

# Concentrations of urinary arsenic species in relation to rice and seafood consumption among children living in Spain

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## 32    **Abstract**

33    Inorganic arsenic (i-As) has been related to wide-ranging health effects in children, leading to  
34    lifelong concerns. Proportionally, dietary i-As exposure dominates in regions with low  
35    arsenic drinking water. This study aims to investigate the relation between rice and seafood  
36    consumption and urinary arsenic species during childhood and to assess the proportion of  
37    urinary i-As metabolites. Urinary arsenic species concentration in 400 4-year-old children  
38    living in four geographical areas of Spain, in addition to repeated measures from 100 children  
39    at 7 years of age are included in this study. Rice and seafood products intake was collected  
40    from children's parents using a validated food frequency questionnaire (FFQ). At 4 years of  
41    age, children's urine i-As and monomethylarsonic acid (MMA) concentrations increased with  
42    rice product consumption ( $p$ -value = 0.010 and 0.018, respectively), and urinary  
43    arsenobetaine (AsB) with seafood consumption ( $p$  = 0.002). Four-year-old children had a  
44    higher consumption of both rice and seafood per body weight and a higher urinary %MMA  
45    ( $p$ -value = 0.001) and lower % dimethylarsinic acid (DMA) ( $p$ -value = 0.017). This study  
46    suggests increased dietary i-As exposure related to rice product consumption among children  
47    living in Spain, and the younger ones may be especially vulnerable to the health impacts of  
48    this exposure also considering that they might have a lower i-As methylation capacity than  
49    older children. In contrast, seafood consumption did not appear to influence the presence of  
50    potentially toxic arsenic species in this population of children.

51    **Keywords:** Food frequency questionnaire; Arsenic speciation; Inorganic arsenic; Rice;  
52    Seafood; Urinary biomarker; Young children; Childhood.

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73    Informed consent was obtained from all participants in each phase, and the hospital ethics  
74    committees in the participating regions approved the study.

75

## 1. Introduction.

There is sufficient evidence in humans for carcinogenicity due to i-As oral exposure (IARC, 2012). Other than cancer, exposure to i-As has been associated with a wide range of adverse health outcomes including neurological, cardiovascular, respiratory and metabolic diseases (Medrano et al., 2010; Naujokas et al., 2013; Tsuji et al., 2014). Fetal and early life exposure to i-As is of particular concern due to the specific vulnerability to the adverse health effects of i-As, which may have a marked impact throughout the lifespan even at low to moderate levels of exposure (Farzan et al., 2016; Gilbert-Diamond et al., 2016; Rodríguez-Barranco et al., 2016; Vahter, 2009). The metabolic pathway of i-As involves a series of reduction and oxidative methylations, in which i-As is sequentially transformed to MMA and DMA. The metabolic process is incomplete such that i-As along with MMA and DMA metabolites are excreted in urine (Tseng, 2009). Health risks associated with exposure to organic arsenic compounds such as DMA and MMA are less certain than i-As, and AsB, which is excreted in the urine unchanged and regarded as non-toxic (Cubadda et al., 2016; Feldmann and Krupp, 2011; Molin et al., 2015).

Diet is the main source of i-As exposure when relatively low arsenic drinking water is available, the maximum level of which is regulated at 0.01 mg/l in the EU (Cubadda et al., 2016; EFSA, 2009; The Council of the European Union, 1998). The 0.01 mg/l is also the WHO guideline value for arsenic in drinking water and thus the legal limit in many countries around the world (WHO, 2011). Rice and rice-based products contain higher levels of i-As compared to other foodstuff, in addition to DMA and traces of MMA, and therefore they are an important i-As dietary source (Meharg et al., 2008; Signes-Pastor et al., 2016a, 2016b; Williams et al., 2007). This is an increasingly recognized issue for infants and young children whose food consumption per body weight is higher compared to adults, and rice products

generally predominate their diet due to the putative organoleptic and nutritional quality, and relatively low allergic potential of these products (Burló et al., 2012; Da Sacco et al., 2013; Karagas et al., 2016; Meharg et al., 2008; Signes-Pastor et al., 2016b). New regulations regarding maximum levels of i-As in rice and rice-based products have been recently established in the European Union (EU) with the most restrictive level of 0.1 mg/kg set for rice-based foods for infants and young children (EC, 2015). The 0.1 mg/kg i-As threshold is also a draft “action level” in the United States (U.S.) (FDA, 2016). As far as seafood is concerned, AsB dominates in fish, whereas other organic arsenic compounds such as arsenosugars and arsenolipids are also found in seafood products, including i-As in bivalves, crustaceans and algae (Feldmann and Krupp, 2011; Fontcuberta et al., 2011; Krishnakumar et al., 2016; Molin et al., 2015; Navas-Acien et al., 2011; Taylor et al., 2016). Thus, if relatively low arsenic drinking water is available, rice and seafood products are principal total arsenic, including potentially toxic arsenic species, dietary sources (Cubadda et al., 2016; EFSA, 2009). However, limited information is as yet available regarding the contribution of these key food items to urinary arsenic species concentration as exposure biomarker during childhood, especially among those with increased and consistent rice and seafood product consumption, such as those living in Spain, whose consumption of rice and seafood products is enhanced by the Spanish gastronomic culture.

