | to | avian | and | mammalian | wildlife | ١. |
|-------------|--------|---------|----------|-------|-------|----|-------|-----|-----------|----------|----|
| | | | | | | | | | | | |

As the development of new livestock water guidelines is a complex process (CCME, 2021), the surface water quality protection goals for wildlife consuming surface water are limited to those defined by AEP (GoA, 2018) and CCME and the surface water benchmarks published by Sample et al. (1996) which is not representative of all identified substances, but it is a first step in protecting wildlife health more broadly. The health risk criteria for the protection of wildlife health from consuming drinking water and fish are provided in Table 3.3.

It is important to note, concentrations of substances required for the protection of wildlife species may be greater than (meaning less conservative than) concentrations associated with toxicological responses in more sensitive receptors (i.e., humans or aquatic biota).

Finally, the health risk criteria for wildlife, should not be adopted unless all other water use categories described in Figure 3.1 have been assessed and identified as not applicable or non-operational (i.e., the surface water being assessed is not used by humans or aquatic biota).

Table 3.3: Health risk criteria for the protection of wildlife species

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | CCME Water Ag | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|--------------------------------|----------------------|--------------------|-----------------|--------------|------------------|-----------------|----------------------------------|-------------------------------|
| 1,1-Dichloroethylene | | | ug/L | | | 929 | 929 | US DOE Wildlife |
| 1,2-Dichloroethane | | | ug/L | 5 | 5 | 4284 | 5 | AEP Water Ag CCME Water Ag |
| Aldicarb | | | ug/L | 11 | 11 | | 11 | AEP Water Ag CCME Water Ag |
| Aldrin | | | ug/L | | | 0.001 | 0.001 | US DOE Wildlife |
| Aluminum | | Total | ug/L | 5000 | 5000 | 18 | 18 | US DOE Wildlife |
| Antimony | | Total | ug/L | | | 161 | 161 | US DOE Wildlife |
| Arsenic | | Total | ug/L | 25 | 25 | 16 | 16 | US DOE Wildlife |
| Atrazine | | | ug/L | 5 | 5 | | 5 | AEP Water Ag CCME Water Ag |
| Benzene | | | ug/L | | | 2293 | 2293 | US DOE Wildlife |
| Benzo(a)pyrene and equivalents | | | ug/L | | | 0.006722 | 0.006722 | US DOE Wildlife |
| Beryllium | | Total | ug/L | 100 | 100 | 136 | 100 | AEP Water Ag CCME Water Ag |
| Boron | | Total | ug/L | 5000 | 5000 | | 5000 | AEP Water Ag CCME Water Ag |
| Bromacil | | | ug/L | 1100 | 1100 | | 1100 | AEP Water Ag CCME Water Ag |
| Bromodichloromethane | | | ug/L | 100 | | | 100 | AEP Water Ag |
| Bromoform | | | ug/L | 100 | | | 100 | AEP Water Ag |
| Bromoxynil | | | ug/L | 11 | 11 | | 11 | AEP Water Ag CCME Water Ag |
| Cadmium | | Total | ug/L | 80 | 80 | 0.2307 | 0.2307 | US DOE Wildlife |
| Calcium | | | $\mathrm{mg/L}$ | 1000 | 1000 | | 1000 | AEP Water Ag CCME Water Ag |
| Captan | | | ug/L | 13 | | | 13 | AEP Water Ag |
| Carbaryl | | | ug/L | 1100 | 110 | | 110 | CCME Water Ag |

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Table 3.3: Health risk criteria for the protection of wildlife species (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | CCME Water Ag | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|--|----------------------|--------------------|-------|--------------|------------------|-----------------|----------------------------------|-------------------------------|
| Carbofuran | | | ug/L | 45 | 45 | | 45 | AEP Water Ag CCME Water Ag |
| Carbon tetrachloride | | | ug/L | 5 | 5 | 913 | 5 | AEP Water Ag CCME Water Ag |
| Chlordane | | | ug/L | 7 | 7 | 0.00889 | 0.00889 | US DOE Wildlife |
| Chloroform | | | ug/L | 100 | 100 | 3439 | 100 | AEP Water Ag CCME Water Ag |
| Chlorophenoxy Herbicide (2,4,5-TP) [Silvex] | | | ug/L | 100 | 100 | | 100 | AEP Water Ag CCME Water Ag |
| Chlorothalonil | | | ug/L | 170 | 170 | | 170 | AEP Water Ag CCME Water Ag |
| Chlorpyrifos | | | ug/L | 24 | 24 | | 24 | AEP Water Ag CCME Water Ag |
| Chromium (III) | | Total | ug/L | 50 | 50 | | 50 | AEP Water Ag CCME Water Ag |
| Chromium (VI) | | Total | ug/L | 50 | 50 | 3593 | 50 | AEP Water Ag CCME Water Ag |
| Cobalt | | Total | ug/L | 1000 | 1000 | | 1000 | AEP Water Ag CCME Water Ag |
| Copper | | Total | ug/L | 500 | 500 | | 500 | AEP Water Ag CCME Water Ag |
| Cyanazine | | | ug/L | 10 | 10 | | 10 | AEP Water Ag CCME Water Ag |
| Cyanide | as free CN | | ug/L | | | 369092 | 369092 | US DOE Wildlife |
| DDT and metabolites | | | ug/L | 30 | | 4.136e-06 | 4.136e-06 | US DOE Wildlife |
| Deltamethrin | | | ug/L | 2.5 | 2.5 | | 2.5 | AEP Water Ag CCME Water Ag |
| Di-n-Butyl Phthalate | | | ug/L | | | 0.15 | 0.15 | US DOE Wildlife |

Table 3.3: Health risk criteria for the protection of wildlife species (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | $\begin{array}{c} \text{CCME Water} \\ \text{Ag} \end{array}$ | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|--------------------------------|----------------------|--------------------|-----------------|--------------|---|-----------------|----------------------------------|-------------------------------|
| Dibromochloromethane | | | ug/L | 100 | 100 | | 100 | AEP Water Ag CCME Water Ag |
| Dicamba | | | $\mathrm{ug/L}$ | 122 | 122 | | 122 | AEP Water Ag CCME Water Ag |
| Dichlorobromomethane | | | ug/L | | 100 | | 100 | CCME Water Ag |
| Dichloromethane | | | ug/L | 50 | 50 | | 50 | AEP Water Ag CCME Water Ag |
| Diclofop-methyl | | | ug/L | 9 | 9 | | 9 | AEP Water Ag CCME Water Ag |
| Dieldrin | | | $\mathrm{ug/L}$ | | | 0.001362 | 0.001362 | US DOE Wildlife |
| Diethyl Phthalate | | | ug/L | | | 210561 | 210561 | US DOE Wildlife |
| Dimethoate | | | ug/L | 3 | 3 | | 3 | AEP Water Ag CCME Water Ag |
| Dinoseb | | | ug/L | 150 | 150 | | 150 | AEP Water Ag CCME Water Ag |
| Dioxin $(2,3,7,8\text{-TCDD})$ | | | ug/L | | | 2.13e-08 | 2.134e-08 | US DOE Wildlife |
| Endosulfan | | | ug/L | | | 1 | 1 | US DOE Wildlife |
| Endrin | | | $\mathrm{ug/L}$ | 0.2 | 0.2 | 0.001313 | 0.001313 | US DOE Wildlife |
| Ethanol | | | ug/L | | | 123377 | 123377 | US DOE Wildlife |
| Ethyl acetate | | | ug/L | | | 136465 | 136465 | US DOE Wildlife |
| Ethylbenzene | | | ug/L | 2.4 | 2.4 | | 2.4 | AEP Water Ag CCME Water Ag |
| Fluoride | | | $\mathrm{mg/L}$ | 1 | 1 | | 1 | AEP Water Ag CCME Water Ag |
| Formaldehyde | | | ug/L | | | 73910 | 73910 | US DOE Wildlife |
| Glyphosate | | | $\mathrm{ug/L}$ | 280 | 280 | | 280 | AEP Water Ag CCME Water Ag |
| Heptachlor | | | ug/L | 3 | 3 | 0.001083 | 0.001083 | US DOE Wildlife |
| | | | | | | | | |

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Table 3.3: Health risk criteria for the protection of wildlife species (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | $\begin{array}{c} \text{CCME Water} \\ \text{Ag} \end{array}$ | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|--|----------------------|--------------------|-----------------|--------------|---|-----------------|----------------------------------|-------------------------------|
| Hexachlorobenzene | | | ug/L | 0.52 | 0.52 | | 0.52 | AEP Water Ag CCME Water Ag |
| Inorganic nitrogen (nitrate and nitrite) | as N | dissolved | $\mathrm{mg/L}$ | 100 | 100 | | 100 | AEP Water Ag CCME Water Ag |
| Lead | | Total | ug/L | 100 | 100 | 168 | 100 | AEP Water Ag CCME Water Ag |
| MCPA | | | ug/L | 25 | 25 | | 25 | AEP Water Ag CCME Water Ag |
| Mercury | | Total | ug/L | 3 | 3 | 0.001576 | 0.001576 | US DOE Wildlife |
| Methanol | | | ug/L | | | 230691 | 230691 | US DOE Wildlife |
| Methoxychlor | | | ug/L | | | 1 | 1 | US DOE Wildlife |
| Methylene chloride | | | ug/L | | | 3990 | 3990 | US DOE Wildlife |
| Metolachlor | | | ug/L | 50 | 50 | | 50 | AEP Water Ag CCME Water Ag |
| Metribuzin | | | ug/L | 80 | 80 | | 80 | AEP Water Ag CCME Water Ag |
| Molybdenum | | Total | ug/L | 500 | 500 | | 500 | AEP Water Ag CCME Water Ag |
| Nickel | | Total | ug/L | 1000 | 1000 | 1438 | 1000 | AEP Water Ag CCME Water Ag |
| Nitrite | as N | dissolved | $\mathrm{mg/L}$ | 10 | 10 | | 10 | AEP Water Ag CCME Water Ag |
| Pentachloronitrobenzene | | | ug/L | | | 4 | 4 | US DOE Wildlife |
| Pentachlorophenol | | | ug/L | | | 0.275 | 0.275 | US DOE Wildlife |
| Phenol | | | ug/L | 2 | 2 | | 2 | AEP Water Ag CCME Water Ag |
| Picloram | | | ug/L | 190 | 190 | | 190 | AEP Water Ag CCME Water Ag |

Table 3.3: Health risk criteria for the protection of wildlife species (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | CCME Water Ag | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|------------------------|----------------------|--------------------|-----------------|--------------|------------------|-----------------|----------------------------------|-------------------------------|
| Selenium | | Total | ug/L | 50 | 50 | 0.2363 | 0.2363 | US DOE Wildlife |
| Simazine | | | ug/L | 10 | 10 | | 10 | AEP Water Ag CCME Water Ag |
| Sulfate | as SO4 | | mg/L | 1000 | 1000 | | 1000 | AEP Water Ag CCME Water Ag |
| Tebuthiuron | | | $\mathrm{ug/L}$ | 130 | 130 | | 130 | AEP Water Ag CCME Water Ag |
| Tetrachloroethylene | | | ug/L | | | 48 | 48 | US DOE Wildlife |
| Thallium | | Total | ug/L | | | 1 | 1 | US DOE Wildlife |
| Toluene | | | ug/L | 24 | 24 | 764 | 24 | AEP Water Ag CCME Water Ag |
| Total dissolved solids | | | $\mathrm{mg/L}$ | 3000 | 3000 | | 3000 | AEP Water Ag CCME Water Ag |
| Toxaphene | | | ug/L | 5 | 5 | 1 | 1 | US DOE Wildlife |
| Triallate | | | ug/L | 230 | 230 | | 230 | AEP Water Ag CCME Water Ag |
| Tribromomethane | | | ug/L | | 100 | | 100 | CCME Water Ag |
| Tributyltin | | | $\mathrm{ug/L}$ | 250 | 250 | | 250 | AEP Water Ag CCME Water Ag |
| Trichloroethylene | | | ug/L | 50 | 50 | 49419 | 22 | US DOE Wildlife |
| Tricyclohexyltin | | | ug/L | 250 | 250 | | 250 | AEP Water Ag CCME Water Ag |
| Trifluralin | | | ug/L | 45 | 45 | | 45 | AEP Water Ag CCME Water Ag |
| Triphenyltin | | | $\mathrm{ug/L}$ | 820 | 820 | | 820 | AEP Water Ag CCME Water Ag |
| Uranium | | Total | $\mathrm{ug/L}$ | 200 | 200 | | 200 | AEP Water Ag CCME Water Ag |

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Table 3.3: Health risk criteria for the protection of wildlife species (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP Water Ag | $\begin{array}{c} \text{CCME Water} \\ \text{Ag} \end{array}$ | US DOE Wildlife | Wildlife Health Risk Criteria | Source |
|---------------------------------------|----------------------|--------------------|-----------------|--------------|---|-----------------|----------------------------------|-------------------------------|
| Vanadium | | Total | $\mathrm{ug/L}$ | 100 | 100 | | 100 | AEP Water Ag CCME Water Ag |
| Vinyl chloride | | | ug/L | | | 78 | 78 | US DOE Wildlife |
| Xylene | | | ug/L | | | 28 | 28 | US DOE Wildlife |
| Zinc | | Total | ug/L | 50 | 50000 | 30 | 30 | US DOE Wildlife |
| gamma-Hexachlorocyclohexane [Lindane] | | | ug/L | 4 | | 9 | 4 | AEP Water Ag |
| Note: AG: Agriculture | | | | | | | | |

Aquatic Ecosystem Health

Indigenous communities identified the health of ecosystems as an indicator of their physical and mental health. Indicators of ecosystem health were identified as the presence and abundance of each of the following groups: invertebrates, fish, amphibians, plants, algae, and wildlife species (birds and mammals).

To evaluate which aquatic biota were considered in development of the CCME PALs (and the majority of GOA 2018 PALs) and understand the level of protection for various aquatic biota within an ecosystem, the technical information sheets for each substance were reviewed. Table 3.4 describes available toxicity data and relative sensitivity for fish, amphibian, invertebrate, plant, and algae species (1 = most sensitive, 4 = least sensitive).

The CCME PALs most frequently included toxicity test species from fish (90%) and invertebrates (76%) classes and less frequently included toxicity data from algae (49%), plant (41%), amphibian (31) species in development of PALs.

Sensitivity is indicated by the number of times (count) a class of species was the most sensitive from exposure to a specific contaminant in comparison to the other species with available toxicity data. If two classes showed similar sensitivity, they were not included in the count (see example for benzene where neither fish nor amphibian were counted). Comparatively, invertebrates were the most sensitive to chemical exposures followed by fish and then primary producers (plants and algae).

Table 3.4: Availability and sensitivity of fish, amphibian, invertebrate, plant and algae species in toxicity data used to derive CCME PAL guidelines (1 = most sensitive, 4 = least sensitive).

| | | | Sensitivity rank* | | |
|-------------------------|-----------------|----------------------|--------------------------|-------------------|------------------|
| Parameter $(n = 29)$ | Fish $(n = 26)$ | Amphibians $(n = 9)$ | Invertebrates $(n = 22)$ | Plants $(n = 12)$ | Algae $(n = 14)$ |
| Acenaphthene | 1 | | | | 2 |
| Ammonia, unionized | 1 | | 2 | 3 | |
| Anthracene | 2 | | 1 | | 3 |
| Benz(a)anthracene | 2 | | | | 1 |
| Benz(a)pyrene | 1 | | | | 2 |
| Benzene | 1 | 1 | | | |
| Boron | 2 | 4 | 3 | 1 | |
| Cadmium | 2 | 4 | 1 | 3 | 3 |
| Chloride | 2 | 3 | 1 | 4 | 4 |
| Chromium, hexavalent | 3 | | 1 | 2 | |

Table 3.4: Availability and sensitivity of fish, amphibian, invertebrate, plant and algae species in toxicity data used to derive CCME PAL guidelines (1 = most sensitive, 4 = least sensitive). (continued)

| | | | Sensitivity rank* | | |
|---|---|----------------------|--------------------------|-------------------|-----------------------|
| Parameter $(n = 29)$ | $\begin{aligned} & \text{Fish} \\ & (n = 26) \end{aligned}$ | Amphibians $(n = 9)$ | Invertebrates $(n = 22)$ | Plants $(n = 12)$ | $ Algae \\ (n = 14) $ |
| Chromium, trivalent | 1 | | 3 | 2 | |
| Ethylbenzene | | | 1 | | 2 |
| Fluoranthene | | | | | |
| Fluorene | | | 1 | | 2 |
| Fluoride | 1 | | 1 | | |
| Manganese | 1 | 3 | 2 | | |
| Mercury | 1 | | 2 | 2 | |
| Molybdenum | 1 | | 3 | | 2 |
| Naphthalene | | | | | |
| Nitrate | 1 | 2 | 3 | | |
| Phenanthrene | 1 | | 1 | | |
| Phenol | 1 | 1 | | 2 | |
| Pyrene | 3 | 3 | 1 | | 2 |
| Silver | 3 | | 1 | | 2 |
| Thallium | 2 | | 3 | 1 | |
| Toluene | 1 | | 2 | | |
| Ammonia (un-ionized) | 1 | | 1 | 1 | |
| Uranium | 3 | | 1 | 2 | 1 |
| Zinc | 2 | 3 | 2 | 1 | 1 |
| Most sensitive class (frequency) * 1 = most sensitive, 4 : | 35% = least sensitiv | - 70 | 42% | 27% | 23% |

Protection of aquatic life guidelines were not available for acrylamide, PHC F1 and F2, naphthenic acids, antimony, barium, lithium, silver, strontium, benzo(a)anthracene, benzo(b)fluoranthene, benzo(k)fluoranthene, chrysene, dibenzo(a,h)anthracene, indeno(1,2,3-cd)pyrene.

The protocol for derivation of surface water quality for the protection of aquatic life is complex and beyond the scope of this project. Recognizing this limitation, health risk criteria for the protection of aquatic ecosystems are proposed in Table 3.5.

While new criteria were not derived guidance is provided on assessment of complex mixtures which may be acting through similar modes of action to illicit toxicological responses (high

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and low MW PAH groups) and overall toxicity (as toxic units). 1942 To assess potential toxicity, results from whole effluent toxicity tests (WET) must be used 1943 and predicted toxicity from water quality modelling is not recommended as toxicity is not a 1944 "conserved substance". If the practitioner is attempting to predict toxicity in ambient environ-1945 ments complex models such as the Biotic Ligand Models (BLMs) for metals or Quantitative 1946 Structure Activity Relationships (QSARs) for organics are required 1947 The health risk criteria presented in Table 3.5 apply to the assessment of aquatic ecosystem 1948 health only and risks to aquatic species may be less than those associated with toxicological 1949 responses in more sensitive receptors (i.e., humans, wildlife species) and other water uses. 1950 1951 As discussed in Section 3.3.3, the US EPA prescribes aquatic life criteria for dissolved fractions which were developed by applying CFs to total recoverable metal concentrations used for 1952 toxicity testing. Comparison of the CFs estimated from laboratory conditions during toxicity 1953 1954 tests differ from conditions in the Athabasca River, therefore the health risk criteria were developed by adopting published guidelines for total recoverable fractions, until site specific CFs 1955 can be developed for the Lower Athabasca River. 1956 1957 1958

However, to better understand the condition of the LAR and potential health risks, the US EPA aquatic life criteria for dissolved metals may be applied, in addition to the health risk criteria for total fractions, when dissolved monitoring data is available. Comparison of trace element monitoring data must be presented for total health risk criteria. If the US EPA aquatic life criteria (dissolved) identified in Table 3.10 are applied to monitoring data, they must be presented alongside comparison with total health risk criteria.

The health risk criteria for aquatic health should not be applied singularly unless all other exposure pathways described in Figure 3.1. have been assessed and identified as not applicable or non-operational (i.e., the surface water being assessed is not used by humans or wildlife).

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Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion).

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|--------------------------------------|----------------------|--------------------|-----------------|--------|--------|------|--------|---|---------------------------------|
| .alphaEndosulfan | | | ug/L | | | | 0.06 | 0.056 | US EPA Aquatic Life Criteria |
| .betaEndosulfan | | | ug/L | | | | 0.06 | 0.056 | US EPA Aquatic Life Criteria |
| 1,1,2-Trichloroethane | | | ug/L | | 21.00 | | | 21 | CCME Water PAL |
| 1,2,3,4- Tetrachlorobenzene | | | ug/L | 1.80 | 1.80 | | | 1.8 | AEP Water PAL CCME Water PAL |
| 1,2,3-Trichlorobenzene | | | ug/L | 8.00 | 8.00 | | | 8 | AEP Water PAL CCME Water PAL |
| 1,2,4-Trichlorobenzene | | | ug/L | 24.00 | 24.00 | | | 24 | AEP Water PAL CCME Water PAL |
| 1,2-Dichlorobenzene | | | ug/L | 0.70 | | | | 0.7 | AEP Water PAL |
| 1,2-Dichloroethane | | | ug/L | 100.00 | 100.00 | | | 100 | AEP Water PAL CCME Water PAL |
| 1,3-Dichlorobenzene | | | ug/L | 150.00 | | | | 150 | AEP Water PAL |
| 1,4-Dichlorobenzene | | | ug/L | 26.00 | | | | 26 | AEP Water PAL |
| 2,4-D | | | ug/L | 4.00 | 4.00 | | | 4 | AEP Water PAL CCME Water PAL |
| 2,4-DB | | | ug/L | 25.00 | | | | 25 | AEP Water PAL |
| 3-Iodo-2-propynyl butyl carbamate | | | ug/L | 1.90 | 1.90 | | | 1.9 | AEP Water PAL CCME Water PAL |
| $Ac enaph the ne^{\dagger}$ | | | $\mathrm{ug/L}$ | 5.80 | 5.80 | | | 5.8 | AEP Water PAL CCME Water PAL |
| Acridine | | | ug/L | 4.40 | 4.40 | | | 4.4 | AEP Water PAL CCME Water PAL |
| | | | | | | | | | |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). (continued)