In this study, the relations between the arsenic species and their summation (excluding AsB) concentrations in urine from 4-year-old children participating in a mother-child prospective cohort study, and rice and seafood product intake were determined. Further, the percentage of urinary i-As, MMA, and DMA were assessed in children at 4 and 7 years of age.

## **2. Methods.**

123 2.1. *Study population.* The study population here belongs to the mother-child pair participants  
124 in the INMA – *Infancia y Medio Ambiente* - Environment and Childhood project, a  
125 prospective population-based birth cohort study conducted in various Spanish regions,  
126 [www.proyectoinma.org](http://www.proyectoinma.org) (Guxens et al., 2012). Women participants of the INMA project were  
127 recruited at the beginning of their pregnancy (2003-2006) at their reference primary health  
128 care centers or public hospitals and were followed-up until delivery ( $n = 2,625$ ). All women  
129 met the inclusion criteria of  $\geq 16$  years old, singleton pregnancy, non-assisted conception,  
130 delivery scheduled at the reference hospital, and adequate knowledge of the Spanish language  
131 to understand and complete the questionnaires accurately. Their children were enrolled at  
132 birth and were followed-up during infancy and childhood.

133 2.2. *Samples collection.* Spot urine samples were collected from a study population of 400  
134 children of 4 years of age comprised of four groups of 100 children evenly distributed  
135 between boys and girls, and randomly chosen from each of the Spanish sub-cohorts of  
136 Asturias, Gipuzkoa, Sabadell, and Valencia, situated in the north and east of the country; Bay  
137 of Biscay and Mediterranean coast, respectively. Additionally, urine samples of the same  
138 children from the Valencia sub-cohort were collected at 7 years of age ( $n = 100$ ).

139 Informed consent was obtained from all participants in each phase, and the hospital ethics  
140 committees in the participating regions approved the study.

141 2.3. *Anthropometric data and questionnaires.* Parental sociodemographic and socioeconomic  
142 characteristics were collected through a maternal questionnaire in their 1<sup>st</sup> trimester of  
143 pregnancy including living area, parents education, and social class according to maternal  
144 and paternal occupation coded using the International Standard Classification of Occupations  
145 (ISCO88) (International Labor Office (ILO), 2012). Trained staff measured children's weight  
146 and height following standard protocols. Children's body mass index (BMI) was calculated in

kg/m<sup>2</sup>. Four-year-old children's dietary intake was assessed with a validated FFQ asking about consumption of 105 food items in the previous year (Vioque et al., 2016). Parents or children's guardians provided the information regarding children's diet, and it was recorded in the FFQ during personal interviews administered by trained nutritionists. The usual daily total energy intake for each child was calculated by multiplying the frequency of the use of each food item by the nutrient content of the portion size specified in the FFQ. Total energy values were primarily obtained from food composition tables from the U.S. Department of Agriculture (USDA, 2016), and other published sources reporting information on content in Spanish food (Palma et al., 2008). Rice products included 3 food items referring to cooked rice and breakfast cereal consumption. Seafood products gathered several food items ( $n = 11$ ) including i) surimi products such as seafood sticks and crab sticks, ii) white fish such as hake, sole and sea bream fried or battered, iii) white fish boiled or grilled, iv) swordfish, v) big oily fish such as tuna, bonito, and salmon, vi) small oily fish such as anchovy, sardine and mackerel, vii) canned tuna in oil, viii) canned tuna in brine, ix) squid, cuttlefish, and octopus, x) clams, mussels, and cockles, and xi) prawns, crab, and lobster.

*2.4. Sample preparation and chemical analysis.* Children's urinary arsenic speciation included AsB, DMA, MMA, and i-As, and the levels of which were previously reported in Signes-Pastor *et al.* (2017) with a median of 9.71 µg/l, 3.97 µg/l, 0.44 µg/l, and 0.35 µg/l for the 400 4-year-old children, respectively. The urine samples were stored at or below -20°C until analysis. Then, they were centrifuged and analytical grade hydrogen peroxide was added to convert any arsenite to arsenate to facilitate subsequent chromatographic detection by IC-ICP-MS. Replicate samples of the urine lyophilized material ClinChek<sup>®</sup> - Control level I (Recipe Chemicals + Instruments GmbH in Munich, Germany), and blank samples were included in each analysis batch as quality control. The specific gravity was measured with a clinical refractometer to normalize urine dilution. The recovery percentages of arsenic

speciation based on several replicate samples of the urine lyophilized material ClinChek® - Control level I ( $n = 33$ ) were  $115 \pm 2\%$  for i-As,  $97 \pm 2\%$  for MMA,  $94 \pm 2\%$  for DMA, and  $90 \pm 2\%$  for AsB. The mean and range concentrations of the arsenic species reference values in the urine lyophilized material ClinChek® - Control level I are as follows: 4.55 (2.73 – 6.37)  $\mu\text{g/l}$  for i-As, 2.50 (1.50 – 3.50)  $\mu\text{g/l}$  for MMA, 9.8 (5.88 – 13.7)  $\mu\text{g/l}$  for DMA, and 16.8 (12.6 – 21.0)  $\mu\text{g/l}$  for AsB. The LOD for arsenic speciation, calculated from DMA calibration, was 0.011  $\mu\text{g/l}$  (Signes-Pastor et al., 2017).

*2.5. Statistical analysis.* Urinary arsenic species concentrations were log-transformed due to their skewed distribution, and their percentages determined based on  $\Sigma\text{As}$ , calculated as the summation of i-As, MMA, and DMA concentrations, excluding AsB. Rice and seafood products consumption were adjusted for energy intake using the residual method, a standard method widely used in nutritional epidemiology. The residuals were calculated from a linear regression with the natural logarithm of the food intake modeled as the dependent variable and the natural logarithm of total energy intake as the independent variable. The mean food intake of the population was then added back to the calculated residuals (Willett et al., 1985). The Wilcoxon test was applied to compare 4- and 7-year-old children's BMI, and consumption of rice and seafood products, and to compare their urinary arsenic species profile as percentage of i-As, MMA, and DMA. The Spearman's correlation (Rho) was used to investigate the relationship between rice and seafood product consumption, and urinary arsenic species concentration. Rice products consumption was right skewed, and thus for sensitivity reasons, the Spearman's correlation analysis was also calculated excluding values higher than 3 standard deviations from the mean and those with urinary As species concentrations below the LOD. All statistical analyses and plots were performed using the R Statistical Software (R Core Team, 2014). The limit of detection (LOD) was calculated as the mean of the blank concentrations plus three times the standard deviation of the blank



concentrations multiplied by the dilution factor. The  $\frac{1}{2}$  LOD value was assigned for statistical analyses of the data and plots when samples were below the LOD.

### 3. Results.

*3.1. Study population.* Nearly the entire study population here had Spanish nationality (>95%) and was living in urban areas (>80%). Further details regarding selected characteristics of the 4- and 7-year-old children study population are reported in **Table 1**.

*3.2. FFQ.* Nearly all children at 4 years of age were reported to be consuming rice and seafood products ( $n = 383$ ). The remaining 17 participants were missing values for both categories. Four-year-old children's median rice intake was 27.2 g/day and seafood consumption was 39.9 g/day with a median percent contribution and range of 87.3% (52.2% – 96.2%) from fish, 4.1% (0.9% - 40.7%) from cephalopods, 3.4% (0.9% - 22.3%) from bivalves, and 3.5% (1.2% - 27.2%) from crustaceans. All 4- and 7-year-old children living in Valencia reported rice and seafood product consumption ( $n = 100$ ). Valencian children's rice product consumption (g/day) increased with age ( $p < 0.001$ ), and a similar trend was found for seafood product consumption (g/day) ( $p = 0.062$ ). However, their rice and seafood product intake decreased with age when adjusted per body weight ( $p = 0.043$ , and  $p < 0.001$ , respectively) (**Table 2**).

*3.3. As exposure.* The urinary AsB concentration was higher than the LOD in all urine samples analyzed. However, 10 (2.5%), 29 (7.2%), and 17 (4.2%) urine samples had DMA, MMA, and i-As concentrations below the LOD for the 4-year-old children dataset, respectively. All the urine samples from 7-year-old children living in Valencia had AsB, DMA, and MMA concentrations higher than the LOD, and only one sample had an i-As level below the LOD. The specific gravity average values for urine samples from children at 4