| Criteria | |
|--|-------------------------|
| Aldrin Aldrin ug/L 0.00 0.00 Alkalinity, total as CaCO3 mg/L Aluminum Total ug/L 100.00 100.00 200.00 AEP Ware Come of the company of the comp | ter PAL Aquatic Life |
| Aldrin | Vater PAL |
| Alkalinity, total as CaCO3 mg/L 20.00 20.00 20 AEP Warden | ter PAL Vater PAL |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | ter PAL Vater PAL |
| Aluminum dissolved ug/L 50.00 50 AEP Water Ammonia mg/L 0.79 0.794 AEP Water Ammonia, unionized mg/L 0.02 0.02 0.016 AEP Water Aniline ug/L 2.20 2.20 2.2 AEP Water | ter PAL Aquatic Life |
| Ammonia mg/L 0.79 0.794 AEP Wa Ammonia, unionized mg/L 0.02 0.02 0.016 AEP Wa Aniline ug/L 2.20 2.20 2.2 AEP Wa | Vater PAL |
| Ammonia, unionized mg/L 0.02 0.02 0.016 AEP Warden Aniline ug/L 2.20 2.20 2.2 AEP Warden AEP Warden Ammonia, unionized 0.016 AEP Ward | ter PAL |
| Aniline ug/L 2.20 2.20 2.2 AEP Wa | ter PAL |
| | ter PAL |
| CCME | ter PAL Vater PAL |
| | ter PAL Vater PAL |
| | ter PAL Vater PAL |
| Arsenic dissolved ug/L 150.00 150 US EPA Criteria | Aquatic Life |
| | ter PAL Vater PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|---------------------------------|----------------------|--------------------|-----------------|----------|----------|------|--------|---|--|
| Azinphos-methyl | | | ug/L | 0.01 | | | 0.01 | 0.01 | AEP Water PAL US EPA Aquatic Life Criteria |
| Benzene | | | ug/L | 40.00 | 370.00 | | | 40 | AEP Water PAL |
| $Benzo(a)anthracene^{\ddagger}$ | | | $\mathrm{ug/L}$ | 0.02 | 0.02 | | | 0.018 | AEP Water PAL CCME Water PAL |
| Benzo(a)pyrene [‡] | | | ug/L | 0.01 | 0.01 | | | 0.015 | AEP Water PAL CCME Water PAL |
| Bisphenol A-d6 | | | ug/L | | | 3.50 | | 3.5 | FEQG Water PAL |
| Boron | | Total | ug/L | 1,500.00 | 1,500.00 | | | 1500 | AEP Water PAL CCME Water PAL |
| Bromacil | | | ug/L | 5.00 | 5.00 | | | 5 | AEP Water PAL CCME Water PAL |
| Bromoxynil | | | ug/L | 5.00 | 5.00 | | | 5 | AEP Water PAL CCME Water PAL |
| Cadmium* | | Total | $\mathrm{ug/L}$ | 0.18 | 0.18 | | | 0.1843828121 | AEP Water PAL CCME Water PAL |
| Cadmium* | | dissolved | ug/L | | | | 0.82 | 0.8237781279 | US EPA Aquatic Life Criteria |
| Captan | | | $\mathrm{ug/L}$ | 1.30 | 1.30 | | | 1.3 | AEP Water PAL CCME Water PAL |
| Carbamazepine | | | ug/L | 10.00 | 10.00 | | | 10 | AEP Water PAL CCME Water PAL |
| Carbaryl | | | ug/L | 0.20 | 0.20 | | 2.10 | 0.2 | AEP Water PAL CCME Water PAL |
| Carbofuran | | | ug/L | 1.80 | 1.80 | | | 1.8 | AEP Water PAL CCME Water PAL |
| | | | | | | | | | |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|--|----------------------|--------------------|-----------------|--------|--------|------|--------|---|---------------------------------|
| Carbon tetrachloride | | | ug/L | 13.30 | 13.30 | | | 13.3 | AEP Water PAL CCME Water PAL |
| Chloramines | | | ug/L | | 0.50 | | | 0.5 | CCME Water PAL |
| Chlordane | | | ug/L | 0.01 | 0.01 | | 0.00 | 0.0043 | US EPA Aquatic Life Criteria |
| Chloride | | | $\mathrm{mg/L}$ | 120.00 | 120.00 | | 230.00 | 120 | AEP Water PAL CCME Water PAL |
| Chlorinated paraffins, long-chain, C18-C20 | | | ug/L | 2.40 | | 2.40 | | 2.4 | AEP Water PAL FEQG Water PAL |
| Chlorinated paraffins, medium-chain, C14-C17 | | | $\mathrm{ug/L}$ | 2.40 | | 2.40 | | 2.4 | AEP Water PAL FEQG Water PAL |
| Chlorinated paraffins, short-chain, C10-C13 | | | ug/L | 2.40 | | 2.40 | | 2.4 | AEP Water PAL FEQG Water PAL |
| Chlorine | | | ug/L | 0.50 | | | 11.00 | 0.5 | AEP Water PAL |
| Chlorobenzene | | | ug/L | 1.30 | | | | 1.3 | AEP Water PAL |
| Chloroform | | | ug/L | 1.80 | 1.80 | | | 1.8 | AEP Water PAL CCME Water PAL |
| Chlorophenol | | | ug/L | 7.00 | 7.00 | | | 7 | AEP Water PAL CCME Water PAL |
| Chlorothalonil | | | ug/L | 0.18 | 0.18 | | | 0.18 | AEP Water PAL CCME Water PAL |
| Chlorpyrifos | | | ug/L | 0.00 | 0.00 | | 0.04 | 0.002 | AEP Water PAL CCME Water PAL |
| Chromium (III)* | | Total | ug/L | 8.90 | 8.90 | | | 8.9 | AEP Water PAL CCME Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|-------------------------------|----------------------|--------------------|-----------------|-------|-------|------|--------|---|--|
| Chromium (III)*§ | | dissolved | ug/L | | | | 100.92 | 100.9185723 | US EPA Aquatic Life Criteria |
| Chromium (VI) | | Total | ug/L | 1.00 | 1.00 | | | 1 | AEP Water PAL CCME Water PAL |
| Chromium (VI) | | dissolved | ug/L | | | 5.00 | 11.00 | 5 | FEQG Water PAL |
| Cobalt* | | Total | ug/L | 1.10 | | 1.10 | | 1.099682588 | AEP Water PAL FEQG Water PAL |
| Copper* | | Total | ug/L | 7.00 | 2.76 | | | 2.763433095 | CCME Water PAL |
| Copper | | dissolved | ug/L | | | 0.53 | | 0.53 | FEQG Water PAL |
| Cyanazine | | | $\mathrm{ug/L}$ | 2.00 | 2.00 | | | 2 | AEP Water PAL CCME Water PAL |
| Cyanide | as free CN | | ug/L | 5.20 | 5.00 | | 5.20 | 5 | CCME Water PAL |
| DDT and metabolites | | | ug/L | 0.00 | 0.00 | | 0.00 | 0.001 | AEP Water PAL CCME Water PAL US EPA Aquatic Life Criteria |
| Deltamethrin | | | ug/L | 0.00 | 0.00 | | | 0.0004 | AEP Water PAL CCME Water PAL |
| Demeton | | | ug/L | 0.10 | | | 0.10 | 0.1 | AEP Water PAL US EPA Aquatic Life Criteria |
| Di(2-ethylhexyl) phthalate | | | $\mathrm{ug/L}$ | 16.00 | 16.00 | | | 16 | AEP Water PAL CCME Water PAL |
| Di-n-Butyl Phthalate | | | ug/L | 19.00 | 19.00 | | | 19 | AEP Water PAL CCME Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|------------------------------------|----------------------|--------------------|-----------------|------------|----------|------|--------|---|--|
| Diazinon | | | ug/L | 0.17 | | | 0.17 | 0.17 | AEP Water PAL US EPA Aquatic Life Criteria |
| Dicamba | | | ug/L | 10.00 | 10.00 | | | 10 | AEP Water PAL CCME Water PAL |
| Dichlorophenol | | | $\mathrm{ug/L}$ | 0.20 | 0.20 | | | 0.2 | AEP Water PAL CCME Water PAL |
| Diclofop-methyl | | | ug/L | 6.10 | 6.10 | | | 6.1 | AEP Water PAL CCME Water PAL |
| Didecyl dimethyl ammonium chloride | | | ug/L | 1.50 | 1.50 | | | 1.5 | AEP Water PAL CCME Water PAL |
| Dieldrin | | | ug/L | 0.00 | 0.00 | | 0.06 | 0.004 | AEP Water PAL CCME Water PAL |
| Diethanolamine | | | ug/L | 450.00 | | | | 450 | AEP Water PAL |
| Diethylene glycol | | | ug/L | 150,000.00 | | | | 150000 | AEP Water PAL |
| Diisopropanolamine | | | $\mathrm{ug/L}$ | 1,600.00 | 1,600.00 | | | 1600 | AEP Water PAL CCME Water PAL |
| Dimethoate | | | ug/L | 6.20 | 6.20 | | | 6.2 | AEP Water PAL CCME Water PAL |
| Dinoseb | | | ug/L | 0.05 | 0.05 | | | 0.05 | AEP Water PAL CCME Water PAL |
| Endosulfan | | | ug/L | 0.00 | 0.00 | | | 0.003 | AEP Water PAL CCME Water PAL |
| Endrin | | | ug/L | 0.00 | 0.00 | | 0.04 | 0.0023 | AEP Water PAL CCME Water PAL |
| Ethinyl estradiol | | | ng/L | 0.50 | | | | 0.5 | AEP Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|-----------------------------|----------------------|--------------------|-----------------|------------|------------|------|--------|---|---------------------------------|
| Ethylbenzene | | | $\mathrm{ug/L}$ | 90.00 | 90.00 | | | 90 | AEP Water PAL CCME Water PAL |
| Ethylene glycol | | | $\mathrm{ug/L}$ | 192,000.00 | 192,000.00 | | | 192000 | AEP Water PAL CCME Water PAL |
| Fluoranthene [†] | | | $\mathrm{ug/L}$ | 0.04 | 0.04 | | | 0.04 | AEP Water PAL CCME Water PAL |
| Fluorene [†] | | | $\mathrm{ug/L}$ | 3.00 | 3.00 | | | 3 | AEP Water PAL CCME Water PAL |
| Fluoride | | | $\mathrm{mg/L}$ | | 0.12 | | | 0.12 | CCME Water PAL |
| Glyphosate | | | $\mathrm{ug/L}$ | 800.00 | 800.00 | | | 800 | AEP Water PAL CCME Water PAL |
| Heptachlor | | | $\mathrm{ug/L}$ | | 0.01 | | 0.00 | 0.0038 | US EPA Aquatic Life Criteria |
| Heptachlor epoxide | | | $\mathrm{ug/L}$ | 0.01 | | | 0.00 | 0.0038 | US EPA Aquatic Life Criteria |
| Hexabromocyclodode- cane | | | $\mathrm{ug/L}$ | 0.56 | | 0.56 | | 0.56 | AEP Water PAL FEQG Water PAL |
| Hexachlorobutadiene | | | $\mathrm{ug/L}$ | 1.30 | 1.30 | | | 1.3 | AEP Water PAL CCME Water PAL |
| Hexachlorocyclohex- ane | | | ug/L | | 0.01 | | | 0.01 | CCME Water PAL |
| Hydrazine | | | $\mathrm{ug/L}$ | 2.60 | | 2.60 | | 2.6 | AEP Water PAL FEQG Water PAL |
| Hydrogen Sulfide | | | $\mathrm{ug/L}$ | | | | 2.00 | 2 | US EPA Aquatic Life Criteria |
| Imidacloprid | | | $\mathrm{ug/L}$ | 0.23 | 0.23 | | | 0.23 | AEP Water PAL CCME Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|-------------------------|----------------------|--------------------|-----------------|----------|--------|----------|----------|---|--|
| Iron | | Total | ug/L | | 300.00 | 4,206.07 | | 300 | CCME Water PAL |
| Iron | | dissolved | ug/L | 300.00 | | | 1,000.00 | 300 | AEP Water PAL |
| Lead* | | Total | $\mathrm{ug/L}$ | 4.01 | 4.01 | | | 4.01275079 | AEP Water PAL CCME Water PAL |
| Lead^* | | dissolved | $\mathrm{ug/L}$ | | | | 3.07 | 3.067487163 | US EPA Aquatic Life Criteria |
| Linuron | | | ug/L | 7.00 | 7.00 | | | 7 | AEP Water PAL CCME Water PAL |
| MCPA | | | $\mathrm{ug/L}$ | 2.60 | 2.60 | | | 2.6 | AEP Water PAL CCME Water PAL |
| Malathion | | | ug/L | 0.10 | | | 0.10 | 0.1 | AEP Water PAL US EPA Aquatic Life Criteria |
| Manganese | | Total | ug/L | | 470.00 | | | 470 | CCME Water PAL |
| Mecoprop | | | ug/L | 13.00 | | | | 13 | AEP Water PAL |
| Mercury (methyl) | | Total | ug/L | 0.00 | | | | 0.001 | AEP Water PAL |
| Mercury (methyl) | | dissolved | ug/L | | 0.00 | | | 0.004 | CCME Water PAL |
| Mercury | | Total | ug/L | 0.00 | 0.03 | | | 0.005 | AEP Water PAL |
| Mercury [§] | | dissolved | ug/L | | | | 0.77 | 0.77 | US EPA Aquatic Life Criteria |
| Methanol | | | ug/L | 1,500.00 | | | | 1500 | AEP Water PAL |
| Methoprene | | | $\mathrm{ug/L}$ | 0.09 | 0.09 | | | 0.09 | AEP Water PAL CCME Water PAL |
| Methoxychlor | | | ug/L | 0.03 | | | 0.03 | 0.03 | AEP Water PAL US EPA Aquatic Life Criteria |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|--------------------------|----------------------|--------------------|-----------------|-------|-----------|------|--------|---|--|
| Methyl tert-butyl ether | | | ug/L | 10.00 | 10,000.00 | | | 10 | AEP Water PAL |
| Methylene chloride | | | ug/L | 98.10 | 98.10 | | | 98.1 | AEP Water PAL CCME Water PAL |
| Metolachlor | | | ug/L | 7.80 | 7.80 | | | 7.8 | AEP Water PAL CCME Water PAL |
| Metribuzin | | | ug/L | 1.00 | 1.00 | | | 1 | AEP Water PAL CCME Water PAL |
| Mirex | | | ug/L | 0.00 | | | 0.00 | 0.001 | AEP Water PAL US EPA Aquatic Life Criteria |
| Molybdenum | | Total | ug/L | 73.00 | 73.00 | | | 73 | AEP Water PAL CCME Water PAL |
| Monochlorobenzene | | | ug/L | 1.30 | 1.30 | | | 1.3 | AEP Water PAL CCME Water PAL |
| Monoethanolamine | | | ug/L | 75.00 | | | | 75 | AEP Water PAL |
| Naphthalene [†] | | | ug/L | 1.00 | 1.10 | | | 1 | AEP Water PAL |
| Nickel* | | Total | ug/L | 60.86 | 109.78 | | | 60.86254826 | AEP Water PAL |
| Nickel*§ | | dissolved | ug/L | | | | 60.68 | 60.67996061 | US EPA Aquatic Life Criteria |
| Nitrate | as N | dissolved | $\mathrm{mg/L}$ | 3.00 | 3.00 | | | 3 | AEP Water PAL CCME Water PAL |
| Nitrite | as N | dissolved | $\mathrm{mg/L}$ | 0.20 | 0.06 | | | 0.06 | CCME Water PAL |
| Nonylphenol | | | ug/L | | | | 6.60 | 6.6 | US EPA Aquatic Life Criteria |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|-------------------------------------|----------------------|--------------------|-----------------|------------|------------|------|--------|---|--|
| Nonylphenol and its ethoxylates | | | $\mathrm{ug/L}$ | 6.60 | 1.00 | | | 1 | CCME Water PAL |
| Parathion | | | $\mathrm{ug/L}$ | 0.01 | | | 0.01 | 0.013 | AEP Water PAL US EPA Aquatic Life Criteria |
| Pentachlorobenzene | | | $\mathrm{ug/L}$ | 6.00 | 6.00 | | | 6 | AEP Water PAL CCME Water PAL |
| Pentachlorophenol | | | $\mathrm{ug/L}$ | 0.50 | 0.50 | | 15.00 | 0.5 | AEP Water PAL CCME Water PAL |
| Perfluorooctanesul- fonate | | | ug/L | | | 6.80 | | 6.8 | FEQG Water PAL |
| Permethrin | | | $\mathrm{ug/L}$ | 0.00 | 0.00 | | | 0.004 | AEP Water PAL CCME Water PAL |
| Phenanthrene [†] | | | $\mathrm{ug/L}$ | 0.40 | 0.40 | | | 0.4 | AEP Water PAL CCME Water PAL |
| Phenol | | | $\mathrm{ug/L}$ | 4.00 | 4.00 | | | 4 | AEP Water PAL CCME Water PAL |
| Picloram | | | $\mathrm{ug/L}$ | 29.00 | 29.00 | | | 29 | AEP Water PAL CCME Water PAL |
| Polychlorinated Biphenyls (PCBs) | | | $\mathrm{ug/L}$ | 0.00 | 0.00 | | 0.01 | 0.001 | AEP Water PAL CCME Water PAL |
| Propylene glycol | | | $\mathrm{ug/L}$ | 500,000.00 | 500,000.00 | | | 500000 | AEP Water PAL CCME Water PAL |
| Pyrene [†] | | | $\mathrm{ug/L}$ | 0.03 | 0.03 | | | 0.025 | AEP Water PAL CCME Water PAL |
| Quinoline | | | ug/L | 3.40 | 3.40 | | | 3.4 | AEP Water PAL CCME Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|--------------------------|----------------------|--------------------|-------------------------|----------|-----------|------|--------|---|---------------------------------|
| Selenium | | Total | ug/L | 2.00 | 1.00 | | | 1 | CCME Water PAL |
| Silver | | Total | ug/L | 0.25 | 0.25 | | | 0.25 | AEP Water PAL CCME Water PAL |
| Simazine | | | ug/L | 10.00 | 10.00 | | | 10 | AEP Water PAL CCME Water PAL |
| Styrene | | | ug/L | 72.00 | 72.00 | | | 72 | AEP Water PAL CCME Water PAL |
| Sulfate | as SO4 | | $\mathrm{mg/L}$ | 309.00 | | | | 309 | AEP Water PAL |
| Sulfide | | | $\mathrm{mg/L}$ | 0.00 | | | | 0.0019 | AEP Water PAL |
| Sulfolane | | | ug/L | 50.00 | 50,000.00 | | | 50 | AEP Water PAL |
| Tebuthiuron | | | ug/L | 1,600.00 | 1.60 | | | 1.6 | CCME Water PAL |
| Tetrabromobisphenol A | | | ug/L | 3.10 | | 3.10 | | 3.1 | AEP Water PAL FEQG Water PAL |
| Tetrachloroethane | | | ug/L | | 13.30 | | | 13.3 | CCME Water PAL |
| Tetrachloroethylene | | | ug/L | 110.00 | 110.00 | | | 110 | AEP Water PAL CCME Water PAL |
| Tetrachlorophenol | | | ug/L | 1.00 | 1.00 | | | 1 | AEP Water PAL CCME Water PAL |
| Thallium | | Total | ug/L | 0.80 | 0.80 | | | 0.8 | AEP Water PAL CCME Water PAL |
| Toluene | | | ug/L | 0.50 | 2.00 | | | 0.5 | AEP Water PAL |
| Toxaphene | | | ug/L | 0.01 | 0.01 | | 0.00 | 0.0002 | US EPA Aquatic Life Criteria |
| Toxicity (acute)¶ | | | Toxic Units (TUa) | 0.30 | | | | | AEP Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|----------------------|----------------------|--------------------|-------------------------|------------|-------|--------|--------|---|---------------------------------|
| Toxicity (chronic)** | | | Toxic Units (TUc) | 1.00 | | | | | AEP Water PAL |
| Triallate | | | ug/L | 0.24 | 0.24 | | | 0.24 | AEP Water PAL CCME Water PAL |
| Tributyltin | | | ug/L | 0.07 | 0.01 | | 0.07 | 0.008 | CCME Water PAL |
| Trichlorfon | | | $\mathrm{ug/L}$ | 0.01 | 0.01 | | | 0.009 | AEP Water PAL CCME Water PAL |
| Trichloroethylene | | | $\mathrm{ug/L}$ | 21.00 | 21.00 | | | 21 | AEP Water PAL CCME Water PAL |
| Trichlorophenol | | | $\mathrm{ug/L}$ | 18.00 | 18.00 | | | 18 | AEP Water PAL CCME Water PAL |
| Triclosan | | | ug/L | | | 0.47 | | 0.47 | FEQG Water PAL |
| Triethylene glycol | | | ug/L | 350,000.00 | | | | 350000 | AEP Water PAL |
| Trifluralin | | | $\mathrm{ug/L}$ | 0.20 | 0.20 | | | 0.2 | AEP Water PAL CCME Water PAL |
| Triphenyltin | | | ug/L | 0.02 | 0.02 | | | 0.022 | AEP Water PAL CCME Water PAL |
| Uranium | | Total | $\mathrm{ug/L}$ | 15.00 | 15.00 | | | 15 | AEP Water PAL CCME Water PAL |
| Vanadium | | Total | ug/L | | | 120.00 | | 120 | FEQG Water PAL |
| Xylene | | | ug/L | 30.00 | | | | 30 | AEP Water PAL |
| Zinc | | Total | ug/L | 30.00 | | | | 30 | AEP Water PAL |
| Zinc^* | | dissolved | ug/L | | 31.35 | | 137.87 | 31.34535401 | CCME Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|--|----------------------|--------------------|-----------------|--------|--------|--------|--------|---|--|
| gamma- Hexachlorocyclohexane [Lindane] | | | ug/L | 0.01 | | | | 0.01 | AEP Water PAL |
| heptaBDE | | | ng/L | 17.00 | | 14.00 | | 14 | FEQG Water PAL |
| hexaBDE | | | ng/L | 120.00 | | 120.00 | | 120 | AEP Water PAL FEQG Water PAL |
| m-Dichlorobenzene | | | ug/L | | 150.00 | | | 150 | CCME Water PAL |
| o-Dichlorobenzene | | | ug/L | 0.70 | 0.70 | | | 0.7 | AEP Water PAL CCME Water PAL |
| octaBDE | | | ng/L | 17.00 | | 14.00 | | 14 | FEQG Water PAL |
| p-Dichlorobenzene | | | ug/L | 26.00 | 26.00 | | | 26 | AEP Water PAL CCME Water PAL |
| рН | | | pH units | 9.00 | 9.00 | | 6.50 | 6.5-9 | AEP Water PAL CCME Water PAL US EPA Aquatic Life Criteria |
| pentaBDE (BDE-100) | | | ng/L | 0.20 | | 0.20 | | 0.2 | AEP Water PAL FEQG Water PAL |
| pentaBDE (BDE-99) | | | ng/L | 4.00 | | 4.00 | | 4 | AEP Water PAL FEQG Water PAL |
| pentaBDE | | | ng/L | 0.20 | | 0.20 | | 0.2 | AEP Water PAL FEQG Water PAL |
| tetraBDE | | | $\mathrm{ng/L}$ | 24.00 | | 24.00 | | 24 | AEP Water PAL FEQG Water PAL |

Table 3.5: Health risk criteria for the protection of aquatic ecosystem health (adopted from GoA (2018); CCME PAL guidelines, Federal Environmental quality Guidelines; US EPA Aquatic Life Criterion). (continued)

| Parameter | Method Speciation | Sample Fraction | Units | AEP | CCME | FEQG | US EPA | Aquatic Ecosystem Health Criteria value | Source |
|-----------|----------------------|--------------------|-------|-------|------|-------|--------|---|---------------------------------|
| triBDE | | | ng/L | 46.00 | | 46.00 | | 46 | AEP Water PAL FEQG Water PAL |

Note:

PAL: Protection of Aquatic Life

* Calculated using modifying factors presented in Table 3.1.

- † Naphthalene applied as surrogate to sum of low molecular weight PAH congeners (Anthracene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) and compare to Naphthalene health risk criteria (adopted as surrogate) (CCME (2010))
- [‡] BaP and equivalents applied as surrogate to sum of high molecular weight PAH congeners (Benzo(a)anthracene, Benzo(b)fluoranthene, Benzo(b)fluoranthene, Benzo(b)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene) should be used for comparison to identified health risk criteria (CCME (2010))
- § Comparison of water quality data must be presented for both dissolved and total fractions
- ¶ Toxic Unit-Acute (TUa) is the reciprocal of the effluent concentration (i.e., TUa = 100/LC50) that causes 50 percent of the organisms to die by the end of an acute toxicity test (US EPA (2000c))
- ** Toxic Unit-Chronic (TUc) is the reciprocal of the effluent concentration (e.g., TUc = 100/NOEC) that causes no observable effect (NOEC) on the test organisms by the end of a chronic toxicity test (US EPA (2000c)).

1966 3.4.5 Water Quality Criteria for the Protection of Indigenous Use 1967 (derived)

The following water use categories are specific to protection of human health. As such, the potential for carcinogenic effects from exposure to chemicals must be considered. Known human carcinogens are identified in each table presenting health risk criteria. For PAHs, a comparison to the BaP health risk criteria requires the practitioner to calculate the BaP equivalent concentration by applying the health Canada (2021) RPFs to measured concentrations of PAH congeners as follows:

1974 Equation (3.4)

1977

1980 1981

1982

1983 1984

1985

1986

1987

1988

1989

1990

1991

1992

1993

1994

$$BaP \ equivalent \ (ug/L) = \sum [PAH \ congener \times BaP \ RPF] \eqno(3.4)$$

Once estimated, the BaP equivalent concentrations should be compared to the risk criteria for BaP in both the traditional foods and surface water and traditional medicine tables.

Local Indigenous Community Food and Medicine Ingestion Rates

1978 Derived health risk criteria for the remaining two water use categories (traditional foods and 1979 drinking water and medicinal plants are described below.

Traditional food consumption surveys were used to identify ingestion rates of culturally important fish and plant species required to develop health risk criteria protective of ACFN, FMFN and MCFN members. Details of the survey methodology and results are provided in Chapter 5. Consumption rates (g/d) for fish and medicinal plants were estimated using methods described in Chan et al. (2016) by multiplying the frequency (servings per year) by serving size (g per serving) and normalizing over the year. The highest calculated ingestion rate for each of fish (as a surrogate for traditional foods) and medicinal plants was adopted to derive the respective health risk criteria.

Modifications were required to address differences in the assumed fish consumption rate (22 g/d) between for the general population that was used to develop the US EPA Ambient Water Quality Criteria for Human Health (US EPA, 2015c) and the fish consumption rates developed in this work for the community members from ACFN, FMFN and MCFN who are consumers of traditional foods as described below.

For each ingestion rate, the upper range (95th percentile) was selected as a representative estimate of the higher range of exposure for members as compared to the 95th percentile upper confidence limit of the mean, which is commonly adopted in risk assessment. This decision

was guided by members from each of the three participating communities. The 95th percentile represents a higher estimate therefore a calorie check was undertaken. The fish consumption rate results in a 1400 kcal/day contribution, as compared to a reference adult value of 2800 kcal/day total, so was deemed possible and appropriate. For reference each of the upper range and mean values are presented in the figures below.

The US EPA HH AWQC for drinking water and fish consumption would protect community members consuming average quantities of fish (up to 22 g/d). However, the community survey data indicates that ACFN, MCFN and FMFN members consume greater quantities of fish than considered in the HH AWQCs. Based on the survey results, community 1 had the highest fish ingestion rate of 0.388 kg/day (Figure 3.2) and this value was adopted to calculate the health risk criteria for fish and water ingestion using Equation (3.2)

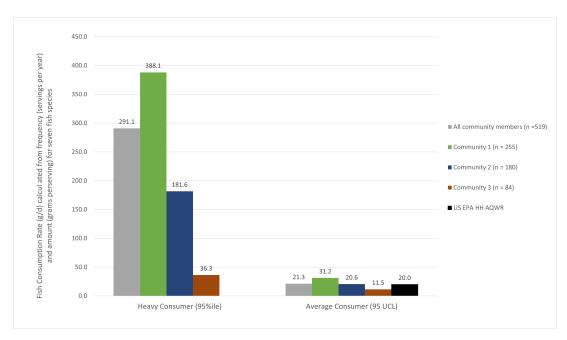


Figure 3.2: Comparison of pooled and individual Indigenous community member plant consumption rates (kg/d) calculated from survey responses for seven traditionally consumed fish species.

Plant Consumption Rates were estimated from the community survey data for wild mint and rat root species. The survey data indicates that rat root consumption (Figure 3.4) was greater than wild mint (Figure 3.3). The rat root consumption rate estimated from the pooled community data (0.0068 kg/d) was adopted as the plant consumption rate in Equation 2 to calculate the medicinal plant health risk criteria which is considered protective of members ingesting either mint or rat root.

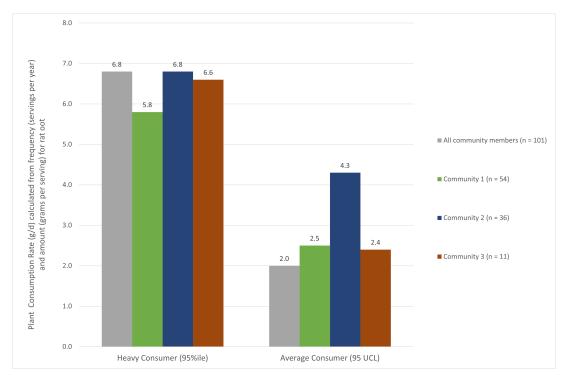


Figure 3.3: Comparison of pooled and individual Indigenous community member plant consumption rates (kg/d) calculated from survey responses for rat root.

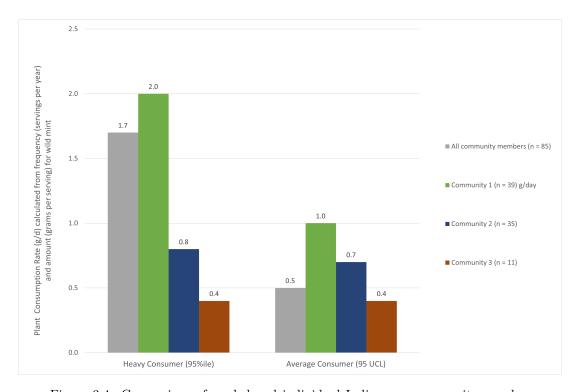


Figure 3.4: Comparison of pooled and individual Indigenous community member plant consumption rates (g/d) calculated from survey responses for wild mint.

Traditional Foods and Drinking Water (adopted and derived)

The health risk criteria for the protection of human health from consuming fish and untreated surface water were derived using fish consumption rates for seven species (0.388 kg/d) and a drinking water ingestion rate of 2.4 L/d. Additional input parameters and calculations are provided in Appendix A.4.

The US EPA HH AWQCs (US EPA, 2015c) are the only ambient water quality criteria which were developed for the protection of human health from consuming surface water (raw) and fish and consider carcinogenicity. As discussed above, the applicability of the HH AWQCs is limited for ACFN, FMFN and MCFN members which consume more fish (Figure 3.2) and more stringent guidelines are required to protect community members as compared to the US population. For certain substances, the guidelines prescribed by Health Canada and the WHO, which not only consider drinking water ingestion but also carcinogenicity, were more protective than the HH ACWR (US EPA) or derived health risk criteria. In these cases, the most stringent guideline was adopted.

The health risk criteria presented in Table 3.6 can be applied to surface water quality data to understand potential risks to human health from consumption of fish and natural/untreated surface water such as lakes, rivers and muskeg.

It is important to note that concentrations of substances required for the protection of humans consuming surface water and traditional foods may be different than concentrations associated with toxicological responses in more sensitive receptors (i.e., wildlife, aquatic biota, ecosystem function) and other water uses.

The health risk criteria for human consumption alone, should not be adopted unless all other exposure pathways described in Table 3.6 have been assessed and identified as not applicable or non-operational (i.e., the surface water being assessed is not used by humans or aquatic biota). The health risk criteria for traditional foods and drinking water may not always be the lowest value so it is important to review the health risk criteria for each water use category to understand risks to humans and ecological receptors.

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water.