years ( $n = 400$ ) and 7 years of age ( $n = 100$ ) were 1.015 and 1.018 g/ml, respectively. At 4 years of age, children's rice product intake positively correlated with urinary i-As ( $Rho = 0.130$ ;  $p = 0.010$ ), MMA ( $Rho = 0.120$ ;  $p = 0.018$ ) and more weakly with DMA ( $Rho = 0.094$ ;  $p = 0.065$ ) and  $\Sigma As$  ( $Rho = 0.088$ ;  $p = 0.083$ ) as depicted in **Figure 1**. When values higher than 3 standard deviations from the mean and those with urinary arsenic species concentration below the LOD were not included in the calculations, the positive correlations between rice product consumption and urinary arsenic species were stronger with urinary i-As ( $Rho = 0.177$ ;  $p = 0.001$ ), similar with urinary MMA ( $Rho = 0.113$ ;  $p = 0.043$ ), and weaker with urinary DMA ( $Rho = 0.077$ ;  $p = 0.156$ ) and  $\Sigma As$  ( $Rho = 0.061$ ;  $p = 0.257$ ) compared to the correlations reported earlier when including the entire dataset in the calculations (See **Supplemental Material, Figure S1**). Further, seafood product consumption had a clear positive correlation with urinary AsB at 4 years of age ( $Rho = 0.153$ ;  $p = 0.002$ ). On the contrary, none of the other urinary arsenic species analyzed was clearly associated with seafood product intake (**Figure 2**). Nor was the urinary  $\Sigma As$  clearly associated with seafood product intake ( $Rho = 0.069$ ;  $p = 0.176$ ). Likewise, the combination of bivalve and crustaceous consumption, excluding the other seafood items, had a clear correlation with urinary AsB at 4 years of age ( $Rho = 0.122$ ;  $p = 0.016$ ); however, none of the other urinary arsenic species was clearly associated with the bivalve and crustacean consumption ( $Rho = 0.020$ ;  $p = 0.687$  for i-As,  $Rho = 0.007$ ;  $p = 0.882$  for MMA,  $Rho = 0.088$ ;  $p = 0.082$  for DMA, and  $Rho = 0.088$ ;  $p = 0.083$  for  $\Sigma As$ ).

In the repeated urinary arsenic species measures from children from Valencia, we observed higher urinary %DMA at 7 years of age (mean, 83%; range, 70 – 96%) than at 4 years (mean, 81%; range, 55 – 96%) ( $p = 0.017$ ), and lower %MMA (mean, 8.7% and 7.4%; range, 0.1 – 21% and 1.6 – 15%, for children of 4 and 7 years of age, respectively) ( $p = 0.001$ ). Children of 4 years of age had a mean and range urinary %i-As of 10% and 1.3 – 28% compared to

245 that of 9.0% and 0.2 – 22% at 7 years ( $p = 0.162$ ) (**Figure 3**).

#### 246 **4. Discussion.**

247 Diet is the main source of i-As and other potential toxic arsenic species in regions with access  
248 to relatively low i-As drinking water. This is especially true for infants and young children  
249 whose diet is generally of lower food diversity, and characterized with a higher food  
250 consumption per body weight compared to adults, which in turn exacerbates dietary exposure  
251 (Cubadda et al., 2016; EFSA, 2009; Karagas et al., 2016). In our study children at 4 years of  
252 age had higher rice and seafood product consumption per body weight than at 7 years. A  
253 strong influence of the Spanish gastronomic culture was reflected in a high regular  
254 consumption of rice and seafood products allowing us to assess the contribution of these  
255 products to children's urinary arsenic species concentration.

256 One of the limitations of our study is the lack of individual data on household water  
257 concentrations of i-As. However, in Spain, each autonomous regional public health  
258 administration is in charge of monitoring tap water quality and maintaining arsenic levels  
259 below the maximum EU standard. Indeed, published records indicate very low arsenic  
260 concentrations with a median  $< 1 \mu\text{g/l}$ , and a high percentage of water samples compliance  
261 with the arsenic drinking water regulation ranging from 97% to 100% between 1996 and  
262 2011 (Espejo-Herrera et al., 2013; Palau Miguel and Guevara Alemany, 2011; The Council  
263 of the European Union, 1998). Likewise, levels far below  $10 \mu\text{g/l}$  have also been reported in  
264 bottled water sold in Spain despite bottled natural mineral water, the most consumed type of  
265 bottled water in Spain, not being subjected to any physicochemical treatment (Espejo-Herrera  
266 et al., 2013; The Council of the European Union, 1998). Also, urinary arsenic species  
267 concentration is commonly used as a biomarker of short-term exposure to i-As. However, it  
268 can be a reliable source of long-term exposure among individuals with consistent intake

patterns (Kile et al., 2009; Marchiset-Ferlay et al., 2012; Navas-Acien et al., 2009). Further, in this study, the children's usual diet was measured in personal interviews with parents and children's guardians using a validated FFQ, which although reliable may be subject to misclassification (Vioque et al., 2016).

Only a few studies have assessed the association between childhood rice products consumption and urinary arsenic species concentration, as recently reviewed (Davis et al., 2017). Consumption of rice has been related to an increase of urinary DMA and MMA in children at 6 to 11 years old from the U.S. belonging to the National Health and Nutrition Examination Survey (NHANES) (Davis et al., 2014; deCastro et al., 2014). Further, in children 5 to 8 years old with low i-As exposure from water residing in Uruguay, an increase of consumption of 5 g of rice per day, assessed via 24-h dietary recall, was associated with a 1.2- and 1.1-fold increment of urinary  $\Sigma$ As and %DMA, respectively (Kordas et al., 2016). In our current study of 4-year-old children, consumption of rice and rice products was higher than in Uruguay where the median consumption was 0 g/day (Kordas et al., 2016), and we found consumption to be positively correlated with urinary i-As and MMA concentrations. We also found a positive trend between rice product intake and urinary DMA, but not as clearly as that reported in previous studies focused on older children or adults (Cascio et al., 2011; Davis et al., 2017; Kordas et al., 2016; Meharg et al., 2014). This may have resulted from the consumption of rice with a relatively high percentage of i-As, such as that grown in the Iberian Peninsula, including Portugal and Spain, with a median i-As percentage and range of 57% (14% - 95%) (Signes-Pastor et al., 2016a).

The relative proportion of urinary arsenic species is a widely used measure of an individual's i-As methylation capacity, which is determined at least in part by genetic factors such as polymorphisms in the arsenic methyltransferase (AS3MT) gene (Jansen et al., 2016; Skröder

293 Löveborn et al., 2016; Tseng, 2009). Children's ability to methylate i-As has been reported to  
294 be more efficient than their mothers (Skröder Löveborn et al., 2016). However, a marked  
295 increase in urinary %MMA has been found in children aged 18 months compared to the same  
296 at 3 months, suggesting a lower i-As methylation efficiency during early life (Fängström et  
297 al., 2009). Despite the relatively small sample size, our study suggests that the urinary arsenic  
298 species content pattern of children of 4 and 7 years old showed higher i-As methylation  
299 capacity with age, with higher mean urinary %DMA and lower mean %MMA and %i-As in  
300 older children. However, it is important to bear in mind that this study did not take into  
301 account potential confounding factors such as direct exposure to DMA from rice  
302 consumption and DMA from arsenosugars and arsenolipids metabolism that may weaken the  
303 measurement of i-As methylation capacity as a relative proportion of the urinary i-As  
304 metabolites concentration (Molin et al., 2014; Signes-Pastor et al., 2016a)

305 Spain is among the highest seafood consumers in Europe (Navarro Serrano et al., 2016;  
306 Taylor et al., 2016), and consequently seafood products were in almost all children's diets  
307 included in this study. The 4-year-old children's total seafood intake was primarily fish  
308 consumption, which contains predominantly AsB (Molin et al., 2015; Taylor et al., 2016),  
309 and thus seafood intake was clearly correlated with the urinary excreted unchanged AsB.

310 There was no evidence of correlation between seafood consumption and i-As or other urinary  
311 arsenic species that have been related to potential health effects (Choi et al., 2010; Fort et al.,  
312 2014; Navarro Serrano et al., 2016; Navas-Acien et al., 2011). Notably there was no evidence  
313 of correlation between the combination of bivalves and crustaceans consumption with urinary  
314 i-As or its metabolites despite higher i-As content being commonly reported in bivalves and  
315 crustaceans compared to fish and cephalopods including those marketed in Spain  
316 (Fontcuberta et al., 2011; Krishnakumar et al., 2016; Taylor et al., 2016). The lack of  
317 association between consumption of bivalves and crustaceans with urinary i-As and its

metabolites may be related to the relatively low and potentially irregular 4-year-old children's consumption of these marine products, which only counted for a median percentage of 7.2% of the total seafood intake. Therefore, seafood consumption generally does not appear to be a major source of arsenic exposure aside from AsB in this population of children.

## **5. Conclusions.**

Our findings suggest that in regions with low i-As drinking water and with a traditional diet of rice and seafood such as in Spain, children's rice and seafood products are key sources of total arsenic dietary exposure including potentially toxic arsenic species. Seafood consumption was associated with an increase of urinary AsB, an unmetabolized form of arsenic that is considered non-toxic. Conversely, urinary i-As and MMA concentrations were clearly related to children's rice intake, consumption of which was higher for younger children (4 years old) compared to older children (7 years old) when adjusted per body weight. Future studies need to focus on the association between rice product dietary i-As exposure during childhood and lifelong adverse health outcomes, especially for younger children (4 years old) who might have a lower i-As methylation capacity compared to older children (7 years old) as suggested in the present study. However, due to the limitations of the study design and the potential confounding effect of dietary DMA this cannot be unequivocally demonstrated here.

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523 **Table 1:** Selected characteristics of the 4- and 7-year-old children study population.