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|-------------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|--|
| 1,1,1- Trichloroethane | | | ug/L | | | 200 | | 10000 | 2e+05 | | 200 | US EPA DWR |
| 1,1,2,2- Tetrachloroethan | e^* | | $\mathrm{ug/L}$ | | | | | 2 | 30 | | 2 | HH DW+Org (US EPA) |
| $1{,}1{,}2{-}$ Trichloroethane * | | | $\mathrm{ug/L}$ | | | 3 | | 5.5 | 89 | | 3 | US EPA DWR |
| 1,1- Dichloroethylene | | | $\mathrm{ug/L}$ | | 14 | 7 | | 300 | 20000 | | 7 | US EPA DWR |
| 1,2,3,4- Tetrachlorobenze | | | ug/L | | | | | 0.03 | 0.03 | | 0.03 | HH DW+Org (US EPA) USEPA WQC HH Org |
| 1,2,4- Trichlorobenzene | | | $\mathrm{ug/L}$ | | | 70 | | 0.071 | 0.76 | | 0.071 | HH DW+Org (US EPA) |
| 1,2-Dibromo-3- chloropropane | | | $\mathrm{ug/L}$ | | | 0.2 | | | | 1 | 0.2 | US EPA DWR |
| 1,2- Dibromoethane | | | $\mathrm{ug/L}$ | | | | | | | 0.4 | 0.4 | WHO DW |
| 1,2- Dichlorobenzene | | | ug/L | | | | | 1000 | 3000 | 1000 | 1000 | HH DW+Org (US EPA) WHO DW |
| $1,2-\\ \text{Dichloroethane}^*$ | | | $\mathrm{ug/L}$ | | 5 | 5 | | 99 | 6500 | 30 | 5 | Health Canada DW US EPA DWR |
| 1,2- Dichloroethene | | | $\mathrm{ug/L}$ | | | | | | | 50 | 50 | WHO DW |
| 1,2- Dichloropropane* | • | | $\mathrm{ug/L}$ | | | 5 | | 9 | 310 | 40 | 5 | US EPA DWR |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|---------------------------------|
| 1,2- Diphenylhydrazir | | | $\mathrm{ug/L}$ | | | | | 0.3 | 2 | | 0.3 | HH DW+Org (US EPA) |
| 1,3- Dichlorobenzene | | | $\mathrm{ug/L}$ | 13.33 | | | | 7 | 10 | | 7 | HH DW+Org (US EPA) |
| 1,3- Dichloropropene* | | | ug/L | | | | | 2.7 | 120 | 20 | 2.7 | HH DW+Org (US EPA) |
| 1,4- Dichlorobenzene | | | ug/L | | | | | 300 | 900 | 300 | 300 | HH DW+Org (US EPA) WHO DW |
| 1,4-Dioxane | | | ug/L | | | | | | | 50 | 50 | WHO DW |
| 2,3,4,6- Tetrachloropheno | ol | | $\mathrm{ug/L}$ | | 100 | | 1 | | | | 1 | USEPA WQC AO |
| 2,3- Dichlorophenol | | | $\mathrm{ug/L}$ | | | | 0.04 | | | | 0.04 | USEPA WQC AO |
| 2,4,5- Trichlorophenol | | | $\mathrm{ug/L}$ | | | | 1 | 300 | 600 | 9 | 1 | USEPA WQC AO |
| 2,4,6-Trichlorophenol* | | | ug/L | | 5 | | 2 | 15 | 28 | 200 | 2 | USEPA WQC AO |
| 2,4-D | | | ug/L | 451.29 | 100 | 70 | | 1300 | 12000 | 30 | 30 | WHO DW |
| 2,4-DB | | | ug/L | | | | | | | 90 | 90 | WHO DW |
| 2,4- Dichlorophenol | | | $\mathrm{ug/L}$ | | 900 | | 0.3 | 10 | 60 | | 0.3 | USEPA WQC AO |
| 2,4- Dimethylphenol | | | $\mathrm{ug/L}$ | | | | 400 | 100 | 3000 | | 100 | HH DW+Org (US EPA) |
| 2,4- Dinitrophenol | | | ug/L | 12.82 | | | | 10 | 300 | | 10 | HH DW+Org (US EPA) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|-----------------------|
| 2,4- Dinitrotoluene* | | | $\mathrm{ug/L}$ | | | | | 0.49 | 17 | | 0.49 | HH DW+Org (US EPA) |
| 2,5- Dichlorophenol | | | $\mathrm{ug/L}$ | | | | 0.5 | | | | 0.5 | USEPA WQC AO |
| 2,6- Dichlorophenol | | | ug/L | | | | 0.2 | | | | 0.2 | USEPA WQC AO |
| 2- Chloronaphthalei | ne | | $\mathrm{ug/L}$ | | | | | 800 | 1000 | | 800 | HH DW+Org (US EPA) |
| 2-Chlorophenol | | | ug/L | | | | 0.1 | 30 | 800 | | 0.1 | USEPA WQC AO |
| 2-Methyl-4,6- Dinitrophenol | | | $\mathrm{ug/L}$ | | | | | 2 | 30 | | 2 | HH DW+Org (US EPA) |
| 2-Methyl-4- Chlorophenol | | | ug/L | | | | 1800 | | | | 1800 | USEPA WQC AO |
| 3,3'- Dichlorobenzidin | e* | | $\mathrm{ug/L}$ | | | | | 0.49 | 1.5 | | 0.49 | HH DW+Org (US EPA) |
| 3,4- Dichlorophenol | | | ug/L | | | | 0.3 | | | | 0.3 | USEPA WQC AO |
| 3-Chlorophenol | | | $\mathrm{ug/L}$ | | | | 0.1 | | | | 0.1 | USEPA WQC AO |
| 3-Methyl-4- Chlorophenol | | | $\mathrm{ug/L}$ | | | | 3000 | 500 | 2000 | | 500 | HH DW+Org (US EPA) |
| 3-Methyl-6- Chlorophenol | | | $\mathrm{ug/L}$ | | | | 20 | | | | 20 | USEPA WQC AO |
| 4-Chlorophenol | | | $\mathrm{ug/L}$ | | | | 0.1 | | | | 0.1 | USEPA WQC AO |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|---------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|--|
| Acenaphthene [‡] | | | $\mathrm{ug/L}$ | 4.79 | | | 20 | 70 | 90 | | 4.79 | HH DW+Org (derived) |
| Acrolein | | | $\mathrm{ug/L}$ | 2.87 | | | | 3 | 400 | | 2.87 | HH DW+Org (derived) |
| Acrylamide | | | $\mathrm{ug/L}$ | 0.07 | | 0.5 | | | | 0.5 | 0.07 | HH DW+Org (derived) |
| Acrylonitrile* | | | $\mathrm{ug/L}$ | 0.53 | | | | 0.61 | 70 | | 0.53 | HH DW+Org (derived) |
| Alachlor | | | ug/L | | | 2 | | | | 20 | 2 | US EPA DWR |
| Aldicarb | | | ug/L | | | | | | | 10 | 10 | WHO DW |
| Aldrin* | | | ug/L | 1e-05 | | | | 7.7e-06 | 7.7e-06 | | 7.7e-06 | HH DW+Org (US EPA) USEPA WQC HH Org |
| Aldrin and dieldrin | | | $\mathrm{ug/L}$ | | | | | | | 0.03 | 0.03 | WHO DW |
| Aluminum | | Total | $\mathrm{ug/L}$ | | | | | | | 200 | 200 | WHO DW |
| Ammonia | | | mg/L | 0.67 | | | | | | 35 | 0.67 | HH DW+Org (derived) |
| Anthracene [‡] | | | $\mathrm{ug/L}$ | 20.07 | | | | 300 | 400 | | 20.07 | HH DW+Org (derived) |
| Antimony | | Total | $\mathrm{ug/L}$ | 4.59 | 6 | 6 | | 5.6 | 640 | 20 | 4.59 | HH DW+Org (derived) |
| Arsenic* | | Total | $\mathrm{ug/L}$ | 0.03 | 10 | 10 | | 0.18 | 1.4 | 10 | 0.03 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--|-------------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|--|
| Asbestos | | | $\mathrm{ug/L}$ | | | 7 | | 7 | | | 7 | US EPA DWR HH DW+Org (US EPA) |
| Atrazine | | | ug/L | | 5 | 3 | | | | | 3 | US EPA DWR |
| Atrazine and its chloro-s-triazine metabolites | | | ug/L | | | | | | | 100 | 100 | WHO DW |
| Azinphos- methyl | | | $\mathrm{ug/L}$ | | 20 | | | | | | 20 | Health Canada DW |
| Barium | | Total | ug/L | 1147.74 | 1000 | 2000 | | 1000 | | 1300 | 1000 | Health Canada DW HH DW+Org (US EPA) |
| Benzene* | | | $\mathrm{ug/L}$ | 2.11 | 5 | 5 | | 5.8 | 160 | 10 | 2.11 | HH DW+Org (derived) |
| Benzidine* | | | $\mathrm{ug/L}$ | 0.001 | | | | 0.0014 | 0.11 | | 0.001 | HH DW+Org (derived) |
| Benzo(a)anthrac | cene ^{*†} | | $\mathrm{ug/L}$ | 0.001 | | | | 0.012 | 0.013 | | 0.001 | HH DW+Org (derived) |
| $\begin{array}{l} Benzo(a)pyrene\\ and\\ equivalents \\ ^{*\dagger} \end{array}$ | | | ug/L | 1e-04 | 0.04 | 0.2 | | 0.001 | 0.0013 | 0.7 | 1e-04 | HH DW+Org (derived) |
| Benzo(b)fluoran | ${ m thene}^{*\dagger}$ | | $\mathrm{ug/L}$ | 0.001 | | | | 0.012 | 0.013 | | 0.001 | HH DW+Org (derived) |
| Benzo(k)fluoran | t | | $\mathrm{ug/L}$ | 0.01 | | | | 0.12 | 0.13 | | 0.01 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|---|
| Beryllium | | Total | ug/L | 3.27 | | 4 | | | | | 3.27 | HH DW+Org (derived) |
| Bis(2-Chloro- 1-methylethyl) Ether | | | ug/L | 127.99 | | | | 200 | 4000 | | 127.99 | HH DW+Org (derived) |
| Bis(2- Chloroethyl) Ether* | | | ug/L | 0.25 | | | | 0.3 | 22 | | 0.25 | HH DW+Org (derived) |
| Bis(2- Ethylhexyl) Phthalate | | | $\mathrm{ug/L}$ | 0.21 | | | | 0.32 | 0.37 | | 0.21 | HH DW+Org (derived) |
| Bis(Chloromethy Ether* | 1) | | $\mathrm{ug/L}$ | 0.001 | | | | 0.002 | 0.17 | | 0.001 | HH DW+Org (derived) |
| Boron | | Total | ug/L | 1333.33 | 5000 | | | | | 2400 | 1333.33 | HH DW+Org (derived) |
| Bromate | | | $\mathrm{ug/L}$ | | 10 | 10 | | | | 10 | 10 | Health Canada DW US EPA DWR WHO DW |
| Bro- modichlorometha | | | ug/L | 6.33 | | | | | | 60 | 6.33 | HH DW+Org (derived) |
| Bromoform | | | ug/L | 38.22 | | | | 7 | 120 | 100 | 7 | HH DW+Org (US EPA) |
| Bromoxynil | | | ug/L | | 5 | | | | | | 5 | Health Canada DW |
| Butylbenzyl Phthalate [*] | | | ug/L | 0.06 | | | | 1 | 1 | | 0.06 | HH DW+Org (derived) |
| Cadmium | | Total | ug/L | 0.002 | | 5 | | | | 3 | 0.002 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|-------------------------------|
| Carbaryl | | | $\mathrm{ug/L}$ | | 90 | | | | | | 90 | Health Canada DW |
| Carbofuran | | | ug/L | | 90 | 40 | | | | 7 | 7 | WHO DW |
| Carbon tetrachloride | | | $\mathrm{ug/L}$ | 1.9 | 2 | 5 | | 4 | 50 | 4 | 1.9 | HH DW+Org (derived) |
| Chloramines | | | ug/L | | 3000 | 4000 | | | | | 3000 | Health Canada DW |
| Chlorate | | | $\mathrm{ug/L}$ | | 1000 | | | | | 700 | 700 | WHO DW |
| Chlordane | | | ug/L | 0.001 | | 2 | | 0.003 | 0.0032 | 0.2 | 0.001 | HH DW+Org (derived) |
| Chloride | | | $\mathrm{mg/L}$ | | 250 | | | | | 250 | 250 | Health Canada DW WHO DW |
| Chlorine dioxide | | | ug/L | | | 800 | | | | | 800 | US EPA DWR |
| Chlorite | | | $\mathrm{ug/L}$ | | 1000 | 800 | | | | 700 | 700 | WHO DW |
| Chlorobenzene | | | ug/L | 40.85 | 80 | 100 | | 100 | 800 | | 40.85 | HH DW+Org (derived) |
| Chlorodibro- momethane | | | $\mathrm{ug/L}$ | | | | | 8 | 210 | | 8 | HH DW+Org (US EPA) |
| Chloroform | | | ug/L | 45.89 | | | | 60 | 2000 | 300 | 45.89 | HH DW+Org (derived) |
| Chlorophenoxy Herbicide (2,4,5-TP) [Silvex] | | | ug/L | 20.55 | | 50 | | 100 | 400 | | 20.55 | HH DW+Org (derived) |
| Chlorotoluron | | | $\mathrm{ug/L}$ | | | | | | | 30 | 30 | WHO DW |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|---------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|--|
| Chlorpyrifos | | | ug/L | | 90 | | | | | 30 | 30 | WHO DW |
| Chromium (III) | | Total | $\mathrm{ug/L}$ | 10000 | | | | 100 | 100 | | 100 | HH DW+Org (US EPA) USEPA WQC HH Org |
| Chromium (VI) | | Total | ug/L | 13.47 | | | | 100 | 100 | | 13.47 | HH DW+Org (derived) |
| Chromium | | Total | ug/L | | 50 | 100 | | | | 50 | 50 | Health Canada DW WHO DW |
| Chrysene*† | | | ug/L | 0.07 | | | | 1.2 | 1.3 | | 0.07 | HH DW+Org (derived) |
| Copper* | | Total | ug/L | | 2000 | 1300 | 1000 | 13000 | | 2000 | 1000 | USEPA WQC AO |
| Cyanazine | | | ug/L | | | | | | | 0.6 | 0.6 | WHO DW |
| Cyanide | as free CN | | ug/L | 3.62 | 200 | 200 | | 4 | 400 | | 3.62 | HH DW+Org (derived) |
| Cyanobacterial toxins | | | ug/L | | 1.5 | | | | | | 1.5 | Health Canada DW |
| DDT and metabolites* | | | ug/L | | | | | 3e-04 | 3e-04 | 1 | 3e-04 | HH DW+Org (US EPA) USEPA WQC HH Org |
| Dalapon | | | ug/L | | | 200 | | | | | 200 | US EPA DWR |
| Di(2- ethylhexyl) adipate | | | ug/L | | | 400 | | | | | 400 | US EPA DWR |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|-----------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|------------------------|
| Di(2- ethylhexyl) phthalate | | | ug/L | | | 6 | | | | 8 | 6 | US EPA DWR |
| Di-n-Butyl Phthalate | | | $\mathrm{ug/L}$ | 1.42 | | | | 20 | 30 | | 1.42 | HH DW+Org (derived) |
| Diazinon | | | $\mathrm{ug/L}$ | | 20 | | | | | | 20 | Health Canada DW |
| Dibenzo(a,h)an | tł | | $\mathrm{ug/L}$ | 1e-04 | | | | 0.001 | 0.0013 | | 1e-04 | HH DW+Org (derived) |
| Dibromoacetoni | itrile* | | $\mathrm{ug/L}$ | | | | | | | 70 | 70 | WHO DW |
| Dibro- mochlorometha: | n | | $\mathrm{ug/L}$ | 5.21 | | | | | | 100 | 5.21 | HH DW+Org (derived) |
| Dicamba | | | $\mathrm{ug/L}$ | | 120 | | | | | | 120 | Health Canada DW |
| Dichloroac- etate | | | $\mathrm{ug/L}$ | | | | | | | 50 | 50 | WHO DW |
| Dichloroace- tonitrile | | | $\mathrm{ug/L}$ | | | | | | | 20 | 20 | WHO DW |
| Dichlorobro- momethane | | | $\mathrm{ug/L}$ | | | | | 9.5 | 270 | | 9.5 | HH DW+Org (US EPA) |
| Dichloromethan | ue* | | $\mathrm{ug/L}$ | | 50 | 5 | | | | 20 | 50 | Health Canada DW |
| Dichlorprop | | | ug/L | | | | | | | 100 | 100 | WHO DW |
| Diclofop- methyl | | | $\mathrm{ug/L}$ | | 9 | | | | | | 9 | Health Canada DW |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|---|
| Dieldrin* | | | $\mathrm{ug/L}$ | 1e-05 | | | | 1e-05 | 1.2e-05 | | 1e-05 | HH DW+Org (derived) HH DW+Org (US EPA) |
| Diethyl Phthalate | | | ug/L | 35.61 | | | | 600 | 600 | | 35.61 | HH DW+Org (derived) |
| Dimethoate | | | ug/L | | 20 | | | | | 6 | 6 | WHO DW |
| Dimethyl Phthalate | | | $\mathrm{ug/L}$ | 102.91 | | | | 2000 | 2000 | | 102.91 | HH DW+Org (derived) |
| Dinitrophenols | | | $\mathrm{ug/L}$ | 10.72 | | | | 10 | 1000 | | 10 | HH DW+Org (US EPA) |
| Dinoseb | | | ug/L | | | 7 | | | | | 7 | US EPA DWR |
| Dioxin (2,3,7,8- TCDD) | | | ug/L | | | 3e-05 | | 5e-08 | 5.1e-08 | | 5e-08 | HH DW+Org (US EPA) |
| Diquat | | | ug/L | | 70 | 20 | | | | | 20 | US EPA DWR |
| Diuron | | | ug/L | | 150 | | | | | | 150 | Health Canada DW |
| Edetic acid | | | ug/L | | | | | | | 600 | 600 | WHO DW |
| Endosulfan Sulfate | | | ug/L | 2.63 | | | | 20 | 40 | | 2.63 | HH DW+Org (derived) |
| Endothall | | | ug/L | | | 100 | | | | | 100 | US EPA DWR |
| Endrin | | | ug/L | 0.01 | | 2 | | 0.03 | 0.03 | 0.6 | 0.01 | HH DW+Org (derived) |
| Endrin Aldehyde | | | ug/L | 0.11 | | | | 1 | 1 | | 0.11 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|------------------------|
| Epichlorohy- drin | | | $\mathrm{ug/L}$ | | | 200 | | | | 0.4 | 0.4 | WHO DW |
| Ethylbenzene | | | $\mathrm{ug/L}$ | 8.54 | 140 | 700 | | 68 | 130 | 300 | 8.54 | HH DW+Org (derived) |
| Ethylene dibromide | | | ug/L | | | 0.05 | | | | | 0.05 | US EPA DWR |
| Fenoprop | | | ug/L | | | | | | | 9 | 9 | WHO DW |
| Fluoranthene [‡] | | | ug/L | 1.09 | | | | 20 | 20 | | 1.09 | HH DW+Org (derived) |
| Fluorene [‡] | | | $\mathrm{ug/L}$ | 6.98 | | | | 50 | 70 | | 6.98 | HH DW+Org (derived) |
| Fluoride | | | mg/L | 0.4 | 1.5 | 4 | | | | 1.5 | 0.4 | HH DW+Org (derived) |
| Glyphosate | | | $\mathrm{ug/L}$ | | 280 | 700 | | | | | 280 | Health Canada DW |
| Haloacetic acids | | | $\mathrm{ug/L}$ | | 80 | 60 | | | | | 60 | US EPA DWR |
| ${\bf Heptachlor}^*$ | | | $\mathrm{ug/L}$ | 4e-05 | | 0.4 | | 6e-05 | 5.9e-05 | | 4e-05 | HH DW+Org (derived) |
| $\begin{array}{l} {\rm Heptachlor} \\ {\rm epoxide}^* \end{array}$ | | | ug/L | 1e-04 | | 0.2 | | 0.00032 | 0.00032 | | 1e-04 | HH DW+Org (derived) |
| Hexachlorobenzer | ne* | | $\mathrm{ug/L}$ | 1e-04 | | 1 | | 0.001 | 0.00079 | | 1e-04 | HH DW+Org (derived) |
| Hexachlorobutad | | | $\mathrm{ug/L}$ | 0.001 | | | | 0.1 | 0.1 | 0.6 | 0.001 | HH DW+Org (derived) |
| Hexachlorocycloh | exane* | | $\mathrm{ug/L}$ | 0.01 | | | | 0.066 | 0.1 | | 0.01 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|------------------------|
| Hexachlorocy- clopentadiene | | | $\mathrm{ug/L}$ | 0.4 | | 50 | 1 | 4 | 4 | | 0.4 | HH DW+Org (derived) |
| Hexachloroethan | e* | | $\mathrm{ug/L}$ | 0.02 | | | | 1 | 1 | | 0.02 | HH DW+Org (derived) |
| Hydroxya- trazine | | | $\mathrm{ug/L}$ | | | | | | | 200 | 200 | WHO DW |
| Indeno(1,2,3-cd)pyrene | | | $\mathrm{ug/L}$ | 0.001 | | | | 0.012 | 0.013 | | 0.001 | HH DW+Org (derived) |
| Iron | | Total | ug/L | | | | 300 | | | | 300 | USEPA WQC AO |
| $Isophorone^*$ | | | $\mathrm{ug/L}$ | 268.41 | | | | 340 | 18000 | | 268.41 | HH DW+Org (derived) |
| Isoproturon | | | ug/L | | | | | | | 9 | 9 | WHO DW |
| Lead | | Total | $\mathrm{ug/L}$ | | 5 | 15 | | | | 10 | 5 | Health Canada DW |
| MCPA | | | ug/L | | 100 | | | | | | 100 | Health Canada DW |
| Malathion | | | $\mathrm{ug/L}$ | | 190 | | | | | | 190 | Health Canada DW |
| Manganese | | Total | ug/L | 933.33 | 120 | | | 50 | 100 | | 50 | HH DW+Org (US EPA) |
| Mecoprop | | | ug/L | | | | | | | 10 | 10 | WHO DW |
| Mercury (methyl) | | Total | $\mathrm{ug/L}$ | 0.67 | | | | | | | 0.67 | HH DW+Org (derived) |
| Mercury | | Total | $\mathrm{ug/L}$ | | 1 | 2 | | | | 6 | 1 | Health Canada DW |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|------|--|---|
| Methoxychlor | | | $\mathrm{ug/L}$ | 0.001 | | 40 | | 0.02 | 0.02 | 20 | 0.001 | HH DW+Org (derived) |
| Methyl Bromide | | | $\mathrm{ug/L}$ | 111.66 | | | | 100 | 10000 | | 100 | HH DW+Org (US EPA) |
| Methylene chloride* | | | $\mathrm{ug/L}$ | 32.62 | | | | 200 | 10000 | | 32.62 | HH DW+Org (derived) |
| Metolachlor | | | $\mathrm{ug/L}$ | | 50 | | | | | 10 | 10 | WHO DW |
| Metribuzin | | | $\mathrm{ug/L}$ | | 80 | | | | | | 80 | Health Canada DW |
| Microcystin- LR | | | $\mathrm{ug/L}$ | | | | | | | 1 | 1 | WHO DW |
| Molinate | | | ug/L | | | | | | | 6 | 6 | WHO DW |
| Molybdenum | | Total | $\mathrm{ug/L}$ | 33.33 | | | | | | | 33.33 | HH DW+Org (derived) |
| Monochlo- ramine | | | $\mathrm{ug/L}$ | | | | | | | 3000 | 3000 | WHO DW |
| Monochloroac- etate | | | $\mathrm{ug/L}$ | | | | | | | 20 | 20 | WHO DW |
| Monochloroben- zene | | | $\mathrm{ug/L}$ | | | | 20 | | | | 20 | USEPA WQC AO |
| N-Nitrosodi-n- Propylamine* | | | ug/L | 0.05 | | | | 0.05 | 5.1 | | 0.05 | HH DW+Org (derived) HH DW+Org (US EPA) |
| N- Nitrosodimethyla | 3 | | ug/L | 0.01 | 0.04 | | | 0.007 | 30 | 0.1 | 0.007 | HH DW+Org (US EPA) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-------|--|--|
| N- Nitrosodiphenyla | amine^* | | ug/L | 68.03 | | | | 33 | 60 | | 33 | HH DW+Org (US EPA) |
| Naphthalene [‡] | | | ug/L | 133.33 | | | | | | | 133.33 | HH DW+Org (derived) |
| Nickel | | Total | $\mathrm{ug/L}$ | 7.35 | | | | 610 | 4600 | 70 | 7.35 | HH DW+Org (derived) |
| Nitrate | as N | dissolved | ${ m mg/L}$ | 10.1 | 10 | 10 | | 10 | | 11.3 | 10 | Health Canada DW US EPA DWR HH DW+Org (US EPA) |
| Nitrilotriacetic acid | | | $\mathrm{ug/L}$ | | 400 | | | | | 200 | 200 | WHO DW |
| Nitrite | as N | dissolved | $\mathrm{mg/L}$ | | 1 | 1 | | | | 0.912 | 0.912 | WHO DW |
| Nitrobenzene | | | $\mathrm{ug/L}$ | 9.72 | | | 30 | 10 | 600 | | 9.72 | HH DW+Org (derived) |
| Nitrosamines | | | ug/L | | | | | 0.008 | 12.4 | | 0.008 | HH DW+Org (US EPA) |
| Nitrosodibuty- lamine | | | ug/L | 0.05 | | | | 0.063 | 2.2 | | 0.05 | HH DW+Org (derived) |
| Nitrosodiethy- lamine | | | ug/L | 0.002 | | | | 0.008 | 12.4 | | 0.002 | HH DW+Org (derived) |
| Nitrosopyrroli- dine | | | $\mathrm{ug/L}$ | 0.16 | | | | 0.16 | 340 | | 0.16 | HH DW+Org (derived) HH DW+Org (US EPA) |
| Oxamyl (Vydate) | | | ug/L | | | 200 | | | | | 200 | US EPA DWR |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|---|------------------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|------------------------|
| Paraquat | as paraquat dichloride | | $\mathrm{ug/L}$ | | 10 | | | | | | 10 | Health Canada DW |
| Pendimethalin | | | ug/L | | | | | | | 20 | 20 | WHO DW |
| Pentachlorobenz | zene* | | $\mathrm{ug/L}$ | 0.01 | | | | 0.1 | 0.1 | | 0.01 | HH DW+Org (derived) |
| Pen- tachlorophenol | | | ug/L | 0.1 | 60 | 1 | 30 | 0.3 | 0.4 | 9 | 0.1 | HH DW+Org (derived) |
| Perchlorate | | | ug/L | | | | | | | 70 | 70 | WHO DW |
| Perfluorooc- tanesulfonate | | | ug/L | | 0.6 | | | | | | 0.6 | Health Canada DW |
| Perfluorooctanoic acid | | | $\mathrm{ug/L}$ | | 0.2 | | | | | | 0.2 | Health Canada DW |
| Phenanthrene | | | $\mathrm{ug/L}$ | 200 | | | | | | | 200 | HH DW+Org (derived) |
| Phenol | | | $\mathrm{ug/L}$ | 1609.58 | | | 300 | 4000 | 3e+05 | | 300 | USEPA WQC AO |
| Phorate [‡] | | | $\mathrm{ug/L}$ | | 2 | | | | | | 2 | Health Canada DW |
| Picloram | | | $\mathrm{ug/L}$ | | 190 | 500 | | | | | 190 | Health Canada DW |
| Polychlori- nated Biphenyls (PCBs) | | | ug/L | | | 0.5 | | 0.001 | 0.00064 | | 0.00064 | USEPA WQC HH Org |
| Pyrene | | | $\mathrm{ug/L}$ | 1.43 | | | | 20 | 30 | | 1.43 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. *(continued)*

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|-------------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-------|--|------------------------|
| Selenium | | Total | ug/L | 18.77 | 50 | 50 | | 170 | 4200 | 40 | 18.77 | HH DW+Org (derived) |
| Silver [‡] | | Total | ug/L | 33.33 | | | | | | | 33.33 | HH DW+Org (derived) |
| Simazine | | | ug/L | | 10 | 4 | | | | 2 | 2 | WHO DW |
| Sodium dichloroisocya- nurate | | | ug/L | | | | | | | 40000 | 40000 | WHO DW |
| Solids Dissolved and Salinity | | | ug/L | | | | | 250000 | | | 250000 | HH DW+Org (US EPA) |
| Strontium | | Total | ug/L | 4000 | 7000 | | | | | | 4000 | HH DW+Org (derived) |
| Styrene | | | ug/L | | | 100 | | | | 20 | 20 | WHO DW |
| Sulfate | as SO4 | | $\mathrm{mg/L}$ | | | | | | | 250 | 250 | WHO DW |
| Terbufos | | | ug/L | | 1 | | | | | | 1 | Health Canada DW |
| Terbuthylazine | | | ug/L | | | | | | | 7 | 7 | WHO DW |
| Tetra- chloroethylene | | | ug/L | 4.48 | 10 | 5 | | 100 | 290 | 40 | 4.48 | HH DW+Org (derived) |
| Thallium | | Total | ug/L | 0.02 | | 0.5 | | 0.24 | 0.47 | | 0.02 | HH DW+Org (derived) |
| Toluene | | | ug/L | 191.93 | 60 | 1000 | | 57 | 520 | 700 | 57 | HH DW+Org (US EPA) |
| Toxaphene | | | ug/L | 0.001 | | 3 | | 0.007 | 0.0071 | | 0.001 | HH DW+Org (derived) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. *(continued)*

| | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|---|
| Trans-1,2- Dichloroethylene | | | ug/L | | | | | 100 | 4000 | | 100 | HH DW+Org (US EPA) |
| Trichloroac- etate | | | $\mathrm{ug/L}$ | | | | | | | 200 | 200 | WHO DW |
| Trichloroethy- lene | | | ug/L | 1.38 | 5 | 5 | | 6 | 70 | 20 | 1.38 | HH DW+Org (derived) |
| Trifluralin | | | ug/L | | 45 | | | | | 20 | 20 | WHO DW |
| Tri- halomethanes | | | ug/L | | 100 | 80 | | | | | 80 | US EPA DWR |
| Uranium* | | Total | $\mathrm{ug/L}$ | 20 | 20 | 30 | | | | 30 | 20 | HH DW+Org (derived) Health Canada DW |
| Vinyl chloride | | | ug/L | 0.18 | 2 | 2 | | 0.22 | 16 | 0.3 | 0.18 | HH DW+Org (derived) |
| Xylene | | | ug/L | 114.15 | 90 | | | | | 500 | 90 | Health Canada DW |
| Xylenes (total) | | | ug/L | | | 10000 | | | | | 10000 | US EPA DWR |
| Zinc* | | Total | ug/L | 12.72 | | | 5000 | 7400 | 26000 | | 12.72 | HH DW+Org (derived) |
| alpha- Endosulfan | | | ug/L | 1.82 | | | | 20 | 30 | | 1.82 | HH DW+Org (derived) |
| alpha- Hexachlorocyclohe | exane | | $\mathrm{ug/L}$ | 2e-04 | | | | 0.0036 | 0.0039 | | 2e-04 | HH DW+Org (derived) |
| beta- Endosulfan | | | ug/L | 2.87 | | | | 20 | 40 | | 2.87 | HH DW+Org (derived) |

and drinking water. (continued)

Units Derived Health DWR US WQC HH Org HH DW WH

Table 3.6: Health risk criteria for the protection of community consumers of fish

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|--|
| beta- Hexachlorocyc | lohexane | | ug/L | 0.01 | | | | 0.08 | 0.14 | | 0.01 | HH DW+Org (derived) |
| cis-1,2- Dichloroethyle: | ne | | ug/L | | | 70 | | | | | 70 | US EPA DWR |
| gamma- Hexachlorocyc [Lindane] | lohexane | | $\mathrm{ug/L}$ | 0.4 | | 0.2 | | 4.2 | 4.4 | 2 | 0.2 | US EPA DWR |
| o- Dichlorobenzer | ne | | ug/L | | 200 | 600 | | | | | 200 | Health Canada DW |
| p,p - Dichlorodipher (DDD)* | nyldichloroethane | e | $\mathrm{ug/L}$ | | | | | 0.001 | 0.0012 | | 0.001 | HH DW+Org (US EPA) |
| p,p - Dichlorodipher (DDE)* | nyl | | $\mathrm{ug/L}$ | | | | | 2e-04 | 0.00018 | | 0.00018 | USEPA WQC HH Org |
| p- Dichlorobenzer | ne | | ug/L | | 5 | 75 | | | | | 5 | Health Canada DW |
| рН | | | pH units | | 7 | | | 5 | | | 44751 | Health Canada DW HH DW+Org (US EPA) |

Table 3.6: Health risk criteria for the protection of community consumers of fish and drinking water. (continued)

| Parameter | Method Speciation | Sample Fraction | Units | Derived | Health Canada | DWR US EPA | WQC AO US EPA | HH Org US EPA | HH DW Org US EPA | WHO | Traditional Foods and Drinking Water Criteria Value | Source |
|--------------------------------|----------------------|--------------------|-----------------|---------|------------------|---------------|---------------------|------------------|------------------------|-----|--|------------|
| trans-1,2- Dichloroethylene | 2 | | $\mathrm{ug/L}$ | | | 100 | | | | | 100 | US EPA DWR |

Note:

Known carcinogens, US EPA HH ACWR (DW+C) were adjusted to reflect 10⁻⁵ ILCR levels (Alberta Health (2019))

* Known human carcinogen via oral exposure route (Health Canada (2021))

[†] The following known human carcinogens and must be converted to Provisional Benzo[a]pyrene RPF and summed as per Health Canada (2021) then compared to the Benzo(a)pyrene and equivalents health risk criteria: Anthanthrene, Benzo[c]chrysene, Benzo[g]chrysene, Benzo[c]phenanthrene, Cyclopenta[c,d]pyrene, Dibenzo[a,e]fluoranthene Dibenzo[a,e]pyrene, Dibenzo[a,h]pyrene, Dibenzo[a,i]pyrene, Dibenzo[a,l]pyrene, 9,10- Dimethylanthracene, 7,12- Dimethylbenzo[a]anthracene, 1,2- Dimethylbenzo[a]pyrene, 3,6- Dimethylbenzo[a]pyrene, 4,5- Dimethylbenzo[a]pyrene, 5,6- Dimethylchrysene, 5,11- Dimethylchrysene, 1,4- Dimethylphenanthrene, 4,10- Dimethylphenanthrene, 5- Ethylchrysene, Fluoranthene, 7- Methylbenzo[a]anthracene, Methylbenzo[a]anthracene, 11- Methylbenzo[b]fluorene, Methylbenzo[a]pyrene, Methylbenzo[a]pyren

[‡] Naphthalene applied as surrogate to sum of low molecular weight PAH congeners (Anthracene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) and compare to Naphthalene health risk criteria (adopted as surrogate) (CCME (2010))

Traditional Medicines (derived)

The health risk criteria for the protection of human health from consuming traditional medicines were derived using consumption rates for rat root (0.0068 kg/d) and are provided in Table 3.7. Additional input parameters and calculations are provided in Appendix A.4.

These criteria were developed using modifications to the (US EPA, 2000b) methodology aligning with human health risk assessment protocols where BCFs for sediment to plants are adopted to predict the uptake of contaminants by aquatic plants.

Due to this uncertainty and lack of BCF data for culturally important aquatic plant species (i.e. fresh rat root), the health risk criteria identified in Table 3.7 should be considered interim until discussions with health agencies can confirm modifications and BCFs for rat root and wild mint should be applied to medicinal plants.

Table 3.7: Health risk criteria for the protection of community consumers of medicinal plants.

| Parameter Name | Units | Value |
|-------------------------------|-----------------|-------|
| Acenaphthene | $\mathrm{mg/L}$ | 0 |
| Anthracene | $\mathrm{mg/L}$ | 0 |
| Antimony | m mg/L | 9 |
| Arsenic* | $\mathrm{mg/L}$ | 2 |
| Barium | $\mathrm{mg/L}$ | 3137 |
| Benzene | $\mathrm{mg/L}$ | 0 |
| Benzo(a)anthracene* | $\mathrm{mg/L}$ | 8 |
| Benzo(a)pyrene* | $\mathrm{mg/L}$ | 0 |
| Benzo(b)fluoranthene* | $\mathrm{mg/L}$ | 16 |
| Benzo(k) fluoranthene * | $\mathrm{mg/L}$ | 160 |
| Cadmium | $\mathrm{mg/L}$ | 3 |
| Chrysene* | $\mathrm{mg/L}$ | 862 |
| Copper | $\mathrm{mg/L}$ | 0 |
| Chromium (VI) | $\mathrm{mg/L}$ | 941 |
| Chromium (III) | $\mathrm{mg/L}$ | 0 |
| Cyanide | $\mathrm{mg/L}$ | 0 |
| $Dibenzo(a,h) anthracene^*\\$ | $\mathrm{mg/L}$ | 3 |
| Ethylbenzene | $\mathrm{mg/L}$ | 0 |
| Fluoranthene | $\mathrm{mg/L}$ | 0 |
| Fluorene | $\mathrm{mg/L}$ | 0 |
| Indeno(1,2,3-cd)pyrene* | $\mathrm{mg/L}$ | 41 |
| Lead | $\mathrm{mg/L}$ | 7320 |
| Manganese | $\mathrm{mg/L}$ | 0 |
| | | |

Table 3.7: Health risk criteria for the protection of community consumers of medicinal plants. *(continued)*

| Parameter Name | Units | Value |
|----------------|-----------------|----------|
| Mercury | m mg/L | 19 |
| Nickel | m mg/L | 1471 |
| Phenol | $\mathrm{mg/L}$ | 0 |
| Pyrene | m mg/L | 0 |
| Selenium | $\mathrm{mg/L}$ | 735 |
| Thallium | m mg/L | 4 |
| Toluene | $\mathrm{mg/L}$ | 0 |
| Zinc | m mg/L | > 10,000 |
| * ~ | | |

^{*} Substances are known carcinogens in humans and cannot be assessed using non-carcinogenic thresholds.

3.5 Discussion

The health risk criteria which were developed in this project recognize both western science environmental assessment methods and Indigenous community world views and knowledge systems.

The conceptual model identified Indigenous water uses and exposure pathways that are not explicitly considered or protected through application of provincial or federal surface water quality guidelines.

A key finding of this project which informed method development was the consideration that water use protection goals (described in 3.8) of ACFN, FMFN and MCFN community members are holistic, require protection of human receptors, and include more water uses than considered under the provincial and federal processes for defining surface water quality guidelines.

Members shared that understanding the health of water (and all-connected components) is experiential, relational, and directly informs their sense of personal health and wellbeing. As such, water cannot be managed as a single component broken off from the environment or communities. Water is the giver of life and must be protected using traditional knowledge and now due to industrial development, western science methods must also be relied on. Members also communicated that western science water management practices were unnecessary prior to industrial development in the Lower Athabasca Region (personal communications).

Table 3.8: Indigenous community water uses and health protection goals used to define health risk criteria.

| Indigenous water use | Protection Goal |
|--------------------------------------|---|
| Traditional foods and drinking water | Safe foods consumption |
| | Safe natural surface water consumption |
| Traditional medicines | Safe medicine consumption |
| Aquatic ecosystem health | Aquatic community consumption unchanged |
| | Robust populations |
| | Natural behaviours and patterns |
| Wildlife health | Healthy wildlife |
| | Robust populations |
| | Natural behaviours and patterns |
| | Good quality pelts |

The review of water quality guidelines prescribed across North American and internationally indicate that ambient surface water guidelines have been derived for the protection of ecological and human receptors. Adaptation of the identified water guidelines used in Alberta (GoA, 2018) to consider the protection of human health can be achieved by supplementing the current protection of aquatic life focused regime with human health guidelines specifically developed for consumption of ambient water and organisms (US EPA, 2015a) and integrated available drinking water quality standards (Health Canada (2020a); World Health Organization (WHO) (2017); US EPA DWRs).

The consumption rates used to develop the regulatory guidelines are generally representative of the average consumption rates of fish and surface water reported for ACFN, FMFN and MCFN members but would not protect members who are heavier consumers of fish.

Modifications of the existing guidelines were used to achieve a higher degree of protection for by deriving health risk criteria that will protect consumers of traditional foods based on the upper range of fish (388 g/d) and medicinal plant (6.8 g/d) consumption.

Further integrating water quality benchmarks to protect piscivorous wildlife species (Sample et al., 1996) and water use pathways developed for agricultural purposes (GoA, 2018), specifically, livestock watering, would offer a degree of protection to wildlife species consuming surface water and being consumed used as traditional foods.

A comparison of the health risk criteria developed for various water uses and protection

goals aligns with the multi-use system developed by GOA and CCME in that some water uses require a higher degree of protection than other uses. This is due to the sensitivity of receptors being exposed, toxicological, chemical, and physical properties of the contaminants and likelihood of exposure. Similar to the application of existing guidelines the various use specific criteria can be selectively applied based on how Indigenous communities are interacting with a specific waterbody or the most protective criteria (i.e. lowest value) can be selected to ensure all other uses are protected.

In general terms, the two most sensitive water uses identified in this research were traditional foods/drinking water supply and aquatic ecosystem health protection.

The toxicity, persistence, and bioaccumulation of contaminants drives risk potential of contaminants in aquatic ecosystems and each substance should be evaluated rather than assessing water quality by use, as is common practice in Alberta (i.e. PAL guidelines to screen surface water quality data regardless of contaminants).

Risk is also driven by the sensitivity of the receptor and chemical, physical and toxicological properties of each substance, therefore a single use protection category cannot meet each of the Indigenous water protection goals for human and ecological receptors. Application of criteria for a single water use will limit protection and underestimate potential risks particularly for carcinogens (i.e., arsenic, high MW PAHs).

Recognizing that human and ecological health risks are a function of exposure and inherent toxicity of the contaminants, it is recommended that the health risk criteria shown in Table 3.9 be used to assess the quality of water in surface water that is being developed for Indigenous use purposes or currently being used by Indigenous communities. The generic use protection category is equivalent to the Tier 1 category within the tiered system used by Alberta (AEP, 2019) for assessing contamination and developing remediation/ treatment programs of soils and groundwater.

For parameters that did not have published guidelines, it is recommended that the current condition for open water season at the Athabasca River location be adopted (see Chapter 2).

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use.