Characteristics	4-year-old children from Spain <sup>a</sup>	Followed up children from Valencia at 4 and 7 years of age
	N = 400 , No. (%)	N = 100
<i>Gender</i>		
Female	200 (50%)	51
Male	200 (50%)	49
<i>Nationality</i>		
Spanish	377 (96%)	95
Others	15 (4%)	5
NA <sup>b</sup>	8	
<i>Living area</i>		
Urban	319 (81%)	81
Semiurban	62 (16%)	17
Rural	11 (3%)	2
NA	8	
<i>Level of education</i>		
<i>Mother</i>		
Primary	76 (19%)	23
Secondary	159 (41%)	48
University	156 (40%)	29
NA	9	
<i>Father</i>		
Primary	128 (33%)	43
Secondary	173 (44%)	38
University	89 (23%)	19
NA	10	
<i>Social class</i>		
<i>Mother</i>		
Upper - I+II	93 (24%)	18
Middle - III	112 (28%)	28
Lower - IV+V	188 (48%)	54
NA	7	
<i>Father</i>		
Upper - I+II	89 (23%)	16
Middle - III	69 (18%)	23
Lower - IV+V	226 (59%)	61
NA	16	

524

525 <sup>a</sup>Include 4-year-old children from Asturias, Gipuzkoa, Sabadell, and Valencia.

526 <sup>b</sup>NA = Missing values

**Table 2:** Children's rice and seafood product consumption, body weight, and BMI according to their age.

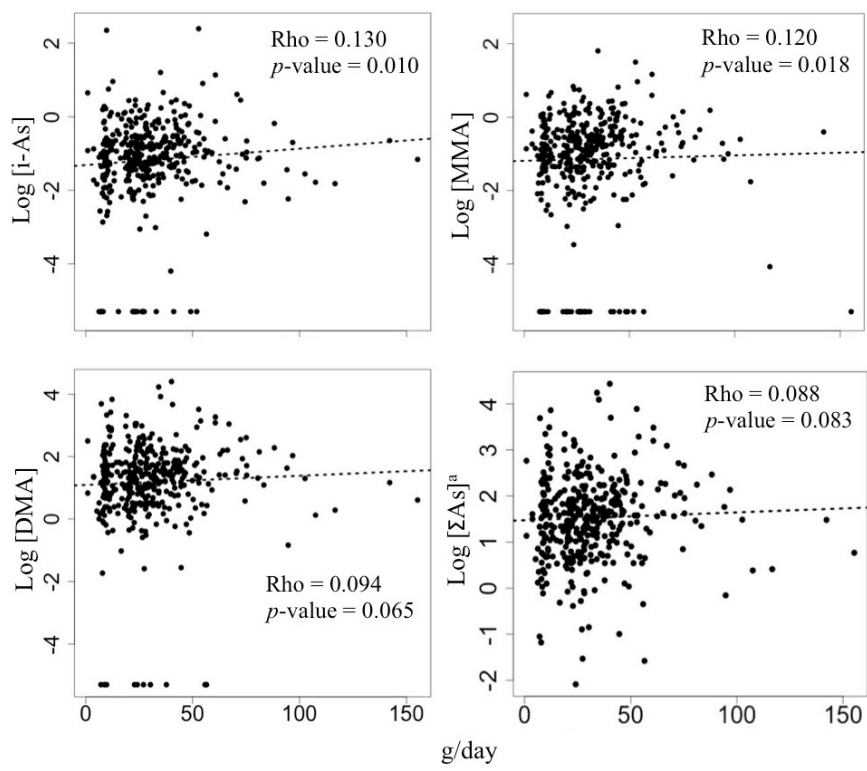
Study population	Children from Spain <sup>a</sup>	Children from Valencia		
	4-year-old n = 400	4-year-old n = 100	7-year-old n = 100	<i>p</i> -value between 4- and 7-year-old.
Body weight (kg)	17.8 (16.4 - 19.8) <sup>b</sup>	17.7 (16.5 - 18.6) <sup>#c</sup>	27.0 (23.4 - 31.0)*	<0.001
BMI (kg/m <sup>2</sup> )	16.0 (15.2 - 17.2)	15.9 (15.2 - 16.8) <sup>#</sup>	16.5 (15.5 - 18.8)*	<0.001
Rice products (g/day)	27.2 (18.28 - 39.6)	28.2 (22.5 - 39.2) <sup>#</sup>	35.6 (28.3 - 55.0)*	<0.001
Seafood products (g/day)	39.9 (31.4 - 49.5)	37.3 (26.7 - 47.1)*	42.4 (30.8 - 52.1)*	0.062
Rice products (g/day) / Body weight (kg)	1.48 (0.86 - 2.16)	1.60 (1.21 - 2.30)*	1.28 (0.99 - 2.00) <sup>#</sup>	0.043
Seafood products (g/day) / Body weight (kg)	2.17 (1.68 - 2.77)	2.10 (1.51 - 2.75)*	1.52 (1.12 - 1.95) <sup>#</sup>	<0.001

<sup>a</sup>Include children living in Asturias, Gipuzkoa, Sabadell, and Valencia.

<sup>b</sup>Values reported as median (1<sup>st</sup> percentile – 3<sup>rd</sup> percentile)).