| | | | | Generic health | Risk Criteria | Specific | Water Use C | ategory Health R | isk Criteria |
|-----------------------------------|--------------------|--------------|-------|--------------------|---|-------------------------------|--------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines |
| .alphaEndosulfan | | ug/L | 0.056 | aquatic biota | US EPA Aquatic Life Criteria | 0.056 | | | |
| .betaEndosulfan | | $_{ m ug/L}$ | 0.056 | aquatic biota | US EPA Aquatic Life Criteria | 0.056 | | | |
| 1,1,1- Trichloroethane* | | ug/L | 200 | human | US EPA DWR | | | 200 | |
| 1,1,2,2- Tetrachloroethane* | | ug/L | 2 | human | HH DW+Org (US EPA) | | | 2 | |
| 1,1,2-Trichloroethane | | ug/L | 3 | human | US EPA DWR | 21 | | 3 | |
| 1,1-Dichloroethylene | | ug/L | 7 | human | US EPA DWR | | 929.00 | 7 | |
| 1,2,3,4- Tetrachlorobenzene | | ug/L | 0.03 | human | USEPA WQC HH Org HH DW+Org (US EPA) | 1.8 | | 0.03 | |
| 1,2,3- Trichlorobenzene | | $_{ m ug/L}$ | 8 | aquatic biota | AEP Water PAL CCME Water PAL | 8 | | | |
| 1,2,4- Trichlorobenzene | | ug/L | 0.071 | human | HH DW+Org (US EPA) | 24 | | 0.071 | |
| 1,2-Dibromo-3- chloropropane | | ug/L | 0.2 | human | US EPA DWR | | | 0.2 | |
| 1,2-Dibromoethane | | ug/L | 0.4 | human | WHO DW | | | 0.4 | |
| 1, 2- Dichlorobenzene | | ug/L | 0.7 | aquatic biota | AEP Water PAL | 0.7 | | 1000 | |
| $1,2\text{-Dichloroethane}^*$ | | ug/L | 5 | human wildlife | Health Canada DW AEP Water Ag CCME Water Ag (limited) US EPA DWR | 100 | 5.00 | 5 | |
| 1,2-Dichloroethene | | $_{ m ug/L}$ | 50 | human | WHO DW | | | 50 | |
| 1,2-Dichloropropane* | | ug/L | 5 | human | US EPA DWR | | | 5 | |
| 1,2 - Diphenylhydrazine * | | ug/L | 0.3 | human | HH DW+Org (US EPA) | | | 0.3 | |
| 1,3-Dichlorobenzene | | ug/L | 7 | human | HH DW+Org (US EPA) | 150 | | 7 | |
| 1,3-Dichloropropene* | | ug/L | 2.7 | human | HH DW+Org (US EPA) | | | 2.7 | |
| 1,4-Dichlorobenzene | | $_{ m ug/L}$ | 26 | aquatic biota | AEP Water PAL | 26 | | 300 | |
| 1,4-Dioxane | | $_{ m ug/L}$ | 50 | human | WHO DW | | | 50 | |
| 2,3,4,6- Tetrachlorophenol | | $_{ m ug/L}$ | 1 | human | USEPA WQC AO | | | 1 | |
| 2,3-Dichlorophenol | | $_{ m ug/L}$ | 0.04 | human | USEPA WQC AO | | | 0.04 | |
| 2,4,5-Trichlorophenol | | ug/L | 1 | human | USEPA WQC AO | | | 1 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific | Water Use Ca | ategory Health R | isk Criteria |
|--|--------------------|--------------|-------|--------------------|---------------------------------|-------------------------------|--------------------|---|-------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines |
| 2,4,6- Trichlorophenol* | | ug/L | 2 | human | USEPA WQC AO | | | 2 | |
| 2,4-D | | ug/L | 4 | aquatic biota | CCME Water PAL AEP Water PAL | 4 | | 30 | |
| 2,4-DB | | ug/L | 25 | aquatic biota | AEP Water PAL | 25 | | 90 | |
| 2,4-Dichlorophenol | | ug/L | 0.3 | human | USEPA WQC AO | | | 0.3 | |
| 2,4-Dimethylphenol | | ug/L | 100 | human | HH DW+Org (US EPA) | | | 100 | |
| 2,4-Dinitrophenol | | ug/L | 10 | human | HH DW+Org (US EPA) | | | 10 | |
| $2,4\text{-Dinitrotoluene}^{\textstyle *}$ | | ug/L | 0.49 | human | HH DW+Org (US EPA) | | | 0.49 | |
| 2,5-Dichlorophenol | | ug/L | 0.5 | human | USEPA WQC AO | | | 0.5 | |
| 2,6-Dichlorophenol | | ug/L | 0.2 | human | USEPA WQC AO | | | 0.2 | |
| 2-Chloronaphthalene | | ug/L | 800 | human | HH DW+Org (US EPA) | | | 800 | |
| 2-Chlorophenol | | ug/L | 0.1 | human | USEPA WQC AO | | | 0.1 | |
| 2-Methyl-4,6- Dinitrophenol | | $_{ m ug/L}$ | 2 | human | HH DW+Org (US EPA) | | | 2 | |
| 2-Methyl-4- Chlorophenol | | $_{ m ug/L}$ | 1800 | human | USEPA WQC AO | | | 1800 | |
| 3,3'- Dichlorobenzidine | | ug/L | 0.49 | human | HH DW+Org (US EPA) | | | 0.49 | |
| 3,4-Dichlorophenol | | $_{ m ug/L}$ | 0.3 | human | USEPA WQC AO | | | 0.3 | |
| 3-Chlorophenol | | $_{ m ug/L}$ | 0.1 | human | USEPA WQC AO | | | 0.1 | |
| 3-Iodo-2-propynyl butyl carbamate | | $_{ m ug/L}$ | 1.9 | aquatic biota | CCME Water PAL AEP Water PAL | 1.9 | | | |
| 3-Methyl-4- Chlorophenol | | $_{ m ug/L}$ | 500 | human | HH DW+Org (US EPA) | | | 500 | |
| 3-Methyl-6- Chlorophenol | | $_{ m ug/L}$ | 20 | human | USEPA WQC AO | | | 20 | |
| 4-Chlorophenol | | ug/L | 0.1 | human | USEPA WQC AO | | | 0.1 | |
| Acenaphthene [§] | | $_{ m ug/L}$ | 4.79 | human | HH DW+Org (derived) | 5.8 | | 4.79 | |
| Acridine | | $_{ m ug/L}$ | 4.4 | aquatic biota | AEP Water PAL CCME Water PAL | 4.4 | | | |
| Acrolein | | $_{ m ug/L}$ | 2.87 | human | HH DW+Org (derived) | 3 | | 2.87 | |
| Acrylamide | | ug/L | 0.07 | human | HH DW+Org (derived) | | | 0.07 | |
| Acrylonitrile* | | ug/L | 0.53 | human | HH DW+Org (derived) | | | 0.53 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific Water Use Category Health Risk Criteria | | | | |
|--|--------------------|-----------------|-----------|--------------------|--|--|--------------------|---|-------------------------|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines | |
| Alachlor | | ug/L | 2 | human | US EPA DWR | | | 2 | | |
| Alcohol ethoxylates | | $_{ m ug/L}$ | 70 | aquatic biota | FEQG Water PAL | 70 | | | | |
| Aldicarb | | $_{ m ug/L}$ | 1 | aquatic biota | AEP Water PAL CCME Water PAL | 1 | 11.00 | 10 | | |
| Aldrin* | | ug/L | 0.0000077 | human | USEPA WQC HH Org HH DW+Org (US EPA) | 0.004 | 0.00 | 0.0000077 | | |
| Aldrin and dieldrin | | ug/L | 0.03 | human | WHO DW | | | 0.03 | | |
| Alkalinity, total | | $\mathrm{mg/L}$ | 20 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | 20 | | | | |
| alpha-Endosulfan | | ug/L | 1.82 | human | HH DW+Org (derived) | | | 1.82 | | |
| alpha- Hexachlorocyclohexane | * | ug/L | 0.0002 | human | HH DW+Org (derived) | | | 0.0002 | | |
| Aluminum | Total | ug/L | 18 | wildlife | US DOE Wildlife | 100 | 18.00 | 200 | | |
| Aluminum | Dissolved | $_{ m ug/L}$ | 50 | aquatic biota | AEP Water PAL | 50 | | | | |
| Ammonia | | $_{ m mg/L}$ | 0.67 | human | HH DW+Org (derived) | 0.794 | | 0.67 | | |
| Ammonia, unionized | | $_{ m mg/L}$ | 0.016 | aquatic biota | AEP Water PAL | 0.016 | | | | |
| Aniline | | $_{ m ug/L}$ | 2.2 | aquatic biota | AEP Water PAL CCME Water PAL | 2.2 | | | | |
| Anthracene | | $_{ m ug/L}$ | 0.012 | aquatic biota | CCME Water PAL AEP Water PAL | 0.012 | | 20.07 | | |
| Antimony | Total | ug/L | 4.59 | human | HH DW+Org (derived) | | 161.00 | 4.59 | 9,412 | |
| Arsenic* | Total | ug/L | 0.03 | human | HH DW+Org (derived) | 5 | 16.00 | 0.03 | 2,179 | |
| Arsenic*†† | Dissolved | $_{ m ug/L}$ | 150 | aquatic biota | US EPA Aquatic Life Criteria | 150 | | | | |
| Asbestos | | ug/L | 7 | human | US EPA DWR HH DW+Org (US EPA) | | | 7 | | |
| Atrazine | | ug/L | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL | 1.8 | 5.00 | 3 | | |
| Atrazine and its chloro-s-triazine metabolites | | ug/L | 100 | human | WHO DW | | | 100 | | |
| Azinphos-methyl | | ug/L | 0.01 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | 0.01 | | 20 | | |
| Barium | Total | $_{ m ug/L}$ | 1000 | human | HH DW+Org (US EPA) Health Canada DW | | | 1000 | 3,137,255 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific Water Use Category Health Risk Criteria | | | | |
|---------------------------------------|--------------------|--------------|---------|------------------------|---|--|--------------------|---|--------------------------|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines | |
| Benzene* | | ug/L | 2.11 | human | HH DW+Org (derived) | 40 | 2,293.00 | 2.11 | | |
| Benzidine* | | ug/L | 0.001 | human | HH DW+Org (derived) | | | 0.001 | | |
| $Benzo(a) anthracene^*\dagger$ | | $_{ m ug/L}$ | 0.001 | human | HH DW+Org (derived) | 0.018 | | 0.001 | 7,978 | |
| $Benzo(a)pyrene^*\dagger$ | | $_{ m ug/L}$ | 0.0001 | human | HH DW+Org (derived) | 0.015 | 0.01 | 0.0001 | | |
| $Benzo(b) fluoranthene^* \dagger$ | | $_{ m ug/L}$ | 0.001 | human | HH DW+Org (derived) | | | 0.001 | 15,956 | |
| $Benzo(k) fluoranthene^*\dagger$ | | ug/L | 0.01 | human | HH DW+Org (derived) | | | 0.01 | 159,565 | |
| Beryllium | Total | $_{ m ug/L}$ | 3.27 | human | HH DW+Org (derived) | | 100.00 | 3.27 | | |
| beta-Endosulfan | | ug/L | 2.87 | human | HH DW+Org (derived) | | | 2.87 | | |
| beta- Hexachlorocyclohexane | | ug/L | 0.01 | human | HH DW+Org (derived) | | | 0.01 | | |
| Bis(2-Chloro-1- methylethyl) Ether | | ug/L | 127.99 | human | HH DW+Org (derived) | | | 127.99 | | |
| Bis(2-Chloroethyl) Ether* | | ug/L | 0.25 | human | HH DW+Org (derived) | | | 0.25 | | |
| Bis(2-Ethylhexyl) Phthalate | | ug/L | 0.21 | human | HH DW+Org (derived) | | | 0.21 | | |
| Bis(Chloromethyl) Ether* | | ug/L | 0.001 | human | HH DW+Org (derived) | | | 0.001 | | |
| Bisphenol A-d6 | | $_{ m ug/L}$ | 3.5 | aquatic biota | FEQG Water PAL | 3.5 | | | | |
| Boron | Total | ug/L | 1333.33 | human | HH DW+Org (derived) | 1500 | 5,000.00 | 1333.33 | | |
| Bromacil | | ug/L | 5 | aquatic biota | AEP Water PAL CCME Water PAL | 5 | 1,100.00 | | | |
| Bromate | | ug/L | 10 | human | Health Canada DW US EPA DWR WHO DW | | | 10 | | |
| Bro- modichloromethane | | ug/L | 6.33 | human | HH DW+Org (derived) | | 100.00 | 6.33 | | |
| Bromoform | | ug/L | 7 | human | HH DW+Org (US EPA) | | 100.00 | 7 | | |
| Bromoxynil | | $_{ m ug/L}$ | 5 | aquatic biota human | AEP Water PAL CCME Water PAL Health Canada DW | 5 | 11.00 | 5 | | |
| Butylbenzyl * Phthalate | | ug/L | 0.06 | human | HH DW+Org (derived) | | | 0.06 | | |
| Cadmium [‡] | Total | ug/L | 0.002 | human | HH DW+Org (derived) | 0.18 | 0.23 | 0.002 | 3,232 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health I | Risk Criteria | Specific | Water Use C | ategory Health R | isk Criteria |
|---|--------------------|-----------------|-------|--------------------|---|-------------------------------|--------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines |
| Cadmium [‡] †† | Dissolved | ug/L | 0.824 | aquatic biota | US EPA Aquatic Life Criteria | 0.824 | | | |
| Calcium | | $_{ m mg/L}$ | 1000 | wildlife | CCME Water Ag (limited) AEP Water Ag | | 1,000.00 | | |
| Captan | | ug/L | 1.3 | aquatic biota | CCME Water PAL AEP Water PAL | 1.3 | 13.00 | | |
| Carbamazepine | | ug/L | 10 | aquatic biota | CCME Water PAL AEP Water PAL | 10 | | | |
| Carbaryl | | ug/L | 0.2 | aquatic biota | AEP Water PAL CCME Water PAL | 0.2 | 110.00 | 90 | |
| Carbofuran | | ug/L | 1.8 | aquatic biota | CCME Water PAL AEP Water PAL | 1.8 | 45.00 | 7 | |
| Carbon tetrachloride | | ug/L | 1.9 | human | HH DW+Org (derived) | 13.3 | 5.00 | 1.9 | |
| Chloramines | | $_{ m ug/L}$ | 0.5 | aquatic biota | CCME Water PAL | 0.5 | | 3000 | |
| Chlorate | | ug/L | 700 | human | WHO DW | | | 700 | |
| Chlordane | | $_{ m ug/L}$ | 0.001 | human | HH DW+Org (derived) | 0.004 | 0.01 | 0.001 | |
| Chloride | | $\mathrm{mg/L}$ | 120 | aquatic biota | CCME Water PAL AEP Water PAL | 120 | | 250 | |
| Chlorinated paraffins, long-chain, C18-C20 | | ug/L | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL | 2.4 | | | |
| Chlorinated paraffins, medium-chain, C14-C17 | | ug/L | 2.4 | aquatic biota | AEP Water PAL FEQG Water PAL | 2.4 | | | |
| Chlorinated paraffins, short-chain, C10-C13 | | ug/L | 2.4 | aquatic biota | FEQG Water PAL AEP Water PAL | 2.4 | | | |
| Chlorine | | $_{ m ug/L}$ | 0.5 | aquatic biota | AEP Water PAL | 0.5 | | 4000 | |
| Chlorine dioxide | | $_{ m ug/L}$ | 800 | human | US EPA DWR | | | 800 | |
| Chlorite | | ug/L | 700 | human | WHO DW | | | 700 | |
| Chlorobenzene | | $_{ m ug/L}$ | 1.3 | aquatic biota | AEP Water PAL | 1.3 | | 40.85 | |
| Chlorodibro- momethane | | ug/L | 8 | human | HH DW+Org (US EPA) | | | 8 | |
| Chloroform | | ug/L | 1.8 | aquatic biota | AEP Water PAL CCME Water PAL | 1.8 | 100.00 | 45.89 | |
| Chlorophenol | | $_{ m ug/L}$ | 7 | aquatic biota | AEP Water PAL CCME Water PAL | 7 | | | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific Water Use Category Health Risk Criteria | | | | |
|---|--------------------|--------------|----------|--------------------|--|--|--------------------|---|--------------------------|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines | |
| Chlorophenoxy Herbicide (2,4,5-TP) [Silvex] | | ug/L | 20.55 | human | HH DW+Org (derived) | | 100.00 | 20.55 | | |
| Chlorothalonil | | ug/L | 0.18 | aquatic biota | CCME Water PAL AEP Water PAL | 0.18 | 170.00 | | | |
| Chlorotoluron | | ug/L | 30 | human | WHO DW | | | 30 | | |
| Chlorpyrifos | | ug/L | 0.002 | aquatic biota | AEP Water PAL CCME Water PAL | 0.002 | 24.00 | 30 | | |
| Chromium | Total | $_{ m ug/L}$ | 50 | human | WHO DW Health Canada DW | | | 50 | | |
| Chromium (III) [‡] | Total | ug/L | 8.9 | aquatic biota | CCME Water PAL AEP Water PAL | 8.9 | 50.00 | 100 | | |
| Chromium (III) ‡ †† | Dissolved | ug/L | 100.92 | aquatic biota | US EPA Aquatic Life Criteria | 100.92 | | | | |
| Chromium (VI) | Total | ug/L | 1 | aquatic biota | CCME Water PAL AEP Water PAL | 1 | 50.00 | 13.47 | 941,176 | |
| Chromium (VI) | Dissolved | ug/L | 5 | aquatic biota | FEQG Water PAL | 5 | | | | |
| Chrysene*† | | $_{ m ug/L}$ | 0.07 | human | HH DW+Org (derived) | | | 0.07 | 861,820 | |
| cis-1,2- Dichloroethylene | | ug/L | 70 | human | US EPA DWR | | | 70 | | |
| Cobalt [‡] | Total | ug/L | 1.10 | aquatic biota | FEQG Water PAL AEP Water PAL | 1.10 | 1,000.00 | | | |
| Copper*; | Total | $_{ m ug/L}$ | 2.76 | aquatic biota | CCME Water PAL | 2.76 | 500.00 | 1000 | | |
| Copper | Dissolved | $_{ m ug/L}$ | 0.53 | aquatic biota | FEQG Water PAL | 0.53 | | | | |
| Cyanazine | | $_{ m ug/L}$ | 0.6 | human | WHO DW | 2 | 10.00 | 0.6 | | |
| Cyanide | | ug/L | 3.62 | human | HH DW+Org (derived) | 5 | 369,092.00 | 3.62 | | |
| Cyanobacterial toxins | | ug/L | 1.5 | human | Health Canada DW | | | 1.5 | | |
| Dalapon | | $_{ m ug/L}$ | 200 | human | US EPA DWR | | | 200 | | |
| DDT and metabolites* | | ug/L | 0.000004 | wildlife | US DOE Wildlife | 0.001 | 0.00 | 0.0003 | | |
| Deltamethrin | | ug/L | 0.0004 | aquatic biota | AEP Water PAL CCME Water PAL | 0.0004 | 2.50 | | | |
| Demeton | | $_{ m ug/L}$ | 0.1 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | 0.1 | | | | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific | Water Use Ca | ategory Health R | isk Criteria |
|--------------------------------------|--------------------|--------------|---------|--------------------|--|-------------------------------|--------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines |
| Di(2-ethylhexyl) adipate | | $_{ m ug/L}$ | 400 | human | US EPA DWR | | | 400 | |
| Di(2-ethylhexyl) phthalate | | ug/L | 6 | human | US EPA DWR | 16 | | 6 | |
| Di-n-Butyl Phthalate | | ug/L | 0.15 | wildlife | US DOE Wildlife | 19 | 0.15 | 1.42 | |
| Diazinon | | ug/L | 0.17 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | 0.17 | | 20 | |
| $Dibenzo(a,h) anthracen \varepsilon$ | | ug/L | 0.0001 | human | HH DW+Org (derived) | | | 0.0001 | 2,518 |
| Dibromoacetonitrile | | $_{ m ug/L}$ | 70 | human | WHO DW | | | 70 | |
| Dibro- mochloromethane | | ug/L | 5.21 | human | HH DW+Org (derived) | | 100.00 | 5.21 | |
| Dicamba | | ug/L | 10 | aquatic biota | CCME Water PAL AEP Water PAL | 10 | 122.00 | 120 | |
| Dichloroacetate | | ug/L | 50 | human | WHO DW | | | 50 | |
| ${\bf Dichloroacetonitrile}^*$ | | $_{ m ug/L}$ | 20 | human | WHO DW | | | 20 | |
| Dichlorobro- momethane | | $_{ m ug/L}$ | 9.5 | human | HH DW+Org (US EPA) | | 100.00 | 9.5 | |
| ${\bf Dichloromethane}^{*}$ | | $_{ m ug/L}$ | 5 | human | US EPA DWR | | 50.00 | 5 | |
| Dichlorophenol | | $_{ m ug/L}$ | 0.2 | aquatic biota | CCME Water PAL AEP Water PAL | 0.2 | | | |
| Dichlorprop | | $_{ m ug/L}$ | 100 | human | WHO DW | | | 100 | |
| Diclofop-methyl | | $_{ m ug/L}$ | 6.1 | aquatic biota | AEP Water PAL CCME Water PAL | 6.1 | 9.00 | 9 | |
| Didecyl dimethyl ammonium chloride | | ug/L | 1.5 | aquatic biota | CCME Water PAL AEP Water PAL | 1.5 | | | |
| Dieldrin | | $_{ m ug/L}$ | 0.00001 | human | HH DW+Org (derived) HH DW+Org (US EPA) | 0.004 | 0.00 | 0.00001 | |
| Diethanolamine | | $_{ m ug/L}$ | 450 | aquatic biota | AEP Water PAL | 450 | | | |
| Diethyl Phthalate | | ug/L | 35.61 | human | HH DW+Org (derived) | | 210,561.00 | 35.61 | |
| Diethylene glycol | | ug/L | 150000 | aquatic biota | AEP Water PAL | 150000 | | | |
| Diisopropanolamine | | ug/L | 1600 | aquatic biota | AEP Water PAL CCME Water PAL | 1600 | | | |
| Dimethoate | | ug/L | 3 | wildlife | CCME Water Ag (limited) AEP Water Ag | 6.2 | 3.00 | 6 | |
| Dimethyl Phthalate | | ug/L | 102.91 | human | HH DW+Org (derived) | | | 102.91 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific | Water Use Ca | tegory Health R | isk Criteria |
|---|--------------------|--------------|---------------|--------------------|---|-------------------------------|--------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value S | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines |
| Dinitrophenols | | ug/L | 10 | human | HH DW+Org (US EPA) | | | 10 | |
| Dinoseb | | ug/L | 0.05 | aquatic biota | CCME Water PAL AEP Water PAL | 0.05 | 150.00 | 7 | |
| Dioxin $(2,3,7,8\text{-TCDD})$ | | ug/L | 0.00000002134 | wildlife | US DOE Wildlife | | 0.00 | 0.00000005 | |
| Diquat | | ug/L | 20 | human | US EPA DWR | | | 20 | |
| Diuron | | $_{ m ug/L}$ | 150 | human | Health Canada DW | | | 150 | |
| Edetic acid | | ug/L | 600 | human | WHO DW | | | 600 | |
| Endosulfan | | ug/L | 0.003 | aquatic biota | AEP Water PAL CCME Water PAL | 0.003 | 1.00 | | |
| Endosulfan Sulfate | | ug/L | 2.63 | human | HH DW+Org (derived) | | | 2.63 | |
| Endothall | | $_{ m ug/L}$ | 100 | human | US EPA DWR | | | 100 | |
| Endrin | | ug/L | 0.001 | wildlife | US DOE Wildlife | 0.002 | 0.00 | 0.01 | |
| Endrin Aldehyde | | ug/L | 0.11 | human | HH DW+Org (derived) | | | 0.11 | |
| Epichlorohydrin | | ug/L | 0.4 | human | WHO DW | | | 0.4 | |
| Ethanol | | $_{ m ug/L}$ | 123377 | wildlife | US DOE Wildlife | | 123,377.00 | | |
| Ethinyl estradiol | | $_{ m ng/L}$ | 0.5 | aquatic biota | AEP Water PAL | 0.5 | | | |
| Ethyl acetate | | | 136465 | wildlife | US DOE Wildlife | | 136,465.00 | | |
| Ethylbenzene | | ug/L | 2.4 | wildlife | AEP Water Ag CCME Water Ag (limited) | 90 | 2.40 | 8.54 | |
| Ethylene dibromide | | $_{ m ug/L}$ | 0.05 | human | US EPA DWR | | | 0.05 | |
| Ethylene glycol | | ug/L | 192000 | aquatic biota | AEP Water PAL CCME Water PAL | 192000 | | | |
| Fenoprop | | ug/L | 9 | human | WHO DW | | | 9 | |
| Fluoranthene | | ug/L | 0.04 | aquatic biota | AEP Water PAL CCME Water PAL | 0.04 | | 1.09 | |
| Fluorene§ | | ug/L | 3 | aquatic biota | AEP Water PAL CCME Water PAL | 3 | | 6.98 | |
| Fluoride | | $_{ m mg/L}$ | 0.12 | aquatic biota | CCME Water PAL | 0.12 | 1.00 | 0.4 | |
| Formaldehyde | | ug/L | 73910 | wildlife | US DOE Wildlife | | 73,910.00 | | |
| gamma- Hexachlorocyclohexan [Lindane] | e | ug/L | 0.01 | aquatic biota | AEP Water PAL | 0.01 | 4.00 | 0.2 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific | Water Use Ca | ategory Health R | isk Criteria |
|---|--------------------|--------------|---------|------------------------|---|-------------------------------|--------------------|---|-------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines |
| Glyphosate | | ug/L | 280 | human wildlife | AEP Water Ag Health Canada DW CCME Water Ag (limited) | 800 | 280.00 | 280 | |
| Haloacetic acids | | ug/L | 60 | human | US EPA DWR | | | 60 | |
| heptaBDE | | $_{ m ng/L}$ | 14 | aquatic biota | FEQG Water PAL | 14 | | | |
| Heptachlor* | | ug/L | 0.00004 | human | HH DW+Org (derived) | 0.0038 | 0.00 | 0.00004 | |
| Heptachlor epoxide* | | ug/L | 0.0001 | human | HH DW+Org (derived) | 0.0038 | | 0.0001 | |
| hexaBDE | | $_{ m ng/L}$ | 120 | aquatic biota | FEQG Water PAL AEP Water PAL | 120 | | | |
| Hexabromocyclodo- decane | | $_{ m ug/L}$ | 0.56 | aquatic biota | FEQG Water PAL AEP Water PAL | 0.56 | | | |
| Hexachlorobenzene* | | ug/L | 0.0001 | human | HH DW+Org (derived) | | 0.52 | 0.0001 | |
| ${\bf Hexachlorobutadiene}^*$ | | $_{ m ug/L}$ | 0.001 | human | HH DW+Org (derived) | 1.3 | | 0.001 | |
| Hexachlorocyclohexane | | ug/L | 0.01 | aquatic biota human | HH DW+Org (derived) CCME Water PAL | 0.01 | | 0.01 | |
| Hexachlorocyclopen- tadiene | | $_{ m ug/L}$ | 0.4 | human | HH DW+Org (derived) | | | 0.4 | |
| Hexachloroethane* | | ug/L | 0.02 | human | HH DW+Org (derived) | | | 0.02 | |
| Hydrazine | | $_{ m ug/L}$ | 2.6 | aquatic biota | FEQG Water PAL AEP Water PAL | 2.6 | | | |
| Hydrogen Sulfide | | ug/L | 2 | aquatic biota | US EPA Aquatic Life Criteria | 2 | | | |
| Hydroxyatrazine | | $_{ m ug/L}$ | 200 | human | WHO DW | | | 200 | |
| Imidacloprid | | ug/L | 0.23 | aquatic biota | AEP Water PAL CCME Water PAL | 0.23 | | | |
| Indeno(1,2,3-cd)pyrene*† | | $_{ m ug/L}$ | 0.001 | human | HH DW+Org (derived) | | | 0.001 | 41,323 |
| Inorganic nitrogen (nitrate and nitrite) | Dissolved | $_{ m mg/L}$ | 100 | wildlife | CCME Water Ag (limited) AEP Water Ag | | 100.00 | | |
| Iron | Total | ug/L | 300 | aquatic biota human | CCME Water PAL USEPA WQC AO | 300 | | 300 | |
| Iron | Dissolved | ug/L | 300 | aquatic biota | AEP Water PAL | 300 | | | |
| Isophorone* | | $_{ m ug/L}$ | 268.41 | human | HH DW+Org (derived) | | | 268.41 | |
| Isoproturon | | ug/L | 9 | human | WHO DW | | | 9 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific | Water Use Ca | tegory Health R | isk Criteria |
|----------------------------------|--------------------|--------------|--------|--------------------|--|-------------------------------|--------------------|---|-------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines |
| $\operatorname{Lead}^{\ddagger}$ | Total | $_{ m ug/L}$ | 4.01 | aquatic biota | AEP Water PAL CCME Water PAL | 4.01 | 100.00 | 5 | 7,320,261 |
| Lead‡ †† | Dissolved | ug/L | 3.07 | aquatic biota | US EPA Aquatic Life Criteria | 3.07 | | | |
| Linuron | | ug/L | 7 | aquatic biota | CCME Water PAL AEP Water PAL | 7 | | | |
| m-Dichlorobenzene | | ug/L | 150 | aquatic biota | CCME Water PAL | 150 | | | |
| Malathion | | $_{ m ug/L}$ | 0.1 | aquatic biota | AEP Water PAL US EPA Aquatic Life Criteria | 0.1 | | 190 | |
| Manganese | Total | ug/L | 50 | human | HH DW+Org (US EPA) | 470 | | 50 | |
| MCPA | | $_{ m ug/L}$ | 2.6 | aquatic biota | CCME Water PAL AEP Water PAL | 2.6 | 25.00 | 100 | |
| Mecoprop | | ug/L | 10 | human | WHO DW | 13 | | 10 | |
| Mercury | Total | ug/L | 0.0016 | wildlife | US DOE Wildlife | 0.005 | 0.00 | 1 | 18,824 |
| $Mercury^{\dagger\dagger}$ | Dissolved | ug/L | 0.77 | aquatic biota | US EPA Aquatic Life Criteria | 0.77 | | | |
| Mercury (methyl) | Total | ug/L | 0.001 | aquatic biota | AEP Water PAL | 0.001 | | 0.67 | |
| Mercury (methyl) | Dissolved | ug/L | 0.004 | aquatic biota | CCME Water PAL | 0.004 | | | |
| Methanol | | ug/L | 1500 | aquatic biota | AEP Water PAL | 1500 | 230,691.00 | | |
| Methoprene | | ug/L | 0.09 | aquatic biota | AEP Water PAL CCME Water PAL | 0.09 | | | |
| Methoxychlor | | ug/L | 0.001 | human | HH DW+Org (derived) | 0.03 | 1.00 | 0.001 | |
| Methyl Bromide | | ug/L | 100 | human | HH DW+Org (US EPA) | | | 100 | |
| Methyl tert-butyl ether | | ug/L | 10 | aquatic biota | AEP Water PAL | 10 | | | |
| Methylene chloride* | | ug/L | 32.62 | human | HH DW+Org (derived) | 98.1 | 3,990.00 | 32.62 | |
| Metolachlor | | $_{ m ug/L}$ | 7.8 | aquatic biota | AEP Water PAL CCME Water PAL | 7.8 | 50.00 | 10 | |
| Metribuzin | | $_{ m ug/L}$ | 1 | aquatic biota | AEP Water PAL CCME Water PAL | 1 | 80.00 | 80 | |
| Microcystin-LR | | ug/L | 1 | human | WHO DW | | | 1 | |
| Mirex | | $_{ m ug/L}$ | 0.001 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | 0.001 | | | |
| Molinate | | ug/L | 6 | human | WHO DW | | | 6 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific Water Use Category Health Risk Criteria | | | | |
|---|--------------------|-----------------|--------|--------------------|---|--|--------------------|---|-------------------------|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines | |
| Molybdenum | Total | ug/L | 33.33 | human | HH DW+Org (derived) | 73 | 500.00 | 33.33 | | |
| Monochloramine | | ug/L | 3000 | human | WHO DW | | | 3000 | | |
| Monochloroacetate | | ug/L | 20 | human | WHO DW | | | 20 | | |
| Monochlorobenzene | | $_{ m ug/L}$ | 1.3 | aquatic biota | CCME Water PAL AEP Water PAL | 1.3 | | 20 | | |
| Monoethanolamine | | ug/L | 75 | aquatic biota | AEP Water PAL | 75 | | | | |
| N-Nitrosodi-n- Propylamine | | ug/L | 0.05 | human | HH DW+Org (US EPA) HH DW+Org (derived) | | | 0.05 | | |
| $\begin{array}{l} \text{N-} \\ \text{Nitrosodimethylamine}^* \end{array}$ | | ug/L | 0.007 | human | HH DW+Org (US EPA) | | | 0.007 | | |
| $\begin{array}{l} \text{N-} \\ \text{Nitrosodiphenylamine}^* \end{array}$ | | ug/L | 33 | human | HH DW+Org (US EPA) | | | 33 | | |
| Naphthalene§ | | ug/L | 1 | aquatic biota | AEP Water PAL | 1 | | 133.33 | | |
| Naphthenic acids (Lower Athabasca River) | Total | $_{ m ug/L}$ | < 0.05 | | Adopted current condition (OSM Reporting Limit) | | | | | |
| Naphthenic acids (Athabasca River Delta) | Total | ug/L | 230 | | Adopted current condition (50th percentile, high flow) | | | | | |
| Naphthenic acids (Lake Athabasca) | Total | $_{ m ug/L}$ | 140 | | Adopted current condition (50th percentile, open water) | | | | | |
| Nickel [‡] | Total | ug/L | 7.35 | human | HH DW+Org (derived) | 60.86 | 1,000.00 | 7.35 | 1,470,588 | |
| Nickel [‡] †† | Dissolved | ug/L | 60.68 | aquatic biota | US EPA Aquatic Life Criteria | 60.68 | | | | |
| Nitrate | Dissolved | $\mathrm{mg/L}$ | 3 | aquatic biota | CCME Water PAL AEP Water PAL | 3 | | 10 | | |
| Nitrilotriacetic acid | | ug/L | 200 | human | WHO DW | | | 200 | | |
| Nitrite | Dissolved | $_{ m mg/L}$ | 0.06 | aquatic biota | CCME Water PAL | 0.06 | 10.00 | 0.912 | | |
| Nitrobenzene | | ug/L | 9.72 | human | HH DW+Org (derived) | | | 9.72 | | |
| Nitrosamines | | ug/L | 0.008 | human | HH DW+Org (US EPA) | | | 0.008 | | |
| Nitrosodibutylamine | | ug/L | 0.05 | human | HH DW+Org (derived) | | | 0.05 | | |
| Nitrosodiethylamine | | ug/L | 0.002 | human | HH DW+Org (derived) | | | 0.002 | | |
| Nitrosopyrrolidine | | $_{ m ug/L}$ | 0.16 | human | HH DW+Org (US EPA) HH DW+Org (derived) | | | 0.16 | | |
| Nonylphenol | | ug/L | 6.6 | aquatic biota | US EPA Aquatic Life Criteria | 6.6 | | | | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health Risk Criteria | | | Specific Water Use Category Health Risk Criteria | | | | |
|---|--------------------|--------------|---------|------------------------------|--|-------------------------------|--|---|--------------------------|--|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines | | |
| Nonylphenol and its ethoxylates | | ug/L | 1 | aquatic biota | CCME Water PAL | 1 | | | | | |
| o-Dichlorobenzene | | ug/L | 0.7 | aquatic biota | AEP Water PAL CCME Water PAL | 0.7 | | 200 | | | |
| octaBDE | | $_{ m ng/L}$ | 14 | aquatic biota | FEQG Water PAL | 14 | | | | | |
| Oxamyl (Vydate) | | ug/L | 200 | human | US EPA DWR | | | 200 | | | |
| p,p - Dichlorodiphenyldichlor (DDD)* | roethane | ug/L | 0.001 | human | HH DW+Org (US EPA) | | | 0.001 | | | |
| p,p - Dichlorodiphenyldichlor $\left(\mathrm{DDE}\right)^*$ | | ug/L | 0.00018 | human | USEPA WQC HH Org | | | 0.00018 | | | |
| p-Dichlorobenzene | | ug/L | 5 | human | Health Canada DW | 26 | | 5 | | | |
| Paraquat | | ug/L | 10 | human | Health Canada DW | | | 10 | | | |
| Parathion | | ug/L | 0.013 | aquatic biota | US EPA Aquatic Life Criteria AEP Water PAL | 0.013 | | | | | |
| Pendimethalin | | $_{ m ug/L}$ | 20 | human | WHO DW | | | 20 | | | |
| pentaBDE | | $_{ m ng/L}$ | 0.2 | aquatic biota | AEP Water PAL FEQG Water PAL | 0.2 | | | | | |
| pentaBDE (BDE-100) | | $_{ m ng/L}$ | 0.2 | aquatic biota | FEQG Water PAL AEP Water PAL | 0.2 | | | | | |
| pentaBDE (BDE-99) | | ng/L | 4 | aquatic biota | AEP Water PAL FEQG Water PAL | 4 | | | | | |
| Pentachlorobenzene | | $_{ m ug/L}$ | 0.01 | human | HH DW+Org (derived) | 6 | | 0.01 | | | |
| Pentachloronitroben- zene | | | 4 | wildlife | US DOE Wildlife | | 4.00 | | | | |
| Pentachlorophenol | | ug/L | 0.1 | human | HH DW+Org (derived) | 0.5 | 0.28 | 0.1 | | | |
| Perchlorate | | ug/L | 70 | human | WHO DW | | | 70 | | | |
| Perfluorooctanesul- fonate | | ug/L | 0.6 | human | Health Canada DW | 6.8 | | 0.6 | | | |
| Perfluorooctanoic acid | | ug/L | 0.2 | human | Health Canada DW | | | 0.2 | | | |
| Permethrin | | $_{ m ug/L}$ | 0.004 | aquatic biota | AEP Water PAL CCME Water PAL | 0.004 | | | | | |
| | | | | | | | | | | | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health I | Specific Water Use Category Health Risk Criteria | | | | |
|--------------------------------------|--------------------|--------------|---------|---------------------------------|--|-------------------------------|--------------------|---|-------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditiona Medicines |
| рН | | pH units | 7-9 | aquatic biota human human | US EPA Aquatic Life Criteria HH DW+Org (US EPA) AEP Water PAL CCME Water PAL Health Canada DW | 6.5-9 | | 7-9 | |
| Phenanthrene§ | | ug/L | 0.4 | aquatic biota | CCME Water PAL AEP Water PAL | 0.4 | | 200 | |
| Phenol | | ug/L | 2 | wildlife | CCME Water Ag (limited) AEP Water Ag | 4 | 2.00 | 300 | |
| Phorate | | ug/L | 2 | human | Health Canada DW | | | 2 | |
| Picloram | | ug/L | 29 | aquatic biota | CCME Water PAL AEP Water PAL | 29 | 190.00 | 190 | |
| Polychlorinated Biphenyls (PCBs)* | | ug/L | 0.00064 | human | USEPA WQC HH Org | 0.001 | | 0.00064 | |
| Propylene glycol | | $_{ m ug/L}$ | 500000 | aquatic biota | CCME Water PAL AEP Water PAL | 500000 | | | |
| Pyrene§ | | $_{ m ug/L}$ | 0.025 | aquatic biota | CCME Water PAL AEP Water PAL | 0.025 | | 1.43 | |
| Quinoline | | ug/L | 3.4 | aquatic biota | AEP Water PAL CCME Water PAL | 3.4 | | | |
| Selenium | Total | ug/L | 0.24 | wildlife | US DOE Wildlife | 1 | 0.24 | 18.77 | 735,294 |
| Silver | Total | ug/L | 0.25 | aquatic biota | AEP Water PAL CCME Water PAL | 0.25 | | 33.33 | |
| Simazine | | $_{ m ug/L}$ | 2 | human | WHO DW | 10 | 10.00 | 2 | |
| Sodium dichloroisocyanurate | | ug/L | 40000 | human | WHO DW | | | 40000 | |
| Solids Dissolved and Salinity | | ug/L | 250000 | human | HH DW+Org (US EPA) | | | 250000 | |
| Strontium | Total | ug/L | 4000 | human | HH DW+Org (derived) | | | 4000 | |
| Styrene | | $_{ m ug/L}$ | 20 | human | WHO DW | 72 | | 20 | |
| Sulfate | | $_{ m mg/L}$ | 250 | human | WHO DW | 309 | 1,000.00 | 250 | |
| Sulfide | | $_{ m mg/L}$ | 0.0019 | aquatic biota | AEP Water PAL | 0.0019 | | | |
| Sulfolane | | ug/L | 50 | aquatic biota | AEP Water PAL | 50 | | | |
| Tebuthiuron | | ug/L | 1.6 | aquatic biota | CCME Water PAL | 1.6 | 130.00 | | |
| Terbufos | | ug/L | 1 | human | Health Canada DW | | | 1 | |
| Terbuthylazine | | ug/L | 7 | human | WHO DW | | | 7 | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health | Risk Criteria | Specific Water Use Category Health Risk Criteria | | | | |
|---|--------------------|-------------------------|--------|--------------------|---|--|--------------------|---|--------------------------|--|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines | |
| tetraBDE | | $_{ m ng/L}$ | 24 | aquatic biota | FEQG Water PAL AEP Water PAL | 24 | | | | |
| $\begin{array}{c} {\rm Tetrabromobisphenol} \\ {\rm A} \end{array}$ | | $_{ m ug/L}$ | 3.1 | aquatic biota | FEQG Water PAL AEP Water PAL | 3.1 | | | | |
| Tetrachloroethane | | ug/L | 13.3 | aquatic biota | CCME Water PAL | 13.3 | | | | |
| ${\it Tetrachloroethylene}^*$ | | ug/L | 4.48 | human | HH DW+Org (derived) | 110 | 48.00 | 4.48 | | |
| Tetrachlorophenol | | ug/L | 1 | aquatic biota | CCME Water PAL AEP Water PAL | 1 | | | | |
| Thallium | Total | ug/L | 0.02 | human | HH DW+Org (derived) | 0.8 | 1.00 | 0.02 | 4,000 | |
| Toluene | | ug/L | 0.5 | aquatic biota | AEP Water PAL | 0.5 | 24.00 | 57 | | |
| Total dissolved solids | | mg/L | 3000 | wildlife | AEP Water Ag CCME Water Ag (limited) | | 3,000.00 | | | |
| Toxaphene | | $_{ m ug/L}$ | 0.0002 | aquatic biota | US EPA Aquatic Life Criteria | 0.0002 | 1.00 | 0.001 | | |
| Toxicity $(acute)^{\dagger\dagger}*$ | | Toxic Units (TUa) | 0.3 | aquatic biota | AEP Water PAL | 0.3 | | | | |
| Toxicity (chronic) ^{††} ** | | Toxic Units (TUc) | 1 | aquatic biota | AEP Water PAL | 1 | | | | |
| trans-1,2- Dichloroethylene | | ug/L | 100 | human | US EPA DWR | | | 100 | | |
| Triallate | | $_{ m ug/L}$ | 0.24 | aquatic biota | CCME Water PAL AEP Water PAL | 0.24 | 230.00 | | | |
| triBDE | | $_{ m ng/L}$ | 46 | aquatic biota | AEP Water PAL FEQG Water PAL | 46 | | | | |
| Tribromomethane | | $_{ m ug/L}$ | 100 | wildlife | CCME Water Ag (limited) | | 100.00 | | | |
| Tributyltin | | $_{ m ug/L}$ | 0.008 | aquatic biota | CCME Water PAL | 0.008 | 250.00 | | | |
| Trichlorfon | | $_{ m ug/L}$ | 0.009 | aquatic biota | AEP Water PAL CCME Water PAL | 0.009 | | | | |
| Trichloroacetate | | $_{ m ug/L}$ | 200 | human | WHO DW | | | 200 | | |
| ${\it Trichloroethylene}^*$ | | $_{ m ug/L}$ | 1.38 | human | HH DW+Org (derived) | 21 | 22.00 | 1.38 | | |
| Trichlorophenol | | $_{ m ug/L}$ | 18 | aquatic biota | AEP Water PAL CCME Water PAL | 18 | | | | |
| Triclosan | | $_{ m ug/L}$ | 0.47 | aquatic biota | FEQG Water PAL | 0.47 | | | | |
| Tricyclohexyltin | | ug/L | 250 | wildlife | CCME Water Ag (limited) AEP Water Ag | | 250.00 | | | |
| | | | | | | | | | | |