<sup>c</sup>Values with the same symbol were not significantly different at *p*-value < 0.05 for the variable studied due to children's age using the Wilcoxon test.

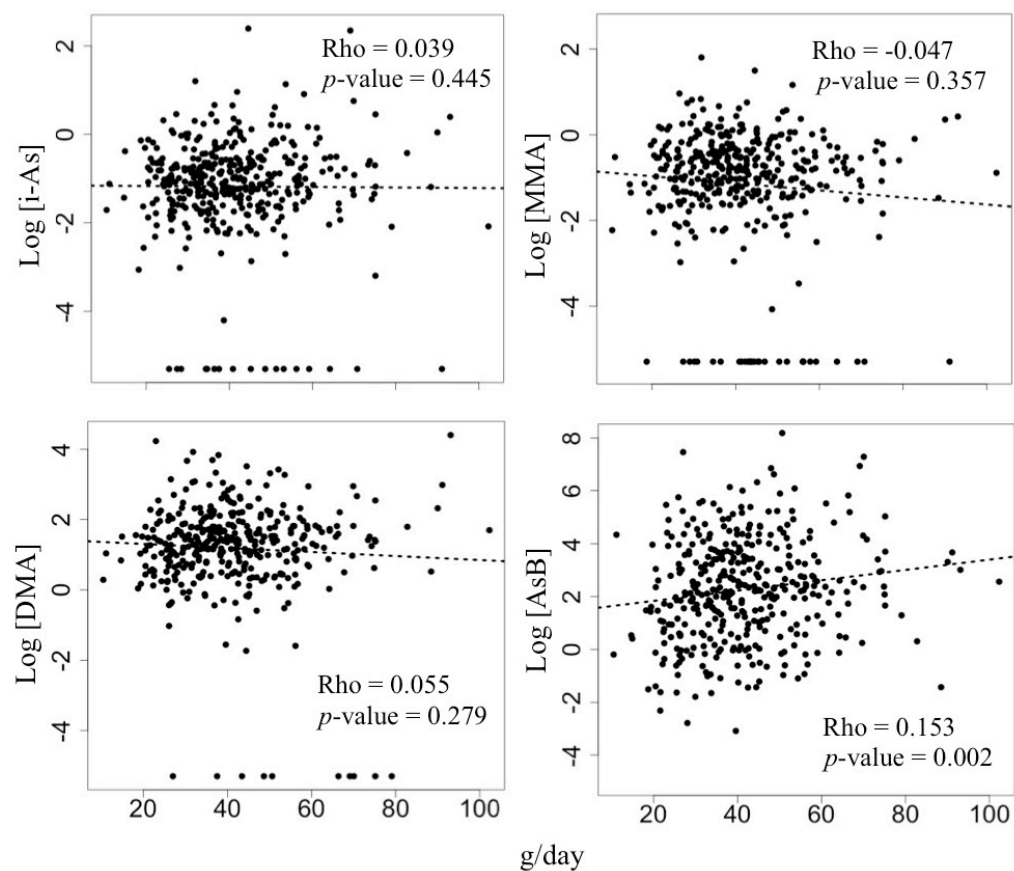
535 **Figure 1:** Spearman correlation between rice products consumption (g/day) and arsenic  
 536 species in urine from 4-year-old children living in Spain.



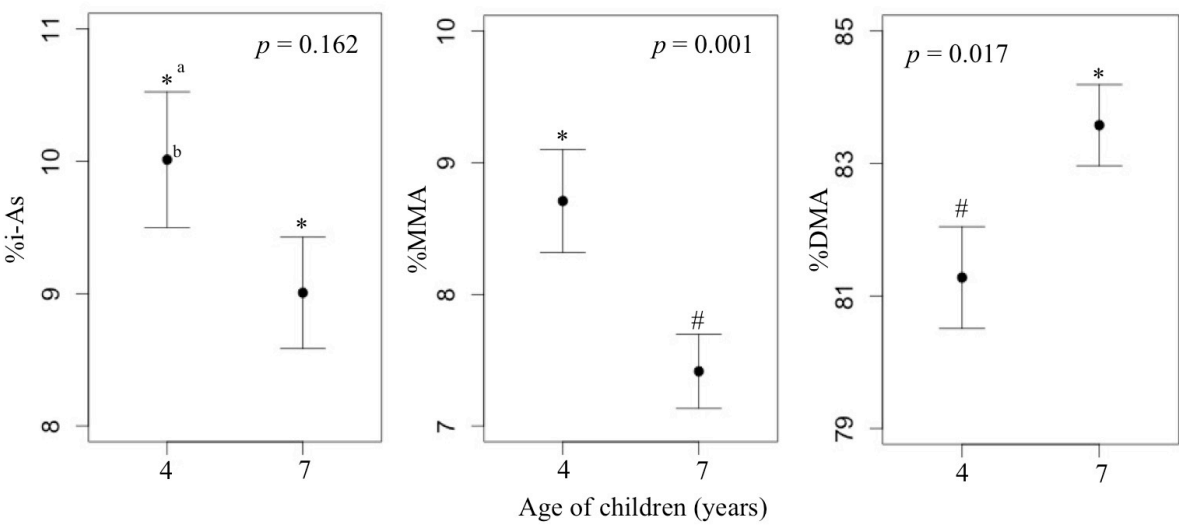
537  
 538 <sup>a</sup>[ $\Sigma$ As] = [i-As + MMA + DMA]