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Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. (continued)

| | | | | Generic health I | Risk Criteria | Specific | Water Use C | ategory Health F | isk Criteria |
|-------------------------------|--------------------|--------------|--------|--------------------|---|-------------------------------|--------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquatic Ecosytem Health | Wildlife Health | Traditional Foods and Drinking Water | Traditional Medicines |
| Triethylene glycol | | $_{ m ug/L}$ | 350000 | aquatic biota | AEP Water PAL | 350000 | | | |
| Trifluralin | | ug/L | 0.2 | aquatic biota | AEP Water PAL CCME Water PAL | 0.2 | 45.00 | 20 | |
| Trihalomethanes | | $_{ m ug/L}$ | 80 | human | US EPA DWR | | | 80 | |
| Triphenyltin | | ug/L | 0.022 | aquatic biota | CCME Water PAL AEP Water PAL | 0.022 | 820.00 | | |
| Uranium | Total | ug/L | 15 | aquatic biota | CCME Water PAL AEP Water PAL | 15 | 200.00 | 20 | |
| Vanadium | Total | ug/L | 100 | wildlife | AEP Water Ag CCME Water Ag (limited) | 120 | 100.00 | | |
| Vinyl chloride* | | ug/L | 0.18 | human | HH DW+Org (derived) | | 78.00 | 0.18 | |
| Xylene | | ug/L | 28 | wildlife | US DOE Wildlife | 30 | 28.00 | 90 | |
| Xylenes (total) | | $_{ m ug/L}$ | 10000 | human | US EPA DWR | | | 10000 | |
| $Zinc^{\ddagger}$ | Total | ug/L | 12.72 | human | HH DW+Org (derived) | 30 | 30.00 | 12.72 | 588,000,000,000,000 |
| Zinc^{\ddagger} | Dissolved | $_{ m ug/L}$ | 31.35 | aquatic biota | CCME Water PAL | 31.35 | | | |
| Low Moelcular Weight PAHs¶ | | | | | | | | | |

Table 3.9: Summary of Generic and Use Specific Health Risk Criteria for protection of Indigenous water use. *(continued)*

| | | | Generic health Risk Criteria | | Spec | ific Water Use (| Category Health F | Risk Criteria | |
|-----------|--------------------|-------|------------------------------|--------------------|--------|----------------------------|-------------------|---|--------------------------|
| Parameter | Sample Fraction | Units | Value | Sensitive Receptor | Source | Aquati Ecosyt Health | | Traditional Foods and Drinking Water | Traditional Medicines |

High Molecular Weight PAHs**

Note:

HH DW + Org and Org were adjusted to reflect carcinogenity of 1 in 1000,000 (1 x 10⁻⁵) ILCR levels (Alberta Health (2019))

HH DW+Org: Human Health (HH) criteria from consuming surface water (SW) and aquatic organisms (O)

AO; Aesthetic Objectives, DW; Drinking Water; PAL; Protection of Aquatic Life, Ag; Agriculture

Aquatic biota; invertebrates, plants and fish

Wildlife; bird and mammalian species

*Known human carcinogen via oral exposure route (Health Canada (2021))

[†] The following known human carcinogens and must be converted to Provisional Benzo[a]pyrene RPF and summed as per Health Canada (2021) then compared to the Benzo(a)pyrene and equivalents health risk criteria: Anthanthrene, Benzo[c]chrysene, Benzo[c]chrysene, Benzo[c]phenanthrene, Cyclopenta[c,d]pyrene, Dibenzo[a,e]fluoranthene Dibenzo[a,e]pyrene, Dibenzo[a,h]pyrene, Dibenzo[a,i]pyrene, Dibenzo[a,l]pyrene, 9,10- Dimethylanthracene, 7,12- Dimethylbenzo[a]anthracene, 1,2- Dimethylbenzo[a]pyrene, 1,6- Dimethylbenzo[a]pyrene, 3,6- Dimethylbenzo[a]pyrene, 4,5- Dimethylbenzo[a]pyrene, 5,6- Dimethylchrysene, 5,7- Dimethylchrysene, 5,11- Dimethylchrysene, 1,4- Dimethylphenanthrene, 4,10- Dimethylphenanthrene, 5- Ethylchrysene, Fluoranthene, 7- Methylbenzo[a]anthracene, Methylbenzo[a]anthracene, 9- Methylbenzo[a]anthracene, 12- Methylbenzo[a]anthracene, 11- Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, Methylbenzo[a]pyrene, 12- Methylbenzo[a]pyrene, 5- Methylchrysene, Methylchrysene, 2- Methylfluoranthene, Phenanthren, 2,9,10- Trimethylanthracene, 2,3,9,10- Tetramethylanthracene, Methylchrysene, Methylchrysene, 2- Methylfluoranthene, Phenanthren, 2,9,10- Trimethylanthracene, 2,3,9,10- Tetramethylanthracene, Methylchrysene, Methylchrysene, 2- Methylfluoranthene, Phenanthren, 2,9,10- Trimethylanthracene, 2,3,9,10- Tetramethylanthracene, Methylchrysene, Methylchrysene, Methylchrysene, Methylchrysene, 2- Methylfluoranthene, Phenanthren, 2,9,10- Trimethylanthracene, 2,3,9,10- Tetramethylanthracene, Methylchrysene, Methylchrys

[‡] Calculated using modifying factors presented in Table reftab:table4.

[§] Sum identified LMW PAH congeners (Anthracene, Acenaphthene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) and compare to Naphthalene health risk criteria (adopted as surrogate) (CCME (2010))

[¶]Sum identified LMW PAH congeners (Anthracene, Acenapthene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) (CCME (2010))

^{**}Sum of identified HMW PAH congeners (Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene) (CCME (2010))

 $^{^{\}dagger\dagger}$ Comparison of water quality data must be presented for both Dissolved and total fractions

 $^{^{\}ddagger \ddagger}$ Toxic Unit-Acute (TUa) is the reciprocal of the effluent concentration (i.e., TUa = 100/LC50) that causes 50 percent of the organisms to die by the end of an acute toxicity test (US EPA (2000c))

^{§§} Toxic Unit-Chronic (TUc) is the reciprocal of the effluent concentration (e.g., TUc = 100/NOEC) that causes no observable effect (NOEC) on the test organisms by the end of a chronic toxicity test (US EPA (2000c)).

Chapter 4

Health Risk Criteria for the

Protection of Sediment to

Support Indigenous Use

- MANDY L. OLSGARD MSC, P. BIOL.
- INTEGRATED TOXICOLOGY SOLUTIONS

4.1 Introduction

- Traditional knowledge of Indigenous communities and modern science both recognize sedi-
- ment as a critical and sustaining component within aquatic ecosystems. Sediments provide 2124
- 2125 substrates for aquatic plants and animals to live and reproduce in, nutrients and minerals that
- maintain local and downstream ecosystems, and through physicochemical processes act as sinks and sources for chemical substances (Palmer, 1997). More recently the role of sediment in sup-2127
- porting ecosystem function has been considered in assessments of ecosystem services (Apitz, 2128
- 2012). 2129

- The Peace Athabasca Delta (PAD), a culturally important area upon which ACFN and 2130
- MCFN cultures and livelihoods depend, was formed through the deposition of sediments, and 2131
- 2132 is sustained by this natural cycle (McLachlan, 2014; Candler et al., 2010).
- Chemicals which enter the aquatic ecosystem (either through natural or human activity) 2133
- may partition into the particulate phase depositing into bed sediments and potentially accumu-2134
- 2135 lating over time (CCME, 2001). As a result, these aquatic systems may act as both a long-term
- sink exposing those organisms living in or having direct contact to potentially harmful levels 2136

$CHAPTER\ 4.\ \ HEALTH\ RISK\ CRITERIA\ FOR\ SEDIMENT\ TO\ PROTECT\ INDIGENOUS\ USE 247$

| 2137 | of contamination and act as a continued source of contamination into the water column. |
|------|--|
| 2138 | As sediments are a crucial component of the aquatic ecosystem, effective assessment of |
| 2139 | sediment quality is necessary to evaluate the potential for adverse effects. Sediment quality |
| 2140 | guidelines provide one such method of evaluating the relationship between chemical concentra- |
| 2141 | tions in sediment and the potential for adverse effects in exposed benthic organisms and plants |
| 2142 | and contamination of overlaying water. |
| 2143 | In Alberta, sediment quality guidelines were primarily adopted from the Canadian Council |
| 2144 | of Ministers of the Environment (CCME), Ontario Ministry of the Environment and Energy |
| 2145 | (OMOEE) with select values sourced from Environment Canada (GoA, 2018). |
| 2146 | Derivation of the CCME Interim Sediment Quality Guidelines (ISQGs) and Probable Effect |
| 2147 | Levels (PELs) was limited by availability of toxicity data and available methodology which |
| 2148 | could consider bioaccumulation of contaminants within food webs. |
| 2149 | These limitations in conjunction with the lack of a recent review and modification to in- |
| 2150 | corporate scientific advancements in sediment toxicity testing may limit the protectiveness of |
| 2151 | GOA and CCME sediment quality guidelines (ISQGs and PELs) for Indigenous water use as |
| 2152 | described in Chapter 3. |
| 2153 | Similar to the water quality criteria developed for Indigenous uses (Table 3.9), Health risk |
| 2154 | sediment quality criteria (SQCs) are required to assess risks to benthic and aquatic inverte- |
| 2155 | brates from contaminants which partition to and may accumulate in sediments from natural |
| 2156 | sources and in surface water receiving OSMW seepage and releases. |
| 2157 | The proposed SQCs are applicable to aquatic environments receiving oil sands mine water |
| 2158 | releases and closure features on oil sands mines (i.e., wetlands, end pit lakes) and can also be |
| 2159 | used to assess the performance of tailings treatment technologies if the treated tailings are to |
| 2160 | be placed in contact with sediments or used to create tailings substrates within aquatic closure |
| 2161 | features. |
| 2162 | The SQC provides a mechanism by which Indigenous communities, government, regulatory |
| 2163 | and industry stakeholders can gauge the potential for adverse effects and through a weight of |
| 2164 | evidence approach, determine logical next steps in addressing the contaminant situation. |
| 2165 | The identified SQCs supplement the Indigenous water use category health risk criteria |
| 2166 | identified in Chapter 3 and application of both criteria form an ecosystem management system |
| 2167 | which considers the protection of Indigenous water use. |

2168 4.2 Objective

- 2169 Review published regulatory guidelines, sediment toxicity data, and guideline derivation meth-
- 2170 ods to identify and when required, derive new, health risk criteria that consider risks to benthic
- 2171 and aquatic biota from partitioning and accumulation of chemicals in sediments and uptake
- 2172 through the aquatic food web.

4.3 Methods

- 2174 The following stages were used to identify and/ or modify existing sediment quality guidelines
- 2175 and when required derive SQCs.
- Identify benthic and aquatic biota sediment exposure pathways for contaminants and
- 2177 community protection goals,
- Identify substances of concern in oil sands mine water and tailings which may partition
- to and accumulate in receiving water body sediments,
- Review and evaluate available sediment quality guidelines by applying criteria that con-
- sider protection of benthic and aquatic biota (biodiversity and toxicity) and biomagnifi-
- cation in aquatic food webs,
- Adopt available sediment quality guidelines as SQCs, when health risks were considered,
- 2184 01
- Identify sediment toxicity data and derive SQCs when health risks were not considered.

2186 4.3.1 Sediment Quality Protection Goals

- 2187 Community members did not identify specific Indigenous uses for sediment, therefore use
- 2188 categories have not been developed for sediment. Rather, sediment protection goals were
- 2189 identified for benthic and aquatic biota and humans which can be exposed to chemicals that
- 2190 partition from surface water to sediments or are naturally occurring.
- 2191 The following protection goals for SQCs were identified:
- Concentrations of chemicals in sediment do not result in toxicological effects to survival,
- health, reproduction, or biodiversity in benthic invertebrate, emergent macrophyte and
- fish populations.
- Concentrations of chemicals in sediment do not result in bioaccumulation of chemicals in
- diet items which are over safe daily intake levels for consumers of benthic invertebrates,
- 2197 emergent macrophytes, and fish.

| 2198 | 4.3.2 Identification of Chemical Substances Related to Oil sands De- |
|------|--|
| 2199 | velopment and Database of Sediment Toxicity Data |
| 2200 | Chemical substances identified in Section 3.4.2 and 3.9 were carried forward and screened |
| 2201 | against available sediment quality guidelines and bioaccumulation data to identify substances |
| 2202 | which require SQCs. |
| 2203 | To support the derivation of SQCs, when required, spiked sediment toxicity study data |
| 2204 | and values were obtained from the Society of Environmental Toxicology and Chemistry (SE- |
| 2205 | ${\it TAC)}\ {\it Sediment}\ {\it Advisory}\ {\it Group}\ ({\it SEDAG})\ {\it database}\ ({\it Society}\ {\it of}\ {\it Environmental}\ {\it Toxicology}\ {\it and}\ $ |
| 2206 | Chemistry Sediment Advisory Group (SETAC SEDAG), 2016). |
| 2207 | 4.3.3 Inventory of Regulatory Sediment Quality Guidelines |
| 2208 | Available sediment quality guidelines developed using various approaches were identified |
| 2209 | through a jurisdictional scan of the following agencies. |
| 2210 | • Federal |
| 2211 | - Canadian Council of Ministers of the Environment [CCME (2001); and updates] |
| 2212 | • Provincial |
| 2213 | - Government of Alberta (GoA, 2018) |
| 2214 | – Nova Scotia Environment (Nova Scotia Environment (NSE), 2014) |
| 2215 | - Ontario Ministry of Environment and Energy (Ontario Ministry of Environment |
| 2216 | (OMOE), 2008) |
| 2217 | - Quebec (Direction du suivi de l'état de l'environment (Environment Canada and |
| 2218 | Ministère du Développement durable de l'Environnement et des Parcs du Québec |
| 2219 | $(\mathrm{DSEE}),2007))$ |
| 2220 | - BC Ministry of Water, Land and Air Protection (MWLAP, 2003) |
| 2221 | • United States Environmental Protection Agency |
| 2222 | - US EPA Assessment and Remediation of Contaminated Sediments Program (ARCS) |
| 2223 | (United States Department of Energy (US DOE), 1997) |
| 2224 | - US EPA Office of Solid Waste and Emergency Response (OSWER) (United States |
| 2225 | Department of Energy (US DOE), 1997) |
| 2226 | - US EPA (Region III) Biological Technical Assistance Group (BTAG) (Environmentus EPA |

- US EPA (Region IV) (United States Department of Energy (US DOE), 1996)

tal Protection Agency Biological Technical Assistance Group (EPA BTAG), 2006)

2227

| 2229 | • United States (State) |
|------|---|
| 2230 | – Minnesota Pollution Control Agency (Minnesota Pollution Control Agency |
| 2231 | (MPCA), 2007) |
| 2232 | - New York State Department of Environmental Conservation of Fish, Wildlife and |
| 2233 | ${\it Marine Resources Bureau of Habitat (New York State Department of Environmental}$ |
| 2234 | Conservation (NYSDEC), 2014) |
| 2235 | - United States Department of Energy (US DOE) Office of Environmental Manage- |
| 2236 | ment (United States Department of Energy (US DOE), 1997) |
| 2237 | – FDEP - Florida Department of Environmental Protection (Florida Department of |
| 2238 | Environmental Protection (FDOEP), 2003) |
| 2239 | - Washington State Department of Environment (Washington State Department of |
| 2240 | Ecology (WS DOE), 2019) |
| 2241 | Jurisdictions throughout North America have developed numerical and objective based |
| 2241 | standards for the protection of freshwater ecosystems. The approaches, listed below, vary |
| 2242 | · |
| 2243 | widely, and may include an empirical and/or theoretical based sediment quality guideline |
| 2244 | (MWLAP, 2003; Florida Department of Environmental Protection (FDOEP), 2003). A de- |
| 2245 | scription of each method is provided in Appendix A.6. |
| 2246 | • Screening Level Concentration Approach (SLCA) |
| 2247 | • Effects Range and Effects Level Approach (ERA, ELA) |
| 2248 | • Apparent Effects Threshold Approach (AETA) |
| 2249 | • Equilibrium Partitioning Approach (EqPA) |
| 2250 | • Logistic Regression Modeling Approach (LRMA) |
| 2251 | • Consensus Approach (CA) |
| 2252 | • Tissue Residue Approach (TRA) |
| | |
| 2253 | 4.3.4 Evaluation of Regulatory Agency Sediment Quality Guidelines. |
| 2254 | ${\bf Numerical\ and\ objective\ based\ sediment\ guidelines\ published\ by\ jurisdictions\ throughout\ North}$ |
| 2255 | America were evaluated against Indigenous water use protection goals established in the |
| 2256 | conceptual model to determine if published regulatory sediment quality guidelines could be |
| 2257 | adopted as SQCs. |

2258 4.3.5 Developing Sediment Quality Criteria for the Protection of In-2259 digenous Water Use

- The approach presented below, adapted from the OMOE (2008) weight of evidence (WoE)
- 2261 methodology, considers overall toxicity, benthos alteration, and biomagnification potential.
- The weight of evidence approach recognizes limitations in published sediment quality guide-
- 2263 line derivation methods and toxicity data and can be used to evaluate potential risks and
- 2264 support decision making regarding sediment contamination and health risks.
- 2265 The selected SQC was identified as the concentration at which limited to no adverse effects
- 2266 would be anticipated to occur and was typically selected from the following published guidelines
- 2267 or derived using toxicity data and prescribed methods.
- Rare Effect Level (REL)
- Spiked-Sediment Toxicity Test Values (Sediment Advisory Group (SEDAG) database)
- Bioaccumulation Sediment Guidance Values (BSGV) and Partitioning Theory Guideline
- Derivations (i.e., higher trophic human and ecological receptors protection)
- Potential for fish-tissue tainting (i.e., adverse taste).

2273 Sediment Quality Criteria (Adopted)

- 2274 The following criteria were used to evaluate published sediment quality guidelines and de-
- 2275 termine if they could be adopted as SQCs. If an available guideline did not meet the most
- 2276 stringent criteria, an SQC was derived, as described in the following section.

2277 Overall Toxicity

- 2278 Overall toxicity is defined as being negligible, minor or major. The following decision criteria
- 2279 were taken directly from the OMOE (2008) guidance document. To adopt the OMOE sediment
- 2280 guideline the sediment guideline must meet negligible or minor criteria

Negligible

2281

- 2282 Reduction of 20% or less in all toxicological test endpoints with only minor effects having been
- 2283 observed in no more than one endpoint.

2284 Minor

- 2285 Statistically significant reduction of more than 20% in one or more toxicological endpoints with
- 2286 multiple tests/endpoints exhibiting minor toxicological effects and no more than one exhibiting

2287 a major effect.

Major

2288

- 2289 Statistically significant reduction of more than 50% in one or more toxicological endpoints with
- 2290 multiple tests/endpoints exhibiting major toxicological effects.

2291 Benthos Alteration

- 2292 Although not explicitly stated within the OMOE guidance document measures of community
- 2293 structure could employ either the Shannon-Wiener or Simpson's index. These approaches are
- 2294 based on the number of species present (the functional group richness of the sample) and their
- 2295 relative abundance (the dominance or evenness of the sample population). One difficulty that
- 2296 may occur during interpretation of the Shannon-Weiner and Simpsons diversity indices is that
- 2297 they do not account for the comparisons of actual species present between reference and sample
- 2298 sites. Instead, the Jaccards similarity index (which acts as a measure of the fraction of shared
- 2299 species between sample sites) can also be calculated. As described by the
- 2300 OMOE (2008) other approaches can also be used (such as multivariate analysis) and description
- 2301 of change in consideration of the diversity, abundance and dominance of species living within
- 2302 the sediment is strongly recommended.

2303 Biomagnification Potential

- 2304 To address the potential risks to both humans and higher trophic aquatic receptors (i.e., fish,
- 2305 mammals, and aquatic birds) an evaluation of the potential for biomagnification is required.
- 2306 Biomagnification is the uptake of one or more contaminants through the food-web resulting in
- 2307 increasing concentrations through three or more trophic levels (Fisheries & Canada, 2019).

2308 Negligible

- 2309 Chemical is not presently known to have bioaccumulating properties or sufficient scientific
- 2310 literature has been established to indicate that the chemical does not readily bioaccumulate
- 2311 (i.e., it is readily metabolized and/or excreted by the body).
- 2312 Consistent with the Canadian Environmental Protection Act (CEPA), 1999 a substance is
- 2313 not considered bioaccumulative under the following considerations:
- Bioaccumulation Factor (BAF) is less than 5,000; or,
- Bioconcentration Factor (BCF) is less than 5,000 (if a BAF cannot be defined); or,
- LogKow is less than 5 (if neither a BAF nor a BCF can be defined)

Possible

2317

Chemical is known to bioaccumulate and/or bioconcentrate within the food web. It is presently unknown whether concentrations measured in sediment presents a confirmed health risk, but conservative modeling assumptions indicate that the potential exists. Non-ionizable, non-polar organic chemicals with one or more of the following characteristics (BAF 5,000 and/or, BCF 5,000 and/or, Log Kow 5) would fit within this category so long as measured concentrations do not exceed known sediment guidelines that are protective of higher trophic receptor effect.

2324 Significant

- Concentrations in sediment exceeds known bioaccumulation-based guidance value and/or there is clear evidence of risk to higher trophic organisms. Chemicals within this category meet one or more of the CEPA (Canadian Environmental Protection Act (CEPA), 1999) considerations for bioaccumulation and/or have a proven impact to higher trophic receptors at concentrations
- 2329 presently exhibited in the sediment chemistry.