539

**Figure 2:** Spearman correlation between seafood products consumption (g/day) and arsenic species in urine from 4-year-old children living in Spain.



544 **Figure 3:** Valencian 4- and 7-year-old children’s i-As methylation capacity assessment based  
545 on urinary percentage of arsenic species excluding AsB.



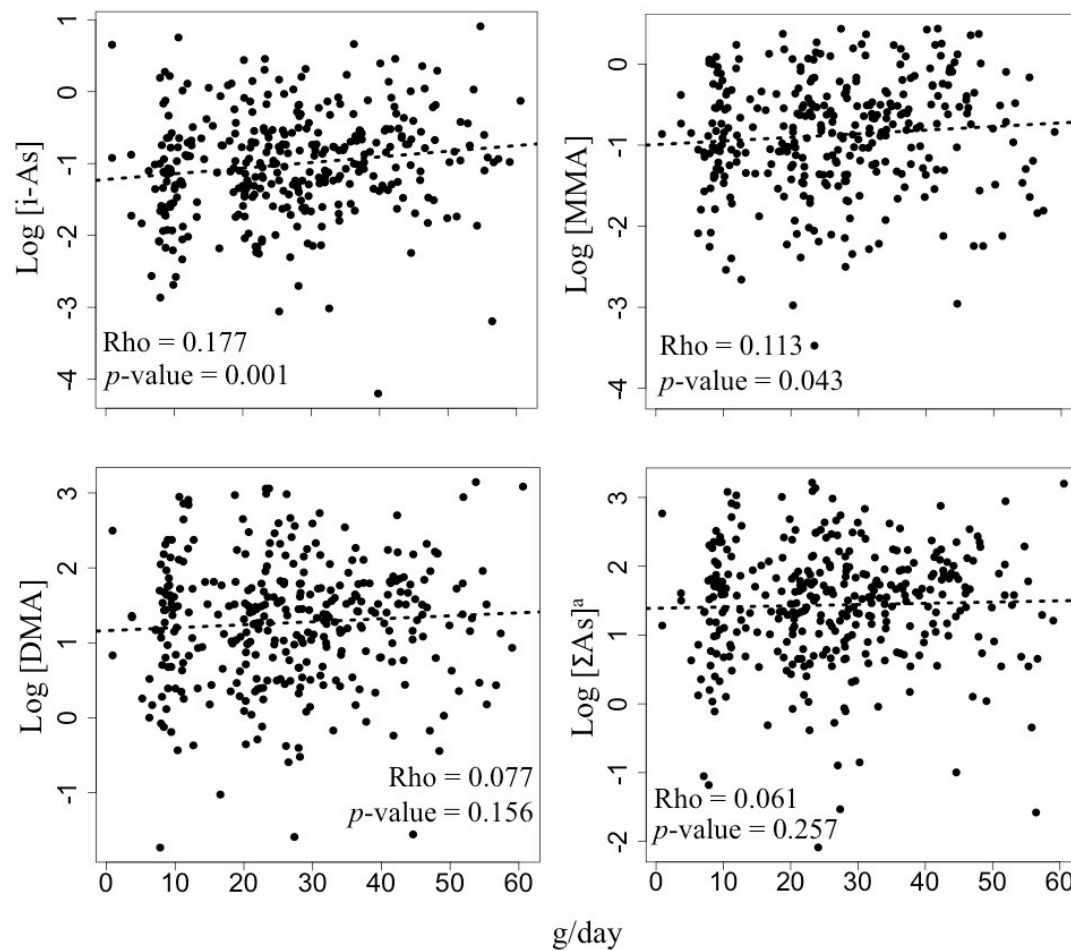
546  
547 <sup>a</sup>Values with the same symbol were not significantly different at  $p$ -value  $<0.05$  for the  
548 variable studied

549 <sup>b</sup>Mean  $\pm$  SE

550

## Supplemental Material

**Figure S1:** Spearman correlation between rice products consumption (g/day) and arsenic species in urine from 4-year-old children living in Spain excluding values outside 3 standard deviations from the mean and those with urinary As species concentration below the LOD.



$$^a[\Sigma As] = [i-As + MMA + DMA]$$