2330 4.3.5.1 Sediment Quality Criteria (Derived)

2331 When available guidelines could not be adopted, SQCs were derived as follows.

2332 US EPA equilibrium partitioning (EqP)

- 2333 The US EPA equilibrium partitioning (EqP) method was used to derive SQCs for non-
- 2334 carcinogenic organic contaminants using the published water quality objective/guideline (US
- 2335 EPA, 2018):
- Equation (4.1): Equation to derive the sediment quality criteria using the equilibrium
- 2337 partitioning method for non carcinogenic organic contaminants (modified US EPA (2018)):

$$SQC = WQO/G \times (K_{oc} \times f_{oc} + (\frac{\theta m}{pw})) \tag{4.1}$$

Where:

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```
SQC = sediment quality criteria ( g/kg)

WQO/G = Water Quality Objective/Guideline ( g/L)

K_{oc} = Organic carbon partitioning coefficient (L/kg)

F_{oc} = fraction organic carbon (%OC/kg sediment (e.g., 2% = 20 g • OC/kg))

pw = 0.9982 density of water at 20°C

\theta = 0.3 (assumed as 30% moisture of sediment by mass)
```

Spiked Sediment Toxicity Test Approach

The spiked-sediment toxicity test (SSTT) approach uses information on the responses of test 2339 organisms to specific sediment associated chemicals under controlled laboratory conditions 2340 (Chapman and Long 1983; Ingersoll 1991; Lamberson and Swartz 1992). Sediments are spiked 2341 with known concentrations of chemicals, either alone or in combination, to establish definitive 2342 cause-and-effect relationships between chemicals and biological responses. At the end of the 2343 2344 test period, the response of the test organism is examined in relation to a biological end point (e.g., mortality, reproduction, growth). As in the development of water quality guidelines in 2345 Canada (Canadian Council of Resource and Environment Ministers (CCREM), 1987) or water 2346 quality criteria in the United States (US EPA, 1986), acute and chronic effect data generated 2347 from sediment toxicity tests can be used to identify concentrations of chemicals in sediment 2348 below which aguatic life would not be adversely affected. 2349 The Spiked Sediment Toxicity Test (SSTT) approach requires a minimum of 4 studies on 2350 2 or more sediment-resident invertebrate species, one of which must be a benthic crustacean, 2351 and one a benthic arthropod and at least 2 of these studies must be partial or full lifecycle 2352 tests of ecologically relevant endpoints (i.e., survival, growth, reproduction) (CCME, 1995). 2353 If the minimum data set requirements are met for the SSTT approach, an SQC can be 2354 derived, preferentially from the lowest-observed-effect level/Concentration (LOEL/C) from a 2355 chronic study using a nonlethal end point. The most sensitive LOEL/C is multiplied by an 2356 2357 appropriate safety factor to derive the SQCs. Applying Safety factors (SFs) to LOECs is a common approach to deriving risk-based 2358 guidelines using published toxicity data when data quality requirements are met. If toxicity 2359 2360 data for a substance met minimum criteria, the LOEC) was multiplied by a SF of 0.2 to derive

the SQC.

2361

The conservative SF (0.2) published by CCME (1995) was derived from published SFs previously used to develop sediment quality guidelines from toxicity data.

Bioaccumulation Based Sediment Guideline Values (BBSGVs)

The approach presented herein is an abbreviation of the work of (Newell et al., 1987) as updated by the works of NYSDEC (1999) and as described in NYSDEC (2014) and the Technical Operational Guidance Series (TOGS) as prepared by the NYSDEC Division of Water.

The first step in derivation of the BBSGV is to identify the Acceptable Daily Intake (ADI) of the receptor (human or wildlife) under consideration. The NYSDEC defines the ADI as the maximum concentration of a chemical in food that the receptor (i.e., bird, animal or human) can consume without exceeding a dietary exposure risk. This varies from the traditional definition of ADIs in risk assessment where DI is usually defined as exposure dose (mg/kgBW/d), also known as Tolerable Daily Intake.

The dietary risk value might be the no observed effect level (NOEL) the lowest observed effect level (LOEL) or another toxicological endpoint. In Canada, typical endpoints associated with wildlife exposures are the daily threshold effect dose (DTED) whereas for humans it is typically referred to as either the oral Tolerable Daily Intake (TDI) (for non-carcinogenic chemicals) or the oral Slope Factor (SF) (for cancer causing chemicals). Note that the slope factor must be converted to a risk specific dose (RsD) utilizing the following equation:

Equation (4.2): Equation to derive the risk specific dose (RsD) using the slope factor (SF) for cancer causing chemicals, and acceptable risk level (ARL).

$$RsD = \frac{ARL}{SF} \tag{4.2}$$

Where:

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RsD = reference dose (mg/kg body-weight/day) ARL = acceptable risk level (10-5) SF = slope factor

Once the ADI is defined the exposure concentration is derived as follows:

Equation (4.3): Equation to derive the baseline bioaccumulation factor (BAF Baseline) using the octanol-water partitioning coefficient and food chain multiplier.

$$BAF_{Baseline} = K_{ow} \times FCM \tag{4.3}$$

Where:

 $BAF_{Baseline}$ = Baseline Bioaccumulation Factor assuming 100% lipid content (trophic level specific) Kow = n-Octanol/Water portioning coefficient FCM = Food Chain Multiplier (as defined in literature based on trophic level)

Once the baseline is established, the wildlife BAF can now be calculated from the baseline BAF. The wildlife BAF is derived from the concentration of the contaminant freely dissolved in pore-water. This concentration is calculated as follows:

Equation (4.4): Equation to derive the concentration of the contaminant freely dissolved in pore-water (f fd) using the concentration of dissolved organic carbon (DOC) and particulate organic carbon (POC) in water.

$$f_{fd} = \frac{1}{1 + \frac{DOC)(K_{ow})}{10} + (POC)(K_{ow})}$$
(4.4)

Where:

 f_{fd} = freely dissolved fraction of a chemical in water

DOC = concentration of dissolved organic carbon in water (kg DOC/L)

POC = concentration of particulate organic carbon in water (kg POC/L)

2391 The value recommended by NYSDEC and applied for DOC is 0.000002 kg/L, and the POC is typically set as 0 (New York State Department of Environmental Conservation (NYSDEC), 2392 2014). Wildlife BAFs must also be adjusted for the lipid content of fish. The values are often 2393 2394 set based on literature derived studies and specified based on trophic level (e.g., 6.46% for trophic level 3 and 10.31 % for trophic level 4 (New York State Department of Environmental 2395 Conservation (NYSDEC), 2014)). Hence, the wildlife BAF for a specific trophic level can be 2396 2397 calculated as follows: Equation (4.5): Equation to derive the wildlife baseline bioaccumulation factor (BAF re-2398

ceptor/trophic level) for a specific trophic level using the BAF Baseline, (f fd) and % lipid in

2399

2400 fish for a given trophic level (%Lipid Trophic Level x Fish).

$$BAF_{TrophLevel_x}^{Receptor} = [(BAF_{Baseline}) \times (\%Lipid_{Trophic\ Level_x\ Fish}) + 1](f_{fd}) \tag{4.5}$$

Where:

 $BAF_{Troph\ Level_x}^{Receptor}=BAF$ for consumption of fish from a specified trophic level $BAF_{Baseline}=Baseline\ Bioaccumulation\ Factor\ (trophic\ level\ specific)\ (L/kg)$ % $Lipid_{Trophic\ Level_x\ Fish}=$ %lipid in fish for a given trophic level $f_{fd}=\text{freely\ dissolved\ fraction\ of\ a\ chemical\ in\ water}$

2401 Once each of the required trophic level BAFs has been derived determination of a 2402 bioaccumulation-based pore-water quality value can be conducted. There are several ways in which this value can be derived and consideration of the various media in which the receptor 2403 can be exposed requires consideration. 2404 The NYSCDEC (2014) defines the fish-flesh criterion (CFF) for protection of wildlife as 2405 the maximum concentration of a chemical that can be present in fish-flesh and not be harm-2406 ful to birds and animals that consume the fish. The NYSCDEC (2014) thus consider the 2407 CFF and ADI wildlife as synonymous. A departure presented herein maintains the assump-2408 tions presented in both CCME (2007) and AEP (2019) whereby an allocation factor (AF) is 2409 incorporated such that protection to the receptor is maintained as the relative proportion of 2410 exposure should include consideration of the various environmental pathways (air, soil, food, 2411 water, and consumer products) by which the receptor may likewise be exposed. As per the 2412 prescribed method, the AF applied incorporates a safety factor, assuming that a substantial 2413 portion of threshold intake will come from sources unrelated to water and sediment. The ADI 2414 also includes an uncertainty factor (UF). When multiplied together, the resulting SQC may be 2415 very conservative. 2416 For simplicity, it is assumed herein that wildlife receptors will have an applied AF of 75% 2417 (0.75) and humans an AF of 20% (0.2) (AEP, 2019; CCME, 2007) in derivation of the SQCOC. 2418 The SQC normalized to organic content of the soil was calculated as: 2419 2420 Equation (4.6): Equation to derive the sediment quality criteria normalized to organic 2421 content of soil (SQC OC) using an applied allocation factor (AF) (AEP, 2019; CCME, 2007).

$$SQC_{OC} = \frac{ADI_{receptor} \times AF}{\sum (BAF_{Trophic\ Level_x}^{Receptor} \times \% diet)} \times 1,000 \times K_{OC} \times \frac{1kg}{1,000gOC} \tag{4.6}$$

Where:

 SQC_{OC} = sediment quality criteria normalized to total organic carbon content (g • gOC)

 $ADI_{receptor} = Acceptable Daily Intake for receptor (mg/kg)$

AF = Allocation Factor (unitless)

 $BAF_{Trophic\ Level}^{Receptor} = BAF$ for fish of specified trophic level (L/kg)

% diet = percent of fish from specified trophic level contribute to diet

1,000 = convert mg/L to g/L

 K_{OC} = Organic carbon partitioning coefficient (L/kg)

Note, an AF does not apply when calculating a human based SQC for a carcinogenic chemical as the RsD already accounts for background exposure. Once the SQCOC has been calculated it can be adjusted (the SQC can be calculated) based on a site-specific TOC using standard equilibrium partitioning assumptions.

2426 **4.4** Results

27 4.4.1 Summary of North America Sediment Quality Guidelines

2428 A summary table of available guidelines from regulatory agencies within North America is 2429 provided in Appendix A.5.

2430 In Alberta, sediment quality guidelines were primarily adopted from the CCME (ISQG and PEL values) and the Ontario Ministry of the Environment and Energy (OMOEE). A select 2431 few chemicals were also sourced from Environment Canada (GoA, 2018). Values obtained 2432 from the OMOEE are listed separately and caution is recommended in their application as 2433 these values were derived over a limited geographic area (AEP 2018). The select few chemicals 2434 adopted from Environment Canada were calculated based on fish tissue guideline levels and 2435 the ratio of the contaminant in fish tissue compared to the concentrations found in sediment 2436 2437 (i.e., biota-sediment accumulation factor (BSAF)) (Environment Canada, 2013).

| 2438 | The effects range approach (ERA), adopted by CCME and GOA (2018) in derivation of both |
|------|--|
| 2439 | the ISQG and PEL guidelines, was formulated to derive SQCs based on assessing the potential |
| 2440 | for various COPCs (as analyzed as part of National Status and Trends Program (NSTP)) to |
| 2441 | illicit adverse effects on sediment-dwelling organisms (CCME, 1995). This process involves |
| 2442 | numerous steps including the acquisition of co-occurrence data. This co-occurrence data (i.e., |
| 2443 | ${\it field-collected sediments\ that\ contain\ chemical\ mixtures)}\ is\ maintained\ within\ Biological\ Effects$ |
| 2444 | Database for Sediment-associated contaminants (BEDS) [Long & Morgan (1990); Long (1992); |
| 2445 | Long & MacDonald (1992); MacDonald (1994); CCME (1995); Long et al. (1995)). Notably |
| 2446 | the CCME utilizes this methodology. |
| 2447 | The BEDs is separated based on measured chemical concentration, location, analysis type |
| 2448 | (or approach), test duration, end point measured, species and life-stage tested, whether associ- |
| 2449 | ated biological effects or no biological effects were observed, and the study reference. The data |
| 2450 | is separated into two specific datasets, one is created for effect data and the other is no effect. |
| 2451 | The effect dataset (E) relates to studies where an observed biological effect was associated |
| 2452 | with a measured chemical concentration. The no effects dataset (NE) comprises studies where |
| 2453 | there were nontoxic, without gradient, small gradient, or no-concordance. Only the effects |
| 2454 | data studies are used to generate SQCs. |
| 2455 | Chemical concentrations between effects and no effects datasets overlap as different species |
| 2456 | and varying site conditions contribute to a range of concentrations where effects and no effects |
| 2457 | data are reported. For these reasons, the effects dataset is sorted in ascending order and specific |
| 2458 | percentiles are selected as an indicator of the likelihood for observation of an adverse effect. |
| 2459 | Limitations in the CCME approach to developing sediment guidelines (adopted by GOA) |
| 2460 | are like those addressed under the OMOE (2008) approach which include lack of ability to |
| 2461 | establish dose-response relationships, absence of community structure consideration and limi- |
| 2462 | tations due to the geographical diversity of the studies used in matching chemistry and benthic |
| 2463 | invertebrate community structure for freshwater ecosystems. |
| 2464 | Based on the paucity of data for chemical dose-response relationships, the fact that the |
| 2465 | BEDs database has not been revisited since the early 1990s, and a general lack of human health |
| 2466 | consideration, it was determined that derivation of sediment quality criteria for application in |
| 2467 | the Lower Athabasca Region would need to be developed. |
| 2468 | In general, the CCME and GOA (2018) ISQGs and PELs do not meet the criteria for |
| 2469 | Indigenous water use protection from sediment associated contaminants. |

2470 4.4.2 Sediment Quality Criteria

- 2471 A summary of adopted and derived SQCs for the protection of Indigenous water use protec-
- 2472 tion goals including human health and carcinogenicity from exposure to bioaccumulative and
- 2473 persistent substances is provided in Table 3.9 along with a comparison to the provincial ISQGs
- 2474 [GoA (2018); CCME].
- Detailed results of the WoE analysis are provided in Appendix A.5. An example of the
- 2476 results for arsenic are presented following Table 4.1, below.

Table 4.1: Risk based sediment quality criteria for the protection of Indigenous use.

| Parameter | Alberta ISQG (mg/kg) | SQC (mg/kg) | Source |
|------------------------|------------------------|-------------|--|
| etals | | | |
| Arsenic* | 5.9 | 4.1 | Quebec (DSEE)-REL |
| Cadmium | _ | 0.33 | Quebec (DSEE)-REL |
| Chromium (total) | 37.3 | 25 | Quebec (DSEE)-REL |
| Copper | 35.7 | 8.6 | SST Benchmark Approach (Derived) |
| Lead | 35 | 25 | Quebec (DSEE)-REL |
| Manganese | | 460 | Ontario (OMOE) LEL |
| Mercury | 0.17 | 0.094 | Quebec (DSEE)-REL |
| Molybdenum | | 718 | SST Benchmark Approach (Derived) |
| Nickel | _ | 16 | Ontario (OMOEE) - LEL |
| Selenium | 2 | 2 | Alberta ISQG |
| Silver | _ | 0.57 | Washington WSDOE |
| Thallium | <u> </u> | 0.86 | Health Canada (2020) |
| Uranium | _ | 0.594 | SST Benchmark Approach (Derived) |
| Vanadium | | 125 | SST Benchmark Approach (Derived) |
| Zinc | 123 | 7.4 | SST Benchmark Approach (Derived) |
| Low MW PAHs | rbons — | 0.552 | US EPA (OSWER)-ER-L |
| High MW PAHs | _ | 0.655 | US EPA (Region IV - FDEP)-TEL |
| Total PAHs | | 1.684 | US EPA (Region IV - FDEP)-TEL |
| Acenaphthene | 0.00671 | 0.0037 | Quebec (DSEE)-REL |
| Acenaphthylene | 0.00587 | 0.0033 | Quebec (DSEE)-REL |
| Anthracene | 0.0469 | 0.0087 | US DOE-EqP secondary |
| Benz[a]anthracene* | 0.0317 | 0.0079 | Derived EqP fish tissue, carcinogenicity |
| Benzo[a]pyrene* | 0.0319 | 6e-04 | Derived EqP fish tissue, carcinogenicity |
| Chrysene* | 0.0571 | 0.079 | Derived EqP fish tissue, carcinogenicity |
| Dibenz[a,h]anthracene* | | 0.00062 | Derived EqP fish tissue, carcinogenicity |
| Fluoranthene | 0.111 | 0.047 | Quebec (DSEE)-REL |
| Fluorene | 0.0212 | 0.01 | Quebec (DSEE)-OEL |
| 2-Methylnaphthalene | _ | 0.016 | Quebec (DSEE)-REL |
| Naphthalene | _ | 0.017 | Quebec (DSEE)-REL |
| Phenanthrene | - | 0.025 | Quebec (DSEE)-REL |
| Pyrene | _ | 0.029 | Quebec (DSEE)-REL |
| Naphthenic acids | _ | 3.3 | Derived (US EPA EqPA method) |

Table 4.1: Risk based sediment quality criteria for the protection of Indigenous use. (continued)

| Parameter | Alberta ISQG (mg/kg) | SQC (mg/kg) | Source |
|-----------|----------------------|-------------|----------------------------------|
| Phenols | _ | 0.23 | Derived EqP fish tissue tainting |

Note:

Sum identified LMW PAH congeners (Anthracene, Acenaphthene, Acenaphthylene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene, Pyrene) (CCME (2010))

Sum of identified HMW PAH congeners (Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Chrysene, Dibenzo(a,h)anthracene, Indeno(1,2,3-cd)pyrene) (CCME (2010))

^{*} Denotes carcinogenic substance

2477 Arsenic

2478 The SQC value of 4.1 mg/kg was adopted from Quebec (DSEE) REL for Arsenic.

2479 Guideline Review

The literature review indicated that sediment guideline values for this chemical range from a low of 4.1 mg/kg (Quebec DSEE) to a high of 120 mg/kg (Washington DSE)).

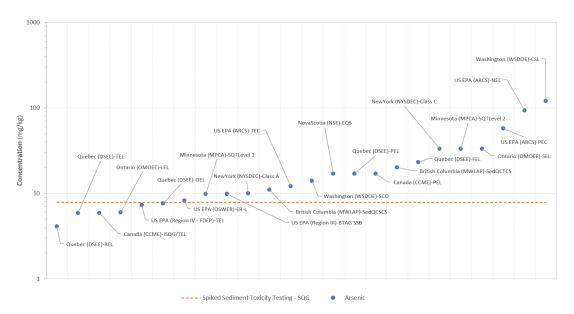


Figure 4.1: Distribution of sediment guideline values based on jurisdiction and associated guideline concentration (blue dots). The orange dashed line indicates a calculated value based on the CCME SST approach (7.8 mg/kg).

SSTT Derivation

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Spiked sediment toxicity values obtained from the Society of Environmental Toxicology and Chemistry (SETAC) Sediment Advisory Group (SEDAG) database (Society of Environmental Toxicology and Chemistry Sediment Advisory Group (SETAC SEDAG), 2016) were used to estimate a SQC based on CCME guidance (1995). The lowest of the lowest observed effect concentration (LOEC) values (39 mg/kg; C. dilutes; survival and growth) was multiplied by an Uncertainty Factor (UF) of 0.2. The calculated value of 7.8 mg/kg is in close agreement with the OEL value (7.6 mg/kg) provided by DSEE (DSEE). However, the data used to derive this SQC does not meet the minimum data-set requirements for derivation of a freshwater SQC for arsenic and confidence in this value is low.

Table 4.2: Spiked sediment toxicity testing results – Arsenic.

| Test Species | Lifestage | Duration (Days) | Endpoint | Effect | Concentra- tion | Units | OCNorm (g/g-OC) | TOC (%) | Citation |
|------------------------------------|-----------|-----------------|----------|--------|--------------------|-------|--------------------|---------|-------------------|
| Chironomus dilutus | juvenile | 10 | survival | NOEC | 39.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | growth | NOEC | 39.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | growth | LOEC | 39.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | survival | LOEC | 116.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | growth | LC25 | 174.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | growth | LC50 | 342.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | survival | NOEC | 462.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | growth | NOEC | 462.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | growth | LC25 | 462.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | growth | LC50 | 462.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | survival | LC25 | 521.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | survival | LC50 | 532.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | survival | LC50 | 642.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Chironomus dilutus | juvenile | 10 | survival | LC25 | 675.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | survival | LOEC | 724.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Hyalella azteca | juvenile | 10 | growth | LOEC | 724.0 | mg/kg | | 7.4 | Liber et al. 2011 |
| Derived guideline (LOEC*UF 0.2) | | | | | 7.8 | mg/kg | | | |

Note:

NA - not applicable

NOEC - no observed effect concentration

LOEC - lowest observed effect concentration

 $\mathrm{LC}25$ - concentration lethal to 25

LC50 - concentration lethal to 50

Biomagnification Check

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There were no biomagnification-based sediment quality guidelines identified. Sediment-to-2494 benthic invertebrate bioconcentration factor reported by the US EPA (1999) is 0.9 (mg COPC 2495 / kg wet tissue per mg COPC / kg dry sediment). Arsenic appears to be bioaccumulated, 2496 through the ingestion of food, but is not biomagnified through food webs (Hepp et al., 2017). 2497 A comparative check in consideration of the potential to cause adverse effect to either 2498 human or ecological (mammalian and avian) receptors was also conducted. An arbitrary 2499 screening concentration of 21 mg/kg for humans and 43 mg/kg for ecological receptors was 2500 identified. It is understood that these values are reflective of terrestrial receptors and terrestrial 2501 exposure scenarios (for which these guidelines were originally intended) but they are presented 2502 2503 here as a simplified check function in an effort to evaluate whether further consideration of 2504 these exposure pathways is warranted. It is considered likely that protection of the aquatic receptors (benthic invertebrates) would inherently be protective of higher trophic organisms 2505 2506 as well.

Derivation Summary

The results of screening existing guidelines, toxicity data and proposed SQC value (mg/kg against Toxicity and Benthos Alteration and Biomagnification Potential criteria are provided in Table 4.3, below.

Table 4.3: Arsenic WoE Evaluation

| Screening Criteria | Proposed SQC value screening results |
|----------------------------|--|
| Toxicity Endpoints | Negligible: Reduction of 20% or less in all toxicological endpoints. |
| Overall Toxicity | Negligible: Minor toxicological effects observed in no more than one endpoint. |
| Benthos Alteration | "equivalent" to reference stations |
| Biomagnification Potential | Negligible: Chemical is unlikely to biomagnify |

4.5 Discussion

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Sediments provide substrates in which aquatic macrophytes root and grow and essential habitats for many sediment-dwelling invertebrates and benthic fish. The nutrients and contaminants in sediments nourish and are accumulated to varying degrees by aquatic

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| 2515 | macrophytes and benthic invertebrates. Importantly, sediments can also provide habitats for |
|------|--|
| 2516 | many wildlife species during portions of their life cycle and a variety of fish species utilize |
| 2517 | sediments for spawning and incubation of their eggs and larvae. The importance of sediment |
| 2518 | in the aquatic ecosystem is substantive and so must the assessment of potential risks from |
| 2519 | contamination of this substrate (MacDonald et al., 2003). |
| 2520 | It has been reported that the use of the CCME ISQG values in establishing sediment |
| 2521 | benchmark concentrations are highly conservative, and their exceedance does not correlate |
| 2522 | with sediment toxicity (Nova Scotia Environment (NSE), 2014). For these reasons, a WoE |
| 2523 | approach to based on benthos alteration, toxicity, and bioaccumulation/ persistence potential |
| 2524 | was used to propose SQCs to meet sediment protection goals. |
| 2525 | When regulatory sediment quality guidelines were not available, spiked sediment toxicity |
| 2526 | test data was used to derive a SQCs using CCME (1995) methods by applying a safety factor |
| 2527 | of 0.2 to the LOEC for that particular substance. |
| 2528 | Within this WoE approach, available guidelines which offered the greatest level of protection |
| 2529 | were adopted as the SQC and proposed as the criteria for assessing sediment contamination |
| 2530 | and protection of Indigenous water use. |
| 2531 | Generally, CCME and GOA (2018) ISQG and PEL values were higher than all other regu- |
| 2532 | latory agencies with published sediment quality guidelines and could not be adopted as SQCs |
| 2533 | as they did not meet Indigenous protection goals for sediment quality (see Appendices 6 and |
| 2534 | 7). |
| 2535 | Table 4.1 provides a summary of the SQCs which together with the Indigenous water use |
| 2536 | category specific criteria provide an ecosystem approach to assessing the quality of surface |
| 2537 | water bodies in the Lower Athabasca Region. The SQCs are intended for application to any |
| 2538 | substrate (i.e. treated tailings in contact with or used to create sediments) that is being used |
| 2539 | to construct a surface water closure feature including EPLs and wetlands. |
| | |

Chapter 5

Community Traditional Food

Survey

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- 2543 THOMAS DYCK PHD
- 2544 Integral Ecology Group

5.1 Introduction

- Consumption of traditional foods and medicines is essential for the health and wellbeing of 2546 Indigenous communities. These resources provide important nutrients and health benefits and 2547 offer a culturally-relevant way for community members to treat specific health conditions and 2548 maintain all aspects of their physical, mental and spiritual health (Kuhnlein & Turner, 1991). 2549 Consumption of traditional resources is essential for Indigenous communities to maintain a con-2550 nection to the land and helps maintain community cohesion. Traditional foods and medicines 2551 are often shared with other family members and elders, promoting stronger social relationships 2552 2553 within the community. Hunting, fishing, and gathering plants are also important practices for communities to exercise their rights as Indigenous peoples. 2554
 - Chapter 5 describes the methods used for the Community Traditional Foods Consumption Survey with a discussion of demographic results, consumption preferences, and barriers to harvesting. The survey's primary role was to gather information from each of the participating Indigenous communities regarding the consumption patterns and ingestion rates for traditional foods and medicines.¹ The information collected was used to inform the risk-based analysis and modelling exercise, which was conducted to determine whether surface water and sediment quality thresholds for the protection of aquatic life (chronic and acute) are protective of

¹Including medicines applied externally to the body (i.e., poultice).

2562 receptors connected through feeding guild interactions or exposures to environmental media.

5.2 Objective

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- 2564 The survey objectives are to:
- 1. Develop a list of community-relevant receptors connected through feeding guild interactions or exposures to environmental media;
- 2567 2. Identify representative community ingestion rates for traditional foods and medicines;
- 3. Identify community consumption preferences and barriers related to consumption of traditional foods and medicines.

The primary method for this component of the project focused on the design and delivery

5.3 Methods

- of a community survey. A survey is a "systematic method for gathering information from (a 2572 sample of) entities for the purpose of constructing quantitative descriptors (statistics) of the 2573 attributes of the larger population of which the entities are members," (Groves et al., 2009). 2574 For this project, using a survey offers three key advantages. First, a survey offers versatility 2575 in its design and format and enables researchers to gather information directly from commu-2576 nity members. Second, a survey involves the collection of responses from a representative 2577 portion of the community's population, meaning that findings can be generalized and applied 2578 2579 to the broader population (i.e., the results are considered statistically representative of the population) (Palys, 1997). In this project, the collection of statistically representative results 2580 enabled the environmental scientist to analyze and calculate community members' ingestion 2581 rates of traditional foods and medicines for the three participating Indigenous communities 2582
- and for different age groups and sex within each community. Third, a survey is an efficient
- 2584 way to collect detailed information from community members about traditional food consump-
- 2585 tion, and enabled the project team to compare and evaluate the survey findings against the
- 2586 Health Canada document Guidance for Evaluating Human Health Impacts in Environmental
- 2587 Assessment: Country Foods (Health Canada, 2017).

5.3.0.1 Survey design and implementation

- 2589 Survey design and implementation consisted of four key elements, summarized below:
- 2590 1. identify and prioritize receptors,

2591 2. survey design,

- 2592 3. planning and preparation, and
- 2593 4. pilot and implementation.
- 2594 The following sub-sections provide details of each element.

5.3.0.2 Identifying and prioritizing receptors

As noted above, information collected in the survey was used to inform the risk-based analysis and modelling exercise. This exercise was used to determine whether surface water and sediment quality thresholds for the protection of aquatic life (chronic and acute) are protective of receptors connected through feeding guild interactions or exposures to environmental media. Receptors are living organisms that could be adversely affected by environmental contaminations released and/or dispersed into the environment from an industrial site.

The first step in developing the survey was to identify and prioritize community relevant receptors, namely, plants and animals that are consumed as food or medicines by members of each community. To identify these receptors, a literature review regarding the consumption of traditional foods and medicines was conducted. Document searches were conducted within internal community databases and online using key words (e.g., Indigenous, ingestion, country foods, traditional foods, rates, consumption) to recover materials from government and organizational sources. Internal sources consisted of a traditional plants book, Indigenous knowledge interview transcripts, and community reports. During this step, a master list of 115 terrestrial and aquatic receptors known to be used by the communities for consumption and medicinal purposes was compiled.

Representatives from each community, along with support from the project technical team (social scientists [Integral Ecology Group Ltd.] and environmental scientists [Integrated Toxicology Solutions Ltd.]), reviewed the master list of receptors and underwent a process to group and prioritize the list of 115 receptors down to 35 receptors and receptor groups. Grouping and prioritizing was necessary to ensure the survey could be completed within each community with a reasonable amount of effort and time. Key steps for grouping and prioritizing the receptors included the following:

Ranking the receptors

- 2620 The receptors were ranked in two ways to help prioritize receptors for including in the survey:
- 1. A frequency table depicting how many times a receptor was mentioned in the community

- documents was compiled to understand how often a particular species was discussed in community documents. Receptors with more mentions ranked higher than receptors with ower mentions. Recognizing that concerns or community importance of a species cannot be fully assumed based on frequency information alone, we used the information as only a guide to estimate concerns and/or importance.
- 2. Available ingestion rates for receptors were reviewed in reports including the First Nations Food, Nutrition, and Environment Study by (Chan et al., 2016), and other internal community traditional foods studies. Receptors were prioritized if they were mentioned in more than three community documents, or if they were reported to be highly consumed in the region as traditional foods (i.e., with a high ingestion rate).
- The results from these two ranking steps were compared and contrasted to develop a single prioritized list of receptors.

Removing terrestrial species

The technical team reviewed the list of priority receptors identified in the ranking exercise and removed a total of 31 terrestrial receptors, or plants and animals that are land-based and/or rely on water primarily for dietary purposes only. Some terrestrial receptors were not removed due to there importance in the community (e.g., moose). Examples of the terrestrial receptors removed at this stage include prickly rose/rose hip, blueberry, high-bush cranberry, pin cherry, and lynx.

Grouping closely related species into receptor groups

The technical team organized the list of priority receptors into individual receptors and receptor groups (i.e., groups of closely related species with similar diets). For example, two receptor groups were created for duck species, based on the differences in their diets. Grouping similar species with similar diets helped to reduce the overall number receptors included in the survey.

The prioritized list of receptors was reviewed by each community for feedback and verifica-

The prioritized list of receptors was reviewed by each community for feedback and verification. Community feedback resulted in the inclusion of new receptors (e.g., lily pads; *Nuphar variegata*) on the list and discussion about other receptors potentially less critical for the study. No receptors were removed at this stage. Following community review, we finalized a list of 35 aquatic receptors, capturing a total of approximately 79 species of mammals, fish, birds, and plants. This list was used as the basis for developing the community survey (see Table 5.1).

Table 5.1: List of the 35 community relevant receptors (including 79 species) for the survey. Note that this is not a comprehensive list of all of the receptors or species that are important to the MCFN, ACFN, or FMFN.

| Receptor | List of species included in receptor |
|--|---|
| Fish and freshwater clan | ns |
| Ling cod (ling, maria, mariah, burbot, loche) or inconnu | Ling cod (ling, maria, mariah, burbot, loche) ($Lota\ lota$), inconnu ($Stenodr leucichthys$) |
| Whitefish or cisco | Mountain whitefish (<i>Prosopium williamsoni</i>), lake whitefish (<i>Coregonus clupeaformis</i>), cisco (<i>Coregonus zenithicus</i>) |
| Arctic grayling | Arctic grayling (Thymallus arcticus) |
| Trout | Rainbow trout (Oncorhynchus mykiss), lake (char) trout (Salvelinus namaycush), brook trout (Salvelinus fontinalis), bull trout (Salvelinus confluentus), cutthroat trout (Oncorhynchus clarki), brown trout (Salmo trutta) |
| Sucker | White sucker (Catostomus commersonii), longnose sucker (Catostomus catostomus) |
| Goldeye | Goldeye (Hiodon alosoides) |
| Walleye (pickerel) | Walleye (pickerel) (Sander vitreus) |
| Great northern pike (jackfish) | Great northern pike (jackfish) (Esox lucius) |
| Freshwater clams ¹ | May include ² giant floater (Anodonta grandis), western floater (Anodonta kennerlyi), creek/brook heelsplitter (Lasmigona compressa), white heelsplitter (Lasmigona complanate), fat mucket (Lampsilis siliquoidea) |
| Mammals Caribou | Woodland caribou (Rangifer tarandus), barren caribou (Rangifer tarandus groenlandicus) |
| Moose | Moose (Alces alces) |
| Deer | White-tailed deer (Odocoileus virginianus), mule deer (Odocoileus hemionus) |
| Elk | Elk (Cervus canadensis) |
| Buffalo or wood bison | Buffalo or wood bison (Bison bison) |
| Bear | Black bear (<i>Ursus americanus</i>), grizzly bear (<i>Ursus arctos horribilis</i>) |
| Beaver | Beaver (Castor canadensis) |
| Muskrat | Muskrat (Ondatra zibethicus) |
| Rabbit or snowshoe hare | Rabbit or snowshoe hare (Lepus americanus) |
| Birds | |
| Duck, group 1 | Mallard (Anas platyrhynchos), green-winged teal (Anas carolinensis), redhead (Aythya americana), ring-necked duck (Aythya collaris) |
| Duck, group 2 | Lesser scaup (Aythya affinis), greater scaup (Aythya marila), canvasback (Aythya valisineria), goldeneye (Bucephala clangula), surf scoter (Melanitt perspicillata), white-winged scoter (Melanitta fusca deglandi), mud hen (Fulica americana), blue-winged teal (Anis discors), northern shoveler (Anas clypeata), northern pintail (Anas acuta), long-tailed (Clangula hyemalis), ruddy (Oxyura jamaicensis), Gadwall duck (Mareca strepera) |
| Goose | Greater white fronted goose (Anser albifrons), snow goose (wavy) (Anser caerulescens), Canada goose (Branta canadensis) |
| Swan | May include trumpeter swan ($Cygnus\ buccinator$), tundra swan ($Cygnus\ columbianus$) |
| Grouse | Blue grouse (<i>Dendragapus obscurus</i>), ruffed grouse (<i>Bonasa umbellus</i>), spruce grouse (<i>Falcipennis canadensis</i>), sharp-tailed grouse (<i>Tympanuchus phasianellus</i>), willow grouse (unknown) |
| Ptarmigan | May include willow ptarmigan (Lagopus lagopus), rock ptarmigan (Lagopu mutus) |
| Prairie chicken | Greater prairie chicken (Tympanuchus cupido pinnatus) |
| Plants | |
| Labrador tea | Labrador tea (Rhododendron groenlandicum) |
| Wild mint | Wild mint (Mentha arvensis) |
| Rat root | Rat root (Acorus americanus) |
| Black spruce | Black spruce (Picea mariana) |
| Bog cranberry | May include bog cranberry (<i>Vaccinium vitis-idaea</i>), small bog cranberry (<i>Vaccinium oxycoccos</i>) |
| Duckweed | Duckweed (Lemna turionifera) |
| Willow | May include red willow (Cornus stolonifera), sandbar willow (Salix exigua) Pacific willow (Salix lucida ssp. lasiandra) |

Table 5.1: List of the 35 community relevant receptors (including 79 species) for the survey. Note that this is not a comprehensive list of all of the receptors or species that are important to the MCFN, ACFN, or FMFN. (continued)

| Receptor | List of species included in receptor |
|----------------------------|---|
| Cattail | Cattail (Typha latifolia) |
| Fiddleheads | May include ostrich fern (Metteuccia struthiopteris), lady fern (Athyrium filix-femina), spinulose shield fern (Dryopteris carthusiana) |
| Lily pads (wild pineapple) | Lily pads (wild pineapple) (Nuphar variegata) |

¹ Freshwater mussels are known locally by Indigenous communities in the Lower Athabasca region as freshwater clams Hopkins et al. (2019). The term "clams" was used in the survey as this is the preferred term among the participating communities.

5.3.0.3 Survey design

The project technical team worked closely with the communities to co-develop the survey questions. The majority of the survey consisted of questions about individual consumption patterns for the 35 receptors, including the frequency of consumption, which parts of the receptor are consumed (e.g. fat, meat/tissue, organs, leaves, flowers, stem, root, eggs), serving or portion size, and preparation methods (e.g., boiled/tea, fried, fresh/raw, baked, dried/smoked, put on skin). An optional set of questions focused on children's consumption patterns, intended for those participants responsible for providing traditional foods and medicines to children (ages 0-18). The survey also covered other topics with relevance to the research questions, including: demographic characteristics, gender, age, changes in the availability of plants and wildlife, barriers to consuming traditional foods, consumption preferences, and the specific waterbodies where traditional foods are harvested within the lower Athabasca region. To achieve the objectives of this study, only demographic results, consumption preferences, and barriers to consumption are discussed (see Section 5.4).

The survey was designed using SoGo Survey², a secure online survey platform that offers survey design tools, multi-channel distribution, and analytics tools. The platform allows potential participants to complete the survey online via computer, tablet or smart phone. The survey included the full survey and once completed and submitted by the participant, responses are saved to an online database. The data collected is always owned by the respective communities. After the survey has been completed and it has been confirmed that all analysis is complete, the results of the survey have been removed from online servers and transferred to respective community servers to be stored and accessed by the community for future use.

Participant consent is an important component of ensuring participants are informed about

² "May include" is used in the table to refer to species that were not listed in the survey questions. These species are thought to be consumed as traditional foods or medicines by community members.

²https://www.sogosurvey.com/

the survey's purpose and how their information will be used. A consent letter and a community handout with information about the survey were developed to accompany the survey (see Appendix A.7). The community handout summarizes the purpose of the survey and reviews the approach for obtaining participant consent. A list of the survey receptors with pictures of key species was also included in the handout as a visual guide for participants completing the survey. The handout and consent letter were tailored for each community and shared with all participants prior to administering the survey. Before finalizing the survey and the accompanying materials (e.g., consent forms and community handouts) a final review was conducted by representatives of each community to ensure the survey questions aligned with community interests and protocols.

5.3.0.4 Planning and preparation

Survey planning and preparation was led by each community according to community-specific protocols for engaging their membership, guided by community leads, community researchers, and input from technical support. With COVID-19 restrictions making it difficult for researchers to meet face-to-face with participants, the research team planned that participants would either selected randomly by the community leads and community researchers or allowed to self-select to participate. Some of the communities identified that identifying participants was necessary due to facilitate access to members that might otherwise not have access to the survey especially with ongoing community and provincial COVID-19 restrictions. A selection criteria was developed to ensure the sample was randomized to the extent possible and that a broad sample of the community was selected. The selection criteria included the following:

- participant is a member of either ACFN, MCFN, or FMFN;
- participant is part of a diverse range of age groups and sexes; and
- participants are from different family groups represented within the community.

All community members had the opportunity to self-select and choose to participate in the survey online via a link provided through local community outlets (e.g. band office Facebook pages, local radio advertisements) or over the telephone via community researcher.

It was important for each community that participants were compensated for taking the time to complete the survey. Honoraria is provided for sharing knowledge and information and is a gift in a show of reciprocity. Honoraria were distributed to survey participants in accordance with protocols within each community. Two of the communities opted to distribute the honoraria as gift cards, while the other community issued payments to survey participants.

A target of approximately 100 surveys per community was set by the project team. This number was determined by communities to be reasonable given the scope of the project and anticipated efforts required by community leads and community researchers to implement the survey. To verify whether the three samples were representative of each community's population, an analysis of demographic results compared to community available profiles were calculated and allowed the researchers to make inferences about the community population.

To support implementation, community researchers were identified and selected by each community. These individuals were members of the participating Indigenous communities and actively participated in the project by attending planning meetings, delivering survey information materials, assisting with survey implementation, and making other planning and implementation related contributions. Remote training sessions with the community researchers were administered by the technical team and focused on interview protocols and survey delivery. The technical team also provided additional support to community researchers throughout the implementation of the survey.

5.3.0.5 Pilot and implementation

A pilot test of the survey was undertaken in late November and early December, 2020 as a first step in survey implementation. The survey pilot was completed by community leads and community researchers, and helped the project team identify inconsistencies, typographical errors, or technical glitches in the survey. Testing the survey with community researchers also helped these individuals gain a sense of familiarity with the online SoGo Survey platform and the flow of questions. Based on the feedback received, the survey was finalized by the research team.

Due to COVID-19 protocols and restrictions at the time when the surveys were being conducted and other restrictions (e.g., poor cellular data service, lack of computer connection or technological support), the research team determined that remote engagement with members was the best approach in order to keep everyone safe and reduce survey access barriers. The surveys were conducted using telephone and online survey methods (Fielding et al., 2008; Hayward et al., 2021; Wolf et al., 2016).

Most members have access to a telephone, and so one-on-one telephone interviews were conducted by the community researchers using a pre-selected randomized list of potential participants developed by the community. Prior to any one-on-one telephone survey, participants were provided with a paper copy of the community handout which included information about the survey and a consent letter to review and confirm within the survey or verbally with the

- interviewee. Using a computer, the community researchers accessed a web-based link to the survey and recorded responses via telephone on behalf of participating individual. The survey was implemented between mid-December, 2020 and mid-February, 2021.
- Participants could also choose to complete the survey via an online link provided through local community outlets. We estimate that approximately 60 surveys were self-conducted via the online link distributed through community outlets.
- To track survey progress, community researchers and community leads accessed a secure link to a Sogo Survey webpage with community-specific survey statistics. This link enabled these individuals to track participation rates within their community in real time for two primary purposes: (i) preparing progress updates about the survey for their department or band office, and (ii) creating a list of honoraria/gift card recipients.

5.3.1 Data Review and management

The raw survey data was compiled into a spreadsheet, stored on researcher computers, and reviewed for quality assurance and quality control by the technical support team. In some 2753 2754 cases, narrative responses were converted into numerical values to assist with data analysis. For example, if a survey participant indicated they consumed whitefish "every two months 2755 in a year," this response was converted to the value of 6 (12/2=6). In addition, community 2756 2757 researchers worked with their membership to develop a list of the approximate average weights for the certain traditional foods noted by participants in the survey (e.g., moose heart, burbot 2758 liver, duck gizzard). Again, these descriptive responses were replaced with numerical average 2759 weight values where possible. When the data review was complete. 2760

761 **5.3.1.1** Limitations

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While the data was being reviewed, the social scientists noticed inconsistencies in the responses to the sub-set questions regarding children's consumption of traditional foods. It was determined that a technical glitch with the Sogo Survey platform was incorrectly recording responses on children consumption questions. This ultimately led to the loss of children consumption data. Once the technical glitch was resolved, the team was able to collect responses for a total of 18 children.

2768 5.3.1.2 Analysis

Data collected by the survey resulted in detailed information about community ingestion rates of traditional foods and medicines, demographic information, and community context that inform community consumption. Ingestion data was analyzed to inform the risk-based analysis and modelling exercise to determine whether surface water and sediment quality thresholds for the protection of aquatic life (chronic and acute) are protective of receptors connected through feeding guild interactions or exposures to environmental media.

Analysis of demographic data and community context information was conducted to better understand the demographic characteristics of survey participants (such as community, age and sex), and to examine key traditional food consumption patterns, including whether members consumed traditional foods in the past year; community preferences for consuming traditional foods; how many members provide traditional foods and medicines to children; and identified barriers to harvesting more traditional foods and medicines.

81 5.4 Results and Discussion

5.4.1 Demographic results

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The survey was implemented between mid-December 2020 and mid-February 2021 and a total of 247 surveys (n=247) were completed by members of the three communities. Approximately 43% of the surveys were completed by members of Athabasca Chipewyan First Nation, 33% were completed by Mikisew Cree First Nation members, and 23% were completed by members of Fort McKay First Nation (see Table 5.2).

Table 5.2: Community survey participation by percentage (n=247).

| Indigenous community | Percent |
|----------------------------------|---------|
| Athabasca Chipewyan First Nation | 43% |
| Fort McKay First Nation | 23% |
| Mikisew Cree First Nation | 33% |

The survey was completed by community members representing different sexes. In total, 2788 58% of the participants were female, 42% were male, and 0.4% identified as "other" (n=247). 2789 2790 Compared to community profiles available for each community, there is a possible gender bias in responses. The reported proportion of female and male across all three communities is 2791 50% compared to 58% female participants surveyed (Indigenous and Northern Affairs Canada, 2792 2793 2016). The survey was completed by community members within four age groups (see Table 5.3). 2794 Participants in the 51 and over age group represent the largest sub-set of survey participants 2795 (48%), followed by participants between 31 and 50 years (29%), and participants between 18 2796 and 30 years (13%). The fewest number of surveys (9%) were completed for children under 18 2797

years (see Section 5.3). Compared to community profiles available for each community, there is a possible bias to persons over 51 years old. The reported proportion of persons 0-19 is 36%, persons 20-64 years old is 56%, and over 65 years old is 9%. (Indigenous and Northern Affairs Canada, 2016). Survey participation by sex and age group was as follows: participants in the 51 and over age group were comprised of 29% female, 19% male, and 0.4% other; participants between 31 and 50 years were comprised of 15% female and 14% male; participants between 18 and 30 years were comprised of 8% female and 5% male; and children under 18 were comprised of 5% female and 5% male individuals.

Table 5.3: Survey participation by age group and sex.

| Sex | Under 18 years | 18 - 30 years | 31 - 50 years | 51 years and over |
|--------|----------------|---------------|---------------|-------------------|
| Female | 4.9% | 8.1% | 15.4% | 29.1% |
| Male | 4.5% | 5.3% | 13.8% | 18.6% |
| Other | 0.0% | 0.0% | 0.0% | 0.4% |
| Total | 9.3% | 13.4% | 29.1% | 48.2% |

5.4.2 Results overview: Community context

The following sub-sections summarize results of the survey regarding consumption of traditional foods and medicines, current and desired future consumption of traditional foods and medicines, providing traditional foods and medicines to children, and barriers to consuming traditional foods and medicines. It is important to note that the findings are presented across the three participating communities and therefore may not align with community-specific results. The results should also not be considered representative of a specific community, the results are representative of all three communities' perspectives and concerns combined.

5.4.2.1 Consumption of traditional foods in the past year

In the past year, 88% of survey participants have eaten or used traditional foods or medicines from the Athabasca River, Peace-Athabasca Delta, Lake Athabasca, or other waterbodies in the surrounding region (n=247; see Table 5.4).

Participants in the 51 years and over and under 18 years age groups represent the largest percentage of individuals who have consumed traditional foods or medicines from within the Athabasca River area (92%, n=119 and 91%, n=23), followed by participants between 31 and 50 years (86%, n=72), and participants between 18 and 30 years (76%, n=33). However, due to the reduced number of survey responses collected for children (n=23), this value (91%) may not be representative of the under 18 years age group. Ultimately, these results highlight that

traditional foods and medicines are important and widely consumed by survey participants within the study area in the past year.

Table 5.4: Percentage of participants who have consumed traditional foods or used traditional medicines in the past year from the Athabasca River, Peace-Athabasca Delta, Lake Athabasca, or other waterbodies in the surrounding region, by age group and sex.

| | | 8 years 23) | |) years = 33) | | 9 years = 72) | 51 years $(n =$ | |
|--------|-----|-------------|-----|------------------|-----|------------------|------------------|----|
| Sex | Yes | No | Yes | No | Yes | No | Yes | No |
| Female | 48% | 4% | 45% | 15% | 43% | 10% | 54% | 7% |
| Male | 43% | 4% | 30% | 9% | 43% | 4% | 37% | 2% |
| Other | 0% | 0% | 0% | 0% | 0% | 0% | 1% | 0% |
| Total | 91% | 9% | 76% | 24% | 86% | 14% | 92% | 9% |

5.4.2.2 Preferences for consuming traditional foods

The majority of participants would like to consume more traditional foods than they currently do across most receptor groups (see Table 5.5). The results suggest that 63% of participants would like to consume more mammals, 54% would like to consume more birds, and 51% of participants indicated they would like to consume more fish and freshwater clams. A slightly smaller percentage of participants (49%) indicated they would like to consume more traditional plants than they currently do. Overall, these results suggest there is a high level of interest among survey participants to consume more traditional foods than they did in the past year.

Table 5.5: Percentage of participants who would like to consume more traditional foods than they currently do, by receptor group

| | Fish and freshwater clams $(n = 220)$ | $\begin{array}{c} \text{Mammals} \\ (n = 225) \end{array}$ | Birds (n = 219) | Plants $(n = 217)$ |
|-----|---------------------------------------|--|------------------|--------------------|
| Yes | 51% | 63% | 54% | 49% |
| No | 49% | 37% | 46% | 51% |

2834 5.4.2.3 Providing traditional foods and medicines to children

A total of 26% of survey participants indicated they are responsible for providing traditional foods or medicines to children under the age of 18 (n=199). Given that just over one quarter of survey participants are responsible for providing traditional foods and medicines to children, this suggests the importance of capturing younger demographics consumption information to ensure their consumption patterns are reflected in determining water quality thresholds for the protection of exposures to environmental media.

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5.4.2.4 Barriers to harvesting more traditional foods and medicines

Participants identified numerous barriers that prevent them from harvesting more traditional 2842 foods and medicines than they currently do (Table 5.6). Fear that a resource may be con-2843 taminated was the most commonly identified barrier, which was reported by participants 224 2844 times or an average of 24% across the four primary receptor groups (i.e., fish, mammals, birds, 2845 plants). The barrier that traditional resources are located too far away was indicated by par-2846 ticipants 122 times or an average of 13% across the four primary receptor groups, and a lack of 2847 tools or equipment was indicated as a major barrier a total of 119 times or reported an aver-2848 age of 13% across the four primary receptor groups. Additional barriers frequently expressed 2849 by participants included (average percentage across receptor groups): changes to water levels 2850 2851 (13%), restricted access to harvesting areas (11%), lack of connection to a harvester (11%), 2852 lack of knowledge of where or how to harvest (11%), lack of transportation (10%), lack of time (8%), concerns that traditional resources are diseased or unhealthy (7%), cost (3%), decreases 2853 2854 in plant or animal populations (2%), lack of experience (1%), medical conditions (1%), being an elder or too old to harvest (1%), as well as several others (10%). 2855 These results may not be comprehensive and likely do not capture all of barriers that 2856 2857 prevent community members from harvesting traditional foods. However, they do suggest that survey participants want to consume more traditional foods and medicines and as a result 2858 estimated consumption patterns of traditional foods may be an underestimate if barriers are 2859 2860 reduced.

³Participants indicated to community researchers that flooding this past year was particularly prohibitive for harvesting traditional foods and medicines.

⁴The 'other' category includes additional barriers identified to a lesser extent (indicated less than 10 times or 1%) by participants included: impacts of wildfires; changes in weather patterns; species migrating to different areas; difficulty finding traditional resources; changes in the taste of traditional resources; impacts of invasive plants; COVID-19-related restrictions; that it is unsafe to travel; that traditional foods are not being provided by the community; being a new member of the community.

Table 5.6: Percentage of participants that identified barriers to harvesting more traditional foods or medicines than they currently do.

| Barrier to harvesting more traditional foods and medicines | Fish and freshwater clams | Mammals | Traditional birds | Traditional plants | Average percentage across primary receptor groups |
|--|---------------------------------|---------------|----------------------|--------------------|--|
| Cost | 3% | 4% | 4% | 1% | 3% |
| Lack of tools or equipment | 12% | 18% | 14% | 8% | 13% |
| Lack of knowledge of where or how to harvest | 10% | 10% | 8% | 14% | 11% |
| Too far away Fear of contamination | $\frac{11\%}{30\%}$ | $16\% \ 28\%$ | $15\% \\ 22\%$ | $10\% \\ 18\%$ | $13\% \ 24\%$ |
| Species appear diseased or unhealthy | 8% | 9% | 5% | 5% | 7% |
| Lack of connection to a harvester | 10% | 14% | 10% | 10% | 11% |
| Medical condition | 2% | 1% | 1% | 1% | 1% |
| Lack of transportation | 10% | 13% | 10% | 7% | 10% |
| Restricted access to harvesting areas | 8% | 15% | 14% | 9% | 11% |
| Lack of time | 7% | 8% | 7% | 8% | 8% |
| Changes to water levels | 14% | 14% | 11% | 11% | 13% |
| Lack of experience | 1% | 1% | 2% | 2% | 1% |
| Decrease in plant or animal populations | 0% | 5% | 3% | 1% | 2% |
| Age related limitations | 1% | 1% | 1% | 0% | 1% |
| Other | 9% | 12% | 8% | 12% | 10% |

${}_{2861} \ \mathbf{Appendix} \ \mathbf{A}$

Linked Appendices

| 2863 | A.1 Data Catalogue | | | |
|----------------------|--|--|--|--|
| 2864 | Data Catalogue – Water and sediment quality data compilation | | | |
| 2865 | https://thompson a quatic.ca/reports/WQCIU/c2a1.pdf | | | |
| 2866 | A.2 Current condition target supplemental information | | | |
| 2867 | Current conditions – Additional information | | | |
| 2868 | https://thompson a quatic.ca/reports/WQCIU/c2a2.pdf | | | |
| 2869 2870 2871 | A.3 Summary of Available Surface Water Quality Guidelines Summary of Available Surface Water Quality Guidelines | | | |
| 2872 | https://thompson a quatic. ca/reports/WQCIU/c3a1.pdf | | | |
| 2873 2874 | A.4 Input Parameters for Derivation of Water Quality Criteria for Indigenous Use Protection | | | |
| 2875 | Input Parameters for Derivation of Water Quality Criteria | | | |
| 2876 | https://thompson a quatic. ca/reports/WQCIU/c3a 2.pdf | | | |

777 A.5 Summary of Sediment Quality Guidelines from

North America

- 2879 Summary of Sediment Quality Guidelines from North America
- ${\tt 2880} \qquad {\tt https://thompsonaquatic.ca/reports/WQCIU/c4a1.pdf}$

2881 A.6 Derivation of Sediment Quality Criteria for Indige-

nous Use Protection

- 2883 Derivation of Sediment Quality Criteria for Proteciton of Indigenous Water Use
- ${\tt 2884} \qquad {\tt https://thompsonaquatic.ca/reports/WQCIU/c4a2.pdf}$

2885 A.7 Consumptive Use Survey Handout

- 2886 Consumptive Use Survey Handout
- ${\it https://thompsonaquatic.ca/reports/WQCIU/c5a1.pdf}$

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